

Regional PM_{2.5} pollution confined by atmospheric internal boundaries in the North China Plain: boundary layer structures and numerical simulation

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1 **Abstract.** This study reveals mesoscale planetary boundary layer (PBL) structures under various
2 pollution categories during autumn and winter in the North China Plain. The role of the atmospheric
3 internal boundaries (AIBs, referring to the discontinuity of meteorological conditions in the lateral
4 direction) in regulating PBL structure and shaping the PM_{2.5} pollution patterns is emphasized. The
5 Weather Research and Forecast model is used to display the three-dimensional meteorological fields, and
6 its performance is evaluated by surface observations and intensive soundings. The evaluation
7 demonstrates that the model reasonably captures the mesoscale processes and the corresponding PBL
8 structures. Based on the reliable simulations, three typical pollution cases are analyzed. Case-1 and Case-
9 2 represent the two main modes of the wind shear category pollution, which is featured with airflow
10 convergence line/zone as AIB and thus is dominated by dynamic effect. Case-1 presents the west-
11 southwest wind shear mode associated with a trough convergence belt. The convergent airflow layer is
12 comparable to the vertical scale of the PBL, allowing PM_{2.5} accumulation to form a high pollution area.
13 Case-2 exhibits another mode with south-north wind shear. A "lying Y-shaped" convergence zone is
14 formed with a thickness of about 3000m, extending beyond the PBL. It defines a clear edge between the
15 southern polluted air mass and the clean air in the north. Case-3 represents the topographic obstruction
16 category, which is characterized by a cold-air damming AIB in front of the mountains. The PBL at the
17 foothills is thermally stable and dynamically stagnant due to the capping inversion and the convergent
18 winds. It is in sharp contrast to the well-mixed/ventilated PBL in the southern plains, especially in the
19 afternoon. At night, this meteorological discontinuity becomes less pronounced. The diurnal variation of
20 the PBL thermal-dynamic structure causes the pollutants to concentrate at the foot of the mountains

21 during the daytime and locally accumulate throughout the entire plain in the evening. These results
22 provide a more complete mesoscale view of the PBL structure and highlight its spatial heterogeneity,
23 which promotes the understanding of air pollution at the regional scale.

24 **Keywords:** Boundary layer structure; atmospheric internal boundaries; PM_{2.5}; modeling

25 **1 Introduction**

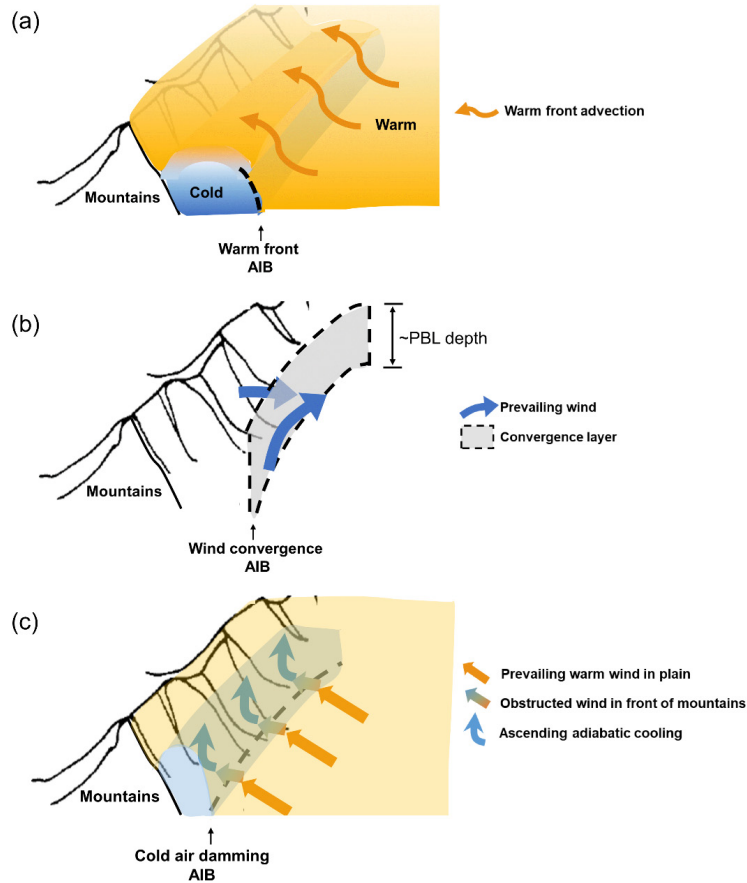
26 The planetary boundary layer (PBL) is the lowest section of the atmosphere that responds directly
27 to the heat and friction from the Earth's surface (Stull, 1988; Garratt, 1992). Most air pollutants are
28 intensively emitted or chemically produced within this layer, and their horizontal transport and vertical
29 mixing are affected by the dynamic flow and thermal stability of the PBL (Tennekes, 1974). Therefore,
30 the PBL structure plays a crucial role in the evolution, magnitude and distribution of air pollution.

31 The PBL structure has been recognized to be strongly dependent on three categories of factors: (i)
32 the single-column vertical property (such as turbulence intensity) forced by the local surface's energy
33 balance; (ii) the lateral-section horizontal variation of wind, temperature and humidity regulated by the
34 mesoscale meteorological process and (iii) the three-dimensional spatial evolution controlled by the
35 large-scale synoptic system (Boutle et al., 2010). The local vertical PBL structure and its impact on air
36 pollution have been widely discussed from different aspects including turbulent mixing (Emeis and
37 Schafer, 2006; Ren et al., 2019), dynamic effect (Dupont et al., 2016), entrainment (Li et al., 2018; Jin et
38 al., 2020), and radiative feedback with aerosol (Petaja et al., 2016). In these studies, the PBL height at a
39 certain site has been the most commonly used indicator to analyze the correlation with pollutant
40 concentration, whether from the time scale of the diurnal cycle, daily variation, or longer period (Bianco
41 et al., 2011; Liu et al., 2019; Miao and Liu, 2019). Moreover, some studies investigate the PBL spatial
42 structure under the large-scale force of weather systems (Prezerakos, 1998; Boutle et al., 2010; Mayfield
43 and Fochesatto, 2013). Sinclair et al. (2010) report the three-dimensional PBL structure developed
44 beneath an idealized mid-latitude weather system, which is characterized by a deep convective PBL in
45 the eastern flanks of the anticyclone and a shallow shear-driven PBL in the cyclone's warm sector. The
46 effect of the monsoon trough on the PBL has also been indicated, showing relatively low PBL capped by
47 a stable layer in the western end of the trough line, while a well-defined deep moist layer with active
48 thermal instability in the eastern end (Rajkumar et al., 1994; Narasimha, 1997; Potty et al., 2001). In
49 recent years, synoptic classification has been used to explore the role of different weather circulations on
50 PBL structure and to further analyze air pollution (Peng et al., 2016; Xiao et al., 2020). The movement
51 of the synoptic systems makes the shallow and deep boundary layers develop alternately in a certain area,
52 regulating the periodic evolution of large-scale air pollution.

53 As the intermediate scale, mesoscale systems interact with PBL in more direct and complex ways,

54 since they occur in the lower troposphere with vertical extension comparable with the PBL depth and
55 horizontal scale close to the regional range. Discontinuity of meteorological properties inside and outside
56 these systems presents as atmospheric internal boundaries (AIBs) in the lateral direction, usually
57 manifested as temperature contrast and/or wind shift. Previous studies have emphasized their influence
58 on the initiation of convective storms (Sanders and Doswell, 1995; Hane et al., 2002; Bluestein, 2008).
59 On the other side, as internal lateral boundaries within the low-level atmosphere, the AIBs can lead to
60 the abrupt change of the PBL spatial structure, which is of particular importance to the evolution of
61 regional pollution. The effects of mesoscale sea-land and mountain-valley circulations on the PBL have
62 been clarified, i.e., the thermal internal boundary layer in the coastal area and the depressed PBL close
63 to a mountain base (Garratt, 1990; Lu and Turco, 1995; Talbot et al., 2007; De wekker, 2008; Miao et al.,
64 2015). Some studies discuss the PBL structure under the rule of other types of mesoscale/sub-synoptic
65 scale systems, such as the persistent cold-air pools in the Salt Lake valley (Lareau et al., 2013), foehn
66 winds in the Eastern Alps (Seibert, 1990; Baumann et al., 2001), and leeside troughs and cold-air
67 damming around the Appalachian mountains (Seaman and Michelson, 2000; Bell and Bosart, 1988), as
68 well as the frequent cold and warm fronts in Europe (Berger and Fricke, 1995; Sinclair, 2013). However,
69 there needs more understanding of their impact on the evolution of air pollution.

70 The North China Plain (NCP) is one of the most polluted areas in the world. The dense population
71 and developed industries produce intensive emissions in this region, with most sources located in the
72 plain area and less in the northern and western mountains (their spatial distribution is presented in the
73 supplement material). High-intensity primary emissions are the fundamental cause of air pollution, which
74 directly releases pollutants into the atmosphere and provides precursors for secondary aerosol formation
75 (Lyu et al., 2016; Zhao et al., 2019). In order to improve the air quality, a series of stringent emission
76 reduction policies are implemented from 2013, which make the annual mean PM_{2.5} concentrations
77 decrease by 32% in 2017 (Zhang et al, 2019). However, the severe polluted days still occur frequently,
78 especially in winter (Zhang et al., 2018). During these pollution episodes, adverse meteorological
79 conditions are the dominant factors causing high pollution levels and various spatial patterns, as there
80 are no significant changes in emissions in a short period (e.g., weeks). Extensive studies have been
81 conducted to investigate the meteorological causes of regional pollution in the NCP, such as the local
82 meteorological factors and large-scale synoptic process (Ye et al., 2016; Ren et al., 2019; Li et al., 2020).
83 Nevertheless, the knowledge about the PBL spatial structures under the impact of the mesoscale AIBs is
84 still insufficient, and the role of the special PBL structures plays in the air pollution evolution at a regional
85 scale is even unclear (Bluestein, 2008; McNider and Pour-Biazar, 2020).



86
 87 Figure 1. Schematic diagram showing the conceptual model of PBL spatial structures under three
 88 pollution categories. (a) Frontal category: the blue-shaded and orange-filled areas represent the isolated
 89 and stable cold air mass ahead of the warm front and the warmer well-mixing atmosphere behind the
 90 front. The orange arrows indicate warm front advection. (b) Wind shear category: two blue arrows
 91 represent the airflows ahead of and behind the trough. The gray-filled area indicates the dynamic
 92 convergence layer with a depth comparable to the boundary layer height. (c) Topographic obstruction
 93 category: the light blue filled area indicates the cold air damming at the foot of the windward mountains.
 94 Terrain obstruction disrupts the geostrophic balance so that the southerly warm advection weakens (long
 95 orange arrows) and turns to the easterly cold advection (short gradient-color arrows), and meanwhile, the
 96 air mass accumulates to produce a lift cooling (up blue arrows). Black dashed lines in (a-c) indicate the
 97 warm front AIB, wind convergence AIB, and cold air damming AIB respectively. The PBL spatial
 98 structure under the first category has been revealed by Jin et al. (2021). For the latter two categories, their
 99 PBL three-dimension structures are discussed in Sect 3.3 in this paper.

100 Based on the surface observations, a thorough survey of the PM_{2.5} pollution categories under the
 101 control of the AIBs is carried out by Jin et al. (2022, submitted). It is found that the pollution formation-
 102 maintenance process in the NCP can be classified into three categories, i.e., the frontal category, wind
 103 shear category and topographic obstruction category during the autumn and winter of the investigated 7

104 years (2014–2020). Figure 1 shows the schematic diagram of three pollution categories corresponding to
105 various AIBs. The frontal category represents about 41 % of all 98 pollution episodes, and its PBL spatial
106 structure has been revealed in a previous case study (Jin et al., 2021). It is characterized by an isolated
107 cold air mass, which is laterally confined by mountains and warm front AIB, and vertically covered by a
108 warm dome (Fig. 1a). The strong elevated inversion depresses the PBL height abruptly to 200~300 m
109 within the cold area in contrast to 600~800 m outside the zone, constituting adverse dispersion conditions
110 and resulting in the most serious PM_{2.5} pollution. The wind shear category is associated with airflow
111 convergence AIB (Fig. 1b), which is dominated by dynamic effect and causes lighter PM_{2.5} pollution.
112 West-southwest wind shear and south-north wind shear are the two main modes. The third category
113 occurs when the airflow cannot cross the topographic obstruction and form the cold air damming AIB. A
114 cold and heavy pollution belt develops at the foot of the windward mountains (Fig. 1c), under the
115 synergistic effect of dynamical obstruction and thermal stratification. Although previous studies have
116 classified the air pollution and revealed the spatial characteristics of the first category, the three-
117 dimensional PBL structures interacted with AIBs under the other two categories are not yet clarified,
118 which is responsible for 43% of pollution episodes in the NCP. In order to fulfill this knowledge gap, the
119 present study deeply analyzes representative cases of wind shear category and topographic obstruction
120 category (Detailed analyses in Sect 3.3), and finally provides a complete conceptual model of the PBL
121 spatial structure in the NCP under various pollution categories and corresponding AIBs (Fig.1).

