# A new insight into the vertical differences in NO<sub>2</sub> heterogeneous reaction to produce HONO over inland and marginal seas

Chengzhi Xing<sup>a</sup>, Shiqi Xu<sup>g</sup>, Yuhang Song<sup>b</sup>, Cheng Liu<sup>b,a,c,d,\*</sup>, Yuhan Liu<sup>f</sup>, Keding Lu<sup>e,\*</sup>, Wei Tana, Chengxin Zhangb, Qihou Hua, Shanshan Wangi, Hongyu Wuh, and Hua Linh

- g School of Earth and Space Sciences, University of Science and Technology of China, Hefei, 230026, China.
- h School of Environmental Science and Optoelectronic Technology, University of Science and Technology of China, Hefei, 230026,
- <sup>i</sup> Shanghai Key Laboratory of Atmospheric Particle Pollution and Prevention (LAP<sup>3</sup>), Department of Environmental Science and Engineering, Fudan University, Shanghai, 200433, China

#### **ABSTRACT**

1

2

3

4

5 6 7

8

9

10

11

12

13

14

15

16

17

18

19

20

21

22

23

24

25 26 Ship based multi-axis differential optical absorption spectroscopy (MAX-DOAS) measurements were conducted along the marginal seas of China from 19 April to 16 May 2018 to measure the vertical profiles of aerosol, nitrogen dioxide (NO<sub>2</sub>), and nitrous acid (HONO). Along the cruise route, we found five hot spots with enhanced tropospheric NO<sub>2</sub> VCDs in Yangtze River Delta, Taiwan straits, Guangzhou-Hong Kong-Macao Greater Bay areas, Zhanjiang Port, and Qingdao port. Enhanced HONO concentrations could usually be observed under high-level aerosol and NO<sub>2</sub> conditions, whereas the reverse was not always the case. To understand the impacts of relative humidity (RH), temperature, and aerosol on the heterogeneous reaction of NO<sub>2</sub> to form HONO in different scenes, the Chinese Academy of Meteorological Sciences (CAMS) and Southern University of Science and Technology (SUST) MAX-DOAS stations were selected as the inland and coastal cases, respectively. The RH turning points in CAMS and SUST cases were both ~65% (60–70%), whereas two turning peaks (~60% and ~85%) of RH were found in the sea cases. As temperature increased, the HONO/NO<sub>2</sub> ratio decreased with peak values appearing at ~12.5°C in CAMS, whereas the HONO/NO<sub>2</sub> gradually increased and reached peak values at ~31.5°C in SUST. In the sea case, when the temperature exceeded 18.0°C, the HONO/NO<sub>2</sub> ratio rose with increasing temperature and achieved its peak at ~25.0°C. This indicated that high temperature can contribute to the secondary formation of HONO in the sea atmosphere. In the inland case, the correlation analysis between HONO and aerosol in the near-surface layer showed that the ground surface is more crucial to the formation of HONO via the heterogeneous reaction of NO<sub>2</sub>; however, in the coastal and sea cases, the aerosol surface contributed more. Furthermore, we discovered that the conversion rate of NO<sub>2</sub> to HONO through heterogeneous reaction in the sea case is larger than that in the inland case in higher atmospheric layers (> 600 m). Three typical events were selected to demonstrate three potential contributing factors of HONO production under marine conditions (i.e., transport, NO<sub>2</sub> heterogeneous reaction, and unknown HONO source). This study elucidates the sea-land and vertical differences in the forming mechanism of HONO via the NO<sub>2</sub> heterogeneous reaction and provides deep insights into tropospheric HONO distribution, transforming process, and environmental effects.

# 1 Introduction

27 Nitrous acid (HONO) is an important part of the atmospheric nitrogen cycle and plays a significant role in 28 atmospheric oxidation capacity (Alicke et al., 2003; Kleffmann et al., 2005). Photolysis of HONO in near ultraviolet 29 bands (Eq. 1) is a substantial source of hydroxyl radicals (OH radicals), which are one of the most important oxidants 30

in the tropospheric atmosphere. Earlier studies reported that the contribution of HONO photolysis to OH radicals can

<sup>&</sup>lt;sup>a</sup> Key Lab of Environmental Optics & Technology, Anhui Institute of Optics and Fine Mechanics, Hefei Institutes of Physical Science, Chinese Academy of Sciences, Hefei 230031, China

b Department of Precision Machinery and Precision Instrumentation, University of Science and Technology of China, Hefei, 230026,

<sup>&</sup>lt;sup>c</sup> Center for Excellence in Regional Atmospheric Environment, Institute of Urban Environment, Chinese Academy of Sciences, Xiamen,

d Key Laboratory of Precision Scientific Instrumentation of Anhui Higher Education Institutes, University of Science and Technology of China, Hefei, 230026, China

e State Key Joint Laboratory of Environment Simulation and Pollution Control, College of Environmental Sciences and Engineering, Peking University, Beijing 100871, China

f Department of unclear safety, Chia Institute of Atomic Energy, Beijing, 102413, China

<sup>\*</sup>Corresponding authors: Cheng Liu (chliu81@ustc.edu.cn); Keding Lu (k.lu@pku.edu.cn)

reach 40–60%, while exceeding 80% in the early morning (Michoud et al., 2012; Ryan et al., 2018; Xue et al., 2020). OH radicals can oxidize and destroy most atmospheric pollutants, such as CO, NO<sub>x</sub> (NO+NO<sub>2</sub>), SO<sub>2</sub>, and volatile organic compounds (VOCs), thereby further promoting the formation of secondary pollutants (e.g., ozone (O<sub>3</sub>), peroxyacetyl nitrate (PAN), and secondary aerosols) and leading to serious haze pollution events (Huang et al., 2014). Additionally, as a nitrosating agent, HONO can produce carcinogenic nitrite amines that pose a threat to human health (Zhang et al., 2015). Therefore, a full understanding of the source and formation mechanism of HONO is scientifically significant for the study of tropospheric oxidation and the control of secondary pollution.

Currently, the known sources of HONO mainly include direct emissions from vehicles, ships, biomass burning and soil, the homogeneous reaction of NO and OH radicals (Eq. 2), the nighttime and daytime heterogeneous reaction of NO<sub>2</sub> (Eq. 3) on aerosols, vegetation, ground and other types of surfaces, and the photolysis of nitrate particles (Eq. 4) (Alicke et al., 2003; Stemmer et al., 2006; Indarto et al., 2012; Wang et al., 2015; Salgado and Rossi, 2002; Zhou et al., 2011). Sources of HONO exist that are poorly understood (Fu et al., 2019). The heterogeneous reaction of NO2 as a source of HONO has received continuous attention in recent years. It was found that the heterogeneous reaction of NO<sub>2</sub> is one of the most important sources of HONO in a variety of scenes such as inland, coastal cities, and offshore seas. Liu et al. (2021) reported the contribution of heterogeneous reaction of NO<sub>2</sub> on aerosol surface to HONO is 19.2% in summer, and this contribution on aerosol and ground surfaces to HONO can reach 54.6% in winter in Beijing. Yang et al. (2021) and Zha et al. (2014) found that the generation rate of HONO through the heterogeneous reaction of NO<sub>2</sub> under sea-wind conditions could elevate 3-4 times than that under land-wind conditions in the northern coastal city of Qingdao and the southern coastal city of Hong Kong, respectively. Cui et al. (2019) illustrated that the heterogeneous reaction of NO<sub>2</sub> on aerosol and sea surfaces is an important source of HONO in East China Sea in summer. The process of HONO formed from the heterogeneous reaction of NO<sub>2</sub> is affected by various atmospheric parameters. The relative humidity (RH), temperature, solar radiation intensity (SRI), and aerosol concentration and its relative surface area are the particularly important parameters. Earlier works always used the linear regression relationship between HONO/NO<sub>2</sub> and the above parameters to characterize the influence of these parameters on the formation of HONO through the heterogeneous reaction of NO2. Although this kind of simple linear regression method may lead to artificial correlations and misleading conclusions, considering the vertical evolution of atmospheric parameters. Wen et al. (2019) found that the increased temperature could promote the heterogeneous reaction of NO2 to form HONO in sea conditions. The generation rate of HONO could increase rapidly, when the temperature was greater than 20 °C. Gil et al. (2019) found that the HONO formed from the heterogeneous reaction of NO<sub>2</sub> will increase along with the increase of RH when RH was less than 80% in a case of land park using deep learning forced by measurement results. Fu et al. (2019) reported that RH and SRI were the main parameters driving the heterogeneous reaction of NO<sub>2</sub> to form HONO in Pearl River Delta, and it contributes to 72% of the total source of HONO. Cui et al. (2019) found that the potential of heterogeneous reaction of NO<sub>2</sub> to form HONO will increase with the increase of particle concentration and the specific surface area of single particle in coastal cities.

$$65 HONO + hv \rightarrow NO + OH(\lambda < 400nm) (1)$$

$$66 \cdot OH + NO + M \rightarrow HONO + M (2)$$

$$67 2NO_2 + H_2O \rightarrow HONO + HNO_3 (3)$$

68 
$$HNO_3 / NO_3^- + hv \rightarrow HONO / NO_2^- + O \cdot (\lambda \sim 300nm)$$
 (4)

However, earlier researches generally focused on the near-surface layer of a single scene, and attentions to the influence mechanism of the heterogeneous reaction of NO<sub>2</sub> to form HONO in vertical direction and in different sea and land scenes are insufficient, which limits the comprehensive assessment to understand the sea-land differences and impact mechanism of HONO formed from the heterogeneous reaction of NO<sub>2</sub>. NO<sub>2</sub> could be transported from inland and coastal cities to offshore seas (Tan et al., 2018). This part of NO<sub>2</sub> can promote the HONO formation through heterogeneous reaction on the high-level aerosol and sea surfaces in the atmosphere of sea (Zhang et al., 2020). The formed HONO is likely to be carried to land cities at night by sea breeze, which will affect the atmospheric oxidation and air quality, and even endanger human health. Additionally, the vertical distributions and values of atmospheric meteorology and aerosol parameters are significantly different in land and sea scenes, which provide different conditions for the heterogeneous reaction of NO<sub>2</sub> to form HONO in different height layers. Furthermore, aerosols and NO<sub>2</sub> have complex evolution and transmission characteristics in the vertical direction. The vertical upward transport of aerosol and NO<sub>2</sub> can promote the HONO formation through heterogeneous reaction at high altitude, and the vertical downward transport of HONO will impact the atmospheric environment near the ground. The vertical observations in land-sea scenes are also helpful to distinguish the contribution of the heterogeneous reaction of NO<sub>2</sub> on the aerosol and ground/sea surfaces (Zhang et al., 2020).

