

Reply to comments

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Title: "High frequency of new particle formation events driven by summer monsoon in the central Tibetan Plateau, China"

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I. Reply to Reviewer 1

Reply to Reviewer 1's overall comments:

This manuscript investigates atmospheric new particle formation (NPF) taking place in the central Tibetan Plateau. To my knowledge, there are no prior publications of NPF in this location, so the obtained results can be considered worth publishing. The paper itself is well organized and the conducted analysis appears to be scientifically sound. There are, however, two major issues that require further consideration.

We appreciate the comments from the reviewer on this manuscript. We have answered them point to point in the following paragraphs (the texts italicized are the comments, the texts indented are the responses, and the texts in blue are revised parts in new manuscript). In addition, all changes made are marked in the revised manuscript. Thanks for the reviewer's affirmation on our work.

Reply to Reviewer 1's major comments (2):

1. First, the measurements periods are rather short, about 4 weeks for the pre-monsoon season and less than 2 weeks for the monsoon season. As a result, it remains unclear how representative the obtained results are for this location during these two seasons.

Thanks for the comment. The measurements periods were a little limited as the reviewer described. But our measurements periods can be representative for this location during pre-monsoon season and monsoon season as follows:

1) The intensity of Indian Summer Monsoon during the two measurements periods can represent that in the whole pre-monsoon and monsoon seasons, respectively. The intensity of Indian Summer Monsoon is an important indicator to distinguish the monsoon season. Here the intensity of Indian Summer Monsoon

(ISM) was indicated by the ISM Index, which are defined by the negative outgoing longwave radiation anomalies (with respect to the climatological annual cycle) averaged over the Bay of Bengal–India region (10°–25°N, 70°–100°E) (Wang and Fan, 1999). As shown in Fig. R1, the measurement periods (green boxes) were in the pre-monsoon season (March-May) and monsoon season (June-September), respectively. And the IMS index during the two measurements periods were equivalent to those of the whole pre-monsoon season (average: -19.5 vs -20.7 W m⁻²) and monsoon season (average: 27.0 vs 26.3 W m⁻²), respectively. Therefore, we considered that these two observation periods are representative in the seasonal characteristics in pre-monsoon season and monsoon season, respectively.

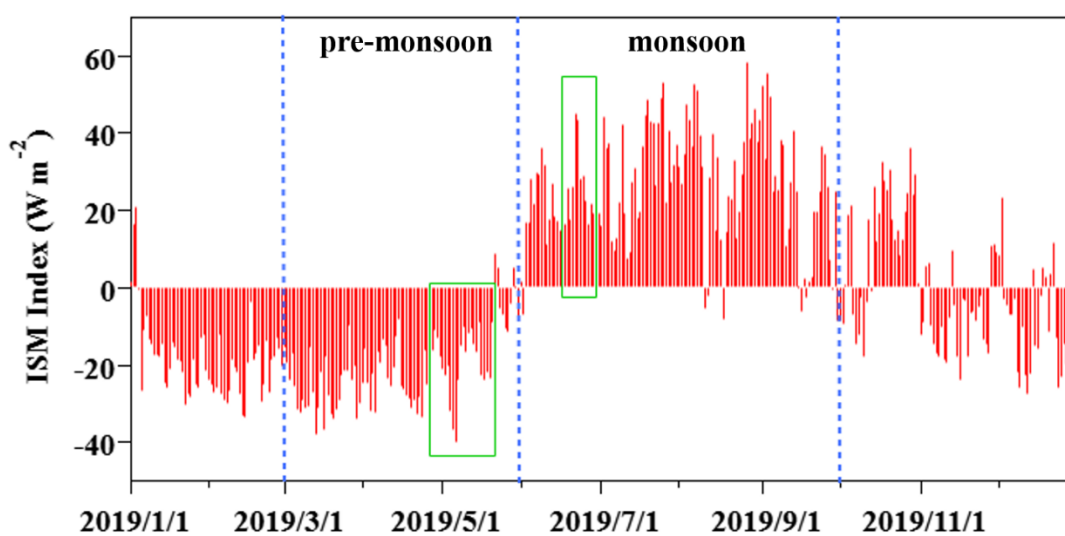
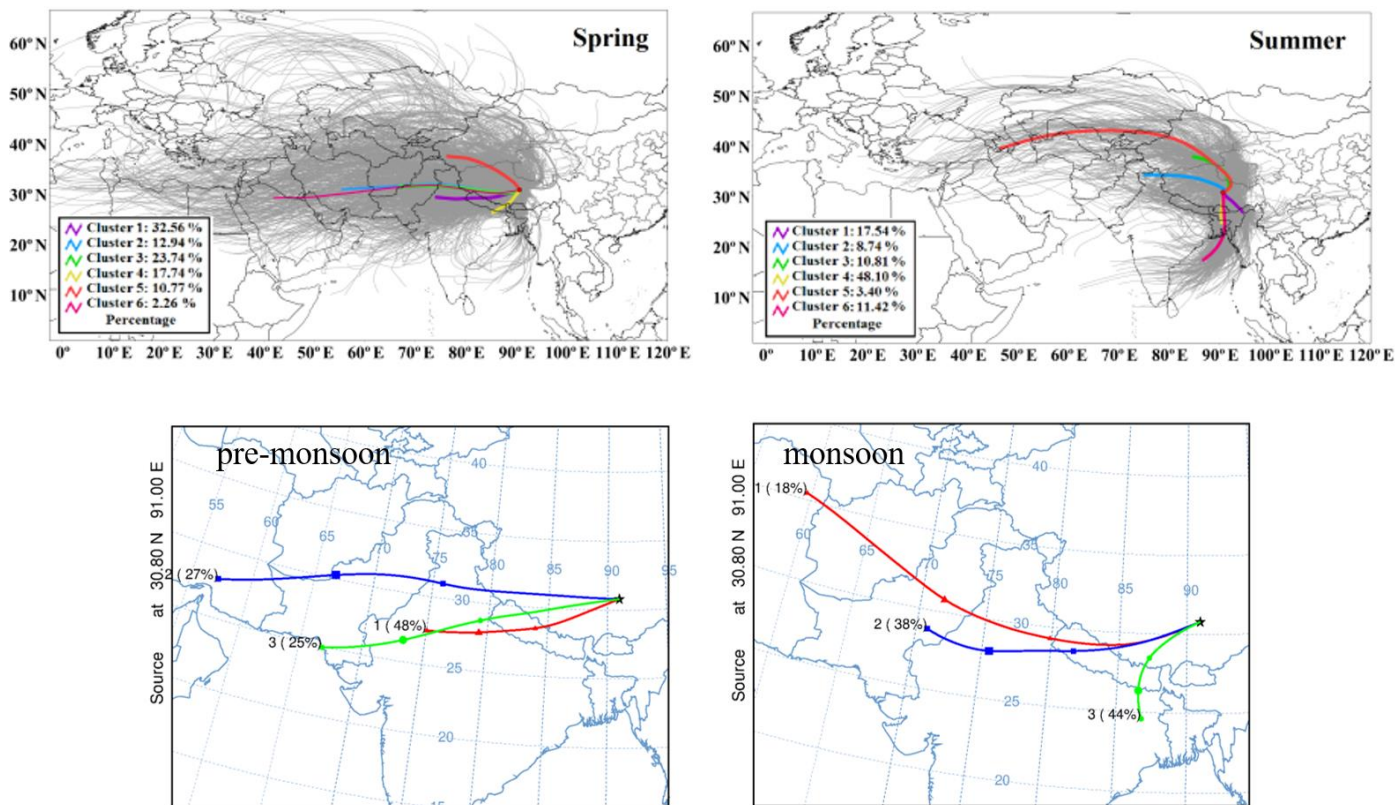


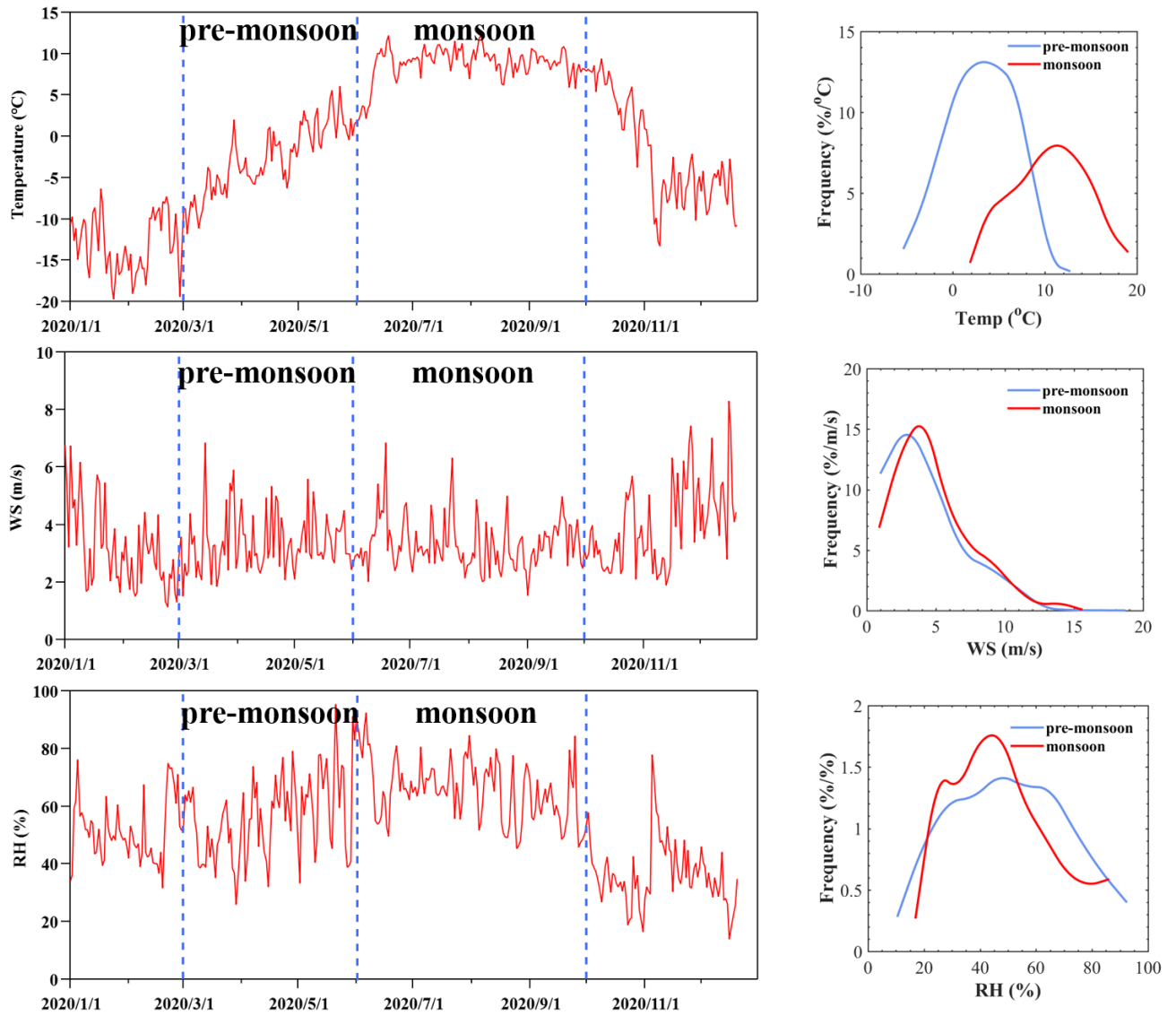
Figure R1. The Indian Summer Monsoon (ISM) Index in 2019. The measurements periods are marked by the green boxes.

2) The characteristics of meteorology and atmospheric pollutants in the two measurements periods was generally in agreement with the previous long-term studies at Nam Co station and other sites in the Tibetan Plateau (TP) (Yin et al., 2017; Cong et al., 2015; Bonasoni et al., 2010; Xu et al., 2018). Both previous research and this study showed that, strong westerlies pass through western Nepal, northwest India and Pakistan in pre-monsoon season, while air masses were mainly derived from Bangladesh and northeast India and brought moisture that originated in the Bay of Bengal in monsoon season (Fig. R2) (Yin et al., 2017). The temperature, WS and RH in the two measurements periods were matched with those in the whole pre-monsoon and monsoon season of 2020 at Nam Co station (Fig. R3) (National Tibetan Plateau Data Center). The average temperature in pre-monsoon and monsoon seasons were around 3 and 10 °C, respectively. WS showed no difference between pre-monsoon and monsoon seasons with the average value of 4 m/s. The average level of RH was similar between the two seasons, but the variation range of RH was larger in pre-monsoon season. In addition, the level of PM, BC and ozone in the two

54 measurements periods were matched with those in the whole pre-monsoon and monsoon season at the
 55 other site of the TP (Fig. R4) (Bonasoni et al., 2010; Xu et al., 2018). The average concentrations of PM₁
 56 (PM_{0.8}) in pre-monsoon and monsoon seasons were around 1 and 2 μg/m³, respectively. The concentration
 57 of BC was at a level of hundreds nanogram per cubic meter in pre-monsoon season, while at a lower level
 58 in monsoon season. Ozone showed the similar pattern with BC. It should be noted that there will be some
 59 differences between this study and other results due to the differences in time resolution. In a word, the
 60 characteristics of meteorology and atmospheric pollutants in the two measurements periods in this study
 61 can well reflect those in pre-monsoon and monsoon seasons.



62 **Figure R2.** Comparison of trajectories between this study and previous study at Nam Co station.
 63 Backward HYSPLIT trajectories for each measurement day (black lines in the maps), and mean back
 64 trajectory for six HYSPLIT clusters (colored lines in the maps) arriving at Nam Co Station in spring
 65 (MAM) and summer (JJA) (Yin et al., 2017) (top). The frequencies of the 48 h back trajectories of air
 66 masses arriving at Nam Co station from different directions during pre-monsoon and monsoon seasons in
 67 this study (bottom).
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Figure R3. Comparison of meteorology between this study and the whole seasons in 2020. Time series of ambient temperature, wind speed and relative humidity at Nam Co station from January 2020 to December 2020 (National Tibetan Plateau Data Center) (left). Comparison in frequency distributions of temperature, WS and RH at Nam Co station in pre-monsoon and monsoon seasons in this study (right).

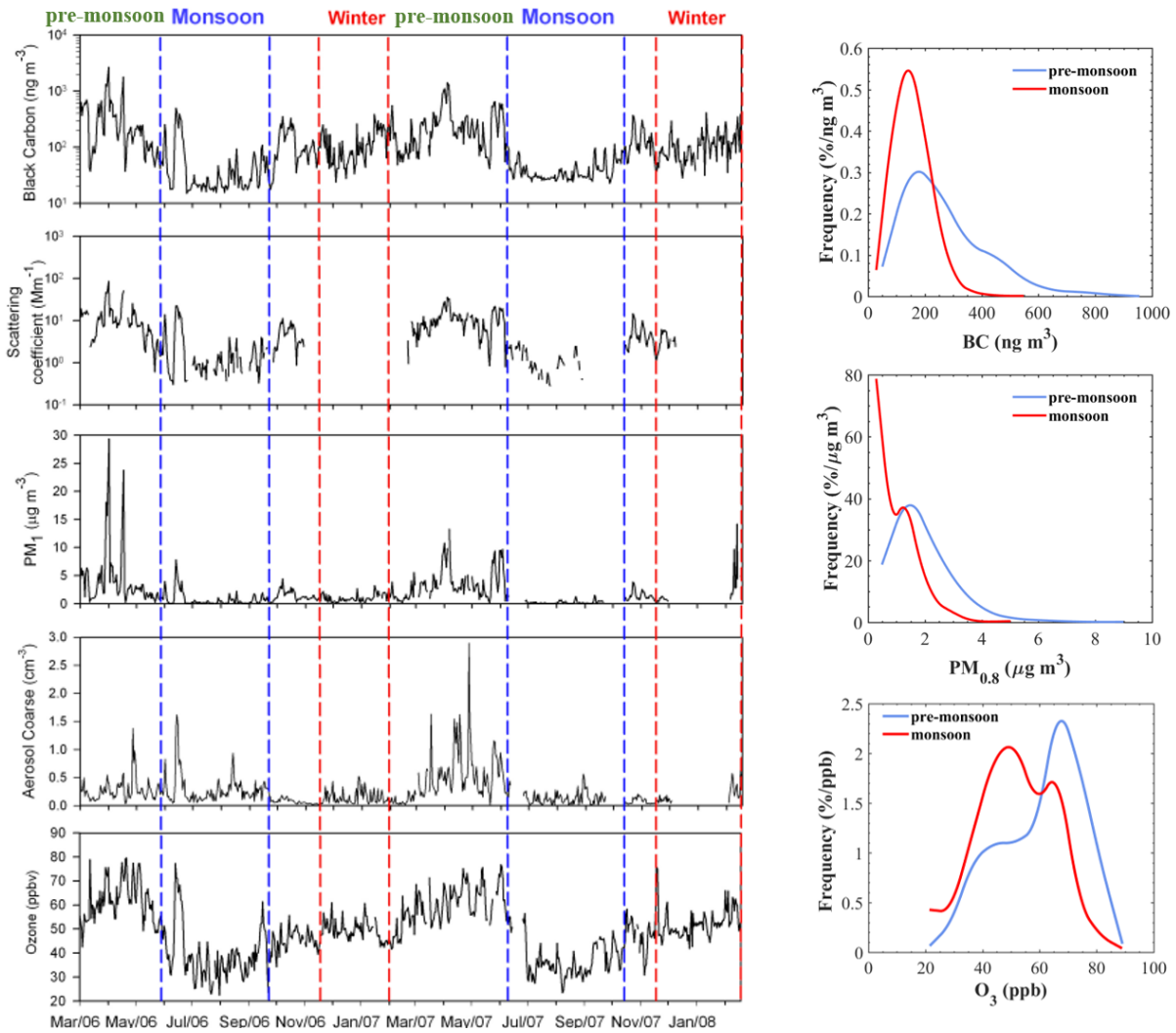


Figure R4. Comparison of atmospheric pollutants between this study and previous TP studies. Time series of f BC, aerosol scattering coefficient, PM₁, coarse particle number and surface ozone at the Nepal Climate Observatory-Pyramid (NCO-P) station from 1 March 2006 to 28 February 2008 (Bonasoni et al., 2010) (left). Comparison in frequency distributions of BC, PM_{0.8} and O₃ at Nam Co station in pre-monsoon and monsoon seasons in this study (right).

