

# The underappreciated role of transboundary pollution in future air quality and health improvements in China

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## Abstract

Studies assessing the achievability of future air quality goals in China have focused on the role of reducing China’s domestic emissions, yet the influence of transboundary pollution of foreign origins has been largely underappreciated. Here, we assess the extent to which future changes in foreign transboundary pollution would affect the achievability of air quality goals in 2030 and 2060 for China. We find that in 2030, under the current-policy scenario in China, transboundary contributions to population-weighted PM<sub>2.5</sub> in China would be reduced by 29% (1.2 μg m<sup>-3</sup>) as foreign countries transition from the fossil fuel-intensive to the low-carbon scenario. By 2060, the difference would be increased to 45% (1.8 μg m<sup>-3</sup>). Adopting the low-carbon instead of the fossil fuel-intensive pathway in foreign countries would avoid 10 million Chinese people from being exposed to PM<sub>2.5</sub> concentrations above China’s Ambient Air Quality Standard (35 μg m<sup>-3</sup>) in 2030 and 5 million Chinese people from being exposed to PM<sub>2.5</sub> concentrations above the World Health Organization Air Quality Guideline (5 μg m<sup>-3</sup>) in 2060. Meanwhile, China adopting the carbon-neutral pathway rather than its current pathway would also be helpful to reduce transboundary PM<sub>2.5</sub> produced from the chemical interactions between foreign-transported and locally-emitted pollutants. In 2060, adopting a low-carbon pathway in China and foreign countries coincidentally would avoid 63% of transboundary pollution and 386,000 associated premature deaths in China, relative to adopting a fossil fuel-intensive pathway in both regions. Thus, the influence of transboundary pollution should be carefully considered when making future air quality expectations and pollution mitigation strategies.

## 1. Introduction

Long-term exposure to ambient fine particulate matter (PM<sub>2.5</sub>, particulate matter smaller than 2.5 μm in aerodynamic diameter) is the largest environmental risk factor for human health, with an estimated 4.1 million attributable deaths worldwide (7.3% of the number of global deaths in 2019; Murray et al., 2020). Countries have taken diverse actions to improve air quality and public health, including setting ambitious future air quality and/or climate goals. The European Commission set the 2030 air quality goal as reducing the number of PM<sub>2.5</sub>-attributable premature deaths by at least 55% compared with 2005 levels, equivalent to reducing PM<sub>2.5</sub> concentrations to below 10 μg m<sup>-3</sup> EU-wide (European Commission, 2022). The United States aims for a 50-52% reduction of greenhouse gas emissions relative to 2005 levels by 2030, which would avoid tens of thousands of premature deaths associated with PM<sub>2.5</sub> as a co-benefit (Burtraw et al.,

1 2021). However, the achievability of air quality goals or the extent of health co-benefits of  
2 climate strategies is subject to large uncertainties as it is affected by a wide range of factors such  
3 as the domestic and global socioeconomic development pathways, energy choices, and air  
4 pollution control measures (Cheng et al., 2021b; O'Neill et al., 2020; Rao et al., 2017).

5 China is a key country to examine factors affecting the achievability of future air quality goals.  
6 On the one hand, China suffers from serious air pollution and adverse health effects, with more  
7 than a quarter of the world's total PM<sub>2.5</sub>-associated premature deaths in 2015 estimated to occur  
8 in China (Zhang et al., 2019). Despite remarkable achievements of the 5-year Clean Air Action  
9 since 2013 (Zhang et al., 2019; Zheng et al., 2018a), annual mean PM<sub>2.5</sub> concentration in China  
10 still exceeded the newly revised World Health Organization (WHO) Air Quality Guideline  
11 (WHO, 2021; AQG; 5 µg m<sup>-3</sup>) by 8 times by the end of the Clean Air Action in 2017. On the  
12 other hand, China has set the most challenging air quality and climate goals among all the  
13 developing countries in the world (Ascensão et al., 2018). In 2018, the Chinese government  
14 proposed the roadmap for The Beautiful China Initiative (The State Council of the People's  
15 Republic of China, 2016), which requires all cities to achieve the national ambient air quality  
16 standards (NAAQS, 35 µg m<sup>-3</sup>) between 2030 and 2035 (The State Council of the People's  
17 Republic of China, 2016). In 2019, China further announced an ambitious climate commitment  
18 to achieve carbon neutrality by 2060. These policies have been regarded as interim steps towards  
19 the WHO AQG. Therefore, investigating potential pathways and factors shaping the future air  
20 quality in China could provide an excellent reference for other countries facing the dual  
21 challenges of economic development and environmental protection.

22  
23 A number of studies seek to answer whether the air quality goals in China can be achieved in  
24 the future. These studies have been primarily focused on potential mitigation pathways to reduce  
25 China's domestic emissions. Cheng et al. (2021b) found that the nation's current emission  
26 control measures could reduce China's PM<sub>2.5</sub> levels to below 30 µg m<sup>-3</sup> by 2030, yet the benefits  
27 of such measures would be mostly exhausted by then (Cheng et al., 2021b; Xing et al., 2020).  
28 Xing et al. (2020) and Tang et al. (2022) suggested that the co-benefits of climate targets alone  
29 (even the most ambitious target of 1.5 °C limit in global warming) were not able to help China  
30 achieve the 35 µg m<sup>-3</sup> target by 2035. Instead, Cheng et al. (2021b) proposed that a combination  
31 of stringent clean air policies and ambitious climate targets (i.e., carbon neutrality by 2060 or a  
32 global warming limit of 1.5 °C) could successfully reduce PM<sub>2.5</sub> to below 35 µg m<sup>-3</sup> by 2030 and  
33 to below 10 µg m<sup>-3</sup> (WHO interim target 4) by 2060. This would require a fundamental  
34 transformation of China's economic-environmental development pathway by phasing out  
35 polluting industries and moving towards renewable energy while implementing strict end-of-pipe  
36 emission control.

37  
38 However, these previous studies have largely neglected how the future changes in foreign  
39 transboundary pollution would affect air quality in China, likely due to the perception that the  
40 relatively short lifetime of PM<sub>2.5</sub> (a few days) does not permit long-distance transport. This  
41 negligence would put into question the confidence of their estimated achievability of air quality  
42 goals in China. Due to large uncertainties in the future socioeconomic development pathways,  
43 environmental commitments, financial supports and technology capabilities, future emissions  
44 from China's neighboring countries might be highly variable, leading to different levels of  
45 transboundary impacts to China's air quality. For example, emissions in South Asia, Central Asia  
46 and Southeast Asia have been estimated to increase in the future by various projections

1 (International Energy Agency, 2021; Koplitz et al., 2017), due to their rapid-economic growth  
2 and a lack of clear commitments on energy choices, climate actions and air pollution control  
3 efforts. In this case, transboundary pollution from neighboring countries to China could  
4 potentially become increasingly important to affect the achievability of air quality goals in  
5 China. Alternatively, these surrounding countries may undergo a sustainable development  
6 pathway, facilitated in part by external financial aids and technology supports, leading to  
7 lowered transboundary pollution to China. Thus, the different prospects of transboundary  
8 pollution could be a significant yet highly uncertain factor in the achievement of future air  
9 quality goals in China.

10  
11 In addition, the mechanism of transboundary pollution poses additional complexity to its  
12 impacts on China's air quality. Foreign emissions can affect air quality in China through direct  
13 transboundary transport in the atmosphere. Moreover, portions of foreign-transported pollution  
14 can also interact chemically with China's locally emitted pollutants. Xu et al. (2023) found that  
15 the transport and transformation of non-methane volatile organic compounds (NMVOCs) from  
16 foreign countries could enhance the atmospheric oxidizing capacity and facilitate the oxidation  
17 of Chinese nitrogen oxides (NO<sub>x</sub>) to form nitric acid (HNO<sub>3</sub>) and nitrate over North China  
18 (referred to as eastern China in that study), leading to considerable foreign contributions to  
19 China's air pollution.

20  
21 The chemical interactions mean that the extent of transboundary pollution will depend on  
22 emission changes in China as well. Therefore, the influence of transboundary pollution on  
23 China's air quality in the future is a complex result of future emissions in China and in foreign  
24 countries, along with their interactions. However, whether the changes in transboundary  
25 pollution via direct transport and chemical interactions could affect the achievement of future air  
26 quality goals in China and to what extent the influence could be remain poorly understood.

27  
28 Here, we assess the potential influences that transboundary pollution could make on the  
29 achievement of future air quality goals in China, considering the changes in direct pollution  
30 transport and in the interactions between transported and China's locally emitted pollution. We  
31 regard the air quality goal stated in The Beautiful China Initiative, which is that all cities have  
32 annual mean PM<sub>2.5</sub> concentrations below 35 μg m<sup>-3</sup> between 2030~2035 (The State Council of  
33 the People's Republic of China, 2016), as China's 2030 air quality goal. We also regard the  
34 WHO AQG (annual mean PM<sub>2.5</sub> below 5 μg m<sup>-3</sup>) as China's 2060 air quality goal. We then  
35 discuss the likelihood of the achievement under currently proposed development pathways in  
36 China and foreign countries. As detailed in Method, given the large uncertainties on future  
37 emissions of a country (Rao et al., 2017), we consider three anthropogenic emission scenarios  
38 (low: SSP119, medium: SSP245, and high: SSP370; O'Neill et al., 2014) for foreign countries  
39 and two anthropogenic emission scenarios (current-policy and carbon-neutral; Cheng et al.,  
40 2021b; Tong et al., 2020) for China (Table 1), so as to understand transboundary pollution in  
41 China from the present (2015) to the future (2030 and 2060) under a wide range of plausible  
42 futures. We do not consider the effects of physical climate change (e.g., temperature and  
43 precipitation) on future transboundary pollution because Liu et al. (2021) found that future PM<sub>2.5</sub>  
44 changes in China driven by anthropogenic emission reduction was 7 times more than the changes  
45 due to meteorological fields by 2050 in both SSP126 and SSP585 scenarios. Other studies have

1 also indicted that the impact of climate change on future PM<sub>2.5</sub> were smaller than those of  
2 anthropogenic emissions of pollutants (Hong et al., 2019; Jiang et al., 2013; Silva et al., 2017).

