Identification of highly oxygenated organic molecules and their role in

aerosol formation in the reaction of limonene with nitrate radical

- 3 Yindong Guo¹, Hongru Shen¹, Iida Pullinen^{2, a}, Hao Luo^{1,3}, Sungah Kang², Luc Vereecken², Hendrik Fuchs², Mattias
- 4 Hallquist⁴, Ismail-Hakki Acir^{2,b}, Ralf Tillmann², Franz Rohrer², Jürgen Wildt², Astrid Kiendler-Scharr², Andreas
- 5 Wahner², Defeng Zhao^{1,5,6*}, Thomas F. Mentel^{2*}
- 6 Department of Atmospheric and Oceanic Sciences & Institute of Atmospheric Sciences, Fudan University, 200438,
- 7 Shanghai, China
- 8 ²Institute of Energy and Climate Research, IEK-8: Troposphere, Forschungszentrum Jülich GmbH, 52425, Jülich,
- 9 Germany

18

24

25

2627

29

33

34

1

- ³IRDR ICoE on Risk Interconnectivity and Governance on Weather/Climate Extremes Impact and Public Health,
- 11 Fudan University, Shanghai 200438, China
- ⁴Department of Chemistry and Molecular biology, University of Gothenburg, Göteborg, 41258, Sweden
- 13 Shanghai Frontiers Science Center of Atmosphere-Ocean Interaction, Fudan University, Shanghai 200438, China
- ⁶Institute of Eco-Chongming (IEC), 20 Cuiniao Rd., Chongming, Shanghai, 202162, China
- ^a Now at: Department of Applied Physics, University of Eastern Finland, Kuopio, 70210, Finland.
- b Now at: Institute of Nutrition and Food Sciences, University of Bonn, Bonn, 53115, Germany.
- 17 Correspondence to: Defeng Zhao (dfzhao@fudan.edu.cn), Thomas F. Mentel (t.mentel@fz-juelich.de)

Abstract. Nighttime NO₃-initiated oxidation of biogenic volatile organic compounds (BVOC) such as monoterpenes is important for the atmospheric formation and growth of secondary organic aerosol (SOA), which has significant impact on climate, air quality and human health. In such SOA formation and growth, highly oxygenated organic molecules (HOM) may be crucial, but their formation pathways and role in aerosol formation have yet to be clarified.

23 Among monoterpenes, limonene is of particular interest for its high emission globally and high SOA yield. In this

work, HOM formation in the reaction of limonene with nitrate radical (NO₃) was investigated in the SAPHIR

chamber (Simulation of Atmospheric PHotochemistry In a large Reaction chamber). About 280 HOM products were

identified, grouped into 19 monomer families, 11 dimer families and 3 trimer families. Both closed-shell products

and open-shell peroxy radicals (RO2•) were observed, and many of them have not been reported previously.

28 Monomers and dimers accounted for 47 % and 47 % of HOM concentrations, respectively, with trimers making up

the remaining 6%. In the most abundant monomer families, C₁₀H₁₅₋₁₇NO₆₋₁₄, carbonyl products outnumbered

30 hydroxyl products, indicating the importance of RO2• termination by unimolecular dissociation. Both RO2•

31 autoxidation and alkoxy-peroxy pathways were found to be important processes leading to HOM. Time-dependent

32 concentration profiles of monomer products containing nitrogen showed mainly second-generation formation

patterns. Dimers were likely formed via the accretion reaction of two monomer RO2•, and HOM-trimers via the

accretion reaction between monomer RO₂• and dimer RO₂•. Trimers are suggested to play an important role in new

particle formation (NPF) observed in our experiment. A HOM yield of 1.5 % was estimated considering only first-generation products. SOA mass growth could be reasonably explained by HOM condensation on particles assuming irreversible uptake of ultra-low volatility organic compounds (ULVOC), extremely low volatility organic compounds (ELVOC) and low volatility organic compounds (LVOC). This work provides evidence for the important role of HOM formed via the limonene + NO₃ reaction in NPF and growth of SOA particles.

1 Introduction

The nitrate radical (NO₃) is an important nighttime oxidant in tropospheric chemistry, and can reach mixing ratios of several hundred pptv during nighttime (Seinfeld and Pandis, 2006). It can react with volatile organic compounds (VOC) and is especially reactive to alkenes, where the nitrate radical can undergo an addition reaction to the C=C double bond (Finlayson-Pitts and Pitts, 1997; Seinfeld and Pandis, 2006). Biogenic monoterpenes (C₁₀H₁₆) are a large contribution to the alkenes in the atmosphere (Klinger et al., 2002; Guenther et al., 2012), and their major nighttime loss pathway is reaction with NO₃ (Beaver et al., 2012; Rollins et al., 2012; Ayres et al., 2015; Fry et al., 2013). The chemistry of monoterpenes with NO₃ has implications on the cycle of reactive nitrogen and thus on ozone formation (Brown and Stutz, 2012). Furthermore, since the NO₃ radical is formed through the reaction of NO₂ with O₃, it is considered to be of anthropogenic origin, and reactions of NO₃ with biogenic VOC (BVOC) thus represent an important interaction between biogenic emissions and anthropogenic emissions.

The reaction of NO₃ with monoterpenes can form secondary organic aerosols (SOA), which can have a large impact on global climate, air quality and human health (Hallquist et al., 2009; Shrivastava et al., 2017). Laboratory studies showed that monoterpenes have high SOA yields in the reaction with NO₃ due to the low volatility of oxidation products (Ng et al., 2008; Rollins et al., 2009; Fry et al., 2013; Fry et al., 2014; Ayres et al., 2015; Jokinen et al., 2015; Zhou et al., 2015; Boyd et al., 2015; Nah et al., 2016; Boyd et al., 2017; Slade et al., 2017; Claffin and Ziemann, 2018; Bates et al., 2022; Dam et al., 2022). Field studies also showed that nighttime NO₃-initiated oxidation of monoterpenes contributes significantly to SOA in forested regions influenced by anthropogenic emissions (Pye et al., 2010; Rollins et al., 2012; Fry et al., 2013; Ayres et al., 2015; Zhou et al., 2015; Xu et al., 2015; Lee et al., 2016; Zhang et al., 2018; Chen et al., 2020) and potentially in urban areas due to the extensive usage of so-called volatile chemical products (VCP) (Nazaroff and Weschler, 2004; Mcdonald et al., 2018). For example, the Southern Oxidant and Aerosol Study (SOAS) showed that the BVOC + NO₃ reactions were a substantial source of SOA (Ayres et al., 2015; Xu et al., 2015; Lee et al., 2016; Massoli et al., 2018). Therefore, accurate predictions and evaluations of SOA concentration and thus its climate and environmental effects require a comprehensive understanding of the reactions of monoterpenes with NO₃.

Recently, a class of organic compounds named highly oxygenated molecules (HOM) have been shown to be critical substances in the SOA formation from BVOC oxidation, particularly monoterpenes, featuring high O/C ratio and low to extremely low volatility (Ehn et al., 2014; Tröstl et al., 2016; Kirkby et al., 2016; Bianchi et al., 2019). HOM here refers to compounds formed in the gas phase via autoxidation which contain at least six oxygen atoms

(Bianchi et al., 2019). Most HOM are classified as ULVOC/ELVOC or LVOC (Bianchi et al., 2019) according to the classification of atmospheric organics based on their volatility (saturation concentration, C*) by Donahue et al. (2012) (extremely low volatility organic compounds (ELVOC), low volatility organic compounds (LVOC), semi-volatile organic compounds (SVOC), intermediate volatility organic compounds (IVOC), volatile organic compounds (VOC)), and a recent update by Schervish and Donahue (2020) (ultra-low volatility organic compounds (ULVOC)). And thus, HOM can be a substantial contribution to growth of SOA particles through gas-particle partitioning.

70

71

72

73

74

75

76

77

78

79

80

81

82

83

84

85

86

87

88

89

90

91

92

93

94

95

96

97

98

99

100

101

102

103104

A better description of the HOM formation chemistry in the oxidation of monoterpenes by NO₃ will improve our understanding of the role of HOM in SOA formation, as well as the relationship between oxidation products, SOA formation and reaction systems. Field observation campaigns as well as laboratory experiments have proven the important contribution of HOM in monoterpene + NO₃ SOA (Lee et al., 2016; Faxon et al., 2018). In the SOAS campaign, HOM-ON (organic nitrates) were identified in both gas and particle phase using a NO₃⁻-Chemical Ionization time-of-flight Mass Spectrometer (CI-APi-TOF) and a High Resolution Time-of-Flight Chemical Ionization Mass Spectrometer (HR-ToF-CIMS) coupled to a Filter Inlet for Gases and AEROsols (FIGAERO). Species with the sum formula C₁₀H_{15,17,19}NO₄₋₁₁ were observed which are formed through the oxidation of monoterpenes by NO₃ (Lee et al., 2016; Massoli et al., 2018). In a campaign in a boreal forest in Hyytiälä, measurement using a NO₃-CI-APi-TOF and positive matrix factor (PMF) analysis showed a nighttime factor of HOM-ON formed via NO₃ oxidation of monoterpenes (Yan et al., 2016). Besides the observations at forested regions, monoterpene-derived HOM via NO₃ oxidation also contribute to organic aerosols in urban regions. For example, Liu et al. (2021) and Nie et al. (2022) have found that HOM derived from monoterpene nighttime chemistry are important in megacities in China, especially during summertime. A number of laboratory studies have reported HOM formation by the oxidation of monoterpenes with NO₃. Boyd et al. (2015) observed C₁₀H₁₇NO_{4/5} and C₁₀H₁₅NO_{5/6} in the gas phase in β-pinene + NO₃ experiments using a quadrupole chemical ionization mass spectrometer with I as the reagent ion (I⁻-CIMS). They proposed possible formation schemes of these ONs. Nah et al. (2016) further detected 5 and 41 HOM-ON in the NO₃ oxidation of α-pinene and β-pinene, respectively, such as C₁₀H_{15/17/19}NO₄₋₉ in the gas- and particle-phase using I-FIGAERO HR-ToF-CIMS. Claflin and Ziemann (2018) provided formation mechanisms for HOM-ON via gas-phase and particle-phase reactions in the β -pinene + NO₃ reaction system, where particle-phase products were analyzed using reversed-phase high-performance liquid chromatography equipped with a UV-vis photodiode array detector (HPLC-UV), Electron-Ionization Thermal Desorption Particle Beam Mass Spectrometer (EI-TDPBMS), Chemical Ionization Finnigan PolarisQ Ion Trap Mass Spectrometer (CI-ITMS), and Electrospray-Ionization Mass Spectrometer (ESI-MS). Recently, Shen et al. (2021) found a large number of HOM (>150 species) in the β-pinene + NO₃ reaction using NO₃-CI-APi-TOF. HOM formed in the reaction of four monoterpenes (αpinene, β-pinene, Δ-3-carene, and α-thujene) with NO₃ were also detected using NO₃-CI-APi-TOF by Dam et al. (2022). Bell et al. (2021) found that dimer dinitrates (C20H32N2O8-13) contribute a large portion to SOA from α-pinene + NO $_3$ and also detected monomer ON such as C10H15NO5-10 and C10H14,16N2O7-11) using FIGAERO-CIMS and an Extractive ElectroSpray Ionization time-of-flight mass spectrometer (EESI-ToF-MS). The detailed speciation

depends on analytical method to some extent, though. Moreover, the HOM composition in the particle-phase was found to depend on aging time and reaction conditions such as dark versus light (Bell et al., 2021; Wu et al., 2021a).

105

106

107

108109

110

111

112

113

114

115

116

117

118119

120

121

122

123

124

125

126

127

128

129130

131

132

133

134

135136

137138

139

Among the monoterpenes, understanding the reaction system of limonene with NO₃ is of specific importance. The emission of limonene makes the 4th largest contribution with an estimated global emission of 11.4 Tg annually, preceded only by α-pinene, trans-β-ocimene and β-pinene (Guenther et al., 2012). Besides its biogenic origin, limonene is also a common additive in cleaning products (Nazaroff and Weschler, 2004) and can even be used as a tracer for fragrances in some places (Gkatzelis et al., 2021). Several studies have shown adverse health effects due to indoor pollution caused by the ozonolysis of limonene (Clausen et al., 2001; Fan et al., 2003; Carslaw et al., 2012; Pagonis et al., 2019). Moreover, limonene stands out with its high reactivity towards the NO₃ radical (with a lifetime of 3 min at 298 K at 20 pptv NO₃) (Ziemann and Atkinson, 2012), and NO₃ oxidation of limonene has high SOA yield (SOA mass yield 15 to 231 %) (Hallquist et al., 1999; Spittler et al., 2006; Fry et al., 2011; Fry et al., 2014; Boyd et al., 2017; Berkemeier et al., 2020; Mutzel et al., 2021). A number of earlier studies have provided valuable insights into the reaction of limonene with NO₃ regarding its main products and their formation pathways, the SOA yield, and the SOA physicochemical properties. For example, Hallquist et al. (1999) measured the SOA mass yield and revealed the dominance of organic nitrates (ON) and carbonyl compounds in the products. Fry et al. (2011) determined the organic nitrate yield and proposed a reaction scheme leading to the formation of ON and carbonyls, and Fry et al. (2014) compared the SOA and ON yields from the NO₃ oxidation of α-pinene, β-pinene, and limonene, and demonstrated why limonene + NO₃ leads to more SOA and ON than α-pinene from a structural perspective. Boyd et al. (2017) found a higher N:C ratio for limonene + NO₃ SOA than for β-pinene + NO₃ SOA. Finally, Peng et al. (2018) studied the optical properties of the limonene + NO₃ SOA.

Regarding the HOM formation in the reaction of limonene with NO₃, Faxon et al. (2018) reported a series of HOM in the particle phase, including C₇₋₁₀ monomers with 3-11 oxygen atoms and C₁₁₋₂₀ dimers with 5-19 oxygen atoms using I⁻FIGAERO HR-ToF-CIMS. However, identification of gas-phase HOM products in the limonene + NO₃ reaction is still lacking and their formation mechanisms remain unclear. Theoretical investigations have revealed that NO₃ addition on the endocyclic C=C double bond is more favorable than the exocyclic one due to a lower energy barrier (Jiang et al., 2009), and this endocyclic double bond of limonene thus tends to be attacked by NO₃ and leads to products including hydroxy-substituted ON or diketone products. The remaining exocyclic double bond can also be attacked by NO₃ in secondary chemistry, leading to more functionalized products (Fry et al., 2011).

The formation of HOM via autoxidation involves a sequence of multiple intramolecular H-shift and O₂ addition reactions, and results in highly oxygenated peroxy radicals (HOM-RO₂•) (Ehn et al., 2014). These HOM-RO₂• can react similarly to traditional RO₂• (Bianchi et al., 2019). The bimolecular reactions of HOM-RO₂• with RO₂•, HO₂• and NO lead to highly oxidized closed shell products including carbonyls, hydroperoxides, alcohols, or organic nitrates as termination groups (R1 to R3), or form accretion products (R4) (Ehn et al., 2014; Mentel et al., 2015). Unimolecular termination reactions of HOM-RO₂• lead to carbonyls or epoxides (R5 to R6) (Crounse et al., 2013). On the other hand, reactions of HOM-RO₂• with NO, RO₂•, NO₃ at nighttime can lead to alkoxy radicals as chain

propagating steps (R7 to R9):

$$141 RO2 + RO2 \rightarrow RHC = O + ROH + O2 (R1)$$

$$RO_{2} \bullet + HO_{2} \bullet \rightarrow ROOH + O_{2}$$
 (R2)

$$RO_{2} \bullet + NO \rightarrow RONO_{2}$$
 (R3)

$$RO_2 \bullet + RO_2 \bullet \rightarrow ROOR + O_2$$
 (R4)

$$RO_2 \rightarrow R = O + OH$$
 (R5)

146
$$RO_{2} \rightarrow epoxide + OH$$
 (R6)

$$RO_{2} \bullet + NO \rightarrow RO \bullet + NO_{2}$$
 (R7)

$$RO_{2} \bullet + RO_{2} \bullet \to RO \bullet + RO \bullet$$
 (R8)

$$RO_2 \bullet + NO_3 \rightarrow RO \bullet + NO_2 + O_2 \tag{R9}$$

If the reactive HOM-RO• products undergo an H-migration reaction, they will again form HOM-RO₂ radicals ("alkoxy-peroxy" pathway) (Mentel et al., 2015), continuing the autoxidation chain. Finally, the HOM-RO• may also fragment leading to small RO₂ radicals, isomerize leading to carbonyls (Bianchi et al., 2019) or react with O₂ to form carbonyls (Ziemann and Atkinson, 2012).

In this study, HOM formation in the NO₃ oxidation of limonene was investigated. We report the identification of gas-phase HOM products, including monomers, dimers and trimers. The formation pathways of dominant products in each category are proposed based on their time profiles in response to multiple additions of limonene in the experiment and on the information in literature. Based on this analysis, we estimated HOM yields and discuss the role of HOM in nucleation and growth of SOA particles.

