Spatio-temporal variation of radionuclide dispersion from nuclear power plant accidents using FLEXPART ensemble modeling

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Abstract.

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We estimate-investigate the spatio-temporal distribution the seasonal and diurnal changes of radionuclides including Iodine-131 (¹³¹I) and Cesium-137 (¹³⁷Cs), transported to Qatar from a-fictitious accidentaccidents at the upwind Barakah Nuclear Power Plant (B-NPP) in the United Arab Emirates (UAE). <u>ToFor</u> modeling the dispersion of radionuclides, we use the

- 15 Lagrangian particle/air parcel dispersion model FLEXible PARTicle (FLEXPART) driven by forecasting and (re)analysis products, and <u>additionally</u> coupled with the Weather Research and Forecasting model (FLEXPART-WRF). A four-member ensemble of meteorological inputs, including one forecast dataset (CFSv2GFS) and three (re)analysis datasets (native resolution and downscaled FNL₂ and downscaled ERA5), is used to investigate the sensitivity of simulations to variations in meteorological inputs. We also investigatestudy the sensitivity of the simulations to accounting for the skewness in the vertical
- 20 velocity distribution. We investigate study the potential risk to populated areas from a nuclear accident in the eatchment region of interest-(Qatar). According to the simulated age spectrum of the Lagrangian particles, radionuclides enter southern Qatar about 20 to 30 hours after the release. Most of the radionuclide transport and deposition in the study area occurs withinin less than 50 and 80 hours after the release, respectively. A higher larger number of long-lived particles is found in the FNL-based simulations, which is interpreted as a greater dispersion of particles atinto the greater distancest part of the study area from the
- 25 emission pointlocation in these simulations. The highest simulated ¹³¹I concentrations and ¹³⁷Cs deposition show a pronounced spatio-temporal pattern. They are mostly found in the south/southeast of Qatar, during the <u>early-daytime</u> development of the boundary layer, and during the cold period of the year. The results show remarkable differences in the spatio-temporal distribution of ¹³¹I and ¹³⁷Cs simulations based on the FNL and GFS datasets, which share a common basic meteorological model, although they produce the simulations with the highest agreement. We conclude that the variation of the meteorological
- 30 <u>inputs</u>, <u>downscalingthe choice of the simulation code</u>, and the turbulence scheme used under convective conditions can collectively lead to_significant deviations_in_the_radionuclide dispersion_simulations. The most populated areas of Qatar

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coincide with moderate ¹³¹I concentration and ¹³⁷Cs depositions while uninhabited areas in southern Qatar receive the highest amounts. Simulated concentrations are found with the same level of consistency as reported for real case studies.

1 Introduction

- 35 Modeling the spatiotemporal distribution of radioactive materials compounds using chemical transport models Modeling the spatio-temporal distribution of radioactive materials through chemical transport models (CTMs), especially after large nuclear events, has received widespread attention especially after major nuclear events, has received widespread attention (Chino et al., 2011, Stohl et al., 2012, Christoudias and Lelieveld, 2013, Evangeliou et al., 2017). Whether explicitly stated or not, all-of-these studies aim to determine the extent and transport of radionuclides at various spatial scales. Thesey studies-also
- 40 aim toresearch activities help improve the performance of CTMs, which are at the core of preparedness programs for potential nuclear accidents (or releases) associated with the growing number of nuclear facilities Whether explicitly stated or implied, all these studies seek to determine the magnitude and transport of radionuclides at different spatial scales. Such research activities help to improve the performance of CTMs as the core of preparedness programs for potential nuclear accidents (or releases) related to the increasing number of nuclear facilities (Farid et al., 2017). However, case studies of reallarge accidents, which typically cover on the order of a few days to a few weeks, are not suitable for studying the effects of diurnal and seasonal
- 45 which typically cover on the order of a few days to a few weeks, are not suitable for studying the effects of diurnal and seasonal (atmospheric) changes on radionuclide dispersion, and the sensitivity to the meteorological (re-)analyses driving the <u>simulations</u>. However, ease studies of real accidents of the order of a few days are not suited to examine the impact of seasonal (atmospheric) changes on the radionuclide dispersion.
- 50 Maurer et al. (2018) found that the performance of dispersion models strongly depends on the successful modeling of boundary layer processes, which vary with season and time of day. Meteorological inputs to the CTMs are generated by atmospheric models, which may exhibit variability when simulating atmospheric conditions over nuclear accident areas (Arnold et al., 2015). For exampleIn addition, Long et al. (2019) studied the effects of the East Asian northeast monsoon on the transport of radionuclides from the Fukushima nuclear power plant accident to the tropical western Pacific and Southeast Asia. They found
- 55 that in these regions, radioactivity levels are is lower than in other regions of the Northern Hemisphere, which is due to the late arrival of the radionuclide plumes carried by the monsoon circulations. That is, the dispersion of radionuclides from this accident could potentially be different under-any other atmospheric conditions, which are only captured by the hypothetical, iterative simulation of this event at different times of the day and year. On the other hand, Maurer et al. (2018) found that the performance of dispersion models strongly depends on the successful modeling of boundary layer processes, which vary with
- 60 season and time of day. In term of local atmospheric conditions, In addition, meteorological inputs to the CTMs are generated by atmospheric models, which in some cases exhibit variability when simulating atmospheric conditions over nuclear accident areas (Arnold et al., 2015). This is mainly due to differences in initial conditions, spatial and temporal resolution, mathematical formulation, and physical parameterization.

While few new nuclear power plants are licensed in the Western world, and most Soviet-era stations are nearing the end-of life decommissioning, severalnew nuclear facilities are planned or proposed, and in the last few years are under construction or becoming operational in the Middle East/North Africa (MENA) region. The Barakah station is the latestmost recent NPP to become operational in a region with unique climatological conditions that were previously void of such developments and where the risk from radionuclide dispersion received little coverageattention in the literature, as opposed in contrast to Europe, Japan and the USA.

70 Using an ensemble of meteorological inputs, t

The main objective of this study is to investigate the spatiotemporal variability of radionuclide transport, surface concentration, and deposition <u>afterduring</u> potential nuclear accidents, <u>using an ensemble of meteorological inputs</u>. <u>Gaseous and aerosol</u>Two radio<u>nuclideactive</u> tracers, <u>including</u> iodine-131 (¹³¹I) and cesium-137 (¹³⁷Cs), are assumed to be released <u>by aina</u> fictitious <u>accidents</u> at the Barakah Nuclear Power Plant (B-NPP) in the United Arab Emirates (UAE). <u>Both radionuclides are</u>

- 75 emitted as gases, but due to its lower volatility, ¹³⁷Cs condenses on aerosol particles shortly after release (Christoudias and Lelieveld, 2013). The amount of ¹³⁷Cs that is deposited on the surface is important due to its relatively long half-life of ¹³⁷Cs (about 30 years)₁₃₇, whereas the ¹³¹L surface-level concentration (about 8 days half-life) is important for human health and in the biosphere in the short term (Tsuruta et al., 2019, Pisso et al., 2019, Takagi et al., 2020, Kinase et al., 2020, Wai et al., 2020) compared to ¹³¹L (about 8 days), gives significance to the amount of ¹³⁷Cs that undergoes surface deposition and to ¹³¹L
- 80 concentration in the biosphere (Tsuruta et al., 2019, Pisso et al., 2019, Takagi et al., 2020, Kinase et al., 2020, Wai et al., 2020).

For the dispersion modeling we use the Lagrangian particle model FLEXible PARTicle (FLEXPART) (Stohl et al., 1998, Stohl et al., 2005, Brioude et al., 2013). The Eulerian CTMs model the transport of air pollutants by numerically solving the equation for the conservation of mass (Moussiopoulos, 1997, Zhang and Chen, 2007), which is computationally expensive (Brioude et al., 2013). Moreover, in an Eulerian model, a particle released from a point source loses its position in the grid box (numerical diffusion). Lagrangian particle dispersion models (LPDMs) instead on the other hand have minimal numerical diffusion because they accurately model the trajectory of each particle (Nabi et al., 2015), instead of modelling the transport of air pollutants by numerically solving the equation for the conservation of mass (Moussiopoulos, 1997, Zhang and Chen, 2007),

90 for modeling particle dispersion because these modelsas they calculate advection and diffusion only for the location of each particle and not for the entire model domain. However, LPDMs also have their limitations as well. They suffer from numerical errors in the interpolations of meteorological fields in space and time. In some cases, the particles may not remain well-mixed during simulation (Brioude et al., 2013). This is mainly due to the treatment of because due to the lack of accuracy required to simulate the stochastic motion of the particles is not treated accurately enough and/or because the mass balance of vertical

as is the casenotably done with in Eulerian CTMs (Brioude et al., 2013). In addition, LPDMs are computationally more efficient

95 velocity is not very accurately mass balanced with the horizontal winds..., particles may not remain well-mixed during the simulation (Brioude et al., 2013). Regardless of which the type of CTM is formalism used, Girard et al. (2016) have shown

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that uncertainties in meteorological fields, namely wind speed and direction and precipitation, and emission rate <u>can</u> contribute significantly to errors in dispersion modeling. <u>According to</u> Galmarini et al. (2004), <u>T</u>there are three common methods for sensitivity analysis of LPDM simulations with respect to meteorological information fields are: i) generating perturbations in the horizontal and vertical position of particles, ii) using a single meteorological model with perturbations in initial conditions and/or model physics, and iii) using a number of different meteorological models (Galmarini et al., 2004). By adopting the third approach, we <u>used use</u> a four-member ensemble in which the performance of the forecast member, forced by 6-hourly data from the Global Forecast System (GFS)NCEP elimate forecast system version 2 (CFSv2), is-being compared to the members based (re)analysis datasets. (Re)analysis-based simulations (<u>unavailable in a real-world scenario</u>) are expected to be
closer to (unavailable in a real-world scenario) actual values than forecast-based simulations (Leadbetter et al., 2022). Two members of the ensemble are forced <u>bythrough</u> dynamically downscaled meteorological inputs to investigate the effects of downscaling on FLEXPARTto higher resolution ion dispersion modeling. We also studied investigate_the sensitivity of the simulations to the choice of turbulence schemes under convective conditions.

