# 1 Measurement report: Atmospheric new particle formation at a peri-

# 2 urban site in Lille, Northern France

3 Suzanne Crumeyrolle<sup>1</sup>, Jenni SS Kontkanen<sup>2,3</sup>, Clémence Rose<sup>4</sup>, Alejandra Velazquez Garcia<sup>1,5</sup>, Eric

4 Bourrianne<sup>1</sup>, Maxime Catalfamo<sup>1</sup>, Véronique Riffault<sup>5</sup>, Emmanuel Tison<sup>5</sup>, Joel Ferreira de Brito<sup>5</sup>, Nicolas

5 Visez<sup>6</sup>, Nicolas Ferlay<sup>1</sup>, Frédérique Auriol<sup>1</sup>, Isabelle Chiapello<sup>1</sup>

6 <sup>1</sup>Univ. Lille, CNRS, UMR 8518 Laboratoire d'Optique Atmosphérique (LOA), 59000 Lille, France

- 7 <sup>2</sup>CSC IT Center for Science, Espoo, Finland
- 8 <sup>3</sup> Institute for Atmospheric and Earth system Research, University of Helsinki, Helsinki, Finland
- <sup>9</sup> <sup>4</sup>Laboratoire de Météorologie Physique, LaMP-UMR 6016, CNRS, Université Clermont Auvergne, 63178, Aubière, France
- 10 <sup>5</sup> IMT Nord Europe, Institut Mines-Télécom, Univ. Lille, Centre for Energy and Environment, F-59000 Lille, France
- <sup>6</sup> Univ. Lille, CNRS, UMR 8516 LASIRE LAboratoire de Spectroscopie pour les Interactions, la Réactivité et
   l'Environnement, F-59000 Lille, France.
- 13
- 14 Correspondence to: Suzanne Crumeyrolle (suzanne.crumeyrolle@univ-lille.fr)

#### 15 Abstract.

Formation of Ultrafine Particles (UFPs) in the urban atmosphere is expected to be less favored than in the 16 17 rural atmosphere due to the high existing particle surface area acting as a sink for newly-formed particles. Despite large Condensation Sink (CS) values, previous comparative studies between rural and urban sites 18 19 reported higher frequency of New Particle Formation (NPF) events over urban sites in comparison to 20 background sites as well as higher particle formation and growth rates attributed to the higher 21 concentration of condensable species. The present study aims at better understanding the environmental 22 factors favoring, or disfavoring, atmospheric NPF over Lille, a large city in the North of France and to analyze their impact on particle number concentration using a 4-year long-term dataset. 23

- 24 The results highlight a strong seasonal variation of NPF occurrences with a maximum frequency observed
- during spring (27 events) and summer (53 events). It was found that high temperature (T > 295 K), low

26 Relative Humidity (RH< 45 %) and high solar radiation are ideal to observe NPF events over Lille. Relatively high CS values (i.e.,  $\sim 2 \times 10^{-2} \text{ s}^{-1}$ ) are reported during event days suggesting that high CS does 27 not inhibit the occurrence of NPF over the ATmospheric Observations in LiLLE (ATOLL) station. 28 Moreover, the particle growth rate was positively correlated with temperatures most probably due to 29 30 higher emission of precursors. Finally, the nucleation strength factor (NSF) was calculated to highlight 31 the impact of those NPF events on particle number concentrations. NSF reached a maximum of 4 in 32 summer, evidencing huge contribution of NPF events to particle number concentration at this time of the 33 vear.

# 34 1 Introduction

New Particle Formation (NPF) leads to the formation of a large number of particles with diameters below 35 20 nm that will contribute significantly to the high levels of fine particles observed in ambient air. These 36 37 particles can have adverse effect on human health as they can penetrate deeply into the pulmonary system (Clifford et al., 2018; Ohlwein et al., 2019). The freshly-formed particles then grow to larger sizes (Dp > 38 39 100 nm) at which they may act as Cloud Condensation Nuclei (CCN, (Pierce and Adams, 2009; Ren et 40 al., 2021; Rose et al., 2017; Spracklen et al., 2006)). NPF events have been observed around the world 41 (Kerminen et al., 2018; Kontkanen et al., 2017; Kulmala et al., 2004) in various environments from the boundary layer at urban locations (Kanawade et al., 2022; Roig Rodelas et al., 2019; Tuch et al., 2006; 42 43 Wehner and Wiedensohler, 2003) as well as remote polar background areas (Dall'Osto et al., 2018) but also within the free troposphere (Rose et al., 2015b, 2015a). NPF events are typically associated to a 44 45 photochemical origin, thus occurring mostly during daytime (Kulmala et al., 2014), with some scarce 46 events being observed during nighttime (Roig Rodelas et al., 2019; Salimi et al., 2017).

47

NPF occurrence depends on various factors including precursor emission strength, number concentration of pre-existing aerosol population, meteorological parameters (in particular solar radiation, temperature and relative humidity (RH)) and the oxidation capacity of the atmosphere (Kerminen et al., 2018). Differences were found in both the seasonality and intensity of NPF events according to the site type (urban, traffic, regional background, rural, polar and high altitude (Dall'Osto et al., 2018; Sellegri et al.,
2019)). This variability seems to be related to environmental conditions specific to each location, which
makes it hard to draw general conclusions on the conditions that trigger NPF events (Berland et al., 2017;
Bousiotis et al., 2021). However, Nieminen et al. (2018) highlighted a common seasonal occurrence of
NPF during spring and summer using datasets from 36 continental sites worldwide.

57 The formation and growth of initial clusters to detectable sizes ( $D_p > 1-3$  nm) compete with their simultaneous removal from the Ultra-Fine Particle (UFP) mode by coagulation with pre-existing particles 58 59 (Kerminen et al., 2001; Kulmala, 2003). For that reason, the number concentration of particles smaller than 20 nm has been observed to be anti-correlated with the aerosol volume and mass concentration over 60 61 a rural area in Northern Italy (Rodríguez et al., 2005). Indeed, the total aerosol volume is rather small during NPF events (Kerminen et al., 2018; Rodríguez et al., 2008). While the negative effect of increased 62 63 pre-existing particle surface area (often described with the Condensation Sink, CS) on the occurrence of 64 these events is widely accepted (Kalkavouras et al., 2017), yet cases are found when NPF events occur 65 on days with higher CS compared to average conditions (Größ et al., 2018; Kulmala et al., 2017).

66 A recent study by Bousiotis et al. (2021) used large datasets (16 sites) over Europe (6 countries) and highlighted that solar radiation intensity, temperature, and atmospheric pressure had a positive 67 relationship with the occurrence of NPF events at the majority of sites (exceptions were found for the 68 69 southern sites), either promoting particle formation or increasing the Growth Rate (GR). Indeed, solar 70 radiation is considered one of the most important factors in the occurrence of NPF events, as it contributes 71 to the production of NPF precursors (Kontkanen et al., 2016). Higher temperatures are considered 72 favorable for the growth of newly formed particles (Dada et al., 2017) as they can be linked to higher 73 concentrations of organic vapor (Wang et al., 2013) that support particulate growth but also reduce the 74 stability of the initial molecular clusters (Deng et al., 2020; Kurtén et al., 2007).

Wind speed, on the other hand, has shown variable effects on the occurrence of NPF events, appearing to depend on the site location rather than their type (Bousiotis et al., 2021). Additionally, the origin of the incoming air masses plays a very important role, since air masses of different origins have different meteorological, physical and chemical characteristics. Therefore, the probability of NPF events occurrence at a given location and time depends not only on local emissions, but also on long range transport (Sogacheva et al., 2007, 2005; Tunved et al., 2006) and on synoptic meteorological conditions
at the European scale (Berland et al., 2017).

Formation of new particles in the urban atmosphere is expected to be less favored than in the rural 82 atmosphere due to the high existing surface area of particles acting as a sink for freshly-formed particles. 83 84 Despite the large CS values, previous comparative studies between rural and urban sites reported higher 85 frequency of NPF events over urban sites in comparison to background sites (Peng et al., 2017) where higher particle formation and higher GR (Nieminen et al., 2018; Salma et al., 2016; Wang et al., 2017) 86 87 were also observed and attributed to the higher concentration of condensable species. This study presents 88 the first observations of NPF events over Lille, a large city in the north of France. Based on a multi-annual 89 dataset (2017-2020), the frequency and intensity of the NPF events are analyzed aiming at better defining 90 the favorable and unfavorable conditions.

# 91 2 Materials and methods

92 The ATmospheric Observations in LiLLE (ATOLL, Figure 1) station is located in Villeneuve d'Ascq, 93 Northern France (50.6114 N, 3.1406 E, 60 m a.s.l.), only 6 km away from the city center of Lille, which 94 is the core of the metropolis (Métropole Européenne de Lille, with more than 1.1 million inhabitants) to which Villeneuve d'Ascq belongs. Low Single Scattering Albedo (SSA) values (0.75 on average within 95 96 the PM<sub>1</sub> fraction, (Velazquez-Garcia et al., 2022)) and large particle number concentrations (6140 cm<sup>-3</sup>) on average) suggest that aerosol measurements performed at ATOLL are comparable to Global 97 98 Atmospheric Watch (GAW) sites classified as urban (Laj et al., 2020; Rose et al., 2021). ATOLL is also 99 part of the Aerosols, Clouds, and Trace gases Research InfraStructure (ACTRIS, http://www.actris.net), 100 providing high-quality long-term atmospheric data in Northern France. This station is under the influence 101 of many anthropogenic sources, e.g. road traffic, residential sector, agriculture and industries (Chen et al., 102 2022), as well as maritime emissions, and episodically under the influence of natural events such as aged 103 volcanic plumes and Saharan dust (Boichu et al., 2019; Bovchaliuk et al., 2016; Mortier et al., 2013).



Figure 1: ATOLL location in Villeneuve d'Ascq (Northern France) and picture of the station on the rooftop of the University of Lille P5 building (© LOA).

105 A large set of *in-situ* and remote sensing instruments are implemented at ATOLL to characterize physical, 106 chemical, optical and radiative properties of particles and clouds. *In-situ* instruments have independent 107 sampling stainless-steel lines located at least 1 meter above the roof top and equipped either with  $PM_1$ 108 cyclone or  $PM_{10}$  inlet. The measurements used for the present study were performed between 1<sup>st</sup> July 109 2017 and 31<sup>st</sup> December 2020 with instruments that are described below.

110

The Scanning Mobility Particle Sizer (SMPS) measures every 5 minutes the particle number size 111 112 distribution between 15.7-800 nm (divided in more than 100 bins) downstream a Nafion membrane as 113 recommended by ACTRIS standards to keep RH below 40 %. The SMPS system consisted of a 114 condensation particle counter (TSI model 3775), differential mobility analyzer (DMA, TSI 3081A) as 115 described by (Villani et al., 2007) and a Nickel aerosol neutralizer (Ni-63 95MBq). The sheath flow rate 116 was controlled with a critical orifice in a closed loop arrangement (Jokinen and Mäkelä, 1997). The scan 117 time was 300 seconds and the particle concentrations were corrected by taking into account charge effects 118 and diffusion calculated using the manufacturer software and algorithms (AIM 10.2.0.11).

119 Accordingly, aerosol number size distribution data from the SMPS measurements were used to classify 120 individual days as NPF event, undefined or non-event days. The classification procedure, presented in 121 Dal Maso et al. (2005), follows the decision criteria based on the presence of UFP (Dp < 25 nm) and their 122 subsequent growth to Aitken mode (Dp < 80 nm). Briefly, event days are identified when sub-25 nm 123 particle formation and growth are observed. Undefined days correspond to days when sub-25 nm particle 124 formation is observed for more than one hour, but those particles are not growing so their diameter 125 remains below 25 nm. On non-event days, the nucleation mode is absent.

SMPS particle number size distributions were also used for CS (Equation 1) and GR (Equation 3)
calculations. The CS estimates the loss rate of the condensable vapors (Kulmala et al., 2001) which were

assumed to have molecular properties similar to sulfuric acid for CS calculation (Dal Maso et al., 2005):

$$129 \quad CS = 2\pi D \sum_{i} \beta_{Mi} D_{p,i} N_i$$

Equation 1

130 where D is the diffusion coefficient of the condensing vapor,  $D_{p,i}$  and  $N_i$  the particle diameter and number 131 concentration for size bin *i*, respectively, and  $\beta_{Mi}$  the transitional correction factor (Fuks and Sutugin, 132 1970) defined in Equation 2:

133 
$$\beta_{Mi} = \frac{1+K_n}{1+0.337Kn + \frac{4}{3}\alpha^{-1}Kn + \frac{4}{3}\alpha^{-1}Kn^2}$$
 Equation 2

134 with  $K_n$ , the Knudsen number, and  $\alpha$  the accommodation coefficient (here, set to unity).

