We thank the referee for the insightful comments and suggestions, which give us new perspectives to describe our work and greatly helped to improve the significance of the paper. Our answers are listed in the following in red, after the referee's comments, which are in black. The modifications in the text are marked in yellow.

# **Reviewer #1**

This paper describes an improved direct measurement system of the photochemical ozone production rate ( $P(O_3)$ ). The authors also performed a field observation of  $P(O_3)$  and evaluated the instrument and observation results using a detailed box model simulation.

 $O_3$  pollution is a crucial problem for atmospheric environment, and a behavior of  $O_3$  is very difficult, so direct measurements of  $P(O_3)$  such as this research are very important and valuable. However, this paper has major concerns about the evaluation by the box model simulation. So, I cannot recommend this paper to be published in Atmospheric Chemistry and Physics in the present form. The authors should resolve the concerns by re-evaluation of the box model simulation or other ways.

Thank you very much for providing useful suggestions. We agree with the reviewer that some of the modeling results in the reference chamber were beyond our expectation. Therefore, we have re-evaluated and carefully checked each step of the box model simulations described in Sect. 3.2, but no error has been found so far. To make the modeling results clearer, we added the budgets of OH, HO<sub>2</sub>, RO<sub>2</sub>, and NO<sub>3</sub> during the 3<sup>rd</sup>-stage 4-min model simulation in the reaction and reference chambers. The main purposes of performing the box model simulation in this study are as follows: (1) how to determine the photolysis frequencies for  $P(O_3)_{net}$  modeling in ambient air by comparing the modeling results with the measured results, and 2 provide a comprehensive understanding of the radical budgets in the reaction and reference chambers. The reactions occurring in the reference chamber are much more complex than those in the reaction chamber, and the results are also different when we use different J value determination methods (labeled as method I and method II). However, the biases of the modeled  $P(O_3)_{net}$  caused by the interferences in the reference chamber were 13.9 % and 22.3 % in method I and method II, respectively, which ensures that the measured  $P(O_3)_{net}$  by the NPOPR detection system should be regarded as the lower limit values of the real  $P(O_3)_{net}$ in the atmosphere. We made some additional box model simulations and answered the reviewer's questions point by point as follows:

Major comments:

• The results of the box model simulation

I suspect such high concentrations of NO<sub>3</sub> in the reference chamber. First, the value of JNO<sub>3</sub> in the reference chamber is about 90% of that in the reaction chamber, which is sufficiently high. Second, the rate constant of the reaction of NO<sub>3</sub> with NO is sufficiently large and the lifetime of NO<sub>3</sub> (at 298 K) is 1.6 s and very short in the presence of NO of 1 ppbv. There might be high concentrations of NO<sub>3</sub> if there are very large sources of NO<sub>3</sub>, but in that case, the authors should mention the evidence. I think the NO<sub>2</sub> + O<sub>3</sub> reaction cannot be a large source of NO<sub>3</sub>. I'm not sure about N<sub>2</sub>O<sub>5</sub>, but it is unlikely that there could be high concentrations of N<sub>2</sub>O<sub>5</sub> under the temperature during the observation period.

To explore the reason for such a high concentration of NO<sub>3</sub> in the reference chamber, we modeled the production and destruction pathways of NO<sub>3</sub> in the reaction and reference chambers (as shown in Fig. S19 for method I and method II). First, we found that even when the  $J(NO_3)$  values were sufficiently large (which were 90% and 100% of that in the reaction chamber for method I and method II, respectively), which led to a high NO<sub>3</sub> photolysis reaction in the reference chamber, the NO<sub>3</sub> concentrations were still sufficiently high due to the high production rates of NO<sub>3</sub> at the same time (the main production pathway of NO<sub>3</sub> is the NO<sub>2</sub>+O<sub>3</sub> reaction, followed by N<sub>2</sub>O<sub>5</sub> decomposition). Second, the NO concentrations were large in the 1<sup>st</sup> minute, which consumes NO<sub>3</sub> very quickly, but as there were continuous NO<sub>3</sub> sources, the net NO<sub>3</sub> production rates ( $P(NO_3)_{net}$ ) were positive, which caused the NO<sub>3</sub> concentration to continue to increase (as shown in Figs. 8d and S20).

The main difference of NO<sub>3</sub> production in the reference chamber compared to that in the reaction chamber is the much higher  $N_2O_5$  decomposition. A proposed mechanism for the decomposition of  $N_2O_5$  is as follows (Kotz et al., 2019):

$$N_{2}O_{5} \xrightarrow{k1} NO_{2}+NO_{3} \qquad \text{step 1 (slow)}$$

$$NO_{2}+NO_{3} \xrightarrow{k2} NO_{2}+O_{2}+NO \qquad \text{step 2 (fast)}$$

$$NO+N_{2}O_{5} \xrightarrow{k3} 3NO_{2} \qquad \text{step 3 (fast)}$$

We proposed that due to the much higher NO<sub>2</sub> concentrations in the reference chamber than in the reaction chamber, NO<sub>3</sub> is consumed to produce NO in step 2, which accelerates step 1 and benefits step 3, thus increasing the steady-state concentration of NO<sub>3</sub>. This is also demonstrated by the increased N<sub>2</sub>O<sub>5</sub> decomposition and NO<sub>2</sub> concentrations along with the increased NO<sub>2</sub>+NO<sub>3</sub> reaction (as shown in Figs. 8-9 and S18-S19). Therefore, the high NO<sub>3</sub> concentrations in the reference chamber were mainly due to the high NO<sub>2</sub> concentrations. On the other hand, although the NO<sub>2</sub>+NO<sub>3</sub> reaction was also one of the dominant NO<sub>3</sub> destruction pathways, NO<sub>3</sub> consumed by the reaction of

 $NO_2$ + $NO_3$  was significantly smaller than NO3 produced by the reaction of NO2+O3 (as shown in Figs. 9 and S19).

