
1 **Markedly different impacts of primary emissions and secondary**
2 **aerosol formations on aerosol mixing states revealed by simultaneous**
3 **measurements of CCNC, V/HTDMA and SP2**

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25 **Abstract**

26 The aerosol mixing state is a crucial physical-chemical property that affects their optical
27 properties and cloud condensation nuclei (CCN) activity. Multiple techniques are commonly employed
28 to determine aerosol mixing states for various applications, and comparisons between these techniques
29 provide insights of the variations in aerosol chemical and physical properties. These techniques include
30 size-resolved CCN activity measurements using a system with CCN counter (CCNC) coupled with a
31 differential mobility analyzer (DMA), a Humidified/Volatility Tandem differential mobility analyzer
32 (H/V-TDMA) which measures aerosol hygroscopicity/volatility distributions, and a single particle
33 soot photometer (SP2) which directly quantifies black carbon (BC) mixing states. This study provides
34 a first time intercomparisons of aerosol mixing state parameters obtained through simultaneous
35 measurements of a DMA-CCNC, a H/VTDMA and a DMA-SP2. The impact of primary aerosols
36 emissions and secondary aerosol formations on the aerosol mixing states and intercomparison results
37 were analyzed. The results showed that differences in mixing state parameters measured by different
38 techniques varied greatly under different conditions. The V-TDMA and DMA-SP2 measurements
39 showed that the non-volatile population identified by the V-TDMA was mainly contributed by BC-
40 containing aerosols. The HTDMA and DMA-SP2 measurements indicated that a substantial proportion
41 of nearly hydrophobic aerosols were not contributed from BC-containing aerosols, but likely
42 originated from fossil fuel combustion and biomass burning emissions. Synthesized comparison results
43 between DMA-CCNC, HTDMA and DMA-SP2 measurements revealed that some of the nearly
44 hydrophobic BC-free particles were CCN-inactive under supersaturated conditions, likely from fossil
45 combustion emissions, while others were CCN-active under supersaturated conditions linked to
46 biomass burning emissions. Fossil fuel combustion-emitted BC-containing aerosols tended to be more
47 externally mixed with other aerosol ~~compositions~~components compared to those emitted from biomass
48 burning activities. These results highlight significant disparities in the mixing states as well as
49 physiochemical properties between aerosol originated from fossil fuel combustion and biomass
50 burning. The formation of secondary nitrate and organic aerosols exerted significant impacts on
51 variations in aerosol mixing states, generally enhancing aerosol hygroscopicity and volatility, while
52 reducing differences in mixing state parameters derived from different techniques, resulting in a
53 reduction in aerosol heterogeneity. The variations in BC-free particle number fractions showed that

54 secondary aerosols tended to form more quickly on BC-free particles than on BC-containing particles.
55 Further comparison of mixing state parameters revealed that the two resolved ~~SOA~~[secondary organic](#)
56 [aerosol](#) factors in this study exhibited remarkably different physical properties, indicating that they
57 were likely formed through different pathways. These findings suggest that intercomparisons among
58 aerosol mixing states derived from different techniques can provide deeper insight into aerosol
59 physical properties and how they are impacted by secondary aerosol formation, aiding the investigation
60 of secondary aerosol formation pathways.

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62

63 1 Introduction

64 The aerosol mixing state is a crucial physicochemical property of aerosol particles (Riemer et al.,
65 2019), exerting a significant impact on their optical properties and cloud condensation nuclei (CCN)
66 activity, affecting their impact on climate and the environment (Fierce et al., 2017; Riemer et al., 2019;
67 Stevens et al., 2022). For instance, variations in the mixing state of black carbon (BC) particles can
68 significantly alter their absorption and radiative effects (Bond et al., 2013; Lack et al., 2012; Zhao et
69 al., 2019; Moffet et al., 2016; Matsui et al., 2018; Peng et al., 2016). Using simple internally mixing
70 state assumptions for aerosol chemical compositions in estimating CCN number concentrations can
71 lead to substantial overestimations ([up to 30%](#), Deng et al., 2013; Farmer et al., 2015; Ren et al., 2018;
72 Ching et al., 2017, 2019; Tao et al., 2021). The aerosol mixing state varies widely due to complex
73 emissions and atmospheric transformations, leading to significant uncertainties in estimating the
74 effects of aerosol based on simplified mixing state assumptions (Ervens, 2015; Wang et al., 2022; Fu
75 et al., 2022).

76 The aerosol mixing state describes the mixture of aerosol chemical components within each single
77 particle and the distribution of these particles in the aerosol population. This property can be directly
78 measured by the single particle chemical composition techniques (Fierce et al., 2017; Riemer et al.,
79 2019), such as single particle soot photometer (SP2) that measures refractory black carbon (rBC) mass
80 concentrations and the mixing state of rBC with other aerosol components or single particle chemical
81 composition measurement techniques (e.g. [Single Particle Aerosol Mass Spectrometer](#), SP-AMS) that
82 have been developed over recent years (Lee et al., 2019; Riemer et al., 2019 and reference therein).
83 Alternatively, the aerosol mixing state can be inferred from indirect measurements of aerosol properties,
84 such as size-resolved aerosol CCN activity (measured by coupling a differential mobility analyzer
85 (DMA) and a CCN counter (CCNC)), size-resolved aerosol hygroscopicity distributions or volatility
86 distributions (measured by a Humidified/Volatility Tandem differential mobility analyzer (H/V-
87 TDMA)).

88 However, each technique yields information on aerosol mixing states based on different aerosol
89 micro-physical properties, thus obtain aerosol mixing states that are different but linked to one another.
90 For instance, while both CCN activity and the hygroscopic growth measurements are associated with
91 aerosol hygroscopicity, intercomparison between CCNC and HTDMA measurements promoted

92 investigations on aerosol hygroscopicity variations under different saturation conditions (Su et al.,
93 2010; Juranyi et al., 2013; Lance et al., 2013; Kawana et al., 2016; Tao et al., 2020; Jiang et al., 2021).
94 Although both SP2 and VTDMA techniques depend on the evaporation of non-refractory compositions,
95 only the rBC remains in SP2 measurements, whereas the non-refractory composition evaporations
96 depend on thermodeuder temperature in VTDMA measurements. Thus, measurements of an SP2 are
97 highly correlated to those of a VTDMA at high temperatures; [\(higher than 200 °C and up to 300 °C\)](#),
98 with their differences reflecting variations in aerosol density, shape or volatility (Philippin et al., 2004;
99 Wehner et al., 2009; Adachi et al., 2018, 2019; Wang et al., 2022). HTDMA and VTDMA can be
100 applied in combination to study the influence of aerosol mixing state on its hygroscopicity and
101 volatility (Zhang et al., 2016; Cai et al., 2017; Wang et al., 2017). Strong correlations were found
102 between hydrophobic and non-volatile particles, suggesting that they might be of similar chemical
103 composition (Zhang et al., 2016). In addition, some studies showed that except for BC, low-volatility
104 particles correlated well with CCN-inactive particles based on VTDMA and CCNC measurements
105 (Kuwata et al., 2007; Kuwata and Kondo, 2008; Rose et al., 2011; Cheng et al., 2012). Therefore,
106 intercomparisons between mixing state parameters measured by distinct techniques provides not only
107 better characterization of the aerosol mixing state itself, but also insight into aerosol physiochemical
108 properties. Previous studies have mainly compared two kinds of aerosol mixing state measurements,
109 lacking comprehensive comparative analysis among SP2, DMA-CCN, and HV-TDMA measurements,
110 hindering the wide applications of derived aerosol mixing states obtained by individual techniques.

111 The mixing state of primary aerosols can vary greatly depending on their type and emission
112 conditions (Cheng et al. 2012; Wang et al. 2017; Wang et al., 2022; Ting et al., 2018; Liu et al., 2021),
113 and can be significantly altered during aging processes or secondary formations (Wehner et al. 2009;
114 Cheng et al. 2012; Wang et al., 2022; Tomlin et al., 2021; Lata et al., 2021). Primary aerosol emissions
115 such as biomass burning, fossil fuel combustion and cooking tend to contribute to weakly hygroscopic
116 (Herich et al., 2008, 2009; Wang et al., 2020; Kim et al., 2020) and low-volatility aerosols (Hong et
117 al., 2017; Saha et al., 2018), while formation of secondary aerosol including aging of BC-containing
118 aerosols and primary organic aerosols mainly contribute to aerosols with strong CCN activity (Mei et
119 al., 2013; Ma et al., 2016; Tao et al., 2021) and high hygroscopicity (Chen et al., 2018; Kim et al.,
120 2020; Wang et al., 2020). To enhance our understanding of the mixing state of aerosols from different

121 emission sources and to improve its characterization in models, it is important to study the impact of
122 specific primary aerosol emissions and secondary aerosol formations on aerosol mixing states and the
123 influence on aerosol mixing state parameters derived from different techniques.

124 The North China Plain (NCP) is among the most polluted regions in China with various primary
125 emission sources and strong secondary aerosol formations playing critical roles in air pollution (Xu et
126 al., 2011; Tao et al., 2012; Liu et al., 2015). The complex mixing state of aerosols in the NCP
127 contributes to uncertainties in evaluating their climate and environmental effects (Zhuang et al., 2013;
128 Nordmann et al., 2014; Zhang et al., 2016; Tao et al., 2020; Shi et al., 2022), particularly with regards
129 to BC particles (Wu et al., 2017; Liu et al., 2019; Zhao et al., 2019; Wang et al., 2011; Zheng et al.,
130 2019). Meteorological conditions can greatly impact the secondary aerosol formation in the NCP,
131 which can be significantly exacerbated during severe pollution events ~~(Kuang et al., 2020), and~~
132 ~~consequently affect the aerosol mixing states (Tao et al., 2021).~~ The secondary aerosol formations
133 under low RH conditions, mainly taken place in gaseous phase, would change to that mainly taken
134 place in aqueous phase under high RH conditions (Kuang et al., 2020). As secondary aerosols formed
135 through different mechanisms have different chemical compositions and add mass to different particle
136 groups, secondary aerosol formations under different meteorological conditions can affect the aerosol
137 mixing states (Tao et al., 2021). This study obtained aerosol mixing state through concurrent
138 measurements of CCN activity, hygroscopicity, volatility and BC particle at a regional site in the NCP,
139 using CCNC, HTDMA, VTDMA and SP2 instruments. This provides a unique opportunity to perform
140 a most comprehensive intercomparison of aerosol mixing state among different techniques for insights
141 into the impact of primary aerosol emissions and secondary aerosol formations on the observed aerosol
142 mixing states.

143

144 **2 Materials and methods**

145 **2.1 Campaign information and instruments set-up**

146 From 16th October to 16th November 2021, aerosol mixing states were continuously and
147 concurrently monitored using different techniques at the Gucheng site in Dingxing county, Hebei
148 province, China, as part of the campaign of investigating AQueous Secondary aerOsol formations in
149 Fogs and Aerosols and their Radiative effects in the NCP (AQ-SOFAR). The observation site, located

150 at 39°09'N, 115°44'E, is an Ecological and Agricultural Meteorology Station of the Chinese Academy
151 of Meteorological Sciences, situated between the megacities of Beijing (approximately 100 km away)
152 and Baoding (approximately 40 km away), and surrounded by farmlands and small towns. This site
153 provides a representative view of the background conditions of atmospheric pollution in the NCP
154 (Kuang et al., 2020; Li et al., 2021).

155 Different measurement techniques were used during the campaign to simultaneously obtain the
156 aerosol mixing state through CCN activity, hygroscopicity, volatility, and BC particle observations (to
157 be discussed in the next section). In addition to aerosol mixing state measurements, the AQ-SOFAR
158 campaign also included measurements of aerosol number size distribution, chemical composition,
159 aerosol scattering and absorption properties. ~~The total rBC~~The aerosol number size distribution
160 covered diameter range of 13 nm to 4 μm are measured by the scanning mobility particle sizer (13-
161 550 nm) and the aerodynamic aerosol classifier (100 to 4 μm), and they are merged together by
162 assuming the aerosol density of 1.6 g/cm³. The total BC mass concentrations were determined using
163 an aethalometer (Magee, AE33; Drinovec et al., 2015), with more information about the correction of
164 the absorption measurements and mass concentration calculations available in Luo et al. (2022). All
165 aerosol measurement instruments were housed in a temperature-controlled container set at 24 °C. The
166 inlet was switched among three impactors: TSP (Total Suspended Particles), PM_{2.5} (Particulate matter
167 with aerodynamic diameter less than 2.5 μm) and PM₁ (Particulate matter with aerodynamic diameter
168 less than 1 μm). Inlet changes would affect the dry state aerosol sampling due to aerosol hygroscopic
169 growth or activation. However, the aerosol mixing state and aerosol chemical composition
170 measurements were made on submicron aerosols, inlet change almost does not affect those
171 measurements under conditions of RH less than 90%, and this would be discussed very carefully in
172 our next paper. The sampled aerosol was dried by two parallelly assembled Nafion dryers with length
173 of 1.2 m. During autumn and winter in the NCP, ambient air temperature (lower than 20 °C and can
174 down to 0 °C) can be significantly lower than the room temperature (~24 °C), this dryer can maintain
175 the RH of sampled aerosols to below 20%. Meteorological data, such as temperature, pressure, wind
176 speed and direction, relative humidity (RH), were obtained from an automatic weather station operated
177 by the station.

178 The chemical composition of submicron aerosols was analyzed using a High-Resolution Time-
179 of-Flight Aerosol Mass Spectrometer (HR-ToF-AMS). The ionization efficiency (IE) was calibrated

180 using 300 nm diameter pure NH_4NO_3 particles, following the standard protocols outlined in Jayne et
181 al. (2000) in the middle of the campaign, with the relative ionization efficiency (RIE) of ammonium
182 determined to be 5.26. The RIE of sulfate was 1.28 using pure $(\text{NH}_4)_2\text{SO}_4$ particles, and the default
183 RIEs of 1.4 for organic aerosols, 1.1 for nitrates, and 1.3 for chlorides were used. The composition
184 dependent collection efficiency reported by Middlebrook et al. (2012) was used in this study.
185 Elemental ratios were derived using the “Improved-Ambient (I-A)” method as described in
186 Canagaratna et al. (2015), including hydrogen to carbon (H/C), oxygen to carbon (O/C), and organic
187 mass to organic carbon (OM/OC) ratios. Two POA ([Primary Organic Aerosol](#)) factors and two
188 ~~SOA~~[oxygenated organic aerosol \(OOA\)](#) factors were identified by HR-PMF ([High-Resolution](#)
189 [Positive Matrix Factorization](#), Ulbrich et al., 2009; Paatero and Tapper, 1994). [Summations of the two](#)
190 [OOA factors are used to represent secondary organic aerosols \(SOA\) in this study](#). The mass spectra
191 of the OA factors and their correlations with external species are shown in Figs. S1-S2. The BBOA
192 ([Biomass Burning Organic Aerosol](#)) spectrum was characterized by obvious m/z 60 (mainly $\text{C}_2\text{H}_4\text{O}_2^+$)
193 and 73 (mainly $\text{C}_3\text{H}_5\text{O}_2^+$), two indicators of biomass burning (Mohr et al., 2009). In addition, BBOA
194 was correlated well with $\text{C}_2\text{H}_4\text{O}_2^+$ ($R^2=0.91$) and $\text{C}_3\text{H}_5\text{O}_2^+$ ($R^2=0.90$). Consistent with previous studies
195 in Beijing (Xu et al., 2019), the PMF analysis revealed a mixed factor named FFOA ([Fossil Fuel](#)
196 [Organic Aerosol](#)) that comprises traffic emissions and coal combustions, which was characterized by
197 typical hydrocarbon ion series. It was observed that FFOA had a relatively high f_{44} (0.083) value,
198 which is likely due to aging during regional transportation, similar to the results observed in the winter
199 of 2016 in Beijing (Xu et al., ~~2019~~) and ~~CCOA in Gucheng (Chen et al., 2022)~~. ~~Two SOA~~[2019](#) and
200 [coal combustion organic aerosol in Gucheng \(Chen et al., 2022\)](#). [Secondary organic aerosol formations](#)
201 [originated from volatile organic compounds precursors could be formed in differ formation pathways](#)
202 [such as aqueous phase reactions, heterogeneous reactions or gas phase reactions and also might be](#)
203 [oxidized under different conditions, for example, oxidized under different nitrogen oxide conditions](#)
204 [with different oxidation capacity and oxidants](#). Two resolved OOA factors were found to display
205 different spectral patterns, correlations with tracers and diurnal variations, suggesting that they resulted
206 from different chemical processing. ~~For example,~~ [however, their formation mechanisms remain to be](#)
207 [explored in our future studies](#). In general, the OOA factor 1 (OOA1) had higher $\text{CO}_2^+/\text{C}_2\text{H}_3\text{O}^+$ (3.9)
208 and O/C (0.91) ratios compared to [the OOA factor 2 \(OOA2](#), ~~2.1~~, and 0.78).

