

# Characterization of ultrafine particles and the occurrence of new particle formation events in an urban and coastal site of the Mediterranean area

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**Abstract.** In this work, new particle formation events (NPF) occurring at two locations in Southern Italy, the urban background site of Lecce (ECO station) and the coastal site of Lamezia Terme (LMT station), are identified and analysed. The study aims to compare the properties of NPF events at the two sites located 225 km away from each other and characterized by marked differences in terms of emission sources and local weather dynamics. Continuous measurements of particle number size distributions, in the size range from 10 nm to 800 nm, were performed at both sites by a Mobility Particle Size Spectrometer (MPSS). The occurrence of NPF events, observed throughout the study period that lasted five years, produced different results in terms of frequency of occurrence, 25 % of the days at ECO and 9 % at LMT. NPF events showed seasonal patterns, higher frequency during spring and summer at the urban background site, while at the coastal site during the autumn-winter period. Some of these events happened simultaneously at both sites, indicating the occurrence of the nucleation process on a large spatial scale. Cluster analysis of 72 h back-trajectories showed that during the NPF events the two stations were influenced by similar air masses, most of which originated from the North-Western directions. Local meteorological conditions characterized by high pressure, with a prevalence of clear skies, low levels of relative humidity ( $RH < 52\%$ ), and moderate winds ( $3-4\text{ m s}^{-1}$ ) dominated the NPF events at both sites. Notable differences were observed in  $SO_2$  and  $PM_{2.5}$  concentrations and  $H_2SO_4$  proxy levels, resulting in  $\sim 65\%$ ,  $\sim 80\%$ , and  $50\%$  lower at LMT compared to ECO, respectively. It is likely that the lower level of that recognized as one of the main gas precursors involved in the nucleation process, can be responsible for the smaller NPF frequency of occurrence ( $\sim 60\%$  less than ECO) observed in LMT.

## 1 Introduction

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The formation of new particles (NPF), by nucleation of gas-phase species and consecutive growth, is an important atmospheric process that contributes to producing high levels of ultrafine particles (UFP, diameter  $< 100\text{ nm}$ ). Together with primary

sources, natural and anthropogenic, the nucleation represents a significant source of secondary ultrafine aerosol particles and cloud condensation nuclei (Merikanto et al., 2009; Yu et al., 2020), accounting for about 50 % of the production of particle number concentration on a global scale. Due to their size, high number concentration, and chemical composition, these particles have profound implications on the environment, climate, and public health (Zhang et al., 2015; Sartelet et al., 2022) which emphasize their relevance.

The formation of new particles was observed and investigated in various geographic locations, under different atmospheric and environmental conditions, from which common characteristics and different peculiarities emerged (Baalbaki et al., 2021; Jokinen et al., 2022; Yadav et al., 2021; Franco et al., 2022).

The new particle formation can occur on a local scale, typically characterized by an intense burst (strong intensity) of secondary particle formation of short duration and without subsequent growth (Dai et al., 2017), usually related to local anthropogenic emissions (Hussein et al., 2014), or else it can manifest as a part of an event which takes place on a large spatial scale (regional event). In this case, the nucleation process can originate from several hundred kilometers away from the measurement site (Dall'Osto et al., 2013; Németh and Salma 2014; Zhu et al., 2014), and the newly-formed particles can be contemporarily observed in multi-location measurements. It was noted that most NPF events occurred in analogous conditions, under the influence of air masses with the same origin and enriched with sulfur-rich precursors (Kalkavouras et al., 2020), and exhibiting similar dynamic characteristics. NPF events were usually detected in ground-based measurement stations, but a number of observations were also performed at high altitudes, on mountain tops (Rose et al., 2015; Bianchi et al., 2016), by balloons and through aircraft (Schroder and Strom, 1997; Mirme et al., 2010) leading to hypothesize that the nucleation process can be triggered within the atmospheric column, at different heights (Boulon et al., 2011; Minguillón et al., 2015; Querol et al., 2017), both throughout the boundary layer and in the free troposphere. At upper atmospheric levels, nucleation would be facilitated by the higher dilution of the pollutants (low condensation sink) and by enhanced photochemical conditions. Once nucleated, the particles can be transported over long distances by the circulation of air masses. Where and at which altitude the mechanism is triggered is still unclear (Boulon et al., 2011), but in this process atmospheric dynamics could play a crucial role in the mixing of precursor gases and/or pre-existing particles, influencing their occurrence and spatial distribution (Wehner et al., 2015).

To date, the variability of the conditions and the complexity of the phenomenon make our knowledge of this process still fragmentary in many aspects. While it is well known that solar intensity, low relative humidity, and the availability of high levels of aerosol nucleation precursors (sulfur dioxide, ammonia, amines, VOCs) are common features in the NPF (Wu et al., 2021), the chemical/physical mechanisms involved in the process remain uncertain and further insights are required.

In this paper, we summarize the results of a long-term study on the new particle formation events carried out at two sites in southern Italy, Lecce (ECO station) and Lamezia Terme (LMT station). To the best of our knowledge, this work represents one of the few long-term studies on NPF events conducted in the Mediterranean area and the first one conducted in Italy. Based on a 5-yr period, simultaneous measurements of aerosol particle number size distribution were analysed with the aim to characterize, investigate, and compare the relationship of NPF events with meteorological and pollution data at a coastal and

an urban background site. To support interpretations, simultaneous measurements of SO<sub>2</sub> and PM<sub>2.5</sub> ambient concentrations were carried out at both ground monitoring stations. The role of air mass and local meteorological factors were also investigated.

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## 2 Measurements and methods

### 2.1 Measurement sites and instrumentation

Measurements of particle number size distribution (PNSD) were made from January 2015 to December 2019 at two sites of southern Italy, Lamezia Terme Observatory (LMT, 38.88° N, 16.23° E, 6 m a.m.s.l.) in Calabria region, and Lecce  
75 Observatory (ECO, 40.20° N, 18.07° E, 50 m a.m.s.l.) in Puglia region. The locations of the measuring sites are depicted in Fig. 1. ECO and LMT, both regional stations of GAW/ACTRIS networks, are representative of urban backgrounds and suburban/coastal sites, respectively. ECO observatory is located inside the University Campus (Dinoi et al., 2020), about 5 km southwest of the municipality of Lecce, while the LMT observatory is located about 17 km from the urban city and about 600 m inland from the Tyrrhenian coastline, at an elevation of 6 m above sea level (a.s.l.). The two sites are about 225 km as the  
80 crow flies. The municipalities cover areas of 238 km<sup>2</sup> (Lecce) and 160 km<sup>2</sup> (Lamezia) with populations of 95,000 and 70,452 inhabitants respectively. Their climate is Mediterranean, albeit characterized by different local meteorological conditions. Being on the coast, the local weather of LMT is influenced by a system of “land-see” breezes that guarantees a temperate climate and continuous ventilation throughout the year that favors an effective dilution of air pollutants. Different emission sources (more details from Cristofanelli et al., 2018; Calidonna et al., 2020; Dinoi et al., 2021b) affect the two sites. The main  
85 aerosol sources in the ECO observatory are the emissions of vehicular traffic and biomass combustions, which are added to natural and anthropogenic long-range contributions (Donateo et al., 2018; Conte et al., 2020). The site is characterized by frequent conditions of clear skies during the whole year, with mild autumn-winter seasons and warm spring-summer. The LMT observatory is located far from the urban agglomeration and therefore is not directly affected by the emission deriving from the main anthropogenic activities.

