Impact of fossil and non-fossil sources on the molecular compositions of water soluble humic-like substance in $PM_{2.5}$ at a suburb site of Yangtze River Delta, China

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Abstract

Atmospheric humic-like substances (HULIS) affect global radiation balance due to their~~its~~ strong light absorption at the ultraviolet wavelength. The potential sources and molecular compositions of water soluble HULIS at a suburb site of Yangtze River Delta from 2017 to 2018 were discussed based on the radiocarbon (^{14}C) analysis combining the Fourier Transform Ion Cyclotron Resonance Mass Spectrometry (FT-ICR MS) technique in this study. The ^{14}C results showed that the averaged non-fossil source contributions to HULIS were 39 ± 8 % and 36 ± 6 % in summer and winter, respectively, indicating the significant contributions from fossil sources to HULIS~~indicating that both the fossil and non-fossil sources played important roles in the formation of HULIS~~. The Van Krevelen diagrams obtained from the FT-ICR MS results showed that the proportions of tannins-like and carbohydrates-like groups were higher in summer, suggesting significant contribution of HULIS from biogenic secondary organic aerosols (SOA). The higher proportions of condensed aromatic structures in winter suggested ~~the~~ increasing anthropogenic

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emissions. Molecular composition analysis on the CHO, CHON, CHOS, and CHONS subgroups showed the relatively higher intensities of high O-containing macromolecular oligomers in CHO compounds in summer, further indicating stronger biogenic SOA formation in summer. High-intensity phenolic substances and flavonoids which were related to biomass burning and polycyclic aromatic hydrocarbons (PAHs) derivatives indicating fossil fuel combustion emissions were found in winter CHO compounds. ~~High-intensity phenolic substances and flavonoids which were related to biomass burning and polycyclic aromatic hydrocarbons (PAHs) derivatives indicating fossil fuel combustion emissions were found in winter CHO compounds.~~ Besides, two high-intensity CHO compounds containing condensed aromatic ring structures ($C_9H_6O_7$ and $C_{10}H_5O_8$) identified in summer and winter samples were similar to those from off-road engine samples, indicating that traffic emission was one of the important fossil sources of HULIS at the study site. The CHON compounds were mainly composed of ~~organonitrates or~~ nitro compounds or organonitrates ~~-~~with significant higher intensities in winter, which was associated to biomass burning emission, as well as the enhanced formation of organonitrates due to high NO_x in winter. However, the high-intensity CHON molecular formulas in summer were referring to N-heterocyclic aromatic compounds, which were produced from the atmospheric secondary processes involving reduced N species (e.g., ammonium). The S-containing compounds were mainly composed of organosulfates (OSs) derived from biogenic precursors, long-chain alkane and aromatic hydrocarbon, illustrating the mixed sources of HULIS~~, further illustrating the mixed sources of HULIS~~. Generally, different policies need to be considered for each season due to the different season sources, i.e., biogenic emission in summer and biomass burning in winter for non-fossil source, ~~and both important biogenic and anthropogenic source contributions to HULIS at the study site.~~ traffic emission and anthropogenic SOA formation in both seasons and additional coal combustion in winter. Measures to control emissions from motor vehicles and industrial processes need to be considered in summer. Additional control measures on coal power plants and biomass burning should be concerned in winter. These findings add to our understanding of the interaction between the sources and the molecular compositions of atmospheric HULIS.

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1. Introduction

Atmospheric humic-like substances (HULIS) have been observed worldwide and can be produced from primary combustion of biomass, fossil fuel, as well as various secondary processes

such as photochemical processes of volatile organic compounds (VOCs) and heterogeneous reactions of organic aerosols in the atmosphere (Kuang et al., 2015; Li et al., 2019; Ma et al., 2018; Sun et al., 2021). As important component of brown caron (BrC) aerosols, HULIS species have been widely reported to have a great impact on global radiative budget, contributing to 20-40% of the direct radiative forcing caused by light absorbing aerosols due to its light absorption at the ultraviolet wavelength (Chung et al., 2012; Zhang et al., 2017; Zhang et al., 2020a; Wang et al., ~~2018b~~2018c). HULIS are a highly complex mixture of polar organic compounds composed of aromatic and hydrophobic aliphatic structures containing carboxyl, carbonyl, and hydroxyl function groups (Zheng et al., 2013; Graber and Rudich, 2006; Zhang et al., 2022b; Zhang et al., 2022c). During the atmospheric secondary oxidation processes, the substitutions of hydrophilic functional groups ~~produced~~ increased aerosol hygroscopicity (Huo et al., 2021; Jiang et al., 2020). Polycarboxylic acids in HULIS are surface-active and play an important role in the cloud condensation nuclei (CCN) activity (Tsui and McNeill, 2018). N-base compounds can promote the generation of atmospheric reactive oxygen species (ROS) which have a great impact on human health (Wang et al., ~~2017b~~2017c; De Haan et al., 2018; Song et al., 2022). Identifying the molecular compositions of HULIS is a challenge due to complex mixtures contained in HULIS and can help to a better understanding of the processes involving organic compounds in atmosphere (Noziere et al., 2015; Laskin et al., 2018).

The Fourier-Transform Ion Cyclotron Resonance Mass Spectrometry (FT-ICR MS) coupled with electrospray ionization (ESI) ion source have been widely used in identifying the chemical structure of HULIS, providing high mass accuracy and can determine molecular formulas from mixed compounds (Chen et al., 2016; Wang et al., 2019b; Lin et al., 2012a; Jiang et al., 2020). Typical molecular formulas composed of C, H, and O atoms in HULIS were observed being abundant in carboxylic acids, lig~~n~~ngin-derived products, and polycyclic aromatic hydrocarbons (PAHs) or their derivatives (Lin et al., 2012a; Sun et al., 2021; Jiang et al., 2020; Huo et al., 2021; Song et al., 2018). In addition, the HULIS formation of N and S containing precursors was~~ere~~ also widely detected. The N-containing compounds such as nitroaromatics were important chromophores in HULIS in aged biomass burning organic aerosols (BBOA), as well as in ambient aerosols influenced by biomass burning (BB), while reduced N compounds such as N-heterocyclic aromatic compounds were found to be important chromophores in fresh BBOA (Wang et al., 2019b; Song et al., 2022; Jiang et al., 2020; Wang et al., ~~2017b~~2017c). Recent laboratory

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simulation experiments showed that the photooxidation of various anthropogenic VOCs (e.g., naphthalene, benzene, toluene, and ethylbenzene) would be promoted under high NO_x condition, producing strongly light absorbing nitroaromatics (Yang et al., 2022; Aiona et al., 2018; Siemens et al., 2022; Xie et al., 2017). Otherwise, nighttime oxidation of biogenic or anthropogenic VOCs, such as benzene/toluene, isoprene (C_5H_8) and monoterpenes ($C_{10}H_{16}$) by NO_3 radicals lead to substantial organonitrates formation, where the VOCs oxidation is strongly affected by NO_x (He et al., 2021; Shen et al., 2021; Wang et al., 2020; Zheng et al., 2021).

The organosulfates (OSs) and nitrooxy organosulfates (nitrooxy-OSs) have also been found to widely exist in HULIS in different atmospheric environment (Lin et al., 2012b; Lin et al., 2012a; Sun et al., 2021). Field study and laboratory smog chamber experiments have confirmed that OSs and nitrooxy-OSs in the atmosphere mainly come from the O_3, OH, or NO_3 oxidation of biogenic VOCs such as isoprene, α/β-pinene as well as aromatic hydrocarbon in the presence of H_2SO_4/SO_2 (Surratt et al., 2008; Glasius et al., 2021; Yang et al., 2020; Lin et al., 2012b; Huang et al., 2020). Coal combustions were found to be important sources of the aromatic OSs and nitrooxy-OSs in HULIS (Song et al., 2018). Besides, the long-chain alkanes were found to be important precursor of OSs in atmospheric aerosol samples from urban area which was related to vehicle emissions (Wang et al., 2019a; Tao et al., 2014).

Nanjing is one of the main cities in the Yangtze River Delta (YRD), which is one of the most developed areas in China. Organic matter can account for 20-40 % of $PM_{2.5}$ in the YRD area due to the impact of complicated sources, especially anthropogenic emissions (Wang et al., 2017a; Wang et al., 2016a). Studies have reported that BrC is an important contributor to aerosol light absorption in Nanjing and exhibited obvious seasonal variations, with peaks in wintertime, owing to emissions from biomass burning, fossil fuel combustion, and secondary formation (Chen et al., 2018; Cui et al., 2021; Xie et al., 2020; Wang et al., 2018a). Recently, works on the field observation of nitrated aromatic compounds (NACs) were conducted to explore the light absorption contributions of NACs to BrC and help to better understand the links between the optical properties and molecular compositions of BrC (Gu et al., 2022; Cao et al., 2023). However, as far as we know, understanding of the sources of atmospheric HULIS at molecular levels was still limited. In this work, ~~Previous study reported by our laboratory have found significant HULIS formation at Nanjing, China influenced by both biogenic and anthropogenic emissions (Bao et al., 2022). t~~The molecular compositions of water soluble HULIS isolated from $PM_{2.5}$ samples

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collected in summertime and wintertime from 2017 to 2018 at Nanjing, China, were investigated combining the FT-ICR MS and radiocarbon (^{14}C) analysis. We aim to obtain the molecular characteristic differences of water soluble HULIS in summertime and wintertime and to get a better understanding of the influence of different sources on the molecular compositions of HULIS.

