- 1 Markedly different impacts of primary emissions and secondary
- 2 aerosol formations on aerosol mixing states revealed by
- 3 simultaneous measurements of CCNC, V/HTDMA2 and SP2
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Abstract

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The aerosol mixing state is a crucial physical-chemical physicochemical property that affects their the optical properties and cloud condensation nuclei (CCN) activity. Multiple techniques are commonly employed to determine the aerosol mixing states for various applications, and comparisons between these techniques provide insights of onto the variations in aerosol chemical and physical properties. These techniques include size-resolved CCN activity measurements using a system with a CCN counter (CCNC) coupled with a differential mobility analyzer (DMA), a Humidified/Volatility Tandem differential mobility analyzer Differential Mobility Analyzer (H/V-TDMA) which that measures aerosol hygroscopicity and volatility distributions, and a single-particle soot photometer (SP2) whichthat directly quantifies black carbon (BC) mixing states. This study provides athe first time intercomparisons comparison of aerosol mixing state parameters obtained through simultaneous measurements of ausing DMA-CCNC, a-H/VTDMA, and a-DMA-SP2. The impactimpacts of primary aerosolsaerosol emissions and secondary aerosol formations (SA) formation on the aerosol mixing states and the intercomparison results were analyzed. The results showed that the differences in the mixing—state parameters measured byusing different techniques varied greatlysignificantly under different conditions. The V-TDMA and DMA-SP2 measurements showed that the non-volatile population identified by the V-TDMA was mainly contributed by the BC-containing aerosols. The HTDMA and DMA-SP2 measurements indicated that a substantial proportion of nearly hydrophobic aerosols weredid not contributed originate from BC-containing aerosols, but likely originated from fossil fuel combustion and biomass-burning emissions. Synthesized The synthesized comparison results between the DMA-CCNC, HTDMA, and DMA-SP2 measurements revealed that some of the nearly hydrophobic BC-free particles were CCN-inactive under supersaturated conditions, likely from fossil fuel combustion emissions, while. In contrast, others were CCN-active under supersaturated conditions linked to biomass-burning emissions. Fossil fuel combustion-emitted BC-containing aerosols tendedemitted from fossil fuel combustion tend to be more externally mixed with other aerosol compositions compared to components than those emitted from by biomass-burning activities. These results highlight significant disparities in the mixing states as well as and physiochemical properties between aerosol originated of aerosols from fossil fuel combustion and biomass burning. The formation of secondary nitrate and organic aerosols exerted significant impacts on significantly affects variations in aerosol mixing states, generally enhancing aerosol hygroscopicity and volatility; while reducing differences in mixing state parameters derived from different techniques, resulting in a reduction inreducing aerosol heterogeneity. The variations in Variations in the number of BC-free particle number fractions showed that secondary aerosolsSAs tended to form more quickly on BC-free particles than on BC-containing particles. Further comparison of the mixing state parameters revealed that the two resolved SOAsecondary organic aerosol factors in this study exhibited remarkably different physical properties, indicating that they were likely formed through different pathways. These findings suggest that intercomparisonscomparisons among aerosol mixing states derived from different techniques can provide deeper insightinsights into aerosolthe physical properties of aerosols and how they are impacted affected by secondary aerosolSA formation, aiding the investigation of secondary aerosolSA formation pathways.

1 Introduction

The aerosol mixing state is a crucial physicochemical property of aerosol particles (Riemer et al., 2019), exerting a significant impact on their optical properties and cloud condensation nuclei (CCN) activity, thus affecting their impact on the climate and the environment (Fierce et al., 2017; Riemer et al., 2019; Stevens et al., 2022). For instanceexample, variations in the mixing state of black carbon (BC) particles can significantly alter their absorption and radiative effects (Bond et al., 2013; Lack et al., 2012; Zhao et al., 2019; Moffet et al., 2016; Matsui et al., 2018; Peng et al., 2016). Using simple internally internal mixing state assumptions for aerosol chemical compositions in estimating to estimate CCN number concentrations can lead to substantial overestimations (up to 30%; Deng et al., 2013; Farmer et al., 2015; Ren et al., 2018; Ching et al., 2017, 2019; Tao et al., 2021). The aerosol mixing state varies widely due to complex emissions and atmospheric transformations, leading to significant uncertainties in estimating the effects of aerosolaerosols based on simplified mixing state assumptions (Ervens, 2015; Wang et al., 2022; Fu et al., 2022).

The aerosol mixing state describes the mixture of aerosol chemical components within each single particle and the distribution of these particles in the aerosol population. This property can be directly measured by theusing single—particle chemical composition techniques (Fierce et al., 2017; Riemer et al., 2019), such as the single—particle soot photometer (SP2) that), which measures refractory black carbon (rBC) mass concentrations and the mixing state of rBC with other aerosol components, or single—particle chemical composition measurement techniques (e.g., single-particle aerosol mass spectrometer, SP-AMS) that have been developed overin recent years (Lee et al., 2019; Riemer et al., 2019 and reference therein). Alternatively, the aerosol mixing state can be inferred from indirect measurements of aerosol properties, such as size-resolved aerosol CCN activity (measured by coupling a differential mobility analyzer (DMA) and a CCN counter (CCNC)), size-resolved aerosol hygroscopicity distributions, or volatility distributions (measured by a Humidified/Volatility Tandem differential mobility analyzer (H/V-TDMA)).

However, each technique yields information on aerosol mixing states based on different aerosol micro-physical properties, thus obtainobtaining aerosol mixing states that are different but linked to one another. For instance, while both CCN activity and the hygroscopic growth measurements are associated with aerosol hygroscopicity, an intercomparison between CCNC and

HTDMA measurements promoted investigations oninto aerosol hygroscopicity variations under different saturation conditions (Su et al., 2010; Juranyi et al., 2013; Lance et al., 2013; Kawana et al., 2016; Tao et al., 2020; Jiang et al., 2021). Although beththe SP2 and VTDMA techniques depend on the evaporation of non-refractory compositions, only the rBC remains in the SP2 measurements, whereas the. In contrast, non-refractory composition evaporations dependevaporation depends on thermodenuderthe thermodynamic temperature in the VTDMA measurements. Thus, measurements of an SP2 are highly correlated to those of a VTDMA at high temperatures, (200 °C-300 °C), with their differences reflecting variations in aerosol density, shape, or volatility (Philippin et al., 2004; Wehner et al., 2009; Adachi et al., 2018, 2019; Wang et al., 2022). HTDMA and VTDMA can be applied in combination combined to study the influence of the aerosol mixing state on its hygroscopicity and volatility (Zhang et al., 2016; Cai et al., 2017; Wang et al., 2017). Strong correlations were found between the hydrophobic and non-volatile particles, suggesting that they might be of have similar chemical composition compositions (Zhang et al., 2016). In addition, some studies showedhave shown that, except for BC, low-volatility particles correlated well with CCNinactive particles based on VTDMA and CCNC measurements (Kuwata et al., 2007; Kuwata and Kondo, 2008; Rose et al., 2011; Cheng et al., 2012). Therefore, intercomparisons between mixing state parameters measured by distinct techniques provides not only provide a better characterization of the aerosol mixing state itself, but also and insight into aerosol physiochemical properties. Previous studies have mainly compared two kindstypes of aerosol mixing state measurements, lacking and <u>lacked a</u> comprehensive comparative analysis among SP2, DMA-CCN, and HV-TDMA measurements, hindering the wide applications application of derived aerosol mixing states obtained by individual techniques.

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The mixing state of primary aerosols can vary greatly depending on their type and emission conditions (Cheng et al., 2012; Wang et al., 2017; Wang et al., 2022; Ting et al., 2018; Liu et al., 2021), and can be significantly altered during aging processes or secondary formations (Wehner et al., 2009; Cheng et al., 2012; Wang et al., 2022; Tomlin et al., 2021; Lata et al., 2021). Primary aerosol emissions, such as biomass burning, fossil fuel combustion, and cooking, tend to contribute to weakly hygroscopieweak hygroscopicity (Herich et al., 2008, 2009; Wang et al., 2020; Kim et al., 2020) and low-volatility aerosols (Hong et al., 2017; Saha et al., 2018), while). The formation of

secondary aerosolaerosols (SAs), including the aging of BC-containing aerosols and primary organic aerosols, mainly contributes to aerosols with strong CCN activity (Mei et al., 2013; Ma et al., 2016; Tao et al., 2021) and high hygroscopicity (Chen et al., 2018; Kim et al., 2020; Wang et al., 2020). To enhance our understanding of the mixing state of aerosols from different emission sources and to improve its characterization in models, it it is important to study the impact of specific primary aerosol emissions and secondary aerosol formations SA formation on aerosol mixing states and the influence onof aerosol mixing state parameters derived from different techniques to enhance our understanding of the mixing state of aerosols from different emission sources and improve their characterization in models.

The North China Plain (NCP) is among the most polluted regions in China, with various primary emission sources and strong secondary aerosolSA formations playingthat play critical roles in air pollution (Xu et al., 2011; Tao et al., 2012; Liu et al., 2015). The complex mixing state of aerosols in the NCP contributes to uncertainties in evaluating their climate and environmental effects (Zhuang et al., 2013; Nordmann et al., 2014; Zhang et al., 2016; Tao et al., 2020; Shi et al., 2022), particularly with regards to regarding BC particles (Wu et al., 2017; Liu et al., 2019; Zhao et al., 2019; Wang et al., 2011; Zheng et al., 2019). Meteorological conditions can greatly impact the secondary aerosol affect SA formation in the NCP, which and can be significantly exacerbated during severe pollution events. SA formation under low relative humidity (RH) conditions, mainly through the condensation of gaseous-phase oxidation products, would change to that mainly occurring in the aqueous phase under high RH conditions (Kuang et al., 2020), and consequently). Because SAs formed through different mechanisms, have different chemical compositions and add mass to different aerosol populations, SA formation under different meteorological conditions can affect the aerosol mixing states (Tao et al., 2021). differently (Tao et al., 2021). This study obtained the aerosol mixing state through concurrent measurements of the CCN activity, hygroscopicity, volatility, and BC particle at a regional site in the NCP, using CCNC, HTDMA, VTDMA, and SP2 instruments. This provides a unique opportunity to perform a most comprehensive intercomparison inter-comparison of the aerosol mixing statestates among different techniques for insights to gain insight into the impact of primary aerosol emissions and secondary aerosol formations SA formation on the observed aerosol mixing states.

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2 Materials and methods

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2.1 Campaign information and instruments set-upsetup

From the 16th of October to the 16th of November 2021, aerosol mixing states were continuously and concurrently monitored using different techniques at the Gucheng site in Dingxing county County, Hebei province Province, China, as part of thea campaign of investigating investigate AQueous Secondary aerOsol formationsaerosol formation in Fogsfog and Aerosolsaerosols and their Radiative aerfects in the NCP (AQ-SOFAR). The observation site, located at 39°09'N, 115°44'E, is an Ecological and Agricultural Meteorology Station of the Chinese Academy of Meteorological Sciences, situated between the megacities of Beijing (approximately 100 km away) and Baoding (approximately 40 km away), and surrounded by farmlands and small towns. This site provides a representative view of the background conditions of atmospheric pollution conditions in the NCP (Kuang et al., 2020; Li et al., 2021).

