

Quantitative impacts of vertical transport on long-term trend of nocturnal ozone increase over the Pearl River Delta region during 2006-2019

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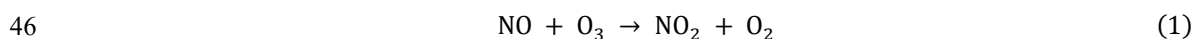
Abstract. The Pearl River Delta (PRD) region in southern China has been subject to severe ozone (O₃) pollution during daytime, and anomalous nocturnal O₃ increase (NOI) during nighttime. In this study, the spatiotemporal variation of NOI events in the PRD region from 2006 to 2019 is comprehensively analyzed and the role of vertical transport in the occurrence of NOI events is quantified based on observed surface and vertical O₃ and the fifth-generation European Centre for Medium-Range Weather Forecasts (ECMWF) reanalysis (ERA5) dataset. The results show that the average annual frequency of NOI events in the whole PRD region during the 14-year period is estimated to be $53 \pm 16 \text{ d yr}^{-1}$, with an average of $58 \pm 11 \mu\text{g m}^{-3}$ for the nocturnal O₃ peak (NOP) concentration. Low-level jets (LLJs) are the main meteorological processes triggering NOI events, explaining on average 61% of NOI events. Annual NOI events exhibit an upward trend before 2011 (4.70 d yr^{-1}) and a downward trend thereafter (-0.72 d yr^{-1}), which is consistent with the annual variation of LLJs ($r=0.88$, $p<0.01$). Although the contribution of convective storms (Conv) to NOI events is relatively small with an average value of 11%, Conv-induced NOI events steadily increased at a rate of 0.26 d yr^{-1} during this 14-year period due to the impact of urbanization. Seasonally, relatively higher frequency of NOI events is observed in spring and autumn, which is consistent with the seasonal pattern of LLJs and maximum daily 8-h average (MDA8) O₃. Spatially, NOI events are frequent in the eastern PRD, which agrees well with the spatial distribution of the frequency of LLJs and partially overlaps with the distribution of MDA8 O₃ concentration, suggesting that vertical transport plays a more important role in NOI events than daytime O₃ concentration. The WRF-CMAQ model and the observed vertical O₃ profiles are further applied to illustrate the mechanisms

31 of NOI formation caused by LLJs and Conv. The results confirm that both LLJs and Conv trigger NOI
32 events by inducing downdrafts with the difference being that LLJs induce downdrafts by wind shear
33 while Conv by compensating downdrafts. Through observational and modeling analysis, this study
34 presents the long-term (2006-2019) trends of NOI events in the PRD region and quantifies the
35 contribution of meteorological processes for the first time, emphasizing the importance of vertical
36 transport as well as daytime O₃ concentration for the occurrence of NOI events.

37 **Keywords:** Nocturnal ozone increase; Ozone profile; Low-level jets; Convective storms; Pearl River
38 Delta; Long-term trend

39 **1 Introduction**

40 As a secondary pollutant, surface ozone(O₃) is formed via photochemical reactions involving nitric oxide
41 (NO_x) and volatile organic compounds (VOCs) in the presence of sunlight. Therefore, O₃ shows
42 significant diurnal variation, with concentration peaks observed during daytime (Kleinman et al., 1994;
43 Zhang et al., 2004). During nighttime, O₃ production ceases owing to the absence of sunlight, and dry
44 deposition and NO titration (Eq. (1)) remove O₃ directly from the atmosphere, lead to relatively low O₃
45 concentrations at night (Jacob, 2000; Brown et al., 2006).



47 However, O₃ concentrations do not always remain at low levels during nighttime, and frequent nocturnal
48 O₃ increase (NOI) events have been observed in various countries in Asia, Europe, North America, etc.
49 in different topographies (plains, valleys, mountains, etc.) (Kuang et al., 2011; Kulkarni et al., 2013;
50 Klein et al., 2019; Udina et al., 2019; Zhu et al., 2020). Kulkarni et al. (2015) found that NOI events were
51 observed around 03:00 (LT) in the UK, with concentrations as high as 118 μg m⁻³, which was much
52 higher than the monthly average daytime O₃ concentration (69 ± 10 μg m⁻³). Yusoff et al. (2019) also
53 reported that frequent NOI events were observed in some cities in Malaysia, and the annual trend of
54 nocturnal O₃ concentration was found to on the increase based on 11 years of ground-based
55 measurements. High nocturnal surface O₃ concentrations have adverse effects on crops and vegetation,
56 leading to plant water loss, stomatal sluggishness, and reduction in plant production (Caird et al., 2007;
57 Cirelli et al., 2016; Yue et al., 2017) as well as on human health (Kurt et al., 2016; Carré et al., 2017).

58 Because there is no photochemical production of O₃ at night, NOI events are likely to be due to
59 meteorological processes (Salmond and McKendry, 2002). It has been widely recognized that low-level
60 jets (LLJs) are one of the most important meteorological processes that cause NOI events (Salmond and
61 McKendry, 2002; Kuang et al., 2011; Sullivan et al., 2017). After sunset, radiative cooling and
62 subsequent weakened turbulence result in a stratified nocturnal boundary layer (NBL) with an altitude of
63 400-500 m (Stull, 1988; Sugimoto et al., 2009; Fan et al., 2022). A residual layer (RL) exists above the
64 NBL, which contains residual O₃ produced during daytime. When an LLJ occurs during nighttime, it can
65 break the delamination between the NBL and the RL by wind shear and bring the O₃ from the RL to the
66 surface, leading to an accumulation of ground-level O₃. An analysis of aircraft data from California has
67 shown that LLJs promote the mixing between the NBL and the RL and transport O₃ from the RL to the
68 surface, leading to NOI events (Caputi et al., 2019). Convective storms (Conv) are another
69 meteorological process that contributes to NOI events, especially at the equator and in tropical areas that
70 have a higher frequency of convection (Prtenjak et al., 2013; Zhu et al., 2020; Wu et al., 2020). Dias-
71 Junior et al. (2017) revealed that downdrafts induced by Conv play an important role in triggering NOI
72 events in the Amazon region of Brazil based on 1-yr observations. Jain et al. (2007) noted that NOI
73 events in India are often accompanied by thunderstorms and stable boundary layer conditions. Other
74 meteorological processes that are highly dependent on topography, such as sea-land breezes and
75 mountain-valley breezes, also contribute to NOI events (Salmond and McKendry, 2002; Nair et al., 2002).
76 Seibert et al. (2000) pointed out that nocturnal O₃ concentrations are elevated during foehn events in the
77 Eastern Alps.

78 Therefore, NOI events are not an exception and can occur worldwide as a result of certain meteorological
79 processes (LLJs, thunderstorms, foehn, etc.) (Hu et al., 2013; Caputi et al., 2019; Klein et al., 2019; Udina
80 et al., 2019; Shith et al., 2021). LLJs and Conv are important factors influencing the generation of NOI
81 events; however, their relative contribution to NOI events has not yet been quantified. Most previous
82 studies are focused on the analysis of a single NOI event or NOI events at limited monitoring sites for
83 short periods (Jain et al., 2007; Hu et al., 2013; He et al., 2021). Consequently, it is of great importance
84 to investigate the long-term trends of NOI events on a larger scale to further quantify the impacts of
85 meteorological processes, such as LLJs and Conv, on NOI events.

86 In China, ground-level O₃ pollution has been deteriorated in recent years, especially in the Pearl River
87 Delta (PRD) region (Wang et al., 2017). Liao et al. (2021) investigated ozonesonde profiles recorded in

88 Hong Kong during 2000-2019 and indicated that O₃ concentrations in the lower troposphere have
89 increased substantially at a rate of 0.618 ppbv yr⁻¹, indicating a continuous deterioration of O₃ pollution
90 in the PRD region over the last 20 years. The PRD region is the first urban agglomeration in China to
91 change its main pollutant from particulate matter with an aerodynamic diameter of less than or equal to
92 2.5 μm (PM_{2.5}) to O₃. Numerous studies in the PRD region have investigated the daytime O₃
93 characteristics, such as the long-term trends (Xue et al., 2014; Li et al., 2022), the nonlinear response of
94 O₃ to precursor emissions (Lu et al., 2010; Mao et al., 2022), the source apportionment of O₃ (Shen et
95 al., 2015; Liu et al., 2020), and the relative contributions of precursor emissions and meteorology to O₃
96 (Yang et al., 2019b; Chen et al., 2020). In terms of nighttime O₃, Tong and Leung (2012) observed a
97 double-peak pattern of diurnal O₃ variation in Hong Kong during 1990-2005, and found that nocturnal
98 O₃ peaks are sometimes higher than daytime maxima. He et al. (2021) studied an NOI event at the city
99 Shaoguan in Guangdong Province, and found that nocturnal mountain-valley breezes from the Nanling
100 Mountains transported O₃ from the RL to the surface. However, studies on the spatio-temporal
101 distribution of nocturnal O₃ concentration in the PRD region and the factors influencing it are still lacking.
102 There is an urgent need to comprehensively study the characteristics of NOI events in the PRD region as
103 it is frequently affected by special meteorological processes (such as LLJs and Conv) that favor NOI
104 events due to its special topography with the coast to the south and the mountains to the north. In addition,
105 high population densities and increasing number of people active at night in the PRD region make NOI
106 events an important potential risk to human health (Kurt et al., 2016; Carré et al., 2017; Yang et al.,
107 2019a; Zhang et al., 2021).

108 In this study, the long-term trends and spatial distribution of NOI are presented via in-situ hourly O₃
109 concentration data collected from 16 air quality monitoring sites in the PRD region during 2006-2019.
110 In addition, the relative contributions of LLJs and Conv to NOI events are quantified based on the ERA5
111 reanalysis dataset. Finally, the observed vertical profile of O₃ and the Weather Research and Forecasting
112 (WRF) model coupled with the Community Multiscale Air Quality (CMAQ) model are applied to further
113 elaborate the impacts of LLJs and Conv on the selected typical NOI events. This study provides a
114 comprehensive analysis of NOI events and the meteorological factors influencing them in the PRD region
115 over a 14-year period for the first time, expanding our knowledge of the meteorological role in NOI
116 events.

117 **2 Data and methods**

118 **2.1 Data sources**

119 The dataset used in this study is summarised in Table 1. In brief, the observed hourly O₃ concentrations
120 at the 16 air quality monitoring sites in the PRD region from 2006 to 2019 are provided by the
121 Guangdong-Hong Kong-Macao Pearl River Delta Regional Air Quality Monitoring Network (HKEPD,
122 2017) (Fig. 1). More detailed information of these sites can be found in Table S1. The observed hourly
123 O₃ data were used for subsequent NOI and NOP analyses, and evaluation of O₃ simulations.

