

1 **Measurement report: Brown Carbon Aerosol in Polluted Urban Air of North China Plain:**
2 **Day-night Differences in the Chromophores and Optical Properties**

3 **Yuquan Gong^{1,2}, Ru-Jin Huang^{1,2,3,4}, Lu Yang^{1,2}, Ting Wang¹, Wei Yuan^{1,2}, Wei Xu¹,**
4 **Wenjuan Cao¹, Yang Wang^{4,5,6}, Yongjie Li^{6,7}**

5 ¹State Key Laboratory of Loess and Quaternary Geology, CAS Center for Excellence in
6 Quaternary Science and Global Change, Institute of Earth Environment, Chinese Academy of
7 Sciences, 710061 Xi'an, China

8 ²University of Chinese Academy of Sciences, Beijing 100049, China

9 ³Institute of Global Environmental Change, Xi'an Jiaotong University, Xi'an 710049, China

10 ~~⁴Laoshan Laboratory, Qingdao 266061, China~~

11 ^{4,5}School of Geographical Sciences, Hebei Normal University, Shijiazhuang, China

12 ^{5,6}State Key Joint Laboratory of Environmental Simulation and Pollution Control, Beijing,
13 China

14 ^{6,7}Department of Civil and Environmental Engineering, Faculty of Science and Technology,
15 University of Macau, Taipa, Macau SAR 999078, China

16 Correspondence: E-mail: rujin.huang@ieecas.cn (R.-J.H)

17 **Abstract.** Brown carbon (BrC) aerosol is light-absorbing organic carbon that affects radiative
18 forcing and atmospheric photochemistry. The BrC chromophoric composition and its linkage
19 to optical properties at the molecular level, however, are still not well characterized. In this
20 study, we investigate the day-night differences in the chromophoric composition (38 species)
21 and optical properties of water-soluble and water-insoluble BrC fractions (WS-BrC and WIS-
22 BrC) in aerosol samples collected in Shijiazhuang, one of the most polluted cities in China. We
23 found that the light absorption contribution of WS-BrC to total BrC at 365 nm was higher during
24 the day ($62 \pm 8\%$) than during the night ($47 \pm 26\%$), which is in line with the difference in
25 chromophoric polarity between daytime (more polar nitrated aromatics) and nighttime (more
26 less-polar polycyclic aromatic hydrocarbons, PAHs). The high polarity and water solubility of
27 BrC in daytime suggests the enhanced contribution of secondary formation to BrC during the
28 day. There was a decrease of the mass absorption efficiency of BrC from nighttime to daytime
29 (2.88 ± 0.24 vs. 2.58 ± 0.14 for WS-BrC and 1.43 ± 0.83 vs. 1.02 ± 0.49 $\text{m}^2 \text{g}^{-1} \text{C}^{-1}$ for WIS-BrC,
30 respectively). Large polycyclic aromatic hydrocarbons (PAHs) with 4–6-rings PAHs and
31 nitrophenols contributed to 76.7% of the total light absorption between 300–420 nm at night
32 time, while nitrocatechols and 2–3-ring oxygenated PAHs accounted for 52.6% of the total light
33 absorption at day. The total mass concentrations of the identified chromophores showed larger
34 day-night difference during the low-pollution period (day-to-night ratio of 4.3) than during the
35 high-pollution period (day-to-night ratio of 1.8). The large day-night difference in BrC
36 composition and absorption, therefore, should be considered when estimating the sources,
37 atmospheric processes and impacts of BrC.

38 **1 Introduction**

39 Light-absorbing organic carbon aerosols, also termed brown carbon (BrC) aerosol, are
40 ubiquitous in the atmosphere (Iinuma et al., 2010; Yuan et al., 2016; Huang et al., 2021).
41 Growing evidence has shown that BrC can reduce atmospheric visibility, affect atmospheric
42 photochemistry, and change regional and global radiation balance (Kirchstetter et al., 2004;
43 Laskin et al., 2015; Hammer et al., 2016). Besides, some components in BrC, such as polycyclic
44 aromatic hydrocarbons (PAHs) are highly toxic and carcinogenic, which can adversely impact
45 human health (Alcanzare, 2006; Zhang et al., 2009; Huang et al., 2014). The extent of these
46 effects is closely related to the optical properties and chemical composition of BrC, which are
47 still not well understood.

48 BrC is often classified into water-soluble (WS-BrC) and water-insoluble (WIS-BrC)
49 fractions because these two fractions are largely different in chemical composition and light
50 absorption. For example, abundant nitrophenols were detected in WS-BrC, while polycyclic
51 aromatic hydrocarbons (PAHs) were the main component of WIS-BrC (Huang et al., 2018;
52 Huang et al., 2020). The difference in BrC chemical composition is associated with the emission
53 sources. For example, methyl nitrocatechols are specific to biomass burning, while PAHs are
54 mainly emitted by fossil fuel combustion (Kitanovski et al., 2012; Dat and Chang, 2017).
55 Atmospheric oxidation can further complicate the BrC chromophores dynamically, leading to
56 light-absorbing enhancement or bleaching. For example, Li et al. (2020) reported that the mass
57 absorption efficient (MAE) of some nitroaromatic compounds (e.g., nitrocatechols) from the
58 biomass burning can enhance about 2–3 times by oxidation to generate secondary
59 chromophores. Yet, prolonged photo-oxidation reactions (exposure to sunlight for few hours)
60 of these nitroaromatic can generate small fragment molecules (e.g., malonic acid, glyoxylic
61 acid) and rapidly reduce the particle absorption (Hems and Abbatt, 2018; Wang et al., 2019b;
62 Li et al., 2020). The complexity in composition and sources, as well as the dynamics in their
63 atmospheric processing limit our understanding in BrC chromophores and their links to light
64 absorption.