Finally, t²Fo study the potential risks to the local population, the radionuclide simulations were-are examined in relation to the population density of the catchment region of interest (Qatar). To the best of our knowledge, this is the first time such a study has been conducted for potential radionuclide releases in the study region, and we anticipate that our results may contribute to the formulation of preparedness plans. The paper is organized as follows: section 2 describes the LPDM, meteorological inputs, study area, and source term of radionuclides used in this study. The results are presented and discussed in section 3. The conclusions of the study are presented in section 4.

115 2 Ensemble model configuration

We use FLEXPART and FLEXPART driven by the Weather Research and Forecasting model (FLEXPART WRF) to model the dispersion of ¹³¹ I and ¹³⁷Cs fictitiously released from the B-NPP. A brief description of the FLEXPART modeling structure follows.

2.1 FLEXPART and FLEXPART-WRF dispersion modeling

- 120 We refer the reader to The details of the modeling of FLEXPART and FLEXPART WRF can be found in_inStohl et al. (2005) Stohl et al. (2005), (Pisso et al., (2019), and Brioude et al. (2013), respectively for detailed descriptions of the FLEXPART and FLEXPART-WRF models. HereIn this subsection, only the principles of FLEXPART Lagrangian modeling are discussed to facilitate the presentation of the results in the next section. FLEXPART was developed as a LPDM based on the zero acceleration scheme (Eq. 1). It solves a Langevin equation (Eq. 2) to model the trajectories of Lagrangian particles. The new position of the particles is influenced by large-scale winds, local turbulence (stochastic component), and mesoscale motions.
 - position of the particles is influenced by large-scale winds, focal turbulence (stochastic component), and mesoscale motion $X(t + \Delta t) = X(t) + v(X, t)\Delta t (1)$

$$\frac{dX}{dt} = v[X(t)]$$

where t is time, Δt is the time increment, X is the position vector, and $v = \bar{v} + v_t + v_m$ is the wind vector composed of the 130 grid-scale wind \bar{v} , the turbulent wind fluctuations v_t , and the mesoscale wind fluctuations v_m . FLEXPART also quantifies changes in the mass, or mixing ratio, of transported particles(Stohl and Thomson, 1999) carried away by calculating various removal processes processes(Stohl et al., 2005, Grythe et al., 2017) (Stohl and Thomson, 1999, Cassiani et al., 2015, Tipka et al., 2020). Turbulent motions v_{t_i} for wind components *i* are parameterized assuming a Markov process based on the Langevin equation (Eq. 2).

and wet deposition (Equation 5).

$$dv_{t_i} = \alpha_i(x, v_t, t)dt + b_{ij}(x, v_t, t)dW_j$$
(2)

<u>w</u>Where α is the drift term, *b* is the diffusion term, and dW_j are incremental components of a Wiener process with mean zero and variance dt, that are uncorrelated in time (Stohl et al., 2010). The minimum value of Δt_i is 1 second. Δt_i is used only for the horizontal turbulent wind components of Equation 3.

$$\Delta t_i = \frac{1}{ctl} min(\Delta \tau_{L_{\omega}}, \frac{h}{2\omega}, \frac{\frac{0.5}{\partial \sigma_{\omega}}}{\partial z}) (3)$$

- 140 where *h* is the height of the atmospheric boundary layer (ABL). To solve the Langevin equation for the vertical wind component, a shorter time step $\Delta t_{\omega} = \frac{\Delta t_l}{ifine}$ is used. Under convective conditions, when the turbulence is skewed, larger areas are occupied by downdrafts (instead of updrafts). This can lead to higher surface concentrations (deposition) in areas near pollution sources (Pisso et al., 2019). To investigate the sensitivity of radionuclide quantities to the turbulence scheme used, FLEXPART-WRF members were are re-run after replacing the standard Gaussian turbulence model (GTM) with the skewed 145 turbulence model (STM) (Luhar et al., 1996, Cassiani et al., 2015). The results of these two schemes are used to calculate the vertical velocity component of the drift term in Eq. 2. The implementation of STM requires shorter time steps, *dt* in Eq. 2, to better resolve turbulence in the convective planetary boundary layer. Therefore, we used–use *ctl*=10 and *ifine* = 10 as recommended by Brioude et al. (2013) and a tenfold finer time step, as recommended by Pisso et al. (2019), in sensitivity runs. For the computation of σ_{v_i} and $\Delta \tau_{L_i}$, FLEXPART uses the parameterization scheme proposed by Hanna (1982). To account 150 for mesoscale motions, a method similar to that of Maryon (1998) is used.
 - In <u>radionuclide</u> dispersion modeling, particle mass reduction of occurs primarily through three <u>processes</u>es<u>processes</u>: radioactive decay, dry deposition, and wet deposition. The following exponential equations characterize radioactive decay (Equation 4)

$$m(t + \Delta t) = m(t)exp(-\Delta t/\beta)$$
 (4)

155 where *m* is the particle mass and the time constant $\beta = \frac{T_{\frac{1}{2}}}{\ln(2)}$ is determined from already calculated the radionuclide half-life $T_{\frac{1}{2}}$.

$$m(t + \Delta t) = m(t)exp(-\Lambda \Delta t)$$
(5)

The scavenging coefficient Λ is calculated differently depending on whether the particles are in the aerosol or gas phase and whether the scavenging takes place inside or below the clouds (Stohl et al., 2005).

- 160 In this study, we have used use FLEXPART 10.4 and a modified version of FLEXPART 9.02 to input meteorological simulations from the Weather Research and Forecasting (WRF) model (Grell et al., 2005) (hereafter referred to as FLEXPART and FLEXPART-WRF, respectively). Compared to previous versions, FLEXPART has undergone a significant revision of the wet deposition calculation. The latest updates incorporate the dependence of wet deposition on aerosol particle size and precipitation type has been incorporated (Grythe et al., 2017). In addition, FLEXPART calculates wet deposition only for
- 165 cloudy <u>pixels grid cells</u> where the precipitation rate exceeds 0.01 mm h⁻¹. Therefore, the accuracy of cloud pixel detection plays a critical role in the accuracy of the location and amount of wet deposition simulated by FLEXPART. In previous versions, <u>(including version 9.02</u> used in the development of FLEXPART-WRF 3.2), in-cloud grid cells were are defined as <u>cells-those</u>-with <u>precipitation and</u> relative humidity above 80%.<u>-and anyThe</u> grid <u>cells</u> below <u>grid cells the in-cloud grid cells</u> up to the surface are defined as below-cloud grid cells beneath these grid cells as below-cloud ones. (Seibert and Arnold, 2013,
- 170 Pisso et al., 2019). Because of recent updates, In recent updates to the FLEXPART's source code, the above threshold has been modified using the 3D cloud water mixing ratio (q_c) fields. The threshold of q_c FLEXPART now uses 3D fields of cloud water mixing ratio > 0 ($q_c = 0$) now identifies grid cells within the cloud (below the cloud) (Pisso et al., 2019), and the relative fractions of ice and liquid water (type of precipitation). Therefore, the differences in approaches into calculating scavenging between FLEXPART and FLEXPART-WRF may_partially_it is expected that part of the explain difference the
- 175 inconsistency discrepancies between of between the FLEXPART and FLEXPART WRFresulting simulations in for the wet seasonis due to the differences in the schemes used to calculate the scavenging. Pisso et al. (2019) have discussed this and other updates in more detail in the latest version of FLEXPART. For dry deposition, for all particles below twice the reference height (*h_{ref}*), the reduction in particle mass is calculated using Eq. 6:

$$\Delta m(t) = m(t) \left[1 - exp\left(\frac{-v_d(h_{ref})\Delta t}{2h_{ref}}\right)\right] (6)$$

180 v_d is the dry deposition velocity calculated as the ratio of $v_d(z) = \frac{F_C}{C(z)}$ for h_{ref} of 15 m. F_C and $C(z) \in$ are the flux and concentration of a species at height z within the layer at constant flux. If the necessary information to parameterize of dry deposition of gases and particles is not available, v_d can be assumed to be constant. The concentration of a given gas or aerosol species in a grid cell is equal to the weighted average of the total particle mass within the grid cell divided by the volume of the grid cell (Pisso et al., 2019), as defined in Equation_7.

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 $C(z) = \frac{1}{v} \sum_{i}^{N} (m_i f_i)$ (7)

where V is the volume of the grid cell, m_i is the particle mass, N is the total number of particles, and f_i is the fraction (weight) of the mass of particle *i* assigned to the particular grid cell. The amount of dry and wet deposition for over the given grid cell i is accumulated by default over the time dimension of the output, unless the age composition of the air parcels is required, as was is the case in this study. A unique feature of FLEXPART/FLEXPART-WRF is the grouping of simulations based on the

- 190 age of the Lagrangian particles. This means that the number of concentrations and deposition at each time step is obtained from aggregating the simulations over the additional dimension of particle age. In this study, we used use the age spectrum of Lagrangian particles to estimate investigate the age composition the transport of radioactive materials from the source to the receptors. Air parcel ages are studied at an hourly resolution. As a result, the output grid, which has a horizontal resolution of 10 km and 14 vertical levels from 5 to 5000 m agl, gains an additional dimension with a length of 96 (hours). Thickness-
- 195 weighted averages of simulated concentrations, hereafter referred to as near-surface concentrations, are calculated from concentrations within the bottom four model layers between 5 and 100 m agl (with layer thicknesses of 5 m, 5 m, 40 m, and 50 m). To statistically compare the age distributions of the ensemble members, we use the maximum normalized difference according to Equation 8. Assuming that aa and bb are two of smooth density estimates of air parcel age distributions that have been smoothed with the Gaussian kernel (Chung, 2020), their maximum normalized difference is calculated as the maximum value of the absolute differences between a and b divided by the maximum value of a and b. (Chung, 2020)Larger variations

in the distributions are indicated by higher values of this indicator (Jin and Kozhevnikov, 2011).

maximum normalized difference $= \frac{\max(abs(a-b))}{\max(\max(a),\max(b))}$ (8)

Modeled concentrations are vertically integrated through all model levels in analyses related to air parcel ages (subsection 3.1) and inter-member evaluation (subsection 3.3). The thickness-weighted averages of simulations in the lowest four model levels between 5 and 100 m agl (with layer thicknesses of 5 m, 5 m, 40 m, and 50 m) are used for the spatial analysis (subsection 3.2).