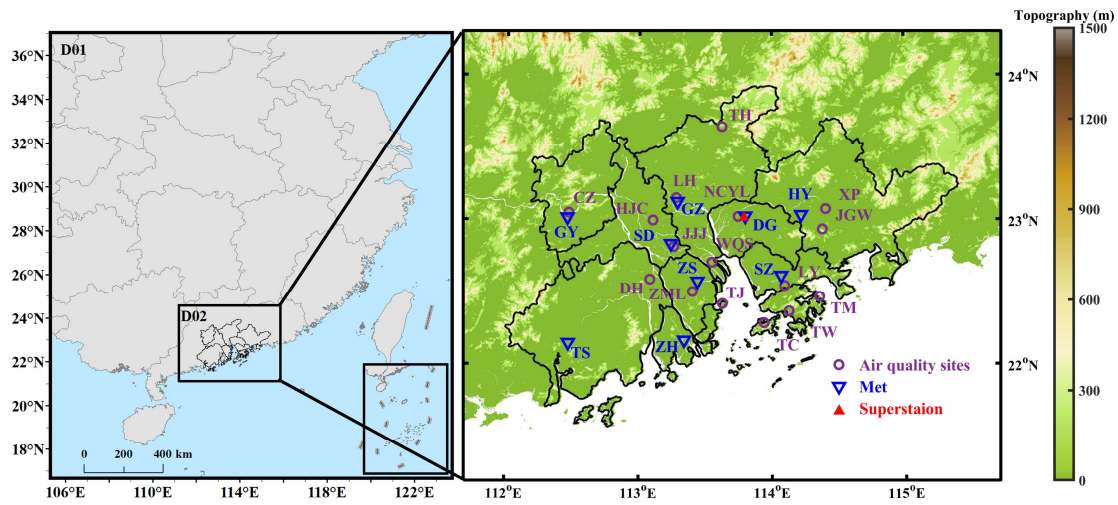
124 The vertical distribution of O₃ concentrations observed at the Dongguan superstation (23.02° N, 113.79°
125 E) in 2019 is also used to investigate the impact of Conv on a particular NOI event. The vertical profile
126 of O₃ is measured using an O₃ lidar (Model: LIDAR-G-2000). The detection height of the O₃ lidar is 3
127 km, with a vertical spatial resolution of 7.5 m and a temporal resolution of 12 mins.

128 The observed meteorological variables at the 9 meteorological sites (Fig. 1) in the PRD region are
129 obtained from the Chinese National Meteorological Centre (CNMC, <http://www.cma.gov.cn/>, last
130 accessed on February 10, 2022), including temperature (T2), relative humidity (RH), and wind speed
131 (WS10). The observed meteorological data were used to evaluate the performance of the model. More
132 detailed information of the 9 meteorological sites can be found in Table S2.

133 To investigate the impacts of meteorological processes on NOI events, the ERA5 reanalysis dataset
134 (<https://cds.climate.copernicus.eu/cdsapp#!/home>, last accessed on February 10, 2022) provided by the
135 European Centre for Medium-Range Weather Forecasts (ECMWF) is used in this study. The ERA5
136 reanalysis dataset, which currently covers the period from 1979 to present, is provided on regular
137 latitude–longitude grids at approximately 0.25° × 0.25° and up to 1 h frequency. Vertically, ERA5
138 resolves the atmosphere using 137 levels from the surface to an altitude of 0.01 hPa. The performance of
139 ERA5 is evaluated in previous studies and has been shown to be adequate for further analysis (Olauson,
140 2018; Hersbach et al., 2020). The ERA5 reanalysis dataset includes wind speed, precipitation,
141 temperature, and vertical wind velocity. Since the ERA5 reanalysis dataset was gridded, the nearest-
142 neighbour interpolation method is used to obtain site-specific meteorological variables at the 16 air
143 quality monitoring sites.

144 The observed cloud-top temperature (CTT) data for 2019 obtained from the Fengyun-2G satellite
145 (<http://satellite.nsmc.org.cn/>, last accessed on August 31, 2022) are used to indicate the occurrence of

146 convection. The CTT data cover the East Asia region with a spatial resolution of 0.1° and the temporal
 147 resolution of 1 h.



148
 149 **Figure 1. Model domains and locations of 16 air quality monitoring sites (purple dots), 9 meteorological**
 150 **stations (blue triangles), and the Dongguan superstation (red triangles). The figure on the right shows the**
 151 **elevation of the terrain (m).**

152 **Table 1. Summary of the dataset used in this study**

| Description | Period | Sites | Temporal resolution | Spatial resolution | Purpose |
|---|-----------------------|-----------------------|---------------------|--------------------|---|
| Observed O ₃ data | 2006-2019 | 16 sites | 1 h | - | Spatiotemporal analysis of NOI and NOP, model performance |
| Observed vertical O ₃ data | 2019 | Dongguan superstation | 12 min | - | Analysis of an NOI event caused by Conv |
| Observed meteorological data | 2017.09.08-2017.09.15 | 9 sites | 1 h | - | Model performance |
| Observed Cloud-top Temperature (CTT) data | 2019 | Gridded data | 1 h | 0.1° | Indicator of the occurrence of convection |
| ERA5 reanalysis dataset | 2006-2019 | Gridded data | 1 h | 0.25° | Definition of LLJs and Conv |

153

154 2.2 Definition of NOI and NOP

155 For our analysis, we define a nocturnal O₃ increase (NOI) event as O₃ concentrations peaked at night
 156 (from 21:00 LT to 06:00 LT the next day), with an increase in levels of at least $10 \mu\text{g m}^{-3}$ compared to
 157 the previous hour (include 20:00 LT) and a decrease of less than $10 \mu\text{g m}^{-3}$ in the next hour (include
 158 07:00 LT). The corresponding nighttime peak concentration of O₃ is referred to as the nocturnal O₃ peak
 159 (NOP) (Zhu et al., 2020). In this study, based on the above observed hourly O₃ data at the 16 air quality
 160 monitoring sites, NOI events are identified at each site, yet only one NOI event is recorded per night,

161 regardless of how many NOI events occur in a single night. In addition, the regional values of NOI and
162 NOP from the 16 air quality monitoring sites were averaged.

163 **2.3 Definition of LLJs and Conv**

164 Low-level jets (LLJs) and convective storms (Conv) are defined in this study based on the above site-
165 specific ERA5 reanalysis dataset. According to Banta et al. (2002) and Hodges and Pu (2019), LLJs are
166 defined as when vertical wind speed maxima occur below 800hPa and exhibit a decrease of at least 1.5
167 m s⁻¹ at vertical levels both above and below the levels of the maxima. **We assume that LLJs cause**
168 **downdrafts because of the vertical wind shear the jets induce, which creates mechanical turbulence.**

169 Conv is defined by the following criterion: the mean K index (KI) is greater than 30 °C within 3 hours
170 prior to an NOI event (George, 1960; Johnson, 1982). The KI is calculated as follows:

$$171 \quad KI = (T_{850} - T_{500}) + Td_{850} - (T_{700} - Td_{700}) \quad (1)$$

172 where T_{850} , T_{700} , and T_{500} are the temperature (°C) at 850 hPa, 700 hPa, and 500 hPa, respectively,
173 and Td_{850} and Td_{700} are the dew point temperature at 850 hPa and 700 hPa, respectively.

174 Cloud-top temperature (CTT) was also introduced as an indicator of the occurrence of convective systems
175 and further used to evaluate the applicability of KI. The lower the CTT, the higher the probability of
176 convection event. According to the work of Ai et al. (2016), CTT lower than -35 °C indicates the
177 occurrence of convection. We randomly selected 10 nights with $KI > 30$ °C (Table S3) and 10 nights
178 with $KI < 30$ °C (Table S4) and examined the corresponding CTT values. In the cases with $KI > 30$ °C,
179 the CTT values were lower than -35 °C in 10 out of 10 nights (Table S3). And the spatial distribution of
180 CTT showed that they had a distinct circular area with lower value over the selected sites, indicating the
181 occurrence of convective systems (Fig. S1). For the cases with $KI < 30$ °C, 6 out of 10 nights were with
182 CTT higher than -35 °C, while the rest 4 nights had no CTT data due to cloudless weather (Table S4).
183 The spatial distribution of CTT did not show the features of a convective system (Fig. S2), suggesting
184 that convection was not observed for the selected 10 cases with $KI < 30$ °C. The above results suggest
185 that the $KI > 30$ °C criterion is a valid metric to capture the occurrence of convection.

186 In this study, an NOI event at each air quality site was classified into four categories: caused by LLJs
187 only, caused by Conv only, caused by LLJs and Conv (LLJs+Conv) at the same time, and caused by
188 other factors.

189 **2.4 Trend analysis**

190 In this study, the nonparametric Mann-Kendall (M-K) test (Mann, 1945) is used to determine the
191 statistical significance (p values) associated with the annual trends of NOI, NOP, MDA8 O₃, LLJs and
192 Conv, etc. A significance level of $p < 0.05$ was used to test the significance of the inter-annual trend. The
193 magnitude of a given trend is calculated by the nonparametric Theil-Sen (T-S) estimator (Sen, 1968).
194 The advantage of the M-K test and the T-S estimator is that they do not require prior assumptions of the
195 statistical distribution for the data and are resistant to outliers. The M-K test and the T-S estimator have
196 been widely used in previous O₃ trend studies (Wang, et al., 2019; Lu et al., 2020; Li et al., 2022).

197 **2.5 WRF-CMAQ model configuration**

198 Due to the lack of observed vertical profiles of wind speed, the WRF-CMAQ model is employed to
199 investigate the effects of LLJs on a selected NOI event. The NOI event induced by LLJs that occurred at
200 the Nancheng Yuanling (NCYL) site in Dongguan on September 13-14, 2017 is selected as a typical case.

201 The simulation was conducted during September 6-14 by using the WRF-CMAQ-IPR model with first
202 2 days used as model spin-up to eliminate the impact of IC (Jiménez et al., 2007).

203 The Weather Research and Forecasting model (WRFv3.9.1) is used to provide meteorological inputs to
204 drive the Community Multiscale Air Quality (CMAQ v5.3.1) model. The initial meteorological
205 conditions (IC) and boundary conditions (BC) are provided by the National Centers for Environmental
206 Prediction (NCEP) Final Analyses (FNL) dataset, with a spatial resolution of $1^\circ \times 1^\circ$ and a temporal
207 resolution of 6 h. The main physics options used for the WRF model are shown in Table 2. Two-nested
208 domains are used in the WRF simulations, with 38 vertical layers from the surface to 100 hPa. Figure 1
209 shows the two nested modeling domains, with spatial resolutions of $27 \text{ km} \times 27 \text{ km}$ and $3 \text{ km} \times 3 \text{ km}$ for
210 the coarse (D01) and inner (D02) domains, respectively. D01 covers most regions of China and D02
211 covers the whole PRD region.

212 The CMAQ model is used to simulate the O₃ concentrations in the PRD region. The SAPRC07 and
213 AERO6 aerosol modules are used for gas-phase and particulate matter chemical mechanisms,
214 respectively (Carter, 2010; Wyatt Appel et al., 2018). The chemical IC and BC for D01 are derived from
215 a global chemical transport model, the Model for Ozone and Related chemical Tracers, version 4
216 (MOZART4) (Emmons et al., 2010), and those for D02 are provided by the simulated results from D01.
217 The anthropogenic emissions for D01 are based on the 2016 Multi-resolution Emission Inventory for

218 China (MEIC), which has a grid resolution of $0.25^\circ \times 0.25^\circ$ (Zheng et al., 2018). Those used for D02 are
 219 based on the 2017 high-resolution emission inventory of the PRD region with a grid resolution of 3 km
 220 \times 3 km (Zhong et al., 2018), which includes the emission sectors of agriculture, biomass combustion,
 221 incineration, dust, industrial processes, nonroad, solvent, storage, transportation, and waste disposal.
 222 Biogenic emissions are calculated using the Model of Emissions of Gases and Aerosols from Nature
 223 (MEGAN) v2.1 that was integrated into the CMAQ model (Guenther et al., 2006; Wang et al., 2011).

224 **Table 2. Model configurations**

| Model | Physical process | Parameterization scheme | Reference |
|-------|--------------------------|-------------------------|----------------------------|
| WRF | Microphysics | Lin | Lin et al. (1983) |
| | Longwave radiation | RRTMG | Iacono et al. (2008) |
| | Shortwave radiation | RRTMG | Iacono et al. (2008) |
| | Surface layer | Monin-Obukhov | Monin and Obukhov (1954) |
| | Planetary boundary layer | MYJ | Nakanishi and Niino (2006) |
| | Cumulus parameterization | Grell-3 | Grell and Dévényi (2002) |
| CMAQ | Land surface | Noah land-surface model | Chen and Dudhia (2001) |
| | Gas-phase chemistry | SAPRC 07 | Carter (2010) |
| | Aerosol chemistry | AERO6 | Carlton et al. (2010) |

225
 226 In order to interpret the underlying atmospheric mechanisms for NOI events, the Integrated Process Rates
 227 (IPR) analysis tool embedded in the WRF-CMAQ model is used to identify and quantify the contribution
 228 of various physical and chemical processes to O_3 . The processes include horizontal transport (HTRA),
 229 vertical transport (VTRA), gas-phase chemistry (CHEM), dry deposition (DDEP), and cloud processes
 230 (CLDS). Horizontal transport is the sum of horizontal advection and diffusion, and vertical transport is
 231 the sum of vertical advection and diffusion. More details on the IPR analysis tool can be found in previous
 232 work (Liu et al., 2010; Wang et al., 2010).