2. Materials and methods

2.1 Sample collection

The 24 h $PM_{2.5}$ samples were collected on the roof of Wende building, which was about 21 m height from the ground at Nanjing University of Information Science and Technology (32.2° N, 118.7° E) using a high-volume sampler (KC-1000, Qingdao, China) at a flow rate of 300 L min^{-1}. The study site was located in~~at~~ the north suburb area of Nanjing, adjacent to G205 State Road and surrounded by an industrial park and residential area. Generally, the study site was affected by human activity, industrial emission, and traffic emission. The sample collection was conducted in summer from 12 August 2017 to 26 August 2017 and in winter from 31 December 2017 to 31 January 2018. A heavy haze event occurred from 31 December 2017 to 3 January 2018, thus the sample frequency was adjusted to 2 h in daytime and 8 h in nighttime. Field blank filters were performed before and after sample collection for each season. More details about the sample collection can be found in previous research reported by Bao et al. (2022). ~~reported by our laboratory (Bao et al., 2022).~~ The air pollutants data including $PM_{2.5}$, SO_2 and NO_2 were provided by China National Environmental Monitoring Centre. Twelve samples were selected for further chemical analysis and the details about the sample selection are described in Section 3.1 in this study.

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2.2 Chemical analysis

~~,~~The solid phase extraction (SPE) cartridge (Oasis HLB, 30 μm, 60 mg/cartridge, Waters, USA) was performed to isolate the water soluble HULIS in this study. Briefly, ~~,and t~~the prepared water extracts ~~,~~passed through the pre-conditioned HLB cartridge firstly, then the retained HULIS on the HLB cartridge were eluted with 2% (v/v) ammonia/methanol and evaporated to dryness under a gentle stream of nitrogen gas, then re-dissolved in ultrapure water for the measurement. The carbon fraction in HULIS (HULIS-C) were determinated using a total carbon analyzer (Shimadzu-TOC-VCPH, Shimadzu, Japan) with standard deviation of reproducibility test less than 3.5 % and detection limit of 0.14 μg C m^{-3}. More details about the HULIS isolation and measurement have been described in ~~(~~Bao et al.~~,~~ (2022).

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The mass concentrations of ~~levoglucosan~~ water soluble ions including NO_3^-, NH_4^+ and SO_4^{2-} were measured using an ion chromatography (Dionex ICS-5000+, ThermoFisher Scientific, USA) separated on an AS11 column (4*250 mm, Dionex). Potassium hydrate (KOH) was used as the gradient eluent for anion determination. The levoglucosan concentrations were analyzed using the same ion chromatograph equipped with a CarboPac MA1 analytical column (4*250 mm, Dionex) and an electrochemical detector. Sodium hydroxide (NaOH) was used as the gradient eluent for levoglucosan determination. All data were blank corrected in this study. ~~and~~ More~~the~~ details of the methods have been described previously (Liu et al., 2019). ~~All data were blank corrected in this study. More details about the HULIS isolation and measurement have been described in (Bao et al., 2022).~~

2.3 Radiocarbon analysis

For the radiocarbon measurement of the HULIS samples, the organic solvents were firstly evaporated under a gentle flow of ultrapure N_2 for 30-40 minutes in tin cups. After that, the tin cups were wrapped into balls and more than 50 μg of carbon from the HULIS samples was combusted into CO_2 using an elemental analyzer (EA, model vario micro, elemental, Germany), then reduced into graphite targets for ^{14}C determination at the State Key Laboratory of Organic Geochemistry, Guangzhou Institute of Geochemistry, Guangzhou, China (Jiang et al., 2020). Detailed descriptions of the ^{14}C data processing can be found in previous study (Mo et al., 2018). Briefly, the ^{14}C values were expressed as the modern carbon (f_m) fraction after correcting for the $\delta^{13}C$ fractionation. The f_m was converted into non-fossil carbon (f_{nf}) fraction with the correction factor of 1.06±0.07 based on the long-term time series of $^{14}CO_2$ sampled at the background station in this study (Levin et al., 2013; Levin and Kromer, 2004). ^{14}C analysis of the oxalic acid standard (IAEA-C7) was conducted in this study (Xu et al., 2021). No field blank correction was performed for the carbon isotope analysis since the carbon content in the field blanks was negligible.

2.4 High-resolution FT-ICR MS analysis

The ultrahigh resolution mass spectra of the HULIS samples were obtained through a SolariX XR FT-ICR MS (Bruker Daltonics, GmbH, Bremen, Germany) equipped with a 9.4 T superconducting magnet (Gamry Instruments, Warminster, USA) and a Paracell analyzer cell (Brucker Daltonik GmbH, Bremen, Germany) in the negative ESI mode. The detection mass range was set as m/z 150 to 800 and the ion accumulation time was set as 0.65 s. A total of 100 continuous

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4M transient data points were superposed to enhance the signal to noise ratio and dynamic range. The mass spectrum was externally calibrated with a standard solution of arginine and internal recalibration was performed using typical O_6S_1 chemical species in DataAnalysis ver. 4.4 software (Bruker Daltonics) (Mo et al., 2018; Tang et al., 2020; Jiang et al., 2020). Field blank filters were analyzed as same as the samples and all the sample data were blank corrected. More details about the data processing can be found in Text S1 in the supporting information.

3. Results and discussion

3.1 General temporal characteristics during the sampling periods

Figure 1 displays the temporal variations of non-fossil contributions to HULIS-C, the mass concentrations of HULIS-C, levoglucosan, NO_3^-, SO_4^{2-}, NH_4^+, SO_2, NO_2, and $PM_{2.5}$, as well as the relative humidity and temperature during the study periods corresponding to the 12 samples. The 12 samples were named as S1-S6 (summer) and W1-W6 (winter) in chronological order corresponding to the six samples in summer and winter, respectively in this study. The averaged mass concentrations of $PM_{2.5}$ in summer and winter during the selected periods were 21.05 ± 8.05 μg m^{-3} and 445.67 ± 275.00 μg m^{-3}, respectively, indicating the serious pollution level in winter. The daily $PM_{2.5}$ mass concentrations in summer were all below the daily averaged Chinese National Ambient Air Quality Standard (NAAQS) of 35 μg m^{-3} ~~for the first grade~~, while the daily $PM_{2.5}$ mass concentrations in winter all exceeded the daily averaged NAAQS of 35 μg m^{-3} ~~for the first grade~~, of which the $PM_{2.5}$ mass concentrations of W1-W3 and W6 exceeded 200 μg m^{-3}.

~~As shown in Fig. 1, the mass concentrations of HULIS-C, levoglucosan, water soluble secondary inorganic aerosols (SIA), and air pollutants showed similar trends in winter, suggesting the influence of BB and anthropogenic emissions in winter (Wu et al., 2019b).~~ The averaged mass concentrations of HULIS in summer and winter during the selected periods were 1.83 ± 0.27 μg m^{-3} and 4.52 ± 2.29 μg m^{-3}, respectively. The averaged HULIS concentration in summer was comparable with those measured in other cities in China, i.e., 1.70 μg m^{-3} in Guangzhou,1.61 μg m^{-3} in Shanghai and 1.50 μg m^{-3} in Xi'an. Compared with those measured in winter samples in other cities, our result was comparable with those in Xi'an (4.50 μg m^{-3}), a little lower than those in the megacity of Shanghai (5.31 μg m^{-3}) and higher than those in the southern coastal city of Guangzhou (3.6 μg m^{-3}) (Fan et al., 2016; Zhang et al., 2020b; Zhao et al., 2016).