3) Based on the above discussion, the two measurements periods in this study can respectively represent the pre-monsoon and monsoon seasons at Nam Co station, so the NPF characteristics of the two observation periods can also be considered as the NPF characteristics in pre-monsoon and monsoon seasons. In addition, the phenomenon of higher NPF frequency in monsoon season than pre-monsoon season was also found in the other sites in the Tibetan Plateau (TP). A 16-month measurements from 2006 to 2007 at Himalayan Nepal Climate Observatory at Pyramid (NCO-P) site on the southern TP showed NPF frequency of 38% in pre-monsoon season and 57% in monsoon season (Venzac et al., 2008). At Mt. Yulong on the southeastern TP, the NPF frequency was only 14% during pre-monsoon season (Shang et al., 2018). The NPF frequency of 15% in pre-monsoon season and 80% in monsoon season at Nam Co

90 station was consistent with these studies, with more significant seasonal differences. The significant
91 seasonal differences may be due to the fact that the occurrence of NPF is more sensitive to the monsoon
92 in extremely clean background areas (such as Nam Co station and Mt. Yulong). In summary, our study
93 emphasized the seasonal differences in NPF frequencies at Nam Co station, and the results was reliable.

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117 Mt. Yulong (3410 m), China, *Atmos. Chem. Phys.*, 18, 15687-15703, 10.5194/acp-18-15687-2018, 2018.

118 Due to harsh conditions and logistical limitations, our observation periods were limited. However, our
119 conclusions are obvious and representative, and we will carry out more detailed observations for a longer
120 period in the future if possible. To illustrate the representativeness of the observation periods, we have
121 made supplements in the revised manuscript as follows:

122 **“2.1 Measurement site**

123 The measurement was conducted from 26 April to 22 May, 2019 and 15 June to 25 June, 2019, and can be
124 representative of the pre-monsoon season and the summer monsoon season, respectively (Text S1) (Bonasoni et al.,
125 2010; Cong et al., 2015).”

126 **“Text S1 The representativeness of the observation periods**

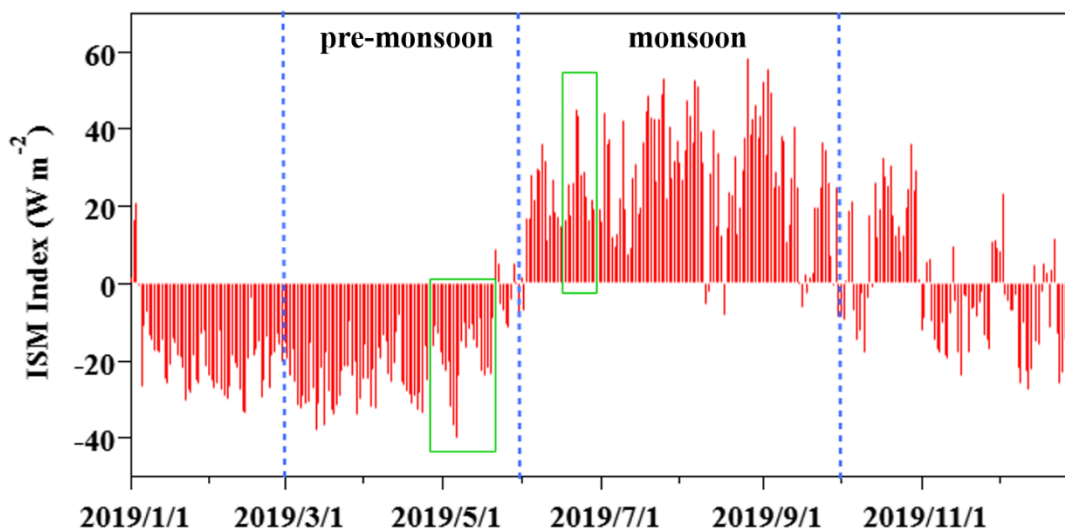
127 The measurement was conducted from 26 April to 22 May, 2019 and 15 June to 25 June, 2019, and can be
128 representative of the pre-monsoon season and monsoon season, respectively.

129 Firstly, the intensity of Indian Summer Monsoon during the two measurements periods can represent that in

130 the whole pre-monsoon and monsoon seasons, respectively. The intensity of Indian Summer Monsoon is an
131 important indicator to distinguish the monsoon season. Here the intensity of Indian Summer Monsoon (ISM) was
132 indicated by the ISM Index, which are defined by the negative outgoing longwave radiation anomalies (with respect
133 to the climatological annual cycle) averaged over the Bay of Bengal–India region (10° – 25° N, 70° – 100° E) (Wang
134 and Fan, 1999). As shown in Fig. S1, the measurement periods (green boxes) were in the pre-monsoon season
135 (March–May) and monsoon season (June–September), respectively. And the IMS index during the two
136 measurements periods were equivalent to those of the whole pre-monsoon season (average: -19.5 vs -20.7 W m^{-2})
137 and monsoon season (average: 27.0 vs 26.3 W m^{-2}), respectively.

138 Secondary, the characteristics of meteorology and atmospheric pollutants in the two measurements periods
139 was generally in agreement with the previous long-term studies at Nam Co station and other sites in the Tibetan
140 Plateau (TP) (Yin et al., 2017; Cong et al., 2015; Bonasoni et al., 2010; Xu et al., 2018). That is, the characteristics
141 of meteorology (temperature, WS and RH) and atmospheric pollutants (PM, BC and ozone) in the two
142 measurements periods were matched with those in the whole pre-monsoon and monsoon season at Nam Co station
143 and other sites in the TP.

144 Therefore, the two observation periods are representative in the seasonal characteristics in pre-monsoon season
145 and monsoon season, respectively.



146
147 **Figure S1.** The Indian Summer Monsoon (ISM) Index in 2019. The measurements periods are marked by the green boxes.”
148

149 **2. Second, many of the conclusions made in the paper rely on SO_2 and VOC concentrations. Unfortunately,**
150 **there are very limited measurements on these 2 trace gases (only VOCs during the monsoon season), instead**
151 **their concentrations were estimated from large-scale model simulations. The simulated concentrations may**
152 **have large uncertainties, which are not quantified by any means in the paper.**

153 Thanks for the comment. The model evaluation including the meteorological fields and air pollutants has

154 been added in the revised manuscript. In general, the results of model simulation of VOC and SO₂ show
155 good performance in statistical parameters of model evaluation and correlation analysis with other tracers.
156 The modelled VOC and SO₂ could be used for the NPF analysis. The detailed information can be found
157 in the revised manuscript as follow:

158 “2.3 Model simulation

159 Considering the limited measurements on SO₂ and VOC in this observation (only VOC during pre-monsoon
160 season), Weather Research and Forecasting/Community Multiscale Air Quality (WRF/CMAQ) modeling system
161 was adopted to simulate the level of SO₂ and VOC in the whole observation period, to assist in the analysis of the
162 role of sulfuric acid and organics in NPF events.

163 Weather Research and Forecasting (WRF) (version 4.2.1) model was used to simulate the meteorological
164 conditions with the FNL reanalysis dataset. The 6 h FNL data were obtained from the U.S. National Centre for
165 Atmospheric Research (NCAR), with a spatial resolution of 1.0° × 1.0° (<http://rda.ucar.edu/datasets/ds083.2/>, last
166 accessed on 28 April 2022). The Community Multiscale Air Quality version 5.3.2 (CMAQv5.3.2) model, being one
167 of the three-dimensional chemical transport models (CTMs) (Appel et al., 2021), configured with the gas-phase
168 mechanism of SAPRC07tic and the aerosol module of AERO6i, was employed in this study to simulate the air
169 quality over Tibet in the observation period (26 April to 22 May and 15 June to 25 June in 2019). Air quality
170 simulations were performed with a horizontal resolution of 12 km. The corresponding domain covered Tibet and
171 the surrounding countries and regions with 166 × 166 grids (Fig. S2), with the 18 layers in vertical resolution.
172 Detailed information about the model setting is provided in Text S2.

173 The Multi-resolution Emission Inventory for China version 1.3 (MEICv1.3) (<http://www.meicmodel.org>) and
174 Regional Emission inventory in ASia (REASv3.2) (<https://www.nies.go.jp/REAS/>) were used to provide the
175 anthropogenic emissions from China and neighboring countries and regions, respectively. The MEICv1.3 emissions
176 of the year 2019 were used. For REAS, the emission inventory in the year 2015 was used for 2019 as no emission
177 inventory was released for the years after 2015. Although emission inventories are usually released 3 years behind,
178 we acknowledge that this may cause additional uncertainties in the simulation. Biogenic emissions were generated
179 using the Model for Emissions of Gases and Aerosols from Nature (MEGANv2.1) (Guenther et al., 2012). The open
180 biomass burning emissions were processed using the Fire Inventory for NCAR (FINN) during the entire study
181 period (Wiedinmyer et al., 2011).

182 The model evaluation is introduced in Text S2. The WRF and CMAQ models successfully reproduced the
183 meteorological fields and air pollutants including PM and O₃ with model performance indices meeting the suggested
184 benchmarks. For VOC, the observed VOC and predicted VOC in pre-monsoon season were compared to examine
185 the model performance. The benchmarks for VOC had not been reported, but the statistical metrics of MFB (mean
186 fractional bias, -0.47) and MFE (mean fractional error, 0.49) in this study are within the range reported in previous

VOC modelling result (Hu et al., 2017). The correlation coefficient (R) between simulated and observed VOC is 0.41, which reflected that the model can fairly simulate the variation of VOC concentration. It should be noted that VOC was underpredicted on the whole, which may due to the uncertainty of the emission inventory as mentioned before. For SO₂, the WRF/CMAQ models have been successfully reproduced SO₂ in major regions in China with R of 0.25-0.79 (Mao et al., 2022). The simulated SO₂ level in the model domain is comparable with that measured at Mt. Yulong (Shang et al., 2018), with average values of 0.03±0.02 ppbv and 0.06±0.05 ppbv. At the same time, considering that both BC and SO₂ are mainly emitted from coal combustion and biomass burning, BC could be a good indicator for SO₂ especially for pristine environment without local anthropogenic source emissions. As shown in Fig.S6, a good correlation between SO₂ and BC measured at Mt. Yulong was found with correlation coefficient (R) of 0.79 (Shang et al., 2018). In this study, the modelled SO₂ and measured BC also showed good correlation with R of 0.58 (Fig. S6). In general, the results of model simulation showed good performance in statistical parameters and correlation analysis with other tracers. The modelled VOC and SO₂ may be helpful for the NPF analysis.”

“2.3 Text S2 Model simulation

Model Configurations

The meteorological conditions were simulated using the Weather Research and Forecasting (WRF) (version 4.2.1) model with the FNL reanalysis dataset. The 6 h FNL data were obtained from the U.S. National Centre for Atmospheric Research (NCAR), with a spatial resolution of 1.0° × 1.0° (<http://rda.ucar.edu/datasets/ds083.2/>, last accessed on 28 April 2022). The physical parameterizations used in this study are the Thompson microphysical process, RRTMG longwave/shortwave radiation scheme; Noah land-surface scheme; MYJ boundary layer scheme; and modified Tiedtke cumulus parameterization scheme. The detailed configuration settings could be found in the works of Hu et al. (2016), Mao et al. (2022), Wang et al. (2021a).

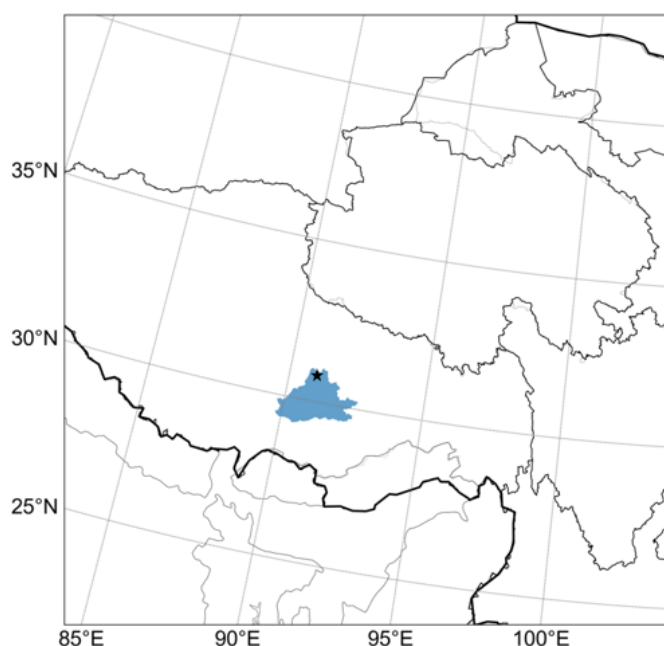
The Community Multiscale Air Quality version 5.3.2 (CMAQv5.3.2) model, being one of the three-dimensional chemical transport models (CTMs) (Appel et al., 2021), configured with the gas-phase mechanism of SAPRC07tic and the aerosol module of AERO6i, was employed in this study to simulate the air quality over Tibet from 24 April to 24 May and 13 June to 27 June in 2019, which contains the observation period. Air quality simulations were performed with a horizontal resolution of 12 km. The corresponding domain covered Tibet and the surrounding countries and regions with 166 × 166 grids (Fig. S2), with the 18 layers in vertical resolution. The initial and boundary conditions were provided by the default profiles. The simulated results of the first two days were not included in the model analysis, which served as a spin-up and reduced the effects of the initial conditions on the simulated results.

Model Evaluation

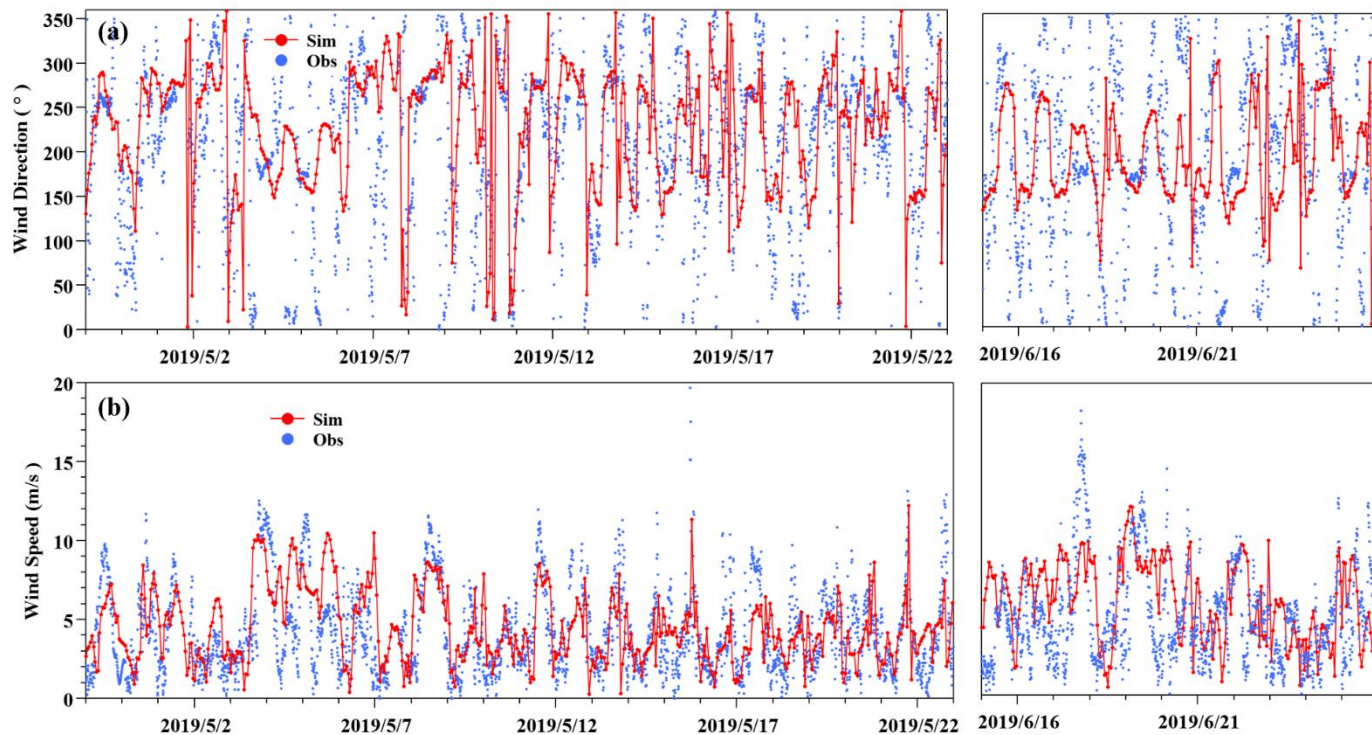
Previous studies have investigated the impacts of meteorological conditions on the formation, transportation,

220 and dissipation of air pollutants (Hu et al., 2016; Hua et al., 2021; Mao et al., 2022; Sulaymon et al., 2021b;
221 Sulaymon et al., 2021a). Therefore, the evaluation of the WRF model performance was carried out before the usage
222 of its meteorological fields in the CMAQ simulations. The evaluation of the WRF model was achieved by
223 comparing the predicted wind speed (WS, m/s), wind direction (WD, °) at 10 m above the surface, RH (%) and
224 temperature (T, °C) to the observed values. Fig. S3 showed that WS was well simulated both in pre-monsoon and
225 monsoon seasons. WD was well simulated in pre-monsoon season, and there seems to be some deviation in the
226 simulation of north wind in monsoon season. The main reason about the deviation in WD may be due to the poor
227 terrain and complicated weather conditions. Nevertheless, both simulations and measurements showed more
228 frequent southerly winds during monsoon season. RH and temperature were well simulated in the whole periods
229 (Fig. S4). The good model performance with the statistical metrics of WS, RH and temperature meeting the
230 suggested benchmarks are shown in Table S1. Generally, the simulated meteorological fields were qualified and can
231 be further utilized in driving the CMAQ model

232 Fig. S5 showed the comparison of simulated hourly mean concentration about PM, O₃ and VOC in observation
233 site, which were simulated by CMAQ. The statistical indices used in evaluating the CMAQ model were present in
234 Table S2. It can be seen that PM and O₃ meet the suggested benchmarks, which reflect the good model performance.
235 The observed VOC and predicted VOC in pre-monsoon season were compared to examine the model performance.
236 The benchmarks for VOC had not been reported, but the MFB (mean fractional bias) and MFE (mean fractional
237 error) values are within the range reported in previous VOC modelling result (Hu et al., 2017). The correlation
238 coefficient (R) between simulated and observed VOC is 0.41, which reflected that the model can fairly simulate the
239 variation of VOC concentration. It should be noted that VOC was underpredicted on the whole, which may due to
240 the uncertainty of the emission inventory.