Currently, a variety of HONO measurement techniques have been developed, which in principle can be roughly divided into wet chemical, spectroscopy, and mass spectrometry methods (Cheng et al., 2013; Bernard et al., 2016;

Gil et al., 2019; Guo et al., 2020; Jordan et al., 2020). However, these technical methods can only measure the HONO

information near the surface layer. Taking tower and aircraft as platforms, these techniques were performed to measure HONO vertical profiles, and it was found that the peak values of HONO usually appeared under 200 m at urban and suburban areas (Kleffmann et al., 2003; Stemmler et al., 2006; Zhang et al., 2009; Wong et al., 2012; Meng et al., 2020; Zhang et al., 2020). These studies also revealed that the heterogeneous reaction of NO<sub>2</sub> on multiple surfaces (ground and aerosol etc.) was an important source of HONO under planetary boundary layer (PBL), especially in haze days. Furthermore, they also reported that the HONO/NO2 ratios usually decreased with the increase of height under 200 m at inland and coastal areas. However, the cost of above techniques used to measure HONO vertical profiles was too high, and the real-time and continuous measurement cannot be realized. Multi-axis differential optical absorption spectroscopy (MAX-DOAS), as a ground-based ultra-hyperspectral remote sensing technology, was widely used for vertical observation of atmospheric pollutants in the past two decades. In the past five years, several researchers carried out campaigns based on MAX-DOAS to measure the vertical profile of HONO in inland and coastal areas, and revealed their vertical characteristics, sources, and the contribution to atmospheric oxidation at different height layers (Garcia-Nieto et al., 2018; Ryan et al., 2018; Wang et al., 2020; Xing et al., 2021; Xu et al., 2021; He et al., 2023). Few studies were conducted on the sources of HONO at different height layers in sea conditions. In this study, MAX-DOAS is used for the first time to study the spatiotemporal distribution and the sources of HONO along the Chinese coastline, and to learn the differences of the HONO formed from the heterogeneous reaction of NO<sub>2</sub> in different height layers and land-sea scenes.

## 2 Methods and methodologies

## 2.1 Measurement cruise

87

88

89

90

91

92

93

94 95

96

97

98

99

100

101

102

103

104

105

106

107

108

109

110

111

112

113

114

115

116

125

126

136

137

138

139

140

141

142

143

The ship-based atmospheric observation campaign along the marginal seas of China was conducted from 19 April to 16 May 2018. The latitude and longitude ranges of the entire campaign covered 21.12°N–35.89°N and 110.67°E–122.16°E. The detailed voyage records of the observation ship are shown in Table 1. An integrated and fully automated MAX-DOAS instrument was installed aboard the stern deck of the ship (Figure S1(a)). To ensure that the instrument is always kept in a horizontal position, a photoelectric gyro was used. The angle between the observation and heading directions of the ship was always maintained at 135° during the whole campaign. The telescope unit of the instrument pointed towards sea during cruise NO.3 and NO.6. The telescope unit pointed towards inland during cruise NO.1, NO.4, and NO.5. During cruise NO.2, the observation telescope always pointed to Chongming island. The measurement ship only sailed in daytime from 19 April to 02 May, and continuously sailed in all the daytime and nighttime from 3 May to 16 May 2018. The ship docked in Daishan port on 9–10 May and no observations were conducted during these two days.

117 The aim of this campaign was to learn the vertical differences of NO<sub>2</sub> heterogeneous reaction to produce HONO in 118 marginal seas of China and compare the influence mechanism of that in inland cities. To fully understand the 119 differences of the impacts of RH, temperature, and aerosol on the HONO secondary formation in land and sea 120 conditions, the Chinese Academy of Meteorological Sciences (CAMS) and Southern University of Science and 121 Technology (SUST) MAX-DOAS stations were selected as inland and coastal areas for analysis, respectively. CAMS 122 is located in the urban of Beijing (116.32°E, 39.94°N), and SUST is located in Shenzhen (114.00°E, 22.60°N) (Figure 123 S2). This study will provide scientific guidance for understanding regional oxidation capacity and controlling the 124 secondary air pollution.

#### 2.2 MAX-DOAS measurements

#### 2.2.1 Instrument setup

127 The compact instrument consists of an ultraviolet spectrometer (AvaSpec-ULS2048L-USB2, 300-460 nm spectral 128 range, 0.6 nm spectral resolution) at a 20°C fixed temperature with a deviation of < 0.01°C, a one-dimensional CCD 129 detector (Sony ILX511, 2048 individual pixels) and a telescope unit driven by a stepper motor to collect scattered 130 sunlight from different elevation angles. The accuracy of elevation angle is < 0.1° and the telescope field of view 131 (open angle) is  $< 0.3^{\circ}$ . A full scanning sequence consists of 11 elevation angles (1°, 2°, 3°, 4°, 5°, 6°, 8°, 10°, 15°, 30°, 132 and 90°). The integration time of one individual spectrum was set to 30 s, and each scanning sequence took about 5.5 133 min. Besides, the controlling electronic devices and connecting fiber are mounted inside. The instrument is equipped 134 with a high-precision Global Position System (GPS) to record the real-time coordinated positions of the cruise ship. 135 The detailed description of the setup of MAX-DOAS in CAMS and SUST can be found in Liu et al. (2021).

## 2.2.2 Data processing and filtering

The MAX-DOAS measurements could be influenced by the exhaust from the measurement ship. Therefore, the data contaminated by the exhaust were filtered out. As shown in Figure S1(b), the direction and speed of the plume exhausted from the ship depends on the ship and the true wind speeds/directions. Individual measurements taken under unfavorable plume directions (plume directions between 45 and 135° with respect to the heading of the ship) were discarded. To avoid the strong influence of the stratospheric absorption, the spectra measured with solar zenith angle (SZA) lager than 75° were filtered out. Under these two filtering criteria, 4.9 and 8.3% of all data were rejected before DOAS analysis (Xing et al., 2017, 2019, 2020).

#### 2.2.3 DOAS analysis

The MAX-DOAS measured spectra were analyzed using the software QDOAS which is developed by BIRA-IASB (http://uv-vis aeronomie.be/software/QDOAS/). The DOAS fit results are the differential slant column densities (DSCDs), i.e. the difference of the slant column density (SCD) between the off-zenith and the corresponding zenith reference spectra. Details of the DOAS fit settings are listed in Table1. A typical DOAS retrieval example for the oxygen dimer (O<sub>4</sub>), nitrogen dioxide (NO<sub>2</sub>), and nitrous acid (HONO) are shown in Figure 1. The stratospheric contribution was approximately eliminated by taking the zenith spectra of each scan as reference in the DOAS analysis. Before profile retrieval, DOAS fit results of O<sub>4</sub>, NO<sub>2</sub>, and HONO with root mean square (RMS) of residuals larger than  $3 \times 10^{-3}$  were filtered. Furthermore, the SCD data under the color index (CI) being < 10% of the thresholds obtained through fitting a fifth-order polynomial to CI data which is a function of time was filtered out to ensure a high signal-to-noise ratio (SNR) of the spectra. This filtering criteria remove 2.1, 3.9, and 5.3% for O<sub>4</sub>, NO<sub>2</sub>, and HONO, respectively.

# 2.3 Vertical profile retrieval

Aerosol and trace gases (i.e., NO<sub>2</sub> and HONO) vertical profiles are retrieved from MAX-DOAS measurements using the algorithm reported by Liu et al. (2021). The inversion algorithm is developed based on the Optical Estimation Method (OEM) (Rodgers, 2000), which employs the radiative transfer model VLIDORT as the forward model. The detailed retrieval procedure is displayed in Appendix I and Figure S3.

In this study, an exponential decreasing a priori with a scale height of 1.0 km was used as the initial profile for both the aerosol and trace gases retrieval (Figure S4). The surface concentrations of aerosol, NO<sub>2</sub>, and HONO were set to 0.2 km<sup>-1</sup>, 3.0 ppb, and 1.0 ppb, respectively. We assume a fix set of aerosol optical properties with asymmetry parameter of 0.69, a single scattering albedo of 0.90, and ground albedo of 0.05. Furthermore, the uncertainty of the aerosol and trace gases a priori profile was set to 100% and the correlation length was set to 0.5 km. The averaging kernels indicated that the sensitivity of the profile retrieval tended to decrease with increasing altitude, and was especially sensitive to the layers within 0–1.5 km (Figure S5). The sum of the diagonal elements in the averaging kernel matrix is the degrees of freedom (DOF), which denotes the number of independent pieces of information contained in the measurements.

#### 2.4 Error analysis

For profile retrieval, the error sources can be divided into four different types: smoothing error, measurement noise error, forward model error, and model parameter error (Rodgers, 2004). However, in terms of this classification, some errors are difficult to be calculated or estimated. For example, forward model error, which is caused by an imperfect representation of the physics of the system, is hard to be quantified due to the difficulty of acquiring an improved forward model. Given calculation convenience and contributing ratios of different errors in total error budget, we mainly took into account error sources based on the following classification, which were smoothing and noise errors, algorithm error, cross section error, and uncertainty related to the aerosol retrieval (only for trace gas). Here, we estimated the contribution of different error sources to the trace gas vertical column densities (VCDs) and AOD, and near-surface (0–200 m) trace gas concentrations and aerosol extinction coefficients (AECs), respectively. The detailed demonstrations and estimation methods are displayed below, and the final results are summarized in Table 3.

- a. Smoothing errors arise from the limited vertical resolution of profile retrieval. Measurement noise errors denote the noise in the spectra (i.e., the fitting error of DOAS fits). They can be quantified by averaging the error of retrieved profiles, as the error of the retrieved state vector equals the sum of these two independent errors. We calculated the sum of smoothing and noise errors on near-surface concentrations and column densities, which were 14 and 5 % for aerosols, 16 and 17 % for NO<sub>2</sub>, and 20 and 22 % for HONO, respectively in the sea scene. The corresponding values were 13 and 5 % for aerosols, 14 and 16 % for NO<sub>2</sub>, and 18 and 20 % for HONO, respectively at SUST and 13 and 5 % for aerosols, 15 and 17 % for NO<sub>2</sub>, and 19 and 21 % for HONO at CAMS.
- b. Algorithm error is the discrepancy between the measured and modelled DSCDs. This error contains forward model error from an imperfect approximation of forward function (e.g., spatial inhomogeneities of absorbers and aerosols), forward model parameter error from selection of parameters, and error not related to the forward function parameters, such as detector noise (Rodgers, 2004). Algorithm error is a function of the viewing angle, and it is difficult to assign this error to each altitude of profile. Usually, the algorithm errors on the near-surface values and column densities are estimated by calculating the average relative differences between the measured and modeled DSCDs at the minimum and maximum elevation angle (except 90 %), respectively (Wagner et al., 2004). Considering its trivial role in the total error budget, we estimated these errors on the near-surface values and the column densities at 4 and 8 % for aerosols, 3 and 11 % for NO<sub>2</sub>, and 20 and 20 % for HONO, according to Wang et al. (2017) and Wang et al. (2020).
- c. Cross section error is the error arising from an uncertainty in the cross section. According to Thalman and Volkamer, (2013), Vandaele et al. (1998), and Stutz et al. (2000), we adopted 4, 3, and 5 % for  $O_4$  (aerosols),  $NO_2$ , and HONO, respectively.
- d. The trace gas profile retrieval error represents the one, which is sourced from aerosol extinction profile retrieval and propagated to retrieved trace gas profile. This error could be roughly estimated based on a linear propagation of the total error budgets of the aerosol retrievals. The errors of trace gases were roughly estimated at 15% for

VCDs and 10% for near-surface concentrations for the two trace gases in the sea scene. The corresponding values were 14 and 10% for near-surface concentrations and VCDs, respectively at SUST, and 14 and 10% at CAMS.