209 Note that losses in inlet line and sampling systems are not considered in this study. reasons are
210 listed below: (1) investigated mixing state parameters are represented by number fractions of different
211 diameters which are much less affected by losses in sampling systems compared with absolute number
212 concentrations; (2) good consistency was achieved between measurements of particle number size
213 distributions (PNSD) by and mass concentrations measured by AMS, with the average ratio between
214 volume concentration derived from AMS and rBC measurements and volume concentration derived
215 from PNSD measurements is 0.79 (R=0.97, as shown in Fig. S3), which is consistent with previous
216 reports due to that AMS cannot detect aerosol components such as dust (Kuang et al., 2021). This
217 means that almost same aerosol populations were sampled by AMS and instruments of measuring
218 aerosol mixing states.

219 **2.2 Aerosol mixing states measurement techniques**

220 **2.2.1 DMA-CCNC measurements**

221 The CCN activity of particles under supersaturated conditions was measured using a DMA-
222 CCNC system, which consisted of a differential mobility analyzer (DMA, model 3081 by TSI, Inc.,
223 MN USA), a condensation particle counter (CPC, model 3756 by TSI, Inc., MN USA), and a
224 continuous-flow CCNC (model CCN100 by Droplet Measurement Technologies, USA). The system
225 operated in a size-scanning mode and provided the Size-resolved Particle Activation Ratio (SPAR) by
226 combining the CPC and CCNC measurements at different particle sizes. ~~Three~~In order to perform
227 intercomparisons among instruments, three supersaturations (SSs) of 0.08%, 0.14% and 0.22% were
228 applied in a single cycle of about 15 minutes. CCN measurement under these three SSs reveals mainly
229 CCN activity of aerosols reside in accumulation mode aerosol with diameter range of about 100-200
230 nm, which are close to diameters of HV-TDMA measurements, and higher SSs would reveal CCN
231 activity of smaller aerosol particles (<100 nm) where DMA-SP2 measurement is not available. The
232 sample and sheath flow rates of the DMA were set at 1 lpm and 5 lpm, respectively, resulting in a
233 measured particle diameter range of 9 nm to 500 nm with a running time of 5 minutes per cycle.
234 Supersaturations in the CCNC were calibrated with monodisperse ammonium sulfate particles (Rose
235 et al., 2008) both before and after the campaign. The flow rates were also calibrated before and after
236 the campaign and were checked daily to minimize uncertainties in the droplet counting and
237 supersaturation formed in the column (Roberts and Nenes, 2005; Lance et al., 2006). The SPAR

238 deviations due to multiple-charge particles were corrected using a modified algorithm based on Hagen
239 and Alofs (1983) and Deng et al. (2011). More details about the system can be found in Ma et al. (2016)
240 and Tao et al. (2021).

241 2.2.2 H/V-TDMA measurements

242 The mixing state of aerosols in terms of hygroscopicity and volatility was measured using a
243 Hygroscopicity/Volatility Tandem Differential Mobility Analyzer (H/V-TDMA; Tan et al., 2013). The
244 H/V-TDMA consists of two DMA (Model 3081L, TSI Inc.), with the first DMA (DMA1) selecting
245 dried particles without conditioning and the second DMA (DMA2) selecting conditioned particles. The
246 H/V-TDMA can operate in either H-mode or V-mode, which is controlled by a three-way solenoid
247 valve. In H-mode, a Nafion humidifier is used to condition the selected dry particles to 90% relative
248 humidity (RH) equilibrium. The number size distribution of the humidified particles (D_p) is measured
249 by DMA2 and a CPC (Model 3772, TSI Inc.). The RH-dependent hygroscopic growth factor (GF) at
250 a certain dry diameter (D_d) is calculated as follows:

$$251 \quad GF = \frac{D_p}{D_d} \frac{D_p(RH)}{D_d} \quad (1)$$

252 Where $D_p(RH)$ is the particle diameter undergo humidification. In this mode, four dry electrical
253 mobility diameters (50, 100, 150 and 200 nm) were measured. The instrument was regularly calibrated
254 using Standard Polystyrene Latex Spheres (PSL) and Ammonium Sulfate particles.

255 In the V-mode, a heated tube was used to evaporate the volatile coatings from the previously
256 selected dry particles. Six temperature settings were used for the heated tube, ranging from 25 to 200°C.
257 The number size distributions of the heated particles were measured using DMA2 and CPC. Besides
258 the four particle sizes measured in H-mode, three additional particle sizes (250, 300 and 350 nm) were
259 measured in the V mode. The temperature dependent shrinkage factor (SF), which is the ratio of the
260 heated particle size to the dry particle size, without heating (D_d), is defined as:

$$261 \quad SF = \frac{D_p}{D_d} \frac{D_p(T)}{D_d} \quad (2)$$

262 Where $D_p(T)$ is the particle diameter undergo heating. A complete cycle of both H-mode
263 measurements at one relative humidity (RH) condition and V-mode measurements at six temperature
264 conditions took approximately 3 hours. A Probability Density Function (PDF) of GF (or SF) can be

265 calculated from the measured density function using the inversion algorithm described by Stolzenburg
266 and McMurry (2008).

267 **2.2.3 DMA-SP2 measurements**

268 The size-resolved BC mixing states were measured by the SP2 (Droplet Measurement
269 Technology, Inc., USA) after a DMA (Model 3081, TSI, USA). The DMA selected aerosols at various
270 dry particle sizes, which were then introduced to the SP2. The DMA-SP2 setup was able to measure
271 the mixing states of aerosols at diameters of 100 nm, 120 nm, 160 nm, 200 nm, 235 nm, 270 nm, 300
272 nm, 335 nm, 370 nm, 400 nm, 435 nm, 470 nm, 500 nm, 535 nm, 570 nm, 600 nm, 635 nm, 670 nm,
273 700 nm within 20 minutes, when it wasn't placed after a thermodenuder-bypass switch system- [\(13th-](#)
274 [24th October, 09:00 am of 5th November to 09:00 am of 8th November\)](#). However, it only measured
275 mixing states at diameters of 120 nm, 160 nm, 200 nm, 250 nm, 300 nm, 400 nm, and 500 nm when it
276 was placed after a thermodenuder-bypass switch system- [\(11:00 am 24th October to 08:00 am 5th](#)
277 [November, and 09:00 am of 8th November to 06:00 pm of 17th November\)](#). [Because the measurements](#)
278 [of HTDMA and VTDMA are conducted solely by a single H/VTDMA system working in different](#)
279 [mode, the time needed for a single particle size measurement of HTDMA and VTDMA is much longer](#)
280 [than that of DMA-SP2 system. Thus, more particle sizes are selected in the measurement DMA-SP2](#)
281 [system for acquiring BC mass concentration and mixing state at more diameters, compared to those of](#)
282 [HTDMA and VTDMA.](#)

283 The SP2 chamber has a continuous Nd: YAG laser beam at a wavelength of 1064 nm. The BC-
284 containing particles passing through the laser beam will be incandescent through absorbing the
285 radiation. By measuring the intensity of the emitted incandescent light, the mass concentration of BC
286 can be calculated. The sheath flow/sample flow ratio was maintained at 10 for DMA to reduce the
287 width of diameter distribution of selected monodisperse aerosols. Additionally, the flow rate of the
288 SP2 was changed from 0.1 to 0.12 L/min starting on October 22nd- [\(flow rate range of SP2: 0.03 to](#)
289 [0.18 L/min\)](#). The SP2 was calibrated using Aquadag soot particles, as reported by Gysel et al. (2011).
290 Further details about the calibrations are introduced in Section 1 of the supplement.

291

292 **2.3 Derivations of mixing state parameters**

293 **2.3.1 Fitting SPAR curves measured by the DMA-CCNC system**

294 The measured SPAR curves can be parameterized with a sigmoidal function with three parameters.
295 As shown in Fig. S4, the measured SPAR is generally characterized as a sigmoidal curve. This
296 parameterization assumes that the aerosol is an external mixture of hydrophilic particles that are CCN-
297 active and hydrophobic particles that are CCN-inactive (Rose et al., 2010). The formula used to
298 parameterize SPAR ($R_a(D_d)$) for a specific SS is as follows (Rose et al., 2008):

$$299 \quad R_a(D_d) = \frac{\text{MAF}}{2} \left(1 + \text{erf} \left(\frac{D_d - D_a}{\sqrt{2\pi}\sigma} \right) \right) \quad (7)$$

300 where erf is the error function. MAF (Maximum Activation Fraction) is an asymptote of the
301 measured SPAR curve at large particle sizes as shown in Fig. S4, and it represents the number fraction
302 of CCNs to total particles. D_a is the midpoint activation diameter and is linked to the hygroscopicity
303 of CCNs, and indicates the diameter where SPAR equals the half of the MAF value. σ is the standard
304 deviation of the cumulative Gaussian distribution function and characterizes the heterogeneity of CCN
305 hygroscopicity. In Fig. S4, the parameter corresponds to the slope of steep increase of SPAR curves
306 when diameter is close to D_a . Generally, hydrophilic particles larger than D_a can become CCN, so
307 these three parameters can be used to characterize the hygroscopicity of those hydrophilic particles.
308 This study did not consider the impact of nearly hydrophobic particles on SPAR, as deviations from
309 this parameterization scheme due to this impact are negligible at low SSs, as stated in Tao et al. (2020).

310

311 2.3.2 Classification of particle type based on hygroscopicity or volatility

312 In this study, ambient aerosol particles were classified into two groups based on their
313 hygroscopicity (hydrophobic and hydrophilic) and two groups based on their volatility (non-volatile
314 and volatile) based on the measurements from H/V-TDMA (Wehner et al., 2009; Liu et al., 2011;
315 Zhang et al., 2016). Each group can be defined using a critical value of GF or SF as the following:

316 Hydrophobic population: $\text{GF} < \text{GF}_C$;

317 Hydrophilic population: $\text{GF} \geq \text{GF}_C$;

318 Non-volatile population: $\text{SF} \geq \text{SF}_C$;

319 Volatile population: $\text{SF} < \text{SF}_C$.

320 The critical values of GF (GF_C) and SF (SF_C) in H/V-TDMA depend on particle size and working
321 conditions such as relative humidity and heating temperature.

322 During this campaign, the SF_C was set at 0.85 for all seven measured particle sizes at a temperature of
323 200 °C. The GF_C for the four measured particle sizes [of 50 nm, 100 nm, 150 nm and 200 nm](#) were 1.1,
324 1.15, 1.175 and 1.2, respectively, and the corresponding hygroscopicity parameter, κ , was
325 approximately 0.07. These values of GF_C and SF_C divide the probability density functions (PDFs) of
326 SF and GF into two modes, consistent with prior studies in the NCP (Liu et al., 2011; Zhang et al.,
327 2016). The number fraction (NF) for the hydrophilic group (NF_H) and volatile group (NF_V) can be
328 calculated as:

$$329 \quad NF_H = \int_{GF_C}^{\infty} GF PDF(GF) dGF \quad (7)$$

$$330 \quad NF_V = \int_0^{SF_C} SF PDF(SF) dSF \quad (8)$$

331 where GFPDF and SFPDF are the PDFs of GF and SF, respectively, derived from H/V-TDMA
332 measurements.

333 **2.3.3 Classification of particle type based on DMA-SP2 measurements**

334 The BC-containing aerosols can be categorized into two groups based on the coating thickness:
335 bare BC/thinly coated BC particles and thickly coated BC particles. [For measurement of coated BC
336 particles in SP2, the incandescence signals is generally detected later than the scattering signals and
337 the time difference between the occurrences of the peaks of the incandescence and scattering signals
338 is defined as the lag time \(Moteki & Kondo, 2007; Sedlacek et al., 2012; Subramanian et al., 2010\).](#)

339 The coating thickness of the BC-containing aerosols in the SP2 measurement can be
340 ~~determined~~[indicated](#) by the lag time between the peak of the scattering signal and the incandescence
341 signal (Moteki and Kondo, 2007; Schwarz et al., 2006; Sedlacek et al., 2012; Subramanian et al., 2010;
342 Metcalf et al., 2012), [which exhibits a clear two-mode distribution in former studies \(Zhang et al.,
343 2018; Zhao et al., 2021\).](#) A critical lag-time threshold can be used to differentiate between the different
344 types of BC-containing aerosols and calculate the number fraction of bare BC particles and coated BC
345 particles in the total identified aerosols. In this study, a two-mode distribution of the lag-time (Δt) was
346 observed, ~~and~~ [As the lag time is positively correlated to the coating thickness,](#) a critical value of 0.8 μ

347 s was used to classify the BC-containing particles into thinly coated (or bare) BC ($\Delta t < 0.8 \mu\text{s}$) and
 348 thickly coated BC ($\Delta t \geq 0.8 \mu\text{s}$), respectively. [In addition, the definitions of all abbreviations are listed](#)
 349 [in Table 1.](#)

350 **Table 1. Definition and description of the abbreviations.**

<u>Abbreviations</u>	<u>Full name and/or Definition</u>
	<u>Biomass Burning Organic Aerosol</u>
<u>BBOA</u>	<u>In this study, characterized by obvious m/z 60 (mainly $\text{C}_2\text{H}_4\text{O}_2^+$) and 73 (mainly $\text{C}_3\text{H}_5\text{O}_2^+$), which are two indicators of biomass burning</u>
	<u>Fossil Fuel Organic Aerosol</u>
<u>FFOA</u>	<u>A mixed factor in this study that comprises traffic emissions and coal combustions, which was characterized by typical hydrocarbon ion series</u>
	<u>Oxygenated Organic Aerosol</u>
<u>OOA1 and OOA2</u>	<u>Two OOA factors resolved from the PMF analysis</u>
	<u>Secondary Organic Aerosol</u>
<u>SOA</u>	<u>Summation of OOA1 and OOA2</u>
	<u>Primary Organic Aerosol</u>
<u>POA</u>	<u>Summation of BBOA and FFOA</u>
<u>PM_{2.5}</u>	<u>Particulate Matter with aerodynamic diameter less than 2.5 μm</u>
<u>PM₁</u>	<u>Particulate Matter with aerodynamic diameter less than 1 μm</u>
<u>NR-PM₁</u>	<u>Non-refractory PM₁</u>
<u>MF</u>	<u>Mass Fraction</u>
<u>D_p</u>	<u>Particle diameter after humidification or heating</u>
<u>D_d</u>	<u>Particle diameter under dry conditions without humidification or heating</u>
<u>κ</u>	<u>Hygroscopicity parameter</u>

<u>SS</u>	<u>Super-saturation</u>
<u>SPAR</u>	<u>Size-resolved Particle Activation Ratio</u> <u>Size-dependent CCN activity under a specific SS</u>
<u>MAF</u>	<u>Maximum Activation Fraction</u> <u>An asymptote of the measured SPAR curve at large particle sizes and represents the number fraction of CCNs to total particles</u>
<u>Da</u>	<u>Midpoint activation diameter</u> <u>Linked to the hygroscopicity of CCNs</u>
<u>GF</u>	<u>Growth factor</u> <u>The ratio between particle with and without humidification, and is linked to aerosol hygroscopicity</u>
<u>SF</u>	<u>Shrinkage Factor</u> <u>The ratio between particle with and without heating, and is linked to aerosol volatility</u>
<u>PDF</u>	<u>Probability Distribution Function</u>
<u>NF_H</u>	<u>Number Fraction of Hydrophilic aerosol whose hygroscopicity parameter is higher than ~0.07.</u>
<u>NF_V</u>	<u>Number Fraction of Volatile aerosol whose Shrink Factor at 200 °C is lower than 0.85.</u>
<u>NF_{noBC}</u>	<u>Number Fraction of BC-free particles</u>
<u>NF_{CBC}</u>	<u>Number Fraction of thickly coated BC particles</u>
<u>R_{exBC}</u>	<u>Number concentration ratio of externally BC particles in total BC-containing particles.</u> <u>Externally BC particles are defined as identified bare/thinly coated BC-containing particles</u>
<u>NF_A-NF_B</u> <u>(NF_{noBC}-NF_H, NF_V-NF_H, NF_{noBC}-NF_V, NF_V-MAF, NF_{noBC}-MAF)</u>	<u>The difference between the number fraction of A and B.</u>

352 3 Results and discussions

353 3.1 Campaign Overview

354 The time series of meteorological parameters, [aerosol mixing state measurements by different](#)
355 [techniques and mass concentrations of aerosol chemical compositions](#) are shown in Fig. 1. In detail,
356 [the measurements of aerosol mixing states include SPAR \(Size-resolved Particle Activation Ratio\) at](#)
357 [SS of 0.08%,% by DMA-CCNC, GF-PDF \(PDF of Growth Factor\) at 200 nm by HTDMA, SF-PDF](#)
358 [\(PDF of Shrink Factor\) at 200 nm and 200 °C by VTDMA, lag-time PDF of 200 nm BC-containing](#)