A high CS indicates the presence of a large particulate surface area onto which NPF precursors cancondensate.

The particle GR from 15.7 to 30 nm (GR<sub>15.7-30nm</sub>) was calculated based on the maximum-concentration method described in (Kulmala et al., 2012). For each event, the NPF starting time (T<sub>1</sub>) was first identified when the newly formed mode was observable in the first bin of the SMPS ( $D_{p,1} = 15.7$  nm). Then, the time (T<sub>2</sub>) when the number concentration of particles with diameter ( $D_{p,2}$ ) of 30 nm (N<sub>30</sub>) peaked was also manually identified. Particle GR<sub>15.7-30nm</sub> was then calculated by linear regression of particle size vs. time span from the NPF start until the time when N<sub>30</sub> reaches a maximum:

143 
$$GR_{15.7-30nm} = (D_{p,2} - D_{p,1}) / (T_2 - T_1)$$
 Equation 3

An Aerosol Chemical Speciation Monitor (ACSM, Aerodyne Research Inc.) equipped with a PM<sub>2.5</sub> cutoff inlet (URG Cyclone 2000-30EH, Chapel Hill, NC, USA) and with a primary flow of 3 L min<sup>-1</sup> was used to monitor the aerosol chemical composition at ATOLL. The chemical characterization of nonrefractory submicron particles (NR-PM<sub>1</sub>), that is to say material vaporizing around 600 °C under closeto-vacuum conditions, was performed online and in real time every 30 minutes. This instrument is based on the same principle as the Aerosol Mass Spectrometers (AMS), without providing aerosol size distribution information. A full description of the instrument is available in Rivellini et al. (2017). Under ambient conditions, mass concentrations of particulate organics, sulfate, nitrate, ammonium, and chloride are obtained with a detection limit < 0.2  $\mu$ g m<sup>-3</sup> for 30 min of signal averaging. An algorithm (Middlebrook et al., 2012) was applied to ACSM mass concentrations to obtain a time-dependent correction of the Collection Efficiency ranging from 0.45 to 0.83.

- 155 Absorption coefficients ( $\sigma_{abs}$ ) were continuously measured with a seven-wavelength aethalometer (AE33, 156 Magee Scientific Inc., Cuesta-Mosquera et al. (2020)). According to ACTRIS current guidelines 157 (https://actris-ecac.eu/particle-light-absorption.html),  $\sigma_{abs}$  coefficients at each wavelength have been 158 recalculated by 1) multiplying equivalent Black Carbon (eBC) by the mass-specific absorption coefficient 159 (MAC) and then 2) dividing by the suitable harmonization factor to account for the filter multiple 160 scattering effect, i.e., 2.21 (M8020 filter tape) in 2017 and 1.76 (M8060 filter tape) afterwards. The 161 aethalometer samples at 5 L min<sup>-1</sup> downstream a PM<sub>1</sub> cyclone (BGI SCC1.197, Mesa Labs). The spectral 162 dependency of  $\sigma_{abs}$  was used to determine the contributions of traffic (fossil fuel - BC<sub>ff</sub>) and wood burning 163 (BC<sub>wb</sub>) to eBC via a source apportionment model (Sandradewi et al., 2008).
- 164 Meteorological data including temperature and water vapour mixing ratio were also measured every 165 minute at the sampling site using a weather station (DAVIS Inc weather station, Vantage Pro 2). Solar 166 radiation at the surface was measured every minute at the sampling site using a set of Kipp & Zonen pyranometers (CM22, for diffuse fluxes using a sphere shadower) and Normal Incidence pyrheliometer 167 168 (CH1 for direct fluxes), with the solar radiation being then calculated as the sum of the diffuse and direct 169 fluxes. The cloud cover was estimated from the Findclouds algorithm, provided by the manufacturer, and 170 applied on a sky imager (Cloudcam, CMS) pictures by comparing the different values of the red, green 171 and blue components of each pixel of the image taken (Shukla et al., 2016).
- Three-day air mass back-trajectories of the air masses arriving at ATOLL at half the boundary layer height
  between July 1, 2017 and December 31, 2020 were computed every hour using the Hybrid Single-Particle
  Lagrangian Integrated Trajectory (HYSPLIT version 5.1.0) transport and dispersion model from the

175 NOAA Air Resources Laboratory (Rolph et al., 2017; Stein et al., 2015) and meteorological input from

176 the Global Data Assimilation System (GDAS) at 1×1° resolution, resulting in 30719 back-trajectories.

#### 177 **3** Results

## 178 **3.1** NPF event frequency

179 The seasonal distribution of NPF events at ATOLL is displayed in Figure 2. SMPS missing data (in Figure 180 2) reach up to about 40 % from January to April due to the yearly calibrations at the manufacturer premises 181 and a few laboratory campaigns (Oct 2018 – Jun 2019). Over the 4 years of measurements (2017-2020), 182 96 days (11%) were classified as NPF event days (Ev), 355 (40 %) as undefined days (Un) and 432 (49 183 %) as non-event days (No). One can also note that most of the NPF events identified were observed during 184 spring (March-April-May, with 27 events corresponding to 15 % of days when observations were 185 available during this season over the 4-year period) and summer (June-July-August, with 53 events 186 corresponding to 19%) with a maximum observed in June consistent with a previous study over central 187 Europe (Dall'Osto et al., 2018). During winter, the number of events is extremely limited (only one event 188 observed in February). In the following sections, only observations from spring and summer seasons will 189 be discussed due to the low representativeness of NPF events in fall (n=15) and winter (n=1). Moreover, 190 the undefined event days are seen all year round (frequency around or larger than 20%) with a clear peak 191 in August (frequency at 62 %) consistent with observations over the boreal forest where undefined days 192 were also observed to be most frequent in early fall (Mazon et al., 2009).

Using long-term measurements from 36 sites (polar, rural, high altitude, remote and urban), Nieminen et al. (2018) reported an annual NPF frequency below 15 % for half of the sites (18 sites from all types) and occasionally over 30 % for 10 sites. Moreover, they highlighted a seasonal variation of NPF occurrence with larger (lower) frequency, about 30 % (10 %), during spring (winter). A frequency analysis of NPF occurring only over urban or anthropogenically influenced sites show large site-to-site differences for all seasons. Indeed, Nieminen et al. (2018) reported NPF occurrence frequencies varying from 20 % (Helsinki, Finland; Sao Paulo, Brazil) to 80 % (Beijing, China; Marikana, South Africa) during spring

and from 7 % (Helsinki) to 78 % (Marikana) during winter. Yearly averages of NPF occurrence
frequencies are between 11 % (Helsinki) and more than 60 % (Beijing and Marikana).

202 The ATOLL event frequency (seasonal variation and values) is similar to observations performed in Paris 203 while the frequency of undefined and non-event days is quite different (Dos Santos et al., 2015). Indeed, 204 in Paris the non-event frequency is larger than 60 % except in July and August whereas over ATOLL the 205 non-event frequency shows a clear seasonal pattern with lower frequency (< 40 %) from April to August. Moreover, undefined event frequency in Paris shows a minimum (< 5 %) in May and June and remains 206 207 quite steady during the rest of the year (around 20%). One can note that the frequency of undefined events 208 is much higher over ATOLL all year long with an average of 40 %. The frequency of undefined events 209 observed at ATOLL is clearly larger than the frequencies observed over a more polluted site (Paris) and 210 similar to those observed over pristine sites in Siberian and Finnish Boreal forests (Uusitalo et al., 2021). 211 This could mean that ATOLL is under the influence of air masses or particle and precursor sinks that 212 favor the burst of UFP.



Figure 2: Seasonal distribution of event days (Ev, blue), undefined days (Un, green), and non-event (No, red) days at the ATOLL station, Lille, France, during 2017–2020. Days with missing data are excluded from the total number of days per month and the frequency of missing data are indicated with the black circles.

214

### 215 3.2 Aerosol number size distribution

Hourly-averaged median particle number size distributions (PNSD) obtained from the SMPS are shown in Figure 3 separately for NPF event (around 800 PNSD), undefined (around 2300 PNSD) and non-event (around 1700 PNSD) days observed during the warm period (only spring and summer). For all event days, the PNSD were first sorted for each hour of the day. Then, the median PNSD was calculated for each hour of the day. PNSD shown in Figure 3a is then representative of a "typical" NPF event day (Kulmala et al., 2022). The same data filtering was done for PNSD observed during undefined (Figure 3b) and non-

222 event days (Figure 3c). Atmospheric NPF and subsequent particle growth are seen in Figure 3a as an 223 appearance of new aerosol particles with small diameters followed by the growth of these particles toward 224 larger sizes. If this phenomenon is taking place regionally (few tens of km in radius), a so called 'banana 225 plot' is observed in PNSD as a function of time at a fixed location. The time evolution of the PNSD for 226 "typical" NPF event day (Figure 3a) displays a similar growth pattern for newly formed particles to the 227 one observed for individual NPF event days (see supplementary materials Figures S1 and S2 for examples). Indeed, one can clearly see a UFP mode appearing from 10:00 to 15:00 (UTC) and growing 228 229 during the rest of the day. The NPF starting time and the growth rate will be discussed in the following 230 section. By 23:00 UTC, the newly formed particles reach an average diameter of 50 nm, similar to the 231 median modal diameter of the pre-existing particles observed during the morning (00:00 - 08:00 UTC). The PNSD observed during "typical" undefined days (Figure 3b) highlights a burst of UFP again from 232 233 10:00 to 15:00 UTC that neither grow nor persist over the whole afternoon. The behavior of the median 234 PNSD is again similar to the individual undefined events observed during this period (not shown here). 235 The PNSD observed during "typical" non-event days (Figure 3c) shows no sign of particle growth, as 236 expected.

237







Figure 3: Hourly median particle number size distribution (15.7 nm  $< D_p < 800$  nm) observed during NPF event (a), undefined (b) and non-event (c) days in spring and summer from 2017-2020.

239

#### 240 **3.3** NPF starting time and growth rate

241 Figure 4 shows the monthly variation of the starting time and particle GR<sub>15,7-30nm</sub> of each event observed 242 at ATOLL. Most NPF events start between 09:00 and 14:00 UTC (74 %), with fewer events starting in 243 the early morning (07:30 - 09:00 UTC, 6 %) and late afternoon (15:00 and 19:30 UTC, 20 %). NPF 244 starting time as well as GR<sub>15,7-30nm</sub> strongly depend on the month during which the event is observed. 245 Indeed, the NPF starting time occurs later during spring days on average (also true for fall and winter) 246 while the earliest time was reached in May and June (around 08:20 UTC). Nocturnal events are rarely 247 observed, with only one occurrence in August 2018. No diurnal NPF event was observed after 16:00 UTC 248 in summer. During spring and fall, the average NPF starting time varies between 10:00 and 19:00 UTC. 249 The start time monthly variability seems linked to sunrise and sunset times. In the following section, the 250 relationship between the total solar irradiation and NPF occurrence will be examined.

The event ending time was determined as the time when the growth of the freshly formed particles was over, i.e., when the diameter reached the diameter of the pre-existing particles. The duration of the nucleation events at ATOLL was then estimated and varies from an hour up to 28 hours. On average, NPF duration is shorter from May to August (around 8 hours) and increases up to around 13 hours on average in March. This seasonal behavior could be due to the presence or availability of condensable vapors, air mass origins, and environmental conditions favorable to NPF events (see section 3.2).

- 257
- 258



Figure 4: Monthly variation of NPF events starting time (a) and their Growth Rate (GR<sub>15.7-30nm</sub>) at the ATOLL station during 2017–2020. The boxes and whiskers (bottom to top) represent the 10<sup>th</sup> and 25<sup>th</sup> percentiles, median (red line), and the 75<sup>th</sup> and 90<sup>th</sup> percentiles. Crosses represent the outliers. The grey area represents the period, from 09:00-14:00 UTC, when most of the NPF events occur. The blue area corresponds to the period before the NPF onset (07:00- 09:00 UTC). N represents the number of events observed per month.