Different measurement techniques were used during the campaign to simultaneously obtain the aerosol mixing state through CCN activity, hygroscopicity, volatility, and BC particle observations (to be discussed in the next section). In addition to aerosol mixing state measurements, the AQ-SOFAR campaign also included includes measurements of aerosol number size distribution, chemical composition, aerosol scattering, and absorption properties. Aerosol number size distributions in the diameter range of 13 nm-4 µm were measured by the scanning mobility particle sizer (13–550 nm) and the aerodynamic aerosol classifier (100 nm-4 µm), and they are merged by assuming an aerosol density of 1.6 g/cm³. The total rBCBC mass concentrations were determined using an aethalometer (Magee, AE33; Drinovec et al., 2015), with); more information abouton the correction of the absorption measurements and mass concentration calculations is available in Luo et al. (2022). All aerosol measurement instruments were housed in a temperature-controlled container set at 24 °C. at 24 °C. The inlet was switched among three impactors: TSP (Total Suspended Particles), PM_{2.5} (Particulate Matter with an aerodynamic diameter of less than 2.5 µm), and PM₁ (Particulate Matter with an aerodynamic diameter of less than 1 µm). Inlet changes among impactors affect dry-state aerosol sampling owing to ambient aerosols are enlarged through aerosol hygroscopic growth or activation. However, the aerosol mixing state and aerosol chemical composition measurements were made on submicron aerosols, and the inlet change almost did not affect those measurements under conditions of RH less than 90%. The sampled aerosol was dried by two parallelly assembled Nafion dryers with a length of 1.2 m. Two Nafion driers was used because of the high RH and sample flow rate (~16 L/min) during the campaign to ensure drying efficiency. In addition, during autumn and winter in the NCP, ambient air temperature (<20 °C and sometimes <0 °C) can be significantly lower than the room temperature (~24 °C). Therefore, this dryer system can maintain the RH of sampled aerosols to below 20%. Meteorological data, such as temperature, pressure, wind speed and, wind direction, relative humidity (RH),and RH were obtained from an automatic weather station operated by the station.

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The chemical composition of the submicron aerosols was analyzed using a High-Resolution Time-of-Flight Aerosol Mass Spectrometer (HR-ToF-AMS). The ionization efficiency (IE) was calibrated using 300 nm diameter pure NH₄NO₃ particles, following the standard protocols outlined in Jayne et al. (2000) in the middle of the campaign, with the relative ionization efficiency (RIE) of ammonium determined to be 5.26. The RIE of sulfate was 1.28 using pure (NH₄)₂SO₄ particles, and the default RIEs of 1.4 for organic aerosols, 1.1 for nitrates, and 1.3 for chlorides were used- as the organic aerosols. The composition-dependent collection efficiency reported by Middlebrook et al. (2012) was used in this study. Elemental ratios were derived using the "Improved-Ambient (I-A)" method as described in Canagaratna et al. (2015), including hydrogen to carbon (H/C), oxygen to carbon (O/C), and organic mass to organic carbon (OM/OC) ratios. Two primary organic aerosol (POA factors) and two SOAoxygenated organic aerosol (OOA) factors were identified by High-Resolution Positive Matrix Factorization (HR-PMF-(; Ulbrich et al., 2009; Paatero and Tapper, 1994). This study used the summation of the two OOA factors to represent secondary organic aerosols (SOA). The mass spectra of the organic aerosol (OA) factors and their correlations with external species are shown in Figs. S1- and S2. The Biomass Burning Organic Aerosol (BBOA) spectrum was characterized by obvious m/z 60 (mainly $C_2H_4O_2^+$) and 73 (mainly $C_3H_5O_2^+$), two indicators of biomass burning (Mohr et al., 2009). In addition, BBOA was correlated well with $C_2H_4O_2^+$ ($R^2=0.91$) and $C_3H_5O_2^+$ ($R^2=0.90$). Consistent with previous studies in Beijing (Xu et al., 2019), the PMF analysis revealed a mixed factor named Fossil Fuel Organic Aerosol (FFOA-that), which comprises traffic emissions and coal combustions, which was combustion and is characterized by a typical hydrocarbon ion series. It was observed that FFOA had a relatively high f_{44} (0.083) value, which iswas likely due to aging during regional transportation, similar to the results observed in the winter of 2016 in Beijing (Xu et al., 2019)

and CCOAcoal combustion organic aerosols in Gucheng (Chen et al., 2022). Two SOASecondary organic aerosol formation from volatile organic compound precursors could occur in different formation pathways, such as aqueous-phase, heterogeneous, or gas-phase reactions. It might also be oxidized under different conditions, such as oxidation under different nitrogen oxide conditions with different oxidation capacities and oxidants. The two resolved OOA factors were found to displaydisplayed different spectral patterns, correlations with tracers, and diurnal variations, suggesting that they resulted from different chemical processing. For example, processes. However, their formation mechanisms remain to be explored in future studies. In general, the OOA factor 1 (OOA1 had) has higher CO2+/C2H3O+ (3.9) and O/C (0.91) ratios than OOA factor 2 (OOA2) with 2.1 and 0.78, respectively.

This study did not consider losses in the inlet line and sampling systems for the following reasons: (1) investigated mixing state parameters are represented by number fractions (NFs) of different diameters, which are much less affected by losses in sampling systems compared to OOA2 (2.1, 0.78with absolute number concentrations; and (2) good consistency was achieved between measurements of particle number size distributions (PNSD) and mass concentrations measured by AMS. The average ratio between volume concentration derived from AMS and rBC measurements (densities of compounds are the same as Kuang et al., 2021) and the volume concentration derived from PNSD measurements was 0.79 (R=0.97, as shown in Fig. S3), consistent with previous reports as AMS cannot detect aerosol components, such as dust (Kuang et al., 2021).

2.2 Aerosol mixing states measurement techniques

2.2.1 DMA-CCNC measurements

The CCN activity of the particles under supersaturated conditions was measured using a DMA-CCNC system, which consisted of a differential mobility analyzer (DMA₅; model 3081-by, TSI, Inc., MN, USA), a-condensation particle counter (CPC₅; model 3756-by, TSI, Inc., MN, USA), and a continuous-flow CCNC (model CCN100-by, Droplet Measurement Technologies, USA). The system was operated in a-size-scanning mode and provided the Size-resolved Particle Activation Ratio (SPAR) by combining the CPC and CCNC measurements at different particle sizes. ThreeTo compare the instruments, three supersaturations (SSs) of 0.08%, 0.14%%, and 0.22% were applied in a single cycle of aboutapproximately 15 minutes.min. CCN measurements under these three SSs revealed that the

CCN activity of aerosols resides in the accumulation mode with an aerosol diameter range of approximately 100–200 nm, which is close to the diameters of the HV-TDMA measurements. Higher SSs would reveal CCN activities of smaller aerosol particles (<100 nm), where the DMA-SP2 measurement is unavailable. The sample and sheath flow rates of the DMA were set at 1 lpm and 5 lpm, respectively, resulting in a measured particle diameter range of 9 nm to 500 nm, with a running time of 5 minutesmin per cycle. Supersaturations Supersaturation in the CCNC werewas calibrated with monodisperse ammonium sulfate particles (Rose et al., 2008) both before and after the campaign. The flow rates were also calibrated before and after the campaign and were checked daily to minimize uncertainties in the droplet counting and supersaturation formed in the column (Roberts and Nenes, 2005; Lance et al., 2006). The SPAR deviations due to multiple-charge particles were corrected using a modified algorithm based on Hagen and Alofs (1983) and Deng et al. (2011). MoreFurther details about the regarding this system can be found in Ma et al. (2016) and Tao et al. (2021).

2.2.2 H/V-TDMA measurements

The mixing state of the aerosols in terms of hygroscopicity and volatility was measured using a Hygroscopicity/Volatility Tandem Differential Mobility Analyzer (H/V-TDMA; Tan et al., 2013). The H/V-TDMA consistsconsisted of two DMA (Model 3081L3081 L, TSI Inc.), with the first DMA (DMA1) selecting dried particles without conditioning (RH ~15%) and the second DMA (DMA2) selecting conditioned particles. The H/V-TDMA can operate in either H-mode or V-mode, which is controlled by a three-way solenoid valve. In H-mode, aA Nafion humidifier iswas used in the H-mode to condition the selected dry particles to 90% relative humidity (RH)RH equilibrium. The number—size distribution of the humidified particles (Dp) iswas measured byusing DMA2 and a CPC (Model 3772, TSI Inc.). The RH-dependent hygroscopic growth factor (GF) at a certain dryspecific diameter is(Dd) was calculated as follows:

$$GF = \frac{D_{\overline{p}}}{D_{\overline{d}}} \frac{D_{p}(RH)}{D_{d}} \tag{1}$$

In this mode, fourwhere D_p(RH) is the size of particles undergoing humidification. Four dry electrical mobility diameters (50, 100, 150, and 200 nm) were measured, in this mode. The instrument was regularly calibrated using Standard Polystyrene Latex Spheres standard polystyrene latex spheres (PSL) and Ammonium Sulfate particles.

In the V-mode, a heated tube was used to evaporate evaporated the volatile coatings from the previously selected dry particles. Six temperature settings were used for the heated tube, ranging from 25-to_200°C. The number_size distributions of the heated particles were measured using DMA2 and CPC. Besides In addition to the four particle sizes measured in the H-mode, three additional particle sizes (250, 300, and 350 nm) were measured in the V-mode (residence time inside the heated tube to be about 1.6 s; Hong et al., 2017). The temperature dependent shrinkage factor (SF), which is the ratio of the heated particle size to the dry particle size, without heating (D_d), is defined as:

$$SF = \frac{\frac{D_p}{D_d} \frac{D_p(T)}{D_d}}{D_d}$$
 (2)

where D_p(T) denotes the particle diameter during heating. A complete cycle of both-H-mode measurements at under one relative humidity (RH) condition and V-mode measurements at six temperature conditions temperatures took approximately 3 hours. Ah. The Probability Density Function (PDF) of the GF (or SF) can be was calculated from the measured density function using the inversion algorithm described by Stolzenburg and McMurry (2008).

2.2.3 DMA-SP2 measurements

The size-resolved BC mixing states were measured by theusing an SP2 (Droplet Measurement Technology, Inc., USA) after a-DMA (Model 3081, TSI, USA). The DMA selected aerosols atoff various dry particle sizes, which were then introduced to the into SP2. The DMA-SP2 setup was able to measure the mixing states of aerosols atwith diameters (detection limit of approximately 80 nm based on the calibration) of 100 nm, 120 nm, 160 nm, 200 nm, 235 nm, 270 nm, 300 nm, 335 nm, 370 nm, 400 nm, 435 nm, 470 nm, 500 nm, 535 nm, 570 nm, 600 nm, 635 nm, 670 nm, and 700 nm within 20 minutes; min when it wasn'twas not placed after a thermodenuderan enuder-bypass switch system-(the 13th to the 24th of October, 09:00 am of the 5th of November to 09:00 am of the 8th of November). However, it only measured mixing states at diameters of 120 nm, 160 nm, 200 nm, 250 nm, 300 nm, 400 nm, and 500 nm when it was placed after a thermodenuder-bypass switch system-(11:00 am of the 24th of October to 08:00 am of the 5th of November, and 09:00 am of the 8th of November to 06:00 pm of the 17th of November). Because the HTDMA and VTDMA measurements were conducted solely by a single H/VTDMA system operating in different modes, the time needed for a single particle size measurement of HTDMA and VTDMA was much longer than that of the DMA-

SP2 system. Thus, more particle sizes were selected in the DMA-SP2 system for acquiring the BC mass concentration and mixing state at larger diameters than HTDMA and VTDMA.