3 For each combination of foreign and Chinese anthropogenic emission scenarios, we simulate  
4 PM<sub>2.5</sub> concentrations at a 0.5° x 0.625° resolution using a chemical transport model (GEOS-  
5 Chem; <http://www.geos-chem.org>) that can represent the complex pollutant emissions and  
6 chemical reactions across a large spatial domain. Then, we correct the systematic bias in the  
7 model simulated PM<sub>2.5</sub> concentrations using a large set of ground-based observations of PM<sub>2.5</sub> in  
8 China; the same correction factor is applied to all present and future concentrations. Finally, we  
9 quantify the transboundary impacts on Chinese PM<sub>2.5</sub> under each emission scenario. We further  
10 quantify the corresponding transboundary impacts on public health in China, measured by  
11 premature deaths, using socio-demographic projections consistent with SSPs and the state-of-  
12 the-art concentration–response relationships (the GEMM model; Burnett et al., 2018).

## 13 14 **2. Method**

15  
16 In this study, we use a set of data and models to investigate future air quality and health burden  
17 in China. Projected air pollutant emissions for foreign countries under SSP-RCP scenarios are  
18 obtained from the International Institute for Applied Systems Analysis (IIASA; Rao et al., 2017;  
19 Riahi et al., 2017) with updates on base year emissions and the harmonization year in this study.  
20 Projected air pollutant emissions for China are developed by Tong et al. (2020) and Cheng et al.  
21 (2021b). Ambient PM<sub>2.5</sub> concentrations under each scenario are simulated by the GEOS-Chem  
22 chemical transport model (<http://www.geos-chem.org>) and further corrected for systematic bias  
23 by a suite of ground-based observations (Xu et al., 2023). PM<sub>2.5</sub>-associated mortalities are  
24 calculated using the Global Exposure Mortality Model (GEMM; Burnett et al., 2018).

### 25 26 **2.1 Scenarios of future anthropogenic air pollutant emissions**

27 Future pollutant emission outcome is a cumulative result of a range of variables including  
28 socio-economic development, technological change, efficiency improvements and policies  
29 directed at pollution control as well as alternative concerns including climate change, energy  
30 access, and agricultural production (Rao et al., 2017). The Shared Socioeconomic Pathways  
31 (SSP) includes 5 five distinctly different pathways about how the future might unfold in terms of  
32 major socioeconomic, demographic, technological, lifestyle, policy and institutional trends  
33 (O'Neill et al., 2014; van Vuuren et al., 2017). We select three scenarios to represent low,  
34 medium and high emission cases: SSP1 - sustainability, SSP2 - middle-of-the-road, SSP3 -  
35 regional rivalry. An assumption about the degree of air pollution control (strong, medium or  
36 weak) is included on top of the baseline pathway. Weak air pollution controls occur in SSP3,  
37 with medium controls in SSP2 and strong controls in SSP1 (Turnock et al., 2020). However, SSP  
38 scenarios do not include explicit climate policies (O'Neill et al., 2020). Instead, the  
39 Representative Concentration Pathways (RCPs) generate climate projections targeting at a range  
40 of climate forcing levels, such as 1.9 W m<sup>-2</sup> (1.5 °C warming), 4.5 W m<sup>-2</sup> (3 °C) and 7.0 W m<sup>-2</sup>  
41 (4 °C) in 2100. Thus, the combination of SSP-RCP scenario framework depicts societal and  
42 climate futures in parallel and explores plausible futures of human activities, the changing  
43 climate and emissions (O'Neill et al., 2020).

1 Here, we use anthropogenic aerosol emissions projected under SSP-RCP scenarios for foreign  
2 countries. We select 3 scenarios to represent low, middle and high air pollutant emissions in  
3 plausible futures: SSP119 (a sustainable development pathway targeting at a rise of the global  
4 mean surface temperature below 1.5 °C from the pre-industrial levels by the end of the century),  
5 SSP245 (a business-as-usual development pathway with 3 °C warming), SSP370 (a regional-  
6 rivalry development pathway with 4 °C warming). Table 1 summarizes the scenario settings in  
7 more detail.

8  
9 The original SSP-RCP anthropogenic emissions future projection starts from 2015. Here, we  
10 update the start year to 2019, which is the latest year that anthropogenic emissions from the  
11 Community Emissions Data System (CEDS; Hoesly et al., 2018) are available. Because CEDS  
12 historical emissions and SSP-RCP future emissions were developed upon different assumptions  
13 and methods, it is important to harmonize the two datasets to ensure smooth transitions between  
14 the two sets of emissions trajectories (Gidden et al., 2019). The update of the harmonization year  
15 from 2015 to 2019 in this study could better represent the actual emission trajectory between  
16 2015 and 2019. We also use the most recently developed CEDS emissions version 2  
17 (<https://data.pnnl.gov/dataset/CEDS-4-21-21>) to harmonize the future SSP-RCP emission  
18 projections as the new emissions can better represent the historical trend of pollutant emissions  
19 ([https://github.com/JGCRI/CEDS/blob/master/documentation/Version\\_comparison\\_figures\\_v\\_2021\\_04\\_21\\_vs\\_v\\_2016\\_07\\_16\(CMIP6\).pdf](https://github.com/JGCRI/CEDS/blob/master/documentation/Version_comparison_figures_v_2021_04_21_vs_v_2016_07_16(CMIP6).pdf)). The final future SSP-RCP emissions for foreign  
20 countries used in this study are presented in Fig. S1a.

22 Future scenarios of anthropogenic pollutant emissions for China were developed by Tong et al.  
23 (2020) and Cheng et al. (2021b). We adopt these emissions instead of SSP-RCP emissions  
24 because Tong et al. (2020) and Cheng et al. (2021b) developed the emissions specifically to meet  
25 China's recent climate goals (i.e., 2030 carbon peak and 2060 carbon neutrality) and used  
26 parameters (i.e., emission factors, energy use, etc.) to reflect the most up-to-date pollutant  
27 control policies (i.e., clean air action since 2013) and technologies in China. Their emissions  
28 have been suggested to better capture China's PM<sub>2.5</sub> concentration decline during 2015–2019  
29 than those driven by SSP-RCP emissions (Cheng et al., 2021a). We select two plausible  
30 scenarios: the current-policy scenario and the carbon-neutral scenario. The current-policy  
31 scenario seeks to achieve China's NDC pledges and the national PM<sub>2.5</sub> air quality goal (i.e. 35 µg  
32 m<sup>-3</sup>) by 2030, elucidating China's future air pollution mitigation pathway towards all the released  
33 and determined upcoming clean air policies since 2015. The carbon-neutral scenario pursues  
34 China's carbon-neutral commitment and the WHO's old PM<sub>2.5</sub> guideline (10 µg m<sup>-3</sup>) by 2060. It  
35 implements the best available end-of-pipe technologies and more stringent pollution control  
36 policies than the current-policy scenario. Future anthropogenic pollutant emissions for China  
37 under these scenarios are developed by firstly simulating China's future energy and  
38 socioeconomic evolution using the Global Change Assessment Model (GCAM-China) and then  
39 translating into pollutant emissions by the Dynamic Projection model for Emissions in China  
40 (DPEC; Tong et al., 2020). More details are summarized in Table 1. The actual future air  
41 pollutant emissions for China used in this study are presented in Fig. S1b.

## 42 **2.2 Simulations of ambient PM<sub>2.5</sub> concentrations**

43

1 We use the GEOS-Chem model to simulate PM<sub>2.5</sub> concentrations in China and other Asian  
2 countries under each emission scenario and year. Detailed descriptions of the model  
3 configurations can be found in Xu et al. (2023). Briefly, we use the Flex-Grid capability of the  
4 GEOS-Chem classic model v13.2.1 to simulate PM<sub>2.5</sub> concentrations over Asia and its adjacent  
5 regions (11° S–60° N, 30°–150° E; covering the entire Asia, eastern Africa and eastern Europe).  
6 We use MERRA-2 assimilated meteorological data provided by the Global Modeling and  
7 Assimilation Office (GMAO) at NASA Goddard Space Flight Center to drive the model. Our  
8 simulation is conducted at a horizontal resolution of 0.5° latitude × 0.625° longitude with 47  
9 vertical levels between the surface and ~ 0.01 hPa. Chemical boundary conditions are taken  
10 from corresponding global simulations under each emission scenario in Table 2 at a resolution of  
11 2° latitude × 2.5° longitude. The regeneration of boundary conditions under each emission  
12 scenario could ensure the inclusion of pollution transported from countries both within and  
13 outside the Flex-Grid domain as transboundary pollution to China. We spin up every simulation  
14 for 1 month to remove the effects of initial conditions.

15  
16 Anthropogenic emissions for the base year (2015) for China are taken from the Multi-  
17 resolution Emission Inventory (MEIC) for 2015 (Zheng et al., 2018b), and for the rest of the  
18 world are taken from the Community Emissions Data System (CEDS) version 2 for 2015  
19 (<https://data.pnnl.gov/dataset/CEDS-4-21-21>). For future simulations, anthropogenic emissions  
20 for China and foreign countries for each scenario are described above and are specified in Table  
21 2. Other emissions are default in GEOS-Chem following Xu et al. (2023) and are fixed in all  
22 present and future scenarios.

23  
24 We conduct simulations for January, April, July and October, and treat the mean of the four  
25 months as annual mean. Our meteorological fields are fixed to 2015 for all scenarios and years to  
26 exclude the influence of climate on the results. More details of our model configurations can be  
27 found in Table 2. We conduct two types of simulations: 1) baseline simulations (simulations with  
28 “Base\_” prefix in Table 2) that include complete anthropogenic emissions for both China and  
29 foreign countries; 2) sensitivity simulations (simulations with “China\_” prefix in Table 2) that  
30 exclude anthropogenic emissions for foreign countries from the baseline simulation. Baseline  
31 simulations calculate PM<sub>2.5</sub> concentrations in China that are driven by both Chinese and foreign  
32 emissions, while sensitivity simulations calculate PM<sub>2.5</sub> concentrations in China that are driven  
33 merely by China’s domestic emissions. The impacts of transboundary pollution on China’s PM<sub>2.5</sub>  
34 is calculated as the difference in China’s PM<sub>2.5</sub> between a baseline simulation and a sensitivity  
35 simulation in a specified year and under a specified Chinese emission scenario.

