2	implications to CCN prediction
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The density of ambient black carbon retrieved by a new method:

Abstract.

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The effective density of black carbon (BC) is a crucial factor relevant to its aging degree that would add uncertainty in evaluating its climate effect. Here, we have developed a new method to retrieve the effective density of internally mixed BC in the atmosphere combining field observations conducted on 15 November -14 December 2016 in urban Beijing with the Köhler theory. The uncertainty of the retrieval method was evaluated within ± 30 %, which is primarily caused by assumptions of the hygroscopic parameter of organics and the fraction of primary organic aerosols in nonhygroscopic or hygroscopic mode. Using the method, we obtain that the ambient internally-mixed BC, accounting for 80±20 % of total BC aerosol particles, is retrieved with campaign mean density of 1.1±0.6 g cm⁻³ during the observed periods. The retrieved result is comparable with that reported in the literatures. By applying a lower (0.14 g cm⁻³) and upper (2.1 g cm⁻³) limit of the retrieved BC density in cloud condensation nuclei (CCN) number concentrations (N_{CCN}) estimation, we derived that neglect of such variations in BC density led to an uncertainty of -28 %~11 % in predicting $N_{\rm CCN}$ at supersaturations of 0.23 % and 0.40 %. We also find that the $N_{\rm CCN}$ is more sensitive to the variations of BC density when it is <1.0 g cm⁻³. This illustrates a necessity of accounting for the effect of BC density on CCN activity closer to source regions where the BC particles are mostly freshly emitted. The CCN closure achieves when introducing the retrieved real-time BC density and mixing state. This study provides a unique way of utilizing field measurements to infer ambient BC density and

highlights the importance of applying variable BC density values in models when predicting CCN and assessing its relevant climate effect.

1 Introduction

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Black carbon (BC) aerosols, as the major absorber of solar radiation, play a vital role in energy budget and climate of the earth-atmosphere system by affecting the radiative forcing and cloud properties (Flanner et al., 2007; Ramanathan and Carmichael, 2008). The light-absorbing capability induced by BC is related to its density and morphology (Zhang et al., 2008; Rissler et al., 2014), which can be modified after mixing with other atmospheric aerosol particles (Khalizov et al., 2009; Xue et al., 2009). Changes in its physicochemical properties would also regulate its ability to serve as cloud condensation nuclei (CCN) and further indirectly affect the radiative balance by affecting the clouds process (Yuan et al., 2008; Wang et al., 2011). Owing to the complex evolution of the mixing state, density and morphology of BC, the contribution of BC particles to CCN budgets is still not well understood. BC particles, with diesel vehicles, industrial and residential coal combustion as major sources, are ubiquitous in urban environments (Bond et al., 2013; Dameto et al., 2017; Li et al., 2017; Liu et al., 2019a). The mixing state of BC describes the distribution of the bare BC and coating masteries among the aerosol population. Typically, freshly generated BC exists in the form of chain aggregates and initially uncoated, which is known as externally mixed BC (Ex-BC). When the BC particles were emitted, they generally mix with other materials by condensation, coagulation, and other processes (Riemer et al., 2004; Zhang et al., 2008; Liu et al., 2013; Zhang et al., 2020a), forming the internally mixed BC (In-BC) particles consisting of BC core and other chemical components (Cheng et al., 2006; Zhang et al., 2016). The BC structure would be more compact with regular shapes (Pagels et al., 2009; Zhang et al., 2008; Wang et al., 2017), and the effective density of internally mixed BC are changed accordingly with the reconstruction (Liu et al., 2019b). The density and morphology of BC particles are closely related to its sources, mobility size, coating thickness, coating material and its chemical composition (Zhang et al., 2008; Pagels et al., 2009; Peng et al., 2016; Zhang et al., 2022). A wide range of BC density has been reported in previous studies (Lide 1992; Mcmurry et al., 2002; Park et al., 2004; Kiselev et al., 2010). Recent field measurements have indicated that the average BC density is ~1.2 g cm⁻³ in the ambient atmosphere (Zhang et al., 2016). Field measurements have also indicated that a considerable fraction of externally mixed/uncoated BC exists (Clarke et al., 2004; Cheng et al., 2012), although a higher proportion of internally mixed/aged BC particles in the ambient atmosphere were observed (Schwarz et al., 2008; Massoli et al., 2015; Chen et al., 2020). In climate models, the BC was generally assumed completely internally-mixed and treated to have a void-free spherical structure and a density value of 1.8 g cm⁻³ (Bond et al., 2013). This may lead to bias in estimating the climate effect driven by BC.

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Previous study based on a case study show that when the aging degree of ambient particles is low, the BC density (~1.8 g cm⁻³) under the spherical assumption will lead to the overestimation of particle hygroscopicity by 40-50 % and the overestimation can

be explained almost 100 % using the effective density of fresh BC (\sim 0.45 g cm⁻³) (Fan et al. 2020). This indicates the importance of using reasonable BC density values in the calculation of particle hygroscopicity. In addition, when estimating the CCN number concentration, a significant bias of -35 % $\sim +20$ % was found due to the assumption of particle mixing state (Ren et al., 2018). However, these studies have not yet accounted for such impact of BC density and mixing state on CCN prediction due to lack of real time measurement data.

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The mixing state and the density of BC particles are usually directly measured by several techniques, such as an integrated system of a volatility tandem differential mobility analyzer and a single particle soot photometer (VTDMA-SP2) (Zhang et al., 2016), or a differential mobility analyzer with a SP2 (DMA-SP2) (Olfert et al., 2007; Rissler et al., 2014; Wu et al., 2019), and a differential mobility analyzer-centrifugal particle analyzer-single-particle soot photometer (DMA-CPMA-SP2) system (Liu et al., 2019b; Yu et al., 2020), etc. However, such techniques or measurements are not available in many previously conducted filed campaigns. In this study, we develop a novel method for retrieving the mixing state and effective density of ambient BC particles by combining field measured hygroscopic growth factor and aerosol chemical composition and Köhler theory (Petters and Kreidenweis, 2007). The uncertainty of the new retrieval method was evaluated. The retrieved results were also compared and validated with existing observations. In addition, the effect of BC density and mixing state on prediction of CCN number concentrations is further evaluated through a sensitivity and closure test by accounting for the retrieved real-time variations of BC

density and mixing state.