We added the related discussion for method I in pages 24-25, lines 551-567 in the modified manuscript:

"Due to sufficiently high  $J(NO_3)$  (~ 90% of that in the reaction chamber) and NO<sub>2</sub> concentrations in the reference chamber, the NO<sub>3</sub> photolysis and NO<sub>2</sub>+NO<sub>3</sub> reaction consumed NO<sub>3</sub> in the reference chamber, but the  $NO_3$  concentrations were still sufficiently high due to high production rates of  $NO_3$  at the same time. The main  $NO_3$  source in the reference chamber was the  $NO_2+O_3$  reaction, followed by  $N_2O_5$  decomposition. The NO concentrations were relatively high in the 1<sup>st</sup> minute and consumed NO<sub>3</sub> very quickly, but due to continuous NO<sub>3</sub> sources, the net NO<sub>3</sub> production rates  $(P(NO_3)_{net})$  were positive (as shown in Fig. 9), which caused the NO<sub>3</sub> concentration to continue to increase (as shown in Fig. 8d). The main difference in NO<sub>3</sub> production in the reference chamber compared to that in the reaction chamber was the much higher N<sub>2</sub>O<sub>5</sub> decomposition, which was mainly due to the high NO<sub>2</sub> concentrations in the reference chamber. On the other hand, although the NO+NO<sub>3</sub> reaction was also one of the dominant NO<sub>3</sub> destruction pathways, NO<sub>3</sub> consumed by the NO+NO<sub>3</sub> reaction was significantly smaller than NO<sub>3</sub> produced by the  $NO_2+O_3$  reaction. Furthermore, in order to check if the  $NO_3+VOCs$  reactions exists, we extracted all the P(ROx) pathways related to NO<sub>3</sub>+VOCs reactions during the 3<sup>rd</sup>-stage 4-min model simulation in the reaction and reference chambers in method I, as shown in Fig. S20. We found that the NO<sub>3</sub>+VOCs reactions are mostly related to the OVOCs (i.e. 6-Ethyl-m-cresol and 3-Ethyl-6methylbenzene-1,2-diol) in Fig. S21. The production and destruction rates of ROx are shown in Fig. S20."

And added the related discussion for method II in page 25, lines 330-343 in the modified supplementary materials:

Due to sufficiently high  $J(NO_3)$  (~ 100 % of that in the reaction chamber) and NO<sub>2</sub> concentrations in the reference chamber, the NO<sub>3</sub> photolysis and NO<sub>2</sub>+NO<sub>3</sub> reaction consumed NO<sub>3</sub> in the reference chamber, but the NO<sub>3</sub> concentrations were still high due to high production rates of NO<sub>3</sub> at the same time. Similar with the results obtained from method I as described in the main manuscript, for method II, the main NO<sub>3</sub> source in the reference chamber was the NO<sub>2</sub>+O<sub>3</sub> reaction, followed by N<sub>2</sub>O<sub>5</sub> decomposition. The NO concentrations were relatively high in the 1<sup>st</sup> minute and consumed NO<sub>3</sub> very quickly, but due to continuous NO<sub>3</sub> sources, the net NO<sub>3</sub> production rates (*P*(NO<sub>3</sub>)<sub>net</sub>) were positive (as shown in Fig. S19b4), which caused the NO<sub>3</sub> concentration to continue to increase (as shown in Fig. S18d). The main difference in NO<sub>3</sub> production in the reference chamber compared to that in the reaction chamber was the much higher  $N_2O_5$  decomposition, which was mainly due to the high NO<sub>2</sub> concentrations in the reference chamber. On the other hand, although the NO+NO<sub>3</sub> reaction was also one of the dominant NO<sub>3</sub> destruction pathways, NO<sub>3</sub> consumed by the NO+NO<sub>3</sub> reaction was significantly smaller than NO<sub>3</sub> produced by the NO<sub>2</sub>+O<sub>3</sub> reaction. The integrated production and destruction rates of ROx are shown in





Figure 9: Production and destruction pathways of OH (a1-b1), HO2 (a2-b2), RO2 (a3-b3), and NO3 (a4-b4) during the 3<sup>rd</sup>-stage 4-min model simulation in the reaction and reference chambers in method I. The related contents for method II (c)-(d) are shown in Fig. S19 in the supplementary materials.

And added Fig. S19 in the modified supplementary materials:

"Due to sufficiently high  $J(NO_3)$  (~ 100 % of that in the reaction chamber) and NO<sub>2</sub> concentrations in the reference chamber, the NO<sub>3</sub> photolysis and NO<sub>2</sub>+NO<sub>3</sub> reaction consumed NO<sub>3</sub> in the reference chamber, but the NO<sub>3</sub> concentrations were still high due to high production rates of NO<sub>3</sub> at the same time. Similar with the results obtained from method I as described in the main manuscript, for method II, the main NO<sub>3</sub> source in the reference chamber was the NO<sub>2</sub>+O<sub>3</sub> reaction, followed by N<sub>2</sub>O<sub>5</sub> decomposition. The NO concentrations were relatively high in the 1<sup>st</sup> minute and consumed NO<sub>3</sub> very quickly, but due to continuous NO<sub>3</sub> sources, the net NO<sub>3</sub> production rates ( $P(NO_3)_{net}$ ) were positive (as shown in Fig. S19b4), which caused the NO<sub>3</sub> concentration to continue to increase (as shown in Fig. S18d). The main difference in NO<sub>3</sub> production in the reference chamber compared to that in the reaction chamber was the much higher N<sub>2</sub>O<sub>5</sub> decomposition, which was mainly due to the high NO<sub>2</sub> concentrations in the reference chamber. On the other hand, although the NO+NO<sub>3</sub> reaction was also one of the dominant NO<sub>3</sub> destruction pathways, NO<sub>3</sub> consumed by the NO+NO<sub>3</sub> reaction was significantly smaller than NO<sub>3</sub> produced by the NO<sub>2</sub>+O<sub>3</sub> reaction."





during the 3<sup>rd</sup>-stag 4-min model simulation in the reaction and reference chambers in method II (c)-(d).