The SP2 chamber hashad a continuous Nd:-YAG laser beam atwith a wavelength of 1064 nm. The BC-containing particles passing through the laser beam will bebecame incandescent throughby absorbing the radiation. By The mass concentration of the BC was calculated by measuring the intensity of the emitted incandescent light, the mass concentration of BC can be calculated. The sheath flow/sample flow ratio was maintained at 10 for the DMA to reduce the width of the diameter distribution of the selected monodisperse aerosols. Additionally, the flow rate of the SP2 was changed from 0.1 to 0.12 L/min starting on Octoberthe 22nd. The of October (allowed flow rate range of SP2: 0.03–0.18 L/min from the specification). SP2 was calibrated using Aquadagquadag soot particles, as reported by Gysel et al. (2011). Further details about regarding the calibrations are introduced provided in Section 1 of the supplementSupplementary Information.

2.3 Derivations of mixing state parameters

2.3.1 Fitting SPAR curves measured by the DMA-CCNC system

The measured-SPAR curves ean bewere parameterized withusing a sigmoidal function with three parameters. As shown in Fig. S4, a sigmoidal curve generally characterized the measured SPAR. This parameterization assumes that the aerosol is an external mixture of CCN-active hydrophilic particles that are CCN-active and CCN-inactive hydrophobic particles that are CCN-inactive (Rose et al., 2010). The formula used to parameterize the SPAR ($R_a(D_d)$) for a specific SS is as follows (Rose et al., 2008):

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$$R_a(D_d) = \frac{MAF}{2} \left(1 + erf\left(\frac{D_d - D_a}{\sqrt{2\pi}\sigma}\right) \right)$$
 (7)

where erf isdenotes the error function. The Maximum Activation Fraction (MAF) is an asymptote of the measured SPAR curve atfor large particle sizes, and it represents the number particles, as shown in Fig. S4, representing the fraction of CCNs relative to the total number of particles. D_a is the midpoint activation diameter and, is linked to the hygroscopicity of the CCNs, and indicates the diameter where the SPAR equals half of the MAF value. The σ is the standard deviation of the cumulative Gaussian distribution function and characterizes the heterogeneity of CCN hygroscopicity. In Fig. S4, the σ indicates the slope of the steep increase in the SPAR curves when the diameter is close to Da. Generally,

hydrophilic particles larger than D_a can become CCN, so. Therefore, these three parameters can be used to characterize the hygroscopicity of those these hydrophilic particles. This study did not consider the impact of nearly hydrophobic particles on SPAR, as deviations from this parameterization scheme due to this impact arewere negligible at low SSs, as stated in Tao et al. (2020).

2.3.2 Classification of particle type based on hygroscopicity or volatility

In this study, ambient aerosol particles were classified into two groups based on their hygroscopicity (hydrophobic and hydrophilic) and two groups based on their volatility (non-volatile and volatile) based on the measurements from H/V-TDMA (Wehner et al., 2009; Liu et al., 2011; Zhang et al., 2016). Each group can be defined using athe critical value of GF or SF as the following: follows: hydrophobic population: GF<GFc; hydrophilic population: GF≥GFc; non-volatile population: SF≥SFc; and volatile population: SF<SFc.

Hydrophobic population: GF < GF_C;

Hydrophilic population: $GF \ge GF_G$;

Non-volatile population: $SF \ge SF_C$;

Volatile population: SF < SF_C.

The critical values of GF (GF_C) and SF (SF_C) in H/V-TDMA depend on <u>the</u> particle size and working conditions₂ such as relative humidity and heating temperature.

During this campaign, the SF_C was set atto 0.85 for all seven measured particle sizes at a temperature of 200 °C. The GF_C for the four measured particle sizes of 50, 100, 150, and 200 nm were 1.1, 1.15, 1.175, and 1.2, respectively, and the corresponding hygroscopicity parameter, κ , was approximately 0.07. These values of GF_C and SF_C divide the probability density functions (PDFs) of SF and GF into two modes, consistent with prior NCP studies in the NCP (Liu et al., 2011; Zhang et al., 2016). The number fraction (NF) for the hydrophilic group (NF_H) and volatile group (NF_V) can be calculated as: follows:

$$NF_{H} = \int_{GF_{c}}^{\infty} GFPDF(GF)dGF \tag{7}$$

NF_V=
$$\int_0^{SF_C} SFPDF(SF) dSF$$
 (8)

where GFPDF and SFPDF are the PDFs of GF and SF, respectively, derived from H/V-TDMA measurements.

2.3.3 Classification of particle type based on DMA-SP2 measurements

The BC-containing aerosols can be categorized into two groups based on the coating thickness: bare BC/thinly coated BC particles and thickly coated BC particles. For the measurement of coated BC particles at SP2, the incandescence signal is generally detected later than the scattering signals and the time difference between the occurrence of the peaks of the incandescence and scattering signals is defined as the lag time (Moteki & Kondo, 2007; Sedlacek et al., 2012; Subramanian et al., 2010). The coating thickness of the BC-containing aerosols in the SP2 measurement can be determined indicated by the lag time between the peak of the scattering signal and the incandescence signal (Moteki and Kondo, 2007; Schwarz et al., 2006; Sedlacek et al., 2012; Subramanian et al., 2010; Metcalf et al., 2012;), which has exhibited a clear two-mode distribution in previous studies (Zhang et al., 2018; Zhao et al., 2021). A critical lag-time threshold can be used to differentiate between the different types of BC-containing aerosols and calculate the number fractionNF of bare BC particles and coated BC particles in the total identified aerosols. In this study, a two-mode distribution of the lag-time (Δt) was observed, and a critical value of 0.8 μs was used to classify the BC-containing particles into thinly coated (or bare) BC (Δt < 0.8 μs) and thickly coated BC (Δt < 0.8 μs), respectively.). The definitions of all abbreviations are listed in Table 1.

3 Results and discussions

3.1 Campaign Overview overview

The time series of the meteorological parameters, aerosol mixing state measurements using different techniques, and mass concentrations of the aerosol chemical compositions are shown in Fig. 1. In detail, the measurements of aerosol mixing states include SPAR at an SS of 0.08%,% by DMA-CCNC, GF-PDF (PDF of GF) at 200 nm by HTDMA, SF-PDF (PDF of SF) at 200 nm and 200 °C, by VTDMA, and lag-time PDF of 200 nm BC-containing particles and mass concentrations of by

DMA-SP2. The SIA (secondary inorganic aerosol), SOA (OOA1 and OOA2), PA (BBOA and FFOA), SOA, POA, and BC mass concentrations are shown in Fig. 1. During the campaign, three (b). Three periods with significantly different aerosol pollution conditions were identified, during the campaign. As shown in Fig. 1(b), before the 23rd of October 23rd (moderately polluted period), the accumulation of aerosols led to SIA mass concentrations greater than < 20 µg/m³, while. In contrast, the highest mass concentrations of SOA, POA, and BC mass all reached 10 µg/m³. The mass concentrations of different chemical compositions aerosol components increased significantly from the 23rd of October 23rd to the 6th of November 6th (heavily polluted period with an average non-refractory PM₁ mass concentration of 49.5±22.5 µg/m³) and decreased totoo much lower levels after the 6th of November 6th (clean period). with a non-refractory PM₁ mass concentration of 5.1±3.3 μg/m³). Two particle groups were identified with regard to concerning the CCN activity, hygroscopicity, volatility, and coating thickness, as demonstrated by the SPAR, GF-PDF, SF-PDF, and the lag-time PDF of BC-containing particles. Significant variations in the aerosol mixing states were also observed during the three periods of with different pollution conditions, as demonstrated by the variations in SF-PDF measured by the VTDMA. For example, the SF of the non-volatile particle group decreased during the heavily polluted period. Aerosol mixing states may have changed due to because of various transformations of existing aerosol particles and distinct secondary formation processes under different pollution conditions (Kuang et al., 2020; Tao et al., 2021; Shi et al., 2022; Yang et al., 2022). This will be analyzed in detail later in the discussion. The diurnal Diurnal variations in the mass concentrations of different aerosol chemical compositions components and mixing states can be seen observed in the variations of in the SPAR measurements, as previously observed in this region (Liu et al., 2011; Ma et al., 2012; Kuang et al., 2015; Tao et al., 2020).

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In-Fig. 2, shows the campaign—averaged SPAR at the three SSs, PDF of the lag-time of BC—containing particles, GFPDF, and SFPDF at 200 °C atfor different particle sizes—are presented. The sigmoidal SPAR curves were characterized by a rapid increase, followed by a gradual increase to unit 1, similar to the measured SPAR curves previously observed in this region (Deng et al., 2011; Zhang et al., 2014; Ma et al., 2016; Tao et al., 2018). At lower SSs, the particle size required for CCN activation was larger; thus, rapid increases in the SPAR curves occurred at larger particle sizes and. In addition, the maximum AR of the SPAR curves becomes smaller decreases as fewer particles

are CCN-active under low SSs. For the three measured SSs, the corresponding fitting parameters, Da, which indicate the center of the particle size range with rapid increases inparticle sizes where SPAR curves, equals approximately 0.5 are approximately 90 nm, 120 nm, and 180 nm, nm for the three SSs of 0.08%, 0.14%, and 0.22%, respectively, consistent with the average Da (see Eq. 7) values of the campaign. The number fractionNF of CCN-active particles in large—diameter ranges (which varies with SS and, for example, is greater than 200 nm for 0.08%) can be indicated by the gradual increase of in the SPAR curves and quantified by the fitting parameter, MAF. (see Eq. 7). The PDFs of the lagtime, GF, and SF arewere all characterized by a bimodal distribution, which indicates two particle groups of BC-containing aerosols with different coating thicknesses, hygroscopicity, and volatility. The variations in the aerosol mixing state will bestates were further analyzed based on the basis of MAF and the number fractions (NF) of hydrophilic particles (NFH), volatile particles (NFV), BC-free particles (NFneBC) and number fractions of thinly coated BC in BC containing aerosols (Respec)-measured mixing state parameters.

3.2 Intercomparisons among aerosol mixing state parameters derived using four techniques

The size-dependence of MAF, NF_H, NF_V (200 °C), NF_{henc}-dependent characteristics of the aerosol mixing state parameters derived from the measurements of the four techniques and R_{expecthe} MFs of different aerosol chemical compositions during the three pollution periods are shown in Fig. 3. In general, the size-dependence-dependent characteristics of MAF, NF_H, NF_{V2} and NF_{noBC} were similar-to-one another, suggesting that they were likely dominated by the same particle group, namely BC-free particles. This particle group had the lowesthighest fraction (>0.7) during the heavily polluted period and the highestlowest fraction (down to 0.5) during the clean period, with the fraction decreasing with increasing particle size. This suggests that primary emissions tend to have higher fractions of BC-containing particles in larger diameter ranges. Since; for example, the fraction of BC-containing particles increases from ~0.1 to ~0.4 as the particle size increases from 200 to 500 nm during the cleaning period. Because the bulk aerosol mass fraction (MF) is mostly contributed by particles larger than >300 nm, there mightmay have been more hydrophilic, volatile, CCN-active, and BC-free particles inwith larger sizes (>300 nm) during the heavily polluted period, due owing to strong secondary aerosolSA formation in larger diameter ranges (Kuang et al., 2020), resulting in a higher

number fraction NF of these particles compared to the clean period. As for R_{exBC} , the small size dependence of R_{exBC} during the moderately polluted period might have been associated with stronger primary emissions, while the decrease of \underline{n} R_{exBC} with increasing particle diameter size in the polluted period confirms the confirmed that SA formation is more efficient secondary aerosol formation infor particles with larger diameter ranges diameters.