36  
37 We correct our simulated PM<sub>2.5</sub> concentrations under each scenario, based on a large set of  
38 ground-based observations of PM<sub>2.5</sub>, because our simulated PM<sub>2.5</sub> concentrations are biased high  
39 by roughly 15.7% (Fig. S2). Descriptions and data screening method of our ground-based  
40 observations can be found in Xu et al. (2023). Further evaluations of PM<sub>2.5</sub> composition  
41 concentrations in China and PM<sub>2.5</sub> concentrations in other Asian countries with ground-based  
42 observations are presented in Xu et al. (2023). We correct simulated PM<sub>2.5</sub> concentrations in  
43 anthropogenic concentration-dominated grid cells (where anthropogenic emissions exceed  
44 natural emissions) by observations located in anthropogenic concentration-dominated grid cells.  
45 Similarly, we correct simulated concentrations in natural concentration-dominated grid cells  
46 (where natural emissions exceed anthropogenic emissions) by observations located in natural

1 concentration-dominated grid cells. Natural and anthropogenic concentration-dominated grid  
2 cells in the model are shown in Fig. S3. To isolate anthropogenic concentration-dominated grid  
3 cells, we conduct a sensitivity that excludes natural emissions in China and in foreign countries.  
4 PM<sub>2.5</sub> concentrations from this sensitivity simulation are referred to as natural PM<sub>2.5</sub>  
5 concentrations. The difference between concentrations from the Base\_2015 simulation (with  
6 complete natural and anthropogenic emissions) in Table 2 and the natural concentrations is the  
7 anthropogenic emission-contributed concentrations (anthropogenic concentrations). We scale  
8 anthropogenic concentrations by the average ratio of observed concentrations and simulated  
9 concentrations at each observation site that falls into an anthropogenically dominated grid cell.  
10 Concentrations before and after the correction for anthropogenic and all grid cells (including  
11 natural grid cells) are shown in Fig. S2. The overestimation in both anthropogenic and total  
12 PM<sub>2.5</sub> concentrations is removed after the correction.

13  
14 We calculate sectoral contributions of foreign anthropogenic emissions to China's PM<sub>2.5</sub>  
15 concentrations under China's current-policy and foreign countries' SSP370 scenario at a  
16 resolution of 2° x 2.5° for January in 2030 and 2060. Sectoral contributions are calculated by  
17 taking the difference of a simulation that includes one sector of SSP370 foreign anthropogenic  
18 emissions (agriculture, industry, energy, traffic, residential combustion, solvent use, waste  
19 burning) at a time and a simulation without foreign anthropogenic emissions  
20 ("China\_current\_2030" and "China\_current\_2060" runs in Table 2).

21  
22 We also conduct sensitivity simulations at a resolution of 2° x 2.5° to separate transboundary  
23 contributions to China's future PM<sub>2.5</sub> concentrations from anthropogenic emissions within or  
24 outside the Flex-Grid domain of the model under the current-policy and SSP370 scenario.  
25 Contributions of foreign anthropogenic emissions from within the Flex-Grid domain are  
26 calculated as the difference of a simulation without foreign anthropogenic emissions outside the  
27 Flex-Grid domain and a simulation without foreign anthropogenic emissions over the globe.  
28 Contributions of foreign anthropogenic emissions from outside the Flex-Grid domain are  
29 calculated as the difference between the total foreign contributions and foreign contributions  
30 from Flex-Grid countries.

### 31 32 **2.3 Health impact assessment**

33  
34 We use the GEMM model (Burnett et al., 2018) to estimate premature deaths attributable to  
35 ambient PM<sub>2.5</sub> exposure for noncommunicable diseases (NCDs) and lower respiratory infections  
36 (LRIs) in China under each scenario. GEMM NCD+LRI calculates premature deaths associated  
37 with ambient PM<sub>2.5</sub> exposure (M) for each population subgroup s (by age and gender) in grid cell  
38 g as:

$$39$$
$$40 M_{s,g}(C_g) = B_s \times AF_s(C_g) \times P_g \quad (1)$$
$$41$$

42 where B<sub>s</sub> is the national baseline mortality rate of NCD+LRI for the exposed population  
43 subgroup s. AF<sub>s</sub>(C<sub>g</sub>) is the attributable fraction of NCD+LRI to PM<sub>2.5</sub> exposure at level C<sub>g</sub> for  
44 population subgroup s. P<sub>g</sub> represents the total exposed population in grid cell g. In particular AF  
45 was calculated as AF = (RR - 1)/RR, where RR is the relative risk of NCD+LRI attributable to

1 ambient PM<sub>2.5</sub> exposure. The dependence of RR of NCD+LRI on PM<sub>2.5</sub> concentrations is  
2 calculated as

3  
4 
$$RR = e^{1+e^{\frac{\theta \times \ln(\frac{z}{\alpha} + 1)}{(-\frac{z-\mu}{v})}}}$$
, where  $z = \max(0, C_g - 2.4)$  (2)

5 where  $\theta$ ,  $\alpha$ ,  $\mu$  and  $v$  are fitted parameters of PM<sub>2.5</sub>–mortality relationships. According to the  
6 GEMM model, the RR of NCD+LRI is calculated by age for adults aged from 25 to greater than  
7 85 years in 5-year intervals.

8 In this study, we use the national baseline mortality data by age group and disease type  
9 from the Global Burden of Disease Results Tool 2017 version (GBD 2017; Institute for Health  
10 Metrics and Evaluation, 2017). For future baseline mortality, we use the age-specific baseline  
11 mortality rates projected by the International Futures (IFs) model v7.89 (Hughes et al., 2011).  
12 Population and age structure data for China for 2015 and future years are obtained from the SSP  
13 dataset (Samir and Lutz, 2017; Riahi et al., 2017), because China’s current-policy scenario is  
14 built upon the SSP2 scenario and the carbon-neutral scenario is built upon the SSP1 scenario  
15 (Cheng et al., 2021b; Tong et al., 2020). The gridded population data for China on a  $0.5^\circ \times 0.5^\circ$   
16 spatial resolution for 2015 and future years under each SSP scenario are developed by Huang et  
17 al. (2019). Future cause-and-effect relationship between PM<sub>2.5</sub> exposure and mortality follows  
18 the PM<sub>2.5</sub>-mortality hazard ratio function in the GEMM model, which has been widely used in  
19 the calculation of PM<sub>2.5</sub>-associated mortality in the future (Hong et al., 2019; Liu et al., 2021;  
20 Yang et al., 2022).

21 With the baseline mortality, the population data and the age-structure data, we calculate, grid  
22 cell by grid cell, the age-specific baseline mortality rate under present and future scenarios. The  
23 impacts of transboundary pollution on mortality in China is calculated as the difference between  
24 mortality associated with PM<sub>2.5</sub> simulated by the full anthropogenic emissions for China and  
25 foreign countries, and mortality associated with PM<sub>2.5</sub> simulated by excluding foreign emissions  
26 from the full anthropogenic emissions.

## 27 **3. Results**

### 28 **3.1 Achievability of future air quality goals in China.**

29 In 2015, the national mean population-weighted PM<sub>2.5</sub> over China is about  $48 \mu\text{g m}^{-3}$  (Fig. 1)  
30 after the observation-based correction, consistent with previous studies ( $48\sim 55 \mu\text{g m}^{-3}$ ) that used  
31 various models (Burnett et al., 2018; Cheng et al., 2021a, 2021b; Tang et al., 2022; Zhang and  
32 Cao, 2015). The correction reduces the overestimation in the model by roughly 16% (Fig. S2);  
33 details about the correction approach are described in Method. From 2015 to 2030 and 2060,  
34 there is a remarkable decreasing trend in China’s annual mean population-weighted PM<sub>2.5</sub>  
35 concentrations under plausible futures (Fig. 1). In 2030, achieving the  $35 \mu\text{g m}^{-3}$  goal on a  
36 national average level would be feasible even under the fossil fuel-intensive pathways in China  
37 (current-policy) and foreign countries (SSP370), yet the most polluted provinces might not be  
38 able to achieve the goal under such scenarios (upper whiskers in Fig. 1). In 2060, achieving the  
39  
40



1 WHO AQG goal of  $5 \mu\text{g m}^{-3}$  would be highly unlikely under any emission pathway analyzed in  
2 this study (Fig. 1).

3  
4 Considering air quality goals at the city level in China ( $35 \mu\text{g m}^{-3}$  in 2015 and 2030;  $5 \mu\text{g m}^{-3}$  in  
5 2060), the fraction of cities achieving air quality goals increases considerably as China and  
6 foreign countries transition from the fossil fuel-intensive to the low-carbon pathway, yet the  
7 achievement is not fully attainable in all cities. In 2015, only 30% of 365 cities in the national  
8  $\text{PM}_{2.5}$  monitoring network has an annual mean population-weighted  $\text{PM}_{2.5}$  below the  $35 \mu\text{g m}^{-3}$   
9 threshold (Fig. 2a). In 2030, the percentage increases considerably (Fig. 2b). Under the current-  
10 policy emission pathway in China (the current clean air policies and Nationally Determined  
11 Contribution pledges), roughly 65% of Chinese cities (average of the three foreign scenarios) is  
12 able to achieve the goal, doubling the percentage in 2015. Under the carbon-neutral emission  
13 pathway in China (stringent clean air policies and carbon neutrality commitments), the  
14 percentage further increases to about 92% (average of the three foreign scenarios). However,  
15 even the cleanest emission pathway in both China and foreign countries cannot allow all cities to  
16 attain the  $35 \mu\text{g m}^{-3}$  goal. In 2060 (Fig. 2c), the WHO AQG goal of  $5 \mu\text{g m}^{-3}$  is not achievable  
17 for the majority of cities ( $\geq 75\%$ ) in China. Emission pathways adopted by foreign countries can  
18 affect up to 6% of cities achieving the AQG goal, and the influence could be even larger over  
19 border regions (as will be shown in Fig. 3).

### 20 21 **3.2 Transboundary impacts on $\text{PM}_{2.5}$ concentrations in China.**

22  
23 In 2015, transboundary pollution contributes about  $3.8 \mu\text{g m}^{-3}$  population-weighted  $\text{PM}_{2.5}$  to  
24 China (Fig. 3a), accounting for roughly 8% of the total population-weighted  $\text{PM}_{2.5}$  (Fig. 4). In  
25 the future, transboundary pollution becomes increasingly important in China, as the share of  
26 transboundary pollution in China's total population-weighted  $\text{PM}_{2.5}$  increases to 12%~22% in  
27 2060 (Fig. 4).

28  
29 For future  $\text{PM}_{2.5}$  concentrations, the contribution of transboundary pollution to  $\text{PM}_{2.5}$  in China  
30 decreases as foreign countries and China undergo the low-carbon pathways (Fig. 3a). In 2030,  
31 under the current-policy scenario in China, transboundary contributions to  $\text{PM}_{2.5}$  in China would  
32 be reduced by  $1.2 \mu\text{g m}^{-3}$  (29%) as foreign countries transition from the fossil fuel-intensive  
33 (SSP370) to the low-carbon (SSP119) scenario. By 2060, the difference would be increased to  
34  $1.8 \mu\text{g m}^{-3}$  (45%). The transboundary pollution will also depend on Chinese domestic emissions,  
35 because of their chemical interactions with foreign-transported pollution (Xu et al., 2023). In  
36 2030, under the SSP370 scenario in foreign countries, transboundary contributions to  $\text{PM}_{2.5}$  in  
37 China could be reduced by  $0.6 \mu\text{g m}^{-3}$  (14%) as China transitions from the current-policy to the  
38 carbon-neutral scenario. In 2060, the  $\text{PM}_{2.5}$  reduction could be increased to  $1.8 \mu\text{g m}^{-3}$  (45%).