65 In recent years, a growing number of studies have investigated the chromophore
66 composition of BrC and found that nitro-phenols, low ring acids/alcohols, PAHs and carbonyl

67 oxygenated PAHs (OPAHs) were the major chromophores in BrC (Teich et al., 2017; Yuan et
68 al., 2020; Huang et al., 2020). Some chromophores in BrC can be generated from both primary
69 emission and secondary formation. For example, 4-nitrophenol and 4-nitrocatechol can be
70 emitted directly from biomass burning and can also be generated through photo-oxidation
71 reactions (Kitanovski et al., 2012; Yuan et al., 2020). The differences in emission sources or
72 atmospheric oxidation conditions have a significant effect on the chemical composition of BrC
73 chromophores. Previous studies mainly focused on seasonal variations of BrC chromophores
74 (Wang et al., 2018; Kasthuriarachchi et al., 2020; Yuan et al., 2021), and the diurnal variation
75 of WS-BrC in fluorescence and inorganic fractions (Deng et al., 2022; Zhan et al., 2022),
76 however, the research of BrC chemical composition on day-night differences is scarce. In this
77 study, the optical properties and chemical composition of the WS-BrC and WIS-BrC in daytime
78 and nighttime were measured with a high-performance liquid chromatography–photodiode
79 array–high-resolution mass spectrometry platform (HPLC-PDA-HRMS) in PM_{2.5} samples
80 collected in Shijiazhuang, one of the most polluted cities in the Beijing-Tianjin-Hebei region.
81 Besides, the relationship between the concentration and light-absorbing contributions of the
82 BrC subgroups was analyzed. The object of this study is to investigate the day-night differences
83 in the optical properties and chromophore composition of BrC and to explore the effect of
84 primary emissions and atmospheric processes on the light absorption and chemical composition
85 of BrC.

86 **2 Experimental**

87 **2.1 Sample collection.**

88 Day and night PM_{2.5} samples were collected on the quartz-fiber filters (8*10 in., Whatman,
89 QM-A; filters prebaked at 750 °C, over 3 h) through a high-volume air sampler (Hi-Vol PM_{2.5}
90 sampler, Tisch, the velocity of flow ~1.03 m³ min⁻¹, Cleveland, OH) from 17 January to 13
91 February 2014. Daytime samples were collected from 08:30 to 18:30 (~10 hours), and nighttime
92 samples are collected from 18:30 to the next day at 8:30 (~14 hours). After collection, the
93 samples were stored in a freezer (-20°C) until analysis. The sampling site was located on the
94 rooftop of a building (~15 m above ground) in the Institute of Genetics and Developmental

95 Biology, Chinese Academy of Sciences (38.2° N, 114.3° E), which is surrounded by a
96 residential–business mixed zone.

97 **2.2 Light Absorption Measurement.**

98 A portion filter (about 0.526 cm² punch) was taken from collected samples and sonicated
99 for 30 min in 10 mL of ultrapure water (>18.2 MΩ) or methanol (J. T. Baker, HPLC grade), and
100 then the extracts WS-BrC and methanol soluble BrC (MS-BrC) were obtained. The extracts
101 were filtered with a 0.45 μm PVDF (water soluble) or PTFE (water insoluble) pore syringe
102 filter to remove insoluble substances. The light absorption spectra of the filtrate were tested
103 using a UV-VIS spectrophotometer (Ocean Optics) over the range from 250 nm to 700 nm,
104 equipped with a liquid waveguide capillary cell (LWCC-3100, World Precision Instruments,
105 Sarasota, FL, USA), following the method of Hecobian et al. (2010). To ensure reliable
106 absorbance measurements (Absorbance between 0.2 to 0.8 at 300 nm in this study), the filtrate
107 was diluted with appropriate folds before absorption spectra measurements. In this study, the
108 light-absorbing of WIS-BrC is obtained by MS-BrC minus WS-BrC. [As shown in Figure S13,](#)
109 [the summed absorbance of WS-BrC and WIS-BrC is very close to the absorbance of MS-BrC](#)
110 [\(difference less than 5%\). Therefore, the interferences of solvent and pH on the measurement](#)
111 [of WIS-BrC should be very limited. The pH of the water extracts was not adjusted because](#)
112 [highly diluted water extracts was used to measure the light absorption, and little change of pH](#)
113 [was observed for water extracts of different samples.](#) The light absorption coefficient (Abs) and
114 absorption data were calculated following the equation:

$$115 \quad Abs_{\lambda} = (A_{\lambda} - A_{700}) \frac{V_l}{V_b \times l} \times \ln(10)$$

116 where Abs_λ (M·m⁻¹) represents the sample absorption coefficient at wavelength of λ; A_λ is the
117 absorbance recorded (Random wavelength); A₇₀₀ for explaining baseline drift as the reference
118 during data analysis. [To account for baseline drift that may occur during analysis, absorption at](#)
119 [all wavelengths below 700 nm are referenced to that at 700 nm where there is no absorption for](#)
120 [BrC extracts.](#) V_l (ml) is the total volume of solvent (water or methanol) used to extract the
121 quartz-fiber filters; V_b (m³) represents the volume of through the filter sample of the air; l (0.94
122 m) is the optical path length of UV-VIS spectrophotometer and ln (10) is the absorption
123 coefficient with base-e, which is the natural logarithm by using the logarithm conversion with

124 the base-10.

125 About the mass absorption efficiency (MAE) of the filter extracts at wavelength of λ can be
126 defined as:

$$127 \quad MAE_{\lambda} = Abs_{\lambda}/C_{OM}$$

128 where C_{OM} ($\mu\text{g m}^{-3}$) stands for the concentration of water-soluble organic carbon (WSOC) or
129 methanol extracts methanol-soluble organic carbon (MSOC). The concentrations of WSOC
130 were measured with a TOC-TN analyzer (TOC-L, Shimadzu, Japan). The concentration of OC
131 was measured by a thermal-optical carbon analyzer (DRI, Model 2001) with the IMPROVE A
132 protocol (Chow et al., 2011). Note that MSOC is usually replaced with OC because previous
133 studies have shown that methanol has a high extraction efficiency (~90%) for OC. But it is
134 difficult to completely extract the OC by methanol (Chen and Bond, 2010; Cheng et al., 2016;
135 Xie et al., 2019). Here, WISOC is obtained by MSOC minus WSOC.

136 The wavelength dependence that the light absorption chromophore of solution can be
137 characterized by this equation:

$$138 \quad Abs_{\lambda} = K \cdot \lambda^{-AAE}$$

139 where K is the fitting parameter of the extracts which is constant related to the chromophoric
140 concentration; AAE is known as the absorption Ångström exponent, which depends on the
141 types of chromophores in the solution. In this study, AAE was calculated by linear regression
142 of $\log_{10} Abs_{\lambda}$ versus $\log_{10} \lambda$ at 300–400 nm.