90 Aerosol PNSD from 10 to 800 nm, with a time resolution of 5 minutes, were collected using a TROPOS-type custom-built MPSS (Mobility Particle Size Spectrometer), designed and manufactured according to EUSAAR/ACTRIS recommendations (Wiedensohler et al., 2012). It is a closed-loop system, with a 5:1 ratio between the sheath and aerosol flow, where the sample air is drawn into the instrument at a flow rate of 1 l/min, while the sample humidity is regulated below 40 % by a Nafion dryer. The two MPSS consist of a bipolar diffusion charger (Ni-63), a differential mobility analyzer (Vienna-type DMA, length 28  
95 cm), and a condensation particle counter (CPC, model: TSI 3772, TSI Inc., USA). The instruments used are the same at the two stations as well as the calibration procedures, carried out periodically, and data quality control. The quality of measurements of both instruments was routinely checked and all data were corrected for particle losses by diffusion and multiple charged particles according to the recommendations of Wiedensohler et al., (2012). Concentrations of sulphur dioxide (SO<sub>2</sub>) were measured using Thermo Instruments analyzers, TEI 43i, and PM<sub>2.5</sub> mass concentrations using low-volume samplers

100 with a  $\beta$ -ray attenuation method (SWAM 5a Dual Channel Monitor (FAI Instruments). Local meteorological parameters, air temperature (T), relative humidity (RH), precipitation (Rain), wind speed (WS), and wind direction (WD) were monitored by automatic weather stations (Vaisala WXT 520), located in the two observatories while solar radiation (solar flux) was collected by CNR 4 net radiometers (Kipp&Zonen). Air mass back trajectories were calculated using the HYSPLIT 4 model developed by NOAA/ARL, a single particle Lagrangian trajectory dispersion model (Stohl 1998).

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## 2.2 Data analysis

Classification of NPF events was performed by visual inspection of daily contour plots (Dal Maso et al., 2005). Examining the time evolution of the particle number size distribution, three main classes were detected: NPF events, non-events, and  
 110 undefined events. NPF events contain cases where a significant increase in the number concentrations of ultrafine particles and growth toward larger diameters was observed for at least 3 - 4 h continuously, displaying the shape of a “banana”. “Non-events” are the days without new particle formation, while “undefined events” group ambiguous cases, with unclear formation and growth, or the occurrence of newly formed particles below 20 nm without the next phase of growth.

Particle growth rate (GR) was calculated from time evolution of the mean geometric diameter  $D_p$  in the size range of 10-20  
 115 nm, using Eq. (1) (Kulmala et al., 2012):

$$GR(nm h^{-1}) = (D_{p2} - D_{p1}) / (t_2 - t_1) \quad (1)$$

with  $D_{p1}$  and  $D_{p2}$  the geometric diameter at the start time  $t_1$  and end time  $t_2$  of the growth event. Using the maximum concentration method (Lehtinen et al., 2003), we identified the time when the concentration is at the maximum in each size bin of the measured particle number size distributions. The growth rates were obtained as the slope of the linear fit of the times  
 120 with the corresponding geometric mean diameters of the size bin particles.

The condensation sink, CS ( $s^{-1}$ ), quantifies how rapidly a condensable gaseous compound condenses on available aerosol particles (Kerminen et al., 2018), and then it depends on the effective surface area of pre-existing particles. CS was calculated using the methods available in the literature (Dal Maso et al., 2005, and references therein) considering sulphuric acid ( $H_2SO_4$ ) as the condensable species:

$$125 \quad CS = 2\pi D \sum_{D',p} \beta_m(D'_{pi}) D'_{pi} N_i \quad (2)$$

where  $D$  is the diffusion coefficient for  $H_2SO_4$ ,  $N_i$  is the particle number concentration with diameter  $D'_{pi}$  of the size bin  $i$ , and  $\beta_m$  is the transition correction factor (Fuchs et al., 1971).

The formation rate,  $J$  ( $cm^{-3}s^{-1}$ ), defined as the flux of changes in particle number concentration within the nucleation mode size range (10–20 nm), was estimated by (Dal Maso et al., 2005; Kulmala et al., 2012):

$$130 \quad J = \frac{dN_{nuc}}{dt} + CoagS_{nuc} * N_{nuc} + \frac{GR}{\Delta dp} N_{nuc} \quad (3)$$

with ( $dN_{nuc}/dt$ ) the temporal change rate of nucleation mode particle number concentration, ( $CoagS_{nuc} \cdot N_{nuc}$ ) the loss caused by collision process and by growth process ( $GR \cdot N_{nuc} / \Delta dp$ ).

Sulfuric acid ( $H_2SO_4$ ) is considered a key precursor for new particle formation, therefore its concentrations were derived by calculating the  $H_2SO_4$  proxy, without scaling factor, using the method presented by Petäjä et al. (2009):

$$135 \quad [H_2SO_4] \propto \frac{SO_2 \times SRad}{CS} \quad (4)$$

where  $SO_2$  is the sulphur dioxide concentration,  $SRad$  is the solar radiation flux, and  $CS$  the condensation sink.

### 3 Results and discussions

#### 3.1 Classification of NPF events

140 The relative frequency of the different classes, events, non-events, and undefined events was calculated for both sites and summarized in Table 1. Over five years of measurements, we had a data coverage of ~78 %, where the available measurement days were 1423 at ECO and 1440 at LMT. The missing days were due to technical/maintenance problems of the MPSS or measurement failures and mainly concerned the months of February/March for LMT, and January for ECO. Among the available data, 25 % and 9 % of days were identified as NPF events while 4% and 3% as undefined events at ECO and LMT, respectively. We found that the percentage frequency of NPF events is higher in ECO than in LMT, with differences in the monthly and seasonal occurrence of the events (Fig. 2). At the ECO site the highest frequencies of NPF events were observed in March and September (~30 %) and the lowest in November and December (12 % -16 %), showing similar results to those found in a shorter measurement period presented by Dinoi et al. (2021a). At LMT, March (19 %) and October (16 %) showed the highest frequencies while July and August were the lowest ones (2 - 3 %). Consequently, this is reflected in a different seasonality of events, more frequent in spring and summer in ECO, and in autumn and winter in LMT. Of all events detected during the study period, 50 were observed simultaneously at both sites. Due to the complementary frequency, most of them occurred in the colder months, ~ 23 % in winter and spring, 14 % in summer, and 39 % in autumn. The annual frequencies of NPF (9% and 25 %) are in good agreement with frequencies (10 % - 36 %) found in other studies, based on long-term measurements, carried out in the Mediterranean area (Kopanakis et al., 2013; Kalivitis et al., 2019; Hussein et al., 2020; Kalkavouras et al., 2020; Baalbaki et al., 2021). The seasonal variability observed in the ECO site was also similar to that found in several other locations (in Europe and around the world) and can be ascribed to the higher emissions of biogenic aerosol precursor compounds and photochemical processes promoted by the higher temperature during the warmer months (Asmi et al., 2016). However, the opposite trend displayed by the LMT site is not a novelty since it has also been observed, albeit less frequently, in other southern Europe sites. In Spain and Greece, for example, Bousiotis et al. (2021) found the occurrence of a larger number of NPF events just during winter and that might be linked to the specific meteorological conditions of the study area.

#### 3.2 Atmospheric particle number concentration (PNC)

A statistical overview of the concentrations and contributions of three diameter modes particles, nucleation ( $N_{nuc}=N_{10-20}$ ),  
165 Aitken ( $N_{Aitk}=N_{20-100}$ ), and accumulation ( $N_{acc}=N_{100-800}$ ), to total particle number concentration ( $N_{tot}=N_{10-800}$ ) was made. As  
summarized in Table 2 (arithmetic means  $\pm$  1 standard deviation, medians, and percentiles (25th-75th)), the mean value of  
total PNC was  $(4.4 \pm 2.2) \times 10^3 \text{ cm}^{-3}$  at LMT and  $(7.8 \pm 3.4) \times 10^3 \text{ cm}^{-3}$  at ECO. Although significantly higher concentrations  
(about 40 %) were found at the urban background site compared to the coastal site, the contribution of each particle fraction  
170 to total particle number concentration was very similar, 24 %, 49 %, and 27 % in LMT and 20 %, 52 % and 28 % in ECO, for  
nucleation, Aitken, and accumulation mode particles, respectively. In both cases, Aitken mode particles represented the largest  
fraction of particle number concentrations.

