Reviewer #1:

This paper studies the city scale NO_x and CO_2 emissions on a very high spatial and temporal resolution. It provides insights into the real-time and detailed emission quantification and control of NO_x and CO_2 . This work is interesting from a scientific point of view and is well organized and developed. Some minor revisions are suggested:

 Only the photochemical loss of NO₂ is considered in the establishment of the superposition column model, how does the other pathways of NO2 loss? Are they also play a role in NO_x chemistry?

Response: In the daytime, NO_x are mainly subjected to photochemical reaction with the hydroxyl radical (OH) to produce nitric acid (HNO₃), which is quickly converted to nitrate aerosols (NO₃⁻). In some rural regions with substantial VOCs emissions, NO_x may also react with VOCs : $CH_3O_2 + NO_2 + M \leq PAN + M$ (e. g. Fischer et al., 2014). We have compared the GEOS-Chem model simulated HNO₃+NO₃⁻ and PAN concentrations over our study domain in daytime:



Figure S3: GEOS-chem model simulated monthly mean HNO3+NO3- (left) and PAN (right) concentration for July 2020.

The modeled $HNO_3+NO_3^-$ concentration is 5-10 times higher than PAN over Wuhan, indicating that NO_x loss via OH is the driving pathway of NO_x chemistry over our study domain in the daytime.

At night, NO₂ is oxidized to HNO₃ through the formation of N₂O₅ and NO₃ and heterogeneous reactions including water vapor and aerosols (Shah et al., 2020; Lamsal et al., 2010). The overpass time of the satellite is 13:00–13:30 local time when NO_x chemistry over the city is dominated by the photochemical process, so we consider the reaction between NO_x and OH in the superposition column model as the main loss pathway. We have clarified this point in the revised manuscript in Page 5, Line 133-134 and in the Text S2 and Fig. S3 in the revised supplementary material.

2. It is not clear to me how the 'starting background NO₂ value' is determined.

Response: For each day, we calculate the mean NO₂ line density value within the 5 (for summer, spring and autumn) or 10 (for winter) cells upwind of the starting cell of the study domain as the starting background NO₂ value. Please refer to Page 5, Line 144-145 in the revised manuscript.

3. In line 140-145, the authors say that the negative α value reflects the decay of upwind NO₂ pollution along the wind, how come there are still positive α values?

Response: In addition to the upwind NO_x pollution, NO_x emitted from natural sources like soil may also contribute to the background NO₂ value. Soil NO_x emissions are not influenced by wind direction, so on some days, when the background NO₂ value is dominated by soil NO_x rather than the upwind NO_x, we may obtain a positive or zero value of α .

4. The study obtains only 50 out of 365 days of valid data to quantify the NO_x and CO_2 emissions, isn't it too few to estimate the daily variation of NOx and CO2 emissions?

Response: To assure the performance of the model, we filter out the days when cloud fraction is greater than 0.2 and the days when the wind direction shows substantial spatial or temporal variation within the study domain, and obtained 81 clear sky days with full satellite NO₂ coverage within our study period. Then we excluded the days with fluctuating wind direction (if wind direction changes more than 45 degrees within 4 hours before TROPOMI overpass or over the study domain), which would lead to a less good correlation between modeled and observed NO₂ line density (R less than 0.85). Finally, we obtain 50 out of the 365 days with reliable NO_x and CO₂ emissions estimation. The fraction of useful days is comparable to what Lorente et al. (2019) obtained for Paris, which is 27 days in 5 months. These 50 days covers at least 2 days for each month (except for December 2019). For 2019, it includes 9 workdays, 3 weekend days and 2

holiday days, which are enough to investigate the 'weekend reduction effect' and 'holiday reduction effect' in NO_x emissions. It also covers 12 days across the lockdown period and 24 days after that, allowing us to monitor the large reduction and recovery of NO_x and CO₂ emissions from Wuhan due to the COVID lockdown. Therefore, these 50 days provide useful information to investigate the temporal emission patterns of NO_x and CO₂ from Wuhan and help to monitor the effectiveness of emission reductions in large urban centers. We have added this discussion in the revised manuscript in Page 16-17, Line 363-371.

5. Is there a difference in the overpass time of the TROPOMI and OCO-2 satellites? And how is this considered in the study?