39  
40 The direct atmospheric transport and the chemical interactions play different roles in  
41 transboundary pollution over different regions in China. Over North China (outlined in Fig. 3c-  
42 o), the influence of chemical interactions on transboundary pollution is prominent. Xu et al.  
43 (2023) found that foreign-transported NMVOCs could enhance the atmospheric oxidizing  
44 capacity and facilitated the oxidation of Chinese nitrogen oxides ( $\text{NO}_x$ ) to form nitrate over  
45 North China. This feature persists in the future. Due to chemical interactions, North China is the  
46 region that is the most strongly affected by transboundary pollution in most scenarios. In 2030,

1 transboundary pollution contributes 3 to 10  $\mu\text{g m}^{-3}$   $\text{PM}_{2.5}$  concentrations to North China, even  
2 under China's carbon-neutral scenario (Fig. 3b and Fig. 3i-l), greatly increasing the difficulty for  
3 this region to achieve the 35  $\mu\text{g m}^{-3}$  goal. Further reduction of transboundary pollution on North  
4 China requires the reduction of not only foreign anthropogenic emissions but also China's  
5 domestic emission. For example, in 2060, if China adopts the current-policy pathway (Fig. 3f-h),  
6 achieving the WHO AQG goal of 5  $\mu\text{g m}^{-3}$  would be highly unlikely for the majority of North  
7 China, as transboundary pollution alone would contribute roughly 3~10  $\mu\text{g m}^{-3}$  of  $\text{PM}_{2.5}$   
8 concentration (Fig. 3b and Fig. 3f-h). Alternatively, adopting the carbon-neutral rather than the  
9 current-policy pathway in China reduces roughly 41%~58% of transboundary  $\text{PM}_{2.5}$  over North  
10 China in 2060 (Fig. 3b), making it possible for the region to achieve the 5  $\mu\text{g m}^{-3}$  goal. Thus, a  
11 low emission pathway for China has a large co-benefit on reducing transboundary pollution  
12 exerted upon its populous northern area by reducing the aforementioned chemical interactions.  
13

14 The western border provinces of Yunnan and Xinjiang (denoted in Fig. 3c-o) are also  
15 influenced substantially by transboundary pollution, with the magnitude of transboundary  
16 pollution determined predominantly by foreign (but not Chinese) emission pathways. For  
17 example, in 2060, under China's carbon-neutral emission pathway, transboundary  $\text{PM}_{2.5}$  over  
18 Yunnan decreases from 6  $\mu\text{g m}^{-3}$  to 3  $\mu\text{g m}^{-3}$  (a 50% reduction) as the foreign pathway switches  
19 from fossil fuel-intensive (SSP370; Fig. 3m) to low-carbon (SSP119; Fig. 3o), reflecting the  
20 considerable impact of Southern Asian pollution to China (Jiang et al., 2013). Over Xinjiang, the  
21 transboundary pollution driven predominantly by direct atmospheric transport can reach 6  $\mu\text{g m}^{-3}$   
22 in many scenarios (i.e., current-policy plus SSP370 for 2030 and 2060; and carbon-neutral plus  
23 SSP245 for 2030), indicating the important influence of anthropogenic emissions from Central  
24 Asia in the future that has hardly been investigated in previous studies  
25

26 As transboundary contributions are the largest under China's current-policy and foreign  
27 countries' SSP370 emission pathways (Fig. 3a), we further investigate major emission sectors in  
28 foreign countries that contribute to China's future  $\text{PM}_{2.5}$  concentrations. We take January as an  
29 example because Xu et al. (2023) found that transboundary contributions to China's  $\text{PM}_{2.5}$  were  
30 the largest in January. Fig. S4 reveals that agriculture, energy, industry, transportation and  
31 residential combustion emissions in foreign countries are major sources of transboundary  
32 pollution to China's national average future  $\text{PM}_{2.5}$  concentrations in 2030 and 2060 January, and  
33 that these sectors contribute roughly evenly (15%~23% for each source). There is pronounced  
34 spatial heterogeneity in the source attribution. Solvent use emissions in foreign countries make  
35 roughly 3 times larger contributions to North China (18% in 2030 and 2060) than to the national  
36 average (6%-7% in 2030 and 2060). This is in line with the result of Xu et al. (2023) that the  
37 NMVOCs are the primary drivers of transboundary  $\text{PM}_{2.5}$  over North China. Residential  
38 combustion in foreign countries contributes about 16% in 2030 and 10% 2060 more  $\text{PM}_{2.5}$  to  
39 Yunnan province than to the entire China, as Yunnan is mostly affected by emissions from South  
40 Asia (Jiang et al., 2013) where residential combustion is intensive (McDuffie et al., 2020).  
41

42 Transboundary pollution to China could arise from emissions of countries within and outside  
43 the Flex-Grid domain of the model. Thus, we conduct a sensitivity simulation driven by China's  
44 current-policy and foreign countries' SSP370 emissions in 2030 and 2060 as an example to  
45 explore the relative importance of anthropogenic emissions from countries within and outside the  
46 Flex-Grid domain for transboundary pollution to China. Fig. S5 shows that anthropogenic

1 emissions from countries within the Flex-Grid domain dominate the transboundary pollution to  
2 China in all seasons (>80%), and their contributions remain in 2030 and 2060. Contributions of  
3 anthropogenic emissions from countries outside the Flex-Grid domain is larger in April and  
4 October when westerly winds prevail and wet deposition is not as strong as in July (Jiang et al.,  
5 2013; Leibensperger et al., 2011; Ni et al., 2018). North China is more influenced by emissions  
6 outside the Flex-Grid domain (13% in April and 24% in October) than China on average is (7%  
7 in April and 13% in October), indicating that North China is more influenced by emissions  
8 outside Asia, such as Europe and North America. Leibensperger et al. (2011) have also found  
9 that NO<sub>x</sub> emission in the US could lead to an enhancement of PM<sub>2.5</sub> over North China.

### 10 3.3 Health threats by transboundary pollution in China.

12 Figure 5 shows our estimated PM<sub>2.5</sub>-associated premature deaths in 2015 and the future. Our  
13 estimated PM<sub>2.5</sub>-associated premature deaths in China in 2015 (2.03 million) is comparable with  
14 other studies (2 to 2.4 million; Burnett et al., 2018; Geng et al., 2021; Tang et al., 2022). Our  
15 estimated increasing trend of premature deaths in China from SSP370 to SSP119 is also  
16 consistent in previous works (Tang et al., 2022; Yang et al., 2022), which is driven primarily by  
17 population ageing (Fig. S6). Our estimated premature deaths in 2030 and 2060 under the current-  
18 policy emission scenario (2.68 to 2.82 million for 2030; 1.8 to 2.05 million for 2060) is  
19 substantially lower than those in Tang et al. (2022) (3.5 to 4 million for 2030; 6.5 to 7.5 million  
20 for 2060). The difference is primarily because Tang et al. (2022) fixed the baseline mortality  
21 rates in future years at the 2015 level, while our future baseline mortality rates from the IFs are  
22 projected on the basis of income, education and technology advancement, and other factors  
23 (Hughes et al., 2011).

24 The PM<sub>2.5</sub> concentrations contributed by transboundary pollution can lead to an extra number  
25 of people (i.e., excess population) exposed to PM<sub>2.5</sub> concentrations above the targeted air quality  
26 levels in China (35 µg m<sup>-3</sup> in 2015 and 2030; and 5 µg m<sup>-3</sup> in 2060), which may lead to potential  
27 health threats. In the future, the number of excess population due to transboundary pollution  
28 depends on both foreign and China's emission pathways (Fig. 6a). In 2030, with Chinese  
29 emissions following the carbon-neutral scenario, adopting the low-carbon (SSP119) rather than  
30 the fossil fuel-intensive pathway (SSP370) in foreign countries could avoid 10 million Chinese  
31 people from being exposed to PM<sub>2.5</sub> concentrations above 35 µg m<sup>-3</sup>. In 2060, 5 million people  
32 could be avoided from being exposed to PM<sub>2.5</sub> concentrations above 5 µg m<sup>-3</sup> if the foreign  
33 scenario switches from SSP370 to SSP119. These results indicate remarkable health benefits that  
34 the low-carbon emission pathway in foreign countries would bring to China.

35  
36 For a given future year and foreign emission scenario, the excess population due to  
37 transboundary pollution tends to be larger when China adopts the carbon-neutral pathway than  
38 when China adopts the current-policy pathway (Fig. 6a). This reflects the increasing influence of  
39 transboundary pollution on air quality in China as China's overall PM<sub>2.5</sub> concentrations drop  
40 sharply towards the air quality goals under the carbon-neutral pathway (Fig. 1).

41  
42 Another measure of health threat that transboundary pollution could exert upon China is PM<sub>2.5</sub>-  
43 related premature deaths. As shown in Fig. 6b, there is an obvious decreasing trend of  
44 transboundary-contributed PM<sub>2.5</sub>-related premature deaths in China as foreign countries and

1 China transition from the fossil fuel-intensive to the low-carbon pathway. In 2030, adopting the  
2 low-carbon pathway (SSP119) in foreign countries would avoid 41% (178,000 under China's  
3 carbon-neutral pathway) to 45% (207,000 under China's current-policy pathway) of premature  
4 deaths that would occur under the fossil fuel-intensive pathway (SSP370) in foreign countries. In  
5 2060, the avoidance would be as large as 76% (270,000 under China's current-policy pathway)  
6 to 91% (63,000 under China's carbon-neutral pathway). In addition, China's low carbon  
7 emission pathway could also bring considerable health benefits through reducing the chemical  
8 interaction-related transboundary pollution and associated premature deaths in China. In 2030,  
9 adopting the carbon-neutral pathway in China would avoid 99,000 (SSP245) to 211,000  
10 (SSP119) people from transboundary pollution-associated premature deaths relative to adopting  
11 the current-policy emission pathway. In 2060, the avoided deaths would be 76,000 (SSP119) to  
12 283,000 (SSP370). These findings highlight the considerable health benefits to China if foreign  
13 countries and China could adopt the low-carbon emission pathways coincidentally.