233 2.6 Model evaluation

234 The WRF-CMAQ simulation results are evaluated by comparison with available ground-based observed
 235 O_3 and meteorological data. Statistical metrics including mean value (\overline{Obs} and \overline{Sim}), mean bias (MB),
 236 normalized mean bias (NMB), normalized mean error (NME), root mean square error (RMSE),
 237 correlation coefficient (r), and index of agreement (IoA), are calculated as follows to evaluate model
 238 performance.

$$239 \quad MB = \overline{Obs} - \overline{Sim} \quad (1)$$

240
$$NMB = \frac{\sum_{i=1}^n (Sim_i - Obs_i)}{\sum_{i=1}^n Obs_i} \times 100 \quad (2)$$

241
$$NME = \frac{\sum_{i=1}^n |Sim_i - Obs_i|}{\sum_{i=1}^n Obs_i} \times 100 \quad (3)$$

242
$$RMSE = \sqrt{\frac{1}{n} \sum_{i=1}^n (Sim_i - Obs_i)^2} \quad (4)$$

243
$$r = \frac{\sum_{i=1}^n (Sim_i - \overline{Sim})(Obs_i - \overline{Obs})}{\sqrt{\sum_{i=1}^n (Sim_i - \overline{Sim})^2 \sum_{i=1}^n (Obs_i - \overline{Obs})^2}} \quad (5)$$

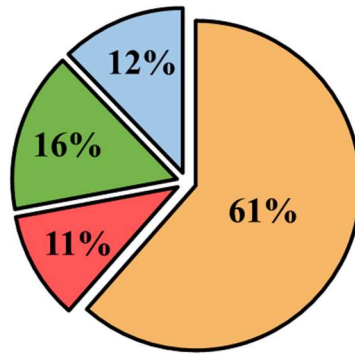
244
$$IoA = 1 - \frac{\sum_{i=1}^n (Sim_i - Obs_i)^2}{\sum_{i=1}^n (|Sim_i - \overline{Obs}| + |Obs_i - \overline{Obs}|)^2} \quad (6)$$

245 The evaluation protocols of the U.S. Environmental Protection Agency (EPA, 2017) are used to evaluate
 246 the performance of the meteorological parameters. The simulated results were accepted when the
 247 statistics met the criteria listed as follows: $MB \leq \pm 0.5$ °C and $IoA \geq 0.8$ for simulated T2; $MB \leq \pm 5\%$
 248 and $IoA \geq 0.6$ for simulated RH; and $MB \leq \pm 0.5$ m/s, $RMSE \leq 2.0$ m/s, and $IoA \geq 0.6$ for simulated
 249 WS10. The evaluation protocols of the Ministry of Environmental Protection of China (MEP, 2015) are
 250 used to evaluate the performance of O₃ and the simulated results were acceptable if the statistics met the
 251 criteria listed below: $-15\% < NMB < 15\%$, $NME < 35\%$, and $r > 0.4$.

252 **3 Results and discussion**

253 **3.1 General characteristics of NOI events**

254 The average annual frequency of NOI events in the 16 sites across the whole PRD region from 2006 to
 255 2019 is estimated to be 53 ± 16 d yr⁻¹, with an average annual NOP concentration of 58 ± 11 µg m⁻³. LLJs
 256 are the primary factor causing NOI events, accounting for about 61%, followed by the combination of
 257 LLJs and Conv (LLJs+Conv) with a value of 16%, while the corresponding value is 11% for Conv (Fig.
 258 2). The remaining 12% of NOI events that cannot be explained by LLJs and Conv may be related to other
 259 meteorological processes, such as mountain-valley breezes and sea-land breezes (Sousa et al., 2011; He
 260 et al., 2021).



■ LLJs ■ Conv ■ LLJs+Conv ■ Other

261

262 **Figure 2. The average relative contribution of different meteorological processes to NOI events during 2006-**
 263 **2019.**

264 3.2 Long-term trends of NOI events

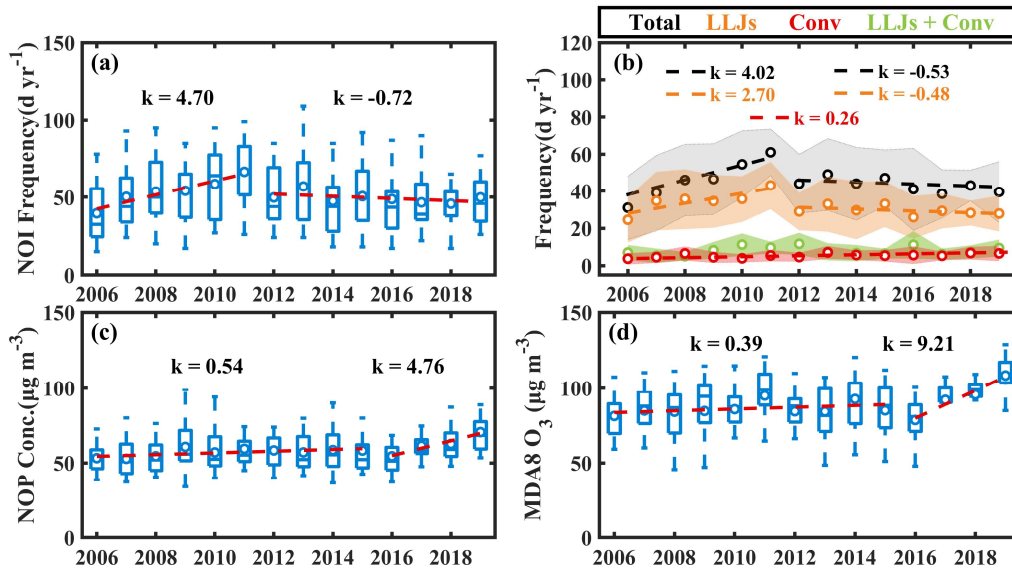
265 As depicted in Fig. 3a, the regional-average annual frequency of NOI events increases from 38 ± 18 d
 266 yr^{-1} in 2006 to as high as 67 ± 18 d yr^{-1} in 2011 at a rate of 4.70 d yr^{-1} ($p < 0.01$) and gradually decreases
 267 after 2012 at a rate of -0.72 d yr^{-1} ($p < 0.05$). A similar annual trend is observed for the frequency of total
 268 downdrafts (sum of LLJs, LLJs+Conv, and Conv) (Fig. 3b). The frequency of total downdrafts increased
 269 at a rate of 4.02 d yr^{-1} ($p < 0.01$) before 2012 and decreased at a rate of -0.53 d yr^{-1} ($p < 0.05$) thereafter,
 270 which is significantly positively correlated with NOI events, with a Pearson correlation coefficient (r) of
 271 0.96 ($p < 0.01$). Among the total downdrafts, LLJs exhibit a similar pattern with NOI events ($r = 0.89$,
 272 $p < 0.01$), further suggesting that LLJs are the predominant driver. Conv presents a continuously
 273 increasing trend over the whole 14-year period, with a rate of 0.26 d yr^{-1} ($p < 0.01$), and the frequency of
 274 LLJs+Conv does not show obvious variation.

275 Both the frequency of NOI and LLJs present increasing trends before 2012 and decreasing trends
 276 thereafter, which **was likely related to** urbanization. Previous studies have shown that urbanization has
 277 large effects on the frequency of LLJs by changing surface conditions (roughness and soil moisture) and
 278 further affecting the turbulence and geostrophic wind speed (McCorcle, 1988; Fast and McCorcle, 1990;
 279 Kallistratova, 2008; Nikolic et al., 2019; Ziemann et al., 2019). Kallistratova (2008) and Nikolic et al.
 280 (2019) pointed out that negative correlation was found between urban areas and the frequency of LLJs.
 281 During 1987-2017, the urban areas in the PRD region grew at an average rate of 8.82% yr^{-1} (Yang et al.,
 282 2019a) and reached maximum urban land expansion growth rate of 6.66% during 2010-2015 (Zhang et

283 al., 2021). Therefore, the trends for the frequency of NOI and LLJs were quite different during these two
284 periods (2006-2011 and 2012-2019).

285 Although the percentage of NOI events caused by Conv alone is relatively small compared to those
286 caused by LLJs (Fig. 2), it is noteworthy that the frequency of Conv-induced NOI events was on the
287 increase during the 14-year period (Fig. 3b), which is also mainly related to the rapid urbanization in the
288 PRD region in recent years (Yang et al., 2019a; Zhang et al., 2021). Surface roughness increase due to
289 city expansion led to the greater frequency and intensity of convection in the form of enhanced
290 mechanical turbulence (Li et al., 2021), thus an increase in the frequency of Conv-induced NOI events.
291 The role of Conv in the occurrence of NOI events is expected to amplify in the future if the urbanization
292 trend in China continues (Seto et al., 2012; Marelle et al., 2020).

293 In contrast to the annual trend of NOI frequency, the nocturnal O₃ peak (NOP) value shows an upward
294 trend during 2006-2019, with a slower growth rate of 0.54 $\mu\text{g m}^{-3} \text{yr}^{-1}$ ($p < 0.05$) before 2015 and a faster
295 growth rate of 4.76 $\mu\text{g m}^{-3} \text{yr}^{-1}$ ($p < 0.01$) thereafter (Fig. 3c). The maximum daily 8-h average (MDA8)
296 O₃ mixing ratio exhibits a similar pattern to NOP, with an increase rate of 0.39 $\mu\text{g m}^{-3} \text{yr}^{-1}$ ($p < 0.05$) before
297 2015 and 9.21 $\mu\text{g m}^{-3} \text{yr}^{-1}$ ($p < 0.01$) thereafter (Fig. 3d). NOP is significantly positively correlated with
298 MDA8 O₃, with r up to 0.88 ($p < 0.01$). This implies that daytime O₃ concentration levels potentially affect
299 NOP concentrations. The variations of NOP and MDA8 O₃ during the two periods (2006-2015 and 2016-
300 2019) are more likely related to the change in precursor emissions. The continuous increase in the
301 emissions of anthropogenic VOCs and NO_x resulted in the gradual increase of O₃ concentrations between
302 2006 and 2012 (Ma et al., 2016; Li et al., 2017; Zhong et al., 2018; Liao et al., 2021). However, since
303 the implementation of Air Pollution Prevention and Control Action Plan (APPCAP) in 2013, NO_x
304 emissions was dramatically decreased by 21% in 2017 compared to 2013 (Feng et al., 2019; Yang et al.,
305 2019b). The weakening of NO titration caused by the dramatic decrease in NO_x emissions and the
306 continuously increasing VOCs emissions due to the lack of controls became important drivers of the
307 sharp rise in O₃ since 2015 (Li et al., 2019; Mousavinezhad et al., 2021; Li et al., 2022). Furthermore,
308 the decreasing PM_{2.5} levels and the increasing atmospheric oxidizing capacity in the PRD region in recent
309 years have also been considered as important contributors to accelerated O₃ growth during 2016-2019
310 (Gong et al., 2018; Li et al., 2019; Han et al., 2019). Consequently, NOP and MDA8 O₃ present slower
311 increase rate before 2015 and higher increase rate thereafter.