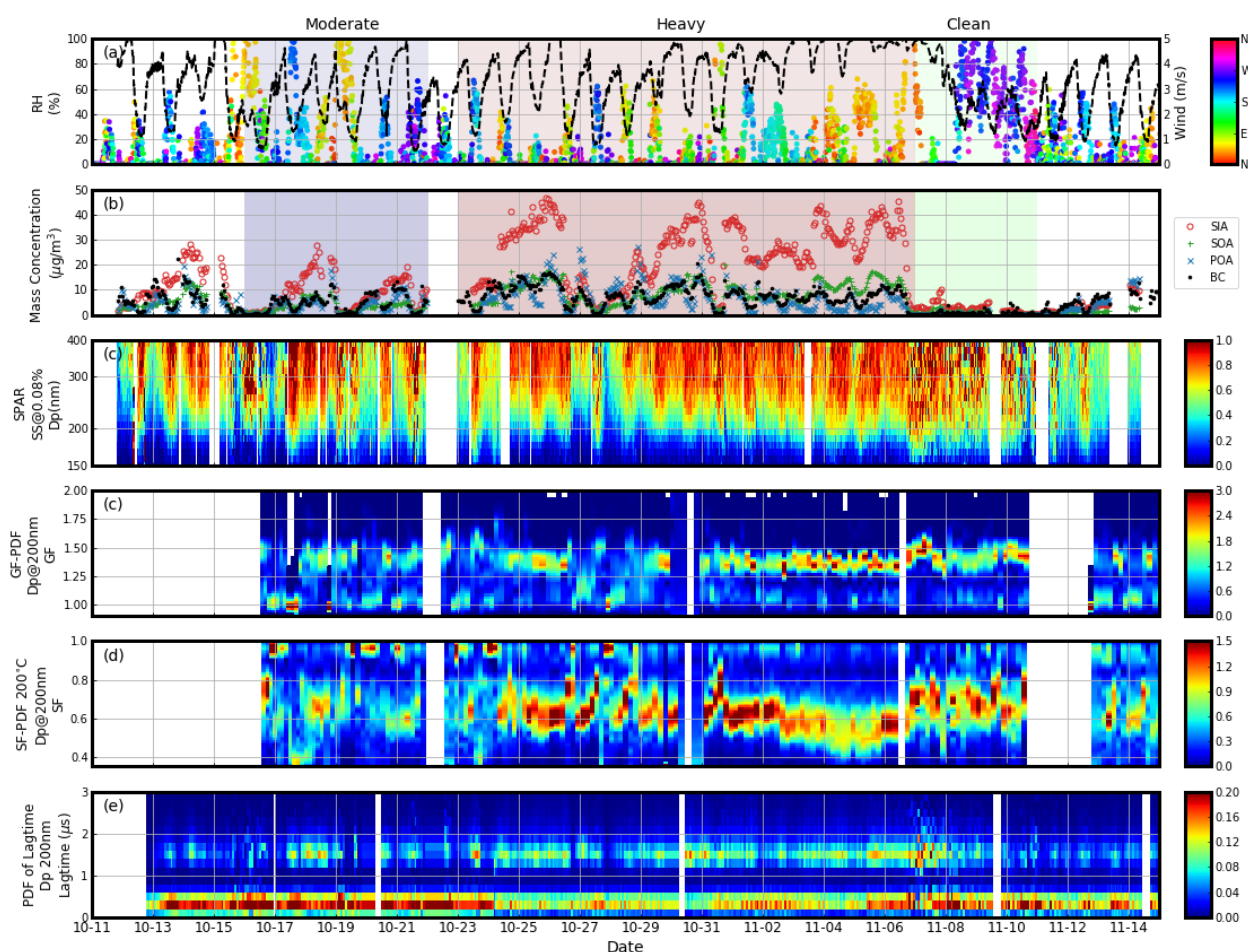


Figure 1. Overview of the measurements during the campaign: (a) meteorological parameters: wind speed (dots) and RH (black line), with colors of dots representing wind direction; (b) mass concentrations of aerosol chemical compositions: secondary inorganic aerosols (SIA, red circle), secondary organic aerosols (SOA, green plus), primary organic aerosols (POA, blue x) and black carbon (BC, black dots); (c) Size-resolved Particle Activation Ratio (SPAR) under supersaturation (SS) of 0.08% observed by DMA-CCN, with warmer colors corresponding to higher value; (d) PDF of growth factor (GFPDF) at 200 nm observed by HTDMA; (e) PDF of shrinkage factor (SFPDF) at 200 nm and 200 °C observed by VTDMA; (f) PDF of lag time at 200 nm observed by DMA-SP2. The blue, red and green shaded periods represent the three periods with moderate pollution, heavy pollution and clean condition, respectively.

359 particles ~~and~~ by DMA-SP2. The mass concentrations of SIA (~~secondary inorganic aerosol~~ Secondary
360 Inorganic Aerosols, include nitrate, sulfate and ammonium), SOA (Secondary Organic Aerosols,
361 include two Oxygenated Organic Aerosol factors: OOA1 ~~and~~ OOA2), PA (~~BBOA~~, POA (Primary
362 Organic Aerosol, include Biomass Burning Organic Aerosol and ~~FFOA~~ Fossil Fuel Organic Aerosol)
363 and BC are shown ~~in Fig. 1~~. During the campaign, three periods with significantly different aerosol
364 pollution conditions were identified. As shown in Fig. 1(b), before October 23rd (moderately polluted
365 period), the accumulation of aerosols led to SIA mass concentrations greater than 20 $\mu\text{g}/\text{m}^3$, while the
366 highest concentrations of SOA, POA and BC mass all reached 10 $\mu\text{g}/\text{m}^3$. The mass concentrations of
367 different ~~chemical compositions~~ aerosol components increased significantly from October 23rd to
368 November 6th (heavily polluted period with average non-refractory PM₁ mass concentration of
369 49.5±22.5 $\mu\text{g}/\text{m}^3$) and decreased to much lower levels after November 6th (clean period) ~~with non-~~
370 refractory PM₁ mass concentration of 5.1±3.3 $\mu\text{g}/\text{m}^3$. Two particle groups were identified with regard
371 to CCN activity, hygroscopicity, volatility and coating thickness, as demonstrated by SPAR, GF-PDF,
372 SF-PDF, and the lag-time PDF of BC-containing particles. Significant variations in aerosol mixing
373 states were also observed during the three periods of different pollution conditions, as demonstrated
374 by the variations in SF-PDF measured by the VTDMA. For example, the SF of the non-volatile particle
375 group decreased during the heavily polluted period. Aerosol mixing states may have changed due to
376 various transformations of existing aerosol particles and distinct secondary formation processes under
377 different pollution conditions (Kuang et al., 2020; Tao et al., 2021; Shi et al., 2022; Yang et al., 2022).
378 This will be analyzed in detail later in the discussion. The diurnal variations in mass concentrations of
379 different aerosol chemical ~~compositions~~ components and mixing states can be seen in the variations of
380 SPAR measurements, as previously observed in this region (Liu et al., 2011; Ma et al., 2012; Kuang
381 et al., 2015; Tao et al., 2020).

382 In Fig. 2, the campaign averaged SPAR at the three SSs, PDF of lag-time of BC containing
383 particles, GFPDF and SFPDF at 200 °C at different particle sizes are presented. The sigmoidal SPAR
384 curves were characterized by a rapid increase followed by a gradual increase to unit 1, similar to
385 measured SPAR curves previously observed in this region (Deng et al., 2011; Zhang et al., 2014; Ma
386 et al., 2016; Tao et al., 2018). At lower SSs, the rapid increases in SPAR curves occur at larger particle
387 sizes ~~and~~, since particle size need for CCN activation is larger. In addition, as SPAR in particle size

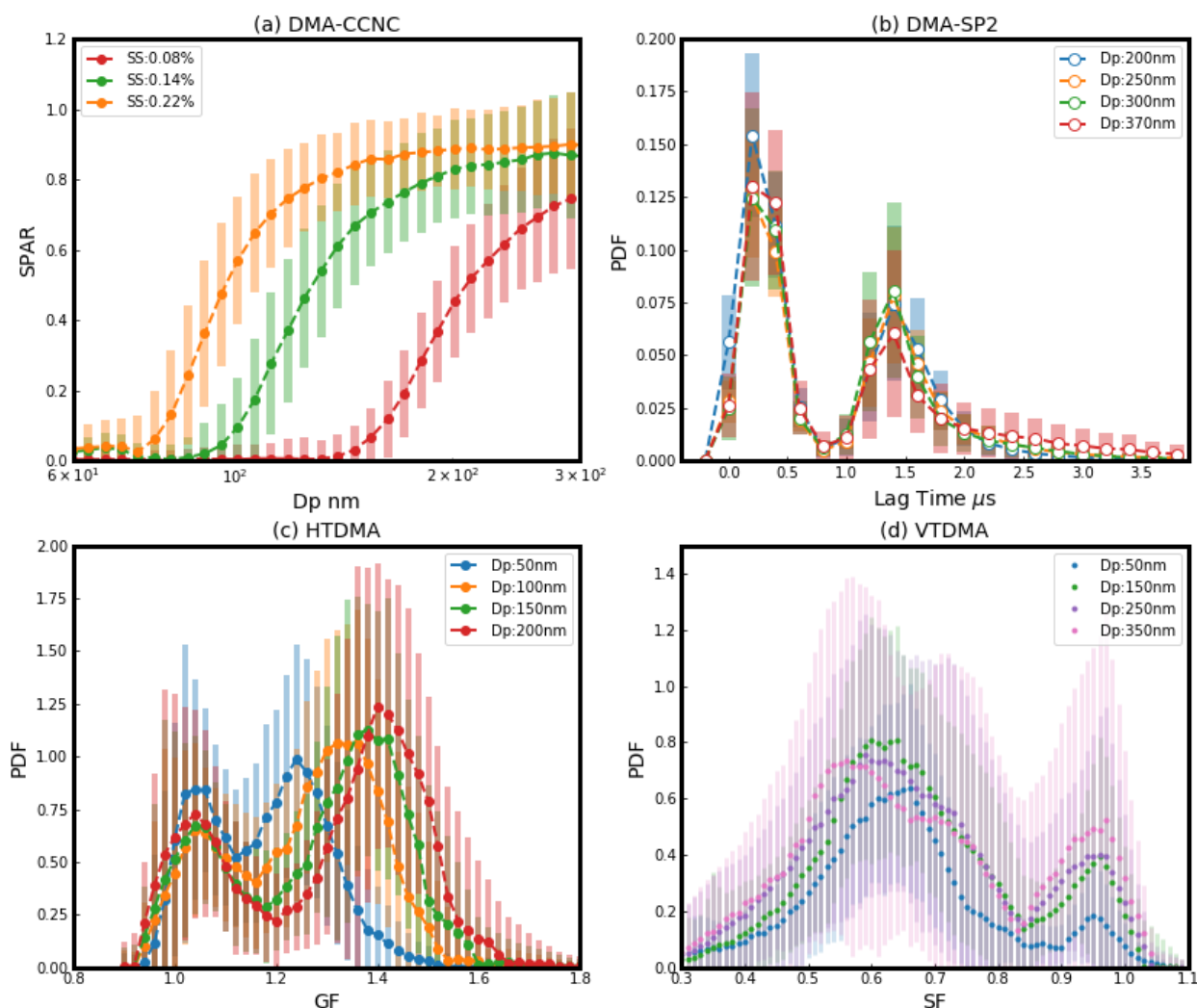


Figure 2. The campaign average of (a) Size-resolved Particle Activation Ratio (SPAR) curves measured by DMA-CCNC at the three supersaturations (SSs, represented by different colors and markers), (b) PDF of lag time measured by DMA-SP2 at four particle sizes (represented by different colors and markers), (c) PDF of growth factor GF (GFPDF) measured by HTDMA at four particle sizes (represented by different colors and markers), (d) PDF of shrinkage factor SF (SFPDF) measured by VTDMA under the temperature of 200 °C at five particle sizes (represented by different colors and markers). The shaded areas indicate the standard deviations.

388 lower than 300 nm is presented, the maximum AR of SPAR curves becomes smaller, as there was less
 389 particle to be CCN active under low SSs. For the three measured SSs, the corresponding fitting
 390 parameters, D_a , which indicate the center of the particle size range with rapid increases in where SPAR
 391 curves, equals about 0.5 are approximately 90 nm, 120 nm and 180 nm, nm for the three SSs of 0.08%,
 392 0.14% and 0.22%, respectively. These particle size agree with the value of the fitting parameter D_a
 393 (midpoint activation diameter, see Eq.7) during the campaign, as the fitting parameter MAF
 394 (Maximum Activation Fraction, an asymptote of the measured SPAR curve at large particle sizes) is

395 [close to 1.](#) The number fraction of CCN-active particles in large ~~diameter ranges~~[diameters range](#)
396 (which varies with SS and, for example, is greater than 200 nm for 0.08%) can be indicated by the
397 gradual increase of SPAR curves and quantified by the fitting parameter, MAF. The PDFs of the lag-
398 time, GF, and SF are all characterized by a bimodal distribution, which indicates two particle groups
399 of BC-containing aerosols with different coating thicknesses, hygroscopicity and volatility. The
400 variations in aerosol mixing ~~state~~[states](#) will be further analyzed on the basis of ~~MAF and the number~~
401 ~~fractions (NF) of hydrophilic particles (NF_H), volatile particles (NF_V), BC free particles (NF_{noBC}) and~~

number fractions of thinly coated BC in BC-containing aerosols (R_{exBC})-measured mixing state parameters.

3.2 Intercomparisons among aerosol mixing state parameters derived using four techniques

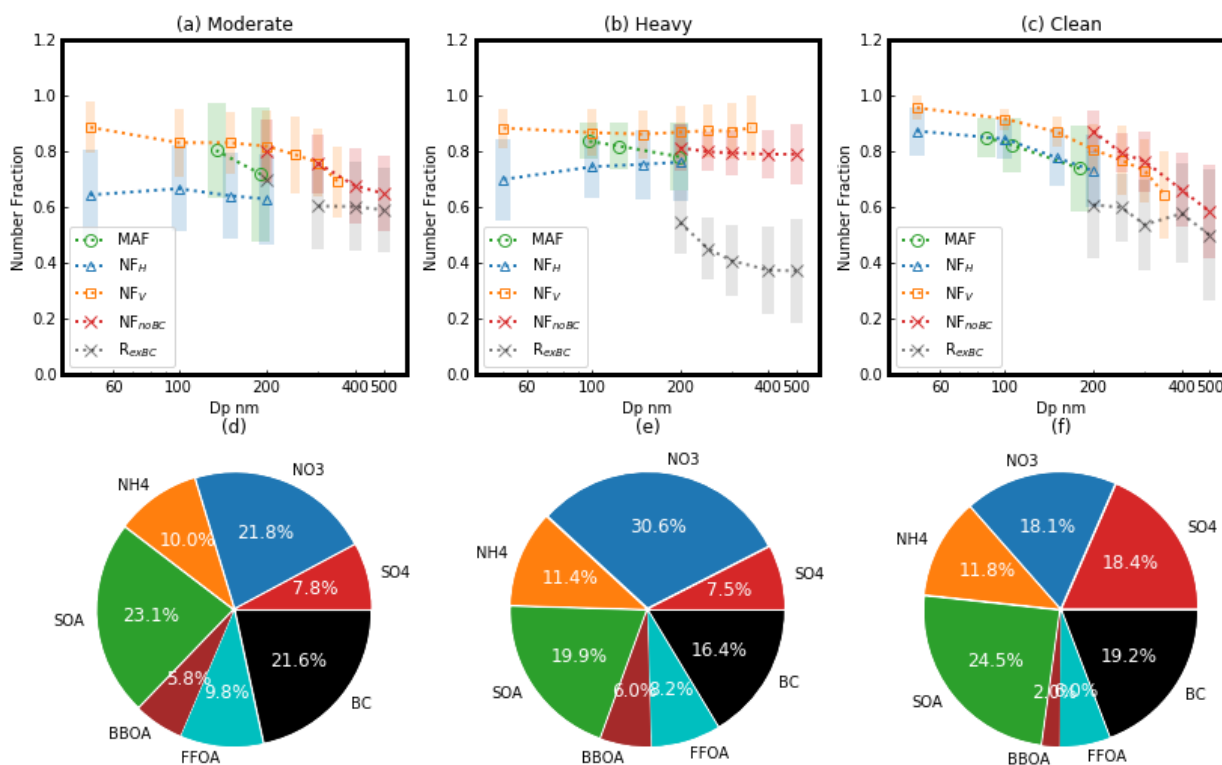


Figure 3. (a-c): Size dependence of MAF (green circle), NF_H (blue triangle), NF_V (yellow square), NF_{noBC} (red x) and R_{exBC} (black plus) during the three periods, MAF (Maximum Activation Fraction): An asymptote of the measured SPAR curve at large particle. NF_H : Number Fraction of Hydrophilic aerosol whose hygroscopicity parameter is higher than ~ 0.07 . NF_V : Number Fraction of Volatile aerosol whose Shrink Factor at $200^\circ C$ is lower than 0.85. NF_{noBC} : Number Fraction of BC-free particles. R_{exBC} : Number fraction of externally BC particles in total BC-containing particles. **(d-f):** Corresponding mass fractions of aerosol chemical compositions (identified by colors) during the three periods, including secondary organic aerosols (SOA), biomass burning organic aerosol (BBOA), fossil fuel organic aerosols (FFOA), and inorganic ions including sulfate (SO_4), nitrate (NO_3) and ammonium (NH_4).