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2 Field measurements and methodology

2.1 Field measurements

Measurements in this study were conducted from 15 November to 14 December 2016 at a typical urban site of Beijing (39.97°N, 116.37°E, 49 m above sea level). The site locates at the Institute of Atmospheric Physics, Chinese Academy of Sciences, which is mainly influenced by the surrounding cooking, road traffic and residential coal burning emissions during the home heating periods (Sun et al., 2016). The detailed information about the sampling site was presented in previous studies (Sun et al., 2015; Zhang et al., 2019). The number concentration of condensation nuclei (CN) at each size was measured by a scanning mobility particle sizer, which is equipped with a differential mobility analyzer (DMA; model 3081, TSI) and a condensation particle counter (CPC; model 3772, TSI). Subsequently, the mono-dispersed particles were introduced into a Droplet Measurement Technologies CCN counter (CCNc, DMT; Lance et al., 2006) to measure CCN number concentration. A hygroscopic tandem differential mobility analyzer (HTDMA) system was used to measure the hygroscopic growth factor (Gf) (Tan er al., 2013). Here, four diameters of 40, 80, 110, 150, and 200 nm are selected in the campaign. Gf is defined as the ratio of the mobility diameter at the given RH to the dry diameter (Petters and Kreidenweis, 2007). The nonrefractory submicron aerosol chemical composition was measured by an Aerodyne highresolution time-of-flight aerosol mass spectrometer (HR-AMS; Xu et al., 2019),

including sulfate, nitrate, ammonium, chloride, and organics. Two factors, including a non-hygroscopic primary organic aerosol (POA) and hygroscopic secondary organic aerosol (SOA) were classified by positive matrix factorization (PMF) with PMF algorithm (v4.2) method (Paatero and Tapper, 1994) and followed the procedures reported in Ulbrich et al. (2009). The refractory black carbon mass loading was measured by an aethalometer (model AE33, Magee Scientific Corporation). Both the nonrefractory materials and BC mass concentration were measured with diameters < 1.0 μ m. The detailed description of the instrument operation and data process have been described in details elsewhere (Ren et al., 2018; Xu et al., 2019; Zhang et al., 2019; Fan et al., 2020).

2.2 Retrieving the mixing state and density of BC

2.2.1 Retrieving the mixing state of BC

The Gf probability distribution function (Gf-PDF) for a specified diameter can be retrieved firstly based on the TDMAinv algorithm (Gysel et al., 2009). The κ -PDF can be further calculated based on the Gf-PDF (Fan et al., 2020). Size-resolved κ is derived using κ -Köhler theory based on hygroscopic growth factor (Gf) (Petters and Kreidenweis, 2007),

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$$\kappa_{gf} = (Gf^3 - 1) \cdot \left[\frac{1}{RH} \exp\left(\frac{4\sigma_{s/a}M_w}{RT\rho_w D_d Gf}\right) - 1 \right]$$
 (1)

where Gf is hygroscopic growth factor, RH is the relative humidity in the HTDMA (90 %), D_d is the dry diameter, $\sigma_{s/a}$ is assumed to be the surface tension of pure water,

R is the universal gas constant, T is the temperature, $M_{\rm w}$ and $\rho_{\rm w}$ is the molecular mass, 153 and the density of water, respectively.

The κ -PDF patterns of particles in different sizes always present two modes: nearly hydrophobic (NH) mode with $\kappa_{\rm gf} \leq 0.1$ and more hygroscopic (MH) mode with $\kappa_{\rm gf} > 0.1$ (Fig. S1). Firstly, based on the κ -PDF patterns, the number fraction (NF) of the total nearly hydrophobic group with the boundary of [0, 0.1] was calculated according to the following equation:

$$NF = \int_0^{0.1} c(\kappa, D_p) d\kappa \tag{2}$$

here, the κ -PDF, represented by c (κ , D_p), was normalized as $\int c(\kappa, D_p) d\kappa = 1$, where κ can be replaced by $\kappa_{\rm gf}$, D_p is the selected electrical mobility diameter in the campaign. The nearly hydrophobic mode consists of both externally mixed POA (Ex-POA or bare POA) and externally mixed BC (Ex-BC). Since the number fraction of the nearly-hydrophobic POA would change with the emission and aging processes, in this study, we have applied different values for the number fractions of hydrophobic POA (NH-POA) under clean (91 %), moderately polluted (70 %), and heavily polluted conditions (31 %) by referring the literature (Liu et al., 2021a), as shown in Fig. S2. The number concentration of Ex-BC was then calculated using the total number fraction of NH

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$$N_{POA-containing} = N_{total} \times NF_{POA-containing}$$
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$$N_{bare-POA} = N_{POA-containing} \times NF_{bare-POA}$$
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$$N_{Ex-BC} = N_{NH} - N_{bare-POA}$$
(3)

mode minus the number of NH-POA.

where $N_{POA-containing}$ and $NF_{POA-containing}$ are the number concentration and fraction of

- POA-containing particles, N_{total} is the total number concentration, $N_{\text{bare-POA}}$ and $NF_{\text{bare-POA}}$ are the number concentration and fraction of bare POA particles, and N_{NH} is the number of nearly- hydrophobic group.
- The number size distribution of the externally mixed BC (n_{Ex-BC} (log D_p)) can be calculated based on the particle number size distribution (PNSD) and the number fraction of the hydrophobic mode of BC (NF_{Ex-BC}) as follows:

$$n_{\text{Ex-BC}}(\log Dp) = NF_{\text{Ex-BC}} \times n (\log Dp)$$
 (4)