On the other hand, I also suspect the reaction of NO<sub>3</sub> with VOCs as a major source of RO<sub>2</sub> in the reference chamber. The rate constants of the reactions of NO<sub>3</sub> with VOCs are not so large, and it is questionable that NO<sub>3</sub> and RO<sub>2</sub> concentrations change in minutes by the reactions of NO<sub>3</sub> with VOCs unless there are extremely high concentrations of VOCs.

The reviewer is correct. We actually described it incorrectly here, and we apologize for this mistake. The major RO<sub>2</sub> sources in the reference chamber were OH+VOC oxidation, followed by OVOC photolysis (i.e.,  $C_3H_4O_2$ ,  $C_2H_2O_2$ ,  $C_4H_6O_2$ , etc.) and  $O_3$ +VOC reactions. To better understand the radical chemistry, we included the

production and destruction pathways of OH, HO<sub>2</sub>, RO<sub>2</sub>, and NO<sub>3</sub> during the 3<sup>rd</sup>stage 4-min model simulation in the reaction and reference chambers in Figs. 9 and S19 (as shown above) and moved the production and destruction pathways of ROx to Fig. S20.



Figure S20: Production and destruction pathways of RO<sub>X</sub> during the 3<sup>rd</sup>-stage 4-min model simulation in the reaction and reference chambers. (PAN: Peroxyacetyl Nitrate; PNs: formations of all peroxynitrate (including CH3O2NO2 and PAN; X: PAN and the net loss of OH+NO to form HONO (usually small)).

Accordingly, we modified the related description in page 24, lines 542-552 in the modified manuscript:

"OH, HO<sub>2</sub>, RO<sub>2</sub>, and NO<sub>3</sub> concentrations greatly impact the O<sub>3</sub> production and destruction rate. To better understand the factors that drive the OH, HO<sub>2</sub>, RO<sub>2</sub>, and NO<sub>3</sub> concentration changes, we have added their production and destruction pathways in Fig. 9. We found that the decrease in HO<sub>2</sub> and RO<sub>2</sub> concentrations in the reference chamber in the 1<sup>st</sup> half minute was mainly due to NO titration effects, as high NO mixing ratios existed during the 1<sup>st</sup> half minute. The increase in HO<sub>2</sub> concentrations afterward was largely attributable to RO+O<sub>2</sub> reaction/RO decomposition, OH+CO/VOCs reaction, OVOCs photolysis (i.e., C<sub>3</sub>H<sub>4</sub>O<sub>2</sub>, C<sub>2</sub>H<sub>2</sub>O<sub>2</sub>, C<sub>4</sub>H<sub>6</sub>O<sub>2</sub>), and NO<sub>3</sub>+VOCs reaction, and the increase in RO<sub>2</sub> concentrations afterward were largely attributable to OH+VOCs oxidation, OVOCs photolysis and O<sub>3</sub>+VOCs reaction. The main OH sources in the reference chamber were both HO<sub>2</sub>+NO in method I and method II."

And in pages 24-25, lines 318-329 in the modified supplementary materials:

"OH, HO<sub>2</sub>, RO<sub>2</sub>, and NO<sub>3</sub> concentrations greatly impact the O<sub>3</sub> production and destruction rate. To better understand the factors that drive the OH, HO<sub>2</sub>, RO<sub>2</sub>, and NO<sub>3</sub> concentration changes in method II, we have added their production and destruction pathways in Fig. S19. We found that the decrease in HO<sub>2</sub> and RO<sub>2</sub> concentrations in the reference chamber in the 1<sup>st</sup> half minute was mainly due to NO titration effects, as high NO mixing ratios existed during the 1<sup>st</sup> half minute. The HO<sub>2</sub> and RO<sub>2</sub> concentrations were became stable afterwards, the main production pathway for HO<sub>2</sub> was RO+O<sub>2</sub> reaction/RO decomposition, followed by OH+ VOCs reaction, OVOCs photolysis (i.e., C<sub>3</sub>H<sub>4</sub>O<sub>2</sub>, C<sub>2</sub>H<sub>2</sub>O<sub>2</sub>, C<sub>4</sub>H<sub>6</sub>O<sub>2</sub>), and NO<sub>3</sub>+VOCs reaction; while the main production pathway for RO<sub>2</sub> was OH+ VOCs reaction, followed by OVOCs photolysis (i.e., C<sub>3</sub>H<sub>4</sub>O<sub>2</sub>, C<sub>2</sub>H<sub>2</sub>O<sub>2</sub>, C<sub>4</sub>H<sub>6</sub>O<sub>2</sub>), OH+CO, NO<sub>3</sub>+VOCs reaction, etc.; the main destruction pathways for HO<sub>2</sub> and RO<sub>2</sub> were HO<sub>2</sub>+NO and RO<sub>2</sub>+NO, respectively. The main OH production and destruction pathways in the reference chamber was HO<sub>2</sub>+NO reaction and OH+ VOCs reaction, respectively."

Furthermore, in order to check if the NO<sub>3</sub>+VOCs reactions exists, we extracted all the P(ROx) pathways related to NO<sub>3</sub>+VOCs reactions during the 3<sup>rd</sup>-stage 4-min model simulation in the reaction and reference chambers in method I, as shown in Fig. S20. We found that the NO<sub>3</sub>+VOCs reactions are mostly related to the OVOCs (i.e. 6-Ethyl-m-cresol and 3-Ethyl-6-methylbenzene-1,2-diol), as shown in Fig. S21.

