Vertical distribution of black carbon and its mixing state in urban boundary layer in summer

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Abstract. The vertical distribution of black carbon (BC) as well as its mixing state is of great concern due to BC's strong regional climatic and environmental effects. In this study, vertical measurements were conducted through a moveable container based on a meteorology tower in Beijing urban area during June and July. A total of 112 vertical profiles (0-240 m), including the concentrations of BC, O_3, NO_x and the optical properties of aerosols, were obtained. Based on BC concentration, the vertical profiles could be classified into four categories: uniform, gradual decrease, sharp decrease, and sudden increase. The uniform type indicates strong vertical mixing with similar pollutant concentrations along the vertical direction. The gradual/sharp decrease types indicate stable vertical conditions with higher pollutant concentrations on the ground and lower concentrations at higher altitudes. Due to the strong radiation in summer, the vertical profiles exhibited a clear diurnal variation in which ~80% of profiles were uniform during the daytime and ~40%-90% of profiles were gradual/sharp decrease types at night. O_3 is an exception, and its concentration generally increases with height even under strong vertical mixing conditions. The size distribution of BC core varied slightly along the vertical direction, and the coating thickness, denoted by the diameter ratio between the BC-containing particle and BC core (D_p/D_c), of BC increased with height under stable conditions. Although the coating thickness could increase the absorption ability with an average absorption enhancement of 1.25 at 23:00, the vertical difference of D_p/D_c (2%) was much lower than that of BC concentration (~35%). The vertical variation of absorption ability was mainly caused by the variation of BC concentration. In addition, O_3 and D_p/D_c occasionally increased during 6:00-8:00 but remained stable during 8:00-10:00. Vertical mixing and transportation from upper heights, such as the residual layer, could significantly influence the pollutant properties on the surface during early mornings. This study exhibits a continuous vertical

picture of BC and its mixing state in urban areas, which would be helpful for understanding BC's regional environmental effect.

1.Introduction

35 Environmental effects associated with black carbon (BC) has attracted much attention (Ramanathan and Carmichael, 2008; Bond et al., 2013; Li et al., 2022). In addition to being one of the most toxic components of atmospheric pollutants (Dominguez-Rodriguez et al., 2015; Lin et al., 2011; Xue et al., 2021), BC could influence the boundary layer height, which can aggregate air pollution (Ding et al., 2016). Such depressing of the boundary layer is called the "dome effect" in which the BC in the upper boundary layer could heat the surrounding atmosphere, increasing the stability of the boundary layer. Through unmanned

40 aerial vehicle measurements, Wilcox et al. (2016) found turbulent kinetic energy reduction with increasing BC concentration leading to a shallower boundary layer. Further studies demonstrated that the dome effect is relevant to the vertical distribution of BC (Wang et al., 2018b). BC could even be attributed to the increase in the boundary layer if the BC concentration in the surface layer was too high. Thus, the precise measurement of BC vertical distribution is essential to evaluate its environmental effect.

45 The present studies mainly focused on the vertical distribution of BC mass concentrations (Wang et al., 2018a; Samad et al., 2020; Wu et al., 2021a; Guan et al., 2022). For instance, Lu et al. (2019) found that the vertical profile of BC mass concentration could be classified into four types, and the variation in the vertical profile was related to the boundary layer variation and local emissions. Based on the measured BC vertical concentration, the precise evaluation of BC's radiative property distribution requires additional BC related information, such as size distribution and mass absorption cross section, which is a property

50 that is commonly assumed in previous studies (Wu et al., 2021a). In fact, the optical properties could be significantly influenced by BC's microphysical properties. For example, the coating material could increase the absorption ability through the lensing effect. Through laboratory study, Shiraiwa et al. (2010) found that BC's absorption ability after coating could reach two times greater than BC without coating. Such absorption enhancement of BC due to coatings was also observed in the ambient with typical factors of 1.0-1.5 (Liu et al., 2017). In addition, the size distribution and morphology of BC have shown to be important

55 factors related to BC's optical and radiative properties (He et al., 2015; Matsui et al., 2018; Luo et al., 2022; Liu et al., 2020b). Such microphysical properties of BC vary greatly from the reported ground-based measurements (Shiraiwa et al., 2008; Pan et al., 2017; Liu et al., 2020c; Zhao et al., 2022), generally depending on the varied emission sources and complex aging processes. However, vertical measurements of such microphysical properties are still rare.

The present vertical measurement approaches include tower, balloon, aircraft, and unmanned aerial vehicle (UAV) approaches.

60 UAVs have the advantages of high flexibility, but large precise instruments cannot be loaded by UAVs due to their low carrying capacity. UAV measurements generally focus on mass concentrations by using light instruments or sensors (Liu et al., 2020a; Lu et al., 2020; Pikridas et al., 2019; Kwak et al., 2020). Aircraft is another vertical measurement approach that combines flexibility and preciseness. Limited vertical measurements of BC and its mixing state using the aircraft approach have been

conducted (Katich et al., 2018; Schulz et al., 2019; Schwarz et al., 2013; Schwarz et al., 2017). Generally, the BC concentration decreases with altitude, but the vertical distribution of size distribution and coating thickness could vary a lot due to the regional transportation (Hu et al., 2020), pollution levels (Ding et al., 2019; Zhao et al., 2020), biomass burning plumes (Ditas et al., 2018) and other factors. However, as summarized in Table S2, aircraft measurement mainly focuses on the vertical difference between boundary layer, upper troposphere and lower stratosphere. The data points in the boundary layer were limited due to flight height restrictions. In addition, continuous aircraft vertical measurements are difficult to conduct due to cost issues. The balloon measurements (Liu et al., 2012; Wang et al., 2022) have requirements for the location, which is typically conducted in rural or remote areas. Among the measurement methods, tower measurements are the most suitable way to study the vertical distribution of BC microphysical characteristics as well as its relationship with the boundary layer in urban areas considering the cost, accuracy, and observation continuity.

BC emissions are substantial in urban areas (Li et al., 2017), exerting a significant impact on the urban environment and climate. The microphysical properties of BC may be highly variable under the influence of fresh emissions, aging and regional transportation. Due to the advantages of high data continuity, low cost and high observation frequency, a tower-based observation campaign was conducted in Beijing during summer. The vertical distributions of BC concentration, size distribution and mixing state were measured. In addition, aerosol absorption, extinction, and pollutant gases (CO, NO_2 and O_3) were measured simultaneously. Such continuous vertical measurements of BC microphysical characteristics in the urban boundary layer are rare. This research provides insight into the vertical distribution of BC and its properties in the boundary layer and is helpful to reduce the uncertainties in BC environmental and climatic evaluations.

2. Method

2.1 Measurements

The observation was conducted on the tower campus of the Institute of Atmospheric Physics, Chinese Academy of Sciences (longitude: 116.37°E; latitude: 39.97°N) from June 17 to July 16, 2022. The site is located in the urban area in Beijing and is approximately 50 m (380 m) away from the closest road (highway). More detailed descriptions of the observation site can be found in the previous literature (Sun et al., 2016; Pan et al., 2017; Xie et al., 2019). All instruments were placed in a moveable container, and an air conditioner and a UPS were used to supply stable temperature and power for the instruments inside. The container is able to move along the cable supporting the meteorology tower. During a normal run of a vertical observation, the container started at the ground level and rose at a nearly constant speed (8 m/s) to the 240 m height of the tower, stayed at 240 m for 45 minutes and descended to the ground at a speed of 8 m/s. The routine vertical observation was conducted 4 times a day (7:00, 11:00, 18:00 and 22:00), representing the morning, noon, afternoon, and night conditions of the atmosphere. The vertical observation was cancelled during the times with heavy rain or high wind, and the moveable container stayed at the ground when the vertical observation was not conducted. During the whole campaign, 112 vertical profiles were obtained, and 105 vertical profiles were analysed through data quality control.