241 **Figure S2.** WRF/CMAQ modeling domain
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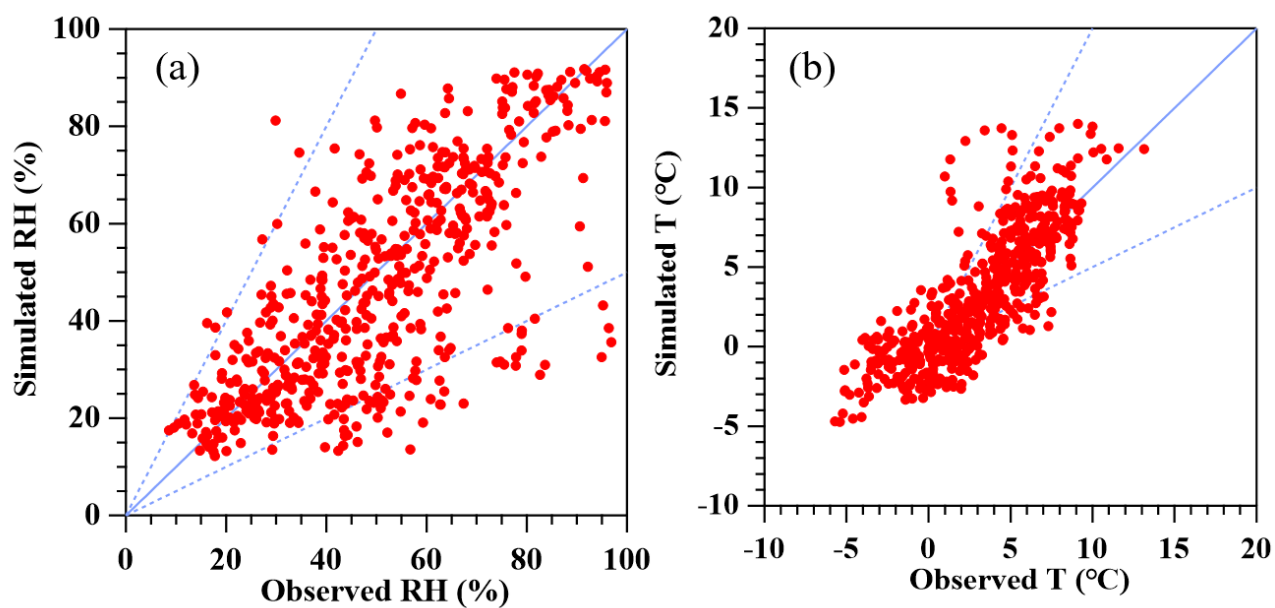
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Figure S3. Comparison of simulated (in red dot-line) and observed (in blue dot) wind direction (WD, °) and wind speed (WS, m/s). Observed is 10 minutes mean data. Simulated is hourly mean data.

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Figure S4. Comparison of simulated and observed RH (%) and temperature (T, °C). RH and temperature are hourly mean data.

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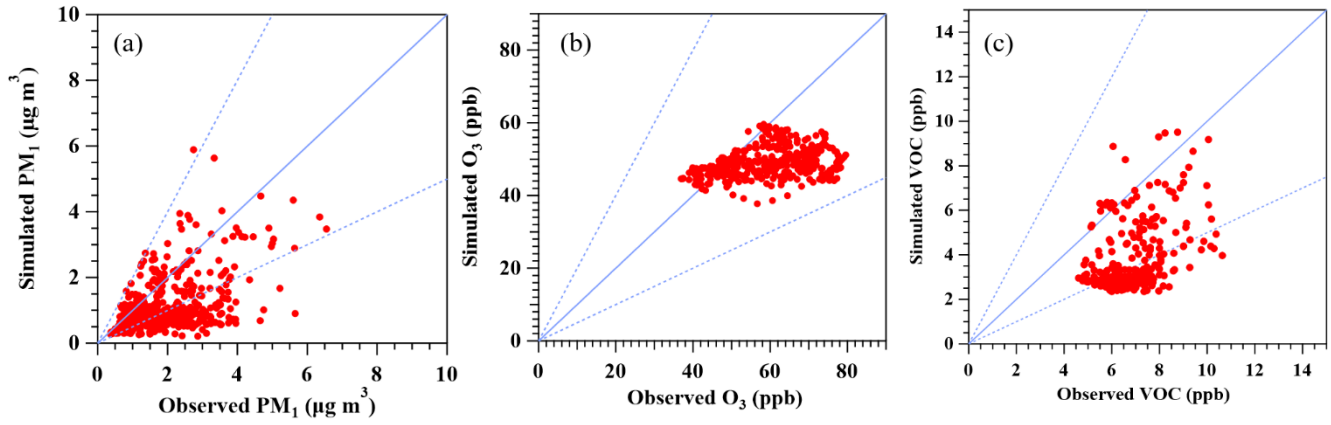


Figure S5. Comparison of simulated and observed PM ($\mu\text{g}/\text{m}^3$), O_3 (ppb) and VOC (ppb). PM, O_3 and VOC are hourly mean concentration.

Table S1. Model performance of meteorological factors at Nam Co station

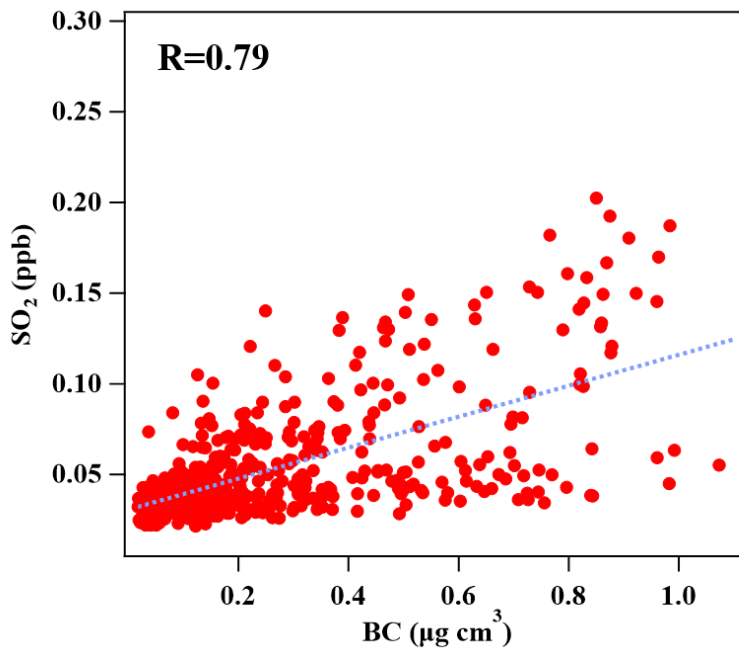
	WS				RH				T			
	MB	ME	RMSE	R	MB	ME	RMSE	R	MB	ME	RMSE	R
Statistic	0.42	0.87	1.20	0.51	-1.38	12.20	16.30	0.67	0.07	1.85	2.43	0.89
Benchmarks	$\leq \pm 0.5$	≤ 2.0	≤ 2.0						$\leq \pm 0.5$	≤ 2.0		

MB: mean bias; ME: mean error; RMSE: root mean square error; R: correlation coefficient. The benchmarks were suggested by Boylan and Russell (2006).

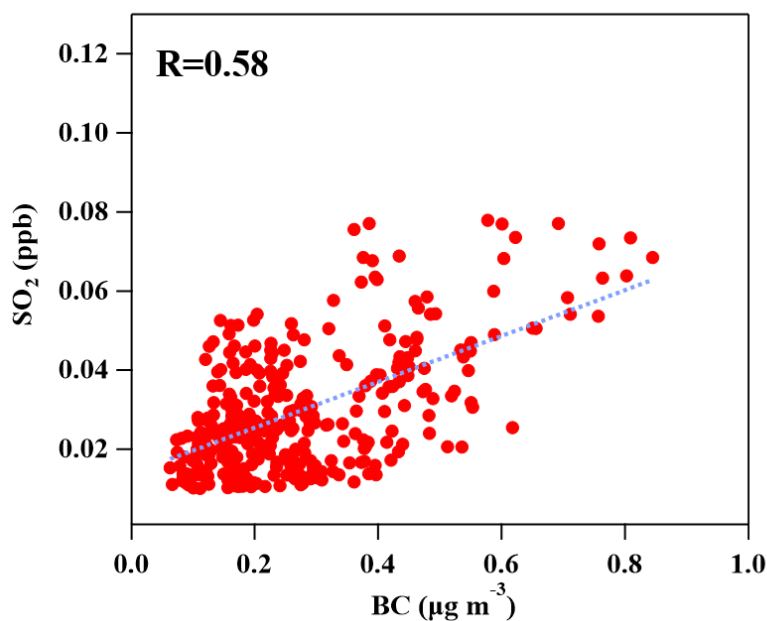
Table S2. Model performance of the air pollutants at Nam Co station

	PM ₁			O ₃			VOC			SO ₂		
	MFB	MFE	R	NMB	NME	R	MFB	MFE	R	NMB	NME	R
Statistic	0.49	0.50	0.72	0.14	0.23	0.51	-0.47	0.49	0.41			
Benchmarks	$< \pm 0.6$	< 0.75	> 0.4	$< \pm 0.15$	< 0.35	> 0.5						
References							$< \pm 0.77$	< 0.74		$< \pm 4.38$	$< \pm 4.38$	0.25- 0.79

NMB: normalized mean bias; NME: normalized mean error; R: correlation coefficient; MFB: mean fractional bias; MFE: mean fractional error. The benchmarks for PM and O₃ were suggested by Emery et al. (2017) and Boylan and Russell (2006), respectively. The references for VOC and SO₂ were from Hu et al. (2017) and Mao et al. (2022), respectively.



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267 **Figure S6.** Relationship between SO₂ and BC at Mt. Yulong in 2015. The correlation coefficient R is 0.79.



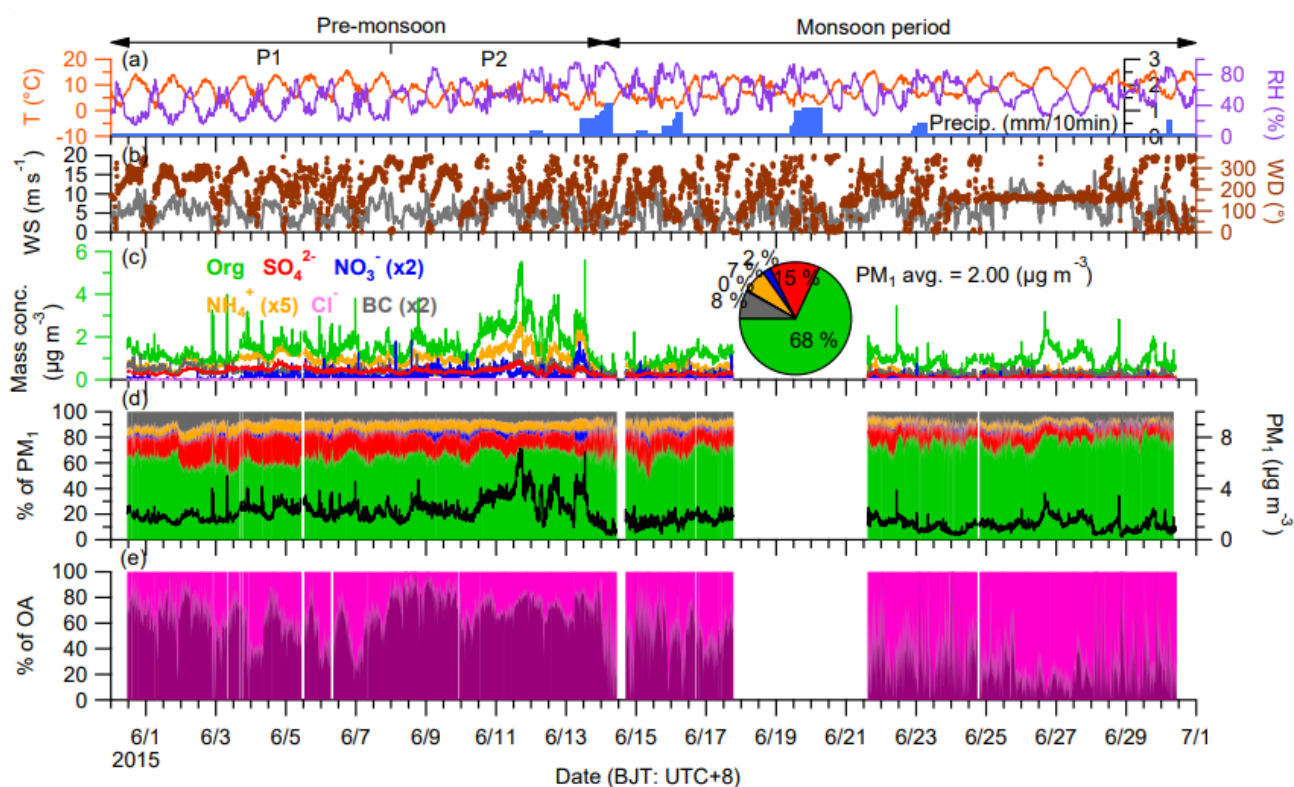
268
269 **Figure S7.** Relationship between modelled SO₂ and BC at Nam Co station. The correlation coefficient R is 0.58.’

270
271 **Reply to Reviewer 1’s important scientific comments (5):**

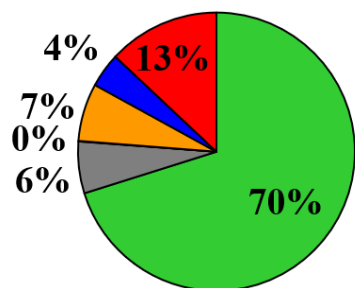
272 *1. The description on how CCN concentrations were calculated (section 2.4) is incomplete. Apparently, the*
 273 *authors used equations 4 and 5 to determine the critical diameters corresponding to different*
 274 *supersaturations, and from these critical diameters one then gets the number of CCN using measured*
 275 *particle number size distributions. However, this calculation cannot be done without knowing the*
 276 *hygroscopicity parameter cappa. Did the authors simply assumed a fixed value for cappa or did they*
 277 *estimate it from some chemical data?*

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Thanks for the comment. The hygroscopicity parameter kappa was assumed to be a constant value of 0.12 throughout the measurement period, according to the previous measurement at Mt. Yulong on the southeastern TP (Shang et al., 2018). This is mainly due to the similar proportions of chemical components between Nam Co station (Xu et al., 2018) and Mt. Yulong with around 70% fraction of organics (Fig R5). And the fractions of chemical components at Nam Co station were stable.



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284

285 **Figure R5.** The combo plot of the data of the Nam Co study including (a) the meteorological conditions (T :
286 air temperature; RH: relative humidity; Precip.: precipitation), (b) the variation of WS (wind speed) colored
287 according to WD (wind direction), (c) the temporal variation of mass concentration of PM₁ species and the
288 average contribution of each species (pie chart), (d) the mass contribution of each PM₁ species and the total
289 mass concentration of PM₁, and (e) the mass contribution of PMF results. Three periods based on the
290 meteorological conditions were marked (Xu et al., 2018) (top). The average contribution of each chemical
291 species at Mt. Yulong on the southeastern TP (Shang et al., 2018) (bottom).