The size dependence of MAF , NF_H , NF_V ($200^\circ C$), NF_{noBC} and R_{exBC} aerosol mixing state parameters derived from measurements of the four techniques as well as the mass fraction of different aerosol chemical compositions during the three pollution periods are shown in Fig 3. In general, the size dependence of MAF, NF_H , NF_V and NF_{noBC} were similar to one another, suggesting they were

410 likely dominated by the same particle group, namely BC-free particles. This particle group had the
411 ~~lowest~~highest fraction (higher than 0.7) during the heavily polluted period and the ~~highest~~lowest
412 fraction (down to 0.5) during the clean period, with the fraction decreasing with particle size. This
413 suggests that primary emissions tend to have higher fractions of BC-containing particles in larger
414 diameter ranges, for example, the fraction of BC-containing particles increases from ~0.1 to ~0.4 as
415 particle size enlarges from 200 nm to 500 nm during the clean period. Since the bulk aerosol mass
416 fraction (MF) is mostly contributed by particles larger than 300 nm, there might have been more
417 hydrophilic, volatile, CCN-active and BC-free particles in larger sizes (> 300 nm) during the heavily
418 polluted period, due to strong secondary aerosol formation in larger diameter ranges (Kuang et al.,
419 2020), resulting in higher number fraction of these particles compared to the clean period. As for R_{exBC} ,
420 the small size dependence of R_{exBC} during the moderately polluted period might have been associated
421 with stronger primary emissions, while the decrease of R_{exBC} with increasing particle diameter size in
422 the polluted period confirms ~~the more efficient~~ secondary aerosol formation into be more efficient on
423 particles with larger diameter ~~ranges~~.

424 As for the difference among the aerosol mixing state parameters, NF_V and NF_{noBC} agreed with
425 each other with a difference smaller than 0.1, and both were higher by at least 0.1 than NF_H in the
426 moderately polluted period. In detail, compared to NF_{noBC} , NF_V was higher during the heavily polluted
427 period, when the nitrate fraction was the highest (~30%) and the SOA fraction was the lowest (~7%)
428 among all three periods, suggesting some BC-containing aerosols in this period were also identified as
429 volatile, which is consistent with the fact that formation of semi-volatile nitrate in the BC-containing
430 particles increases their volatility. However, during the clean period, NF_V was even lower than NF_{noBC} ,
431 suggesting that some BC-free aerosols were characterized as low volatile and non-negligible fractions
432 of BC-free aerosols dominated ~~within~~ these less volatile aerosol components, which were likely less
433 volatile organic aerosols. In addition, the MAF values generally agreed with NF_H during the clean
434 period, but were larger than NF_H during the moderately and heavily polluted periods (by ~0.2) when
435 the POA/SOA fractions were higher- (~40% vs ~35%). POA generally had lower hygroscopicity than
436 SOA. The critical κ of hydrophilic mode aerosols was 0.07, suggesting that a higher fraction of
437 aerosols had κ below 0.07 (i.e. hydrophobic mode aerosol in this study) during the moderately polluted
438 period. However, under supersaturated conditions, they demonstrated enhanced hygroscopicity by

439 becoming CCN-active. The NF_H was consistently lower than NF_V and NF_{noBC} (the average difference
 440 between NF_H and NF_{noBC} was about 0.2), especially). As mentioned above that NF_H was also lower
 441 than MAF during the moderately polluted period, suggesting periods, there may be a significant
 442 fraction of volatile BC-free aerosols had hygroscopicity lower than critical κ of 0.07 but were still
 443 CCN-active and therefore not fully hydrophobic.

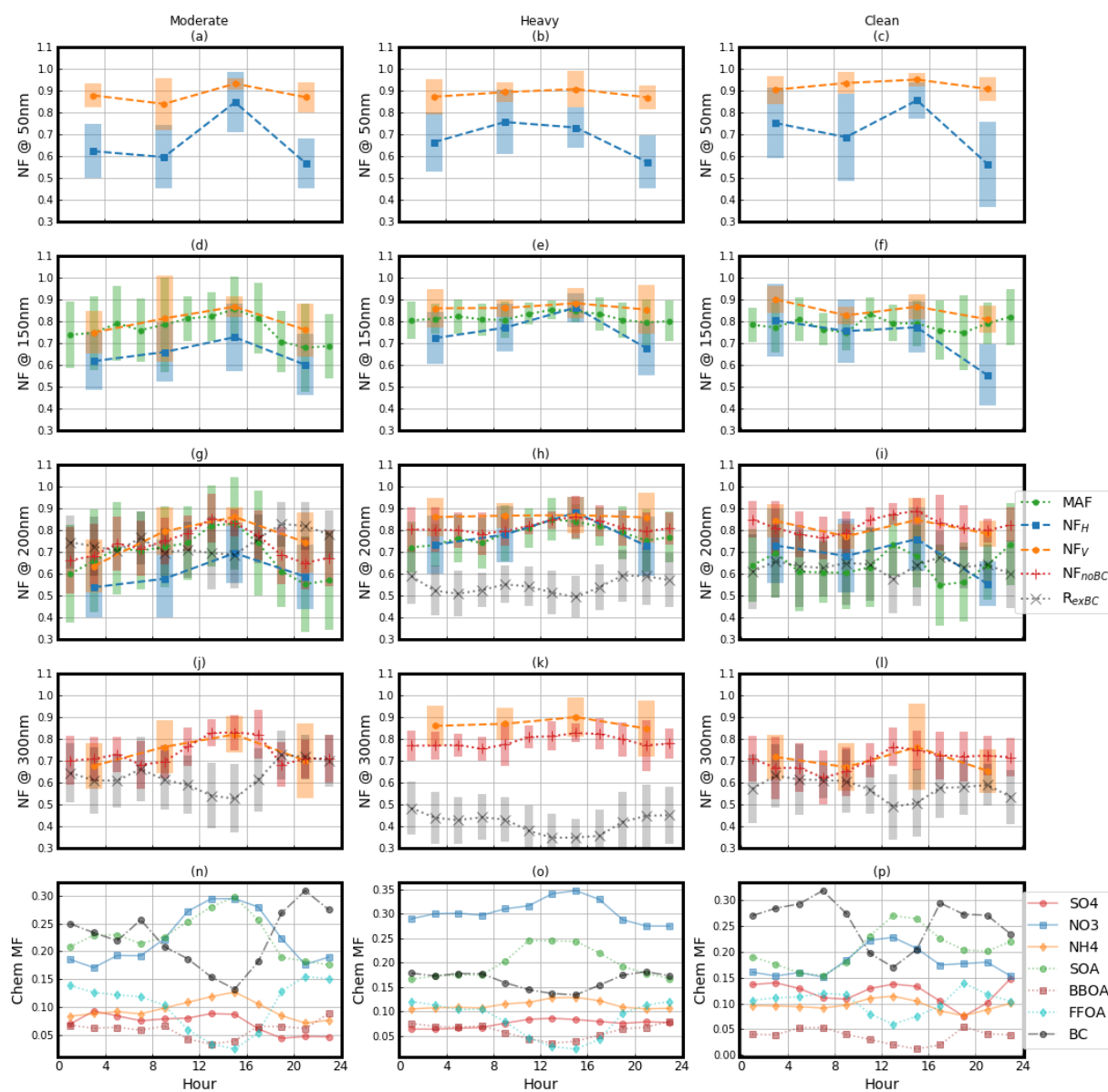


Figure 4. (a-l) Diurnal variations of aerosol mixing state parameters (identified by color and marker) at different particle sizes (50, 150, 200 and 300 nm) during the three periods. The shaded areas indicate the standard deviations. **MAF** (Maximum Activation Fraction): An asymptote of the measured SPAR curve at large particle. **NF_H** : Number Fraction of Hydrophilic aerosol whose hygroscopicity parameter is higher than ~ 0.07 . **NF_V** : Number Fraction of Volatile aerosol whose Shrink Factor at 200 °C is lower than 0.85. **NF_{noBC}** : Number Fraction of BC-free particles. **R_{exBC}** : Number fraction of externally BC particles in total BC-containing particles. **(m-o)** Diurnal variations of mass fractions of aerosol chemical compositions including secondary organic aerosols (SOA), biomass burning organic aerosol (BBOA), fossil fuel organic aerosols (FFOA), and inorganic ions including sulfate (SO₄), nitrate (NO₃) and ammonium (NH₄) (identified by color and marker) during the three periods.

444 The diurnal variations of MAF, NF_H , NF_V and NF_{noBC} along with the mass fractions of aerosol
445 chemical ~~compositions~~components during the three periods, are shown in Fig.4. With the exception of
446 particle size 50 nm, the diurnal variations of these four mixing state parameters were generally similar
447 for all measured sizes. The different diurnal variations at particle size 50 nm may be due to the different
448 effects of emissions and aging processes on different aerosol modes, as particles smaller than 100nm
449 were mainly in a different aerosol mode (Aitken mode) to particles larger than 100nm (Wang et al.,
450 2022). For particle sizes larger than 100 nm, (shown in both Fig. 4 and Fig. S5), there were
451 ~~peaks~~maxima in the afternoon for MAF, NF_H , NF_V and NF_{noBC} , indicative of a peak during this time
452 due to the increase in secondary aerosol compositions like nitrate and SOA, and the decrease of POA
453 and BC. The diurnal variations of aerosol mixing state parameters and aerosol chemical compositions
454 were more pronounced in the moderately polluted period. During the heavily polluted period, the
455 diurnal variation
456 was least pronounced for NF_V and most pronounced for NF_H . In the clean period, there was another
457 ~~peak~~maxima at midnight for MAF and NF_{noBC} , which may be attributed to the diurnal variations of
458 secondary aerosol compositions like sulfate and SOA, and the decrease of BC and FFOA. The average
459 size-dependence of the aerosol mixing state parameters in different time ranges during the heavily
460 polluted period is shown in Fig. ~~S3~~S6. It can be seen that the differences among the four parameters
461 were the least from 12:00 to 18:00, with the most SOA and the least POA. This is consistent with the
462 results in Fig.3, where the difference between MAF and NF_H becomes smaller when POA fractions
463 are the smallest. R_{exBC} tended to be lower during the daytime and its diurnal variation was more
464 significant in larger particle sizes. In general, the diurnal variations for R_{exBC} are opposite to those of
465 NF_{noBC} and agree better with those of the primary aerosol mass fractions. This is because BC particles
466 originate from primary emissions and are mainly externally mixed. After experiencing aging process
467 in the atmosphere, BC particles can be coated by secondary aerosol formed on, resulting in more coated
468 BC particles and less externally mixed BC particles. As the secondary aerosol tends to form on larger
469 particles, the diurnal variations of secondary aerosol formations may affect more significantly on those
470 of mixing state of BC particles and thus R_{exBC} in larger particle sizes.

471 The intercomparisons among MAF, NF_H , NF_V and NF_{noBC} were conducted based on their
472 correlations at different particle sizes as summarized in Table. S1. ~~It's worth noting~~Note that MAF at

473 SSs of 0.08%, 0.14% and 0.22% was used for comparison at particle sizes of 200 nm, 150 nm and
474 100 nm, respectively. ~~In general, there were moderate correlations.~~ This is because that the diameter
475 range of rapid increases in SPAR curves are determined by aerosol hygroscopicity in this particle size
476 range, and the midpoint of rapid increase diameter ranges of SPAR curves at SSs of 0.08%, 0.14% and
477 0.22% are approximately 180 nm, 120 nm and 90 nm (as shown in Fig. 2). In general, there were
478 moderate correlations ($r \sim 0.5$) among MAF, NF_H and NF_V , suggesting the dominance of CCN-active,
479 hygroscopic and volatile aerosols are contributed by a similar particle group (Zhang et al., 2016). The
480 ~~consistency of agreement between~~ MAF- NF_V was slightly higher than that of MAF- NF_H or NF_H - NF_V
481 with similar correlation coefficients (~ 0.65) but smaller systematic differences (slope and intercept
482 were much closer to 1 and 0, respectively), which is consistent with the previous finding that a
483 substantial fraction of volatile but less hygroscopic aerosols is CCN-active. At smaller particle size,
484 the correlation became weaker, ($r \sim 0.4$), while the degree of the reduction was the least for the
485 correlation between MAF and NF_V .

486

487 3.3 Impacts of primary aerosol emissions on aerosol mixing states and parameter 488 intercomparisons

489 In Fig. 5, the correlation between each aerosol mixing state parameter at 200 nm and the mass
490 fraction of each primary organic aerosol composition during the three periods is presented. In general,
491 these four mixing state parameters (MAF, NF_H , NF_V and NF_{noBC}) were negatively correlated with
492 MF_{FFOA} and MF_{BBOA} . However, the correlation with MF_{FFOA} ($-0.45 \sim -0.74$) was much weaker
493 compared to MF_{BBOA} ($-0.10 \sim -0.45$). Biomass burning emissions and fossil fuel emissions are two
494 major sources of BC in the NCP (Yang et al., 2022), and NF_{noBC} was negatively correlated with MF_{FFOA}
495 ($r = -0.49$) and weakly correlated ($r = -0.18$) with MF_{BBOA} , suggesting that fossil fuel emission were the
496 more likely source of BC during this field campaign. The negative correlation between MAF and
497 MF_{FFOA} was even weaker than that of NF_{noBC} with MF_{FFOA} (-0.62 vs. -0.49). Especially, at the same
498 MF_{FFOA} , MAF was lower than NF_{noBC} , demonstrating that some BC-free particles were CCN-inactive,
499 and likely mainly composed of organic aerosols from fossil fuel combustion emission. The negative
500 correlation between NF_V and MF_{FFOA} was slightly weaker than that between NF_{noBC} and MF_{FFOA} ($-$
501 0.56 vs -0.49). At the same MF_{FFOA} , NF_{noBC} was close to NF_V , and considering that BC-containing

502 aerosols were dominated by thinly coated BC during most times (as shown in Fig.5), this demonstrates
 503 that the non-volatile population identified by the V-TDMA was mainly contributed by BC-containing
 504 aerosols. NF_H had the lowest negative correlation with MF_{FFOA} ($r=-0.74$), demonstrating the significant
 505 contributions from fossil fuel emissions to nearly hydrophobic aerosol populations. At the same

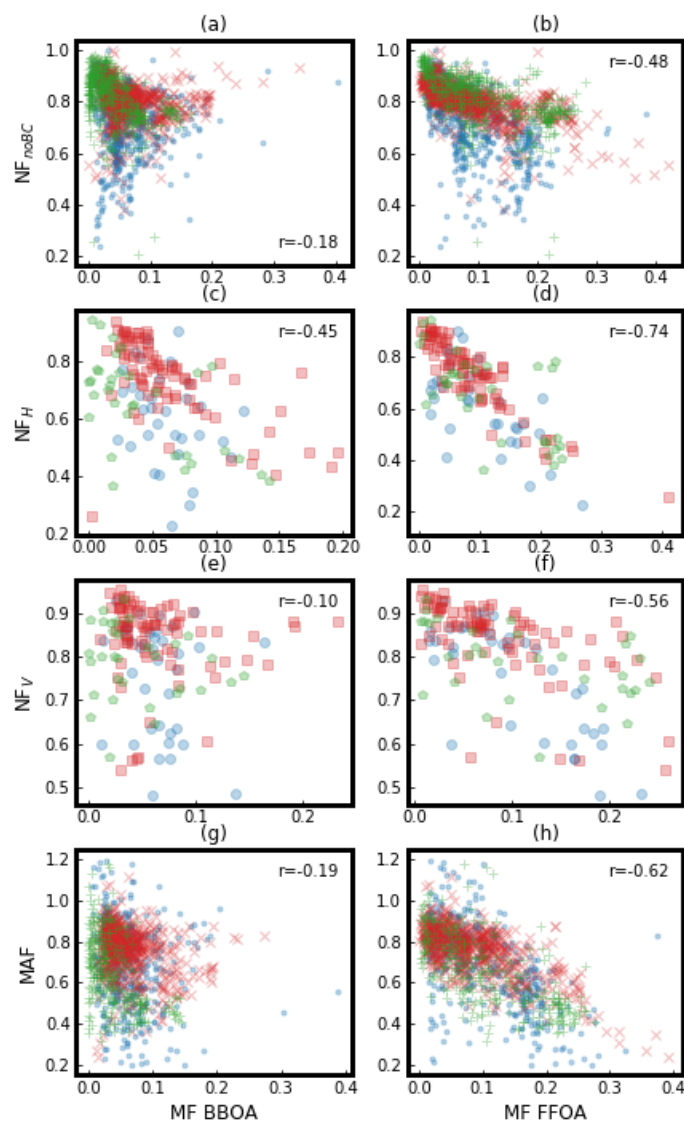


Figure 5. The correlations between aerosol mixing state parameters and mass fractions (MF) of BBOA (Biomass Burning Organic Aerosol) and FFOA (Fossil Fuel Organic Aerosol) during different periods (Moderately polluted period: Blue dot or circle; Heavily polluted period: Red x or square; Clean period: Green plus or pentagon.), with r representing correlation coefficient. **MAF** (Maximum Activation Fraction): An asymptote of the measured SPAR curve at large particle. **NF_H** : Number Fraction of Hydrophilic aerosol whose hygroscopicity parameter is higher than ~ 0.07 . **NF_V** : Number Fraction of Volatile aerosol whose Shrink Factor at $200\text{ }^\circ\text{C}$ is lower than 0.85 . **NF_{noBC}** : Number Fraction of BC-free particles.