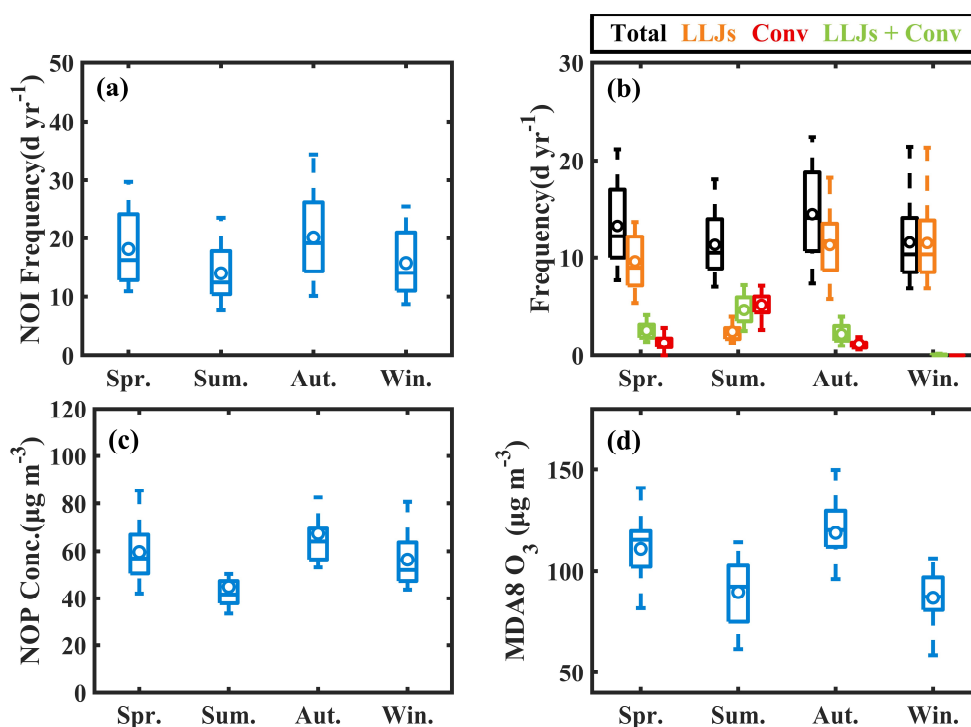
312
 313 **Figure 3. Regional-average annual trends of (a) frequency of NOI events, (b) frequency of total downdrafts**
 314 **(black), LLJs (orange), LLJs+Conv (green), and Conv (red) that can induce NOI events, (c) NOP**
 315 **concentrations, and (d) MDA8 O₃ concentrations in the PRD region during 2006-2019. The units of k (Sen's**
 316 **Slope) are d⁻¹ yr⁻¹ in (a) - (b) and μg m⁻³ yr⁻¹ in (c) - (d). Linear trends significant at the 95% confidence level**
 317 **are illustrated with dashed lines. The error bars indicate the range of deviations for the 16 air quality sites.**

318 3.3 Seasonal variations of NOI events

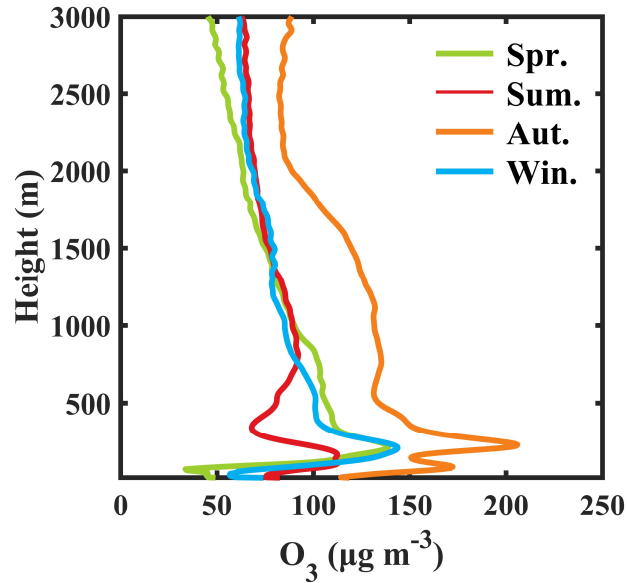
319 NOI events exhibit obvious seasonal variation (Fig. 4a), with relatively higher frequency observed in
 320 spring (18 ± 4 d yr⁻¹) and autumn (20 ± 5 d yr⁻¹) and lower frequency in summer (14 ± 3 d yr⁻¹) and winter
 321 (16 ± 3 d yr⁻¹). LLJs are the dominant inducer of NOI events in spring, autumn, and winter (Fig. 4b),
 322 while in summer, the dominant factors are LLJs+Conv and Conv because convective activity is more
 323 intense during summer (Chen et al., 2014). Given that LLJs can enhance turbulence below the jet and
 324 create favorable formation conditions for Conv (Trier et al., 2017; Du and Chen, 2019), most Conv events
 325 preferentially occur on days when LLJs exist in the PRD region (Chen et al., 2014), which makes
 326 LLJs+Conv the main contributor in summer.

327 In terms of NOP (Fig. 4c), relatively higher concentrations are observed in spring and autumn, with
 328 values of 59 ± 12 μg m⁻³ and 66 ± 10 μg m⁻³, respectively, while the concentration in summer is the
 329 lowest (44 ± 7 μg m⁻³). The MDA8 O₃ has a similar seasonal variation to NOP except in winter (Fig. 4d);
 330 it is high in spring (111 ± 15 μg m⁻³) and autumn (120 ± 13 μg m⁻³) and low in summer (88 ± 15 μg m⁻³).
 331 In winter, surface MDA8 O₃ was the lowest (86 ± 12 μg m⁻³) while NOP remained at relatively high
 332 levels (56 ± 10 μg m⁻³). This is because the higher O₃ concentrations in the lower troposphere in winter
 333 allow more O₃ to be transported downward during the NOI period, resulting in a higher NOP

334 concentration in winter, as shown by the seasonally observed vertical O₃ profile at the Dongguan
 335 superstation (Fig. 5). As illustrated in Fig. 5, higher O₃ concentrations are observed at 200 to 750 m
 336 altitude in winter than in summer. A similar result was also observed in Hong Kong (Liao et al., 2021).
 337 This is mainly due to the typical Asian monsoon circulation, which brings clean marine air to the lower
 338 troposphere of the PRD region in summer and dilutes polluted air masses inland, while it brings pollutant-
 339 laden air from mainland China in winter resulting in higher O₃ concentrations over the PRD region (Wang
 340 et al., 2009).



341
 342 **Figure 4.** Seasonal variation of (a) frequency of NOI events, (b) frequency of total downdrafts (black), LLJs
 343 (orange), LLJs+Conv (green), and Conv (red) that can induce NOI events, (c) NOP concentrations, and (d)
 344 MDA8 O₃ concentrations in the PRD region during 2006-2019. The error bars indicate the range of deviations
 345 for the 16 air quality sites.



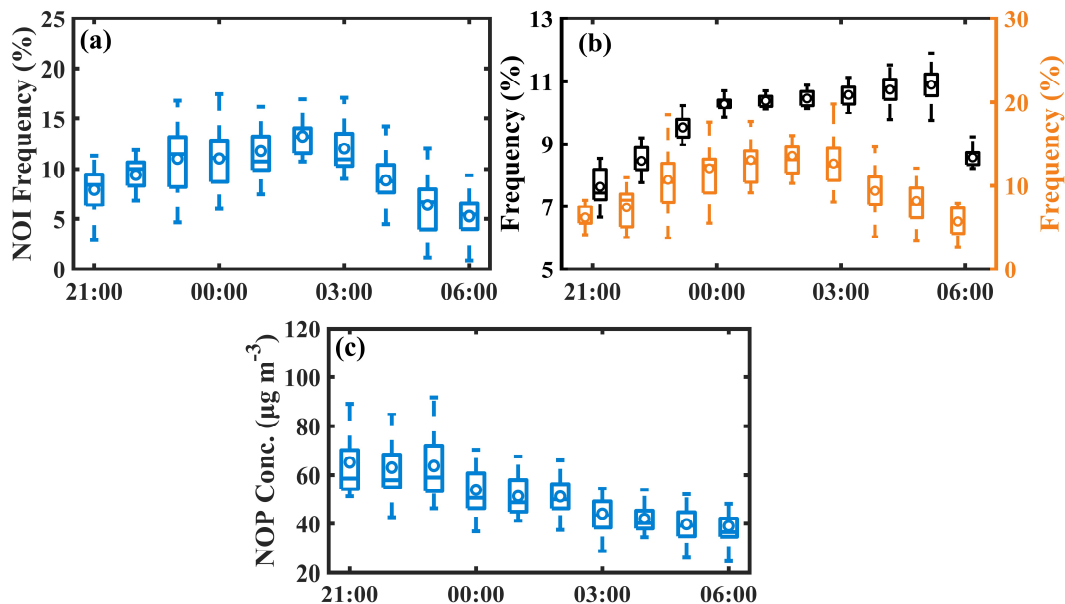
346

347 **Figure 5. Seasonally averaged vertical distribution of O₃ concentrations at the Dongguan superstation from**
 348 **the surface to an altitude of 3000 m in 2019. Spring (Spr.): March-May; Summer (Sum.): June-August;**
 349 **Autumn (Aut.): September-November; Winter (Win.): December-February.**

350 **3.4 Diurnal variation of NOI events**

351 Distinct diurnal variation is observed in NOI events (Fig. 6a) with an increasing trend from 21:00 to
 352 03:00 and a decreasing trend thereafter. It is estimated that about 60% of the events occurred in the
 353 middle of the night (11:00-03:00). The LLJs that can induce NOI events show a similar diurnal variation
 354 to the frequency of NOI events and often occur around at midnight. However, the frequency of total LLJs
 355 differs from the frequency of LLJs that can induce NOI events (Fig. 6b), as it increases steadily from
 356 21:00 to 00:00 and remains stable after 00:00. This suggests that LLJs are not the only factor that
 357 determines whether an NOI event can develop, and O₃ concentration in the RL can also affect the
 358 development of an NOI event. As the sun sets and the daytime boundary layer fades away, the O₃
 359 produced during daytime remains at a relatively high level in the RL during 21:00-03:00. During this
 360 period, the occurrence of LLJs tends to increase the probability of NOI events. After 03:00, the O₃
 361 concentrations in the RL decreased due to horizontal transport to downwind area and vertical transport
 362 (e.g., LLJs, convection, O₃ dry deposition process) during 21:00-03:00, which reduced the amount of O₃
 363 that can be transported downward. Hence, even though the frequency of the total LLJs is relatively high
 364 after 03:00, the lower O₃ content in the RL results in less O₃ being transported downward to form an NOI
 365 event, which ultimately decreases the frequency of NOI events. As illustrated in Fig. 6c, the trends of
 366 NOP concentrations from 21:00 to 06:00 also reflect the fact that O₃ concentrations in the RL are higher

367 during 21:00-00:00 and lower during 00:00-06:00. Therefore, the development of an NOI event is
 368 influenced by the combination of a downdraft induced by meteorological processes, and the level of O₃
 369 concentrations.