143 The MAE of standards samples ($MAE_{S,\lambda}$), e.g., 4-nitrocatechol and 4-nitrophenol, in the
144 water or methanol solvent at a wavelength of λ were calculated as the Laskin et al. (2015)

$$145 \quad MAE_{S,\lambda} = \frac{A_{\lambda} - A_{700}}{l \times C} \ln(10)$$

146 where C ($\mu\text{g mL}^{-1}$) is the concentration of the standards in the extracts.

147 **2.3 BrC Chemical composition analysis.**

148 The main chromophores in WS-BrC and WIS-BrC were identified by the HPLC-PDA-
149 HRMS platform (Thermo Electron, Inc.), and the details are presented in our previous study
150 (Huang et al., 2020). Firstly, the filter samples ($3.5\sim 48.3 \text{ cm}^2$) were ultrasonically extracted
151 with 6 mL of the ultrapure water for 30 min and repeated two times. The extracts were filtered
152 through a PVDF filter ($0.45 \mu\text{m}$) to remove insoluble materials. Then the solution was subjected

153 to an SPE cartridge (Oasis HLB, USA) to remove water-soluble inorganic salt ions. On the
154 other hand, the residual filters were dried and the WIS-BrC fractions were further extracted two
155 times with 6 mL of methanol for 30 min. to extract the WIS-BrC fractions. Afterward, the
156 extracts of WS-BrC and WIS-BrC chromophores were dried with a gentle stream of nitrogen
157 and then redissolved in 150 μ l of ultrapure water and methanol.

158 The BrC factions were analyzed by an HPLC-PDA-HRMS platform (including the Dionex
159 UltiMate system and the high-resolution Q Exactive Plus hybrid quadrupole-Orbitrap mass
160 spectrometer). Here, the extracts were loaded onto a Thermo Accucore RP-MS column by the
161 binary solvent with an aqueous solution containing 0.1% formic acid and methanol solution
162 containing 0.1% formic acid as mobile phases L₁ and L₂, eluting at a flow rate of 0.3 mL min⁻¹.
163 ¹. The process of gradient elution here was set as follows: firstly, aggrandize linearly the
164 concentration of L₂ from 15% to 30% in the preliminary 15 minutes, and then linearly increased
165 to 90% from 15 to 45minutes, held at 90% from 45 to 50 minutes, afterward decreased to 15%
166 from 50 to 52 minutes and held there for 60 minutes. The Q Exactive Plus hybrid quadrupole-
167 Orbitrap mass spectrometer, negative/positive mode ESI (-)/ESI (+) for details usage and data
168 processing can refer to in the article by Huang et al. (2020) and Liu et al. (2016). Briefly,
169 HPLC/PDA/HRMS platform was employed in ESI (-) and ESI (+) mode to acquire BrC
170 fractions that mass range from m/z 100 to 800. Strongly polar aromatic hydrocarbons like
171 nitrophenol and carboxylic acid are preferentially ionized in ESI (-) mode, ~~conversely, ESI (+)~~
172 ~~mode is helpful to detect oxygenated aliphatic compounds and nitrogen bases~~ OPAHs
173 (Oxygenated PAHs) and PAHs fractions (Lin et al., 2017). OPAHs and nitrogen heterocyclic
174 PAHs were quantified in ESI (+) mode, while PAHs were detected by PDA spectroscopic
175 analysis due to their super low ionization efficiency in ESI. The absorption spectra of
176 chromophores were measured by a PDA detector in the wavelength range of 190-700 nm. The
177 results of this study were corrected by a blank.

178 The elemental composition of individual chromatographic peaks was assigned with the
179 molecular formula calculator in Xcalibur 4.0 software using a mass tolerance of ± 3 ppm and
180 the maximum numbers of atoms for the formula calculator were set as 30 ¹²C, 60 ¹H, 15 ¹⁶O, 3
181 ¹⁴N, 1 ³²S, 1 ²³Na. To eliminate the chemically unreasonable formulas, the identified formulas

182 were constrained by setting $0.3 \leq H/C \leq 3.0$, $0.0 \leq O/C \leq 3.0$, $0.0 \leq N/C \leq 0.5$, $0.0 \leq S/C \leq 0.2$
183 in ESI- mode and $0.3 \leq H/C \leq 3.0$, $0.0 \leq O/C \leq 3.0$, $0.0 \leq N/C \leq 1.3$, $0.0 \leq S/C \leq 0.8$ in ESI+
184 mode, as suggested in a previous study (Lin et al., 2012). Further, the calculated neutral
185 molecular formulas that did not fit the nitrogen rule were excluded. In total, 20 WS-BrC
186 chromophores (two quinolines, four 2-3 ring OPAHs, four nitrocatechols, six nitrophenols and
187 four aromatic alcohols & acids) and 18 WIS-BrC chromophores (three 4-ring OPAHs and 15
188 PAHs) were identified and their concentrations were quantified with authentic standards (28
189 species) or surrogates (10 species) (see Table S1). Thereinto, the WS-BrC chromophores,
190 benzanthrone (21#) and benzo[b]fluoren-11-one (22#) were quantified by mass spectrometry
191 analysis in either negative or positive ESI mode, while the rest of WIS-BrC chromophores were
192 quantified by PDA spectroscopic analysis due to their super low ionization efficiency in ESI
193 (see Table S1). In this study, 38 BrC components (20 WS-BrC and 18 WIS-BrC species) were
194 detected by mass spectrometry and PDA spectroscopy (see Table S1). Thereinto, the main
195 absorbing chromophores in ambient aerosol samples (28 species) were identified by standards.
196 Of course, some BrC chromophores that cannot be accurately identified (substances with (*),
197 see Table S1 in Supplemental Information), we will select the more reasonable structure
198 from those that satisfy the molecular weight, the ratios of C_eH_hO_oN_nS_sN_ax (setting $0.3 \leq H/C$
199 ≤ 3.0 , $0.0 \leq O/C \leq 3.0$, $0.0 \leq N/C \leq 0.5$, $0.0 \leq S/C \leq 0.2$), the double bond equivalent (DBE/C
200 ≥ 0.5), and polarity (water soluble/water in-soluble) (Huang et al., 2020). Mass data processed
201 by the Xcalibur 4.0 software which the parameter of molecular mass set in ± 3 ppm and
202 maximum numbers of atoms for the formula calculator set as 30 ¹²C, 60 ¹H, 15 ¹⁶O, 3 ¹⁴N, 1 ³²S,
203 1 ²³Na. The results of this study were corrected by a blank.