- 181 where n (log D_p) is the function of the aerosol number size distribution, D_p is the mobility diameter.
- By assuming that the particles are spherical (Rader and McMurry, 1986), the mass size distribution of Ex-BC ($M_{\rm Ex-BC}$) was obtained as follows:

$$M_{\text{Ex-BC}}(\log Dp) = \frac{\pi}{6} D_p^{\ 3} \rho n_{\text{Ex-BC}}(\log Dp)$$
 (5)

where D_p is the mobility diameter, ρ is the effective density of Ex-BC, and $n_{\text{Ex-BC}}$ (log D_p) is the function of the number size distribution of Ex-BC, respectively. By reviewing and summarizing the existing results about, we show that typical values of density for the freshly emitted or externally mixed BC observed in the winter of urban Beijing or North China Plain spans over 0.14-0.50 g cm⁻³, with mean of ~0.40±0.10 g cm⁻³ (Fig. S3), in the size range of 100 to 300 nm, where the mass concentration of externally mixed BC accounted for a large proportion in urban Beijing (Geller et al., 2006; Peng et al., 2016, 2017; Wu et al., 2019; Liu et al., 2020; Zhao et al., 2022). Therefore, an average $\rho_{\text{Ex-BC}}$ of 0.4 g cm⁻³ was used for calculating the mass concentration of externally-mixed BC in our study. The uncertainty analysis exhibits that the variations

of the $\rho_{\text{Ex-BC}}$ could lead to an average deviation of ± 10 % in the calculating In-BC density (Fig. 3e) by increasing the $\rho_{\text{Ex-BC}}$ from 0.1 to 0.6 g cm⁻³, showing a small impact on the retrieved result. Uncertainty analyses due to the variations of $\rho_{\text{Ex-BC}}$ were given in section 2.3.

The mass size distribution of Ex-BC was fit using the log-normal distribution as shown in Fig. S4 (Wu et al., 2017; Liu et al., 2019a; Zhao et al., 2022). Thus, the bulk mass concentration of Ex-BC can be calculated from the integration of the mass size distribution:

$$m_{\text{Ex-BC}} = \int_{D_{start}}^{D_{end}} M_{\text{Ex-BC}}(\log D_p) \ d \log(D_p)$$
 (6)

$$m_{\text{In-BC}} = m_{\text{BC}} - m_{\text{Ex-BC}} \tag{7}$$

where D_{start} and D_{end} are the lower and upper size limit, $M_{\text{Ex-BC}}$ (log D_{p}) is the function of the Ex-BC mass size distribution. We then obtained the bulk mass concentration of internally mixed BC ($m_{\text{In-BC}}$) by subtracting $m_{\text{Ex-BC}}$ from the bulk BC mass concentration measured by AE33 in equation 7.

2.2.2 Retrieving the density of BC

For retrieval of the density of BC, the principal idea is to use the measured κ_{gf} to calculate the density of BC based on the Zdanovskii–Stokes–Robinson (ZSR) mixing rule (Stokes and Robinson, 1966; Zdanovskii, 1948) with the chemical composition measured by AMS (Petters & Kreidenweis, 2007). In the retrieval, several aspects are concerned. First, since the ZSR rule is to assume the aerosol particles are internally mixed, the κ_{gf} value of the more MH mode (κ_{gf-MH}) is thus applied for retrieving the

density of internally mixed BC. Second, since the size distribution of BC number concentration is usually with peaks between 100 and 200 nm (Liu et al., 2019a; Yu et al., 2020; Zhao et al., 2022), the $\kappa_{\text{ef-MH}}$ value of particles in accumulation mode was averaged and applied for the retrieval. Previous studies showed an independence of $\kappa_{\rm gf}$ _{MH} on particle size when the $D_p > 100$ nm during the campaign period (Fan et al., 2020). Therefore, the average of κ_{gf-MH} in accumulation mode is reasonable for the determination of the In-BC density. In addition, because the inversion including measurements from HTDMA and HR-AMS, a total mass closure of the measured aerosol particles was conducted between the two techniques by comparing the mass concentration of PM1 and the results are well consistent (Fig. S5). The density of internally mixed BC (In-BC), $\rho_{\text{In-BC}}$ is then derived from the following equations:

$$\kappa_{gf-MH} = \kappa_{chem} = \sum_{i} \varepsilon_{i} \kappa_{i} = \frac{v_{inorg}}{v_{total}} \kappa_{inorg} + \frac{v_{SOA}}{v_{total}} \kappa_{SOA} + \frac{v_{In-POA}}{v_{total}} \kappa_{POA} + \frac{v_{In-BC}}{v_{total}} \kappa_{BC}$$
(8)

where $\kappa_{\text{gf-MH}}$ is the hygroscopic parameter of the more hygroscopic (MH) mode, κ_{chem} is the hygroscopic parameter of aerosol particles in the mixed composition and can be calculated based on chemical volume fractions using a simple rule (Stokes and Robinson, 1966; Petters & Kreidenweis, 2007), κ_i is the hygroscopic parameter of each pure composition and ε_i is the volume faction of the individual components in the internal-mixed particle. v_{inorg} , v_{SOA} and $v_{\text{In-POA}}$ are the volume of the inorganic, SOA and internally mixed POA species, and can be calculated as follows: $v_{inorg} = \frac{m_{inorg}}{\rho_{inorg}}$, $v_{SOA} = \frac{m_{SOA}}{\rho_{SOA}}$, and $v_{In-POA} = \frac{m_{In-POA}}{\rho_{POA}}$. v_{total} is the total volume of all the species and can be written as $v_{total} = \frac{m_{inorg}}{\rho_{inorg}} + \frac{m_{SOA}}{\rho_{SOA}} + \frac{m_{In-POA}}{\rho_{POA}} + \frac{m_{In-BC}}{\rho_{In-BC}}$. In equation (8), κ_{BC} and κ_{POA} are assumed to be 0. Then, the ρ_{In-BC} can be calculated based on its mass concentration and