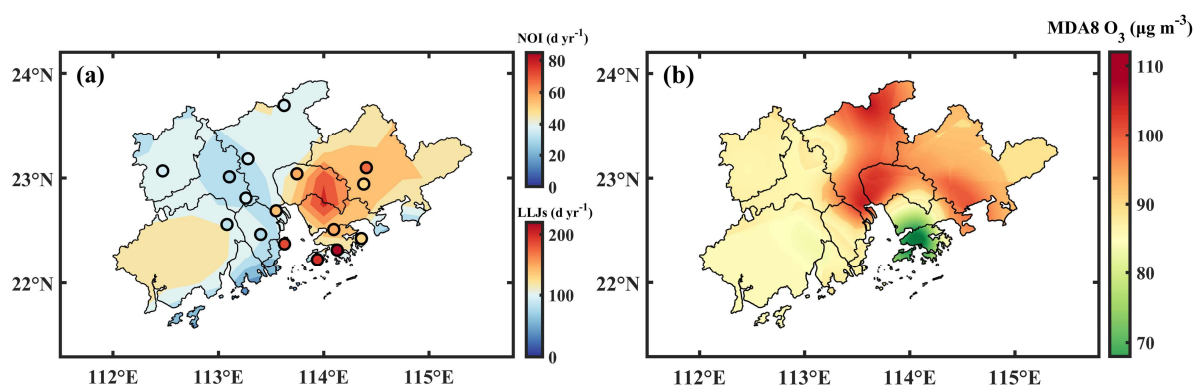


370
 371 **Figure 6. Diurnal variation of (a) the frequency of NOI events, (b) frequency of total LLJs (black) and the**
 372 **LLJs that can induce NOI events (orange), (c) NOP concentrations during 21:00-06:00 (LTC) in the PRD**
 373 **region during 2006-2019. The error bars indicate the range of deviations for the 16 air quality sites.**

374 3.5 Spatial distribution of NOI events

375 As most NOI events are caused by LLJs, LLJs are taken as an example to explore the role of
 376 meteorological processes in the spatial distribution of NOI events. The spatial distribution of the average
 377 annual frequency of NOI events and LLJs in the PRD region from 2006 to 2019 is shown in Fig. 7a, and
 378 the spatial distribution of MDA8 O₃ concentrations obtained by Kriging's interpolation method is shown
 379 in Fig. 7b. Obvious geographical variations are observed for NOI events, with a higher frequency in the
 380 eastern PRD region, coupled with a higher frequency of LLJs, although the MDA8 O₃ concentrations are
 381 relatively lower in these regions. In the central PRD region, despite the highest MDA8 O₃ concentrations,
 382 the frequency of NOI events was the lowest, implying a more important role of vertical transport induced
 383 by meteorological processes in the formation of NOI events. At the three sites located in the southern
 384 part of the PRD regions (TJ, TW, and TC), the frequency of NOI events was the highest while the
 385 frequency of LLJs was not. This is because these three sites were also affected by non-LLJs
 386 (=Conv+(LLJs+Conv)+Other) processes with comparable contributions of LLJs and non-LLJs to the

387 NOI events. And the contributions of LLJs (60-70%) were higher than those of non-LLJs at the rest of
388 sites (Table S5).



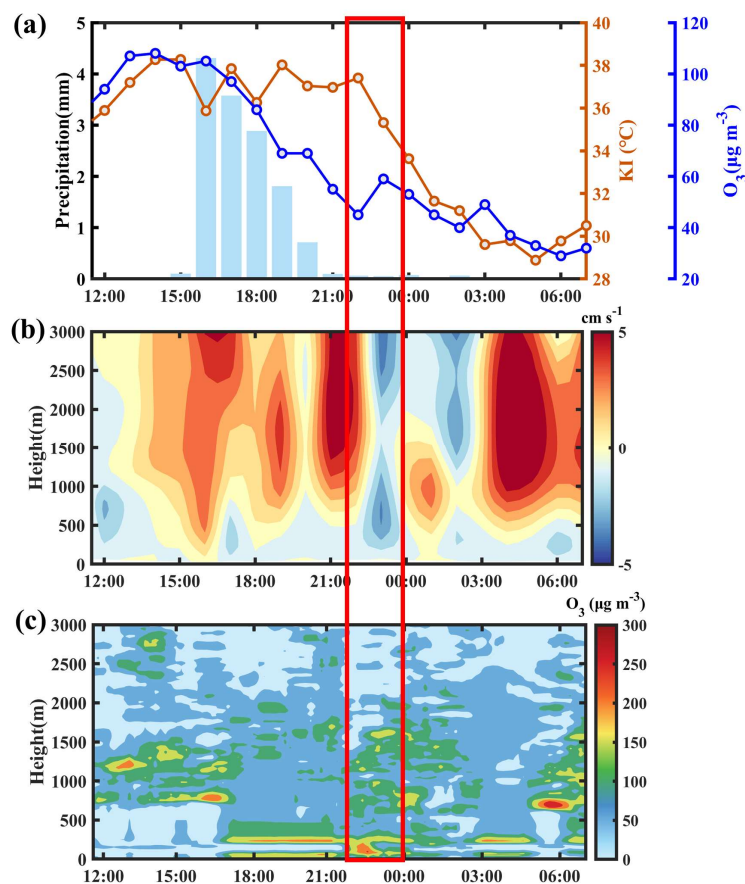
389
390 **Figure 7. Spatial distribution of annual average (a) NOI event frequency (points) and LLJs frequency**
391 **(contours), (b) MDA8 O₃ concentrations.**

392 **3.6 Causative analysis of NOI events: Convective storms trigger**

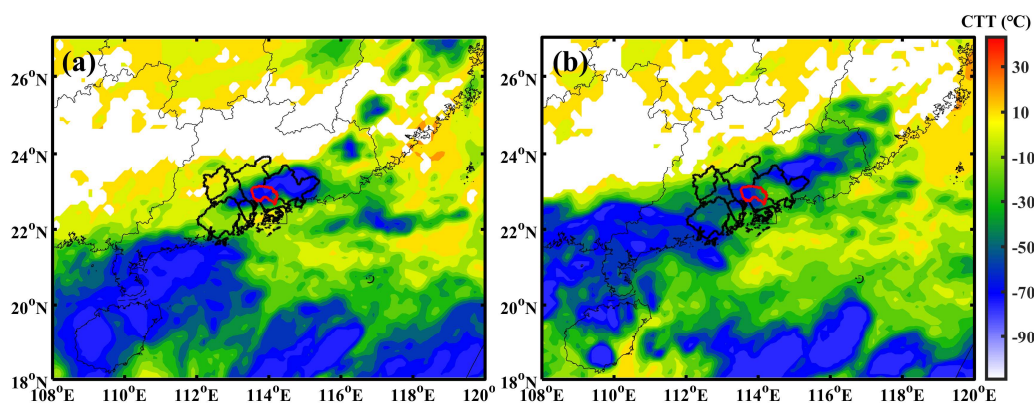
393 In order to elaborate the underlying atmospheric mechanisms for Conv-induced NOI events, a distinct
394 NOI event associated with Conv observed at the Nancheng Yuanling (NCYL) site in Dongguan on
395 September 3-4, 2019 is taken as a typical example for further discussion. The vertical O₃ profile data
396 observed at the Dongguan superstation are used to represent the vertical O₃ distribution at the NCYL site
397 during this NOI event, since the distance between these two stations is only 3 km.

398 The KI remains above 36 °C (Fig. 8a) and the vertical velocity show continuous updraft trends at 1-3 km
399 altitude from 14:00 to 23:00 (Fig. 8b), indicating a high possibility of convection. Although the
400 magnitude of vertical velocity was relatively low, it also has been found in previous studies (Ploeger et
401 al., 2021). In addition, the spatial distribution of CTT show that the CTT value at 18:00 over Dongguan
402 was around -70 °C (Fig. 9a), which was lower than the criterion (-35 °C) for the happening of convection
403 process. The results of KI, the vertical velocity and the CTT indicate that the possibility of a convection
404 process is high. Thus, the precipitation occurs during 16:00-20:00 was a convective precipitation. The
405 effect of rainfall on O₃ removal is relatively small after sunset, because wet deposition of O₃ occurs
406 through the removal of the precursors HNO₃ and H₂O₂ by water vapor under solar radiation, which is
407 indirect and rather peripheral, and the effect of heterogeneous processes on O₃ removal is weak (Jacob,
408 2000; Awang et al., 2015; Zhu et al., 2020). Therefore, the unconsumed O₃ remains stable in the RL. As
409 illustrated in Fig. 8c, higher O₃ concentrations are found in the RL after 18:00, reaching around 200 µg
410 m⁻³. At 21:00, a strong updraft suddenly appears above 1.5 km (Fig. 8b) and a distinct area of low CTT

411 values (around $-66\text{ }^{\circ}\text{C}$) could be observed over Dongguan (Fig. 9b), confirming the happening of updraft.
 412 The updraft subsequently caused a strong compensating downdraft below 1 km at 22:00-23:00 (Fig. 8b).
 413 The downdraft then breaks through the stable nocturnal boundary layer and transports O_3 from the RL to
 414 the surface (Fig. 8c). Hence, an NOI event occurs at 23:00, with O_3 concentrations increased from $45\text{ }\mu\text{g m}^{-3}$
 415 m^{-3} at 22:00 to $59\text{ }\mu\text{g m}^{-3}$ at 23:00. Although the modeled downdraft occurred at 22:00-23:00 (Fig. 8b)
 416 was around half an hour later than the observed O_3 intrusion into the nocturnal boundary layer (Fig. 8c)
 417 due to the model errors, the modeled results can still generally capture the occurrence of convection
 418 processes.



419
 420 **Figure 8. (a) Hourly variations of KI (brown line), O_3 concentrations (blue line), and hourly precipitation**
 421 **amount (blue bar), (b) vertical wind velocity, with positive and negative values related to updrafts and**
 422 **downdrafts, (c) vertical profile of O_3 concentrations at the NCYL site in Dongguan on September 3-4, 2019.**



423

424 **Figure 9. Spatial distribution of cloud-top temperature (CTT) at (a) 18:00 and (b) 21:00 LT on September 3,**
 425 **2019.**

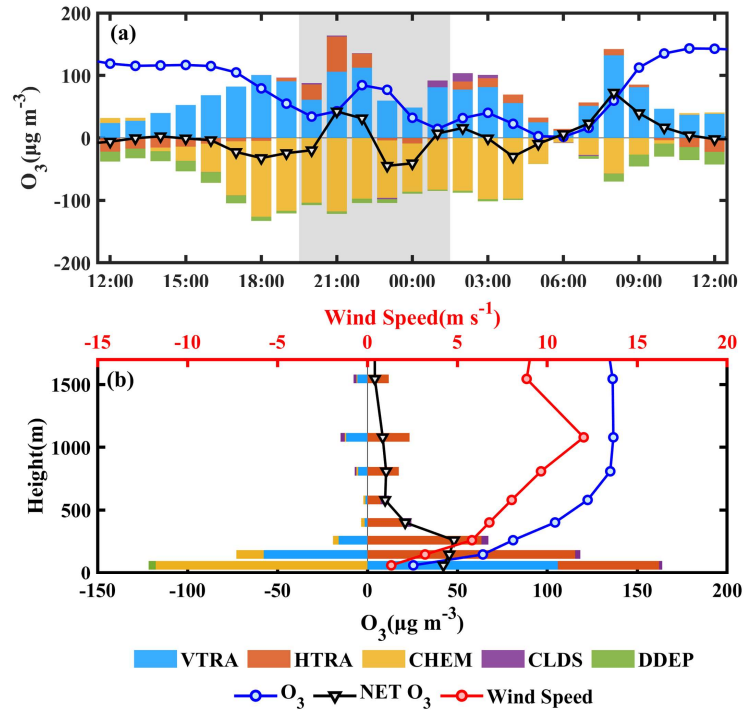
426 **3.7 Causative analysis of NOI events: LLJs trigger**

427 Another typical NOI event induced by LLJs occurred at the NCYL site in Dongguan on September 13-
 428 14, 2017 was simulated using the WRF-CMAQ-IPR model because no vertical profiles of wind speed
 429 were observed. Model performance was first evaluated. Comparisons of the simulated meteorological
 430 parameters with observations for the 9 sites in the PRD region during September 8-14, 2017 are shown
 431 in Fig. S3, with statistical indices reported in Table S6. The results show that WS10 was reasonably well
 432 simulated, as the regional average of MB, RMSE, and IoA met the EPA criteria mentioned in section 2.6.
 433 The simulated regional average of RH and T2 were slightly overestimated (MB=-6.1%) and
 434 underestimated (MB=1.4 °C), respectively, while they performed well at the Dongguan site, where both
 435 the MB (RH=-1.0%, T2=0.5 °C) and IoA (both are 1.0) met the EPA criteria. The MB of the simulated
 436 WS10 at the Dongguan site was slightly underestimated (MB=0.7 m s⁻¹), with the RMSE and IoA met
 437 the EPA criteria.