Response: Both the TROPOMI and OCO-2 satellites overpasses at 13:00 – 13:30 local time, the difference is small, so we didn't consider this difference in the study.

6. According to Fig. S1, the predicted NO_x emission pattern is 'smoother' compared to the bottom-up emissions, do the authors think about the reason?

Response: We agree to the reviewer that the predicted NO_x emission pattern is smoother compared to the bottom-up emissions. It is caused by two reasons. First, the spatial resolution of the bottom-up emission inventories is $1 \text{ km} \times 1 \text{ km}$, while it is $5 \text{ km} \times 5 \text{ km}$ for the estimated emissions. Second, the spatial pattern of the bottom-up emission inventories is used as a first guess for the estimated emissions, but we let it shift along with the wind during the fitting, and the final predicted NO_x emission pattern is determined by the mean of all the valid days. This is discussed further in the Section 1 of the revised supplementary material.

7. Fig. S4 shows that when the study domain is smaller, the estimated NOx lifetime is longer, how come?

Response: In the robustness test with respect to the area of study domain, we chose a smaller domain which encompasses the central part of Wuhan. The NO₂ column density within this smaller domain is higher than outside. The OH radial is the major oxidizing agent to convert primary pollutants to secondary ones in the atmosphere, so the

concentration of OH radical concentration is lower inside because of titration (Tan et

al., 2018; Lorente et al., 2019). Consequently, the lifetime of NO_x inside the smaller domain will be longer. We have added this information in the revised supplementary material in Section 4.

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Reviewer #2:

This study presents a new top-down superposition column model to estimate daily NOx and CO2 emissions from the largest city Wuhan in central China. It gives a very detailed description of this model, and the application to Wuhan clearly demonstrates the promising future. Overall, I think this manuscript is well structured, and the topic is suitable for ACP.

Major comments.

I think the authors need to clearly describe their way of calculating the uncertainty of their estimates. According to Section 2.6, large uncertainties are attached to the parameters used by the model (20% for OH, 10% of NOx/NO2, 30% for CO2-to-NOx, 15% of NOx lifetime). But the reported uncertainty from the text and Figure 3 is much lower (probably less than 10%). Could the authors specify how they calculate the full uncertainty? Have they used a Monte Carlo (or a similar) method to account for the uncertainty of each sub-process together? **Response: Thank you for this suggestion. We ran a test by randomly choosing parameter values within their uncertainty ranges for 20 times to predict an ensemble of NO_x and CO₂ emission outcomes is regarded as the uncertainty on NO_x and CO₂ emission caused by uncertainties in the corresponding parameters. The results in terms of uncertainties on NO_x and CO₂ emissions are listed below:**

factor	uncertainty	Influence on NO _x and (or) CO ₂ emissions
Satellite NO ₂ retrieval	±20 %	20 %
OH concentration	±20 %	3 %
NO_2/NO_x ratio	±10 %	8 %
Wind field	±20 %	17 %
CO ₂ -to-NO _x emission ratio	±30 %	30 %

The areas of the study domain may also lead to uncertainty in NO_x and CO_2 emissions. We have narrowed down our study domain to check the sensitivity of our method to the chosen study domain (see Fig. S7). In the test, the study domain is narrowed down to 84 km in diameter, and, as expected, the result turns out to be structurally different from that with the 186 km diameter domain, for the mean OH concentration is lower in the city center, leading to longer fitted NO_x lifetime. However, the change in fitted NO_x lifetime and NO_x emission is less than $\pm 15\%$. So we give an 15% uncertainty in NO_x emission estimation caused by the size in the area of the study domain. All the uncertainty factors and their influence on NO_x and CO₂ emission estimation are listed in Table S2.

Finally, considering that all these parameters are independent from each other, we use the root mean square sum of the contributions to represent the overall uncertainty estimation, which we quantify for NO_x emission on a single day at ~31 %, and for CO_2 emission at ~43 %. We updated our revised manuscript accordingly on Page 7-8, Line 192-202, Fig. 3 and Fig. 6 in the revised manuscript and Table S2 in the revised supplementary material.

Minor comments.

Line 21. Please specify the uncertainty. I believe the uncertainty of bottom-up inventories should be much larger than 3%.