506 MF_{FFOA} , the NF_H was obviously lower than NF_{noBC} (NF_H and NF_{noBC} are larger and smaller than 0.7

507 when MF_{FFOA} was larger than 0.1, respectively), demonstrating that a substantial portion of nearly
 508 hydrophobic aerosols was not contributed by BC-containing aerosols, but likely by FFOA or BBOA
 509 dominant aerosols (NF_H also had a negative correlation with MF_{BBOA}). However, the markedly
 510 different correlations between MAF with MF_{FFOA} ($r=-0.62$) and between MAF with MF_{BBOA} ($r=-0.2$)
 511 imply that those nearly hydrophobic but CCN-active aerosols were likely contributed by biomass
 512 burning emissions. The correlations between the ratio of thinly coated BC in total BC-containing
 513 particles (R_{exBC}) and the mass fractions of BBOA and FFOA are shown in Fig.6, and weak correlations
 514 ($r<0.3$) between R_{exBC} with MF_{BBOA} and MF_{FFOA} are observed. However, R_{exBC} tended to increase with
 515 MF_{FFOA} , suggesting that BC containing aerosols emitted from fossil fuel combustion tended to be more
 516 externally mixed with other aerosol components than those emitted from biomass burning activities.

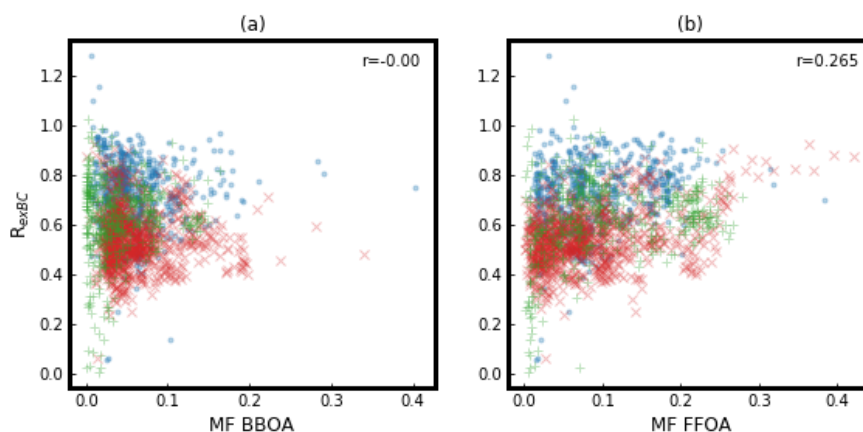


Figure 6. The correlations between the ratio of external mixed BC in total BC particles (R_{exBC}) and mass fractions (MF) of BBOA (Biomass Burning Organic Aerosol) and FFOA (Fossil Fuel Organic Aerosol) during different periods (Moderately polluted period: Blue dot; Heavily polluted period: Red x; Clean period: Green plus), with r representing correlation coefficient.

517 These results demonstrate remarkably different mixing states as well as physical and chemical
 518 properties of fossil fuel combustion aerosols and biomass burning aerosols.

519 The impact of primary emissions on the differences among the four aerosol mixing state
 520 parameters at particle size of 200 nm is analyzed and shown in Fig. 7. The difference between NF_{noBC}
 521 and NF_H and NF_V ($NF_{noBC}-NF_H$ both) had strong significant positive correlations with MF_{FFOA} and
 522 MF_{BBOA} ($r>0.5$), suggesting that a substantial proportion of POA resided in BC-free aerosols and was
 523 volatile but contributed substantially to nearly hydrophobic aerosols. So did the difference between
 524 NF_V and NF_H (NF_V-NF_H). The mass fractions of BBOA and FFOA were poorly linked with the

525 difference between MAF- and NF_V ($MAF-NF_V$), or MAF- and NF_{noBC} or NF_V ($MAF-NF_{noBC}$), or NF_V
 526 and NF_{noBC} (NF_V-NF_{noBC}) (Fig. S5S7). The difference between MAF- NF_H had a positive correlation
 527 with MF_{BBOA} , further suggesting BBOA contributed to nearly hydrophobic aerosols under subsaturated

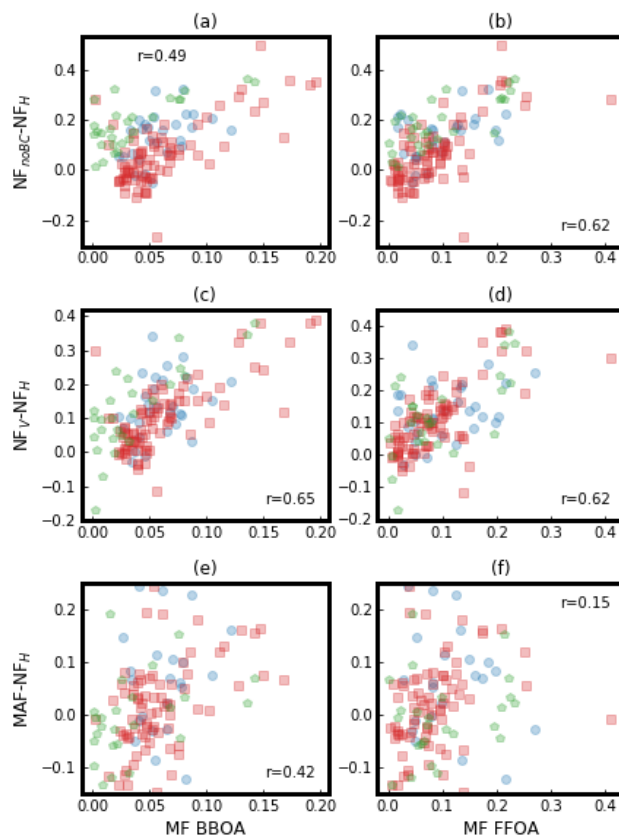


Figure 7. The correlations between the difference among the four aerosol mixing state parameters at particle size of 200 nm and mass fractions (MF) of **BBOA** (Biomass Burning Organic Aerosol) and **FFOA** (Fossil Fuel Organic Aerosol) during different periods (Moderately polluted period: Blue circle; Heavily polluted period: Red square; Clean period: Green pentagon), with r representing correlation coefficient. NF_A-NF_B ($NF_{noBC}-NF_H$, NF_V-NF_H , $MAF-NF_H$): The difference between the number fraction of A and B. **MAF** (Maximum Activation Fraction): An asymptote of the measured SPAR curve at large particle. **NF_H** : Number Fraction of Hydrophilic aerosol whose hygroscopicity parameter is higher than ~ 0.07 . **NF_V** : Number Fraction of Volatile aerosol whose Shrink Factor at 200 °C is lower than 0.85. **NF_{noBC}** : Number Fraction of BC-free particles.

528 conditions, however, their hygroscopicity was enhanced and became CCN-active at supersaturated
 529 conditions. In addition, the correlation between mixing state parameters and primary aerosol
 530 composition during the campaign and different pollution periods were summarized in Fig. S7.

531

532 3.4 Impacts of secondary aerosol formations on aerosol mixing states and parameter 533 intercomparisons

534 The correlations between each aerosol mixing state ~~parameter~~parameters at 200 nm and the mass
535 fraction of each secondary aerosol (SA) component ~~are~~is presented in Fig. 8 for three periods and the
536 entire campaign are presented. To be noted, in order to compare the four aerosol mixing state

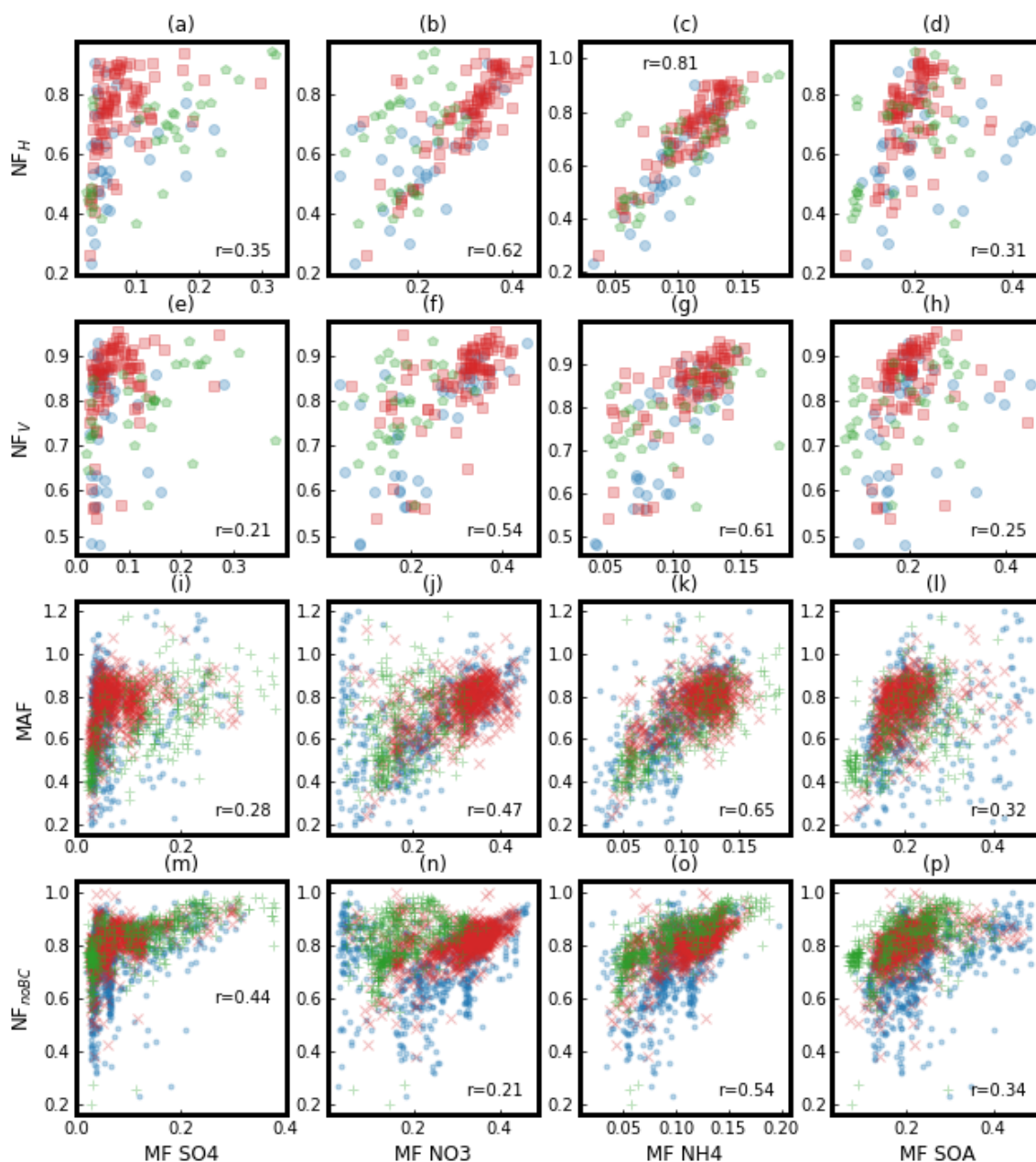


Figure 8. The correlation between the four aerosol mixing state parameters and mass fraction (MF) of secondary aerosol components during different periods (Moderately polluted period: Blue dot or circle; Heavily polluted period: Red x or square; Clean period: Green plus or pentagon.), with r representing correlation coefficient. **MAF** (Maximum Activation Fraction): An asymptote of the measured SPAR curve at large particle. **NF_H**: Number Fraction of Hydrophilic aerosol whose hygroscopicity parameter is higher than ~ 0.07 . **NF_V**: Number Fraction of Volatile aerosol whose Shrink Factor at 200 °C is lower than 0.85. **NF_{noBC}**: Number Fraction of BC-free particles. Secondary aerosol components including secondary organic aerosols (SOA), sulfate (SO₄), nitrate (NO₃) and ammonium (NH₄).

537 parameters as well as their relationships with aerosol chemical compositions at the same time, the
 538 analysis is conducted at only 200 nm where all the four aerosol mixing state parameters were measured.

539 Generally, MAF, NF_H , NF_V and NF_{noBC} had a strong positive correlation with ammonium- MF_{NH_4}
540 ($r > 0.5$). This is likely due to the fact that ammonium was mainly formed through the neutralization of
541 sulfuric and nitric acid with ammonia, so variations in ammonium better represent overall secondary
542 inorganic aerosol formation. As shown in Fig.3, the secondary inorganic aerosol components
543 dominated over SA (about 50% vs about 70%), indicating that secondary aerosol formations were
544 primarily composed of secondary inorganic aerosol formation, which explains the weaker correlations
545 with SOA ($r \sim 0.3$) seen in Fig. 8.

546 During the clean period when the mass fraction of SOA and sulfate were both above 15%, all four
547 parameters had a strong positive correlation with MF_{SO_4} and MF_{SOA} ($r > 0.5$), suggesting that when
548 clean background air mass with higher fractions of sulfate and SOA prevailed, local primary emissions
549 that contributed substantially to BC-containing and less hygroscopic POA aerosols became less
550 significant. The positive correlations between MAF and secondary aerosol components have been
551 extensively discussed by Tao et al. (2021), who found that secondary aerosol formations enhance the
552 hygroscopicity of nearly hydrophobic aerosols, thereby increasing CCN activity. This also explains
553 the highest correlations of NF_H or MAF with ammonium formation. The strong positive correlations
554 between NF_V and secondary aerosol formations ($r \sim 0.6$) are consistent with the fact that nitrate
555 dominates secondary aerosol formations during this campaign and nitrate is semi-volatile. For the first
556 time, the strong positive correlations between NF_{noBC} and secondary aerosol formations were revealed-
557 ($r \sim 0.6$). This is because NF_{noBC} primarily depends on the relative variations of BC-containing and BC-
558 free aerosols. The increase in NF_{noBC} at 200 nm as a function of secondary aerosol mass fraction
559 suggests that secondary aerosol formations migrated a higher fraction of BC-free aerosols smaller than
560 200 nm to 200 nm, highlighting that secondary aerosols tended to form more quickly on BC-free
561 aerosols than on BC-containing aerosols.

562 The effects of secondary aerosol formation on the differences between the four aerosol mixing
563 state parameters were studied and illustrated in Fig. 9. To be noted, two OOA factors (OOA1 and
564 OOA2) resulted from different chemical processing are identified. Differences between NF_{noBC} - and
565 NF_H and NF_V ($NF_{noBC} - NF_H$ both) showed a strong negative correlation with MF_{NH_4} and MF_{NO_3} - (mainly
566 -0.6). So did the difference between NF_V and NF_H ($NF_V - NF_H$). As previously noted, NF_H typically

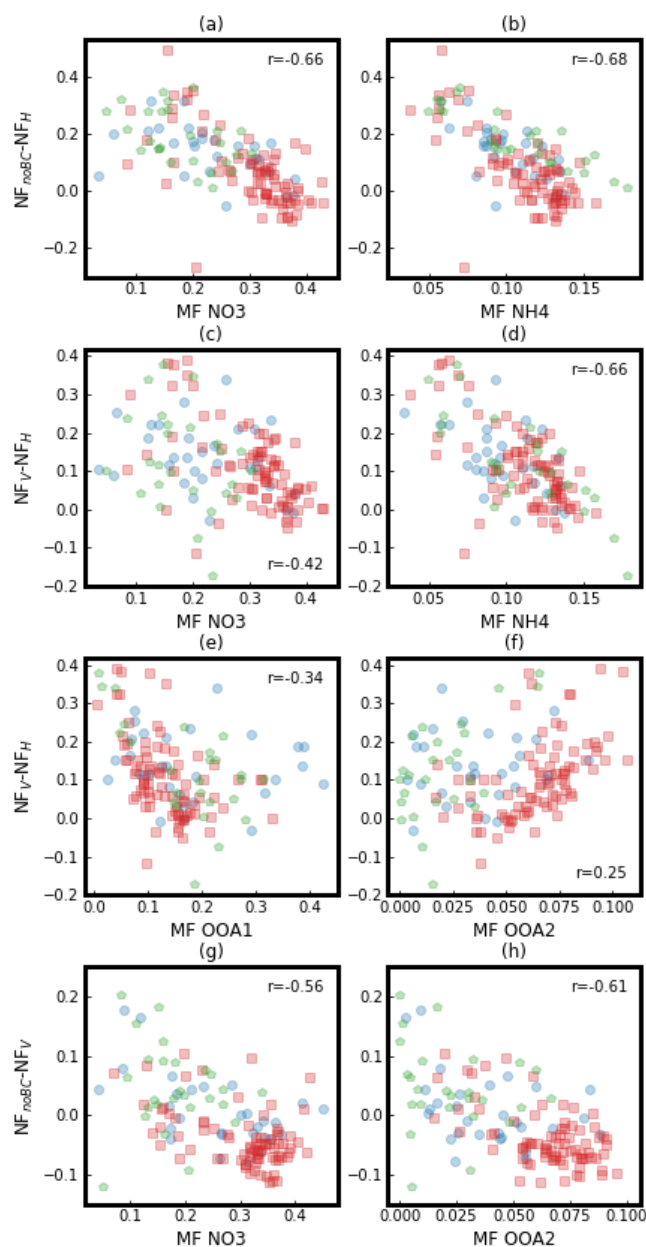


Figure 9. The correlation between the difference among the four aerosol mixing state parameters and mass fractions (MF) of secondary aerosol chemical components during different periods. OOA1 and OOA2 are two SOA factors resolved from AMS measurements using the PMF technique. Moderately polluted period: Blue circle; Heavily polluted period: Red square; Clean period: Green pentagon. $NF_A - NF_B$ ($NF_{noBC} - NF_H$, $NF_V - NF_H$, $NF_{noBC} - NF_V$): The difference between the number fraction of A and B. NF_H : Number Fraction of Hydrophilic aerosol whose hygroscopicity parameter is higher than ~ 0.07 . NF_V : Number Fraction of Volatile aerosol whose Shrink Factor at $200\text{ }^\circ\text{C}$ is lower than 0.85. NF_{noBC} : Number Fraction of BC-free particles.