As shown in Fig. 1, the mass concentrations of HULIS-C, levoglucosan, water soluble secondary inorganic aerosols (SIA), and air pollutants showed similar trends in winter, suggesting

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the influence of BB and anthropogenic emissions in winter (Wu et al., 2019b). The radiocarbon analysis results showed that the f_{nf} of HULIS-C ranged from 30 % to 50 % with an average contribution of 39 ± 8 % in summer and ranged from 32 % to 48 % with an average contribution of 36 ± 6 % in winter, indicating the significant contributions from fossil sources to HULIS at the study site. ~~Significant increasing of the levoglucosan and HULIS-C mass concentrations were found from 31 December 2017 to 1 January 2018, corresponding to the W1-W3 samples, indicating the BB impact during the periods. The maximum of the levoglucosan and HULIS-C mass concentrations were 552.79 μg m^{-3} and 7.40 μg m^{-3}, respectively. Despite the higher levoglucosan mass concentrations in the W1-W3 samples, the radiocarbon analysis results showed that the f_{nf} of HULIS-C ranged from 30 % to 50 % with an average contribution of 39 ± 8 % in summer and ranged from 32 % to 48 % with an average contribution of 36 ± 6 % in winter, indicating that both fossil and non-fossil sources played important roles in the formation of HULIS at the study site. There were other emission sources of HULIS in winter other than BB.~~ ~~T~~Figure ~~S1 shows t~~he 48 h back trajectories (Fig. S1) showed that ~~of each sample during the selected periods.~~ ~~t~~The study site was affected by ~~by the clean air masses from the ocean in summer and~~ the polluted air masses mainly from the northern ~~heating~~ cities in winter, suggesting the coal combustion contributions to HULIS in winter (Ma et al., 2018; Sun et al., 2021)~~.~~. In addition, significant increasing of the levoglucosan and HULIS-C mass concentrations were found from 31 December 2017 to 1 January 2018, corresponding to the W1-W3 samples and the maximum of the levoglucosan and HULIS-C mass concentrations were 552.79 ng m^{-3} and 7.40 μg m^{-3}, respectively, indicating the BB impact during the periods. ~~–~~In summer, the study site was affected by both regional transport from the nearby cities in the north and west of Nanjing and the Donghai Sea. The anthropogenic emissions from the neighboring cities might cause the anthropogenic SOA formation, i.e., secondary N-containing and S-containing compounds with aromatic structures during the atmospheric transport processes, which was discussed in detail in section 3.4 in this study.

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3.2 Mass spectra and molecular formula assignments

Figure S2 and S3 show~~s~~ the negative ion ESI FT-ICR mass spectra of HULIS in summer and winter, respectively. The molecular formulas listed are some of the top ten molecular formulas. Thousands of peaks are present in the spectra in the range from m/z 150 to m/z 600 and the most intense ion peaks are those in the range m/z 200-400 in summer and m/z 150-350 in winter. Our

results are similar to those found for the ultrahigh resolution mass spectra of water-soluble organic compounds in particles produced from BB, coal combustion, vehicle exhaust emissions, as well as in ambient aerosols and cloud water samples, within a reasonable range -(Tang et al., 2020; Sun et al., 2021; Song et al., 2018; Song et al., 2019; Bianco et al., 2018). In this study, the assigned molecular formulas were classified into the following four main subgroups based on their elemental compositions: CHO (compounds containing only C, H, and O), CHON (compounds containing C, H, O and N), CHOS (compounds containing C, H, O, and S), and CHONS (compounds containing C, H, O, N, and S). As shown in Fig. 2, the proportions of the four subgroups accounted for the overall formulas followed as CHO (20 %-27 %), CHON (28 %-43 %), CHOS (19 %-26 %), and CHONS (16 %-26 %) in summer, respectively and CHO (15 %-19 %), CHON (30 %-40 %), CHOS (21 %-32 %), and CHONS (20 %-29 %) in winter, respectively. The average proportions of the CHO, CHON, CHOS, and CHONS compounds in summer were 22 ± 3 %, 36 ± 5 %, 22 ± 3 %, and 20 ± 4 %, respectively. The average proportions of the four subgroups in winter were 17 ± 2 %, 32 ± 4 %, 24 ± 3 %, and 27 ± 4 %, respectively. The CHON groups were the major components of molecular formulas, furthermore, the relative intensity of CHON groups increased significantly in winter (Fig. S2 and Fig. S3). Studies have suggested that HULIS emitted from biomass burning can produce a high abundance of CHON compounds and S-containing compounds were the dominant component for primary HULIS emitted from coal combustion (Zhang et al., 2021; Song et al., 2018). The higher intensity of CHON compounds in winter in this study further indicated the BB contribution. ~~Notably, the~~The contributions of S-containing compounds (CHOS and CHONS groups) increased in winter which might be related to the polluted air masses transported from the northern ~~heating~~ cities with increasing coal combustions emissions in winter (Song et al., 2018). Notably, the relatively higher proportions of CHO and CHON groups in summer were most probably related to the increasing biogenic emissions in summer, resulting in the formation of some high molecular weight oligomers or highly oxidized organonitrates, which was discussed in detail in section 3.4.1 and 3.4.2 in this study.

Table S1 and S2 displays the composition characteristics of atmospheric HULIS in the summer and winter samples, including the relative intensity weighted average values of number, molecular weight (MW_w), elemental ratios (O/C_w and H/C_w), double-bond equivalent (DBE_w), aromaticity index (AI_w), and DBE/C_w. A total of 14387 and 15731 peaks were detected in the summer and winter samples, respectively. The O/C and H/C ratios are commonly calculated to

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evaluate the oxidation degree and saturation degree of the compounds, respectively (Ning et al., 2022). The O/C_w values were in a range of 0.61-0.80 with an average value of 0.71 ± 0.07 for summer samples and in a range of 0.59-0.67 with an average value of 0.62 ± 0.03 for winter samples, respectively. The higher oxidation degree of summer samples than winter samples indicated stronger secondary HULIS formation in summer. The H/C_w values were in a range of 1.38-1.46 with an average value of 1.42 ± 0.03 for summer samples and in a range of 1.33-1.41 with an average value of 1.36 ± 0.04 for winter samples, respectively. The O/C_w and H/C_w of each molecular subgroup followed a changing trend of CHO < CHON < CHOS < CHONS compounds. Most of the S-containing compounds had a O/C value ≥0.7, suggesting the large amounts of highly oxidized OSs in S-containing compounds which contained various functional groups and were mainly from the photochemical oxidation of biogenic or anthropogenic volatile organic compounds (VOCs) (Mutzel et al., 2015). The DBE values were calculated to describe the degree of unsaturation of compounds and restricted the assigned molecular formulas with unreasonably high or low number of rings or double bonds (Kroll et al., 2011). The related parameter DBE/C was the double-bond equivalent of unit carbon which can reflect the condensed ring structures in the compounds (Jiang et al., 2021). The higher DBE_w and DBE/C_w values of CHO and CHON compounds were found in this study, indicating the higher unsaturation degree of these two groups.

Considering that double bonds can be formed by heteroatoms especially O atoms, whereas make no contributions to the aromaticity of the compounds, AI_w was calculated to supplement the DBE results (Song et al., 2018; Ning et al., 2019). AI_w can eliminate the contribution of O, N, and S atoms to the C=C double bond density of molecules. The AI_w values of different compounds groups in HULIS presented the changing trends: AI_w (CHONS) > AI_w (CHON) > AI_w (CHO) > AI_w (CHOS) in summer and AI_w (CHON) > AI_w (CHO) > AI_w (CHONS) > AI_w (CHOS) in winter, respectively. The formulas can be classified into three parts based on AI values proposed by previous studies: aliphatic (AI =0), olefinic (0< AI ≤0.5) and aromatic (AI >0.5) (Koch and Dittmar, 2006). As shown in Fig. S4 and S5, the aliphatic were the main components of S-containing compounds in this study and the olefinic and aromatic were the main components of CHO and CHON compounds. Furthermore, the aromatic proportion of CHO and CHON compounds significantly increased in winter, suggesting the increasing anthropogenic emissions in winter.

3.3 Comparative analysis using Van Krevelen diagrams

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In this study, the Van Krevelen diagrams (Fig. 3) were constructed to display the molecular composition and categorical distribution of the collected samples (Noziere et al., 2015; Patriarca et al., 2018; Li et al., 2022). According to the elemental ratios (O/C and H/C ratios) and AI values, seven major compound classes were classified, including lipids-like species, lignins-like species, proteins-like species, tannins-like species, carbohydrates-like species, condensed aromatics structure, and unsaturated hydrocarbons (Table S3). The Van Krevelen diagrams showed similar distributions in the 12 samples. The CHO and CHON compounds located in the lower left area and the S-containing compounds located in the upper light area with higher O/C and H/C ratios, indicating a higher degree of oxidation and saturation. The condensed aromatic structure mainly consisted in CHO and CHON compounds, further suggesting the influence of anthropogenic emissions on the formation of CHO and CHON compounds.