259

The GR<sub>15.7-30nm</sub> values observed at ATOLL lie within 0.8 to 15.7 nm h<sup>-1</sup> and show a strong monthly variability with the lowest values observed in spring (and fall, not shown here). The largest median values are observed in May and August while the 75<sup>th</sup> percentile highlights larger values of GR<sub>15.7-30nm</sub> during summer (Figure 4b). GR<sub>15.7-30nm</sub> values were in addition plotted as a function of temperature for all years and seasons in Figure 5, which highlights that below 20 °C, GR<sub>15.7-30nm</sub> values are lower than 6 nm h<sup>-1</sup>, while, under warmer conditions (T > 20 °C), GR<sub>15.7-30nm</sub> reach values up to 16 nm h<sup>-1</sup>. These results show

266 a temperature dependence ( $R \approx 0.4$ ) of the particle growth consistent with previous observations over the 267 boreal forest (Liao et al., 2014). Higher temperatures have been shown to favor the emission of biogenic 268 precursors, including monoterpenes known to favor the occurrence of NPF events (Kulmala et al., 2004). 269 Previous studies (Paasonen et al., 2018; Yli-Juuti et al., 2011) have shown that GR usually exhibit larger 270 values during warm periods especially during summer, and that there is a link between GR seasonal 271 pattern and the high abundance of biogenic Volatile Organic Compounds (VOC) during warmer periods (spring and summer) over the boreal forest. Therefore, the observed seasonal variation of GR<sub>15.7-30nm</sub> may 272 273 be related to temperature-dependent emissions of organic compounds in the vicinity of ATOLL (Figure 274 5). This hypothesis is supported by the larger contribution of organics during NPF event days observed 275 in Figure 9. However, over urban areas such as Beijing or Shanghai, GR<sub>15-25nm</sub> showed no clear seasonal 276 variation (Yao et al., 2018). As previously observed in Figure 3a, the median diameter reached at the end 277 of all NPF events is around 50 nm. Moreover, the seasonal variation of the NPF event durations could be 278 related to the GR<sub>15,7-30nm</sub> seasonal variation. The lower GR<sub>15,7-30nm</sub> values are associated with the longer 279 NPF duration. The seasonal variation of NPF duration highlighted earlier could then only be a 280 consequence of the GR<sub>15.7-30nm</sub> seasonal variation.



Figure 5: Growth Rate (GR15.7-30nm) values as a function of ambient temperature for different seasons.

#### 282 3.4 Environmental conditions

283 The cloud fraction was calculated from the sky imager dataset following the method by Shukla et al. 284 (2016) and sorted as a function of event, undefined and non-event days. The effect of cloudiness on NPF 285 event occurrence is shown in Figure 6a, with a specific focus on measurements collected between 09:00 286 and 14:00 UTC, i.e., the period of time where most NPF events tended to start. There is a clear inverse 287 correlation between cloud fraction and NPF occurrences. The average cloud fraction is around 0.47 during event days, 0.68 during undefined days and 0.74 during non-event days. Moreover, the 25th percentiles of 288 289 the cloud fractions for NPF event (0.06), undefined (0.47) and non-event days (0.63) clearly show that 290 the absence of clouds (lower cloud fraction) is mostly associated with NPF event days. This result is 291 consistent with previous studies performed over the boreal forest (Dada et al., 2017) and is linked to the 292 fact that radiation seems essential for NPF during the warmer period (spring and summer), as the events 293 occur almost solely during daylight hours (Kulmala et al., 2004). Figure 6b shows the average diel total

solar radiation observed during NPF events, non-event and undefined days for spring and summer. As
expected, the total solar radiation is on average always larger during event days in comparison to nonevent days, with a more pronounced difference observed during spring.

297



Figure 6: (a) Cloud fraction observed from 09:00 to 14:00 UTC during event (Ev), undefined (Un) and non-event (No) days. The red line represents the median and the circles the average, while the lower and upper edges of the box represent the 25<sup>th</sup> and 75<sup>th</sup> percentiles, respectively. The lower and upper edges of the whisker correspond to the 10<sup>th</sup> and 90<sup>th</sup> percentiles, respectively. (b) Diel variations (UTC) of the mean total solar radiation observed during the event days (Ev; blue squares), undefined days (Un; red dots) and non-event days (No; black triangles) during spring (MAM, top) and summer (JJA, bottom) seasons (b). The error bars correspond to one standard deviation.

298 Other environmental parameters known to influence the occurrence of NPF events, such as temperature 299 and RH were also investigated to highlight diel and seasonal variations (Figure 7). Our results confirm 300 that NPF is favored by low values of ambient RH (Figure 7a), especially during spring, consistently with 301 previous studies (Duplissy et al., 2016; Hamed et al., 2011; Merikanto et al., 2016). A few reasons can explain this tendency: (1) high RH values (RH > 90 %) observed at the surface are usually associated to 302 303 the presence of low altitude clouds reducing the incoming total radiation and then preventing NPF occurrence, (2) at moderately high RH (RH > 40 %), hydrophilic aerosols could grow through 304 305 condensation which will enlarge the sink for possible precursors and (3) high RH values may limit the 306 formation of some semi-volatile VOC through ozonolysis reactions, inhibiting the formation of 307 condensable vapors necessary for nucleation (Fick et al., 2003; Tillmann et al., 2010).

308 Figure 7b shows the diel median temperature conditions (T) during NPF event, non-event and undefined 309 days. NPF events occurred for temperatures ranging between 3 and 33.5 °C. In both seasons, averaged 310 temperatures during event days are always larger than during non-event days, with again larger 311 differences during spring. One should note that days with high temperatures in spring and summer are 312 usually also days with high solar radiation, consistently with conclusions from Figure 6. The temperature 313 difference between undefined days and event days is clearly marked during spring and fades away during 314 summer. As previously discussed, higher temperatures favor the emission of biogenic precursors, 315 including monoterpenes known to favor the occurrence of NPF events (Kulmala et al., 2004). Isoprene 316 emission is also larger at higher temperatures but according to Heinritzi et al. (2020) it is one example 317 where a biogenic compound inhibits NPF events. Moreover, high temperature can also lead to the 318 evaporation of molecular clusters which may inhibit NPF events (Dada et al., 2017; Deng et al., 2020).



Figure 7: Diel variations (UTC) of the mean Relative Humidity (RH, a) and mean temperature (b) observed during the NPF event days (Ev; blue squares), undefined days (Un; red dots) and non-event days (No; black triangles) during spring (MAM) and summer (JJA) seasons. The error bars correspond to one standard deviation.

#### 319 **3.5** Condensation sink

- 320 The CS characterizes the loss rate of atmospheric vapors to aerosol particles. The diel variations of CS
- 321 calculated for spring and summer for NPF event, undefined and non-event days are shown in Figure 8a.
- Hourly averaged CS values are high (larger than  $2 \times 10^{-2}$  s<sup>-1</sup>) during NPF event days occurring during

spring and summer (Figure 8a). CS values ranging from 0.6 up to  $10.7 \times 10^{-2}$  s<sup>-1</sup> were reported during 323 324 NPF event days and over different urban sites (Beijing, Nanjing or Hong Kong) (Xiao et al., 2015). Over pristine sites, such as Hyytiälä (Finland), the CS values are between  $0.05 - 0.35 \times 10^{-2}$  s<sup>-1</sup>. Low values of 325 326 CS, often considered as the major limiting factor in the NPF occurrence, do not inhibit the occurrence of NPF events at ATOLL, consistently with previous observations in similar environments, such as the 327 328 Melpitz (Germany) observatory (Größ et al., 2018) or over Chinese megacities (Xiao et al., 2015). One 329 can assume that the presence of large concentrations of precursors could explain the formation of particles 330 over polluted sites such as ATOLL. Unfortunately, precursors were not measured over the 4-year period 331 therefore this assumption would require further investigation beyond the scope of this study. 332 Nevertheless, recent studies, performed in the CLOUD chamber, demonstrate that the presence of nitric 333 acid (HNO<sub>3</sub>) and ammonia (NH<sub>3</sub>), typical within urban environment and particularly in the north of France 334 (Roig Rodelas et al., 2019), contribute to fresh particle survival by increasing dramatically their growth 335 rate (Marten et al., 2022; Wang et al., 2020).

In the afternoon of NPF event days, the CS increases due to the growth of freshly emitted particles, especially during summer. The contribution of newly formed particles ( $D_p < 50$  nm) to the CS is about 36 % and 27 %, during summer and spring, respectively, while the contribution of pre-existing particles ( $D_p > 150$  nm) to the CS is below 20 % for both seasons. Moreover, during non-event days, the sizeresolved median CS is shifted toward larger particle diameters with a maximum observed around 100 nm for all seasons.

To evaluate the impact of the background CS on NPF occurrences, all CS values observed from 07:00-09:00 UTC ( $CS_{07:00-09:00}$ ), the period before the NPF starting time (green area on Figure 4a), were averaged during NPF event, non-event and undefined days. It was found that the total  $CS_{07:00-09:00}$  was larger (around 16 %) during non-event days in comparison to undefined and event days. Moreover, this difference is mostly due to particles larger than 70 nm according to the size-resolved  $CS_{07:00-09:00}$  (Figure 8b). The difference between non-event and event days is lower than what is usually observed over pristine sites (Lyubovtseva et al., 2005) but significant enough to trigger the NPF event occurrence.



Figure 8: (a) Diel variations of the Condensation Sink (CS) during spring (MAM) and summer (JJA) seasons and (b) Median size-resolved background CS (from 07:00 to 09:00 UTC) for MAM and JJA during event days (Ev, blue squares), undefined (Un; red dots) and non-event days (No, black triangles).

Du et al (2022) studied the chemical composition of particles contributing to the CS when NPF occurrence 350 is at its highest (10:00 - 15:00 local time) over Beijing. They observed a large increase of nitrate and a 351 352 decrease of organics in PM2.5 with increasing CS values for NPF event and non-event days. As the CS 353 observed over ATOLL is largely influenced by the freshly formed particles, the chemical composition of 354 particles as a function of CS can be presented during two specific periods: before (07:00 - 09:00 UTC)and during (09:00 – 14:00 UTC) the period when the NPF occurrence is at its highest for NPF event and 355 356 non-event days (Figure 9). During non-event days, both periods (Figure 9b and Figure 9d) exhibit similar 357 mass fraction for all compounds with on average 41 % of organics, 16 % of nitrate, 21 % of sulfate, 11 358 % of ammonium, less than 1 % of chloride and around 10 % of Black Carbon (BC). As the aerosol sources 359 during non-event days are supposed to be the same throughout the day, this result was actually expected. Larger contribution of organics to the CS is observed for NPF event days during the NPF period (CS<sub>09:00-</sub> 360 361  $_{14:00} \sim 54$  % in average, Figure 9c) in comparison with the period right before the start of the NPF events  $(CS_{07:00-09:00} \sim 46 \%$  in average, Figure 9a). Indeed, for large values of  $CS_{09:00-14:00}$  (> 0.045 s<sup>-1</sup>) the 362 363 contribution of organics is larger than 50 %, reaching a maximum of 69 % for  $CS_{09:00-14:00}$  of 0.085 s<sup>-1</sup>. This result suggests that organic vapors may be involved in the particle growth. 364



Figure 9: Mass fractions of the major compound measured before (07:00 – 09:00 UTC, a and b) and during (09:00 – 14:00 UTC, c and d) NPF period at the ATOLL station during event (top) and non-event (bottom) days. The black dashed line corresponds to 50% mass fraction.

366

Additionally, the correlation coefficients (R) between meteorological parameters and gaseous or particulate pollutants are reported in Table 1 for the entire period of measurements (all seasons). Hourly averages over a time window between 09:00 - 14:00 UTC (NPF event starting time period) of a few variables (total CS, T, RH and BC from wood burning (BC<sub>wb</sub>)) were used to calculate those correlation coefficients (corresponding to 7025 and 35433 data points for NPF event and non-event days, respectively).

The correlation of BC<sub>wb</sub> with the CS during non-event days is high (R = 0.67) and is clearly absent during NPF event days (R = 0.19). One can also note that NO<sub>x</sub> concentrations have a positive correlation (0.30) with CS during NPF non-event days while the same correlation is negative (-0.17) during NPF event days. NO<sub>x</sub> sources over urban areas are mostly anthropogenic ones (residential heating, traffic and industries) which is consistent with its relatively high correlation coefficients with BC<sub>wb</sub> (0.47 and 0.65). As highlighted in Barreira et al. (2020), BC<sub>wb</sub> and NO<sub>x</sub> are evolving through the year showing a minimum in summer and a maximum in winter when sources are stronger due to colder temperatures and residential

380 heating emissions. As non-event days are mostly (62 %) observed during cold months (fall and winter, 381 not shown here) and NPF events are largely (82 %) observed during warmer months (spring and summer), 382 the correlation between  $BC_{wb}$ ,  $NO_x$  and CS during non-event days is expected. However, during spring, 383 air masses observed during NPF events are clearly "cleaner" (in terms of NOx and BC<sub>wb</sub>) than non-event 384 cases. Indeed, NO<sub>x</sub> and BC<sub>wb</sub> concentrations are lowered by 18 % and 36 % respectively during NPF 385 event days occurring in spring in comparison to non-event days. During summer,  $NO_x$  and  $BC_{wb}$ 386 concentrations reach an annual minimum and therefore both pollutant concentrations are similarly 387 correlated between NPF event and non-event days (lowered by -0.04 % and 0.01 % during NPF event 388 days).