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293

In addition, the kappa at Nam Co station in pre-monsoon and monsoon seasons can be estimated by using

the previously measured chemical data at this site (Xu et al., 2018). The kappa of the mixed particles was calculated based on κ -Köhler theory and the Zdanovskii–Stokes–Robinson (ZSR) mixing rule (Petters and Kreidenweis, 2007). The values of κ is 0.48 for $(\text{NH}_4)_2\text{SO}_4$ and 0.58 for NH_4NO_3 (Petters and Kreidenweis, 2007). The κ is assumed to be 0.1 for organics (Wu et al., 2015). We derived the volume fraction of each species by dividing mass concentration by its density. The densities are 1.77 g cm^{-3} for $(\text{NH}_4)_2\text{SO}_4$ and 1.72 g cm^{-3} for NH_4NO_3 . The densities of organics are assumed to be 1.2 g cm^{-3} (Fan et al., 2020). The κ and density of BC are assumed to be 0 and 1.7 g cm^{-3} . It was found that average value of kappa was 0.15 and 0.13 in pre-monsoon and monsoon seasons, respectively. The D_c at S_c levels of 0.6% and 1.2% were 72.4 ± 1.0 and 45.7 ± 0.6 nm in pre-monsoon season, and 69.1 ± 0.9 and 43.6 ± 0.6 nm in monsoon season. It was comparable with that used in this study (κ : 0.12, $S_c=0.6\%$: 73.4 ± 1.3 , $S_c=1.2\%$: 46.3 ± 0.8 nm). And it had little effect on the final result of CCN concentration. Therefore, we finally decided to adopt the fixed κ value of 0.12.

Shang, D., Hu, M., Zheng, J., Qin, Y., Du, Z., Li, M., Fang, J., Peng, J., Wu, Y., Lu, S., and Guo, S.: Particle number size distribution and new particle formation under the influence of biomass burning at a high altitude background site at Mt. Yulong (3410 m), China, Atmos. Chem. Phys., 18, 15687-15703, 10.5194/acp-18-15687-2018, 2018.

Xu, J., Zhang, Q., Shi, J., Ge, X., Xie, C., Wang, J., Kang, S., Zhang, R., and Wang, Y.: Chemical characteristics of submicron particles at the central Tibetan Plateau: insights from aerosol mass spectrometry, Atmos. Chem. Phys., 18, 427-443, 10.5194/acp-18-427-2018, 2018.

Wu, Z. J., Poulain, L., Birmili, W., Größ, J., Niedermeier, N., Wang, Z. B., Herrmann, H., and Wiedensohler, A.: Some insights into the condensing vapors driving new particle growth to CCN sizes on the basis of hygroscopicity measurements, Atmos. Chem. Phys., 15, 13071-13083, 10.5194/acp-15-13071-2015, 2015.

Petters, M. D. and Kreidenweis, S. M.: A single parameter representation of hygroscopic growth and cloud condensation nucleus activity, Atmos. Chem. Phys., 7, 1961-1971, 10.5194/acp-7-1961-2007, 2007.

Fan, X., Liu, J., Zhang, F., Chen, L., Collins, D., Xu, W., Jin, X., Ren, J., Wang, Y., Wu, H., Li, S., Sun, Y., and Li, Z.: Contrasting size-resolved hygroscopicity of fine particles derived by HTDMA and HR-ToF-AMS measurements between summer and winter in Beijing: the impacts of aerosol aging and local emissions, Atmos. Chem. Phys., 20, 915-929, 10.5194/acp-20-915-2020, 2020.

“3.4 Significant increase of CCN in monsoon season

The CCN concentration was estimated following the method introduced in Sect. 2.4. Considering the similar proportions of chemical components between Nam Co station and Mt. Yulong with around 70% fraction of organics, and the stability of the fractions of chemical components, the hygroscopicity parameter κ was assumed to be a constant value of 0.12 throughout the measurement period according to the previous measurement at Mt. Yulong in the TP (Shang et al., 2018).”

2. The discussion on the role of condensation sink (CS) in favoring/disfavoring NPF is not logical. The authors first say that their result differs from those found in earlier studies (lines 243-244), but then mention a few studies which actually agree with their findings (lines 245-247). Please reformulate this part

331 *of the text, as it causes confusion in its present form.*

332 Thanks for the comment. Considering the possible confusion, we have reformulated the discussion in the
333 revised manuscript as follows:

334 “3.3.1 Condensation sink

335 CS is the key factor controlling the occurrence of NPF events especially in urban environment (Yan et al.,
336 2021). At some high-altitude observations at a larger scale, the important role of the transported pre-existing
337 particles in NPF events was also emphasized (Rose et al., 2019; Boulon et al., 2010). Here we analyzed the CS
338 levels in NPF days and non-event days at Nam Co station. As shown in Fig. 3a, the levels of CS in NPF-pre days,
339 NPF-monsoon days and non-event days were approximate during 11:00–18:00 (the occurrence time of NPF events),
340 although the CS in the early morning of NPF-pre days seems to be slightly lower. The CS was mainly in the range
341 of 0.1×10^{-2} – $0.15 \times 10^{-2} \text{ s}^{-1}$ during the NPF occurrence time, which was much lower than that at most locations in
342 China, such as $\sim 0.01 \text{ s}^{-1}$ in urban Beijing (Deng et al., 2021), and 0.1×10^{-2} to $28.4 \times 10^{-2} \text{ s}^{-1}$ at Mt. Tai (Lv et al.,
343 2018), and comparable with that at Mt. Yulong ($\sim 0.2 \times 10^{-2} \text{ s}^{-1}$) (Shang et al., 2018). The result varied from previous
344 studies which reported much lower CS during NPF days than that in non-event days (Zhou et al., 2021; Lv et al.,
345 2018). It indicated that CS was not the decisive factor controlling the occurrence of NPF events at Nam Co station.”

346
347 *3. The discussion on the role of VOCs (lines 266-277) could be improved as well. First, considering the*
348 *typical variability of VOC concentrations in ambient measurements, I would think a 20% higher VOC*
349 *concentration is slightly rather than noticeably higher (line 268). It is also confusing that for the pre-*
350 *monsoon season the VOC concentration difference is given as % while for the monsoon season it is given*
351 *as an absolute value (ppb).*

352 Thanks for the comment. This part has been modified in the revised manuscript as follows:

353 “3.3.1 Gas precursors

354 In addition to sulfuric acid, organics are also considered to be an important factor of NPF events. Observations
355 and laboratory experiments have found that organics may participate in or even dominate the nucleation and growth
356 process in NPF events in pristine environments and the preindustrial atmosphere. For example, CLOUD (Cosmics
357 Leaving Outdoor Droplets) experiments observed obvious NPF events from highly oxidized organics without the
358 involvement of sulfuric acid (Kirkby et al., 2016). At the high-altitude sites of Jungfrauoch and Himalaya, NPF
359 events occurred mainly through the condensation of highly oxygenated molecules (HOMs) (Bianchi et al., 2016;
360 Bianchi et al., 2021). Due to instrument status, VOC measurement was only available in pre-monsoon season. The
361 concentration of VOC (total VOC) showed a higher value (20%) during 11:00-18:00 on NPF-pre days compared
362 with non-event days (Fig. 3c). Aromatics, which can be used as the indicator of anthropogenic emissions, also
363 exhibited a higher level (20%) during NPF-pre days (Fig. 3d). This suggested that VOC may be the key factor in

364 determining the occurrence of NPF events. In order to further evaluate the role of VOC, we used WRF/CMAQ
365 models to simulate the spatial distribution of VOC concentration in pre-monsoon and monsoon seasons. The
366 detailed information about the model setting and evaluation can be found in Text S2. As shown in Fig.4, the
367 simulated VOC levels in NPF days including NPF-pre and NPF-monsoon days were higher than those in non-event
368 days. The average modelled VOC concentrations in NPF-pre days and NPF-monsoon days were 25% and 88%
369 higher than those in non-event days, respectively. Therefore, we further considered that the organic matter could be
370 the key factor to determine whether NPF event occurred. Nucleation of pure organics or organics involved could be
371 the dominant mechanism at Nam Co station. In addition, higher organic concentrations were observed in monsoon
372 season (NPF-monsoon days) compared with those in pre-monsoon season (NPF-pre days and non-event days). The
373 result is consistent with one recent research which has found that the concentration of monoterpene-derived HOMs
374 in East Asia was higher in summer (June-August) than that in Spring (March-May) by using GEOS-Chem global
375 chemical transport model (Xu et al., 2022). This means that the frequent NPF events in monsoon season could be
376 caused by the higher organic matters.

377
378 ***4. Wind direction is a very local quantity, and does not necessary tell correctly air pollutant sources or***
379 ***transport pathways. I wonder whether the authors have information on air mass trajectories which would***
380 ***provide more direct support for their statements on lines 283-295.***

381 Thanks for the comment. The air mass trajectories have been added in the revised manuscript as follpws:

382 “3.3.1 Air mass origins and meteorology

383 While the concentrations of organic precursors could be the most possible reason for the occurrence of NPF
384 events, the external factors driving the difference in VOC levels between the NPF and non-event days and other
385 conditions that may affect the characteristics of NPF were still unknown. This indicated that a further investigation
386 into other NPF-related variables was still required.

387 There are almost no local anthropogenic source emissions at Nam Co station. The air pollutants at the
388 observation site mainly brought by air mass transmission. Backward trajectories were utilized to identify the air
389 mass origins associated with NPF events. The frequencies of the 48 h back trajectories of air masses arriving at
390 Nam Co station during the occurrence time of nucleation (11:00-18:00) in non-event days, NPF-pre days and NPF-
391 monsoon days were present in Fig. 5. It can be found that the dominant air masses in non-event days were from the
392 west (almost 100%) and passed by western Nepal, northwest India and Pakistan. In comparison, air masses in NPF-
393 pre days and NPF-monsoon days mainly come from the south and southwest (57% and 75%) and originated in the
394 northeast India. WD, which can reflect the local situation for air masses and the source of air pollutants, showed a
395 high frequency of strong southerly wind in NPF-pre days and NPF-monsoon days and westerly wind in non-event
396 days (Fig. S14). The example of the spatial distribution of wind field in non-event days, NPF-pre days and NPF-

monsoon days displayed the similar phenomenon on a larger scale (Fig. S15). These results suggested that the occurrence of NPF events was related to the southerly and southwesterly air masses. When the southerly air mass occurred, it may bring organic precursors from the southern region (northeast India) to Nam Co station, driving the occurrence of NPF events in this area. As for the significantly higher NPF frequency in monsoon season, it resulted from the more frequent southerly air masses (summer monsoon) in monsoon season in comparison with pre-monsoon season as introduced in Sec. 3.1. The summer monsoon can bring the higher organic concentrations in monsoon season (NPF-monsoon days) compared with those in pre-monsoon season (NPF-pre and non-event days) (Fig.4), thus triggered almost daily NPF events. Similar result was found in the recent study which showed that the Indian summer monsoon acted as a facilitator for transporting gaseous pollutants (Yin et al., 2021).

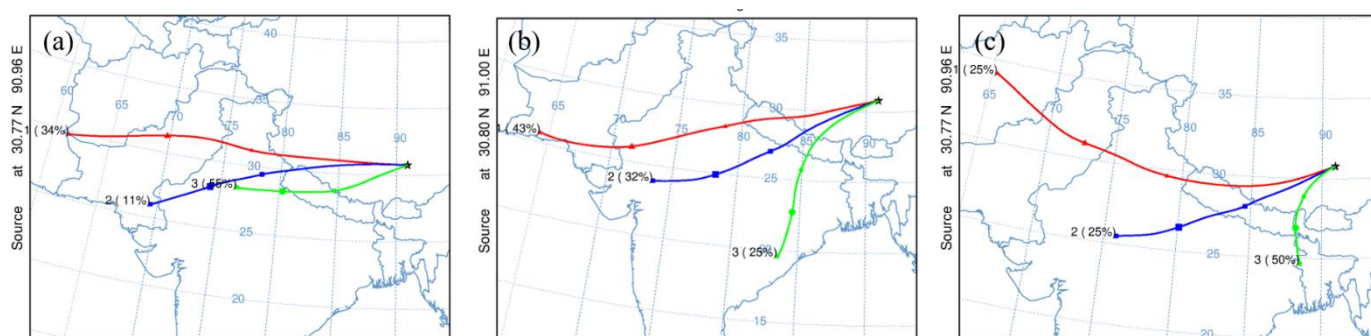


Figure 4. The frequencies of the 48 h back trajectories of air masses arriving at Nam Co station from different directions during the occurrence time of nucleation (11:00-18:00) in (a) non-event days, (b) NPF-pre days and (c) NPF-monsoon days.”

5. The enhancements of CCN concentrations due to NPF is reported in 3 different ways in section 3.4: 1) using an enhancement factor EF, 2) using percentage increases, and 3) stating that something is N times higher than... This is confusing. I highly recommend the authors to unify this discussion.

Thanks for the comment. Considering the possible confusion, we have unified the discussion (N times higher than) in the revised manuscript as follows:

“3.4 Significant increase of CCN in monsoon season

The high-frequency NPF events in the summer monsoon period markedly increased the number concentration of atmospheric aerosols. The average PNSD during pre-monsoon and monsoon seasons were plotted in Fig. 7a with much higher number concentrations observed during monsoon season. The mean total particle number concentration (PN₄₋₇₀₀) in monsoon season was $3647 \pm 2671 \text{ cm}^{-3}$, which was more than 2 times higher than that of pre-monsoon season ($1163 \pm 1026 \text{ cm}^{-3}$). Although the measured particle size ranges were not the same with other studies, the results can still be comparable as the background particles were mainly distributed in tens to hundreds of nanometers. As shown in Table 1, PN₄₋₇₀₀ at Nam Co station in pre-monsoon season was comparable with other high-altitude sites around the world, while PN₄₋₇₀₀ in monsoon season was much higher due to the frequent NPF events. The atmospheric particles contributed by new particle nucleation and growth in monsoon season were mainly concentrated below 100 nm. Among them, the

425 concentration of Nucleation-mode particles in monsoon season was about 2 times higher than that in pre-monsoon season,
426 and the concentration of Aitken-mode particles was 3.5 times higher than that in pre-monsoon season. In contrast,
427 Accumulation-mode particles (>100 nm) which were related to the secondary formation process and long-range transport
428 (Wang et al., 2013; Vu et al., 2015), were nearly at the same level in the two seasons (around 200 cm⁻³). As for CCN, the
429 average number concentrations of CCN at S_c of 0.6% and 1.2% in monsoon season were 434±242 and 863±628 cm⁻³,
430 respectively (Fig. 8). The results were about 0.1 and 0.6 times higher than those in pre-monsoon season at S_c of 0.6%
431 (396±177 cm⁻³) and 1.2% (552±261 cm⁻³).

432 In addition to the average particle number concentration in the two seasons, the important impact of NPF events is
433 more reflected in the increased number concentration of aerosol and CCN in a short time, that is, the aerosol/ CCN
434 production. The aerosol production and CCN production during an NPF event can be obtained by comparing the particle
435 number concentration at the beginning of the increase of the target particles (N_{init}) with the maximum number
436 concentration (N_{max}), as introduced by Rose et al. (2017). The N_{init} and N_{max} are hourly average concentrations as shown
437 in Fig. S16. Fig. 9 showed the average daily variation of PNSD, three modes particles and CCN in the whole periods in
438 pre-monsoon season and monsoon season. In monsoon season, Nucleation-mode particles (4-25 nm) started to increase
439 quickly at around 11:00 when nucleation occurred. The freshly nucleated particles grew to larger sizes due to
440 condensation and coagulation of the pre-existing particles within several hours, which contributed to the increase of
441 Aitken-mode particles (25-100 nm) from 15:00. The average daily aerosol production of Nucleation-mode particles and
442 Aitken-mode particles in monsoon season was around 3400 cm⁻³ and 1200 cm⁻³, respectively. As for CCN, the average
443 production of CCN at S_c of 0.6% and 1.2% in monsoon season was 180 and 518 cm⁻³, respectively. The production of
444 CCN were lower than previous studies because they only considered NPF days and their environment was relatively
445 polluted (Rose et al., 2017; Shen et al., 2016). The production of aerosols and CCN was much lower in pre-monsoon
446 season, although the particles and CCN at around 11:00 were comparable with monsoon season. The average daily
447 aerosol production of Nucleation-mode particles and Aitken-mode particles in pre-monsoon season were around 500 cm⁻³
448 and 300 cm⁻³, respectively. And the CCN production at S_c of 0.6% and 1.2% was 160 and 286 cm⁻³, respectively. The
449 average daily production of Nucleation-mode particles, Aitken-mode particles and CCN at S_c of 1.2% in monsoon
450 season was 5.8, 3 and 0.8 times higher than that in pre-monsoon season, respectively.”

451 452 **Reply to Reviewer 1’s minor comments (5):**

453 *1. line 49: I suppose that the authors mean quantities like the particle formation and growth rate as*
454 *referring to parameters of NPF. I do not feel that parameter is a good wording here, rather suggesting*
455 *something as characteristics of NPF.*

456 Thanks for the help comment. We have made correction in the revised manuscript.

457 **2. lines 127-128: Classification of a NPF event seems untypical. Has the performance of this classification**
458 **method tested and has it been used in other studies besides Fang et al. (2020)? The word obviously does**
459 **not fit into this context.**

460 Thanks for the comment. The classification method of NPF events has been used in Tang et al (2021) and
461 Shang et al (2022). Considering the possible misunderstanding, we have described the classification
462 method in more detail in the revised manuscript.

463 Tang, L., Shang, D., Fang, X., Wu, Z., Qiu, Y., Chen, S., Li, X., Zeng, L., Guo, S., and Hu, M.: More Significant Impacts From
464 New Particle Formation on Haze Formation During COVID-19 Lockdown, *Geophysical Research Letters*, 48,
465 e2020GL091591, <https://doi.org/10.1029/2020GL091591>, 2021.