204 **3 Results and discussion**

205 **3.1 Optical properties of BrC during the day and night.**

206 Figure 1 (a) shows the average absorption spectra of WS-BrC and WIS-BrC at the
207 wavelength range between 300 and 500 nm during the day and the night. It can be seen that the
208 light absorption of both WS-BrC and WIS-BrC sharply increased toward the short wavelength.
209 The average absorbance of WS-BrC is $46.04 \pm 35.92 \text{ Mm}^{-1}$ (at 365 nm) during the day that is

210 higher than the night ($35.68 \pm 35.50 \text{ Mm}^{-1}$). However, the light absorption of WIS-BrC at 365
211 nm is much lower during the night ($27.90 \pm 24.80 \text{ Mm}^{-1}$) than during the day (40.89 ± 23.42
212 Mm^{-1}). The day-night differences of light absorption of WS-BrC and WIS-BrC indicates the
213 difference in water solubility and polarity of the chromophores. The average AAE of WS-BrC
214 ($\text{AAE}_{\text{WS-BrC}}$) and WIS-BrC ($\text{AAE}_{\text{WIS-BrC}}$) during the day are 5.10 ± 0.28 and 6.36 ± 0.45 ,
215 respectively, which are lower than those of the night (5.51 ± 0.40 and 6.97 ± 0.80 , respectively).
216 Note that both during the day and night the $\text{AAE}_{\text{WS-BrC}}$ is lower than $\text{AAE}_{\text{WIS-BrC}}$, which is
217 different from findings in previous studies (see Table S2). For example, Huang et al. (2020)
218 found that the $\text{AAE}_{\text{WS-BrC}}$ was higher (8.2 ± 1.0 and 8.2 ± 1.0 in Beijing and Xi'an, respectively)
219 than that of $\text{AAE}_{\text{WIS-BrC}}$ (5.7 ± 0.2 and 5.4 ± 0.2 in Beijing and Xi'an, respectively). Besides,
220 MAE_{365} of WS-BrC are 2.0- and 2.5-fold of WIS-BrC during the day (2.88 ± 0.24 vs. $1.43 \pm$
221 $0.83 \text{ m}^2 \text{ g C}^{-1}$) and night (2.58 ± 0.14 vs. $1.02 \pm 0.49 \text{ m}^2 \text{ g C}^{-1}$), respectively, which is opposed
222 to the results of previous studies. For example, the MAE_{365} of WS-BrC are 0.7- and 0.5-fold of
223 WIS-BrC in winter of Beijing (1.22 ± 0.11 vs. $1.66 \pm 0.48 \text{ m}^2 \text{ g C}^{-1}$) (Chen and Bond, 2010)
224 and Xi'an (1.00 ± 0.18 vs. $1.82 \pm 1.06 \text{ m}^2 \text{ g C}^{-1}$) (Li et al., 2020), respectively. This result
225 indicates that the chemical composition of BrC in the most polluted city, Shijiazhuang, is
226 different from other urban areas on primary sources and secondary aging process. However,
227 both WS-BrC and WIS-BrC have higher MAE_{365} and average AAE values during the day than
228 the night. This suggests that the day-night differences of AAE and MAE_{365} of BrC fractions are
229 likely associated with the different primary emissions and atmospheric aging processes (Cheng
230 et al., 2016; Wang et al., 2019a; Wang et al., 2020). For example, the AAE and MAE_{365} of BrC
231 emitted from biomass burning (AAE ~ 7.31 , and $\text{MAE}_{365} \sim 1.01 \text{ m}^2 \text{ g C}^{-1}$, respectively) (Siemens
232 et al., 2022) showed large differences with that from vehicle emissions (AAE ~ 10.5 , and
233 $\text{MAE}_{365} \sim 0.32 \text{ m}^2 \text{ g C}^{-1}$) (Xue et al., 2018). Besides, photochemical oxidation of fresh BrC from
234 coal combustion resulted in considerable changes in AAE and MAE_{365} , e.g., the AAE and
235 MAE_{365} of fresh coal combustion emission are 7.2 and $0.84 \pm 0.54 \text{ m}^2 \text{ g C}^{-1}$, much higher than
236 those in aged samples (6.4 and $0.14 \pm 0.08 \text{ m}^2 \text{ g C}^{-1}$, respectively) (Ni et al., 2021).

237 Figure 1 (b) shows the light absorption contributions of WS-BrC and WIS-BrC to total
238 BrC over the wavelength range of 300–500 nm. It is obvious that the absorption contribution

239 of WS-BrC is increased from 53.8% at 300 nm to 87.4% at 500 nm during the day, and from
240 38.4% to 61.5% during the night. The higher absorption contributions of WS-BrC at longer
241 wavelengths during the day compared to that of the night may be related to photo-oxidation
242 reaction in day time (Wang et al., 2019b; Chen et al., 2021). The absorption contribution of
243 WS-BrC accounts for $62 \pm 8\%$ to total BrC absorption at 365 nm during the day, but only $47 \pm$
244 8% during the night. The large difference in BrC light absorption between samples from the
245 day and those from the night observed in this study is comparable with previous studies (Shen
246 et al., 2019; Li et al., 2020), and indicates the significant day-night difference in chemical
247 composition.

248 **3.2 Composition and absorption contribution of BrC during day and night.**

249 In total, 38 major chromophores were quantified in WS-BrC and WIS-BrC with HPLC-
250 PDA-HRMS analysis, and the concentrations of these chromophores are shown in Table S3.
251 According to the characteristics of the molecular structures and absorption spectra, these
252 chromophores are divided into ten subgroups, including two quinolines, four nitrocatechols, six
253 nitrophenols, four aromatic alcohols/acids, four 2–3-ring OPAHs, three 4-ring OPAHs, two 3-
254 ring PAHs, four 4-ring PAHs, five 5-ring PAHs, and four 6-ring PAHs. Detailed information
255 about these chromophores is listed in Table S43. Figure 2 shows the chemical composition of
256 the identified BrC components during the day and night. The total concentration of these
257 chromophores during the day (169.8 ng/m^3) is similar to that at night (171.8 ng/m^3), and the
258 chemical composition of the BrC subgroups is clearly different between the day and night. For
259 example, nitrocatechols, aromatic alcohols/acids and 2–3-rings OPAHs are the major
260 contributors to the total mass concentration of identified BrC chromophores during the day
261 (accounting for 23.3%, 22.3%, and 16.6%, respectively). These BrC chromophores, however,
262 are the minor components during the night (accounting for 12.1% and 15.6%, and 6.9%,
263 respectively). This result indicates the enhanced formation of these chromophores during the
264 day. On the contrary, the relative contributions of nitrophenols and 4–6-ring PAHs are much
265 lower during the day (15.3% and 15.2%, respectively) than those during the night (35.8% and
266 24.0%, respectively). During the night, 4-nitrophenol (4NP) contributes 24.4% of the total
267 concentration, followed by 2-methyl-4-nitrophenol, fluoranthene, and chrysene (2M4NP 4.7%,