2.2 Meteorological data

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FLEXPART and FLEXPART-WRF are drivenwork in the offline bymode with gridded meteorological fields from numerical weather prediction models on a grid scale. In this study, we obtain meteorological productsinputs from the NCEP elimate forecast system version 2 (CFSv2GFS (NCEP, 2015a)) (Saha et al., 2011, Saha et al., 2014), the NCEP final analysis (FNL) (NCEP, 2015b), and the ECMWF reanalysis fifth generation (ERA5) (Hersbach et al., 2020). The CFSv2GFS , which entered into the operational mode in March 2011, is a fully coupled elimate model that represents the interactions between oceans, land, and atmosphere. The CFSv2GFS can bedataset, available since 2015, used to provides 6-hourly forecast inputs at 03, 09, 15, and 21 hours for FLEXPART with a spatial resolution of 0.25 degreesonward. The FNL dataset is produced using the same baseie meteorological model that produces the GFS dataset. The former_provides the three-hourly combination of analysis (at 00, 06, 12, and 18) and forecast meteorological fields (at 03, 09, 15, and 21) at a spatial resolution of 0.25 degrees starting in 2015. ERA5 is the latest generation of ECMWF reanalysis data, covering the period from January 1, 1950 to nearly the present. It is They are generated in hourly time steps with a spatial resolution of about 0.25 degrees. CFSv2GFS and FNL are used to force the FLEXPART directly. In addition, WRF 4.2 is applied to dynamically downscale the ERA5 and FNL toso that they

220 can be used as the inputs to <u>FLEXPART</u>downscaled members. They <u>are are used</u>then fed into FLEXPART-WRF with <u>higher</u> downscaled the new spatial and temporal resolutions of 10 km and hourly and temporal resolution, respectively. <u>To the best</u>

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of our knowledge, this is the first time that the ERA5 reanalysis has been used with the to implement FLEXPART-WRF model

<u>setup.</u> Table 1 provides an overview of the <u>WRF</u> model <u>configuration</u>.Table 1 Overview of the WRF model configuration.

Dynamics	Non-hydrostatic				
Initial and boundary condition data	FNL/ERA5				
Temporal interval of boundary data	<u>3 h/1 h</u>				
Resolution	<u>10 km x 10 km</u>				
Extent of domain	17°N-33°N and 40°E-60°E				
grid-nudging	<u>On</u>				
PBL Scheme	YSU				
Cumulus parameterization	Grell 3D ensemble scheme				
Surface layer parameterization	Noah land surface scheme				
Terrain and land use data	<u>USGS</u>				

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Table 2 provides an overview of the input data and the corresponding simulation codes. For comparison of the above meteorological inputs with observations, daily total precipitation and daily averages of wind speed and temperature data are obtained from 157 climate stations within the model domain. These observations are freely available from the Global Surface Summary of the Day (GSOD) database which is maintained by the National Oceanic and Atmospheric Administration (NOAA).

230 (NOAA).

Table 24 Summary of meteorological inputs used to run FLEXPART and FLEXPART-WRF.

Inputs	Simulation code	Spatial	<u>T</u> tempora	<u>T</u> time	<u>D</u> downscale	<u>T</u> type of	<u>W</u> wet
		resolution	1	coverag	d	input	deposition
			resolutio	e			scheme
			n				
CFSv	FLEXPART_CFSv2	0.5 degrees	6-hourly	2011-	<u>X-</u>	forecast	Grythe et al.
2 <u>GFS</u>				present			(2017)
FNL	FLEXPART_FNL/FLEXP	0.25 degrees/	<u>3-hourly/</u>	2015-	<u></u> X-/√	analysis	Grythe et al.
	ART-WRF	10 km	hourly	present	_		(2017)/Seiber
							t and Arnold
							(2013)
ERA5	FLEXPART-WRF	10 km	hourly	1950-	\checkmark	reanalys	Seibert and
				present		is	Arnold
							(2013)

2.3 Emission scenario and study area

We have simulated asimulate fictitious nuclear accidentaccidents at the B-NPP (Fig. 1-A) in which 22 PBq (6.9 kg) of ¹³⁷Cs and 192 PBq (0.042 kg) of ¹³¹I are released during the first 24 h of each 96-hour simulation period. These amounts are the upper bounds of emissions estimated by Babukhina et al. (2016) for the Fukushima Daiichi nuclear accident in May-March 2011. Although Wwe simulated simulate a fictitious release of radioactivity at a level comparable to the Fukushima nuclear accident. However, our study does not intend to replicate past accidents or simulates a specific real-world case.; These source terms are used only because havingto provide a real-world comparison that gives the reader a tangible point of reference.reference from which to make sense of the results. rather we designed this analysis to determine what time of the day and year is associated with particular probabilities of transport of dense radionuclide plumes from any hypothetical release in the study area.

<u>The optimal-number of Lagrangian particles required for a given task is dictated</u> <u>byon the specific problem at handand</u> <u>needs</u>. To choose the optimal number of Lagrangian particles, we have followed the literature (Papagiannopoulos et al., 2020,

- 245 Thompson et al., 2015, Fast and Easter, 2006). The computational time scales increase linearly with the increasing number of particles, while the statistical error of the simulations decreases with the square root of the particle density (Pisso et al., 2019). In this study, aA total number of 10⁴ Lagrangian particles were are released almost uniformly during the emission period (~417 particles per hour). To assesstest if this rate-number of particles is sufficient, we to study the dry deposition process, which is directly influenced by the boundary layer conditions.² Wwe performed a preliminary test run with GFS data with a 10-fold
- 250 <u>increase</u> in <u>particles</u>. As shown in S1, the simulated <u>dry deposition of</u> ¹³⁷Cs and ¹³¹I and the <u>wet deposition of</u> ¹³⁷Cs do not undergo significant changes with the increase in <u>particles of an order of magnitude</u>.

<u>Regarding the height of the releases</u>, This result shows the sufficiency of the particle abundance to study the transport and deposition of particles near the surfaceTo choose the optimal number of particles, we have followed the literature (Papagiannopoulos et al., 2020, Thompson et al., 2015, Fast and Easter, 2006). The computational time scales linearly increase with the increasing number of particles, while the statistical error of simulations decreases with the square root of the particle density (Pisso et al., 2019). The particles are initially distributedemitted at height-model heights-levels between 100 and 300 m above ground level (agl) abovet the emission point. Each simulation starts at the beginning of each day and lasts 96 hours (Fig. 1-B). We This experiment has been run this scenarioed for every day of the year 2019, resulting in a total of 1460 simulation days (365 days of the year x 4 forward simulation days). In the following analysis, wWhen reporting particle ages associated with the <u>simulationsFor the diurnal and seasonal stratification of simulations</u>, we always refer to the time when a particle is released (not to when it travels or reaches receptors). The simulation domain <u>extendsis limited</u> between <u>17°N-33°N and 40°E-60°E</u>, and for post-processing the-we analyze <u>simulations geographical box</u>-over Qatar, between from 24.25°N-to-26.35°N and from 50.65°E-to 51.75°E, has been analysed.



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Figure 1 <u>A</u>: Study area encompassing the B-NPP (red square) and the <u>State of Qatar</u>. The base map and <u>overlay information are</u> from <u>Google Earth</u>. B: Schematic representation of the <u>LPDM simulation cycle</u>. The <u>original image isA is the study area embracing</u> the B-NPP (red square) and the state of Qatar. The base map and overlaying information are taken from <u>Google Earth</u>. B is the schematic illustration of the <u>LPDM simulation cycle</u>. The original figure is available at <u>https://www.janicke.de/en/lasat.html</u> (last accessed: <u>4 October</u>-February 1, <u>20222023</u>).

3 FLEXPART/FLEXPART-WRF simulationsResults and discussion:

3.1 Meteorological inputs vs. observations

275 <u>In this section, Wwe compare the relevant meteorological input fieldss including daily mean surface wind speed and temperature and daily total precipitation with observations from 157 elimate stations spanningwithin the model domain (Fig. 2).</u>

A relatively The-largerst discrepancy between observations and model inputs is observedattained for precipitation. 280 section, we compare the daily average of surface wind speed and temperature and total precipitation with observations measured in climate stations within the model domain... The highest inconsistency with observations was found in precipitation simulations (red dots in figure 1). The used datasets yielded The averaged Spearman correlation coefficient (r) of about ~0.41 and root mean square error (RMSE) of 3.5 indicate a moderate relationship between precipitation inputs and observations in all members a Spearman correlation coefficient of about 0.4 (with the highest in ERA5-WRF) and a RMSE of 3.5 (with the lowest in FNL dataset). The ERA5-WRF and FNL datasets have the highest correlation (0.44) and lowest RMSE (3) compared towith observations._All datasets, especially the downscaled ones, underestimate precipitation amounts, resulting in an averaged mean bias error (MBE) of -0.22<u>All datasets, especially downscaled ones, underestimated precipitation amounts.</u> This is to be expected since-Pprecipitation is the result of complex sub-grid-scale processes, most of which are parameterized in atmospheric models, and it suffers from the multiplicative effect of errors in thermodynamics and dynamics. This is to be expected as precipitation results from complex processes that are mostly parametrized in atmospheric models and it suffers the multiplicative effect of errors in both thermodynamics and dynamics. Furthermore, precipitation can be spatially inconsistent due to the complex topography of the study area<u>In addition, in contrast with the variables such as temperature and wind speed</u>, precipitation can be spatially patchy due to presence of complex topography (Tapiador et al., 2019).