Figure 4 presents the averaged relative contributions of the number of molecular formulas from the seven categories in summer and winter samples, respectively. Lignins-like species accounted for the highest proportion of CHO compounds with average contributions of 58 % and 61 % in summer and winter, respectively, followed by CHON compounds with average contributions of 48 % and 57 % in summer and winter, respectively. Lignins are mainly composed of carboxyl groups, alicyclic rings, aromatic rings, and other O-containing groups. Previous studies have reported that lignin was a complex phenolic polymer which usually came from direct biological emissions or combustions of biofuel (Ning et al., 2019; Boreddy et al., 2021; Sun et al., 2021). Lignins pyrolysis products and other lignins derived molecules have been shown to be oxidized into light absorbing BrC chromophore under certain conditions (Fleming et al., 2020).

Tannins-like species accounted for 21 %, 27 %, 23 %, and 30 % of CHO, CHON, CHOS, and CHONS compounds, respectively in summer which were higher than those in winter with contributions of 13 %, 16 %, 16 %, and 23 % to CHO, CHON, CHOS, and CHONS compounds, respectively. Tannins-like species are a series of polyphenolic compounds containing hydroxyls and carboxylic groups which have been widely reported in fogs, cloud water and aerosol samples, attributing to highly oxidized organic compounds such as OSs or nitrooxy-OSs produced from the nighttime chemistry between the biogenic VOCs with the NO_3 (Altieri et al., 2009; Bianco et al., 2018; Ning et al., 2019; Altieri et al., 2008; Shen et al., 2021). Carbohydrates-like species which contain monosaccharide, alditols, and anhydrosugars mainly consisted in CHONS compounds which also had a relative higher proportion of 33 % in summer than that of 29 % in winter (Sun et

al., 2021). $C_{10}H_{16}NO_{7\text{-}9}S$, as monoterpene nitrooxy-OSs, showing high relative intensities, were typical carbohydrates-like species detected in this study which represented biogenic secondary organic aerosols (SOA) (Ning et al., 2019; Surratt et al., 2008; Wang et al., 2020). Both the higher proportions of tannins-like and carbohydrates-like classes in summer indicated stronger biogenic SOA formation in this study.

Proteins-like classes mainly consisted in CHOS compounds with average proportions of 29 % and 38 % in summer and winter, respectively. Proteins contain peptide-like structures formed by dehydration with different kinds of amino acids and consist of short chains of amino acid residues (Bianco et al., 2018). These compounds are associated with photochemical oxidation processing in aerosols, thus resulting in the significant formation of OSs from biogenic or anthropogenic precursors in this study (Bigg and Leck, 2008).

Higher condensed aromatics were detected in winter with average proportions of 14 % in CHO compounds and 8 % in CHON compounds, respectively which were 2-2.5 times of those in summer. Condensed aromatics are important components of PAHs which were usuall~~l~~y emitted from incomplete combustion of fossil fuels (Ma et al., 2020). The increase of the proportion of condensed aromatics in winter indicated the stronger influence of anthropogenic sources on HULIS formation. The unsaturated hydrocarbons and lipids-like species showed the lowest molecular number percentage of less than 1 % in this study. Previous studies have shown that the lipids-like species were the main components of water insoluble organic compounds in aerosols and could be attributed to monocarboxylic acids (Ning et al., 2022; Wozniak et al., 2008).

In summary, both the summer and winter samples were mainly composed of compounds from biogenic origins (lignins-like, tannins-like, proteins-like, and carbohydrates-like species). More tannins-like and carbohydrates-like species were detected in summer including large amounts of highly oxidized OSs or nitrooxy-OSs, indicating biogenic SOA formation. More condensed aromatic structures in CHO and CHON compounds were detected in winter, owing to increasing anthropogenic emissions. It is noted that ESI ionization technology is more sensitive for the identification of polar compounds. Therefore, the low polar or nonpolar compounds, such as PAHs or their derivatives from fossil sources, were probably underestimated in this study (Jiang et al., 2014; Lin et al., 2018).

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3.4 Molecular composition of HULIS

3.4.1 Molecular characteristics of CHO compounds

The O/C_w and H/C_w ratios for the CHO compounds were 0.45-0.56 and 1.15-1.30 for the summer samples and 0.42-0.48 and 0.90-1.02 for the winter samples (Table S1 and S2). The summer samples showed higher oxidation degree and saturation degree. We firstly plotted the Van Krevelen diagrams of the four molecular subgroups showing relative intensities for all the 12 samples and similar distributions of the high-intensity compounds were found in the 6 summer samples and the 6 winter samples, respectively. Then we combined all the data in summer and winter, respectively. As shown in Fig. 5a and 5d, the CHO compounds in summer with high relative abundance were located at the area within $0.2 \leq O/C \leq 1.0$ and $1.0 \leq H/C \leq 1.7$, mainly including lignins-like species and tannins-like species which were closely related to biogenic emissions. On the contrary, the condensed aromatics showed high relative abundance in winter, suggesting ~~the~~ obviously different sources of HULIS in summer and winter. The DBE values increased with the increasing of the C numbers (Fig. 5b and 5e). The high-intensity CHO compounds in HULIS had DBE values between 3-7 with C numbers from 10 to 20 for summer samples. In winter, the high-intensity CHO compounds had DBE values between 7-11 with C numbers from 5 to 15. As mentioned above, the aromatic (AI >0.5) proportion of CHO compounds significantly increased in winter, the higher DBE values in winter further indicated the consists of more highly unsaturated aromatic compounds which reflected the anthropogenic emissions.

The CHO compounds were classified according to the number of oxygen atoms to evaluate the oxygen content. As shown in Fig. 5c and 5f, the high-intensity CHO compounds with 6-11 oxygen atom were detected in summer, such as $C_{15}H_{24}O_6$, $C_{15}H_{22}O_{10}$, $C_{18}H_{26}O_8$, and $C_{18}H_{26}O_9$, these highly oxygenated organic molecules with high molecular weight have also been detected in laboratory α-pinene ozonolysis SOA (Pospisilova et al., 2020). We further classified the CHO compounds by different carbon atom numbers. As shown in Fig. S6, the C_{17}-C_{22} compounds were the main components of the CHO compounds, accounting for more than 50 % of the total number of CHO molecular formulas in both summer and winter seasons. However, the total relative intensities of the CHO compounds in summer were significantly higher than those in winter, of which the C_{23}-C_{26} and C_{27}-C_{32} compounds were enriched in summer. These high molecular weight compounds were probably oligomers formed from various biogenic precursors, such as isoprene, sesquiterpene, and monoterpene (Daellenbach et al., 2019; Berndt et al., 2018). The high intensities of these compounds in summer further indicated the stronger biogenic SOA formation in summer compared with that in winter.

High-intensity CHO compounds with 4-9 oxygen atom were detected in winter (Fig. 5c) of which the $C_{14}H_{10}O_4$ formula with a DBE value of 10 appeared the highest intensity, which was probable functional PAHs and have been reported in HULIS from coal combustion smoke particles (Song et al., 2019). As shown in Fig. S2 and S3, the $C_{14}H_{10}O_4$ formula appeared high intensity in all the winter samples, providing the evidence of coal combustion emissions in winter. Some other high-intensity compounds in winter, such as $C_{14}H_8O_4$ and $C_{14}H_8O_5$ both with DBE values of 11, and $C_{13}H_8O_2$, $C_{13}H_8O_5$, and $C_{13}H_8O_6$ with DBE values of 10, might refer to hydroxyl substitutions derived from anthracenedione and xanthone, respectively, which have been reported in secondary wood combustion products (Bruns et al., 2015). $C_{15}H_{10}O_6$, $C_{15}H_8O_6$, and $C_{16}H_{12}O_7$ which had DBE values of 11, 12, and 11, respectively, might be flavonoids which had flavone backbone, the key structure of plant pigments, widely existing in plants in nature and could be important sources of BrC chromophores in aged BBOA (Fleming et al., 2020; Lin et al., 2016; Huang et al., 2021). Phenolic substances derived from phenol, guaiacol, and syringol are also widely existed in BBOA, usually from the pyrolysis of lignins in wood, which also play an important role in aqueous-phase SOA formation (Boreddy et al., 2021). For instance, $C_{13}H_{10}O_3$ and $C_{13}H_{10}O_5$ are guaiacol derivatives, $C_{15}H_{16}O_8$ are syringol derivatives and $C_{18}H_{14}O_6$ and $C_{18}H_{14}O_7$ are phenol derivatives (Sun et al., 2021). As shown in Fig. S7, the relative intensities of the CHO compounds mentioned above produced from BB were found to have similar trends with the mass concentrations of levoglucosan, which were significantly higher in W1-W3 samples, corresponding to the BB period from 31 December 2017 to 1 January 2018, providing the evidence of BB influence on HULIS formation in winter.