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As for the difference among the aerosol mixing state parameters, NF_V and NF_{noBC} agreed with each other with a ≤ 0.1 difference smaller than 0.1, and both were higher than NF_H by at least 0.1 than NF_H in the moderately polluted period. In detail, compared to Compared with NF_{noBC}, NF_V was higher during the heavily polluted period, when the nitrate fraction was the highest and the (~30%). The SOA fraction was the lowest (\sim 7%) among all three periods, suggesting that some BC-containing aerosols in this period were also identified as volatile, which is consistent with the fact that the formation of semi-volatile nitrate in the BC-containing particles increases their volatility. However, during the eleancleaning period, NFv was even lower than NF_{noBC}, suggesting that some BC-free aerosols were characterized as low volatile and non-negligible fractions of BC-free aerosols dominated within these less volatile aerosol components, which were likely less volatile organic aerosols. (not likely contributed by BC-containing particles with a BC smaller than the SP2 detection limit, because the SF of this type of volatile BC-containing aerosols has an SF lower than 80/200, which is substantially lower than the threshold SF of 0.85 for NF_V calculation). In addition, the MAF values generally agreed with the NF_H during the clean period, but. However, they were larger than the NF_H during the moderately and heavily polluted periods (by ~ 0.2) when the POA/SOA fractions were higher. ($\sim 40\%$ vs. ~35%). POA generally hadhas a lower hygroscopicity than SOA. The critical κ of hydrophilic mode aerosols was 0.07, suggesting that a higher fraction of aerosols had κ below 0.07 (i.e., hydrophobic mode aerosols in this study) during the moderately polluted period. However, under supersaturated conditions, they demonstrated enhanced hygroscopicity by becoming CCN-active. The NF_H was consistently lower than NF_V and NF_{noBC} (the average difference between NF_H and NF_{noBC} was about approximately 0.2), especially). As mentioned above, NF_H was also lower than MAF during the moderately polluted period, suggesting periods, and there may be a significant fraction of volatile BC-free aerosols hadwith hygroscopicity lower than the critical κ value of 0.07-but; however, they were still CCN-active and therefore not fully hydrophobic.

The diurnal variations of MAF, NF_H, NF_V and NF_{noBC} along with the mass fractions MFs of the aerosol chemical compositions components during the three periods, are shown in Fig. 4. With the exception of Except for a particle size of 50 nm, the diurnal variations of these four mixing state parameters were generally similar for all measured sizes. The different diurnal variations at a particle size of 50 nm may be due to the different effects of emissions and aging processes on the different aerosol modes, as particles smaller than 100nm were 100 nm mainly in a different aerosol mode (reside in the Aitken mode) to, which is where particles larger than 100nm>100 nm mainly reside in the accumulation mode (Wang et al., 2022). For particle sizes larger than particles > 100 nm, (Fig. 4 and S5), there were peaks was a maximum in the afternoon for MAF, NF_H, NF_V, and NF_{noBC}, indicating a peak during this time due to thean increase in secondary aerosolSA compositions like, such as nitrate and SOA, and thea decrease of potential POA and BC. The diurnal Diurnal variations of the aerosol mixing state parameters and aerosol chemical compositions were more pronounced induring the moderately polluted period. During the heavily polluted periodperiods, the diurnal variation was least pronounced for NF_V and most pronounced for NF_H. In the clean-air period, there was another peakmaximum at midnight for MAF and NF_{noBC}, which may be attributed to the diurnal variations of secondary aerosolin <u>SA</u> compositions <u>like</u>, <u>such as</u> sulfate and SOA, and the decrease <u>ofin</u> BC and FFOA. The average_ size-dependence of the aerosol mixing state parameters inover different time ranges during thea heavily polluted period is shown in Fig. \$3\$6. It can be seen that the differences among the four parameters were the least from 12:00 to 18:00, with the most SOA and the least POA. This is consistent with the results shown in Fig. 3, where the difference between the MAF and NF_H becomes smaller <u>decreased</u> when the POA fractions are were the smallest. RexBC tended to be lower during the daytime, and its diurnal variation was more significant for larger particle sizes. In general, the diurnal variations for R_{exBC} are were opposite to those of NF_{noBC} and agree agreed better with those of the primary aerosol mass fractions.

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The intercomparisons among MAF, NF_H, NF_V and NF_{noBC} were conducted based on their correlations at differentMFs. This is because BC particles originate from primary emissions and are mainly mixed externally. After aging in the atmosphere, BC particles can be coated by SAs, resulting in more coated BC particles and fewer externally mixed BC particles. As SAs tend to form on larger particles, the diurnal variations in SA formation may significantly affect the RexBC of larger particle

sizes as.

As summarized in Table. S1. It's worth noting that, the comparison among MAF, NFH, NFV, and NFnoBC was conducted based on their correlations with different particle sizes. Note that the MAF at SSs of 0.08%, 0.14%, and 0.22% waswere used for comparison at particle sizes of 200 nm, 150-nm, and 100 nm particle sizes. This is because the diameter range of rapid increases in the SPAR curves is determined by aerosol hygroscopicity in this particle size range. The midpoints of the rapidly increasing diameter ranges of the SPAR curves at SSs of 0.08%, 0.14%, and 0.22% were approximately 180 nm, 120 nm, and 90 nm, respectively- (as shown in Fig. 2). In general, there were moderate correlations among (r=~0.5) between MAF, NF_H, and NF_V, suggesting that a similar particle group contributed to the dominance of CCN-active, hygroscopic, and volatile aerosols are contributed by a similar particle group (Zhang et al., 2016). The consistency of agreement between MAF- and NFv was slightly higher than that of between MAF-and NFH or between NFH-and NFV with similar correlation coefficients (~0.65) but). However, smaller systematic differences (slope and intercept) were much closer to 1 and 0, respectively), which. This is consistent with the previous finding that a substantial fraction of volatile but less hygroscopic aerosols isare CCN-active. AtFor smaller particle sizesizes, the correlation became weaker, while $(r=\sim0.4)$, whereas the degree of reduction was the least lowest for the correlation between MAF and NF_V.

3.3 Impacts of primary aerosol emissions on aerosol mixing states and parameter intercomparisons

In Fig. 5, presents the correlation between each aerosol mixing state parameter at 200 nm and the mass fractionMF of each primary organic aerosol composition during the three periods is presented. In general, these. The four mixing state parameters (MAF, NF_H, NF_V, and NF_{noBC}) were negatively correlated with MF_{FFOA} and MF_{BBOA}. However, the correlationanticorrelation with MF_{FFOA} (-0.45~-0.74) was much weaker compared tostronger than MF_{BBOA}, (-0.10~-0.45). Biomass—burning emissions and fossil fuel emissions are the two major sources of BC in the NCP (Yang et al., 2022), and NF_{noBC} was negatively correlated with MF_{FFOA} (r=-0.49) and weakly correlated (r=-0.18) with MF_{BBOA}, suggesting that fossil fuel emissionemissions were the more likely the dominant source of BC during

this field campaign. The negative correlation between MAF and MF_{FFOA} was even weaker than that of NF_{noBC} with MF_{FFOA} (-0.62 vs. -0.49). Especially In particular, at the same MF_{FFOA}, the MAF was lower than NF_{noBC}, demonstrating that some BC-free particles were CCN-inactive, and were likely mainly composed of organic aerosols from fossil fuel combustion emissions. The negative correlation between NF_V and MF_{FFOA} was slightly weaker than that between NF_{noBC} and MF_{FFOA} (-0.56 vs_. -0.49). At the same MF_{FFOA}, NF_{noBC} was close to NF_V, and considering that BC-containing aerosols were dominated by thinly coated BC during most times (as shown in of the time (Fig. 5), this demonstrates that the non-volatile population identified by the V-TDMA was mainly contributed by BC-containing aerosols. NF_H had the lowest negative correlation with MF_{FFOA} (r=-0.74), demonstrating the significant contributions from fossil fuel emissions to nearly hydrophobic aerosol populations. At the same MF_{FFOA}, the NF_H was obviously lower than NF_{noBC} (NF_H and NF_{noBC} were larger and smaller than 0.7 when MF_{FFOA} was larger than 0.1, respectively), demonstrating that a substantial portion of nearly hydrophobic aerosols was not contributed by BC-containing aerosols, but likely by FFOA or BBOA (BC-containing aerosols of 200 nm with BC core smaller than 80 nm which is smaller than the detection limit of SP2 likely to be quite aged in the air, thus not possible to be nearly hydrophobic), but likely by FFOA- or BBOA-dominant aerosols (NF_H also had a negative correlation with MF_{BBOA}). However, the markedly different correlations between MAF withand MF_{FFOA} (r=-0.62) and between MAF withand MF_{BBOA} (r=-0.2) imply that those-nearly hydrophobic but CCN-active aerosols were likely contributed by biomass-burning emissions. The correlations between the ratio of thinly coated BC in the total BC-containing particles (R_{exBC}) and the mass fractions MFs of BBOA and FFOA are shown in Fig. 6, and weak. Weak correlations (r<0.3) between R_{exBC} withand MF_{BBOA} and MF_{FFOA} arewere observed. However, R_{exBC} tended to increase with MF_{FFOA}, suggesting that BC-containing aerosols emitted from fossil fuel combustion tended to be more externally mixed with other aerosol components than those emitted from biomass burning activities. These results demonstrate remarkably different mixing states as well as and the physical and chemical properties of fossil fuel combustion aerosols and biomass-burning aerosols.

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The impact of primary emissions on the differences among the four aerosol mixing state parameters is at a particle size of 200 nm was analyzed and is shown in Fig. 7. The difference between NF_{noBC} and NF_H and NF_V (NF_{noBC} - NF_H both had strong positive correlations) was significantly

positively correlated with MF_{FFOA} and MF_{BBOA} (r>0.5), suggesting that a substantial proportion of POA resided in BC-free aerosols and was volatile, but contributed substantially to nearly hydrophobic aerosols—; as did the differences between NF_V and NF_H (NF_V-NF_H). The mass fractions MFs of BBOA and FFOA were poorly linked correlated with the difference differences between the MAF-and NF_V or (MAF-NF_V), MAF and NF_{noBC} or NF_V(MAF-NF_{noBC}), and NF_V and NF_{noBC} (NF_V-NF_{noBC}) (Fig. S5S7). The difference between MAF-NF_H had a positive correlation was positively correlated with MF_{BBOA}, further suggesting that BBOA contributed to nearly hydrophobic aerosols under subsaturated conditions—; however, their hygroscopicity was enhanced, and they became CCN-active atunder supersaturated conditions. The correlations between the mixing-state parameters and primary aerosol composition during the campaign and different pollution periods are summarized in Fig. S7.