2 Experimental and Methods

2.1 Experimental setup

The limonene + NO₃ experiment was performed in the atmospheric simulation chamber SAPHIR (Simulation of Atmospheric PHotochemistry In a large Reaction chamber) at the Forschungszentrum Jülich, Germany. SAPHIR is a 270 m³ double-wall cylindrical Teflon chamber with a surface-to-volume ratio of ~1 m² m⁻³. Details of SAPHIR have been described before (Rohrer et al., 2005; Zhao et al., 2015a; Zhao et al., 2015b; Zhao et al., 2018) and are only summarized here. Detailed experimental procedures can be found in Fig. 1a. Before each experiment, SAPHIR was flushed for about 4 h at a flow rate of 370 m³ h⁻¹ with high-purity synthetic air (purity >99.9999 % O₂ and N₂) in order to clean the chamber. To simulate nighttime conditions for the NO₃ chemistry the chamber roof remained closed throughout the experiment. The experiment was performed under dry conditions (RH <2 %) at a temperature of 302 ± 3 K. No seed aerosols were used in the experiments. A fan was used for active mixing in the chamber, leading to a typical mixing time of ~1 min (Fuchs et al., 2013).

NO₃ radicals were generated via the reaction of ozone with nitrogen dioxide:

$$NO_2 + O_3 \rightarrow NO_3 + O_2$$
 (R10)

$$174 NO_2 + NO_3 \leftrightarrow N_2O_5 (R11)$$

Therefore, O₃ and NO₂ were first added to the chamber to form N₂O₅ and NO₃ with mixing ratios of ~2 ppbv and ~0.15 ppbv, respectively. About 20 min later, 5 ppbv of limonene was added to start the organic chemistry. Five more additions of limonene followed, with added concentrations of about 3 ppbv, 3 ppbv, 2 ppbv, 2 ppbv, and finally 8 ppbv, which divided the experiment into six periods (P1 to P6) (Fig. 1a). For period P3 and P5, NO₂ and O₃ were also added to compensate for the loss of NO₃ and N₂O₅ (Fig. 1a). The concentrations of NO₂ and O₃ were maintained around 20 to 70 ppbv throughout the experiment, ensuring the major loss of limonene was by reaction with NO₃ rather than with O₃ (Fig. S1 in the SI). In the first ten min of reaction (named period P1a hereafter, Fig. 1a), NO₃ accounted for 86 % of the chemical loss of limonene.

2.2 Instrumentation

Gas-phase HOM were detected by a Chemical Ionization time-of-flight Mass Spectrometer (CI-APi-TOF, Aerodyne Research Inc., USA) with a resolution $(m/z)/(\Delta m/z)$ of ~3800 using $^{15}NO_3^-$ as the reagent ion, which is capable of detecting organic molecules with high oxygen content (Eisele and Tanner, 1993; Jokinen et al., 2012). The mass spectra were analyzed using the software Tofware (Tofwerk/Aerodyne) in Igor Pro (WaveMetrics, Inc.). Peak identification was conducted by a high-resolution analysis (examples shown in Fig. S2). We observed several peaks which were obviously products from the isoprene + NO₃ reaction, such as $C_5H_{10}N_2O_8^{-15}NO_3^-$ at m/z 289. Such peaks were present before the limonene oxidation reaction started, suggesting that these compounds preexisted in the chamber. These isoprene oxidation products were likely formed in an isoprene + NO₃ experiment performed two days before (Zhao et al., 2021) and released slowly from chamber walls due to their semi-volatile character. Their total concentration is less than 1 ppt. All the isoprene-HOM observed ($C_5H_9NO_{7,10}$, $C_5H_8N_2O_{8-10}$, $C_5H_{10}N_2O_8$, $C_5H_9N_3O_{9,10}$) are saturated and do not contain C=C double bond. The isoprene-HOM will not influence the reaction of limonene with NO₃ in this study. Therefore, they are not discussed as products from the limonene oxidation in our experiment. (However, we cannot exclude that they were partly generated from fragmentation in the limonene + NO₃ reaction.)

A set of instruments were used to measure other gas-phase species, including VOC, NO_x, O₃, NO₃ and N₂O₅ (Shen et al., 2021). Concentrations of NO₃ and N₂O₅ were measured in-situ using a home-built diode laser-based, cavity ring-down spectrometer similar to the instrument described in the work by Wagner et al. (2011). The concentrations of limonene were measured using a Proton Transfer Reaction Time-of-Flight Mass Spectrometer (PTR-TOF-MS, Ionicon Analytik, Austria). The SOA number concentration, surface concentration and size distribution were detected by an SMPS (TSI DMA3081/TSI CPC3786) and a CPC (TSI 3785). Temperature and relative humidity were continuously monitored throughout the experiment.

2.3 Determination of HOM concentration and "primary" HOM yield

HOM concentrations were obtained from the normalized signals to the total signals of the mass spectra (nc, normalized counts) by applying a calibration coefficient (C) of 2.5×10¹⁰ molecule cm⁻³ nc⁻¹. C was determined using H₂SO₄ as the charging efficiency of HOM and H₂SO₄ are considered to be equal (Ehn et al., 2014; Pullinen et al., 2020; Shen et al., 2021). The details of determination of the calibration coefficient are shown in the supplement S1.

A mass-independent transmission efficiency was used according to our previous study, which causes an additional uncertainty of 14 % (Pullinen et al., 2020). In this previous study, the transmission efficiency curve of nitrate CI-APi-TOF was determined and found to monotonously decrease with increasing mass of ions but only slightly depend on the mass range (14 % change). As we used the same setting as our previous study, we have included the slight dependence of transmission on m/z in the uncertainties. The concentrations of HOM were corrected for chamber wall losses, which were determined for a number of HOM similar to our previous study (Zhao et al., 2018), with details described in the supplement. When the chamber is actively mixed, the wall loss was determined to be $(2.2\pm0.2)\times10^{-3}$ s⁻¹. As the HOM yield was determined during the first 3 min of the experiment, we considered the wall loss rate to be constant $(2.2\times10^{-3} \text{ s}^{-1})$ during this period. Sensitivity analysis showed that the HOM yield in this study is not very sensitive to the wall loss rate and is changing by only +0.88 % and -0.44 % if the wall loss rate is varied by +100 % or -50 %.

The HOM yield was calculated as:

$$Y = \frac{[HOM]}{[VOC]_r} = \frac{I(HOM) \cdot C}{[N_2 O_5]_r}$$
 (Eq. 1)

where [HOM] is the concentration of HOM, I(HOM) is the total signal intensity of HOM, C is the calibration factor, and $[VOC]_r$ and $[N_2O_5]_r$ stand for the concentrations of limonene and N_2O_5 reacted, respectively. We used the reacted concentration of N_2O_5 rather than the measured reacted limonene concentration as a large fraction of limonene was already reacting away during the VOC injection before it was homogeneously mixed in the chamber. During this part of the experiment, the high limonene concentration resulted in a rapid loss of NO_3 , such that every NO_3 formed from the decomposition of N_2O_5 reacted with limonene:

$$N_2O_5 \rightarrow NO_2 + NO_3 \tag{R12}$$

limonene +
$$NO_3 \rightarrow Products$$
 (R13)

The initial NO₃ concentration before the limonene injection was small compared to the time-integrated loss of N₂O₅, and other NO₃ loss processes were negligible right after the limonene injection, so that the observed decrease in the N₂O₅ concentration equals indeed the consumption of limonene. The wall loss rate constant of N₂O₅ in the SAPHIR chamber is 7.2×10^{-5} s⁻¹ (Fry et al., 2009). As the HOM yield determination is based on the first 3 min, the wall loss of N₂O₅ can be ignored compared to the loss via the reaction of NO₃ with limonene.

The uncertainty of the HOM yield was estimated to be -55 %/ +117 % based on the combined uncertainties of the HOM-ON peak intensities (\sim 10 %), the limonene concentration (\sim 15 %), the transmission efficiency (-0 %/+14 %) and the calibration factor (-52 %/ +101 %) using error propagation (Zhao et al., 2021). The first 3 min after the injection of limonene were used to calculate the HOM yield, when most of the first-generation oxidation products were produced and negligible particles were yet formed. The HOM yield thus reflects the "primary" HOM yield.

2.4 Determination of HOM condensation on SOA

The SOA mass from the condensation of HOM was calculated to evaluate the role of HOM for the SOA mass growth.

Detailed estimation methods are described in the supplement, including the determination of particle wall loss and dilution loss rate (Sect. S2). In brief, the growth rate of SOA through HOM vapor condensation is based on the

245 collision rate of vapor molecules with aerosols in the kinetic regime. The Fuchs-Sutugin approach is applied to 246 describe the correction for transition from the kinetic to the diffusion regime (Fuchs and Sutugin, 1971; Ehn et al., 247 2014). Based on the volatility of HOM, we considered two scenarios for HOM condensation. In Scenario 1, all HOM 248 were assumed to irreversibly condense on the surface of particles leading to particle mass growth. In Scenario 2, only 249 the irreversible uptake of LVOC and ULVOC/ELVOC compounds were considered to contribute to the growth of 250 SOA particles in order to examine the role of LVOC and ELVOC while IVOC and SVOC were not included, although 251 they may also contribute to SOA. The calculation of saturation concentration C^* (in $\mu g/m^3$) of each HOM was done 252 based on their molecular compositions using two different parameterizations considering the uncertainties in 253 estimating volatility (Wu et al., 2021b):

1. an updated version of the parameterization of Donahue et al. (2011) by Mohr et al. (2019) (Scenario 2a):

$$log_{10}(C) = (25 - n_C) \times 0.475 - (n_O - 3n_N) \times 0.2 - 2 \frac{(n_O - 3n_N)n_C}{(n_C + n_O - 3n_N)} \times 0.9 - n_N \times 2.5$$
 (Eq. 2)

- where n_C , n_O , n_N and n_H are the number of carbon, oxygen, nitrogen and hydrogen atoms of the compound, respectively.
- 258 2. a parameterization based on HOM from α-pinene ozonolysis by Peräkylä et al. (2020) (Scenario 2b):

$$log_{10}(C) = n_C \times 0.18 - n_H \times 0.14 - n_O \times 0.38 + n_N \times 0.80 + 3.1$$
 (Eq. 3)

with similar parameter notation.

254

261262

263

264

265

266

267

268269

270

271

272

273

274275

276

277

278

279

2.5 Simulations of the RO2• loss pathway based on the Master Chemical Mechanism (MCM)

The RO₂• loss pathways were estimated based on MCM simulations (http://mcm.york.ac.uk/). The gas-phase reactions of limonene + NO₃ under dark condition were simulated using iChamber, an open-source program (https://sites.google.com/view/wangsiyuan/models?authuser=0) (Wang and Pratt, 2017). The default chemistry of limonene + NO₃ in the MCM was applied in this study (Saunders et al., 2003b). Photolysis reactions were excluded by setting the zenith angle to 90°. Concentrations of O₃, NO₃, NO₂ and N₂O₅ as well as temperature and relative humidity were constrained to the experimental data with a time resolution of 1 min. The chamber dilution rate of 1.5×10⁻⁵ s⁻¹ was applied to all species. The P1 period was simulated using the above conditions and the initial concentrations of limonene were added in the model according to the experimental procedures. The sum of all 140 RO₂• in the limonene subset of MCM v3.3.1 were used in the usual way to estimate the loss rates of RO₂• bimolecular reactions. The reaction rate constants are provided in Table S3, and calculated loss rates are shown in Fig. S3. We note that the MCM reaction schemes do not include the accretion reactions between HOM-RO₂•. Berndt et al. (2018a) determined the rate constant of accretion reaction of $C_{10}H_{15}O_4$ • formed via α -pinene ozonolysis to be $\sim 1\times 10^{-11}$ cm³ molecule⁻¹ s⁻¹, which is of the same order as the upper limit for RO₂• + RO₂• reactions used in the MCM schemes for functionalized peroxy radicals such as acyl peroxy radicals (Jenkin et al., 1997; Saunders et al., 2003a). However, currently we do not see a reliable updated set of rate coefficients that are applicable to the reaction system in this study. If the rate constants of some $RO_2^{\bullet} + RO_2^{\bullet}$ reactions were higher than those used in MCM, the concentrations of RO2• would be lower and relative importance of RO2• + RO2• in RO2• fate would increase. Several simulation results are shown in Fig. S4, including NO₃, N₂O₅, limonene, RO₂•, reaction rate of limonene with NO₃

(k×limonene×NO₃), and examples of 1st and 2nd-generation RO₂•.

In the early stage of each period, RO_2^{\bullet} mainly reacted with RO_2^{\bullet} and NO_3 , although in the later stage the reaction with NO_2 also contributed to a significant fraction of RO_2^{\bullet} loss (Fig. S3, showing period P1 as an example). During the period P1a when our peak assignment was based on, the RO_2^{\bullet} loss was dominated by $RO_2^{\bullet} + RO_2^{\bullet}$ and $RO_2^{\bullet} + NO_3$.

285

286

287

288

289

290

291

292

293

294

295

296

297

298

299

300

301

302

303

304305

306

307

308

309

310

311

312313

314

280

281

282

283

284

3 Results and discussion

3.1 Experiment overview and observed HOM

After each limonene addition, the concentration of limonene rose first and then rapidly declined, while the concentrations of NO₃ and N₂O₅ rapidly decreased due to the fast reaction between limonene and NO₃ and gradually increased when limonene had been consumed (Fig. 1a). About 10 min after the first limonene addition, new particles had already formed and quickly grew in size (Fig. 1b). Therefore, we used the first 10 min reaction time (period P1a) to identify gas-phase HOM products, and the whole experiment to examine the contribution of HOM to SOA.

During period P1, HOM were quickly formed. We identified about 280 HOM compounds, including monomers $(C_7-C_{10}, \sim 280-460 \text{ Th})$, dimers $(C_{17}-C_{20}, \sim 490-700 \text{ Th})$, and trimers $(C_{26}-C_{30}, \sim 720-960 \text{ Th})$ (Fig. 2a). Their detailed formulas can be found in Table S1. HOM on the horizontal lines of the Kendrick mass defect plot (O-based) (Fig. 3 and Fig S5, S6) share the same number of C, N and H atoms, with the number of oxygen atoms increasing from left to right. Such HOM compounds are defined as a family. We notice that most monomer peroxy radical families are each related to two monomer closed-shell product families, with one H atom more or one H atom less, which are the expected termination products of RO₂• + RO₂• reactions, or if HO₂• is present, RO₂• + HO₂• termination products. These three related families are defined as a "series", with the same number of C and N number, such as $C_{10}H_{15}$ $_{17}NO_{6-14}$. In total, we identified 6 monomer series ($C_{10}H_{15-17}NO_{6-14}$, $C_{10}H_{14-16}N_{2}O_{9-15}$, $C_{10}H_{14-16}O_{7-12}$, $C_{9}H_{13-15}NO_{7-14}$, $C_8H_{11-13}NO_{6-13}$ and $C_7H_{9-11}NO_{7-11}$) and 1 monomer family ($C_{10}H_{17}N_3O_{12-16}$), 11 dimer families ($C_{20}H_{31}NO_{10-15}$, $C_{20}H_{33}NO_{12\text{-}16}, \quad C_{20}H_{32}N_2O_{9\text{-}20}, \quad C_{20}H_{31}N_3O_{14\text{-}20}, \quad C_{20}H_{33}N_3O_{12\text{-}20}, \quad C_{20}H_{34}N_4O_{15\text{-}20}, \quad C_{20}H_{32}O_{13\text{-}16}, \quad C_{19}H_{29}NO_{10\text{-}13}, \quad C_{10}H_{29}NO_{10\text{-}13}, \quad C_{10}H_$ $C_{19}H_{31}NO_{10-15}$, $C_{19}H_{30}N_2O_{10-18}$ and $C_{19}H_{31}N_3O_{15-19}$), and 3 trimer families ($C_{30}H_{47}N_3O_{18-24}$, $C_{30}H_{48}N_4O_{16-24}$ and $C_{29}H_{46}N_4O_{19-24}$). The monomer family $C_{10}H_{17}N_3O_{12-16}$ are not classified as series because the supposedly relevant families are not clearly identified. Compounds containing at least one nitrogen atom accounted for more than 90 % of the identified HOM products. We assume that compounds containing nitrogen atoms are organic nitrates, because other N-containing species such as amines or nitro compounds are very unlikely to be formed from the reaction of limonene with NO₃. Organic nitrates formed in this study could be alkylnitrates, or (acyl)peroxynitrates formed via the reaction of RO₂• with NO₂.

During period P1a, in the absence of particles, both HOM monomers and oligomers were observed, including monomers (47 %), dimers (47 %) and trimers (6 %) (Fig. 2a). Concentrations of gas-phase dimers and trimers decreased evidently after particle formation (Fig. 2b, 5, 6), indicating a fast gas-particle condensation and strong tendency of oligomers to condense on particles.

