Comparison of aqueous SOA product distributions from guaiacol oxidation by non-phenolic and phenolic methoxybenzaldehydes as photosensitizers in the absence and presence of ammonium nitrate

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12 Abstract. Aromatic carbonyls (e.g., methoxybenzaldehydes), an important class of photosensitizers, are abundant in the 13 atmosphere. Photosensitization and nitrate-mediated photo-oxidation can occur simultaneously, yet studies about their 14 interactions, particularly for aqueous secondary organic aerosol (aqSOA) formation, remain limited. This study compared 15 non-phenolic (3,4-dimethoxybenzaldehyde, DMB) and phenolic (vanillin, VL) methoxybenzaldehydes as photosensitizers 16 for aqSOA formation via guaiacol (GUA) oxidation in the absence and presence of ammonium nitrate (AN) under 17 atmospherically relevant cloud and fog conditions. GUA oxidation by triplet excited states of DMB (³DMB*) (GUA+DMB) was ~4 times faster and exhibited greater light absorption than oxidation by ³VL* (GUA+VL). Both GUA+DMB and 18 19 GUA+VL formed aqSOA composed of oligomers, functionalized monomers, oxygenated ring-opening species, and N-20 containing products in the presence of AN. The observation of N-heterocycles such as imidazoles indicates the participation 21 of ammonium in the reactions. The majority of generated aqSOA are potential brown carbon (BrC) chromophores. 22 Oligomerization and functionalization dominated in GUA+DMB and GUA+VL, but functionalization appeared to be more 23 important in GUA+VL due to contributions from VL itself. AN did not significantly affect the oxidation kinetics, but it had 24 distinct effects on the product distributions, likely due to differences in the photosensitizing abilities and structural features 25 of DMB and VL. In particular, the more extensive fragmentation in GUA+DMB than in GUA+VL likely generated more N-26 containing products in GUA+DMB+AN. In GUA+VL+AN, the increased oligomers may be due to VL-derived phenoxy 27 radicals induced by 'OH or 'NO₂ from nitrate photolysis. Furthermore, increased nitrated products observed in the presence 28 of both DMB or VL and AN than in AN alone implies that photosensitized reactions may promote nitration. This work 29 demonstrates how the structural features of photosensitizers affect aqSOA formation via non-carbonyl phenol oxidation. 30 Potential interactions between photosensitization and AN photolysis were also elucidated. These findings facilitate a better 31 understanding of photosensitized aqSOA formation and highlight the importance of AN photolysis in these reactions.

32 1 Introduction

Photosensitized reactions involving triplet excited states of organic compounds (³C*) are efficient pathways for the 33 34 formation of secondary organic aerosol in the aqueous phase (aqSOA; Smith et al., 2014, 2015, 2016; Yu et al., 2014, 2016; 35 Chen et al., 2018; Lu et al., 2019; Ye et al., 2019; Chen et al., 2020; Liu et al., 2020; Jiang et al., 2021; Ma et al., 2021; 36 Misovich et al., 2021; Ou et al., 2021; F. Li et al., 2022; X. Li et al., 2022; Aregahegn et al., 2022; Mabato et al., 2022; 37 Wang et al., 2022). Upon irradiation by solar radiation, photosensitizers form an excited triplet state that directly reacts with substrates (e.g., phenols), and can generate singlet oxygen ($^{1}O_{2}$), superoxide (O_{2}^{\cdot}) or hydroperoxyl ($^{\cdot}HO_{2}$) radicals, and 38 39 hydroxyl radicals (OH) upon reactions with O_2 and substrates (George et al., 2018; Chen et al., 2020), thereby facilitating 40 the oxidation of rather volatile species and contributing to aqSOA formation. An important class of photosensitizers is aromatic carbonyls (e.g., methoxybenzaldehydes) which are abundant in aerosol particles, cloud waters, and fog waters 41 42 (Anastasio et al., 1997; Felber et al., 2021). Aromatic carbonyls can be emitted from anthropogenic sources and biomass 43 burning (BB; Lipari et al., 1984; Edye and Richards, 1991; Hawthorne et al., 1992; Simoneit et al., 1993, 1999; Anastasio et 44 al., 1997; Felber et al., 2021), or formed via atmospheric oxidation of aromatic hydrocarbons (Hoshino et al., 1978; Calvert 45 and Madronich, 1987; Anastasio et al., 1997; Felber et al., 2021). BB is also a significant source of phenols through lignin pyrolysis (Simpson et al., 2005). Phenolic carbonyls have a hydroxyl (-OH) group on the aromatic ring, whereas non-46 47 phenolic carbonyls do not. BB smoke has been reported to have comparable concentrations of phenolic and non-phenolic carbonyls (Simoneit et al., 1993; Anastasio et al., 1997). 48

49 Most previous studies on agSOA formation via photosensitized non-carbonyl phenol oxidation have examined 3.4-50 dimethoxybenzaldehyde (DMB), a non-phenolic methoxybenzaldehyde, as the photosensitizer (Smith et al., 2014, 2015; Yu 51 et al., 2014, 2016; Ye et al., 2019; Chen et al., 2020; Jiang et al., 2021; Ma et al., 2021; Misovich et al., 2021; Ou et al., 52 2021; X. Li et al., 2022). By contrast, phenolic carbonyls have been mainly studied as aqSOA precursors via 'OH-, nitrate-, 53 nitrite-, and ³DMB*-mediated oxidation (Li et al., 2014; Huang et al., 2018; Pang et al., 2019; Jiang et al., 2021; Misovich et al., 2021). However, strongly light-absorbing phenolic carbonyls (e.g., molar absorptivity above 300 nm $\ge 7 \times 10^3$ M⁻¹ cm⁻¹) 54 55 can also serve as photosensitizers to promote aqSOA formation (Smith et al., 2016; Mabato et al., 2022). For instance, the direct photosensitized oxidation of phenolic carbonyls (i.e., oxidation of phenolic carbonyls by their ³C* or ³C*-derived 56 57 oxidants) such as vanillin (VL; another methoxybenzaldehyde) efficiently form low-volatility products, with aqSOA mass 58 yields of up to 140% (Smith et al., 2016). Moreover, the aqSOA mass yields from the oxidation of syringol by ³DMB* and 59 ³VL* are similar (111% and 114%, respectively; Smith et al., 2014, 2016). In addition, we recently reported that the direct 60 photosensitized oxidation of VL and guaiacol oxidation by ³VL* yield similar products (oligomers, functionalized monomers, and oxygenated ring-opening products) as observed with ³DMB* (Yu et al., 2014: Mabato et al., 2022). Guaiacol 61 62 is a non-carbonyl BB methoxyphenol with an emission rate from fireplace wood combustion in the range of 172 to 279 63 mg/kg (Schauer et al., 2001; Simoneit, 2002). The atmospheric reactivity of methoxyphenols has recently been reviewed 64 (Liu et al., 2022). However, our previous experiments (Mabato et al., 2022) were performed at a concentration (0.1 mM VL) higher than what was typically used for DMB (0.005 to 0.01 mM; Smith et al., 2014, 2015; Yu et al., 2014, 2016). Therefore,
direct comparisons between photosensitization by ³DMB* and ³VL* cannot be made. Despite the above findings, much is
still unknown about how aqSOA formation proceeds in systems using phenolic carbonyls as photosensitizers.

68 BB aerosols are typically internally mixed with other aerosol components, such as ammonium nitrate (AN; 69 Zielinski et al., 2020). Hence, aromatic carbonyls and phenols may coexist with AN in BB aerosols. Nitrate and ammonium 70 facilitate the formation of aqSOA and brown carbon (BrC) via a number of pathways. Nitrate photolysis can produce 'OH 71 and nitrating agents (e.g., 'NO2: Minero et al., 2007; Huang et al., 2018; Mabato et al., 2022; Wang et al., 2022; Yang et al., 72 2022), and ammonium reacts with carbonyls to yield N-containing heterocycles (e.g., imidazoles) and oligomers capable of 73 UV-Vis light absorption (De Haan et al., 2009, 2011; Nozière et al., 2009, 2010, 2018; Shapiro et al., 2009; Yu et al., 2011; 74 Lee et al., 2013; Powelson et al., 2014; Gen et al., 2018; Grace et al., 2019; Mabato et al., 2019). Furthermore, nitrate 75 photolysis may be an important process for SO_2 oxidation and SOA formation in the particle phase (Gen et al., 2019a, 76 2019b, 2022; Zhang et al., 2020, 2021, 2022), and it can potentially modify the morphology of atmospheric viscous particles 77 (Liang et al., 2021). Yet, understanding of the effects of inorganic nitrate on aqSOA formation remains limited. In addition, 78 aqSOA formation studies involving aromatic carbonyls and phenols have probed either photosensitization or nitrate-79 mediated photo-oxidation, but these reactions can occur simultaneously. For instance, we previously reported nitrated 80 compounds, including a potential imidazole derivative from the direct photosensitized oxidation of VL in the presence of AN (Mabato et al., 2022). Accordingly, investigations on reaction systems including both photosensitizers and AN may provide 81 82 further insights into the aqueous-phase processing of BB aerosols.

83 In this work, we compared aqSOA formation from photosensitized guaiacol (GUA) oxidation by ³C* of non-84 phenolic and phenolic methoxybenzaldehydes under identical conditions (simulated sunlight and concentration) relevant to 85 cloud and fog waters. The effects of AN on photosensitized aqSOA formation were also examined. In this study, the dominant aqSOA precursor is GUA (Henry's law constant of 9.2×10^2 M atm⁻¹; Sagebiel et al., 1992), and DMB and VL 86 87 were used as photosensitizers to oxidize GUA. DMB and VL (Henry's law constants of 7.3×10^3 M atm⁻¹ and 4.7×10^5 M atm⁻¹, respectively; Yaws, 1994; EPI Suite version 4.1, 2012; Felber et al., 2021), which are also abundant in BB emissions 88 89 (Schauer et al., 2001; Li et al., 2014; Chen et al., 2017; Pang et al., 2019; Mabato et al., 2022) and whose structures differ 90 only by one functional group (-OCH₃ for the former and -OH for the latter, Fig. 1), represented non-phenolic and phenolic 91 methoxybenzaldehydes, respectively. The structures of GUA, DMB, and VL are provided in Figure 1. Based on their 92 quantum yield of ${}^{3}C^{*}$ formation, DMB and VL have been classified as moderate and poor photosensitizers, respectively 93 (Felber et al., 2021). The photosensitized oxidation of GUA by ³DMB* or ³VL* in the absence (and presence) of AN are 94 referred to as GUA+DMB(+AN) and GUA+VL(+AN), respectively. GUA photo-oxidation by AN alone (GUA+AN) was 95 also explored for comparison with GUA+DMB+AN and GUA+VL+AN. The molar absorptivities of GUA, DMB, VL, and 96 nitrate are shown in Figure 1. The precursor and photosensitizer decay kinetics, detected products, and absorbance 97 enhancement were used to characterize the reactions. However, it should be noted that we mainly focused on the analyses of 98 the reaction products and product distribution.

99 While several studies on photo-oxidation of BB emissions are available, this work focuses on the comparison 100 between non-phenolic and phenolic methoxybenzaldehydes as photosensitizers in the absence and presence of AN for 101 aqSOA formation. We found that GUA oxidation by ³DMB* was faster and exhibited greater light absorption relative to 102 GUA+VL. These are likely attributed to the stronger photosensitizing ability of DMB and the -OH group of VL, making it 103 more prone to oxidation and more reactive towards electrophilic aromatic substitution. Oligomerization and functionalization 104 dominated in GUA+DMB and GUA+VL, but functionalization appeared to be more significant in GUA+VL due to VL 105 transformation products. Although AN did not significantly influence the oxidation kinetics due to the predominant role of 106 photosensitizer chemistry compared to nitrate, AN promoted the formation of N-containing products. These include Nheterocycles (e.g., imidazoles), suggesting the participation of ammonium in the reactions. Moreover, the product 107 distributions indicate distinct interactions between photosensitization by ³DMB* and ³VL* and AN photolysis. In particular, 108 109 AN generated more N-containing products in GUA+DMB+AN than in GUA+VL+AN, and increased the oligomers in 110 GUA+VL+AN. Furthermore, increased nitrated compounds in GUA+DMB+AN and GUA+VL+AN compared to GUA+AN 111 suggest that photosensitized reactions may promote reactions by nitrate photolysis.