The total uncertainty was calculated by adding all the error terms in the Gaussian error propagation, and the final results were listed in the bottom row of Table 3. We found that the sum of smoothing and noise errors played a dominant role in the total uncertainty.

# 2.5 Ancillary data

204

205

206

207

208

209

220

221

222

223

224

225

226

227

228

229

230

231

232

233

234

235

236

237

238

239

240241

242

243

244

245246

247

248

249

250

251

252

253

254

255

256

257

258

259

260

261

262

Meteorological data (including temperature, pressure, relative humidity, visibility, solar radiation intensity, wind speed, and wind direction) with a temporal resolution of 1 min was measured in the weather station installed on the ship. NO was measured using NO analyzer (Thermo Scientific model 42i) with a 1 min resolution. The speed of the

ship was calculated referring to the GPS data.

- The temperature and relative humidity of two ground-based stations (i.e., CAMS and SUST) were collected from
- Weather Underground website, temporal resolution of which is around 3 hours.
- The backward trajectory was calculated using HYSPLIT (Hybrid Single-Particle Lagrangian Integrated Trajectory)
- developed by the National Oceanic and Atmospheric Administration Air Resource Laboratory (NOAA-ARL). The
- 218 meteorological data with a 1°×1° spatial resolution and 24 layers were collected from the Global Data Assimilation
- 219 System (GDAS).

# 3 Results and Discussion

# 3.1 Overview of the MAX-DOAS observation over marginal seas of China

A radiative transfer model SCIATRAN was used to convert SCDs of NO<sub>2</sub> and HONO to their tropospheric VCDs. The vertical profiles of aerosol, NO<sub>2</sub>, and HONO retrieved from MAX-DOAS, the temperature and pressure vertical profiles simulated using a dynamical-chemical model (WRF-Chem), and the geo-position data collected by GPS were introduced as inputs in SCIATRAN for the NO<sub>2</sub> and HONO air mass factor (AMF) calculation. Missing data are due to power and instrument system failure, interference of ship plume, unfavorable weather condition (i. e., heavy rain), and night sailing. During the cruise of Chongming to Zhanjiang, NO<sub>2</sub> VCDs varied from  $1.05 \times 10^{14}$  to  $4.02 \times 10^{16}$  molec.cm<sup>-2</sup> with an averaged value of  $3.90 \times 10^{15}$  molec. cm<sup>-2</sup>. From Zhanjiang to Qingdao, NO<sub>2</sub> VCDs varied from  $1.08 \times 10^{14}$  to  $2.60 \times 10^{16}$  molec.cm<sup>-2</sup> with an averaged value of  $4.27 \times 10^{15}$  molec. cm<sup>-2</sup>. From Chongming to Zhanjiang, HONO VCDs varied from  $1.00 \times 10^{14}$  to  $2.58 \times 10^{15}$  molec. cm<sup>-2</sup> with a mean value of  $2.39 \times 10^{14}$ molec. cm<sup>-2</sup>. From Zhanjiang to Qingdao, HONO VCDs varied from  $1.01 \times 10^{14}$  to  $2.61 \times 10^{15}$ molec. cm<sup>-2</sup> with a mean value of  $2.74 \times 10^{14}$  molec. cm<sup>-2</sup>.

Figure 2 showed the spatial distribution of NO<sub>2</sub> and HONO VCDs over the marginal seas of China. Five enhanced tropospheric NO2 VCDs hot spots were observed during the whole campaign, i.e., the coastal areas of Yangtze River Delta, Taiwan straits, Guangzhou-Hong Kong-Macao Greater Bay areas, Zhanjiang Port, and Qingdao port. In the coastal areas of Yangtze River Delta, the hot spots were mainly distributed in the Yangtze River estuary, Hangzhou Bay, Ningbo port, Taizhou port, and Wenzhou port. These areas are mostly important shipping channels or shipping ports, and are great NO<sub>2</sub> emission sources. The averaged NO<sub>2</sub> VCDs in above five areas reached 1.07×10<sup>16</sup>, 1.30×10<sup>16</sup>, 7.27×10<sup>15</sup>, 5.34×10<sup>15</sup>, and 3.12×10<sup>15</sup> molec. cm<sup>-2</sup>, respectively (Figure S6(a)). HONO exhibited similar spatial distribution characteristics as NO<sub>2</sub>, and the averaged HONO VCDs in above five hot-spot areas reached 1.01×10<sup>15</sup>,  $7.91 \times 10^{14}$ ,  $6.02 \times 10^{14}$ ,  $5.36 \times 10^{14}$ , and  $5.17 \times 10^{14}$  molec. cm<sup>-2</sup>, respectively (Figure S6(b)). It indicates that NO<sub>2</sub> is an important precursor of HONO. Earlier studies reported that HONO can be generated from NO<sub>2</sub> through heterogeneous reaction on the surface of aerosol and sea (Yang et al., 2021). However, there are obvious differences in the concentration distribution of HONO and NO<sub>2</sub> in the southeast coastal area of Jiangsu (from Qidong to Dongtai). In this area, NO<sub>2</sub> showed a higher concentration (1.66×10<sup>16</sup> molec. cm<sup>-2</sup>, which is 4 times higher than the mean NO<sub>2</sub> VCD), while HONO showed a lower concentration (2.06×10<sup>14</sup> molec. cm<sup>-2</sup>, which is ~80% of the mean HONO VCD). It may be the fresh ship emission plume on the route enhancing the NO<sub>2</sub> concentration and HONO has not been fully formed from NO<sub>2</sub> heterogeneous reaction in time, since the observations from ship-based MAX-DOAS are instantaneous.

The surface concentration of NO<sub>2</sub> and HONO were extracted from their corresponding vertical profiles. As shown in Figure 3, the total averaged near-surface NO<sub>2</sub> concentrations under sea-oriented and land-oriented measurements were 8.46 and 11.31 ppb, respectively. The total averaged near-surface HONO concentrations were 0.23 and 0.27 ppb under sea-oriented and land-oriented measurements. The total averaged near-surface HONO/NO<sub>2</sub> ratios in sea-oriented and land-oriented measurements were 0.027 and 0.024, respectively. Earlier studies reported that vehicle and ship emissions were the main primary HONO sources on land and sea, respectively, and NO<sub>2</sub> heterogeneous reaction on the surfaces of ground, sea, vegetation, and aerosol were the important secondary HONO sources (Liu et al., 2021). Additionally, they found that the surface HONO concentration under the sea case was lower than that under the land case, especially in the morning and evening (Yang et al., 2021). Figure 4 showed the time series of AOD, the surface concentrations of NO<sub>2</sub> and HONO, and the surface HONO/NO<sub>2</sub> during the whole campaign. We found that the time series of AOD and NO<sub>2</sub> were similar. The high AOD and NO<sub>2</sub> usually appeared in busy shipping channels and ports, and the obvious high-value areas were the coast of the Yangtze River Delta, the Taiwan Strait, Xiamen port, Zhanjiang port, and Qingdao port (with mean AOD and NO<sub>2</sub> of 1.28 and 18.90 ppb, respectively). HONO always

appeared under high AOD and NO<sub>2</sub> conditions, however, high AOD and NO<sub>2</sub> were not necessarily accompanied with high HONO concentration. This was because the heterogeneous formation of HONO requires suitable meteorological conditions (i.e., RH and temperature) in addition to its precursor (NO<sub>2</sub>) and the reaction surface (aerosol) (Liu et al., 2019). The high HONO/NO<sub>2</sub> values were found on O2, 13, and 14 May with an average value of 0.45. Furthermore, we found the high values of HONO/NO<sub>2</sub> always appeared from 11:00 to 14:00 during a whole day.

# 3.2 Relationship between HONO/NO2 with RH, Temperature, and aerosol in land and sea

Sun et al. (2020) reported that HONO concentrations could increase up to 40-100% over the shipping routes and international ports, and Huang et al. (2017) reported vehicle exhaust could contribute to ~12-49% of the atmospheric HONO budget. Since the direct emissions of the measurement ship were removed before data analysis, the primary source of HONO during the whole campaign was mainly from the direct emissions of cargo ships. By subtracting the average marine background of NO<sub>x</sub> and HONO from the ship plume emission values, the impact of background values is reduced and the emission ratio of  $\Delta HONO/\Delta NOx$  can be obtained, and this emission ratio can be used for quantifying the primary HONO (Sun et al., 2020; Xu et al., 2015). In this study, we used an averaged 0.46±0.31% emission ratio of  $\Delta HONO/\Delta NOx$  referring to Sun et al. (2020) to understand the primary source of HONO on the sea surface during the campaign. The NO was measured using in situ instrument, and sea-surface NO<sub>2</sub> was extracted from the retrieved  $NO_2$  vertical profiles ( $NO_x = NO + NO_2$ ). Additionally, the calculation method of emission ratios of ΔHONO/ΔNOx in CAMS and SUST was referred from Xu et al. (2015), Liu et al. (2018), and Xing et al. (2021) (Appendix II). The averaged emission ratios in CAMS and SUST were 0.82 ±0.34% and 0.79 ±0.31%, respectively. The direct emissions were deduced in the following study of the secondary formation of HONO. The ratios of HONO/NO2 in CAMS, SUST, and the ship-based campaign could be found in Figure S7. Furthermore, the main secondary formation pathway of HONO is considered as the heterogeneous reaction of NO<sub>2</sub> on the surface. The linear regression between HONO and NO<sub>2</sub> in land and static sea scenarios is shown in Figure 5. We found the fitting slopes in static sea scenes was ~8-10 times larger than that in land scenes, especially on sea-oriented measurements under static weather condition (slope  $\approx 0.06$ ). The correlation coefficients (R) in inland and static sea scenarios were all > 0.62, except in SUST (R = 0.58), which indicates the formation rate of secondary HONO from NO<sub>2</sub> heterogeneous reaction in static sea scenarios may be faster than that in land scenarios. The corresponding temperature and RH conditions of each spot are displayed in Figure S8, which roughly reveals the impact of RH and temperature on the process of NO<sub>2</sub> forming HONO through heterogeneous reactions.

#### 3.2.1 RH dependence on HONO formation

The scatter plots of HONO/NO<sub>2</sub> against RH in different land and sea conditions are illustrated in Figure 6. The highest values can represent varying range of data in each interval and reveal concentration levels of data distribution. To eliminate the influence of other factors, the average of the six highest HONO/NO<sub>2</sub> in each 10% RH interval is calculated to reflect the distribution range of data in each interval (Liu et al., 2019). The dependence of the averaged top-6 HONO/NO<sub>2</sub> on RH reveal an overall variation tendency of HONO/NO<sub>2</sub> against RH. In the inland (CAMS) and coastal (SUST) cases, the RH turning points are both ~65% (60–70%), where increasing trend switches to decreasing tendency. The HONO/NO<sub>2</sub> increases along with RH when RH is less than 65%, and the HONO/NO<sub>2</sub> will decrease when RH is larger than 65%, which implies that it contributes to the HONO formation from the heterogeneous reaction of NO<sub>2</sub> on wet surfaces with the gradual increase of RH until 65%. The decrease of HONO/NO<sub>2</sub> with RH larger than 65% is presumably due to the efficient uptake of HONO on wet surfaces, and the wet surfaces being less accessible or less reactive to NO<sub>2</sub> when RH being larger than 65% (Liu et al., 2019). However, two turning peaks of RH were found in the sea cases. The first RH turning peak occurred in ~60%, which is the similar to that under the inland and coastal cases, and another RH turning peak appeared in ~85% (80-90%). This implies that high RH also could increase the HONO formation in sea cases, Additionally, the HONO/NO2 decreased sharply when RH was larger than 95%, because the reaction surface will asymptotically approach a water droplet state to limit the formation of HONO with RH larger than 95%.