268 FLU 4.6%, CHR 4.6%, respectively). The higher contributions of nitrophenols and 4–6-rings
269 PAHs at night are likely caused by enhanced primary emissions (Lin et al., 2020; Chen et al.,
270 2021). Our previous study has found that the emitted organic aerosols from coal combustion
271 had a clearly increase at midnight in Shijiazhuang (Huang et al., 2019; Lin et al., 2020). Thus,
272 the large contribute of nitrophenols and 4–6-rings PAHs to total mass concentration at night
273 that may be impacted by emissions from the coal combustion.

274 To investigate the source of the BrC chromophores, the mass concentrations (these
275 concentrations of chromophores are OC normalized) of the day and night were compared. The
276 day-to-night ratios of identified BrC compounds in mass concentrations is shown in Figure 3.
277 It can be seen that the average day-to-night ratios of WS-BrC chromophores are 4.87 for
278 quinolines, 3.49 for 2–3-ring OPAHs, 3.47 for nitrocatechols, 0.48 for nitrophenols, and 2.53
279 for aromatic alcohols/acids, respectively. Previous studies have found that quinolines are
280 important products of fossil fuel combustion, and were used as tracers of the vehicular exhaust
281 (Banerjee and Zare, 2015; Xue et al., 2018; Lyu et al., 2019). Thus, the higher day-to-night ratio
282 of quinolines may be due to increased primary emissions from vehicles during the day. Nitro-
283 phenols and vanillin are typical biomass burning tracers for atmospheric aerosols (Harrison et
284 al., 2005; Scaramboni et al., 2015; Huang et al., 2021). Previous studies have identified
285 secondary formation as an important source of phthalic acid (PA) and methyl-nitrocatechols
286 (Chow et al., 2015; Zhang and Hatakeyama, 2016; Liu et al., 2017). In this study, vanillin,
287 phthalic acid, and three methyl-nitrocatechols (including 4M5NC, 3M6NC, and 3M5NC)
288 isomers have high day-to-night ratios (4.16, 3.75 and 3.28, respectively). The high day-to-night
289 ratios of these BrC chromophores suggest that biomass burning and secondary formation likely
290 play important roles in the daytime source of BrC.

291 ~~Both nitrophenols and nitrocatechols can be emitted from biomass burning, while their daily~~
292 ~~variation has surprising differences in this study. Different from the nitrocatechols, However,~~
293 ~~the average day-to-night ratio (~0.48) of nitrophenols is smaller than one, and this result. The~~
294 ~~day to night ratio of nitrophenols is similar to values in~~ previous studies (Yuan et al., 2016;
295 Schnitzler and Abbatt, 2018). ~~Besides, previous studies found that emissions from residential~~
296 ~~coal fired heating are significant sources of nitrophenols (Wang et al., 2018; Lu et al., 2019).~~

297 [Although both nitrophenols and nitrocatechols can be emitted from biomass burning, they show](#)
298 [largely different day-night variation patterns. The higher concentrations of nitrocatechols](#)
299 [during daytime indicate enhanced secondary formation, which is similar to the results observed](#)
300 [in urban Beijing \(Cheng et al., 2021\). In addition, previous studies found that emissions from](#)
301 [residential coal-fired heating are significant sources of nitrophenols \(Wang et al., 2018; Lu et](#)
302 [al., 2019\). The higher concentrations of nitrophenols during nighttime, however, suggest that](#)
303 [they are mainly emitted from primary emission sources such as residential heating during winter](#)
304 [in North China. This suggests that coal combustion emissions have important effect on the](#)
305 [nocturnal concentration of nitrophenols. Furthermore, Cheng et al. \(2021\) also reported a](#)
306 [different source mechanism for nitrocatechols and nitrophenols. The nitrocatechols peaked at](#)
307 [noon, while the nitrophenols peaked in the evening. Their results indicated that the decrease in](#)
308 [nitrocatechols was associated with an increase in NO₂ concentration at night. In this study, we](#)
309 [also observed an increase in NO₂ concentration at night \(see Figure S3\), which is similar to the](#)
310 [results of Cheng et al. \(2021\).](#) Compared with the WS-BrC chromophores (the day-to-night
311 ratio > 2.53), the day-to-night ratios of the WIS-BrC chromophores approach or below one,
312 with average ratios of 1.46 for 3-ring PAHs, 1.34 for 4-ring OPAHs, 0.74 for 4-ring PAHs, 0.91
313 for 5-ring PAHs, and 0.79 for 6-ring PAHs, respectively. A number of studies showed that coal
314 combustion was the dominant source of PAHs (Wang et al., 2018; Xie et al., 2019; Yuan et al.,
315 2020). Thus, the local emissions may be responsible for the majority of 4–6-rings PAHs during
316 the night.

317 Figure S21 (see Supplemental Information) shows the light absorption contributions of the
318 BrC subgroups to total BrC subgroups in the wavelength range between 300 and 420 nm (the
319 absorptions above 420 nm are too low to exactly estimate the contributions), exhibiting large
320 day-night difference. For example, quinolines show evident absorption below 340 nm (3.6% at
321 310 nm during the day), but negligible contribution above 360 nm. Nitrophenols exhibit a
322 maximum contribution at about 350 nm, while nitrocatechols show higher absorption in the
323 wavelength range from 360 to 400 nm. For PAHs, the absorption maxima shift to a longer
324 wavelength with the increase of the aromatic rings (e.g., 320 nm for 4-ring PAHs and 400 nm
325 for 6-ring PAHs). Overall, the combined light absorption contributions of nitrophenols,

326 nitrocatechols, and PAHs are 86.5% and 80.1% (averaged between 300 and 420 nm) at night
327 and day, respectively. This result is similar to previous studies, in which PAHs and nitro
328 aromatic compounds were identified as the major chromophores (Huang et al., 2020; Yuan et
329 al., 2020).