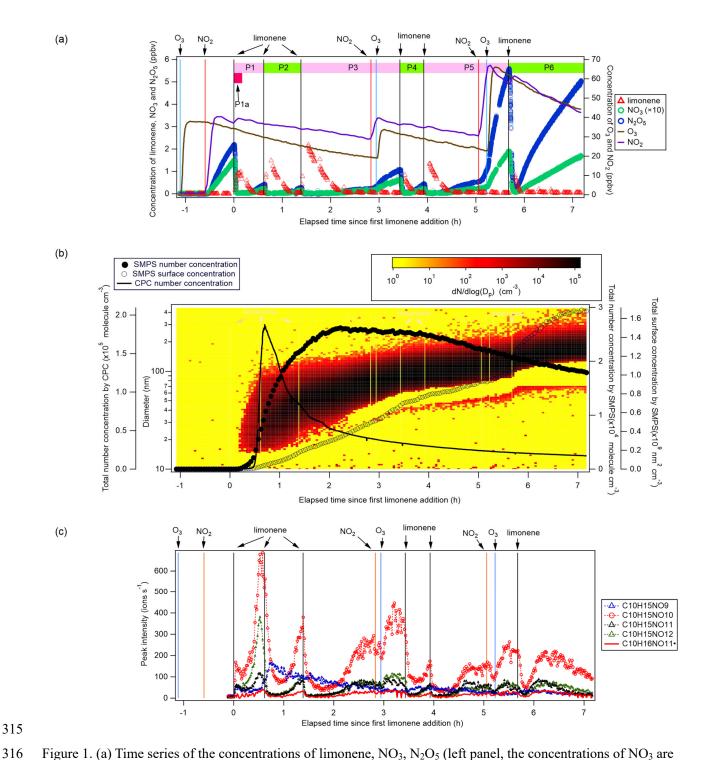
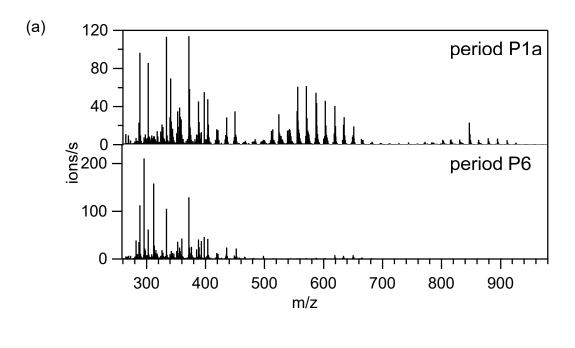


Figure 1. (a) Time series of the concentrations of limonene, NO_3 , N_2O_5 (left panel, the concentrations of NO_3 are magnified by a factor of 10 for clarity), as well as O_3 and NO_2 (right panel). The time periods P1 to P6 as well as P1a are also marked. (b) Total particle concentration and its size distribution during the whole period of experiment detected by SMPS and CPC. The solid and hollow black circles refer to total number concentration and total surface concentration detected by SMPS, respectively. Colors represent particle number concentration distribution based on $log(D_p)$. The solid black line refers to total number concentration detected by CPC. (c) Time series of peak intensity of typical products of the $C_{10}H_{15}NO_x$ family and $C_{10}H_{16}NO_{11}$ as a representative of the $C_{10}H_{16}NO_x$

family. Vertical lines indicate the time of O₃ and NO₂ additions, as well as six limonene injections.

Based on their typical time series (Fig. 1c), products can be classified as first-generation or second-generation products. Generally, the concentrations of first-generation products, which result from the direct reaction of limonene with NO₃, are expected to quickly increase after the limonene addition, followed by a steady decline due to wall loss or chemical reactions. Concentrations of typical second-generation products, which result from further reactions of first-generation products, are expected to show a gradually increasing concentration pattern after a limonene addition and reach their maximum concentration later than first-generation products. These general expectations are modified in our case, since the particle concentration increased in our experiment (Fig. 1b) and the condensational sink of HOM products became stronger over time. Thus, an increase in concentration suggests that the increasing condensational sink was exceeded by increasing production with time, i.e. from second-generation pathways.

To sum up, gas-phase HOM formed in the limonene + NO₃ system were dominated by HOM monomers and dimers. Time series patterns of the products indicate multiple generations of reaction pathways.



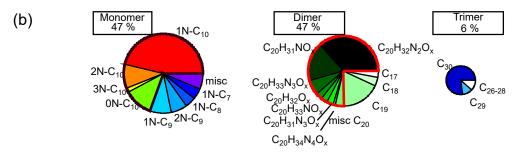


Figure 2. (a) Average mass spectra of the first 10 min reaction time after the first addition of limonene (period P1a, upper panel) and the last limonene addition period till particles reached maximum mass concentration (period P6,

lower panel). (b) Pie charts (from left to right: rainbow, green, and blue colors) representing the relative contributions of identified families to HOM monomers, dimers, and trimers, respectively, during the P1a period. The area of each pie is in proportion to their concentrations during the P1a period.

3.2 Monomers and their formation pathways

3.2.1 Overview of HOM monomers

A number of HOM monomer families were detected with an increasing oxygenation pattern at 16 Th intervals (Fig. 3). Such a pattern is attributed to autoxidation of RO_2 • (with 32 Th interval for each O_2 addition) plus the alkoxyperoxy pathway (shifted by 16 Th compared with exclusive autoxidation) as discussed below. During period P1a, the most abundant HOM monomers are C_{10} compounds (64 %), such as peroxy radicals $C_{10}H_{16}NO_x$ • and closed-shell products $C_{10}H_{15}NO_x$ and $C_{10}H_{17}NO_x$, which are carbonyl compounds and hydroxyl or hydroperoxy compounds from the termination reactions of $C_{10}H_{16}NO_x$ •, respectively (R14).

$$C_{10}H_{16}NO_{x} \bullet + C_{10}H_{16}NO_{x} \bullet \to C_{10}H_{15}NO_{x} + C_{10}H_{17}NO_{x}$$
(R14)

According to the nitrogen atoms contained, C_{10} -HOM monomers can be classified into 1N-, 2N-, 3N-monomers, and monomers without nitrogen atoms. While 1N- C_{10} HOM monomers were likely formed by direct NO₃ addition to limonene, C_{10} HOM monomers containing multiple N atoms were likely formed via multiple reaction steps. Besides C_{10} HOM monomers, C_{6-9} HOM monomers were also observed. These C_{6-10} families are discussed below in the order of their contributions to HOM monomers.

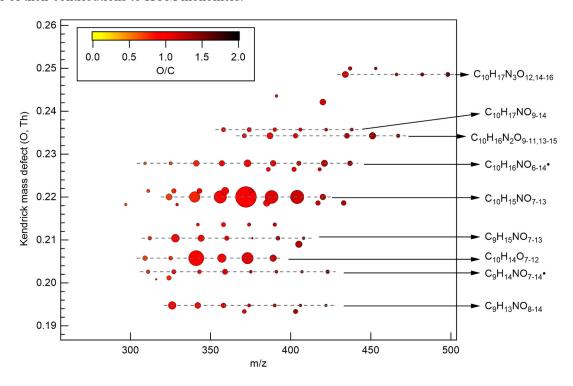


Figure 3. Kendrick mass defect plot (O-atom-based) of major monomer products. The area of the circles is proportional to the average intensity of each compound during the P1a period with the largest circle representing $C_{10}H_{15}NO_{10}$. The color denotes O/C ratios. Dashed lines indicate major product families. For clarity, the reagent ions

¹⁵NO₃⁻ is omitted from molecular formula. The calculation of O-atom-based Kendrick mass defect includes two steps. First, the IUPAC mass scale (based on the ¹²C atomic mass as exactly 12 Da) is rescaled to Kendrick mass: Kendrick mass = IUPAC mass × (16/15.9949), which converts the mass of O from 15.9949 to exactly 16. Then, Kendrick mass defect is given by: Kendrick mass defect = nominal Kendrick mass – exact Kendrick mass. Thus, compounds with the same number of each kind of atom except for O have equal O-atom-based Kendrick mass defect, and are shown in a horizontal line in the O-atom-based Kendrick mass defect plot.

3.2.2 1N-C₁₀ monomers

Among C_{10} HOM monomers, the $1N-C_{10}$ families were most abundant and included stable closed-shell products $C_{10}H_{15}NO_x$ (x=7-13) and $C_{10}H_{17}NO_x$ (x=9-14) and peroxy radicals $C_{10}H_{16}NO_x^{\bullet}$ (x=6-14). The concentration of $C_{10}H_{16}NO_{11}^{\bullet}$ increased in the later phase of each limonene addition period (Fig. 1c), showing mostly a time profile of a second-generation product, similar as most of the other radicals in the $C_{10}H_{16}NO_x^{\bullet}$ family (Fig. S7). However, the time series of $C_{10}H_{15}NO_x$ compounds showed an overlaying pattern of first- and second-generation products dominated by a second-generation time profile with the exception of $C_{10}H_{15}NO_9$ (Fig. 1c). Due to sensitivity restrictions of CIMS, the primary peroxy radical $C_{10}H_{16}NO_5^{\bullet}$ was not detected, which was supposed to show 1^{st} -generation pattern. The absence of first-generation characteristics of the time profile of most HOM peroxy radicals $C_{10}H_{16}NO_{x(x\geq 6)}^{\bullet}$ may be attributed to two possible reasons. They either did not undergo efficient autoxidation, or they underwent immediate conversion including autoxidation and/or bimolecular reactions with other RO_2^{\bullet} or RO_3 forming closed-shell products such as dimers or continuing the radical chain forming RO_1^{\bullet} . The instantaneous increase of 2N-dimers and trimers after the first limonene addition shown below suggests that $C_{10}H_{16}NO_{x(x\geq 6)}^{\bullet}$ were indeed formed efficiently via autoxidation. Therefore, the latter reason is more likely. At this time, we do not have a reasonable explanation for the trend of $C_{10}H_{15}NO_9$, though we should consider that there are many isomers at play, which may have very different chemical pathways (un)available.

Since the C₁₀H₁₅NO_x family showed an overlaying pattern of the first-generation and second-generation products, they likely contained multiple isobaric substances produced through different pathways. Based on the literature, possible formation pathways of these products were tentatively proposed (Seinfeld and Pandis, 2006; Vereecken and Peeters, 2010; Mentel et al., 2015; Vereecken and Nozière, 2020). As an example of the pathways to form first-generation products, C₁₀H₁₆NO_{2x-1}• (with an odd number of oxygen atom) and their corresponding termination products can be formed via autoxidation of the first peroxy radical C₁₀H₁₆NO₅• (R1OO), showing C₁₀H₁₆NO₉• (R3OO) as an example (Scheme 1a, first-generation products). C₁₀H₁₆NO_{2x}• (with an even number of oxygen atom) can be formed via alkoxy-peroxy channels. For example, the ring-opening of the alkoxy radical C₁₀H₁₆NO₄• (R1O), which was formed via the reaction of C₁₀H₁₆NO₅• (R1OO) with another RO₂• or NO₃ radical (Scheme 1a, first-generation products). Ring-opening of R1O leads to C₁₀H₁₆NO₆• (R4OO), which can undergo autoxidation forming C₁₀H₁₆NO_{2x}•. In addition, the alkoxy radical C₁₀H₁₆NO₄• (R1O) is susceptible to ring-opening reactions (Novelli et al., 2021), which can lead to a first-generation stable product 3-isopropenyl-6-oxoheptanal

(endolim, TP1) after C-C bond cleavage followed by the elimination of a NO₂ fragment (Scheme 1b, second-generation products). Endolim (TP1) has been detected as a major product in previous limonene + NO₃ studies (Hallquist et al., 1999; Spittler et al., 2006).

As an example of second-generation chemistry, the remaining double bond of endolim could react with NO₃ to form RO₂•, followed by the autoxidation to form second-generation $C_{10}H_{16}NO_x$ • (with odd number of oxygen atoms). Similar to first-generation pathways, second-generation $C_{10}H_{16}NO_x$ • with even number of oxygen atoms can be formed via alkoxy-peroxy channel. From the time profile of $C_{10}H_{15}NO_x$, the second-generation pathway (Scheme 1b) was expected to play a more important role, in agreement with the theoretical result by Kurtén et al. (2017), in which the two bond-cleavage pathways of limonene-derived RO• radical were considered. It is worth mentioning that the reaction products of limonene with O_3 may also react with NO_3 , forming $C_{10}H_{16}NO_x$ • (Scheme S1). However, as shown above, this was a minor pathway in our experiment (Sect. 2.1). We would like to note that to simplify the scheme, only the reaction of NO_3 with the endocyclic double bond is presented, since this reaction is faster than that with the exocyclic double bond (Jiang et al., 2009; Fry et al., 2011).

 $C_{10}H_{16}NO_x^{\bullet}$ with both even and odd number of oxygen atoms as well as their termination products had comparable abundance, which suggests that the alkoxy-peroxy pathway was important for RO_2^{\bullet} formation in this reaction. This finding is analogous to the findings in the reaction of a number of alkenes with O_3 and in the reaction of isoprene and β -pinene with NO_3 (Mentel et al., 2015; Zhao et al., 2021; Shen et al., 2021).

Scheme 1. Illustrative scheme for HOM formation in the limonene + NO_3 reaction. (a) Example formation pathways leading to first-generation $1N-C_{10}$ HOM-RO₂ radicals ($C_{10}H_{16}NO_{x^{\bullet}}$ with even or odd numbers of O-atoms). (b) Second-generation scheme involving the formation of endolim. (c) Scheme of intramolecular termination of RO_2^{\bullet}

radicals forming carbonyl products taking the $C_{10}H_{16}NO_9$ • radical as an example. Note that the depicted reactions may not be the dominant pathways.

418 419

420

421

422

423

424

425

426

427

428

429

430

431

432

433

434

435

436

437

438

439

440

441

442

443

444

445

446

447

448

449

450

451

417

Among 1N-C₁₀ monomers, concentrations of carbonyl compounds were much higher than the sum of hydroxyand hydroperoxy-substituted compounds (Table 1). According to Hyttinen et al. (2015), for nitrate CI-APi-TOF, HOM containing two hydrogen bond donors (such as -OOH and -OH group) have strong binding energy with NO₃. Additional hydrogen bond donors only enhance the binding energy marginally. If we compare HOM carbonyl product (such as $C_{10}H_{15}NO_{10}$) with the corresponding hydroxy product ($C_{10}H_{17}NO_{10}$), they only differ in one functional group. As both are highly functionalized, it is likely that HOM carbonyl have a quite similar sensitivity with HOM alcohol. If the sensitivity of carbonyl HOM were lower, this would result in even more dominance of carbonyl HOM over hydroxyl HOM. Thus, we conclude that carbonylnitrates are more abundant than hydroxynitrates or hydroperoxynitrates. This finding is likely attributed to unimolecular termination reactions of RO2•, although reaction paths via RO• also cannot be excluded. Smaller unbranched RO• tend to react with O2 forming carbonyl compounds while for larger or branched RO•, isomerization can also form carbonyl compounds and is a more energetically favorable and thus faster pathway compared with the reaction with O2 (Ziemann and Atkinson, 2012). The importance of unimolecular termination reactions of HOM-RO2 and the resulting high ratio of carbonyl compounds to hydroxyl/hydroperoxyl compounds has also been found in the reaction system of β -pinene + NO₃ (Shen et al., 2021; Dam et al., 2022). This high ratio is also consistent with findings in the ozonolysis of alkenes (Mentel et al., 2015), where unimolecular termination reactions were also proposed to be the likely explanation (Crounse et al., 2013; Rissanen et al., 2014). As discussed in our previous study by Shen et al. (2021), this higher abundance of carbonylnitrates is not likely to be explained by the reaction of alkoxy RO• + O2 forming carbonyls and HO2•, decomposition of β-nitrooxyperoxynitrate or self-reactions of RO₂• via the Bennett and Summers mechanism forming carbonyls and H₂O₂. Reactions between RO₂• in general should produce overall equal amounts of carbonyl and hydroxyl compounds. The decomposition of β-nitrooxyperoxynitrate is slow in the gas-phase. The reaction of alkoxy RO• with O₂ for large RO• is generally slower than isomerization and decomposition (Vereecken and Peeters, 2009, 2010). Thus, the higher abundance of carbonylnitrates compared to hydroxynitrates may be attributed to unimolecular termination of HOM-RO₂•. In addition, isomerization of RO• forming carbonyl compounds may also contribute to this finding. Our result thus further emphasizes that unimolecular termination reactions of RO₂ radicals are important pathways in the formation of HOM monomers derived from the reactions of monoterpenes with NO₃ (Shen et al., 2021). Scheme 1c shows this unimolecular termination process using a C₁₀H₁₆NO₉• radical as an example. C₁₀H₁₆NO₉• undergoes a 1,4-H-shift and O₂ addition to form a C₁₀H₁₆NO₁₁• radical. The C₁₀H₁₆NO₁₁• radical further undergoes an H-shift of the α-OOH H-atom, which produces a carbonyl closed-shell product as well as an OH• radical.

For 1N- C_{10} HOM monomers, the products detected in this study generally agree with previous laboratory and field studies on the reaction of limonene and other monoterpenes. Faxon et al. (2018) also observed $C_{10}H_{15}NO_x$ as

the most prevalent products in the particle phase from limonene + NO₃. In the SOAS campaign, both $C_{10}H_{15}NO_x$ and $C_{10}H_{17}NO_x$ products were detected and were believed to be products of nighttime chemistry (Lee et al., 2016). The high abundance of 1N-C₁₀ HOM monomers is consistent with the finding that $C_{10}H_{15}NO_x$ and $C_{10}H_{17}NO_x$ dominate the chemical composition of SOA formed via NO₃ oxidation of α -pinene and β -pinene, as shown in previous chamber studies (Takeuchi and Ng, 2019).