The concentration of pollutant gases (NO₂, O₃, CO), the concentration and mixing state of BC and optical properties (light extinction and absorption) of aerosol were measured through the instruments inside the container. In addition to the container, meteorological parameters, including temperature (T), relative humidity (RH), wind speed (WS) and wind direction (WD), at 15 different heights were obtained through the instruments equipped on the tower. All instrument models and manufacturers are listed in Table S3.

The height of boundary layer height (BLH) was determined (Tang et al., 2015; Tang et al., 2016) using the vertical profile of the attenuated backscatter coefficient measured by a single-lens ceilometer (CL51, Vaisala, Finland) at the same observation site. The cloud amount data was extracted from the ECMWF Reanalysis v5 data product (ERA5, https://cds.climate. copernicus.eu/cdsapp#!/home) with spatial resolution of 0.25°*0.25° and time resolution of 1 hour.

2.2 Data analysis

2.2.1 Coating thickness and size distribution of BC

The BC microphysical properties were obtained through a single particle soot photometer (SP2, Droplet Measurement Technology, Inc., Boulder, CO, USA). SP2 obtained the mass equivalent diameter (D_c) of the BC core in every single BC-containing particle, and the optical diameter (D_p) of BC-containing particles was derived through the Leading edge only (LEO) fitting method (Gao et al., 2007) and a look-up table (table of precalculated scattering cross section of BC-containing particles with variable D_c and D_p through Mie scattering theory) method (Taylor et al., 2015). The refractive indices of the BC core and coatings in the derivation of D_p were 2.26-1.26i (Moteki and Kondo, 2008) and 1.48-0i (Taylor et al., 2015), respectively, at a wavelength of 1064 nm. The optical properties of BC is less sensitive to its morphology at wavelength of 1064 nm and Liu et al. (2015) estimate the bias of D_p/D_c due to the core-shell assumption in the look-up table method would be less than 6%.

Aquadag (Acheson Inc., USA) and polystyrene latex sphere (PSL, Nanosphere Size Standards, Duke Scientific Corp., USA) aerosols were used to calibrate the incandescence and scattering signals of SP2. Because of the higher incandescence sensitivity of ambient BC than Aquadag (Moteki and Kondo, 2010), the calibration was scaled following the recommendation from (Laborde et al., 2012). The data acquisition mode of SP2 was set to record 1 particle every 2 particles to guarantee the high resolution of the vertical profile. More specific details regarding calibration and data processing can be found in our previous literature (Liu et al., 2020b). In general, the size distribution of the BC core (D_c) and coating thickness (D_p/D_c) could be determined through SP2. The mass median diameter (MMD) was used to quantify the size distribution of the BC core and calculated from the D_c distribution below and above which the BC mass was equal (Liu et al., 2019a). The D_p/D_c of bulk BC aerosols was calculated using the D_p/D_c of every single BC particle in a certain time window:

$$\frac{D_p}{D_c} = \sqrt[3]{\frac{\sum_i D_{p,i}^3}{\sum_i D_{c,i}^3}} \tag{1}$$

125 The hygroscopicity of BC-containing particles was calculated based on the volume-weighted Zdanovskii-Stokes-Robinson (ZSR) rule (Stokes and Robinson, 1966) by assuming that the hygroscopicity parameters of BC and coatings were 0 and 0.3, respectively (Pringle et al., 2010). A detailed calculation of the hygroscopicity can be found in (Hu et al., 2021a).

2.2.2 Optical properties

The absorption efficiency (b_{abs}) at a wavelength of 880 nm was directly measured by an Aethalometer (AE33, Magee Scientific
130 Corp.) The scattering efficiency (b_{sca}) at a wavelength of 870 nm was measured by a photoacoustic extinctiometer (PAX, Droplet Measurement Technologies). A PM$_{2.5}$ cyclone with a supporting pump was placed before AE33 and PAX; thus, the measured b_{abs} and b_{sca} are characteristics of PM$_{2.5}$. The single scattering albedo (SSA) at a wavelength of 870 nm was calculated as follows:

$$SSA = \frac{b_{sca,870\ nm}}{b_{sca,870\ nm} + b_{abs,870\ nm}} \qquad (2)$$

135

The b_{abs} at a wavelength of 880 nm was transferred to b_{abs} at a wavelength of 870 nm through Equation 3:

$$b_{abs,870\ nm} = b_{abs,880\ nm} * \left(\frac{870}{880}\right)^{-AAE} \qquad (3)$$

The absorption Ångström exponent (AAE) is assumed to be 1 in the calculation since brown carbon has little effect at long wavelengths.
140 The absorption enhancement (E_{abs}) of BC aerosols at a wavelength of 550 nm is calculated based on the information of D_c and D_p from SP2's measurement by using the Mie core-shell approach. The refractive indices used in the calculation are 1.95+0.79i for the BC core (Bond and Bergstrom, 2006) and 1.50+0i for the coatings (Liu et al., 2015). The E_{abs} may be overestimated in this study due to the core-shell assumption in the calculation (Liu et al., 2017; Wu et al., 2018; Liu et al., 2019b) because of the complex morphology of BC (Wang et al., 2017; Hu et al., 2021). Liu et al. (2020c) estimated that such overestimates could
145 be ~17% in summer in Beijing.

3. Results

3.1 Overview of the observation

The time series of the major observation parameters during the campaign are illustrated in Fig. 1. The height of the moveable container was also shown in Fig. 1a. Thus, the time series in Fig. 1 also include information on the vertical distribution of
150 pollutant concentrations. The BLH ranged between 50-2000 m (Fig. 1f) during the observation. The variation of BLH followed a clear diurnal pattern (Fig. S1) with an average value of 400-500 m at night and 1000-1200 m at noon. The observation height (0-240 m) was generally in the mixing layer. However, occasionally at night and in the morning, the BLH could be below 240 m, and the moveable container could detect the BC properties in the residual layer in such cases.

155   The average concentration of BC was 0.34 μg/m^3, and the average mixing ratio of O$_3$ was 33 ppb. The relative humidity and temperature were high during the summer, with average values of 67% and 27 °C, respectively. In general, the air condition is relatively clean during the observation period due to the frequent rainy event, except for a heavy O$_3$ pollution event that occurred on the 25th of June with a maximum hourly mixing ratio of 124 ppb.

The mass size distribution of the BC core followed a lognormal distribution with a mass median diameter (MMD) of 173 nm

160   and a geometric standard deviation (GSD) of 1.57, as shown in Fig. S2, which is consistent with the MMD value (171 nm) observed in Beijing in the summer of 2018 (Liu et al., 2020b). The MMD value remained stable between 160-190 nm. The coating thickness (D$_p$/D$_c$) of the BC bulk varied largely from 1.1 to 1.6, with an average value of 1.33. The coating thickness of BC could be as high as 1.4-1.5 even in the relatively clean conditions during July 1st to the 5th. Fig. S3 exhibits the vertical meteorological conditions. The wind direction from July 1st to 5th was dominated by the south wind. The wind was mainly

165   from the north during other periods. The transportation of aged BC from the south may lead to an increase in coating thickness even under clean conditions. Such an increase in D$_p$/D$_c$ was also found from the 10th of July to the 12th, during which the wind also came from the south. Thus, the coating thickness may be highly dependent on the air mass. Zhang et al. (2018) reported that the transportation of aged BC from regional sources outside Beijing in winter and BC from transportation could account for 63% of total BC aerosols. Such transportation may also be efficient in summer. It is also noted that the D$_p$/D$_c$ value

170   could increase quickly from 1.1 to 1.4 during the ozone pollution day (25th June). Previous studies have suggested that BC aging could occur more rapidly in the presence of increased O$_3$ concentrations (Zhang et al., 2022). Furthermore, Liu et al. (2020c) found the D$_p$/D$_c$ could increase at a faster rate during the daytime under high O$_x$ condition. These findings suggest that severe photochemical processes and high atmospheric oxidation rates could contribute to the aging of BC and the subsequent increase in its coating thickness.