295 All meteorological datasets are closer to the observations when evaluating the simulations of wind speed and temperature (S2). All datasets are found much closer to observations considering the simulations of wind speed (Fig. 1) and temperature (S1). The ERA5-WRF dataset shows the best agreement with the wind speed observations in terms of correlation (0.65) and RMSE (1.5), but .-However, this dataset underestimates theunderestimates -wind speed more to a greater extent than the GFS and FNL datasets. All datasets are significantlywell correlated with the temperature observations and with each other, as shown by the distribution of data points along the identity line. The ERA5 WRF dataset yielded the highest agreement with wind speed observations, in terms of correlation (0.65) and RMSE (1.5). However, it has underestimated the wind speed more than GFS and FNL datasets. Regarding the air temperature, as expected, all datasets are well correlated with observations and with each other indicated by a fair distribution of the data points along the unity line. From these above results, weit can be concluded that the downscaled datasets have relatively better performance in terms of correlate betterion with observations.

IThe intercomparison of the input datasets shows that the FNL and GFS simulations have the highestbest agreement, as can be expected since they are produced using the same data assimilation and forecasting system. According to the above results, it can be concluded that the downscaled datasets have a relatively better performance in terms of correlation with observations.
310 However, the increase in their association with the observations has come at the price of increase in the error levels, especially in wind speed simulations. However, we keep both datasets for the ensemble analysis because the smallest difference in the meteorological input leads to sequential deviations in the transport of air pollutants, as will be shown later. It is worth to notice that the mere comparison of meteorological inputs against surface observations is not adequate to identify the best meteorological dataset for the modeling of radionuclide dispersion (and it is not the purpose of this study either). We only have access to daily meteorological data at surface at a few points whereas transport modeling is carried out in several layers of the atmosphere in hourly time step throughout the study area. In addition The inter comparison of datasets shows that FNL and GFS simulations have the highest agreement because they are products from the same data assimilation and forecast system. However, we decided to use both datasets for the ensemble analysis on the grounds that subtle differences in meteorological

inputs can have additive impacts on transport and distribution of air pollutants, discussed later. The relatively higherbetter
 agreement found between the ERA5-WRF and FNL-WRF datasets than between the latter and the FNL dataset shows the impact of the downscaling on the homogeneity of the meteorological inputs. The relatively higher agreement of ERA5-WRF with FNL WRF dataset than with FNL dataset indicated the increased homogeneity of two global scale scale datasets due to dvnamical downscaling. It should be noted, although this is not the goal of this study, that a systematicsimple comparison of meteorological datasets with sporadica representative sample of surface observations is not-neededsufficient to determine the optimal choice ofbest meteorological inputs for forcing a CTM for different regions. In addition, Here we have access to daily surface meteorological data, while transport modeling is performed in several layers of the atmosphere at hourly timesteps or finershorter, in cases where the forecast of radionuclide dispersion is of interest, not many forecast datasets, like GFS, are left to choose for making a forecast run.



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Figure 2 <u>Scatter plot of observed daily total precipitation (mm, shown as blue circles) and daily mean wind speed (m/s, shown as red squares)</u> against <u>FLEXPART/FLEXPART WRF</u> inputs. The scatter plot of observed and simulated <u>daily surface temperature (k)</u> is shown in <u>S2</u>. The observations are obtained from 157 climate stations in the study area<u>Scatter plot between observed daily total</u> precipitation (mm, shown in red squres) and <u>FLEXPART/FLEXPART/FLEXPART</u>.

335 WRF inputs. The scatter plot of daily surface temperature (k) is shown in S1. Observations are obtained from 157 climate stations in the study area.

3.21 Age composition of radionuclide plumes

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It should be noted that due to high solubility and the relatively long half-life of ¹²⁷Cs, further emphasis is placed on the spatiotemporal distribution of ¹²⁷Cs deposition. Conversely, the ambient surface-level concentration of ¹²¹I over the study area will be considered since it resides mostly in the gaseous phase and has a short half-life of 8 days.

In order to investigate the diurnal and seasonal variations in the transport of radionuclides, we examined the simulated age of the Lagrangian particles entering the study area by the season and time of day in which they were released from B-NPP. The seasonal distributions of <u>the</u> air parcel ages for ¹³¹I and ¹³⁷Cs are shown in Figures 3 and S3, respectively. The <u>high degree of significant</u> similarity <u>betweenof</u> the age distributions of ¹³¹I and ¹³⁷Cs <u>is</u> due to the fact that the <u>removal process</u> in LPDMs

- 345 only affects mass concentrations and does not affect particle positions, but masses, indicates that differences in the transport characteristics of these radionuclides, such as the wet and dry deposition rate and radioactive decay, are not large enough to cause the abundance of cases where ¹³¹Land ¹³⁷Cs particles are not present in a common grid. The close dispersion of ¹³¹L and ¹³²Cs concentrations, especially at low intensities, and, consequently, the age of corresponding air parcels can be attributed to the lack of heavy precipitation in the region. This can be seen in S2 where a positively skewed distribution is found for the
- 350 ratio of ¹³³Cs wet deposition to total ¹³⁷Cs deposition. Therefore, the following discussion of the ages of the ¹³¹I particles is also valid for the <u>ages</u> of the ¹³⁷Cs concentrations particles (shown in the supplement). All ensemble members, in particular <u>especially</u> GFS and FNL, in fall, in all seasons simulate predominantlyan abrupt increase (or a peak) in the number of air parcels with lifetimes of less than 20 to 30 hours-in the autumn seasonin all seasons. This indicates that, in most cases, the regardless of the time of year, radionuclide clouds begin to cross the boundaries of the study area within a few hours after the
- 355 emission. Other thanExcept for fall, where the GFS and FNL have a stronger peak in the above time interval, the age distribution of the particles is relatively similar in all other seasons. As will be discussed in the subsection 3.2_s, the frequency of radionuclide transfer to the study area in all ensemble members is much lower in fall than in winter and spring. Therefore, we believe that a seasonal nunusual atmospheric pattern simulated in the GFS and FNL datasets is responsible for the striking difference in particle age distributions in fall. Compared to the other ensemble members, the FNL-based simulations show
- 360 <u>athe most</u> delayed appearance of the second peak in air particle ages. This could be due to 4-a decrease in transport speed to the same receptors and/or 2-a higher <u>numberfrequency</u> of air parcels reaching areas away from the source (discussed later). The maximum normalized distance of age distributions (Fig. 3) shows larger similarity between FNL WRF and ERA5-WRF than between the former and FNL in all seasons other than the fall (0.3, 0.19, 0.25, and 0.38 vs. 0.3, 0.32, 0.32, and 0.41 in fall, spring, summer, winter). This may to be due to the use of meteorological inputs with the same spatio temporal resolution
- 365 and a common simulation code and, consequently, similar modeling schemes for the two former members. Although the base model used for the production of FNL, the Global Forecast System (GFS), is also the atmospheric component of CFSv2, the distribution differences were found to be lower between the FNL age distribution and those from ERA5– and FNL WRF for

most seasons. In addition to having finer spatial resolution than CFSv2, FNL assimilates observations like ERA5. Compared to other members, air parcel ages are distributed in a wider range in all seasons in FNL (note the location of the first and last
 peaks). This could be because: FNL simulated 1 – a decrease in transport speed to the same receptors and/or 2 – a higher frequency of air parcels reaching further receptors from the source (discussed later).

To understand how-the radionuclide concentrations transport-may vary with transport timeswith the level of radionuclide concentrations, beside the previous categorization, air parcel ages were are divided into three groupslevels representing corresponding to low (the top row in S4-A), moderate (the bottom row in S4-A), and high (the bottom row in Fig. 3-A) concentrations near-surface concentrations of ¹³¹I (see S3 and S5 for ¹³⁷Cs). These categories are determined for each member based on on column (mass) densities, computed from the vertical integration of concentrations. The low, moderate, and high concentrations values below the 33rd percentile, between the 33rd and 66th percentiles, and above the 66th percentile, are respectively. The short time interval required for therapid transport of dense radionuclide clouds within the region of interest to the study area is demonstrated by the strongly positively skewed age distributions of particles corresponding to moderate and high radionuclide concentrations in all members and all seasons. Given the proximity of the source (B-NPP) to the southeast of the study area, these results are expected. The variation dissimilarity of the-age distributions for high and moderate concentrations in summer can be attributed to the low frequency of radionuclide transfer to Qatar and to an unusual transport pattern in this season (as what observed for the air parcel ages corresponding to all concentrations in fall).