In summary, 1N-C_{10} HOM monomers are mainly formed via second-generation pathways, and unimolecular termination of RO_2 • likely plays an important role leading to higher abundance of carbonyl HOM-ON ($C_{10}H_{15}NO_x$) than hydroxy/hydroperoxy HOM-ON ($C_{10}H_{17}NO_x$).

Table 1. Observed $C_{10}H_{16}NO_x$ • radicals (m) and their termination products, including carbonyl compounds (m-17), hydroxyl compounds (m-15), and hydroperoxy compounds (m+1). Their concentrations during period P1a are normalized to that of $C_{10}H_{15}NO_{10}$, which had the highest concentration among the families of 1N- C_{10} monomers. Their relative signal intensities during the P1a period are shown in the second line of each cell.

Peroxy radical m	Carbonyl m-17	Hydroxy m-15	Hydroperoxy m+1
C ₁₀ H ₁₆ NO ₆ •			
1.5 %			
C ₁₀ H ₁₆ NO ₇ •			
2.0 %			
C ₁₀ H ₁₆ NO ₈ •	C ₁₀ H ₁₅ NO ₇		
6.7 %	8.0 %		
C ₁₀ H ₁₆ NO ₉ •	$C_{10}H_{15}NO_{8}$		C ₁₀ H ₁₇ NO ₉
6.0 %	25.2 %		3.7 %
C ₁₀ H ₁₆ NO ₁₀ •	C ₁₀ H ₁₅ NO ₉	C ₁₀ H ₁₇ NO ₉	$C_{10}H_{17}NO_{10}$
10.2 %	34.6 %	3.7 %	3.6 %
C ₁₀ H ₁₆ NO ₁₁ •	C ₁₀ H ₁₅ NO ₁₀	$C_{10}H_{17}NO_{10}$	C ₁₀ H ₁₇ NO ₁₁
6.6 %	100.0 %	3.6 %	3.0 %
C ₁₀ H ₁₆ NO ₁₂ •	$C_{10}H_{15}NO_{11}$	C ₁₀ H ₁₇ NO ₁₁	C ₁₀ H ₁₇ NO ₁₂
4.1 %	39.0 %	3.0 %	2.3 %
	$C_{10}H_{15}NO_{12}$	$C_{10}H_{17}NO_{12}$	$C_{10}H_{17}NO_{13}$
	41.2 %	2.3 %	1.5 %
C ₁₀ H ₁₆ NO ₁₄ •	$C_{10}H_{15}NO_{13}$	$C_{10}H_{17}NO_{13}$	C ₁₀ H ₁₇ NO ₁₄
4.7 %	6.7 %	1.5 %	1.8 %
		$C_{10}H_{17}NO_{14}$	
		1.8 %	

466 3.2.3 2N and 3N-C₁₀ monomers

C₁₀ monomers with 2 and 3 nitrogen atoms accounted for 27 % and 1 % of HOM monomers, respectively. They were likely formed via the reaction of a second attack of NO₃ to the first-generation products as the 1N-C₁₀ closed-shell

products formed via the reactions shown in Scheme 1a should contain a remaining limonene C=C double bond. Typical 2N- and 3N-HOM showed a second-generation time profile (Fig. 4). For clarity, only periods P1 to P3 are shown. This time profile is consistent with the pathways with multiple NO₃ attacks. Scheme 2 shows possible formation pathways of 2N- and 3N-C₁₀ monomers. 2N-C₁₀ HOM were likely to be formed from NO₃ oxidation of 1N-C₁₀ monomers (C₁₀H₁₅NO_x and C₁₀H₁₇NO_x), resulting in C₁₀H₁₅N₂O_{x*} and C₁₀H₁₇N₂O_{x*} (Scheme 2a, 2b). While C₁₀H₁₅N₂O_{x*} (x=9-12) were observed, C₁₀H₁₇N₂O_{x*} could not be uniquely identified because the peaks of the C₁₀H₁₇N₂O_{x*} and C₁₀H₁₅NO_x families are too close in the mass spectra to be separated based on the resolution of our mass spectrometer. 3N-C₁₀ monomers, C₁₀H₁₇N₃O_x, were expected to be formed from limonene via two steps of NO₃ oxidation to the double bonds and an addition of NO₂ to an RO₂ radical, leading to a peroxynitrate or peroxyacylnitrate. NO₂ addition reactions may also contribute to the formation of 2N-C₁₀ monomers. The addition of NO₂ to RO₂ radicals could occur either before (Scheme 2d) or after (Scheme 2c) the second NO₃ attack.

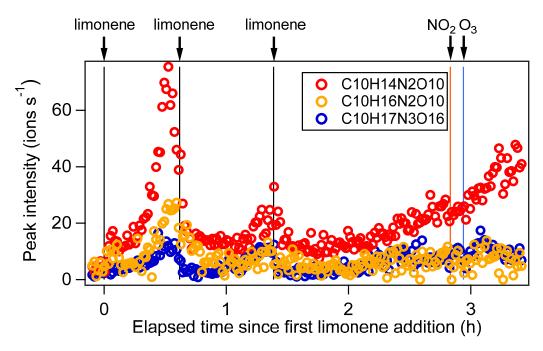


Figure 4. Time series of peak intensity of several monomers $C_{10}H_{14}N_2O_{10}$, $C_{10}H_{16}N_2O_{10}$ and $C_{10}H_{17}N_3O_{16}$ as the representatives of multiple N monomers during the periods P1-P3.

(a)
$$C_{10}H_{15}NO_n \xrightarrow{NO_3\cdot , +O_2} C_{10}H_{15}N_2O_{n+7} \xrightarrow{\text{termination}} C_{10}H_{14}N_2O_{n+6} \text{ or } C_{10}H_{16}N_2O_{n+6/n+7}$$

(c)
$$C_{10}H_{17}NO_n \xrightarrow{NO_3$$
, $+O_2$
 H -shiff, $+O_2$
 $C_{10}H_{17}N_2O_{n+7}$. NO_2
 $C_{10}H_{17}N_3O_{n+9}$

(d)
$$C_{10}H_{16}NO_n$$
: O_2 $C_{10}H_{16}N_2O_{n+2}$ O_{n+2} O_3 , O_3 , O_4 O_5 O_6 O_7 O_8 O_8

Scheme 2. Possible formation pathways of C₁₀-monomers containing 2 nitrogen atoms (a, b) and 3 nitrogen atoms (c, d). Termination denotes reactions of RO₂• with other RO₂• or HO₂, or unimolecular reactions, leading to closed-shell products.

3.2.4 Formation pathways of C₁₀ monomers without N-atoms and monomers with less than 10 C-atoms

Besides C_{10} products containing nitrogen atoms, HOM monomers without nitrogen atoms were also identified. Among these products, $C_{10}H_{14}O_x$ (x=7-12) were the most prevalent family, which were also detected in limonene ozonolysis (Jokinen et al., 2015). The $C_{10}H_{14}O_x$ family showed a time series typical of first-generation products (Fig. S8). $C_{10}H_{14}O_x$ and $C_{10}H_{16}O_x$ could be formed from limonene + NO₃ with $C_{10}H_{16}NO_x$ • terminating their autoxidation by migration of the α -NO₃ H-atom, eliminating an NO₂ fragment (Scheme S2) (Novelli et al., 2021). Alternatively, these products could be formed via the reaction of O₃ with limonene (Scheme S2). Either way, $C_{10}H_{14}O_x$ and $C_{10}H_{16}O_x$ were formed via first-generation pathways.

We also observed monomers with carbon atom number less than 10. During the P1a period, C_9 monomer families were the most abundant contributors to C<10 HOM monomers, followed by C_8 families. The majority of C_9 monomers were $C_9H_{15}NO_x$ (x=7-13) (time series shown in Fig. S9) and $C_9H_{13}NO_x$ (x=8-14). The loss of one carbon atom may follow the mechanism shown in Scheme S3 (Fry et al., 2011; Bianchi et al., 2019). The major product family in C_8 monomers is $C_8H_{11}NO_x$ (x=6, 7, 9-13). While during period P1a $C_8H_{11}NO_x$ compounds could be hardly observed, their concentrations increased considerably in the later periods (Fig. S10). The gas-phase concentration of $C_8H_{11}NO_7$ was even the highest among all compounds in later periods (highest intensity signal in Fig. 2b). This is partly attributed to the relatively high volatility of C_8 compounds compared with C_{10} HOM species and accretion products, which tend to condense on particles. The major family in C_7 monomers, $C_7H_9NO_x$ (x=6-13), showed a time series pattern similar to $C_8H_{11}NO_x$ compounds (Fig. S11). Such a time profile indicates that C_7 and C_8 products were likely a result of multi-generation gas-phase reactions.

3.3 Dimers and their formation

Among dimers, C_{20} products were the most abundant, followed by C_{19} products. Among C_{20} and C_{19} dimers, the most prevalent families included $C_{20}H_{32}N_2O_x$ (x=9-20), $C_{20}H_{33}N_3O_x$ (x=12-20), $C_{20}H_{31}NO_x$ (x=10-15), $C_{20}H_{31}N_3O_x$ (x=14-20), $C_{20}H_{34}N_4O_x$ (x=15-20), and $C_{19}H_{30}N_2O_x$ (x=10-18) (Fig. S5). The O/C ratio of dimers did not exceed one, while that of monomers was as high as two. This could be due to oxygen atom loss and participation of less oxygenated RO_2 • in the dimer formation as discussed below. Time series of dimers also showed different behavior compared to monomers. For example, compounds of the $C_{20}H_{32}N_2O_x$ family only reached a considerable peak intensity in period P1 and decreased rapidly thereafter, while the signal intensity in periods P2 to P6 were low (Fig. 5). Generally, other dimers showed similar patterns (Fig. S12-S14), though the difference of their concentration between P2-P6 and P1 were not as large as for the $C_{20}H_{32}N_2O_x$ family. The time when signals of several dimers (e.g. $C_{20}H_{32}N_2O_x$, $C_{20}H_{33}N_3O_x$, $C_{20}H_{34}N_4O_x$) dropped substantially matched the time of new particle formation (NPF) and the onset of particle growth, indicating that some dimers were likely involved in the early growth of particles. Such a behavior is expected since dimers have a much lower volatility than monomers. This observation is consistent with the limonene + NO₃ laboratory study by Faxon et al. (2018) that found a significant fraction of HOM dimer derived in the particle phase.

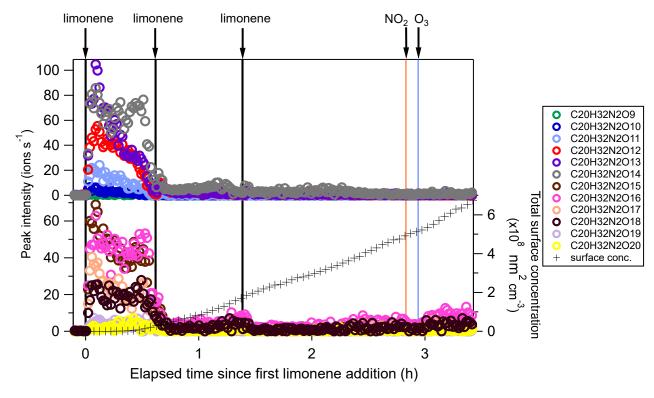


Figure 5. Time series of peak intensity of the $C_{20}H_{32}N_2O_x$ family compounds during the periods P1 to P3. The cross markers (lower right y-axis) indicate total particle surface concentration.

In general, $C_{20}H_{32}N_2O_x$ showed an overlaying time profile of first- and second-generation products (Fig. 5). $C_{20}H_{32}N_2O_x$ were likely formed via the accretion reaction between two monomer RO_2 • ($C_{10}H_{16}NO_x$ •):

$$C_{10}H_{16}NO_{x1} \bullet + C_{10}H_{16}NO_{x2} \bullet \to C_{20}H_{32}N_{2}O_{x1+x2-2} + O_{2}$$
(R15)

Since $C_{10}H_{16}NO_x$ • can be first- or second-generation products, the resulting dimers $C_{20}H_{32}N_2O_x$ can also be first- or second-generation products. The time series show that $C_{20}H_{32}N_2O_x$ with lower O number presented more of a first-generation product time profile (Fig. 5), while the relative contribution of second-generation formation was observed to increase with oxygen number.

We compared the observed dimer formula with those expected based on accretion reactions of HOM-RO2•. X in the $C_{20}H_{32}N_2O_x$ observed was \geq 9; however, according to the accretion mechanism and the observed $C_{10}H_{16}NO_x$ • $(x \geq 6)$, x in $C_{20}H_{32}N_2O_x$ should be \geq 10 (6+6-2=10). Moreover, as the most abundant RO_2 • within the $C_{10}H_{16}NO_x$ • family was $C_{10}H_{16}NO_{10}$ • (Table 1), the most abundant $C_{20}H_{32}N_2O_x$ would have an oxygen number of 18 if they were exclusively formed by the accretion reaction of HOM RO_2 •. This contradicted the fact that the most abundant molecule among the $C_{20}H_{32}N_2O_x$ family was $C_{20}H_{32}N_2O_{13}$. The findings above could only be explained by the participation of less oxygenated RO_2 • such as $C_{10}H_{16}NO_5$ • in the accretion reaction (Berndt et al., 2018a; Berndt et al., 2018b; Mcfiggans et al., 2019). $C_{10}H_{16}NO_5$ • was not detected by our CI-APi-TOF, which is attributed to the lower detection sensitivity of molecules with O number \leq 5 in the NO_3 -CIMS (Riva et al., 2019). Still, $C_{10}H_{16}NO_5$ • is the first RO_2 radical formed in the limonene + NO_3 reaction (Scheme 1a) so a high mass flux has to pass through this RO_2 •. If we assume that the abundance of $C_{10}H_{16}NO_5$ • was high, and considering that the concentration of $C_{10}H_{16}NO_{10}$ • was the highest in the $C_{10}H_{16}NO_x$ • family, their accretion reaction (R16) could form $C_{20}H_{32}N_2O_{13}$ and support that $C_{20}H_{32}N_2O_{13}$ was the most abundant C_{20} dimer product:

$$C_{10}H_{16}NO_{5} + C_{10}H_{16}NO_{10} \rightarrow C_{20}H_{32}N_{2}O_{13} + O_{2}$$
(R16)

Time series of dimers with different numbers of N atoms were different, indicating different formation pathways. For example, the $C_{20}H_{31}NO_x$ family were mainly first-generation products (Fig. S12), which may be formed via the following reaction:

$$C_{10}H_{16}NO_{x1} \bullet + C_{10}H_{15}O_{x2} \bullet \to C_{20}H_{31}NO_{x1+x2-2} + O_2$$
(R17)

 $C_{10}H_{15}O_X^{\bullet}$ were first-generation radicals (Sect. 3.2.4), while $C_{10}H_{16}NO_X^{\bullet}$ were mainly second-generation radicals. $C_{10}H_{16}NO_X^{\bullet}$ could also be formed via first-generation pathway as discussed above (Scheme 1a), but that was not borne out by the time profile, suggesting a fast termination of first-generation $C_{10}H_{16}NO_X^{\bullet}$ radicals. Reaction R17 could be one of the termination pathways of first-generation $C_{10}H_{16}NO_X^{\bullet}$ based on the first-generation time profile of $C_{20}H_{31}NO_X$. In the study by Faxon et al. (2018), the formation of $1N-C_{20}$ dimers was explained by a mechanism involving two $1N-RO_2$ radicals which produced $4NO_3$ as a by-product. However, C_{10} $4N-C_{20}$ dimers through R17.

On the other hand, $C_{20}H_{33}N_3O_x$ and $C_{20}H_{34}N_4O_x$ were mainly second-generation products (Fig. S13, S14). $C_{20}H_{33}N_3O_x$ and $C_{20}H_{34}N_4O_x$ were likely to be formed via NO₃ oxidation of dimers containing less nitrogen atoms, and were thus second-generation products. The related radicals were also detected, such as $C_{20}H_{32}N_3O_x^{\bullet}$ (x=16-19) and $C_{20}H_{31}N_2O_x^{\bullet}$ (x=13-16). Possible formation pathways of dominant oligomer families are displayed in Table 2. We cannot exclude that the formation pathway of $C_{20}H_{33}NO_x$, $C_{20}H_{34}N_4O_x$ and $C_{19}H_{31}NO_x$ may also involve limonene

oxidation by OH• (Table 2), which can be formed in the ozonolysis of limonene as a minor pathway. In addition, the high abundance of $C_{20}H_{31}NO_x$ (x=10-15) among the dimers may be partly attributed to a contribution of the reaction of limonene with O_3 .