volume as follows:

$$\rho_{In-BC} = \frac{\frac{m_{In-BC}}{\frac{m_{inorg}}{\rho_{IOOTg}} \kappa_{inorg} + \frac{m_{SOA}}{\rho_{SOA}} \kappa_{SOA}}}{\frac{m_{inorg}}{\kappa_{gf-MH}} - \frac{m_{inorg}}{\rho_{inorg}} \frac{m_{SOA}}{\rho_{SOA}} \frac{m_{In-POA}}{\rho_{POA}}})$$
(9)

where, $m_{\text{In-BC}}$ is the mass concentration of internally mixed BC, m_{inorg} and m_{SOA} are the mass concentrations of the inorganic species and SOA, which are measured by the AMS. $m_{\text{In-POA}}$ is the mass concentrations of internally mixed POA and can be calculated subtracting the mass fraction of NH-POA from the total mass concentrations of POA. ρ_{inorg} , ρ_{SOA} and ρ_{POA} are the density of the inorganic species, SOA and POA. Since the AMS measures the concentrations of the organic and inorganic ions, including SO_4^{2-} , NO_3^- , NH_4^+ , Cl^- . Here inorganic species were derived by applying a simplified ion pairing scheme (Gysel et al., 2007) to convert mass concentrations of ions to the inorganic salts as follows:

$$251 n_{\rm NH_4NO_3} = n_{\rm NO_3^-}$$

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$$n_{\text{NH}_4\text{HSO}_4} = \min(2n_{\text{SO}_4^{2-}} - n_{\text{NH}_4^+} + n_{\text{NO}_3^-}, n_{\text{NH}_4^+} - n_{\text{NO}_3^-})$$

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$$n_{(NH_4)_2SO_4} = \max(n_{NH_4^+} - n_{NO_3^-} - n_{SO_4^{2-}}, 0)$$

$$250 n_{\text{H}_2\text{SO}_4} = \max(0, n_{\text{SO}_4^{2-}} - n_{\text{NH}_4^+} + n_{\text{NO}_3^-}) (10)$$

where n represents the number of moles, then the mass concentrations were obtained by the number of moles times the molar mass of each inorganic salts. Because the maximum value of the $n_{\rm H_2SO_4}$ was zero in this campaign. Three inorganic salts including NH₄HSO₄, (NH₄)₂SO₄, and NH₄NO₃ were applied in our study. The densities for inorganic salts were taken from previous studies (Gysel et al., 2007; Wu et al., 2016). Here the densities for three inorganics are 1.78, 1.77 and 1.72 g cm⁻³, respectively. By summarizing the previous studies (Gysel et al., 2007; Dinar et al., 2006), 1.4 g cm⁻³ was

selected as the density of SOA (ρ_{SOA}). The density of POA (ρ_{POA}) is assumed to be 1.0 g cm⁻³ for urban environments, which is similar to the lubricating oil (Wu et al., 2016). Considering the cooking organic aerosols represent a high contribution to POA in urban environments, a density of 0.85 g cm⁻³ chosen as the mean density for the rapeseed oil and oleic acid (Reyes-Villegas et al., 2018) was also used to evaluated the result as shown in section 2.3. The values of κ for inorganic components are 0.56 for NH₄HSO₄, 0.48 for (NH₄)₂SO₄ and 0.58 for NH₄NO₃, along with the best-fit values for the three inorganic salts (Petters & Kreidenweis, 2007 and Gunthe et al., 2009). The κ_{SOA} is assumed to be 0.15 according to the field studies in urban areas (Chang et al., 2010; Kawana et al., 2016).

Note that, this method fails to retrieve the BC density when organics account for a large fraction (>60 %). This is because that a higher fraction of OA usually corresponds to lower total volume of all the species (Fig. S6), yielding negative values for $v_{\text{In-BC}}$ introduced in equation 11. As a result, 61 % of the data observed during the campaign were valid for calculating the BC density.

$$v_{In-BC} = \frac{v_{inorg} \kappa_{inorg} + v_{SOA} \kappa_{SOA}}{\kappa_{gf-MH}} - v_{inorg} - v_{SOA} - v_{In-POA}$$
 (11)

Similarly, the bulk density of BC ($\rho_{bulk-BC}$) is calculated with the same method as that for calculating the ρ_{In-BC} . When calculating the $\rho_{bulk-BC}$, the bulk κ_{gf} value measured by HTDMA is applied with the assumption of all the aerosol particles are internally mixed.

2.3 Uncertainties and limitations

For the retrieval, the assumptions on the values of κ_{SOA} , ρ_{POA} , ρ_{SOA} and ρ_{Ex-BC} as
well as the fraction of primary organic aerosols in non-hygroscopic or hygroscopic
mode would add uncertainty in the inferred values of ambient internally mixed BC
density. For example, the freshly emitted POA particles might consistently be coated
with the secondary particles during the aging process, resulting in changes of the $NF_{\rm NH-}$
$_{POA}$. However, a real-time variation of the NF_{NH-POA} is not yet available due to the lack
of such measurements data. Applying the rough fractions of hydrophobic POA only
under three different atmospheric conditions could still cause uncertainties. Also, the
densities of POA and SOA may differ due to their precursors, emission sources and the
formation mechanisms in ambient atmosphere (Alfarra et al., 2006; Reyes-Villegas et
al., 2018). And the density of Ex-BC is generally characterized by the morphology and
size (Wu et al., 2019). In addition, the value of κ_{SOA} spans largely due to the variability
in the emissions of gas precursors and formation processes under different atmospheric
conditions (Zhang et al., 2015; Liu et al., 2021b). Therefore, we examined the
sensitivities of In-BC density to the variations of these factors, as exhibited in Fig. 1
and Fig.2.
The figures show that the In-BC density gradually decreases with the increment of
the $NF_{ m NH-POA}$, implying the higher fraction of bare POA particles correspond to the early
aging stage of aerosol particles. With increase of κ_{SOA} , the In-BC density is generally

of assumptions of κ_{SOA} on the retrieved BC density. In addition, the In-BC density

reduced, but with small fluctuations (Fig.1a, Fig. 2b). This suggests a complex impact

decreases very slightly as $\rho_{\text{Ex-BC}}$ increases (Fig. 2e), suggesting applying a larger $\rho_{\text{Ex-BC}}$ would derive smaller values for In-BC density. The In-BC density is insensitive to the changes of the density of POA and SOA, showing an almost negligible effect on the retrieved results (Fig. 2c and d).