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It is also-worth noting that all members of the ensemble-model (including the forecast member) have a greater agreement for air parcel ages corresponding to the moderate and high levels of ¹³¹I and ¹³⁷Cs concentrations than for those corresponding to low concentrations<u>concentrations</u>. This is consistent with the principles of FLEXPART dispersion modeling. According to Pisso et al. (2019), the error rate of the simulations decreases with the square root of the particle density. As a result, all four members simulate the age of Lagrangian particles corresponding to moderate and high concentrations in better agreementmore closely and, most likely₇ more accurately near the source. The age distributions corresponding to the low concentrations are almost the same as those corresponding to all concentrations. Using the maximum normalized distance, athe quantitative comparison of the particle age distributions, corresponding to all concentrations, is shown in Figure 4. The age distribution in the ERA5-WRF- and FNL-WRF-based simulations shows a greater similarity than to that of the FNL-based simulations in all seasons, except fall. This is due to the use of dynamically downscaled meteorological inputs with the same spatio-temporal resolution-and to a common simulation code for ERA5 and FNL-WRF. The age distributions based on the FNL and GFS

inputs, which are generated by a similar base model, do not show the same degree of similarity in all seasons. They have the smallest difference in fall and spring and the greatera larger differencebut not in winter and summer, where the distance between the age distributions based oncontrary to the intercomparing FNL and FNL-WRF-datasets is the smallest. Figure 2 B
 (S1-B) shows the categorization of age distributions based on the release time of particles carrying. ¹³¹L (¹³²Cs) in the first 24

hours of each simulation period. This panel is similar to 2-A, but shows the age distributions plotted separately for particles released in 6-hour periods of the first day of the simulations. For example, the upper left figure shows the age distribution of the particles that is released within the first 6 hours of simulations. This new categorization reveals that the number of longlived air parcels (including all intensities of concentrations) increased in all members when particles are released between 12 405 and 6 p. m. (top row in Fig. 2 B and S1 B). This temporal pattern may be due to the coincidence of particle release and the development of the planetary boundary layer in the afternoon that caused the transport of radionuclides to longer distances, represented by longer Lagrangian particle ages. Because particles that are released at the end of the day have less time to travel by the end of simulation period, a sharper fall is observed in the right tail of the age distributions during the second half of the day. However, the lifetime of most simulated particles, especially of those that caused moderate (bottom row in Fig. S3-B and 410 S4-B) and high intensities (bottom row in Fig. 2-B and S1-B), is found to be between 20 and 70 hours after release. Consequently, the difference in release time of Lagrangian particles is not significantly affected by their age spectrum. Compared to ERA5 WRF and FNL WRF. CFSv2 and FNL have simulated a larger number of shorter aged air parcels after sunset (between 6 and 11 p.m.). It is because the spatio temporal resolution of inputs used in these two members is too coarse to resolve the gradual reduction of boundary layer height resulting in fewer particles being transported to remote areas.



distribution, the diurnal variations of air parcel ages for high concentrations are very similar in all members. They all simulated a higher number of shorter-lived parcels when they are released in the first half of the day. Although a firm conclusion would require further studies at very high resolution of land sea circulations in the region, onshore winds passing the emission point from the south may cause the abundance of shorter-lived parcels in southern Qatar.

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Figure 3 A: Smoothed age Ddistributions of the air parcels corresponding to all ¹³¹I intensities <u>concentrations</u> of ¹³¹I column densities</sup> (top row) and of those above the 66th percentile (bottom row). B: the same as A, but for four times of the day. The curves represent simulations from each member of the ensemble, y-and x-axes show density values and air parcel ages, respectively.



425 Figure 4 Seasonal maximum normalized difference of air parcel age distributions.

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Due to the relatively long half-life of ¹³⁷Cs, its deposition rate in the affected areas is of great importance. To analyze the relationship between the age composition of air parcels and the amount of ¹³⁷Cs deposition, the deposition values cumulatively aggregated across time steps (jj) and age spectr<u>a.e.(ii)</u> are normalized to the total amount of ¹³⁷Cs deposition simulated in each grid cell (k) at the end of_-each simulation run (l).

$$\frac{137}{\text{CS}_{klna_{norm_depso}}} = \begin{cases} \frac{\sum_{j=1}^{k} \sum_{j=1}^{k} \sum_{l=1}^{k} \sum_{j=1}^{k} \sum_{l=1}^{k} \sum_{l=1}^{k}$$

435 where n is the time step with a maximum of 96 (the last time step) and a is the given particle age with a maximum of n-1. ${}^{137}Cs_{klij} = 0$ in two conditions (1) $n \ge 26$ if $i \in [1, n - 25]$ and (2) $i \ge j$.

Figure 5 shows the normalized deposition amounts ($^{137}Cs_{klnta_{norm_depso}}$) in winter, when both dry and wet deposition occur in the study area. It is worth noting that Similar deposition patterns are obtained seen form other seasons (S6). As shown in S7, the main reason for the small difference between the seasonal deposition patterns is the lack of precipitation and subsequent wet deposition in the region. Although the spatial pattern of the deposition varies considerably, as indicated by the range of quartiles, the median of the normalized deposition shows that about 80 percent of the deposition occurs within 80 hours after <u>anthe hypothetical</u> accidents. The <u>fall in the</u> cumulative deposition at the end of the simulation period is <u>mostly inrelated to</u> the areas <u>farthestar</u> from the source, <u>andbut the total</u> deposition reaches 100% as it approaches the end of the 96-hour simulation period. This <u>is clearly evidentcan be better understood</u> in S8, where the ages of the <u>deposited</u> particles simulating the ¹³⁷Cs deposition peak around 20-30 hours, <u>withand</u> a rather small number of them are older thanafter 80 hours.



Figure 5 <u>T</u>the normalized cumulative.¹³⁷Cs deposition to the total <u>¹³⁷Cs deposition</u> of <u>amount</u> each 96-hour simulation period at each grid cell <u>in winter</u>. This Results for <u>other seasons are</u> shown <u>in S6</u>, <u>plot (S5) shows simulations in winter (other seasons)</u>. The x-axis shows the age of the Lagrangian particles and the y-axis is the normalized deposition. The boxes show the quartiles of the normalized deposition. The whiskers represent the range between 1.5 times the interquartile range above the upper quartile and below the lower

450 deposition. The whiskers represent the range between 1.5 times the interquartile range above the upper quartile and below the lowe quartile.

3.2 Spatio-temporal distribution of radionuclides

For the analysis of near-surface radionuclide concentrations, the heightthickness-weighted average of the simulations in the bottom four layers of the model between 5 and 100 meters agl has been is used. Because of the high radioactivity of ¹³¹I and the solubility and relatively long half-life of ¹³⁷Cs, we focus in this part on ¹³¹I concentrations and ¹³⁷Cs deposition. In view of 455 the serious health risks, including thyroid cancer, that the exposure to ¹³¹I radiation may cause (National Research Council, 1999), Using the model proposed by WHO (2012) for internal dose from inhalation, the seasonal median of the 96-hour integrated ¹³¹I concentrations (¹³¹I^{intg_conc_seas}, (in units of Bq m⁻³) is converted to the thyroid internal dose from inhalation effective dose from inhalation of ¹³¹I (TIDI, in units of µSv) (WHO, 2012),. The model conversion coefficients inputs are as 460 specified defined by by WHO (2012) for three age groups: 1 year-old-infants, 10 year-old children, and adults (11 years and older). Considering thatSince approximately 90% of the population of Qatar is in the adult age group (UNStats, 2020), the inhalation doses calculated for this age group are discussed here (Fig. 6) and those for two others are available in the supplement (Figures S7 and S8). As reference values for comparison, TIDI values between about 2000 and 20000 µSv are measured in the first year following the Fukushima accident in areas close to the emission point (see Table 4 in WHO (2012)). The ensemble 465 simulations show that the largest advance of TIDI above 2500 µSv into the study area occurs in the cold period of the year, especially in fall, in the simulations of all members. This may be due in part to the lower (higher) PBLH in the cold (warm) seasons and to the synoptic patterns that bring the radionuclides into the study area. An exception to this rule, however, is the distribution of TIDI in the FNL based simulations during the summer. It appears that an unusual weather pattern caused abnormal TIDI values to occur in central Qatar during the summer. This may change as the modeling period is extended. In 470 terms of spatial distribution, the TIDI above 2500 µSv occur close to the source in the southeastern part of the study area, and the intensity of the TIDI decreases with increasing distance to the north of Qatar. In other words, southeastern Qatar is the first area to be affected by dense ¹³¹I clouds in the event of a a nuclear accident, especially during the cold period of the year. The comparison of the TIDI simulations between the members of the ensemble shows that, despite the similarities mentioned above, there are also clear differences. The advance of TIDI above 2500 µSv to the southeast of Qatar in simulations based on 475 FNL inputs occurs in both fall and winter, but is observed only in winter in the GFS-based simulations. TIDI values peak in the fall in both downscaled runs, with inputs from the ERA5- and FNL-WRF datasets, but the extent of high TIDI values to the southeast of Oatar in fall is much larger in the ERA5-WRF run. This is also true when comparing the FNL-WRF and FNLbased simulations. The above mentioned differences between the members have resulted in TIDI values varying by a factor of 2 to 10 in the south of Qatar. To investigate the influence of the particle release time on the radionuclide dispersion, the 480 seasonal median of the particle release time (hours in local time (LT)) coinciding with the maximum concentration of ¹³¹I and the completion of ¹³⁷Cs deposition is considered (contours in Figures 6 and 7). To identify the highest possible level of pollution at each point, regardless of its frequency, local maxima are calculated only from non-zero intensities. The results show that the highest ¹³¹I concentrations is coincide with particles released between 9 a.m. and 2 p.m. LT in most parts of the study area.

This is the time of day when the development of the planetary boundary layer, the intensification of the land-sea thermal

485 gradient, and the resulting daytime onshore winds coincide in the region. Among the members, the earliest particle release times (between 6 and 9 a.m. LT) leading to the highest ¹³¹I concentrations are observed in simulations based on ERA5-WRF data. On the contraryIn contrast, it can be seen that the particles released after 6 p.m. LT lead to the highest concentrations in the vicinity of B-NPP in the southeast of Qatar in the simulations based on GFS and FNL-datasets, especially in fall. This can be attributed to the decrease in boundary layer height at this time of day, which leads to an increase in radionuclide concentrations in areas close to the accident site.