#### 3.2.2 Temperature dependence on HONO formation

The scatter plots of HONO/NO<sub>2</sub> against temperature in different land and sea conditions are shown in Figure 7. Similar to the scatter plots of HONO/NO<sub>2</sub> against RH, we also adopted the averaged top-6 HONO/NO<sub>2</sub> values in each 5°C interval to represent a general variation tendency of HONO/NO<sub>2</sub> against temperature. In the inland condition (CAMS), the HONO/NO<sub>2</sub> decreased along with the increase of temperature, and the highest values of HONO/NO<sub>2</sub> appeared at ~12.5°C. However, we found that HONO/NO<sub>2</sub> increased along with the increase in temperature, and the highest values of HONO/NO<sub>2</sub> appeared at ~31.5°C in coastal condition (SUST), which indicates that the HONO formation from NO<sub>2</sub> heterogeneous reaction will be accelerated under lower and higher temperature in the inland and coastal conditions, respectively. In the sea condition, the HONO/NO<sub>2</sub> increased along with the increase in temperature with a high value under ~25.0°C when the atmospheric temperature was larger than 18.0°C, and simultaneously, a ~1.9 averaged HONO/NO<sub>2</sub> high value was found under ~15.0°C (14.0–17.0°C). Furthermore, we found that the appearance of HONO/NO<sub>2</sub> high values under lower temperature (14.0–17.0°C) was usually accompanied by land

268

269

270

271

272

273

274

275

276

277

278

279

280

281

282

283

284

285

286

287

288

289

290

291

292

293

294

295

296

297

298

299

300

301

302

303

304

305

306

307

308

309

310

311

312

313

314

315

316

317

318

breeze. Wen et al. (2019) also reported that relatively high temperature could contribute to the formation of HONO in the sea condition.

## 3.2.3 Impact of aerosol on HONO formation

To further understand the HONO formation from NO<sub>2</sub> heterogeneous reaction on aerosol surface, several correlation analyses were conducted. As shown in Figure 8, the linear regression plot between HONO and aerosol in land and sea conditions was performed. It was found that the correlation coefficient (R) between HONO and aerosol varied in the order of coastal (0.55) > sea (0.51) > inland (0.14). Additionally, the fitting slopes under coastal and sea conditions (0.07) are about 2.3 times larger than that under inland condition (0.03), which implies that the ground surface maybe more important than aerosol surface during the process of HONO formed from NO<sub>2</sub> heterogeneous reaction in the ground surface layer of the inland. In the coastal and sea conditions, the aerosol and sea are both important in providing heterogeneous reaction surface for NO<sub>2</sub> to form HONO (Cui et al., 2019; Wen et al., 2019; Yang et al., 2021). Additionally, we found the averaged values of HONO/NO<sub>2</sub> were 0.011±0.004, 0.014±0.006, 0.008±0.003, and 0.007±0.003 when aerosol extinctions are 0–0.3, 0.3–0.6, 0.6–0.9 and > 0.9 km<sup>-1</sup> in the inland case, respectively (Figure 8(b)). As shown in Figure 8, the high values of HONO/NO<sub>2</sub> were mainly under aerosol extinction being less than 1.0 km<sup>-1</sup> with averaged values of 0.012±0.006 and 0.090±0.004 in the coastal and sea cases, respectively. It indicates that aerosol surface plays a more important role in forming HONO through NO<sub>2</sub> heterogeneous reaction in the sea condition than that in the land condition.

#### 3.3 Vertical distributions of HONO/NO2 under different aerosol condition in land and sea

To further investigate the height dependence of HONO/NO<sub>2</sub> under land and sea conditions, two cases in Pearl River Delta (PRD) were selected from the whole campaign. As shown in Figure 9, "A" and "B" were under similar aerosol level (the extinction coefficients in surface layer being 0.45–0.60 km<sup>-1</sup>) and vertical distribution structure, and were all observed from 10:00 to 11:00 hrs. The instrument viewed sea accompanied with sea wind in "A" named sea scene, and the instrument viewed land accompanied with land wind in "B" named land scene. The NO<sub>2</sub> in the sea and land scenes have a similar vertical structure, and the NO<sub>2</sub> concentration in land scene are larger than that in sea scene except on the surface layer. The HONO have the same vertical distribution structure in the above two scenes, and the HONO concentration in the land scene is always larger than that in the sea scene. In Figure 9(e), we found that HONO/NO<sub>2</sub> under 0–400 m in the land scene is higher than that in the sea scene, however, the HONO/NO<sub>2</sub> values are obviously lower in the land scene than that in the sea scene above 400 m. Furthermore, the growth rate of HONO/NO<sub>2</sub> with the increase of height in the sea scene is significantly faster than that in the land scene above 400 m. This indicates the generation rates of HONO sourced from NO<sub>2</sub> heterogeneous reaction on aerosol surface in the sea scene is larger than that in the land scene above 400 m. Under 400 m, the HONO generation rates in the land scene is larger than that in the sea scene.

Additionally, we selected inland cases (CAMS) to learn the difference of height dependence of HONO/NO2 compared with sea scenes under different aerosol loads. As shown in Figure 10, the sea and inland scenes had similar aerosol levels (low aerosol level: < 0.2 km<sup>-1</sup>) and vertical structure. Furthermore, the NO<sub>2</sub> and HONO in the sea and inland scenes have similar vertical structure, although their concentrations in the sea scene are all larger than that in the inland scene. In Figure 10(d), we found that the HONO/NO2 in the sea scene was obviously larger than that in the inland scene above 400 m. The HONO/NO2 in the sea scene was about 4.5 times larger than that in the inland scene especially above 600 m. As shown in Figure 11, the aerosols under the sea and inland scenes exhibited similar extinction levels (relatively high level: ~0.8 km<sup>-1</sup>) and vertical structure. The NO<sub>2</sub> concentration in the sea scene was higher than that in the inland scene but with a similar vertical structure. The HONO concentration in the sea scene was lower than that in the inland scene under 400 m, while the concentration in the sea scene was larger than that in the inland scene above 400 m. In Figure 11 (d), we found the HONO/NO<sub>2</sub> in the inland scene was larger than that in the sea scene under 600 m, while the HONO/NO2 in the sea scene was about 2 times larger than that in the inland scene above 600 m. All the above cases indicated that the HONO generation rate from NO<sub>2</sub> heterogeneous reaction in the sea scene was larger than that in the inland scene in higher atmospheric layers above 400-600 m. The high-altitude (> 400-600 m) atmospheric parameters in the sea scene were more conductive to promote the HONO formation through the heterogeneous reaction of NO<sub>2</sub>. As shown in Figure S9, the ratio of HONO/NO<sub>2</sub> also generally increased with the increase in height above 0.2 km during the whole ship-based campaign. The greatest sensitivity under 1.5 km and the high degree of freedom (DOF) for aerosol, NO<sub>2</sub>, and HONO gave confidence in the retrieval results (Figure S10).

# 3.4 Case study

- The important factors and precursors to drive the formation of HONO through heterogeneous reaction had complex
- evolution and transport characteristics. To further clarify the role of these parameters in the heterogeneous process of
- NO<sub>2</sub> to form HONO, three typical processes were selected to reveal the favorable conditions for HONO formation at
- 374 the sea scene.

## 3.4.1 20 April: A typical transport event

As shown in Figure 12, the aerosol mainly distributed in 0–200 m with a mean extinction coefficient larger than 0.74 km<sup>-1</sup>. NO<sub>2</sub> was mainly distributed near the ground surface with a mean concentration of 28.54 ppb before 13:20. The NO<sub>2</sub> during this period may come from local ship emissions, as this area is a main shipping channel. From 14:25 to 17:10, a high-concentration NO<sub>2</sub> air mass (averaged 13.29 ppb) was found at ~2.0 km. To understand the source of this high-altitude NO<sub>2</sub> air mass, we further investigated the possible influence of transport by using the backward trajectories. We calculated 24 h backward trajectories of air masses at 500, 1000, and 2000 m using HYSPLIT (Figure S11). In Figure S11, we found that the dominant wind direction during this period was southeast at all heights, i.e., 500, 1000, and 2000 m. The transport of air masses carried NO<sub>2</sub> emitted by ships in Ningbo and Zhoushan ports to main cargo ports of China and Shanghai. Furthermore, the concentration of NO<sub>2</sub> was low (averaged 2.32 ppb) near the ground surface from 14:25 to 17:10. As shown in Figure 12 (e) and (g), a low pressure (< 1020 hPa), north dominant wind direction with the wind speed > 12 m/s appeared at the ground surface during this period, which implies that the clean air from north reduced the local surface NO<sub>2</sub>. The HONO was mainly distributed near the surface with a mean concentration of 0.07 ppb, and the two peaks were found in the early morning (averaged 0.15 ppb) and at 12:15 (averaged 0.11 ppb), respectively (Figure S12). The relatively high concentration of HONO appearing in the early morning was possibly attributed to the accumulation with the stabilization of boundary layer and attenuation of solar radiation after sunset the day before (Xing et al., 2021). The HONO peak appearing at 12:15 may be sourced from the heterogeneous reaction of NO<sub>2</sub> on the aerosol surface under a ~80% RH, 18.5°C temperature, and 1×10<sup>3</sup> W/m<sup>2</sup> SRI conditions.

# 3.4.2 28 April: A typical event of HONO produced from NO<sub>2</sub> heterogeneous reaction

From a typical port observation case, the measurement ship moored at Xiamen port on 28 April. As shown in Figure 13, we found two peaks for aerosol and NO<sub>2</sub> from 09:00–11:00 and 14:00–16:00, respectively (averaged aerosol extinction coefficient > 0.8 km<sup>-1</sup>, averaged NO<sub>2</sub> concentration > 12.0 ppb). NO<sub>2</sub> was mainly distributed near the sea surface layer 0–200 m, and a high-concentration NO<sub>2</sub> air mass was found from 1.0–2.0 km during 13:00–14:00 due to the short distance transport of NO<sub>2</sub> emitted from ships in Xiamen port (Figure S13). However, aerosol appeared in the range of 0.0–2.0 km during 09:00–11:00 and 14:00–16:00. In Figure 13 (g), we found that the wind speeds in the above two peak periods were obviously higher than that in other periods. From 09:00–11:00, the wind speed was ~5.0 m/s with a northwest dominant direction (urban), and the wind speed was ~6.0 m/s with a southeast dominant direction (port gateway) during 14:00–16:00, which indicates that the short-distance high-altitude transport caused the appearance of high-extinction aerosol mass during the above two periods.