175   **3.2 Classification of the vertical profile**

Fig. S4-Fig. S22 exhibit all measured vertical profiles of different parameters during the observation. Based on the vertical profiles of BC concentration, the profiles were classified into four categories, and Fig. 2 shows the typical profiles of the four categories.

3.2.1 Uniform type

180   The vertical profile at 12:00 on 30th June was selected to represent the uniform type. As shown in Fig. 2a, the concentration of BC varied little from the surface to 240 m in height in this type. In addition to BC concentration, the microphysical properties (such as MMD and D$_p$/D$_c$) and optical properties (such as b$_{abs}$, b$_{sca}$ and SSA) also varied little. The uniform aerosol vertical distribution indicates sufficient mixing in the lower boundary layer. In this case, the temperature of the surface was ~4°C higher than that at 240 m height. Such a large temperature difference would promote the vertical motion of the air mass and

185   contribute to vertical mixing. Although vertical mixing is efficient in this case, the ozone concentration was found to be lower in the near-surface area, and the vertical profile exhibits a uniform distribution above ~50 m height. The presence of substantial

NO emissions from traffic sectors would decrease the concentration of ozone through the titration reaction between ozone and NO, leading to the decreasing trend of ozone concentration with height.

3.2.2 Gradual decrease type

190 The vertical profile at 18:00 on 2nd July was selected to represent the gradual decrease type. In this type, the BC concentration gradually decreases with height. Such a vertical profile is mainly attributed to the stable boundary layer as well as weak convection and turbulence. The continuous emission of BC from the surface leads to a higher concentration near the surface. Fresh BC emissions would also lead to a decrease in D_p/D_c near the surface, resulting in an increasing trend of D_p/D_c with height. The aging process is efficient in increasing the coating thickness but not the size of the BC core. The MMD value

195 remained stable at all heights in this case. Studies have shown that the size of the BC core varies depending on the emission source. For instance, BC emitted from traffic sources typically has a lower MMD of about 150-180 nm, while that from solid fuel burning sources has a higher MMD of about 170-230 nm (Schwarz et al., 2008; Pan et al., 2017; Holder et al., 2014). In Beijing and London, (Liu et al., 2014; Liu et al., 2019a) found that the MMD of BC was much larger in winter than in summer, likely due to increased solid fuel burning for heating purposes. The nearly stable MMD value in this case may suggest similar

200 sources of BC in the lower boundary layer, but the BC in the upper heights had a longer aging time (larger D_p/D_c). The concentration of O_3 increased with increasing height at a sharper slope than that in the uniform case.

3.2.3 Sharp decrease type

The vertical profile at 23:00 on 27th June was selected to represent the sharp decrease type. This type is also referred to as the 'stratified type' in other studies (Guimaraes et al., 2019; Lei et al., 2021). The concentration of pollutants decreased sharply at

205 a certain height but remained uniform or slightly decreased with height above a certain height. Profiles of this type mainly appeared in the early morning and midnight. The D_p/D_c was larger and $\triangle BC/\triangle CO$ was smaller at the upper height, indicating more aged BC at the upper height.

The vertical profiles at 22:00 on 27th June and 23:00 on 27th June are simultaneously shown in Fig. S8. As shown, the vertical profile at 22:00 exhibited a gradual decreasing trend. However, the profile changed into a sharp decrease type 1 hour later.

210 Comparing the vertical profiles at these two hours, the BC concentration was similar above 200 m, but the concentration significantly increased below 200 m at 23:00. The D_p/D_c was also similar above 200 m and significantly decreased below 200 m at 23:00. The BC concentration and D_p/D_c characteristics in these two profiles suggest that freshly emitted BC was trapped at lower heights (below 200 m) at 23:00. Such a stratified vertical distribution is also found in previous studies (Wang et al., 2018a; Guimaraes et al., 2019; Lei et al., 2021). Lei et al. (2021) found that in the stratified type, the BLH measured by

215 ceilometer was relatively low (~110-250 m), which is consistent with the findings of this study. The BLH measured by ceilometer in the stratified type ranged between 100-300 m, which is lower than the average value of ~400-500 m. On the night of 27th June, the BLH measured by ceilometer even reached 80 m. It is known that the absolute value of BLH determined from different methods can vary significantly. Wang et al. (2018a) found that in the stratified type, the BLH determined from

ceilometer would always be higher than the sudden concentration transition point and suggested that it is more precise to
220 determine the BLH by using the sudden concentration transition point in such cases. In this study, the top of the BLH was regarded as the sudden concentration transition point in the sharp decrease type, and the profiles were used to investigate the differences in pollutants between the nocturnal boundary layer and residual layer in the following section.

3.2.4 Sudden increase type

225 The vertical profile at 11:00 on the 11th of July was selected to represent the sudden increase type. As shown in Fig. 2d, The BC mass concentration rapidly increased from 0.4 to 1.2 μg/m^3 at 80 m above the ground and quickly decreased to the background level above 90 m. The concentration of NO$_x$ and the mixing state of BC also exhibited a sudden change. The MMD and D$_p$/D$_c$ experienced sudden decreases, while the NO$_x$ concentration increased sharply. The characteristics of MMD, D$_p$/D$_c$ and NO$_x$ indicate that the sudden increase in this case may come from fresh emissions from traffic sources, since BC
230 from traffic sources generally has a small BC core and thin coatings (Liu et al., 2017; Holder et al., 2014), and the NO$_x$ emissions from traffic sources are also substantial.

Among the 7 sudden increase cases, 4 sudden increases occurred at ground level, and 3 sudden increases were found in mid-air. In addition to the emissions from traffic sources (characterized by the low MMD, D$_p$/D$_c$ and simultaneous increase in NO$_x$), the emissions from other sources could also lead to a sudden increase. The sudden increase that occurred at 11:00 on the 9th
235 of July, as shown in Fig. S16, was accompanied by an increase in MMD, a decrease in D$_p$/D$_c$ and no significant variation in NO$_x$ concentration. The sudden increase in this case may be influenced by the emission from solid fuel burning, since the MMD is generally higher for BC from solid fuel burning than that of fossil fuel burning (Schwarz et al., 2008; Pan et al., 2017; Liu et al., 2019a).

3.2.5 Other profiles in special cases

240 Fig. S7 shows the profiles on the ozone pollution day (maximum hourly ozone mixing ratio reached 124 ppb), the vertical profiles of BC, including its concentration, mixing state and ozone presented in a uniform type. Since the formation of ozone is always accompanied by high temperature and radiation leading to highly sufficient mixing in the boundary layer, the uniform vertical distribution of pollutants may be the common situation on ozone pollution days.

Another special profile is the profile (19:00 1st July, shown in Fig. S12) after a heavy rainy event. The moveable container
245 started rising at 18:00 and reached a height of 240 m at 18:30. When the container stayed at the tower, heavy rain happed. The rainy event lasted for nearly 80 minutes, and the container started moving to the ground as soon as the rain ended. Thus, the vertical profile between 18:00 on 1st July and 19:00 on 1st July could be compared to analyze the effect of wet scavenging. As shown, these two vertical profiles both belong to the uniform type, indicating that the wet scavenging efficiency may be vertically similar in the lower boundary layer or that there may be quick vertical mixing during the daytime in summer. It is
250 noted that the MMD decreased largely from 190 nm to 170 nm within 1 hour. The bulk D$_p$/D$_c$ value remained invariant, but

the D_p/D_c for D_c = 180 nm decreased from 1.32 to 1.25. The variation in MMD and D_p/D_c suggests that wet scavenging may prefer to remove large BC cores with thicker coatings. The preference of wet scavenging for BC with large sizes has also been reported in previous research (Liu et al., 2020b , Wang et al., 2015 , Taylor et al., 2014). As also exhibited in Fig. 1c, the MMD tended to decrease during other rainy events when the container was at ground level, but the D_p/D_c showed less variation.