Figure 7 shows the seasonal median of total ¹³⁷Cs deposition (¹³⁷Cs^{tot_depos_seas}) simulated by four ensemble members. As defined by the International Atomic Energy Agency (IAEA, 2009), any area covered by radioactive substances that emit beta particles and gamma rays, such as ¹³¹I and ¹³⁷C, in quantities greater than 40 kBqm² is considered to be "contaminated". Accordingly, 495 allAccordingly, all members have simulated deposition above this threshold over a significant portion of the study area during the cold period of the year. As with TIDL, tThe highest ¹³⁷Cs^{tot_depos_seas} are observed in the south and southeast.-of Oatar near the emission point. Compared to the FNL- and GFS-based simulations, the simulations based on the FNL- and ERA5-WRF datasets show the much greater extent of the contamination of the ¹³⁷C deposition over almost the whole of Qatar in the winter and, to a lesser extent, in the fall. Among the ensemble members, the highest ¹³⁷Cs^{tol_depos_seas} in the southeast-of Oatar are seen 500 in the simulations based on ERA5-WRF inputs, followed by the FNL-WRF-based simulations, in fall, While the ERA5-WRFbased simulations of ¹³⁷Cs^{tot_depos_seas} exceed 300 kBqm², the deposition in GFS-based simulations is are up to ten times lower at the same location and same period. The ¹³⁷Cs^{tot_depos_seas} in the warm period of the year, except for the simulations based on the FNL dataset in summer, are mostly either close to or below the threshold. The higher ¹³⁷Cstot_depos_seas in the cold period of the year, especially in winter, can be largely attributed to the seasonal increase of ¹³⁷Cs transport by northerly winds, discussed 505 later, and to the relative increase in wet deposition (S7). The above results indicate that if a nuclear accident occurs during the cold period of the year, the magnitude of extreme ¹³¹I concentrations and total ¹³⁷Cs deposition in the south of Qatar may be up to 10 times greater than during the warm seasons. A similar order of magnitude-is seen in simulated variability is seen across the ensemble members. With respect to the particle release times, ¹³⁷Cs^{tot_depos_seas} occur mainly when particles are released between 10 a.m. and 5 p.m. LT, in the presence of a turbulent boundary layer. The analysis of the age spectra in subsection 3.2 shows the higher frequency of long-lived air parcels in the FNL-based simulations (see top row in Figure 3). 510

In view of the serious health risks, including thyroid cancer, thatwhich can be caused by the exposure to ¹³¹I radiation-may cause (National Research Council, 1999), the seasonal median of the 96-hour integrated ¹³¹I concentrations (¹³¹I^{intg_conc_seas}, in units of Bq m⁻³) is converted to the thyroid internal dose from inhalation (TIDI, in units of µSv). We use the coefficients
 defined by WHO (2012) to calculate dosages for three age groups: 1-year-old infants, 10-year-old children, and adults (11)

years and older). Since approximately 90% of the population of Qatar is in the adult age group (UNStats, 2020), the inhalation doses calculated for this age group are discussed here (Fig. 6). As reference values for comparison, TIDI values between about

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2,000 and 20,000 µSv weare measured in the first year following the Fukushima accident in areas close to the power plant (see Table 4 in WHO (2012)). Our ensemble simulations show that the largest -TIDI above 2500 µSv -occurs in the cold period of

- 520 the year, in fall, in the simulations of all members. This may be due in part to the lower (higher) PBLH in the cold (warm) seasons and to the synoptic conditions. An exception- is the distribution of TIDI in the FNL-based simulations during the summer. This may change as the modeling period is extended. In terms of spatial distribution, the TIDI above 2500 µSv occur close to the source in the southeastern part of the domain, and the intensity of the TIDI decreases with distance to the north. Southeastern Qatar is the first area to be affected by dense ¹³¹I clouds in the event of a nuclear accident, especially during the
- 525 cold period of the year. The advance of TIDI above 2500 µSv to the southeast of Qatar in simulations based on FNL inputs occurs in both fall and winter, but is observed only in winter in the GFS-based simulations. TIDI values peak in the fall in both downscaled runs, with inputs from the ERA5- and FNL-WRF datasets, but the extent of high TIDI values to the southeast of Qatar in fall is much larger in the ERA5-WRF run. This is also true the case when comparing the FNL-WRF and FNL-based simulations. The differences- have resulted in TIDI values varying by a factor of 2 to 10 in the south of the area of interest
- 530 between ensemble members.

To further examine these results, the spatial distribution of the full-year median of the 96-hour integrated ¹³¹I concentrations, converted to TIDI, and the full-year median of the air parcel ages coinciding with the maximum concentration of 131 are shown in Figure 8. As expected, the age of the Lagrangian particles decreases southward with the decreasing distance from proximity 535 to the source. A we see that all ensemble members simulate the age of particles ages loweress than 25 hours along with relatively high levels of TIDI (131 concentrations) at the southeastern edgecorner of the study area. The longer-lived particles (above 35 hours) are found in northern latitudes Qatar. In simulations based on the FNL dataset, particle ages in northern Qatar exceed 40 to 50 hours, while <u>falling below they do not exceed</u> 40 hours in simulations from other ensemble members. Furthermore, we observe obtain the larger extent of relatively higher TIDI to central Qatar in these simulations compared to simulations from other ensemble members (thesee both seasonal and full-yearly median of TIDI is shown in Figures 6 and 8).

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Figure 6 the color scale shows the Sseasonal median of 96-hour integrated ¹³¹I concentrations (¹³¹I^{intg_conc_seas}) converted to thyroid internal dose from inhalation (TIDI, in units of μSv) for the adult age group. The contour lines (in <u>local time</u>, the unit of hours of the dayin local time) depictare the seasonal median of the Lagrangian particle release time coinciding with the maximum ⁵⁴⁵
 ¹³¹I^{intg_conc_seas} found in each 96-hour run.



Figure 7 Same as 6, but for the seasonal median total 137 Cs deposition (137 Cs^{tot_depos_seas}) in the units of Bqm⁻². The contour lines are the seasonal median Lagrangian particle release time coinciding with the completion of 137 Cs deposition found in each 96-hour run.

550 IFrom the above two points, it can be concluded that a higher numberfrequency of air parcels reaching areas away from the source (and not a lower decrease in transport speed speed to the same receptors) which seems to causeleads to the abundance of longer-lived particles in FNL-based simulations. Because of the high importance of the population being exposed to radionuclides, TIDI and ¹³⁷Cs^{tot_depos_seas} are analyzed against the population density of Qatar (Fig. 9). The spatial distribution of the population shows that tThe desert areas of southern and southeastern Qatar, over whichas the area of entry of 555 radionuclides enneter in the countryregion, host a small number of people or are almost uninhabited (Fig. 9-A). Figures 9-B and C show that the extremely high levels of TIDI (greater than 10000 µSv) and ¹³⁷Cs^{tot_depos_seas} (greater than 150 kBqm²) occur mostly in areas with a population density of less than five persons per arc-second. In the populated areas (with a density of more than 8 persons per arc second) the TIDI and ¹³⁷Cs^{tot_depos_seas} do not exceed more than 5000 µSv and 100 kBqm²₄ in most cases. Due to the exceptional weather pattern that occurs in the simulations based on the FNL dataset in the eastern part 560 of Qatar (where the most densely populated areas are located) in summer, the highest values of TIDI in densely populated areas are observed-in these simulations. Otherwise, all ensemble members simulate the highest TIDI during the cold seasons. This is agreement with the previous results. The highest ¹³⁷Cs^{tot_depos_seas} in the same areas are observed in the simulations of



565 Figure 8 A: **#**The full-year median TIDI (μSv) for the adult age group and B: The full-year median ¹³⁷Cs^{tot_depos_seas} (kBqm⁻²). The contour lines are the full-year median age spectra coinciding with the maximum ¹³¹I^{intg_conc_seas} and the completion of ¹³⁷Cs deposition found in each 96-hour run.

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Figure 9 <u>A: Tthe subplot A shows the gridded population density of Qatar at one arc-second resolution in 2020. The relationship 570 between TIDI (μSv) for the adult age group and ¹³⁷Cs^{tot_depos_seas} and the population density of Qatar are shown in B and C, respectively. The shapes and colors of the markers represent seasons and ensemble members.</u>

In addition to the magnitude of the transported radioactive materials, the temporal distribution of the extreme events in the affected area is also of importance for the preparedness programs. Figure 10-A shows the frequency of occurrence (FoO) of 575 ¹³¹I concentrations above the 66th percentile. In all the members of the ensemble, more than half of the- eventsevents in most parts of Oatar, especially in the northern part, take place in the winter. While the FoO of high ¹³¹I varies between around 15 and 30% in spring, the lowest FoO is observed in the summer when less than about 10% of the extreme events occur-in Qatar15% of extreme events have reached receptors in Qatar. The seasonal FoO of 137Cs column densities concentrations (S9) also show highly very similars almost identical results. A case where ERA5-WRF simulated the northwestward movement of the near-surface ¹³¹I concentrations (Bq/m3) inon 14 January at noon is shown in the upper panel of Figure 10-B. We have 580 observedfind a similar pattern in a large number of events in which high levels of radionuclides are transported to Qatar. This synoptic pattern seems to be due is related to the juxtaposition of low and high-pressure cells located to the west and east of the region. The resulting pressure gradient cause strong south/southeasterly winds to develop between two cyclonic and anticyclonic cells, bringing the dense ¹³¹I clouds into the study area. This pattern occurs mainly occurs in the late winter and early 585 spring, coinciding with the southward movement of the westerlies and the eastward movement of the Saudi Arabian subtropical

high pressure system occurs mainly in the late winter early spring period simultaneous with the southward movement of westerlies and the eastward movement of the Saudi Arabian subtropical high pressure (De Vries et al., 2016). The lower panel of Fig. 10-B shows the summertime mean of near-surface ¹³¹I concentrations (Bq/m3), obtained from simulations based on the ERA5-WRF database. The near-surface atmospheric circulation is superimposed on these simulations. The seasonal pattern

590 found here illustrates well why very few extreme events are observed in the study area during the summer. The northwest-southeast winds, known as the Shamal winds (Yu et al., 2016), cause the simulated radionuclides to move away from the study area. Therefore, the highest (lowest) contamination risk to the population of Qatar from the occurrence of a nuclear accident and the resulting release of radionuclides is expected to occur in winter (summer).