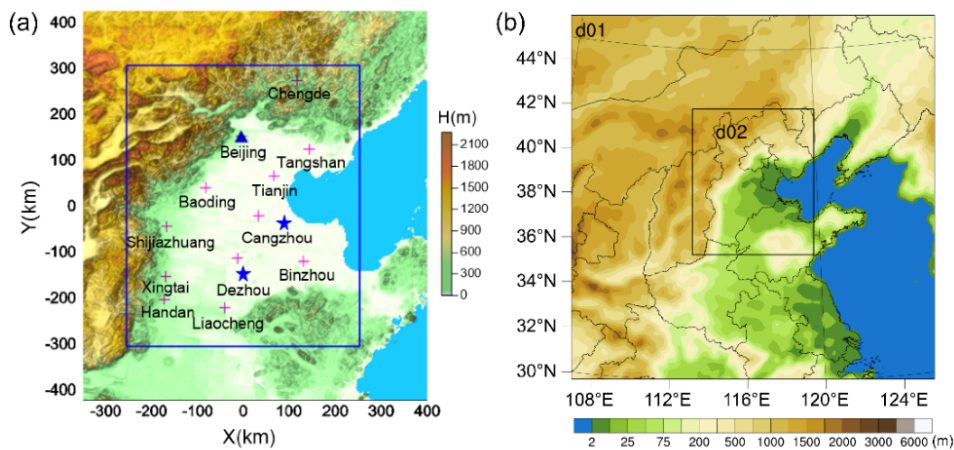
122 The mesoscale meteorological models, such as the Weather Research and Forecast (WRF) with the
123 high spatial and temporal resolution, are plausible tools to capture the mesoscale systems and display
124 detailed spatial structures in the lower atmosphere, including the AIBs and the PBL (Jimenez et al., 2016,
125 Pielke and Uliasz, 1998; Seaman, 2000; Hanna and Yang, 2001; McNider and Pour-Biazar, 2020). The
126 present study aims to reveal the thermal and dynamic structures of the PBL and their evolution associated
127 with different AIBs in the condition of pollution episodes, by using the WRF model. For this purpose,
128 the model performance is at first evaluated with detailed sounding data from intensive experiments, to
129 ensure the model's ability in reproducing the meteorological fields and their three-dimensional structures
130 in the concerned region. The article is organized as follows. The following section describes the PBL
131 sounding observations as well as the WRF model settings. Section 3 provides an overview of
132 representative pollution cases and the evaluation of the model performance. Furthermore, the PBL spatial
133 structures under various pollution categories are analyzed. Finally, the conclusions are presented and the
134 uncertainty of the mesoscale meteorological model is discussed in Sect. 4.

135 **2 Data and methods**

136 **2.1 Observations and data analysis**

137 **Intensive GPS (Global Positioning System) sounding data:** Two periods of field experiments
138 were carried out to evaluate the meteorological model and explore wintertime PBL structure in the NCP:
139 at Cangzhou (38°13' N, 117°48' E, Fig. 2a) from January 8 to 28, 2016 and at Dezhou (37°16' N, 116°43'
140 E, Fig. 2a) from December 25, 2017, to January 24, 2018. GPS radiosonde (Beijing Changzhi Sci & Tech
141 Co. Ltd., China) was used to obtain profiles of wind speed, wind direction, temperature, and relative
142 humidity with a vertical resolution of approximately 1 s (3~5 m). Eight soundings were taken on each
143 day, at 0200, 0500, 0800, 1100, 1400, 1700, 2000 and 2300 LT (i.e., Local Time = Universal Time
144 Coordinated + 8). The reliability of the GPS sounding data has been systematically evaluated by Li et al.
145 (2020) and Jin et al. (2020, 2021).

146 **Routine radiosonde sounding data:** Routine sounding data from the meteorological station of
147 Beijing (39°56' N, 116°17' E, Fig. 2a) were collected during October 7–12, 2014, in the absence of
148 intensive PBL observation. The data were obtained from Wyoming University, USA
149 (<http://weather.uwyo.edu.html>), and the original observation data with higher vertical resolution were
150 provided by the China Meteorological Administration. The routine soundings were taken 2 times a day,
151 at 0800 and 2000 LT.



152
153 Figure 2. Geographical map of the (a) observation area and (b) WRF model domain. Intensive GPS
154 soundings at Dezhou and Cangzhou (star), routine radiosonde sounding at Beijing (triangle) and air
155 quality stations (plus) are indicated in (a). The rectangle in (a) is the same as the model inner domain d02
156 in (b).

157 **PBL height and vertical profiles:** During the two periods of intensive field experiments, 160 and
158 240 datasets were collected at these two sites, including vertical profiles of temperature, relative humidity,
159 wind speed, and wind direction. We carried out quality control on the original sounding data and

160 eliminated outliers and then calculated the profiles of potential temperature. All the profiles were
161 smoothed by a three-point moving average method and were interpolated to obtain a vertical resolution
162 of 10 m. The PBL height was derived via the potential temperature profile method and the detailed
163 calculation followed the mathematical method established by Liu and Liang (2010). Sounding data were
164 used to evaluate the model performance and to analyze the three-dimensional thermal and dynamic
165 spatial structure of the PBL.

166 In addition to the PBL sounding data, the routine meteorological observation and air quality
167 monitoring data were used to obtain the surface meteorological field and pollutant concentration field.
168 The spatial distributions of sea level pressure, 10 m wind vector, potential temperature, and the
169 corresponding PM_{2.5} concentration were obtained by data interpolation or diagnostic model, details of
170 the methods were referred to Jin et al. (2021).

171 **2.2 Model simulations**

172 The WRF model was used to investigate the vertical and horizontal structures of the PBL. Two
173 nested domains (Fig. 2b) were employed with horizontal grid resolutions of 15 and 5 km. Each domain
174 had 37 vertical layers extending from the surface to 100 hPa, with 25 layers within 2 km (with the
175 respective height of about 9 m, 25 m, 50 m, 85 m, 120 m, 160 m, 200 m, 240 m, 290 m, 350 m, 420 m,
176 500 m, 580 m, 660 m, 740 m, 820 m, 900 m, 980 m, 1080 m, 1200 m, 1350 m, 1550 m, 1700m, 1850 m,
177 and 2000 m) to resolve the PBL structure. The meteorological initial and boundary conditions were set
178 using the United States National Center for Environmental Prediction Final Analysis (NCEP-FNL)
179 dataset. The physics parameterization schemes applied in this study were the same as Jin et al. (2021).

180 **2.3 Representative cases**

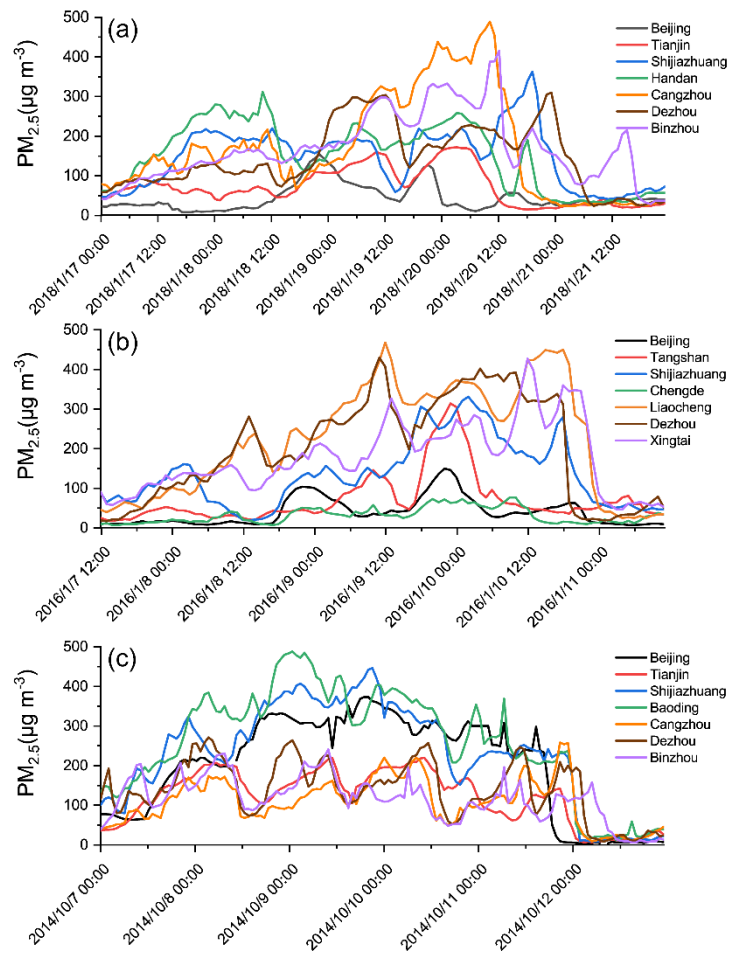
181 As mentioned above, PM_{2.5} pollution episodes in the NCP are identified in the frontal category, wind
182 shear category, and topographic obstruction category, according to their association with the mesoscale
183 AIBs (Jin et al., 2022 submitted). The present study tries to reveal the PBL structures modified by the
184 AIBs under various pollution categories. Among them, the first category has been investigated previously
185 (Jin et al., 2021). We focus on the representative cases under the other two categories in this paper. For
186 the wind shear category, there are two main shear modes: west-southwest wind shear and south-north
187 wind shear. Therefore a total of three typical cases are selected to respectively represent these two
188 pollution categories, i.e., Case-1 for west-southwest wind shear mode: during January 17–21, 2018;
189 Case-2 for south-north wind shear mode: during January 7–11, 2016; and Case-3 for topographic
190 obstruction category: during October 7–12, 2014. The temporal and spatial evolution of their PM_{2.5}
191 concentrations and the corresponding surface meteorological conditions would be analyzed based on
192 routine observations, and their PBL spatial structures would be revealed by the WRF model simulations.

193 3 Results

194 3.1 Basic features of the cases

195 The surface observations for these three cases are presented firstly. According to the temporal
196 evolution of PM_{2.5} concentration at different stations in the NCP (Fig. 3), all of these three pollution
197 episodes went through the stages of formation, maintenance and diffusion. As shown in Fig. 3a, Case-1
198 was characterized by two main concentration peaks (300 μg m⁻³ at Handan vs 500 μg m⁻³ at Cangzhou)
199 in the formation-maintenance stage (January 17–20, 2018), with the latter being higher than the former.
200 From noon on January 20, 2018, pollution in Tianjin-Cangzhou-Shijiazhuang diffused successively and
201 all sites reached a clean level on the afternoon of January 21, 2018. For Case-2, the pollution formed in
202 the first two days, maintained over the next day and was cleaned on the night of January 10, 2016 (Fig.
203 3b). The southern sites such as Liaocheng and Dezhou were the most polluted (reaching 450 μg m⁻³) and
204 the northern cities such as Beijing and Chengde were the least polluted (less than 150 μg m⁻³). Pollution
205 in Case-3 experienced the formation process on October 7–8, 2014, maintained for the successive three
206 days, and ended on October 12, 2014 (Fig. 3c). During this period, the piedmont sites (Baoding, Beijing
207 and Shijiazhuang) kept always a high concentration regardless of day and night (about 400 μg m⁻³), while
208 the southeast sites (Binzhou, Dezhou and Cangzhou) had lighter pollution and obvious diurnal cycle
209 (lower than 250 μg m⁻³).

210 The spatial patterns of PM_{2.5} pollution, from the formation (Fig. 4i), maintenance (Fig. 4ii-iv), to
211 the diffusion stage (Fig. 4v), are illustrated for each case. In the formation stage, the polluted air mass of
212 Case-1 and Case-3 built up along the mountains from the southwest of the NCP, with the latter being
213 more concentrated and the former spreading southwestward (Fig. 4a-i, c-i). While the pollution in Case-
214 2 first developed from the south (Fig. 4b-i). During the pollution maintenance process, Case-1 was
215 featured with widespread PM_{2.5} flooding the NCP, making the eastern region gradually covered by heavy
216 pollution (Fig. 4a, ii-iv); in Case-2, a polluted air mass has been advancing northward with a clear edge,
217 but it did not reach the northern mountainous area (Fig. 4b, ii-iv); the spatial distribution of PM_{2.5} of
218 Case-3 was characterized by the day-night contrast, manifested as pollution filling the entire plain area
219 at night while concentrating in front of the mountains with a distinct edge on the southeast side during
220 the daytime (Fig. 4c, ii-iv). Finally, these pollution cases were diffused in different ways. In Case-1, the
221 clean air first occupied the northern parts of the NCP with a large concentration gradient on the front
222 edges (Fig. 4a, v). As for Case-2, PM_{2.5} was restored to a clean level from the northeast (Fig. 4b, v).
223 Pollution in the northwest was earliest removed in Case-3, with Beijing acting like a
224 loophole/passageway in the cleaning process (Fig. 4c, v). These cases presented various pollution
225 distributions, however, all of them were characterized by clear edges or distinct heavy pollution cores.