3.3 Diurnal variation in the vertical profile

Fig. 3 shows the number fractions of vertical profile types at different times during the day. The uniform type is the dominant type, accounting for ~ 80% during the daytime (8:00–18:00). The number fraction of the uniform type started dropping at 19:00 and only accounted for 14% at 23:00 and rose back after sunrise. In contrast, the number fraction of the gradual decrease type started rising at night and dropped back ~10% during the daytime. The diurnal variation in the boundary layer (Fig. S1) leads to such vertical profile type changes. Due to the much higher heat capacity of the surface than that of air, the temperature of the surface responds to radiation at a much faster rate, leading to the boundary layer's thermal structure and stability change, especially in summer with strong radiation. Thus, the vertical motion and mixing during the daytime is more severe than that at night, and a uniform profile is the most common case during the daytime. In fact, several gradual decrease cases observed during the daytime, such as 11:00-12:00 on 2nd July (Fig. S13) and 19:00 on 27th June (Fig. S8), generally occurred on cloudy days when the radiation was relatively weak. As shown in Fig.1, the total amounts of clouds were nearly 100% for the two cases and the radiation was relatively low compared with that of the entire observation except several rainy days. The sudden increase type always occurred during the daytime, partially caused by the more frequent human activity at the time, resulting in occasional plume transportation to the measurement site. The sharp decrease type, indicating the low nocturnal boundary layer height, occurred five times during the whole observation and was found only at midnight (23:00) or the early morning (6:00-7:00).

In addition to the BC concentration, the average vertical profiles of other parameters at 7:00, 12:00, 18:00 and 23:00 are counted to further study the diurnal variation, as illustrated in Fig. 4 (the averaged profiles with error bars could be found in Fig. S27). The average value and the shape of vertical profiles both have clear diurnal changes.

For the average value, the mass concentration of BC is higher and D_p/D_c is lower at night and in the morning than during the daytime. The changes are mainly attributed to the fresh BC emissions and low night time boundary layer height. Fresh BC tends to accumulate in the nocturnal boundary layer. The development of the boundary layer and photochemical aging with the condensation process during the daytime would result in a decrease in the concentration and an increase in D_p/D_c. Since O_3 is the product of photochemical reactions, the concentration of O_3 is higher at noon and significantly decreases at night. The concentration of NO_x exhibits the opposite variation trend with O_3.

For the shape of the vertical profile, the average profiles of all parameters during the daytime (12:00 and 18:00) generally follow a uniform distribution except O_3. The O_3 concentration exhibits a slight decrease at the surface level at 12:00, and the degree of decrease extends at 18:00.

Compared with the uniform shape during the daytime, the vertical gradients of the average profiles at night and in the morning (7:00 and 23:00) are larger. For BC aerosols, the average concentration decreases with height at a rate of -0.075 μg/m^3 per 100 m at 23:00. The D_p/D_c is nearly uniform below 160 m and slightly increases with height above 160 m. The transition at 160 m is mainly affected by the two sharp decrease profiles at 23:00, which both occurred at ~160 m height. Excluding the two sharp decrease types, the average D_p/D_c shows little variation with height. The MMD also remained nearly invariable at 0-240 m. The variation tendency of the BC concentration, D_p/D_c and MMD at 7:00 was similar to that at 23:00, but the vertical gradient at 7:00 was lower. The vertical profile of NO_2 shows a very similar variation trend with that of BC concentration, partially caused by the similar emission sources of these two pollutants in the urban area. The concentration of NO was near zero at 12:00, 18:00 and 23:00 and significantly increased at 7:00 due to the low O_3 concentration at the time as well as the large emissions from the morning commute. The vertical profile of NO at 7:00 shows a decreasing trend with increasing height. The average vertical profile of O_3 increased with increasing height at 7:00 and 23:00.

3.4 Pollutant properties in the residual layer

The BC properties, O_3 and NO_x at the 240 m height and ground are counted for the 5 sharp decrease profiles to study the pollutant difference between the ground and residual layer. For BC aerosols, the concentration is much higher on the ground than that in the residual layer. The BC concentration on the ground was 84.6% higher than that in the residual layer at 23:00 on the 8th of July. BC in the residual layer generally had thicker coatings, indicating a higher aging degree. The maximum difference in D_p/D_c was found at 23:00 on 27th June, and the D_p/D_c in the residual layer was 8.4% higher than that on the ground. The difference in BC concentration and D_p/D_c in the two layers could be explained by the accumulation of fresh BC in the nocturnal boundary layer. In addition, as exhibited in Fig. S23, the number frequency of another BC type with an apparently higher D_p/D_c, denoted by the red circle, increased at 240 m height. The transportation of aged BC from other areas in the residual layer may also lead to an increase in D_p/D_c. The MMD values were relatively stable (~170-190 nm) on the ground, while the MMD in the residual layer was more variable, ranging between 210 nm and 165 nm. Since the BC core is inertial in the atmosphere, the size distribution of the BC core is mainly affected by the emission sources and has been used as an index for source apportionment (Wu et al., 2017). The stable MMD may indicate stable emission sources in the boundary layer, and the MMD value on the ground is close to the reported MMD value observed at the urban site, which is mainly influenced by traffic emissions (Liu et al., 2020c; Wu et al., 2021b). The variable MMD value in the residual layer may indicate more complicated sources. Other physical properties, such as absorption ability and hygroscopicity, are calculated and exhibited in Fig. 5e-f. Due to the thicker coating, BC in the residual layer generally has a higher absorption ability and hygroscopicity. At 23:00 on June 27th, the absorption enhancement (E_{abs}) in the residual layer is 8.6% higher than that on the ground, and the critical supersaturation point in the residual layer is 25.0% lower than that on the ground.

The O_3 concentration showed an opposite and much more significant difference between the ground and residual layer compared with the concentration of BC. In the case at 23:00 on June 27th, the concentration of O_3 at 240 m height was 226.6%

315 higher than that on the ground, since the higher NO_x concentration on the ground would decrease the O_3 in the nocturnal boundary layer at night.

Previous studies found that the higher O_3 concentration in the residual layer could increase the O_3 concentration on the ground in the early morning with the development of the mixing layer and entrainment of O_3 from the residual layer (Xu et al., 2018; Hu et al., 2018). The transportation of O_3 from the residual layer to the ground was also observed in this study. As shown in

320 Fig. 6, the O_3 concentration was nearly zero before 6:00. Two vertical measurements were conducted during 6:00-8:00, and a much higher O_3 concentration was observed in the upper height. The O_3 concentration significantly increased at 8:00 when the moveable container returned to the ground compared with the O_3 concentration at 6:00, but the concentration stayed stable between 8:00 and 10:00 and rose again after 10:00. The stable concentration of O_3 between 8:00-10:00 means that the photochemical reaction may not be very efficient before 10:00 in this case. The increase in the O_3 concentration on the ground

325 between 6:00 and 8:00 may not be mainly attributed to photochemical reactions but to transportation from the upper height with high O_3 concentrations. Such development of the boundary layer and vertical mixing was caused by the quicker heating of the ground (Fig. 6e) and would decrease the O_3 concentration in the upper heights in turn. As denoted by the grey dashed line in Fig. 6d, the O_3 concentration remained invariable at the beginning and then started to decrease when the container was located at 240 m height during 6:30-7:15.