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3.4 Impacts of secondary acrosol formations SA formation on acrosol mixing states and parameter intercomparisons

The correlations between eachthe acrosol mixing state parameter at 200 nm and the mass fraction MF of each secondary aerosol-SA component are presented in Fig. 8 for three periods. and the entire campaign areis presented. The analysis is conducted at only 200 nm, where all four aerosol mixing state parameters were measured to compare the four aerosol mixing state parameters and their relationships with aerosol chemical compositions simultaneously. Generally, MAF, NF_H, NF_{V₂} and NF_{noBC} had aexhibited strong positive correlationcorrelations with ammonium. MF_{NH4} (r>0.5). This is likely due to the fact that because ammonium was mainly formed through the neutralization of neutralizing sulfuric and nitric acidacids with ammonia, so; therefore, variations in ammonium better represent overall secondary inorganic aerosol formation. As shown in Fig. 3, the secondary inorganic aerosol components dominated over SA (aboutapproximately 50% vs-about. approximately 70%), indicating that secondary aerosol formations were SA formation was primarily of secondary composed inorganic aerosol formation, which explains the weaker correlations with SOA seen(r=~0.3), as shown in Fig. 8.

During the clean-air period, when the mass fractionMFs of SOA and sulfate were both above 15%, all four parameters had a strong positive correlation with MF_{SO4} and MF_{SOA} (r>0.5), suggesting that when a clean background airmass with higher fractions of sulfate and SOA prevailed, the

local primary emissions that contributed substantially to BC-containing and less hygroscopic POA aerosols became less significant. The positive correlations between the MAF and secondary aerosols components have been extensively discussed by Tao et al. (2021), who found that secondary aerosol formations enhances the hygroscopicity of nearly hydrophobic aerosols, thereby increasing CCN activity. This also explains the highest correlations of strong correlation between the NF_H or MAF withand ammonium formation. The strong positive correlations between NF_V and secondary aerosol formationsSA formation (r=~0.6) are consistent with the fact that nitrate dominates secondary aerosol formationsSA formation during this campaign and nitrate is semi-volatile. For the first time, the strong positive correlations between NF_{noBC} and secondary aerosol formationsSA formation were revealed. This is because observed (r=0.6). NF_{noBC} depends primarily depends on the relative variations of variation between BC-containing and BC-free aerosols. The increase in NF_{noBC} at 200 nm as a function of secondary aerosol mass fractionthe SA MF suggests that secondary aerosol formationsSAs migrated to a higher fraction of BC-free aerosols smaller than 200 nm to particle size of 200 nm, highlighting that secondary aerosolsSAs tended to form more quickly on BC-free aerosols than on BC-containing aerosols.

The effects of secondary aerosolSA formation on the differences between the four aerosol mixing state parameters were studied and are illustrated in Fig. 9. Differences between The two OOA factors (OOA1 and OOA2) were formed through different chemical pathways. The difference between NF_{noBC}-and NF_H and NF_V(NF_{noBC}-NF_H-both) showed a strong negative correlation with MF_{NH4} and MF_{NO3}-(mainly -0.6), as did the differences between NF_V and NF_H (NF_V-NF_H). As previously noted-, NF_H typically hadhas smaller values than NF_V and NF_{noBC}. Thus, a negative correlation withbetween the mass fractionMFs of ammonium and nitrate indicates indicated that the formation of secondary nitrate results resulted in a smaller difference between these mixing state parameters. TheAn increase in the fraction of ammonium nitrate, a pure—scattering semi-volatile compound with strong hygroscopicity, can render the aerosol population more dominated by particles with strong volatility and hygroscopicity. ThisAs the secondary inorganic aerosol components increase, this can result in a smaller difference between NFnoBC, NFH, and NFV-as the secondary inorganic aerosol components increase.

Furthermore, the difference between NF_V-and NF_H showed a positive correlation with MF_{OOA2} and a negative correlation with MF_{OOA1}, indicating different volatility and hygroscopicity of the two SOA factors. The variations in the differencedifferences between NF_V and NF_H with concerning the mass fraction MF of OOA1 and OOA2 are shown in Fig. 9(e) and (f)-), respectively. As previously noted, NF_V was generally higher than NF_H, and the difference between the two decreases decreased with an increase inincreasing MF_{OOA1}, which werewas generally smaller than 0.3. This suggests that the formation of OOA1 enhances the hygroscopicity of volatile particles, which aligns with the highest oxidation state of OOA1 (higher O/C but lower H/C compared to OOA2), which had) and has a significant and overall positive impact on aerosol hygroscopicity (Cerully et al., 2015; Thalman et al., 2017; Zhang et al., 2023). A positive correlation is seen was observed between NF_V and MF_{OOA2} (r=~0.25), whereas).

In contrast, the correlation between NF_H and MF_{OOA2} iswas weak (R iswas close to 0), implying that OOA2 might be semi-volatile but only weakly hygroscopic, which could contribute to NF_V being higher than NF_H as OOA1 increases. The difference between NF_{noBC}-and NF_V is(NF_{noBC}-NF_V) was negatively correlated with MF_{NO3}, which is consistent with the semi-volatile nature of nitrate. The negative correlation between NF_{noBC}-NF_V and MF_{OOA2} indicates that the difference is smaller when there is more OOA2, implying that OOA2 are also a semi-volatile compounds compound and were likely formed mainly on BC-free particles. The correlations between the difference differences between NF_V-MAF and NF_{noBC}-MAF and the mass fraction MF of each secondary aerosol SA composition are muchwere very weak. The impacts of secondary aerosol SA formation on BC mixing states are depicted shown in Fig. S4S8. In general, the number fraction NF of thinly coated BC has a negative correlation with SIA and a weak association with SOA, suggesting that SIA formation mainly enhances the thickness of BC coating the Coating. The correlations between the mixing state parameters and SA composition during the campaign and different pollution periods are summarized in Fig. S9.

In addition to the changes in the mass fractions MFs of secondary aerosol SA compositions, the accumulation of SA pollution due to secondary aerosols may also provide insightinsights into the impact of secondary aerosol formations SA formation on aerosol mixing states. As shown in Fig. 10 (a), during the heavily polluted periods, there were two distinct pollution accumulation processes from Octthe 23rd-to Octthe 27th of October and from Oct 28rd-the 28th to Octthe 31st of October,

respectively. During the pollution accumulation process, the mass concentration of secondary aerosolsSAs increased by approximately three times-fold, indicating the rapid formation of secondary compositions and causing a significant rise in PMincrease in non-refractory PM₁ (NR-PM₁) mass concentration. Fig. 10-(b) and (c) illustrates illustrate that this increase in secondary aerosols led to a significant enhancement of SAs significantly enhanced aerosol mixing state parameters, including MAF, NF_V, NF_{H₂} and NF_{noBC}, which reseincreased from about approximately 0.5 to about 0.8 with evident diurnal variations. This highlights the impact of secondary aerosol formations on SA formation on the aerosol mixing states and the importance of studying the pollution accumulation processes of secondary aerosols. SAs. The enhancements of the different aerosol mixing state parameters during the pollution accumulation process were not uniform. MAF and NF_H initially showed exhibited lower values compared tothan NF_v and NF_{noBC}, but; however, their later enhancement was stronger than that of NF_{noBC}. Fig. 10-(d) and (e) show the difference between NF_{noBC} and NF_V at 200 nm as a function of secondary aerosolSA mass concentrations during these two pollution periods, which clearly displays how, during secondary aerosol formations SA formation, NF_V became higher than NF_{noBC} while NF_V keptremained close to NF_{noBC} plus the number fraction NF of thickly coated BC-containing aerosols. The (NF_{CBC}) plus NF_{noBC} (NF_{CBC}+NF_{CBC}). These results suggest that secondary aerosolSA formation increases the volatility of BC-free and BC-containing aerosols, leading to an increased NF_V compared towith NF_{noBC}. And almost Almost all BC-free particles and some BCcontaining aerosols became become volatile during the pollution accumulation process of pollution.

4. Conclusions

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The aerosol mixing state is one of the most important physicochemical properties of aerosol particles, which has significant impacts on the and significantly affects their optical properties and the CCN activity of aerosol particles. AerosolThe aerosol mixing state varies largely understates vary significantly with complex aerosol emissions and atmospheric transformations. In this study, aerosol mixing states derived from CCN activity, hygroscopicity, volatility, and BC particlesparticle observations, along with their relationship to primary aerosolsaerosol emissions and secondary aerosol formationsSA formation, were systematically analyzed based on simultaneous measurements of a CCNC, a-H/VTDMA, and a-SP2. Statistical analysis demonstrated that the number fractionNFs of

CCN-active, hygroscopic, and volatile particles were generally positively correlated to one another and were mainly contributed mainly by BC-free aerosols. Therefore, NF_{noBC}, MAF, NF_H, and NF_V four mixing state parameters were all negatively correlated to either MF_{FFOA}the MFs of BBOA or MF_{BBOA}, FFOA because fossil fuel combustion and biomass burning were the two major sources of BC-containing aerosols during this field campaign. However, the differences among between these mixing state parameters varied much vary significantly under different conditions.

The intercomparison results highlight the differences in the aerosol mixing states and physiochemical physicochemical properties caused by fossil fuel combustion and biomass-burning emissions. The CCN and SP2 measurements showed that CCN-inactive, BC-free particles were mainly produced by fossil fuel combustion. On the other hand, the In contrast, a comparison between the VTDMA and SP2 measurements indicated that the non-volatile aerosols were mostly composed ofcomprised BC-containing particles. The comparison between HTDMA and SP2 measurements revealed that a significant proportion of nearly hydrophobic aerosols was not from BC-containing particles, but from fossil fuel combustion or biomass-burning-dominated organic aerosols. The correlation between the CCNC and HTDMA measurements also showed that nearly hydrophobic, BCfree aerosols eancould become CCN-active under supersaturated conditions and wereare tightly linked to biomass—burning emissions. This suggests that biomass—burning aerosols mightmay exhibit different hygroscopicityhygroscopicities under sub- and super- saturated conditions (Bougiatioti et al., 2016). Furthermore, the correlation analysis between SP2 measurements and mass fractions the MFs of BBOA and FFOA indicated that BC-containing aerosols from fossil fuel combustion tendtended to be more externally mixed with other aerosol compositions compared to than those from biomass burning activities.

Other thanIn addition to primary aerosol emissions, secondary aerosol formations SA formation also exerted significants impacts on variations of a aerosol mixing states. During this campaign, the secondary aerosolSA formation was dominated by nitrate and SOA production, which have had markedly different impacts on the aerosol mixing states due owing to their different physical properties and formation pathways. NFnoBC, MAF, NFH, and NFV were all positively correlated with the nitrate and SOA mass fractionsMFs, revealing much stronger correlations to with SOA than with nitrate. This is consistent with the semi-volatile but highly hygroscopic properties of nitratenitrates.

The high correlation coefficient between NF_{noBC} and the mass fractionsMFs of secondary aerosolsSAs at 200 nm and 300 nm suggests that secondary aerosolSA formation led to the migration of BC-free aerosols towards larger diameters more quickly than that of BC-containing aerosols. This outcomercult reveals that secondary aerosols-SAs formed more rapidly on BC-free aerosols than on BC-containing aerosols, which is in linealigns with the BC-containing aerosols' hydrophobic nature of BC-containing aerosols that doesdo not favor aqueous secondary aerosolSA formation. Moreover, as the mass fractionsMFs of nitrate or ammonium increased, the differences among between the mixing state parameters (NF_{noBC}, MAF, NF_{H2} and NF_V) mostly decreased due to because of the hygroscopic and semi-volatile nature of ammonium nitrate. However, the two resolved SOA factors exhibited different impacts on NF_V-the differences between NF_V and NF_{H7} (NF_V-NF_H), and their correlations with NF_V and NF_H revealed that OOA1 was more hygroscopic but less volatile, suggesting distinct formation mechanisms for the those two OOA factors during the field campaign.