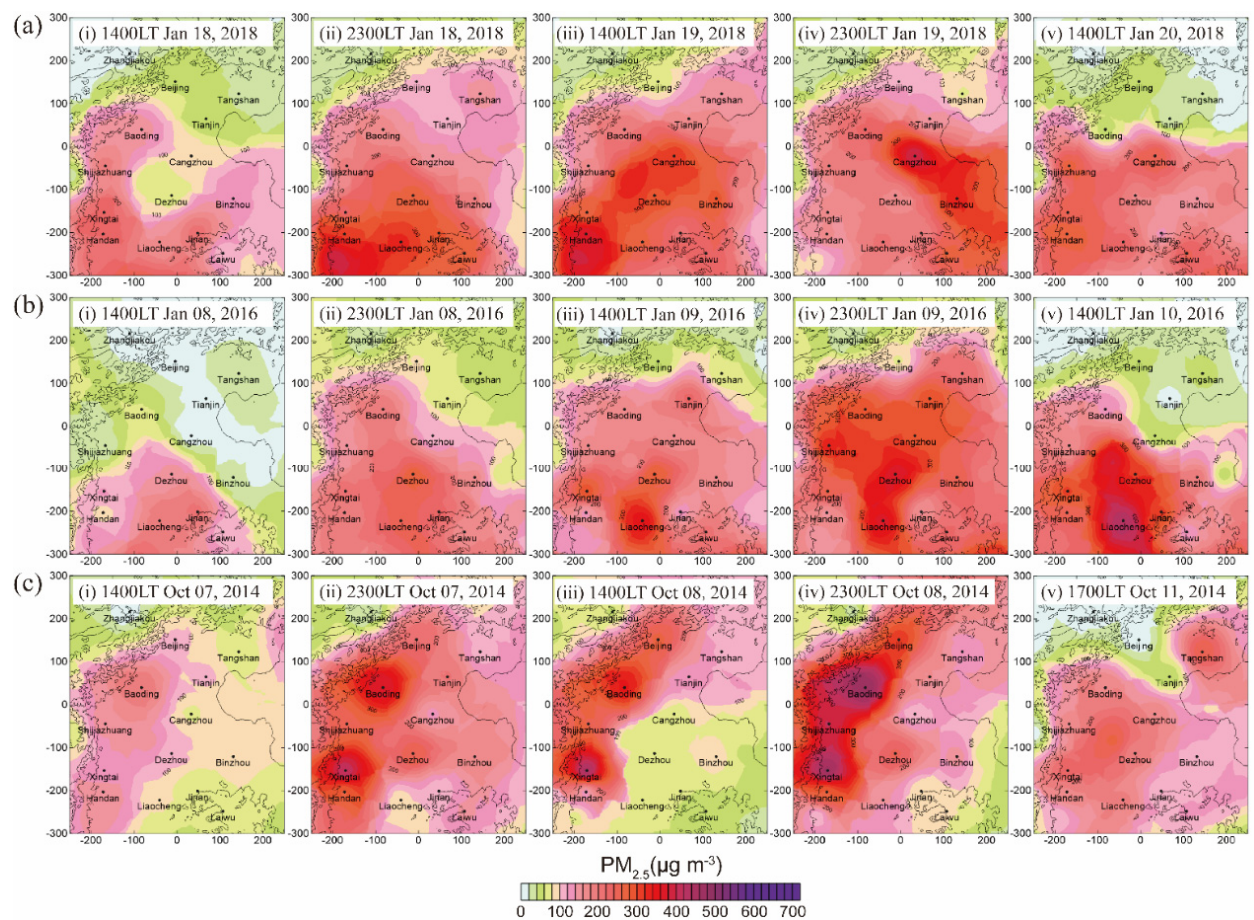
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227 Figure 3. Temporal evolution of $PM_{2.5}$ concentrations during Case1–3, respectively represent (a) west-

228 southwest wind shear mode (January 17–21, 2018), (b) south-north wind shear mode (January 7–11,

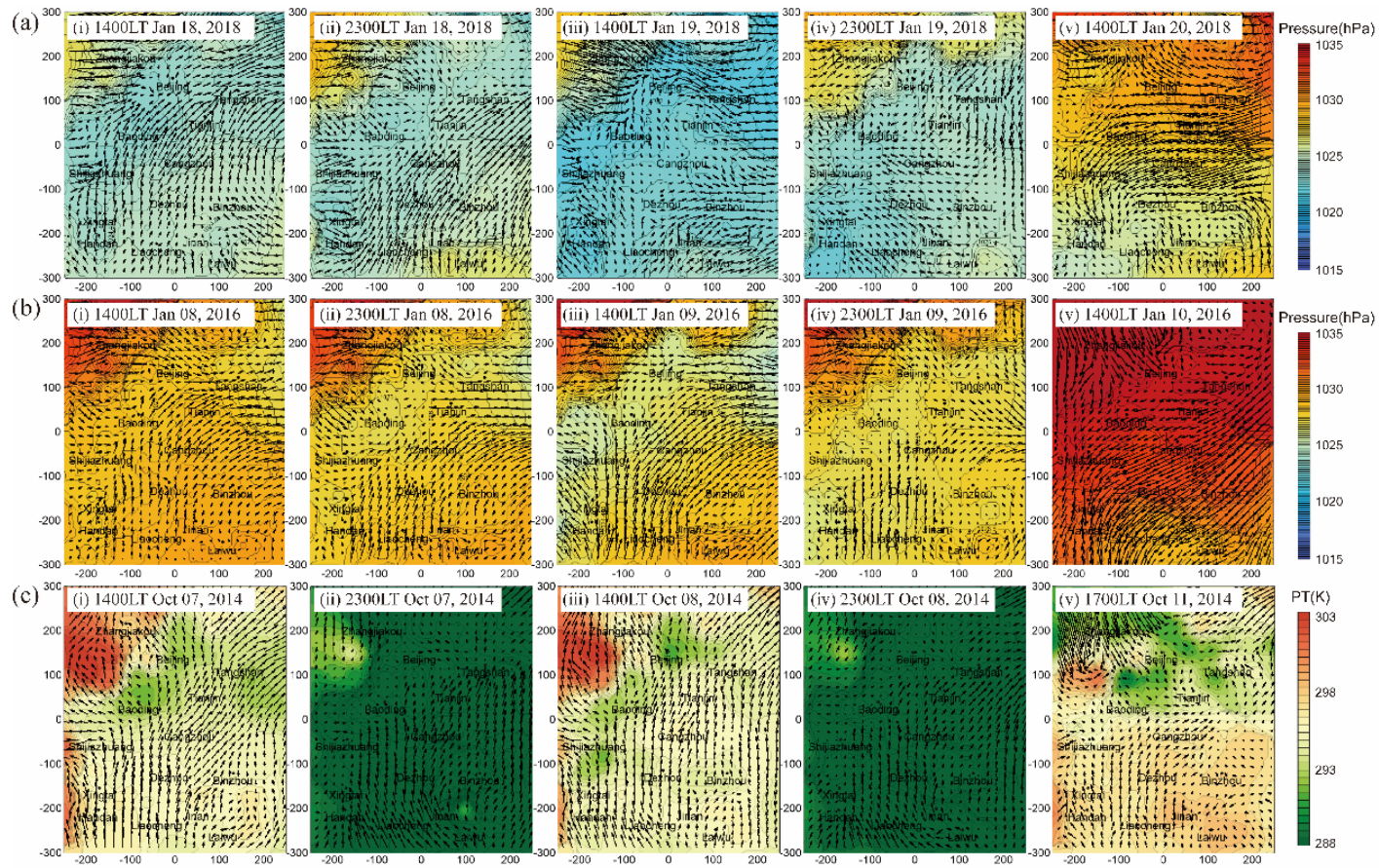
229 2016), and (c) topographic obstruction category (October 7–12, 2014). The locations of these $PM_{2.5}$

230 stations are marked in Fig. 2a.



231

232 Figure 4. Spatial distributions of observed surface $PM_{2.5}$ concentrations (shaded colors) at the pollution stages of (i) formation, (ii-iv) maintenance, and (v) diffusion during
 233 representative Case1–3 under (a) west-southwest wind shear mode, (b) south-north wind shear mode, and (c) topographic obstruction category. Values shown in x- and y-axis
 234 denote the distances (km) to the domain center. The $PM_{2.5}$ concentration fields are derived from spatial interpolation of pollution observed data at monitoring stations.



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237

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Figure 5. Observed sea level pressure/potential temperature and wind vectors at the pollution stages of (i) formation, (ii-iv) maintenance, and (v) diffusion during representative Case1-3 under (a) west-southwest wind shear mode, (b) south-north wind shear mode, and (c) topographic obstruction category. The shaded colors represent the sea level pressure in (a-b) and the potential temperature in (c). The arrows indicate wind vectors. Values shown in x- and y-axis denote the distances (km) to the domain center.

239 The correspondent surface meteorological fields of the three cases are shown in Fig. 5. Case-1 and
240 Case-2 are the two main modes of wind-shear category, for which dynamic AIB plays a dominant role,
241 and thus the observed sea level pressure and wind fields are discussed (Fig. 5a-b). Case-3 belongs to the
242 topographic obstruction category affected by the AIB created by the cold air damming, and its potential
243 temperature and wind fields are displayed to focus on the combined action of the thermal and dynamic
244 properties (Fig. 5c). As shown in Fig. 5a, i-iii, the pollution formation and maintenance processes of
245 Case-1 were dominated by a leeward trough, which induced the westerly airflow shear to the southwest
246 wind and produced a convergence belt at the trough axis. As the trough broadened and moved eastward,
247 the wind convergence zone also moved (Fig. 5a, i-iii). On the evening of January 19, 2018, the leeward
248 trough temporarily evolved into an inverted trough under the force of the approaching high-pressure,
249 creating a cyclonic convergence (Fig. 5a, iv). This explains why the heavy pollution expands eastward
250 in this episode (refer to Fig. 4a, i-iv). Until January 20, 2018, a high-pressure system invaded the NCP
251 from the northeast, bringing strong northeast winds (Fig. 5a, v), which made the pollution disperse
252 southward in turn (refer to Fig. 4a, v). During Case-2, a saddle-shaped pressure field persisted in the
253 pollution formation-maintenance stage and induced the prevailing northerly winds in the northern NCP
254 against the dominant southerly flows in the southern area (Fig. 5b, i-iv). As a result, the polluted air mass
255 was prevented from advancing northward to the mountains, causing a strong contrast in pollution
256 concentration between the northern and southern parts of the domain (refer to Fig. 4b i-iv). Its pollution
257 diffusion process was also associated with a northeast high-pressure system, by strong northeasterly
258 airflows cleaning up the PM_{2.5} (Fig. 5b, v). As for the Case-3 under the topographic obstruction category,
259 there was a narrow area with low potential temperature and weak southerly wind at the foot of the
260 mountains on the windward side in the daytime, but this feature became fuzzy at night (Fig. 5c, i-iv).
261 This diurnal variation repeatedly occurred during the formation and maintenance stage, which
262 corresponded excellently to the day-night difference in pollution distribution (refer to Fig. 4c i-iv). In the
263 end, the strong flows and cold air bursting like a jet stream through a pathway across Zhangjiakou-
264 Beijing-Tianjin (Fig. 5c, v), made pollutants begin to be swept out from the northwest (refer to Fig. 4c,
265 v).

266 3.2 Evaluation of simulated meteorological field

267 To reveal the PBL three-dimensional structure of these representative cases, numerical simulations
268 are conducted using the WRF model. It is necessary to evaluate the model reliability before analyzing
269 the simulated results. The model-observation comparisons in the previous studies usually focus on the
270 time series of surface meteorological elements, such as 10 m wind speed and direction, 2 m temperature
271 and humidity (Rogers et al., 2013; Bei et al., 2018; Qu et al., 2021). The model performance of their
272 spatial fields is often ignored, and the simulated PBL vertical structure is rarely evaluated. But the