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Figure 10 A: Frequency of occurrence (%) of ¹³¹I column densities <u>concentrations</u> above the respective 66th percentile. B: <u>The figure</u> <u>at the top shows</u> the simulation of <u>near-surface¹³¹I concentrations</u> based on the <u>ERA5-WRF</u> dataset on January 14, 2019<u>. The <u>r</u> at <u>noon (top)</u> and the figure at the bottom is the summer average of near-surface ¹³¹I concentrations</u>. The wind field at 12:00 on the first day of the 96-hour simulation period is used to represent the atmospheric circulation during the radionuclide transport.

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Concerning With respect to the sensitivity of the radionuclide simulations to the choice of turbulence scheme, <u>ourthe</u> results shows that the use of STM leads to a more frequent occurrence of high $\frac{131}{131}$ concentrations and $\frac{137}{131}$ concentrations within the

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study area (Fig. 11). The difference between the seasonal (and full-year) median of the simulations in the STM- and GTMbased concentrations is trivialnot noticeable. However, the upper quartiles of the simulations, especially in the cold seasons, show a significant increase after applying the STM scheme. The upper quartiles of the STM-based ¹³¹I concentrations are 15.5, 16, and 21.2% higher than those of the GTM-based simulations in winter, fall, and year-round, respectively. The quartiles of STM-based deposition simulations also increase by 20, 20, and 12% over the same periods in comparison to GTM-based deposition simulations. The application of STM has comparable effects on concentration and deposition simulations based on the FNL-WRF dataset (shown in S10). The increase in ¹³¹I concentrations and ¹³⁷Cs deposition values <u>can perhaps beis</u> expected because under the skewed turbulence condition, downdrafts are more frequent than updrafts, resulting in higher surface concentrations and deposition in areas near pollution sources Pisso et al. (2019).

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3.3 Inter-comparison of FLEXPART and FLEXPART-WRF runs (effect of downscaling)

620 Radionuclide simulations based on (re)analysis datasets are assumed to be a better approximation of the actual situation <u>atmospheric conditions</u> than those based on the forecasting dataset (GFS, in this study) (Leadbetter et al., 2022). The <u>evaluation</u> statistical <u>evaluation</u> metrics used here are recommended by Maurer et al. (2018), who <u>searched for the bestassessed</u> transport models to simulatinge Xe-133 using measurements from six International Monitoring System (IMS) stations: <u>Among other</u> <u>evaluation metrics</u>, they used the fractional bias (FB) and the fraction within a factor of 5 (F5). FB, in the range -2 and 2, is

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- 625 the bias of the <u>simulation simulated</u> mean <u>values</u> normalized <u>by the sum of the simulation and measurement means by the sum</u> of the two means and multiplied by 2. F5 is the fraction of simulations that are at most one factor larger (5) or smaller (0.2) than the reference values. The Spearman correlation coefficient (r) calculated between the simulations of ¹³¹I column densities <u>concentrations</u> shows that the <u>CFSv2-simulationGFS-based simulations</u> are <u>well-closely</u> associated with those of the (re)analysis-based members (red circles in the bottom row in Figure 12). The <u>CFSv2GFS</u>-based simulations <u>have-attain</u> the
- 630 highest correlation with the FNL-based simulations (0.85) followed by FNL-WRF-based simulations (0.73). While the CFSv2 simulation<u>GFS-based simulations</u> have (on average) a <u>smalltrivial</u> FB (0.07) compared to the FNL<u>-based</u> simulations, they are positively biased compared to the simulations based on the FNL-WRF (-0.99) and ERA5-WRF (-0.98) datasets (note that the placement of the (re)analysis members on the x-axis causes the FB to have a negative sign). According to FB5F5, 74.06% and 55.69% of the CFSv2 simulationGFS-based simulations are respectively within a factor of 5 of the FNL and FNL-WRF-based
- 635 simulations, respectively, whereas F5 decreases to 54.72% between the GFS- and ERA5-WRF-based simulations. The RMSE between simulations based on the CFS+2GFS and FNL datasets (68443.95 Bq/m³) is smaller than that found between the former and the FNL- (108001.33 Bq/m³) and WRF-based (110632.06 Bq/m³) simulations. As for the otherwith the previous metrics, the CFS+2 simulationGFS-based simulations produce the lowest and highest NMSE, the normalized RMSE which is less sensitive to extreme values, against the FNL (1.24) and ERA5-WRF (8.78) simulations, respectively. In short, the GFS-based simulations is produced by the the text of text of text of text of text of text of the text of the text of tex of text of te
- 640 and FNL-based simulations have the highest agreement due to the <u>largesignificant</u> similarities between their meteorological inputs (see subsection 3.1). Among all ensemble members, the largest difference is occurs between simulations based on ERA5-WRF and FNL datasets (r=0.55, FB=-0.92, FB5F5=50.6%, RMSE=120282.62 Bq/m³, and NMSE=11.18). In <u>contrastOn the other hand</u>, the downscaling of the inputs and the application of the same simulation code lead to ayield closer higher agreement of the simulations based on the FNL-WRF and ERA5-WRF datasets than to those based on the FNL and FNL and FNL and FNL applies and the application of the same simulation code lead to ayield closer higher agreement of the simulations based on the FNL-WRF and ERA5-WRF datasets than to those based on the FNL and FNL applies applies and the application of the same simulation code lead to ayield closer higher agreement of the simulations based on the FNL-WRF and ERA5-WRF datasets than to those based on the FNL and FNL applies applies applies applies and the application of the same simulation code lead to ayield closer higher agreement of the simulations based on the FNL-WRF and ERA5-WRF datasets than to those based on the FNL and FNL applies appli

645 FNL-WRF datasets (r=0.7 vs. 0.67, FB=0.01 vs. -0.93, FB5F5=69.83% vs. 53.3%, RMSE=36012.82 Bq/m3 vs. 113888.80 Bq/m3, and NMSE=2.74 vs. 10.08).

The evaluation metrics <u>ofealculated between the simulations of</u> ¹³⁷Cs column densities <u>concentrations</u> show-lead to the same results as discussed above. For example, the simulations based on the GFS dataset <u>have show</u> the highest agreement with the 650 FNL-based simulations (r=0.85, FB=0.07, FB5F5=74.98%, RMSE= 8485.22 Bq/m³, and NMSE=1.19). Similarly, simulations based on FNL and ERA5-WRF datasets and simulations based on FNL-WRF and ERA5-WRF datasets have the lowest and highest agreement among the (re)analysis-based members. According to the unitless For all metrics, including r (with an exception between GFS and ERA5 WRF based simulations), FB, F5 and NMSE, the GFS based simulations have leads to <u>closea higher</u> agreement with those based on the (re)analysis data sets when comparing the ¹³⁷Cs concentration simulations than the ¹³¹I concentration simulations. The comparison of the simulated ¹³⁷Cs deposition shows-yields lowerpoorera general degreeses in the agreement with those particle argument of the simulated ¹³⁷Cs deposition shows-yields lowerpoorera general

decrease in the agreement between all ensemble members (Fig. 13). The inconsistencies between wet deposition <u>rates</u> simulations are much more pronounced than <u>between for</u> dry deposition <u>simulations</u>. <u>This is to be expected as the The</u>

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precipitation occurrence and rate and cloud water content, which are required to calculate wet deposition., are among the most error prone meteorological fields to predicthaveare associated with a high degree of uncertaitnty, inwhich additioning to the existing uncertainties in the deposition parameterisationformulation (Gudiksen et al., 1988). In additionMoreover, FLEXPART and FLEXPART-WRF use different scavenging schemes (see subsection 2.1). However, as withsimilar to the simulated concentrations, CFSv2GFS-based simulations of ¹³⁷Cs (dry/wet) deposition show the greatestclosest agreement with FNL-based simulations. We also observe-find that the largest difference between the simulated dry deposition of ¹³⁷Cs from all ensemble members is seenoccurs between those forced by the FNL and ERA5-WRF datasets. The higher Better agreement between the simulations based on the FNL-WRF and ERA5-WRF datasets than between those based on the FNL and FNL-WRF datasets is also the trueobtained for the simulations of ¹³⁷Cs dry deposition, but not for the simulations of ¹³⁷Cs wet deposition. The above results also apply to the simulations of ¹³¹I deposition (Fig. S11), except that in this study ¹³¹I is assumed to be <u>in-the-insoluble</u>, remaining in the a-gaseous phaseradionuclide and that does not undergo-subject to wet deposition.



670 Figure 12 The inter-comparison of the 96-hour integrated simulations of <u>near-surface</u>¹³¹I (red circles) and ¹³⁷Cs (blue squares) column densities <u>concentrations</u> (Bq/m2<u>m3</u>) simulated at each grid point at each day of 2019. Solid and dotted lines show regression and identity lines, respectively. The asterisk next to the Spearman correlation coefficient (r) indicates a statistically significant correlation at p<0.05.



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Figure 13 Same as Figure 12 but for ¹³⁷Cs dry (in red) and wet (in blue) deposition.