It is noted that the top compounds $C_9H_6O_7$ and $C_{10}H_6O_8$ were detected both in the summer and winter samples (Fig. S2 and S3), which had DBE values of 7 and 8, respectively, containing abundant condensed aromatic ring structures with high O numbers. Their peaks were also detected in the HFO (heavy-fuel-oil)-fueled off-road engine samples reported before, suggesting the traffic emission contributions to HULIS (Cui et al., 2019). This supported the radiocarbon analysis results in this study and gave further information that the traffic emissions were important fossil sources in both summer and winter seasons, which was also found in previous research which reported the sources of HULIS based on the positive matrix factorization (PMF) model by Bao et al. (2022).

3.4.2 Molecular characteristics of CHON compounds

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The O/C_w of CHON compounds in summer and winter were 0.57-0.71 and 0.52-0.56, respectively, while the H/C_w were 1.20-1.32 and 1.00-1.11, respectively (Table S1 and S2). Compared with the summer CHON compounds, the winter CHON compounds presented significantly higher ion abundance (Fig. 6a and 6d). The most abundant CHON subgroups had DBE values of 4-7 and 3-10 in summer and winter, respectively (Fig. 6b and 6e). Similar with the CHO compounds, the higher DBE values of high-intensity CHON compounds in HULIS in winter indicated a high prevalence of double bonds or ring structures. According to the N and O number, the CHON compounds were classified into N_1O_x (N_1O_1-N_1O_{15}) and N_2O_x (N_2O_2-N_2O_{14}) subgroups in summer and N_1O_x (N_1O_1-N_1O_{12}) and N_2O_x (N_2O_2-N_2O_{12}) subgroups in winter, respectively (Fig. 6c and 6f). $NO_{8\text{-}12}$ and $NO_{6\text{-}9}$ compounds were mostly enriched subgroups in summer and winter, respectively. More oxygen-enriched CHON compounds containing O number above 9 were detected in summer, implying the higher oxidation degree for summer samples. In addition, the N_1O_x were both the major compounds represented average of 64 ± 4 % and 61 ± 6 % of the CHON molecular formulas in summer and winter, respectively, indicating the presence of more single nitro/amino substituents in CHON compounds in this study.

Among the CHON compounds, 95 ± 1 % and 86 ± 3 % CHON compounds had O/N values ≥3 in summer and winter, respectively in this study, indicating these compounds contained large amounts of oxidized nitrogen functional groups such as nitro compounds ($-NO_2$) and/or organonitrates ($-ONO_2$) and excess oxygen atoms indicated the existence of other oxygen-containing functional groups (Laskin et al., 2009). The organonitrates formation from NO_3 oxidation of biogenic or anthropogenic VOCs can affect the interactions between anthropogenic and natural emissions (He et al., 2021; Shen et al., 2021; Wang et al., 2020). Organonitrates were found to be important species contributing to SOA formation in the polluted urban environment, which were enhanced under high NO_x level (Zheng et al., 2021). The significant higher relative intensities of CHON compounds in winter indicated that the high NO_x environment in winter promoted the formation of organonitrates and highlighted the importance of orgnonitrates for SOA control in polluted environment.

Furthermore, we found that the increase of the relative abundance of CHON compounds in winter was particularly significant in W1-W3 samples (Fig. S2 and S3), corresponding to the BB episode. Phenols produced from the pyrolysis of lignins can react with NO_3 radicals in the atmosphere, producing nitrophenols, which have been shown to be important BrC chromophore

in BBOA (Wang et al., ~~2017b~~2017c; Lin et al., 2016; Cai et al., 2020). It was reported that the gas-phase reactions of NO_3 radicals with phenolic substances took place at least 4 orders of magnitude faster than those with aromatic hydrocarbon and even faster in the aqueous phase (Lin et al., 2017). Among the top CHON compounds with high relative abundance in W1-W3 samples, such as $C_6H_4N_2O_6$ and $C_7H_6N_2O_6$ both with a DBE value~~s~~ of ~~5 and 6, respectively~~, were refer to nitrophenols containing one or two nitrogen-containing functional groups, which have been widely reported in aged BBOA, indicating the increasing of the CHON compounds relative intensity in W1-W3 samples were closely related to BB (Lin et al., 2017; Cai et al., 2020; Mohr et al., 2013; Kourtchev et al., 2016; Lin et al., 2016). Some other top CHON compounds in winter samples such as $C_9H_4NO_4$ and $C_{10}H_6NO_4$ with low O/C and H/C ratios most likely indicated the presence of condensed aromatic structures in the compounds. The $C_9H_4NO_4$ compounds were most likely emitted from vehicle emissions which have previously been reported (Cui et al., 2019).

It is worth noting that some high-intensity CHON compounds with low O/C and H/C ratios were detected in summer samples in this study (Fig. 6a), which were closely related to aromatic compounds from anthropogenic emissions. The top compounds with molecular formulas of $C_8H_5N_2O_2$ and $C_{19}H_{11}N_2O_4$, which had O/N of ~~2~~1 and ~~1~~2, respectively, both were reduced N compounds referring to N-heterocyclic compounds. Previously studies~~y~~ have found that the N-heterocyclic aromatic compounds can be formed through the aldehyde−ammonia reactions (De Haan et al., 2018; Zhang et al., 2022a). This indicated the important role of reduced N species (e.g., ammonium) in the formation of anthropogenic SOA in summer. Our results were consistent with previous study conducted in Xi'an, China which also found formation of reduced N compounds in light-absorbing aerosols through ammonia involved reactions in summer (Zeng et al., 2021).

3.4.3 Molecular characteristics of S-containing compounds (CHOS and CHONS compounds)

The O/C_w of CHOS compounds in summer and winter were 0.60-0.79 and 0.56-0.67, respectively, while the H/C_w were 1.50-1.54 and 1.53-1.72, respectively. The O/C_w of CHONS compounds in summer and winter were 0.82-1.01 and 0.76-0.94, respectively, while the H/C_w were 1.57-1.65 and 1.58-1.66, respectively (Table S1 and S2). As shown in Fig. 7a, 7d, 8a, and 8d, the high-intensity S-containing compounds in summer and winter were both located at the area where O/C >0.5 and H/C >1.5, respectively. In addition, the relative intensity of S-containing compounds increased with the O/C ratios, suggesting the S-containing compounds were highly oxidized. A small number of high-intensity S-containing compounds with O/C <1.0 and H/C <1.0

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were also found in winter in this study, which might be related to OSs and nitrooxy-OSs produced from the oxidation of aromatic hydrocarbon. The CHOS compounds presenting high relative abundance were rich in $O_{6\text{-}9}S$ and $O_{5\text{-}7}S$ groups in summer and winter, respectively, of which the DBE values were all below 4. The CHONS compounds were rich in $O_{8\text{-}10}S$ and $O_{7\text{-}9}S$ groups in summer and winter, respectively, of which the DBE values were all below 6 (Fig. 7b, 7e, 7c, 7f, 8b, 8e, 8c, and 8f). Compared with those of the CHO and CHON compounds, the DBE values of S-containing compounds were significantly lower.

Among the S-containing compounds, more than 95 % of the CHOS, $CHON_1S$, and $CHON_2S$ formulas had O/S ratios greater than 4, 7, and 10, respectively, implying these compounds may contain organic sulfate functional groups ($-OSO_3$) or one or two organic nitrate groups ($-ONO_2$) and these compounds ~~are~~ were more likely OSs or nitrooxy-OSs, presenting lower DBE values and higher O/C and H/C ratios (Table S5 and S6) (O'Brien et al., 2014). The high-intensity CHONS compounds observed in this study, such as $C_{10}H_{16}NO_{7\text{-}9}S$, $C_{10}H_{18}NO_{8\text{-}9}S$, $C_{10}H_{18}N_2O_{11}S$, and $C_9H_{14}NO_{8\text{-}9}S$ could be nitrooxy-OSs derived from monoterpenes such as limonene and ~~,~~ α-terpinene, ~~and monoterpene~~of which we found the formulas in summer contained more oxygen atoms, indicating the higher oxidation degree of these nitrooxy-OSs in summer (Figure S2 and S3) (Sun et al., 2021; Bruggemann et al., 2020; Wang et al., 2020; Wang et al., ~~2018c~~2018d).