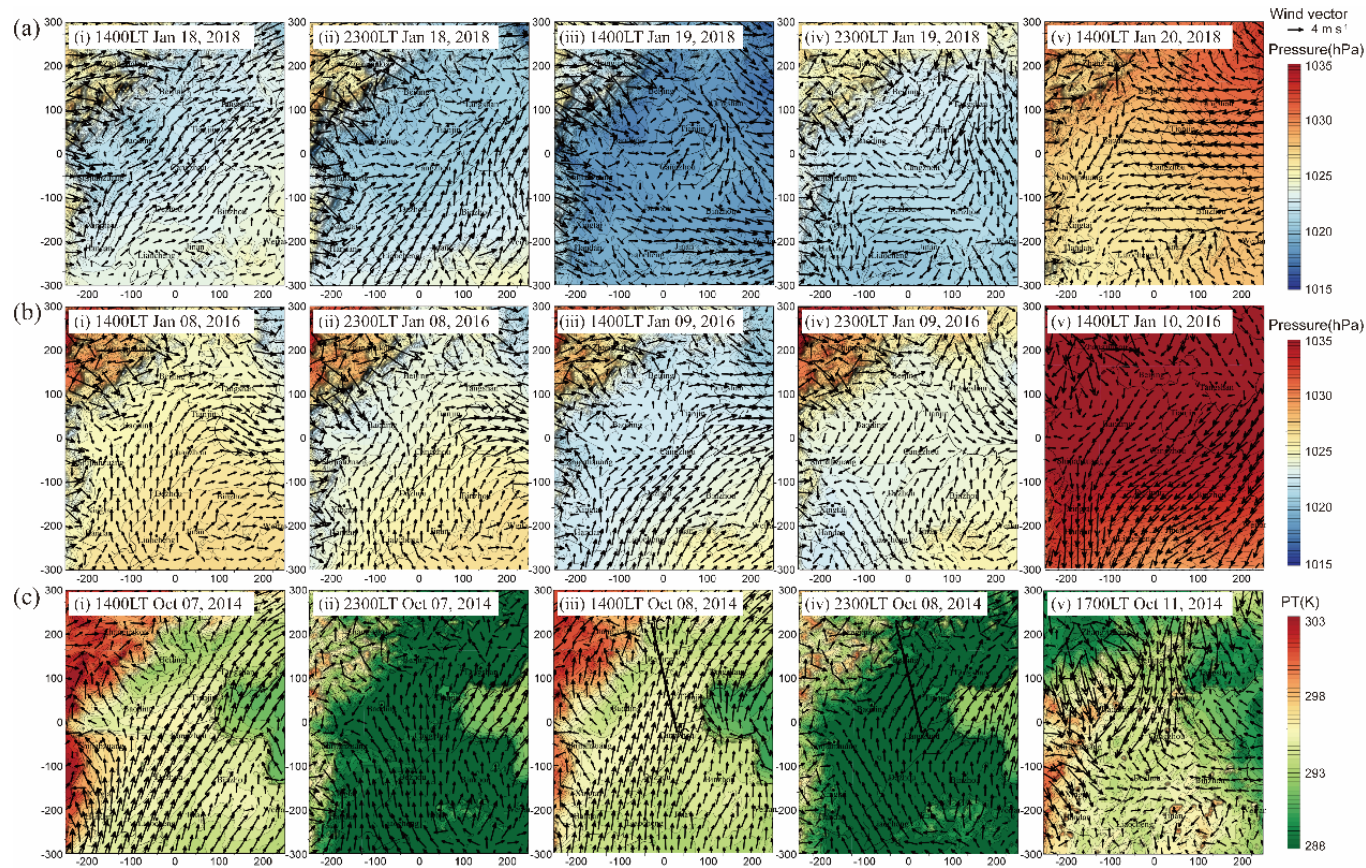
273 regional distribution and vertical structure of wind, temperature, and humidity are crucial for air pollution.
 274 And of course, the PBL height is a key parameter in characterizing air pollution ventilation conditions.
 275 In this study, the evaluation is carried out from three perspectives: i) the temporal evolution and ii) the
 276 spatial pattern of near-surface potential temperature and wind speed, as well as iii) the vertical-temporal
 277 structure of these two variables at the sounding sites, in addition to the temporal variation of PBL height.

278 For the temporal evolution of the near-surface potential temperature and wind speed, the hourly
 279 observations and simulations of 13 key cities (Beijing, Tianjin, Shijiazhuang, Baoding, Handan,
 280 Tangshan, Cangzhou, Dezhou, Jinan, Weifang, Binzhou, Chengde and Zhangjiakou) evenly distributed
 281 in the NCP are compared during these three pollution cases. The model outputs are extracted from the
 282 grid points nearest to the observed sites. As shown in Table 1, the correlation coefficients of the simulated
 283 and observed hourly evolution of potential temperature and wind speed are 0.80~0.91 and 0.54~0.64
 284 ($p < 0.01$), respectively. In order to exclude the influence of the diurnal cycle on the correlation, the daily
 285 averages are also calculated and the obtained correlation coefficients are as high as 0.65~1 and 0.62~1
 286 ($p < 0.01$) for potential temperature and wind speed, respectively (Table S1). The statistical results
 287 demonstrate that the major variations in the time series of the surface observations are reproduced well
 288 by the model, which has also been recognized in previous studies (Rogers et al., 2013; Bei et al., 2018;
 289 Qu et al., 2021).

290 Table 1. Statistics of model performance for the hourly evolution of near-surface potential temperature
 291 and 10 m wind speed for selected 13 cities during the representative cases.

| | Case-1 | | | | Case-2 | | | | Case-3 | | | |
|----------------|-------------|-------------|-------------------------|-------------|-------------|-------------|-------------------------|-------------|-------------|-------------|-------------------------|-------------|
| | PT (K) | | WS (m s ⁻¹) | | PT (K) | | WS (m s ⁻¹) | | PT (K) | | WS (m s ⁻¹) | |
| | R | RMSE | R | RMSE | R | RMSE | R | RMSE | R | RMSE | R | RMSE |
| Beijing | 0.80 | 2.20 | 0.62 | 1.15 | 0.87 | 2.60 | 0.61 | 1.69 | 0.91 | 2.20 | 0.73 | 1.65 |
| Tianjin | 0.89 | 2.40 | 0.66 | 1.48 | 0.85 | 1.90 | 0.63 | 1.97 | 0.92 | 2.10 | 0.61 | 2.13 |
| Shijiazhuang | 0.77 | 2.80 | 0.52 | 2.02 | 0.82 | 2.50 | 0.66 | 1.69 | 0.88 | 2.20 | 0.58 | 1.95 |
| Baoding | 0.83 | 2.50 | 0.60 | 1.34 | 0.85 | 2.40 | 0.61 | 1.53 | 0.89 | 2.30 | 0.60 | 1.97 |
| Handan | 0.93 | 1.40 | 0.48 | 1.36 | 0.78 | 3.20 | 0.56 | 2.27 | 0.95 | 1.30 | 0.66 | 1.94 |
| Tangshan | 0.69 | 4.00 | 0.62 | 1.44 | 0.81 | 3.30 | 0.53 | 1.64 | 0.85 | 3.00 | 0.46 | 2.24 |
| Cangzhou | 0.85 | 3.00 | 0.64 | 1.23 | 0.79 | 2.50 | 0.60 | 1.92 | 0.94 | 2.10 | 0.75 | 1.45 |
| Dezhou | 0.78 | 3.70 | 0.51 | 1.69 | 0.87 | 1.50 | 0.63 | 2.82 | 0.90 | 2.30 | 0.55 | 2.97 |
| Jinan | 0.76 | 2.80 | 0.49 | 2.96 | 0.74 | 2.40 | 0.63 | 2.45 | 0.91 | 2.10 | 0.56 | 3.10 |
| Weifang | 0.79 | 2.10 | 0.53 | 1.42 | 0.78 | 2.50 | 0.71 | 1.99 | 0.94 | 2.10 | 0.85 | 1.40 |
| Binzhou | 0.81 | 2.50 | 0.51 | 1.97 | 0.83 | 2.30 | 0.86 | 1.29 | 0.92 | 2.00 | 0.81 | 1.47 |
| Chengde | 0.75 | 5.10 | 0.47 | 2.06 | 0.63 | 6.50 | 0.47 | 2.60 | 0.84 | 3.70 | 0.56 | 1.74 |
| Zhangjiakou | 0.90 | 5.40 | 0.33 | 2.23 | 0.77 | 5.30 | 0.47 | 3.13 | 0.96 | 4.80 | 0.54 | 2.50 |
| Average | 0.81 | 3.07 | 0.54 | 1.72 | 0.80 | 2.99 | 0.61 | 2.08 | 0.91 | 2.47 | 0.64 | 2.04 |

292 Case-1: west-southwest wind shear mode (January 17–21, 2018); Case-2: south-north wind shear mode
 293 (January 7–11, 2016); Case3: topographic obstruction category (October 7–12, 2014).



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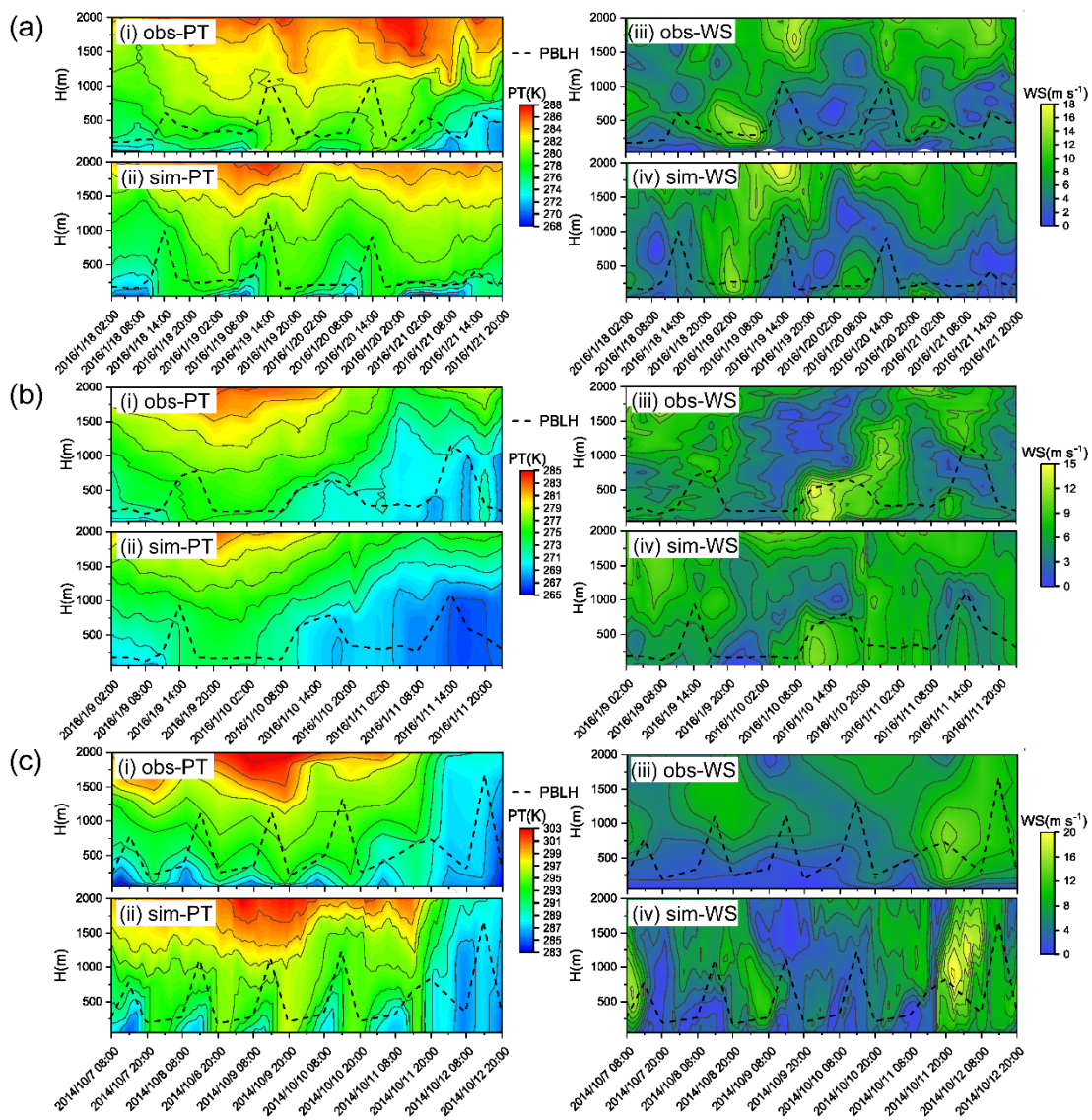
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Figure 6. Simulated sea level pressure/potential temperature and wind vectors at the pollution stages of (i) formation, (ii-iv) maintenance, and (v) diffusion during representative Case1-3 under (a) west-southwest wind shear mode, (b) south-north wind shear mode, and (c) topographic obstruction category. The shaded colors represent the sea level pressure in (a-b) and the surface potential temperature in (c). The arrows indicate wind vectors. Values shown in x- and y-axis denote the distances (km) to the domain center. Lines C_1C_1' in (c) refer to the cross-sections of the potential temperature in Fig. 11.

299 Compared with Fig. 5, the simulated surface meteorological fields during the three cases are
300 displayed in Fig. 6. In Case-1 and Case-2, the leeward trough and saddle-shaped pressure field, as well
301 as the corresponding west-southwest wind shear and south-north wind shear are reproduced in the
302 simulated fields (Fig. 6a-b, i-iv vs Fig. 5a-b, i-iv). Also, their movement and evolution during the
303 pollution formation-maintenance processes are captured by the WRF model, although there are small
304 deviations in the specific positions. At the diffusion stage, the simulated northeastern high-pressure and
305 the prevailing easterly/northeasterly winds are comparable with the observed fields (Fig. 6a-b, v vs Fig.
306 5a-b, v). As for Case-3, the modeling result of surface meteorological fields successfully reflect the
307 narrow cold zone and stagnant wind belt at the foot of the mountains, as well as their diurnal variation
308 and sustainability in the pollution formation-maintenance stage (Fig. 6c, i-iv vs Fig. 5c, i-iv). In the
309 simulation field, the cold zone is shorter at its south end on the afternoon of October 08, 2014, and there
310 is an overestimate of the potential temperature in the northwest mountains and the Bohai Sea at night. At
311 the end of this episode, a strong northerly cold airflow similar to the observation appears in the simulation
312 field (Fig. 6c, v vs Fig. 5c, v). Generally, the main features of the surface distributions of meteorological
313 observations during these three cases are reflected well in the simulated fields.