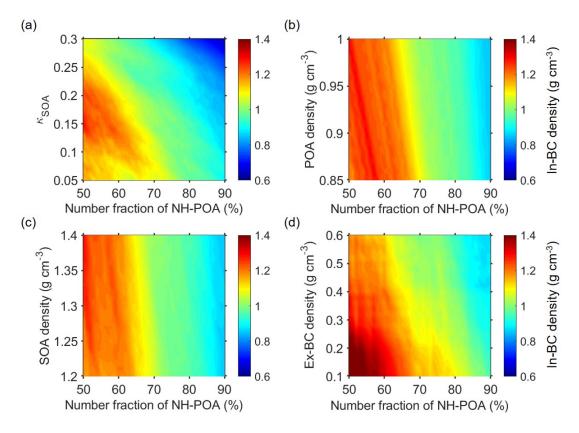


Figure 1. Sensitivities of In-BC density to the variations in the number fraction of nearly hydrophobic (NH) POA and hygroscopic parameter of OA (k_{SOA}) (a), POA density (b), SOA density (c) and the externally mixed BC density (d).

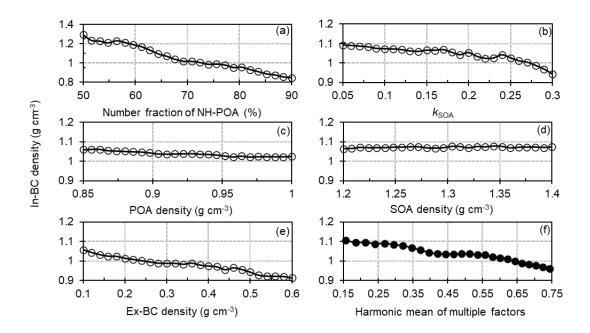


Figure 2. Sensitivity of the In-BC density to variations in the number fraction of nearly hydrophobic (NH) POA (a), the hygroscopic parameter of SOA (b), the POA density (c), the SOA density (d), the externally mixed BC density (e) and the harmonic mean of multiple factors (f).

The uncertainty analysis shows that, by comparing the results based on the mean fractions of the $NF_{\rm NH-POA}$ with a typical atmospheric observed range of 50-90 % for the $NF_{\rm NH-POA}$ (Liu et al., 2021a), we show that the assumption on $NF_{\rm NH-POA}$ can lead to relative deviations (uncertainty) of -17 %-+27 % for the retrieved BC density (Fig.3a). In addition, unlike inorganics (eg., NH₄HSO₄, (NH₄)₂SO₄ and NH₄NO₃), which the hygroscopicity has been already well-understood (Petters and Kreidenweis, 2007), the hygroscopicity of organic species varies largely due to the complexity in organic aerosol constituents. Therefore, the assumption of the values of $\kappa_{\rm SOA}$ will add the uncertainty in the calculation of BC density. Previous studies have suggested that the organics has a wide range of κ values ranging from 0.05 to 0.3 (Jimenez et al., 2009;

Mei et al., 2013). Thus, the sensitivity test has also been done to examine the effect due to changes in κ_{SOA} on calculating the density of BC (Fig. 1a). The result shows that the assumption of κ_{SOA} value can cause an average relative deviation of -10 %-+3 % in calculating the density of In-BC (Fig. 3b).

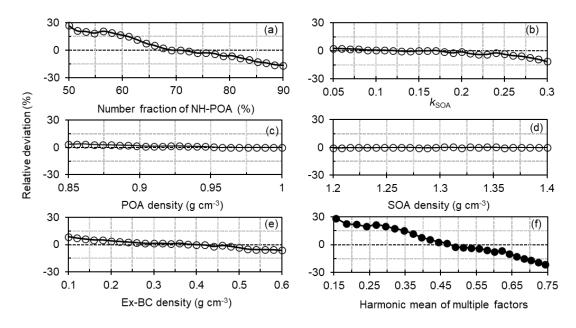


Figure 3. Relative deviations of the number fraction of nearly hydrophobic (NH) POA to the In-BC density (a), the hygroscopic parameter of OA to the In-BC density (b), the POA density to the In-BC density (c), the SOA density to the In-BC density (d), the externally-mixed BC density to In-BC density (e) and the combined deviations based on multiple factors mentioned above (f).