The CHOS compounds with high intensity abundance, such as typical isoprene epoxydiols (IEPOX) derived OSs with molecular formulas of $C_5H_8O_7S$ and $C_5H_{10}O_7S$ were both detected in the summer and winter samples, of which the relative intensity of $C_5H_8O_7S$ were over 80 % in S1, S2, S5, and S6 samples, indicating the significant isoprene SOA formation in summer (Kourtchev et al., 2016; Kourtchev et al., 2013). The results were consistent with the ~~previous research on the sources of HULIS based on positive matrix factorization (~~PMF~~)~~ ~~mode~~results~~l~~ reported by ~~our laboratory (~~Bao et al.~~,~~ (2022). The monoterpenes derived OSs such as $C_8H_{14}O_6S$, $C_8H_{14}O_8S$, $C_{10}H_{18}O_8$, $C_{10}H_{14}O_6$, and $C_{11}H_{16}O_7$ were detected in both summer and winter samples in this study, which could refer to monoterpene-OSs derived from α-pinene, α-terpinene, and limonene (Wang et al., 2020). Moreover, OSs with high carbon numbers (C ≥14) such as $C_{14}H_{22}O_7S$, $C_{14}H_{22}O_8S$, $C_{14}H_{24}O_7S$, $C_{15}H_{26}O_7S$, $C_{15}H_{24}O_7S$, $C_{15}H_{24}O_8S$, and $C_{16}H_{28}O_7S$ were also observed in both summer and winter samples. Long-chain alkanes emitted from vehicle emissions might be precursors of these OSs which was consistent with the molecular structures of OSs collected in urban areas affected by traffic emissions such as Shanghai, Los Angeles, and Beijing (Wang et al.,

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2019a; Tao et al., 2014; Wang et al., 2016b). The aromatic OSs such as naphthalene derived OSs with molecular formulas of $C_{10}H_{10}O_6S$, $C_{10}H_{10}O_7S$, and $C_{10}H_{12}O_7S$, 2-methylnaphthalene derived OSs with molecular formulas of $C_9H_{12}O_6S$, $C_{11}H_{12}O_7S$, and $C_{11}H_{14}O_7S$, and hydroxybenzene derived OSs with molecular formulas of $C_6H_6O_5S$ were also observed in this study (Qi et al., 2021; Riva et al., 2015; Blair et al., 2017). Figure S8 further displays the ternary plot of the relative intensities of OSs from biogenic precursors (e.g., isoprene and monoterpenes), long-chain alkanes and aromatic hydrocarbon. As shown in Fig. S8, the biogenic OSs and long-chain alkanes OSs formation were comparable in summer and winter, demonstrating both biogenic and anthropogenic emission contributions to HULIS. The aromatic OSs presented higher relative intensities in winter, further indicating the increasing anthropogenic emissions in winter. The presence of long-chain alkanes derived OSs in both summer and winter seasons provided another evidence that the traffic emission was one of the important fossil sources of HULIS in this study.

3.5 Comparison with organic compounds in source and atmospheric aerosol samples

The O/C and H/C ratios of water soluble HULIS in this study were compared with those of water soluble organic compounds reported in source samples from BB, coal combustions, and vehicle emissions (Tang et al., 2020; Song et al., 2018; Cui et al., 2019; Song et al., 2019), cloud water samples (Bianco et al., 2018; Zhao et al., 2013), rainwater samples (Altieri et al., 2009), fog samples (Brege et al., 2018), as well as aerosol samples collected in Beijing (Jang et al., 2020; Wu et al., 2019a; Wang et al., 2018a), Tianjin (Han et al., 2022), Baoding (Sun et al., 2021), Shanghai (Wang et al., ~~2017a~~2017b), Guangzhou (Jiang et al., 2021), respectively in China, Mainz (Wang et al., ~~2018a~~2018b), Cork city (Kourtchev et al., 2014), and Bologna (Brege et al., 2018), respectively in Europe, and Bakersfield (O'Brien et al., 2014) and Virginia (Willoughby et al., 2014), respectively in the United States (Fig. 9). The O/C ratios were obviously higher than those detected in primary BB, coal combustion, and vehicle emission samples. The H/C ratios of the CHO and CHON compounds were comparable with the source samples, indicating the organics in HULIS experienced atmospheric secondary process and the mixed sources of HULIS in this study. The H/C ratios of the S-containing compounds were much higher than those of source samples which could be attributed to the significant organosulfates formation in the atmosphere.

The O/C ratios reported in this study were also higher than those reported in aerosol samples in urban area in China, further indicating the serious secondary pollution at Nanjing, China. Among the CHO and CHON compounds, we found that the highest H/C ratio values were observed

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in the southern city of Guangzhou, followed by those in Nanjing and Shanghai, and the lowest values were observed in the northern cities such as Beijing, Tianjin, and Baoding, indicating the higher unsaturation degree of the aerosol samples collected from the northern ~~heating~~ cities, which were also considered as the heavy industrial region in China. The higher H/C ratios of aerosol samples collected in Europe and the United States indicated the less anthropogenic emissions such as industrial emissions from those areas.

4. Conclusions

This study focuses on the sources and molecular characteristics differences of water soluble HULIS in summertime and wintertime from 2017 to 2018 at a suburb site of the YRD, China based on the radiocarbon analysis and FT-ICR MS measurement with ESI ion source in negative mode. The carbon isotope analysis results highlight ~~both~~ the important fossil ~~and non-fossil~~ source contributions to HULIS at the study site. A total of 14387 and 15731 peaks were detected in the summer and winter samples, respectively based on the FT-ICR MS results. The assigned molecular formulas were classified into CHO, CHON, CHOS, and CHONS subgroups according to their elemental compositions. The Van Krevelen diagrams showed that more tannins-like and carbohydrates-like species were detected in summer indicating biogenic SOA formation. Whereas more compounds containing condensed aromatic structures were detected in winter which were derived from anthropogenic emissions. The total relative intensity of CHO compounds in summer were significantly higher than those in winter, containing lots of macromolecular oligomers derived from biogenic precursors. The high-intensity CHO compounds in winter were mainly aromatic compounds such as phenolic substances and flavonoids which were related to aged BBOA and oxidized PAHs most probably from fossil fuel combustion. On the contrary, the total relative intensity of CHON compounds significantly increased in winter, mainly composed of nitro compounds or organonitrates. The enhanced formation of nitrophenols in winter indicated the BB influence. The increasing organonitrates formation in winter highlighted the secondary N-containing compounds formation via NO_3 radical-initiated oxidation processes. It is worth noting that the top CHON compounds in summer were referring to aromatic reduced N compounds produced from the aldehyde−ammonia reactions. The S-containing compounds were mainly composed of highly oxidized OSs. The monoterpenes derived OSs and long-chain alkanes derived OSs were widely observed in both summer and winter samples, while the aromatic OSs formation were found to be more significant in winter. The presence of long-chain alkanes derived OSs

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supported the radiocarbon results, ~~proving~~ indicating that the traffic emission was the important fossil sources at the study site. The presence of aromatic secondary N-containing and S-containing compounds provided evidence for the substantial contributions from anthropogenic SOA formation to fossil sources at the study site. These results further verified the work reported before by Bao et al. (2022) based on the PMF model which have found the significant anthropogenic SOA and fossil fuel combustion contributions to HULIS in urban area in China at molecular level. In addition, strong biogenic emission in summer and BB in winter were found in this study, ~~Our results~~ highlighting ~~ed~~ the ~~equal~~ importance of different control policies for each season in the future.~~future reduction in both fossil and non-fossil emissions on atmospheric pollution control.~~

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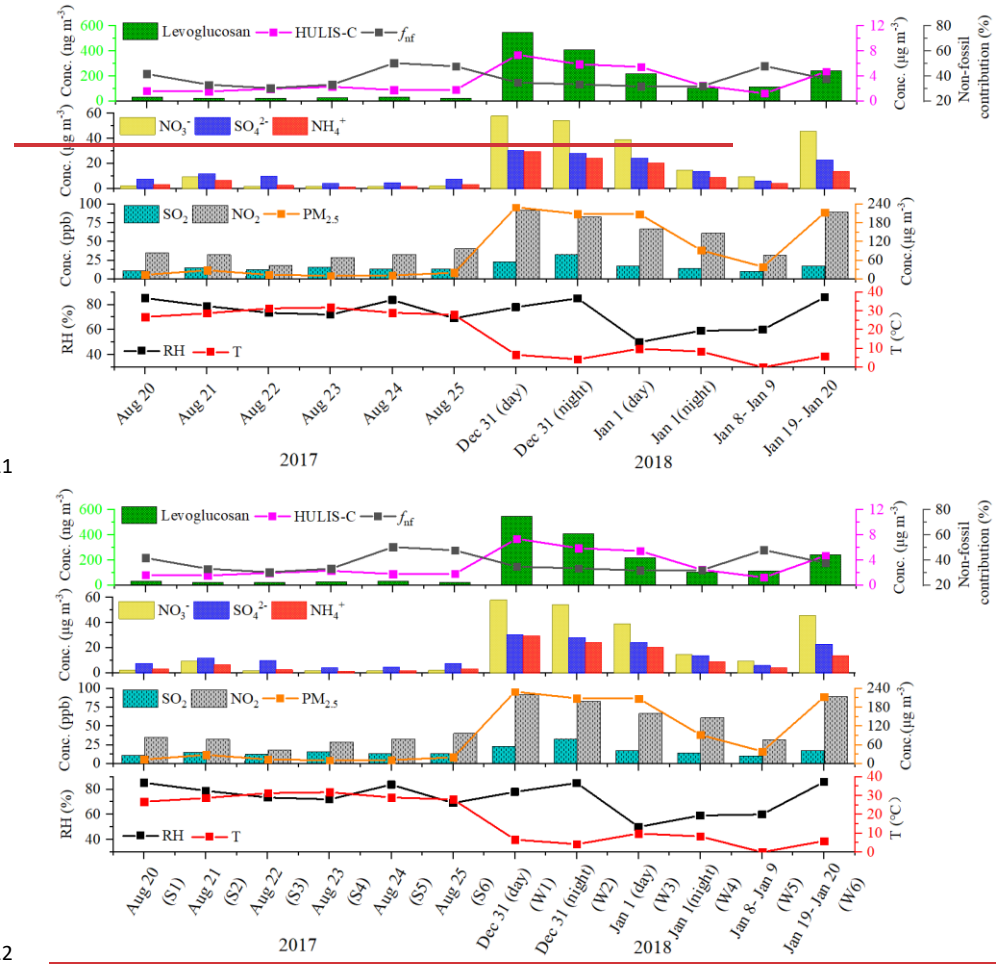