The initial drop of the products (dimers and monomers) in Fig. 1, Fig. S8, and Fig. S12 during P1 (the characteristic time of the fastest decay was 15 min, 10 min, and 13 min, respectively) is attributed to the balance of their sources via the reaction of limonene with NO₃, their wall loss, and their potential loss by the reaction with NO₃. The characteristic time of the fastest decay of the HOM over the 2nd limonene addition in Fig. 1, 4, 5, S12, and S13 are 4-8 min. These decays can be explained by the wall loss rate (characteristic time ~8 min) and condensation sink of vapor loss to particles according to the study of Kulmala et al. (2012) (characteristic time ~13 min). The characteristic times of the fastest decay of the HOM at the end of P2 in Fig. S12 and S13 are 1.4-3.4 min, which can also be well explained by the updated wall loss rate and condensation sink of vapor loss to particles at the end of P2 (characteristic time ~ 1.4 min). In addition, for some HOM (e.g., $C_{10}H_{15}NO_{10}$ (Fig. 1), $C_{10}H_{14}N_2O_{10}$ (Fig. 4), C₂₀H₃₃N₃O₁₇ (Fig. S13), C₂₀H₃₄N₄O₁₇ (Fig. S14)), the times of limonene additions (except for the first time) matched the time when HOM signals dropped rapidly. This phenomenon implies a sudden decrease of the source of these HOM at limonene additions as sinks including the losses to walls and particles were largely invariant in such a short time. The decrease of the source may be attributed to the rapid depletion of NO₃ at limonene injections (Fig. 1). As many of these HOM are 2nd-generation products, i.e. formed via the reactions of 1st-generation products with NO₃, the depletion of NO₃ could lead to sudden decreases of the source of the 2nd-generation HOM, which accounted for most HOM in the study.

In summary, HOM dimers are likely to be formed via accretion reactions of monomer RO₂•, and some dimers can undergo secondary oxidation by NO₃. Some dimers were likely involved in the early growth of SOA particles.

Table 2. Major dimer and trimer families and their possible formation pathways.

Dimer/Trimer family	Possible formation pathways	
$C_{20}H_{32}N_2O_x$	$C_{10}H_{16}NO_x \bullet + C_{10}H_{16}NO_x \bullet$	
$C_{20}H_{33}N_3O_x / C_{20}H_{31}N_3O_x$	$C_{20}H_{32}N_2O_x + NO_3 + HO_2 \bullet /RO_2 \bullet$	
$C_{20}H_{31}NO_x$	$C_{10}H_{16}NO_x \bullet + C_{10}H_{15}O_x \bullet$	
$C_{20}H_{33}NO_x$	$C_{10}H_{16}+OH_{}^{\bullet}+C_{10}H_{16}NO_{x}^{\bullet}$	
$C_{20}H_{34}N_4O_x\\$	$(C_{10}H_{16}N_2O_x + OH^{\bullet}) + (C_{10}H_{16}N_2O_x + OH^{\bullet})$	
$C_{19}H_{30}N_2O_x$	$C_{10}H_{16}NO_x \bullet + C_9H_{14}NO_x \bullet$	
$C_{19}H_{31}N_3O_x$	$C_{19}H_{30}N_2O_x + NO_3 + HO_2 \bullet /RO_2 \bullet$	
$C_{19}H_{29}NO_x$	$C_9H_{14}NO_x{\color{red}\bullet} + C_{10}H_{15}O_x{\color{red}\bullet}$	
$C_{19}H_{31}NO_x$	$C_{10}H_{16}+OH^{\bullet}+C_{9}H_{14}NO_{x}^{\bullet}$	
$C_{30}H_{48}N_4O_x$	$C_{20}H_{32}N_3O_x{\color{red}\bullet} + C_{10}H_{16}NO_x{\color{red}\bullet}$	
$C_{30}H_{47}N_3O_x$	$C_{20}H_{31}N_2O_x{\color{red}\bullet} + C_{10}H_{16}NO_x{\color{red}\bullet}$	

565

566

567

568

569

570

571572

573

574

575

576

577

578579

580

581

582

583

584 585

586587

3.4 Trimers and their formation

Trimers (C_{26-30}) were dominated by C_{30} compounds (Fig. S6, Table S1). To the best of our knowledge, this is the first study that identified gas-phase trimers in the limonene + NO₃ reaction. The O/C ratio of trimers were lower than that of monomers and dimers, suggesting possible multiple accretion reactions in their formation pathways, which lose 2 oxygen atoms in each reaction. As each accretion reaction terminates the peroxy radical chain, the observation of trimers also implies that some dimers could further react with NO₃, creating dimer RO₂•. The most prevalent product families were $C_{30}H_{48}N_4O_x$ (x=16-24) and $C_{30}H_{47}N_3O_x$ (x=18,19,21,23,24), which were likely formed via the most abundant monomer RO₂ radicals - $C_{10}H_{16}NO_x$ • and the most abundant dimer RO₂ radicals - $C_{20}H_{32}N_3O_x$ • and $C_{20}H_{31}N_2O_x$ •. Trimers from other monoterpenes + NO₃ have been observed in previous laboratory studies. For example, $C_{30}H_{48}N_4O_{16}$ and $C_{30}H_{47}N_3O_{16}$ were observed in the mass spectra of α-pinene + NO₃ SOA by Wu et al. (2021a), and $C_{30}H_{47}N_3O_{13}$ was identified in β-pinene + NO₃ SOA by Claflin and Ziemann (2018).

Similar to their precursors $C_{20}H_{32}N_2O_x$, $C_{30}H_{48}N_4O_x$ showed negligible signal except in period P1, and presented an overlaying time profile of first- and second-generation product pattern (Fig. 6). For comparison, gas-phase trimer products were not observed in the β -pinene + NO₃ reaction (Shen et al., 2021), and the trimers observed in SOA from β -pinene + NO₃ are likely formed via particle phase reactions (Claflin and Ziemann, 2018). An efficient gas-phase trimer production via subsequent accretion reactions between peroxy radicals requires that the precursor dimer has a high enough reactivity to create a dimer RO₂•, e.g. via NO₃ reaction to a double bond. This suggests that the VOC containing at least two double bonds are likely more favorable to form trimers, which is consistent with our previous findings that trimers were formed in the NO₃ reaction with isoprene which also contains two double bonds (Zhao et al., 2021) while they were not observed in the reaction of NO₃ with β -pinene which contains only one double bond (Shen et al., 2021).

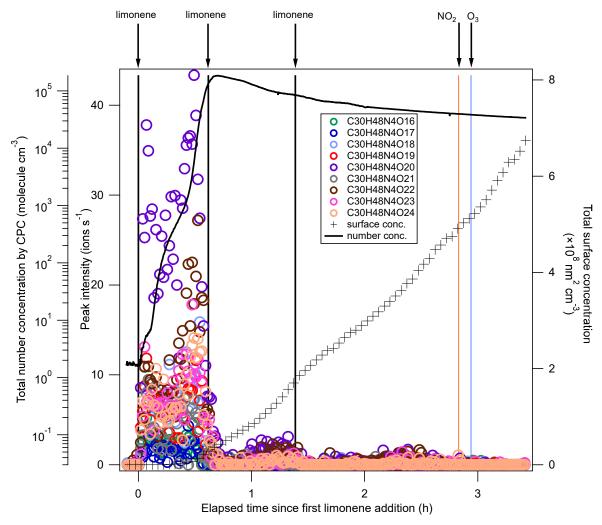


Figure 6. Time series of peak intensity of the $C_{30}H_{48}N_4O_x$ family compounds during the periods P1 to P3. The solid black line refers to total number concentrations detected by CPC. The cross markers (right y-axis) indicate total particle surface concentration.

3.5 "Primary" incremental HOM yields

We chose period P1 for the calculation of HOM yields in order to minimize the influence of the condensational sink on HOM concentration. However, both first-generation and second-generation products existed in this period, as discussed in Sect. 3.2 through 3.4 and supported by the time-behavior of the total HOM concentration (Fig. S15). Period P1 can be roughly divided into three phases based on the trends in HOM concentration. Shortly after the limonene injection, large quantities of HOM were produced (first-production phase) followed by a balanced intermediate phase when HOM concentrations stopped increasing. After the intermediate phase, HOM concentrations began to increase again (second-production phase). The first-production phase overlapped with the time span where limonene, NO₃ and N₂O₅ concentrations decreased, implying the dominance of first-generation HOM production process. During the second production period, wall loss was compensated by second-generation HOM formation, leading to another rise of the total HOM concentrations. Therefore, we use the first-production phase to estimate

primary HOM production, determined over the first 3 min of the experiment. The calculated "primary" HOM molar yield is 1.5 % This value is significantly lower than the HOM yield of 5 to 17 % in earlier limonene ozonolysis experiments (Ehn et al., 2014; Jokinen et al., 2015; Pagonis et al., 2019). It should be emphasized that second-generation HOM, which contributed greatly to the limonene + NO₃ reaction system, is not included in this primary HOM yield.

3.6 Contribution of HOM to particle formation and growth

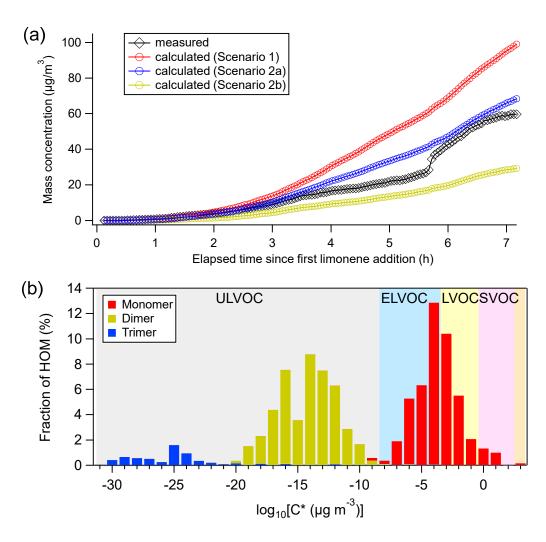


Figure 7. (a) Comparison of measured particle concentrations (black) against those predicted from condensation of measured HOM on aerosol particles, where red markers were calculated under Scenario 1, blue and yellow markers were calculated under Scenario 2 (only considering the condensation of ULVOC, ELVOC and LVOC), with the volatility calculated using the method by Mohr et al. (2019) (Scenario 2a) and Peräkylä et al. (2020) (Scenario 2b) respectively. (b) HOM volatility distribution using formula provided by Mohr et al. (2019). Average concentrations of HOM in the P1 period were used to calculate the fraction of HOM.

We observed nucleation and growth of SOA particles in the limonene $+\,NO_3$ reaction. We calculated the contribution

of HOM to SOA formation and particle growth and compared it to the measured particle growth (Fig. 7a). We assumed different scenarios of HOM uptake on aerosol particles, using the calculation methods described in the literature (Ehn et al., 2014; Seinfeld and Pandis, 2006; Nieminen et al., 2010). The assumption that all HOM irreversibly condense on the particles (Scenario 1) resulted in a strong overestimation of particle mass growth (red markers in Fig. 7a). Applying the parameterizations of Mohr et al. (2019) (Scenario 2a) or Peräkylä et al. (2020) (Scenario 2b) for classification and accounting only LVOC- and ULVOC/ELVOC-HOM for irreversible uptake framed the observed values (blue and yellow markers in Fig. 7a). While Scenario 2a agreed quite well with the observations and only slightly overestimated SOA concentrations after 7 h by +11%, Scenario 2b underestimated the SOA concentration at the end by -53%. The agreement between the modeled and observed SOA concentration suggests that HOM, and particularly LVOC- and ULVOC/ELVOC-HOM play a major role in growth of SOA particles in this study. This is consistent with the work by Faxon et al. (2018) who found that many of the dimers are ELVOC, which is also supported by our calculation result based on the method of Mohr et al. (2019).

642

643

644

645

646

647 648

649

650

651

652

653

654

655656

657

658

659

660 661

662

663

664

665

666

667

668 669

670 671

672

673

674

675676

Since neither SO₂ nor H₂SO₄ was added in our experiment, new particle formation (NPF) could be attributed to the nucleation initiated by HOM of low volatility. HOM trimers with as many as 30 carbon atoms were identified in the early stage of this study, and their sudden loss matched the onset of rapid formation of SOA. Trimers identified in our experiment are classified as ULVOC/ELVOC, with much lower volatility than monomers and dimers (Fig. 7b). Therefore, NPF in the current study can more likely be attributed to HOM trimers since they have the strongest potential of initiating nucleation, although we cannot rule out some contributions of dimers in the NPF. In contrast, in an earlier experiment investigating the NO₃-initiated oxidation of β-pinene also conducted in the SAPHIR chamber under similar conditions, new particles were barely formed (<20 cm⁻³) (Shen et al., 2021). As already mentioned above, no trimer HOM products were observed in that study, and only molecules with C≤20 were detected (Sect. 3.4). Extremely low volatile organic vapors formed in α -pinene ozonolysis have been shown to induce nucleation and drive initial particle growth in the atmosphere (Tröstl et al., 2016; Kirkby et al., 2016). Since our experiment of NO₃ oxidation of limonene was performed under near atmospheric conditions, such NPF events induced by the oxidation of limonene by NO₃ could also occur in the ambient atmosphere. Although monoterpene concentrations in this study (0-0.92 ppbv) are higher than in most ambient regions, they are still in the range of ambient concentrations ($\sim 0.01-1$ ppby) (e.g. Coggon et al., 2021; Wang et al., 2022), especially for forested regions (e.g. Xu et al., 2015; Kontkanen et al., 2016; Janson, 1992). Assuming that dimers react with NO₃ at a rate similar to limonene, and that they have a condensation sink similar to H₂SO₄ (10⁻³-10⁻¹ s⁻¹) (Dada et al., 2020), the lifetime with respect to NO₃ at an NO₃ concentration of 5-300 ppt and to condensation on particles are ~0.1-10 min and ~0.1-20 min, respectively. Therefore, although aerosols may scavenge HOM dimers in the ambient atmosphere, dimers can still react with NO₃ at nighttime, forming trimers. Such reactions are particularly important when the ambient aerosol concentration is low. Several field observations have shown NPF events taking place at nighttime where biogenic emissions dominate (Kammer et al., 2018; Huang et al., 2019). The work by Ortega et al. (2012) demonstrated an important role of monoterpene ozonolysis products in nocturnal NPF events in chamber experiments. In a previous laboratory study, limonene +

 NO_3 appears more effective at initiating nucleation than the limonene + O_3 reaction (Fry et al., 2014), which supports that limonene + NO_3 can play a significant role in nighttime nucleation. Our study suggests that NO_3 oxidation of limonene could contribute to the nighttime NPF via HOM trimer formation. In contrast, we infer that NO_3 reactions with other monoterpenes containing only one double bond such as α -pinene and β -pinene are less likely candidates for nighttime NPF, because gas-phase trimers are not observed.

4 Conclusion and implications

HOM formation in the reaction of limonene with NO₃ was investigated in the SAPHIR chamber. About 280 gasphase HOM products were identified, including monomers (C_{6-10} , O_{6-16} , N_{0-3}), dimers (C_{17-20} , O_{7-20} , N_{0-4}) and trimers (C_{27-30} , O_{16-25} , N_{1-6}). Nitrogen-containing products dominated the HOM, with compounds of the $C_{10}H_{15-17}NO_{6-14}$ series being the most prevalent. Dimers contributed 47 % in the early stage of the experiment when particle surface concentration was rather low ($< 6 \times 10^4 \text{ nm}^2 \text{ cm}^{-3}$), which was similar to monomers (47 %). Tentative formation pathways of major families were proposed in this work based on their time-dependent concentration profiles.

In HOM monomers, the abundance of carbonyl compounds significantly exceeded that of hydroxy or hydroperoxy compounds, indicating the significance of unimolecular termination of HOM-RO₂ radicals. Both RO₂• autoxidation and alkoxy-peroxy pathways were found to be important in the formation of HOM monomers. Monomers with 1 nitrogen atom (1N-monomers) contained both first- and second-generation products, which could be formed via NO₃ oxidation of limonene and its first-generation products with the latter being more important. Monomers with 2 nitrogen atoms were classified as second-generation products, which could be formed via NO₃ oxidation of the remaining C=C double bond of 1N-monomers.

Dimers showed both first- and second-generation time pattern. Dimers were mostly formed via accretion reactions between monomer RO₂ radicals, resulting in a decrease in O/C ratio compared to monomers. The initial less oxygenated RO₂•, including the C₁₀H₁₆NO₅• radical that cannot be observed in our instrument, likely played an important role in dimer formation based on the comparison of measured dimers against expected dimer identity and concentrations according to accretion monomer RO₂• reactions. Trimers were likely formed via accretion reactions between monomer RO₂ and dimer RO₂ radicals formed from secondary reactions of dimers with NO₃. Trimer formation is thus linked to the presence of two double bonds in limonene, of which the first reacts with NO₃ leading to dimer products while the remaining C=C double bond provides a reactive site for further oxidation of the dimers by NO₃, forming dimer RO₂ radicals.

A "primary" HOM molar yield of 1.5 $\%_{0.7\%}^{+1.7\%}$ in the limonene + NO₃ reaction was estimated, including only the first-generation HOM. Second-generation HOM contributed greatly to monomers, dimers and trimers, and hence the HOM yield we obtained is a lower limit of the total HOM yield, and is likewise much lower than the total HOM yield in the reaction of limonene with ozone (5 to 17 %) (Ehn et al., 2014; Jokinen et al., 2015; Pagonis et al., 2019).