330 A similar variation tendency was also found for the D_p/D_c value, as shown in Fig. 6c. The aged BC in the residual layer could also increase the D_p/D_c on the ground in the early morning through entrainment and vertical mixing, even though the fresh BC emissions from traffic sources are substantial at the time. Another typical case of the increase in D_p/D_c and O_3 through vertical mixing occurred on 11th July. A similar increase in the early morning and stable condition at the later hours was found as exhibited in Fig. S24. In the cases with enough photochemical reactions, stable conditions after the early morning increase

335 may not be found, and the concentrations of O_3 and D_p/D_c continuously increase from the early morning. However, it is reasonable to infer that vertical mixing could also contribute to the increase in O_3 and D_p/D_c in the early morning, since O_3 and D_p/D_c were higher in the residual layer in most cases. From the profiles of O_3 and D_p/D_c on July 13th (Fig. S20), the pollutants from the residual layer firstly influenced at the top of boundary layer during the daytime denoted by the large difference of O_3 and D_p/D_c between 240 m and the surface at 6:00. Then, with the development of vertical mixing due to the sun rise, the

340 pollutants at higher height reached to the surface. The O_3 and D_p/D_c at the surface increased and vertical profiles became uniform at 7:00. Such phenomenon could be also reflected on the averaged profiles (Fig. 4), since the averaged value of O_3 and D_p/D_c were higher at 240 m at 7:00. The presence of aerosol, especially the absorbing aerosols, would depress the development of mixing layer (Ding et al., 2016; Wang et al., 2018b). The higher D_p/D_c of BC in the residual layer may amplify the absorbing of BC in the upper height in the early mornings, resulting in a more stable structure and deteriorating the air

345 pollution.

The distinguished pollutants in the residual layer could also influence the chemical composition of particulate matter in the boundary layer (Lei et al., 2021). Due to the limitation of the observation height, the pollutant conditions at the top of the boundary during the daytime layer could not be detected. The D_p/D_c above the boundary layer is reported to be higher than

that on the ground under clean conditions (Ding et al., 2019). Liu et al. (2019b) reported rapid BC aging at the top of the

350   boundary layer in summer due to intensive actinic flux. Besides the high actinic flux, the higher O_3 concentration may also contribute to the D_p/D_c increase in the upper height. Thus, the D_p/D_c increase in the boundary layer due to vertical transportation may not only occur in the early morning, but entrainment from the lower troposphere may also lead to a D_p/D_c increase in the daytime.

3.5 Optical properties of black carbon

355   The optical properties of BC aerosols are of most concern due to BC's strong absorption ability. Fig. S25 exhibits the relationship between BC concentration and b_{abs} during the whole observation period. In general, b_{abs} increased with BC concentration. For the same BC concentration, b_{abs} increased with D_p/D_c through the so-called "lensing effect". The "lensing effect" also appeared in the vertical distribution. As shown in Fig. 4, the vertical distribution of E_{abs} shows a similar vertical variation in D_p/D_c with a uniform distribution during the daytime and a slight increase at 150-240 m at night. The E_{abs} at 240

360   m height is ~1-2% higher than that on the ground on average at night. The vertical difference in E_{abs} caused by thicker coatings is significantly lower than that of the BC concentration, as shown in Fig. 7. The vertical difference in BC concentration reached -40% - 40% but was only -5%-5% for D_p/D_c. Thus, the vertical distribution of b_{abs} is mainly determined by the variation in BC concentration, as proven by the similar average vertical profiles of BC concentration and b_{abs}. For the averaged profiles at 23:00 (the profiles with most significant vertical difference), b_{abs} at 240 m is 29.3% lower than that at ground. Such decrease

365   could reach 30.3%, if the same coating thickness was used at all heights. In other words, the increase in absorption enhancement at upper height due to coatings could slightly decrease the vertical difference of b_{abs}, but couldn't compensate for the decrease of b_{abs} at upper height due to the much more significant decrease of BC concentration.

Single scattering albedo (SSA) is another key optical parameter for aerosols since the influence of aerosols on surface temperature could change from cooling to heating when the SSA is lower than the critical point of 0.90-0.93 (Hansen et al.,

370   1997; Cook and Highwood, 2004). Although the "lensing effect" could increase the absorption ability for BC-containing aerosols, the scattering of BC-containing aerosols simultaneously increases with D_p/D_c, leading to an increase in SSA, as shown in Fig. S26. For the bulk aerosols (including BC and non-BC aerosols) in the ambient atmosphere, the SSA also increases with increasing D_p/D_c (Fig. S25b). The increase in SSA for bulk aerosols with D_p/D_c may be influenced by the higher SSA for BC-containing particles with higher D_p/D_c. Another reason is that the BC fraction in bulk aerosols may be lower in

375   the airmass with a higher D_p/D_c, since D_p/D_c is an indicator for secondary aging processes and BC is the product of primary emissions. The vertical distribution of SSA exhibited a uniform type during the daytime. At night time, b_{sca} and b_{abs} generally decrease with increasing height, but b_{abs} decreases at a quicker rate, resulting in an increase in SSA with height at night. Such an SSA variation trend has also been reported in previous vertical measurements (Li et al., 2022b).

4 Conclusion and remarks

380 The mixing state of black carbon (BC) is crucial for estimating its climatic and environmental impact, but there is limited research on the vertical distribution of BC mixing state. In the summer of 2022, continuous vertical measurements (0-240 m) of BC mixing state were conducted in the urban area of Beijing using a meteorological tower. The goal of this study was to fill the observation gap of BC mixing state distribution in an urban area with significant fresh BC emissions and explore the aging process of BC at the initial emission stage. The major findings are summarized below:

385 1) Based on the vertical distribution of BC concentrations, the profiles were classified into four categories: uniform, gradual decrease, sharp decrease, and sudden increase. The uniform type indicates sufficient vertical mixing in the boundary layer and accounts for ~80% of the total profiles during the daytime. The gradual decrease and sharp decrease types suggest stable vertical conditions and account for ~40-90% of the total profiles at midnight and early morning. The vertical variation of O_3 is different from other pollutants, and the O3 concentration generally increases with height even in the uniform type for other

390 pollutants due to the large emission of NO on the ground.

2) The mass median diameter (MMD) and the diameter ratio between the BC-containing particle and BC core (D_p/D_c) were used to represent the size distribution of the BC core and coating thickness. The MMD and D_p/D_c followed a generally uniform vertical distribution in the boundary layer. Under some stable conditions, such as a nocturnal boundary layer or cloudy days, D_p/D_c increased with height. However, the vertical difference in BC concentration is much more significant than that in D_p/D_c

395 under stable conditions. For the average profiles at 23:00, the BC concentration at 240 m height could be 34.6% lower than that on the ground, but the D_p/D_c at 240 m height was only ~2% higher than that on the ground, leading to 1-2% absorption enhancement (E_{abs}) difference. Although coatings could increase absorption, they increase the absorption in a similar degree in the lower boundary layer. The vertical variation in BC concentration plays a more critical role in the vertical difference of BC's absorption ability.

400 3) The study obtained BC properties in both the boundary layer and residual layer in special cases. Compared with the ground, the residual layer generally has BC with a higher D_p/D_c, lower concentration, and more variable MMD. In some cases, the D_p/D_c in the residual layer could be 8.4% higher than that on the ground, which could lead to an 8.6% absorption increase and 25.0% critical supersaturation decrease. Pollutants with distinct properties in the residual layer would influence the pollutants in the boundary layer in the early morning through entrainment and vertical mixing with the development of the boundary

405 layer. It was observed that aged BC and high concentration O_3 in the residual layer could lead to a quick increase in D_p/D_c and O_3 concentrations on the ground in the early morning (6:00–8:00).

Our study has revealed that the mixing state of BC is nearly vertically uniform within the boundary layer due to strong vertical mixing. Therefore, assuming constant microphysical properties of BC in the boundary layer for model simulations such as radiative transfer models is acceptable without additional information. However, with higher actinic flux and O_3 concentration,

410 the aging process of BC may be more efficient at the top of the boundary layer. Additionally, fresh BC tends to accumulate in the boundary layer at night. The difference in D_p/D_c between the boundary layer and residual layer could be significant in some

cases and should not be ignored. Such a difference could lead to a larger absorption enhancement of BC in the top of the boundary layer, potentially suppressing boundary layer development in the morning.