Furthermore, we found the high-concentration HONO only appeared at 14:00-16:00 with a 0.57 ppb averaged concentration under 0.9 km, while it was only about 0.14 ppb during 09:00-11:00 period. The slight increase of RH and temperature (Tem) at 14:00-16:00 (RH: ~75.0%, Tem:  $23.7^{\circ}$ C) may contribute to HONO formation through heterogeneous reaction of  $NO_2$  on the aerosol surface than that at 09:00-11:00 (Figure 13 (d)-(e), Section 3.2). Contrarily, the solar radiation intensity (SRI) (~600 W/m²) at 09:00-11:00 was obviously larger than that (~250 W/m²) at 14:00-16:00 (Figure 13 (f)). The higher SRI accelerated the photolysis of HONO during 09:00-11:00 period (Kraus et al., 1998). Therefore, the lower formation rate and higher photolysis rate lead to a significantly lower HONO concentration at 09:00-11:00 than that at 14:00-16:00.

#### 3.4.3 03 May: A typical event with unknown HONO source

The measurement ship conducted observation in the sea area near Zhanjiang on 03 May, 2018. As shown in Figure 14, we found that there was an obvious sinking process for aerosol from ~1.0 km during 09:00-16:00, and eventually accumulated near the sea surface with a high extinction coefficient > 0.92 km<sup>-1</sup>. The NO<sub>2</sub> was mainly concentrated near the sea surface layer (0–400 m) with an averaged concentration of 8.93 ppb from 08:00 to 09:00. Thereafter, with the rise the planetary boundary layer (PBL) height after sunrise, NO<sub>2</sub> was gradually mixed and spread throughout the PBL from 09:00–13:00. During this period, it was accompanied by the increase of the NO<sub>2</sub> concentration (averaged 11.2 ppb) under PBL (Figure S14). It is due to the contribution of ship emissions near the sea surface. Contrarily, the regional transport of NO<sub>2</sub> from land also increased the NO<sub>2</sub> concentration in this area of the sea, with wind speed increasing from 2.5 to 7.8 m/s with a north wind direction from 10:00 to 16:00 (Figure 14 (g)).

Several HONO peaks (> 0.2 ppb) at 0.5–1.0 km were found from 09:45 to 13:00, and the aerosol and NO<sub>2</sub> high values were also observed at this height layer, simultaneously, which implies that the heterogeneous reaction of NO<sub>2</sub> on aerosol surface is more important than that on the sea surface for HONO source under sea atmosphere. Additionally, HONO concentration obviously elevated after 14:00, especially during 14:00–16:00 (> 0.4 ppb). It may be sourced from heterogeneous reaction of NO<sub>2</sub> on the aerosol surface, under RH a being ~92.5% (Figure 14 (d)). The photolysis of HONO also decreased with SRI < 150 W/m² (Figure 14 (f)) during this period. Furthermore, a HONO peak (> 0.32 ppb) was observed during 16:40–17:10. However, the NO<sub>2</sub> concentration always kept low (< 1.5 ppb) after 16:00, and the temperature was lower than 17 °C (Figure 14 (e)), which indicates the heterogeneous reaction of NO<sub>2</sub> not being the source of the observed HONO peak. The wind was north dominant with an average speed at 7.8 m/s after 15:00,

433 which implies that the regional transport may not be the source of the observed high-concentration of HONO.

434 Furthermore, the SRI was lower than 87.5 W/m<sup>2</sup>, and it shows the photolysis of nitrate aerosol also not being the 435

source of the elevated HONO. The unknown HONO source in this area of the sea need to be further explored.

## **4 Summary and Conclusions**

436

437

438

439

440

441

442

443

444

445

446

447

448

449

450

451

452

453

454

455

456

457

458

459

460

461

462

463

464

465

466

467

468

469

470

471 472

473

474

475

476

477

478

479

480 481

482

483 484 Currently, many uncertainties in the study of the HONO forming mechanism through the heterogeneous reaction of NO<sub>2</sub> exist. Earlier studies mostly focused on the near-surface layer, and the assessment of the contribution of NO<sub>2</sub> heterogeneous reaction to HONO formation in the vertical direction of the boundary layer is insufficient. Therefore, we aim to learn the sea-land and vertical differences of the HONO forming mechanism from NO<sub>2</sub> heterogeneous reaction and provide deep insights into the distribution characteristics, transforming process, and environmental effects of tropospheric HONO. Ship based MAX-DOAS observations along the marginal seas of China were performed from 19 April to 16 May 2018. Simultaneously, two ground-based MAX-DOAS observations were conducted in the inland station CAMS and the coastal station SUST to measure the aerosol, NO2, and HONO vertical profiles.

Along the cruise route, we found five hot spots with enhanced tropospheric NO<sub>2</sub> VCDs in Yangtze River Delta, Taiwan straits, Guangzhou-Hong Kong-Macao Greater Bay areas, Zhanjiang Port, and Qingdao port. Under high-level NO<sub>2</sub> conditions in the above five hot spots, we also observed enhanced HONO levels. Contrastingly, the low-concentration HONO accompanied high-level NO2 in the southeast coastline of Jiangsu province. When peak AOD and NO<sub>2</sub> conditions were observed, enhanced HONO were observed, although the reverse was not always the case.

To understand the impacts of RH, temperature, and aerosol on the heterogeneous reaction of NO<sub>2</sub> to produce HONO, the emission ratios of  $\Delta HONO/\Delta NOx$  were calculated to quantify the contribution of the primary HONO source to the total production of HONO. We found that the RH turning points in CAMS and SUST cases were both ~65% (60–70%), whereas two turning peaks (~60% and ~85%) of RH were found in the sea cases. This implied that high RH could contribute to the secondary formation of HONO in sea atmosphere. With increase in temperature, the HONO/NO<sub>2</sub> decreased with peak values appearing at ~12.5°C in CAMS, whereas the HONO/NO<sub>2</sub> gradually increased and reached peak values at ~31.5°C in SUST. In the sea case, when the temperature exceeded 18.0°C, the HONO/NO<sub>2</sub> increased with the increasing temperature and achieved peak at ~25.0℃. This indicated that high temperature could promote the secondary formation of HONO in the sea and coastal atmosphere. Additionally, the correlation analysis under different sea-land conditions indicated that the ground surface is more crucial to the formation of HONO from NO<sub>2</sub> heterogeneous reaction in the inland case, whereas the aerosol surface contributed more in the coastal and sea cases.

Furthermore, we found that the HONO/NO<sub>2</sub> in the sea case was about 4.5 times larger than that in the inland case above 600 m when AEC was ~0.2 km<sup>-1</sup>, and the HONO/NO<sub>2</sub> ratio in the sea case was about 2 times larger than that in the inland case above 600 m when AEC was ~0.8 km<sup>-1</sup>, which implied that the generation rate of HONO from NO<sub>2</sub> heterogeneous reaction in the sea case is larger than that in the inland case in higher atmospheric layers (> 600 m). To have a deep understanding of three potential contributing factors of HONO production under marine condition, we selected three typical events, which represented the impacts of transport, NO<sub>2</sub> heterogeneous reaction, and unknown HONO source, respectively.

# Acknowledgements

This research is supported by the National Natural Science Foundation of China (42207113), the Anhui Provincial Natural Science Foundation (2108085QD180), the Presidential Foundation of the Hefei Institutes of Physical Science, Chinese Academy Sciences (YZJJ2021QN06), the National Natural Science Foundation of China (41941011, 51778596, 41575021 and 41977184), the Strategic Priority Research Program of the Chinese Academy of Sciences (XDA23020301), the National High-Resolution Earth Observation Project of China (No. 05-Y30B01-9001-19/20-3). We would like to also thank Fudan University (Professor Jianmin Chen's group) for organizing the ship-based campaign and providing meteorological data.

# **Compliance with ethics guidelines**

All authors (the name of author) declare that they have no conflict of interest or financial conflicts to disclose.

#### References

- 485 [1] Alicke B, Geyer A, Hofzumahaus A, Holland F, Konrad S, Patz H W, Schafer J, Stutz J, Volz-Thomas A, Platt U.
- 486 OH formation by HONO photolysis during the BERLIOZ experiment, J. Geophys. Res.-Atmos., 2003, 108, 8247.
- 487 [2] Kleffmann J, Gavriloaiei T, Hofzumahaus A, Holland F, Koppmann R, Rupp L, Schlosser E, Siese M, Wahner A.
- 488 Daytime formation of nitrous acid: A major source of OH radicals in a forest, Geophys. Res. Lett., 2005, 32, L05818.