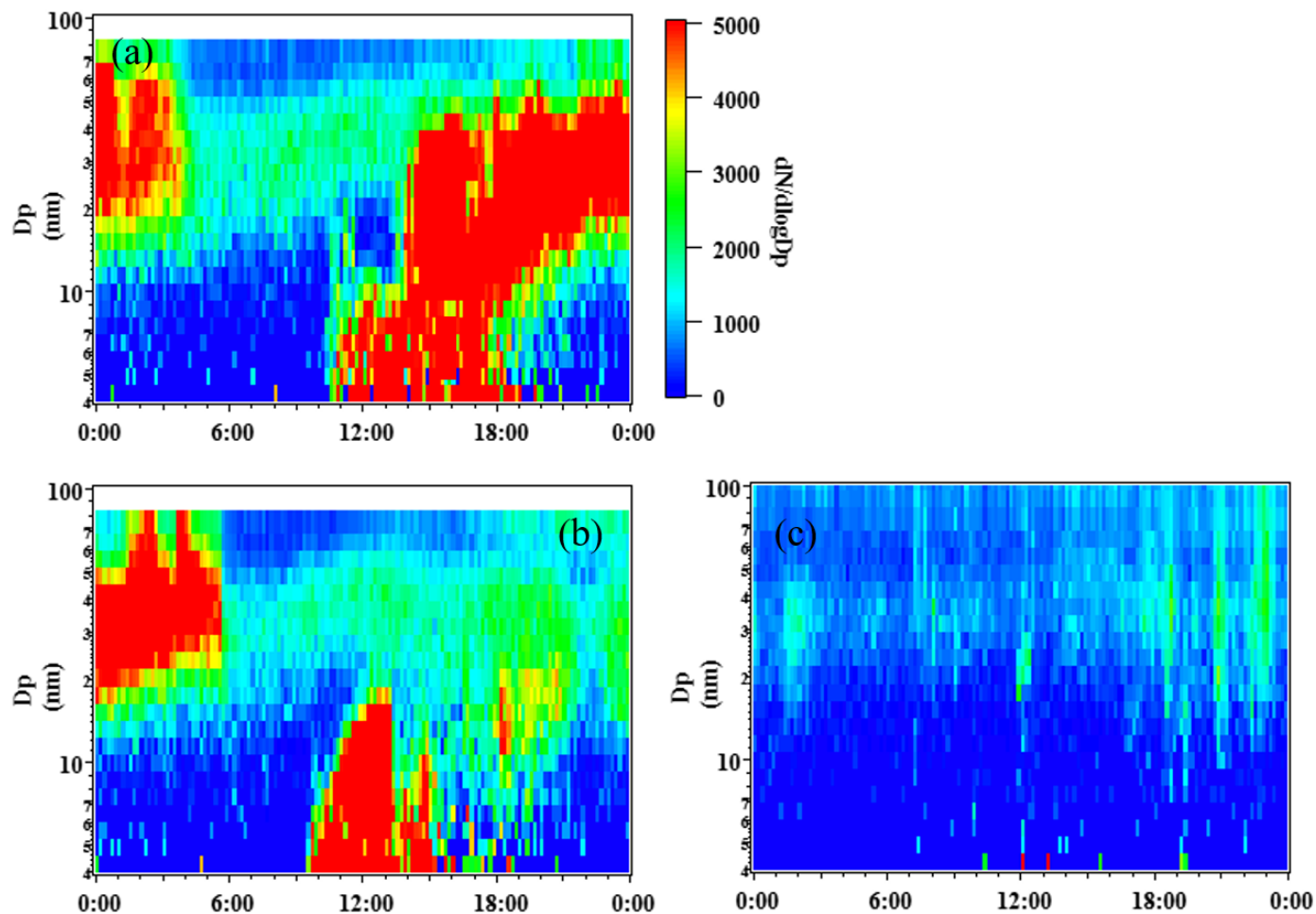
466 Shang, D., Tang, L., Fang, X., Wang, L., Yang, S., Wu, Z., Chen, S., Li, X., Zeng, L., Guo, S., and Hu, M.: Variations in source
467 contributions of particle number concentration under long-term emission control in winter of urban Beijing,
468 *Environmental Pollution*, 304, 119072, <https://doi.org/10.1016/j.envpol.2022.119072>, 2022.

469 **“2.3 Parameterization of NPF**

470 In this study, a typical NPF event was defined by that there is a burst in the 3-10 nm particles number concentration
471 (PN₃₋₁₀) and subsequent growth of these newly formed particles (Fang et al., 2020; Dal Maso et al., 2005). The days
472 without newly particle formation were defined as non-event days. Days in which the increase of PN₃₋₁₀ was observed
473 without the particles growing to the larger size, or days when we can see the later phase of a mode growing in the Aitken
474 mode size range were treated as undefined days (Dal Maso et al., 2005). The examples of the classification of NPF events
475 are shown in Fig.S8.

476 Fang, X., Hu, M., Shang, D., Tang, R., Shi, L., Olenius, T., Wang, Y., Wang, H., Zhang, Z., Chen, S., Yu, X., Zhu,
477 W., Lou, S., Ma, Y., Li, X., Zeng, L., Wu, Z., Zheng, J., and Guo, S.: Observational Evidence for the
478 Involvement of Dicarboxylic Acids in Particle Nucleation, *Environmental Science & Technology Letters*, 7,
479 388-394, [10.1021/acs.estlett.0c00270](https://doi.org/10.1021/acs.estlett.0c00270), 2020.

480 Dal Maso, M., Kulmala, M., Riipinen, I., and Wagner, R.: Formation and growth of fresh atmospheric aerosols:
481 Eight years of aerosol size distribution data from SMEAR II, Hyytiälä, Finland, *Boreal Environment Research*,
482 10, 323-336, 2005.



483 **Figure S8.** Typical particle number size distributions of (a) a NPF day (23 June, 2019), (b) an undefined day (24 June, 2019),
 484 and (c) a non-event day (15 May, 2019).”
 485
 486

487 **3. lines 225-233: The authors list a number of things that potentially affect the occurrence of NPF. The list**
 488 **misses one highly relevant quantity: the intensity of solar radiation. This quantity should be mentioned**
 489 **here.**

490 Thanks for the comment. We have made correction in the revised manuscript as follows:

491 “Whether an NPF event can occur is mainly related to 1) the CS, which mainly referred to the scavenging rate of
 492 precursors, clusters, and newly formed particles by background aerosols. High CS can lead to the continual reduction in
 493 newly formed particle number concentration, and inhibit the occurrence of NPF; 2) the gaseous precursors that can
 494 participate in nucleation and growth, including sulfuric acid (Kulmala et al., 2013), dimethylamine (Yao et al., 2018),
 495 ammonia (Xiao et al., 2015) and VOC (Tröstl et al., 2016; Fang et al., 2020; Qiao et al., 2021). A sufficiently high
 496 concentration of low volatility vapors (precursors) can contribute to persistent nucleation and generating new
 497 atmospheric particles; 3) air mass origins and meteorological factors including WD, RH, temperature, the intensity of
 498 solar radiation, etc, which can influence the occurrence and intensity of NPF events by directly or indirectly affecting
 499 the source and sink parameters.”

502 **4. Unlike the Aitken mode, the nucleation mode is usually written using a lower-case letter (lines 326, 359,**
503 **383)**

504 Thanks for the comment. The nucleation mode is written using a lower-case letter in the revised
505 manuscript:

506 “nucleation mode particles”

507
508 **5. In several places (lines 225, 254, 258, 261, 284, 296, 299, 340), the use of tense is somewhat wrong, or at**
509 **least uncommon. Please reconsider which tense to use in these places.**

510 Thanks for the comment. We have made correction in the revised manuscript as follows:

511 “Whether an NPF event can occur is mainly related to 1) the CS, which mainly referred to the scavenging rate of
512 precursors, clusters, and newly formed particles by background aerosols.”

513 “Gaseous sulfuric acid is identified as the key precursor for nucleation and initial growth due to its low
514 volatility (Kulmala et al., 2013; Qiao et al., 2021).”

515 “In addition to sulfuric acid, organics are also considered to be an important factor of NPF events. Observations
516 and laboratory experiments have found that organics may participate in or even dominate the nucleation and growth
517 process in NPF events in pristine environments and the preindustrial atmosphere.”

518 “The frequencies of the 48 h back trajectories of air masses arriving at Nam Co station during the occurrence time
519 of nucleation (11:00-18:00) in non-event days, NPF-pre days and NPF-monsoon days are present in Fig. 5.”

520 “In Fig. 6, we show the diurnal variations of meteorological factors during NPF-pre days, NPF-monsoon days and
521 non-event days at Nam Co station.”

522 “The similar temperature in NPF-pre days and non-event days suggests that temperature is not a crucial factor for
523 NPF event occurrence.”

524 “The average PNSD during pre-monsoon and monsoon seasons are plotted in Fig. 7a with much higher number
525 concentrations observed during monsoon season.”

526
527
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531
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533

534 II. Reply to Reviewer 2

535 Reply to Reviewer 2's overall comments:

536 *New particle formation (NPF) at high altitudes is crucial to understand sources of aerosol and CCN in the*
537 *free troposphere. In this study, the authors conducted intensive measurements at Nam Co station (4379 m*
538 *a.s.l) in the central TP to understand the new particle formation during pre-monsoon and monsoon seasons.*
539 *They identified the frequency of NPF during monsoon seasons was significantly higher than during pre-*
540 *monsoon seasons. This study did provide valuable observation data. But the explanation that higher VOCs*
541 *triggered the frequent NPF during monsoon season is unconvincing. Therefore, the manuscript is not*
542 *recommended to be published on ACP unless the authors can address the following major concerns.*

543 We appreciate the comments from the reviewer on this manuscript. We have answered them point to point
544 in the following paragraphs (the texts italicized are the comments, the texts indented are the responses,
545 and the texts in blue are revised parts in new manuscript). In addition, all changes made are marked in the
546 revised manuscript.

547 Reply to Reviewer 2's comments (3):

548 *1. Sulfuric acid (SA), Ultra/Extremely Low Volatility Organic Compounds (U/ELVOCs), and bases, e.g.,*
549 *NH₃ or DMA, are known as the essential precursor of NPF. Their concentrations determine whether NPF*
550 *can occur, as well as the intensity. However, in this study, all these key precursors were not measured. Even*
551 *the precursors of these "direct precursors", e.g. SO₂ and VOCs who can form low-volatile oxidation*
552 *products, were not well measured either. First, the simulated concentration of SO₂ used in this study*
553 *without any verification by observation data is not convincing. Since SO₂ is a very reactive species, one*
554 *needs to use the simulated value very carefully. Second, although 99 types of VOCs were measured during*
555 *pre-monsoon using a GC-MS/FID, they are key precursors of ozone formation and are not suitable as*
556 *indicator precursors for ELVOCs. The author, at least, needs to provide the concentration of monoterpenes,*
557 *which are well-known sources of ELVOCs. In addition, the simulation of VOCs during monsoon is needed*
558 *to be verified. In summary, the authors need to provide more solid evidence to support their main conclusion*
559 *that higher VOCs triggered the frequent NPF during monsoon season.*

560 Thanks for the comment. The measurements of the NPF precursors including SO₂ and VOC are limited
561 due to the harsh conditions and logistical limitations, with only VOC in pre-monsoon season. So we
562 utilized WRF/CMAQ modeling system to simulate the levels of SO₂ and VOC in the whole observation
563 period, to assist in the analysis of the role of sulfuric acid and organics. As for the verification of the
564 simulated SO₂ and VOC, we have added the analysis on statistical parameters of model evaluation and
565 correlation analysis with other tracers in the revised manuscript. Firstly, the WRF/CMAQ models

566 successfully reproduced the meteorological fields and air pollutants including PM and O₃ with model
567 performance indices meeting the suggested benchmarks, which means the simulated meteorological fields
568 and particulate and gaseous pollutants are qualified. For VOC, the observed VOC and predicted VOC in
569 pre-monsoon season were compared to examine the model performance. The benchmarks for VOC had
570 not been reported, but the statistical metrics of MFB (mean fractional bias, -0.47) and MFE (mean
571 fractional error, 0.49) in this study are within the range reported in previous VOC modelling result (Hu et
572 al., 2017). The correlation coefficient (R) between simulated and observed VOC is 0.41, which reflected
573 that the model can fairly simulate the variation of VOC concentration. For SO₂, the WRF/CMAQ models
574 have been successfully reproduced SO₂ in major regions in China with R of 0.25-0.79 (Mao et al., 2022).
575 The simulated SO₂ level in the model domain is comparable with that measured at Mt. Yulong (Shang et
576 al., 2018), with average values of 0.03 ± 0.02 ppbv and 0.06 ± 0.05 ppbv. At the same time, considering that
577 both BC and SO₂ are mainly emitted from coal combustion and biomass burning, BC could be a good
578 indicator for SO₂ especially for pristine environment without local anthropogenic source emissions. A
579 good correlation between SO₂ and BC measured at Mt. Yulong was found with correlation coefficient (R)
580 of 0.79 (Shang et al., 2018). In this study, the modelled SO₂ and measured BC also showed good
581 correlation with R of 0.58. In general, the results of model simulation showed good performance in
582 statistical parameters and correlation analysis with other tracers. The modelled VOC and SO₂ could be
583 used for the NPF analysis.

584 As for monoterpene such as α -pinene, unfortunately, it was not measured in this study. And there can be
585 some other reference compound such as OVOC for understanding new particle formation from tree
586 emissions as indicated by the plant chamber experiment (Mentel et al., 2009). In addition, recent studies
587 showed that aromatic compounds such as benzene, toluene, and naphthalene, and C₆–C₁₀ alkanes can
588 produce considerable amounts of highly oxygenated products through multi-generation OH oxidation or
589 autoxidation (Garmash et al., 2020; Wang et al., 2021), which may trigger the occurrence of NPF events.
590 Therefore, we prefer that different VOC can affect the occurrence of NPF, and we mainly use the
591 concentration of total VOC for analysis.

592 On the whole, we examined the potential reasons for the distinct NPF frequency using the measured CS,
593 precursors, meteorology and simulated SO₂ and VOC. The comprehensive analysis points to the
594 important role of organics. The higher NPF frequency driven by the higher organics concentration in
595 monsoon season can be supported in one recent research which has found that the concentration of
596 monoterpene-derived HOMs in East Asia was higher in summer (June-August) than that in Spring
597 (March-May) by using GEOS-Chem global chemical transport model (Xu et al., 2022).

- 598 Hu, J., Chen, J., Ying, Q., and Zhang, H.: One-year simulation of ozone and particulate matter in China using WRF/CMAQ
599 modeling system, *Atmos. Chem. Phys.*, 16, 10333-10350, 10.5194/acp-16-10333-2016, 2016.
- 600 Shang, D., Hu, M., Zheng, J., Qin, Y., Du, Z., Li, M., Fang, J., Peng, J., Wu, Y., Lu, S., and Guo, S.: Particle number size
601 distribution and new particle formation under the influence of biomass burning at a high altitude background site at Mt.
602 Yulong (3410 m), China, *Atmos. Chem. Phys.*, 18, 15687-15703, 10.5194/acp-18-15687-2018, 2018.
- 603 Mao, J., Li, L., Li, J., Sulaymon, I. D., Xiong, K., Wang, K., Zhu, J., Chen, G., Ye, F., Zhang, N., Qin, Y., Qin, M., and Hu, J.:
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619 “2.3 Model simulation

620 Considering the limited measurements on SO₂ and VOC in this observation (only VOC during pre-monsoon
621 season), Weather Research and Forecasting/Community Multiscale Air Quality (WRF/CMAQ) modeling system
622 was adopted to simulate the level of SO₂ and VOC in the whole observation period, to assist in the analysis of the
623 role of sulfuric acid and organics in NPF events.

624 Weather Research and Forecasting (WRF) (version 4.2.1) model was used to simulate the meteorological
625 conditions with the FNL reanalysis dataset. The 6 h FNL data were obtained from the U.S. National Centre for
626 Atmospheric Research (NCAR), with a spatial resolution of 1.0° × 1.0° (<http://rda.ucar.edu/datasets/ds083.2/>, last
627 accessed on 28 April 2022). The Community Multiscale Air Quality version 5.3.2 (CMAQv5.3.2) model, being one
628 of the three-dimensional chemical transport models (CTMs) (Appel et al., 2021), configured with the gas-phase
629 mechanism of SAPRC07tic and the aerosol module of AERO6i, was employed in this study to simulate the air
630 quality over Tibet in the observation period (26 April to 22 May and 15 June to 25 June in 2019). Air quality
631 simulations were performed with a horizontal resolution of 12 km. The corresponding domain covered Tibet and
632 the surrounding countries and regions with 166 × 166 grids (Fig. S2), with the 18 layers in vertical resolution.
633 Detailed information about the model setting is provided in Text S2.

634 The Multi-resolution Emission Inventory for China version 1.3 (MEICv1.3) (<http://www.meicmodel.org>) and
635 Regional Emission inventory in ASia (REASv3.2) (<https://www.nies.go.jp/REAS/>) were used to provide the
636 anthropogenic emissions from China and neighboring countries and regions, respectively. The MEICv1.3 emissions
637 of the year 2019 were used. For REAS, the emission inventory in the year 2015 was used for 2019 as no emission

638 inventory was released for the years after 2015. Although emission inventories are usually released 3 years behind,
639 we acknowledge that this may cause additional uncertainties in the simulation. Biogenic emissions were generated
640 using the Model for Emissions of Gases and Aerosols from Nature (MEGANv2.1) (Guenther et al., 2012). The open
641 biomass burning emissions were processed using the Fire Inventory for NCAR (FINN) during the entire study
642 period (Wiedinmyer et al., 2011).