112 2 Methods

113 2.1 Aqueous phase photo-oxidation experiments

114 Procedures for the photo-oxidation experiments are presented in detail in our previous study (Mabato et al., 2022). Experimental solutions were prepared using 0.1 mM guaiacol (GUA, Sigma Aldrich, >98.0%) and 0.01 mM 3.4-115 dimethoxybenzaldehyde (DMB, Acros Organics, 99+%) or 0.01 mM vanillin (VL, Acros Organics, 99%, pure), in the 116 117 absence and presence of ammonium nitrate (1 mM; AN, Acros Organics, 99+%, for analysis). These GUA and 118 methoxybenzaldehydes concentrations are within the values expected in cloud or fog drops in areas with significant wood 119 combustion (Anastasio et al., 1997; Rogge et al., 1998; Nolte et al., 2001). The AN concentration represents values usually 120 observed in cloud and fog waters (Munger et al., 1983; Collett et al., 1998; Zhang and Anastasio, 2003; Li et al., 2011; 121 Giulianelli et al., 2014: Bianco et al., 2020). It must be noted that this study did not intend to identify the AN concentrations 122 that would affect the kinetics but attempted to analyze the effects of AN on photosensitized aqSOA formation. A solution 123 composed of 0.1 mM GUA and 1 mM AN (GUA+AN) was also examined for comparison with GUA+DMB+AN and 124 GUA+VL+AN. Sulfuric acid (H₂SO₄; Acros Organics, ACS reagent, 95% solution in water) was used to adjust the pH of the 125 solutions to 4, which is within typical cloud pH values (2-7; Pye et al., 2020) and pH values observed in wood burning-126 impacted cloud and fog waters (Collett et al., 1998; Raja et al., 2008). The solutions (initial volume of 500 mL) were 127 bubbled with synthetic air (0.5 dm³/min) for 30 min before irradiation and throughout the reactions to achieve air-saturated 128 conditions (Du et al., 2011; Chen et al., 2020) and were continuously magnetically stirred. In this study, the reactions can generate ${}^{3}DMB^{*/3}VL^{*}$ and secondary oxidants (${}^{1}O_{2}$, $O_{2}^{*}/{}^{*}HO_{2}$, ${}^{\circ}OH$) but not ozone. Solutions contained in a quartz 129 130 photoreactor were irradiated using a xenon lamp (model 6258, Ozone free xenon lamp, 300 W, Newport) equipped with a

longpass filter (20CGA-305 nm cut-on filter, Newport) to eliminate light below 300 nm. The reaction temperatures were 131 132 maintained at 27 ± 2 °C using cooling fans positioned around the photoreactor and lamp housing. The averaged initial photon flux in the reactor measured from 300 to 380 nm was $\sim 3 \times 10^{15}$ photons cm⁻² s⁻¹ nm⁻¹ (Fig. 1), similar to our previous 133 work (Mabato et al., 2022). Samples were collected every 30 min for 180 min for offline analyses of (1) GUA, DMB, and 134 135 VL concentrations using ultra-high-performance liquid chromatography with photodiode array detector (UHPLC-PDA) and 136 (2) absorbance measurements using UV-Vis spectrophotometry. Moreover, the samples collected before and after irradiation 137 (180 min) were analyzed for (3) reaction products using UHPLC coupled with heated electrospray ionization Orbitrap mass 138 spectrometry (UHPLC-HESI-Orbitrap-MS) operated in positive and negative ion modes and (4) concentrations of small 139 organic acids using ion chromatography (IC). Each experiment was repeated independently at least three times. The reported 140 decay rate constants, small organic acids concentration, and absorbance enhancement were averaged from triplicate 141 experiments, and the corresponding errors represent one standard deviation. The pseudo-first-order rate constant (k') for 142 GUA decay was determined using the following equation (Huang et al., 2018):

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$$ln\left([\text{GUA}]_t/[\text{GUA}]_0\right) = -k't \tag{Eq. 1}$$

144 where $[GUA]_t$ and $[GUA]_0$ are GUA concentrations at time t and 0, respectively. DMB or VL decay rate constants were 145 calculated by replacing GUA with DMB or VL in Eq. 1. The decay rate constants were normalized to the photon flux measured for each experiment through dividing k' by the measured 2-nitrobenzaldehyde (2NB; a chemical actinometer) 146 147 decay rate constant, i(2NB) (Mabato et al., 2022). In addition, the decay rate constants were corrected for the internal light 148 screening due to DMB, VL, and AN (Leifer, 1988; Zhang and Anastasio, 2003; Smith et al., 2014, 2015, 2016). The values 149 of the internal light screening factor (S_{λ}) determined around the peak in the light absorption action spectrum (DMB: 310-335 150 nm, VL: 304-364 nm, nitrate: 300-331 nm) (Smith et al., 2014, 2015, 2016) for an 8.5 cm cell were 0.95 for GUA+AN, 0.51 151 for GUA+DMB, 0.54 for GUA+DMB+AN, 0.57 for GUA+VL, and 0.59 for GUA+VL+AN. Moreover, two independently 152 prepared samples for each reaction condition were analyzed using UHPLC-HESI-Orbitrap-MS. Only peaks that were 153 reproducibly detected in both sets of samples were considered. For clarity, the formulas discussed in this work correspond to 154 neutral analytes (e.g., with H^+ or NH_4^+ removed from the ion formula). The details of the analytical procedures are provided 155 in the Supplement (Sects. S1 to S4).

156 2.2 Calculation of normalized abundance of products

Several recent studies have used comparisons of relative abundance of products based on peak areas from mass spectrometry (MS) results (e.g., Lee et al., 2014; Romonosky et al., 2017; Wang et al., 2017; Fleming et al., 2018; Song et al., 2018; Klodt et al., 2019; Ning et al., 2019) to show the relative importance of different types of compounds (K. Wang et al., 2021). However, comparisons of relative abundance among different compounds can be subject to uncertainties as ionization efficiencies in soft ionization, such as ESI, may significantly vary between different compounds (Kebarle, 2000; Schmidt et al., 2006; Leito et al., 2008; Perry et al., 2008; Kruve et al., 2014). In our previous work (Mabato et al., 2022), we introduced the normalized abundance of products ([P], unitless) (Eq. 2) as a semi-quantitative analysis that gives an overview of how the signal intensities changed under different experimental conditions but not the quantification of the absolute product concentration. The calculation assumes equal ionization efficiencies of different compounds, which is commonly used to estimate O:C ratios of SOA (Bateman et al., 2012; Lin et al., 2012; Laskin et al., 2014; De Haan et al., 2019):

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$$[P] = \frac{A_{P,t}}{A_{GUA,t}} \cdot \frac{[GUA]_t}{[GUA]_0}$$
(Eq. 2)

169 where $A_{P,t}$ and $A_{GUA,t}$ are the extracted ion chromatogram (EIC) peak areas of the product P and GUA from UHPLC-HESI-170 Orbitrap-MS analyses at time t, respectively; $[GUA]_t$ and $[GUA]_0$ are the GUA concentrations (μ M) determined using 171 UHPLC-PDA at time t and 0, respectively. Note that the normalized abundance of products has intrinsic uncertainties due to 172 the variability in ionization efficiencies for various compounds. Moreover, it should be noted that the normalized abundance 173 of products was calculated using only the positive ion mode data as the GUA signal from the negative ion mode was weak 174 and thus may present large uncertainties during normalization. Therefore, products that may not give signals or may have 175 weak signals in the positive ion mode were possibly underestimated in the normalized product abundance. Nevertheless, it 176 enables the comparison of MS results among different experiments. As demonstrated in our previous work (Mabato et al., 177 2022) and the current study, a higher normalized abundance of products generally correlates with higher efficiency of 178 oxidation. The reported uncertainties were propagated from the changes in [GUA] measured using UHPLC-PDA and the MS 179 signal intensities.

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181 3 Results and Discussion

Using kinetics data, MS analyses, and absorbance enhancement data, we first examined the differences between GUA+DMB and GUA+VL (Sect. 3.1). Then, we analyzed GUA+DMB+AN, GUA+VL+AN, and GUA+AN (Sect. 3.2) to explore the effects of nitrate photolysis and ammonium on photosensitized aqSOA formation.

185 3.1 Comparison of photosensitized GUA oxidation by non-phenolic (³DMB*) and phenolic (³VL*) 186 methoxybenzaldehydes

Prior studies have reported that photosensitized non-carbonyl phenol oxidation in the presence of 3,4dimethoxybenzaldehyde (DMB) and vanillin (VL) (separately) was mainly driven by ³DMB* and ³VL*, respectively (Smith et al., 2014; Mabato et al., 2022), while contributions from secondary oxidants such as ${}^{1}O_{2}$ and 'OH were likely minor. However, both ³DMB* and ³VL* are efficiently quenched by O₂, suggesting that energy transfer should be considered in evaluating photosensitized processes involving these methoxybenzaldehydes (Felber et al., 2021). Moreover, it was found that ${}^{3}DMB*$, ${}^{1}O_{2}$, and O_{2}^{-} were the major contributors to the photosensitized oxidation of 4-ethylguaiacol (Chen et al., 2020). Recently, the oxidation of guaiacyl acetone (a non-conjugated phenolic carbonyl) in the presence of DMB has been reported to be initiated by ³DMB*, ¹O₂, 'OH, or methoxy radical ('OCH₃) (Misovich et al., 2021). Further studies are thus required to identify the specific oxidants in these reaction systems. In this study, reactions initiated in the presence of DMB or VL are collectively referred to as photosensitized reactions. The reaction conditions, initial guaiacol (GUA) and DMB or VL decay rate constants, normalized product abundance, and the chemical characteristics of aqSOA formed in this work are summarized in Table 1.

199 3.1.1 Kinetic analysis of photosensitization by ³DMB* and ³VL*

200 No significant loss of GUA or photosensitizers was observed for dark experiments (p > 0.05). Figure S1 shows the decay of 201 GUA, DMB, and VL under different experimental conditions. Upon irradiation, the GUA decay rate constant in GUA+DMB 202 was ~4 times higher than in GUA+VL. In GUA+DMB, the decay rate constant of GUA was ~8 times higher than that of 203 DMB, consistent with a previous study (Smith et al., 2014). Contrastingly, the decay rate constant of VL was 2.4 times 204 higher than that of GUA in GUA+VL. This VL consumption was also observed in our earlier work using 0.1 mM GUA + 205 0.1 mM VL (Mabato et al., 2022). These trends could be explained by the following reasons. First, DMB has a stronger photosensitizing ability than VL based on its higher quantum yield of ³C* formation and longer lifetime of ³DMB* 206 207 compared to ³VL* (Felber et al., 2021). Second, VL is also a phenolic compound similar to GUA, and is therefore highly reactive towards oxidation. For instance, its -OH group can be oxidized by ${}^{3}VL*$ via H-atom abstraction to form phenoxy 208 209 radicals which can undergo coupling to form oligomers (Kobayashi and Higashimura, 2003; Sun et al., 2010; Mabato et al., 210 2022). The faster consumption of VL than GUA suggests a competition between ground-state VL and GUA for reaction with 211 $^{3}VL^{*}$. Moreover, compared to a –OCH₃ group (in DMB), an –OH group (in VL) has a stronger electron-donating ability and is thus more activating towards electrophilic aromatic substitution. It should be noted that the differences in the GUA decay 212 213 rate constants among different reaction systems are not quantitatively equivalent to photosensitizing efficiencies, and a 214 detailed quantitative analysis of which is beyond the scope of this study. Nonetheless, these results suggested that GUA 215 oxidation in GUA+DMB was overall more efficient than in GUA+VL. Our kinetic analysis focused on the decay rate 216 constants of the agSOA precursor (GUA) and the photosensitizers (DMB and VL) during photosensitization under the same 217 experimental conditions (same aqSOA precursor and concentration, same photosensitizer concentration, and same lamp 218 photon flux). The effects of other factors (e.g., intersystem crossing efficiency) on the rate constants were not examined. 219 Explicit kinetic studies (e.g., Smith et al., 2014, 2015) that measure second-order rate constants should be conducted in the 220 future to extend the applicability of the kinetic parameters to other conditions.