314 Moreover, the simulated and observed height-time cross sections of potential temperature and wind
315 speed, as well as the PBL height, are compared to reveal the model's ability to capture the atmospheric
316 vertical structure of each case (Fig. 7). The observation data of Case-1 and Case-2 are obtained from
317 intensive sounding experiments at the Dezhou site and Cangzhou site, respectively. While the observation
318 information during Case-3 is provided by routine soundings at the Beijing site. As for Case-1, the model
319 successfully reproduces the thermal structure evolution in the pollution formation-maintenance period,
320 while the final uplift of the inversion layer and the growth of PBL are not well captured with an
321 underestimation of about 200-300 m (Fig. 7a, i-ii). In comparison, the dynamic structures, the dominant
322 roles in this category, are simulated much better. The vertical location and temporal transition of the
323 strong and weak wind layers are comparable with observations (Fig. 7a, iii-iv). The correlation
324 coefficient (R) between simulated and observed PBL height is about 0.68 ($p < 0.01$). The model
325 performance during Case-2 is satisfactory both for cross-sections of the potential temperature and wind
326 speed. The formation and decay of upper temperature inversion and the development of the cold
327 convective PBL are consistent between observation and simulation, though there are some
328 underestimations in the modeled results (Fig. 7b, i-ii). The weak wind layer presented in the maintenance
329 stage and vertical wind shear that occurred in the diffusion stage are also captured by the model with
330 smaller gradients (Fig. 7b, iii-iv). Meanwhile, observed and simulated PBL heights show a consistent
331 evolution with a correlation coefficient as high as 0.78 ($p < 0.01$). Both of their PBL heights are lower
332 during the pollution formation-maintenance stage and increase by more than 1000 m in the diffusion
333 stage. In Case-3, the WRF reproduces the observed diurnal cycle of the potential temperature in the low-

334 level and the continuous warming at the upper layer during the formation-maintenance process, as well
 335 as the replacement of a well-mixed cold air mass in the last phase (Fig. 7c, i-ii). The evolution of the
 336 simulated wind speed is roughly similar to the observation, including the maintenance of the calm wind
 337 layer in the first four days and the appearance of the final strong wind layer (Fig. 7c, iii-iv).
 338 Correspondingly, the PBL height is characterized by typical diurnal variations during the polluted period,
 339 and begins to abruptly develop in the evening of October 12, 2014, associated with the cold air mass and
 340 strong wind, both in observation and simulation ($R=0.81$, $p<0.01$). Even so, there are some
 341 inconsistencies in the details of observation and simulation evolution, which may result from the coarse
 342 resolution of routine soundings in time and vertical direction, in addition to the uncertainties of model
 343 simulation.



344
 345 Figure 7. Observed and simulated time-height cross sections of potential temperature (left) and wind
 346 speed (right) during representative Case1–3 under (a) west-southwest wind shear mode (January 18–21,
 347 2018), (b) south-north wind shear mode (January 9–11, 2016), and (c) topographic obstruction category

348 (October 7–12, 2014). The dashed lines indicate the PBL heights. The observation data in (a-b) and in
349 (c) are obtained from intensive sounding experiments and routine soundings, respectively.

350 Overall, the model shows the ability to capture the observed mesoscale systems and atmospheric
351 thermal-dynamic structures reasonably both at the surface and in the vertical direction. With confidence
352 in the model results, we now proceed to a detailed investigation of the PBL spatial structure affected by
353 mesoscale AIBs under various pollution categories.

354 **3.3 PBL spatial structure**

355 We analyze the simulated vertical cross-sections of the mesoscale systems and AIBs to reveal the
356 three-dimensional structure of the PBL. Two key parameters, potential temperature and wind divergence,
357 are used to respectively indicate the atmospheric thermal stability and dynamic convergence, in addition
358 to another important parameter: the PBL depth. They directly affect the vertical mixing and horizontal
359 diffusion of PM_{2.5}, and are critical for pollution formation and distribution.

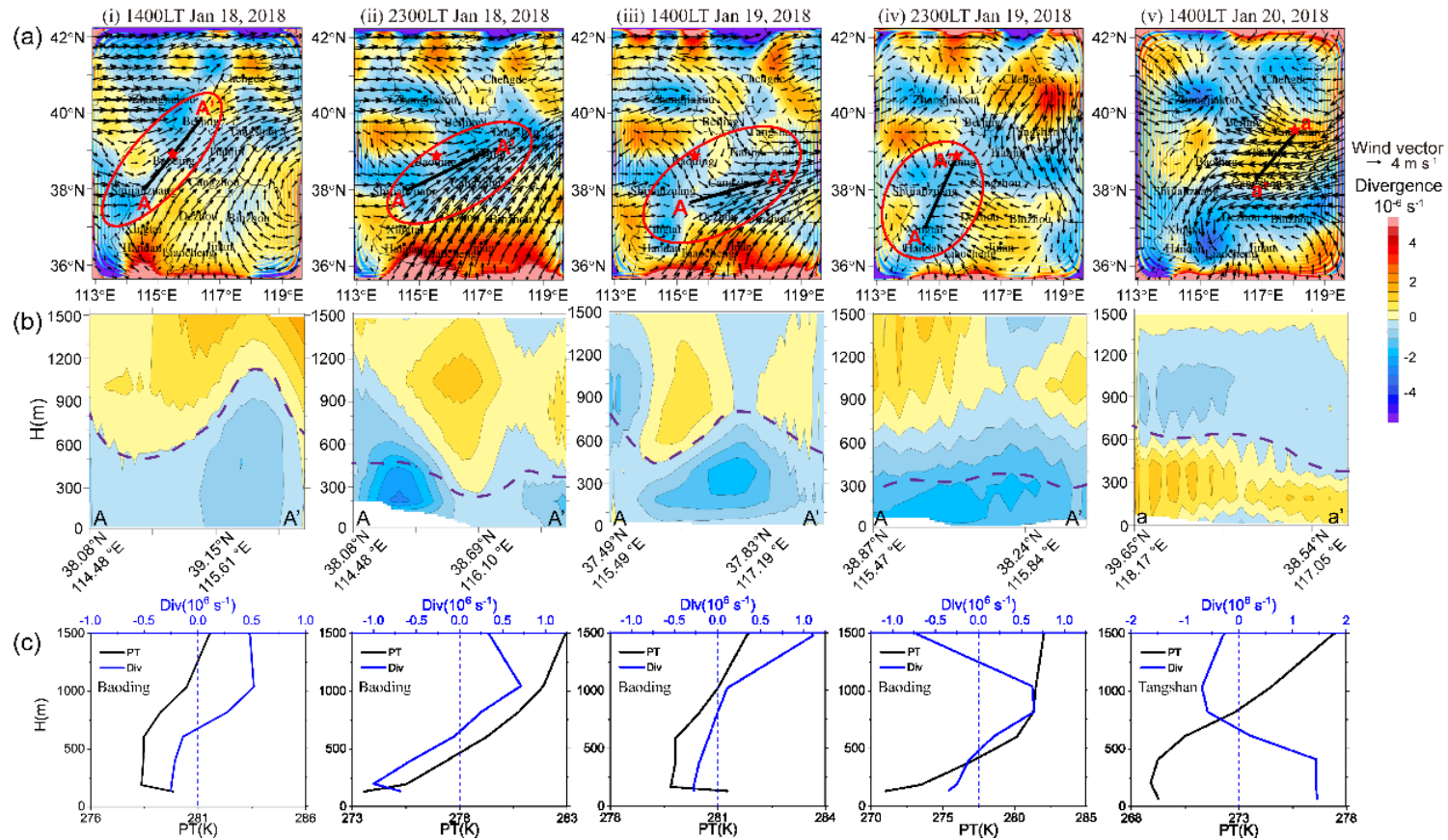
360 **Wind shear category**

361 This pollution category, mainly involving two modes of west-southwest wind shear and south-north
362 wind shear, is driven by dynamic flows. Therefore, for the corresponding Case-1 and Case-2, the wind
363 divergence sections are analyzed in detail in the following (Figs. 8-9). The potential temperature sections
364 are presented in the supplementary material (Figs. S2-3), which illustrates that there is no significant
365 thermal discontinuity.

366 Figure 8 displays the PBL dynamic structure of Case-1. During the pollution formation-maintenance
367 stage, with the establishment of a low-pressure trough, westerly winds shifted to southwesterly winds at
368 the trough axis and thus formed a convergence belt at the surface with a divergence of $-2\sim-4\times 10^{-6}\text{ s}^{-1}$
369 (Fig. 8a, i). As a consequence, a mass of pollutants were transported here and further accumulated to
370 form a pollution zone (refer to Fig. 4a, i). This trough-convergence belt continued to move to the
371 southeast, and evolved into a cyclonic-convergence center at the end of the maintenance phase (Fig. 8a,
372 i-iv). During this process, its affected area was expanded, so that the large range of NCP was filled with
373 pollutants (refer to Fig. 4a, ii-iv). The vertical section across the surface convergence belt shows that the
374 depth of the convergence layer did not exceed 1000 m, with a compensating divergence layer
375 immediately above it, being consistent with the evolution of the PBL (Fig. 8b, i-iv). Furthermore, the
376 vertical profiles of the wind divergence and potential temperature at the Baoding site located in the
377 convergence belt are extracted to illustrate the PBL dynamic structure more clearly. It shows that the
378 mutation of divergence value and the jump of potential temperature roughly appeared at the same height
379 (Fig. 8c, i-iv), which demonstrated the vertical scale of the wind convergence belt was equivalent to the
380 depth of the PBL. This phenomenon reveals that the west-southwest wind convergence caused by the
381 trough mainly occurred within the PBL, reflecting its mesoscale property. In the process of pollution

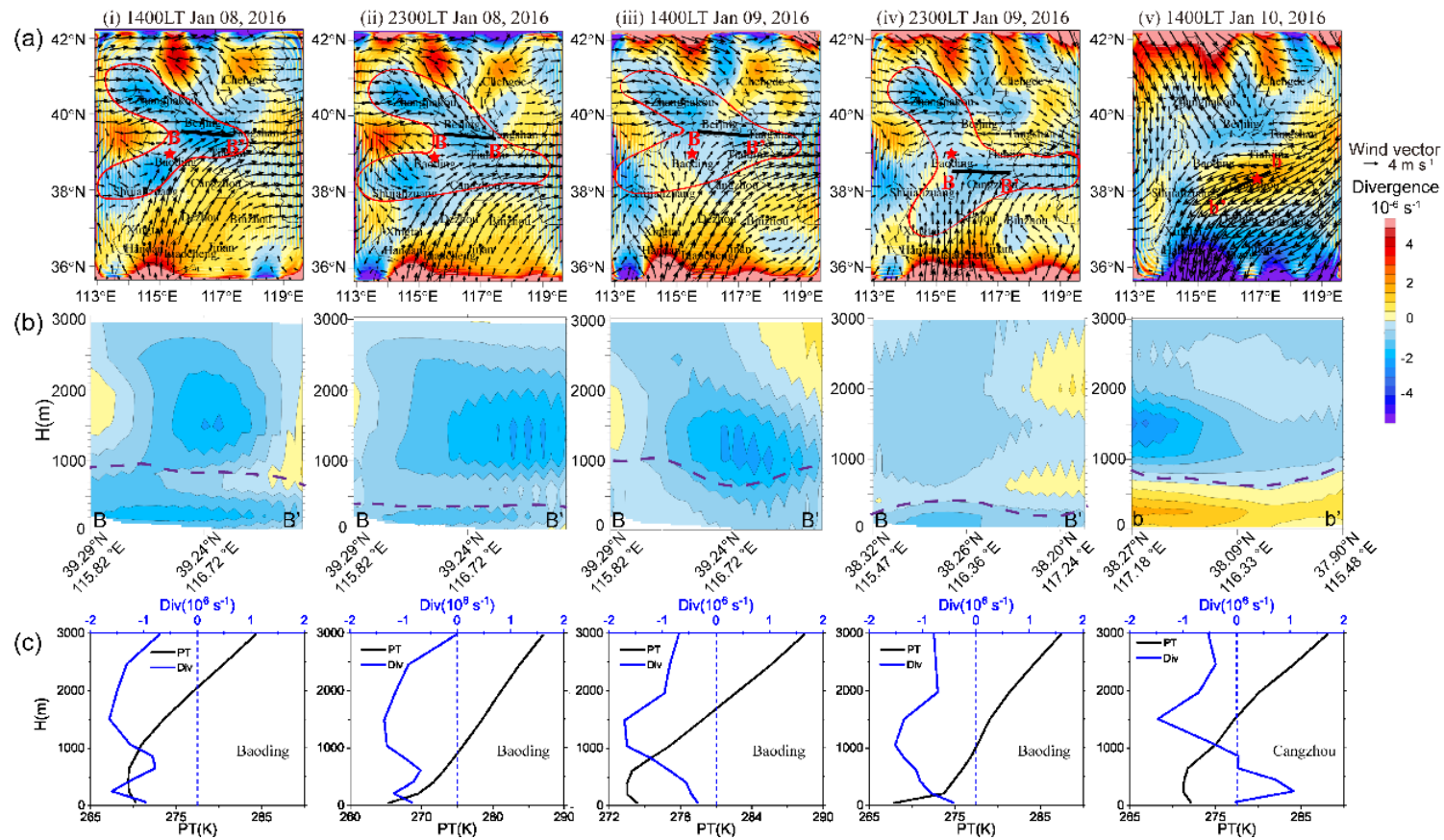
382 diffusion, with the advent of a northeast high-pressure system, divergent wind fields occurred
383 correspondently (Fig. 8a, v), which made this part of the pollutants cleaned quickly (refer to Fig. 4a, v).
384 The vertical cross-section of this divergent layer and vertical profiles at the Tangshan site show that the
385 northeast wind divergence layer was relatively thin with a thickness of no more than 600 m (Fig. 8b-c,
386 v), implying that the removal of pollutants only occurred within the PBL.