NPF observed in this work was likely related to the trimer formation due to much lower volatility of trimers compared to monomers and dimers. The SOA concentration in the limonene + NO₃ reaction could be explained by

the condensation of the HOM belonging to LVOC and ULVOC/ELVOC classes assuming irreversible uptake, indicating an important role of HOM for growth of SOA particles in this reaction system. To our knowledge, this work is the first identifying trimer products from the limonene + NO₃ reaction system, suggesting that limonene + NO₃ is a possible crucial source of new particles formed in nighttime biogenic emission-dominated areas (Kammer et al., 2018; Huang et al., 2019). Our work highlights the need to consider the role of limonene + NO₃ in NPF in models simulating nighttime aerosols formation in biogenic-emission dominated areas, especially with large limonene emissions. In addition, comparison with the reactions of NO₃ with isoprene (Zhao et al., 2021) and other monoterpenes (Shen et al., 2021) reveals a strong dependence of HOM products on the molecular structure of the VOC species in NO₃-initiated chemistry.

The concentration of limonene and NO₃ in this study were on the order of few ppb and a few to one hundred ppt, respectively, which are similar to the ambient levels in rural and forest regions affected by anthropogenic emissions (Brown and Stutz, 2012). The chemical lifetime of RO₂• was of the order of 50 to 500 s, which is also similar to ambient conditions at nighttime (Fry et al., 2018). The RO₂• loss pathway in our study was dominated by the reactions RO₂• + NO₃ and RO₂• + RO₂•, which is relevant for the RO₂• fate in urban areas and forested areas influenced by an urban plume at nighttime. However, in more pristine forested regions, the RO₂• fate is mostly determined by RO₂• + HO₂ and RO₂• + RO₂•, as shown by Bates et al. (2022) for the example of a Southeast US forest. As NO₃ concentration is generally enhanced with increased anthropogenic emissions, RO₂• + NO₃ will become more important going from remote to urban areas. Therefore, the HOM products and their formation process in our study are relevant for rural and forested regions influenced by anthropogenic plumes and ambient urban regions with high volatile commercial products emissions as limonene is a typical component of volatile chemical products (VCP) (Nazaroff and Weschler, 2004). In these regions, HOM from monoterpene + NO₃ reactions can be major components of nighttime SOA. As nitrooxy-RO₂ fate can strongly affect the oxidation product distribution and SOA yield as shown for the reaction of α-pinene with NO₃ (Bates et al., 2022), more studies of HOM formation by NO₃ at various RO₂• fates are needed to be representative of various environment including (remote) forested regions.

This study also highlights the important role of second-generation chemistry in HOM formation, which needs to be further investigated and should be included in chemical mechanism used in numerical models. Additional work is also needed to investigate the role of different HOM formed via NO_3 -initiated BVOC oxidation reactions in NPF and growth of SOA particles in order to better constrain the climatic and environmental effects of BVOC + NO_3 chemistry.

Acknowledgement

743 Y. Guo, H. Shen, H. Luo, and D. Zhao would like to thank the funding support of Science and Technology

Commission of Shanghai Municipality (No. 20230711400), National Natural Science Foundation of China (No.

745 41875145), and Shanghai International Science and Technology Partnership Project (No. 21230780200). Sungah

Kang, Astrid Kiendler-Scharr, and Thomas F. Mentel acknowledge the support by the EU Project FORCeS (grant

747 agreement no. 821205).

References

748

- Ayres, B. R., Allen, H. M., Draper, D. C., Brown, S. S., Wild, R. J., Jimenez, J. L., Day, D. A., Campuzano-Jost, P., Hu, W.,
- de Gouw, J., Koss, A., Cohen, R. C., Duffey, K. C., Romer, P., Baumann, K., Edgerton, E., Takahama, S., Thornton,
- J. A., Lee, B. H., Lopez-Hilfiker, F. D., Mohr, C., Wennberg, P. O., Nguyen, T. B., Teng, A., Goldstein, A. H.,
- Olson, K., and Fry, J. L.: Organic nitrate aerosol formation via NO₃ + biogenic volatile organic compounds in the southeastern United States, Atmos. Chem. Phys., 15, 13377-13392, 10.5194/acp-15-13377-2015, 2015.
- Bates, K. H., Burke, G. J. P., Cope, J. D., and Nguyen, T. B.: Secondary organic aerosol and organic nitrogen yields from the nitrate radical (NO₃) oxidation of alpha-pinene from various RO₂ fates, Atmos. Chem. Phys., 22, 1467-1482,
- 756 10.5194/acp-22-1467-2022, 2022.
- 757 Beaver, M. R., Clair, J. M. S., Paulot, F., Spencer, K. M., Crounse, J. D., LaFranchi, B. W., Min, K. E., Pusede, S. E.,
- Wooldridge, P. J., Schade, G. W., Park, C., Cohen, R. C., and Wennberg, P. O.: Importance of biogenic precursors
- to the budget of organic nitrates: observations of multifunctional organic nitrates by CIMS and TD-LIF during BEARPEX 2009, Atmos. Chem. Phys., 12, 5773-5785, 10.5194/acp-12-5773-2012, 2012.
- Bell, D. M., Wu, C., Bertrand, A., Graham, E., Schoonbaert, J., Giannoukos, S., Baltensperger, U., Prevot, A. S. H., Riipinen,
- I., El Haddad, I., and Mohr, C.: Particle-phase processing of α-pinene NO3 secondary organic aerosol in the dark,
 Atmos. Chem. Phys. Discuss., 2021, 1-28, 10.5194/acp-2021-379, 2021.
- Berkemeier, T., Takeuchi, M., Eris, G., and Ng, N. L.: Kinetic modeling of formation and evaporation of secondary organic
- 765 aerosol from NO₃ oxidation of pure and mixed monoterpenes, Atmos. Chem. Phys., 20, 15513-15535, 10.5194/acp-20-15513-2020, 2020.
- Berndt, T., Mender, B., Scholz, W., Fischer, L., Herrmann, H., Kulmala, M., and Hansel, A.: Accretion Product Formation
- from Ozonolysis and OH Radical Reaction of α-Pinene: Mechanistic Insight and the Influence of Isoprene and Ethylene, Environ. Sci. Technol., 52, 11069-11077, 10.1021/acs.est.8b02210, 2018a.
- 770 Berndt, T., Scholz, W., Mentler, B., Fischer, L., Herrmann, H., Kulmala, M., and Hansel, A.: Accretion Product Formation
- from Self- and Cross-Reactions of RO₂ Radicals in the Atmosphere, Angew. Chem. Int. Edit., 57, 3820-3824, 10.1002/anie.201710989, 2018b.
- High Bianchi, F., Kurtén, T., Riva, M., Mohr, C., Rissanen, M. P., Roldin, P., Berndt, T., Crounse, J. D., Wennberg, P. O., Mentel,
- T. F., Wildt, J., Junninen, H., Jokinen, T., Kulmala, M., Worsnop, D. R., Thornton, J. A., Donahue, N., Kjaergaard,
- 775 H. G., and Ehn, M.: Highly Oxygenated Organic Molecules (HOM) from Gas-Phase Autoxidation Involving
- Peroxy Radicals: A Key Contributor to Atmospheric Aerosol, Chem. Rev., 119, 3472-3509,
- 777 10.1021/acs.chemrev.8b00395, 2019.
- 778 Boyd, C. M., Nah, T., Xu, L., Berkemeier, T., and Ng, N. L.: Secondary Organic Aerosol (SOA) from Nitrate Radical
- Oxidation of Monoterpenes: Effects of Temperature, Dilution, and Humidity on Aerosol Formation, Mixing, and Evaporation, Environ. Sci. Technol., 51, 7831-7841, 10.1021/acs.est.7b01460, 2017.
- Boyd, C. M., Sanchez, J., Xu, L., Eugene, A. J., Nah, T., Tuet, W. Y., Guzman, M. I., and Ng, N. L.: Secondary organic
- aerosol formation from the β-pinene+NO₃ system: effect of humidity and peroxy radical fate, Atmos. Chem. Phys.,
- 783 15, 7497-7522, 10.5194/acp-15-7497-2015, 2015.
- Brown, S. S. and Stutz, J.: Nighttime radical observations and chemistry, Chem. Soc. Rev., 41, 6405-6447, 10.1039/c2cs35181a, 2012.
- Carslaw, N., Mota, T., Jenkin, M. E., Barley, M. H., and McFiggans, G.: A significant role for nitrate and peroxide groups on indoor secondary organic aerosol, Environ. Sci. Technol., 46, 9290-9298, 10.1021/es301350x, 2012.
- 788 Chen, Y., Takeuchi, M., Nah, T., Xu, L., Canagaratna, M. R., Stark, H., Baumann, K., Canonaco, F., Prévôt, A. S. H., Huey,
- 789 L. G., Weber, R. J., and Ng, N. L.: Chemical characterization of secondary organic aerosol at a rural site in the
- southeastern US: insights from simultaneous high-resolution time-of-flight aerosol mass spectrometer (HR-ToF-
- AMS) and FIGAERO chemical ionization mass spectrometer (CIMS) measurements, Atmos. Chem. Phys., 20,

- 792 8421-8440, 10.5194/acp-20-8421-2020, 2020.
- 793 Claflin, M. S. and Ziemann, P. J.: Identification and Quantitation of Aerosol Products of the Reaction of β-Pinene with NO₃
- Radicals and Implications for Gas- and Particle-Phase Reaction Mechanisms, J. Phys. Chem. A, 122, 3640-3652, 10.1021/acs.jpca.8b00692, 2018.
- Clausen, P. A., Wilkins, C. K., Wolkoff, P., and Nielsen, G. D.: Chemical and biological evaluation of a reaction mixture of R-(+)-limonene/ozone Formation of strong airway irritants, Environ. Int., 26, 511-522, 10.1016/s0160-4120(01)00035-6, 2001.
- Coggon, M. M., Gkatzelis, G. I., McDonald, B. C., Gilman, J. B., Schwantes, R. H., Abuhassan, N., Aikin, K. C., Arend,
 M. F., Berkoff, T. A., Brown, S. S., Campos, T. L., Dickerson, R. R., Gronoff, G., Hurley, J. F., Isaacman-VanWertz,
- G., Koss, A. R., Li, M., McKeen, S. A., Moshary, F., Peischl, J., Pospisilova, V., Ren, X., Wilson, A., Wu, Y.,
- Trainer, M., and Warneke, C.: Volatile chemical product emissions enhance ozone and modulate urban chemistry, Proc. Nat. Acad. Sci. U.S.A., 118, 10.1073/pnas.2026653118, 2021.
- Crounse, J. D., Nielsen, L. B., Jørgensen, S., Kjaergaard, H. G., and Wennberg, P. O.: Autoxidation of Organic Compounds in the Atmosphere, J. Phys. Chem. Lett., 4, 3513-3520, 10.1021/jz4019207, 2013.
- Dada, L., Ylivinkka, I., Baalbaki, R., Li, C., Guo, Y., Yan, C., Yao, L., Sarnela, N., Jokinen, T., Daellenbach, K. R., Yin, R., Deng, C., Chu, B., Nieminen, T., Wang, Y., Lin, Z., Thakur, R. C., Kontkanen, J., Stolzenburg, D., Sipilä, M.,
- Hussein, T., Paasonen, P., Bianchi, F., Salma, I., Weidinger, T., Pikridas, M., Sciare, J., Jiang, J., Liu, Y., Petäjä,
- Tusselii, 1., Taasolieli, 1., Blailelli, 1., Sallia, 1., Weldinger, 1., Tikhdas, W., Serare, 3., Slang, 3., Etd, 1., Tetaja,

 T., Kerminen, V.-M., and Kulmala, M.: Sources and sinks driving sulfuric acid concentrations in contrasting
- environments: implications on proxy calculations, Atmos. Chem. Phys., 20, 11747-11766, 10.5194/acp-20-11747-
- 811 2020, 2020.
- Dam, M., Draper, D. C., Marsavin, A., Fry, J. L., and Smith, J. N.: Observations of gas-phase products from the nitrateradical-initiated oxidation of four monoterpenes, Atmos. Chem. Phys., 22, 9017-9031, 10.5194/acp-22-9017-2022, 2022.
- Donahue, N. M., Epstein, S. A., Pandis, S. N., and Robinson, A. L.: A two-dimensional volatility basis set: 1. organic-aerosol mixing thermodynamics, Atmos. Chem. Phys., 11, 3303-3318, 10.5194/acp-11-3303-2011, 2011.
- Donahue, N. M., Kroll, J. H., Pandis, S. N., and Robinson, A. L.: A two-dimensional volatility basis set Part 2: Diagnostics of organic-aerosol evolution, Atmos. Chem. Phys., 12, 615-634, 10.5194/acp-12-615-2012, 2012.
- 819 Ehn, M., Thornton, J. A., Kleist, E., Sipilä, M., Junninen, H., Pullinen, I., Springer, M., Rubach, F., Tillmann, R., Lee, B.,
- Lopez-Hilfiker, F., Andres, S., Acir, I. H., Rissanen, M., Jokinen, T., Schobesberger, S., Kangasluoma, J.,
- Kontkanen, J., Nieminen, T., Kurtén, T., Nielsen, L. B., Jørgensen, S., Kjaergaard, H. G., Canagaratna, M., Dal Maso, M., Berndt, T., Petäjä, T., Wahner, A., Kerminen, V. M., Kulmala, M., Worsnop, D. R., Wildt, J., and Mentel,
- T. F.: A large source of low-volatility secondary organic aerosol, Nature, 506, 476-485, 10.1038/nature13032, 2014.
- Eisele, F. L. and Tanner, D. J.: Measurement of the gas phase concentration of H₂SO₄ and methane sulfonic acid and estimates of H₂SO₄ production and loss in the atmosphere, J. Geophys. Res.-Atmos., 98, 9001-9010, 10.1029/93jd00031, 1993.
- Fan, Z. H., Lioy, P., Weschler, C., Fiedler, N., Kipen, H., and Zhang, J. F.: Ozone-initiated reactions with mixtures of volatile organic compounds under simulated indoor conditions, Environ. Sci. Technol., 37, 1811-1821, 10.1021/es026231i, 2003.
- Faxon, C., Hammes, J., Le Breton, M., Pathak, R. K., and Hallquist, M.: Characterization of organic nitrate constituents of secondary organic aerosol (SOA) from nitrate-radical-initiated oxidation of limonene using high-resolution chemical ionization mass spectrometry, Atmos. Chem. Phys., 18, 5467-5481, 10.5194/acp-18-5467-2018, 2018.
- Finlayson-Pitts, B. J. and Pitts, J. N.: Tropospheric air pollution: Ozone, airborne toxics, polycyclic aromatic hydrocarbons, and particles, Science, 276, 1045-1052, 10.1126/science.276.5315.1045, 1997.