- 489 [3] Michoud V, Kukui A, Camredon M, Colomb A, Borbon A, Miet K, Aumont B, Beekmann M, Durand-Jolibois R,
- 490 Perrier S, Zapf P, Siour G, Ait-Helal W, Locoge N, Sauvage S, Afif C, Gros V, Furger M, Ancellet G, Doussin J F.
- 491 Radical budget analysis in a suburban European site during the MEGAPOLI summer field campaign. Atmos. Chem.
- 492 Phys., 2012, 12, 11951–11974.
- 493 [4] Ryan RG, Rhodes S, Tully M, Wilson S, Jones N, Frieß U, Schofield R. Daytime HONO, NO2 and aerosol 494 distributions from MAX-DOAS observations in Melbourne. Atmos. Chem. Phys., 2018, 18, 13969–13985.
- 495 [5] Xue C, Zhang C, Ye C, Liu P, Catoire V, Krysztofiak G, Chen H, Ren Y, Zhao X, Wang J, Zhang F, Zhang C,
- 496 Zhang J, An J, Wang T, Chen J, Kleffmann J, Mellouki A, Mu Y. HONO Budget and Its Role in Nitrate Formation in 497 the Rural North China Plain. Environ. Sci. Technol., 2020, 54, 11048-11057.
- 498 [6] Huang R, Zhang Y, Bozzetti C, Ho K F, Cao J J, Han Y, Daellenbach K R, Slowik J G, Platt S M, Canonaco F,
- 499 Zotter P, Wolf R, Pieber S M, Bruns E A, Crippa M, Ciarelli G, Piazzalunga A, Schwikowski M, Abbaszade G,
- 500 Schnelle-Kreis J, Zimmermann R, An Z, Szidat S, Baltensperger U, Haddad I E, Prevot A S H. High secondary 501 aerosol contribution to particulate pollution during haze events in China. Nature, 2014, 514, 218-222.
- 502 [7] Zhang B, Wang Y, Hao J. Simulating aerosol-radiation-cloud feedbacks on meteorology and air quality over 503 eastern China under severe haze conditions in winter. Atmos. Chem. Phys., 2015, 15, 2387-2404.
- 504 [8] Wang L, Wen L, Xu C, Chen J, Wang X, Yang L, Wang W, Yang X, Sui X, Yao L, Zhang Q. HONO and its 505 potential source particulate nitrite at an urban site in North China during the cold season. Sci. Total Environ., 2015, 506 538, 93-101.
- 507 [9] Fu X, Wang T, Zhang L, Li O, Wang Z, Xia M, Yun H, Wang W, Yu C, Yue D, Zhou Y, Zheng J, Han R. The 508 significant contribution of HONO to secondary pollutants during a severe winter pollution event in southern China. 509 Atmos. Chem. Phys., 2019, 19, 1-14.
- 510 [10] Liu J, Liu Z, Ma Z, Yang S, Yao D, Zhao S, Hu B, Tang G, Sun J, Cheng M, Xu Z, Wang Y. Detailed budget analysis of HONO in Beijing, China: Implication on atmosphere oxidation capacity in polluted megacity. Atmos. 511
- 512 Environ., 2021, 244, 117957.
- 513 [11] Yang Y, Li X, Zu K, Lian C, Chen S, Dong H, Feng M, Liu H, Liu J, Lu K, Li S, Ma X, Song D, Wang W, Yang
- 514 S, Yang X, Yu X, Zhu Y, Zeng L, Tan Q, Zhang Y. Elucidating the effect of HONO on O3 pollution by a case study 515 in southwest China. Sci. Total Environ., 2021, 756, 144127.
- 516 [12] Zha Q, Xue L, Wang T, Xu Z, Yeung C, Louie P K K, Luk C W Y. Large conversion rates of NO2 to HNO2
- 517 observed in air masses from the South China Sea: Evidence of strong production at sea surface? Geophys. Res. Lett., 518 2014, 41, 7710-7715.
- 519 [13] Cui L, Li R, Fu H, Li Q, Zhang L, George C, Chen J. Formation features of nitrous acid in the offshore area of 520 the East China Sea. Sci. Total Environ., 2019, 682, 138–150.
- 521 [14] Wen L, Chen T, Zheng P, Wu L, Wang X, Mellouki A, Wang W. Nitrous acid in marine boundary layer over 522 eastern Bohai Sea, China: Characteristics, sources, and implications. Sci. Total Environ., 2019, 670, 282–291.
- 523 [15] Gil G, Kim J, Lee M, Lee G, Lee D, Jung J, An J, Hong J, Cho S, Lee J, Long R. The role of HONO in O3 524 formation and insight into its formation mechanism during the KORUS-AO Campaign, Atmos. Chem. Phys. Discus...
- 525 2019, https://doi.org/10.5194/acp-2019-1012.
- 526 [16] Cui L, Li R, Zhang Y, Meng Y, Fu H, Chen J. An observational study of nitrous acid (HONO) in Shanghai,
- 527 China: The aerosol impact on HONO formation during the haze episodes. Sci. Total Environ., 2018, 630, 1057–1070.
- 528 [17] Tan W, Liu C, Wang S, Xing C, Su W, Zhang C, Xia C, Liu H, Cai Z, Liu J. Tropospheric NO2, SO2, and 529 HCHO over the East China Sea, using ship-based MAX-DOAS observations and comparison with OMI and OMPS
- 530 satellite data. Atmos. Chem. Phys., 2018, 18, 15387-15402.
- 531 [18] Zhang W, Tong S, Jia C, Wang L, Liu B, Tang G, Ji D, Hu B, Liu Z, Li W, Wang Z, Liu Y, Wang Y, Ge M. Different HONO Sources for Three Layers at the Urban Area of Beijing. Environ. Sci. Technol., 2020, 54,
- 532 533 12870-12880.
- 534 [19] Cheng P, Cheng Y, Lu K, Su H, Yang Q, Zou Y, Zhao Y, Dong H, Zeng L, Zhang Y. An online monitoring 535 system for atmospheric nitrous acid (HONO) based on stripping coil and ion chromatography. J. Environ. Sci., 2013,
- 536
- 537 [20] Bernard F, Cazaunau M, Grosselin B, Zhou B, Zheng J, Liang P, Zhang Y, Ye X, Daele V, Mu Y, Zhang R,
- 538 Chen J, Mellouki A. Measurements of nitrous acid (HONO) in urban area of Shanghai, China. Environ. Sci. Pollut.
- 539 Res., 2016, 23:5818-5829.
- 540 [21] Jordan N, Osthoff HD. Quantification of nitrous acid (HONO) and nitrogen dioxide (NO2) in ambient air by
- 541 broadband cavity-enhanced absorption spectroscopy (IBBCEAS) between 361 and 388 nm. Atmos. Meas. Tech., 542 2020, 13, 273–285.
- 543
- [22] Guo Y, Wang S, Gao S, Zhang R, Zhu J, Zhou B. Influence of ship direct emission on HONO sources in channel
- 544 environment. Atmos. Environ., 2020, 242, 117819.
- 545 [23] Meng F, Qin M, Tang K, Duan J., Fang W., Liang S., Ye K., Xie P., Sun Y., Xie C., Ye C., Fu P., Liu J., Liu W.
- 546 High-resolution vertical distribution and sources of HONO and NO2 in the nocturnal boundary layer in urban Beijing,
- 547 China. Atmos. Chem. Phys., 2020, 20, 5071–5092.
- 548 [24] Garcia-Nieto D, Benavent N, Saiz-Lopez A. Measurements of atmospheric HONO vertical distribution and
- 549 temporal evolution in Madrid (Spain) using the MAX-DOAS technique. Sci. Total Environ., 2018, 643, 957–966.

- 550 [25] Wang Y, Dörner S, Donner S, Bohnke S, Smedt I D, Dickerson R R, Dong Z, He H, Li Z, Li Z, Li D, Liu D, Ren
- X, Theys N, Wang Y, Wang Z, Xu H, Xu J, Wagner T. Vertical profiles of NO2, SO2, HONO, HCHO,
- 552 CHOCHO and aerosols derived from MAX-DOAS measurements at a rural site in the central western North China
- 553 Plain and their relation to emission sources and effects of regional transport. Atmos. Chem. Phys., 2019, 19, 554 5417–5449.
- 555 [26] Xing C, Liu C, Wang S, Chan K L, Gao Y, Huang X, Su W, Zhang C, Dong Y, Fan G, Zhang T, Chen Z, Hu Q,
- Su H, Xie Z, Liu J. Observations of the vertical distributions of summertime atmospheric pollutants and the corresponding ozone production in Shanghai, China. Atmos. Chem. Phys. 2017, 17, 14275–14289.
- 558 [27] Xing C, Li C, Wang S, Hu Q, Liu H, Tan W, Zhang W, Li B, Liu J. A new method to determine the aerosol
- optical properties from multiple-wavelength O4 absorptions by MAX-DOAS observation. Atmos. Meas. Tech., 2019, 12, 3289–3302.
- [28] Xing C, Liu C, Hu Q, Fu Q, Lin H, Wang S, Su W, Wang W, Javed Z, Liu J. Identifying the wintertime sources
- of volatile organic compounds (VOCs) from MAX-DOAS measured formaldehyde and glyoxal in Chongqing, southwest China. Sci. Total Environ., 2020, 715, 136258.
- 564 [29] Vandaele AC, Hermans C, Simon PC, Carleer M, Colin R, Fally S, Merienne M F, Jenouvrier A, Coquart B.
- Measurements of the NO2 absorption cross-section from 42,000 cm-1 to 10,000 cm-1 (238–1000 nm) at 220K and
- 566 294 K. J. Quant. Spectrosc. Ra., 1998, 59, 171–184.
- 567 [30] Serdyuchenko A, Gorshelev V, Weber M, Chehade W, Burrows J P. High spectral resolution ozone absorption cross-sections –part 2: temperature dependence. Atmos. Meas. Tech., 2014, 7, 625–636.
- 569 [31] Thalman R, Volkamer R. Temperature dependent absorption cross-sections of O2–O2 collision pairs between 340 and 630 nm and at atmospherically relevant pressure. Phys. Chem. Chem. Phys., 2013, 15, 15371–15381.
- 571 [32] Meller R, Moortgat GK. Temperature dependence of the absorption cross sections of formaldehyde between 223 and 323 K in the wavelength range 225–375 nm. J. Geophys. Res., 2000, 105, 7089–7101.
- 573 [33] Rothman L, Gordon I, Barber R, Dothe H, Gamache R R, Goldman A, Perevalov, Tashkun S A, Tennyson J.
- HITEMP, the high-temperature molecular spectroscopic database. J. Quant. Spectrosc. Ra., 2010, 111, 2139–2150.
- 575 [34] Fleischmann OC, Hartmann M, Burrows JP, Orphal J. New ultraviolet absorption cross-sections of BrO at atmospheric temperatures measured by time-windowing Fourier transform spectroscopy. J. Photoch. Photobio., 2004,
- 577 A 168, 117–132.
- 578 [35] Stutz J, Kim ES, Platt U, Bruno P, Perrino C, Febo A. UV–visible absorption cross sections of nitrous acid. J.
- 579 Geophys. Res.-Atmos., 2000, 105, 14585–14592.
- 580 [36] Aliwell S, Van Roozendael M, Johnston P, Richter A, Wagner T, Arlander D W, Burrows J P, Fish D J, Jones R
- L, Tornkvist K K, Lambert J C, Pfeilsticker K, Pundt I. Analysis for BrO in zenith-sky spectra: an Intercomparison exercise for analysis improvement. J. Geophys. Res., 2002, 107, 4199.
- 583 [37] Liu C, Xing C, Hu Q, Li Q, Liu H, Hong Q, Tan W, Ji X, Lin H, Lu C, Lin J, Liu H, Wei S, Chen J, Yang K,
- Wang S, Liu T, Chen Y. Ground-based hyperspectral stereoscopic remote sensing network: A promising strategy to learn coordinated control of O3 and PM2.5 over China. Engineering, 2021.
- 586 [38] Liu Y, Nie W, Xu Z, Wang T, Wang R, Li Y, Wang L, Chi X, Ding A. Semi-quantitative understanding of
- source contribution to nitrous acid (HONO) based on 1 year of continuous observation at the SORPES station in eastern China. Atmos. Chem. Phys., 2019, 19, 13289–13308.
- 589 [39] Sun L, Chen T, Jiang Y, Zhou Y, Sheng L, Lin J, Li J, Dong C, Wang C, Wang X, Zhang Q, Wang W, Xue L.
- Ship emission of nitrous acid (HONO) and its impacts on the marine atmospheric oxidation chemistry. Sci. Total Environ., 2020, 735, 139355.
- 592 [40] Huang R, Yang L, Cao J, Wang Q, Tie X, Ho K F, Shen Z, Zhang R, Li G, Zhu C, Zhang N, Dai W, Zhou J, Liu
- 593 S, Chen Y, Chen J, O'Dowd C D. Concentration and sources of atmospheric nitrous acid (HONO) at an urban site in
- 594 Western China. Sci. Total Environ., 2017, 593–594, 165–172.
- 595 [41] Xu Z, Wang T, Wu J, Xue L, Chan J, Zha Q, Zhou S, Louie P K K, Luk C W Y. Nitrous acid (HONO) in a polluted subtropical atmosphere: Seasonal variability, direct vehicle emissions and heterogeneous production at
- 597 ground surface, Atmos. Environ., 2015, 106, 100–109.
- 598 [42] Xing C, Liu C, Hu Q, Fu Q, Wang S, Lin H, Zhu Y, Wang S, Wang W, Javed Z, Ji X, Liu J. Vertical
- distributions of wintertime atmospheric nitrogenous compounds and the corresponding OH radicals production in
- 600 Leshan, southwest China. J. Environ. Sci., 2021, 105, 44–55.
- 601 [43] Song Y, Xing C, Liu C, Lin J, Wu H, Liu T, Lin H, Zhang C, Tan W, Ji X, Liu H, Li Q. Evaluation of transport
- processes over North China Plain and Yangtze River Delta using MAX-DOAS observations 10.5194/egusphere-2022-653.
- 604 [44] Kleffmann J, Kurtenbach R, Lorzer J, Wiesen P, Kalthoff N, Vogel B, Vogel H. Measured and simulated vertical profiles of nitrous acid-Part I: Field measurements, Atmos. Environ., 2003, 37, 2949-2955.
- [45] Stemmler K, Ammann M, Donders C, Kleffmann J, George C. Photosensitized reduction of nitrogen dioxide on humic acid as a source of nitrous acid, nature, 2006, 440, 195-198.
- [46] Indarto A. Heterogeneous reactions of HONO formation from NO2 and HNO3: a review, 2012, 38, 1029-1041.
- 609 [47] Wang Y, Apituley A, Bais A, Beirle S, Benavent N, Borovski A, Bruchkouski H, Chan K L, Donner S,
- Drosoglou T, Finkenzeller H, Friedrich M M, Frieß U, Garcia-Nieto D, Gómez-Mart ń, L, Hendrick F, Hilboll A, Jin