389

Table 1: Correlation coefficients between different meteorological parameters (T, RH), nitrogen oxides (NO<sub>x</sub>), black carbon concentrations from wood burning (BC<sub>wb</sub>) and the total condensation sink (CS) during NPF event and non-event days for the 4-year

392 period (2017- 2020) and over the 09:00 – 14:00 UTC time window. High positive or negative correlations are marked in bold.

		CS	Т	RH	NO <sub>x</sub>	$BC_{wb}$
Event days	CS	1				
	Т	0.55	1			
	RH	-0.39	-0.40	1		
	NO <sub>x</sub>	-0.17	-0.24	0.48	1	
	$BC_{wb}$	0.19	-0.04	0.11	0.47	1
Non- event days	CS	1				
	Т	0.06	1			
	RH	-0.03	-0.50	1		
	NOx	0.30	-0.44	0.44	1	
	$BC_{wb}$	0.67	-0.37	0.28	0.65	1

393

Moreover, during NPF event days the temperature is positively correlated (0.55) with the CS, while during non-event days, this correlation is clearly not observed (0.06). As discussed previously, this coupling was expected as it could be related to larger VOCs emissions leading to enhanced particle growth and to higher concentrations of larger particles (Section 3.4).

#### 399 3.6 Air mass trajectories

Environmental parameters such as CS, T and RH (Figures 6, 7 and 8) observed during undefined events exhibit values mostly between those observed during event and non-event days. A deeper analysis of undefined days is therefore needed to evaluate if the UFP observed during those days are coming from failed events or from pollution-related peaks (Buenrostro Mazon et al., 2016). A first analysis on undefined days reveals that on these days, particle growth stopped suddenly due to (i) a decrease of the total irradiance due to a cloud passing over the site (20 % of cases), (ii) a shift of the wind direction (17% of cases), (iii) or both parameters changing simultaneously (35 % of cases).

407 The shift of the wind orientation leading to a stop of the particle growth indicates that NPF events are 408 associated with certain wind directions or air mass origins. To investigate this, HYSPLIT back-409 trajectories were first sorted as a function of NPF event, non-event and undefined days. Only back-410 trajectories arriving between 09:00-14:00 UTC (period of NPF high occurrences) were selected for further 411 analysis. During the NPF events, the predominant air masses were tracked back along the Eastern North 412 Sea region (Figure 10a). Comparing these results to back-trajectories for non-event days highlight a more 413 continental influence. Indeed, most of the back-trajectories during non-event days pass over large cities 414 (Dunkirk, Paris, London, Rotterdam) before reaching Lille metropolis (Figure 10b). Those air masses 415 might then have been slightly enriched in primary precursor vapors. This result is consistent with previous results showing that "cleaner" air masses are associated with higher probability of NPF events (Bousiotis 416 417 et al., 2019).



Figure 10: 3-day hourly back-trajectories arriving at ATOLL between 09:00-14:00 UTC during (a) NPF event and (b) non-event days. Back-trajectories were calculated for each hour at ATOLL arriving at half the boundary layer height using GDAS  $1^{\circ} \times 1^{\circ}$  meteorological data. The color contour represents the back-trajectory crossing counts in each grid cell (resolution  $0.2^{\circ} \times 0.2^{\circ}$ ).

419

#### 420 **3.7** Nucleation strength factor

421 The Nucleation Strength Factor (NSF<sub>15.7-100</sub>) is calculated as the ratio of fine ( $15.7 < D_p < 100 \text{ nm}$ ) to 422 accumulation ( $100 < D_p < 800 \text{ nm}$ ) particle concentrations observed during NPF event days over the same

423 ratio observed during non-event days (Salma et al., 2017). Fine and accumulation mode particle number

424 concentrations ( $N_{15.7-100}$  and  $N_{100-800}$ ) were retrieved from the SMPS data (Figure 11a). The limited

425 atmospheric residence time of fine particles (typically lower than 10 hours, Seinfeld and Pandis, (2016)) 426 means that a large portion of the N<sub>157-100</sub> concentrations can be related to local emissions and/or formation 427 processes, including NPF events. On the contrary, due to a longer residence time within the atmosphere 428 (up to 10 days, Seinfeld and Pandis, (2016)),  $N_{100-800}$  is more related to large spatial and temporal scales. 429 Therefore, the numerator represents the increase of  $N_{15,7-100}$  relative to  $N_{100-800}$  caused by all sources while 430 the denominator represents the same property due to all sources except NPF. The NSF method is based 431 on the hypothesis that aerosol sources are similar from day to day and from season to season, excepting 432 the sporadic occurrence of NPF. Considering the large number of NPF event (96) and non-event (432) 433 days used to calculate NSF<sub>15,7-100</sub>, one can assume that the sporadic/occasional (i.e., not observed on daily 434 basis) sources of UFP other than NPF events (e.g., volcanic plumes) have little impact on the NSF<sub>15,7-100</sub> in comparison to the sources always active (such as traffic, industries, etc). 435

436 NSF is generally used to better assess the contribution of NPF to fine particle number concentrations 437 relative to the regional background particle number concentrations. If the NSF  $\approx$  1, then the relative 438 contribution of NPF to particle number concentration with respect to other sources is negligible, as 439 observed at the Granada (Spain) urban site (Casquero-Vera et al., 2021). Moreover, Salma et al. (2017) 440 also defined two thresholds for  $NSF_{6-100}$  to describe NPF contribution as a single source: a considerable contribution ( $1 < NSF_{6-100} < 2$ ) or larger than all other sources together ( $NSF_{6-100} > 2$ ). One should keep 441 442 in mind that these thresholds were defined accordingly to the lower cut-off diameter originally set at 6 443 nm. As the lower cut-off diameter used in this study is a bit larger (15.7 instead of 6 nm) than the one 444 used by Salma et al. (2017), the calculated NSF<sub>15.7-100</sub> would necessarily be underestimated in comparison to NSF<sub>6-100</sub> from Salma et al. (2017). The hourly median value of fine-to-accumulation particle 445 concentration ratio was computed for NPF event and non-event days. Figure 10 shows the NSF<sub>15.7-100</sub> diel 446 447 variation observed at ATOLL over 4 years of measurements.

During spring, the NSF<sub>15.7-100</sub> factor remains quite constant (about 1.5) during night and morning and peaks at 16:00 UTC to reach a maximum of 2.5 (Figure 11b). This indicates that NPF has a significant effect on particle number concentration only a few (2-3) hours after the averaged NPF starting time. During summer, the tendency of the NSF<sub>15.7-100</sub> is quite similar with a unique peak at 13:00 UTC (again 2-3 hours after the averaged NPF starting time). At that time the median NSF<sub>15.7-100</sub> values reach 4, while

- 453 from 21:00 to 06:00 UTC the NSF $_{15.7-100}$  remains low (averaged at 1.08). Therefore, during summer, the 454 NPF contribution to particle number concentration is extremely high from 10:00 to 18:00 UTC and then 455 negligible for the rest of the day in comparison to other sources.
- Such NSF<sub>10-100</sub> diel variations were observed in other European cities (Budapest, Vienna and Prague) with 456 457 maximum values reaching 2.7, 2.3 and 3.4, respectively, for a lower cut-off diameter set at 10 nm (Németh 458 et al., 2018). Moreover, Salma et al. (2017) reported NSF<sub>6-100</sub> peaks at midday varying from 2.2 and 2.7 for Budapest city center and from 2 to 7.2 for near city background for each season with NSF<sub>6-100</sub> 459 460 maximum reached during winter. The nucleation frequency during winter in Budapest is low (< 10 %), 461 similarly to our observations, however, the impact of these limited number of NPF events on particle 462 number concentrations is high. It may be mentioned that the NSF<sub>15.7-100</sub> factor over ATOLL peaked at 3.5 and 2.3 during winter and fall, respectively. 463
- 464 As previously shown by Sebastian et al. (2021), NPF events can also play a major role on the Earth's 465 radiative budget when the newly formed particles grow to climate-relevant sizes (50-100 nm). In order to 466 understand the NPF influence on these particles the NSF<sub>50-100</sub> was also calculated (see supplementary 467 Figure S3). The results show a large increase up to 1.6 of the NSF<sub>50-100</sub> in the early afternoon for both 468 seasons. This suggests a potential impact on CCN concentrations that needs to be further studied with 469 proper instrumentation.
- 470
- 471





Figure 11: (a) Diel variations of fine (15.7  $< D_p < 100$  nm) and accumulation (100  $< D_p < 1000$  nm) mode particle number concentrations (N<sub>15.7-100</sub> in black and N<sub>100-1000</sub> in red) during MAM and JJA at the ATOLL site during the 2017-2020 period. The dots correspond to NPF event days while the line correspond to non-event days. (b) Diel variations of the Nucleation Strength Factor (NSF<sub>15.7-100</sub>) for each season calculated from N<sub>15.7-100</sub> and N<sub>100-800</sub> observed during the 2017-2020 period.

#### 472 4 Conclusions

473 This study was based on a 4-year (2017-2020) dataset collected at the ATOLL site, in the close vicinity 474 of the city of Lille, Northern France, to study NPF occurrence over a peri-urban site. The results highlight 475 a strong seasonal variation of the NPF event frequency, with a maximum occurrence observed during 476 spring (15 %) and summer (19 %). The undefined cases, which correspond to bursts of UFP that do not 477 grow, are much more frequent (40 % on average) than NPF events all year long. The highest frequency 478 (68 %) is observed in August and the lowest one (17 %) in February. The interruption of particle growth 479 during undefined events can be mostly attributed to changes in environmental conditions (irradiance and 480 wind direction).

481 The seasonal variation of NPF parameters was also clearly observed and associated with environmental 482 parameters. High temperature (T > 295 K), low RH (RH< 45 %) and high solar radiation favor the 483 occurrence of NPF events at ATOLL. The presence of clouds, linked to a decrease of solar radiation, also 484 limits the NPF event occurrences. Moreover, NPF events start earlier in the morning from May to September most probably related to variations in sunrise time. The GR calculated between 15.7 and 30 485 486 nm (GR<sub>15.7-30nm</sub>) ranges from 1.8 nm h<sup>-1</sup> in March up to 10.9 nm h<sup>-1</sup> in July. The GR<sub>15.7-30nm</sub> was also found 487 to be positively correlated with temperature. This correlation might be related to larger emissions of 488 biogenic precursors at higher temperatures, including monoterpenes known to favor the occurrence of 489 NPF events (Kulmala et al., 2004).

490 Relatively high values of CS (>  $2 \times 10^{-2}$  s<sup>-1</sup>) are reported during NPF events as well as during non-event 491 days. These results suggest that high CS values are not limiting the NPF event occurrence, consistent with 492 recent studies focusing on NPF events over urban sites (Deng et al., 2020; Hussein et al., 2020; 493 Pushpawela et al., 2018). Looking more closely before the NPF onset (from 07:00 – 09:00 UTC), CS<sub>07:00</sub>-494 <sub>09:00</sub> values are larger by 16 % during non-event days. Interestingly, CS tends to increase during NPF 495 event days (especially in summer) and size-resolved CS clearly shows a peak shift from 150 nm during
496 non-event days to 50 nm during NPF event days, thus highlighting the strong contribution of newly
497 formed particles on CS.

Air mass back-trajectories (HYSPLIT) arriving over ATOLL during NPF event days revealed a specific path along the Eastern North Sea region with only a small fraction passing over any continental area and therefore not crossing many anthropogenic sources, while most of the back-trajectories during non-event days pass over large cities (Dunkirk, Paris, London, Rotterdam) before reaching Lille. The precursor vapor concentrations and probably their nature might differ from both "clean" and "polluted" air masses and therefore promote or inhibit NPF event occurrences, a point which requires further investigation.

504

505 The impact of NPF events on particle number concentrations has been estimated through the nucleation 506 strength factor (NSF; Salma et al., 2017). The NSF<sub>15.7-100nm</sub> diel variation was calculated for spring and 507 summer occurring 2 to 3 hours after the average NPF starting time and reaching 1.5 and 4 during spring 508 and summer respectively. The extremely large NSF<sub>15,7-100nm</sub> values observed during summer evidence the very high NPF contribution to the fine particle ( $D_p < 100$  nm) number concentrations in comparison to 509 510 other regional sources. Recently, Ren et al. (2021) highlighted the strong impact of newly formed particles 511 from NPF on CCN especially at sites close to anthropogenic sources, such as ATOLL. In future studies, 512 the impact of local vertical dynamics such as the effect of the boundary layer dynamics as in Lampilahti 513 et al. (2021, 2020) as well as the CCN enhancement factor will be analyzed.