However, the sensitivity test shows that the impact of both the ρ_{POA} and ρ_{SOA} variations on the BC density estimation is very small or even negligible (Fig. 1b, c). By varying the ρ_{POA} from 0.85 to 1.0 g cm⁻³ and the ρ_{SOA} from 1.2 to 1.4 g cm⁻³ according to the literatures (Noureddini et al., 1992; Alfarra et al., 2006; Reyes-Villegas et al., 2018), the retrieval uncertainties in the BC density are within ± 5 % and ± 1 % respectively (Fig. 3c, d). For ρ_{Ex-BC} , it exhibits that the evolution of the ρ_{Ex-BC} could lead

to an average deviation of ± 10 % in calculating In-BC density (Fig. 3e) when increasing the values of $\rho_{\text{Ex-BC}}$ from 0.1 to 0.6 g cm⁻³, which represents a typical range in ambient atmosphere (Wu et al., 2019; Liu et al., 2020). A combined uncertainty (δ) caused by the multiple factors (δ_i), which is calculated by equation 12, is -21 %-+29 % as shown in Fig. 3f.

$$\delta = \sqrt{\sum_{i=1}^{n} \delta_i^2} \tag{12}$$

3 Results and Discussion

3.1 Retrieved mixing state and density of BC: comparison and validation

Figure 4a shows retrieved time series of the mixing state of ambient BC during the campaign. Large temporal variations of the mass fraction of internally and externally mixed BC are presented during the observed period at the sites. The temporal changes should be related to the atmospheric aging process or diurnal variations of emissions (Liu et al., 2019a; Fan et al., 2020). Statistically, the average mass fraction of externally and internally mixed BC is 20±18 % and 80±20 % respectively, showing that most of the BC particles were aged and internally mixed with other components. Previous studies at urban sites have shown that the co-existence of the externally mixed BC in the ambient atmosphere (Schwarz et al., 2008; Cheng et al., 2012; Chen et al., 2020) due to continuous combustion processes (e.g., vehicle exhaust and residential sector) (Wang et al., 2017; Liu et al., 2019a). Our results are basically comparable with those

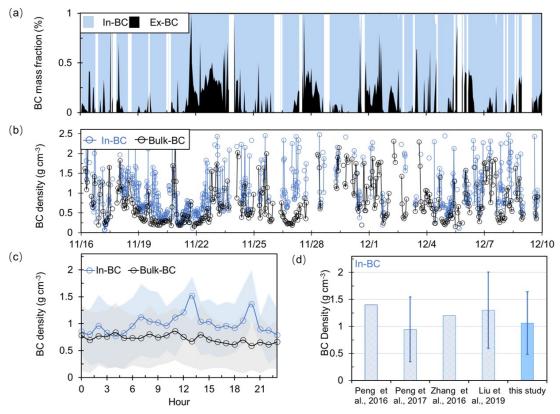


Figure 4. (a) Time series of the mass fraction of the retrieved internal- and external-mixed BC; (b) Time series of the retrieved density of the bulk and internal- mixed BC (In-BC); (c) Diurnal variation of the retrieved density of bulk and In-BC; (d) Comparison of the results of the derived In-BC density in this study with that reported in literatures.

directly measured or indirectly retrieved previously reported results. For example, Chen et al., 2020 found that the mass fraction of internally mixed BC particles was nearly to be ~80–90 % in summer of Beijing based on VTDMA measurements. Liu et al. (2020), using a tandem system with an aerodynamic aerosol classifier and SP2, reported that the mass fraction of internally BC-containing particles would increase with increasing

size and reach ~70 % in Beijing. Overall, the mass fraction obtained in our study is

comparable with those reported in urban Beijing. Previous studies also displayed that

the significant diversity of the BC mixing state among emission conditions and coating process (Shiraiwa et al., 2008; Pan et al., 2017; Zhang et al., 2020b). Accordingly, the densities of the bulk and internally mixed BC present apparent fluctuations as shown in Fig. 4b, which is significantly affected by the variations of BC emission sources and its rapid aging process. The density of the In-BC during daytime was generally higher than that at night (Fig. 4c). The elevated BC density during daytime is likely due to that the strong photochemical processes promote the aging of BC particles, which resulted in a conversion from uncompacted structure to compact and regular spherical shapes of BC (Qiao et al., 2018; Liu et al., 2019b; Zhou et al., 2022). The lift in BC density around 20:00 LT might indicate that the BC particles would be rapidly coated with the SIA particles and continuously aged in the polluted period due to the heterogeneous reactions of SIA in urban regions (Zhang et al., 2016; Peng et al., 2017). Actually, following the haze evolution, the fraction of nearly hydrophobic group reduced rapidly (Fig. S7). Consequently, the average density of In-BC increased obviously from the clean conditions to the polluted periods (Fig. S8). A slight decrease was observed in the bulk BC density during traffic hours. This is likely associated with the continues emissions (e.g., vehicle exhaust) that lead to uncoated or uncompacted BC particles in this period. The diurnal cycle in In-BC density is consistent with the coating thickness measured by a tandem CPMA-SP2-DMA-SP2 (Liu et al., 2020), demonstrating that the new method can derive the density of ambient BC particles reasonably. Averagely, the bulk and internally mixed BC densities are with campaign averaged values of 0.7±0.5 and 1.1±0.6 g cm⁻³ respectively, which are much less than 1.8 g cm⁻³, implying that the

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BC particles is not a void-free spheres in the urban atmosphere. The results of In-BC density are comparable with that observed at the other sites in North China Plain (NCP) as shown in Fig. 4d, illustrating that the BC effective density retrieved by this method is within the range of field measurements.

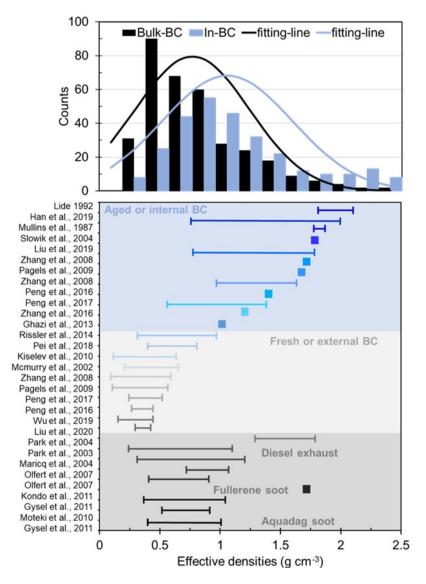


Figure 5. The probability distribution function (PDF) of the retrieved density of bulk and In-BC and the measured density distribution spectrum of BC from different sources reported in literatures.