330 The light absorption contribution of these BrC subgroups exhibits obvious day-night
331 differences. For example, the absorption contribution of 2–3-rings OPAHs and nitrocatechols
332 at 365 nm increased by ~2.0 and ~3.5 times during the day compared to that during the night
333 (see Figure S32). This result differs from previous studies (Kampf et al., 2012; Gao et al., 2022),
334 which indicated that light absorption of BrC compounds were enhanced after exposure to photo-
335 oxidation. On the other hand, the absorption contributions of nitrophenols and 4–6-rings PAHs
336 at 365 nm are ~1.6 times and ~2.2 times higher at night than at day, respectively. The day-night
337 difference of light absorption of nitrophenols is comparable with previous studies (Harrison et
338 al., 2005; Wang et al., 2020). High absorbance of nitrophenols at night is closely related to their
339 higher mass fraction at night. The absorption characteristics of 4–6-ring PAHs are significantly
340 different from the nitro-phenols, and their absorption per unit mass is larger than that of nitro-
341 phenols. The per unit mass absorbance of PAHs much higher than the low-ring aromatic
342 hydrocarbons (e.g., aromatic alcohols/acids) are due to their strongly conjugated systems. It is
343 worth noting that the absorption contributions of some BrC compounds (including quinolines,
344 aromatic alcohols/acids, 4-ring OPAHs, 3-ring PAHs four subgroups) are much lower than
345 those of the above-mentioned BrC compounds because of their lower mass concentration or
346 light absorption coefficient.

347 **3.3 Comparisons between the low and high pollution period.**

348 The relative contributions of day-night subgroups of BrC chromophores in light absorption
349 and mass concentration were further investigated for different pollution levels. The sampling
350 campaign was classified into low-pollution period ($PM_{2.5} < 150 \mu g m^{-3}$) and high-pollution
351 period ($PM_{2.5} > 250 \mu g m^{-3}$). Figure 4 (a) shows the mass fractional contributions of the
352 identified subgroups during these periods, which show an evidently different during the day
353 and night. For example, the mass fraction of quinolines during the day (~24.1%) is much higher
354 than during the night (3.4%) at low-pollution period, which may be related to increased vehicle

355 emissions at day (Rogge et al., 1993; Lyu et al., 2019). Moreover, during the low-pollution
356 period, with good atmospheric dispersion conditions during the day, the fractional concentration
357 of BrC is only 56.9 ng m⁻³ much lower than the nights and high-pollution periods. In the high-
358 pollution period, however, the mass concentration of quinolines is much lower than other BrC
359 chromophores and there is no evident difference between day and night. The mass fraction of
360 aromatic alcohols/acids during the day (35.4%) is much higher than during the night (12.0%)
361 at low-pollution period. For high-pollution period, the mass fraction of aromatic alcohols/acids
362 shows little difference between the day and night. However, their mass concentration during
363 the day (55.5 ng m⁻³) is higher than that during the night (31.9 ng m⁻³). Thereinto, the mass
364 concentration of phthalic acid (a tracer from photochemical oxidation) contributes more than
365 60% to the aromatic alcohols/acids at day for low and high-pollution period (Zhang and
366 Hatakeyama, 2016). This evidence may be suggesting that there is stronger photo-chemical
367 oxidation for aromatic alcohols/acids during the day, especially at low-pollution period.

368 The mass fractional contribution of nitrocatechols is lower during the day than the night
369 at low-pollution period, while there is obvious secondary formation during the day for high-
370 pollution period. This likely suggests that the daytime conditions of the high-pollution period
371 are conducive for the generation of nitrocatechols. The mass fractional contribution of PAHs
372 during the day is much lower than the night at low-pollution period. At night, residential coal
373 heating is an important source of PAHs, and therefore the daytime contributions of PAHs are
374 much lower than nighttime (Wang et al., 2017; Ni et al., 2021). While there is no day-night
375 difference for PAHs at high-pollution period, which is related to the stable sources and stagnant
376 weather conditions (Huang et al., 2019; Lin et al., 2020). It is noteworthy that the mass
377 contributions of the nitrophenols (nighttime is 2–3 times more than daytime) and 2–3-rings
378 OPAHs (daytime is ~2 times more than nighttime) is opposite between the day and night. This
379 demonstrates that they have stable sources compared to other BrC subgroups even during the
380 low-pollution period and high-pollution period. The higher mass fractional contribution of 2–
381 3-rings OPAHs during the day is related to photochemical oxidation. Nitrophenols exhibit a
382 higher mass fractional contribution during the night than the day, indicating a significant
383 contribution from primary emissions (Lu et al., 2019; Lin et al., 2020). Besides, previous

384 investigations have shown that NO_x concentrations and relative humidity are higher at night in
385 Shijiazhuang, which may have accelerated the formation of nitrophenols in the dark (Yuan et
386 al., 2016; Huang et al., 2019). This result exhibits a clear day-night difference during the low-
387 pollution period than high-pollution period, which indicates that the low-pollution period is
388 easily influenced by the external environment (e.g., solar radiation and wind speed).

389 The day-night light absorption contribution of WS-BrC and WIS-BrC chromophores in
390 different pollution periods is shown in Figure 4 (b). For the low-pollution period, the light
391 absorption contribution of the ten BrC subgroups shows a large difference during the day and
392 night. Thereinto, the WS-BrC chromophores (e.g., quinolines, nitrophenols and nitrocatechols)
393 is the main contributor (accounting for ~75% at 365 nm) of total identified BrC during the day.
394 While, the WIS-BrC chromophores (e.g., 4–6-rings PAHs) become an abundance contributor
395 (accounting for ~65% at 365 nm) during the night. There is an obvious day-night differences in
396 light absorption at low-pollution period, which is consistent with the difference in their mass
397 concentration contribution. Different from the low-pollution period, the light absorption
398 contribution of the total WS-BrC and WIS-BrC chromophores showed no significant day-night
399 differences during the high-pollution period. However, the absorption contributions of
400 subgroups in WS-BrC chromophores have a significant day-night difference (e.g., nitrocatechol
401 and nitrophenols) during the high-pollution period, which is due to the change of the mass
402 contributions. WS-BrC chromophores have stronger light absorption both during the day and
403 night compared to the WIS-BrC chromophores at high pollution period. Specifically, the
404 absorption contribution of nitrocatechols and nitrophenols combined accounts for 66.1% at day
405 and 60.7% at night at 365 nm, respectively, which depend on the different emission sources or
406 formation mechanisms between during the day and night. Our results show a significant day-
407 night differences in mass contributions and absorption contributions of BrC components at
408 different pollution levels. This suggests that the variation of BrC chromophores in different
409 pollution periods may be caused by different sources and weather conditions.