The related explanations are added in page 25, line 562-567 in the modified manuscript:

"Furthermore, in order to check if the NO<sub>3</sub>+VOCs reactions exists, we extracted all the *P*(ROx) pathways related to NO<sub>3</sub>+VOCs reactions during the 3<sup>rd</sup>-stage 4-min model simulation in the reaction and reference chambers in method I, as shown in Fig. S20. We found that the NO<sub>3</sub>+VOCs reactions are mostly related to the OVOCs (i.e. 6-Ethyl-m-cresol and 3-Ethyl-6-methylbenzene-1,2-diol) in Fig. S21. The production and destruction rates of ROx are shown in Fig. S20."

And in page 27, lines 354-357 in the modified supplementary materials:



Figure S21: The P(ROx) pathways related to NO<sub>3</sub>+VOCs reactions during the 3<sup>rd</sup>-stage 4-minute model simulation in the reaction and reference chambers in method I (a)-(b) and method II (c)-(d).

For Figs. 9(b) and S10(b), photodissociation of OVOCs is a significant part of ROx sources in the reference chamber. Why such a large fraction in the absence of UV? The authors should mention the evidence.

The reference chamber was covered by a UV protection Ultem film (SH2CLAR, 3M, Japan) to block the sunlight with the wavelengths < 390 nm, so the film has the transmission of lights > 390 nm, thus the photolysis frequencies of some species still existed, as shown in Tables S7 and S13. Therefore, we still observed OVOCs photodissociation in the reference chamber., i.e.,  $C_3H_4O_2$ ,  $C_2H_2O_2$ ,  $C_4H_6O_2$  and  $C_4H_4O_4$ , etc. According to MCMv3.3.1, there are in total 979 OVOCs species photolysis in the reference chamber, some of which are listed as follows:

Photolysis reactions from IUPAC Task Group on Atmospheric Chemical Kinetic Data Evaluation (http://iupac. pole-ether.fr/):

*hv*: 225-410 nm → CH<sub>3</sub>CO<sub>3</sub> +CO+H<sub>2</sub>O IUPAC (2006) 1. C<sub>3</sub>H<sub>4</sub>O<sub>2</sub> (Methylglyoxal) *hv*: 230.5-462.0 nm → CO+CO+H<sub>2</sub> IUPAC(2006) 2. (1) $C_2H_2O_2$  (Glyoxal) – *hv*: 230.5-462.0 nm CO+CO+HO<sub>2</sub>+HO<sub>2</sub> IUPAC(2006)  $(2)C_2H_2O_2(Glyoxal)$ *hv*: 230.5-462.0 nm  $(3)C_2H_2O_2$  (Glyoxal) -→ HCHO+CO IUPAC(2006) *hv*: 206-493 nm → CH<sub>3</sub>CO<sub>3</sub>+CH<sub>3</sub>CO<sub>3</sub> IUPAC(2011) 3. C<sub>4</sub>H<sub>6</sub>O<sub>2</sub> (Biacetyl) – hv: 225-410 nm 4. C<sub>4</sub>H<sub>6</sub>O<sub>2</sub> (Ethylglyoxal)- $\rightarrow$  C<sub>2</sub>H<sub>5</sub>CO<sub>3</sub> +CO+HO<sub>2</sub> IUPAC(2006)

We added the evidence in page 24, lines 546-550 in the modified manuscript:

"The increase in HO<sub>2</sub> concentrations afterward was largely attributable to  $RO+O_2$  reaction/RO decomposition, OH+CO/VOCs reaction, OVOCs photolysis (i.e.,  $C_3H_4O_2$ ,  $C_2H_2O_2$ ,  $C_4H_6O_2$ ), and  $NO_3+VOCs$  reaction, and the increase in RO<sub>2</sub> concentrations afterward were largely attributable to OH+VOCs oxidation, OVOCs photolysis and  $O_3+VOCs$  reaction."

I entertain doubts about the results of the present box model simulation. I think the authors should revalidate the appropriateness of the model thoroughly.

We have re-evaluated the model simulations and added more model simulations to explain some abnormal phenomena in the reference chamber, such as the production and destruction pathways of OH, HO<sub>2</sub>, RO<sub>2</sub>, and NO<sub>3</sub> during the 3<sup>rd</sup>stage 4-min model simulation in the reaction and reference chambers, the  $NO_3$ +VOC reactions, etc. We found that the radicals and gas species in the reaction chamber of the NPOPR detection system were similar to those in the ambient air, while these radicals also unexpectedly existed in the reference chamber. This was mainly because the UV protection film used by the reference chamber did not completely filter out sunlight, which led to the low transmittance of light ranging from 390 nm to 790 nm. We evaluated the interference of the unexpected reactions in the reference chambers and found that the biases of the modeled  $P(O_3)_{net}$  caused by this interference using method I and method II were 13.9 % and 22.3 %, respectively; therefore, we noted that ozone photochemical production still existed in the reference chamber, and the measured  $P(O_3)_{net}$  by the NPOPR system should be regarded as the lower limit values, as described in page 29, in lines 650-660 in the modified manuscript:

<sup>&</sup>quot;In conclusion, modeling tests demonstrated that the radicals and gas species in the reaction

chamber of the NPOPR detection system were similar to those in genuine ambient air, while these radicals also unexpectedly existed in the reference chamber. This was mainly because the UV protection film used by the reference chamber did not completely filter out sunlight, which led to the low transmittance of light ranging from 390 nm to 790 nm. The  $P(O_3)_{net}$  biases caused by this interference modeled in method I and method II were 13.9 % and 22.3 %, respectively, which ensured that the measured  $P(O_3)_{net}$  by the NPOPR detection system should be regarded as the lower limit values of real  $P(O_3)_{net}$  in the atmosphere. We recommend that the *J* values obtained from method I should be used in the model simulation, which can better explain the photochemical formation of  $O_3$  in the actual atmosphere, but if direct *J* value measurements cannot be achieved during field observations, the *J* values obtained from method II would also be acceptable in modeling studies."