567 had smaller $_$ values than NF_V and NF_{noBC} . Thus, a negative correlation with the mass fraction of
 568 ammonium and nitrate indicates that the formation of secondary nitrate results in a smaller difference
 569 between these mixing state parameters. The increase in the fraction of ammonium nitrate, a pure
 570 scattering semi-volatile compound with strong hygroscopicity, can render the aerosol population more

571 dominated by particles with strong volatility and hygroscopicity. This can result in a smaller difference
572 between NF_{noBC} , NF_H , and NF_V as the secondary inorganic aerosol components increases. Furthermore,
573 the difference between $NF_V - NF_H$ showed a positive correlation with MF_{OOA2} and a negative correlation
574 with MF_{OOA1} , indicating different volatility and hygroscopicity of the two SOA factors. The variations
575 in the difference between NF_V and NF_H with the mass fraction of OOA1 and OOA2 are shown in Fig.
576 9(e) and (f). As previously noted, NF_V was generally higher than NF_H , and the difference between the
577 two decreases with an increase in MF_{OOA1} , which were generally smaller than 0.3. This suggests that
578 the formation of OOA1 enhances the hygroscopicity of volatile particles, which aligns with the highest
579 oxidation state of OOA1 (higher O/C but lower H/C compared to OOA2), which had significant and
580 overall positive impact on aerosol hygroscopicity (Cerully et al., 2015; Thalman et al., 2017; Zhang et
581 al., 2023). A positive correlation is seen between NF_V and MF_{OOA2} ($r \sim 0.25$), whereas the correlation
582 between NF_H and MF_{OOA2} is weak (R is close to 0), implying that OOA2 might be semi-volatile but
583 only weakly hygroscopic, which could contribute to NF_V being higher than NF_H as OOA1 increases.
584 The difference between NF_{noBC} and NF_V ($NF_{noBC} - NF_V$) is negatively correlated with MF_{NO3} , which is
585 consistent with the semi-volatile nature of nitrate. The negative correlation between $NF_{noBC} - NF_V$ and
586 MF_{OOA2} indicates that the difference is smaller when there is more OOA2, implying that OOA2 are
587 also semi-volatile compounds and were likely formed mainly on BC-free particles. The correlations
588 between the difference between $NF_V - MAF$ and $NF_{noBC} - MAF$ and the mass fraction of each secondary
589 aerosol composition are much weak. The impacts of secondary aerosol formation on BC mixing states
590 are depicted in Fig. S4S8. In general, the number fraction of thinly coated BC has a negative correlation
591 with SIA and a weak association with SOA, suggesting that SIA formation mainly enhances the
592 thickness of BC coating. [Besides, the correlation between mixing state parameters and secondary](#)
593 [aerosol composition during the campaign and different pollution periods were summarized in Fig. S9.](#)

594 In addition to the changes in the mass fractions of secondary aerosol compositions, the
595 accumulation of pollution due to secondary aerosols may also provide insight into the impact of
596 secondary aerosol formations on aerosol mixing states. As shown in Fig. 10 (a), during the heavily
597 polluted periods, there were two distinct pollution accumulation processes from Oct 23rd to Oct 27th

598 and from Oct 28rd to Oct 31st, respectively. During the pollution accumulation process, the mass
 599 concentration of secondary aerosols increased by approximately three times, indicating rapid
 600 formation of secondary compositions and causing a significant rise in **PM_{non-refractory} PM₁** (NR-

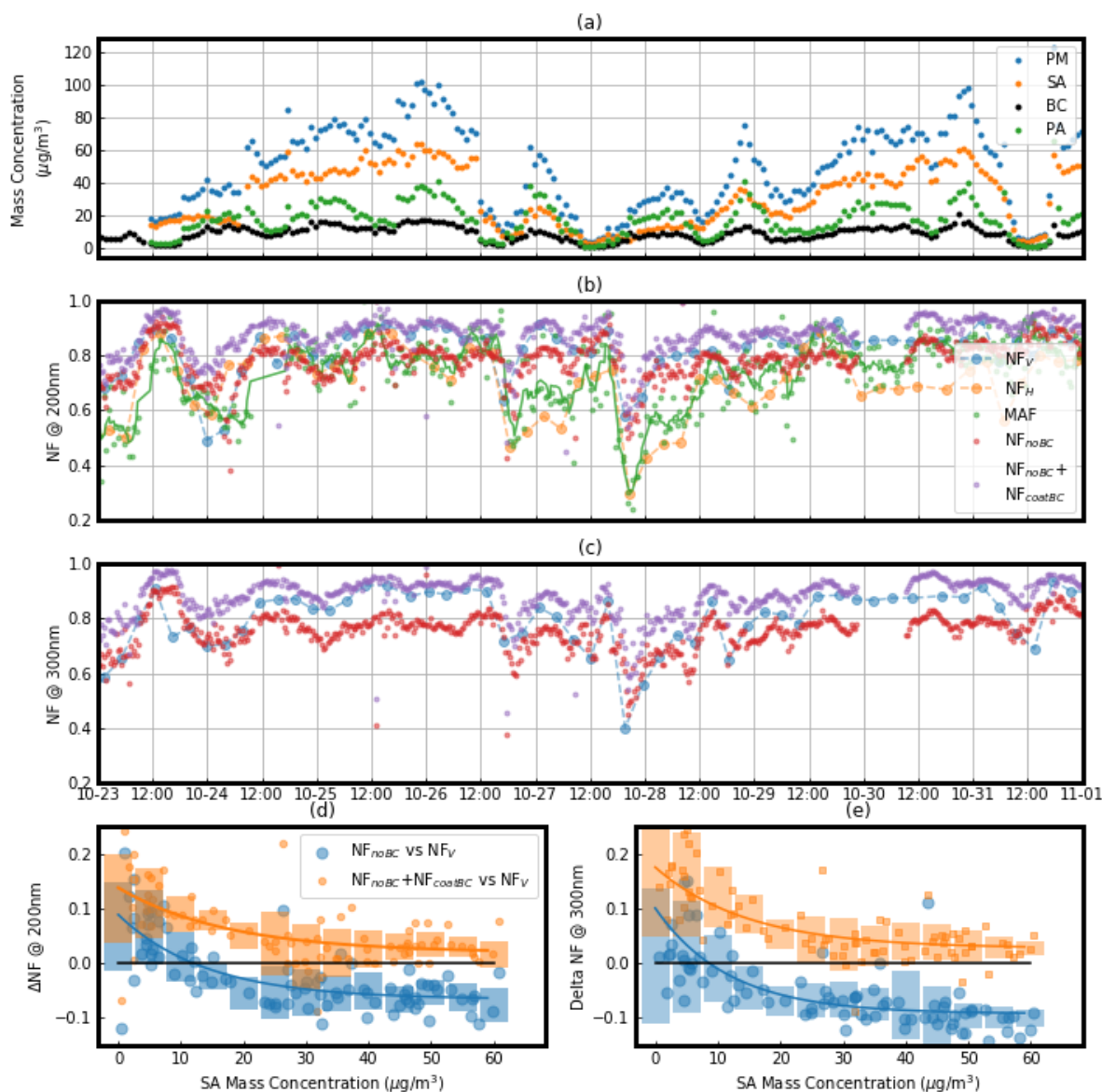


Figure 10. The variations of different aerosol mixing state parameters during the pollution accumulation process. **(a)** The time series of mass concentrations of non-refractory PM_{10} (NR- PM_{10}), secondary aerosols (SA, including inorganic ions and SOA), primary organic aerosols (POA) and BC (identified by colors and markers). **(b and c)** The variations of different aerosol mixing state parameters (identified by colors and markers) at particle size of 200nm **(b)** and 300nm **(c)**; **(d and e)** The variations of the difference between NF_V and NF_{noBC} ($\text{NF}_V - \text{NF}_{noBC}$, blue large circle) and the difference between NF_V and $\text{NF}_{noBC} + \text{NF}_{coatedBC}$ ($\text{NF}_V - (\text{NF}_{noBC} + \text{NF}_{coatedBC})$, yellow small circle) with the mass concentration of SA at particle size of 200nm **(d)** and 300nm **(e)**. **MAF** (Maximum Activation Fraction): An asymptote of the measured SPAR curve at large particle. **NF_H** : Number Fraction of Hydrophilic aerosol whose hygroscopicity parameter is higher than ~ 0.07 . **NF_V** : Number Fraction of Volatile aerosol whose Shrink Factor at 200 °C is lower than 0.85. **$\text{NF}_{coatedBC}$** : Number Fraction of thickly coated BC particles.

601 [PM₁](#)) mass concentration. Fig. 10 (b and c) illustrates that this increase in secondary aerosols led to a
602 significant enhancement of aerosol mixing state parameters, including MAF, NF_V, NF_H and NF_{noBC},
603 which rose from about 0.5 to about 0.8 with evident diurnal variations. This highlights the impact of
604 secondary aerosol formations on aerosol mixing states and the importance of studying the pollution
605 accumulation processes of secondary aerosols. The enhancements of different aerosol mixing state
606 parameters during the pollution accumulation process were not uniform. MAF and NF_H initially
607 showed lower values compared to NF_V and NF_{noBC}, but their later enhancement was stronger than that
608 of NF_{noBC}. Fig. 10 (d) and (e) show the difference between NF_{noBC} and NF_V at 200 nm and 300 nm as
609 a function of secondary aerosol mass concentrations during these two pollution periods, which clearly
610 displays how during secondary aerosol formations NF_V became higher than NF_{noBC} while NF_V kept
611 close to ~~NF_{noBC}~~ plus the number fraction of thickly coated BC containing aerosols: [\(NF_{CBC}\) plus](#)
612 [NF_{noBC} \(NF_{CBC}+NF_{CBC}\)](#). The results suggest that secondary aerosol formation increases the volatility
613 of BC-free and BC-containing aerosols, leading increased NF_V compared to NF_{noBC}. And almost all
614 BC-free particles and some BC-containing aerosols became volatile during the pollution accumulation
615 process.

616

617 4. Conclusions

618 The aerosol mixing state is one of the most important physicochemical properties of aerosol
619 particles, which has significant impacts on the optical properties and CCN activity of aerosol particles.
620 Aerosol mixing ~~state varies~~[states vary](#) largely under complex aerosol emissions and atmospheric
621 transformations. In this study, aerosol mixing states derived from CCN activity, hygroscopicity,
622 volatility and BC particles observations, along with their relationship to primary aerosols emissions
623 and secondary aerosol formations, were systematically analyzed based on simultaneous measurements
624 of a CCNC, a H/VTDMA and a SP2. Statistical analysis demonstrated that the number fraction of
625 CCN-active, hygroscopic and volatile particles ~~were~~[was](#) generally positively correlated to one another
626 and were contributed mainly by BC-free aerosols. Therefore, [the four mixing state parameters \(NF_{noBC}:
627 \[Number Fraction of BC-free particles\]\(#\), MAF; \[Maximum Activation Fraction\]\(#\). NF_H; ~~and~~: \[Number
628 \\[Fraction of Hydrophilic aerosol whose hygroscopicity parameter is higher than ~0.07\\]\\(#\\). NF_V: \\[Number
629 \\\[Fraction of Volatile aerosol whose Shrink Factor at 200 °C is lower than 0.85\\\]\\\(#\\\).\\\)\\]\\(#\\) were all negatively\]\(#\)](#)

630 correlated to either MF_{FFOA} mass fractions of BBOA (Biomass Burning Organic Aerosols) or
631 MF_{BBOA} FFOA (Fossil Fuel Organic Aerosols), because fossil fuel combustion and biomass burning
632 were two major sources of BC-containing aerosols during this field campaign. However, differences
633 among these mixing state parameters varied much under different conditions.

634 The intercomparison results highlight the differences in aerosol mixing states and physiochemical
635 properties caused by fossil fuel combustion and biomass burning emissions. CCN and SP2
636 measurements showed that CCN-inactive, BC-free particles were mainly produced by fossil fuel
637 combustion. On the other hand, the comparison between VTDMA and SP2 measurements indicated
638 that non-volatile aerosols were mostly composed of BC-containing particles. The comparison between
639 HTDMA and SP2 measurements revealed that a significant proportion of nearly hydrophobic aerosols
640 was not from BC-containing particles, but from fossil fuel combustion or biomass burning dominated
641 organic aerosols. The correlation between CCNC and HTDMA measurements also showed that nearly
642 hydrophobic, BC-free aerosols can become CCN-active under supersaturated conditions and were
643 tightly linked to biomass burning emissions. This suggests that biomass burning aerosols might exhibit
644 different hygroscopicity under sub- and super- saturated conditions (Bougiatioti et al., 2016).
645 Furthermore, the correlation analysis between SP2 measurements and mass fractions of BBOA and
646 FFOA indicated that BC-containing aerosols from fossil fuel combustion tend to be more externally
647 mixed with other aerosol compositions compared to those from biomass burning activities.

648 Other than primary aerosol emissions, secondary aerosol formations also exerted significant
649 impacts on variations of aerosol mixing states. During this campaign, the secondary aerosol formation
650 was dominated by nitrate and SOA production, which have markedly different impacts on aerosol
651 mixing states due to their different physical properties and formation pathways. NF_{noBC} , MAF, NF_H ,
652 and NF_V were all positively correlated with nitrate and SOA mass fractions, revealing much stronger
653 correlations to SOA than with nitrate. This is consistent with the semi-volatile but highly hygroscopic
654 properties of nitrate. The high correlation coefficient between NF_{noBC} and the mass fractions of
655 secondary aerosols at 200 nm and 300 nm suggests that secondary aerosol formation led to the
656 migration of BC-free aerosols towards larger diameters more quickly than BC-containing aerosols.
657 This outcome reveals that secondary aerosols formed more rapidly on BC-free aerosols than on BC-
658 containing aerosols, which is in line with the BC-containing aerosols' hydrophobic nature that does

659 not favor aqueous secondary aerosol formation. Moreover, as the mass fractions of nitrate or
660 ammonium increased, the differences among mixing state parameters (NF_{noBC} , MAF, NF_H and NF_V)
661 mostly decreased due to the hygroscopic and semi-volatile nature of ammonium nitrate. However, the
662 two resolved SOA factors exhibited different impacts on ~~NF_V~~ [the difference between \$NF_V\$ and \$NF_H\$](#) ;
663 [\(\$NF_V - NF_H\$ \)](#), and their correlations with NF_V and NF_H revealed that OOA1 was more hygroscopic but
664 less volatile, suggesting distinct formation mechanisms for the two OOA factors during the field
665 campaign.

666 The findings of this study highlight markedly different effects of primary emissions and
667 secondary aerosol formations on aerosol mixing states and suggest that comparisons of aerosol mixing
668 states obtained using various techniques are useful for gaining insights into the hygroscopicity,
669 volatility, and CCN activity of different aerosols. These comparisons can also indicate the impacts of
670 secondary aerosol formations on aerosol physical properties, which can help to understand the
671 pathways of secondary aerosol formation. However, it is important to be cautious in the application of
672 aerosol mixing state parameters, as the suitability of VTDMA-derived mixing state parameters in
673 representing BC mixing states is largely dependent on the composition and mass of secondary aerosols.

674

675 **Data availability.** The data used in this study are available from the corresponding author upon request
676 Ye Kuang (kuangye@jnu.edu.cn) and Li Liu (liul@gd121.cn)

677 **Competing interests.** The authors declare that they have no conflict of interest.

678

679 **Author Contributions.**

680 YK and WY planned this campaign and YK designed the aerosol experiments and conceived this
681 research together with JC, and JC wrote the manuscript. JC performed measurements of CCNC, BL
682 performed measurements of SP2 and analyzed SP2 datasets with the help of GZ, WQ and YL
683 performed AMS measurements, LL performed HV-TDMA measurements and conducted post-data
684 processing as well as some of data analysis. BX, HX, MMZ, HZ and SR participated this campaign
685 and helped instruments maintenance. GZ provided full support for the campaign. All authors
686 contributed to discussions and revisions of this paper.