14

#### 15 **4. Discussion**

16

17 This study reveals the increasingly important role that transboundary pollution would play in  
18 the achievement of future air quality goals and the protection of public health in China. The  
19 magnitude of transboundary pollution depends on both Chinese and foreign emissions, given the  
20 direct pollution transport and the indirect impact through chemical interactions between  
21 transported and China's locally emitted pollutants. Adopting the low-carbon (SSP119) instead of  
22 the fossil fuel-intensive (SSP370) pathway in foreign countries would avoid millions of Chinese  
23 people (Fig. 6a) from being exposed to PM<sub>2.5</sub> concentrations above the targeted air quality levels  
24 in 2030 (35 μg m<sup>-3</sup>) and 2060 (5 μg m<sup>-3</sup>), and would avoid 63,000~270,000 of transboundary  
25 PM<sub>2.5</sub>-associated mortality in China in 2060 (Fig. 6b). Adopting the carbon-neutral instead of  
26 current-policy pathway in China would avoid 76,000~283,000 premature mortality associated  
27 with transboundary pollution in 2060 (Fig. 6b). If China and foreign countries undergo the low-  
28 carbon pathways coincidentally, transboundary pollution in China would be reduced by 63%  
29 relative to adopting a fossil fuel-intensive emission pathway in both regions (Fig. 4a), and could  
30 avoid 386,000 premature deaths in China (Fig. 6b). Cutting foreign emissions are particularly  
31 effective at reducing transboundary pollution upon the western border provinces of Yunnan and  
32 Xinjiang that are dominated by direct transport. Fully achieving the WHO AQG goal of 5 μg m<sup>-3</sup>  
33 over the populous North China would be possible only when both China and foreign countries  
34 adopt the low-carbon pathways (carbon-neutral and SSP119, respectively).

35

36 The importance of transboundary pollution is not confined in China. In the future, significant  
37 emission changes are expected in many developing countries, affecting air quality locally and in  
38 the downwind regions. These developing countries are often financially and/or technologically  
39 less capable to control emissions by themselves. Thus, enhanced external aids would be essential  
40 for these developing countries to undergo a low-carbon development in the future, which in turn  
41 would benefit air quality and public health of the entire globe. External aids could be done  
42 through global cooperation programs such as the Paris Agreement or through inter-regional  
43 collaboration such as the Belt and Road Initiative.

44

45 Uncertainties arise from several factors in this study. The future development pathway of a  
46 country is highly uncertain, leading to a wide spread of projected emission trajectories in the

1 future. We thereby use a set of emission projections to represent the plausible range of future  
2 emissions. The simulation of PM<sub>2.5</sub> is subject to uncertainties in aerosol chemical and physical  
3 processes, such as the wet deposition of nitrate (Luo et al., 2020) and the simplified secondary  
4 organic aerosol formation scheme (Pai et al., 2020). Our correction to the simulated PM<sub>2.5</sub>  
5 concentrations using ground-based observations could reduce the uncertainty by 15% to 18%  
6 (Fig. S2). Future population and age structure change are projected based on their historical  
7 relationships with GDP and urbanization (O'Neill et al., 2020; Riahi et al., 2017). Thus, they  
8 may introduce biases if the future development of global GDP and urbanization deviates from  
9 the historical path (e.g., due to the emergence of anti-globalization (Dür et al., 2020) and regional  
10 rivalry (O'Neill et al., 2014; van Vuuren et al., 2014). There are additional uncertainties from  
11 PM<sub>2.5</sub>-related death estimates due to the limited epidemiology evidence and statistical estimation  
12 of the GEMM model, such as the influences of particulate species and size on health outcomes  
13 (Burnett et al., 2018). We estimate the overall uncertainties of PM<sub>2.5</sub>-related death in each  
14 scenario as 95% CI in the main text. Besides, we do not consider potential influences of climate  
15 change and the change of natural emissions on PM<sub>2.5</sub> and transboundary pollution, yet their  
16 influences are found small compared to the influence of anthropogenic emissions (Hong et al.,  
17 2019; Jiang et al., 2013; Liu et al., 2021; Silva et al., 2017).

### 18 19 **Data availability**

20  
21 The global SSP-RCP emissions data for 2015 and future scenarios, and the area-weighted  
22 PM<sub>2.5</sub> concentrations in China for 2015 and future scenarios are available upon request to the  
23 corresponding author. All other data used in this study are publicly available and can be  
24 downloaded from the following links. (1) China's future emission scenarios 2015–  
25 2060: <http://www.meicmodel.org/dataset-dpec.html>. (2) Chinese future population data:  
26 [https://www.scidb.cn/en/detail?dataSetId=73c1ddbd79e54638bd0ca2a6bd48e3ff&dataSetType=](https://www.scidb.cn/en/detail?dataSetId=73c1ddbd79e54638bd0ca2a6bd48e3ff&dataSetType=organization)  
27 [organization](https://www.scidb.cn/en/detail?dataSetId=73c1ddbd79e54638bd0ca2a6bd48e3ff&dataSetType=organization). (3) 2015 baseline mortality rate: <https://gbd2017.healthdata.org/gbd-search/>. (4)  
28 Future baseline mortality rate projection: [https://www.ifs.du.edu/IFs/frm\\_MortCohorts/](https://www.ifs.du.edu/IFs/frm_MortCohorts/).

### 29 30 **Code availability**

31  
32 The GEOS-Chem model v13.2.1 source code used for PM<sub>2.5</sub> concentration simulations is  
33 available at: <http://www.geos-chem.org>. The SSP-RCP emission harmonization source code is  
34 available at <http://software.ene.iiasa.ac.at/aneris/>. All computer codes generated during this study  
35 are available from the corresponding authors upon reasonable request.

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38  
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### 42 43 **Author contributions**

44  
45 J.L. led the study. J.X. and J.L. designed the study. J.X. performed the model simulations and  
46 conducted the data analysis. D.T. provided China's future emissions data. L.C. provided SSP-

1 RCP emission harmonization and health impact assessment methods. J.X. wrote the manuscript  
2 with inputs from J.L. All authors commented on the manuscript.