514

## 515 Acknowledgements

516 This research was supported by the French national research agency (ANR) under the MABCaM (ANR-16-CE04-0009) 517 contract. Part of the instrumental system has been financially supported by the CaPPA project (Chemical and Physical 518 Properties of the Atmosphere), which is funded by the French National Research Agency (ANR) through the PIA (Programme 519 d'Investissement d'Avenir) under contract "ANR-11-LABX-0005-01", and by the Regional Council "Hauts-de-France". 520 ATOLL is a French component of the Aerosol, Clouds and Trace Gases Research Infrastructure (ACTRIS, 521 https://www.actris.eu/) and the particle chemical composition measurements are also supported by the CARA program of the 522 LCSQA funded by the French Ministry of Environment. The authors also thank the Région Hauts-de-France, the Ministère de 523 l'Enseignement Supérieur et de la Recherche (CPER Climibio), and the European Fund for Regional Economic Development for their financial support. The authors gratefully acknowledge the NOAA Air Resources Laboratory (ARL) for the provision of the HYSPLIT transport and dispersion model and/or READY website (https://www.ready.noaa.gov) used in this publication. We thank François Thieuleux for ECMWF data sharing during this work.

527

# 528 **Data availability**

ATOLL measurements are available through the EBAS database (<u>https://ebas.nilu.no</u>) and SMPS data before 2020 through <u>https://doi.org/10.5281/zenodo.6794562</u> (Crumeyrolle et al., 2022). GDAS files for back-trajectory calculations are available at <u>https://www.arl.noaa.gov/hysplit/</u>. NOx data can be found through the ATMO open data website: <u>https://data-atmo-hdf.opendata.arcgis.com</u>.

- 533
- 534
- 535

# 536 **References**

537 Barreira, L.M.F., Helin, A., Aurela, M., Teinilä, K., Friman, M., Kangas, L., Niemi, J.V., Portin, H.,

538 Kousa, A., Pirjola, L., Rönkkö, T., Saarikoski, S., Timonen, H., 2020. In-depth characterization of

submicron particulate matter inter-annual variations at a street canyon site in Northern Europe (preprint).

540 Aerosols/Field Measurements/Troposphere/Chemistry (chemical composition and reactions).

541 https://doi.org/10.5194/acp-2020-908

542 Berland, K., Rose, C., Pey, J., Culot, A., Freney, E., Kalivitis, N., Kouvarakis, G., Cerro, J.C., Mallet, M.,

543 Sartelet, K., Beckmann, M., Bourriane, T., Roberts, G., Marchand, N., Mihalopoulos, N., Sellegri, K.,

544 2017. Spatial extent of new particle formation events over the Mediterranean Basin from multiple ground-

545 based and airborne measurements. Atmospheric Chem. Phys. 17, 9567–9583. 546 https://doi.org/10.5194/acp-17-9567-2017

547 Boichu, M., Favez, O., Riffault, V., Petit, J.-E., Zhang, Y., Brogniez, C., Sciare, J., Chiapello, I., Clarisse,

548 L., Zhang, S., Pujol-Söhne, N., Tison, E., Delbarre, H., Goloub, P., 2019. Large-scale particulate air

549 pollution and chemical fingerprint of volcanic sulfate aerosols from the 2014–2015 Holuhraun flood lava

550 eruption of Bárðarbunga volcano (Iceland). Atmospheric Chem. Phys. 19, 14253-14287.

- 551 https://doi.org/10.5194/acp-19-14253-2019
- 552 Bousiotis, D., Brean, J., Pope, F.D., Dall'Osto, M., Querol, X., Alastuey, A., Perez, N., Petäjä, T.,
- 553 Massling, A., Nøjgaard, J.K., Nordstrøm, C., Kouvarakis, G., Vratolis, S., Eleftheriadis, K., Niemi, J.V.,
- 554 Portin, H., Wiedensohler, A., Weinhold, K., Merkel, M., Tuch, T., Harrison, R.M., 2021. The effect of
- 555 meteorological conditions and atmospheric composition in the occurrence and development of new
- 556 particle formation (NPF) events in Europe. Atmospheric Chem. Phys. 21, 3345-3370.
- 557 https://doi.org/10.5194/acp-21-3345-2021
- 558 Bousiotis, D., Dall'Osto, M., Beddows, D.C.S., Pope, F.D., Harrison, R.M., 2019. Analysis of new
- 559 particle formation (NPF) events at nearby rural, urban background and urban roadside sites. Atmospheric
- 560 Chem. Phys. 19, 5679–5694. https://doi.org/10.5194/acp-19-5679-2019
- 561 Bovchaliuk, V., Goloub, P., Podvin, T., Veselovskii, I., Tanre, D., Chaikovsky, A., Dubovik, O., Mortier,
- 562 A., Lopatin, A., Korenskiy, M., Victori, S., 2016. Comparison of aerosol properties retrieved using
- 563 GARRLiC, LIRIC, and Raman algorithms applied to multi-wavelength lidar and sun/sky-photometer
- 564 data. Atmos Meas Tech 9, 3391–3405. https://doi.org/10.5194/amt-9-3391-2016
- 565 Buenrostro Mazon, S., Kontkanen, J., Manninen, H.E., Nieminen, T., Kerminen, V.-M., Kulmala, M.,
- 566 2016. A long-term comparison of nighttime cluster events and daytime ion formation in a boreal forest.
- 567 Boreal Environ. Res. 21, 242–261.
- 568 Casquero-Vera, J.A., Lyamani, H., Titos, G., Minguillón, M.C., Dada, L., Alastuey, A., Querol, X.,
- 569 Petäjä, T., Olmo, F.J., Alados-Arboledas, L., 2021. Quantifying traffic, biomass burning and secondary
- 570 source contributions to atmospheric particle number concentrations at urban and suburban sites. Sci. Total
- 571 Environ. 768, 145282. https://doi.org/10.1016/j.scitotenv.2021.145282
- 572 Chen, G., Canonaco, F., Tobler, A., Aas, W., Alastuey, A., Allan, J., Atabakhsh, S., Aurela, M.,
- 573 Baltensperger, U., Bougiatioti, A., De Brito, J.F., Ceburnis, D., Chazeau, B., Chebaicheb, H.,
- 574 Daellenbach, K.R., Ehn, M., El Haddad, I., Eleftheriadis, K., Favez, O., Flentje, H., Font, A., Fossum, K.,
- 575 Freney, E., Gini, M., Green, D.C., Heikkinen, L., Herrmann, H., Kalogridis, A.-C., Keernik, H., Lhotka,
- 576 R., Lin, C., Lunder, C., Maasikmets, M., Manousakas, M.I., Marchand, N., Marin, C., Marmureanu, L.,
- 577 Mihalopoulos, N., Močnik, G., Nęcki, J., O'Dowd, C., Ovadnevaite, J., Peter, T., Petit, J.-E., Pikridas,
- 578 M., Matthew Platt, S., Pokorná, P., Poulain, L., Priestman, M., Riffault, V., Rinaldi, M., Różański, K.,

- 579 Schwarz, J., Sciare, J., Simon, L., Skiba, A., Slowik, J.G., Sosedova, Y., Stavroulas, I., Styszko, K.,
- 580 Teinemaa, E., Timonen, H., Tremper, A., Vasilescu, J., Via, M., Vodička, P., Wiedensohler, A., Zografou,
- 581 O., Cruz Minguillón, M., Prévôt, A.S.H., 2022. European Aerosol Phenomenology 8: Harmonised
- Source Apportionment of Organic Aerosol using 22 Year-long ACSM/AMS Datasets. Environ. Int.
   107325. https://doi.org/10.1016/i.envint.2022.107325
- 584 Clifford, S., Mazaheri, M., Salimi, F., Ezz, W.N., Yeganeh, B., Low-Choy, S., Walker, K., Mengersen,
- 585 K., Marks, G.B., Morawska, L., 2018. Effects of exposure to ambient ultrafine particles on respiratory
- 586 health and systemic inflammation in children. Environ. Int. 114, 167–180. 587 https://doi.org/10.1016/j.envint.2018.02.019
- 588 Crumeyrolle, S., Kontkanen, J., Rose, C., Velasquez Garcia, A., Bourrianne, E., Catalfamo, M., Riffault,
- 589 V., Tison, E., Ferreira de Brito, J., Visez, N., Ferlay, N., Auriol, F., Chiapello, I., 2022. Measurement
- 590 report: Atmospheric new particle formation in a peri-urban site in Lille, Northern France. Atmospheric
- 591 Chem. Phys. Discuss. 1–35. https://doi.org/10.5194/acp-2022-436
- 592 Cuesta-Mosquera, A., Močnik, G., Drinovec, L., Müller, T., Pfeifer, S., Minguillón, M., Björn, B.,
- 593 Buckley, P., Dudoitis, V., Fernández-García, J., Fernández Amado, M., Brito, J., Flentje, H., Heffernan,
- 594 E., Kalivitis, N., Kalogridis, C., Keernik, H., Marmureanu, L., Luoma, K., Wiedensohler, A., 2020.
- Intercomparison and characterization of 23 Aethalometers under laboratory and ambient air conditions:
  Procedures and unit-to-unit variabilities. https://doi.org/10.5194/amt-2020-344
- 597 Dada, L., Paasonen, P., Nieminen, T., Buenrostro Mazon, S., Kontkanen, J., Peräkylä, O., Lehtipalo, K.,
- 598 Hussein, T., Petäjä, T., Kerminen, V.-M., Bäck, J., Kulmala, M., 2017. Long-term analysis of clear-sky
- 599 new particle formation events and nonevents in Hyytiälä. Atmospheric Chem. Phys. 17, 6227–6241.
- 600 https://doi.org/10.5194/acp-17-6227-2017
- Dal Maso, M., Kulmala, M., Riipinen, I., Wagner, R., Hussein, T., Aalto, P.P., Lehtinen, K.E.J., 2005.
- 602 Formation and growth of fresh atmospheric aerosols: Eight years of aerosol size distribution data from
- 503 SMEAR II, Hyytiälä, Finland. Boreal Environ. Res. 10, 323–336.
- 604 Dall'Osto, M., Beddows, D.C.S., Asmi, A., Poulain, L., Hao, L., Freney, E., Allan, J.D., Canagaratna,
- 605 M., Crippa, M., Bianchi, F., de Leeuw, G., Eriksson, A., Swietlicki, E., Hansson, H.C., Henzing, J.S.,
- 606 Granier, C., Zemankova, K., Laj, P., Onasch, T., Prevot, A., Putaud, J.P., Sellegri, K., Vidal, M., Virtanen,

- A., Simo, R., Worsnop, D., O'Dowd, C., Kulmala, M., Harrison, R.M., 2018. Novel insights on new
  particle formation derived from a pan-european observing system. Sci. Rep. 8, 1482.
  https://doi.org/10.1038/s41598-017-17343-9
- Deng, C., Fu, Y., Dada, L., Yan, C., Cai, R., Yang, D., Zhou, Y., Yin, R., Lu, Y., Li, X., Qiao, X., Fan,
  X., Nie, W., Kontkanen, J., Kangasluoma, J., Chu, B., Ding, A., Kerminen, V.-M., Paasonen, P.,
  Worsnop, D.R., Bianchi, F., Liu, Y., Zheng, J., Wang, L., Kulmala, M., Jiang, J., 2020. Seasonal
- 613 Characteristics of New Particle Formation and Growth in Urban Beijing. Environ. Sci. Technol. 54, 8547–
- 614 8557. https://doi.org/10.1021/acs.est.0c00808
- 615 Dos Santos, V.N., Herrmann, E., Manninen, H.E., Hussein, T., Hakala, J., Nieminen, T., Aalto, P.P.,
- 616 Merkel, M., Wiedensohler, A., Kulmala, M., Petäjä, T., Hämeri, K., 2015. Variability of air ion
- 617 concentrations in urban Paris. Atmospheric Chem. Phys. 15, 13717–13737. https://doi.org/10.5194/acp-
- 618 15-13717-2015
- 619 Duplissy, J., Merikanto, J., Franchin, A., Tsagkogeorgas, G., Kangasluoma, J., Wimmer, D., Vuollekoski,
- 620 H., Schobesberger, S., Lehtipalo, K., Flagan, R.C., Brus, D., Donahue, N.M., Vehkamäki, H., Almeida,
- 521 J., Amorim, A., Barmet, P., Bianchi, F., Breitenlechner, M., Dunne, E.M., Guida, R., Henschel, H.,
- Junninen, H., Kirkby, J., Kürten, A., Kupc, A., Määttänen, A., Makhmutov, V., Mathot, S., Nieminen, T.,
- 623 Onnela, A., Praplan, A.P., Riccobono, F., Rondo, L., Steiner, G., Tome, A., Walther, H., Baltensperger,
- 624 U., Carslaw, K.S., Dommen, J., Hansel, A., Petäjä, T., Sipilä, M., Stratmann, F., Vrtala, A., Wagner, P.E.,
- 625 Worsnop, D.R., Curtius, J., Kulmala, M., 2016. Effect of ions on sulfuric acid-water binary particle
- 626 formation: 2. Experimental data and comparison with QC-normalized classical nucleation theory:
- 627 BINARY PARTICLE FORMATION EXPERIMENTS. J. Geophys. Res. Atmospheres 121, 1752–1775.
- 628 https://doi.org/10.1002/2015JD023539
- 629 Fick, J., Pommer, L., Nilsson, C., Andersson, B., 2003. Effect of OH radicals, relative humidity, and time
- on the composition of the products formed in the ozonolysis of  $\alpha$ -pinene. Atmos. Environ. 37, 4087–
- 631 4096. https://doi.org/10.1016/S1352-2310(03)00522-3
- 632 Fuks, N.A., Sutugin, A.G., 1970. Highly Dispersed Aerosols. Ann Arbor Science Publishers.
- 633 Größ, J., Hamed, A., Sonntag, A., Spindler, G., Manninen, H.E., Nieminen, T., Kulmala, M., Hõrrak, U.,
- 634 Plass-Dülmer, C., Wiedensohler, A., Birmili, W., 2018. Atmospheric new particle formation at the