Author contributions.

Hang Liu: Conceptualization, Methodology, Software. Xiaole Pan: Conceptualization. Shandong Lei: Investigation. Yuting ZHANG: Investigation. Aodong Du: Investigation. Weijie Yao: Investigation. Guiqian Tang: Investigation. Tao Wang: Visualization. Jinyuan XIN: Resources. Jie Li: Visualization. Yele Sun: Resources. Junji CAO: Resources. Zifa Wang: Resources, Supervision.

Data availability

To request the data given in this study, please contact Dr. Hang Liu at the Institute of Atmospheric Physics, Chinese Academy of Sciences, via email (liuhang@mail.iap.ac.cn).

Competing interests

The authors declare that they have no conflict of interest.

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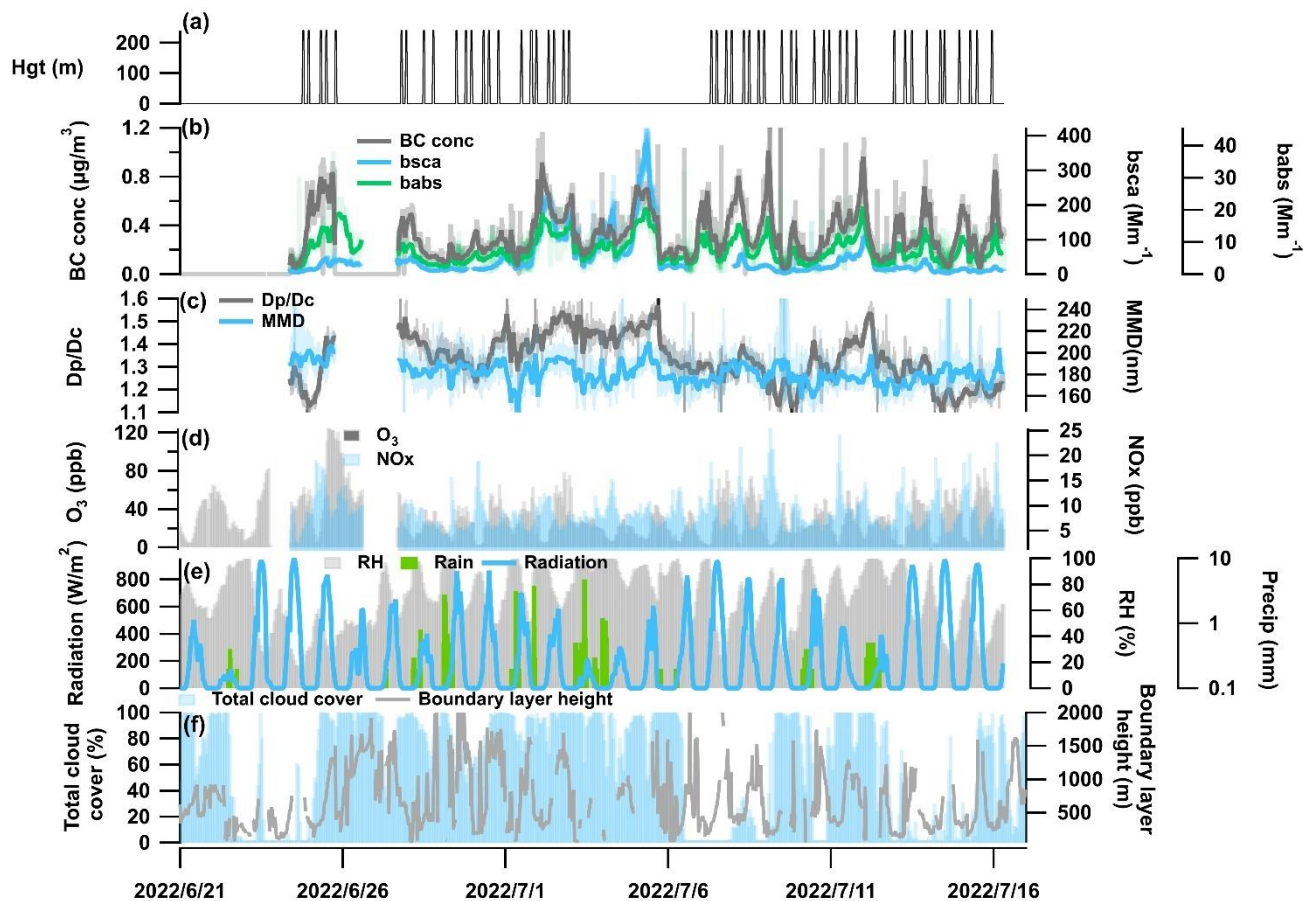
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Figure. 1 Time series of the major parameters during the observation. (a) The height of the moveable container. (b) Mass concentration of black carbon and absorption (b_{abs}) and scattering coefficient (b_{sca}), the solid and light-coloured lines denote the data with 1-hour and 1-min average. (c) The coating thickness (D_p/D_c) and mass median diameter (MMD) of BC, the solid and light-coloured lines denote the data with 1-hour and 1-min average. (d) The 1-min

630 **averaged mixing ratio of O_3 and NO_x. (e)(f) The meteorological conditions during the observation period.**
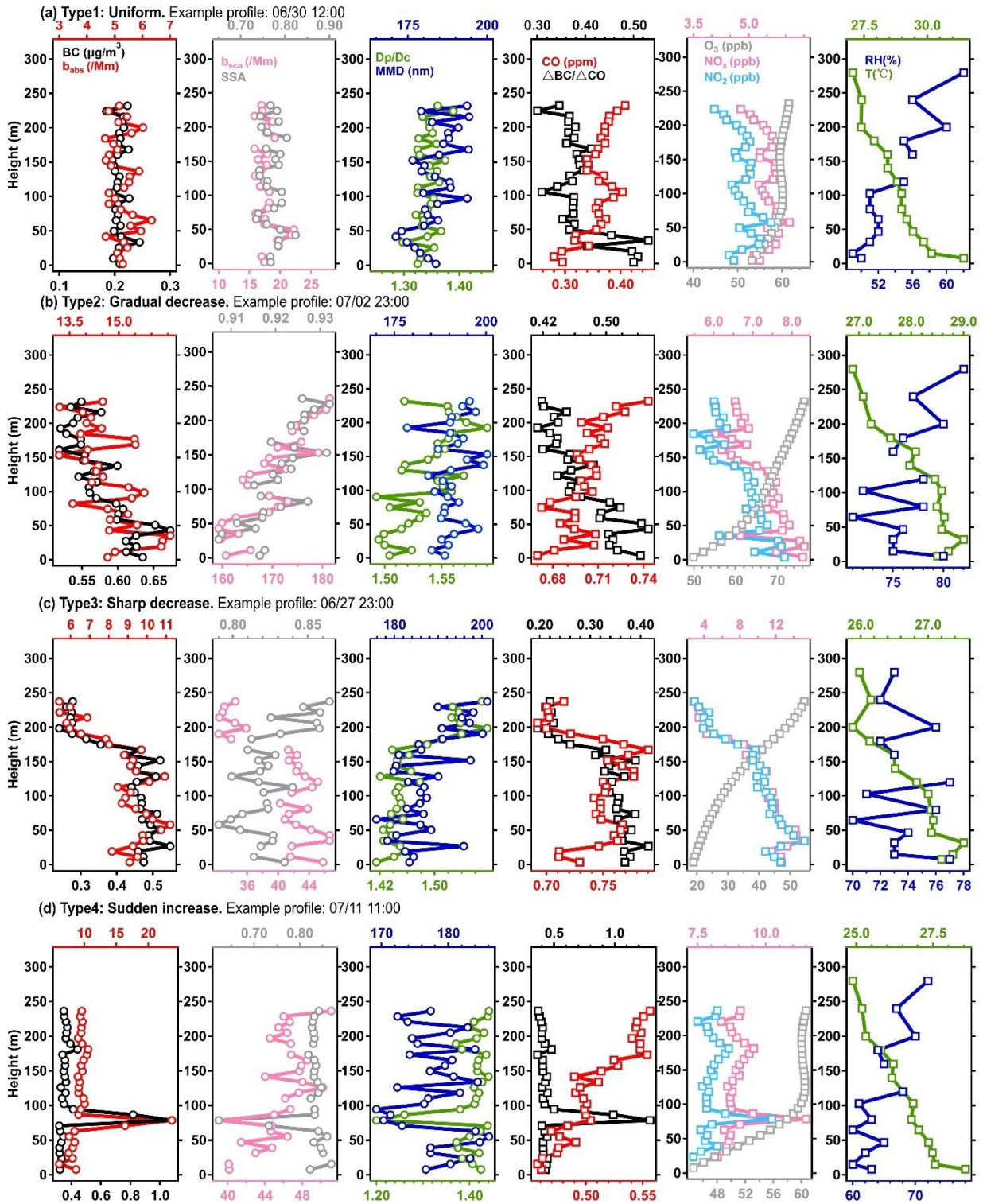
Figure. 2 Typical profiles of the pollutants and parameters of the four types. The four types were classified based on the vertical profiles of BC mass concentration. The exhibited parameters from left to right are BC mass concentration, aerosol absorption coefficient (b_{abs}), aerosol scattering coefficient (b_{sca}), single scattering albedo (SSA), BC coating thickness (D_p/D_c), mass median diameter of BC core (MMD), CO mixing ratio, $\triangle BC/\triangle CO$, mixing ratio of O_3 and NO_x, as well as relative humidity (RH) and temperature (T).