221 3.1.2 Product distributions and chemical characteristics of aqSOA from photosensitization by ³DMB* and ³VL*

The products detected using UHPLC-HESI-Orbitrap-MS were used to characterize the aqSOA formed in this work. The signal-weighted distributions of aqSOA calculated from combined positive (POS) and negative (NEG) ion modes MS results are summarized in Figure 2. The signal-weighted distributions calculated separately from POS and NEG ion modes MS results are available in Figures S2 and S3. It should be noted that in this work, the product distributions for all experiments 226 were based on the same irradiation time of 180 min. An irradiation time of 180 min was chosen as it was sufficient to show 227 the differences in the extent of reaction of GUA among the reaction systems studied. For reaction systems with precursors of 228 different reactivities, chemical analysis at a fixed reaction time may be looking at different generations of products of each 229 precursor, as Yu et al. (2014) reported. Measuring the product distribution at a fixed time might have missed the information 230 on what/how many products are formed at the similar amounts of precursors reacted. The situation could be even more 231 complicated if different precursors had major differences in pathways and dominant intermediates. However, comparing the 232 product distributions after a certain time of light exposure, as is the case for this study, is useful to evaluate what products 233 would form after a certain time of photosensitization. Oligomers and derivatives of GUA dominated both GUA+DMB and 234 GUA+VL, in agreement with pronounced oligomerization from triplet-mediated oxidation of relatively high phenol 235 concentration (e.g., 0.1 to 3 mM; Li et al., 2014; Yu et al., 2014, 2016; Slikboer et al., 2015; Ye et al., 2019; Mabato et al., 2022). Figure 3 schematically depicts the main differences between photosensitized GUA oxidation by ³DMB* and ³VL* in 236 the absence and presence of AN. As shown in Fig. 3, ³DMB* and ³VL* can oxidize GUA via H-atom abstraction to form 237 238 phenoxy radicals which undergo coupling to form oligomers (Kobayashi and Higashimura, 2003; Sun et al., 2010; Mabato et 239 al., 2022). The higher oligomer contribution in GUA+DMB is likely due to the better photosensitizing ability of DMB than VL and partly the lower abundance of ³VL* due to fast VL consumption. VL was consumed faster than DMB during GUA 240 241 oxidation ascribable to the -OH group of VL, making it more susceptible to oxidation and more reactive towards 242 electrophilic aromatic substitution. In addition, the normalized product abundance for GUA+DMB was ~4 times higher than 243 that for GUA+VL (Table 1), further suggesting more efficient photosensitized GUA oxidation by ³DMB* than by ³VL*. The 244 oxidation of GUA or transient organic intermediates by secondary oxidants (e.g., ${}^{1}O_{2}$ and ${}^{\circ}OH$) from ${}^{3}DMB^{*}$ or ${}^{3}VL^{*}$ and the 245 fragmentation of larger compounds generate highly oxidized ring-opening products (Yu et al., 2014; Huang et al., 2018; 246 Chen et al., 2020). GUA+DMB had a higher contribution of ring-opening products than GUA+VL, likely due to the greater 247 availability of secondary oxidants in the former and fast VL consumption lowering the production of these species in 248 GUA+VL. The IC analyses also indicate the formation of small organic acids (e.g., formic acid), which appeared to have 249 higher concentrations in the presence of DMB than in VL (Fig. S4). Although no data is available for the concentration 250 changes (every 30 min) of small organic acids during the reaction, it is likely that an increasing trend would be observed as 251 fragmentation, which leads to the decomposition of initially formed oligomers and the generation of smaller oxygenated 252 products, becomes important at longer irradiation times (Huang et al., 2018). This trend has also been observed in our 253 previous work on the direct photosensitized oxidation of VL (Mabato et al., 2022), as well as other studies on 254 photosensitized oxidation of non-carbonyl phenols and phenolic carbonyls (e.g., Yu et al., 2016; Jiang et al., 2021). The 255 reactions of secondary oxidants or ring-opening products with GUA can form functionalized products. Notably, the 256 contribution of monomers in GUA+VL was almost twice as high as in GUA+DMB, ascribable to VL transformation 257 products. We previously showed that for the direct photosensitized oxidation of VL, functionalization prevails over 258 oligomerization at 0.01 mM VL, the [VL] used in this work, while oligomerization dominates at higher [VL] (0.1 mM; 259 Mabato et al., 2022).

It has been reported that oligomerization could occur during the electrospray ionization process (Yasmeen et al., 2010). In this work, it was confirmed that the oligomers observed were generated in the solutions via aqueous reactions instead of being artefacts of HESI-MS. This is based on the absence of dimers and higher oligomers in the HESI mass spectra of dark control solutions acquired by direct infusion (Yu et al., 2016).

264 The major GUA+DMB and GUA+VL products (Tables S1-S2) are mostly oligomers which can be formed through the coupling of phenoxy radicals (Kobayashi and Higashimura, 2003; Sun et al., 2010; Mabato et al., 2022). GUA+DMB 265 products matched those reported in previous works on ³DMB*- and/or 'OH-mediated phenol oxidation (Yu et al., 2014, 266 267 2016). These include GUA dimers and trimers (e.g., $C_{14}H_{14}O_4$ and $C_{21}H_{18}O_8$, #1 and 19; Table S1), aldehydes ($C_7H_6O_4$, #13; 268 Table S1), and esters ($C_{16}H_{18}O_6$, #14; Table S1). Functionalized products include $C_{11}H_{12}O_5$ and $C_{10}H_{12}O_3$ (#8 and 12; Table 269 S1). More than half of the major GUA+VL products are the same oligomers detected from GUA+DMB (e.g., $C_{13}H_{10}O_4$ and $C_{20}H_{18}O_6$, #4 and 21; Table S1). The rest are mainly functionalized species such as $C_7H_8O_4$ and $C_8H_8O_5$ (#28 and 35; Table 270 271 S2), corresponding to a hydroxylated GUA and hydroxylated VL, respectively.

272 The average elemental ratios and elemental distribution of the products (Fig. S5a-d) were consistent with those in 273 previous studies on similar reaction systems (Yu et al., 2014, 2016; Mabato et al., 2022). The majority of the GUA+DMB 274 and GUA+VL products had H:C \leq 1.0 and O:C \leq 0.5, typical for aromatic species (Mazzoleni et al., 2012; Kourtchev et al., 2014; Jiang et al., 2021). GUA+DMB had more compounds with higher O:C (≥0.6), in agreement with higher contributions 275 276 of ring-opening products than in GUA+VL (Fig. 2). The higher $\langle OS_C \rangle$ for GUA+VL than in GUA+DMB (Table 1) was 277 probably due to the significant functionalization in the former. Moreover, the distributions of OS_C and carbon number (Fig. 278 S6a-d) show that these aqSOA products have similar elemental composition to those of low-volatility oxygenated organic 279 aerosols (LV-OOA), semi-volatile oxygenated organic aerosols (SV-OOA), and slightly with biomass burning organic 280 aerosols (BBOA) (Kroll et al., 2011). Further discussions on van Krevelen diagrams (Fig. S5a-d) and OS_C vs. carbon 281 number plots (Fig. S6a-d) for GUA+DMB and GUA+VL agSOA are available in the Supplement (Sect. S5). In brief, 282 ³DMB*-initiated GUA oxidation was faster and yielded higher normalized product abundance than oxidation by ³VL*. This 283 is likely due to the stronger photosensitizing ability of DMB than VL and the -OH group of VL facilitating its rapid 284 consumption. In addition, oligomerization and functionalization dominated in both GUA+DMB and GUA+VL, as reported 285 in similar studies (Yu et al., 2014, 2016; Chen et al., 2020; Jiang et al., 2021; Misovich et al., 2021; Mabato et al., 2022). 286 However, functionalization was more prominent in the latter, attributable to the transformation of VL. Nonetheless, it must 287 be noted that for phenolic aqSOA, fragmentation will ultimately be more predominant at longer irradiation times (Huang et 288 al., 2018; Yu et al., 2016; Mabato et al., 2022).

289 3.1.3 Light absorption of aqSOA from photosensitization by ³DMB* and ³VL*

The absorbance enhancement of phenolic aqSOA generated via reactions with ${}^{3}C^{*}$ has been linked to the formation of conjugated structures due to oligomerization and functionalization (e.g., additions of hydroxyl and carbonyl groups; Yu et al., 2014, 2016; Smith et al., 2016; Ye et al., 2019; Chen et al., 2020; Jiang et al., 2021; Misovich et al., 2021; Ou et al., 293 2021; F. Li et al., 2022; X. Li et al., 2022; Mabato et al., 2022; Wang et al., 2022). Moreover, the aqueous-phase photo-294 oxidation of BB emissions can enhance BrC absorbance via the formation of aromatic dimers and functionalized products 295 (Hems et al., 2020). The increase in light absorption throughout 180 min of irradiation and the change in the rate of sunlight 296 absorption (ΔR_{abs}) (Jiang et al., 2021) from 350 to 550 nm at 180 min during typical clear and haze days in Beijing, China for 297 all the reaction systems studied are provided in Figure 4. Figure S7 shows the absorption spectra after 180 min of irradiation 298 for each reaction system studied. In this work, the absorbance enhancement of GUA+DMB and GUA+VL (Fig. 4a) could be 299 due to oligomers and functionalized monomers, which are the highest contributors to the product signals. Identifying the 300 chromophores responsible for the absorbance enhancement may be beneficial in understanding the impact of aqSOA on the 301 Earth's radiative balance and determining the reactions that affect light absorption by aqSOA (Mabato et al., 2022). 302 However, the detected products did not exhibit distinct peaks in the UHPLC-PDA chromatograms, likely due to the 303 concentration of the chromophores being below the detection limit of PDA. Nevertheless, the higher absorbance 304 enhancement and ΔR_{abs} for GUA+DMB than GUA+VL was probably due to the higher contribution and normalized 305 abundance (by ~6 times) of oligomers in the former.

306 Additional information about aqSOA lightabsorption can be deduced from the plots of the double bond equivalent 307 (DBE) values vs. carbon number (n_c) (Lin et al., 2018). Figure S8 shows these plots along with the DBE reference values of 308 fullerene-like hydrocarbons (Lobodin et al., 2012), cata-condensed polycyclic aromatic hydrocarbons (PAHs; Siegmann and 309 Sattler, 2000), and linear conjugated polyenes with a general formula $C_{x}H_{x+2}$. The shaded area indicates a sufficient level of 310 conjugation for visible light absorption, and species within this region are potential BrC chromophores. GUA+DMB and 311 GUA+VL aqSOA exhibited a significant overlap in the DBE vs. n_C space; nearly all products from both systems, including 312 the high-relative-abundance species, are potential BrC chromophores. GUA+DMB had more oligomeric products with high 313 relative abundance ($n_c \ge 12$ and DBE ≥ 8). For GUA+VL, high-relative-abundance products also include monomeric species $(n_c = 7-8 \text{ and } 4-5 \text{ DBE})$ corresponding to hydroxylated products (e.g., $C_7H_8O_4$ and $C_8H_8O_5$; 28 and 35; Table S2). These 314 315 observations further indicate the importance of oligomerization and functionalization for the absorbance enhancement of 316 aqSOA generated via photosensitization by ³DMB* and ³VL*. In summary, ³DMB* and ³VL* can oxidize GUA resulting in 317 aqSOA and BrC formation, but GUA+DMB products exhibited stronger light absorption. In GUA+VL, the extent of GUA 318 oxidation was limited by significant VL consumption.

319 **3.2** Comparison of photosensitized GUA oxidation by non-phenolic (³DMB*) and phenolic (³VL*) 320 methoxybenzaldehydes in the presence of AN

321 3.2.1 Kinetic analysis of photosensitization by ³DMB* and ³VL* in the presence of AN

322 Ammonium nitrate (AN) did not significantly affect (p > 0.05) the decay rate constants of GUA, DMB, and VL for both 323 GUA+DMB+AN and GUA+VL+AN (Table 1), likely due to the higher molar absorptivities of the photosensitizers 324 compared to that of nitrate. This implies that the chemistry of ³DMB* and ³VL* dominated that of nitrate. In this work, the 325 GUA decay rate constants decreased in the order of GUA+DMB/GUA+DMB+AN > GUA+VL/GUA+VL+AN > GUA+AN (Table 1). Note that as the molar absorptivities of the photosensitizers are higher than that of nitrate, the kinetics data were also analyzed on a per-photon-absorbed basis for a more appropriate comparison of reaction efficiency (Sect. S6). The apparent quantum efficiency of GUA photodegradation (φ_{GUA}) in the presence of nitrate (GUA+AN: $0.17 \pm 3.8 \times 10^{-2}$) was ~2 and ~7 times higher than that in the presence of DMB ($0.10 \pm 2.9 \times 10^{-3}$) or VL ($0.026 \pm 7.2 \times 10^{-3}$), respectively. This suggests that nitrate-mediated GUA photo-oxidation is more efficient than photosensitization by ³DMB* or ³VL* on a perphoton-absorbed basis.