### 5 **Competing interests**

7 The authors declare that they have no conflict of interest.

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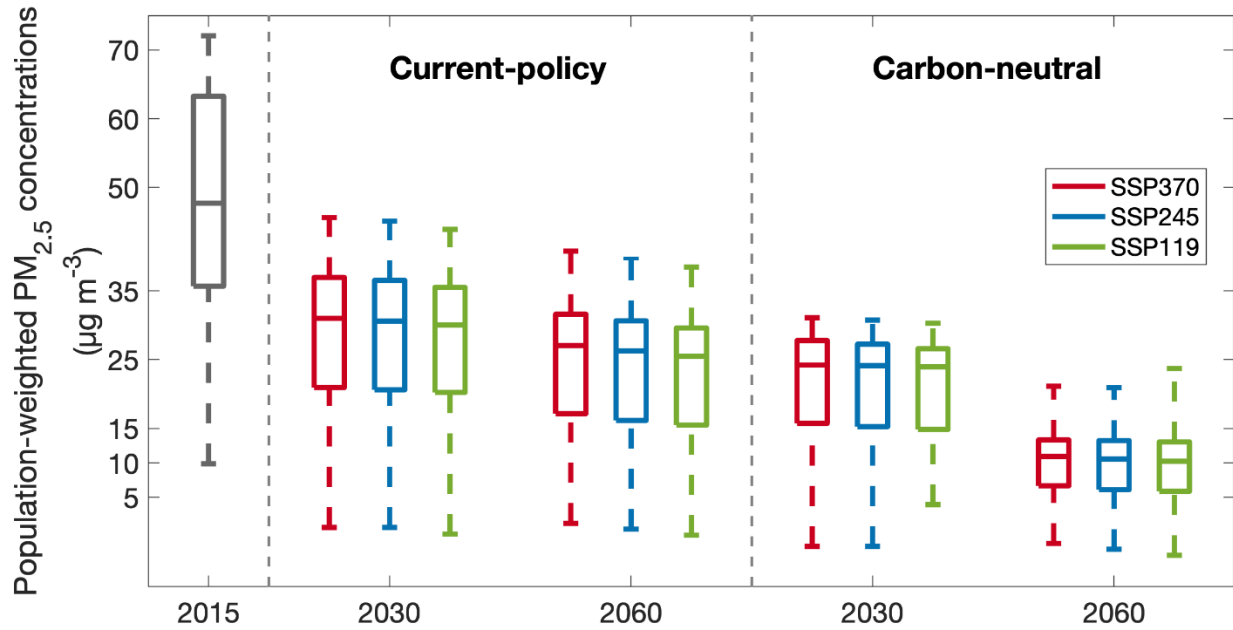
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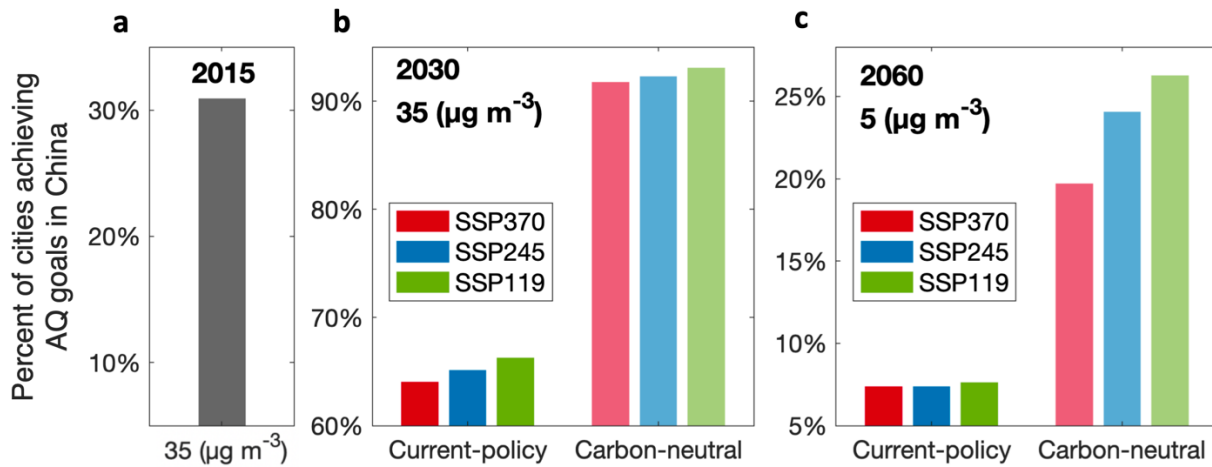
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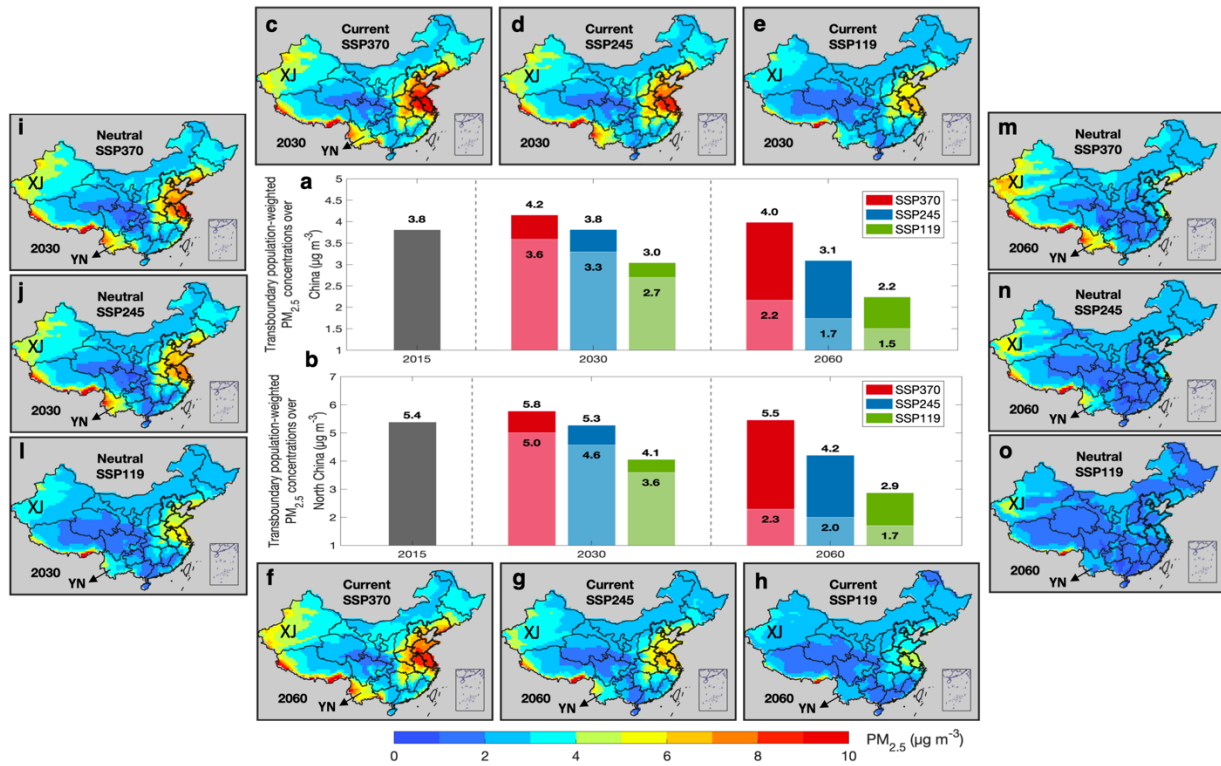
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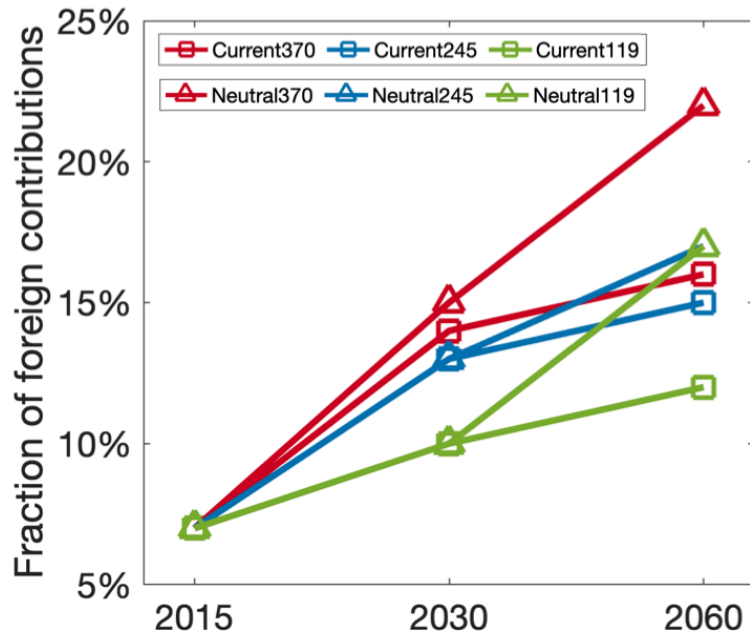
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 2 **Figure 1.** Population-weighted PM<sub>2.5</sub> concentrations over China. Box-and-whisker plots  
 3 represent 5th, 25th, 75th, and 95th percentiles of provincial population-weighted PM<sub>2.5</sub>  
 4 concentrations in 2015, 2030 and 2060. Lines in the middle of each box represent the national  
 5 mean population-weighted PM<sub>2.5</sub> concentrations. Future emission scenarios in China are labeled  
 6 as text at the top and in foreign countries are represented by colors according to the legend.  
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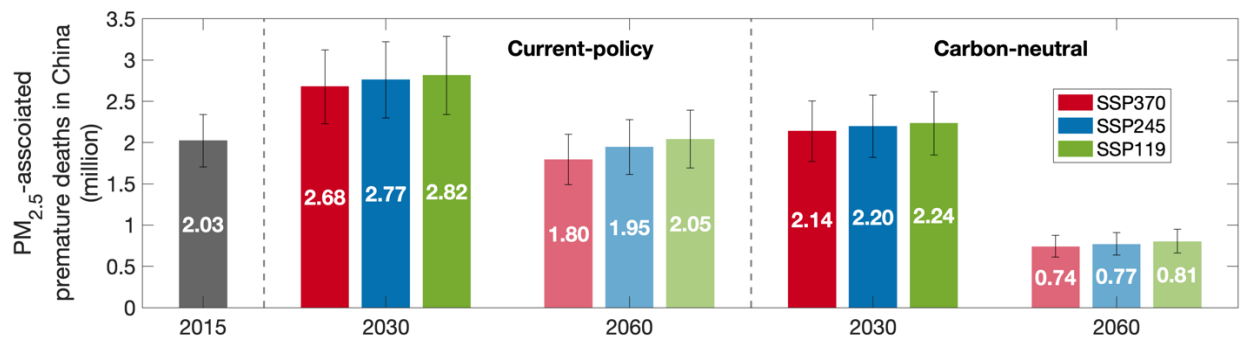
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 10 **Figure 2.** Percent of cities achieving air quality goals in China. **(a)** Percent of cities in China  
 11 with an annual mean population-weighted PM<sub>2.5</sub> concentration below 35 µg m<sup>-3</sup> in 2015. **(b-c)**  
 12 Percent of cities with an annual mean population-weighted PM<sub>2.5</sub> concentration achieving the 35  
 13 µg m<sup>-3</sup> goal in 2030 (b) and the 5 µg m<sup>-3</sup> goal in 2060 (c). PM<sub>2.5</sub> concentrations are simulated  
 14 under different future emission scenarios in China (current-policy and carbon-neutral) and  
 15 foreign countries (represented by colors following the legend).  
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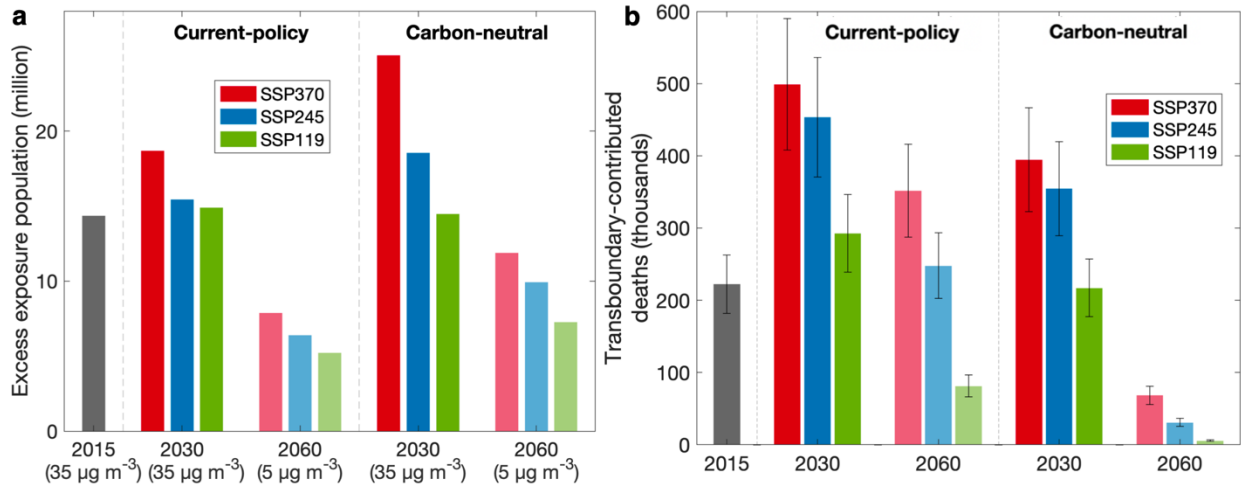
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3 **Figure 3.** Contributions of transboundary pollution to PM<sub>2.5</sub> concentrations over China. (a)  
4 Transboundary contributions to national annual mean population-weighted PM<sub>2.5</sub> in China in  
5 2015, 2030 and 2060. Future scenarios are estimated by different emission scenarios in China  
6 represented by light (carbon-neutral scenario) and dark shadings (current-policy scenario), along  
7 with different emission scenarios in other countries (SSP370, SSP245, SSP119) represented by  
8 colors according to the legend. Text on top of each bar represents the transboundary-contributed  
9 population-weighted PM<sub>2.5</sub> under China's current-policy emission scenarios. Text in the light  
10 shading of each bar represents the transboundary-contributed population-weighted PM<sub>2.5</sub> under  
11 China's carbon-neutral emission scenarios. (b) Same as a, but for North China. (c-o) Spatial  
12 distributions of transboundary-contributed annual mean PM<sub>2.5</sub> concentrations over China in 2030  
13 and 2060 under different emission scenarios in China and in other countries. YN represents  
14 Yunnan province. XJ represents Xinjiang province. North China is outlined by thick black lines.  
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3 **Figure 4.** Fractional transboundary contributions to PM<sub>2.5</sub> concentrations over China. The  
4 percentage fraction of transboundary-contributed PM<sub>2.5</sub> in China's total population-weighted  
5 PM<sub>2.5</sub> in 2015, 2030 and 2060. Future anthropogenic emission scenarios are represented by  
6 different colors and markers following the legend.  
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11 **Figure 5.** PM<sub>2.5</sub>-associated premature deaths in China. Total PM<sub>2.5</sub>-associated premature deaths  
12 in China for 2015, 2030 and 2060 under each emission scenario in China (denoted as text at the  
13 top) and in other countries (represented by colors in the legend). Numbers denote the estimated  
14 deaths in each scenario. Error bars represent the 95% confidence interval of the RR function.  
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**Figure 6.** Potential health threats associated with transboundary pollution in China. **(a)** Population exposed to an annual mean population-weighted PM<sub>2.5</sub> concentration above the goals (35 μg m<sup>-3</sup> for 2015 and 2030; 5 μg m<sup>-3</sup> for 2060) due to transboundary contributions of PM<sub>2.5</sub> in China under different emission scenarios in China (denoted as text at the top) and in other countries (represented by colors in the legend). **(b)** Transboundary-contributed PM<sub>2.5</sub>-related premature mortality in China under different emission scenarios in China (denoted as text at the top) and in other countries (represented by colors in the legend). Error bars represent the 95% confidence interval of the RR function.