The levels measured at two sites are within the range reported for similar locations observed in other European sites (Putaud  
et al., 2010; Asmi et al., 2011; Kalivitis et al., 2019; Casquero-Vera et al., 2020; Kalkavouras et al., 2020). The monthly  
variability of PNC, averaged over the whole study period, is shown in Fig. 3. In ECO and LMT the highest monthly average  
175 concentrations of Aitken mode particles were observed during winter ( $\sim 5.6 \times 10^3 \text{ cm}^{-3}$  and  $\sim 2.7 \times 10^3 \text{ cm}^{-3}$ ), approximately  
45 % in ECO and 35 % in LMT higher than in summer ( $\sim 3.2 \times 10^3 \text{ cm}^{-3}$  and  $\sim 1.7 \times 10^3 \text{ cm}^{-3}$ ). In ECO accumulation mode  
particles ranged from  $\sim 1.7 \times 10^3 \text{ cm}^{-3}$  in spring to  $\sim 3.1 \times 10^3 \text{ cm}^{-3}$  in winter, while in LMT from  $\sim 9.8 \times 10^2 \text{ cm}^{-3}$  in autumn to  
 $\sim 1.4 \times 10^3 \text{ cm}^{-3}$  in summer. The increase in  $N_{acc}$  concentration during the summer can be ascribed to the dry conditions that  
characterize the Mediterranean summers (Pikridas et al., 2018). Nucleation mode particles displayed a similar seasonal pattern  
180 to Aitken mode particles in LMT (from  $\sim 8.0 \times 10^2 \text{ cm}^{-3}$  in summer to  $\sim 1.4 \times 10^3 \text{ cm}^{-3}$  in winter), while in ECO values slightly  
higher were observed during spring/summer,  $\sim 1.9 \times 10^3 \text{ cm}^{-3}$ , than in winter  $\sim 1.3 \times 10^3 \text{ cm}^{-3}$ . The higher concentrations  
recorded during cold months are generally associated with enhanced anthropogenic emissions from fossil fuel combustion and  
biomass burning, due to residential heating, and unfavourable meteorological conditions for pollution dispersion.

A clear diurnal pattern in each mode particle number concentration was observed in every season. Fig. 4 shows the trend of  
185 each mode fraction considering separately the days of NPF events (E, solid line) and the days of non-events (NE, dashed line).  
The timing of measurements is expressed in solar time (UTC + 1). Nucleation, Aitken, and accumulation mode particles have  
very similar behavior during non-events and, except for the different concentrations, both sites show a pronounced diurnal  
cycle with a morning and evening peak. The two peaks are shifted by one hour between spring-summer and autumn-winter  
because of daylight savings time and are mainly linked to vehicular emissions, most intense during the morning and evening  
190 rush hour. In addition, the evening peaks of Aitken and accumulation mode particles in winter and autumn can be linked to  
domestic heating emissions, mainly biomass burning, considered an important source of ultrafine particles in urban sites. In  
cold months these particles tend to accumulate during the night due to the reduced boundary layer compared to the daytime  
layer. These peaks are also present in LMT but are less intense due to the greater distance of the site from the urban centre.  
Regarding event days, in both sites, together with the two peaks of rush hours, nucleation mode particles present further peaks  
195 around noon, more marked in summer, spring, and winter in ECO, and spring, summer, and autumn in LMT. Less pronounced  
are instead those in winter and autumn in LMT and ECO respectively.

Similar observations have been reported in Cusack et al. (2013), Kalivitis et al. (2019), Kalkavouras et al. 2020, Dinoi et al. (2020, 2021a), for the western Mediterranean sites where the diurnal variation in nucleation mode particles presents a clear maximum at noon under both polluted and clean conditions. The contribution of the NPF process to the number concentration is also observed in the Aitken mode particles, more noticeable in the LMT site with 30 % in autumn-winter and 41 % in spring-summer, and with 21 % only in spring-summer in the ECO site. In Table 3 is reported mean values  $\pm$  1 standard deviation of nucleation, Aitken and accumulation of PNC ( $\text{cm}^{-3}$ ) related to event days (E) and non-event days (NE) for each season (winter W, spring Sp, summer S, and autumn A) in both sites. Nucleation mode particles show an increase of 52 %, 65 %, 61 %, and 49 % in winter, spring, summer, and autumn in LMT, and of 47 %, 52 %, 55 %, and 39 % in ECO. These results highlight that the formation of new particles contributes to the overall particle population more in the warm months and more significantly in the coastal site than in the urban background site, probably because the urban site is also affected by local emission of ultrafine particles that tend to suppress the NPF process.

No contribution is observed in the concentration of accumulation mode particles where especially in the first half of the day, the concentrations were higher on non-event than event days, especially at the ECO site. This could explain the different frequency of events that characterized the two sites in these seasons, assuming that the NPF events were favored on those days with lower particle number concentrations (Salma et al., 2017).

The relative increase in particle number concentration due to the NPF process was also quantified with the nucleation strength factor (NFS) proposed by Salma et al. (2016, 2017, 2019). It measures the effects of nucleation events on ultrafine particles at a site considering two factors,  $\text{NSF}_{\text{NUC}}$  that provides a measure of the concentration increment on nucleation days exclusively caused by NPF, and  $\text{NSF}_{\text{GEN}}$  that gives a measure of the overall contribution of NPF over a longer time span. In this work we considered only  $\text{NSF}_{\text{NUC}}$ , calculated following Eq. (5):

$$\text{NFS}_{\text{NUC}} = \frac{(N_{10-100})/(N_{100-800})_E}{(N_{10-100})/(N_{100-800})_{NE}} \quad (5)$$

with  $N_{10-100} / N_{100-800}$  the concentration ratios for nucleation days at the numerator and the  $N_{10-100} / N_{100-800}$  concentration ratios for not nucleation days at the denominator. Depending on the value,  $\text{NFS}_{\text{NUC}} < 1.0$ ,  $1.0 < \text{NFS}_{\text{NUC}} < 2.0$ , or  $\text{NFS}_{\text{NUC}} > 2.0$ , the nucleation process can be considered negligible, comparable to the other sources or the main contributor. In LMT the mean values of  $\text{NFS}_{\text{NUC}}$  were 2.0, 2.1, 2.3, and 1.9 while in ECO 1.7, 1.6, 1.9, and 1.9 in winter, spring, summer, and autumn, respectively. From the coastal to urban background site, we found a decrease in the contribution of NPF events to particle number, like what was observed by Salma et al. (2017) between the near city background (2.3) and the city center (1.6) of Budapest over 5 years. In the study of Bousiotis et al. (2021), on 13 sites from five countries in Europe it was found that for almost all rural background sites  $\text{NFS}_{\text{NUC}}$  was greater than 2 and reached 4 in a very clean site of Finland. Nemet et al. (2018) found lower values of  $\text{NFS}_{\text{NUC}}$ , 1.58, 1.54, and 2.01, in the cities of Budapest, Vienna and Prague, respectively, while in Granada urban site  $\text{NFS}_{\text{NUC}}$  was 1.05 (Casquero-Vera et al., 2021). The decrease in the contribution of NPF events to

230 particle number, moving from a less polluted to a more polluted site, may be related to the higher contribution to particle number concentrations of other sources, i.e. traffic and heating, and the associated increased condensation sink. These results point out that the particles produced by the NFS<sub>NUC</sub> process can become the dominant sources in a clean environment compared to more polluted areas.

### 235 3.3 Factors associated with NPF events

The growth rate (GR) and the particle formation rate (J) were analyzed to investigate the dynamic properties of NPF events. At the ECO site the growth rate values ranged from 3 nm h<sup>-1</sup> to 14 nm h<sup>-1</sup> (average 7.5 ± 3.3 nm h<sup>-1</sup>) and J from 0.6 to 8.6 cm<sup>-3</sup> s<sup>-1</sup> (average 3.3 ± 3.1 cm<sup>-3</sup> s<sup>-1</sup>); while at LMT the GR varied from 2.5 to 10 nm h<sup>-1</sup> (average 6.1 ± 2.3 nm h<sup>-1</sup>) and the J from 0.3 to 6.2 cm<sup>-3</sup> s<sup>-1</sup> (average 2.4 ± 1.8 cm<sup>-3</sup> s<sup>-1</sup>). The values of both parameters are comparable with what was