The findings of this study highlight the markedly different effects of primary emissions and secondary aerosol formations SA formation on aerosol mixing states and suggest that comparisons of aerosol mixing states obtained using various techniques are useful for gaining insights into the hygroscopicity, volatility, and CCN activity of different aerosols. These comparisons ean also indicate the impacts impact of secondary aerosol formations SA formation on aerosol physical properties, which can help to understand the pathways of secondary aerosol SA formation. However, it is important to be cautious in the application of aerosol mixing state parameters, as because the suitability of VTDMA-derived mixing state parameters in for representing BC mixing states is largely dependent on the composition and mass of secondary aerosolsthe SAs.

- 718 **Data availability**. The data used in this study are available from the corresponding author upon request
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- 720 **Competing interests**. The authors declare that they have no conflict of interest.
- 721
- 722 Author Contributions.
- YK and WY planned this campaign and YK designed the aerosol experiments and conceived this
- research together with JC, and JC wrote the manuscript. JC performed measurements of CCNC, BL
- performed measurements of SP2 and analyzed SP2 datasets with the help of GZ, WQ and YL
- 726 performed AMS measurements, LL performed HV-TDMA measurements and conducted post-data
- processing as well as some of data analysis. BX, HX, MMZ, HZ and SR participated this campaign
- and helped instruments maintenance. GZ provided full support for the campaign. All authors
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Table 1. Definition and description of abbreviations.

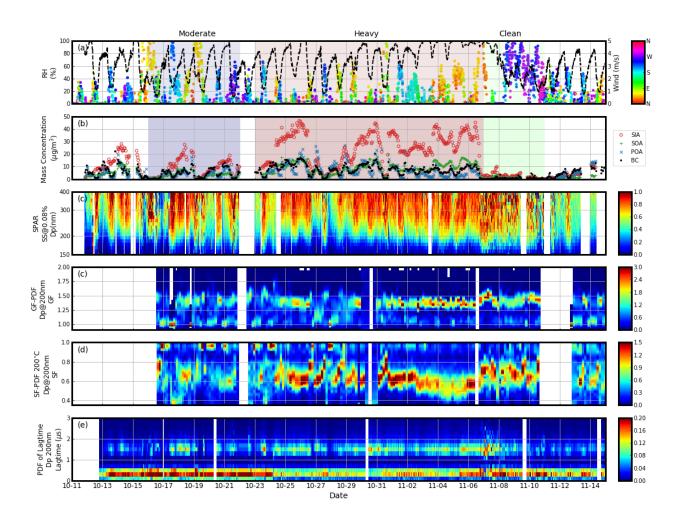
Abbreviation	Full name and/or Definition
BBOA	Biomass Burning Organic Aerosol Characterized by obvious m/z 60 (mainly C ₂ H ₄ O ₂ ⁺) and 73 (mainly C ₃ H ₅ O ₂ ⁺), which are two indicators of biomass burning
<u>FFOA</u>	Fossil Fuel Organic Aerosol A mixed factor that comprises traffic emissions and coal combustion, which was characterized by typical hydrocarbon ion series
<u>OOA</u>	Oxygenated Organic Aerosol
OOA1 and OOA2	Two OOA factors resolved from the PMF analysis
SOA	Secondary Organic Aerosol Summation of OOA1 and OOA2
<u>POA</u>	Primary Organic Aerosol Summation of BBOA and FFOA
SIA	Secondary Inorganic Aerosols, including nitrate, sulfate, and ammonium
<u>PM_{2.5}</u>	Particulate Matter with an aerodynamic diameter <2.5 μm
$\underline{PM_1}$	Particulate Matter with an aerodynamic diameter <1 μm
NR-PM ₁	Non-refractory PM ₁
<u>MF</u>	Mass Fraction
$\underline{D}_{\mathtt{p}}$	Particle diameter after humidification or heating
<u>D</u> d	Particle diameter under dry conditions without humidification or heating
<u>K</u>	Hygroscopicity parameter
<u>SS</u>	<u>Supersaturation</u>
SPAR	Size-resolved Particle Activation Ratio

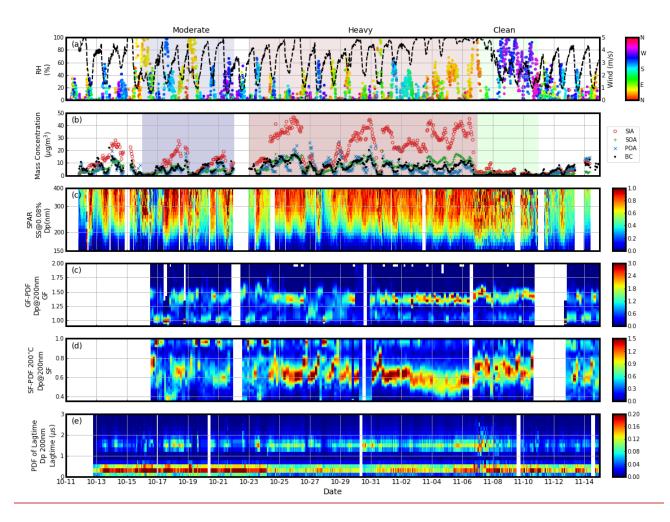
Size-dependent CCN activity under a specific SS

	Maximum Activation Fraction
MAF	An asymptote of the measured SPAR curve at large particle
	sizes and represents the number fraction of CCNs to total
	<u>particles</u>
<u>D</u> a	Midpoint activation diameter
	Linked to the hygroscopicity of CCNs
<u>GF</u>	Growth factor
	The ratio between particles with and without humidification
	and is linked to aerosol hygroscopicity
<u>SF</u>	Shrinkage Factor
	The ratio between particles with and without heating and is
	linked to aerosol volatility
<u>PDF</u>	Probability Distribution Function
NF_{H}	Number Fraction of Hydrophilic aerosol whose
	hygroscopicity parameter is >~0.07
<u>NF</u> _V	Number Fraction of Volatile aerosol whose Shrinkage Factor
	at 200 °C is <0.85
NF _{noBC}	Number Fraction of black carbon (BC)-free particles
<u>INI noBC</u>	Number Traction of black carbon (BC)-nec particles
NF _{CBC}	Number Fraction of thickly coated BC particles
RexBC	The number concentration ratio of externally mixed BC
	particles in total BC-containing particles
	Externally mixed BC particles are defined as identified
	bare/thinly coated BC-containing particles
NF_A-NF_B	
F _H , NF _V -NF _H , NF _{noBC} -	The difference between the number fraction of A and B

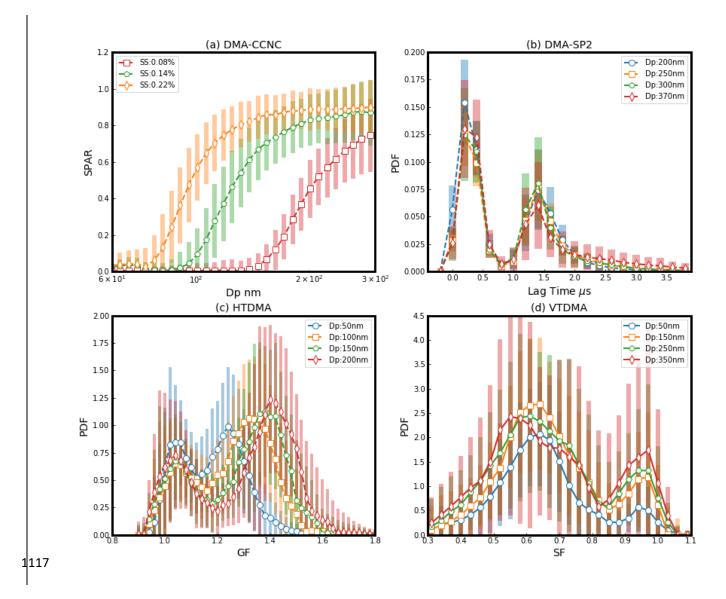
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(NF_{noBC}-NF_H, NF_V-NF_H, NF_{noBC}-NF_V, NF_V-MAF, NF_{noBC}-MAF)





FigFigure 1. Overview of the measurements during the campaign: (a) dots representmeteorological parameters: wind speed with color indicating(dots) and relative humidity (RH) (black line), with colors of dots representing wind direction, and black lines represent RH; (b) mass concentrations of aerosol chemical compositions: secondary inorganic aerosols (SIA, red circle), secondary organic aerosols (SOA, green plus), primary organic aerosols (POA, blue x) and black carbon (BC, black dots represent mass concentration of SIA, SOA, POA and BC, respectively;); (c) Size-resolved Particle Activation Ratio (SPAR) under supersaturation (SS) of 0.08%;% observed by the DMA-CCNC, with warmer colors corresponding to higher values; (d) PDF of GF (GFPDF) at 200 nm; (e) Probability Density Function (PDF) of SF (SFPDF growth factor (GF-PDF) at 200-nm observed by the HTDMA; (e) PDF of shrinkage factor (SF-PDF) at 200 nm and 200-°C observed by the VTDMA; (f) PDF of lag-time at 200-nm-observed by the DMA-SP2. The blue, red, and green shaded periods represent the three periods with moderate pollution, heavy pollution, and clean conditionconditions, respectively.



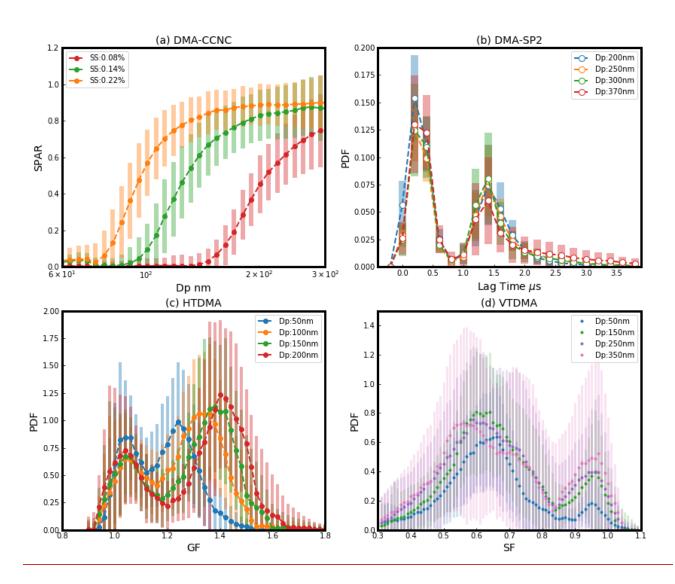
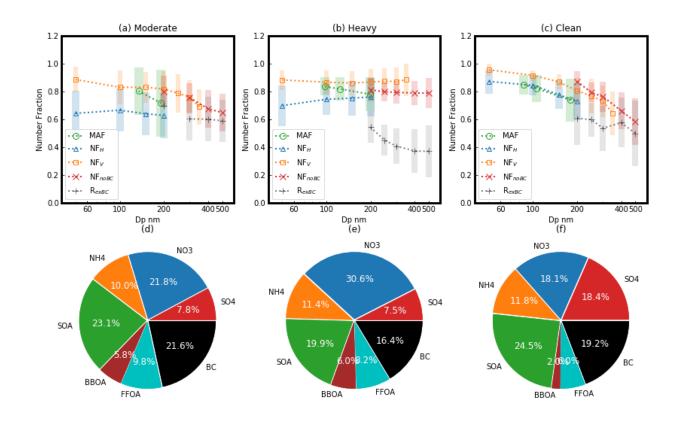


Fig.Figure 2. The campaign average of (a) SPARSize-resolved Particle Activation Ratio (SPAR) curves measured by DMA-CCNC at the three supersaturations (SSs-(identified, represented by different colors and markers), (b) Probability Density Function (PDF) of lag-time measured by DMA-SP2 at four particle sizes (identified represented by different colors and markers), (c) PDF of growth factor (GF-(GFPDF)-) measured by HTDMA at four particle sizes (identified represented by different colors and markers), (d) PDF of SF (SFPDF shrinkage factor (SF) measured by VTDMA under the temperature of 200-_°C at five particle sizes (identified represented by different colors and markers). The shaded areas indicate the standard deviations.