1 **Table 1.** Description of future scenario settings

Scenarios	Socioeconomic pathway	Climate target	Pollution control strength	PM <sub>2.5</sub> emission level	Key features
Foreign					
SSP119	Sustainability	1.9 W m <sup>-2</sup> (1.5 °C)	Strong	Low	Strong economic growth via sustainable pathway. Incomes increase substantially and inequality within and between countries is greatly decreased. Significantly lower demand for energy- and resource-intensive agricultural commodities. Effective pollution controls result in substantial reductions in air pollutant emissions.
SSP245	Middle-of-the-road	4.5 W m <sup>-2</sup> (3 °C)	Medium	Medium	An intermediate case between SSP1 and SSP3.
SSP370	Regional rivalry	7.0 W m <sup>-2</sup> (~4 °C)	Weak	High	High inequity between countries. GDP growth is low and concentrated in current high-income nations, while population growth is focused in low- and middle- income countries. Energy system is coal-intensive. The implementation of pollution controls is delayed and less effective.
China					
Current-policy	Middle-of-the-road (SSP2)	4.5 W m <sup>-2</sup> (3 °C)	Medium	Medium	Achieve China's Nationally Determined Contribution (NDC) pledges and the national PM <sub>2.5</sub> air quality standard (i.e. 35 µg m <sup>-3</sup> ) by 2030, elucidating China's future air pollution mitigation pathway towards all the released and determined upcoming clean air policies since 2015.
Carbon-neutral	Sustainability (SSP1)	Net-zero CO <sub>2</sub> emissions in 2060	Strong	Low	Pursue China's carbon-neutral commitment and the WHO's old PM <sub>2.5</sub> guideline (10 µg m <sup>-3</sup> ) by 2060. It implements the best available end-of-pipe technologies and more stringent pollution control policies than the current-policy.

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1 **Table 2.** Simulation configurations

Simulation type	Anthropogenic Emissions		Emission year	Met field year
	China	Foreign countries		
Base_2015 China_2015	MEIC MEIC	CEDS None	2015	2015
Base_current_SSP119_2030 Base_current_SSP245_2030 Base_current_SSP370_2030 China_current_2030	Current-policy Current-policy Current-policy Current-policy	SSP119 SSP245 SSP370 None	2030	
Base_neutral_SSP119_2030 Base_neutral_SSP245_2030 Base_neutral_SSP370_2030 China_neutral_2030	Carbon-neutral Carbon-neutral Carbon-neutral Carbon-neutral	SSP119 SSP245 SSP370 None		
Base_current_SSP119_2060 Base_current_SSP245_2060 Base_current_SSP370_2060 China_current_2060	Current-policy Current-policy Current-policy Current-policy	SSP119 SSP245 SSP370 None	2060	
Base_neutral_SSP119_2060 Base_neutral_SSP245_2060 Base_neutral_SSP370_2060 China_neutral_2060	Carbon-neutral Carbon-neutral Carbon-neutral Carbon-neutral	SSP119 SSP245 SSP370 None		

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1 **Supplementary Information for**  
2 **The underappreciated role of transboundary pollution in future air quality and health**  
3 **improvements in China**

4  
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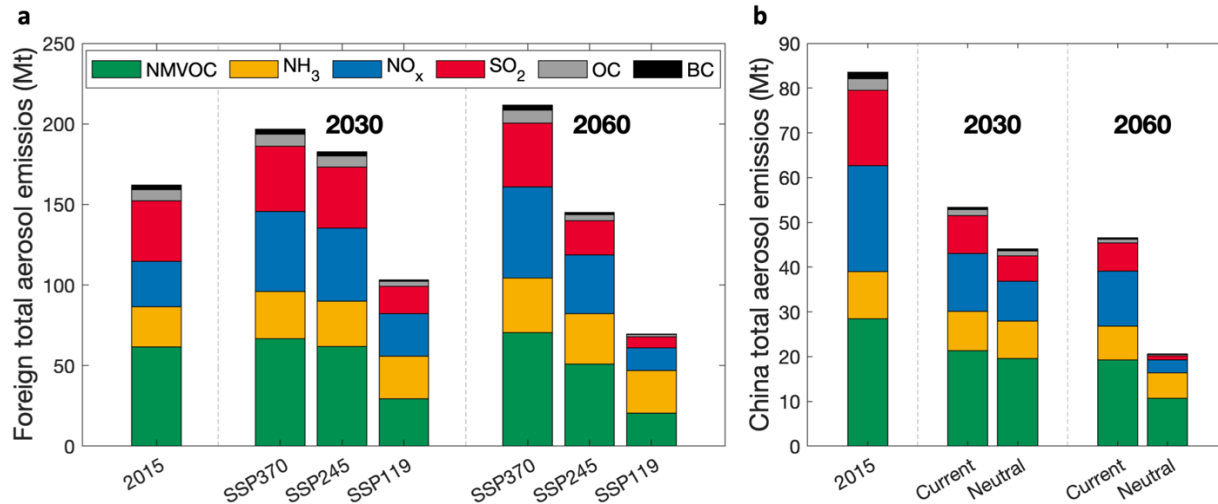
9 <sup>2</sup>Department of Earth System Science, Ministry of Education Key Laboratory for Earth System  
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12 Correspondence: Jintai Lin (linjt@pku.edu.cn).

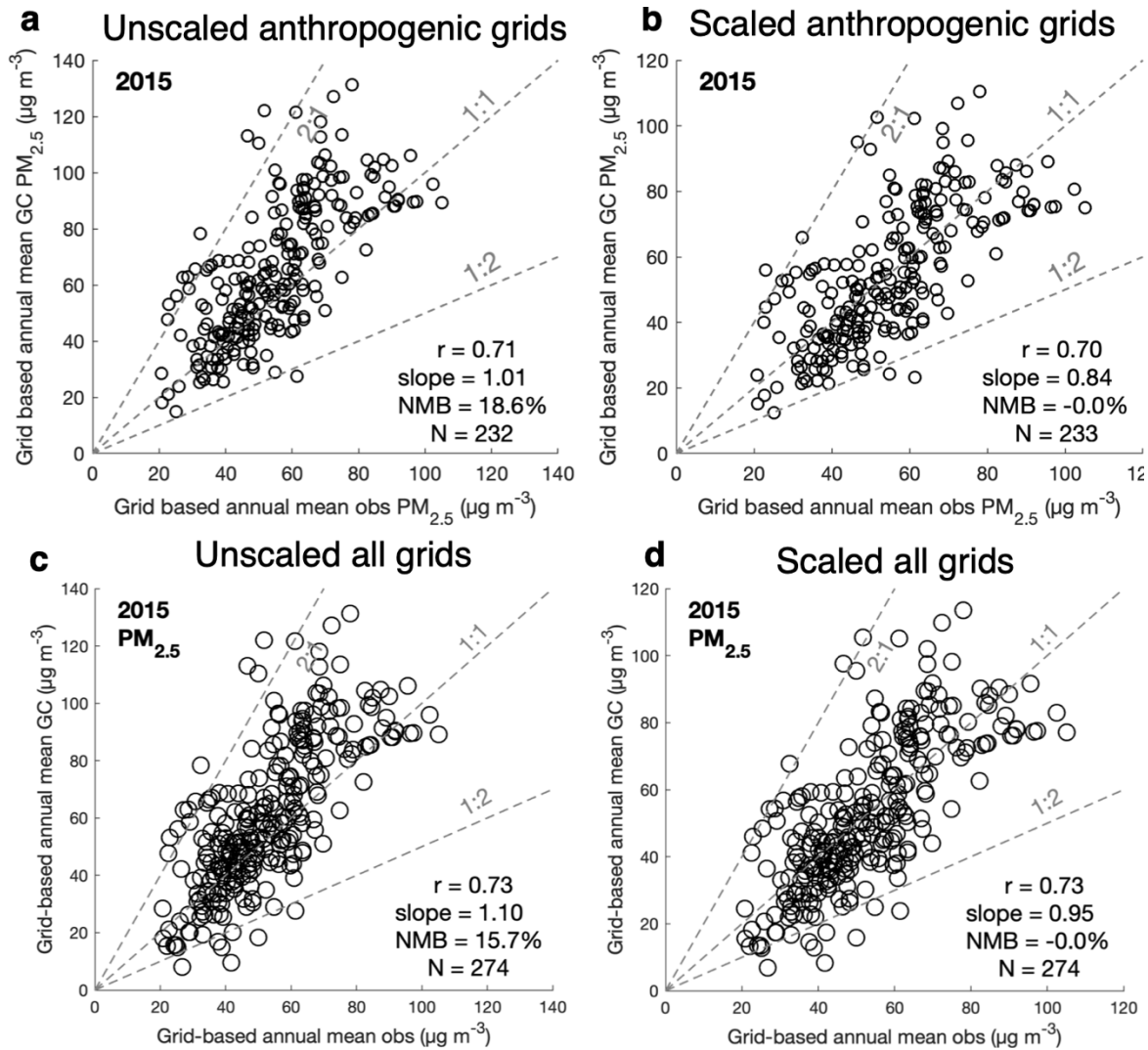
13 This Supplemental Information document presents the additional data to support the primary  
14 document.