438 Comparisons of simulated O₃ with hourly observations during September 8-14, 2017 are shown in Fig.
 439 S3, with statistical indices reported in Table S7. The simulated O₃ showed a good performance in the
 440 PRD region, with NMB (-4.8%), NME (17.7%) and r (1.0) all met the MEP criteria, while it was slightly
 441 underestimated at the NCYL site in Dongguan but still met the MEP criteria (NMB=-12.7%,
 442 NME=29.8%, r=0.8). Therefore, the simulation results of the meteorological parameters and O₃ are
 443 reasonable and reliable for further analysis.

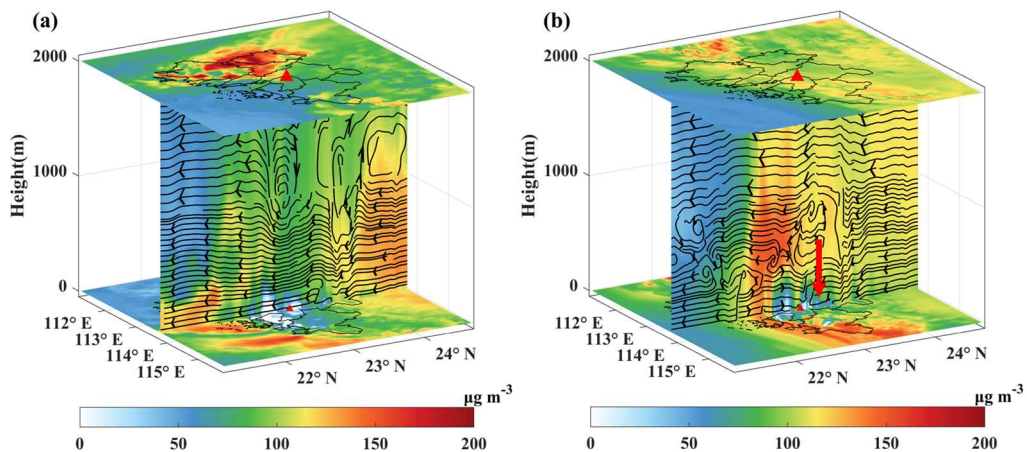
444 Figure 10a shows the time series of simulated O₃ concentrations and the contributions of different
 445 processes to surface O₃ concentrations at the NCYL site in Dongguan during September 13-14, 2017.

446 During the night on September 13, the O₃ concentration increased from 45 μg m⁻³ at 21:00, reached the
447 peak of 85 μg m⁻³ around 22:00 and dropped to 15 μg m⁻³ at 00:00. Before 21:00, the magnitude of the
448 negative contribution of chemical processes to O₃ was greater than the positive contribution of vertical
449 transport, resulting in net O₃ depletion. This suggests that gas-phase chemistry processes such as NO
450 titration are the main pathway for O₃ loss at night. At 21:00, the vertical and horizontal transport
451 contribution increased abruptly by 48 μg m⁻³ and 27 μg m⁻³, respectively, while the chemical depletion
452 remained constant. At this point, the net O₃ concentration turned from loss to production (51 μg m⁻³). In
453 terms of vertical distribution (Fig. 10b), a positive contribution of both vertical and horizontal transport
454 can be found at the surface, while vertical transport became negative in the upper layers and horizontal
455 transport remained positive, indicating the occurrence of a downdraft. In addition, the wind profile
456 showed a typical LLJs characteristic (Fig. 10b), with a maximum wind speed of about 12 m s⁻¹ at 1 km
457 altitude and a wind speed difference of more than 3 m s⁻¹ above and below. Figure 11 further presents
458 the process of vertical transport during an NOI event. Compared to normal days, the nocturnal boundary
459 layer during the NOI event was more unstable and turbulent, with significant upward and downward
460 transport. At around 1 km, there was a straight stream over the NCYL site during the NOI event (Fig.
461 11b). This suggested that LLJs broke the stable structure between the nocturnal boundary layer and the
462 RL and enhanced the strength of turbulence (Caputi et al., 2019). The LLJs-induced turbulence promoted
463 mixing between the upper and lower layers and continuously transported O₃ from the upper layer to the
464 surface, causing an unusual surge in O₃ at the surface and leading to an NOI event. As a result, the LLJs
465 process contributed as much as 40 μg m⁻³ O₃ from the upper layer to the surface during this NOI event.



466

467 **Figure 10. Contribution of individual processes to (a) hourly O_3 concentration near the surface during**
 468 **September 13-14, 2017 and (b) vertical O_3 concentration at 21:00 on September 13, 2017. VTRA: vertical**
 469 **transport, the net effect of vertical advection and diffusion; HTRA: horizontal transport, the net effect of**
 470 **horizontal advection and diffusion; CHEM: gas-phase chemistry; CLDS: cloud processes; DDEP: dry**
 471 **deposition; NET: the net change in O_3 due to all atmospheric processes.**



472

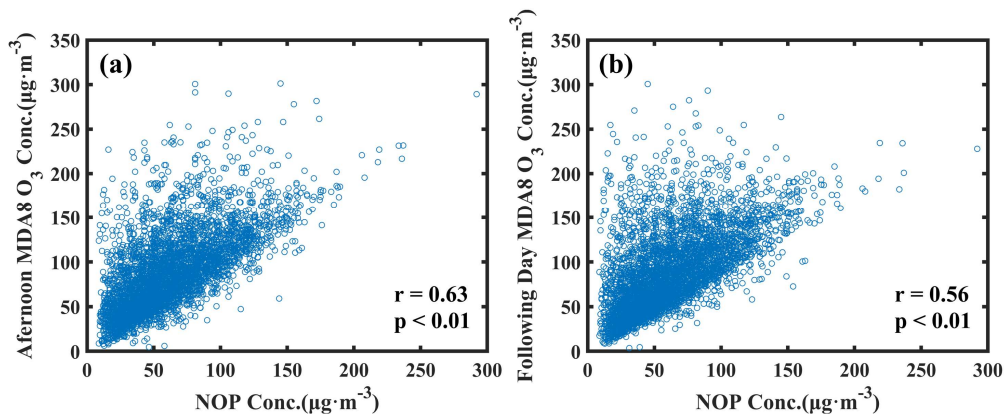
473 **Figure 11. Vertical profiles of O_3 concentrations at 21:00 during (a) a normal day (September 12, 2017) and**
 474 **(b) an NOI event (13 September 2017). Red triangles represent the NCYL site in Dongguan, contours**
 475 **represent O_3 concentrations ($\mu g m^{-3}$), and black lines and arrows indicate vertical airflow and its direction.**

476 3.8 Comparison and Prospects

477 Zhu et al. (2020) identified an NOI event frequency of 16-19 $d yr^{-1}$ in the summers from 2014 to 2015 in
 478 Beijing, China, with nocturnal O_3 maxima ranging from 45 to 85 $\mu g m^{-3}$, which is comparable to our

479 result of NOI frequency ($14 \pm 3 \text{ d yr}^{-1}$) and slightly higher than our NOP concentration ($44 \pm 7 \mu\text{g m}^{-3}$)
480 in summer. Sousa et al. (2011) analyzed nocturnal O₃ maxima events (maxima higher than the average
481 nocturnal O₃ concentration of $10 \mu\text{g m}^{-3}$) during 2005-2007 in northern Portugal and found that the
482 frequencies of nocturnal O₃ maxima were between 40% and 50% in urban areas and 15% in rural areas,
483 which is higher than our NOI frequency (53 d yr^{-1} , 14.5%). Other studies focusing on short-term nocturnal
484 O₃ maxima cases found that NOP concentrations were 30-50 $\mu\text{g m}^{-3}$ in the Lower Fraser Valley, British
485 Columbia Canada (Salmond and McKendry, 2002), 20-60 $\mu\text{g m}^{-3}$ in Senegal (Grant et al., 2008), and 40-
486 80 $\mu\text{g m}^{-3}$ in North America (Kuang et al., 2011; Hu et al., 2013; Sullivan et al., 2017), values that are
487 comparable to our results ($58 \pm 11 \mu\text{g m}^{-3}$).

488 Our study emphasizes the importance of meteorological processes as well as daytime O₃ concentration
489 in the occurrence of NOI events, implying that higher NOP may occur during a severe daytime O₃
490 pollution period under the effect of meteorological processes. The occurrence of NOI events is likely to
491 impact the O₃ levels on the following day, which makes O₃ prevention more complex and challenging
492 (Ravishankara, 2009; Sullivan et al., 2017). However, the relationship between NOI events and the
493 following daytime O₃ pollution remains unclear and controversial. Kuang et al. (2011) and Sullivan et al.
494 (2017) revealed that NOI events led to a higher increasing rate of O₃ and worse air quality on the
495 following day, while Klein et al. (2019) and Caputi et al. (2019) observed lower O₃ levels during the
496 daytime following NOI events. To further explore the relationship between the daytime MDA8 O₃ and
497 nighttime NOP in the PRD region, we display the correlation between the MDA8 O₃ and the following
498 night's NOP (shorthand MDA8-NOP) (Fig. 12a) and the NOP and the following MDA8 O₃ (shorthand
499 NOP-MDA8) (Fig. 12b), respectively. The results show that MDA8 O₃ was positively correlated with
500 NOP with a correlation coefficient of 0.63 ($p < 0.01$) and 0.56 ($p < 0.01$) for MDA8-NOP and NOP-MDA8,
501 respectively, suggesting an interplay between daytime O₃ and NOP in the PRD region.



502

503 **Figure 12. Correlation between (a) the afternoon MDA8 O₃ concentration and the following night's NOP**
 504 **concentration and (b) the NOP concentration and the following afternoon MDA8 O₃ concentration.**

505 **4 Conclusion**

506 In this study, based on in-situ O₃ concentrations, observed vertical profiles of O₃, ERA5 reanalysis
 507 datasets, and the WRF-CMAQ model, the spatial and temporal characteristics of NOI events are
 508 comprehensively presented and the role of vertical transport in NOI events in the PRD region from 2006
 509 to 2019 is further quantified.

510 The average annual frequency of NOI events is estimated to be $53 \pm 16 \text{ d yr}^{-1}$ from 2006 to 2019, with
 511 an annual average of $58 \pm 11 \mu\text{g m}^{-3}$ for nocturnal O₃ peak (NOP). LLJs are the dominant factors causing
 512 NOI events (61%), followed by the combination of LLJs and Conv (LLJs+Conv) with a value of 16%.
 513 The high correlation between NOI events and the frequency of LLJs in the annual trend ($r=0.89, p<0.01$)
 514 supports the important influence of LLJs on the occurrence of NOI events. Although the contribution of
 515 Conv to NOI events is relatively small, Conv-induced NOI events steadily increased at a rate of 0.26 d
 516 yr^{-1} during this 14-year period due to the impact of urbanization. Moreover, the significant positive
 517 correlation between NOP and maximum daily 8-h average (MDA8) O₃ in annual ($r=0.88, p<0.01$) and
 518 seasonal trends ($r=0.80, p<0.01$) and a higher NOI frequency (60%) during the first half of the night
 519 imply that daytime O₃ concentrations are also an important factor influencing the formation of NOI
 520 events.