410 **4 Conclusions**

411 In general, our study shows the large day-night differences in optical properties and
412 chemical composition of the bulk BrC in urban atmosphere. Thereinto, WS-BrC is the main

413 light-absorbing contributors during the day, while WIS-BrC is main light-absorbing compound
414 at night. The polar WS-BrC has higher MAE₃₆₅ compared to the less-polar WIS-BrC, mainly
415 due to the different conjugate systems and functional groups in the two fractions. ~~During the~~
416 ~~day, polar components such as nitroaromatic have a higher absorption percentage and lower~~
417 ~~MAE, while non polar components such as PAHs have a higher absorption percentage and~~
418 ~~higher MAE during the night.~~ Different types of the identified BrC chromophores exhibit
419 unique characteristics of day-night differences, reflecting their particular sources and formation
420 pathways. For example, nitrocatechols and 2–3-rings OPAHs are important contributors to mass
421 concentration and light absorption during the day, while 4–6-rings PAHs and nitrophenols
422 become the significant contributors at night.

423 Day-night differences of BrC chromophores are associated with different sources during
424 day (mainly secondary formation and vehicle emission) and night (mainly emissions from
425 residential heating) as well as the dynamic development of planetary boundary layer height.
426 ~~Day-night differences of BrC chromophores are associated with different sources during day (more~~
427 ~~secondary formation and vehicle emission) and night (more primary sources such as residential~~
428 ~~heating emissions).~~ —Moreover, these day-night differences are largely affected by the air
429 pollution level, which determines the concentrations of BrC precursors (e.g., aromatic
430 hydrocarbon and phenols) and oxidants (e.g., NO_x, NO₃[·] and OH), as well as meteorological
431 conditions (e.g., solar irradiation and RH) (Liu et al., 2012; Laskin et al., 2015; Wang et al.,
432 2019). For example, our results found that the day-night difference of BrC fractions is more
433 pronounced in chemical composition and light absorption during the low-pollution period than
434 high-pollution period. These factors may show different effects on the formation and
435 photobleaching of different types of the identified chromophores. However, our current
436 understanding of the formation mechanisms of and influencing factors on these identified
437 chromophores is still incomplete (Huang et al., 2018; Yuan et al., 2020). Therefore, a
438 combination of more laboratory and field studies is needed to (1) make comprehensive
439 characterization of the chromophore composition BrC in ambient aerosol; (2) explore
440 thoroughly the formation mechanisms of different types of BrC chromophore. This will
441 significantly enhance our understanding of atmospheric BrC formation mechanisms and

442 therefore improve the accuracy of the atmospheric effects of BrC in air quality and climate
443 models.

444 **Data availability.** Detailed data can be obtained from <https://doi.org/10.5281/zenodo.7690230>.

445 **Author contributions.** Ru-jin Huang designed the study. Data analysis was done by Yuquan Gong
446 and Rujin Huang. Yuquan Gong and Rujin Huang interpreted data, prepared the display items and
447 wrote the manuscript. Lu Yang, Ting Wang, Wei Yuan, Wei Xu, Wenjuan Cao, Yang Wang, and
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449 **Competing interests.** The authors declare that they have no conflict of interest.

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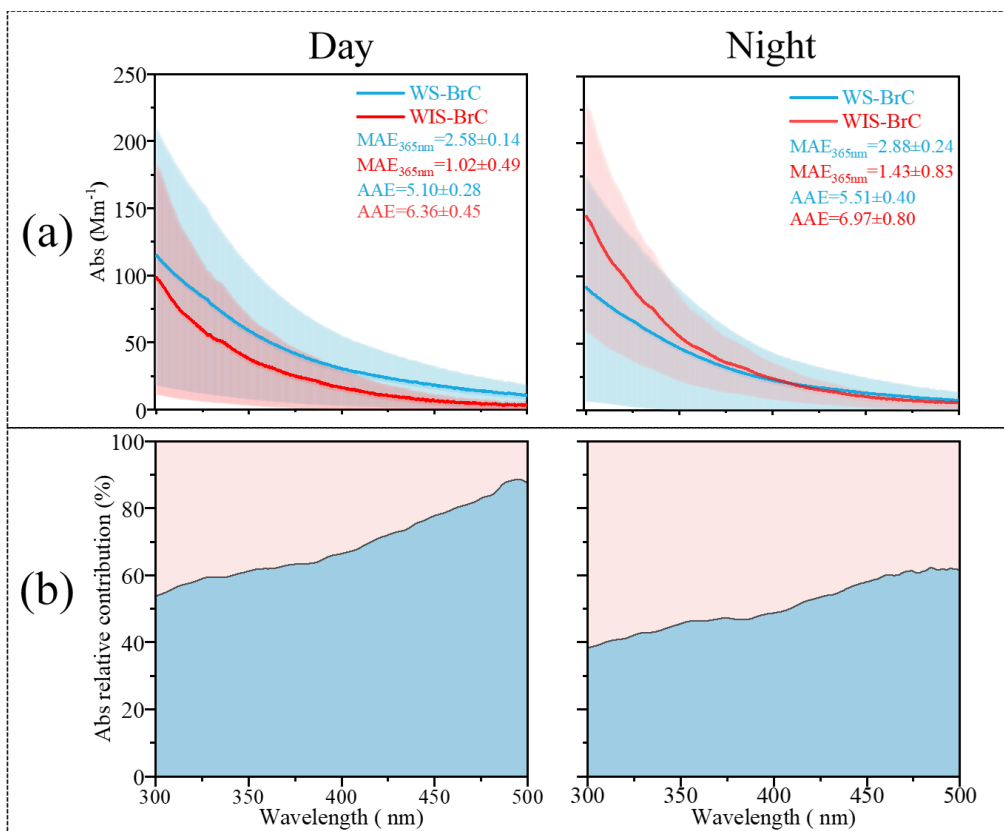


Figure 1. (a) Day-night absorption spectra (Abs, in the wavelength range of 300–500 nm), mass absorption efficiency (MAE, determined at 365 nm), and absorption Ångström exponent (AAE, calculated between 300 and 400 nm) of water-soluble/insoluble BrC (WS-/WIS-BrC) in Shijiazhuang. (b) Light-absorbing proportion of WS-BrC and WIS-BrC between 300 to 500 nm.