Figure 1. Time series of non-fossil contributions to HULIS-C, the mass concentrations of HULIS-C, Levoglucosan, NO_3^-, SO_4^{2-}, NH_4^+, SO_2, NO_2, and $PM_{2.5}$, relative humidity, and temperature during the study periods.

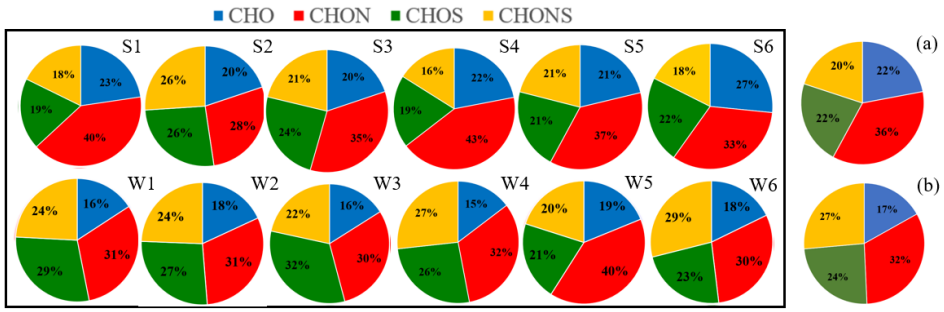


Figure 2. Pie graph of the number percentages of each elemental formula group for the 12 samples plotted in the box and the averaged number percentages of each elemental formula group for the summer samples (a) and winter samples (b).

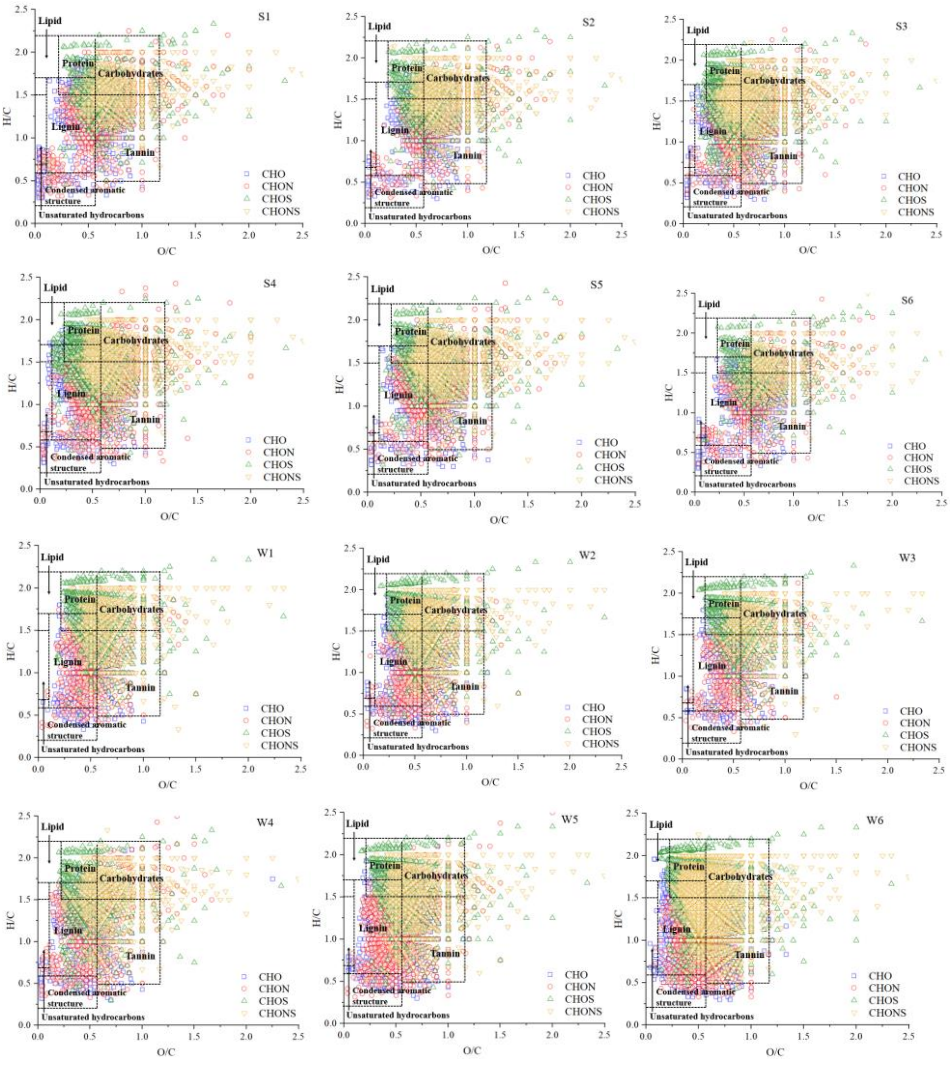


Figure 3. Van Krevelen diagrams of the 12 samples.

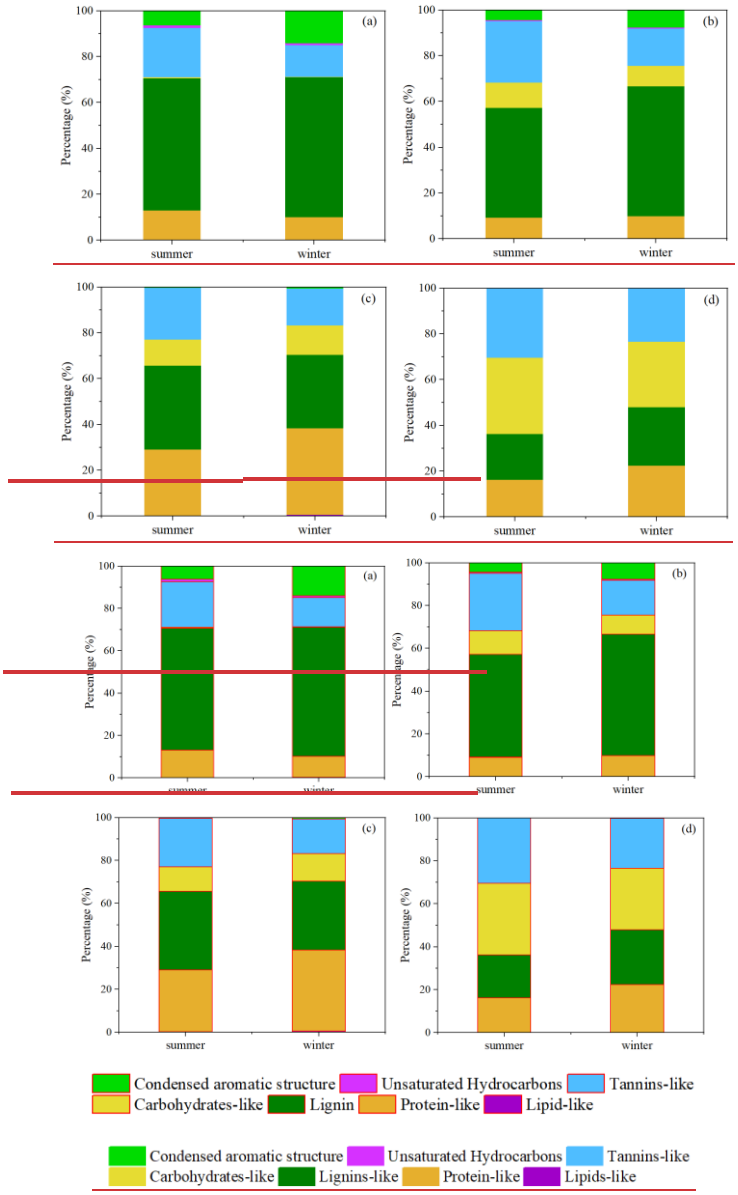


Figure 4. Contributions of seven categories in CHO (a), CHON (b), CHOS (c), and CHONS (d) compounds.