521 Two typical NOI events caused by LLJs and Conv, respectively, further demonstrate that downdrafts
 522 from enhanced turbulence are the direct cause of NOI events, as these can transport O₃ from the RL to

523 the surface. The difference is that LLJs induce downdrafts by a fast-moving air mass enhancing shear
524 below, whereas Conv induce downdraft by compensating downdrafts.

525 This study emphasizes the importance of vertical transport induced by LLJs and Conv, and daytime O₃
526 concentration in the formation of NOI events and highlights the key role of vertical transport in linking
527 daytime and nighttime O₃ pollution. This study not only provides a new perspective and better
528 understanding to reconceptualize the role of meteorology in daytime and nighttime O₃ pollution, but also
529 provides a reference for other regions with ground-level O₃ pollution.

530

531 *Data availability.* In-situ hourly O₃ concentrations at 16 stations across the PRD region from 2006 to
532 2019 can be downloaded from <http://113.108.142.147:20047>; the observed hourly meteorological data
533 at the 9 sites across the PRD region can be downloaded from <http://www.cma.gov.cn/>; the ERA5
534 reanalysis dataset can be downloaded from <https://cds.climate.copernicus.eu/cdsapp#!/home>; and the
535 vertical O₃ profile data available upon request.

536

537 *Author contributions.* YW and WC designed the research. YW did the data analysis and simulation
538 work and prepared the draft with support and editing from WC. YY and QX contributed to data analysis.
539 SJ and XW contributed to paper revision.

540

541 *Competing interests.* The authors declare that they have no conflict of interest.

542

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552 **References**

- 553 Ai, Y., Li, W., Meng, Z., & Li, J.: Life cycle characteristics of MCSs in middle east China tracked by
554 geostationary satellite and precipitation estimates, *Mon. Weather Rev.*, 144(7), 2517-2530,
555 <https://doi.org/10.1175/MWR-D-15-0197.1>, 2016.
- 556 Awang, N. R., Ramli, N. A., Yahaya, A. S., and Elbayoumi, M.: High nighttime ground-level ozone
557 concentrations in Kemaman: NO and NO₂ concentrations attributions, *Aerosol Air Qual. Res.*, 15, 1357-
558 1366, <https://doi.org/10.4209/aaqr.2015.01.0031>, 2015.
- 559 Banta, R. M., Newsom, R. K., Lundquist, J. K., Pichugina, Y. L., Coulter, R. L., and Mahrt, L.: Nocturnal
560 low-level jet characteristics over Kansas during CASES-99, *Boundary-Layer Meteorol.*, 105, 221-252,
561 <https://doi.org/10.1023/A:1019992330866>, 2002.
- 562 Brown, S. S., Neuman, J. A., Ryerson, T. B., Trainer, M., Dubé, W. P., Holloway, J. S., ... &
563 Ravishankara, A. R.: Nocturnal odd-oxygen budget and its implications for ozone loss in the lower
564 troposphere, *Geophys. Res. Lett.*, 33(8), <https://doi.org/10.1029/2006GL025900>, 2006.
- 565 Caird, M. A., Richards, J. H., and Donovan, L. A.: Nighttime stomatal conductance and transpiration in
566 C₃ and C₄ plants, *Plant Physiol.*, 143, 4-10, <https://doi.org/10.1104/pp.106.092940>, 2007.
- 567 Caputi, D. J., Faloona, I., Trousdell, J., Smoot, J., Falk, N., and Conley, S.: Residual layer ozone, mixing,
568 and the nocturnal jet in California's San Joaquin Valley, *Atmos. Chem. Phys.*, 19, 4721-4740,
569 <https://doi.org/10.5194/acp-19-4721-2019>, 2019.
- 570 Carlton, A. G., Bhave, P. V., Napelenok, S. L., Edney, E. O., Sarwar, G., Pinder, R. W., Pouliot, G. A.,
571 and Houyoux, M.: Model representation of secondary organic aerosol in CMAQv4.7, *Environ. Sci.*
572 *Technol.*, 44, 8553-8560, <https://doi.org/10.1021/es100636q>, 2010.
- 573 Carré, J., Gatimel, N., Moreau, J., Parinaud, J., and Leandri, R.: Influence of air quality on the results of
574 in vitro fertilization attempts: A retrospective study, *Eur. J. Obstet. Gynecol. Reprod. Biol.*, 210, 116-
575 122, <https://doi.org/10.1016/j.ejogrb.2016.12.012>, 2017.
- 576 Carter, W. P. L.: Development of the SAPRC-07 chemical mechanism, *Atmos. Environ.*, 44, 5324-5335,
577 <https://doi.org/10.1016/j.atmosenv.2010.01.026>, 2010.
- 578 Chen, F. and Dudhia, J.: Coupling an advanced land surface–hydrology model with the Penn State–
579 NCAR MM5 modeling system. Part I: Model implementation and sensitivity, *Mon. Weather Rev.*, 129,
580 569-585, [https://doi.org/10.1175/1520-0493\(2001\)129<0569:CAALSH>2.0.CO;2](https://doi.org/10.1175/1520-0493(2001)129<0569:CAALSH>2.0.CO;2), 2001.

581 Chen, X., Zhao, K., and Xue, M.: Spatial and temporal characteristics of warm season convection over
582 Pearl River Delta region, China, based on 3 years of operational radar data, *J. Geophys. Res. Atmos.*,
583 119, 12,447-465, <https://doi.org/10.1002/2014jd021965>, 2014.

584 Chen, X., Zhong, B., Huang, F., Wang, X., Sarkar, S., Jia, S., Deng, X., Chen, D., and Shao, M.: The
585 role of natural factors in constraining long-term tropospheric ozone trends over southern China, *Atmos.*
586 *Environ.*, 220, 117060, <https://doi.org/10.1016/j.atmosenv.2019.117060>, 2020.

587 Cirelli, D., Equiza, M. A., Liefvers, V. J., and Tyree, M. T.: *Populus* species from diverse habitats
588 maintain high night-time conductance under drought, *Tree Physiol.*, 36, 229-242,
589 <https://doi.org/10.1093/treephys/tpv092>, 2016.

590 Dias-Junior, C. Q., Dias, N. L., Fuentes, J. D., and Chamecki, M.: Convective storms and non-classical
591 low-level jets during high ozone level episodes in the Amazon region: An ARM/GOAMAZON case
592 study, *Atmos. Environ.*, 155, 199-209, <https://doi.org/10.1016/j.atmosenv.2017.02.006>, 2017.

593 Du, Y. and Chen, G.: Heavy rainfall associated with double low-level jets over southern China. Part II:
594 Convection initiation, *Mon. Weather Rev.*, 147, 543-565, <https://doi.org/10.1175/mwr-d-18-0102.1>,
595 2019.

596 Emmons, L. K., Walters, S., Hess, P. G., Lamarque, J. F., Pfister, G. G., Fillmore, D., Granier, C.,
597 Guenther, A., Kinnison, D., Laepple, T., Orlando, J., Tie, X., Tyndall, G., Wiedinmyer, C., Baughcum,
598 S. L., and Kloster, S.: Description and evaluation of the Model for Ozone and Related chemical Tracers,
599 version 4 (MOZART-4), *Geosci. Model. Dev.*, 3, 43-67, <https://doi.org/10.5194/gmd-3-43-2010>, 2010.

600 EPA.: Guidance on the use of models and other analyses for demonstrating attainment of air quality goals
601 for ozone, PM_{2.5}, and regional haze. [https://www.epa.gov/sites/default/files/2020-10/documents/final-](https://www.epa.gov/sites/default/files/2020-10/documents/final-03-pm-rh-guidance.pdf)
602 [03-pm-rh-guidance.pdf](https://www.epa.gov/sites/default/files/2020-10/documents/final-03-pm-rh-guidance.pdf), 2017.

603 Fast, J. D., & McCorcle, M. D.: A two-dimensional numerical sensitivity study of the Great Plains low-
604 level jet, *Mon. Weather Rev.*, 118(1), 151-164, [https://doi.org/10.1175/1520-](https://doi.org/10.1175/1520-0493(1990)118<0151:ATDNSS>2.0.CO;2)
605 [0493\(1990\)118<0151:ATDNSS>2.0.CO;2](https://doi.org/10.1175/1520-0493(1990)118<0151:ATDNSS>2.0.CO;2), 1990.

606 Fan, X., Xia, X., Chen, H., Zhu, Y., Li, J., Yang, H., & Luo, H.: Baseline of Surface and Column-
607 Integrated Aerosol Loadings in the Pearl River Delta Region, China, *Front. Environ. Sci.*, 574,
608 <https://doi.org/10.3389/fenvs.2022.893408>, 2022.

609 Feng, Y., Ning, M., Lei, Y., Sun, Y., Liu, W., & Wang, J.: Defending blue sky in China: Effectiveness
610 of the “Air Pollution Prevention and Control Action Plan” on air quality improvements from 2013 to
611 2017, *J. Environ. Management*, 252, 109603, <https://doi.org/10.1016/j.jenvman.2019.109603>, 2019.

612 George, J. J.: Weather forecasting for aeronautics. Academic press, [https://doi.org/10.1016/C2013-0-](https://doi.org/10.1016/C2013-0-12567-6)
613 12567-6, 1960.

614 Gong, D., Wang, H., Zhang, S., Wang, Y., Liu, S. C., Guo, H., Shao, M., He, C., Chen, D., He, L., Zhou,
615 L., Morawska, L., Zhang, Y., and Wang, B.: Low-level summertime isoprene observed at a forested
616 mountaintop site in southern China: Implications for strong regional atmospheric oxidative capacity,
617 *Atmos. Chem. Phys.*, 18, 14417-14432, <https://doi.org/10.5194/acp-18-14417-2018>, 2018.

618 Grant, D. D., Fuentes, J. D., DeLonge, M. S., Chan, S., Joseph, E., Kucera, P., Ndiaye, S. A., and Gaye,
619 A. T.: Ozone transport by mesoscale convective storms in western Senegal, *Atmos. Environ.*, 42, 7104-
620 7114, <https://doi.org/10.1016/j.atmosenv.2008.05.044>, 2008.

621 Grell, G. A. and Dévényi, D.: A generalized approach to parameterizing convection combining ensemble
622 and data assimilation techniques, *Geophys. Res. Lett.*, 29, 381-384,
623 <https://doi.org/10.1029/2002gl015311>, 2002.

624 Guenther, A., Karl, T., Harley, P., Wiedinmyer, C., Palmer, P. I., and Geron, C.: Estimates of global
625 terrestrial isoprene emissions using MEGAN (Model of Emissions of Gases and Aerosols from Nature),
626 *Atmos. Chem. Phys.*, 6, 3181-3210, <https://doi.org/10.5194/acp-6-3181-2006>, 2006.

627 Han, C., Liu, R., Luo, H., Li, G., Ma, S., Chen, J., and An, T.: Pollution profiles of volatile organic
628 compounds from different urban functional areas in Guangzhou China based on GC/MS and PTR-TOF-
629 MS: Atmospheric environmental implications, *Atmos. Environ.*, 214, 116843,
630 <https://doi.org/10.1016/j.atmosenv.2019.116843>, 2019.

631 He, Y., Wang, H., Wang, H., Xu, X., Li, Y., and Fan, S.: Meteorology and topographic influences on
632 nocturnal ozone increase during the summertime over Shaoguan, China, *Atmos. Environ.*, 256, 118459,
633 <https://doi.org/10.1016/j.atmosenv.2021.118459>, 2021.