240 reported for NPF events in other urban and coastal sites (Hussein et al., 2020, Kalivitis et al. 2019; Salma et al. (2019), Kalkavouras et al., 2020; Nieminen et al. 2018). In particular, similarities were found with some Mediterranean sites such as the coastal station of Finokalia GR~ 5 nmh<sup>-1</sup>, J~ 0.9 cm<sup>-3</sup> s<sup>-1</sup> (Pikridas et al., 2012), the coastal/rural/suburban station of Akrotiri, GR~ 6 nmh<sup>-1</sup>, J~13 cm<sup>-3</sup> s<sup>-1</sup> (Kopanakis et al., 2013) and the Cyprus Island, GR~ 2.8–5 nmh<sup>-1</sup>, J~ 5–11.4 cm<sup>-3</sup> s<sup>-1</sup> (Debevec et al., 2018). The mean values of GR and J turned out to be higher at the ECO site than at the LMT site, and they

245 showed a clear seasonal pattern with higher values during warm months, not observed at LMT where both parameters did not show distinctive features. As reported by Nieminen et al. (2018), the production rate of nucleation particles is generally higher in urban sites than in remote/clean ones because of the greater availability of precursors deriving from greater anthropogenic activity. In particular, the higher values of GR and J in warm months can be attributed to the intensification of the photochemical activity and abundance of SO<sub>2</sub> which acts as a precursor of sulfuric acid. In this perspective, we investigated

250 the effects of PM<sub>2.5</sub>, CS, and SO<sub>2</sub> concentrations on the occurrence of NPF at both sites where visible differences in pollution load (Fig. 5) were found. The average PM<sub>2.5</sub> and SO<sub>2</sub> concentrations during event days were 2.5±1.5 µg m<sup>-3</sup> (ranging from 2.0 to 3.8 µg m<sup>-3</sup>) and 0.4±0.3 ppb (from 0.2 to 0.6 ppb) for LMT, while they were 12.0±6.0 µg m<sup>-3</sup> (from 9.5 to 15.3 µg m<sup>-3</sup>) and 1.1±0.4 ppb (from 0.8 to 1.3 ppb) for ECO. These values were compared with mean values of PM<sub>2.5</sub> and SO<sub>2</sub> measured on non-event days, 6.1±2.5 µg m<sup>-3</sup> and 0.2±0.2 ppb in LMT, 15.5±8.0 µg m<sup>-3</sup> and 0.9±0.4 ppb in ECO. In both sites statistically

255 significant differences (p-value <0.05 level, by Wilcoxon-Mann-Whitney test) emerged, highlighting the different role they may have played in the new particle formation. CS, estimated from measurements of the particle number size distribution between 10 and 800 nm, was (0.7 ± 0.3) × 10<sup>-2</sup> s<sup>-1</sup> (0.005 - 0.009 s<sup>-1</sup>) for LMT and (1.0 ± 0.6) × 10<sup>-2</sup> s<sup>-1</sup> (0.008 - 0.01 s<sup>-1</sup>) for ECO, within the range of coastal and urban environments (Salma et al., 2016; Kalivitis et al., 2019; Baalbaki et al., 2021). In this case, the comparison with values measured during non-event days, (0.7±0.4) × 10<sup>-2</sup> s<sup>-1</sup> in LMT and (1.3 ±0.5) × 10<sup>-2</sup> s<sup>-1</sup> in

260 ECO showed no noteworthy differences. The two sites displayed evident differences in both concentrations, with SO<sub>2</sub> ~65 %, PM<sub>2.5</sub> ~80 %, and CS ~30 % lower in LMT than ECO. This underlines how the two areas are affected in different ways by anthropogenic contributions, and the lower levels of SO<sub>2</sub> could be responsible for the reduced NPF frequency (~65 % less than



ECO) observed in LMT. In general, particulate matter concentration has a direct effect on the condensation sink that quantifies how rapidly a condensable gaseous compound condenses on available aerosol particles (Kerminen et al., 2018). As known, new particle formation is not favoured by the high concentration of pre-existing particles because they increase condensation sink (CS) for vapours. Therefore, lower CS would favour nucleation proportionally to the amount of condensable vapours available (Saha et al., 2018). Sulphuric acid is identified as one of the key components directly connected to NPF process (Sipila et al., 2010). Because no direct measurements of it were done in this study, we investigate its role, considering the proxy of sulphuric acid (Eq.4), without scaling factor (Petaja et al. 2009). The proxy only allows us to estimate the order of the average concentration levels of  $\text{H}_2\text{SO}_4$  and although the results obtained are subject to uncertainties, they can still provide indications of trends (Salma et al., 2019). The average monthly values of  $\text{H}_2\text{SO}_4$  proxy showed substantial differences between the two sites on event days (Fig 5d),  $40 \times 10^3 \text{ ppbWm}^{-2}\text{s}$  (ranging from  $18 \times 10^3$  to  $61 \times 10^3 \text{ ppbWm}^{-2}\text{s}$ ) at ECO and  $20 \times 10^3 \text{ ppbWm}^{-2}\text{s}$  (from  $11 \times 10^3$  to  $38 \times 10^3 \text{ ppbWm}^{-2}\text{s}$ ) at LMT. These values are about 35% higher than non-events days in both cases. The proxy values of  $\text{H}_2\text{SO}_4$  are larger in warm months and are substantially higher, by a factor of 2, in the urban background than in the coastal site, mainly due to the values of CS and  $\text{SO}_2$ . These last results are in accordance with GR and J values obtained for the two sites, where the low J and GR values at LMT could be associated with the low concentrations of  $\text{H}_2\text{SO}_4$  (or other precursors). The conditions for the occurrence of NPF events are mainly driven by the ratio of the source and sink terms for the condensing vapors, therefore a greater availability of this gas precursor could have favored the occurrence of NPF events at ECO, although the higher values of CS, as well as the lower levels could have limited its development at LMT.

### 3.4 Meteorological conditions and air masses associated with NPF events

The comparison of the main meteorological parameters was carried out to better characterize the occurred NPF also in terms of the local climate of the two study areas. The monthly average values of solar flux, relative humidity, and wind speed were calculated starting from the daily averages. All NPF events happened in comparable weather conditions featured by high pressure, with a prevalence of clear skies, typical of the Mediterranean climate. Monthly mean values showed an intense solar flux both in LMT (from 260 to  $512 \text{ W m}^{-2}$ ) and in ECO (from 230 to  $510 \text{ W m}^{-2}$ ) during events (columns). These values were very similar to those measured during non-event periods except at the LMT site where higher mean values are observed in the colder months, differences which however are not statistically significant. During event days, the RH was comparable at the two sites:  $52 \pm 4 \%$  in LMT and  $50 \pm 7 \%$  in ECO, and it was about 25 % lower than that on non-event days (Fig. 6b). This result is not a novelty because lower RH is usually observed during NPF events in both clean and polluted environments (Kerminen et al., 2018). On average, monthly wind speeds were higher during events in both sites,  $4 \pm 1 \text{ m s}^{-1}$  for LMT (from 2.9 to  $5.4 \text{ m s}^{-1}$ ) and  $3 \pm 1 \text{ m s}^{-1}$  for ECO (from 2.6 to  $4.3 \text{ m s}^{-1}$ ), respectively about 10% and 40% higher than the wind speed observed on non-event days (Fig. 6c). The difference is statistically significant only for ECO (p-value <0.05). Regarding the wind directions (Fig. 7) in ECO the dominating wind direction was mainly from N–NNW, with a frequency of occurrence of

295 75 %, while inLMT from E-NE and W-NW, with a frequency of occurrence of 30 % and 55 %, respectively, corresponding to a land-sea breeze system. The two prevalent wind directions together with the higher wind speed (both on events and on non-event days) are typical of a coastal area where generally, thanks to the circulation of the sea-land breeze, wind speed is approximately 20 % greater than in the innermost (inland) sites and almost constant during the whole year (Jensen et al., 2017). The breezes influence the diffusion of atmospheric pollutants and their dilution, thus explaining the low and almost constant  
300 (lack of seasonality) levels of PM measured at LMT.