332 3.2.2 Product distributions and chemical characteristics of aqSOA from photosensitization by ³DMB* and ³VL* in 333 the presence of AN

334 For both GUA+DMB+AN and GUA+VL+AN, AN had no significant effect on the normalized product abundance (Table 1), 335 but it induced the formation of N-containing products composed of N-heterocycles (e.g., imidazoles and pyridines) and 336 oligomers, as well as nitrated species. Similarly, we previously reported a potential imidazole derivative from the direct 337 photosensitized oxidation of VL in the presence of AN, which was attributed to the reaction of ring-opening products with 338 dissolved ammonia (Mabato et al., 2022). Oligomers remained the highest signal contributors in the presence of AN (Fig. 2), 339 but interactions between photosensitization by ³DMB* and ³VL* and AN photolysis were distinct. First, nitrated species had 340 similar contributions in both cases, but the contribution and normalized abundance of all N-containing products in 341 GUA+DMB+AN were 2 and ~14 times higher, respectively, than in GUA+VL+AN. This difference can be attributed to the higher contribution of N-heterocycles and N-containing oligomers in GUA+DMB+AN. Compared to GUA+VL, 342 343 GUA+DMB had a higher contribution of ring-opening products which can react with ammonia, as discussed earlier (Figs. 2 344 and 3). Second, the decrease in oligomers in GUA+DMB+AN may be due to their fragmentation induced by 'OH from 345 nitrate photolysis, then conversion to N-containing products. Correspondingly, the contribution of possibly ring-retaining Ncontaining products in GUA+DMB+AN (18.6%) was ~3 times higher than that in GUA+VL+AN (6.5%). While 346 347 fragmentation of oligomers likely occurred in GUA+VL+AN as well, the increase in oligomers suggests that other reactions 348 have taken place. For GUA+VL+AN, 'OH or 'NO₂ from nitrate photolysis may have initiated H-atom abstraction from the – 349 OH group of VL, generating phenoxy radicals which can undergo coupling to form more oligomers (Kobayashi and Higashimura, 2003; Sun et al., 2010; Mabato et al., 2022). This may also explain the more significant decrease of monomers 350 351 in GUA+VL+AN (~3 times) compared to GUA+DMB+AN (~2 times). Similarly, we previously observed an increase in 352 oligomers during the direct photosensitized oxidation of 0.01 mM VL (Mabato et al., 2022), the [VL] used in this work, 353 upon adding 1 mM AN. These findings indicate that photosensitization by non-phenolic and phenolic 354 methoxybenzaldehydes may interact differently with AN photolysis.

GUA+AN mainly formed oligomers analogous to 'OH-mediated phenol oxidation (Yu et al., 2014, 2016), followed by N-containing products. The normalized product abundance of GUA+AN was the lowest among all experiments, likely due to the lower GUA decay constant relative to photosensitized oxidation. Moreover, the normalized abundance of Ncontaining products in GUA+AN was ~12 times lower than that in GUA+DMB+AN but comparable to that in 359 GUA+VL+AN. This discrepancy for GUA+VL+AN might be due to the weaker signals of its N-containing products in the positive compared to the negative ion mode. As previously mentioned, the normalized product abundance was calculated 360 361 using only the positive ion mode data as the GUA signal from the negative ion mode was weak and thus may present large 362 uncertainties during normalization. Interestingly, the contributions from nitrated species in GUA+DMB+AN and 363 GUA+VL+AN were higher than in GUA+AN, suggesting possible enhancement of nitration reactions. This is likely due to the increased formation of 'NO₂, for instance, via the reactions of 'OH and O₂⁻⁻ (from ³DMB* or ³VL*) with NO₂⁻ (Pang et 364 al., 2019; Mabato et al., 2022). Similarly, we previously reported enhanced nitration via the direct photosensitized oxidation 365 366 of VL in the presence of AN under air-saturated conditions (O₂ is present) relative to nitrogen-saturated conditions (Mabato et al., 2022). These imply that photosensitization may promote reactions induced by nitrate photolysis. 367

368 The major products from GUA+DMB+AN, GUA+VL+AN, and GUA+AN (Tables S3-S5) include oligomers and 369 functionalized monomers detected in GUA+DMB and GUA+VL (Tables S1-S2). The N-heterocycles from 370 GUA+DMB+AN include C₆H₆N₄ (#41: Table S3), which may be 2.2'-biimidazole (BI), a reaction product from glyoxal + 371 reduced nitrogenous compounds (e.g., ammonium salts) (De Haan et al., 2009; Galloway et al., 2009; Nozière et al., 2009; 372 Shapiro et al., 2009; Yu et al., 2011; Kampf et al., 2012; Gen et al., 2018; Mabato et al., 2019). The nitrated products include 373 $C_{12}H_{11}N_3O_3$ and $C_{15}H_{10}N_4O_3$ (#42 and 49; Table S3), which possibly have a nitrated imidazole moiety and a nitrophenol 374 moiety, respectively. For GUA+VL+AN, oligomers (C₁₄H₁₂O₆ and C₂₀H₁₆O₇; #55 and 59, Table S4) which were not among 375 the major products in GUA+VL were noted. $C_{10}H_8O_2$ likely has a furanone group (#50; Table S4); furanones are the primary 376 products of the reaction of 'OH with toluene and other aromatic hydrocarbons (Smith et al., 1999). Moreover, C₁₁H₉N₃O₃ 377 (#57; Table S4) has a nitrated imidazole moiety. Among the N-containing compounds in GUA+AN is $C_4H_3N_3O_3$ (#69; Table 378 S5), which may be a nitrated imidazole-2-carboxaldehyde. Imidazole-2-carboxaldehyde is also a reaction product from 379 glyoxal + reduced nitrogenous compounds (e.g., ammonium salts) (De Haan et al., 2009; Galloway et al., 2009; Nozière et 380 al., 2009; Shapiro et al., 2009; Yu et al., 2011; Kampf et al., 2012; Gen et al., 2018; Mabato et al., 2019).

381 The (O:C) for GUA+DMB+AN and GUA+VL+AN were lower than those in the absence of AN (Table 1), possibly 382 due to the formation of N-heterocycles, altering the elemental ratios. The (O:C) and (H:C) were comparable in 383 GUA+DMB+AN and GUA+VL+AN, but the (N:C) for the former was higher, implying a greater extent of reactions 384 involving AN. Relative to GUA+DMB+AN and GUA+VL+AN, GUA+AN had a higher (N:C), as can be expected given 385 that AN was the only oxidant source. The lower (OS_c) of GUA+DMB+AN and GUA+VL+AN compared to GUA+AN may 386 be attributed to triplet-initiated oxidation generating higher-molecular-weight products with less fragmentation compared to 387 'OH-mediated oxidation (Yu et al., 2014; Chen et al., 2020). Nonetheless, AN generally increased the (OS_C) for both 388 GUA+DMB and GUA+VL, with a more noticeable increase for the former, suggesting more oxidized products. Similarly, in 389 a previous work, the more oxygenated and oxidized aqSOA from the photo-oxidation of phenolic carbonyls in AN solutions 390 than in ammonium sulfate solutions has been ascribed to nitrate photolytic products promoting the reactions (Huang et al., 391 2018). Furthermore, GUA+DMB+AN and GUA+VL+AN aqSOA had mainly similar features in the OS_C vs. n_C plots as 392 those observed in the absence of AN (Fig. S6). More information on van Krevelen diagrams (Figs. S5e-h and S9) and OS_C 393 vs. n_C plots (Figs. S6e-h and S10) for GUA+DMB+AN, GUA+VL+AN, and GUA+AN aqSOA are provided in the 394 Supplement (Sect. S7). In essence, AN had no significant effect on the decay kinetics ascribable to photosensitizer chemistry 395 prevailing over nitrate, but it induced the formation of N-containing products. Moreover, AN modified the product 396 distributions, albeit in different ways (Figs. 2 and 3). In particular, N-containing products were more abundant in 397 GUA+DMB+AN, probably due to more extensive fragmentation in GUA+DMB than in GUA+VL. In GUA+VL+AN, AN 398 promoted oligomer formation likely via the -OH group of VL. Furthermore, GUA+DMB+AN and GUA+VL+AN had more 399 nitrated products than GUA+AN, suggesting that photosensitized reactions may promote nitrate photolysis-initiated 400 reactions.

401 3.2.3 Light absorption of aqSOA from photosensitization by ³DMB* and ³VL* in the presence of AN

The presence of AN also did not appreciably affect the absorbance enhancement and ΔR_{abs} for both GUA+DMB+AN and GUA+VL+AN (Fig. 4). For GUA+DMB+AN, the N-containing products may have offset the decrease in oligomers to maintain the absorbance enhancement observed from GUA+DMB. Wang et al. (2022) reported that nitration might contribute significantly to absorbance enhancement for methoxyphenols in sodium nitrate. In GUA+VL+AN, the decrease in monomers may have counteracted the increased oligomers and the generated N-containing products. Compared to GUA+DMB+AN, the N-containing products from GUA+VL+AN probably had less impact on the absorbance enhancement based on their smaller signal contribution.

409 Similar to experiments without AN, CHO species from GUA+DMB+AN and GUA+VL+AN were mainly 410 overlapped in the DBE vs. n_c space (Fig. S8c,d) and were mostly potential BrC chromophores. In both systems, GUA dimers 411 were the products with the highest relative abundance. For GUA+DMB+AN, products with high relative abundance also 412 include a CHN species, while two CHON species had high n_c (18,20) and DBE (16,14) values. In GUA+VL+AN, products with high relative abundance include a CHON species ($n_c = 11$ and 9 DBE). Approximately 30% and 43% of the N-413 414 containing products for GUA+DMB+AN and GUA+VL+AN, respectively, were among the potential BrC chromophores. 415 This suggests the possible significance of N-containing products for light absorption of aqSOA from photosensitization by 416 methoxybenzaldehydes and AN photolysis. Correspondingly, nitroaromatic compounds and N-heterocycles are frequently 417 noted in BBOA (Iinuma et al., 2010; Kitanovski et al., 2012; Kourtchev et al., 2016) and have been proposed to be potential 418 contributors to BrC light absorption (Laskin et al., 2015). Relative to GUA+DMB+AN and GUA+VL+AN, only 19% of the 419 N-containing products in GUA+AN were potential BrC chromophores (Fig. S8e,f), and these did not include CHN species. 420 These indicate that the N-containing products formed in the presence of both photosensitizers and AN may be more 421 significant contributors to the light absorption of phenolic aqSOA than those formed in AN only.

422 4 Conclusions and atmospheric implications

423 The photosensitized oxidation of guaiacol (GUA) by triplet excited states of 3,4-dimethoxybenzaldehyde (³DMB*) and 424 vanillin (³VL^{*}) (separately) in the absence and presence of ammonium nitrate (AN) were compared under identical 425 conditions (simulated sunlight and concentration) relevant to atmospheric cloud and fog waters. Compared to GUA+VL, 426 faster GUA oxidation and stronger light absorption by the products were observed in GUA+DMB. Moreover, VL was 427 consumed faster relative to DMB, limiting the extent of GUA oxidation in GUA+VL. These differences are rooted in DMB 428 having a better photosensitizing ability than VL and the -OH group of VL, making it more susceptible to oxidation and more 429 reactive towards electrophilic aromatic substitution. Both GUA+DMB and GUA+VL generated aqSOA (including potential 430 BrC chromophores) composed of oligomers, functionalized monomers, oxygenated ring-opening products, and N-containing 431 products in the presence of AN. The major agSOA formation processes for GUA+DMB and GUA+VL were oligomerization 432 and functionalization, but functionalization appeared to be more significant in GUA+VL due to VL transformation products. 433 The photochemical evolution of aqSOA from GUA+DMB has been reported by Yu et al. (2016). Similar experiments for 434 aqSOA from GUA+VL should be conducted in the future to better understand photosensitized reactions involving phenolic 435 carbonyl photosensitizers.

AN did not significantly affect the decay kinetics due to the predominant effect of ³DMB* and ³VL* chemistry 436 437 compared to nitrate, but it promoted the formation of N-containing products; these are composed of N-heterocycles (e.g., imidazoles) and oligomers and nitrated species. The observation of N-heterocycles agrees with our previous findings that 438 439 ammonium participates in photosensitized oxidation of phenolic compounds in the presence of AN (Mabato et al., 2022). 440 These results also suggest that photosensitized oxidation of phenolic compounds in the presence of AN might be an 441 important source of N-heterocycles and nitrated products. Identifying the sources of N-heterocycles and nitrated compounds 442 is important due to their environmental and health impacts (Laskin et al., 2009). Moreover, photosensitized reactions by non-443 phenolic and phenolic methoxybenzaldehydes may be differently influenced by AN photolysis. For instance, the more 444 extensive fragmentation in GUA+DMB than in GUA+VL possibly resulted in more N-containing products in 445 GUA+DMB+AN. Furthermore, the increased oligomers in GUA+VL+AN may be due to VL-derived phenoxy radicals induced by 'OH or 'NO₂ from nitrate photolysis. In addition, more nitrated compounds observed in GUA+DMB+AN and 446 447 GUA+VL+AN than in GUA+AN imply that photosensitized reactions may promote nitrate-mediated photolytic reactions. 448 On a related note, the significance of photosensitization by BrC (via formation of solvated electrons; Y. Wang et al., 2021) 449 and marine dissolved organic matter (via O₂[•] formation; Garcia et al., 2021) in enhanced nitrite production from nitrate 450 photolysis have been reported. A recent study from our group has shown that glyoxal photo-oxidation mediated by both 451 nitrate photolysis and photosensitization can significantly enhance the atmospheric sink of glyoxal (Zhang et al., 2022). 452 Further studies are needed to improve our understanding of the interplay between photosensitized reactions and nitrate 453 photolysis.