- 635 research station Melpitz, Germany: connection with gaseous precursors and meteorological parameters.
- 636 Atmospheric Chem. Phys. 18, 1835–1861. https://doi.org/10.5194/acp-18-1835-2018
- 637 Hamed, A., Korhonen, H., Sihto, S.-L., Joutsensaari, J., Järvinen, H., Petäjä, T., Arnold, F., Nieminen,
- 538 T., Kulmala, M., Smith, J.N., Lehtinen, K.E.J., Laaksonen, A., 2011. The role of relative humidity in
- 639 continental new particle formation. J. Geophys. Res. 116, D03202.
- 640 https://doi.org/10.1029/2010JD014186
- 641 Heinritzi, M., Dada, L., Simon, M., Stolzenburg, D., Wagner, A.C., Fischer, L., Ahonen, L.R.,
- 642 Amanatidis, S., Baalbaki, R., Baccarini, A., Bauer, P.S., Baumgartner, B., Bianchi, F., Brilke, S., Chen,
- 643 D., Chiu, R., Dias, A., Dommen, J., Duplissy, J., Finkenzeller, H., Frege, C., Fuchs, C., Garmash, O.,
- 644 Gordon, H., Granzin, M., El Haddad, I., He, X., Helm, J., Hofbauer, V., Hoyle, C.R., Kangasluoma, J.,
- 645 Keber, T., Kim, C., Kürten, A., Lamkaddam, H., Laurila, T.M., Lampilahti, J., Lee, C.P., Lehtipalo, K.,
- 646 Leiminger, M., Mai, H., Makhmutov, V., Manninen, H.E., Marten, R., Mathot, S., Mauldin, R.L.,
- 647 Mentler, B., Molteni, U., Müller, T., Nie, W., Nieminen, T., Onnela, A., Partoll, E., Passananti, M., Petäjä,
- 548 T., Pfeifer, J., Pospisilova, V., Quéléver, L.L.J., Rissanen, M.P., Rose, C., Schobesberger, S., Scholz, W.,
- 649 Scholze, K., Sipilä, M., Steiner, G., Stozhkov, Y., Tauber, C., Tham, Y.J., Vazquez-Pufleau, M., Virtanen,
- 650 A., Vogel, A.L., Volkamer, R., Wagner, R., Wang, M., Weitz, L., Wimmer, D., Xiao, M., Yan, C., Ye,
- 651 P., Zha, Q., Zhou, X., Amorim, A., Baltensperger, U., Hansel, A., Kulmala, M., Tomé, A., Winkler, P.M.,
- 652 Worsnop, D.R., Donahue, N.M., Kirkby, J., Curtius, J., 2020. Molecular understanding of the suppression
- 653 of new-particle formation by isoprene. Atmospheric Chem. Phys. 20, 11809–11821. 654 https://doi.org/10.5194/acp-20-11809-2020
- 655 Hussein, T., Atashi, N., Sogacheva, L., Hakala, S., Dada, L., Petäjä, T., Kulmala, M., 2020.
- 656 Characterization of Urban New Particle Formation in Amman—Jordan. Atmosphere 11, 79.
- 657 https://doi.org/10.3390/atmos11010079
- 558 Jokinen, V., Mäkelä, J.M., 1997. Closed-loop arrangement with critical orifice for DMA sheath/excess
- 659 flow system. J. Aerosol Sci. 28, 643–648. https://doi.org/10.1016/S0021-8502(96)00457-0
- 660 Kalkavouras, P., Bossioli, E., Bezantakos, S., Bougiatioti, A., Kalivitis, N., Stavroulas, I., Kouvarakis,
- 661 G., Protonotariou, A.P., Dandou, A., Biskos, G., Mihalopoulos, N., Nenes, A., Tombrou, M., 2017. New
- b62 particle formation in the southern Aegean Sea during the Etesians: importance for CCN production and

- cloud droplet number. Atmospheric Chem. Phys. 17, 175–192. https://doi.org/10.5194/acp-17-175-2017
- Kanawade, V.P., Sebastian, M., Hooda, R.K., Hyvärinen, A.-P., 2022. Atmospheric new particle
  formation in India: Current understanding and knowledge gaps. Atmos. Environ. 270, 118894.
- 666 https://doi.org/10.1016/j.atmosenv.2021.118894
- Kerminen, V.-M., Chen, X., Vakkari, V., Petäjä, T., Kulmala, M., Bianchi, F., 2018. Atmospheric new
  particle formation and growth: review of field observations. Environ. Res. Lett. 13, 103003.
  https://doi.org/10.1088/1748-9326/aadf3c
- 670 Kerminen, V.-M., Pirjola, L., Kulmala, M., 2001. How significantly does coagulational scavenging limit
- atmospheric particle production? J. Geophys. Res. Atmospheres 106, 24119–24125.
  https://doi.org/10.1029/2001JD000322
- 673 Kontkanen, J., Järvinen, E., Manninen, H.E., Lehtipalo, K., Kangasluoma, J., Decesari, S., Gobbi, G.P.,
- 674 Laaksonen, A., Petäjä, T., Kulmala, M., 2016. High concentrations of sub-3nm clusters and frequent new
- 675 particle formation observed in the Po Valley, Italy, during the PEGASOS 2012 campaign. Atmospheric
- 676 Chem. Phys. 16, 1919–1935. https://doi.org/10.5194/acp-16-1919-2016
- 677 Kontkanen, J., Lehtipalo, K., Ahonen, L., Kangasluoma, J., Manninen, H.E., Hakala, J., Rose, C., Sellegri,
- K., Xiao, S., Wang, L., Qi, X., Nie, W., Ding, A., Yu, H., Lee, S., Kerminen, V.-M., Petäjä, T., Kulmala,
- 679 M., 2017. Measurements of sub-3 nm particles using a particle size magnifier in different environments:
- 680 from clean mountain top to polluted megacities. Atmospheric Chem. Phys. 17, 2163-2187.
- 681 https://doi.org/10.5194/acp-17-2163-2017
- Kulmala, M., 2003. Atmospheric science. How particles nucleate and grow. Science 302, 1000–1001.
- 683 https://doi.org/10.1126/science.1090848
- Kulmala, M., Dal Maso, M., Mäkelä, J.M., Pirjola, L., Väkevä, M., Aalto, P., Miikkulainen, P., Hämeri,
- 685 K., O'Dowd, C.D., 2001. On the formation, growth and composition of nucleation mode particles. Tellus
- 686 Ser. B Chem. Phys. Meteorol. 53, 479–490. https://doi.org/10.1034/j.1600-0889.2001.530411.x
- 687 Kulmala, M., Junninen, H., Dada, L., Salma, I., Weidinger, T., Thén, W., Vörösmarty, M., Komsaare, K.,
- 588 Stolzenburg, D., Cai, R., Yan, C., Li, X., Deng, C., Jiang, J., Petäjä, T., Nieminen, T., Kerminen, V.-M.,
- 689 2022. Quiet New Particle Formation in the Atmosphere. Front. Environ. Sci. 10.
- 690 Kulmala, M., Kerminen, V.-M., Petäjä, T., Ding, A.J., Wang, L., 2017. Atmospheric gas-to-particle

- 691 conversion: why NPF events are observed in megacities? Faraday Discuss. 200, 271-288. 692 https://doi.org/10.1039/C6FD00257A
- Kulmala, M., Petäjä, T., Ehn, M., Thornton, J., Sipilä, M., Worsnop, D.R., Kerminen, V.-M., 2014. 693
- 694 Chemistry of Atmospheric Nucleation: On the Recent Advances on Precursor Characterization and
- 695 Atmospheric Cluster Composition in Connection with Atmospheric New Particle Formation. Annu. Rev.
- 696 Phys. Chem. 65, 21–37. https://doi.org/10.1146/annurev-physchem-040412-110014
- 697 Kulmala, M., Petäjä, T., Nieminen, T., Sipilä, M., Manninen, H.E., Lehtipalo, K., Maso, M.D., Aalto,
- 698 P.P., Junninen, H., Paasonen, P., Riipinen, I., Lehtinen, K.E.J., Laaksonen, A., Kerminen, V.M., 2012.
- 699 Measurement of the nucleation of atmospheric aerosol particles. Nat. Protoc. 7, 1651–1667. 700 https://doi.org/10.1038/nprot.2012.091
- 701 Kulmala, M., Vehkamäki, H., Petäjä, T., Dal Maso, M., Lauri, A., Kerminen, V.-M., Birmili, W.,
- 702 McMurry, P.H., 2004. Formation and growth rates of ultrafine atmospheric particles: a review of 703 observations. J. Aerosol Sci. 35, 143–176. https://doi.org/10.1016/j.jaerosci.2003.10.003
- 704 Kurtén, T., Torpo, L., Ding, C.-G., Vehkamäki, H., Sundberg, M.R., Laasonen, K., Kulmala, M., 2007. 705 A density functional study on water-sulfuric acid-ammonia clusters and implications for atmospheric 706 cluster formation. J. Geophys. Res. Atmospheres 112. https://doi.org/10.1029/2006JD007391
- Laj, P., Bigi, A., Rose, C., Andrews, E., Lund Myhre, C., Collaud Coen, M., Lin, Y., Wiedensohler, A., 707
- 708 Schulz, M., Ogren, J.A., Fiebig, M., Gliß, J., Mortier, A., Pandolfi, M., Petäja, T., Kim, S.-W., Aas, W.,
- 709
- Putaud, J.-P., Mayol-Bracero, O., Keywood, M., Labrador, L., Aalto, P., Ahlberg, E., Alados Arboledas,
- 710 L., Alastuev, A., Andrade, M., Artíñano, B., Ausmeel, S., Arsov, T., Asmi, E., Backman, J.,
- 711 Baltensperger, U., Bastian, S., Bath, O., Beukes, J.P., Brem, B.T., Bukowiecki, N., Conil, S., Couret, C.,
- 712 Day, D., Dayantolis, W., Degorska, A., Eleftheriadis, K., Fetfatzis, P., Favez, O., Flentje, H., Gini, M.I.,
- 713 Gregorič, A., Gysel-Beer, M., Hallar, A.G., Hand, J., Hoffer, A., Hueglin, C., Hooda, R.K., Hyvärinen,
- A., Kalapov, I., Kalivitis, N., Kasper-Giebl, A., Kim, J.E., Kouvarakis, G., Kranjc, I., Krejci, R., Kulmala, 714
- 715 M., Labuschagne, C., Lee, H.-J., Lihavainen, H., Lin, N.-H., Löschau, G., Luoma, K., Marinoni, A.,
- 716 Martins Dos Santos, S., Meinhardt, F., Merkel, M., Metzger, J.-M., Mihalopoulos, N., Nguyen, N.A.,
- 717 Ondracek, J., Pérez, N., Perrone, M.R., Petit, J.-E., Picard, D., Pichon, J.-M., Pont, V., Prats, N., Prenni,
- 718 A., Reisen, F., Romano, S., Sellegri, K., Sharma, S., Schauer, G., Sheridan, P., Sherman, J.P., Schütze,