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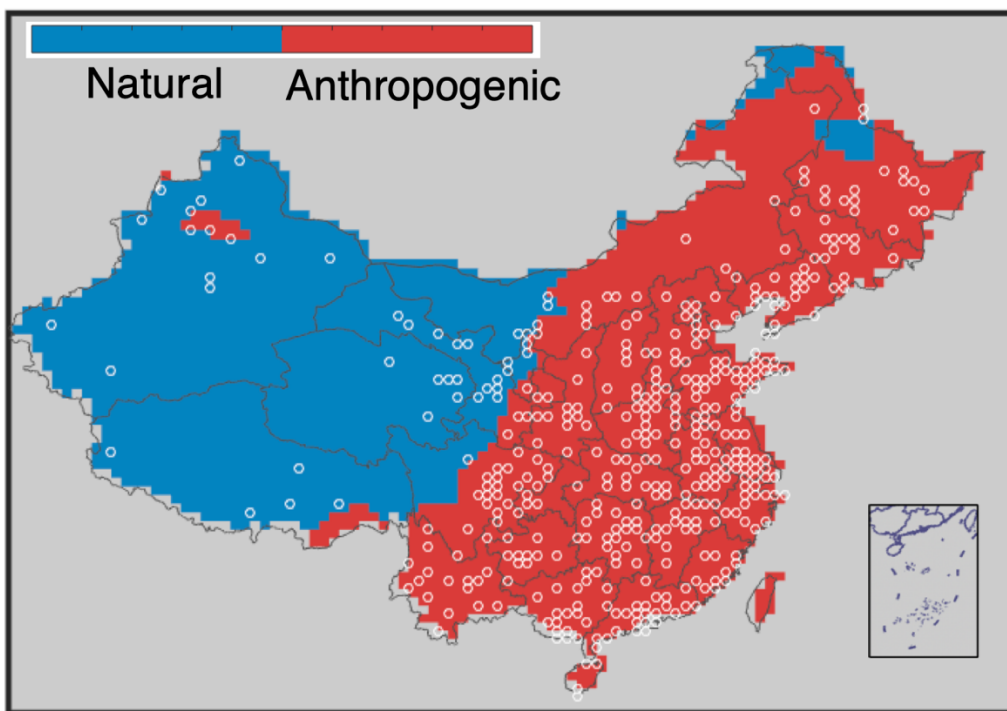


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 2 **Figure S1.** Future emissions of PM<sub>2.5</sub>-related pollutants in China and in other countries. **(a)**  
 3 Emissions for foreign countries within our Flex-Grid simulation domain (11° S–60° N, 30°–150°  
 4 E) for future years are projected under SSP-RCP scenarios, with updates on base year emissions  
 5 and the harmonization year in this study. Colors represent the emissions of different pollutants  
 6 following the legend. **(b)** Emissions for China for future years are projected with the current-  
 7 policy scenario and the carbon-neutral scenario.  
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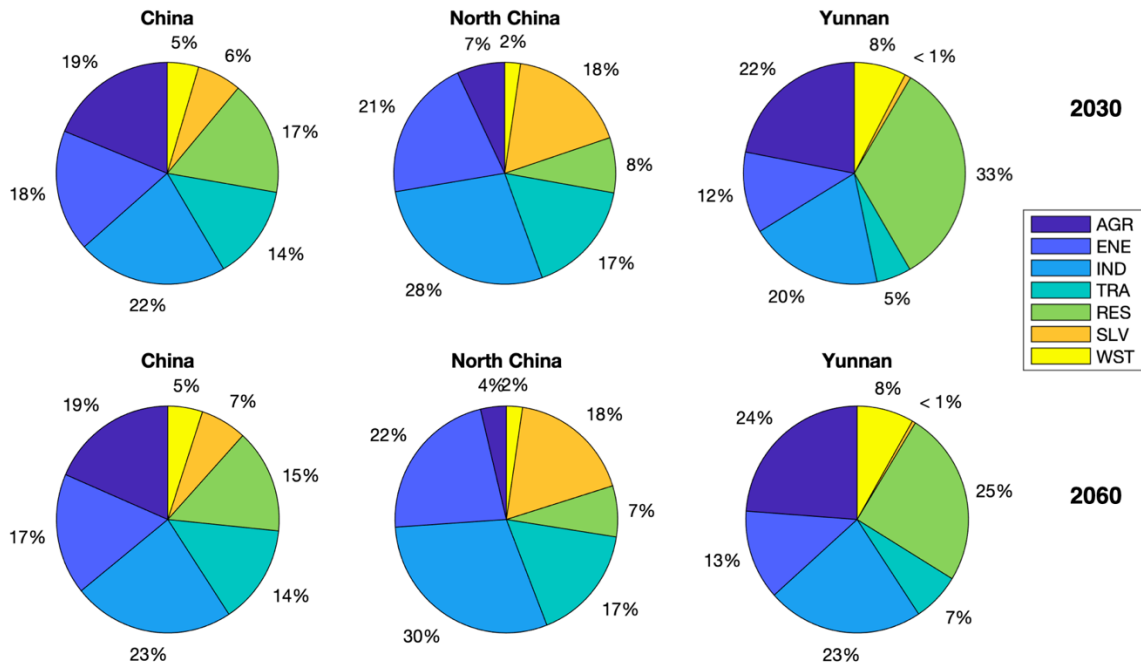


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**Figure S2.** Correction and evaluation of simulated  $PM_{2.5}$  over China. **(a)** Comparison of the observed and unscaled simulated annual mean concentrations  $PM_{2.5}$  at collocated grids for 2015. Both the observed and the simulated  $PM_{2.5}$  were in grids where anthropogenic  $PM_{2.5}$  exceeds natural  $PM_{2.5}$ .  $r$  refers to the correlation coefficient. NMB refers to normalized mean bias.  $N$  indicates the number of grids shown in the figure. **(b)** Same as (a), but for scaled simulated  $PM_{2.5}$  concentrations to remove the systematic bias. **(c)** Same as (a), but for all grids (including both anthropogenic pollution-dominated and natural pollution-dominated grids). **(d)** Same as (b), but for all grids.

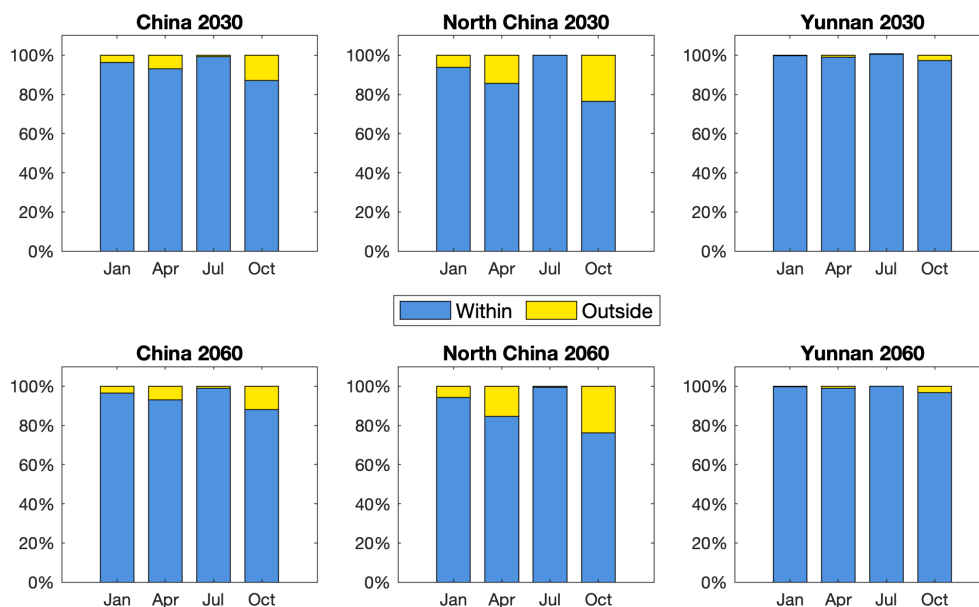


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2 **Figure S3.** Anthropogenic emission-dominated grid cells (red) and natural emission-dominated  
3 grid cells (blue) at a 0.5° x 0.625° resolution in China. Circles indicate locations of PM<sub>2.5</sub>  
4 observations used in this study.  
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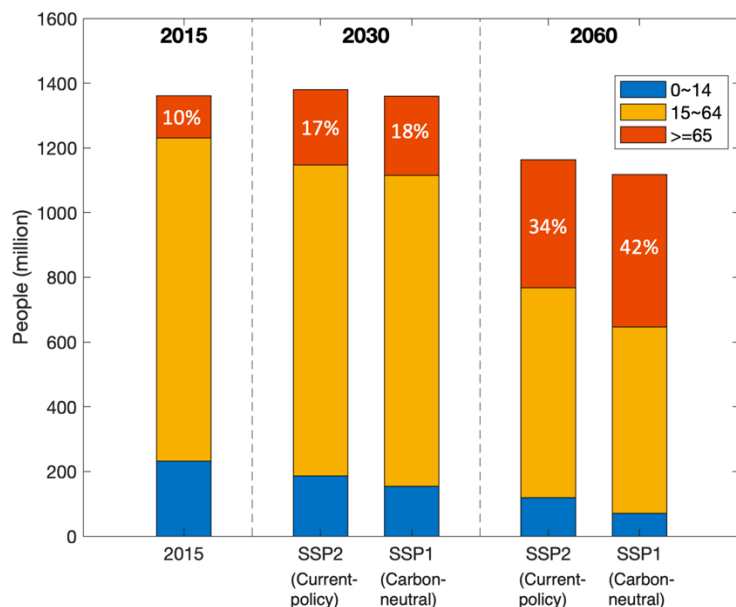
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2 **Figure S4.** Simulated sectoral contributions of foreign anthropogenic emissions to PM<sub>2.5</sub>  
3 concentrations over the entire China, North China and Yunnan province for January in 2030  
4 (top) and in 2060 (bottom) under the current-policy scenario in China and SSP370 scenario in  
5 foreign countries. Sectors include agriculture (AGR), energy (ENE), industry (IND),  
6 transportation (TRA), residential combustion (RES), solvent use (SLV), and waste burning  
7 (WST). The simulation is conducted at a resolution of 2° x 2.5°.

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 2 **Figure S5.** Fractional contributions of foreign anthropogenic emissions from countries within  
 3 and outside the Flex-Grid domain ( $11^{\circ}\text{S}$ – $60^{\circ}\text{N}$ ,  $30^{\circ}$ – $150^{\circ}\text{E}$ , including the entire Asia, eastern  
 4 Africa and eastern Europe) to  $\text{PM}_{2.5}$  concentrations over China, North China and Yunnan  
 5 province in 2030 (top) and in 2060 (bottom) under the current-policy scenario in China and  
 6 SSP370 scenario in foreign countries. The simulation is conducted at a resolution of  $2^{\circ} \times 2.5^{\circ}$ .

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 10 **Figure S6.** Future population and age structure in China. Total populations in China in 2015 and  
 11 future years projected under different SSP scenarios (SSP1 corresponds to China’s carbon-  
 12 neutral scenario; SSP2 corresponds to China’s current-policy scenario). Colors represent  
 13 different age groups according to the legend. White numbers indicate the fraction of population  
 14 above the age of 65 in the total population.