454 This study demonstrates that the structural features of photosensitizers affect aqSOA formation via non-carbonyl phenol oxidation. The VL results are broadly relevant to other phenolic carbonyls, but the effects of different functional 455 456 groups should still be considered. For instance, the aldehyde/ketone pair of syringaldehyde and acetosyringone, both 457 phenolic carbonyls, have been reported to have equal reactivity towards direct photosensitized oxidation. This is due to the 458 greater light absorption by the aldehyde form but higher quantum efficiency for loss for the ketone form (Smith et al. 2016). 459 However, more aqSOA was observed from syringaldehyde than acetosyringone (in either AN or ammonium sulfate; Huang et al., 2018). Our findings also imply that while the contributions of photosensitization by ${}^{3}VL^{*}$ (and other phenolic 460 carbonyls) to aqSOA formation would be relatively less compared to that of ³DMB* (and other non-phenolic carbonyls), 461 these are not negligible. As both non-phenolic and phenolic carbonyls such as the methoxybenzaldehydes examined in this 462 463 work are emitted in large amounts from biomass burning, future experiments should probe the aqSOA contribution of a 464 wider variety of photosensitizers. Moreover, further experiments on photosensitized reactions in authentic particulate matter 465 (PM) samples should be conducted in the future. Multicomponent reactions such as GUA+DMB+AN and GUA+VL+AN should also be explored for a more accurate simulation of ambient conditions. These would be useful in assessing the overall 466 impact of photosensitized reactions and AN photolysis on aqSOA formation in areas impacted by biomass burning and high 467 468 AN concentrations, and for their better representation in aqSOA models.

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470 Data availability.

The data used in this publication are available to the community and can be accessed by request to the corresponding author.

472 Author contributions.

473 BRGM designed and conducted the experiments; BRGM and CKC wrote the paper. All co-authors contributed to the 474 discussion of the manuscript.

475 *Competing interests.*

476 The authors declare that they have no conflict of interest.

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484 **References**

Anastasio, C., Faust, B. C., and Rao, C. J.: Aromatic carbonyl compounds as aqueous-phase photochemical sources of
hydrogen peroxide in acidic sulfate aerosols, fogs, and clouds. 1. Non-phenolic methoxybenzaldehydes and
methoxyacetophenones with reductants (phenols), Environ. Sci. Technol., 31, 218–232, https://doi.org/10.1021/es960359g,
1997.

490 Aregahegn, K. Z., Felber, T., Tilgner, A., Hoffmann, E. H., Schaefer, T., and Herrmann, H.: Kinetics and mechanisms of 491 aqueous-phase reactions of triplet-state imidazole-2-carboxaldehyde and 3,4-dimethoxybenzaldehyde with α , β -unsaturated 492 carbonyl compounds, J. Phys. Chem. A, 126, 8727–8740, https://doi.org/10.1021/acs.jpca.2c05015, 2022.

493

503

Bateman, A. P., Laskin, J., Laskin, A., and Nizkorodov, S. A.: Applications of high-resolution electrospray ionization mass
spectrometry to measurements of average oxygen to carbon ratios in secondary organic aerosols, Environ. Sci. Technol., 46,
8315–832, https://doi.org/10.1021/es3017254, 2012.

497
498 Bianco, A., Passananti, M., Brigante, M., and Mailhot, G.: Photochemistry of the cloud aqueous phase: a review, Molecules,
499 25, 423, https://doi.org/10.3390/molecules25020423, 2020.

500
501 Calvert, J. G. and Madronich, S.: Theoretical study of the initial products of the atmospheric oxidation of hydrocarbons, J.
502 Geophys. Res., 92, 2211–2220, https://doi.org/10.1029/JD092iD02p02211, 1987.

504 Chen, H., Ge, X., and Ye, Z.: Aqueous-phase secondary organic aerosol formation via reactions with organic triplet excited 505 states – a short review, Curr. Pollut. Rep., 4, 8–12, https://doi.org/10.1007/s40726-018-0079-7, 2018. 506

507 Chen, Y., Li, N., Li, X., Tao, Y., Luo, S., Zhao, Z., Ma, S., Huang, H., Chen, Y., Ye, Z., and Ge, X.: Secondary organic
508 aerosol formation from ³C*-initiated oxidation of 4-ethylguaiacol in atmospheric aqueous-phase, Sci. Total Environ., 723,
509 137953, https://doi.org/10.1016/j.scitotenv.2020.137953, 2020.

511 Chen, Z. and Anastasio, C.: Concentrations of a triplet excited state are enhanced in illuminated ice, Environ. Sci.: Processes
512 Impacts, 19, 12–21, https://doi.org/10.1039/C6EM00534A, 2017.

513

510

Collett, J. L. Jr., Hoag, K. J., Sherman, D. E., Bator, A., and Richards, L. W.: Spatial and temporal variations in San Joaquin
Valley fog chemistry, Atmos. Environ., 33, 129–140, https://doi.org/10.1016/S1352-2310(98)00136-8, 1998.

516

517 De Haan, D. O., Tolbert, M. A., and Jimenez, J. L.: Atmospheric condensed-phase reactions of glyoxal with methylamine,
518 Geophys. Res. Lett., 36, No. L11819, https://doi.org/10.1029/2009GL037441, 2009.
519

De Haan, D. O., Hawkins, L. N., Kononenko, J. A., Turley, J. J., Corrigan, A. L., Tolbert, M. A., and Jimenez, J. L.:
Formation of nitrogen-containing oligomers by methylglyoxal and amines in simulated evaporating cloud droplets, Environ.
Sci. Technol., 45, 984–991, https://doi.org/10.1021/es102933x, 2011.

523

De Haan, D. O., Pajunoja, A., Hawkins, L. N., Welsh, H. G., Jimenez, N. G., De Loera, A., Zauscher, M., Andretta, A. D.,
Joyce, B. W., De Haan, A. C., Riva, M., Cui, T., Surratt, J. D., Cazaunau, M., Formenti, P., Gratien, A., Pangui, E., and
Doussin, J-F.: Methylamine's effects on methylglyoxal-containing aerosol: chemical, physical, and optical changes, ACS
Earth Space Chem., 3, 1706–1716, https://doi.org/10.1021/acsearthspacechem.9b00103, 2019.

528

529 Du, Y., Fu, Q. S., Li, Y., and Su, Y.: Photodecomposition of 4-chlorophenol by reactive oxygen species in UV/air system, J.

530 Hazard. Mater., 186, 491–496, https://doi.org/10.1016/j.jhazmat.2010.11.023, 2011.

- Edye, L. A. and Richards, G. N.: Analysis of condensates from wood smoke. components derived from polysaccharides and lignins, Environ. Sci. Technol., 25, 1133–1137, https://doi.org/10.1021/es00018a018, 1991.
- 534
- Felber, T., Schaefer, T., He, L., and Herrmann, H.: Aromatic carbonyl and nitro compounds as photosensitizers and their photophysical properties in the tropospheric aqueous phase, J. Phys. Chem. A, 125, 5078–5095, https://doi.org/10.1021/acs.jpca.1c03503, 2021.
- 538

Fleming, L. T., Lin, P., Laskin, A., Laskin, J., Weltman, R., Edwards, R. D., Arora, N. K., Yadav, A., Meinardi, S., Blake, D.
R., Pillarisetti, A., Smith, K. R., and Nizkorodov, S. A.: Molecular composition of particulate matter emissions from dung
and brushwood burning household cookstoves in Haryana, India, Atmos. Chem. Phys., 18, 2461–2480,
https://doi.org/10.5194/acp-18-2461-2018, 2018.

543

Galloway, M. M., Chhabra, P. S., Chan, A. W. H., Surratt, J. D., Flagan, R. C., Seinfeld, J. H., and Keutsch, F. N.: Glyoxal
uptake on ammonium sulphate seed aerosol: reaction products and reversibility of uptake under dark and irradiated
conditions, Atmos. Chem. Phys., 9, 3331–3345, https://doi.org/10.5194/acp-9-3331-2009, 2009.

547

554

557

Garcia, S. L. M., Pandit, S., Navea, J. G., and Grassian, V. H.: Nitrous acid (HONO) formation from the irradiation of
aqueous nitrate solutions in the presence of marine chromophoric dissolved organic matter: comparison to other organic
photosensitizers, ACS Earth Space Chem., 5, 3056–3064, https://doi.org/10.1021/acsearthspacechem.1c00292, 2021.

- 552 Gen, M., Huang, D. D., and Chan, C. K.: Reactive uptake of glyoxal by ammonium-containing salt particles as a function of 553 relative humidity, Environ. Sci. Technol., 52, 6903–6911, https://doi.org/10.1021/acs.est.8b00606, 2018.
- 555 Gen, M., Zhang, R., Huang, D. D., Li, Y. J., and Chan, C. K.: Heterogeneous SO₂ oxidation in sulfate formation by 556 photolysis of particulate nitrate, Environ. Sci. Technol. Lett., 6, 86–91, https://doi.org/10.1021/acs.estlett.8b00681, 2019a.
- Gen, M., Zhang, R., Huang, D. D., Li, Y. J., and Chan, C. K.: Heterogeneous oxidation of SO₂ in sulfate production during
 nitrate photolysis at 300 nm: effect of pH, relative humidity, irradiation intensity, and the presence of organic compounds,
 Environ. Sci. Technol., 53, 8757–8766, https://doi.org/10.1021/acs.est.9b01623, 2019b.
- Gen, M., Liang, Z., Zhang, R., Mabato, B. R. G., and Chan, C. K.: Particulate nitrate photolysis in the atmosphere, Environ.
 Sci.: Atmos., 2, 111–127, https://doi.org/10.1039/d1ea00087j, 2022.
- George, C., Brüggemann, M., Hayeck, N., Tinel, L., and Donaldson, J.: Interfacial photochemistry: physical chemistry of
 gas-liquid interfaces, in: Developments in Physical & Theoretical Chemistry, edited by: Faust, J. A. and House, J. E.,
 Elsevier, 435–457, https://doi.org/10.1016/B978-0-12-813641-6.00014-5, 2018.
- 568

- Giulianelli, L., Gilardoni, S., Tarozzi, L., Rinaldi, M., Decesari, S., Carbone, C., Facchini, M. C., and Fuzzi, S.: Fog
 occurrence and chemical composition in the Po valley over the last twenty years, Atmos. Environ., 98, 394–401,
 https://doi.org/10.1016/j.atmosenv.2014.08.080, 2014.
- 572
- Grace, D. N., Sharp, J. R., Holappa, R. E., Lugos, E. N., Sebold, M. B., Griffith, D. R., Hendrickson, H. P., and, Galloway,
 M. M.: Heterocyclic product formation in aqueous brown carbon systems, ACS Earth Space Chem., 3, 2472–2481,
 https://doi.org/10.1021/acsearthspacechem.9b00235, 2019.
- 576
- 577 Hawthorne, S. B., Miller, D. J., Langenfeld, J. J., and Krieger, M. S.: PM-10 High-volume collection and quantitation of 578 semi- and nonvolatile phenols, methoxylated phenols, alkanes, and polycyclic aromatic hydrocarbons from winter urban air 579 smoke emissions, Environ. Sci. Technol., 26, 2251 - 2262, and their relationship to wood 580 https://doi.org/10.1021/es00035a026, 1992.
- 581

- 582 Hems, R. F., Schnitzler, E. G., Bastawrous, M., Soong, R., Simpson, A. J., and Abbatt, J. P. D.: Aqueous photoreactions of 583 wood smoke brown carbon, ACS Earth Space Chem., 4, 1149–1160, https://doi.org/10.1021/acsearthspacechem.0c0011, 584 2020.
- 585
- 586 Hoshino, M., Akimoto, H., and Okuda, M.: Photochemical oxidation of benzene, toluene, and ethylbenzene initiated by OH 587 radicals in the gas phase, Bull. Chem. Soc. Jpn., 51, 718–724, https://doi.org/10.1246/bcsj.51.718, 1978.
- 589 Huang, D. D., Zhang, O., Cheung, H. H. Y., Yu, L., Zhou, S., Anastasio, C., Smith, J. D., and Chan, C. K.: Formation and 590 evolution of agSOA from aqueous-phase reactions of phenolic carbonyls: comparison between ammonium sulfate and 591 ammonium nitrate solutions, Environ, Sci. Technol., 52, 9215–9224, https://doi.org/10.1021/acs.est.8b03441, 2018.
- 592