387 As for the south-north wind shear mode, the surface divergence fields displayed a "lying Y shaped"
388 convergence zone with the opening to the west during the pollution formation-maintenance stage of
389 Case-2 (Fig. 9a, i-iv), which was caused by the meeting of the southerly winds and the northerly winds
390 and then turning to the easterly winds. This convergence mode made the distribution of pollutants in a
391 pattern of much higher concentration in the south and lower in the north, with a clear edge between these
392 two air masses (refer to Fig. 4b, i-iv). Although the southerly winds in the southern NCP kept the
393 pollutants transported northward, they never reached the northernmost part due to the opposite airflow
394 there. The vertical cross-sections of this special convergence zone exhibited a depth extending upwards
395 for more than 3000 m, with a peak between 1000 m and 2000 m above the PBL top (Fig. 9b, i-iv).
396 Referring to the vertical profiles of wind divergence and potential temperature at the Baoding site, it can
397 be seen that the depth of the convergence layer far exceeded the height of the PBL, whether it was in the
398 daytime or at night (Fig. 9c, i-iv). These phenomena demonstrate that the south-north wind shear created
399 by the saddle-shaped pressure field is of much larger vertical and horizontal scales. The dynamic feature
400 was no longer limited to the PBL, but extended to the sub-synoptic scales. In the pollution diffusion stage
401 of this case, the PBL structure was the same as in Case-1 (Fig. 9a-c, v), and has been described in the
402 above paragraph.



403

404 Figure 8. (a) Surface spatial distributions, (b) vertical cross-sections and (c) vertical profiles of the simulated wind divergence at the pollution stages of (i) formation, (ii-iv)
 405 maintenance, and (v) diffusion during representative Case-1 under west-southwest wind shear mode. The red ellipses, black lines, and red stars in (a) indicate the convergence
 406 belt, the section lines in (b), and the profile sites in (c), respectively. The purple dashed lines in (b) indicate the PBL heights. The potential temperature profiles are presented
 407 in (c) to indicate the boundary layer top at the representative sites.



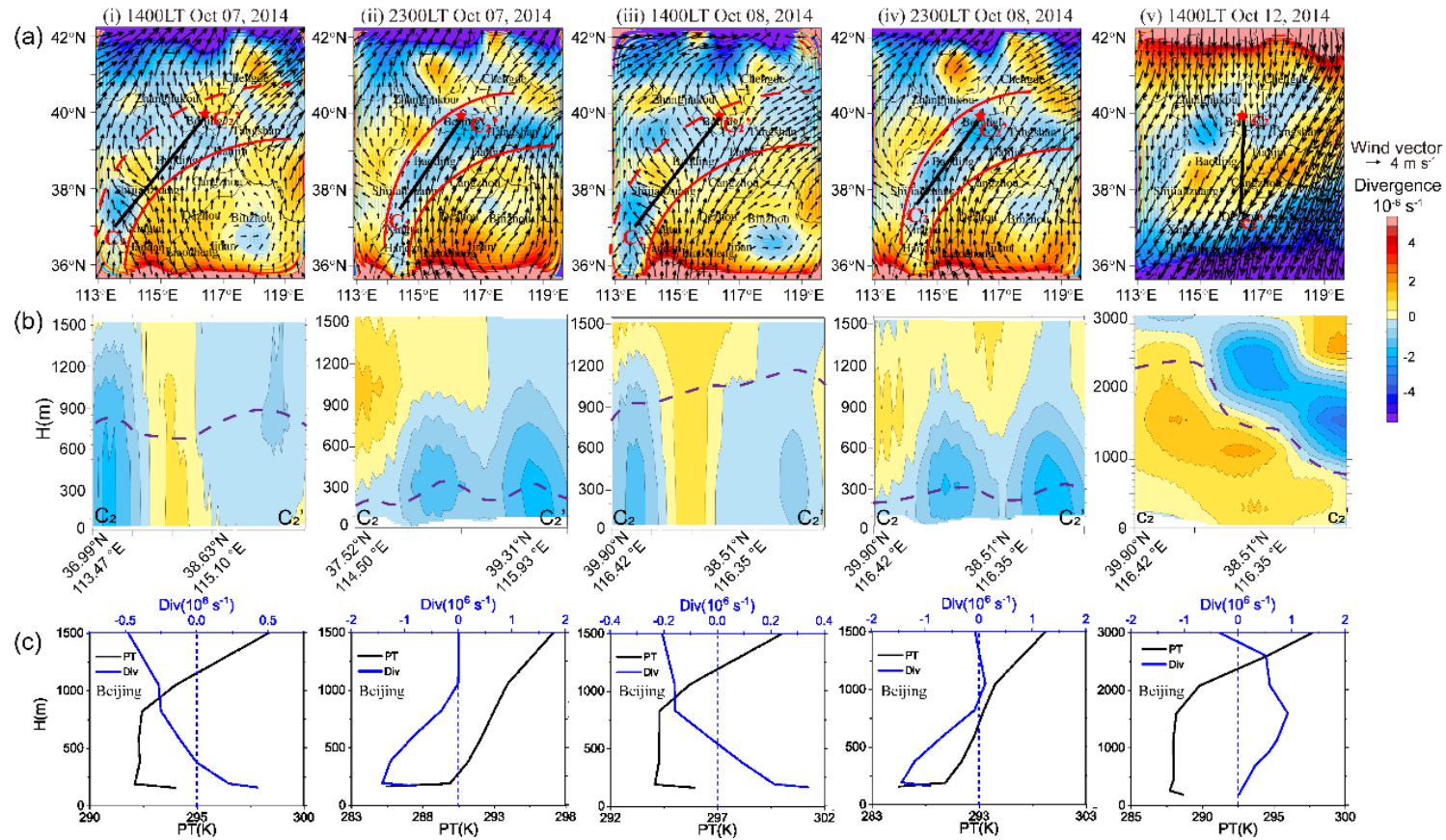
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409 Figure 9. Same as Fig. 8, but for representative Case-2 under south-north wind shear mode. The red lying-Y shapes, black lines, and red stars in (a) indicate the convergence
 410 belt, the section lines in (b), and the profile sites in (c), respectively. The purple dashed lines in (b) indicate the PBL heights. The potential temperature profiles are presented
 411 in (c) to indicate the boundary layer top at the representative sites.

412 **Topographic obstruction category**

413 As an outcome of a mixture of the thermal and dynamic effects, the topographic obstruction
414 category pollution is analyzed from the perspectives of both the wind divergence and potential
415 temperature to reveal the thermal and dynamic structure of the PBL.

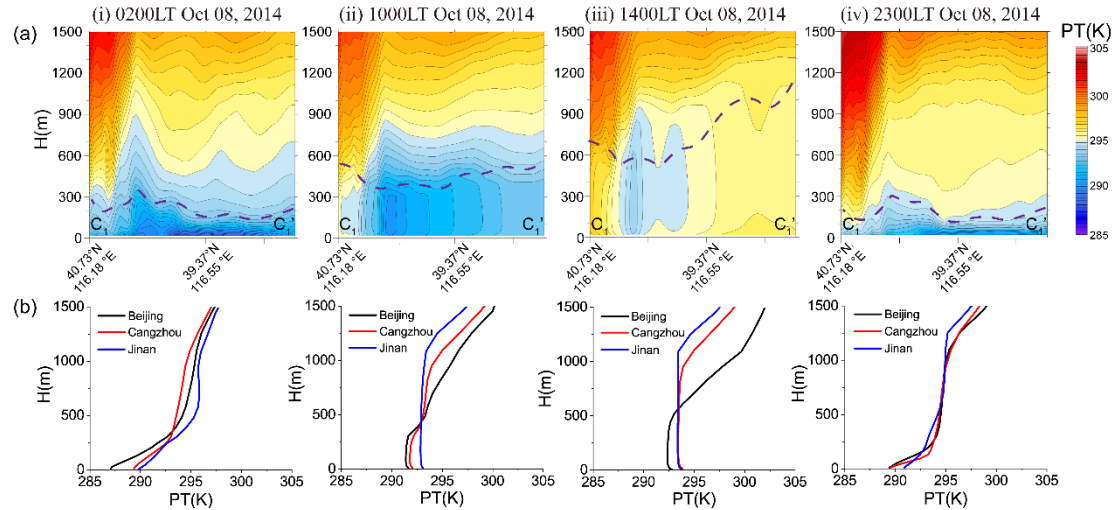
416 Figure 10 shows the dynamic characteristics of the PBL during Case-3. In the pollution formation-
417 maintenance stage, there was an arc-shaped convergence belt at the foot of the mountains on the
418 windward, due to the momentum loss in the northward flow under the action of topographic obstruction
419 (Fig. 10a, i-iv). The shape of this convergent belt was more regular at night (Fig. 10a, ii, iv) but had some
420 breakages at the northern edge during the day when there was a local southeast wind around Beijing and
421 Shijiazhuang (Fig. 10a, i, iii). The vertical sections also displayed the general features and diurnal
422 difference, showing an integral convergence layer at night with a depth of the mountain height (Fig. 10b,
423 ii, iv), and an isolated divergent layer emerged during the daytime (Fig. 10b, i, iii). The vertical profiles
424 of the wind divergence and potential temperature at Beijing were further shown in Fig. 10c, i-iv. In the
425 evening, the atmosphere below 1200 m was convergent with the peak appearing near the surface of about
426 $-1.5 \times 10^{-6} \text{ s}^{-1}$. In the afternoon, there was a weak divergence layer with a strength of about $0.5 \times 10^{-6} \text{ s}^{-1}$
427 and a thickness of about 200~300 m within the PBL. We infer that the day-night variation may be the
428 consequence of the mountain-valley circulation, since the northwestward daytime valley winds
429 developed along the mountain gorges near Beijing and Shijiazhuang leading to flow divergence, and
430 downslope winds formed at night strengthening the surface convergence. During the pollution diffusion
431 stage, the northern part of the domain was in a strong divergence condition (Fig. 10a, v). The
432 corresponding cross-section shows that the north wind divergence layer was very deep (nearly 3000 m),
433 even extending beyond the boundary layer. It gradually thins from north to south with the decrease of
434 the PBL height (Fig. 10b, v). Moreover, the vertical profiles of the divergence and potential temperature
435 at the Beijing site show that the PBL was well developed up to 2000 m, accompanied by strong horizontal
436 divergence throughout the layer (Fig. 10c, v). Both of them indicate extremely favorable ventilation
437 conditions.



438

439 Figure 10. Same as Fig. 8, but for representative Case-3 under the topographic obstruction category. The red curves, black lines, and red stars in (a) indicate the convergence
 440 belt, the section lines in (b), and the profile sites in (c), respectively. The purple dashed lines in (b) indicate the PBL heights. The potential temperature profiles are presented
 441 in (c) to indicate the boundary layer top at the representative sites.

442 The thermal properties and their evolution, especially diurnal variation, play an important role in
443 this pollution pattern, which has been presented in the above surface analysis. Hence, we further explore
444 the three-dimension thermal structure of the PBL, taking the vertical cross-sections of potential
445 temperature across the characteristic cold area in the pollution maintenance stage (October 8, 2014, the
446 location of the cross-section is shown in Fig. 6c) as an illustration. In the early hours of the morning,
447 although there were surface inversions across the whole region, the cold air masses in front of the
448 mountains were much thicker (Fig. 11a, i). After sunrise, the convective boundary layer developed both
449 in the front of the mountains and in the plain due to the surface heating, but the temperature in the
450 southern plain was higher and the PBL was slightly deeper (Fig. 11a, ii). In the afternoon, a deep, well-
451 mixed warm PBL (with a height of more than 1000 m) has formed in the southern plain while a cold air
452 mass capped by strong inversion (at the height of about 600-1000 m) still remained in the northern
453 piedmont area (Fig. 11a, iii). At night, large amounts of cold air accumulated at the foot of the mountains
454 again (Fig. 11a, iv). The vertical profiles of the simulated potential temperature of the three cities from
455 south to north, Jinan, Cangzhou and Beijing, also support this thermal evolution process. At 0200 LT,
456 there were surface inversions at all three cities, and Beijing had the strongest inversion intensity of about
457 2 K per 100 m (Fig. 11b, i). By 1000 LT, the PBL height in Jinan had increased to 1100 m, while the
458 convective boundary layers in Beijing and Cangzhou were shallow (about 400 m, Fig. 11b, ii). In the
459 afternoon, the PBL was fully developed with the height from the south to the north site ranging from
460 1150 m to 650 m, and there was still a thick inversion layer above Beijing (Fig. 11b, iii). At 2300 LT, the
461 surface inversion at the three sites has formed again (Fig. 11b, iv). The persistent cold air mass in front
462 of the mountains is similar to the cold air damming on the eastern side of the Appalachian (Bell and
463 Bosart, 1988). The prevailing southerly warm airflows were blocked by the mountains and the
464 geostrophic balance was disrupted, so that the heat cannot reach the foothills and the air was further
465 cooled due to the turning easterly wind. Meanwhile, the air mass accumulated and ascended with
466 adiabatic cooling in front of the mountains. It should be noted that the southeast edge of this cold area
467 was more pronounced during the daytime (Fig. 5c, Fig. 11), in comparison to that at night. This is
468 reasonable given that the nocturnal boundary layer was stable over the whole domain and more
469 susceptible to the local property, such as surface heterogeneity, meandering motions, and gravity waves
470 (Mahrt, 1998). Although the AIB was relatively unclear at the surface during nighttime, the nocturnal
471 cold layer at the foothills was deeper than the southern plain area, probably due to the cold drainage
472 flows along the sidewall of the mountains to form a cold air pool. This diurnal cycle of the PBL thermal
473 structure can well explain the day-night difference in pollution distribution pattern (refer to Fig. 4e).