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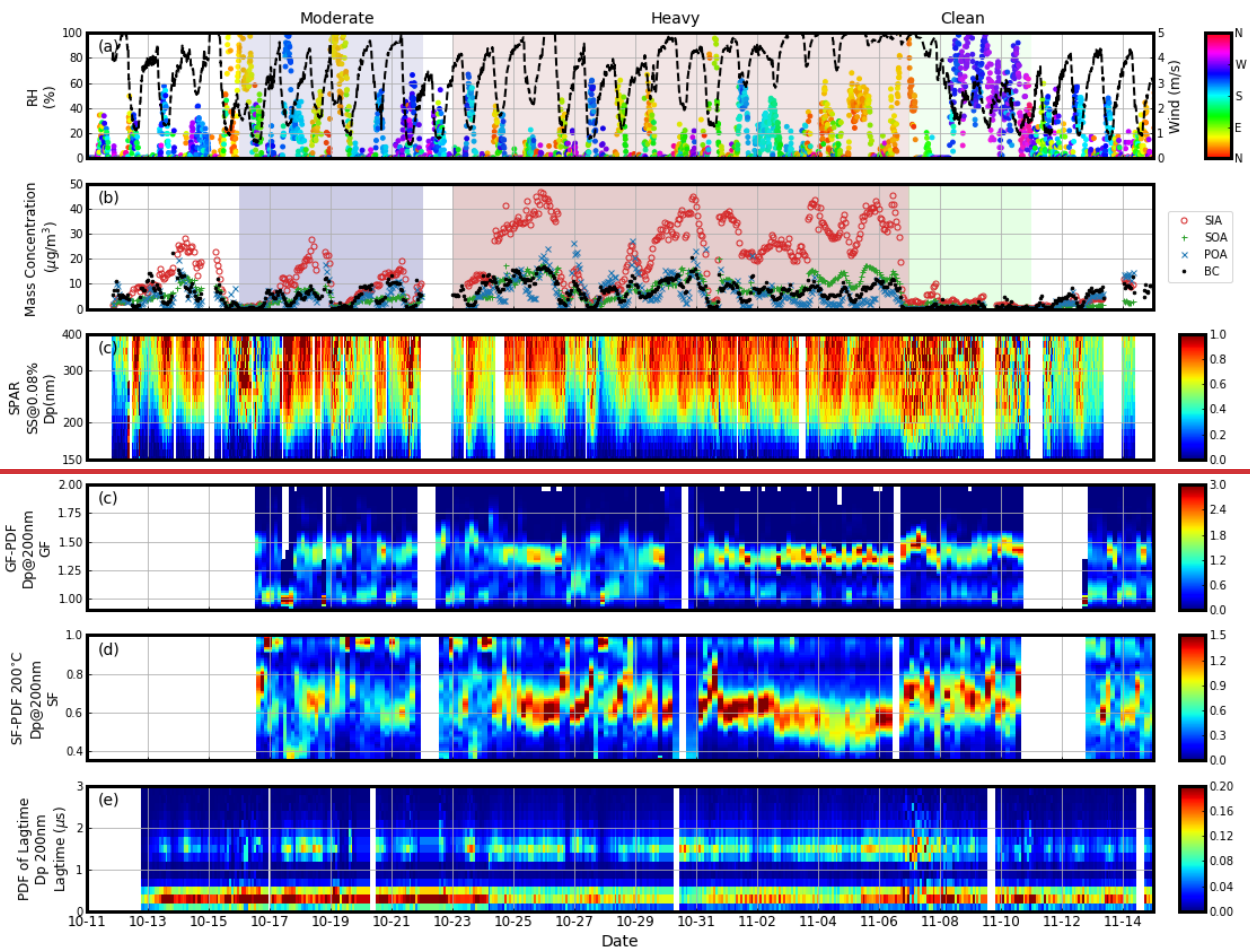


Fig 1. Overview of the measurements during the campaign: (a) dots represent wind speed with color indicating wind direction, and black lines represent RH; (b) red circle, green plus, blue x and black dots represent mass concentration of SIA, SOA, POA and BC, respectively; (c) SPAR under SS of 0.08%; (d) PDF of GF (GFPDF) at 200 nm; (e) PDF of SF (SFPDF) at 200 nm and 200 °C; (f) PDF of lag time at 200 nm. The blue, red and green shaded periods represent the three periods with moderate pollution, heavy pollution and clean condition, respectively.

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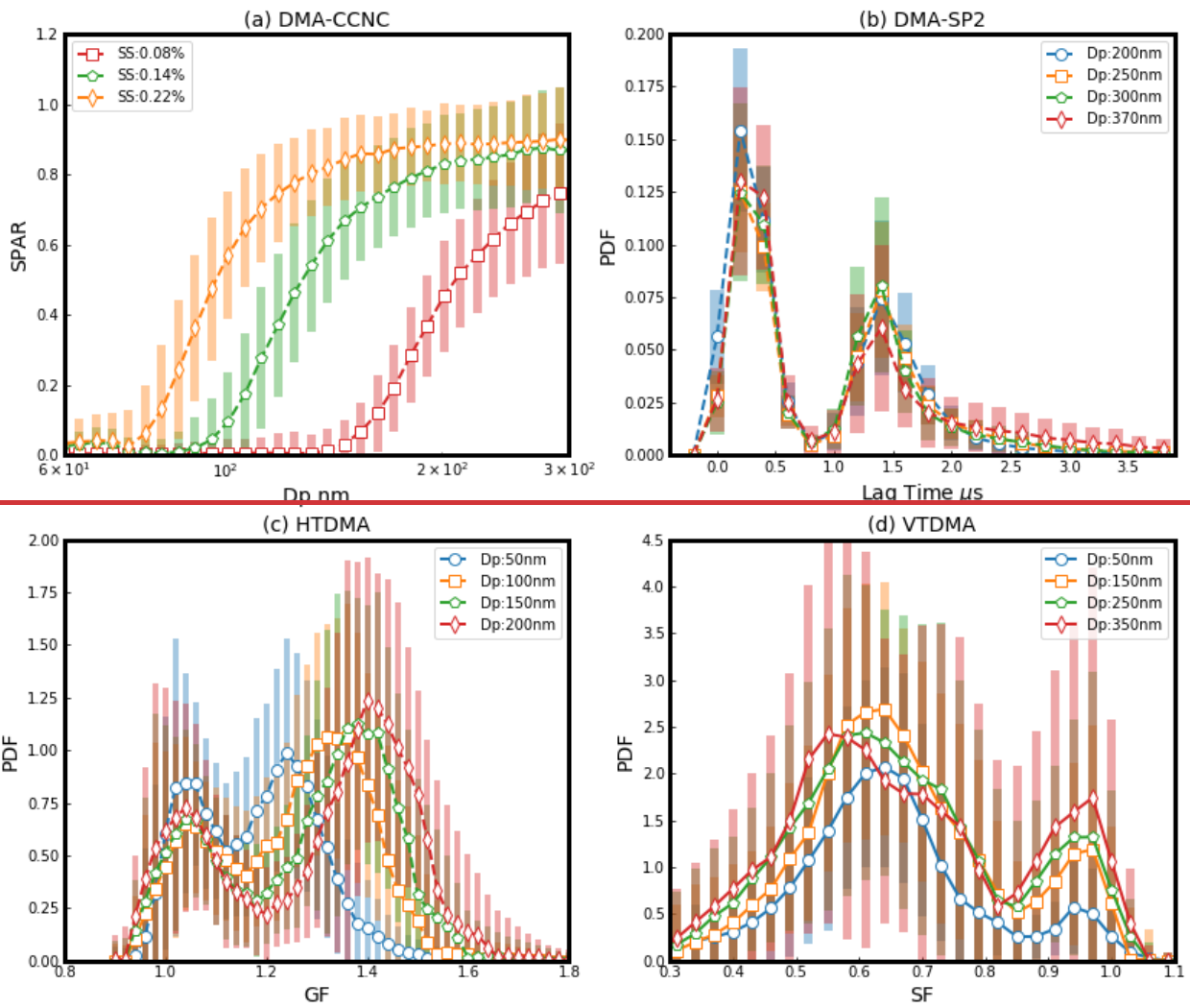


Fig. 2. The campaign average of (a) SPAR curves measured by DMA-CCNC at the three SSs (identified by colors and markers), (b) PDF of lag-time measured by DMA-SP2 at four particle sizes (identified by colors and markers), (c) PDF of GF (GFPDF) measured by HTDMA at four particle sizes (identified by colors and markers), (d) PDF of SF (SFPDF) measured by VTDMA under the temperature of 200 °C at five particle sizes (identified by colors and markers).

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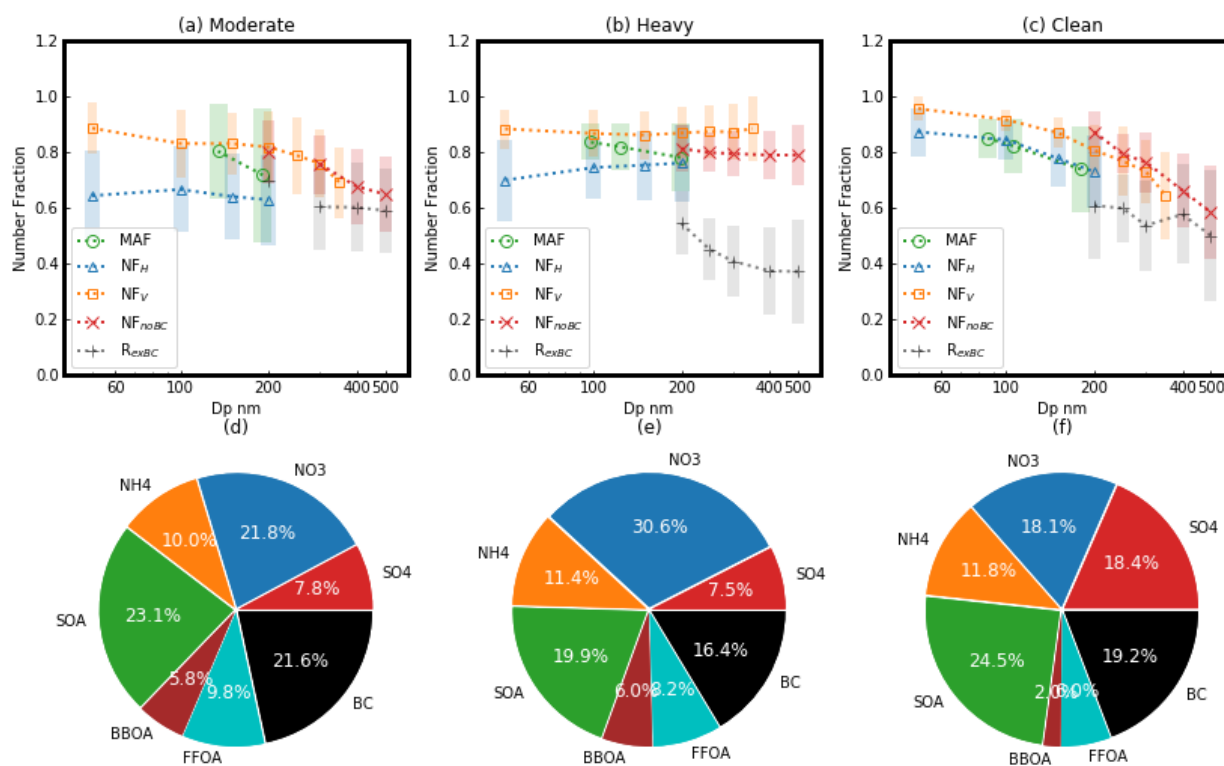


Fig. 3. (a-c): Size dependence of MAF (green circle), NF_H (blue triangle), NF_V at 200 °C (yellow square), NF_{noBC} (red x) and R_{exBC} (black plus) during the three periods. (d-f): Corresponding mass fractions of aerosol chemical compositions (identified by colors) during the three periods.

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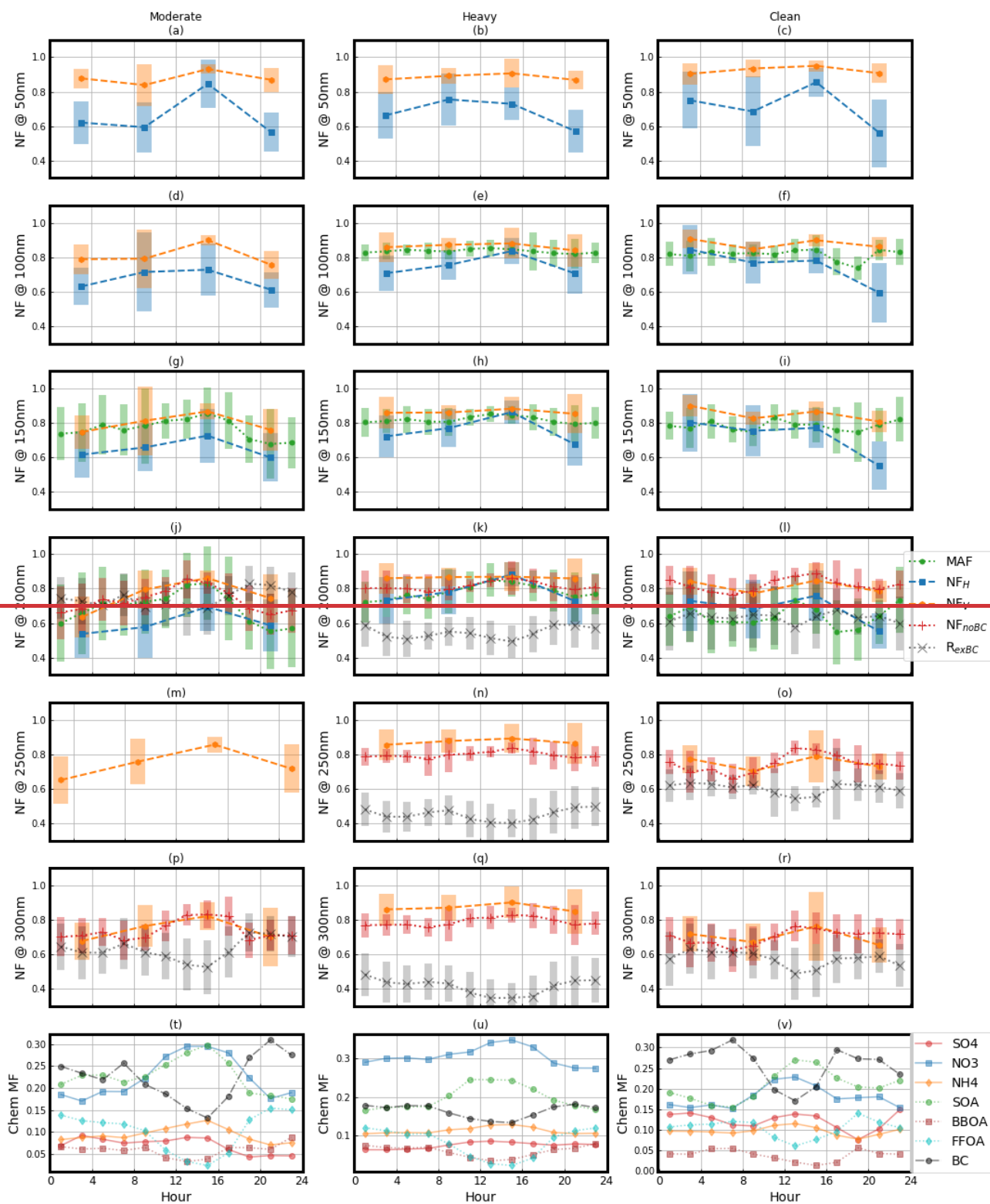


Fig. 4. (a-r) Diurnal variation of aerosol mixing state parameters (identified by color and marker) at different particle size during the three periods. (t-v) Diurnal variation of mass fraction of aerosol chemical compositions (identified by color and marker) during the three periods.