- 719 M., Schwerin, A., Sohmer, R., Sorribas, M., Steinbacher, M., Sun, J., Titos, G., Toczko, B., Tuch, T.,
- 720 Tulet, P., Tunved, P., Vakkari, V., Velarde, F., Velasquez, P., Villani, P., Vratolis, S., Wang, S.-H.,
- 721 Weinhold, K., Weller, R., Yela, M., Yus-Diez, J., Zdimal, V., Zieger, P., Zikova, N., 2020. A global
- analysis of climate-relevant aerosol properties retrieved from the network of Global Atmosphere Watch
- 723 (GAW) near-surface observatories. Atmospheric Meas. Tech. 13, 4353-4392.
- 724 https://doi.org/10.5194/amt-13-4353-2020
- Lampilahti, J., Leino, K., Manninen, A., Poutanen, P., Franck, A., Peltola, M., Hietala, P., Beck, L., Dada,
- L., Quéléver, L., Öhrnberg, R., Zhou, Y., Ekblom, M., Vakkari, V., Zilitinkevich, S., Kerminen, V.-M.,
- 727 Petäjä, T., Kulmala, M., 2021. Aerosol particle formation in the upper residual layer. Atmospheric Chem.
- 728 Phys. 21, 7901–7915. https://doi.org/10.5194/acp-21-7901-2021
- 729 Lampilahti, J., Manninen, H.E., Leino, K., Väänänen, R., Manninen, A., Buenrostro Mazon, S.,
- 730 Nieminen, T., Leskinen, M., Enroth, J., Bister, M., Zilitinkevich, S., Kangasluoma, J., Järvinen, H.,
- 731 Kerminen, V.-M., Petäjä, T., Kulmala, M., 2020. Roll vortices induce new particle formation bursts in
- the planetary boundary layer. Atmospheric Chem. Phys. 20, 11841–11854. https://doi.org/10.5194/acp20-11841-2020
- Liao, L., Kerminen, V.-M., Boy, M., Kulmala, M., Dal Maso, M., 2014. Temperature influence on the
  natural aerosol budget over boreal forests. Atmospheric Chem. Phys. 14, 8295–8308.
  https://doi.org/10.5194/acp-14-8295-2014
- Lyubovtseva, Y.S., Sogacheva, L., Maso, M.D., Bonn, B., Keronen, P., Kulmala, M., 2005. Seasonal
  variations of trace gases, meteorological parameters, and formation of aerosols in boreal forests 10, 18.
- 739 Marten, R., Xiao, M., Rörup, B., Wang, M., Kong, W., He, X.-C., Stolzenburg, D., Pfeifer, J., Marie, G.,
- 740 Wang, D.S., Scholz, W., Baccarini, A., Lee, C.P., Amorim, A., Baalbaki, R., Bell, D.M., Bertozzi, B.,
- 741 Caudillo, L., Chu, B., Dada, L., Duplissy, J., Finkenzeller, H., Carracedo, L.G., Granzin, M., Hansel, A.,
- 742 Heinritzi, M., Hofbauer, V., Kemppainen, D., Kürten, A., Lampimäki, M., Lehtipalo, K., Makhmutov,
- 743 V., Manninen, H.E., Mentler, B., Petäjä, T., Philippov, M., Shen, J., Simon, M., Stozhkov, Y., Tomé, A.,
- 744 Wagner, A.C., Wang, Y., Weber, S.K., Wu, Y., Zauner-Wieczorek, M., Curtius, J., Kulmala, M., Möhler,
- 745 O., Volkamer, R., Winkler, P.M., Worsnop, D.R., Dommen, J., Flagan, R.C., Kirkby, J., Donahue, N.M.,
- 746 Lamkaddam, H., Baltensperger, U., Haddad, I.E., 2022. Survival of newly formed particles in haze

- conditions. Environ. Sci. Atmospheres 2, 491–499. https://doi.org/10.1039/D2EA00007E
- 748 Mazon, S.B., Riipinen, I., Schultz, D.M., Valtanen, M., Maso, M.D., Sogacheva, L., Junninen, H.,
- Nieminen, T., 2009. Classifying previously undefined days from eleven years of aerosol-particle-size
  distribution data from the SMEAR II station, Hyytia "la", Finland. Atmos Chem Phys 10.
- 751 Merikanto, J., Duplissy, J., Määttänen, A., Henschel, H., Donahue, N.M., Brus, D., Schobesberger, S.,
- 752 Kulmala, M., Vehkamäki, H., 2016. Effect of ions on sulfuric acid-water binary particle formation: 1.
- 753 Theory for kinetic- and nucleation-type particle formation and atmospheric implications: BINARY
- PARTICLE FORMATION THEORY. J. Geophys. Res. Atmospheres 121, 1736–1751.
  https://doi.org/10.1002/2015JD023538
- 756 Middlebrook, A.M., Bahreini, R., Jimenez, J.L., Canagaratna, M.R., 2012. Evaluation of Composition-
- 757 Dependent Collection Efficiencies for the Aerodyne Aerosol Mass Spectrometer using Field Data.
- 758 Aerosol Sci. Technol. 46, 258–271. https://doi.org/10.1080/02786826.2011.620041
- 759 Mortier, A., Goloub, P., Podvin, T., Deroo, C., Chaikovsky, A., Ajtai, N., Blarel, L., Tanre, D., Derimian,
- Y., 2013. Detection and characterization of volcanic ash plumes over Lille during the Eyjafjallajökull
  eruption. Atmos Chem Phys 13, 3705–3720. https://doi.org/10.5194/acp-13-3705-2013
- 762 Németh, Z., Rosati, B., Zíková, N., Salma, I., Bozó, L., Dameto de España, C., Schwarz, J., Ždímal, V.,
- Wonaschütz, A., 2018. Comparison of atmospheric new particle formation events in three Central
  European cities. Atmos. Environ. 178, 191–197. https://doi.org/10.1016/j.atmosenv.2018.01.035
- 765 Nieminen, T., Kerminen, V.-M., Petäjä, T., Aalto, P.P., Arshinov, M., Asmi, E., Baltensperger, U.,
- 766 Beddows, D.C.S., Beukes, J.P., Collins, D., Ding, A., Harrison, R.M., Henzing, B., Hooda, R., Hu, M.,
- 767 Hõrrak, U., Kivekäs, N., Komsaare, K., Krejci, R., Kristensson, A., Laakso, L., Laaksonen, A., Leaitch,
- 768 W.R., Lihavainen, H., Mihalopoulos, N., Németh, Z., Nie, W., O'Dowd, C., Salma, I., Sellegri, K.,
- 769 Svenningsson, B., Swietlicki, E., Tunved, P., Ulevicius, V., Vakkari, V., Vana, M., Wiedensohler, A.,
- 770 Wu, Z., Virtanen, A., Kulmala, M., 2018. Global analysis of continental boundary layer new particle
- formation based on long-term measurements. Atmospheric Chem. Phys. 18, 14737-14756.
- 772 https://doi.org/10.5194/acp-18-14737-2018
- 773 Ohlwein, S., Kappeler, R., Kutlar Joss, M., Künzli, N., Hoffmann, B., 2019. Health effects of ultrafine
- particles: a systematic literature review update of epidemiological evidence. Int. J. Public Health 64, 547–

- 775 559. https://doi.org/10.1007/s00038-019-01202-7
- 776 Paasonen, P., Peltola, M., Kontkanen, J., Junninen, H., Kerminen, V.-M., Kulmala, M., 2018.
- Comprehensive analysis of particle growth rates from nucleation mode to cloud condensation nuclei in
  boreal forest. Atmospheric Chem. Phys. 18, 12085–12103. https://doi.org/10.5194/acp-18-12085-2018
- 779 Peng, Y., Dong, Y., Li, X., Liu, X., Dai, J., Chen, C., Dong, Z., Du, C., Wang, Z., 2017. Different
- 780 Characteristics of New Particle Formation Events at Two Suburban Sites in Northern China. Atmosphere
- 781 8, 258. https://doi.org/10.3390/atmos8120258
- 782 Pierce, J.R., Adams, P.J., 2009. Uncertainty in global CCN concentrations from uncertain aerosol
- nucleation and primary emission rates. Atmospheric Chem. Phys. 9, 1339–1356.
  https://doi.org/10.5194/acp-9-1339-2009
- Pushpawela, B., Jayaratne, R., Morawska, L., 2018. Temporal distribution and other characteristics of
- new particle formation events in an urban environment. Environ. Pollut. 233, 552–560.
  https://doi.org/10.1016/j.envpol.2017.10.102
- 788 Ren, J., Chen, L., Fan, T., Liu, J., Jiang, S., Zhang, F., 2021. The NPF Effect on CCN Number
- 789 Concentrations: A Review and Re-Evaluation of Observations From 35 Sites Worldwide. Geophys. Res.
- 790 Lett. 48, e2021GL095190. https://doi.org/10.1029/2021GL095190
- 791 Rivellini, L.-H., Chiapello, I., Tison, E., Fourmentin, M., Féron, A., Diallo, A., N'Diaye, T., Goloub, P.,
- 792 Canonaco, F., Prévôt, A.S.H., Riffault, V., 2017. Chemical characterization and source apportionment of
- submicron aerosols measured in Senegal during the 2015 SHADOW campaign. Atmos Chem Phys 17,
- 794 10291–10314. https://doi.org/10.5194/acp-17-10291-2017
- 795 Rodríguez, S., Cuevas, E., González, Y., Ramos, R., Romero, P.M., Pérez, N., Querol, X., Alastuey, A.,
- 796 2008. Influence of sea breeze circulation and road traffic emissions on the relationship between particle
- number, black carbon, PM1, PM2.5 and PM2.5–10 concentrations in a coastal city. Atmos. Environ. 42,
- 798 6523–6534. https://doi.org/10.1016/j.atmosenv.2008.04.022
- Rodríguez, S., Van Dingenen, R., Putaud, J.-P., Martins-Dos Santos, S., Roselli, D., 2005. Nucleation
- 800 and growth of new particles in the rural atmosphere of Northern Italy-relationship to air quality
- 801 monitoring. Atmos. Environ. 39, 6734–6746. https://doi.org/10.1016/j.atmosenv.2005.07.036
- 802 Roig Rodelas, R., Chakraborty, A., Perdrix, E., Tison, E., Riffault, V., 2019. Real-time assessment of

- 803 wintertime organic aerosol characteristics and sources at a suburban site in northern France. Atmos.
- 804 Environ. 203, 48–61. https://doi.org/10.1016/j.atmosenv.2019.01.035
- Rolph, G., Stein, A., Stunder, B., 2017. Real-time Environmental Applications and Display sYstem:
  READY. Environ. Model. Softw. 95, 210–228. https://doi.org/10.1016/j.envsoft.2017.06.025
- 807 Rose, C., Collaud Coen, M., Andrews, E., Lin, Y., Bossert, I., Lund Myhre, C., Tuch, T., Wiedensohler,
- 808 A., Fiebig, M., Aalto, P., Alastuey, A., Alonso-Blanco, E., Andrade, M., Artíñano, B., Arsov, T.,
- 809 Baltensperger, U., Bastian, S., Bath, O., Beukes, J.P., Brem, B.T., Bukowiecki, N., Casquero-Vera, J.A.,
- 810 Conil, S., Eleftheriadis, K., Favez, O., Flentje, H., Gini, M.I., Gómez-Moreno, F.J., Gysel-Beer, M.,
- 811 Hallar, A.G., Kalapov, I., Kalivitis, N., Kasper-Giebl, A., Keywood, M., Kim, J.E., Kim, S.-W.,
- 812 Kristensson, A., Kulmala, M., Lihavainen, H., Lin, N.-H., Lyamani, H., Marinoni, A., Martins Dos
- 813 Santos, S., Mayol-Bracero, O.L., Meinhardt, F., Merkel, M., Metzger, J.-M., Mihalopoulos, N., Ondracek,
- J., Pandolfi, M., Pérez, N., Petäjä, T., Petit, J.-E., Picard, D., Pichon, J.-M., Pont, V., Putaud, J.-P., Reisen,
- 815 F., Sellegri, K., Sharma, S., Schauer, G., Sheridan, P., Sherman, J.P., Schwerin, A., Sohmer, R., Sorribas,
- 816 M., Sun, J., Tulet, P., Vakkari, V., van Zyl, P.G., Velarde, F., Villani, P., Vratolis, S., Wagner, Z., Wang,
- 817 S.-H., Weinhold, K., Weller, R., Yela, M., Zdimal, V., Laj, P., 2021. Seasonality of the particle number
- 818 concentration and size distribution: a global analysis retrieved from the network of Global Atmosphere
- 819 Watch (GAW) near-surface observatories. Atmospheric Chem. Phys. 21, 17185–17223.
- 820 https://doi.org/10.5194/acp-21-17185-2021
- 821 Rose, C., Sellegri, K., Asmi, E., Hervo, M., Freney, E., Colomb, A., Junninen, H., Duplissy, J., Sipilä,
- 822 M., Kontkanen, J., Lehtipalo, K., Kulmala, M., 2015a. Major contribution of neutral clusters to new
- 823 particle formation at the interface between the boundary layer and the free troposphere. Atmospheric
- 824 Chem. Phys. 15, 3413–3428. https://doi.org/10.5194/acp-15-3413-2015
- 825 Rose, C., Sellegri, K., Freney, E., Dupuy, R., Colomb, A., Pichon, J.-M., Ribeiro, M., Bourianne, T.,
- 826 Burnet, F., Schwarzenboeck, A., 2015b. Airborne measurements of new particle formation in the free
- troposphere above the Mediterranean Sea during the HYMEX campaign. Atmospheric Chem. Phys. 15,
- 828 10203–10218. https://doi.org/10.5194/acp-15-10203-2015
- 829 Rose, C., Sellegri, K., Moreno, I., Velarde, F., Ramonet, M., Weinhold, K., Krejci, R., Andrade, M.,
- 830 Wiedensohler, A., Ginot, P., Laj, P., 2017. CCN production by new particle formation in the free