643 The model evaluation is introduced in Text S2. The WRF and CMAQ models successfully reproduced the
644 meteorological fields and air pollutants including PM and O₃ with model performance indices meeting the suggested
645 benchmarks. For VOC, the observed VOC and predicted VOC in pre-monsoon season were compared to examine
646 the model performance. The benchmarks for VOC had not been reported, but the statistical metrics of MFB (mean
647 fractional bias, -0.47) and MFE (mean fractional error, 0.49) in this study are within the range reported in previous
648 VOC modelling result (Hu et al., 2017). The correlation coefficient (R) between simulated and observed VOC is
649 0.41, which reflected that the model can fairly simulate the variation of VOC concentration. It should be noted that
650 VOC was underpredicted on the whole, which may due to the uncertainty of the emission inventory as mentioned
651 before. For SO₂, the WRF/CMAQ models have been successfully reproduced SO₂ in major regions in China with
652 R of 0.25-0.79 (Mao et al., 2022). The simulated SO₂ level in the model domain is comparable with that measured
653 at Mt. Yulong (Shang et al., 2018), with average values of 0.03±0.02 ppbv and 0.06±0.05 ppbv. At the same time,
654 considering that both BC and SO₂ are mainly emitted from coal combustion and biomass burning, BC could be a
655 good indicator for SO₂ especially for pristine environment without local anthropogenic source emissions. As shown
656 in Fig.S6, a good correlation between SO₂ and BC measured at Mt. Yulong was found with correlation coefficient
657 (R) of 0.79 (Shang et al., 2018). In this study, the modelled SO₂ and measured BC also showed good correlation
658 with R of 0.58 (Fig. S6). In general, the results of model simulation showed good performance in statistical
659 parameters and correlation analysis with other tracers. The modelled VOC and SO₂ may be helpful for the NPF
660 analysis.”

661 “2.3 Text S2 Model simulation

662 Model Configurations

663 The meteorological conditions were simulated using the Weather Research and Forecasting (WRF) (version
664 4.2.1) model with the FNL reanalysis dataset. The 6 h FNL data were obtained from the U.S. National Centre for
665 Atmospheric Research (NCAR), with a spatial resolution of 1.0° × 1.0° (<http://rda.ucar.edu/datasets/ds083.2/>, last
666 accessed on 28 April 2022). The physical parameterizations used in this study are the Thompson microphysical
667 process, RRTMG longwave/shortwave radiation scheme; Noah land-surface scheme; MYJ boundary layer scheme;
668 and modified Tiedtke cumulus parameterization scheme. The detailed configuration settings could be found in the
669 works of Hu et al. (2016), Mao et al. (2022), Wang et al. (2021a).

670 The Community Multiscale Air Quality version 5.3.2 (CMAQv5.3.2) model, being one of the three-

671 dimensional chemical transport models (CTMs) (Appel et al., 2021), configured with the gas-phase mechanism of
672 SAPRC07tic and the aerosol module of AERO6i, was employed in this study to simulate the air quality over Tibet
673 from 24 April to 24 May and 13 June to 27 June in 2019, which contains the observation period. Air quality
674 simulations were performed with a horizontal resolution of 12 km. The corresponding domain covered Tibet and
675 the surrounding countries and regions with 166×166 grids (Fig. S2), with the 18 layers in vertical resolution. The
676 initial and boundary conditions were provided by the default profiles. The simulated results of the first two days
677 were not included in the model analysis, which served as a spin-up and reduced the effects of the initial conditions
678 on the simulated results.

679 **Model Evaluation**

680 Previous studies have investigated the impacts of meteorological conditions on the formation, transportation,
681 and dissipation of air pollutants (Hu et al., 2016; Hua et al., 2021; Mao et al., 2022; Sulaymon et al., 2021b;
682 Sulaymon et al., 2021a). Therefore, the evaluation of the WRF model performance was carried out before the usage
683 of its meteorological fields in the CMAQ simulations. The evaluation of the WRF model was achieved by
684 comparing the predicted wind speed (WS, m/s), wind direction (WD, °) at 10 m above the surface, RH (%) and
685 temperature (T, °C) to the observed values. Fig. S3 showed that WS was well simulated both in pre-monsoon and
686 monsoon seasons. WD was well simulated in pre-monsoon season, and there seems to be some deviation in the
687 simulation of north wind in monsoon season. The main reason about the deviation in WD may be due to the poor
688 terrain and complicated weather conditions. Nevertheless, both simulations and measurements showed more
689 frequent southerly winds during monsoon season. RH and temperature were well simulated in the whole periods
690 (Fig. S4). The good model performance with the statistical metrics of WS, RH and temperature meeting the
691 suggested benchmarks are shown in Table S1. Generally, the simulated meteorological fields were qualified and can
692 be further utilized in driving the CMAQ model

693 Fig. S5 showed the comparison of simulated hourly mean concentration about PM, O₃ and VOC in observation
694 site, which were simulated by CMAQ. The statistical indices used in evaluating the CMAQ model were present in
695 Table S2. It can be seen that PM and O₃ meet the suggested benchmarks, which reflect the good model performance.
696 The observed VOC and predicted VOC in pre-monsoon season were compared to examine the model performance.
697 The benchmarks for VOC had not been reported, but the MFB (mean fractional bias) and MFE (mean fractional
698 error) values are within the range reported in previous VOC modelling result (Hu et al., 2017). The correlation
699 coefficient (R) between simulated and observed VOC is 0.41, which reflected that the model can fairly simulate the
700 variation of VOC concentration. It should be noted that VOC was underpredicted on the whole, which may due to
701 the uncertainty of the emission inventory.

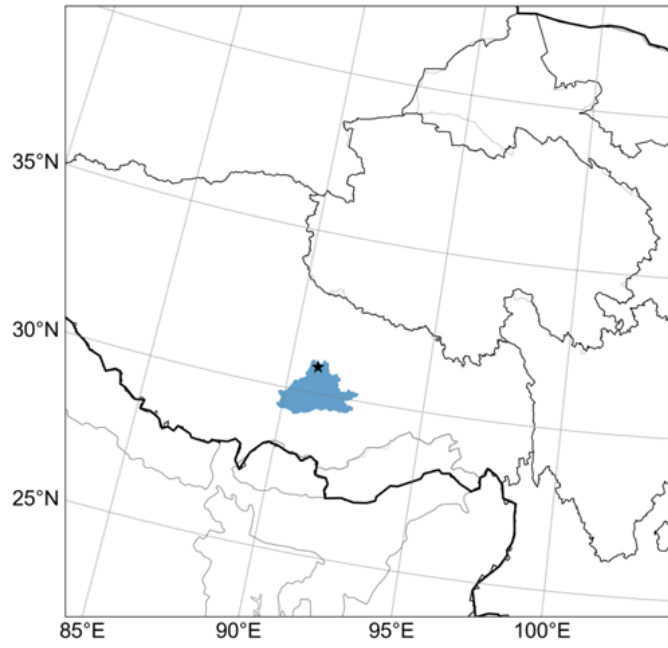


Figure S2. WRF/CMAQ modeling domain

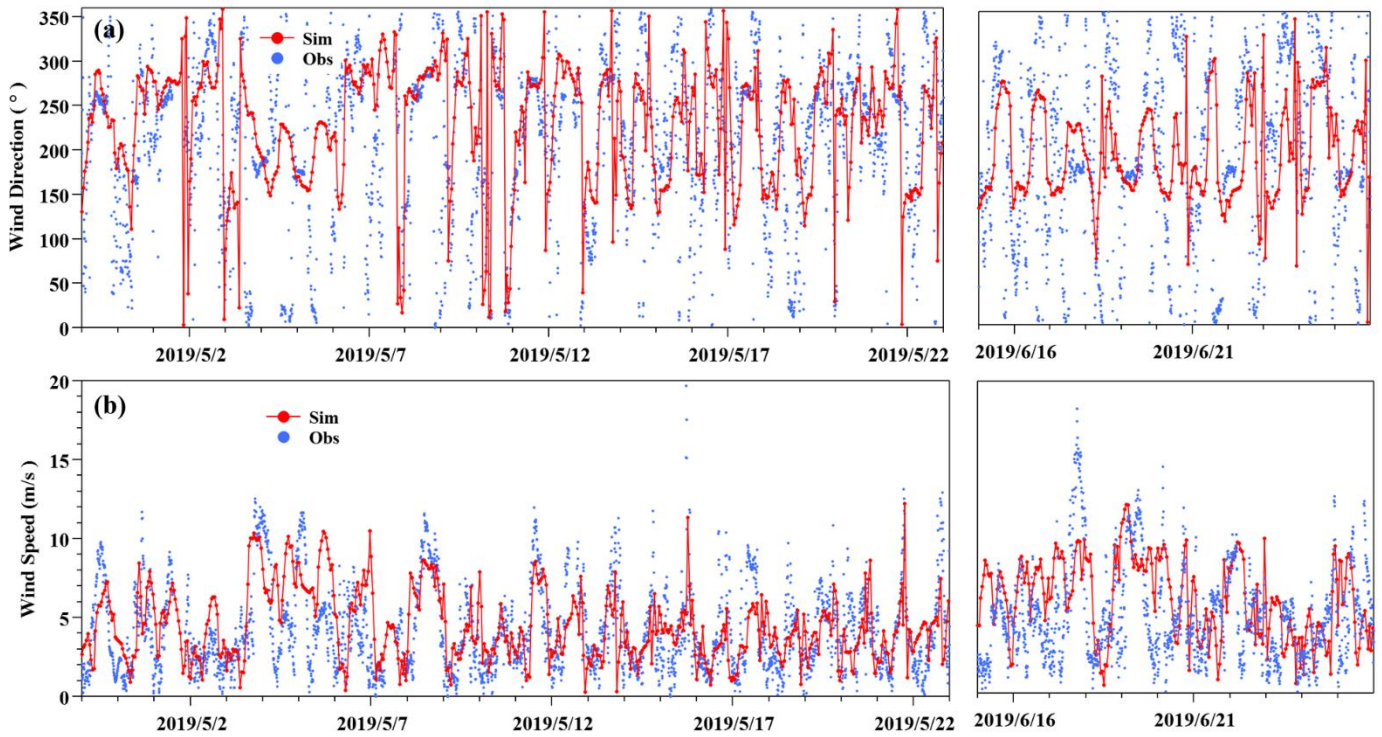


Figure S3. Comparison of simulated (in red dot-line) and observed (in blue dot) wind direction (WD, °) and wind speed (WS, m/s). Observed is 10 minutes mean data. Simulated is hourly mean data.

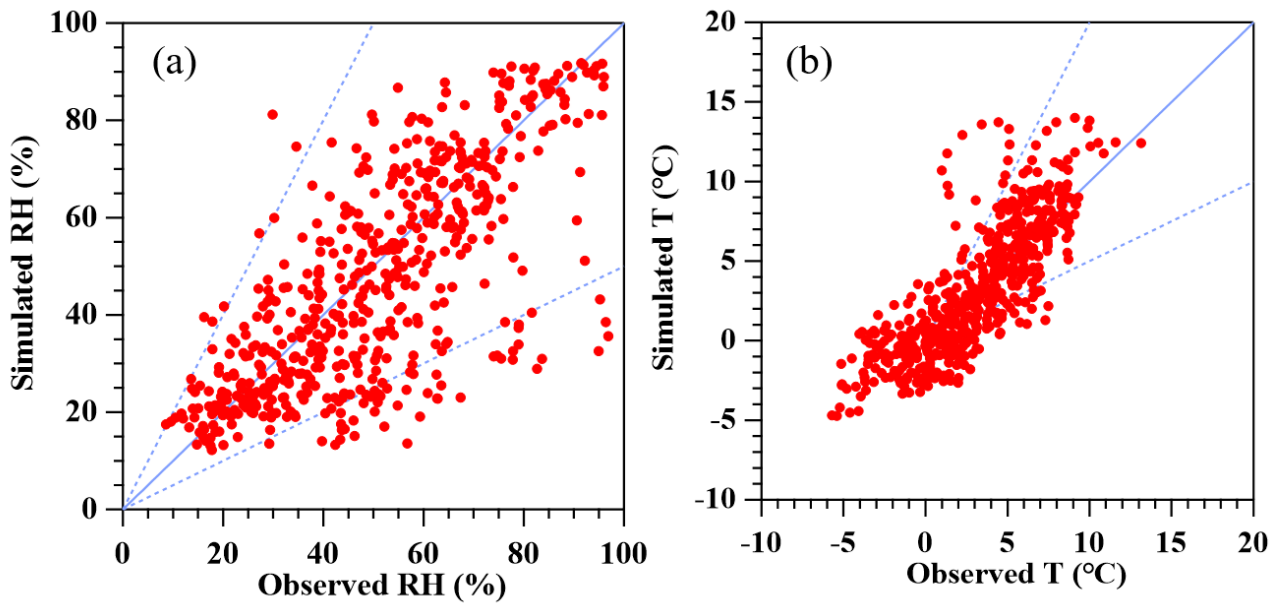


Figure S4. Comparison of simulated and observed RH (%) and temperature (T, °C). RH and temperature are hourly mean data.

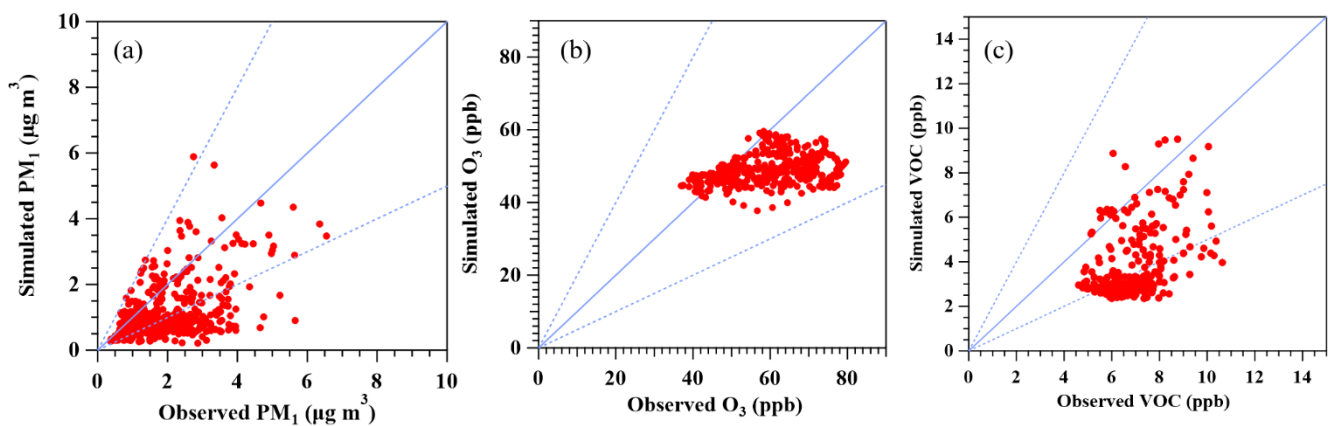


Figure S5. Comparison of simulated and observed PM ($\mu\text{g}/\text{m}^3$), O_3 (ppb) and VOC (ppb). PM, O_3 and VOC are hourly mean concentration.

Table S1. Model performance of meteorological factors at Nam Co station

	WS				RH				T			
	MB	ME	RMSE	R	MB	ME	RMSE	R	MB	ME	RMSE	R
Statistic	0.42	0.87	1.20	0.51	-1.38	12.20	16.30	0.67	0.07	1.85	2.43	0.89
Benchmarks	$\leq \pm 0.5$	≤ 2.0	≤ 2.0						$\leq \pm 0.5$	≤ 2.0		

MB: mean bias; ME: mean error; RMSE: root mean square error; R: correlation coefficient. The benchmarks were suggested by Boylan and Russell (2006).