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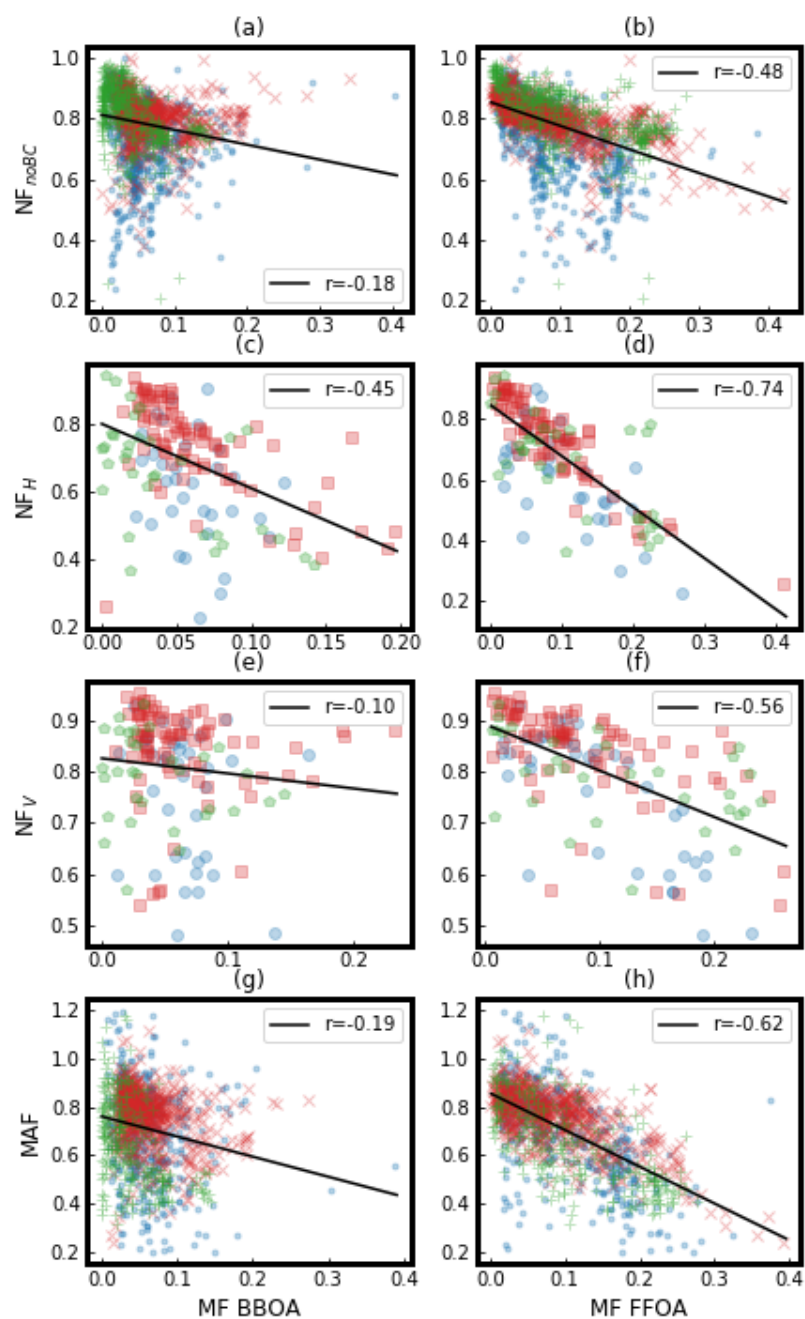


Fig. 5. The correlation between aerosol mixing state parameters and MF of primary organic aerosol-chemical composition during different periods. Moderately polluted period: Blue dot or circle; Heavily polluted period: Red x or square; Clean period: Green plus or pentagon.

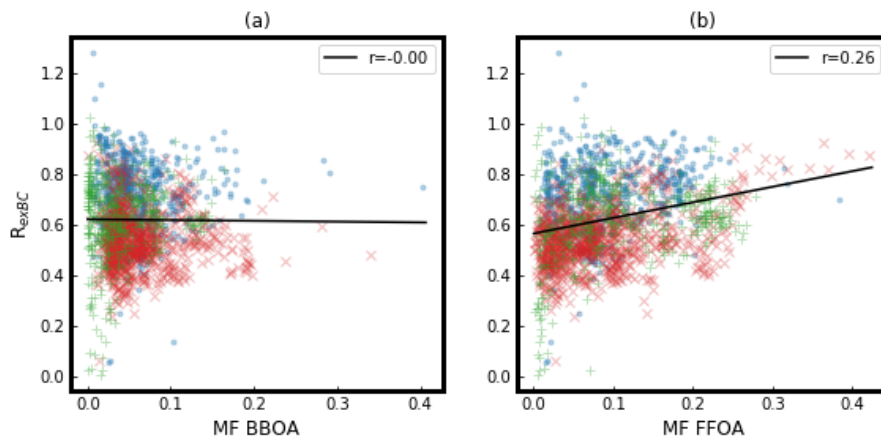


Fig. 6. The correlation between the ratio of external mixed BC in total BC particles and MF of primary organic aerosol chemical composition during different periods. Moderately polluted period: Blue dot; Heavily polluted period: Red x; Clean period: Green plus.

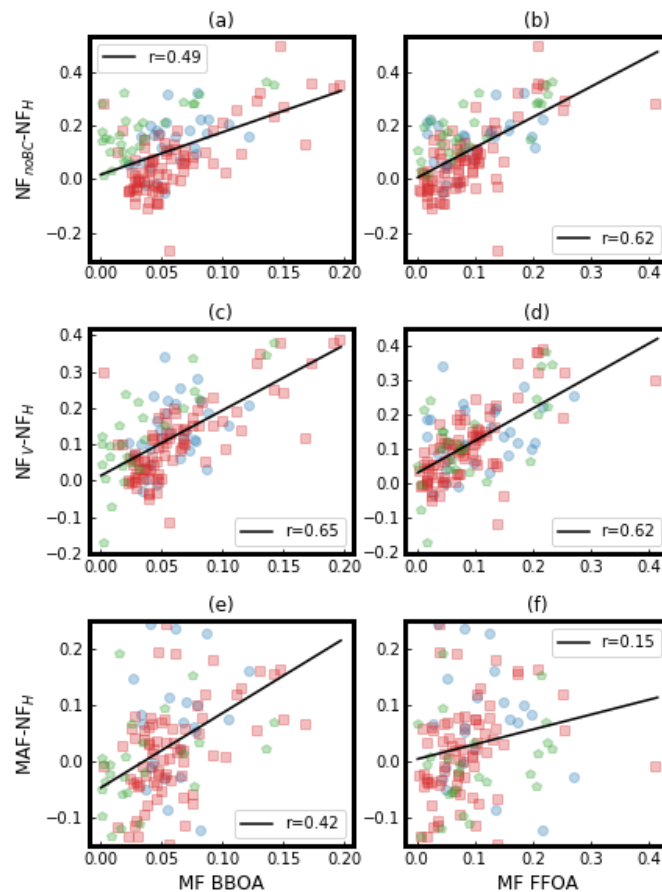


Fig. 7. The correlation between the difference among the four aerosol mixing state parameters and MF of primary organic aerosol chemical composition during different periods. Moderately polluted period: Blue circle; Heavily polluted period: Red square; Clean period: Green pentagon.

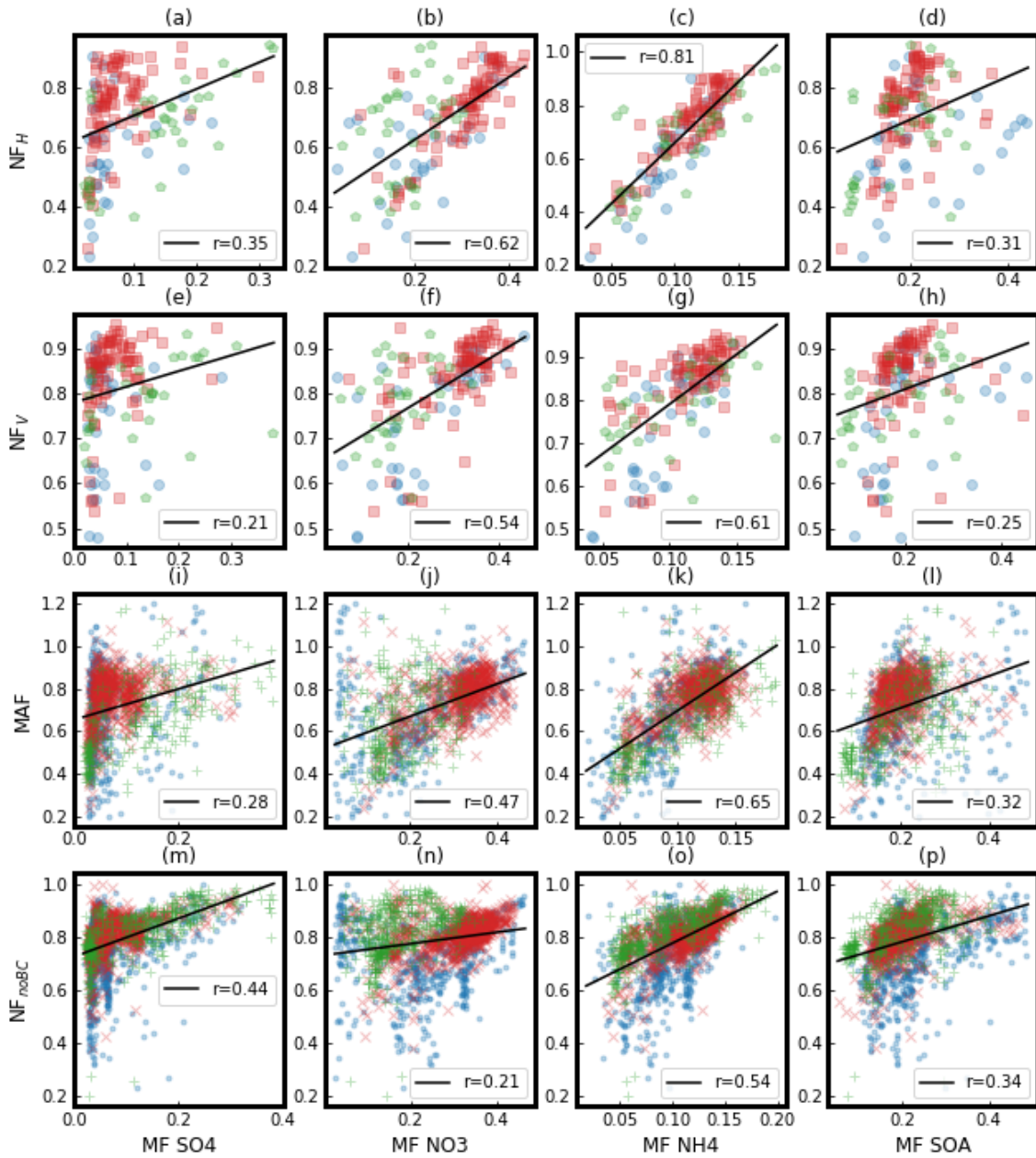


Fig. 8. The correlation between the four aerosol mixing state parameters and MF of secondary aerosol chemical composition during different periods. Moderately polluted period: Blue dot or circle; Heavily polluted period: Red x or square; Clean period: Green plus or pentagon.

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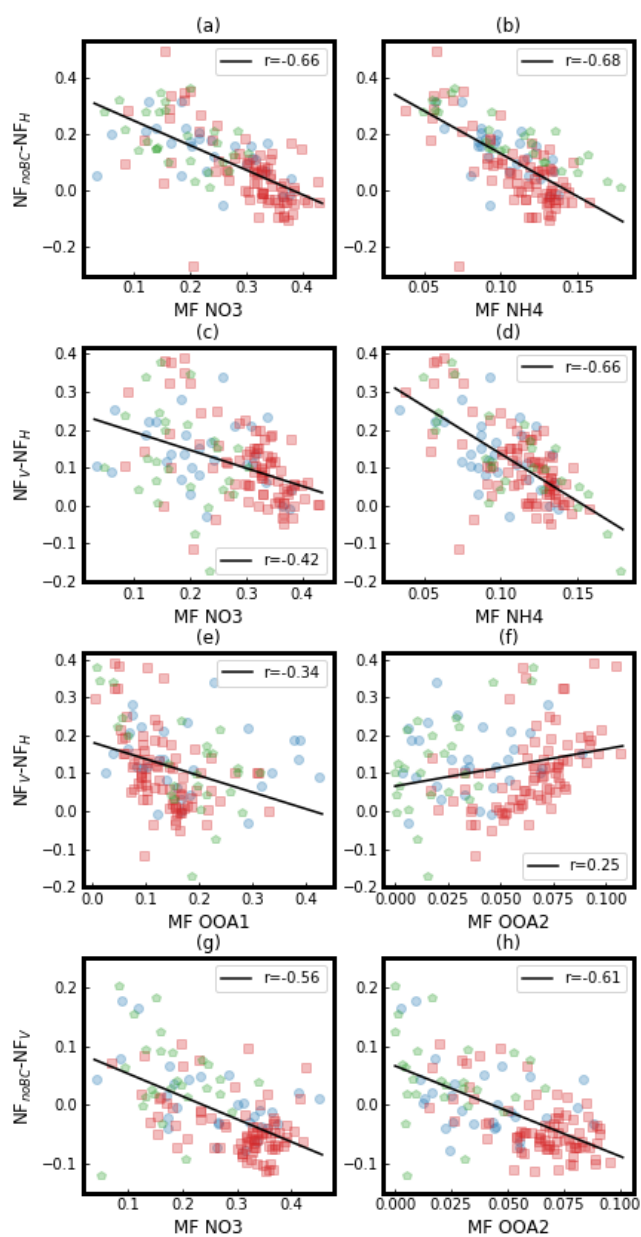


Fig. 9. The correlation between the difference among the four aerosol mixing state parameters and MF of secondary aerosol chemical composition during different periods. Moderately polluted period: Blue circle; Heavily polluted period: Red square; Clean period: Green pentagon.

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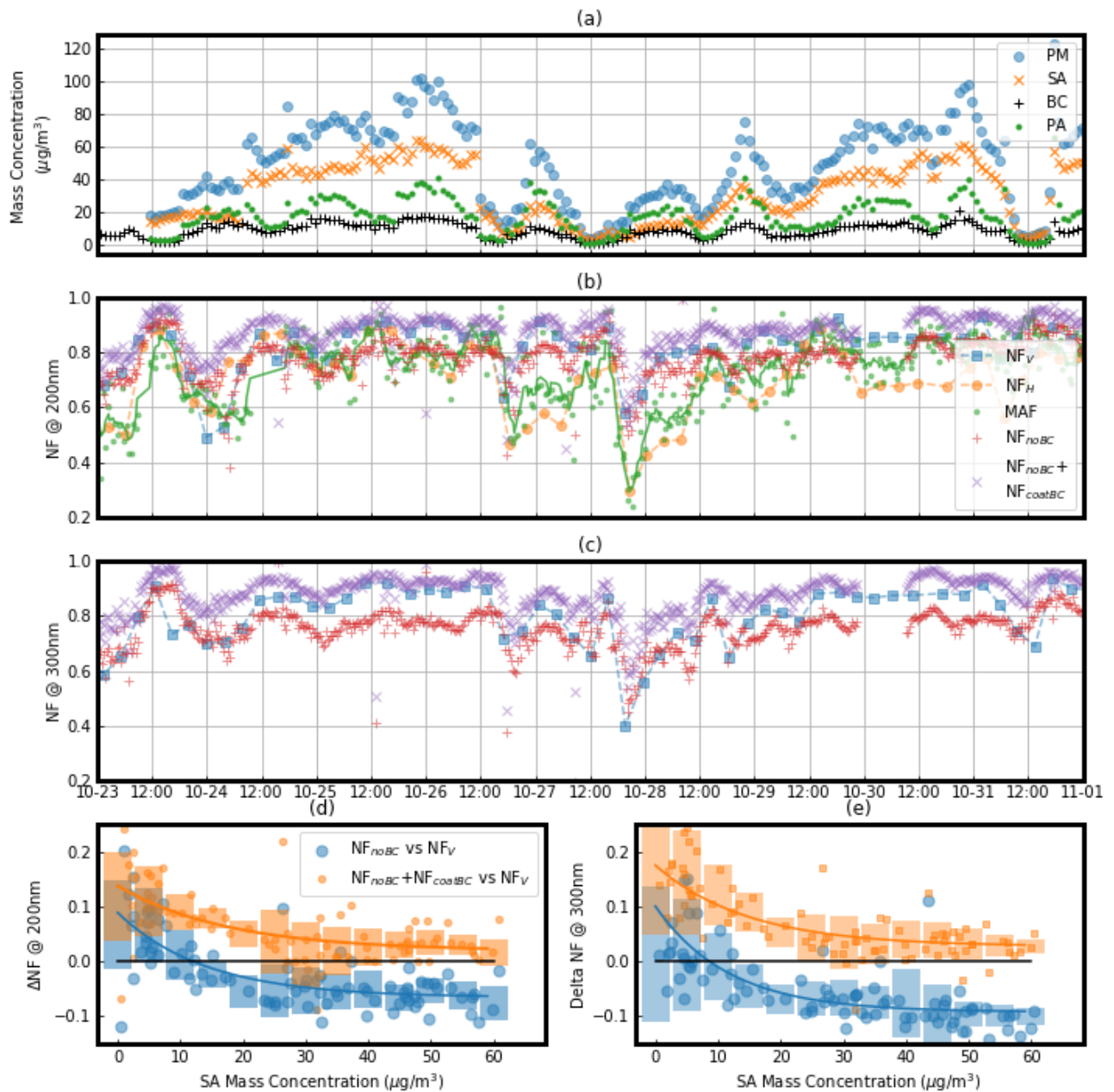


Fig. 10. The variations of different aerosol mixing state parameters during the pollution accumulation process. (a) The mass concentration of PM, SA, PA and BC (identified by colors and markers). (b and c) The variations of different aerosol mixing state parameters (identified by colors and markers) at particle size of 200nm (b) and 300nm (c); (d and e) The variations of the difference between $\text{NF}_V - \text{NF}_{noBC}$ (blue large circle) and $\text{NF}_V - \text{NF}_{noBC} + \text{NF}_{coastBC}$ (yellow small circle) with the mass concentration of SA at particle size of 200nm (d) and 300nm (e).