Also, the origin and pathways of air masses can contribute significantly to the occurrence of NPF processes (Wonaschutz et al., 2015). Depending on their advection pattern, air masses can be affected by emission sources and chemical/physical processes able to influence the properties of the pollutants during transportation. With this aim, the role of air masses was investigated through the analysis of the back trajectories of all days identified as NPF events. 72 h back trajectories were  
305 computed each day at 08:00 UTC, with 6 h resolution, at the 500 m height arrival above ground level (AGL), using the HYSPLIT 4, single-particle lagrangian trajectory dispersion Model, developed by NOAA/ARL (Draxler and Hess, 1998). We considered only the back trajectories at an altitude of 500 m because they can be better correlated with ground-based measurements and, therefore, more representative of the atmospheric boundary layer in which air pollutants are well mixed by horizontal and vertical advection. The analysis of the back trajectories allows us to gain qualitative information on the paths  
310 of the air masses associated with NPF events, and although they can be affected by an error estimated between 15 % and 30 % (Stohl, 1998; Draxler and Hess, 2004), for the purposes of this paper we considered negligible such inaccuracy. The main transport pathways to the study sites were identified by cluster analysis of back trajectories. Fig. 8 shows 6 centroids, each representative of a group of trajectories, labelled according to the direction of the cluster, frequency of occurrence, and length. Short trajectories are indicative of slow-moving air masses while long trajectories are for fast flows; different lengths can have implications on a load of pollutants transported (associated) by air mass. As represented in Fig. 8, during the  
315 occurrence of NPF events the two stations were influenced by similar air masses, most of which were enclosed in the north-western quadrant. Air masses arriving at LMT and originating from the North-Northwester sector exhibited an occurrence rate of 38 % (17 % long and 20.5 % medium trajectories), followed by 62 % of air masses from the West-Northwest sector (9.8 % long, 19.6 % medium and 33 % short trajectories). At ECO, air masses originating from the North-Northwester sector showed a percentage of 20 % (medium trajectories), those from the West-Northwest sector an occurrence of 55 % (21.6 % long and 33.4 % short trajectories), while another important contribution, 25 % (short trajectory), was associated to Eastern sector. All the air masses flew over both the European continental areas highly anthropized and the marine environment before reaching the receptor points. It is interesting to note the lack of air masses associated with NPF events coming from the southern sector. These air masses, because of their pathway, are usually characterized by high levels of dust and humidity able to counter the  
325 NPF process because can suppress photochemical activity by scavenging reactive gases and condensable vapors (De Reus et al., 2000; Ndour et al., 2009).

We also focused on “common” NPF events considering only the air masses pathway related to these days detected during the study period. From 3-days back-trajectory analysis, emerged that the back-trajectories of common events exhibit similar

characteristics in terms of origin and pathways to what already observed, but in addition, two different cases were detected, in  
330 the first the trajectory passes from ECO before reaching LMT, while in the other case, vice versa the air mass passes from  
LMT and then reaches ECO. Out of 50 common NPF events, 31 were in the first case and 19 in the second. Two representative  
examples of back-trajectories for each case are depicted in Fig. 9 and show that there is not a preferential path that can  
characterize them, because in both cases we find trajectories that come from both Eastern and Western Europe. These events  
occurred almost synchronously at the two sites, with a difference in starting time not greater than 30 minutes.

335 The factors that characterized the concomitant events of NPF (Table 4) are similar to those observed for the non-concomitant  
events, with values of  $PM_{2.5}$ ,  $SO_2$ , CS and  $H_2SO_4$  proxy in LMT lower than ECO, and similar meteorological conditions. The  
simultaneous observation of these events indicates that the formation of new particles has a wide horizontal extension and can  
be seen as a large-scale phenomenon. It is probable that the air masses already contain particles that have been formed by  
nucleation somewhere and then transported. Or during their travel, the air masses are enriched with gaseous precursors deriving  
340 from anthropogenic emissions and/or from biogenic production such as to foster (potentially) NPF processes, even in locations  
far from the sources.

These results emphasize the role that air masses have not only in terms of transport of precursors but also in synoptic  
atmospheric conditions to them associated. In general, these results underline the importance of specific atmospheric  
conditions (temperature, solar radiation, RH, origin of air mass, pollutant concentrations) under which the NPF events have  
345 occurred and emphasize how the observed differences are associated with the levels of pollution found in them. The more  
frequent NPF events at the urban background compared to the coastal site can be ascribed to a greater abundance of  
condensable species, deriving from anthropogenic emissions, which favour the growth of particles increasing their chance of  
survival. Regarding the different seasonality of the events, while the trend at the urban site of ECO can be associated with the  
increased photochemical activity and the higher concentrations of different precursors during the warm months, the seasonality  
350 at the coastal site of LMT is more difficult to explain. Along with the lower availability of precursors, local conditions could  
play an additional role as well. These may include synoptic systems such as increased turbulence during warm months and the  
different atmospheric composition (related to the proximity to the sea and the effects of land-sea breezes) due to which the  
newly formed particles could be more effectively suppressed preventing further growth.

#### 355 **4 Conclusions**

New particle formation events were studied and compared at two sites in southern Italy, ECO (Lecce) and LMT (Lamezia  
Terme) observatories, over a period of five years. The nucleation events occurred with a different frequency, 25 %, and 9 %, and seasonality, highest in spring-summer and autumn-winter at ECO and LMT, respectively.

Throughout the investigation period, 50 simultaneous NPF events (14 % of the cases in ECO and 40 % in LMT) were identified  
360 at both sites, many of which occurred in the colder months, indicating that the NPF process was affecting a large spatial extent.  
The NPF days were characterized by lower  $PM_{2.5}$  concentrations (~60 % and 22 %) and higher  $SO_2$  concentrations (~50 %

and 20 %) compared to non-event days in LMT and ECO respectively, while the condensation sink seems not to be significantly different during events and non-event days. Marked differences in PM<sub>2.5</sub>, SO<sub>2</sub>, H<sub>2</sub>SO<sub>4</sub> proxy, and CS levels were observed between the two sites, indicating the minor anthropogenic influence to which LMT is subjected. Common meteorological features were observed during NPF events that occurred in conditions of high pressure, low RH (~52 %), and moderate wind speed (3-4 m s<sup>-1</sup>). The study of the back trajectories associated with the events also highlighted a common origin of the air masses, both of continental and marine origin, from the North-Northwest directions, suggesting that the chemical compounds involved in the NPF could have been transported by the air masses. Our findings let us assume that the lower levels of SO<sub>2</sub> and H<sub>2</sub>SO<sub>4</sub> proxy (-60 % and -50% with respect to ECO) together with a different chemical composition of the aerosols and different local meteorology might be the reason for the lower frequency of events occurring in LMT, -60% with respect to ECO. The results presented in this paper are a contribution toward a better understanding of the complex NPF phenomenon in central Mediterranean area which will require further investigations and measurements of different precursors (such as ammonia, amines, VOCs) involved in the process that were not considered in this work.

**Data Availability:** Data is available upon request.

**Author contributions:** AD and DC conceptualized the study design. AD, DG and KW collaborated to data collection and post-processing. AD and DC prepared the draft. AD, DG, KW, IA, CC, AW and DC collaborated to interpretation of results, wrote, read, commented, and approved the final manuscript.

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## Figures



Figure 1: Map of the examined area with the locations of the two environmental-climate observatories, ECO and LMZ. The map was retrieved from Google (Map data © 2022).