Table S2. Model performance of the air pollutants at Nam Co station

	PM ₁			O ₃			VOC			SO ₂		
	MFB	MFE	R	NMB	NME	R	MFB	MFE	R	NMB	NME	R
Statistic	0.49	0.50	0.72	0.14	0.23	0.51	-0.47	0.49	0.41			
Benchmarks	<±0.6	<0.75	>0.4	<±0.15	<0.35	>0.5						
References							<±0.77	<0.74		<±4.38	<±4.38	0.25- 0.79

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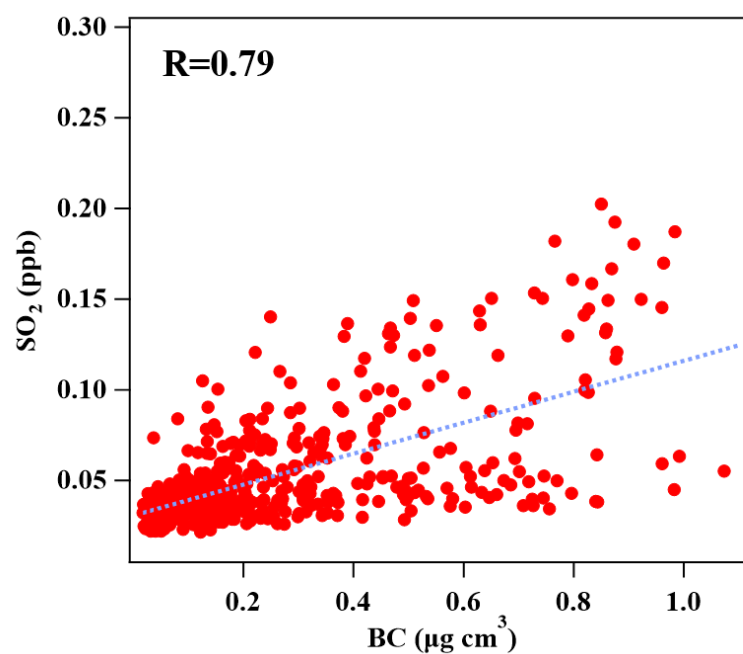
NMB: normalized mean bias; NME: normalized mean error; R: correlation coefficient; MFB: mean fractional bias; MFE:

722

mean fractional error. The benchmarks for PM and O₃ were suggested by Emery et al. (2017) and Boylan and Russell (2006),

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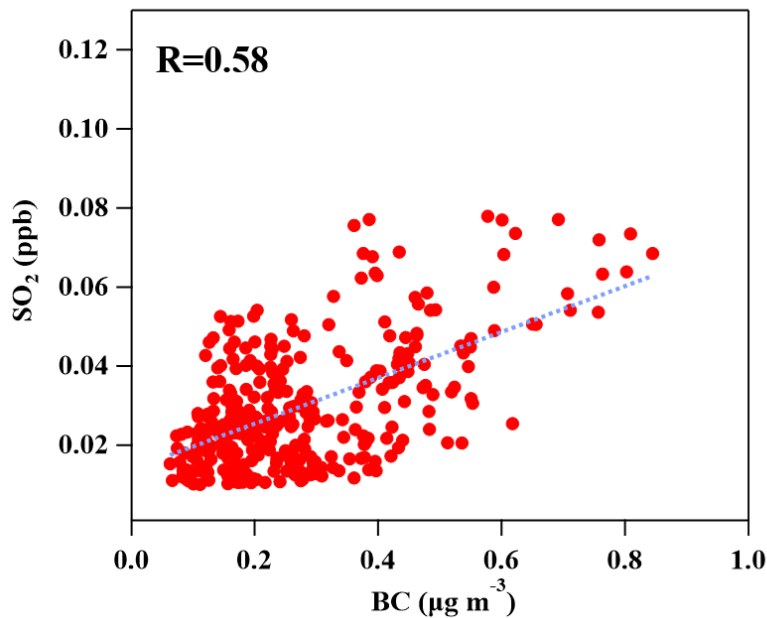
respectively. The references for VOC and SO₂ were from Hu et al. (2017) and Mao et al. (2022), respectively.



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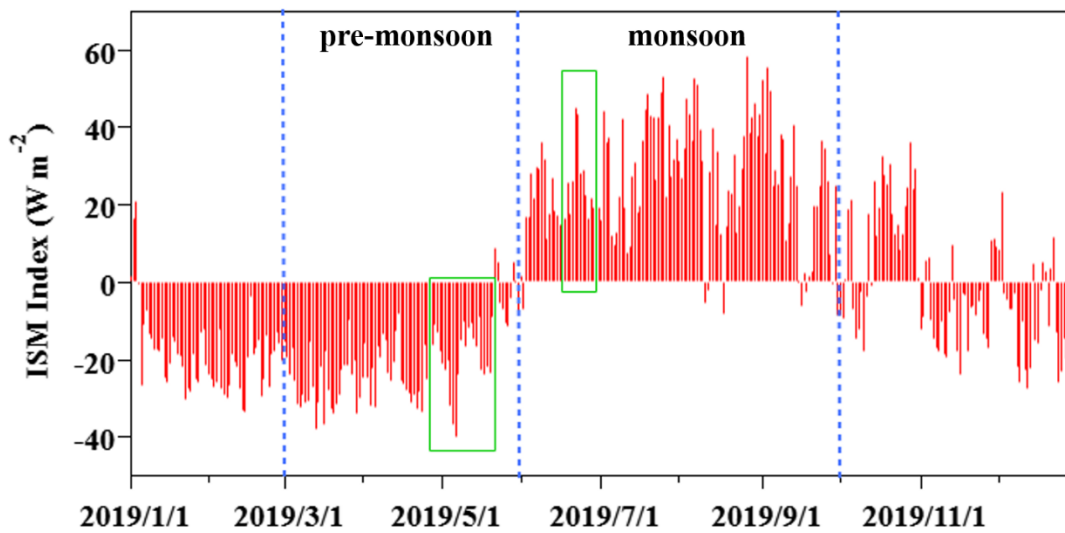
Figure S6. Relationship between SO₂ and BC at Mt. Yulong in 2015. The correlation coefficient R is 0.79.



726
727 **Figure S7.** Relationship between modelled SO₂ and BC at Nam Co station. The correlation coefficient R is 0.58.”
728

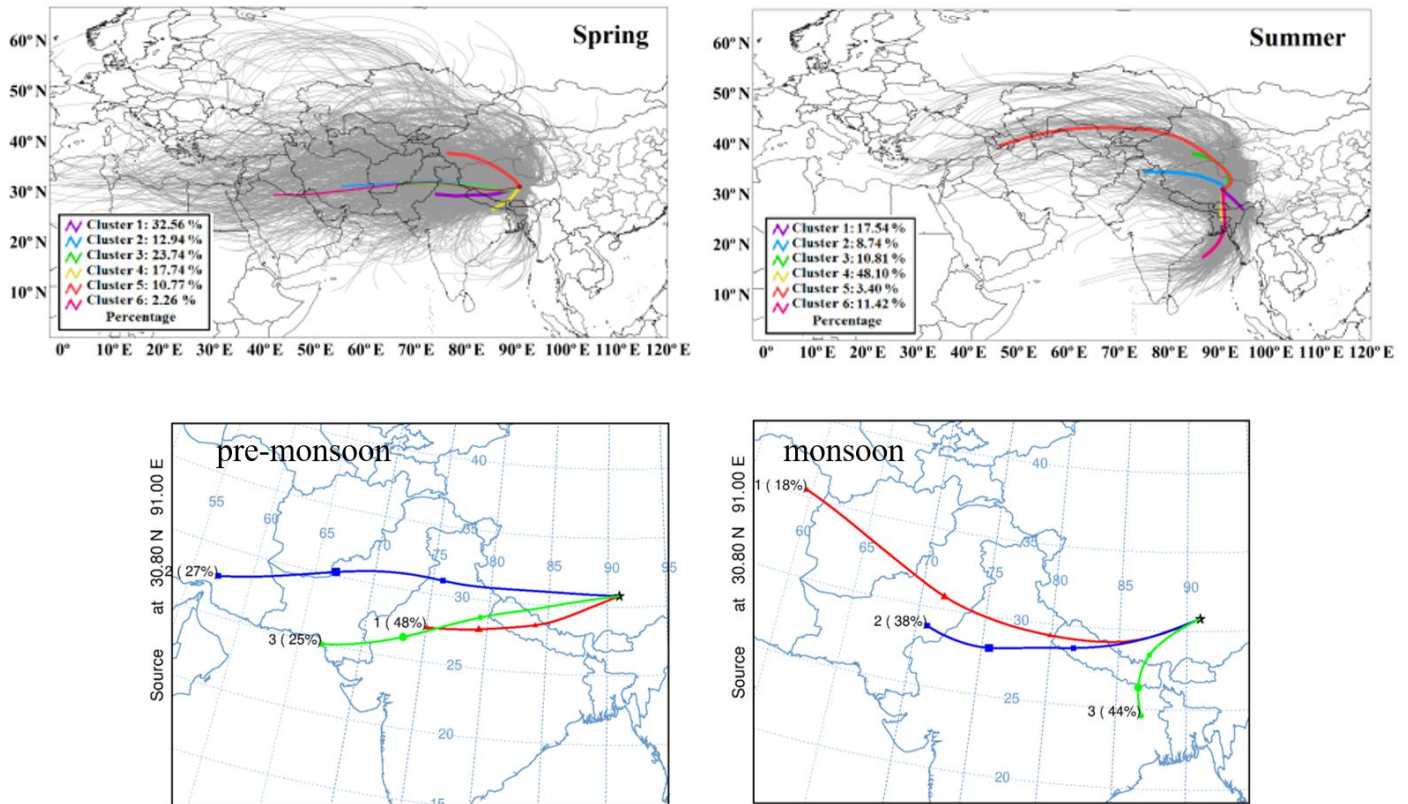
729 **2. The observation period is a bit too short, especially, with only 10 days during monsoon. One cannot be**
730 **sure that the high NPF frequency observed during this 10-day observation can be representative of the**
731 **entire monsoon period.**

732 Thanks for the comment. The measurements periods were a little short as the reviewer described, with
733 about 4 weeks for the pre-monsoon season and 10 days for the monsoon season. But our measurements
734 periods can be representative for this location during pre-monsoon season and monsoon season as follows:
735 1) The intensity of Indian Summer Monsoon during the two measurements periods can represent that in
736 the whole pre-monsoon and monsoon seasons, respectively. The intensity of Indian Summer Monsoon is
737 an important indicator to distinguish the monsoon season. Here the intensity of Indian Summer Monsoon
738 (ISM) was indicated by the ISM Index, which are defined by the negative outgoing longwave radiation
739 anomalies (with respect to the climatological annual cycle) averaged over the Bay of Bengal–India region
740 (10°–25°N, 70°–100°E) (Wang and Fan, 1999). As shown in Fig. R1, the measurement periods (green
741 boxes) were in the pre-monsoon season (March-May) and monsoon season (June-September),
742 respectively. And the IMS index during the two measurements periods were equivalent to those of the
743 whole pre-monsoon season (average: -19.5 vs -20.7 W m⁻²) and monsoon season (average: 27.0 vs 26.3
744 W m⁻²), respectively. Therefore, we considered that these two observation periods are representative in
745 the seasonal characteristics in pre-monsoon season and monsoon season, respectively.

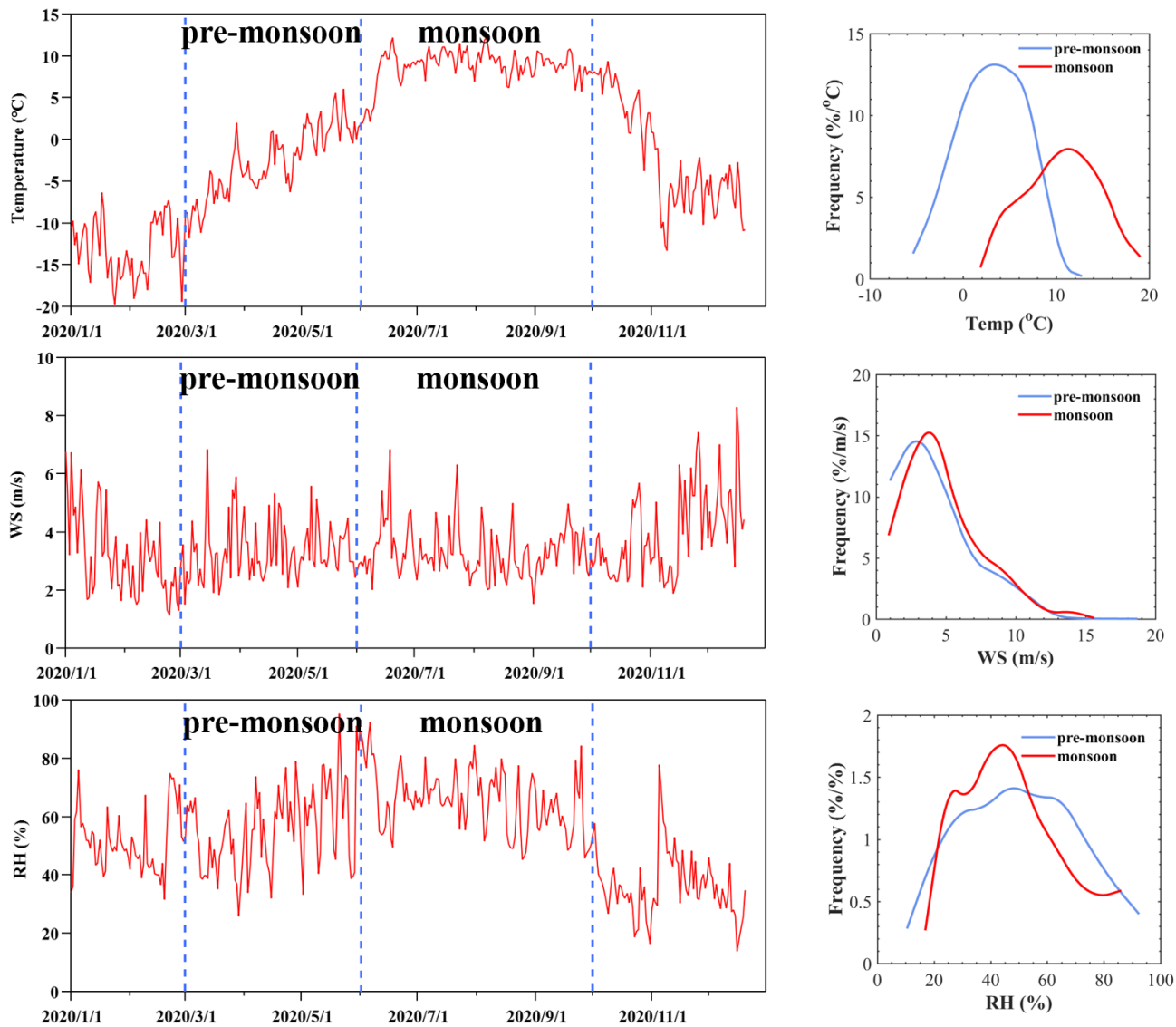


746
747 **Figure R1.** The Indian Summer Monsoon (ISM) Index in 2019. The measurements periods are marked
748 by the green boxes.

749
750 2) The characteristics of meteorology and atmospheric pollutants in the two measurements periods was
751 generally in agreement with the previous long-term studies at Nam Co station and other sites in the Tibetan
752 Plateau (TP) (Yin et al., 2017; Cong et al., 2015; Bonasoni et al., 2010; Xu et al., 2018). Both previous
753 research and this study showed that, strong westerlies pass through western Nepal, northwest India and
754 Pakistan in pre-monsoon season, while air masses were mainly derived from Bangladesh and northeast
755 India and brought moisture that originated in the Bay of Bengal in monsoon season (Fig. R2) (Yin et al.,
756 2017). The temperature, WS and RH in the two measurements periods were matched with those in the
757 whole pre-monsoon and monsoon season of 2020 at Nam Co station (Fig. R3) (National Tibetan Plateau
758 Data Center). The average temperature in pre-monsoon and monsoon seasons were around 3 and 10 °C,
759 respectively. WS showed no difference between pre-monsoon and monsoon seasons with the average
760 value of 4 m/s. The average level of RH was similar between the two seasons, but the variation range of
761 RH was larger in pre-monsoon season. In addition, the level of PM, BC and ozone in the two
762 measurements periods were matched with those in the whole pre-monsoon and monsoon season at the
763 other site of the TP (Fig. R4) (Bonasoni et al., 2010; Xu et al., 2018). The average concentrations of PM₁
764 (PM_{0.8}) in pre-monsoon and monsoon seasons were around 1 and 2 µg/m³, respectively. The concentration
765 of BC was at a level of hundreds nanogram per cubic meter in pre-monsoon season, while at a lower level
766 in monsoon season. Ozone showed the similar pattern with BC. It should be noted that there will be some
767 differences between this study and other results due to the differences in time resolution. In a word, the
768 characteristics of meteorology and atmospheric pollutants in the two measurements periods in this study
769 can well reflect those in pre-monsoon and monsoon seasons.



770
 771 **Figure R2.** Comparison of trajectories between this study and previous study at Nam Co station.
 772 Backward HYSPLIT trajectories for each measurement day (black lines in the maps), and mean back
 773 trajectory for six HYSPLIT clusters (colored lines in the maps) arriving at Nam Co Station in spring
 774 (MAM) and summer (JJA) (Yin et al., 2017) (top). The frequencies of the 48 h back trajectories of air
 775 masses arriving at Nam Co station from different directions during pre-monsoon and monsoon seasons in
 776 this study (bottom).



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Figure R3. Comparison of meteorology between this study and the whole seasons in 2020. Time series of ambient temperature, wind speed and relative humidity at Nam Co station from January 2020 to December 2020 (National Tibetan Plateau Data Center) (left). Comparison in frequency distributions of temperature, WS and RH at Nam Co station in pre-monsoon and monsoon seasons in this study (right).