### Answers to all of the reviewer's concerns are listed in the following:

• Measurements of NOx using a chemiluminescence NOx monitor.

Did the authors use a commercially available chemiluminescence NOx monitor without further modification? If so, accuracy of NO<sub>2</sub> and NOx concentrations would be low because NOz such as HNO<sub>3</sub> and PANs interfere observed values of NO<sub>2</sub> and NOx. I think this interference could affect discussion in this paper, so the authors should evaluate the interference quantitatively. In the fresh air masses, the interference by descendent pollutants of NOx such as HNO<sub>3</sub> and PANs might be small, but that by HONO could be large instead.

Yes, we used a commercially available chemiluminescence NOx monitor without further modification. To investigate the interference of HNO<sub>3</sub> and PANs, we added an evaluation of the effect of the NO<sub>2</sub> measurement bias of the chemiluminescence NOx monitor on  $P(O_3)_{net}$ . We compared the NO<sub>2</sub> measured by the chemiluminescence NOx monitor with that measured by the CAPS (which is regarded as the trustable NO<sub>2</sub> measurement technique without chemical interference) and revealed that a 5% bias could be caused by the chemiluminescence NOx monitor (will be published elsewhere). Therefore, we simulated  $P(O_3)_{net}$  by reducing and increasing the mixing ratios of NO<sub>2</sub> by 5% to check the interference caused by using the chemiluminescence NOx monitor to  $P(O_3)_{net}$ . The results show that increasing and decreasing NO<sub>2</sub> by 5% resulted in a decrease in  $P(O_3)_{net}$  by 1.64% and 3.68%, respectively. We added this explanation in the modified manuscript in pages 29-30, lines 661-671:"

"Furthermore, because the NO<sub>2</sub> data used here were measured by a commercially available

chemiluminescence NOx monitor, the NO2 and NOx mixing ratios would be overestimated due to NOz

interference (i.e., HNO<sub>3</sub>, PANs, HONO, etc.) (Dunlea et al., 2007). According to our test, the chemiluminescence technique could bias NO<sub>2</sub> by 5 % compared to the CAPS technique, which is regarded as a trustworthy NO<sub>2</sub> measurement technique without chemical interference. Therefore, we simulated the interference of NO<sub>2</sub> measured by a chemiluminescence NOx monitor in method I as follows: reducing and increasing the ambient NO<sub>2</sub> mixing ratios by 5 % in the 3<sup>rd</sup>-stage 4-min simulation in the reaction and reference chambers. The results show that increasing and decreasing NO<sub>2</sub> by 5 % resulted in a decrease in  $P(O_3)_{net}$  by 1.64 % and 3.68 %, respectively (as shown in Fig. S23), which is much smaller than the bias caused by  $P(O_3)_{net}$  in the reference chambers (which were 13.9 % and 22.3 % for method I and method II, respectively)."

And added Fig. S23 in the modified supplementary materials:





#### NO2.

Other minor comments:

Some mathematical formulae: The authors should use italic and roman letters correctly.

We have changed the font of the formula to italic and roman letters in the modified manuscript (in lines 56-57, 167, 292, 322, 364, 382, 395-396 and 481-483) and supplementary materials (in line 26, 28, 53-54, 206, 217, 235, 286 and Table S12).

L65: at 424 nm → at < 424 nm

We revised it in lines 65-66 in the modified manuscript: "The specific process of the photochemical reaction is the photolysis of NO<sub>2</sub> at < 420 nm to generate O(<sup>3</sup>P) atoms, thereby promoting the formation of O<sub>3</sub> (Sadanaga et al., 2017)."

L72: are proportional to  $\rightarrow$  affect?

We wanted to express that the surface deposition and advection of  $O_3$  **are proportional to** ambient  $O_3$  mixing ratios,  $[O_3]$ , which is mainly generated by local photochemistry (Carzorla et al., 2010). We have added the reference in line 73 in the modified manuscript.

Fig. 1: Why do the authors use critical orifices instead of mass flow controllers? Is the temperature of orifices controlled to keep a constant flow rate?

According to our tests, using critical orifices will make the sampling air flow rate in the reaction and reference chambers more stable than using mass flow controllers. We ensured that the air flow rates during the measurement period were constant in the reaction and reference chambers by checking the air flow rates every day during the campaign. Furthermore, the temperature of the orifices did not increase during the sampling time; thus, we assumed that the temperature of the orifices did not affect the air flow rate.

Fig. 1: Are inner walls of the reaction and reference chambers coated with Teflon? If so, please indicate the kind of the Teflon coat.