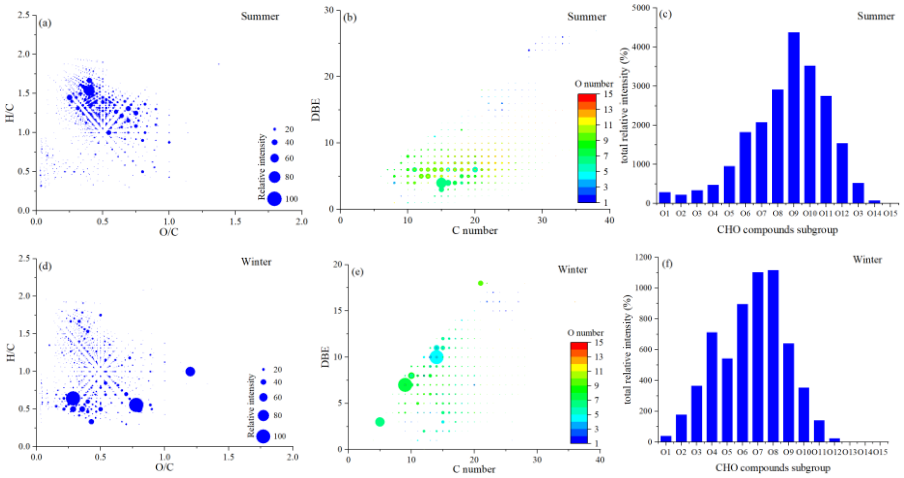


Figure 5. Van Krevelen diagram ((a) and (d)), plot of DBE values vs carbon atom numbers ((b) and (e)), and the total relative intensity of each subgroup ((c) and (f)) for the CHO compounds in summer and winter.

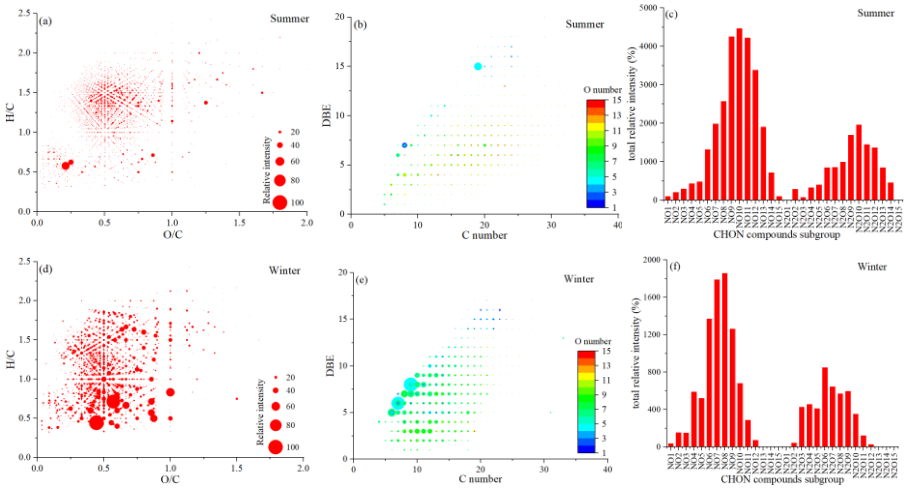


Figure 6. Van Krevelen diagram ((a) and (d)), plot of DBE values vs carbon atom numbers ((b) and (e)), and the total relative intensity of each subgroup ((c) and (f)) for the CHON compounds in summer and winter.

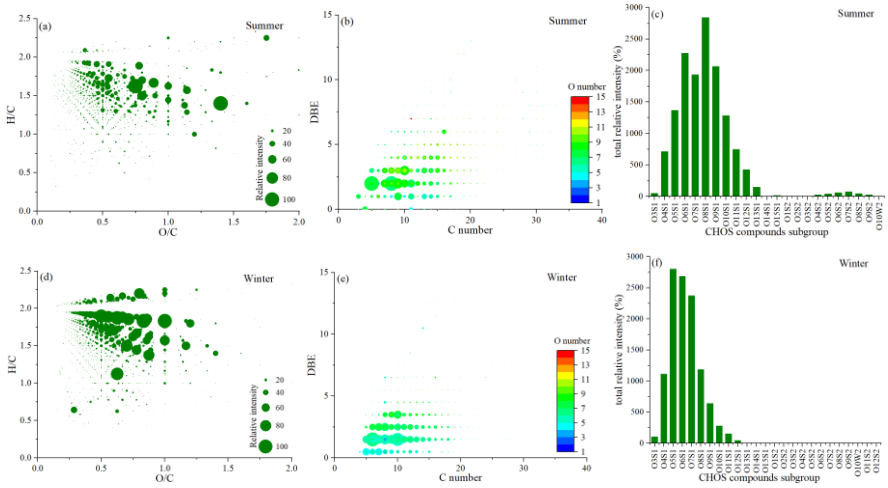


Figure 7. Van Krevelen diagram ((a) and (d)) , plot of DBE values vs carbon atom numbers ((b) and (e)), and the total relative intensity of each subgroup ((c) and (f)) for the CHOS compounds in summer and winter.

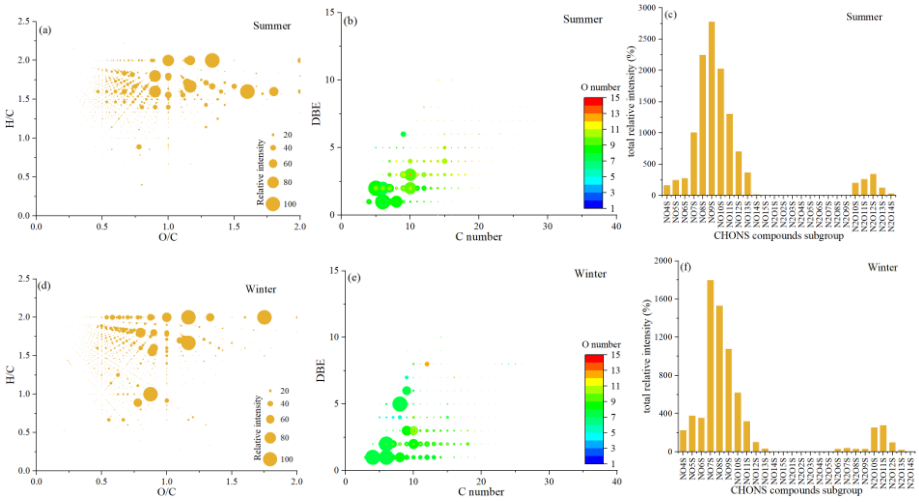
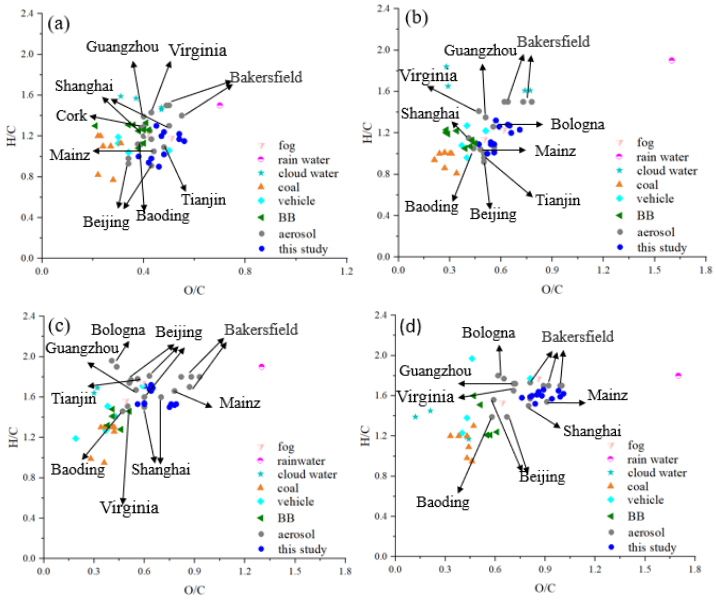


Figure 8. Van Krevelen diagram ((a) and (d)), plot of DBE values vs carbon atom numbers ((b) and (e)), and the total relative intensity of each subgroup ((c) and (f)) for the CHONS compounds in summer and winter.



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1056 Figure 9. Comparison of O/C and H/C ratios of water soluble organic compounds in different
1057 atmospheric media in CHO (a), CHON (b), CHOS (c), and CHONS (d) compounds.