Based on both field measurements (e.g. Lide 1992; Zhang et al., 2016; Wu et al., 2019; Liu et al., 2019b) and laboratory studies (e.g. McMurry et al., 2002; Park et al.,

2003, 2004; Olfert et al., 2007; Kiselev et al., 2010; Gysel et al., 2011, 2012), the BC density from diverse combustion sources or representing different aging degree has been obtained and ranges widely from 0.14 to 2.1 g cm⁻³, as has been summarized and shown in Fig. 5. Mean probability distribution function (PDF) of the density of bulk and In-BC retrieved by this study is also presented in Fig. 5. It shows that the retrieved density of bulk BC exhibits a dominant mode with a peak value of 0.7 g cm⁻³, which is situated between the typical density range of those externally mixed and internally mixed BC measured previously. For the In-BC, the PDF is with a peak value at 1.1 g cm⁻³, but ranges widely from ~0.5 to 2.5 g cm⁻³, which indicates various morphologies, different aging degree and compositions of ambient BC particles due to the complex impact of multiple local sources and aging processes during the observed period in urban Beijing. Overall, the retrieved values for In-BC fall within the range of typical internal mixed BC reported in the literatures, verifying the reliability of our inversion results.

3.2 Sensitivity of predicted N_{CCN} to changes of BC density

A previous study showed that the use of an inaccurate density value of BC particles would result in large bias in estimating κ of ambient aerosol particles with the ZSR mixing rule (Fan et al., 2020), as would further lead to uncertainties in prediction of $N_{\rm CCN}$ and relevant climate effects. Considering the large variation range of BC density during the campaign, which is closely associated with its morphology or degree of its aging, we further examine the sensitivity of critical supersaturation (S_c), critical

diameter (D_{cut}) and predicted N_{CCN} to variations of BC density (Fig. 6). Here, we use the critical diameter and particle number size distribution to calculate N_{CCN} . The method to derive the critical diameter is based on Köhler theory and ZSR rule.

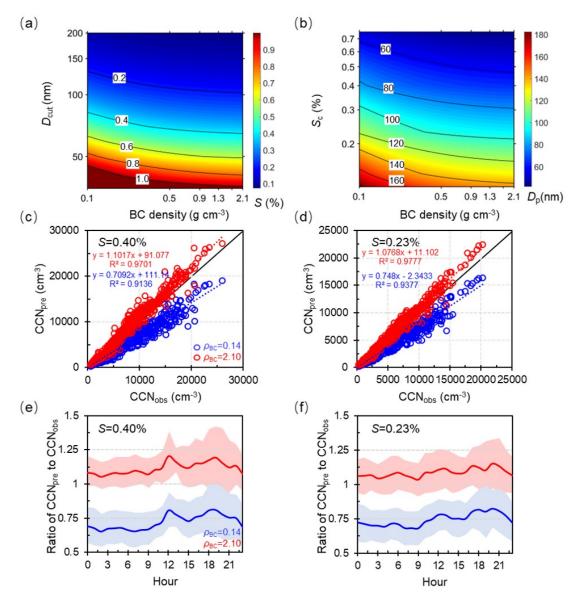


Figure 6. Sensitivity of critical supersaturation (S_c) (a) and diameter (D_{cut}) (b) to the variations in BC density; Predicted N_{CCN} as a function of measured N_{CCN} by varying the density from 0.14 to 2.1 g cm⁻³ at S=0.40 % (c) and S=0.23 % (d), the black solid line is the 1:1 line; Diurnal variations in the ratio of predicted-to-measured N_{CCN} at S=0.40 % (e) and S=0.23 % (f).

The results show that, by varying the value of density from 0.14 to 2.1 g cm⁻³ that represents the lower and upper limit of BC density in the atmosphere, the D_{cut} reduces apparently at a given supersaturation (S) (Fig. 6a), or similarly, the S_c decreases rapidly for a given particle size (Fig. 6b). The results show that the changes of the D_{cut} and S_c are more sensitive when the BC density is below 1.0 g cm⁻³. And the effects on the D_{cut} and S_c both gradually weakened with the increase of BC density. This shows that it is critical to apply more accurate BC density for the aerosol particles with low aging degree in predicting CCN and its climate effect. Accordingly, the ratios of predicted-to-measured N_{CCN} ranged from 0.72 to 1.11 by varying the BC density from 0.14 to 2.1 g cm⁻³ at the typical S of 0.23 % and 0.40 % (Fig. 6c, 6d), showing an estimation uncertainty of -28 %-11 % in N_{CCN} prediction.

The diurnal variations in the ratio of predicted-to-measured $N_{\rm CCN}$ at S=0.40 % and 0.23 % are shown to examine the response of the BC density on $N_{\rm CCN}$ prediction at different time periods (Fig. 6e, 6f). By applying the lower limit of density value of 0.14 g cm⁻³, the prediction is much worse compared to the use of the density of 2.1 g cm⁻³ at nighttime (00:00-06:00 LT), when the latter is much closer to the real density of ambient BC (Fig. 4c). The prediction is improved substantially by applying the value of 0.14 g cm⁻³ during evening rush hours (18:00-20:00 LT), during which the ambient BC particles is disturbed by the traffic emissions (Fig. 4c). And now, the prediction becomes worse by applying the value of 2.1 g cm⁻³, and an obvious overestimation by up to ~40 % is shown. The results further illustrate that it is critical to account for the real-time mixing state and density of BC particles in $N_{\rm CCN}$ prediction, particularly in those regions

with heavy traffic and residential coal emissions.

It should be noted that the assumption of the surface tension of water would overestimate the critical diameter and underpredict CCN number concentration. While the surface tension depression might be more obvious for the small size particles (<60 nm), as the fraction of organics are higher at small particles size (Meng et al., 2014; Cai et al., 2018). Here, in this study, we calculated the critical diameters at supersaturations of 0.40 % and 0.23 %, typical values in cloud, corresponding to larger sizes (> 70 nm and 90 nm) of aerosols. Therefore, the uncertainties from the application of the surface tension of pure water should be negligible (< 10 %). Here, three schemes were assumed to evaluate the effect of BC density and mixing state on prediction of CCN number concentrations. The detailed calculation methods are presented in the supporting information (SI: Methods) or referenced from Ren et al., 2018.