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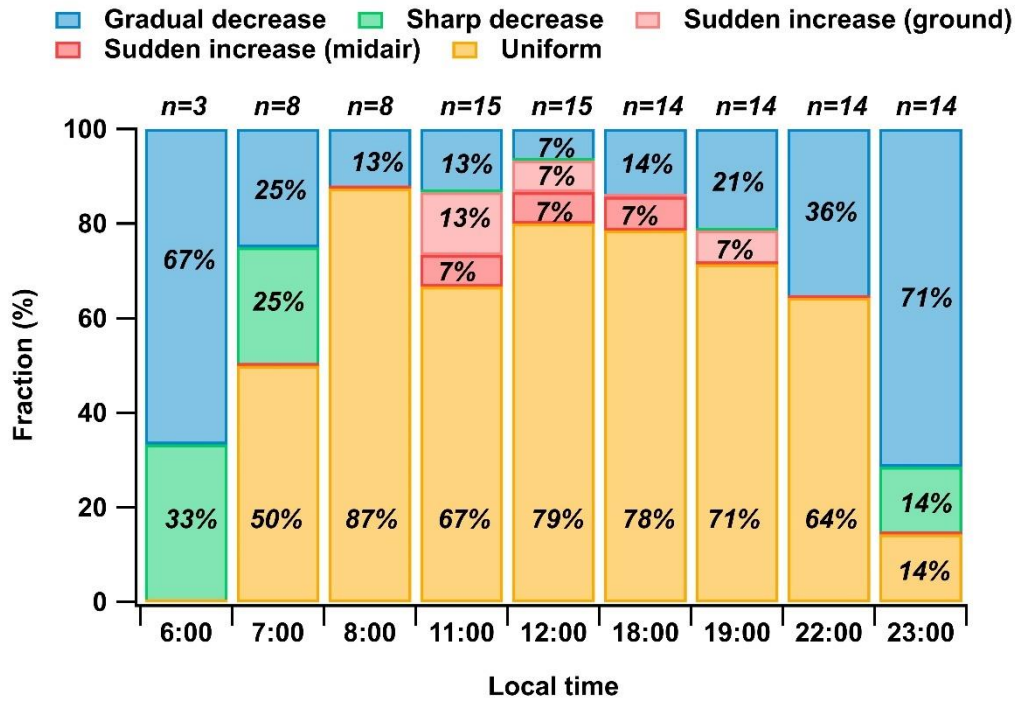
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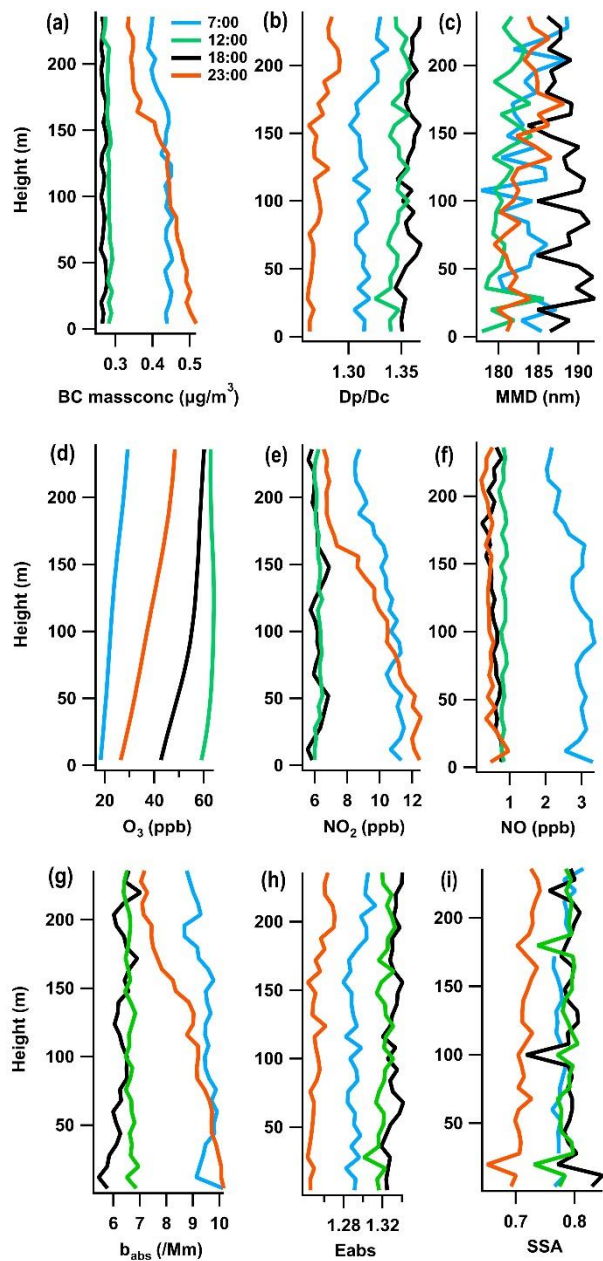
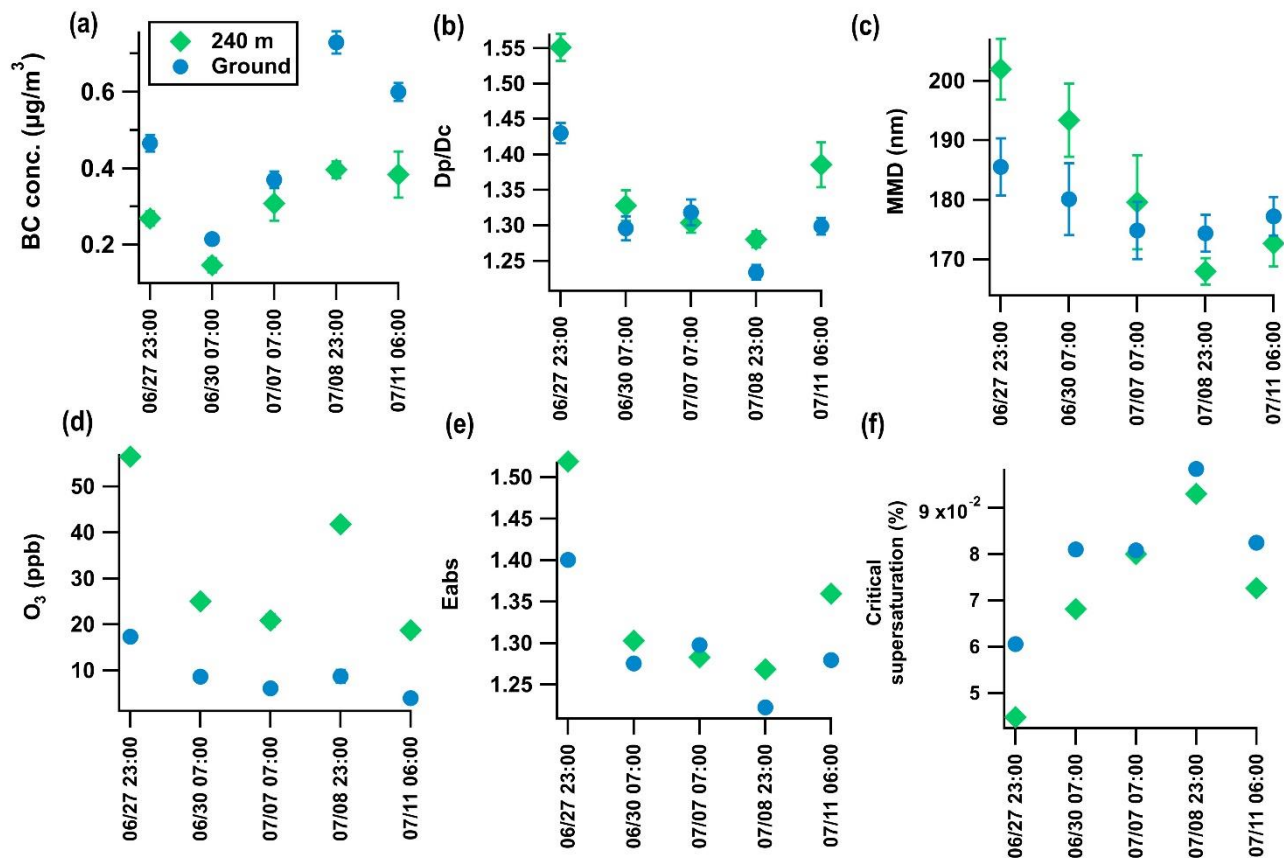
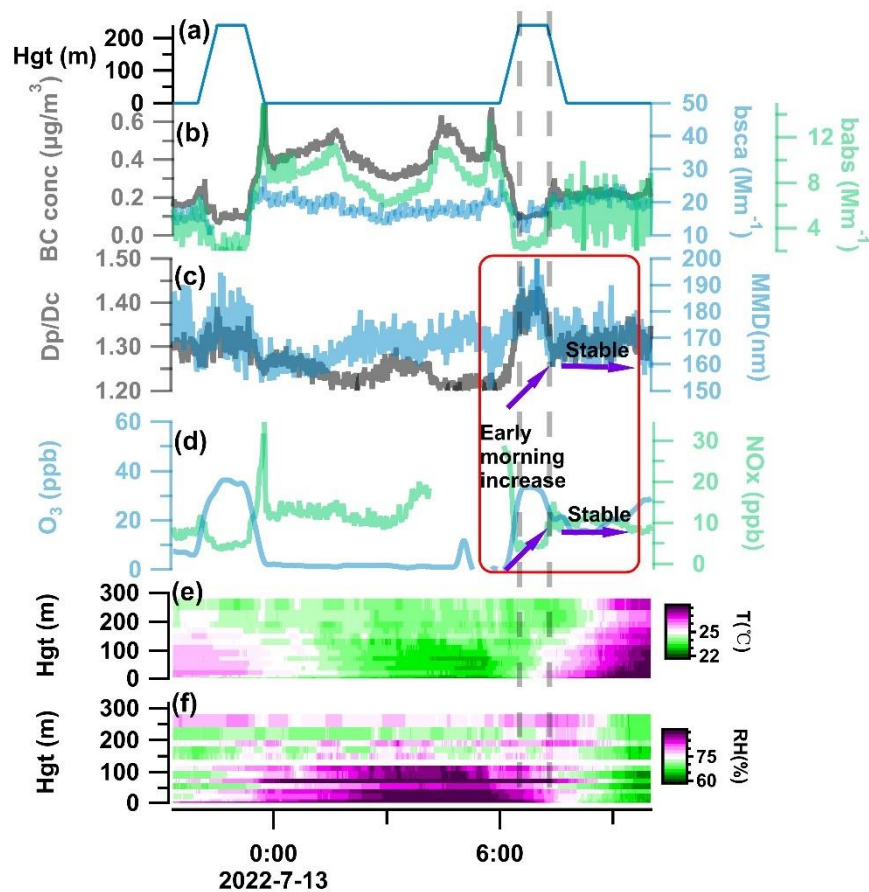
Figure. 3 Number fraction of the different vertical profile types at various times. The sudden decrease type was further classified into two types based on the sudden increase location. *n* above each bars denotes the total measured profile numbers at the time.