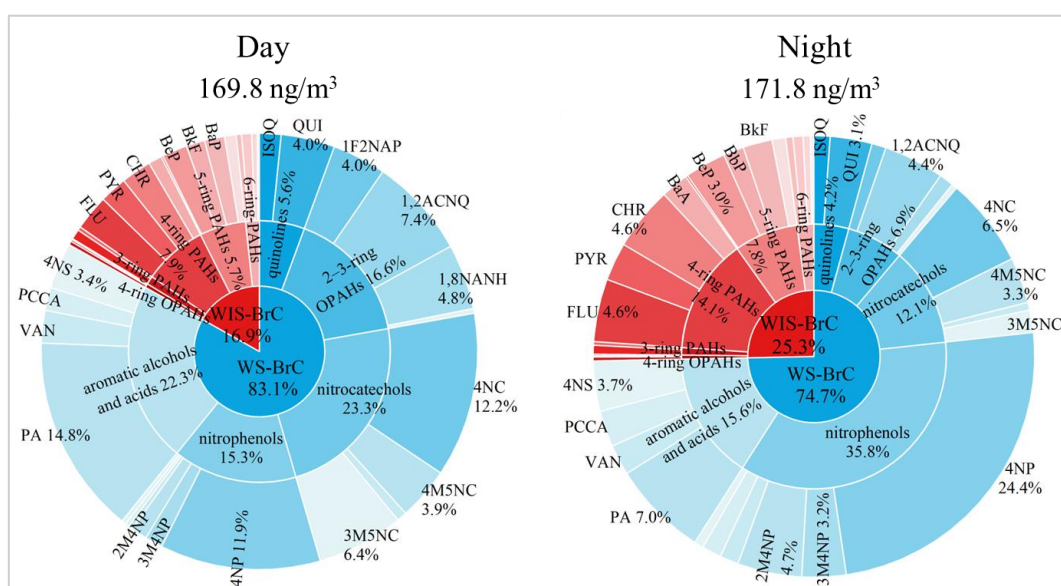


Figure 2. Mass fraction of the identified BrC chromophores during the day and night (details

of the identified BrC chromophores are shown in Table S1).

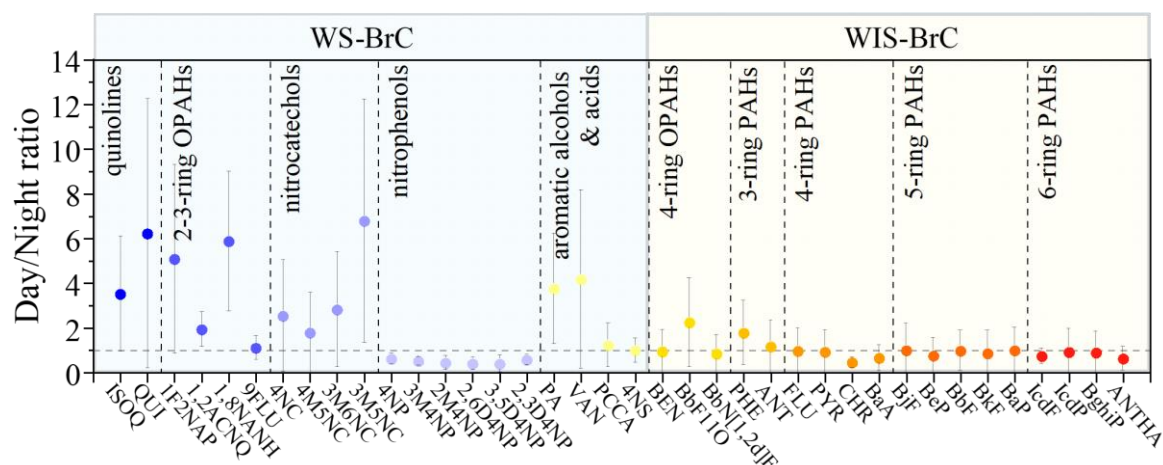


Figure 3. Day-to-night ratios of the concentrations of different BrC chromophores.

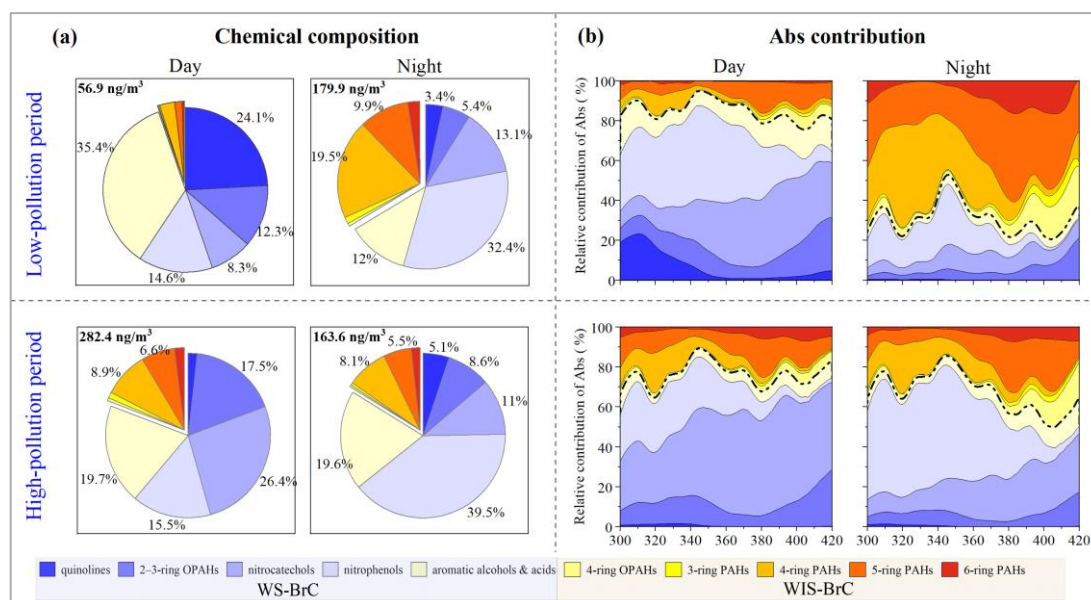


Figure 4. Day-night fractional contributions of mass concentrations (a) and light absorption (b) of the ten BrC subgroups in low-pollution period and high-pollution period. Here the BrC chromophore is the main chromophore substance that has been identified. In (b) WS-BrC is below the dotted line and WIS-BrC is above the dotted line.