- Fry, J. L., Draper, D. C., Barsanti, K. C., Smith, J. N., Ortega, J., Winkler, P. M., Lawler, M. J., Brown, S. S., Edwards, P.
- M., Cohen, R. C., and Lee, L.: Secondary organic aerosol formation and organic nitrate yield from NO₃ oxidation
- of biogenic hydrocarbons, Environ. Sci. Technol., 48, 11944-11953, 10.1021/es502204x, 2014.
- Fry, J. L., Kiendler-Scharr, A., Rollins, A. W., Wooldridge, P. J., Brown, S. S., Fuchs, H., Dubé, W., Mensah, A., dal Maso,
- M., Tillmann, R., Dorn, H. P., Brauers, T., and Cohen, R. C.: Organic nitrate and secondary organic aerosol yield
- from NO₃ oxidation of β-pinene evaluated using a gas-phase kinetics/aerosol partitioning model, Atmos. Chem.
- Phys., 9, 1431-1449, 10.5194/acp-9-1431-2009, 2009.
- Fry, J. L., Kiendler-Scharr, A., Rollins, A. W., Brauers, T., Brown, S. S., Dorn, H. P., Dubé, W. P., Fuchs, H., Mensah, A.,
- Rohrer, F., Tillmann, R., Wahner, A., Wooldridge, P. J., and Cohen, R. C.: SOA from limonene: role of NO₃ in its
- generation and degradation, Atmos. Chem. Phys., 11, 3879-3894, 10.5194/acp-11-3879-2011, 2011.
- Fry, J. L., Draper, D. C., Zarzana, K. J., Campuzano-Jost, P., Day, D. A., Jimenez, J. L., Brown, S. S., Cohen, R. C., Kaser,
- L., Hansel, A., Cappellin, L., Karl, T., Hodzic Roux, A., Turnipseed, A., Cantrell, C., Lefer, B. L., and Grossberg,
- N.: Observations of gas- and aerosol-phase organic nitrates at BEACHON-RoMBAS 2011, Atmos. Chem. Phys.,
- 849 13, 8585-8605, 10.5194/acp-13-8585-2013, 2013.
- Fry, J. L., Brown, S. S., Middlebrook, A. M., Edwards, P. M., Campuzano-Jost, P., Day, D. A., Jimenez, J. L., Allen, H. M.,
- Ryerson, T. B., Pollack, I., Graus, M., Warneke, C., de Gouw, J. A., Brock, C. A., Gilman, J., Lerner, B. M., Dubé,
- W. P., Liao, J., and Welti, A.: Secondary organic aerosol (SOA) yields from NO₃ radical + isoprene based on
- nighttime aircraft power plant plume transects, Atmos. Chem. Phys., 18, 11663-11682, 10.5194/acp-18-11663-
- 854 2018, 2018.
- Fuchs, H., Hofzumahaus, A., Rohrer, F., Bohn, B., Brauers, T., Dorn, H. P., Häseler, R., Holland, F., Kaminski, M., Li, X.,
- Lu, K., Nehr, S., Tillmann, R., Wegener, R., and Wahner, A.: Experimental evidence for efficient hydroxyl radical
- regeneration in isoprene oxidation, Nat. Geosci., 6, 1023-1026, 10.1038/ngeo1964, 2013.
- Fuchs, N. A. and Sutugin, A. G.: Topics in Current Aerosol Research (Part 2), Pergamon, New York, 1971.
- 659 Gkatzelis, G. I., Coggon, M. M., McDonald, B. C., Peischl, J., Aikin, K. C., Gilman, J. B., Trainer, M., and Warneke, C.:
- Identifying Volatile Chemical Product Tracer Compounds in U.S. Cities, Environ. Sci. Technol., 55, 188-199,
- 861 10.1021/acs.est.0c05467, 2021.
- Guenther, A. B., Jiang, X., Heald, C. L., Sakulyanontvittaya, T., Duhl, T., Emmons, L. K., and Wang, X.: The Model of
- Emissions of Gases and Aerosols from Nature version 2.1 (MEGAN2.1): an extended and updated framework for
- modeling biogenic emissions, Geosci. Model Dev., 5, 1471-1492, 10.5194/gmd-5-1471-2012, 2012.
- Hallquist, M., Wängberg, I., Ljungström, E., Barnes, I., and Becker, K. H.: Aerosol and product yields from NO₃ radicalinitiated oxidation of selected monoterpenes, Environ. Sci. Technol., 33, 553-559, 10.1021/es980292s, 1999.
- Hallquist, M., Wenger, J. C., Baltensperger, U., Rudich, Y., Simpson, D., Claeys, M., Dommen, J., Donahue, N. M., George,
- 868 C., Goldstein, A. H., Hamilton, J. F., Herrmann, H., Hoffmann, T., Iinuma, Y., Jang, M., Jenkin, M. E., Jimenez,
- J. L., Kiendler-Scharr, A., Maenhaut, W., McFiggans, G., Mentel, T. F., Monod, A., Prévôt, A. S. H., Seinfeld, J.
- H., Surratt, J. D., Szmigielski, R., and Wildt, J.: The formation, properties and impact of secondary organic aerosol:
- 871 current and emerging issues, Atmos. Chem. Phys., 9, 5155-5236, 10.5194/acp-9-5155-2009, 2009.
- Huang, W., Saathoff, H., Shen, X., Ramisetty, R., Leisner, T., and Mohr, C.: Chemical Characterization of Highly
- Functionalized Organonitrates Contributing to Night-Time Organic Aerosol Mass Loadings and Particle Growth,
- 874 Environ. Sci. Technol., 53, 1165-1174, 10.1021/acs.est.8b05826, 2019.
- 875 Hyttinen, N., Kupiainen-Määttä, O., Rissanen, M. P., Muuronen, M., Ehn, M., and Kurtén, T.: Modeling the Charging of
- Highly Oxidized Cyclohexene Ozonolysis Products Using Nitrate-Based Chemical Ionization, J. Phys. Chem. A,
- 877 119, 6339-6345, 10.1021/acs.jpca.5b01818, 2015.
- 878 Janson, R.: Monoterpene concentrations in and above a forest of scots pine, J. Atmos. Chem., 14, 385-394,
- 879 10.1007/bf00115246, 1992.

- Jenkin, M. E., Saunders, S. M., and Pilling, M. J.: The tropospheric degradation of volatile organic compounds: A protocol for mechanism development, Atmos. Environ., 31, 81-104, 10.1016/s1352-2310(96)00105-7, 1997.
- Jiang, L., Wang, W., and Xu, Y. S.: Theoretical investigation of the NO₃ radical addition to double bonds of limonene, Int.

 J. Mol. Sci., 10, 3743-3754, 10.3390/ijms10093743, 2009.
- Jokinen, T., Sipilä, M., Junninen, H., Ehn, M., Lönn, G., Hakala, J., Petäjä, T., Mauldin, R. L., Kulmala, M., and Worsnop,
 D. R.: Atmospheric sulphuric acid and neutral cluster measurements using CI-APi-TOF, Atmos. Chem. Phys., 12,
 4117-4125, 10.5194/acp-12-4117-2012, 2012.
- Jokinen, T., Berndt, T., Makkonen, R., Kerminen, V. M., Junninen, H., Paasonen, P., Stratmann, F., Herrmann, H., Guenther, A. B., Worsnop, D. R., Kulmala, M., Ehn, M., and Sipilä, M.: Production of extremely low volatile organic compounds from biogenic emissions: Measured yields and atmospheric implications, P. Natl. Acad. Sci. USA, 112, 7123-7128, 10.1073/pnas.1423977112, 2015.
- Kammer, J., Perraudin, E., Flaud, P. M., Lamaud, E., Bonnefond, J. M., and Villenave, E.: Observation of nighttime new particle formation over the French Landes forest, Sci. Total Environ., 621, 1084-1092, 10.1016/j.scitotenv.2017.10.118, 2018.
- 894 Kirkby, J., Duplissy, J., Sengupta, K., Frege, C., Gordon, H., Williamson, C., Heinritzi, M., Simon, M., Yan, C., Almeida, 895 J., Tröstl, J., Nieminen, T., Ortega, I. K., Wagner, R., Adamov, A., Amorim, A., Bernhammer, A. K., Bianchi, F., 896 Breitenlechner, M., Brilke, S., Chen, X., Craven, J., Dias, A., Ehrhart, S., Flagan, R. C., Franchin, A., Fuchs, C., 897 Guida, R., Hakala, J., Hoyle, C. R., Jokinen, T., Junninen, H., Kangasluoma, J., Kim, J., Krapf, M., Kürten, A., 898 Laaksonen, A., Lehtipalo, K., Makhmutov, V., Mathot, S., Molteni, U., Onnela, A., Peräkylä, O., Piel, F., Petäjä, 899 T., Praplan, A. P., Pringle, K., Rap, A., Richards, N. A., Riipinen, I., Rissanen, M. P., Rondo, L., Sarnela, N., 900 Schobesberger, S., Scott, C. E., Seinfeld, J. H., Sipilä, M., Steiner, G., Stozhkov, Y., Stratmann, F., Tomé, A., 901 Virtanen, A., Vogel, A. L., Wagner, A. C., Wagner, P. E., Weingartner, E., Wimmer, D., Winkler, P. M., Ye, P., 902 Zhang, X., Hansel, A., Dommen, J., Donahue, N. M., Worsnop, D. R., Baltensperger, U., Kulmala, M., Carslaw, 903 K. S., and Curtius, J.: Ion-induced nucleation of pure biogenic particles, Nature, 533, 521-526, 904 10.1038/nature17953, 2016.
- Klinger, L. F., Li, Q. J., Guenther, A. B., Greenberg, J. P., Baker, B., and Bai, J. H.: Assessment of volatile organic compound emissions from ecosystems of China, J. Geophys. Res.-Atmos., 107, 10.1029/2001jd001076, 2002.
- Kontkanen, J., Paasonen, P., Aalto, J., Bäck, J., Rantala, P., Petäjä, T., and Kulmala, M.: Simple proxies for estimating the concentrations of monoterpenes and their oxidation products at a boreal forest site, Atmos. Chem. Phys., 16, 13291-13307, 10.5194/acp-16-13291-2016, 2016.
- Kulmala, M., Petäjä, T., Nieminen, T., Sipilä, M., Manninen, H. E., Lehtipalo, K., Dal Maso, M., Aalto, P. P., Junninen, H.,
 Paasonen, P., Riipinen, I., Lehtinen, K. E. J., Laaksonen, A., and Kerminen, V.-M.: Measurement of the nucleation of atmospheric aerosol particles, Nat. Protoc., 7, 1651-1667, 10.1038/nprot.2012.091, 2012.
- Kurtén, T., Møller, K. H., Nguyen, T. B., Schwantes, R. H., Misztal, P. K., Su, L., Wennberg, P. O., Fry, J. L., and Kjaergaard,
 H. G.: Alkoxy Radical Bond Scissions Explain the Anomalously Low Secondary Organic Aerosol and
 Organonitrate Yields From α-Pinene + NO₃, J. Phys. Chem. Lett., 8, 2826-2834, 10.1021/acs.jpclett.7b01038,
 2017.
- Lee, B. H., Mohr, C., Lopez-Hilfiker, F. D., Lutz, A., Hallquist, M., Lee, L., Romer, P., Cohen, R. C., Iyer, S., Kurten, T., Hu, W., Day, D. A., Campuzano-Jost, P., Jimenez, J. L., Xu, L., Ng, N. L., Guo, H., Weber, R. J., Wild, R. J.,
- Brown, S. S., Koss, A., de Gouw, J., Olson, K., Goldstein, A. H., Seco, R., Kim, S., McAvey, K., Shepson, P. B.,
- 920 Starn, T., Baumann, K., Edgerton, E. S., Liu, J., Shilling, J. E., Miller, D. O., Brune, W., Schobesberger, S.,
- 921 D'Ambro, E. L., and Thornton, J. A.: Highly functionalized organic nitrates in the southeast United States:
- Contribution to secondary organic aerosol and reactive nitrogen budgets, P. Natl. Acad. Sci. USA, 113, 1516-1521,
- 923 10.1073/pnas.1508108113, 2016.

- 924 Liu, Y., Nie, W., Li, Y., Ge, D., Liu, C., Xu, Z., Chen, L., Wang, T., Wang, L., Sun, P., Qi, X., Wang, J., Xu, Z., Yuan, J.,
- Yan, C., Zhang, Y., Huang, D., Wang, Z., Donahue, N. M., Worsnop, D., Chi, X., Ehn, M., and Ding, A.: Formation
- of condensable organic vapors from anthropogenic and biogenic volatile organic compounds (VOCs) is strongly
- 927 perturbed by NO_x in eastern China, Atmos. Chem. Phys., 21, 14789-14814, 10.5194/acp-21-14789-2021, 2021.
- 928 Massoli, P., Stark, H., Canagaratna, M. R., Krechmer, J. E., Xu, L., Ng, N. L., Mauldin, R. L., Yan, C., Kimmel, J., Misztal,
- P. K., Jimenez, J. L., Jayne, J. T., and Worsnop, D. R.: Ambient Measurements of Highly Oxidized Gas-Phase
- Molecules during the Southern Oxidant and Aerosol Study (SOAS) 2013, ACS Earth Space Chem., 2, 653-672,
- 931 10.1021/acsearthspacechem.8b00028, 2018.
- 932 McDonald, B. C., de Gouw, J. A., Gilman, J. B., Jathar, S. H., Akherati, A., Cappa, C. D., Jimenez, J. L., Lee-Taylor, J.,
- Hayes, P. L., McKeen, S. A., Cui, Y. Y., Kim, S.-W., Gentner, D. R., Isaacman-VanWertz, G., Goldstein, A. H.,
- Harley, R. A., Frost, G. J., Roberts, J. M., Ryerson, T. B., and Trainer, M.: Volatile chemical products emerging as
- largest petrochemical source of urban organic emissions, Science, 359, 760-764, 10.1126/science.aaq0524, 2018.
- McFiggans, G., Mentel, T. F., Wildt, J., Pullinen, I., Kang, S., Kleist, E., Schmitt, S., Springer, M., Tillmann, R., Wu, C.,
- 257 Zhao, D. F., Hallquist, M., Faxon, C., Le Breton, M., Hallquist, A. M., Simpson, D., Bergström, R., Jenkin, M. E.,
- Ehn, M., Thornton, J. A., Alfarra, M. R., Bannan, T. J., Percival, C. J., Priestley, M., Topping, D., and Kiendler-
- Scharr, A.: Secondary organic aerosol reduced by mixture of atmospheric vapours, Nature, 565, 587-593,
- 940 10.1038/s41586-018-0871-y, 2019.
- 941 Mentel, T. F., Springer, M., Ehn, M., Kleist, E., Pullinen, I., Kurtén, T., Rissanen, M., Wahner, A., and Wildt, J.: Formation
- 942 of highly oxidized multifunctional compounds: autoxidation of peroxy radicals formed in the ozonolysis of
- 943 alkenes deduced from structure–product relationships, Atmos. Chem. Phys., 15, 6745-6765, 10.5194/acp-15-
- 944 6745-2015, 2015.
- Mohr, C., Thornton, J. A., Heitto, A., Lopez-Hilfiker, F. D., Lutz, A., Riipinen, I., Hong, J., Donahue, N. M., Hallquist, M.,
- Petäjä, T., Kulmala, M., and Yli-Juuti, T.: Molecular identification of organic vapors driving atmospheric
- 947 nanoparticle growth, Nat. Commun., 10, 4442, https://doi.org/10.1038/s41467-019-12473-2, 2019.
- 948 Mutzel, A., Zhang, Y., Böge, O., Rodigast, M., Kolodziejczyk, A., Wang, X., and Herrmann, H.: Importance of secondary
- organic aerosol formation of α-pinene, limonene, and m-cresol comparing day- and nighttime radical chemistry,
- 950 Atmos. Chem. Phys., 21, 8479-8498, 10.5194/acp-21-8479-2021, 2021.
- 951 Nah, T., Sanchez, J., Boyd, C. M., and Ng, N. L.: Photochemical Aging of α-pinene and β-pinene Secondary Organic
- Aerosol formed from Nitrate Radical Oxidation, Environ. Sci. Technol., 50, 222-231, 10.1021/acs.est.5b04594,
- 953 2016.
- Nazaroff, W. W. and Weschler, C. J.: Cleaning products and air fresheners: exposure to primary and secondary air pollutants,
- 955 Atmos. Environ., 38, 2841-2865, 10.1016/j.atmosenv.2004.02.040, 2004.
- 956 Ng, N. L., Kwan, A. J., Surratt, J. D., Chan, A. W. H., Chhabra, P. S., Sorooshian, A., Pye, H. O. T., Crounse, J. D., Wennberg,
- P. O., Flagan, R. C., and Seinfeld, J. H.: Secondary organic aerosol (SOA) formation from reaction of isoprene
- 958 with nitrate radicals (NO₃), Atmos. Chem. Phys., 8, 4117-4140, 10.5194/acp-8-4117-2008, 2008.
- 959 Nie, W., Yan, C., Huang, D. D., Wang, Z., Liu, Y., Qiao, X., Guo, Y., Tian, L., Zheng, P., Xu, Z., Li, Y., Xu, Z., Qi, X., Sun,
- 960 P., Wang, J., Zheng, F., Li, X., Yin, R., Dallenbach, K. R., Bianchi, F., Petäjä, T., Zhang, Y., Wang, M., Schervish,
- 961 M., Wang, S., Qiao, L., Wang, Q., Zhou, M., Wang, H., Yu, C., Yao, D., Guo, H., Ye, P., Lee, S., Li, Y. J., Liu, Y.,
- Chi, X., Kerminen, V.-M., Ehn, M., Donahue, N. M., Wang, T., Huang, C., Kulmala, M., Worsnop, D., Jiang, J.,
- 963 and Ding, A.: Secondary organic aerosol formed by condensing anthropogenic vapours over China's megacities,
- 964 Nat. Geosci., 15, 255-261, 10.1038/s41561-022-00922-5, 2022.
- Nieminen, T., Lehtinen, K. E. J., and Kulmala, M.: Sub-10 nm particle growth by vapor condensation effects of vapor
- 966 molecule size and particle thermal speed, Atmos. Chem. Phys., 10, 9773-9779, 10.5194/acp-10-9773-2010, 2010.
- 967 Novelli, A., Cho, C., Fuchs, H., Hofzumahaus, A., Rohrer, F., Tillmann, R., Kiendler-Scharr, A., Wahner, A., and Vereecken,