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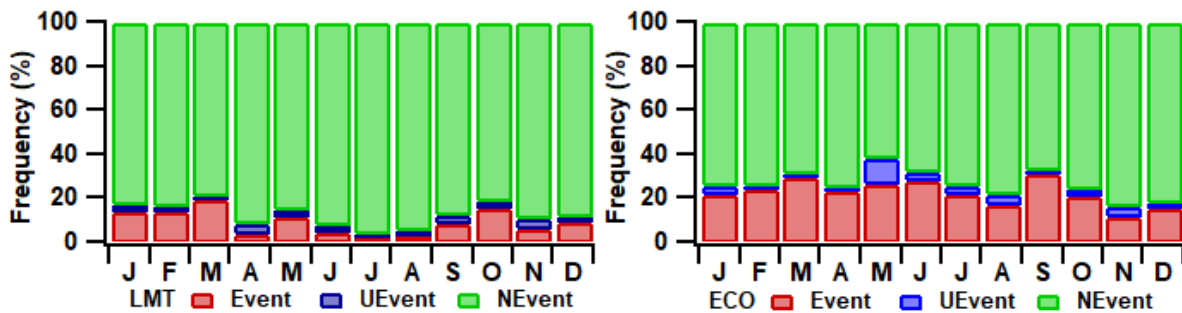
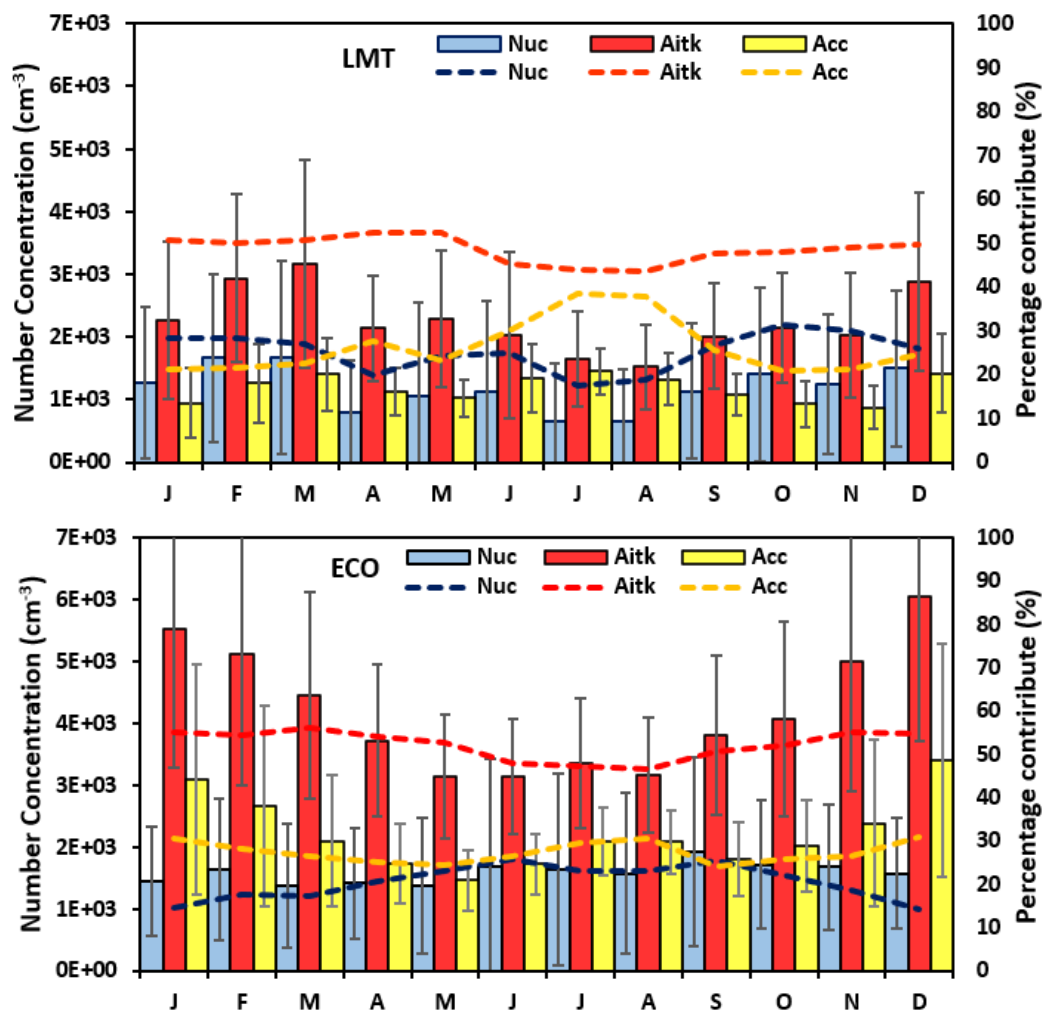


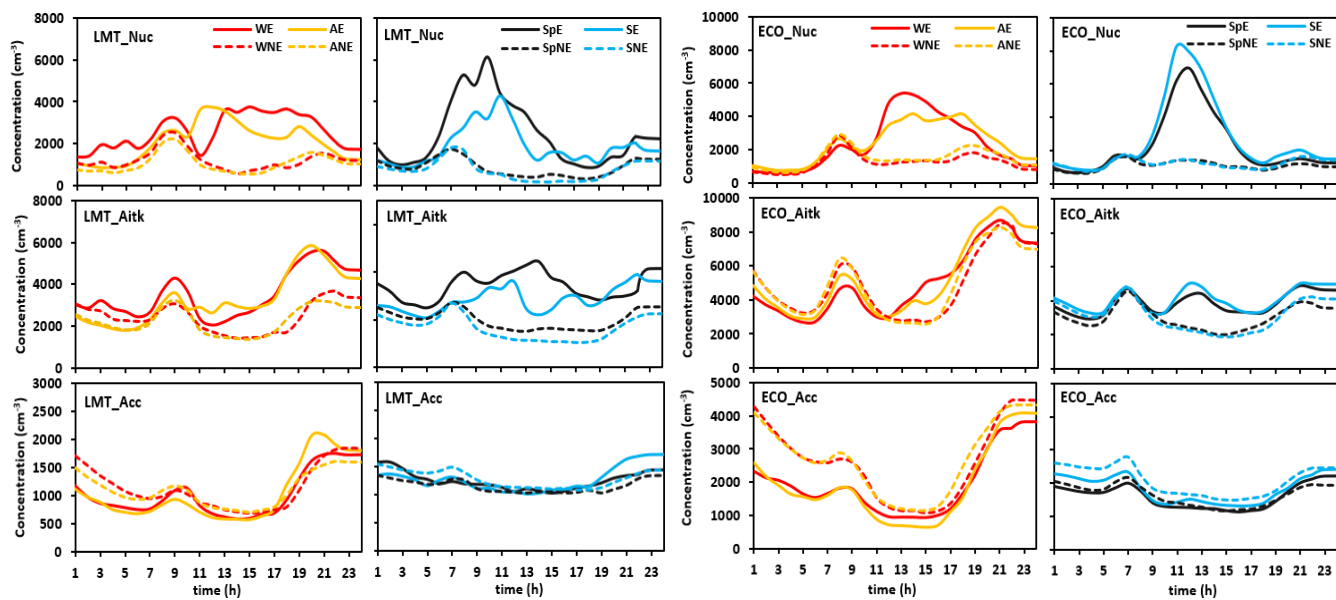
Fig. 2. Monthly percentage of occurrence of Events (red bars), Undefined events (blue bars), and Non-Events (green bars) related to available measurement days at LMT and ECO.