474

475 Figure 11. (a) Vertical cross-sections and (b) vertical profiles of the simulated potential temperature at
 476 (i) 0200 LT, (ii) 1000 LT, (iii) 1400 LT and (iv) 2300 LT on October 08, 2014 in Case-3 under the
 477 topographic obstruction category. The cross-sections along the line C_1C_1' are shown in Fig. 6c, iii, iv.
 478 The purple dashed lines in (a) indicate the PBL heights.

479 4 Summary and discussion

480 This study investigated the three-dimensional PBL structures modified by mesoscale AIBs under
 481 various pollution categories by using the mesoscale meteorological model WRF. Based on the
 482 classification of pollution episodes in the NCP (Jin et al., 2022 submitted), representative pollution cases
 483 of wind shear category and orographic obstruction category were analyzed. The WRF model was
 484 comprehensively evaluated for its reliability, by comparison with observed PBL vertical structure, as
 485 well as the temporal series and spatial distribution of the surface meteorological fields. The evolution of
 486 the PBL spatial structures and their interaction with the mesoscale AIBs during the pollution episodes
 487 were fully revealed, from both thermal and dynamic perspectives.

488 The results of this paper, together with a previous systematic classification study (Jin et al., 2022
 489 submitted) and a detailed case study for frontal category (Jin et al., 2021), depict a more complete and
 490 clearer view of the PBL spatial structures during pollution episodes in the regional scale of NCP (as
 491 schematically shown in Fig.1). All the pollution conditions during the autumn and winter were classified
 492 into three categories. The most prominent was the frontal category. With an isolated cold air mass
 493 laterally bounded by the warm frontal AIB on one side and mountains on another side, the PBL was
 494 vertically suppressed by a dome-like warm cap. Typically, the intensity of the frontal inversion can be as
 495 large as 3~6 K per 100 m. As a consequence, the PBL in this cold area was very shallow (as low as
 496 200~300 m) and kept stable stratification, in sharp contrast to the deep and well-mixing boundary layer
 497 outside this zone (Fig. 1a). This explained why $PM_{2.5}$ accumulated rapidly in this enclosed and stable

498 space and formed a laterally clearly defined polluted air mass. Diurnally, the nocturnal PBL in this
499 category was less typical than its daytime counterpart. The thermal structure of the PBL played a leading
500 role in this category, resulting in the most severe pollution level.

501 The wind shear category, with two main modes: west-southwest wind shear and south-north wind
502 shear, was featured with airflow convergence AIB and dominated by dynamic processes. The first mode
503 was characterized by a low-pressure trough. A convergence layer lay in the wind shear zone with the
504 thickness of the PBL depth (Fig. 1b), and a typical near-surface divergence of $-2\sim-4\times 10^{-6} \text{ s}^{-1}$. It is
505 accompanied by a compensating divergence layer above the PBL, reflecting the mesoscale property of
506 the trough AIB. The latter mode displayed a "lying Y shaped" convergence layer from the surface
507 extending upwards to about 3000 m, with a convergence peak above the PBL top (not shown in Fig. 1).
508 This implied the sub-synoptic scale features. In this category of both modes, the boundary layer was
509 dominated by dynamic convergence effects, which resulted in pollutants transport and accumulation, and
510 correspondent to relatively light pollution in the NCP.

511 The topographic obstruction pollution category was characterized by a cold air damming AIB at the
512 foot of the windward side of the mountains. It usually occurred when the southerly winds were too weak
513 to cross the terrain barrier and the northward flows were blocked. In response, the geostrophic balance
514 was adjusted, which made the southerly warm advection weaken and further turned to easterly cold
515 advection. All these factors allowed air masses to accumulate and ascend with adiabatic cooling at the
516 foothills. The PBL air was cold and capped by a strong inversion in the damming area, in contrast with
517 well-mixed warm PBL in the southern plain. Meanwhile, the air flows were convergent in front of the
518 mountains. These general characteristics are shown in Fig. 1c. In more detail, the thermal discontinuity
519 became indistinct at night due to the surface inversion over the whole domain, while the nocturnal wind
520 convergence belt was more pronounced. The diurnal variation of the PBL dynamic and thermal structure
521 made the pollutants concentrate at the foot of the mountains during the daytime while local pollution
522 formed throughout the entire plain at night.

523 The present study focuses on the characteristic mesoscale PBL structures under pollution conditions,
524 and emphasizes their role in shaping regional pollution patterns. The analysis of pollution evolution is
525 based on the $\text{PM}_{2.5}$ concentration fields interpolated or diagnosed from monitoring data, relying on
526 densely distributed stations. However, the PBL spatial structure is presented by numerical simulation,
527 due to the scarcity and limitation of sounding data. Evaluation from the spatial-temporal variation of the
528 surface meteorological field and PBL vertical structure indicates that the model performance is good.
529 WRF can capture mesoscale systems and AIBs, as well as their overall evolution process and diurnal
530 variation. It should be noted that, it is still difficult to reproduce the precise timing of the buildup and
531 breakup as well as the exact location and range of these systems. This deficiency should be concerned
532 seriously when simulated meteorological fields are used to drive air quality models, since a small position

533 bias and time deviation of the AIBs can significantly alter pollution levels at a certain site (Seaman, 2000;
534 McNider and Pour-Biazar, 2020). Accurate capture of mesoscale AIBs is a necessary prerequisite for
535 reliable simulation of pollution evolution. Besides, successful reproduction and forecast of air quality by
536 the chemical transport models also involves other factors, such as the accuracy of source inventories and
537 the complexity of chemical mechanisms (Travis et al., 2016; Bouarar et al., 2019; Wang et al., 2021),
538 which are beyond the scope of this study. The aim of the present work is to provide a clear cognition of
539 these typical PBL structures reproduced by numerical simulations. This goal is achieved satisfactorily.

540 At last, the pollution categories presented in this study can still be rough or oversimplified, and the
541 real processes may be more complex and atypical as analyzed. However, this work, to the authors'
542 knowledge, is the first trial to reveal the various PBL structures over the vast scale of the NCP, and to
543 clarify their role in regional PM_{2.5} pollution. Modulation of the PBL by mesoscale meteorological
544 processes, particularly the AIBs, is clearly demonstrated. Extending the view of the PBL from local
545 vertical properties to mesoscale three-dimension structures may be a step toward a better understanding
546 of the meteorological effects on regional-scale PM_{2.5} pollution.

547

548 **Data availability**

549 The data in this study are available from the corresponding author (xhcai@pku.edu.cn).

550 **Author contribution**

551 XHC and XPJ designed the research. MYY and HSZ collected the data. XPJ performed the simulations
552 and wrote the paper. XHC reviewed and commented on the paper. YS, XSW and TZ participated in the
553 discussion of the article.

554 **Competing interests**

555 The authors declare that they have no conflict of interest.

556 **Acknowledgements**

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560 **References**

- 561 Baumann, K., Maurer, H., Rau, G., Piringer, M., Pechinger, U., Prevo, A., Furger, M., Neining, B. and
562 Pellegrini, U.: The influence of south Foehn on the ozone distribution in the Alpine Rhine valley -
563 results from the MAP field phase, *Atmos. Environ.*, 35(36), 6379-6390, doi:10.1016/s1352-
564 2310(01)00364-8, 2001.
- 565 Bei, N. F., Zhao, L. N. Wu, J. R., Li, X., Feng, T. and Li, G. H.: Impacts of sea-land and mountain-valley
566 circulations on the air pollution in Beijing-Tianjin-Hebei (BTH): A case study, *Environ. Pollut.*, 234,
567 429-438, doi:10.1016/j.envpol.2017.11.066,2018.
- 568 Bell, G. D. and Bosart, L. F.: Appalachian cold-air damming. *Mon. Wea. Rev.*, 116, 137–161,
569 doi:10.1175/1520-0493(1988)116,0137:ACAD.2.0.CO;2, 1988.
- 570 Berger, B. W. and Friehe, C. A.: Boundary-layer structure near the cold-front of a marine cyclone during
571 “ERICA”, *Bound.-Layer Meteor.*, 317-317, doi:10.1007/BF0071125874, 1995.
- 572 Bluestein, H. B.: Surface boundaries of the Southern Plains: Their role in the initiation of convective
573 storms, in: *Synoptic-dynamic meteorology and weather analysis and forecasting: a tribute to Fred*
574 *Sanders*, edited by Bosart, L.F., Bluestein, H.B., American Meteorological Society, Boston, pp5-33,
575 2008.
- 576 Bianco, L., Djalalova, I. V., King, C. W., and Wilczak, J. M.: Diurnal Evolution and Annual Variability
577 of Boundary-Layer Height and Its Correlation to Other Meteorological Variables in California's
578 Central Valley, *Bound.-Layer Meteor.*, 140, 491-511, doi:10.1007/s10546-011-9622-4, 2011.
- 579 Bouarar, I., Brasseur, G., Petersen, K., Granier, C., Fan, Q., Wang, X.M., Wang, L.L., Ji, D. S., Liu, Z.R.,
580 Xie, Y., Gao, W., and Elguindi, N.: Influence of anthropogenic emission inventories on simulations
581 of air quality in China during winter and summer 2010. *Atmos. Environ.* 198, 236–256.
582 <https://doi.org/10.1016/j.atmosenv.2018.10.043>, 2019.
- 583 Boutle, I. A., Beare, R. J., Belcher, S. E., Brown, A. R., and Plant, R. S.: The Moist Boundary Layer
584 under a Mid-latitude Weather System, *Bound.-Layer Meteor.*, 134, 367-386, doi:10.1007/s10546-
585 009-9452-9, 2010.
- 586 De Wekker, S. F. J.: Observational and numerical evidence of depressed convective boundary layer
587 heights near a mountain base, *J. Appl. Meteorol. Climatol.*, 47, 1017-1026,
588 doi:10.1175/2007jamc1651.1, 2008.
- 589 Dupont, J. C., Haefelin, M., Badosa, J., Elias, T., Favez, O., Petit, J. E., Meleux, F., Sciare, J., Crenn, V.,
590 Bonne, J. L.: Role of the boundary layer dynamics effects on an extreme air pollution event in Paris.
591 *Atmos. Environ.*, 141, 571–579, doi:10.1016/j.atmosenv.2016.06.061, 2016.
- 592 Emeis, S., and Schafer, K.: Remote sensing methods to investigate boundary-layer structures relevant to
593 air pollution in cities, *Bound.-Layer Meteor.*, 121, 377-385, doi:10.1007/s10546-006-9068-2, 2006.
- 594 Garratt, J. R: *The Atmospheric Boundary Layer*. Cambridge University Press, Cambridge, 1992.

595 Garratt, J. R.: The internal boundary-layer-A review, *Bound.-Layer Meteor.*, 50, 171-203,
 596 doi:10.1007/bf00120524, 1990.

597 Hane, C. E., Rabin, R. M., Crawford, T. M., Bluestein, H. B. and Baldwin, M. E.: A case study of severe
 598 storm development along a dryline within a synoptically active environment. Part II: Multiple
 599 boundaries and convective initiation. *Mon. Wea. Rev.*, 130, 900–920, doi: 10.1175/1520-
 600 0493(2002)130<0900:ACSOSS>2.0.CO;2,2002.

601 Hanna, S. R. and Yang, R.: Evaluations of mesoscale models' simulations of near-surface winds,
 602 temperature gradients, and mixing depths. *J. Appl. Meteorol.* 40 (6):1095–104, doi:10.1175/1520-
 603 0450(2001)040<1095:EOMMSO>2.0.CO;2, 2001.

604 Jimenez, P. A., de Arellano, J. V. G., Dudhia, J., and Bosveld, F. C.: Role of synoptic- and meso-scales
 605 on the evolution of the boundary-layer wind profile over a coastal region: the near-coast diurnal
 606 acceleration, *Meteorol. Atmos. Phys.*, 128, 39-56, doi:10.1007/s00703-015-0400-6, 2016.

607 Jin, X. P., Cai, X. H., Yu, M. Y., Song, Y., Wang, X. S., Kang, L. and Zhang, H. S.: Diagnostic analysis
 608 of wintertime PM_{2.5} pollution in the North China Plain: The impacts of regional transport and
 609 atmospheric boundary layer variation. *Atmos. Environ.*, 224, 117346, doi:
 610 10.1016/j.atmosenv.2020.117346, 2020.

611 Jin, X. P., Cai, X. H., Yu, M. Y., Wang, X. B., , Song, Y., Wang, X. S., Zhang, H. S. and Zhu, T.: Regional
 612 PM_{2.5} pollution confined by atmospheric internal boundaries in the North China Plain: Analysis
 613 based on surface observations, *Sci. Total Environ.*, (submitted), 2022.