We did not coat the inner walls of the reaction and reference chambers with Teflon. The reasons are listed as follows: 1) Sadanaga et al. (2017) reported 8–10% dark losses of O<sub>3</sub> on uncoated quartz surfaces for a residence time of 21 min in the chambers, which is consistent with the reported dark loss of less than 5% for O<sub>3</sub>conditioned flow chambers and a residence time of 4.5 min in Sklaveniti et al. (2018). Sadanaga et al. (2017) indicated that the values of [NO<sub>2</sub>]<sub>out</sub> were in agreement with [NO<sub>2</sub>]<sub>in</sub> within one standard deviation under both dry (0%) and humidified (80%) conditions. The NO<sub>2</sub> loss is lower than 5% in both flow chambers and is close to 3% on average in Sklaveniti et al. (2018) under dark conditions. In our study, the wall losses of NO<sub>2</sub> were lower than 4% and 2% in the reaction and reference chambers, respectively, and the wall losses of  $O_3$  were both lower than 3% in the reaction and reference chambers, as shown in Tables S2 and S3; thus, we assumed this would not cause much bias in our measurement results under dark conditions. 2) Sklaveniti et al. (2018) thought Teflon coating seemed to remove or reduce the photolytic loss of ozone to a negligible level on their instrument because they thought the instrument design reported by Sadanaga et al. (2017) did not seem to be significantly impacted by a photolytic loss of ozone on the quartz flow chambers whose inner surface was coated with Teflon. In Sadanaga et al. (2017), wall losses of  $O_3$  were found to be approximately 10% for both chambers without clear Teflon coating, but wall losses decreased to less than 1.5% when the chambers were coated with Teflon. In our study, we calibrated photolytic  $O_3$  losses by performing a set of laboratory and ambient experiments (see sec. S2, pages 18-20 lines 207-237 in the supplementary materials), the results after photolytic  $O_3$  loss correction compensated for the photolytic  $O_3$  loss interference in the measurement results. 3) We tried to apply Teflon film to the inner walls of the reaction and reference chambers but found that there were some particles produced from the coated wall, which may have been due to bad coating techniques. According to previous studies, particles will take part in the RO<sub>2</sub>/HO<sub>2</sub> heterogeneous reactions, thus influencing photochemical  $O_3$  production. Taking these reasons into consideration, we did not coat the inner walls of the reaction and reference chambers with Teflon; instead, we did the photolytic  $O_3$  losses calibration to correct the data, which we think will make our measurement results more accurate.

Table S1: The authors should add standard deviation to average residence times.

Thanks for the suggestions. We have tested in total three sets of experiments at each flow rate, we now added the standard deviation to average residence times in Table S1, we found that for all air flow rates, the standard deviations were <1%, and this may be caused by different operation conditions during the experiments, thus we didn't take this into account when estimate the measured  $P(O_3)_{net}$  error, as described in page 17 lines 393-396:



Fig. S5: The regression lines have non-zero intercepts. These are significant? If so, why? The regression lines for ozone have negative intercepts. In this case, there are large losses of ozone in the high concentration of ozone? For NO<sub>2</sub>, how about relative humidity in the experiment? Is there no loss of NO<sub>2</sub> at high relative humidity?

Thanks for the questions, our answers for the different questions are listed as follows:

(1) The regression lines have non-zero intercepts. These are significant? If so, why? The regression lines for ozone have negative intercepts. In this case, there are large losses of ozone in the high concentration of ozone?

The non-zero intercept is not significant. We added the fittings without an intercept and compared the results with those with an intercept. We found that the wall losses of  $O_3$  and  $NO_2$  were not much different, and the wall losses affected by the fitting intercepts for  $NO_2$  and  $O_3$  at an air flow rate of 5 L min<sup>-1</sup> were all below 4% (as shown in Tables S4 and S5). According to the abovementioned results, we found that when  $O_3$  exhibited negative intercepts, the  $O_3$  wall losses were still below 4 % at ambient  $O_3$  mixing ratios of 0-200 ppb, which was not significant.

We added this explanation in the modified supplementary materials in pages 5-6, lines 73-79:

"The regression lines have non-zero intercepts but not significant. We added the regression fittings without intercept, and compared the regression fitting results with and without intercept (as shown in Figs. S5 and S6). We found that the O<sub>3</sub> and NO<sub>2</sub> wall losses were not much different (as shown in Tables S2 and S3), and the wall loss affected by the fitting intercepts for NO<sub>2</sub> (at ambient mixing ratios of 0-100 ppbv) and O<sub>3</sub> (at ambient mixing ratios of 0-200 ppbv) at the air flow rate of 5 L min<sup>-1</sup> were all below 4 % (as shown in Tables S4 and S5). We found that when the O<sub>3</sub> have negative intercepts, the O<sub>3</sub> wall losses are still below 4 %, which is not significant."



Figure S5: Relationship between (a,b) [O<sub>3</sub>]<sub>in</sub> and [O<sub>3</sub>]<sub>out</sub> and (c,d) [NO<sub>2</sub>]<sub>in</sub> and [NO<sub>2</sub>]<sub>out</sub> in the reaction and reference chambers with intercepts at the flow rates of 1.3, 2, 3, 4, and 5 L min<sup>-1</sup>, respectively, the solid lines represent the linear fitting of the O<sub>3</sub> or NO<sub>2</sub> mixing ratios at the inlet and outlet of the chambers.



Figure S6: Relationship between (a, b) [O3]<sub>in</sub> and [O3]<sub>out</sub> and (c,d) [NO2]<sub>in</sub> and [NO2]<sub>out</sub> in the reaction and reference chambers without intercepts at the flow rates of 1.3, 2, 3, 4, and 5 L min<sup>-1</sup>, respectively, the solid lines represent the linear fitting of the O3 or NO2 mixing ratios at the inlet and outlet of the chambers.

Flow rate of air (L min <sup>-1</sup> )	Wall losses of O <sub>3</sub> (%)		Wall losses of NO <sub>2</sub> (%)	
	Reaction chamber	Reference chamber	Reaction chamber	Reference chamber
1.3	<mark>2.0</mark>	2.0	<mark>4.0</mark>	2.0
2	0.0	1.0	<mark>4.0</mark>	1.0
<mark>3</mark>	0.0	1.0	2.0	<mark>0.0</mark>
<mark>4</mark>	0.0	0.0	<mark>0.0</mark>	<mark>0.0</mark>
5	<mark>0</mark>	0.7	<mark>0.3</mark>	<mark>0.6</mark>

Table S2. Wall losses of O3 and NO2 of the reaction and reference chambers with intercepts.