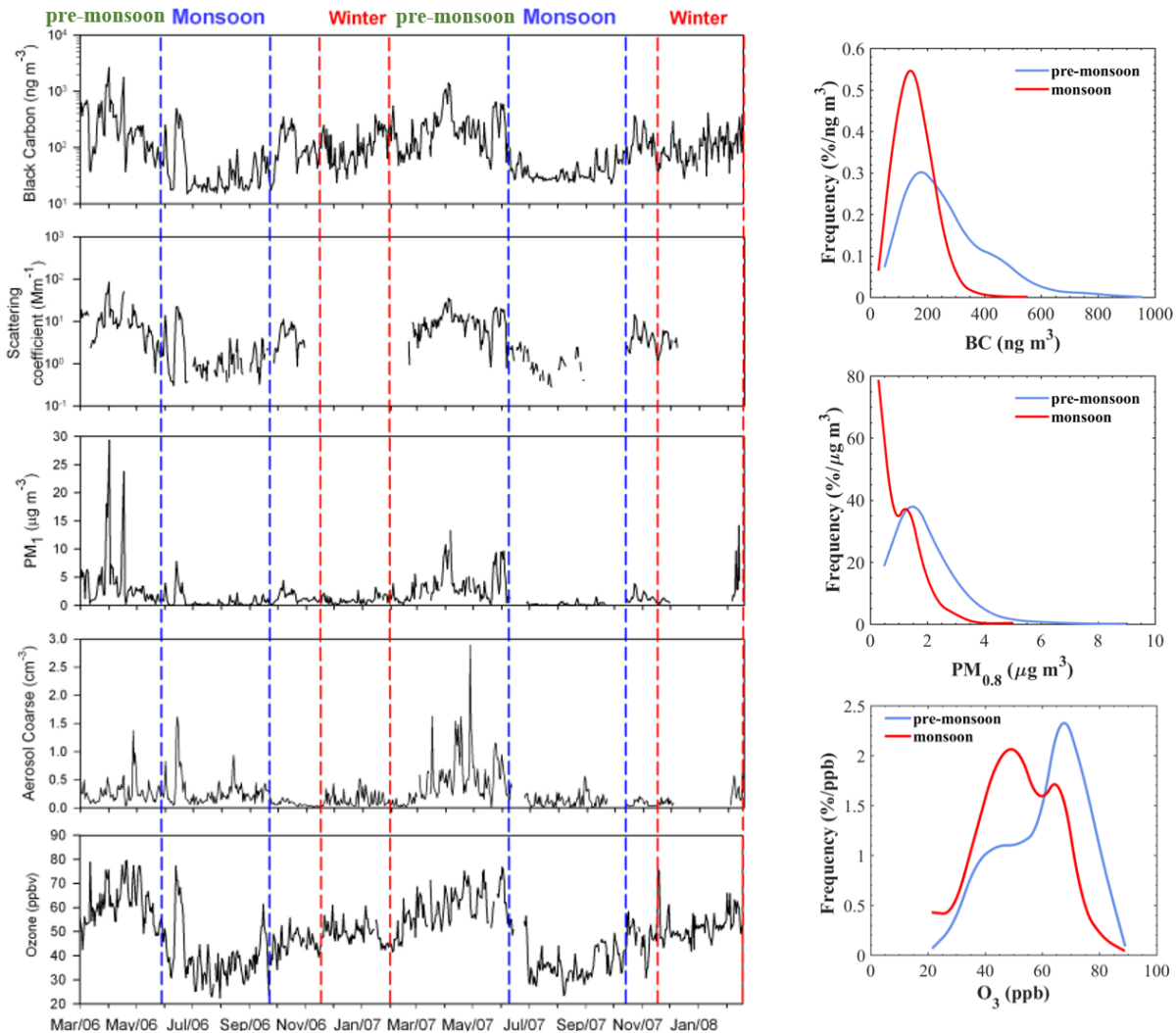


Figure R4. Comparison of atmospheric pollutants between this study and previous TP studies. Time series of f BC, aerosol scattering coefficient, PM₁, coarse particle number and surface ozone at the Nepal Climate Observatory-Pyramid (NCO-P) station from 1 March 2006 to 28 February 2008 (Bonasoni et al., 2010) (left). Comparison in frequency distributions of BC, PM_{0.8} and O₃ at Nam Co station in pre-monsoon and monsoon seasons in this study (right).

3) Based on the above discussion, the two measurements periods in this study can respectively represent the pre-monsoon and monsoon season at Nam Co station, so the NPF characteristics of the two observation periods can also be considered as the NPF characteristics in pre-monsoon and monsoon seasons. It is true that we can not be sure that the extremely high NPF frequency observed during the 10-day observation period can be found in the entire monsoon season. But the difference of NPF frequency between the two seasons is very notable. Here we do not emphasize the absolute value of NPF frequency, but the significant difference of NPF frequency between two seasons. In addition, the phenomenon of higher NPF frequency in monsoon season than pre-monsoon season was also found in the other sites in the Tibetan Plateau (TP). A 16-month measurements from 2006 to 2007 at Himalayan Nepal Climate

798 Observatory at Pyramid (NCO-P) site on the southern TP showed NPF frequency of 38% in pre-monsoon
799 season and 57% in monsoon season (Venzac et al., 2008). At Mt. Yulong on the southeastern TP, the NPF
800 frequency was only 14% during pre-monsoon season (Shang et al., 2018). The NPF frequency of 15% in
801 pre-monsoon season and 80% in monsoon season at Nam Co station was consistent with these studies,
802 with more significant seasonal differences. The significant seasonal differences may be due to the fact
803 that the occurrence of NPF is more sensitive to the monsoon in extremely clean background areas (such
804 as Nam Co station and Mt. Yulong). In summary, our study emphasized the seasonal differences in NPF
805 frequencies at Nam Co station, and the results was reliable.

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829 Mt. Yulong (3410 m), China, *Atmos. Chem. Phys.*, 18, 15687-15703, 10.5194/acp-18-15687-2018, 2018.

830 Due to harsh conditions and logistical limitations, our observation periods were limited. However, our
831 conclusions are obvious and representative, and we will carry out more detailed observations for a longer
832 period in the future if possible. To illustrate the representativeness of the observation periods, we have
833 made supplements in the revised manuscript as follows:

834 “2.1 Measurement site

835 The measurement was conducted from 26 April to 22 May, 2019 and 15 June to 25 June, 2019, and can be
836 representative of the pre-monsoon season and the summer monsoon season, respectively (Text S1) (Bonasoni et al.,
837 2010; Cong et al., 2015).”

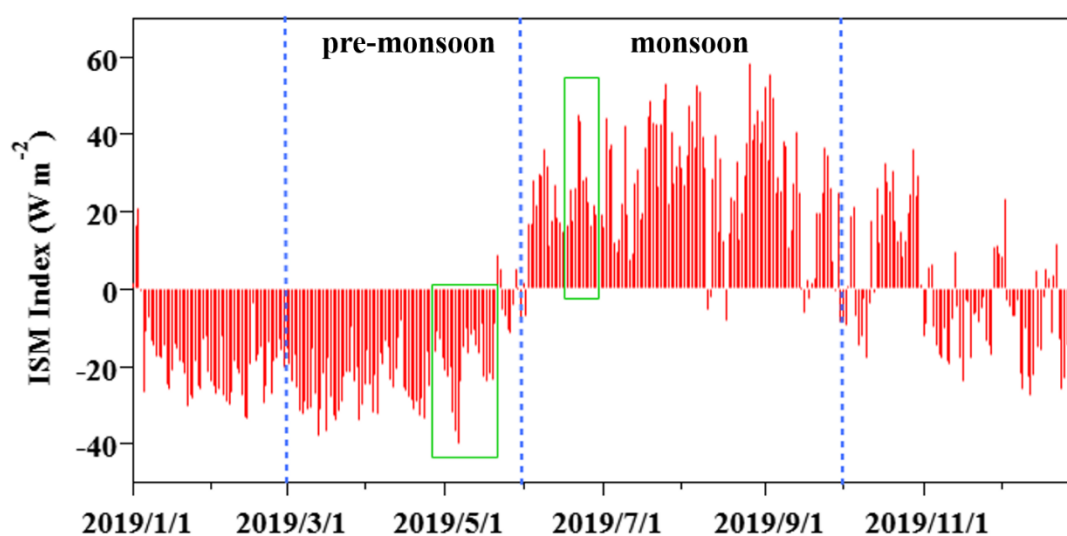
838 **“Text S1 The representativeness of the observation periods**

839 The measurement was conducted from 26 April to 22 May, 2019 and 15 June to 25 June, 2019, and can be
840 representative of the pre-monsoon season and monsoon season, respectively.

841 Firstly, the intensity of Indian Summer Monsoon during the two measurements periods can represent that in
842 the whole pre-monsoon and monsoon seasons, respectively. The intensity of Indian Summer Monsoon is an
843 important indicator to distinguish the monsoon season. Here the intensity of Indian Summer Monsoon (ISM) was
844 indicated by the ISM Index, which are defined by the negative outgoing longwave radiation anomalies (with respect
845 to the climatological annual cycle) averaged over the Bay of Bengal–India region (10°–25°N, 70°–100°E) (Wang
846 and Fan, 1999). As shown in Fig. S1, the measurement periods (green boxes) were in the pre-monsoon season
847 (March-May) and monsoon season (June-September), respectively. And the IMS index during the two
848 measurements periods were equivalent to those of the whole pre-monsoon season (average: -19.5 vs -20.7 W m⁻²)
849 and monsoon season (average: 27.0 vs 26.3 W m⁻²), respectively.

850 Secondary, the characteristics of meteorology and atmospheric pollutants in the two measurements periods
851 was generally in agreement with the previous long-term studies at Nam Co station and other sites in the Tibetan
852 Plateau (TP) (Yin et al., 2017; Cong et al., 2015; Bonasoni et al., 2010; Xu et al., 2018). That is, the characteristics
853 of meteorology (temperature, WS and RH) and atmospheric pollutants (PM, BC and ozone) in the two
854 measurements periods were matched with those in the whole pre-monsoon and monsoon season at Nam Co station
855 and other sites in the TP.

856 Therefore, the two observation periods are representative in the seasonal characteristics in pre-monsoon season
857 and monsoon season, respectively.



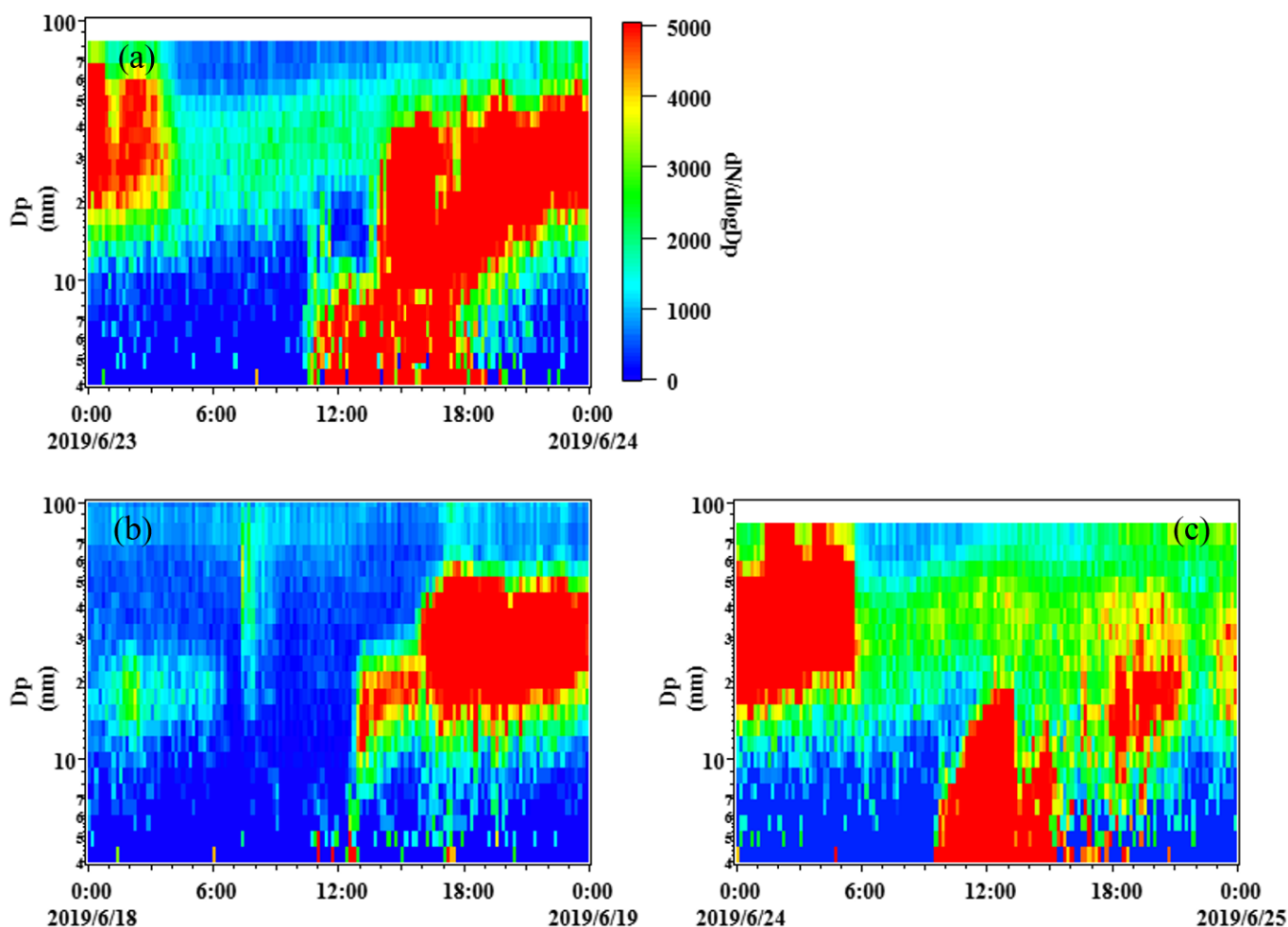
858 **Figure S1.** The Indian Summer Monsoon (ISM) Index in 2019. The measurements periods are marked by the green boxes.”
859
860

861 **3. The authors grouped their data into “NPF-pre days”, “NPF-monsoon days” and “non-event days” in**

862 **Fig. 3- Fig. 7 and related discussions. I may suggest separating “non-event days” into pre-monsoon and**
863 **monsoon non-event days.**

864 Thanks for the comment. There were no typical “non-event days” in monsoon season according to the
865 classification of NPF events. The two days which were not “NPF days” in monsoon season were
866 “undefined days”, as in which the increase of PN_{3-10} was observed without the particles growing to the
867 larger size (24 June, 2019), or we can see the later phase of a mode growing in the Aitken mode size range
868 (18 June, 2019) (Dal Maso et al., 2005). The particle number size distributions in 18 June and 24 June,
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873 **Figure R5.** The particle number size distribution in (a) 23 June, 2019 (NPF day), (b) 18 June, 2019
874 (undefined day), and (c) 24 June, 2019 (undefined day).
875

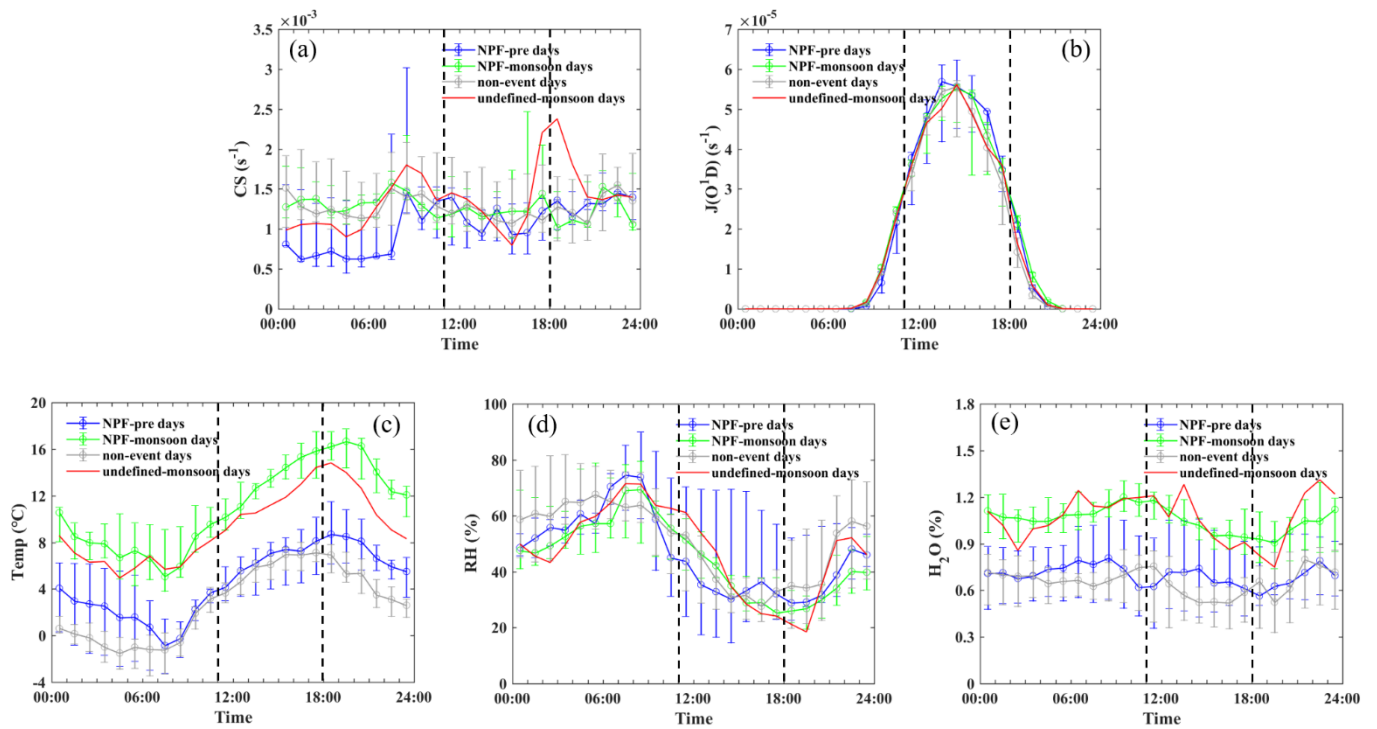
876
877 At the same time, we also analyzed the atmospheric characteristics in “undefined days” in monsoon
878 season (Fig. R6). And we found that the atmospheric characteristics including CS, $J(O^1D)$, temperature,
879 RH, wind speed, water content in “undefined days” in monsoon season were comparable with those in

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“NPF days” in monsoon season. Considering that the purpose is to find the difference between “NPF

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days” and “non-event days”, we did not add the two “undefined days” in the manuscript.



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Figure R6. The Diurnal variations of (a) condensation sink (CS), (b) JO^1D , (c) temperature, (d) RH, and (e) H_2O in NPF-pre days, NPF-monsoon days, non-event days and undefined days in monsoon season (undefined-monsoon days).

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