614 Jin, X. P., Cai, X. H., Yu, M. Y., Wang, X. S., Song, Y., Kang, L., Zhang, H. S. and Zhu, T.: Mesoscale
 615 structure of the atmospheric boundary layer and its impact on regional air pollution: A case study,
 616 *Atmos. Environ.*, 258, doi:10.1016/j.atmosenv.2021.118511, 2021.

617 Lareau, N.P., Crosman, E., Whiteman, C. D., Horel, J. D., Hoch, S.W., Brown, W.O.J. and Horst, T.W.:
 618 The persistent cold-air pool study, *Bull. Amer. Meteor. Soc.*, 94, 51-63, doi: 10.1175/BAMS-D-11-
 619 00255.1, 2013.

620 Li, J., Sun, J. L., Zhou, M. Y., Cheng, Z. G., Li, Q. C., Cao, X. Y. and Zhang, J. J.: Observational analyses
 621 of dramatic developments of a severe air pollution event in the Beijing area. *Atmos. Chem. Phys.*
 622 18, 3919-3935, doi: 10.5194/acp-18-3919-2018, 2018.

623 Li, Q. H., Wu, B. G., Liu, J. L., Zhang, H. S., Cai, X. H. and Song, Y.: Characteristics of the atmospheric
 624 boundary layer and its relation with PM_{2.5} during haze episodes in winter in the North China Plain.
 625 *Atmos. Environ.* 223, 117265, doi:org/10.1016/j.atmosenv.2020.117265, 2020.

626 Liu, N., Zhou, S., Liu, C. S., and Guo, J. P.: Synoptic circulation pattern and boundary layer structure
 627 associated with PM_{2.5} during wintertime haze pollution episodes in Shanghai, *Atmos. Res.*, 228,
 628 186-195, doi:10.1016/j.atmosres.2019.06.001, 2019.

629 Liu, S. and Liang, X. Z.: Observed diurnal cycle climatology of planetary boundary layer height, *J.*

630 Climate, 23(21), 5790–5809, doi:10.1175/2010jcli3552.1, 2010.

631 Lu, R. and Turco, R. P.: Air pollution transport in a coastal environment II: 3-dimension simulations over
632 Los-Angeles basin, *Atmos. Environ.*, 29(13), 1499-1518, doi:10.1016/1352-2310(95)00015-q,
633 1995.

634 Lyu, W., Li, Y., Guan, D. B., Zhao, H. Y., Zhang, Q., and Liu, Z.: Driving forces of Chinese primary air
635 pollution emissions: an index decomposition analysis, *Journal of Cleaner Production*, 133, 136-144,
636 doi:10.1016/j.jclepro.2016.04.093,2016.

637 Mahrt, L.: Stratified atmospheric boundary layers and breakdown of models, *Theoretical and*
638 *Computational Fluid Dynamics*, 11, 263-279, doi:10.1007/s001620050093,1998.

639 Mayfield, J. A. and Fochesatto, G. J.: The Layered Structure of the Winter Atmospheric Boundary Layer
640 in the Interior of Alaska, *J. Appl. Meteorol. Climatol.*, 52, 953-973, doi:10.1175/jamc-d-12-01.1,
641 2013.

642 McNider, R. T. and Pour-Biazar, A.: Meteorological modeling relevant to mesoscale and regional air
643 quality applications: a review, *J. Air Waste Manage. Assoc.*, 70, 2-43,
644 doi:10.1080/10962247.2019.1694602, 2020.

645 Miao, Y. C. and Liu, S. H.: Linkages between aerosol pollution and planetary boundary layer structure
646 in China, *Sci. Total Environ.*, 650, 288-296, doi:10.1016/j.scitotenv.2018.09.032, 2019.

647 Miao, Y. C., Hu, X. M., Liu, S. H., Qian, T. T., Xue, M., Zheng, Y. J. and Wang, S.: Seasonal variation of
648 local atmospheric circulations and boundary layer structure in the Beijing-Tianjin-Hebei region and
649 implications for air quality, *J. Adv. Model. Earth Syst.*, 7(4), 1602-1626,
650 doi:10.1002/2015ms000522, 2015.

651 Narasimha, R., Sikka, D. R. and Prabhu, A.: The Monsoon Trough Boundary Layer. *Indian Academy of*
652 *Sciences*, 422 pp, 1997.

653 Peng, H. Q., Liu, D. Y., Zhou, B., Su, Y., Wu, J. M., Shen, H., Wei, J. S., and Cao, L.: Boundary-Layer
654 Characteristics of Persistent Regional Haze Events and Heavy Haze Days in Eastern China, *Adv.*
655 *Meteorol.*, doi:10.1155/2016/6950154, 2016.

656 Petaja, T., Jarvi, L., Kerminen, V. M., Ding, A. J., Sun, J. N., Nie, W., Kujansuu, J., Virkkula, A., Yang,
657 X. Q., Fu, C. B., Zilitinkevich, S. and Kulmala, M: Enhanced air pollution via aerosol-boundary
658 layer feedback in China. *Sci Rep* 6, 6, doi:10.1038/srep18998, 2016.

659 Pielke, R. A. and Uliasz, M.: Use of meteorological models as input to regional and mesoscale air quality
660 models—Limitations and strengths. *Atmos. Environ.*, 32(8):1455–66, doi:10.1016/S1352-
661 2310(97)00140-4, 1998.

662 Potty, K. V. J., Mohanty, U. C., and Raman, S.: Simulation of boundary layer structure over the Indian
663 summer monsoon trough during the passage of a depression, *J. Appl. Meteorol.*, 40, 1241-1254,
664 doi:10.1175/1520-0450(2001)040<1241:soblso>2.0.co;2, 2001.

665 Prezerakos, N. G.: Lower tropospheric structure and synoptic scale circulation patterns during prolonged
666 temperature inversions over Athens, Greece, *Theor. Appl. Climatol.*, 60, 63-76,
667 doi:10.1007/s007040050034, 1998.

668 Qu, K., Wang, X. S., Yan, Y., Shen, J., Xiao, T., Dong, H. B., Zeng, L. M. and Zhang, Y. H.: A comparative
669 study to reveal the influence of typhoons on the transport, production and accumulation of O₃ in
670 the Pearl River Delta, China, *Atmos. Chem. Phys.*, 21, 11593-11612, doi:10.5194/acp-21-11593-
671 2021, 2021.

672 Rajkumar, G., Saraswat, R. S., and Chakravarty, B.: Thermodynamic structure of the monsoon boundary-
673 layer under the influence of a large-scale depression, *Bound.-Layer Meteor.*, 68, 131-137,
674 doi:10.1007/bf00712667, 1994.

675 Ren, Y., Zhang, H. S., Wei, W., Wu, B. G., Cai, X. H. and Song, Y.: Effects of turbulence structure and
676 urbanization on the heavy haze pollution process, *Atmos. Chem. Phys.*, 19, 1041-1057,
677 doi:10.5194/acp-19-1041-2019, 2019.

678 Rogers, R. E., Deng, A. J., Stauffer, D. R., Gaudet, B. J., Jia, Y. Q., Soong, S. T. and Tanrikulu, S.:
679 Application of the Weather Research and Forecasting Model for Air Quality Modeling in the San
680 Francisco Bay Area, *J. Appl. Meteorol. Climatol.*, 52, 1953-1973, doi:10.1175/jamc-d-12-
681 0280.1,2013.

682 Sanders, F., and Doswell, C. A.: A case for detailed surface-analysis, *Bull. Amer. Meteorol. Soc.*, 76(4),
683 505-521, doi:10.1175/1520-0477(1995)076<0505:acfdsa>2.0.co;2,1995.

684 Seaman, N. L. and Michelson, S. A.: Mesoscale meteorological structure of a high-ozone episode during
685 the 1995 NARSTO-Northeast study, *J. Appl. Meteorol.*, 39, 384-398, doi:10.1175/1520-
686 0450(2000)039<0384:mmssoah>2.0.co;2, 2000.

687 Seaman, N. L.: Meteorological modeling for air-quality assessments. *Atmos. Environ.* 34 (12–14):2231–
688 59, doi:10.1016/S1352-2310(99)00466-5., 2000.

689 Seibert, R.: South foehn studies since the ALPEX experiment, *Meteorol. Atmos. Phys.*, 43, 91–103,
690 doi:10.1007/BF01028112, 1990.

691 Sinclair, V. A., Belcher, S. E., and Gray, S. L.: Synoptic Controls on Boundary-Layer Characteristics,
692 *Bound.-Layer Meteor.*, 134, 387-409, doi:10.1007/s10546-009-9455-6, 2010.

693 Sinclair, V. A.: A 6-yr Climatology of Fronts Affecting Helsinki, Finland, and Their Boundary Layer
694 Structure, *J. Appl. Meteorol. Climatol.*, 52, 2106-2124, doi:10.1175/jamc-d-12-0318.1, 2013.

695 Stull, R.: *An Introduction to Boundary Layer Meteorology*. Springer, New York, 1988.

696 Talbot, C., Augustin, P., Leroy, C., Willart, V., Delbarre, H., and Khomenko, G.: Impact of a sea breeze
697 on the boundary-layer dynamics and the atmospheric stratification in a coastal area of the North Sea,
698 *Bound.-Layer Meteor.*, 125, 133-154, doi:10.1007/s10546-007-9185-6, 2007.

699 Tennekes, H.: The atmospheric boundary layer. *Phys. Today* 27, 52–63, doi:10.1063/1.3128397, 1974.

700 Travis, K. R., Jacob, D. J., Fisher, J. A., Kim, P. S., Marais, E. A., Zhu, L., Yu, K., Miller, C. C., Yantosca,
701 R. M., Sulprizio, M. P., Thompson, A. M., Wennberg, P. O., Crouse, J. D., St Clair, J. M., Cohen,
702 R. C., Laughner, J. L., Dibb, J. E., Hall, S. R., Ullmann, K., Wolfe, G. M., Pollack, I. B., Peischl, J.,
703 Neuman, J. A., and Zhou, X. L.: Why do models overestimate surface ozone in the Southeast United
704 States?, *Atmos. Chem. Phys.*, 16, 13561-13577, doi:10.5194/acp-16-13561-2016,2016.

705 Wang, W. G., Liu, M. Y., Wang, T. T., Song, Y., Zhou, L., Cao, J. J., Hu, J. N., Tang, G. G., Chen, Z., Li,
706 Z. J., Xu, Z. Y., Peng, C., Lian, C. F., Chen, Y., Pan, Y. P., Zhang, Y. H., Sun, Y. L., Li, W. J., Zhu,
707 T., Tian, H. Z., and Ge, M. F.: Sulfate formation is dominated by manganese-catalyzed oxidation of
708 SO₂ on aerosol surfaces during haze events, *Nat. Commun.*, 12, doi:10.1038/s41467-021-22091-
709 6,2021.

710 Xiao, Z. S., Miao, Y. C., Du, X. H., Tang, W., Yu, Y., Zhang, X., and Che, H. Z.: Impacts of regional
711 transport and boundary layer structure on the PM_{2.5} pollution in Wuhan, Central China, *Atmos.*
712 *Environ.*, 230, doi:10.1016/j.atmosenv.2020.117508, 2020.

713 Ye, X. X., Song, Y., Cai, X. H., Zhang, H. S.: Study on the synoptic flow patterns and boundary layer
714 process of the severe haze events over the North China Plain in January 2013, *Atmos. Environ.*
715 *Times* 124, 129–145. <https://doi.org/10.1016/j.atmosenv.2015.06.011>, 2016.

716 Zhang, Q., Zheng, Y. X., Tong, D., Shao, M., Wang, S. X., Zhang, Y. H., Xu, X. D., Wang, J. N., He, H.,
717 Liu, W. Q., Ding, Y. H., Lei, Y., Li, J. H., Wang, Z. F., Zhang, X. Y., Wang, Y. S., Cheng, J., Liu, Y.,
718 Shi, Q. R., Yan, L., Geng, G. N., Hong, C. P., Li, M., Liu, F., Zheng, B., Cao, J. J., Ding, A. J., Gao,
719 J., Fu, Q. Y., Huo, J. T., Liu, B. X., Liu, Z. R., Yang, F. M., He, K. B., and Hao, J. M.: Drivers of
720 improved PM_{2.5} air quality in China from 2013 to 2017, *Proc. Natl. Acad. Sci. U. S. A.*, 116, 24463-
721 24469, doi:10.1073/pnas.1907956116, 2019.

722 Zhang, Q. Q., Ma, Q., Zhao, B., Liu, X. Y., Wang, Y. X., Jia, B. X., and Zhang, X. Y.: Winter haze over
723 North China Plain from 2009 to 2016: Influence of emission and meteorology, *Environ. Pollut.*, 242,
724 1308-1318, doi:10.1016/j.envpol.2018.08.019, 2018.

725 Zhao, C. Wang, F., Y., Shi, X. Q., Zhang, D. Z., Wang, C. Y., Jiang, J. H., Zhang, Q., and Fan, H.:
726 Estimating the Contribution of Local Primary Emissions to Particulate Pollution Using High-
727 Density Station Observations, *J. Geophys. Res.-Atmos.*, 124, 1648-1661,
728 doi:10.1029/2018jd028888,2019.