- L.: Experimental and theoretical study on the impact of a nitrate group on the chemistry of alkoxy radicals, Phys. Chem. Chem. Phys., 23, 5474-5495, 10.1039/d0cp05555g, 2021.
- Ortega, I. K., Suni, T., Boy, M., Grönholm, T., Manninen, H. E., Nieminen, T., Ehn, M., Junninen, H., Hakola, H., Hellén,
- H., Valmari, T., Arvela, H., Zegelin, S., Hughes, D., Kitchen, M., Cleugh, H., Worsnop, D. R., Kulmala, M., and
- 972 Kerminen, V. M.: New insights into nocturnal nucleation, Atmos. Chem. Phys., 12, 4297-4312, 10.5194/acp-12-973 4297-2012, 2012.
- Pagonis, D., Algrim, L. B., Price, D. J., Day, D. A., Handschy, A. V., Stark, H., Miller, S. L., de Gouw, J. A., Jimenez, J. L.,
 and Ziemann, P. J.: Autoxidation of Limonene Emitted in a University Art Museum, Environ. Sci. Tech. Let., 6,
 520-524, 10.1021/acs.estlett.9b00425, 2019.
- Peng, C., Wang, W., Li, K., Li, J., Zhou, L., Wang, L., and Ge, M.: The Optical Properties of Limonene Secondary Organic Aerosols: The Role of NO3, OH, and O3 in the Oxidation Processes, J. Geophys. Res.-Atmos., 123, 3292-3303, https://doi.org/10.1002/2017JD028090, 2018.
- Peräkylä, O., Riva, M., Heikkinen, L., Quéléver, L., Roldin, P., and Ehn, M.: Experimental investigation into the volatilities of highly oxygenated organic molecules (HOMs), Atmos. Chem. Phys., 20, 649-669, 10.5194/acp-20-649-2020, 2020.
- Pullinen, I., Schmitt, S., Kang, S., Sarrafzadeh, M., Schlag, P., Andres, S., Kleist, E., Mentel, T. F., Rohrer, F., Springer, M.,
 Tillmann, R., Wildt, J., Wu, C., Zhao, D., Wahner, A., and Kiendler-Scharr, A.: Impact of NO_x on secondary
 organic aerosol (SOA) formation from α-pinene and β-pinene photooxidation: the role of highly oxygenated
 organic nitrates, Atmos. Chem. Phys., 20, 10125-10147, 10.5194/acp-20-10125-2020, 2020.
- Pye, H. O. T., Chan, A. W. H., Barkley, M. P., and Seinfeld, J. H.: Global modeling of organic aerosol: the importance of reactive nitrogen (NO_x and NO₃), Atmos. Chem. Phys., 10, 11261-11276, 10.5194/acp-10-11261-2010, 2010.
- Rissanen, M. P., Kurtén, T., Sipilä, M., Thornton, J. A., Kangasluoma, J., Sarnela, N., Junninen, H., Jørgensen, S., Schallhart,
 S., Kajos, M. K., Taipale, R., Springer, M., Mentel, T. F., Ruuskanen, T., Petäjä, T., Worsnop, D. R., Kjaergaard,
 H. G., and Ehn, M.: The formation of highly oxidized multifunctional products in the ozonolysis of cyclohexene,
 J. Am. Chem. Soc., 136, 15596-15606, 10.1021/ja507146s, 2014.
- Riva, M., Rantala, P., Krechmer, J. E., Peräkylä, O., Zhang, Y., Heikkinen, L., Garmash, O., Yan, C., Kulmala, M., Worsnop,
 D., and Ehn, M.: Evaluating the performance of five different chemical ionization techniques for detecting gaseous
 oxygenated organic species, Atmos. Meas. Tech., 12, 2403-2421, 10.5194/amt-12-2403-2019, 2019.
- Rohrer, F., Bohn, B., Brauers, T., Brüning, D., Johnen, F. J., Wahner, A., and Kleffmann, J.: Characterisation of the
 photolytic HONO-source in the atmosphere simulation chamber SAPHIR, Atmos. Chem. Phys., 5, 2189-2201,
 10.5194/acp-5-2189-2005, 2005.
- Rollins, A. W., Browne, E. C., Min, K.-E., Pusede, S. E., Wooldridge, P. J., Gentner, D. R., Goldstein, A. H., Liu, S., Day, D. A., Russell, L. M., and Cohen, R. C.: Evidence for NO_x control over nighttime SOA formation, Science, 337, 1210-1212, 10.1126/science.1221520, 2012.
- Rollins, A. W., Kiendler-Scharr, A., Fry, J. L., Brauers, T., Brown, S. S., Dorn, H. P., Dubé, W. P., Fuchs, H., Mensah, A.,
 Mentel, T. F., Rohrer, F., Tillmann, R., Wegener, R., Wooldridge, P. J., and Cohen, R. C.: Isoprene oxidation by
 nitrate radical: alkyl nitrate and secondary organic aerosol yields, Atmos. Chem. Phys., 9, 6685-6703,
 10.5194/acp-9-6685-2009, 2009.
- Saunders, S. M., Jenkin, M. E., Derwent, R. G., and Pilling, M. J.: Protocol for the development of the Master Chemical
 Mechanism, MCM v3 (Part A): tropospheric degradation of non-aromatic volatile organic compounds, Atmos.
 Chem. Phys., 3, 161-180, 2003a.
- Saunders, S. M., Jenkin, M. E., Derwent, R. G., and Pilling, M. J.: Protocol for the development of the Master Chemical
 Mechanism, MCM v3 (Part A): tropospheric degradation of non-aromatic volatile organic compounds, Atmos.
 Chem. Phys., 3, 161-180, 10.5194/acp-3-161-2003, 2003b.

- Schervish, M. and Donahue, N. M.: Peroxy radical chemistry and the volatility basis set, Atmos. Chem. Phys., 20, 1183-1013 1199, 10.5194/acp-20-1183-2020, 2020.
- Seinfeld, J. H. and Pandis, S. N.: Atmospheric Chemistry and Physics: From Air Pollution to Climate Change, 2nd ed,
 Wiley, John & Sons, New York, 2006.
- Shen, H., Zhao, D., Pullinen, I., Kang, S., Vereecken, L., Fuchs, H., Acir, I.-H., Tillmann, R., Rohrer, F., Wildt, J., Kiendler-Scharr, A., Wahner, A., and Mentel, T. F.: Highly Oxygenated Organic Nitrates Formed from NO₃ Radical-Initiated
- Oxidation of β-Pinene, Environ. Sci. Technol., 10.1021/acs.est.1c03978, 2021.
- 1019 Shrivastava, M., Cappa, C. D., Fan, J., Goldstein, A. H., Guenther, A. B., Jimenez, J. L., Kuang, C., Laskin, A., Martin, S.
- T., Ng, N. L., Petäjä, T., Pierce, J. R., Rasch, P. J., Roldin, P., Seinfeld, J. H., Shilling, J., Smith, J. N., Thornton,
- J. A., Volkamer, R., Wang, J., Worsnop, D. R., Zaveri, R. A., Zelenyuk, A., and Zhang, Q.: Recent advances in understanding secondary organic aerosol: Implications for global climate forcing, Rev. Geophys., 55, 509-559,
- 1023 10.1002/2016rg000540, 2017.
- Slade, J. H., de Perre, C., Lee, L., and Shepson, P. B.: Nitrate radical oxidation of γ-terpinene: hydroxy nitrate, total organic nitrate, and secondary organic aerosol yields, Atmos. Chem. Phys., 17, 8635-8650, 10.5194/acp-17-8635-1026 2017, 2017.
- Spittler, M., Barnes, I., Bejan, I., Brockmann, K. J., Benter, T., and Wirtz, K.: Reactions of NO₃ radicals with limonene and α-pinene: Product and SOA formation, Atmos. Environ., 40, 116-127, 10.1016/j.atmosenv.2005.09.093, 2006.
- Takeuchi, M. and Ng, N. L.: Chemical composition and hydrolysis of organic nitrate aerosol formed from hydroxyl and nitrate radical oxidation of α-pinene and β-pinene, Atmos. Chem. Phys., 19, 12749-12766, 10.5194/acp-19-12749-1031 2019, 2019.
- 1032 Tröstl, J., Chuang, W. K., Gordon, H., Heinritzi, M., Yan, C., Molteni, U., Ahlm, L., Frege, C., Bianchi, F., Wagner, R.,
- Simon, M., Lehtipalo, K., Williamson, C., Craven, J. S., Duplissy, J., Adamov, A., Almeida, J., Bernhammer, A.
- 1034 K., Breitenlechner, M., Brilke, S., Dias, A., Ehrhart, S., Flagan, R. C., Franchin, A., Fuchs, C., Guida, R., Gysel,
- 1035 M., Hansel, A., Hoyle, C. R., Jokinen, T., Junninen, H., Kangasluoma, J., Keskinen, H., Kim, J., Krapf, M., Kürten,
- A., Laaksonen, A., Lawler, M., Leiminger, M., Mathot, S., Möhler, O., Nieminen, T., Onnela, A., Petäjä, T., Piel,
- 1037 F. M., Miettinen, P., Rissanen, M. P., Rondo, L., Sarnela, N., Schobesberger, S., Sengupta, K., Sipilä, M., Smith,
- J. N., Steiner, G., Tomè, A., Virtanen, A., Wagner, A. C., Weingartner, E., Wimmer, D., Winkler, P. M., Ye, P.,
- Carslaw, K. S., Curtius, J., Dommen, J., Kirkby, J., Kulmala, M., Riipinen, I., Worsnop, D. R., Donahue, N. M.,
- and Baltensperger, U.: The role of low-volatility organic compounds in initial particle growth in the atmosphere,
- Nature, 533, 527-531, 10.1038/nature18271, 2016.
- Vereecken, L. and Nozière, B.: H migration in peroxy radicals under atmospheric conditions, Atmos. Chem. Phys., 20, 7429-7458, 10.5194/acp-20-7429-2020, 2020.
- Vereecken, L. and Peeters, J.: Decomposition of substituted alkoxy radicals-part I: a generalized structure-activity relationship for reaction barrier heights, Phys. Chem. Chem. Phys., 11, 9062-9074, 10.1039/b909712k, 2009.
- Vereecken, L. and Peeters, J.: A structure-activity relationship for the rate coefficient of H-migration in substituted alkoxy radicals, Phys. Chem. Chem. Phys., 12, 12608-12620, 10.1039/c0cp00387e, 2010.
- 1048 Wagner, N. L., Dubé, W. P., Washenfelder, R. A., Young, C. J., Pollack, I. B., Ryerson, T. B., and Brown, S. S.: Diode laser-
- 1049 based cavity ring-down instrument for NO₃, N₂O₅, NO, NO₂ and O₃ from aircraft, Atmos. Meas. Tech., 4, 1227-1050 1240, 10.5194/amt-4-1227-2011, 2011.
- 1051 Wang, H., Ma, X., Tan, Z., Wang, H., Chen, X., Chen, S., Gao, Y., Liu, Y., Liu, Y., Yang, X., Yuan, B., Zeng, L., Huang, C.,
- Lu, K., and and Zhang, Y.: Anthropogenic monoterpenes aggravating ozone pollution, Natl. Sci. Rev., nwac103, 10.1093/nsr/nwac103, 2022.
- Wang, S. and Pratt, K. A.: Molecular Halogens Above the Arctic Snowpack: Emissions, Diurnal Variations, and Recycling
 Mechanisms, J. Geophys. Res.-Atmos., 122, 11991-12007, 10.1002/2017jd027175, 2017.

- 1056 Wu, C., Bell, D. M., Graham, E. L., Haslett, S., Riipinen, I., Baltensperger, U., Bertrand, A., Giannoukos, S., Schoonbaert,
- J., El Haddad, I., Prevot, A. S. H., Huang, W., and Mohr, C.: Photolytically induced changes in composition and
- volatility of biogenic secondary organic aerosol from nitrate radical oxidation during night-to-day transition,
- 1059 Atmos. Chem. Phys., 21, 14907-14925, 10.5194/acp-21-14907-2021, 2021a.
- 1060 Wu, R., Vereecken, L., Tsiligiannis, E., Kang, S., Albrecht, S. R., Hantschke, L., Zhao, D., Novelli, A., Fuchs, H., Tillmann,
- 1061 R., Hohaus, T., Carlsson, P. T. M., Shenolikar, J., Bernard, F., Crowley, J. N., Fry, J. L., Brownwood, B., Thornton,
- J. A., Brown, S. S., Kiendler-Scharr, A., Wahner, A., Hallquist, M., and Mentel, T. F.: Molecular composition and
- volatility of multi-generation products formed from isoprene oxidation by nitrate radical, Atmos. Chem. Phys.,
- 1064 21, 10799-10824, 10.5194/acp-21-10799-2021, 2021b.
- 1065 Xu, L., Guo, H., Boyd, C. M., Klein, M., Bougiatioti, A., Cerully, K. M., Hite, J. R., Isaacman-VanWertz, G., Kreisberg,
- N. M., Knote, C., Olson, K., Koss, A., Goldstein, A. H., Hering, S. V., de Gouw, J., Baumann, K., Lee, S.-H.,
- Nenes, A., Weber, R. J., and Ng, N. L.: Effects of anthropogenic emissions on aerosol formation from isoprene
- and monoterpenes in the southeastern United States, P. Natl. Acad. Sci. USA, 112, 37-42,
- 1069 10.1073/pnas.1417609112, 2015.
- 1070 Yan, C., Nie, W., Äijälä, M., Rissanen, M. P., Canagaratna, M. R., Massoli, P., Junninen, H., Jokinen, T., Sarnela, N., Häme,
- 1071 S. A. K., Schobesberger, S., Canonaco, F., Yao, L., Prévôt, A. S. H., Petäjä, T., Kulmala, M., Sipilä, M., Worsnop,
- D. R., and Ehn, M.: Source characterization of highly oxidized multifunctional compounds in a boreal forest
- environment using positive matrix factorization, Atmos. Chem. Phys., 16, 12715-12731, 10.5194/acp-16-12715-
- 1074 2016, 2016.
- 1075 Zhang, H., Yee, L. D., Lee, B. H., Curtis, M. P., Worton, D. R., Isaacman-VanWertz, G., Offenberg, J. H., Lewandowski,
- 1076 M., Kleindienst, T. E., Beaver, M. R., Holder, A. L., Lonneman, W. A., Docherty, K. S., Jaoui, M., Pye, H. O. T.,
- Hu, W., Day, D. A., Campuzano-Jost, P., Jimenez, J. L., Guo, H., Weber, R. J., de Gouw, J., Koss, A. R., Edgerton,
- E. S., Brune, W., Mohr, C., Lopez-Hilfiker, F. D., Lutz, A., Kreisberg, N. M., Spielman, S. R., Hering, S. V.,
- 1079 Wilson, K. R., Thornton, J. A., and Goldstein, A. H.: Monoterpenes are the largest source of summertime organic
- aerosol in the southeastern United States, P. Natl. Acad. Sci. USA, 115, 2038-2043, 10.1073/pnas.1717513115,
- 1081 2018.
- Zhao, D., Pullinen, I., Fuchs, H., Schrade, S., Wu, R., Acir, I.-H., Tillmann, R., Rohrer, F., Wildt, J., Guo, Y., Kiendler-
- 1083 Scharr, A., Wahner, A., Kang, S., Vereecken, L., and Mentel, T. F.: Highly oxygenated organic molecule (HOM)
- formation in the isoprene oxidation by NO₃ radical, Atmos. Chem. Phys., 21, 9681-9704, 10.5194/acp-21-9681-
- 1085 2021, 2021.
- Zhao, D., Schmitt, S. H., Wang, M., Acir, I.-H., Tillmann, R., Tan, Z., Novelli, A., Fuchs, H., Pullinen, I., Wegener, R.,
- Rohrer, F., Wildt, J., Kiendler-Scharr, A., Wahner, A., and Mentel, T. F.: Effects of NO_x and SO₂ on the secondary
- 1088 organic aerosol formation from photooxidation of α-pinene and limonene, Atmos. Chem. Phys., 18, 1611-1628,
- 1089 10.5194/acp-18-1611-2018, 2018.
- 1090 Zhao, D. F., Buchholz, A., Kortner, B., Schlag, P., Rubach, F., Kiendler-Scharr, A., Tillmann, R., Wahner, A., Flores, J. M.,
- Rudich, Y., Watne, Å. K., Hallquist, M., Wildt, J., and Mentel, T. F.: Size-dependent hygroscopicity parameter (κ)
- and chemical composition of secondary organic cloud condensation nuclei, Geophys. Res. Lett., 42, 10920-10928,
- 1093 10.1002/2015gl066497, 2015a.
- Zhao, D. F., Kaminski, M., Schlag, P., Fuchs, H., Acir, I. H., Bohn, B., Häseler, R., Kiendler-Scharr, A., Rohrer, F., Tillmann,
- R., Wang, M. J., Wegener, R., Wildt, J., Wahner, A., and Mentel, T. F.: Secondary organic aerosol formation from
- hydroxyl radical oxidation and ozonolysis of monoterpenes, Atmos. Chem. Phys., 15, 991-1012, 10.5194/acp-15-
- 1097 991-2015, 2015b.
- 1098 Zhou, L., Gierens, R., Sogachev, A., Mogensen, D., Ortega, J., Smith, J. N., Harley, P. C., Prenni, A. J., Levin, E. J. T.,
- Turnipseed, A., Rusanen, A., Smolander, S., Guenther, A. B., Kulmala, M., Karl, T., and Boy, M.: Contribution

1100	from biogenic organic compounds to particle growth during the 2010 BEACHON-ROCS campaign in a Colorado
1101	temperate needleleaf forest, Atmos. Chem. Phys., 15, 8643-8656, 10.5194/acp-15-8643-2015, 2015.
1102	Ziemann, P. J. and Atkinson, R.: Kinetics, products, and mechanisms of secondary organic aerosol formation, Chem. Soc.
1103	Rev., 41, 6582-6605, 10.1039/c2cs35122f, 2012.
1104	