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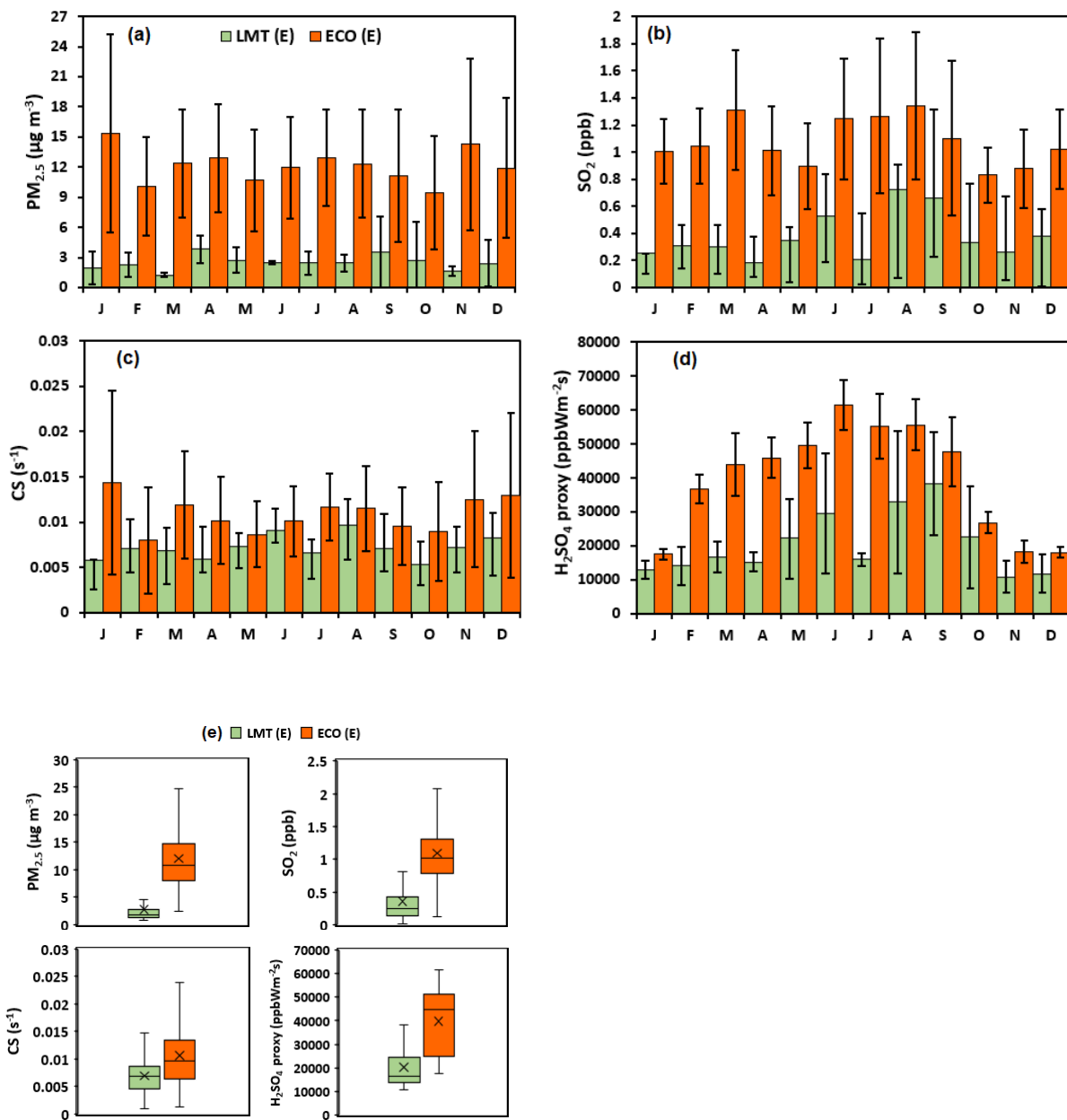
Fig. 3. Average monthly variation (bars) of the nucleation, Aitken, and accumulation mode particle number concentration over the whole period of study at LMT and ECO. The dashed lines represent the monthly percentage contribution of each particle fraction to total particle number concentration.



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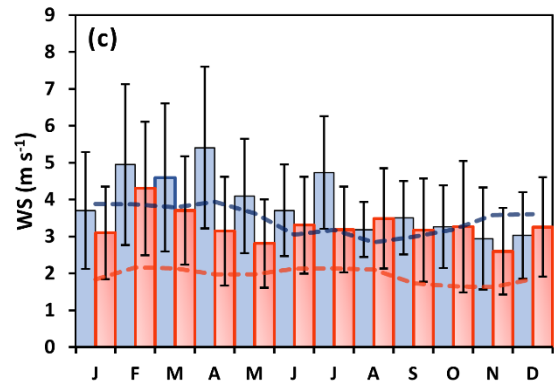
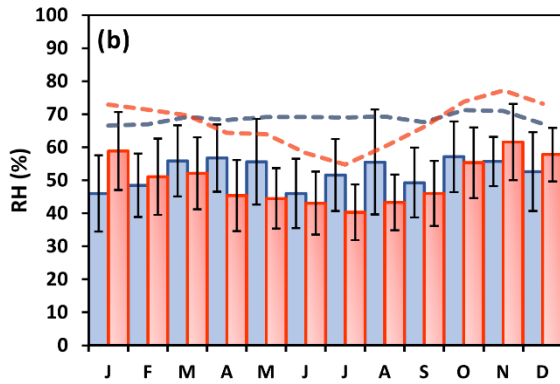
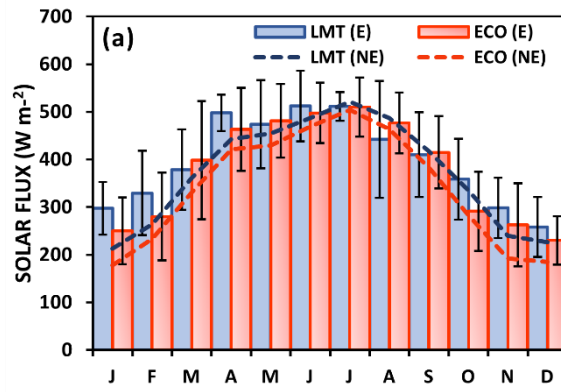
**Fig. 4. Average diurnal variation in nucleation, Aitken, and accumulation mode particle number concentration (from top to bottom) over the whole period of study at LMT and ECO. Solid lines are for NPF events days (E) and dashed lines for non-events days of (NE), red for winter, grey for spring, blue for summer, and yellow for autumn.**

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625 Fig. 5. Monthly means of (a) PM<sub>2.5</sub>, (b) SO<sub>2</sub>, (c) CS and (d) H<sub>2</sub>SO<sub>4</sub> proxy related to event days for LMT (green) and ECO (red). The bars indicate the standard deviation. (e) Box plot of PM<sub>2.5</sub>, SO<sub>2</sub>, CS and H<sub>2</sub>SO<sub>4</sub> proxy associated with event days for LMT (green) and ECO (red). The box shows the quartiles, the median (the line inside the box), the mean (the star), and 90 th and 10 th percentiles (the whiskers).



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**Fig. 6. Monthly means of (a) solar flux, (b) relative humidity, and (c) wind speed related to event days (columns) and non-event days (dashed lines) for LMT (blue) and ECO (red). The bars indicate the standard deviation.**

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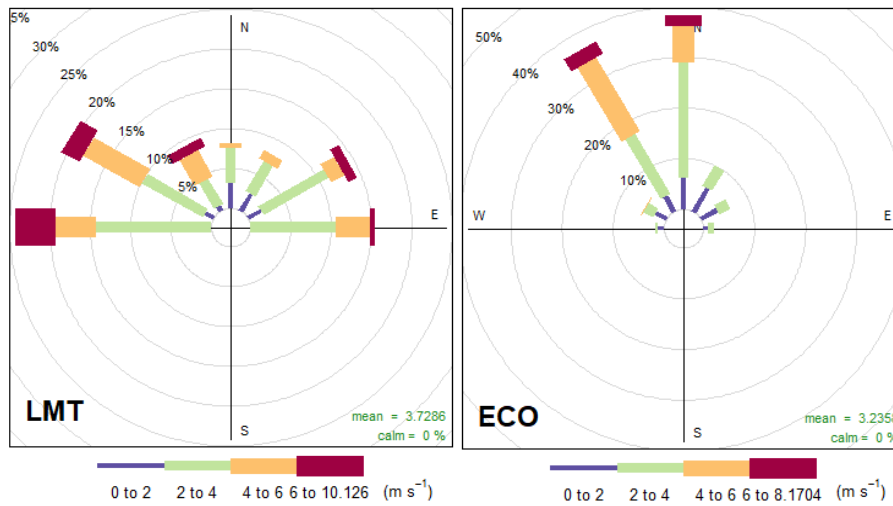


Figure 7. Wind roses related to event days for LMT and ECO observatories.