Conclusions

In this study, we examined the spatio-temporal dispersion of radionuclides, including ¹³¹I and ¹³⁷Cs, in the ficticious event of a nuclear accidents at the Barakah Nuclear Power Plant (B-NPP) in the United Arab Emirates (UAE). The resulting 680 concentrations and deposition of the studied radionuclides in the catchment region of interest (Qatar)we-are simulated using the Lagrangian particle dispersion model FLEXible PARTicle (FLEXPART) and FLEXPART coupled with the Weather Research and Forecasting model (FLEXPART-WRF). To investigate the diurnal and seasonal variations of the radionuclide dispersion, the particles (air parcels) werare released within the first 24 hours of a 96-hour simulation period at B-NPP, between 100 and 300 m above ground level (agl), with a daily iterationiterated daily over the year 2019. The source term is scaled to 685 the maximum estimates of the radioactivity emissionse materials from the Fukushima accident (22 PBq of ¹³⁷Cs and 192 PBq of ¹³¹D to give the reader a tangible point of reference. In the course of examining the study question, Wwe foundpay special attention to the differences in the simulations with respect to the meteorological inputs. We investigated the meteorological uncertainties by constructing an ensemble with three members based on (re)analysis datasets including Final Analysis (FNL) at native resolution, FNL and the ECMWF 5th Generation Reanalysis (ERA5) downscaled by WRF (ERA5- and FNL-WRF), 690 and one member based on the NCEP the Global Forecast System elimate forecast system version 2 (CFSv2GFS) run by NCEP. This ensemble also provides the basis for comparing the simulations based on the forecast and (re)analysis datasets. The FNL- and GFS-simulations weare compared with the simulations based on the FNL-WRF and ERA5-WRF datasets to on the modeled dispersion. We also studiedy the sensitivity of ERA5- and FNL-WRF simulations to the turbulence scheme used under convective conditions. We <u>compared</u> the <u>daily means</u> of <u>surface wind speed</u> and temperature and <u>daily total</u> <u>precipitation</u> from the above datasets <u>with observations</u> measured from <u>157 elimatemonitoring</u> stations within the model domain. We examined the simulations of all four members in relation to Qatar's population density, with the goal of identifying possible risks to populated areas from a nuclear accident in the region. A summary of the results of the study is presented below:

- The transport of radionuclides: to analyze the time interval between emission and observation exposure toof radionuclides, we investigated the age composition of the radionuclide plumes. Analysis of air parcel ages indicates that dense radionuclide clouds arrive in the south of the study area approximately 20 to 30 hours after the emission. A significant portion of ¹³¹I released is transported to the most distant parts of the study area up to 40 to 50 hours after the accidentaccidents. Alm addition, all members simulated that a large fraction of the ¹³⁷Cs deposition occurs within the first 75 to 80 hours after the emission. The largest deposition of longer-livedr contribution of long lived particles wais found in FNL-based simulations foris found in all seasons, except in the fall. We attributed this to the more distant transport of air parcels from the emission point in FNL-based simulations, compared to other members. The time transport of air parcels by the downscaled datasets (FNL-WRF and ERA5-WRF) and generated by the same simulation code (FLEXPART WRF) simulated a more similar distribution of air parcel ages than those simulated by the other two members in all seasons except fall.
- DThe distribution of extremely high concentrations and deposition of radionuclides: we calculated investigate 2the seasonal median of 96-hour integrated ¹³¹I concentrations (¹³¹I^{intg_conc_seas,} -(in units of Bq m⁻³), converted to the thyroid internal dose from inhalation effective dose from inhalation of 131 L(TIDI, in units of μ Sv), and the total 137 Cs deposition (137Cstot_depos_seas) across the study area. As expected, all members simulated much higher TIDI over the 715 south/southeast of Qatar, which is the closest point of the study area to the emission point, than over the more distant points in the northern half. The inter-seasonal comparison of the simulations showeds that the ensemble members simulate the largest advance of TIDI > 2500μ Sv from the source to south/southeast of Qatar in the cold period of the year. The simulations of TIDI between the members of the ensemble differ by a factor of 10. In the simulations based on the FNL dataset, TIDI > 2500μ Sv cover the northern half of Qatar (in summer and fall) and the entire western half 720 of Qatar (in winter). In the simulations based on the FNL-WRF dataset, TIDI > 2500µSv wais observed only over the southeastern-corner of Qatar throughout the year. The differences in the TIDI simulations based on these two datasets demonstrated how the use of different model simulation codes and the downscaling of meteorological inputs can affect the FLEXPART modeling and, consequently, the decisions made based on its simulations after nuclearfor real accidents. Similarly, there exist remarkable differences were found in the spatio-temporal distribution of TIDI 725 simulations based on the FNL and CFSv2GFS datasets. This is the case even though these datasets are produced by the same base meteorological model-and are fed into the same simulation code (FLEXPART). We attribute this toto

the fact that the smallest differences in the meteorological inputs that lead to sequential cumulative deviations in the transport and concentration calculations of air-pollutants atmospheric contaminants. As with TIDI, 137Cs^{tot_depos_seas} simulations from all the ensemble members peaked in the southeastern corner-part of the study area in the cold period of the year when both wet and dry deposition occur. The higheststrongest extent of contaminatedion of areas surfaces with 137Cstot_depos_seas greater than 40 kBqm⁻²we-are found in the simulations based on the ERA5-WRF and FNL-WRF datasets in winter. The highest levels of ¹³⁷Cs^{tot_depos_seas} (above 300 kBqm⁻²) weare found in the simulations based on the ERA5-WRF dataset in the southeastern corner of Qatar in the fall. This region receiveds far less ¹³⁷Cs^{tot_depos_seas} (around 40 kBqm⁻² and less) in the simulations based on the GFS and FNL datasets in the same period. The examination of the release time of the air parcels resulting in the extreme TIDI and ¹³⁷Cs^{tot_depos_seas} showeds that the corresponding particles are mostly released between 9 a.m. and 2 p.m. LT. TH seems that the development of the boundary layer height, the intensification of the thermal gradient between the land and sea, and the resulting onshore winds increase the transport of radionuclides to the study area during at this time of day. The analysis of the frequency with which ¹³¹I and ¹³⁷Cs concentrations above the 66th percentile are transported to the populated areas of eastern Oatar showeds that, form all members, more than 50% of the extreme cases occur in winter and between 15% and 30% in spring. The above results indicate that any nuclear accident-at B-NPP in the winter will be-more likely to be accompanied by the with the highest radionuclide concentrations and deposition within the study area. especially in the southeastern corner of Oatar. This pronounced intra-annual variation distribution is attributed to a seasonal atmospheric pattern in which south/southeasterly winds transport the dense radionuclide clouds to the study area. The collocation of population density-and simulations showeds that the populated areas (with more than 8 persons per arcsecond) receivecoincide with moderate (around 5000 µSv and 100 kBqm⁻²) to low levels of TIDI and ¹³⁷Cs^{tot_depos_seas}. Uninhabited areas in southern Qatar receive the highest levels of TIDI (above $10_{a}000 \ \mu Sv$) and ${}^{137}Cs^{tot_depos_seas}$ (above 150 kBqm²). WIn this study, we also investigated the effect of the turbulence scheme selected under convective conditions on the radionuclide dispersion. The implementation of the Skewed Turbulence Model (STM) instead of the Gaussian Turbulence Model (GTM) increaseds the occurrence of high levels of ¹³¹I concentrations and ¹³⁷Cs deposition. For example, the quartiles of simulations of ¹³¹I concentration and ¹³⁷Cs deposition based on the ERA5-WRF dataset increased by 21-2% and 12%, respectively. According to Pisso et al. (2019), this can be interpreted as the enhancement ofd concentrations and deposition in the areas around the source under skewed turbulence conditions.

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755 <u>3- The iInter-comparison of ensemble members:</u> the simulations of the ensemble members are have been compared, with a focus on the CFSv2GFS based simulations. The simulations of ¹³⁷Cs and ¹³¹I concentrations based on the GFS dataset weare found to have the best-closest agreement with the simulations based on the FNL dataset. CFSv2 simulationThis is because theyese input datasets share a common meteorological base model. The comparison of wind speed, precipitation, and temperature from the GFS and FNL datasets <u>showedseause</u> the highest-best agreement

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760 compared to the other two input datasets. However, the above results do not contradict we also found the existence of-importantsignificant differences in the spatio-temporal distribution of GFS- and FNL-based simulations. This may be due to the sequential cumulative effect of even small minutesmall differences in the meteorological inputs on particle dispersion. In additionFurther, these such differences may be caused by the inconsistency of meteorological parameters (,-which is beyond the scope of this study to examine) all of them at different levels of the atmosphere. 765 Our results showed that the sensitivity of the simulation to the meteorological inputs can be better represented by using both GFS and FNL datasets. The comparison of simulations based on ERA5- and FNL-WRF indicated-shows that the downscaling of the inputs and the application of the same simulation increases the agreement of resulting simulations to an extent that exceeds the degree of similarity agreement between simulations based on FNL and FNL-WRF, with the same origin source of the meteorological inputs. The deposition simulations of all the members showed 770 relatively-a larger inconsistency for both radionuclides. This wais more pronounced for the simulations of ¹³⁷Cs wet deposition. This is in part because wet deposition forcing factors are among the most error-pronechallenging meteorological fields-parameters to model accuratelypredict. In additionMoreover, the recently updated wet deposition scheme implemented in the FLEXPART uses different methods to determine the occurrence of wet deposition than FLEXPART-WRF (Girard et al., 2016, Gudiksen et al., 1988, Evangeliou et al., 2017).

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Data availability. The FLEXPART and FLEXPART-WRF simulations are available upon request. The open-source codes for the FLEXPART 10.4 and FLEXPART-WRF 3.3.2 can be downloaded from https://www.flexpart.eu/downloads (last access: 2022). Qatar's high-resolution population density datasets freely available 27 May are at https://data.humdata.org/dataset/qatar-high-resolution-population-density-maps-demographic-estimates (last access: 27 May 2022).

Author contributions. SON performed the WRF, FLEXPART, and FLEXPART-WRF simulations and led the integration of results and writing. SON and TC designed the experiments. All the co-authors have read the paper and provided professional comments.

785 **Competing interests.** The authors declare that they have no conflict of interest.

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