Flow rate of air (L min <sup>-1</sup> )	Wall losses of O <sub>3</sub> (%)		Wall losses of NO <sub>2</sub> (%)	
	Reaction chamber	Reference chamber	Reaction chamber	Reference chamber
1.3	<mark>3.0</mark>	3.0	3.0	2.0
2	1.0	1.0	3.0	<mark>2.0</mark>
<mark>3</mark>	0.0	2.0	2.0	0.0
<mark>4</mark>	1.0	1.0	0.0	0.0
<mark>5</mark>	2.0	2.0	<mark>0.0</mark>	<b>1.0</b>

### Table S3. Wall losses of O3 and NO2 of the reaction and reference chambers without intercepts.

## Table S4. NO<sub>2</sub> wall loss affected by the intercept.

	Wall loss affected by the intercept (NO <sub>2</sub> , %)		
Ambient NO <sub>2</sub> mixing ratios (ppbv)	Reaction chamber	Reference chamber	
20	2.0	2.0	
<mark>40</mark>	<b>1.0</b>	<mark>1.5</mark>	
<mark>60</mark>	<mark>0.7</mark>	<mark>1.3</mark>	
<mark>80</mark>	<mark>0.5</mark>	1.2	
<mark>100</mark>	<mark>0.4</mark>	<mark>1.2</mark>	

## Table S5. O3 wall loss affected by the intercept.

	Wall loss affected by the intercept (O <sub>3</sub> , %)		
Ambient O <sub>3</sub> mixing ratios (ppbv)	Reaction chamber	Reference chamber	
<mark>50</mark>	3.9	2.9	
<mark>80</mark>	2.1	2.2	
120	1.1	1.8	
<mark>160</mark>	<mark>0.5</mark>	<mark>1.6</mark>	
<mark>200</mark>	0.2	1.5	

(2) For NO<sub>2</sub>, how about relative humidity in the experiment? Is there no loss of NO<sub>2</sub> at high relative humidity?

We added this explanation in the modified supplementary materials in page 6,

lines 80-87: "Sklaveniti et al. (2018) found that the wall loss of NO<sub>2</sub> is significantly less than that of O<sub>3</sub> at higher humidity levels. However, in our O<sub>3</sub> photo-enhanced uptake experiments, the wall loss of O<sub>3</sub> was almost unaffected by humidity at a flow rate of 5 L min<sup>-1</sup>. We also tested the wall losses of NO<sub>2</sub> and O<sub>3</sub> in the chamber at a 5 L min<sup>-1</sup> flow rate at different humidities of 35-75 %, the detailed results are shown in Fig. S7 and S8, which shows that the variation in humidity effected the wall loss of NO<sub>2</sub> and O<sub>3</sub> by 0.03-0.12 % and 1.06-1.19 %, respectively, which is much smaller than the instrument detection error (which is 2 % at ambient NO<sub>2</sub> mixing ratios of 0-100 ppb), thus we didn't count this interference





Figure. S7 (a) and (c) represent the NO<sub>2</sub> wall loss at different humidities for the reaction and reference chambers, respectively, (b) and (d) represent the points fitted to all humidities, respectively. Uncertainty in the regression formula was one standard deviation (1σ).<sup>"</sup>

L218: low  $\rightarrow$  high?

We retained low. We wanted to say the UV film used in this study still has a low light transmission because it could transmit the lights with wavelength>390 nm.

L226: Why the transmittivity of HONO in the reference chamber is lower than that of  $O_3$ ? How about accuracy and precision of the actinic flux spectrometer?

We measured the transmittivities of all species as follows: we simulated sunlight illumination by adjusting the sunlight (SERIC XG-500B) to provide different intensities of illumination to study the solar UV transmittance through the reaction and reference chambers. The photolysis frequencies of NO<sub>2</sub>, O<sub>3</sub>, HONO, etc., inside and outside the reaction and reference chambers were measured using an actinic flux spectrometer (PFS-100; Focused Photonics Inc). The results are shown in Table S7.

The actinic flux spectrometer uses a quartz light receiver head to collect solar radiation from all directions and connects the collected light radiation via optical quartz fibers to the spectrometer, which obtains spectral information in a certain wavelength range and transmits the spectrometer data to the industrial control computer. The computer can convert the signal of the spectral scanning into the actinic flux F $\lambda$  and calculate the light by integrating the actinic flux with the known absorption cross-section  $\sigma(\lambda)$  and quantum yield  $\varphi(\lambda)$ . Therefore, there are no differences when measuring the transmittivities of HONO and O<sub>3</sub> on the method and technique aspects. The lower HONO transmittivities in the reference chamber than that of O<sub>3</sub> may be due to the UV protection Ultem film on the reference chamber blocking sunlight with wavelengths < 390 nm, as the spectral atlas of HONO was under 190-395 nm at 298 K (IUPAC, 2004), while the spectral atlas of O<sub>3</sub> was under 410-750 nm at 298 K (IUPAC, 2004). We added this explanation in the modified manuscript on page 11, lines 254-258:

"The reason for the lower HONO transmittivities in the reference chamber than that of  $O_3$  may be that the UV protection Ultem film on the reference chamber blocks sunlight at wavelengths < 390 nm, where the spectral atlas of HONO was under wavelengths of 190-395 nm at 298 K, while that of  $O_3$  was under wavelengths of 410-750 nm at 298 K (IUPAC, 2004, http://iupac. pole-ether.fr/)."

We realized that the transmittivities of HONO in the reference chamber in Baier et al. (2015) was also lower than that of  $O_3$ , which demonstrated that our testing results are reasonable. The actinic flux spectrometer has high resolution and sensitivity for actinic flux measurements. There are no deviations or accuracies in the results of the photochemical flux spectrometer, as there are no standards for reference, but the deviation in spectral resolution is  $\pm 0.8$  nm for the spectral band range (270-790) nm, which is small.