- troposphere. Atmospheric Chem. Phys. 17, 1529–1541. https://doi.org/10.5194/acp-17-1529-2017
- 832 Salimi, F., Rahman, M.M., Clifford, S., Ristovski, Z., Morawska, L., 2017. Nocturnal new particle
- 833 formation events in urban environments. Atmospheric Chem. Phys. 17, 521–530.
  834 https://doi.org/10.5194/acp-17-521-2017
- 835 Salma, I., Németh, Z., Kerminen, V.-M., Aalto, P., Nieminen, T., Weidinger, T., Molnár, Á., Imre, K.,
- 836 Kulmala, M., 2016. Regional effect on urban atmospheric nucleation. Atmospheric Chem. Phys. 16,
- 837 8715–8728. https://doi.org/10.5194/acp-16-8715-2016
- 838 Salma, I., Varga, V., Németh, Z., 2017. Quantification of an atmospheric nucleation and growth process
- as a single source of aerosol particles in a city. Atmospheric Chem. Phys. 17, 15007-15017.

840 https://doi.org/10.5194/acp-17-15007-2017

- 841 Sandradewi, J., Prévôt, A.S.H., Szidat, S., Perron, N., Alfarra, M.R., Lanz, V.A., Weingartner, E.,
- 842 Baltensperger, U., 2008. Using aerosol light absorption measurements for the quantitative determination
- of wood burning and traffic emission contributions to particulate matter. Environ. Sci. Technol. 42, 3316–
  3323.
- 845 Sebastian, M., Kanawade, V.P., Soni, V.K., Asmi, E., Westervelt, Daniel.M., Vakkari, V., Hyvärinen,
- 846 A.-P., Pierce, J.R., Hooda, R.K., 2021. New Particle Formation and Growth to Climate-Relevant Aerosols
- at a Background Remote Site in the Western Himalaya. J. Geophys. Res. Atmospheres 126,
  e2020JD033267. https://doi.org/10.1029/2020JD033267
- Seinfeld, J.H., Pandis, S.N., 2016. Atmospheric Chemistry and Physics: From Air Pollution to ClimateChange, Wiley. ed.
- 851 Sellegri, K., Rose, C., Marinoni, A., Lupi, A., Wiedensohler, A., Andrade, M., Bonasoni, P., Laj, P., 2019.
- 852 New Particle Formation: A Review of Ground-Based Observations at Mountain Research Stations.
- 853 Atmosphere 10, 493. https://doi.org/10.3390/atmos10090493
- 854 Shukla, K.K., Niranjan Kumar, K., Phanikumar, D.V., Newsom, R.K., Kotamarthi, V.R., Ouarda,
- 855 T.B.M.J., Ratnam, M.V., 2016. Identification of the cloud base height over the centralHimalayan region:
- 856 Intercomparison of Ceilometer and DopplerLidar (preprint). Clouds/Remote Sensing/Validation and
- 857 Intercomparisons. https://doi.org/10.5194/amt-2016-162
- 858 Sogacheva, L., Hamed, A., Facchini, M.C., Kulmala, M., Laaksonen, A., 2007. Relation of air mass

- history to nucleation events in Po Valley, Italy, using back trajectories analysis. Atmos Chem Phys 15.
- 860 Sogacheva, L., Maso, M.D., Kerminen, V.M., Kulmala, M., 2005. Probability of nucleation events and
- aerosol particle concentration in different air mass types arriving at Hyytiala southern Finland, based on
  back trajectories analysis. Boreal Environ. Res. 10, 479–491.
- 863 Spracklen, D.V., Carslaw, K.S., Kulmala, M., Kerminen, V.-M., Mann, G.W., Sihto, S.-L., 2006. The
- 864 contribution of boundary layer nucleation events to total particle concentrations on regional and global
- 865 scales. Atmospheric Chem. Phys. 6, 5631–5648. https://doi.org/10.5194/acp-6-5631-2006
- 866 Stein, A.F., Draxler, R.R., Rolph, G.D., Stunder, B.J.B., Cohen, M.D., Ngan, F., 2015. NOAA's
- 867 HYSPLIT Atmospheric Transport and Dispersion Modeling System. Bull. Am. Meteorol. Soc. 96, 2059–
  868 2077. https://doi.org/10.1175/BAMS-D-14-00110.1
- 869 Tillmann, R., Hallquist, M., Jonsson, Å.M., Kiendler-Scharr, A., Saathoff, H., Iinuma, Y., Mentel, Th.F.,
- 870 2010. Influence of relative humidity and temperature on the production of pinonaldehyde and OH radicals
- from the ozonolysis of  $\alpha$ -pinene. Atmospheric Chem. Phys. 10, 7057–7072. https://doi.org/10.5194/acp-
- 872 10-7057-2010
- 873 Tuch, T.M., Herbarth, O., Franck, U., Peters, A., Wehner, B., Wiedensohler, A., Heintzenberg, J., 2006.
- 874 Weak correlation of ultrafine aerosol particle concentrations <800 nm between two sites within one city.
- 875 J. Expo. Sci. Environ. Epidemiol. 16, 486–490. https://doi.org/10.1038/sj.jes.7500469
- 876 Tunved, P., Hansson, H.-C., Kerminen, V.-M., Ström, J., Dal Maso, M., Lihavainen, H., Viisanen, Y.,
- Aalto, P.P., Komppula, M., Kulmala, M., 2006. High natural aerosol loading over boreal forests. Science
  312, 261–263. https://doi.org/10.1126/science.1123052
- 879 Uusitalo, H., Kontkanen, J., Ylivinkka, I., Ezhova, E., Demakova, A., Arshinov, M., Belan, B.D.,
  880 Davydov, D., Ma, N., Petäjä, T., Wiedensohler, A., Kulmala, M., Nieminen, T., 2021. Occurrence of new
  881 particle formation events in Siberian and Finnish boreal forest (preprint). Aerosols/Field
- 882 Measurements/Troposphere/Physics (physical properties and processes). https://doi.org/10.5194/acp-
- 883 2021-530
- 884 Velazquez-Garcia, Crumeyrolle, S., Brito, J., Tison, E., Bourrianne, E., Chiapello, I., Riffault, V., 2022.
- 885 Deriving composition-dependent aerosol absorption, scattering and extinction mass efficiencies from
- 886 multi-annual high time resolution observations in Northern France. Submitt. Atmospheric Environ.

- Villani, P., Picard, D., Marchand, N., Laj, P., 2007. Design and validation of a 6-volatility tandem
  differential mobility analyzer (VTDMA). Aerosol Sci. Technol. 41, 898–906.
  https://doi.org/10.1080/02786820701534593
- 890 Wang, M., Kong, W., Marten, R., He, X.-C., Chen, D., Pfeifer, J., Heitto, A., Kontkanen, J., Dada, L.,
- 891 Kürten, A., Yli-Juuti, T., Manninen, H.E., Amanatidis, S., Amorim, A., Baalbaki, R., Baccarini, A., Bell,
- 892 D.M., Bertozzi, B., Bräkling, S., Brilke, S., Murillo, L.C., Chiu, R., Chu, B., De Menezes, L.-P., Duplissy,
- J., Finkenzeller, H., Carracedo, L.G., Granzin, M., Guida, R., Hansel, A., Hofbauer, V., Krechmer, J.,
- 894 Lehtipalo, K., Lamkaddam, H., Lampimäki, M., Lee, C.P., Makhmutov, V., Marie, G., Mathot, S.,
- 895 Mauldin, R.L., Mentler, B., Müller, T., Onnela, A., Partoll, E., Petäjä, T., Philippov, M., Pospisilova, V.,
- 896 Ranjithkumar, A., Rissanen, M., Rörup, B., Scholz, W., Shen, J., Simon, M., Sipilä, M., Steiner, G.,
- 897 Stolzenburg, D., Tham, Y.J., Tomé, A., Wagner, A.C., Wang, D.S., Wang, Y., Weber, S.K., Winkler,
- 898 P.M., Wlasits, P.J., Wu, Y., Xiao, M., Ye, Q., Zauner-Wieczorek, M., Zhou, X., Volkamer, R., Riipinen,
- 899 I., Dommen, J., Curtius, J., Baltensperger, U., Kulmala, M., Worsnop, D.R., Kirkby, J., Seinfeld, J.H.,
- El-Haddad, I., Flagan, R.C., Donahue, N.M., 2020. Rapid growth of new atmospheric particles by nitric
  acid and ammonia condensation. Nature 581, 184–189. https://doi.org/10.1038/s41586-020-2270-4
- 902 Wang, Z., Wu, Z., Yue, D., Shang, D., Guo, S., Sun, J., Ding, A., Wang, L., Jiang, J., Guo, H., Gao, J.,
- 903 Cheung, H.C., Morawska, L., Keywood, M., Hu, M., 2017. New particle formation in China: Current
- 904 knowledge and further directions. Sci. Total Environ. 577, 258–266. 905 https://doi.org/10.1016/i.scitotenv.2016.10.177
- 906 Wang, Z.B., Hu, M., Wu, Z.J., Yue, D.L., Zheng, J., Zhang, R.Y., Pei, X.Y., Paasonen, P., Dal Maso, M.,
- Boy, M., Wiedensohler, A., 2013. Investigation of the connections between atmospheric new particle 907 908 formation and organics at an urban site of Beijing (preprint). Aerosols/Field (chemical 909 Measurements/Troposphere/Chemistry composition and reactions).
- 910 https://doi.org/10.5194/acpd-13-3419-2013
- 911 Wehner, B., Wiedensohler, A., 2003. Long term measurements of submicrometer urban aerosols:
- 912 statistical analysis for correlations with meteorological conditions and trace gases. Atmospheric Chem.
- 913 Phys. 3, 867–879. https://doi.org/10.5194/acp-3-867-2003
- 914 Xiao, S., Wang, M.Y., Yao, L., Kulmala, M., Zhou, B., Yang, X., Chen, J.M., Wang, D.F., Fu, Q.Y.,

- 915 Worsnop, D.R., Wang, L., 2015. Strong atmospheric new particle formation in winter in urban Shanghai,
- 916 China. Atmospheric Chem. Phys. 15, 1769–1781. https://doi.org/10.5194/acp-15-1769-2015
- 917 Yao, L., Garmash, O., Bianchi, F., Zheng, J., Yan, C., Kontkanen, J., Junninen, H., Mazon, S.B., Ehn,
- 918 M., Paasonen, P., Sipilä, M., Wang, M., Wang, X., Xiao, S., Chen, H., Lu, Y., Zhang, B., Wang, D., Fu,
- 919 Q., Geng, F., Li, L., Wang, H., Qiao, L., Yang, X., Chen, J., Kerminen, V.-M., Petäjä, T., Worsnop, D.R.,
- 920 Kulmala, M., Wang, L., 2018. Atmospheric new particle formation from sulfuric acid and amines in a
- 921 Chinese megacity. Science 361, 278–281. https://doi.org/10.1126/science.aao4839
- 922 Yli-Juuti, T., Nieminen, T., Hirsikko, A., Aalto, P.P., Asmi, E., Hõrrak, U., Manninen, H.E., Patokoski,
- 923 J., Dal Maso, M., Petäjä, T., Rinne, J., Kulmala, M., Riipinen, I., 2011. Growth rates of nucleation mode
- 924 particles in Hyytiälä during 2003–2009: variation with particle size, season, data analysis method and
- 925 ambient conditions. Atmospheric Chem. Phys. 11, 12865–12886. https://doi.org/10.5194/acp-11-12865-
- 926 2011
- 927 928