588

- 593 Iinuma, Y., Böge, O., Gräfe, R., and Herrmann, H.: Methyl-nitrocatechols: atmospheric tracer compounds for biomass 594 burning secondary organic aerosols, Environ. Sci. Technol., 44, 8453–8459, https://doi.org/10.1021/es102938a, 2010.
- 595
- 596 Jiang, W., Misovich, M. V., Hettiyadura, A. P. S., Laskin, A., McFall, A. S., Anastasio, C., and Zhang, Q.: Photosensitized 597 reactions of a phenolic carbonyl from wood combustion in the aqueous phase-chemical evolution and light absorption 598 properties of aqSOA, Environ. Sci. Technol., 55, 5199-5211, https://doi.org/10.1021/acs.est.0c07581, 2021.
- 599
- 600 Kampf, C. J., Jakob, R., and Hoffmann, T.: Identification and characterization of aging products in the glyoxal/ammonium 601 sulfate system – implications for light-absorbing material in atmospheric aerosols, Atmos. Chem. Phys., 12, 6323–6333, 602 https://doi.org/10.5194/acp-12-6323-2012, 2012. 603
- 604 Kebarle, P. A.: A brief overview of the mechanisms involved in electrospray mass spectrometry, J. Mass Spectrom., 35. 605 804-817, https://doi.org/10.1002/9783527628728.ch1, 2000.
- 607 Kitanovski, Z., Grgić, I., Vermeylen, R., Claevs, M., and Maenhaut, W.: Liquid chromatography tandem mass spectrometry 608 method for characterization of monoaromatic nitro-compounds in atmospheric particulate matter, J. Chromatogr. A, 1268, 609 35–43, https://doi.org/10.1016/j.chroma.2012.10.021, 2012.
- 610

- 611 Klodt, A.L., Romonosky, D.E., Lin, P., Laskin, J., Laskin, A., and Nizkorodov, S.A.: Aqueous photochemistry of secondary 612 organic aerosol of α -pinene and α -humulene in the presence of hydrogen peroxide or inorganic salts, ACS Earth Space 613 Chem., 3, 12, 2736–2746, https://doi.org/10.1021/acsearthspacechem.9b00222, 2019.
- 614
- 615 Kobayashi, S. and Higashimura, H.: Oxidative polymerization of phenols revisited, Prog. Polym. Sci., 28, 1015–1048, 616 https://doi.org/10.1016/S0079-6700(03)00014-5, 2003.
- 617
- 618 Kourtchev, I., Fuller, S. J., Giorio, C., Healy, R. M., Wilson, E., O'Connor, I., Wenger, J. C., McLeod, M., Aalto, J., 619 Ruuskanen, T. M., Maenhaut, W., Jones, R., Venables, D. S., Sodeau, J. R., Kulmala, M., and Kalberer, M.: Molecular 620 composition of biogenic secondary organic aerosols using ultrahigh-resolution mass spectrometry: comparing laboratory and 621 field studies, Atmos. Chem. Phys., 14, 2155–2167, https://doi.org/10.5194/acp-14-2155-2014, 2014.
- 622
- 623 Kourtchev, I., Godoi, R. H. M., Connors, S., Levine, J. G., Archibald, A. T., Godoi, A. F. L., Paralovo, S. L., Barbosa, C. G.
- 624 G., Souza, R. A. F., Manzi, A. O., Seco, R., Sjostedt, S., Park, J., Guenther, A., Kim, S., Smith, J., Martin, S. T., and
- 625 Kalberer, M.: Molecular composition of organic aerosols in central Amazonia: an ultra-high-resolution mass spectrometry 626 study, Atmos. Chem. Phys., 16, 11899–11913, https://doi.org/10.5194/acp-16-11899-2016, 2016.
- 627
- 628 Kroll, J. H., Donahue, N. M., Jimenez, J. L., Kessler, S. H., Canagaratna, M. R., Wilson, K. R., Altieri, K. E., Mazzoleni, L.
- 629 R., Wozniak, A. S., Bluhm, H., Mysak, E. R., Smith, J. D., Kolb, C. E., and Worsnop, D. R.: Carbon oxidation state as a 3. 133 - 139.
- 630 metric for describing the chemistry of atmospheric organic aerosol, Nat. Chem.
- 631 https://doi.org/10.1038/nchem.948, 2011.

- Kruve, A., Kaupmees, K., Liigand, J., and Leito, I.: Negative electrospray ionization via deprotonation: predicting the
 ionization efficiency, Anal. Chem., 86, 4822–4830, https://doi.org/10.1021/ac404066v, 2014.
- 634
- Laskin, A., Smith, J. S., and Laskin, J.: Molecular characterization of nitrogen-containing organic compounds in biomass
 burning aerosols using high-resolution mass spectrometry, Environ. Sci. Technol., 43, 3764–3771,
 https://doi.org/10.1021/es803456n, 2009.
- 638
- Laskin, A., Laskin, J., and Nizkorodov, S. A.: Chemistry of atmospheric brown carbon, Chem. Rev., 115, 4335–4382,
 https://doi.org/10.1021/cr5006167, 2015.
- 641
- Laskin, J., Laskin, A., Nizkorodov, S. A., Roach, P., Eckert, P., Gilles, M. K., Wang, B., Lee, H. J., and Hu, Q.: Molecular
 selectivity of brown carbon chromophores, Environ. Sci. Technol., 48, 12047–12055, https://doi.org/10.1021/es503432r,
 2014.
- 645
- Lee, A. K. Y., Zhao, R., Li, R., Liggio, J., Li, S., and Abbatt, J. P. D.: Formation of light absorbing organo-nitrogen species
 from evaporation of droplets containing glyoxal and ammonium sulfate, Environ. Sci. Technol., 47, 12819–12826,
 https://doi.org/10.1021/es402687w, 2013.
- 649
- Lee, H. J., Aiona, P. K., Laskin, A., Laskin, J., and Nizkorodov, S. A.: Effect of solar radiation on the optical properties and
 molecular composition of laboratory proxies of atmospheric brown carbon, Environ. Sci. Technol., 48, 10217–10226,
 https://doi.org/10.1021/es502515r, 2014.
- Leifer, A.: The Kinetics of environmental aquatic photochemistry: Theory and practice, American Chemical Society, Washington, DC, 1988.
- 656

Leito, I., Herodes, K., Huopolainen, M., Virro, K., Künnapas, A., Kruve, A., and Tanner, R.: Towards the electrospray
ionization mass spectrometry ionization efficiency scale of organic compounds, Rapid Commun. Mass Sp., 22, 379–384,
https://doi.org/10.1002/rcm.3371, 2008.

- Li, F., Zhou, S., Du, L., Zhao, J., Hang, J., and Wang, X.: Aqueous-phase chemistry of atmospheric phenolic compounds: A
 critical review of laboratory studies, Sci. Total Environ., 856, 158895, https://doi.org/10.1016/j.scitotenv.2022.158895,
 2022.
 - 663 664

Li, P., Li, X., Yang, C., Wang, X., Chen, J., and Collett, J. L. Jr.: Fog water chemistry in Shanghai, Atmos. Environ., 45,
4034–4041, https://doi.org/10.1016/j.atmosenv.2011.04.036, 2011.

667

Li, X., Tao, Y., Zhu, L., Ma, S., Luo, S., Zhao, Z., Sun, N., Ge, X., and Ye, Z.: Optical and chemical properties and oxidative
potential of aqueous-phase products from OH and ³C*-initiated photooxidation of eugenol, Atmos. Chem. Phys., 22, 7793–
7814, https://doi.org/10.5194/acp-22-7793-2022, 2022.

- 671
- Li, Y. J., Huang, D. D., Cheung, H. Y., Lee, A. K. Y., and Chan, C. K.: Aqueous-phase photochemical oxidation and direct
 photolysis of vanillin a model compound of methoxy phenols from biomass burning, Atmos. Chem. Phys., 14, 2871–2885,
 https://doi.org/10.5194/acp-14-2871-2014, 2014.
- 675
- Liang, Z., Zhang, R., Gen, M., Chu, Y., and Chan, C. K.: Nitrate photolysis in mixed sucrose–nitrate–sulfate particles at
 different relative humidities, J. Phys. Chem. A, 125, 3739–3747, https://doi.org/10.1021/acs.jpca.1c00669, 2021.
- 679 Lin, P., Yu, J. Z., Engling, G., and Kalberer, M.: Organosulfates in humic-like substance fraction isolated from aerosols at
- seven locations in East Asia: a study by ultra-high-resolution mass spectrometry, Environ. Sci. Technol., 46, 13118–13127,
- 681 https://doi.org/10.1021/es303570v, 2012.

- Lin, P., Fleming, L. T., Nizkorodov, S. A., Laskin, J., and Laskin, A.: Comprehensive molecular characterization of
 atmospheric brown carbon by high resolution mass spectrometry with electrospray and atmospheric pressure
 photoionization, Anal. Chem., 90, 12493–12502, https://doi.org/10.1021/acs.analchem.8b02177, 2018.
- 685

688

691

Lipari, F., Dasch, J. M., and Scruggs, W. F.: Aldehyde emissions from wood-burning fireplaces, Environ. Sci. Technol., 18,
326–330, https://doi.org/10.1021/es00123a007, 1984.

- Liu, C., Chen, D., and Chen, X.: Atmospheric reactivity of methoxyphenols: a review, Environ. Sci. Technol., 56, 2897–
 2916, https://doi.org/10.1021/acs.est.1c06535, 2022.
- Liu, Y., Lu, J., Chen, Y., Liu, Y., Ye, Z., and Ge, X.: Aqueous-phase production of secondary organic aerosols from
 oxidation of dibenzothiophene (DBT), Atmosphere, 11, 151, https://doi.org/10.3390/atmos11020151, 2020.
- 694

697

- Lobodin, V. V., Marshall, A. G., and Hsu, C. S.: Compositional space boundaries for organic compounds, Anal. Chem., 84,
 3410–3416, https://doi.org/10.1021/ac300244f, 2012.
- Lu, J., Ge, X., Liu, Y., Chen, Y., Xie, X., Ou, Y., Ye, Z., and Chen, M.: Significant secondary organic aerosol production from aqueous-phase processing of two intermediate volatility organic compounds, Atmos. Environ., 211, 63– 68, https://doi.org/10.1016/j.atmosenv.2019.05.014, 2019.
- 701
- Ma, L., Guzman, C., Niedek, C., Tran, T., Zhang, Q., and Anastasio, C.: Kinetics and mass yields of aqueous secondary
 organic aerosol from highly substituted phenols reacting with a triplet excited state, Environ. Sci. Technol., 55, 5772–
 5781, https://doi.org/10.1021/acs.est.1c00575, 2021.
- 705
- Mabato, B. R. G., Gen, M., Chu, Y., and Chan, C. K.: Reactive uptake of glyoxal by methylaminium-containing salts as a
 function of relative humidity, ACS Earth Space Chem., 3, 150–157, https://doi.org/10.1021/acsearthspacechem.8b00154,
 2019.
- Mabato, B. R. G., Lyu, Y., Ji, Y., Li, Y. J., Huang, D. D., Li, X., Nah, T., Lam, C. H., and Chan, C. K.: Aqueous secondary
 organic aerosol formation from the direct photosensitized oxidation of vanillin in the absence and presence of ammonium
 nitrate, Atmos. Chem. Phys., 22, 273–293, https://doi.org/10.5194/acp-22-273-2022, 2022.
- 713

Mazzoleni, L. R., Saranjampour, P., Dalbec, M. M., Samburova, V., Hallar, A. G., Zielinska, B., Lowenthal, D. H., and
Kohl, S.: Identification of water-soluble organic carbon in non-urban aerosols using ultrahigh-resolution FT-ICR mass
spectrometry: organic anions, Environ. Chem., 9, 285–297, https://doi.org/10.1071/EN11167, 2012.