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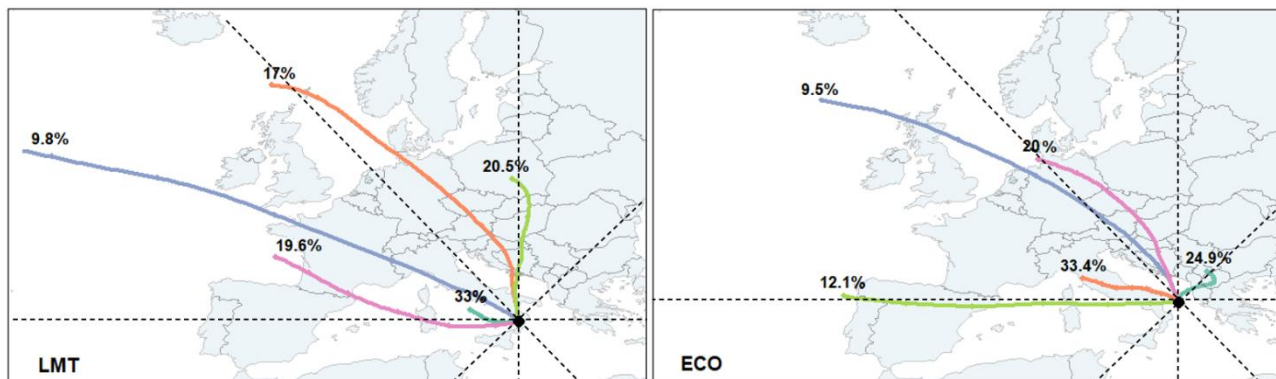
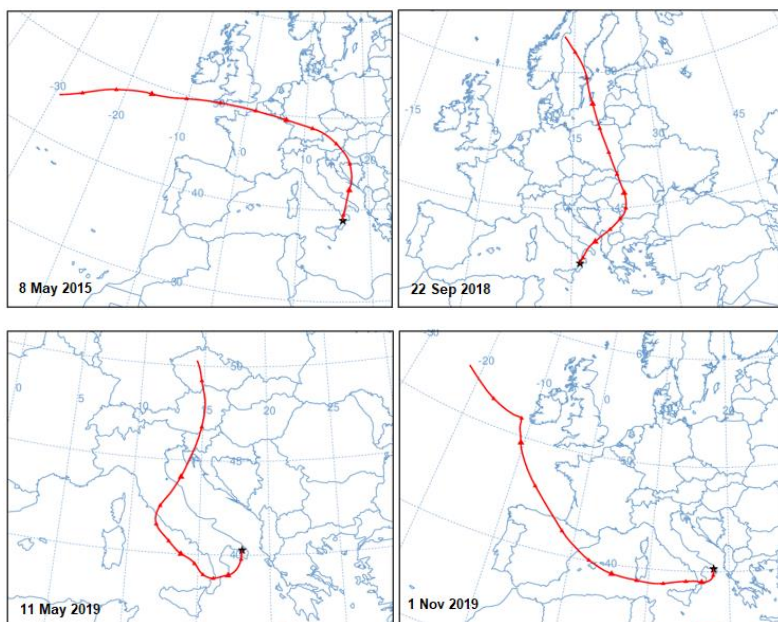


Fig. 8. Cluster centroids retrieved from the 3-day analytical back trajectories reaching ECO and LMT at 500m above sea levels. The maps were plotted using Hysplit integrated with the Openair package.

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Fig. 9. 72 h back-trajectories arriving at LMT (up) and ECO (down) during four simultaneous new-particle formation events. The maps were retrieved from NOAA Hysplit Model.

## Tables



655 **Table 1. Number of available days (N) and relative frequencies (f %) of each class, NPF events, non-events, and undefined events detected in ECO and LMT observatories.**

Day Classification	LMT (Total=1440)		ECO (Total=1423)	
	Number	Frequency (%)	Number	Frequency (%)
NPF events	124	9	352	25
Non-events	1272	88	1019	71
Undefined events	44	3	52	4

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**Table 2. Arithmetic means  $\pm$  1 standard deviation and medians (25th–75th) percentiles of total, accumulation, Aitken, and nucleation PNC ( $\text{cm}^{-3}$ ) for both sites.**

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	Nnuc ( $10^3\text{cm}^{-3}$ )	Naitk ( $10^3\text{cm}^{-3}$ )	Nacc ( $10^3\text{cm}^{-3}$ )	Ntot ( $10^3\text{cm}^{-3}$ )
LMT	1.1 $\pm$ 1.0	2.2 $\pm$ 1.1	1.2 $\pm$ 0.6	4.4 $\pm$ 2.2
	0.7 (0.3-1.5)	1.9 (1.4-2.7)	1.0 (0.7-1.0)	4.0 (2.8-5.6)
ECO	1.6 $\pm$ 1.0	4.1 $\pm$ 1.8	2.1 $\pm$ 1.4	7.8 $\pm$ 3.4
	1.3 (0.9-1.9)	3.7 (2.8-4.8)	1.8 (1.2-2.6)	7.1 (5.5-9.3)

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**Table 3. Arithmetic means  $\pm$  1 standard deviation of nucleation, Aitken and accumulation PNC ( $\text{cm}^{-3}$ ) related to event days (E) and non-event days (NE) for each season (winter W, spring Sp, summer S, and autumn A) and for both sites.**

	LMT			ECO		
	Nnuc ( $10^3\text{cm}^{-3}$ )	Naitk ( $10^3\text{cm}^{-3}$ )	Nacc ( $10^3\text{cm}^{-3}$ )	Nuc ( $10^3\text{cm}^{-3}$ )	Naitk ( $10^3\text{cm}^{-3}$ )	Nacc ( $10^3\text{cm}^{-3}$ )
W (E)	2.5 $\pm$ 0.8	3.5 $\pm$ 1.1	1.0 $\pm$ 0.4	2.3 $\pm$ 1.6	4.9 $\pm$ 1.9	1.9 $\pm$ 0.9
W (NE)	1.2 $\pm$ 0.5	2.4 $\pm$ 0.7	1.1 $\pm$ 0.4	1.2 $\pm$ 0.6	4.8 $\pm$ 1.8	2.6 $\pm$ 1.2
Sp (E)	2.4 $\pm$ 1.5	3.4 $\pm$ 0.6	1.2 $\pm$ 0.2	2.2 $\pm$ 1.8	3.8 $\pm$ 0.6	1.6 $\pm$ 0.3
Sp (NE)	0.8 $\pm$ 0.4	2.0 $\pm$ 0.4	1.2 $\pm$ 0.2	1.1 $\pm$ 0.3	3.0 $\pm$ 0.7	1.6 $\pm$ 0.3
S (E)	1.8 $\pm$ 0.9	2.9 $\pm$ 0.5	1.3 $\pm$ 0.2	2.7 $\pm$ 2.2	4.1 $\pm$ 0.7	1.8 $\pm$ 0.4
S (NE)	0.7 $\pm$ 0.4	1.6 $\pm$ 0.5	1.3 $\pm$ 0.2	1.2 $\pm$ 0.3	3.1 $\pm$ 0.9	2.1 $\pm$ 0.4
A (E)	2.0 $\pm$ 0.9	3.3 $\pm$ 1.2	1.0 $\pm$ 0.5	2.4 $\pm$ 1.2	5.2 $\pm$ 2.2	1.9 $\pm$ 1.1
A (NE)	1.1 $\pm$ 0.4	2.3 $\pm$ 0.6	1.1 $\pm$ 0.3	1.5 $\pm$ 0.6	4.9 $\pm$ 1.9	2.7 $\pm$ 1.1

675 **Table 4. Arithmetic means  $\pm$  1 standard deviation and medians (25 th–75 th) percentiles of different parameters, PM<sub>2.5</sub>, SO<sub>2</sub>, CS, H<sub>2</sub>SO<sub>4</sub>, GR, RH, WS and SF for both sites.**

	PM <sub>2.5</sub> ( $\mu\text{g m}^{-3}$ )	SO <sub>2</sub> (ppb)	CS $\times 10^2$ (s <sup>-1</sup> )	H <sub>2</sub> SO <sub>4</sub> $\times 10^3$ (ppbWm <sup>-2</sup> s <sup>-1</sup> )	GR (nm h <sup>-1</sup> )	RH (%)	WS (m s <sup>-1</sup> )	SF (W m <sup>-2</sup> )
LMT	3.1±2.6 2.0 (1.3-4.2)	0.3±0.3 0.2 (0.2-0.4)	0.6±3.0 0.6 (0.4-0.8)	21±19 15 (10-23)	6.7±2.6 6.2 (5.0-8.2)	48±10 50 (40-56)	4.2±1.4 3.9 (3.0-5.0)	394±110 360 (313-496)
ECO	8.9±4.8 8.0 (5.1-11.0)	0.9±0.3 0.9 (0.8-1.0)	0.9±0.6 0.8 (0.8-1.0)	38±23 39 (26-62)	6.6±2.0 6.6 (5.0-7.6)	49±12 48 (41-56)	3.8±1.7 3.5 (2.2-4.9)	365±133 350 (340-470)