- 611 J, Johnston P, Koenig T K, Kreher K, Kumar V, Kyuberis A, Lampel J, Liu C, Liu H, Ma J, Polyansky O L,
- 612 Postylyakov O, Querel R, Saiz-Lopez A, Schmitt S, Tian X, Tirpitz J-L, Roozendael M V, Volkamer R, Wang Z, Xie
- 613 P, Xing C, Xu J, Yela M, Zhang C, Wagner T. Inter-comparison of MAX-DOAS measurements of tropospheric
- 614 HONO slant column densities and vertical profiles during the CINDI-2 campaign, Atmos. Meas. Tech., 2020, 13,
- 615
- 616 [48] Xu S, Wang S, Xia M, Lin H, Xing C, Ji X, Su W, Tan W, Liu C, Hu Q. Observations by ground-based
- 617 MAX-DOAS of the vertical characters of winter pollution and the influencing factors of HONO generation in
- 618 Shanghai, China, Remote Sens., 2021, 13(17), 3518.
- [49] He S, Wang S, Zhang S, Zhu J, Sun Z, Xue R, Zhou B. Vertical distributions of atmospheric HONO and the 619
- 620 corresponding OH radical production by photolysis at the suburb area of Shanghai, China, Sci. Total Environ., 2023,
- 621 858, 159703.
- 622 [50] Zhang N, Zhou X, Shepson P B, Gao H, Alaghmand M, Stirm B. Aircraft measurement of HONO vertical
- 623 profiles over a forested region, Geophys. Res. Lett., 2009, 36(15), L15820.
- 624 [51] Wong K W, Tsai C, Lefer B, Haman C, Grossberg N, Brune W H, Ren X, Luke W, Stutz J. Daytime HONO 625 vertical gradients during SHARP 2009 in Houston, TX, Atmos. Chem. Phys., 2012, 12, 635-652.
- 626 [52] Salgado, M. S. and Rossi, M. J.: Flame soot generated under controlled combustion conditions: Heterogeneous 627 reaction of NO2 on hexane soot, International Journal of Chemical Kinetics, 34, 620-631, 10.1002/kin.10091, 2002.
- 628 [53] Zhou, X., Zhang, N., TerAvest, M., Tang, D., Hou, J., Bertman, S., Alaghmand, M., Shepson, P. B., Carroll, M.
- 629 A., Griffith, S., Dusanter, S., and Stevens, P. S.: Nitric acid photolysis on forest canopy surface as a source for 630 tropospheric nitrous acid, Nature Geoscience, 4, 440-443, 10.1038/ngeo1164, 2011.
- 631 [54] Kraus, A., Hofzumahaus, A. Field Measurements of Atmospheric Photolysis Frequencies for O3, NO2, HCHO,
- 632 CH3CHO, H2O2, and HONO by UV Spectroradiometry. In: Rudolph, J., Koppmann, R. (eds) Atmospheric
- Measurements during POPCORN Characterisation of the Photochemistry over a Rural Area. Springer, Dordrecht. 633
- 634 https://doi.org/10.1007/978-94-017-0813-5 8.
- 635 [55] Liu, Y., Nie, W., Xu, Z., Wang, T., Wang, R., Li, Y., Wang, L., Chi, X., and Ding, A.: Contributions of different
- 636 sources to nitrous acid (HONO) at the SORPES station in eastern China: results from one-year continuous
- 637 observation., Atmospheric Chemistry and Physics Discussions, 10.5194/acp-2019-219, 2019.
- 638 [56] Rodgers, C. D., Taylor, F. W. (Ed.): Inverse methods for atmospheric sounding, theory and practice, World
- 639 Scientific, 255 pp.2004.
- 640 [57] Stutz, J., Kim, E. S., Platt, U., Bruno, P., Perrino, C., and Febo, A.: UV-visible absorption cross sections of 641 nitrous acid, Journal of Geophysical Research: Atmospheres, 105, 14585-14592, 10.1029/2000jd900003, 2000.
- 642 [58] Wang, Y., Apituley, A., Bais, A., Beirle, S., Benavent, N., Borovski, A., Bruchkouski, I., Chan, K. L., Donner, S.,
- 643 Drosoglou, T., Finkenzeller, H., Friedrich, M. M., Frieß, U., Garcia-Nieto, D., Gómez-Mart n, L., Hendrick, F.,
- 644 Hilboll, A., Jin, J., Johnston, P., Koenig, T. K., Kreher, K., Kumar, V., Kyuberis, A., Lampel, J., Liu, C., Liu, H., Ma, 645 J., Polyansky, O. L., Postylyakov, O., Querel, R., Saiz-Lopez, A., Schmitt, S., Tian, X., Tirpitz, J.-L., Van Roozendael,
- 646 M., Volkamer, R., Wang, Z., Xie, P., Xing, C., Xu, J., Yela, M., Zhang, C., and Wagner, T.: Inter-comparison of
- 647 MAX-DOAS measurements of tropospheric HONO slant column densities and vertical profiles during the CINDI-2 648 campaign, Atmospheric Measurement Techniques, 13, 5087-5116, 10.5194/amt-13-5087-2020, 2020.
- [59] Wang, Y., Lampel, J., Xie, P., Beirle, S., Li, A., Wu, D., and Wagner, T.: Ground-based MAX-DOAS 649
- 650 observations of tropospheric aerosols, NO2, SO2 and HCHO in Wuxi, China, from 2011 to 2014, Atmospheric
- 651 Chemistry and Physics, 17, 2189-2215, 10.5194/acp-17-2189-2017, 2017.

Table 1. Detailed information of the measurement cruise								
Cruise NO.	Periods	Measurement cruise						
NO. 1	08:50 to 21:02 19 Apr.	Daishan port (30.24°N, 122.16°E) to Chongming (31.18°N, 121.82°E)						
NO. 2	05:40 to 17:45 20 Apr.	Sailing around Chongming island						
NO. 3	06:03 21 Apr. to 08:07 03 May	Chongming (31.18°N, 121.82°E) to Zhanjiang port (21.12°N, 110.67°E)						
NO. 4	08:07 03 May to 06:52 09 May	Zhanjiang port (21.12°N, 110.67°E) to Daishan port (30.24°N, 122.16°E)						
NO. 5	05:40 11 May to 05:55 14 May	Daishan port (30.24°N, 122.16°E) to Qingdao (35.89°N, 120.87°E)						
NO. 6	05:55 14 May to 10:00 16 May	Qingdao (35.89°N, 120.87°E) to Daishan port (30.24°N, 122.16°E)						

 $\underline{T}\underline{able~2.~Detailed~retrieval~settings~of~O_4,~NO_2,~and~HONO.}$ 

Parameter	Data source	Fitti	Fitting internals (nm)			
		O <sub>4</sub>	$NO_2$	HONO		
Wavelength range		338-370	338-370	335-373		
$NO_2$	298K, I <sub>0</sub> -corrected, Vandaele et al. (1998)	$\checkmark$	$\checkmark$	$\checkmark$		
$NO_2$	220K, I <sub>0</sub> -corrected, Vandaele et al. (1998)	√	$\checkmark$	$\checkmark$		
$O_3$	223K, I <sub>0</sub> -corrected, Serdyuchenko et al. (2014)	√	$\checkmark$	$\checkmark$		
$O_3$	243K, I <sub>0</sub> -corrected, Serdyuchenko et al. (2014)	√	√	$\checkmark$		
$O_4$	293K, Thalman and Volkamer (2013)	√	$\checkmark$	$\checkmark$		
НСНО	298K, Meller and Moortgat (2013)	√	√	√		
$H_2O$	HITEMP (Rothman et al. 2010)	×	$\times$	$\checkmark$		
BrO	223K, Fleischmann et al. (2004)	$\checkmark$	√	✓		
HONO	296K, Stutz et al. (2000)	×	$\times$	✓		
Ring	Calculated with QDOAS	√	√	$\checkmark$		
Polynomial degree		Order 5	Order 5	Order 5		
Intensity offset		Constant	Constant	Constant		

<sup>\*</sup> Solar  $I_0$  correction; Aliwell et al. (2002).

 **Table 3.** Error budget estimation (in %) of the retrieved near-surface (0–200 m) trace gas concentrations and AECs, and trace gas VCDs and AOD

and trace gas VCDs and AOD.											
			Error source								
			Smoothing and noise errors	Algorithm error	Cross section error	Related to the aerosol retrieval (only for trace gases)					
Cruise route	Near-surface	aerosol	14	4	4	-	15				
		NO <sub>2</sub>	16	3	3	15	22				
		HONO	20	20	5	15	32				
	VCD or AOD	AOD	5	8	4	-	10				
		NO <sub>2</sub>	17	11	3	10	23				
		HONO	22	20	5	10	32				
SUST	Near-surface	aerosol	13	4	4	-	14				
		NO <sub>2</sub>	14	3	3	14	20				
		HONO	18	20	5	14	31				
	VCD or AOD	AOD	5	8	4	-	10				
		NO <sub>2</sub>	16	11	3	10	22				
		HONO	20	20	5	10	30				
CAMS	Near-surface	aerosol	13	4	4	-	14				
		NO <sub>2</sub>	15	3	3	14	21				
		HONO	19	20	5	14	31				
	VCD or AOD	AOD	5	8	4	-	10				
		NO <sub>2</sub>	17	11	3	10	23				
		HONO	21	20	5	10	31				



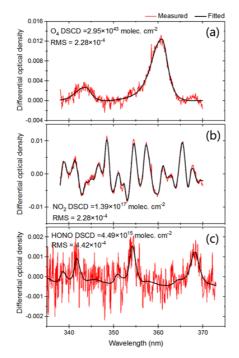


Figure 1. Plots depicting typical DOAS spectral fittings for (a) O<sub>4</sub>, (b) NO<sub>2</sub> and (c) HONO.