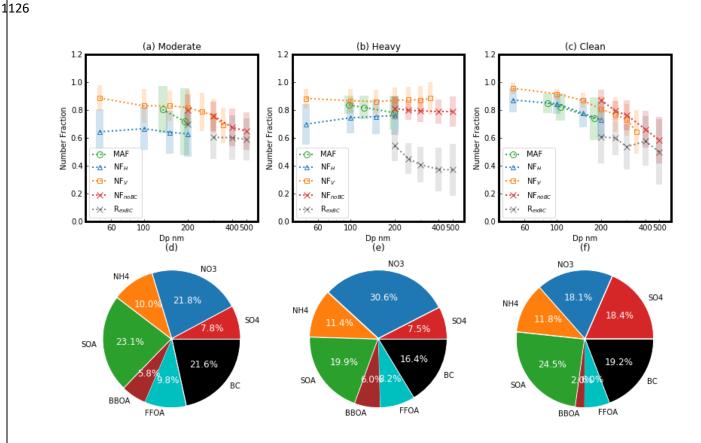
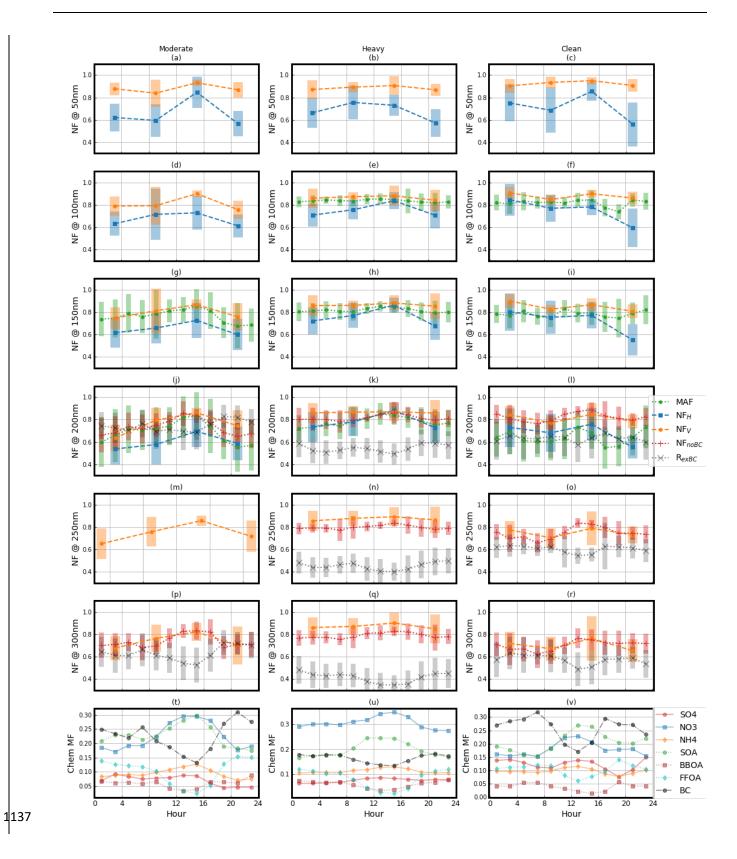


Fig. Figure 3. (a—c):) Size dependence of MAF (green circle), NF_H (blue triangle), NF_V at 200 °C (yellow square-),

NF_{noBC} (red x₎), and R_{exBC} (black plusx) during the three periods. MAF: Maximum Activation Fraction, an asymptote of the measured Size-resolved Particle Activation Ratio (SPAR) curve at large particle. NF_H: Number Fraction of Hydrophilic aerosol whose hygroscopicity parameter is higher than ~0.07. NF_V: Number Fraction of Volatile aerosol whose Shrink Factor at 200 °C is lower than 0.85. NF_{noBC}: Number Fraction of black carbon (BC)-free particles. RexBC: Number fraction of externally mixed BC particles in total BC-containing particles. (d—f):) Corresponding mass fractions (MFs) of aerosol chemical compositions (identified by colors) during the three periods-, including secondary organic aerosols (SOA), biomass burning organic aerosol (BBOA), fossil fuel organic aerosols (FFOA), and inorganic ions including sulfate (SO₄), nitrate (NO₃), and ammonium (NH₄).



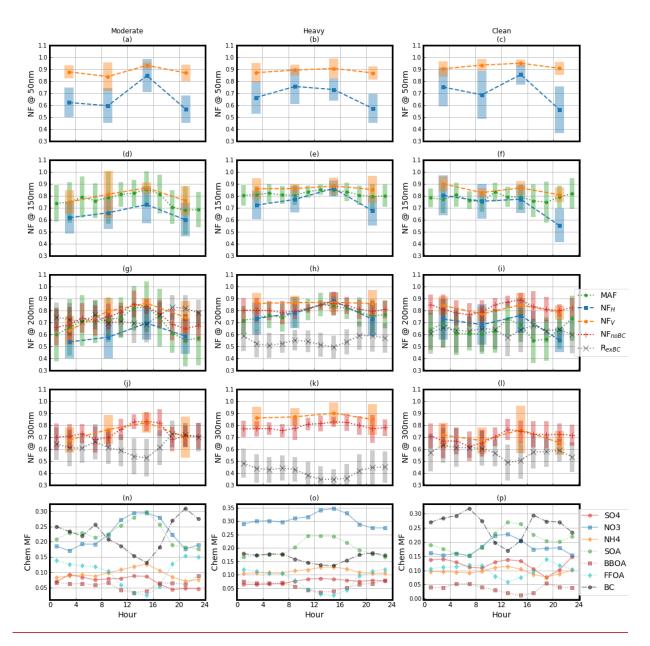
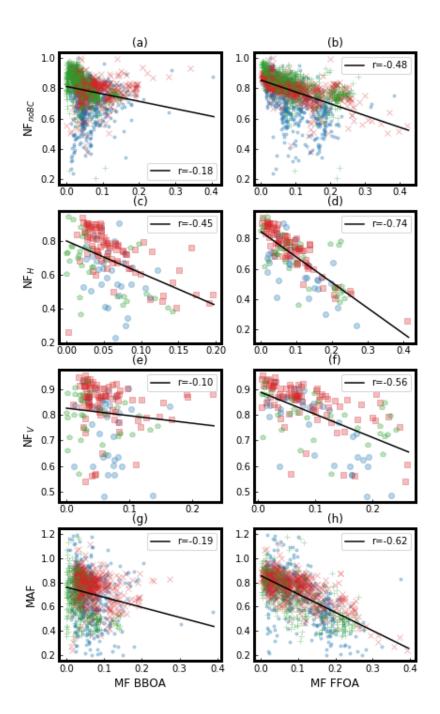


Fig.Figure 4. (a-r_l) Diurnal variation of aerosol mixing state parameters (identified by color and marker) at different particle size sizes (50, 150, 200, and 300 nm) during the three periods. (t-vThe shaded areas indicate the standard deviations. (m-o) Diurnal variation of mass fraction of aerosol chemical compositions fractions (MFs) of aerosol chemical compositions, including secondary organic aerosols (SOA), biomass burning organic aerosol (BBOA), fossil fuel organic aerosols (FFOA), and inorganic ions including sulfate (SO₄), nitrate (NO₃), and ammonium (NH₄) (identified by color and marker) during the three periods.



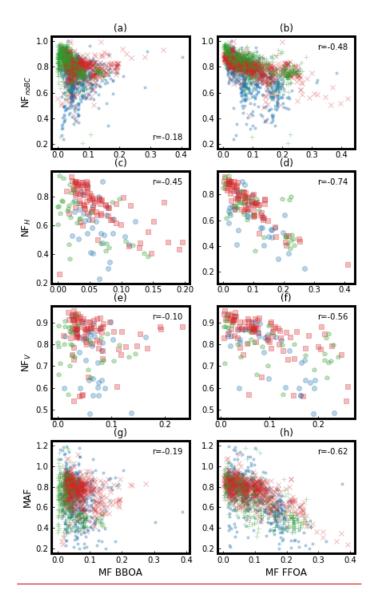


Fig.Figure 5. The correlations between aerosol mixing state parameters and MFmass fractions (MFs) of primarybiomass burning organic aerosol chemical composition (BBOA) and fossil fuel organic aerosols (FFOA) during different periods. Moderately (moderately polluted period: Blueblue dot or circle; Heavilyheavily polluted period: Redred x or square; Cleanclean period: Greengreen plus or pentagon-), with r representing the correlation coefficient.

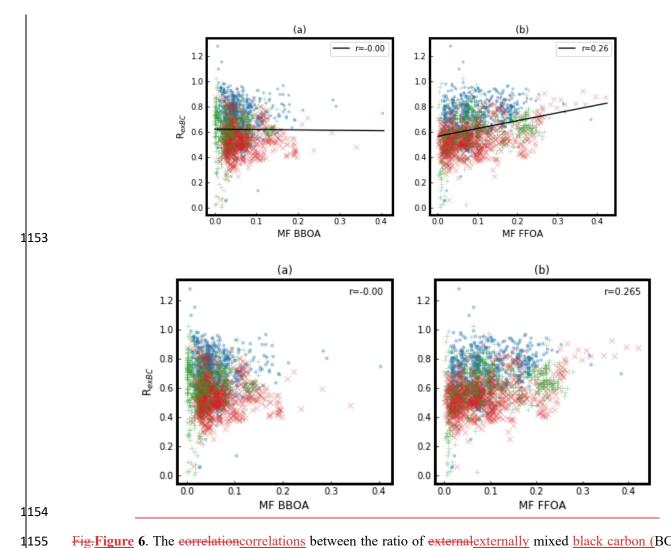


Fig.Figure 6. The correlation between the ratio of external mixed black carbon (BC) in total BC particles (RexBC) and MFmass fractions (MFs) of primary biomass-burning organic aerosol chemical composition (BBOA) and fossil fuel organic aerosols (FFOA) during different periods. Moderately (moderately polluted period: Blueblue dot; Heavily heavily polluted period: Redred x; Cleanclean period: Greengreen plus-), with r representing correlation coefficient.