3.3 Using the real-time variations of BC density and mixing state to predict $N_{\rm CCN}$

Figure 7 exhibits the comparisons between predicted and measured $N_{\rm CCN}$ at S of 0.23 % and 0.40 % by accounting for the retrieved real-time variations of BC density and mixing state. It shows that the $N_{\rm CCN}$ can be well predicted with a slope of 1.01 and 1.02 at S of 0.23 % and 0.40 % respectively (Fig. 7a, 7b), only presenting a slight deviation. The slight deviation is primarily due to the fixed value of the density for the externally mixed BC caused by the retrieved method, especially during noontime and evening rush periods (Fig. 7c and 7d).

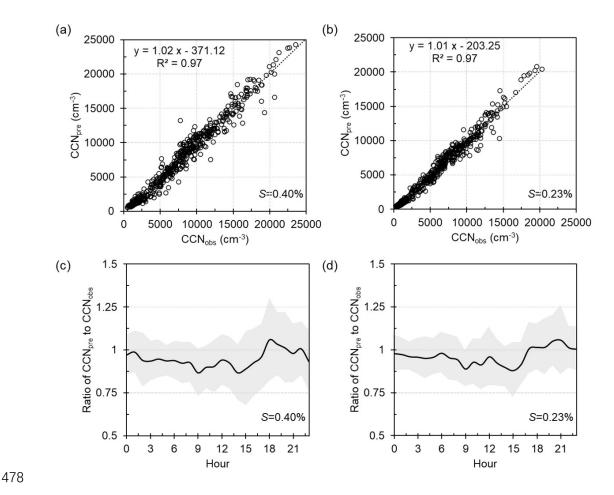


Figure 7. Prediction CCN number concentration using the mixing state and In-BC density derived from HTDMAs at S=0.40 % (a) and S=0.23 % (b). Diurnal variations in the ratio of predicted-to-measured N_{CCN} at S=0.40 % (c) and S=0.23 % (d).

The diurnal variations in the ratio of predicted-to-measured $N_{\rm CCN}$ shows the $N_{\rm CCN}$ can be underestimated by up to 15 % at S=0.40 % during those periods. While, a slightly overrated during the evening traffic hours and nighttime may be due to the underestimation of the number fraction of Ex-BC. Overall, the dependence of the CCN prediction on S is due to the size dependence of κ and mixing state (Zhang et al., 2017; Liu et al., 2020; Xu et al., 2021). The better closure at S=0.23 % is because that the bulk κ of particles is closer to that the critical diameter corresponding to S=0.23 %, with $D_{\rm P}$ of 100-150 nm. Similarly, the effect on CCN prediction induced by the bulk mixing

state would be more critical for smaller particles, corresponding to the critical diameter at high *S*.

Overall, when considering the effective density of BC relevant to its mixing state, the CCN closure achieves. Previous studies have shown that the fresh emitted BC particles may convert from fractal-like aggregates to a compact structure and its density would increase with the aging process (Pagels et al., 2009; Rissler et al., 2014; Peng et al., 2016; Liu et al., 2019b; Zhang et al., 2020a, 2022), but the actual density of In-BC may be lower than 1.8 g cm⁻³ in the ambient atmosphere according to this study. Therefore, the currently applied value represents a density of the void-free structure of BC particles may cause an overestimation in CCN prediction.

4 Conclusions

The mixing state and effective density of BC changed through heterogenous chemistry process and thus would cause uncertainty in evaluating its CCN activity. In this study, we develop a new method to retrieve the mixing state and effective density of ambient BC using field measurements and the Köhler theory. The uncertainty of the new retrieval method was evaluated within ± 30 %, which is primarily caused to assuming the κ_{SOA} and the fraction of primary organic aerosols in non-hygroscopic or hygroscopic mode. The retrieved results show that most of the BC particles were aged and internally mixed with other components, with mean mass fraction of 80 ± 20 %. Averagely, the retrieved densities of the bulk and internal-mixed BC are 0.7 ± 0.5 and 1.1 ± 0.6 g cm⁻³ respectively, but ranges widely from ~ 0.1 to 2.5 g cm⁻³, indicating

various morphologies, different aging degree and compositions of ambient BC particles due to the complex impact of multiple local sources and aging processes during the observed period. The retrieved results are basically comparable with the previous observations in North China Plain.

Further examination shows the $N_{\rm CCN}$ prediction is with uncertainties of -28 %-11 % at the typical S of 0.23 % and 0.40 % by varying the BC density from 0.14 to 2.1 g cm⁻³ that represents the lower and upper limit of ambient BC particles. Moreover, the prediction is found more sensitive to the variability of BC density when it is <1.0 g cm⁻³, suggesting a great significance to account for the effect of BC density for the aerosol particles with low aging degree when evaluating the climate effect. The CCN closure achieves when introducing the retrieved real-time BC density relevant to its mixing state. This work provides a unique way of utilizing field observations to infer ambient BC density and highlights the current assumption of a void-free structure of BC particles in models would cause large uncertainties in CCN prediction and in the relevant climate effect evaluation.

Data availability.

All data needed to evaluate the conclusions in the paper are present in the paper and/or the Supplement. All data used in the study are also available from the corresponding author upon request (zhangfang2021@hit.edu.cn).

Author contributions.

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FZ and JR conceived the conceptual development of the manuscript. JR directed and performed of the experiments with JL, LC, and FZ. JR conducted the data analysis and wrote the draft of the manuscript. All authors edited and commented on the various sections of the manuscript.

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Competing interests.

The contact author has declared that neither they nor their co-authors have any competing interests.

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