- 717
- Minero, C., Bono, F., Rubertelli, F., Pavino, D., Maurino, V., Pelizzetti, E., and Vione, D.: On the effect of pH in aromatic
 photonitration upon nitrate photolysis, Chemosphere, 66, 650–656, https://doi.org/10.1016/j.chemosphere.2006.07.082, 2007.
- Misovich, M. V., Hettiyadura, A. P. S., Jiang, W., Zhang, Q., and Laskin, A.: Molecular-level study of the photo-oxidation
 of aqueous-phase guaiacyl acetone in the presence of ³C*: formation of brown carbon products, ACS Earth Space Chem., 5,
 1983–1996, https://doi.org/10.1021/acsearthspacechem.1c00103, 2021.
- 724
- Munger, J. W., Jacob, D. J., Waldman, J. M., and Hoffmann, M. R.: Fogwater chemistry in an urban atmosphere, J.
 Geophys. Res. [Oceans], 88, 5109–5121, https://doi.org/10.1029/JC088iC09p05109, 1983.
- 727
- Ning, C., Gao, Y., Zhang, H., Yu, H., Wang, L., Geng, N., Cao, R., and Chen, J.: Molecular characterization of dissolved organic matters in winter atmospheric fine particulate matters (PM2.5) from a coastal city of northeast China, Sci. Total
- 730 Environ., 689, 312–321, https://doi.org/10.1016/j.scitotenv.2019.06.418, 2019.
- 731

- Nolte, C. G., Schauer, J. J., Cass, G. R., and Simoneit, B. R. T.: Highly polar organic compounds present in wood smoke and
 in the ambient atmosphere, Environ. Sci. Technol., 35, 1912–1919, https://doi.org/10.1021/es001420r, 2001.
- 734
- Nozière, B., Dziedzic, P., and Córdova, A.: Products and kinetics of the liquid-phase reaction of glyoxal catalyzed by ammonium ions (NH_4^+), J. Phys. Chem. A, 113, 231–237, https://doi.org/10.1021/jp8078293, 2009.
- 737

Nozière, B., Dziedzic, P., and Córdova, A.: Inorganic ammonium salts and carbonate salts are efficient catalysts for aldol
condensation in atmospheric aerosols, Phys. Chem. Chem. Phys., 12, 3864–3872, https://doi.org/10.1039/B924443C, 2010.

740

Nozière, B., Fache, F., Maxut, A., Fenet, B., Baudouin, A., Fine, L., and Ferronato, C.: The hydrolysis of epoxides catalyzed
by inorganic ammonium salts in water: kinetic evidence for hydrogen bond catalysis, Phys. Chem. Chem. Phys., 20,
1583–1590, https://doi.org/10.1039/C7CP06790A, 2018.

744

Ou, Y., Nie, D., Chen, H., Ye, Z., and Ge, X.: Characterization of products from the aqueous-phase photochemical oxidation
 of benzene-diols, Atmosphere, 12, 534, https://doi.org/10.3390/atmos12050534, 2021.

747

Pang, H., Zhang, Q., Lu, X., Li, K., Chen, H., Chen, J., Yang, X., Ma, Y., Ma, J., and Huang, C.: Nitrite-mediated
photooxidation of vanillin in the atmospheric aqueous phase, Environ. Sci. Technol., 53, 14253–14263,
https://doi.org/10.1021/acs.est.9b03649, 2019.

751

Perry, R. H., Cooks, R. G., and Noll, R. J.: Orbitrap mass spectrometry: instrumentation, ion motion and applications, Mass
Spectrom. Rev., 27, 661–699, https://doi.org/10.1002/mas.20186, 2008.

Powelson, M. H., Espelien, B. M., Hawkins, L. N., Galloway, M. M., and De Haan, D. O.: Brown carbon formation by
aqueous-phase carbonyl compound reactions with amines and ammonium sulfate, Environ. Sci. Technol., 48, 985–993,
https://doi.org/10.1021/es4038325, 2014.

758

Pye, H. O. T., Nenes, A., Alexander, B., Ault, A. P., Barth, M. C., Clegg, S. L., Collett, J. L. Jr., Fahey, K. M., Hennigan, C.
J., Herrmann, H., Kanakidou, M., Kelly, J. T., Ku, I., McNeill, V. F., Riemer, N., Schaefer, T., Shi, G., Tilgner, A., Walker,
J. T., Wang, T., Weber, R., Xing, J., Zaveri, R. A., and Zuend, A.: The acidity of atmospheric particles and clouds, Atmos.
Chem. Phys., 20, 4809–4888, https://doi.org/10.5194/acp-20-4809-2020, 2020.

763

Raja, S., Raghunathan, R., Yu, X., Lee, T., Chen, J., Kommalapati, R. R., Murugesan, K., Shen, X., Qingzhong, Y., Valsaraj,
K. T., and Collett, J. L. Jr.: Fog chemistry in the Texas-Louisiana Gulf Coast corridor, Atmos. Environ., 42, 2048–2061,
https://doi.org/10.1016/j.atmosenv.2007.12.004, 2008.

Rogge, W. F., Hildemann, L. M., Mazurek, M. A., and Cass, G. R.: Sources of fine organic aerosol. 9. Pine, oak, and
synthetic log combustion in residential fireplaces, Environ. Sci. Technol., 32, 13–22, https://doi.org/10.1021/es960930b,
1998.

771

772Romonosky, D. E., Li, Y., Shiraiwa, M., Laskin, A., Laskin, J., and Nizkorodov, S. A.: Aqueous photochemistry of773secondary organic aerosol of α-Pinene and α-Humulene oxidized with ozone, hydroxyl radical, and nitrate radical, J. Phys.774Chem. A, 121, 1298–1309, https://doi.org/10.1021/acs.jpca.6b10900, 2017.

775

Sagebiel, J. C., Seiber, J. N., and Woodrow, J. E.: Comparison of headspace and gas-stripping methods for determining the
Henry's law constant (H) for organic compounds of low to intermediate H, Chemosphere, 25, 1763–1768,
https://doi.org/10.1016/0045-6535(92)90017-L, 1992.

- 780Schauer, J. J., Kleeman, M. J., Cass, G. R., and Simoneit, B. R. T.: Measurement of emissions from air pollution sources. 3.781 C_1 - C_{29} organic compounds from fireplace combustion of wood, Environ. Sci. Technol., 35, 1716–1728,782https://doi.org/10.1021/es001331e, 2001.
- 783
- Schmidt, A-C., Herzschuh, R., Matysik, F-M., and Engewald, W.: Investigation of the ionisation and fragmentation
 behaviour of different nitroaromatic compounds occurring as polar metabolites of explosives using electrospray ionisation
 tandem mass spectrometry, Rapid Commun. Mass Sp., 20, 2293–2302, https://doi.org/10.1002/rcm.2591, 2006.
- 787
- Shapiro, E. L., Szprengiel, J., Sareen, N., Jen, C. N., Giordano, M. R., and McNeill, V. F.: Light-absorbing secondary
 organic material formed by glyoxal in aqueous aerosol mimics, Atmos. Chem. Phys., 9, 2289–2300,
 https://doi.org/10.5194/acp-9-2289-2009, 2009.
- 791
- Siegmann, K. and Sattler, K.: Formation mechanism for polycyclic aromatic hydrocarbons in methane flames, J. Chem.
 Phys., 112, 698–709, https://doi.org/10.1063/1.480648, 2000.
- 794
- Simoneit, B. R. T.: Biomass burning a review of organic tracers for smoke from incomplete combustion, Appl.
 Geochem., 17, 129–162, https://doi.org/10.1016/S0883-2927(01)00061-0, 2002.
- 797
- Simoneit, B. R. T., Rogge, W. F., Mazurek, M. A., Standley, L. J., Hildemann, L. M., and Cass, G. R.: Lignin pyrolysis
 products, lignans, and resin acids as specific tracers of plant classes in emissions from biomass combustion, Environ. Sci.
 Technol., 27, 2533–2541, https://doi.org/10.1021/es00048a034, 1993.
- Simoneit, B. R. T., Schauer, J. J., Nolte, C. G., Oros, D. R., Elias, V. O., Fraser, M. P., Rogge, W. F., and Cass, G. R.:
 Levoglucosan, a tracer for cellulose in biomass burning and atmospheric particles, Atmos. Environ., 33, 173–182,
 https://doi.org/10.1016/S1352-2310(98)00145-9, 1999.
- 805

- Simpson, C. D., Paulsen, M., Dills, R. L., Liu, L.-J. S., and Kalman, D. A.: Determination of methoxyphenols in ambient
 atmospheric particulate matter: tracers for wood combustion, Environ. Sci. Technol., 39, 631–637,
 https://doi.org/10.1021/es0486871, 2005.
- 809
- Slikboer, S., Grandy, L., Blair, S. L., Nizkorodov, S. A., Smith, R. W., and Al-Abadleh, H. A.: Formation of light absorbing
 soluble secondary organics and insoluble polymeric particles from the dark reaction of catechol and guaiacol with Fe(III),
 Environ. Sci. Technol., 49, 7793–7801, https://doi.org/10.1021/acs.est.5b01032, 2015.
- 813
- Smith, D. F., Kleindienst, T. E., and McIver, C. D.: Primary product distributions from the reaction of OH with m-, p-xylene,
 1,2,4-and 1,3,5-trimethylbenzene, J. Atmos. Chem., 34, 339–364, https://doi.org/10.1023/A:1006277328628, 1999.
- 816
- Smith, J. D., Sio, V., Yu, L., Zhang, Q., and Anastasio, C.: Secondary organic aerosol production from aqueous reactions of
 atmospheric phenols with an organic triplet excited state, Environ. Sci. Technol., 48, 1049–1057,
 https://doi.org/10.1021/es4045715, 2014.
- 820
- Smith, J. D., Kinney, H., and Anastasio, C.: Aqueous benzene-diols react with an organic triplet excited state and hydroxyl 821 822 organic aerosol, Phys. Chem. Chem. Phys., 10227-10237, radical to form secondary 17, 823 https://doi.org/10.1039/C4CP06095D, 2015.
- 824
- Smith, J. D., Kinney, H., and Anastasio, C.: Phenolic carbonyls undergo rapid aqueous photodegradation to form lowvolatility, light-absorbing products, Atmos. Environ., 126, 36–44, https://doi.org/10.1016/j.atmosenv.2015.11.035, 2016.
- 828 Song, J., Li, M., Jiang, B., Wei, S., Fan, X., and Peng, P.: Molecular characterization of water-soluble humic like substances
- 829 in smoke particles emitted from combustion of biomass materials and coal using ultrahigh-resolution electrospray ionization

- 830 Fourier transform ion cyclotron resonance mass spectrometry, Environ. Sci. Technol., 52, 2575–2585, 831 https://doi.org/10.1021/acs.est.7b06126, 2018.
- 832
- Sun, Y. L., Zhang, Q., Anastasio, C., and Sun, J.: Insights into secondary organic aerosol formed via aqueous-phase
 reactions of phenolic compounds based on high resolution mass spectrometry, Atmos. Chem. Phys., 10, 4809–4822,
 https://doi.org/10.5194/acp-10-4809-2010, 2010.
- 836

US EPA: Estimation Programs Interface Suite[™] for Microsoft[®] Windows, v 4.1, United States Environmental Protection
 Agency, Washington, DC, USA, 2012.

839

840 Wang, K., Huang, R-J., Brüggemann, M., Zhang, Y., Yang, L., Ni, H., Guo, J., Wang, M., Han, J., Bilde, M., Glasius, M., 841 and Hoffmann, T.: Urban organic aerosol composition in eastern China differs from north to south: molecular insight from a 842 liquid chromatography-mass spectrometry (Orbitrap) study, Atmos. Chem. Phys., 21. 9089-9104. 843 https://doi.org/10.5194/acp-21-9089-2021, 2021.

844

845 Wang, X., Hayeck, N., Brüggemann, M., Yao, L., Chen, H., Zhang, C., Emmelin, C., Chen, J., George, C., and Wang, L.: 846 Chemical characterization of organic aerosols in Shanghai: A study by ultrahigh-performance liquid chromatography 847 coupled with orbitrap spectrometry, J. Geophys. Res. Atmos., 122. 11703-11722, mass 848 https://doi.org/10.1002/2017JD026930, 2017.

849

Wang, Y., Huang, D. D., Huang, W., Liu, B., Chen, Q., Huang, R., Gen, M., Mabato, B. R. G., Chan, C. K., Li, X., Hao, T.,
Tan, Y., Hoi, K. I., Mok, K. M., and Li, Y. J.: Enhanced nitrite production from the aqueous photolysis of nitrate in the
presence of vanillic acid and implications for the roles of light-absorbing organics, Environ. Sci. Technol., 55, 15694–15704,
https://doi.org/10.1021/acs.est.1c04642, 2021.

854

Wang, Y., Huang, W., Tian, L., Wang, Y., Li, F., Huang, D. D., Zhang, R., Mabato, B. R. G., Huang, R., Chen, Q., Ge, X.,
Du, L., Ma, Y. G., Gen, M., Hoi, K. I., Mok, K. M., Yu, J. Z., Chan, C. K., Li, X., and Li, Y. J.: Decay kinetics and
absorption changes of methoxyphenols and nitrophenols during nitrate-mediated aqueous photochemical oxidation at 254
and 313 nm, ACS Earth Space Chem., 6, 1115–1125, https://doi.org/10.1021/acsearthspacechem.2c00021, 2022.