645 **Figure. 4 The diurnal variation of the average profiles of BC properties including (a) BC, (b) D_p/D_c, (c)MMD. The average profiles of (d) O_3 and its precursors (e) NO_2 and (f) NO. The averaged profiles of aerosol optical properties including (g) absorption coefficient, (h) absorption enhancement of BC due to coatings, (i) single scattering albedo.**

650   **Figure. 5 Comparison between the surface and 240 m height in the sharp decrease type of major pollutant concentrations (BC, O₃) and BC micro physical characteristics including coating thickness (D$_p$/D$_c$), mass median diameter of BC core (MMD), the absorption enhancement of BC due to coatings (E$_{abs}$) and critical supersaturation point of BC.**

655 **Figure. 6 A typical case of vertical mixing from the residual layer leading to an increase in O₃ and Dₚ/D꜀ in the early morning. (a) The height of moveable container. (b)-(d) Time series (1-min average) of BC mass concentration, b_abs, b_sca, Dₚ/D꜀, MMD, O₃ and NOₓ. (e) 2D (height and time) plot of temperature, the colours denote the quantity of the temperature. (f) The same as e, but for relative humidity.**

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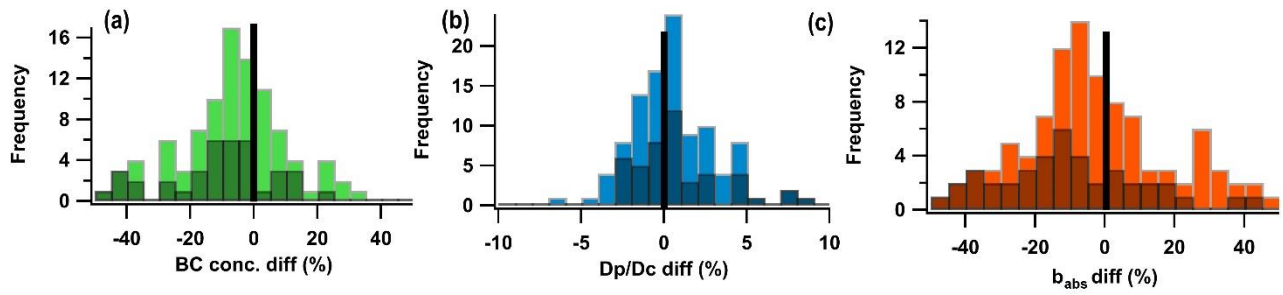
Figure. 7 Histogram of the number frequency of the relative differences between 240 m and the ground for a) BC concentration, (b) D_p/D_c and (c) b_{abs}. The shaded black area denotes the number frequency at night and morning

665 **(6:00,7:00, 8:00, 22:00, 23:00), the black line denotes no difference between 240m and the ground.**