L228: agree  $\rightarrow$  agreement

We changed "agree" to "agreement" in the modified manuscript in line 246.

Table 1: What is Ultem? There are no definitions in the text.

We added the definition of "Ultem" in modified manuscript in lines 128-130 "... an ultraviolet (UV) protection Ultem film (SH2CLAR, 3 M, Japan) was used to cover the outer surface to block sunlight with wavelengths < 390 nm."

Table 1: The values of 0.019±0.011 should be shaded.

Indeed, we revised it in Table 1.

L272-278 (The airtightness of the reaction and reference chambers): It is hard to follow this section. I think the authors should explain using a schematic diagram for the experiment in the supplement.

Sorry for the confusion description. We have added a schematic diagram for the experiment in the supplementary materials (see Fig. S14), and modified the description accordingly in page13, lines 306-308 in the main manuscript:

"We also checked the airtightness of the reaction and reference chambers by passing through gases with different flow rates based on the schematic diagram shown in Fig. S14



Figure S14. Schematic diagram for investigating the airtightness of the reaction and reference chambers, where MFC1 could measure air flow rate and pressure at the chamber inlet, MFC2 could measure air flow rate and pressure at the chamber inlet.

L297-L300: For calibration of NO<sub>2</sub>, it is not appropriate to perform calibration of NO<sub>2</sub> using a NO<sub>2</sub> standard gas because of low reliability. Calibration should be performed using a gas-phase titration method using NO and O<sub>3</sub>.

We apologize for this mistake. We used NO<sub>2</sub> standard gas after we calibrated it using the gas-phase titration method using NO and O<sub>3</sub>. We used the CAPS NO<sub>2</sub> monitor reading as a transition value between the two to obtain the NO<sub>2</sub> standard gas and NO+O<sub>3</sub> mixing ratios corresponding to the same CAPS NO<sub>2</sub> monitor reading. This result showed that the purification of NO<sub>2</sub> standard gas is sufficient to calibrate the CAPS NO<sub>2</sub> monitor, and we added the related experiments in Fig. S15.

We have added this explanation in the modified supplementary materials in pages 17-18, lines 194-201:

"*Calibration of CAPS NO<sub>2</sub> monitor* CAPS NO<sub>2</sub> monitor was used to measure the NO<sub>2</sub> standard gas after we have calibrated it using the gas-phase titration method using NO and O<sub>3</sub>. We used the CAPS-NO<sub>2</sub> monitor reading as a transition value between the two to obtain the NO<sub>2</sub> standard gas and NO+O<sub>3</sub> mixing ratios corresponding to the same CAPS-NO<sub>2</sub> monitor reading. Results showed the purification of NO<sub>2</sub> standard gas was good enough to calibrate CAPS-NO<sub>2</sub> monitor, as shown in Fig. S15"



Figure. S15 Correlation between NO<sub>2</sub> standard gas and the NO<sub>2</sub> generated using the gas-phase titration method (NO +  $O_3$ ).

Reference: The authors should put the list into alphabetical order.

We have revised it in the modified manuscript and supplementary materials.

## References

- Baier, B. C., Brune, W. H., Lefer, B. L., Miller, D. O., and Martins, D. K.: Direct ozone production rate measurements and their use in assessing ozone source and receptor regions for Houston in 2013, Atmos. Environ., 114, 83-91, http://dx.doi.org/10.1016/j.atmosenv.2015.05.033, 2015.
- Cazorla, M., Brune, and W. H.: Measurement of ozone production sensor, Atmos. Meas. Tech., 3, 545-555, https://doi.org/10.5194/amt-3-545-2010, 2010.
- Dunlea, E. J., Herndon, S. C., Nelson, D. D., Volkamer, R. M., Martini, F. S., Sheehy, P. M., Zahniser, M. S., Shorter, H., Wormhoudt, J. C., Lamb, B. K., Allwine, E. J., Gaffney, J. S., Marley, N. A., Grutter, M., Marquez, C., Blanco, S., Cardenas, B., Retama, A., Villegas, C. R. R., Kolb, C. E., Molina, L. T., and Molina, M. J.: Evaluation of nitrogen dioxide chemiluminescence monitors in a polluted urban environment, Atmos. Chem. Phys., 7, 2691– 2704, https://doi.org/10.5194/acp-7-2691-2007, 2007.
- Sadanaga, Y., Kawasaki, S., Tanaka, Y., Kajii, Y., and Bandow, H.: New system for measuring the photochemical ozone production rate in the atmosphere, Environ. Sci. Technol., 51, 2871-2878, https://doi.org/10.1021/acs.est.6b04639, 2017.
- Sklaveniti, S., Locoge, N., Stevens, P. S., Wood, E., Kundu, S., and Dusanter, S.: Development of an instrument for direct ozone production rate measurements: measurement reliability and current limitations, Atmos. Meas. Tech., 11, 741-761, https://doi.org/10.5194/amt-11-741-2018, 2018.
- Zhou J, Sato K, Bai Y, et al. Kinetics and impacting factors of HO<sub>2</sub> uptake onto submicron atmospheric aerosols during the 2019 Air QUAlity Study (AQUAS) in Yokohama, Japan. Atmos. Chem. Phys., 21, 12243-12260, https://doi.org/10.5194/acp-21-12243-2021,2021.
- Zhou J, Fukusaki Y, Murano K, et al. Investigation of HO<sub>2</sub> uptake mechanisms onto multiple-component ambient aerosols collected in summer and winter time in Yokohama, Japan. J. Environ. Sci., 137, 18-29, https://doi.org/10.1016/j.jes.2023.02.030, 2023.
- Kotz, J. C., Treichel, P. M., Townsend, J. and Treichel, D.: Chemistry and Chemical Reactivity. 10th 19 ed., Cengage Learning, 2019.