- 859
 860 Yang, J., Au, W. C., Law, H., Leung, C. H., Lam, C. H, and Nah, T.: pH affects the aqueous-phase nitrate-mediated
 861 photooxidation of phenolic compounds: implications for brown carbon formation and evolution, Environ. Sci.: Processes
 862 Impacts, https://doi.org/10.1039/D2EM00004K, 2022.
- 863

864Yasmeen, F., Vermeylen, R., Szmigielski, R., Iinuma, Y., Böge, O., Herrmann, H., Maenhaut, W., and Claeys, M.:865Terpenylic acid and related compounds: precursors for dimers in secondary organic aerosol from the ozonolysis of α and β-866pinene, Atmos. Chem. Phys., 10, 9383–9392, https://doi.org/10.5194/acp-10-9383-2010, 2010.

867

Yaws, C. L.: Handbook of vapor pressure, volume 3: Organic compounds C8 to C28, Gulf Professional Publishing, 1994.

- Ye, Z., Qu, Z., Ma, S., Luo, S., Chen, Y., Chen, H., Chen, Y., Zhao, Z., Chen, M., and Ge, X.: A comprehensive
 investigation of aqueous-phase photochemical oxidation of 4-ethylphenol, Sci. Total Environ., 685, 976–985,
 https://doi.org/10.1016/j.scitotenv.2019.06.276, 2019.
- 873

Yu, G., Bayer, A. R., Galloway, M. M., Korshavn, K. J., Fry, C. G., and Keutsch, F. N.: Glyoxal in aqueous ammonium
sulfate solutions: products, kinetics and hydration effects, Environ. Sci. Technol., 45, 6336–6342,
https://doi.org/10.1021/es200989n, 2011.

- Yu, L., Smith, J., Laskin, A., Anastasio, C., Laskin, J., and Zhang, Q.: Chemical characterization of SOA formed from
 aqueous-phase reactions of phenols with the triplet excited state of carbonyl and hydroxyl radical, Atmos. Chem. Phys., 14,
 13801–13816, https://doi.org/10.5194/acp-14-13801-2014, 2014.
- 881
- 882 Yu, L., Smith, J., Laskin, A., George, K. M., Anastasio, C., Laskin, J., Dillner, A. M., and Zhang, Q.: Molecular
- transformations of phenolic SOA during photochemical aging in the aqueous phase: competition among oligomerization, functionalization, and fragmentation, Atmos, Chem. Phys., 16, 4511–4527, https://doi.org/10.5194/acp-16-4511-2016.
- 884 885
- 886Zhang, Q. and Anastasio, C.: Conversion of fogwater and aerosol organic nitrogen to ammonium, nitrate, and NO_x during887exposure to simulated sunlight and ozone, Environ. Sci. Technol., 37, 3522–3530, https://doi.org/10.1021/es034114x, 2003.
- 888
- 889 Zhang, R., Gen, M., Huang, D. D., Li, Y. J., and Chan, C. K.: Enhanced sulfate production by nitrate photolysis in the presence 890 of halide ions in atmospheric particles, Environ. Sci. Technol.. 54. 3831-3839, 891 https://doi.org/10.1021/acs.est.9b06445, 2020.
- 892
- Zhang, R., Gen, M., Fu, T-M., and Chan, C. K.: Production of formate via oxidation of glyoxal promoted by particulate
 nitrate photolysis, Environ. Sci. Technol., 55, 5711–5720, https://doi.org/10.1021/acs.est.0c0819, 2021.
- 895
- Zhang, R., Gen, M., Liang, Z., Li, Y. J., and Chan, C. K.: Photochemical reactions of glyoxal during particulate ammonium
 nitrate photolysis: Brown carbon formation, enhanced glyoxal decay, and organic phase formation, Environ. Sci. Technol.,
 56, 1605–1614, https://doi.org/10.1021/acs.est.1c07211, 2022.
- Zielinski, T., Bolzacchini, E., Cataldi, M., Ferrero, L., Graßl, S., Hansen, G., Mateos, D., Mazzola, M., Neuber, R., Pakszys,
 P., Posyniak, M., Ritter, C., Severi, M., Sobolewski, P., Traversi, R., and Velasco-Merino, C.: Study of chemical and optical
 properties of biomass burning aerosols during long-range transport events toward the Arctic in summer 2017, Atmosphere,
 11, 84, https://doi.org/10.3390/atmos11010084, 2020.
- 904
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921 Table 1. Reaction conditions, initial GUA (and DMB or VL) decay rate constants, normalized abundance of products, 922 average elemental ratios, and average carbon oxidation state ((OS_C)) in each experiment. The reaction systems consisted of 923 GUA (0.1 mM), DMB (0.01 mM), VL (0.01 mM), and AN (1 mM) under air-saturated conditions after 180 min of simulated 924 sunlight irradiation. The UHPLC-HESI-Orbitrap-MS data were obtained in both positive (POS) and negative (NEG) ion 925 modes.

Exp no.	Reaction conditions	Initial GUA (and DMB or VL) decay rate constants (min ⁻¹ /s ⁻¹) ^a	Normalized abundance of products ^b	Normalized abundance of N- containing compounds ^b	⟨O:C⟩°	⟨H:C⟩ °	⟨N:C⟩°	⟨OS _C ⟩°
1	GUA+DMB	GUA: 6.3 ± 0.25	376 ± 22	NA	POS: 0.34	0.91	NA	-0.22
		DMB: 0.78 ± 0.10			NEG: 0.40	0.94	NA	-0.15
2	GUA+ DMB+AN	GUA: 5.3 ± 0.50	210 + 4	114	POS: 0.28	0.94	0.12	-0.03
		DMB: 0.69 ± 0.052	310 ± 4		NEG: 0.37	0.91	0.04	-0.05
3	GUA+VL	GUA: 1.5 ± 0.14	94 ± 5	NA	POS: 0.41	0.91	NA	-0.10
		VL: 3.6 ± 0.55	77 - 5		NEG: 0.40	0.94	NA	-0.14
4	GUA+ VL+AN	GUA: 1.6 ± 0.12 VL: 2.9 ± 0.032	100 ± 2	8	POS: 0.31	1.02	0.02	-0.34
					NEG: 0.39	0.91	0.03	-0.02
5	GUA+AN	0.57 ± 0.036	23 ± 1	9	POS: 0.35	0.99	0.16	0.19
					NEG: 0.38	1.01	0.05	-0.08

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928 ^aThe data fitting was performed in the initial linear region. Each value is the average of results from triplicate experiments, corrected for internal light screening due to DMB, VL, and AN, and normalized to the experimental photon flux. Errors 929 930 represent one standard deviation. ^bThe normalized product abundance was calculated using the data from UHPLC-HESI-931 Orbitrap-MS in the positive (POS) ion mode as the GUA signal from the negative (NEG) ion mode was weak, which may introduce significant uncertainties during normalization. The uncertainties were propagated from the changes in [GUA] 932 933 measured using UHPLC-PDA and the MS signal intensities. The samples for experiments without AN (marked with NA) were not analyzed for N-containing compounds. "The average elemental ratios ($\langle O:C \rangle$, $\langle H:C \rangle$, and $\langle N:C \rangle$) and $\langle OS_C \rangle$ were 934 935 based on the UHPLC-HESI-Orbitrap-MS results and estimated using the signal-weighted method (Bateman et al., 2012). 936 The OS_C of GUA, DMB, and VL are -0.57, -0.44, and -0.25, respectively.

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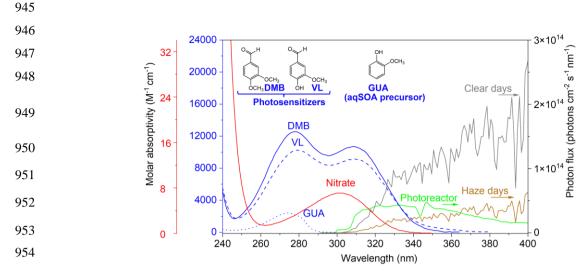


Figure 1. The base-10 molar absorptivities (M^{-1} cm⁻¹) of 3,4-dimethoxybenzaldehyde (DMB, blue solid line), vanillin (VL, blue dashed line), guaiacol (GUA, blue dotted line), and nitrate (red solid line). The green line is the photon flux in the aqueous photoreactor. The gray and brown lines are the photon fluxes on typical clear and haze days, respectively, in Beijing, China (Mabato et al., 2022). The top of the figure also shows the structures of DMB, VL, and GUA.

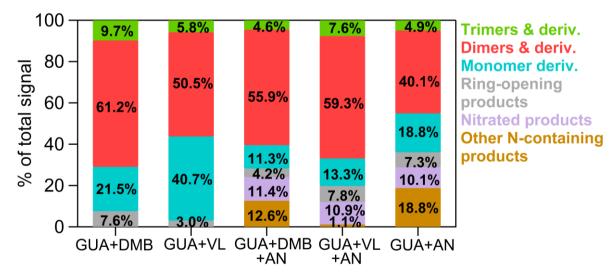


Figure 2. Signal-weighted distributions of aqSOA from GUA+DMB, GUA+VL, GUA+DMB+AN, GUA+VL+AN, and GUA+AN. These product distributions were calculated from combined UHPLC-HESI-Orbitrap-MS data obtained in positive (POS) and negative (NEG) ion modes. The values indicate the contribution of different product classifications to the total signals for each reaction condition.

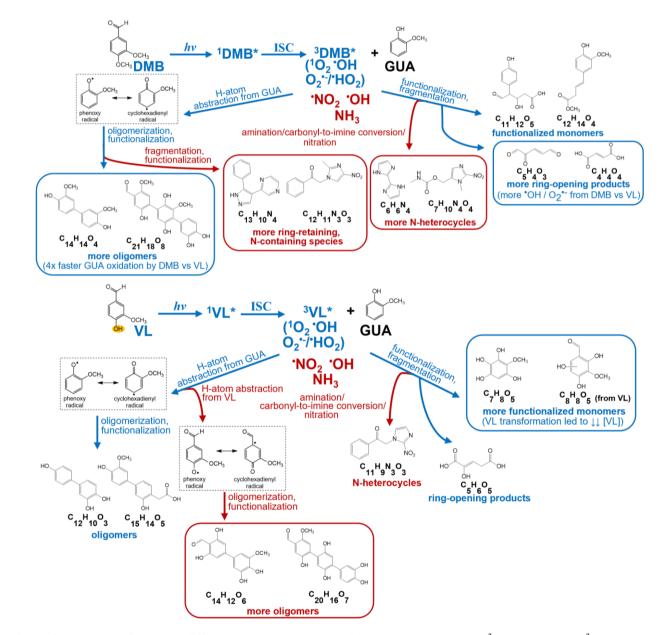


Figure 3. Summary of the main differences between photosensitized GUA oxidation by ³DMB* (top) and ³VL* (bottom) in the absence (blue labels and boxes) and presence (red labels and boxes) of ammonium nitrate at pH 4 under air-saturated conditions. Boxed structures indicate product classifications with notable differences. DMB and VL absorb light and are promoted to their singlet excited states (¹DMB* and ¹VL*), which then undergo intersystem crossing (ISC) to form ³DMB* and ³VL*. Secondary oxidants (¹O₂, O₂*-/*HO₂, 'OH) can be formed from ³DMB* and ³VL* upon reactions with O₂ and GUA (George et al., 2018; Chen et al., 2020; Misovich et al., 2021; Mabato et al., 2022). The structures shown are examples of the major products (Tables S1 to S4) for different product classifications.

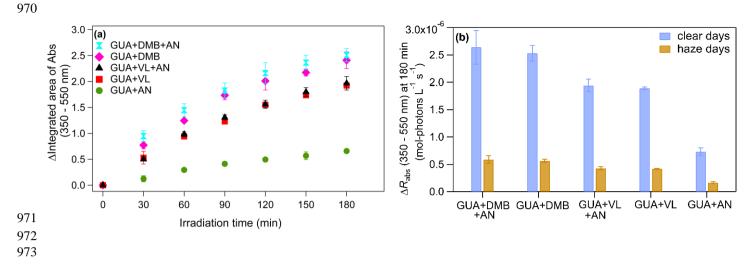


Figure 4. (a) Increase in light absorption throughout 180 min of irradiation for all reaction systems studied and (b) Change in the rate of sunlight absorption (ΔR_{abs}) from 350-550 nm at 180 min during typical clear and haze days in Beijing, China for aqSOA from GUA+DMB+AN, GUA+DMB, GUA+VL+AN, GUA+VL, and GUA+AN. Error bars represent one standard deviation of triplicate experiments.