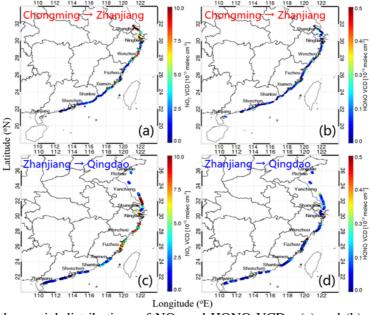


Figure 2. Maps showing the spatial distributions of  $NO_2$  and HONO VCDs. (a) and (b) show the  $NO_2$  and HONO VCDs along the cruise route from Chongming to Zhanjiang. (c) and (d) depict the  $NO_2$  and HONO VCDs along the cruise route from Zhanjiang to Qingdao.

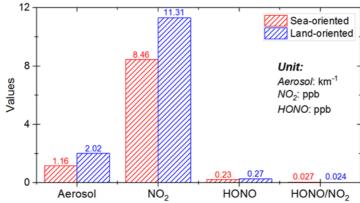


Figure 3. Bar plots of the averaged aerosol extinction, NO<sub>2</sub> concentration, HONO concentration, and HONO/NO<sub>2</sub> ratio during the campaign. The red and blue boxes denote sea-oriented and land-oriented measurements, respectively.

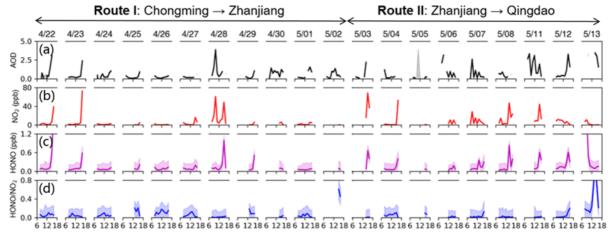


Figure 4. Histograms of the time series of (a) AOD, (b) surface NO<sub>2</sub> concentration, (c) surface HONO concentration, and (f) surface HONO/NO<sub>2</sub> ratios.

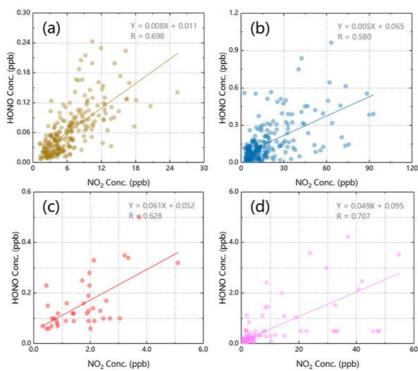
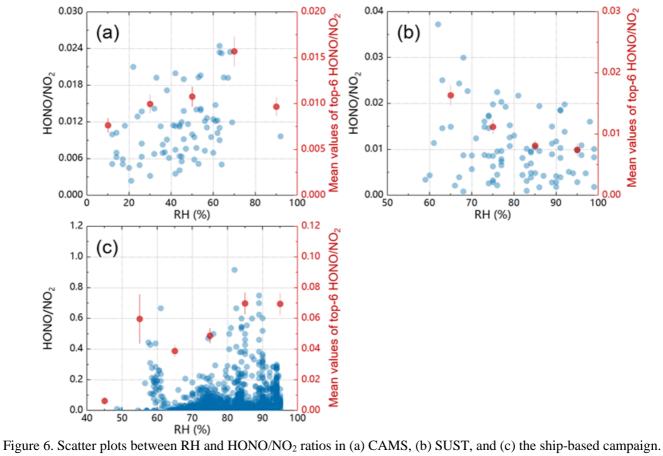


Figure 5. Linear regression plots between surface  $NO_2$  and HONO concentrations in (a) CAMS, (b) SUST, and ship-based measurements of (c) sea-oriented and (d) land-oriented under static weather condition.

689



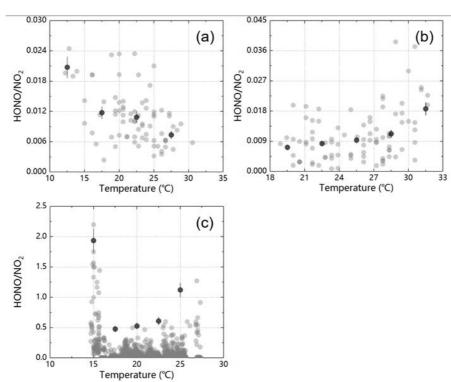


Figure 7. Scatter plots between temperature and HONO/NO2 ratios in (a) CAMS, (b) SUST, and (c) the ship-based campaign.

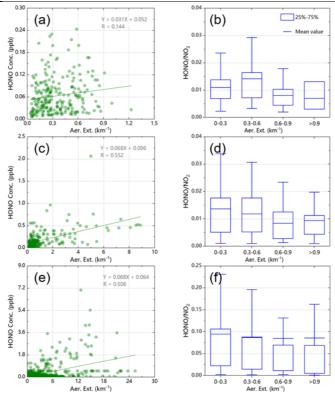


Figure 8. (a), (c), and (e) show the linear regression plots between surface aerosol extinction and HONO concentrations in CAMS, SUST and the ship-based campaign, respectively. Plots (b), (d), and (f) depicts the HONO/NO<sub>2</sub> ratio distribution under different aerosol extinction coefficient conditions in CAMS, SUST and the ship-based campaign.

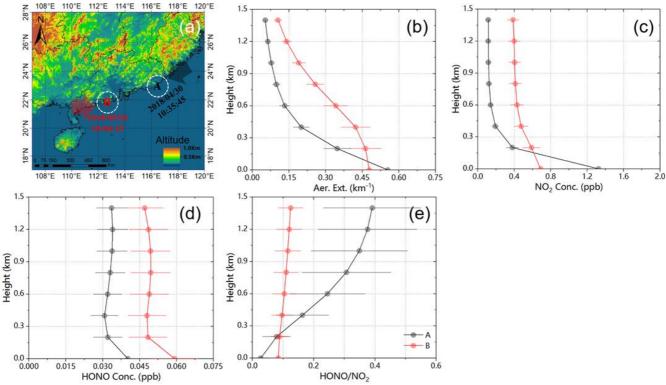


Figure 9. Map (a) shows the two measurement points (A: black, sea-oriented with sea wind; B: red, land-oriented with land wind) during the campaign. Plots (b)–(e) show the vertical profiles of aerosol, NO<sub>2</sub>, HONO, and HONO/NO<sub>2</sub> ratios in the above two measurement points, respectively.

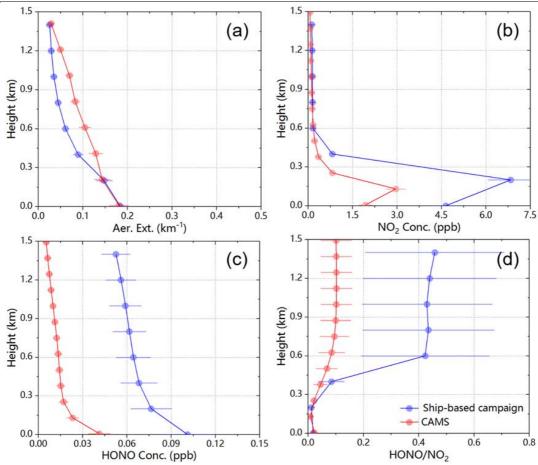
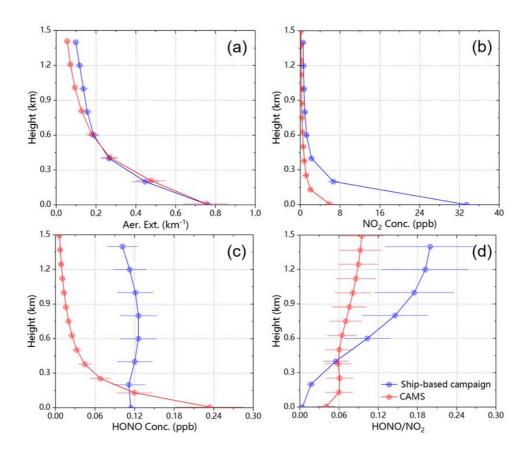


Figure 10. Plots showing the vertical distributions of (a) aerosol extinction, (b)  $NO_2$  concentration, (c) HONO concentration, and (d)  $HONO/NO_2$  ratio. The blue and red lines represent a ship-based campaign case and a CAMS case, respectively.



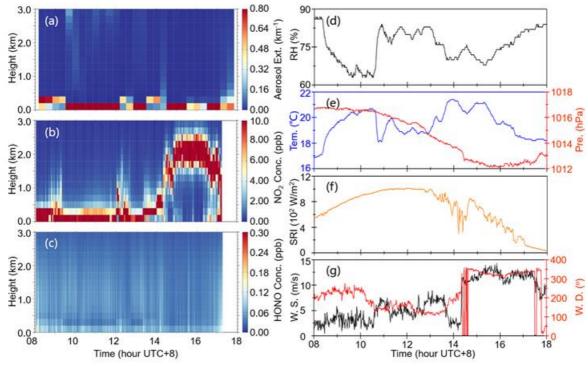
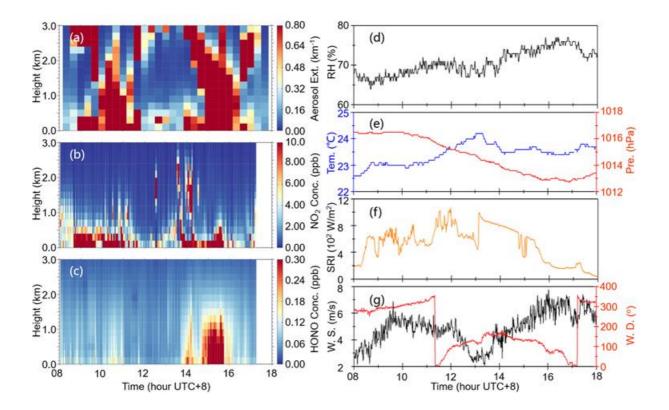


Figure 12. Case of 20 April 2018. Gradient image showing the time series of (a) aerosol extinction, (b) NO<sub>2</sub>, and (c) HONO vertical profiles. Plot (d) shows the time series of surface RH. Plot (e) depicts the time series of surface temperature and pressure. Plot (f) shows the time series of surface SRI. (g) depicted the time series of surface wind speed and wind direction.



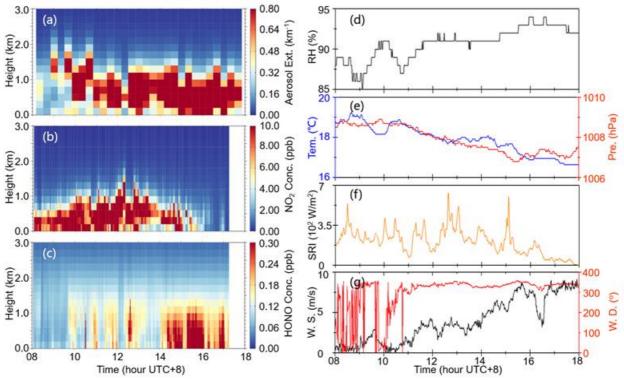


Figure 14. Case of 03 May 2018. Gradient image showing the time series of (a) aerosol extinction, (b) NO<sub>2</sub> and (c) HONO vertical profiles, respectively. Plot (d) shows the time series of surface RH. Plot (e) depicts the time series of surface temperature and pressure. Plot (f) shows the time series of surface SRI. Plot (g) depicts the time series of surface wind speed and wind direction.