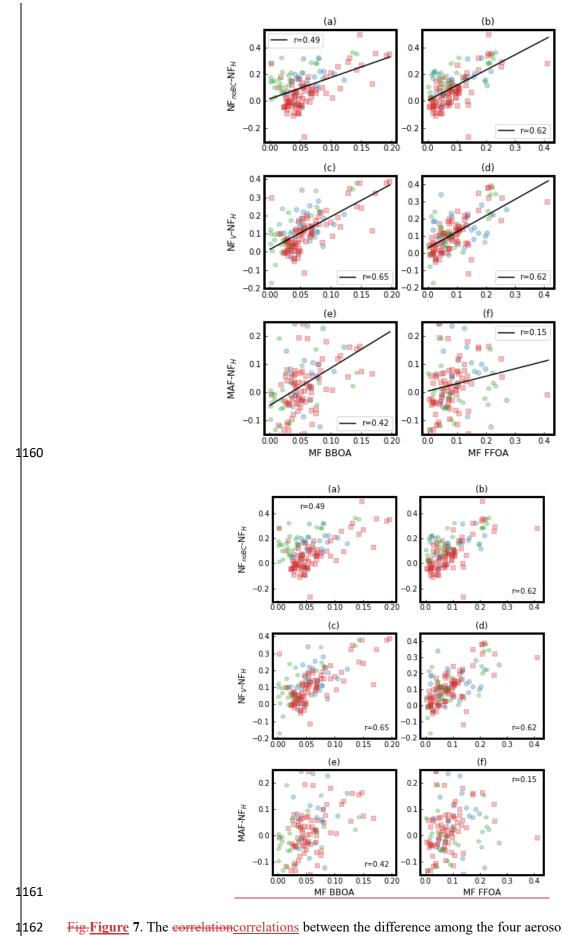
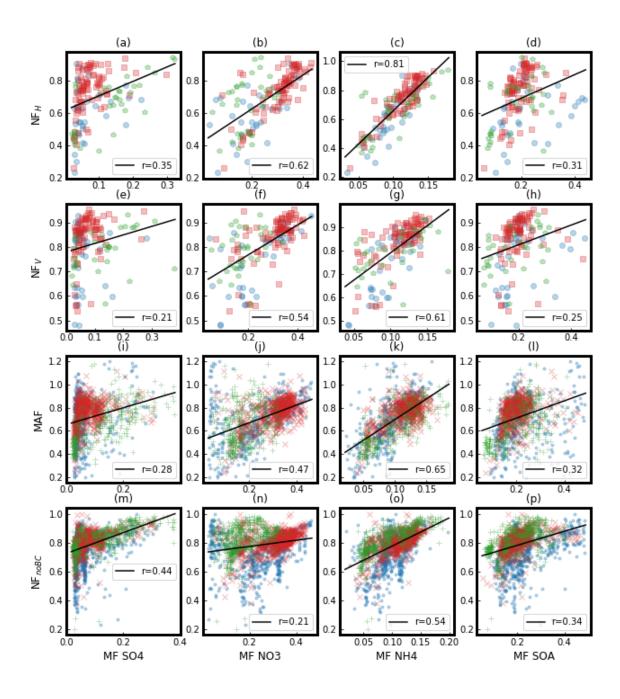


Fig. Figure 7. The correlations between the difference among the four aerosol mixing state parameters at

particle size 200 nm and MFmass fractions (MFs) of primarybiomass burning organic aerosol ehemical composition (BBOA) and fossil fuel organic aerosols (FFOA) during different periods. Moderately moderately polluted period: Blueblue circle; Heavilyheavily polluted period: Redred square; Cleanclean period: Greengreen pentagon.—), with r representing correlation coefficient.



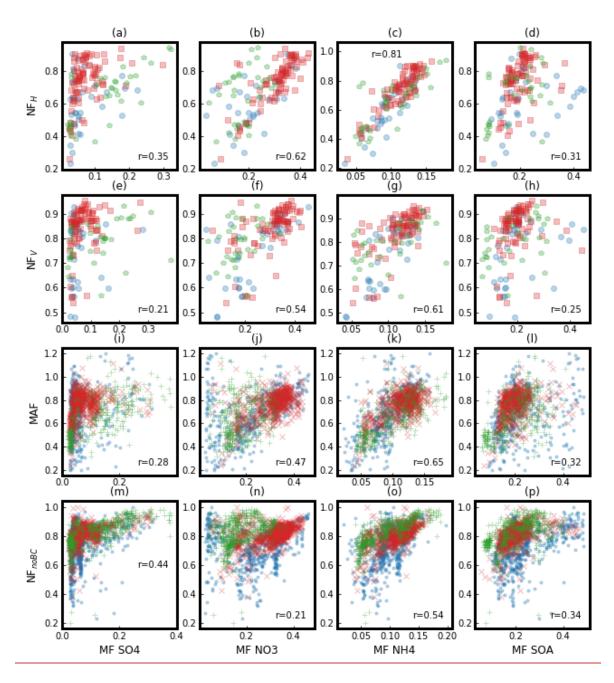
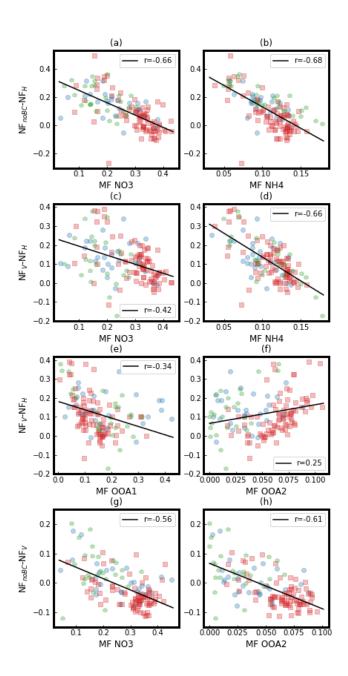


Fig.Figure 8. The correlation between the four aerosol mixing state parameters and mass fraction (MF) of secondary aerosol chemical composition(SA) components during different periods. Moderately (moderately polluted period: Blueblue dot or circle; Heavilyheavily polluted period: Redred x or square; Cleanclean period: Greengreen plus or pentagon-), with r representing correlation coefficient. SA components include secondary organic aerosols (SOA), sulfate (SO₄), nitrate (NO₃), and ammonium (NH₄)



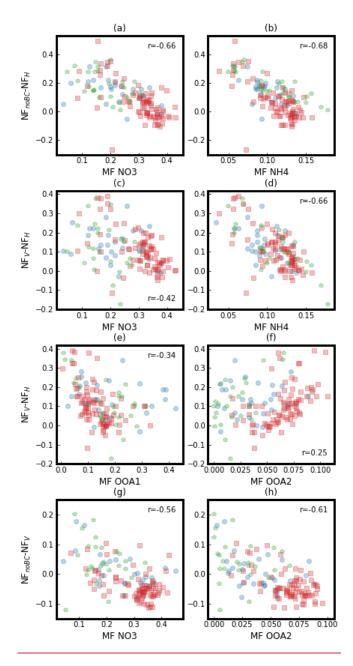


Fig. Figure 9. The correlation between the difference among the four aerosol mixing state parameters and mass fraction (MF) of secondary aerosol (SA) chemical composition components during different periods. OOA1 and OOA2 are two secondary organic aerosol (SOA) factors resolved from aerosol mass spectrometer (AMS) measurements using the Positive Matrix Factorization (PMF) technique. Moderately polluted period: Blueblue circle; Heavilyheavily polluted period: Redred square; Cleanclean period: Greengreen pentagon.

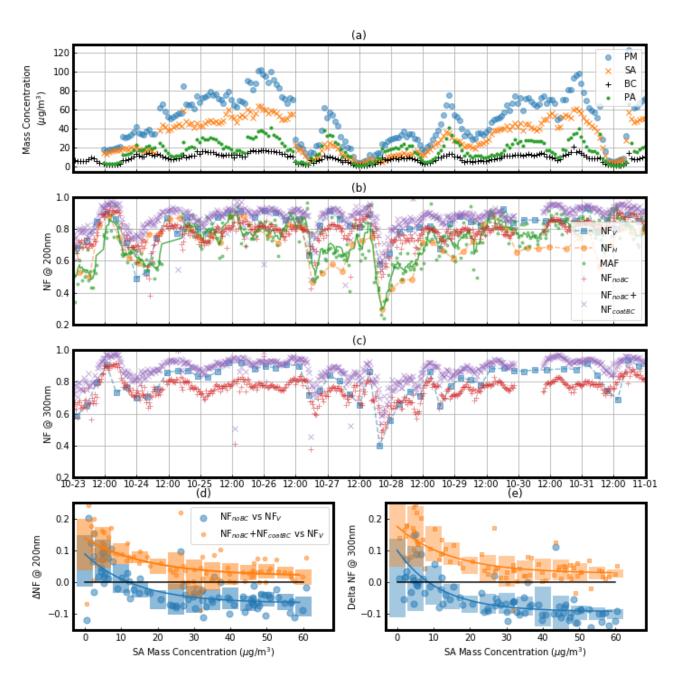


Fig. 10. The variations of different aerosol mixing state parameters during the pollution accumulation process. (a) The mass concentration of PM, SA, PA and BC (identified by colors and markers). (b and c) The variations of different aerosol mixing state parameters (identified by colors and markers) at particle size of 200nm (b) and 300nm (c); (d and e) The variations of the difference between NFV-NFnoBC (blue large circle) and NFV-NFnoBC+CBC (yellow small circle) with the mass concentration of SA at particle size of 200nm (d) and 300nm (e).

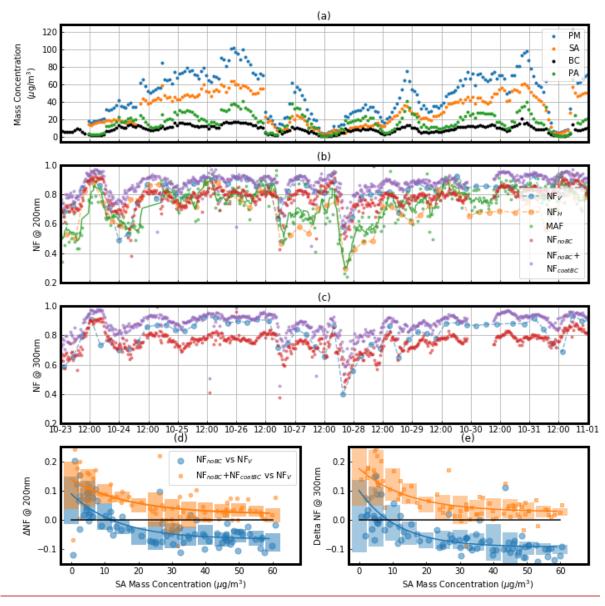


Figure 10. Variations of different aerosol mixing state parameters during the pollution accumulation process. (a) The time series of mass concentrations of non-refractory PM₁ (NR-PM₁), secondary aerosols (SAs) (including inorganic ions and secondary organic aerosols (SOA)), primary organic aerosols (POA) and black carbon (BC) (identified by colors and markers). (b and c) The variations of different aerosol mixing state parameters (identified by colors and markers) at particle size 200 nm (b) and 300 nm (c). (d and e) The variations of the difference between NF_V and NF_{noBC} (NF_V-NF_{noBC}, blue large circle) and the difference between NF_V and NF_{noBC}+NF_{CBC} (NF_V-(NF_{noBC}+NF_{CBC}), yellow small circle) with the mass concentration of SA at particle size 200 nm (d) and 300 nm (e) NF_{CBC}: Number Fraction of thickly coated black carbon (BC) particles.