Response: We agree. Differences between emissions from two inventories for the year 2019 were <3%, but the uncertainty is arguably larger than that. We rephrased the sentences: 'We estimated daily NO_x and CO₂ emissions from Wuhan between September 2019 to October 2020 with uncertainties of 31% and 43% respectively. Our estimated NO_x and CO₂ emissions are verified against bottom-up inventories with small mutual deviations (< 3 % for 2019 mean, ranging from -20 % to 48 % on a daily basis).'

Line 77. The retrieval methods can considerably affect the column concentrations. So it is better if the authors can comment on the related effects if using a newer version of TROPOMI data.

Response: This is exactly why we used two contemporary versions of the operational TROPOMI NO₂ products (v1.3 and v2.3.1) and evaluated the impact of the retrieval version on our emission estimates. Improved (residual) cloud pressures correct the low bias of v1.3 data compared to OMI and ground-based measurements over east China

(Wang et al., 2020; Liu et al., 2020). In addition, an improved treatment for the surface albedo increases the columns for cloud-free scenes (Van Geffen et al., 2022). Compared to the earlier version, the v2.3.1 dataset has 10-40 % higher tropospheric NO₂ columns over polluted scenes due to the improved cloud retrieval and other algorithm updates (Van Geffen et al., 2020; Riess et al., 2022). Over Wuhan we find an average increase in tropospheric NO₂ of about 25%, but there are also differences between the two versions in terms of spatial and temporal distribution (Fig. S2). According to Fig. S2, the increase in v2.3.1 is much stronger over a polluted area (city center) and polluted period (9 September and 3 October 2019). The estimated NO_x lifetime and emissions from the two datasets for the whole study period are presented in Fig. S5. On average, using the **TROPOMI-v1.3** data leads to 13% lower NO_x emissions from Wuhan than using the TROPOMI-v2.3.1 data. The NO_x lifetime estimated from TROPOMI-v1.3 data is 5% shorter than that from TROPOMI-v2.3.1, which may be attributed to the fact that the TROPOMI-v2.3.1 data has a higher ratio between city center to the background. This information is added in the revised manuscript in Page 3 Line 79-91 and Page 9 Line226-231, and the revised Fig. S6.



Figure S6. Estimated NO_x (a) emissions and (b) lifetime over Wuhan during the study period based on the TROPOMI-v1.3 (blue bars) and TROPOMI-v2.3.1 (red bars) datasets. The error bars denote the corresponding uncertainty.

Line 88. Area-weighted sampling?

Response: The satellite data is sampled in the regular longitude-latitude grid at a resolution of 0.05° (lon) $\times 0.05^{\circ}$ (lat), and it is approximately 6 km \times 6 km in the area of each grid. We didn't use an area-weighted sampling method, and a single pixel is only considered when its center falls in the cell. The NO₂ column density in each grid is calculated as the arithmetic mean of all the considered pixels.

Line 105. Please specify the version of GEOS-Chem and give the reference. Also, give the full names of ERA5 and ECMWF

Response: added. We use the version 12.1 of GEOS-Chem model with a horizontal resolution of $0.25^{\circ} \times 0.3125^{\circ}$ (~ $30 \times 37.5 \text{ km}^2$) to provide the a priori guesses for chemical parameters relevant to daytime NO_x. The wind field is from ERA5 (ECMWF Reanalysis v5), the fifth generation ECMWF (European Centre for Medium-Range Weather Forecasts) atmospheric reanalysis of the global climate.

Line 148. The correlation coefficient of pixels along the wind direction is not a very useful metric here because these pixels are not independent of each other. It is better to show the mean bias along with the correlation coefficients.

Response: The reviewer's point is well-taken. The mean bias is now also included in the revised manuscript in Page 6 Line 158-161, and is displayed in the revised Fig. 1. Both the correlation coefficient and the mean bias are useful metrics to evaluate the performance of the superposition model. The correlation coefficient quantifies the success with which the fitting model reproduces the observed line densities, and the mean bias describes the deviation of fitted line densities from the observations. Compared to fitting results with a constant background value, we obtain a better correlation (up to 25%) and lower bias (nearly 50% lower) between fitted and observed NO₂ line densities then when fitting with a linearly changing background value.

Line 342. Please specify the resolutions here.

Response: This sentence is rephrased as "Compared to previous studies, our work shows that satellite measurements can provide detailed information on sub-city scale NO_x and CO_2 emissions on daily basis".

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