634 Hersbach, H., Bell, B., Berrisford, P., Hirahara, S., Horányi, A., Muñoz-Sabater, J., Nicolas, J., Peubey,
635 C., Radu, R., Schepers, D., Simmons, A., Soci, C., Abdalla, S., Abellan, X., Balsamo, G., Bechtold, P.,
636 Biavati, G., Bidlot, J., Bonavita, M., Chiara, G., Dahlgren, P., Dee, D., Diamantakis, M., Dragani, R.,
637 Flemming, J., Forbes, R., Fuentes, M., Geer, A., Haimberger, L., Healy, S., Hogan, R. J., Hólm, E.,
638 Janisková, M., Keeley, S., Laloyaux, P., Lopez, P., Lupu, C., Radnoti, G., Rosnay, P., Rozum, I.,

639 Vamborg, F., Villaume, S., and Thépaut, J. N.: The ERA5 global reanalysis, *Q. J. R. Meteorol. Soc.*, 146,
640 1999-2049, <https://doi.org/10.1002/qj.3803>, 2020.

641 HKEPD.: Pearl River Delta Regional Air Quality Monitoring Report for Year 2017.
642 https://www.epd.gov.hk/epd/sites/default/files/epd/english/resources_pub/publications/files/PRD_2017
643 [_report_en.pdf](#), 2017.

644 Hodges, D. and Pu, Z. X.: Characteristics and variations of low-level jets and environmental factors
645 associated with summer precipitation extremes over the Great Plains, *J. Clim.*, 32, 5123-5144,
646 <https://doi.org/10.1175/jcli-d-18-0553.1>, 2019.

647 Hu, X. M., Klein, P. M., Xue, M., Zhang, F., Doughty, D. C., Forkel, R., Joseph, E., and Fuentes, J. D.:
648 Impact of the vertical mixing induced by low-level jets on boundary layer ozone concentration, *Atmos.*
649 *Environ.*, 70, 123-130, <https://doi.org/10.1016/j.atmosenv.2012.12.046>, 2013.

650 Iacono, M. J., Delamere, J. S., Mlawer, E. J., Shephard, M. W., Clough, S. A., and Collins, W. D.:
651 Radiative forcing by long-lived greenhouse gases: Calculations with the AER radiative transfer models,
652 *J. Geophys. Res. Atmos.*, 113, <https://doi.org/10.1029/2008jd009944>, 2008.

653 Jacob, D. J.: Heterogeneous chemistry and tropospheric ozone, *Atmos. Environ.*, 34, 2131-2159,
654 [https://doi.org/10.1016/S1352-2310\(99\)00462-8](https://doi.org/10.1016/S1352-2310(99)00462-8), 2000.

655 Jain, S. L., Arya, B. C., Kumar, A., Ghude, S. D., and Kulkarni, P. S.: Observational study of surface
656 ozone at New Delhi, India, *Int. J. Remote Sens.*, 26, 3515-3524,
657 <https://doi.org/10.1080/01431160500076616>, 2007.

658 Jiménez, P., Parra, R., and Baldasano, J. M.: Influence of initial and boundary conditions for ozone
659 modeling in very complex terrains: A case study in the northeastern Iberian Peninsula, *Environ. Model.*
660 *Softw.*, 22, 1294-1306, <https://doi.org/10.1016/j.envsoft.2006.08.004>, 2007.

661 Johnson, D. L.: A stability analysis of AVE-IV severe weather sounding, NASA Tech. Paper, 2045-2126,
662 1982.

663 Kallistratova M A.: Investigation of low-level-jets over rural and urban areas using two sodars, *IOP Conf.*
664 *Ser.: Earth and Environ. Sci.*, 1(1): 012040, <https://doi.org/10.1088/1755-1315/1/1/012040>, 2008.

665 Klein, A., Ravetta, F., Thomas, J. L., Ancellet, G., Augustin, P., Wilson, R., Dieudonné, E., Fourmentin,
666 M., Delbarre, H., and Pelon, J.: Influence of vertical mixing and nighttime transport on surface ozone
667 variability in the morning in Paris and the surrounding region, *Atmos. Environ.*, 197, 92-102,
668 <https://doi.org/10.1016/j.atmosenv.2018.10.009>, 2019.

669 Kleinman, L., Lee, Y. N., Springston, S. R., Nunnermacker, L., Zhou, X., Brown, R., Hallock, K., Klotz,
670 P., Leahy, D., and Lee, J. H.: Ozone formation at a rural site in the southeastern United States, *J. Geophys.*
671 *Res. Atmos.*, 99, 3469-3482, <https://doi.org/10.1029/93JD02991>, 1994.

672 Kuang, S., Newchurch, M. J., Burris, J., Wang, L., Buckley, P. I., Johnson, S., Knupp, K., Huang, G.,
673 Phillips, D., and Cantrell, W.: Nocturnal ozone enhancement in the lower troposphere observed by lidar,
674 *Atmos. Environ.*, 45, 6078-6084, <https://doi.org/10.1016/j.atmosenv.2011.07.038>, 2011.

675 Kulkarni, P. S., Bortoli, D., and Silva, A. M.: Nocturnal surface ozone enhancement and trend over urban
676 and suburban sites in Portugal, *Atmos. Environ.*, 71, 251-259,
677 <https://doi.org/10.1016/j.atmosenv.2013.01.051>, 2013.

678 Kulkarni, P. S., Bortoli, D., Silva, A. M., and Reeves, C. E.: Enhancements in nocturnal surface ozone
679 at urban sites in the UK, *Environ. Sci. Pollut. Res.*, 22, 20295-20305, [https://doi.org/10.1007/s11356-](https://doi.org/10.1007/s11356-015-5259-z)
680 [015-5259-z](https://doi.org/10.1007/s11356-015-5259-z), 2015.

681 Kurt, O. K., Zhang, J., and Pinkerton, K. E.: Pulmonary health effects of air pollution, *Curr. Opin. Pulm.*
682 *Med.*, 22, 138-143, <https://doi.org/10.1097/MCP.0000000000000248>, 2016.

683 Li, M., Liu, H., Geng, G., Hong, C., Liu, F., Song, Y., ... & He, K.: Anthropogenic emission inventories
684 in China: a review, *Natl. Sci. Rev.*, 4(6), 834-866, <https://doi.org/10.1093/nsr/nwx150>, 2017.

685 Li, K., Jacob, D. J., Liao, H., Zhu, J., Shah, V., Shen, L., ... & Zhai, S.: A two-pollutant strategy for
686 improving ozone and particulate air quality in China, *Nat. Geosci.*, 12(11), 906-910,
687 <https://doi.org/10.1038/s41561-019-0464-x>, 2019.

688 Li, Y., Wang, W., Chang, M., and Wang, X.: Impacts of urbanization on extreme precipitation in the
689 Guangdong-Hong Kong-Macau Greater Bay Area, *Urban Clim.*, 38, 100904,
690 <https://doi.org/10.1016/j.uclim.2021.100904>, 2021.

691 Li, X. B., Yuan, B., Parrish, D. D., Chen, D., Song, Y., Yang, S., Liu, Z., and Shao, M.: Long-term trend
692 of ozone in southern China reveals future mitigation strategy for air pollution, *Atmos. Environ.*, 269,
693 118869, <https://doi.org/10.1016/j.atmosenv.2021.118869>, 2022.

694 Liao, Z., Ling, Z., Gao, M., Sun, J., Zhao, W., Ma, P., Quan, J., and Fan, S.: Tropospheric ozone
695 variability over Hong Kong based on recent 20 years (2000–2019) ozonesonde observation, *J. Geophys.*
696 *Res. Atmos.*, 126, e2020JD033054, <https://doi.org/10.1029/2020jd033054>, 2021.

697 Lin, Y., Farley, R. D., and Orville, H. D.: Bulk parameterization of the snow field in a cloud model, J.
698 Appl. Meteorol. Climatol., 22, 1065-1092, <https://doi.org/10.1175/1520->
699 0450(1983)022<1065:BPOTSF>2.0.CO;2, 1983.

700 Liu, H., Zhang, M., and Han, X.: A review of surface ozone source apportionment in China, Atmos.
701 Ocean. Sci. Lett., 13, 470-484, <https://doi.org/10.1080/16742834.2020.1768025>, 2020.

702 Liu, X. H., Zhang, Y., Xing, J., Zhang, Q., Wang, K., Streets, D. G., Jang, C., Wang, W., and Hao, J. M.:
703 Understanding of regional air pollution over China using CMAQ, part II. Process analysis and sensitivity
704 of ozone and particulate matter to precursor emissions, Atmos. Environ., 44, 3719-3727,
705 <https://doi.org/10.1016/j.atmosenv.2010.03.036>, 2010.

706 Lu, K., Zhang, Y., Su, H., Shao, M., Zeng, L., Zhong, L., Xiang, Y., Chang, C., Chou, C. K. C., and
707 Wahner, A.: Regional ozone pollution and key controlling factors of photochemical ozone production in
708 Pearl River Delta during summer time, Sci. China Chem., 53, 651-663, <https://doi.org/10.1007/s11426->
709 010-0055-6, 2010.

710 Lu, X., Zhang, L., Wang, X., Gao, M., Li, K., Zhang, Y., ... & Zhang, Y.: Rapid increases in warm-
711 season surface ozone and resulting health impact in China since 2013, Environ. Sci. Technol. Lett., 7(4),
712 240-247, <https://doi.org/10.1021/acs.estlett.0c00171>, 2020.

713 Mann, H. B.: Nonparametric tests against trend, Econometrica: Journal of the econometric society, 245-
714 259, <https://doi.org/10.2307/1907187>, 1945.

715 Monin, A. S. and Obukhov, A. M.: Basic laws of turbulent mixing in the surface layer of the atmosphere,
716 Contrib. Geophys. Inst. Acad. Sci. USSR, 151, e187,
717 https://gibbs.science/efd/handouts/monin_obukhov_1954.pdf, 1954.

718 Mccorcle, M. D.: Simulation of surface-moisture effects on the Great Plains low-level jet, Mon. Weather
719 Rev., 116(9), 1705-1720, [https://doi.org/10.1175/1520-0493\(1988\)116<1705:SOSMEO>2.0.CO;2](https://doi.org/10.1175/1520-0493(1988)116<1705:SOSMEO>2.0.CO;2),
720 1988.

721 MEP.: The ministry of environmental protection of China. Guidelines for selection of ambient air quality
722 models (Trial).
723 https://english.mee.gov.cn/Resources/standards/Air_Environment/quality_standard1/201605/t2016051
724 1_337502.shtml, 2015.

725 Ma, Z., Xu, J., Quan, W., Zhang, Z., Lin, W., and Xu, X.: Significant increase of surface ozone at a rural
726 site, north of eastern China, *Atmos. Chem. Phys.*, 16, 3969-3977, [https://doi.org/10.5194/acp-16-3969-](https://doi.org/10.5194/acp-16-3969-2016)
727 2016, 2016.

728 Marelle, L., Myhre, G., Steensen, B. M., Hodnebrog, Ø., Alterskjær, K., and Sillmann, J.: Urbanization
729 in megacities increases the frequency of extreme precipitation events far more than their intensity,
730 *Environ. Res. Lett.*, 15, 124072, <https://doi.org/10.1088/1748-9326/abcc8f>, 2020.

731 Mousavinezhad, S., Choi, Y., Pouyaei, A., Ghahremanloo, M., & Nelson, D. L.: A comprehensive
732 investigation of surface ozone pollution in China, 2015–2019: Separating the contributions from
733 meteorology and precursor emissions, *Atmos. Res.*, 257, 105599,
734 <https://doi.org/10.1016/j.atmosres.2021.105599>, 2021.

735 Mao, J., Yan, F., Zheng, L., You, Y., Wang, W., Jia, S., Liao, W., Wang, X., and Chen, W.: Ozone
736 control strategies for local formation- and regional transport-dominant scenarios in a manufacturing city
737 in southern China, *Sci. Total Environ.*, 813, 151883, <https://doi.org/10.1016/j.scitotenv.2021.151883>,
738 2022.

739 Nair, P. R., Chand, D., Lal, S., Modh, K. S., Naja, M., Parameswaran, K., Ravindran, S., and
740 Venkataramani, S.: Temporal variations in surface ozone at Thumba (8.6 °N, 77 °E) - a tropical coastal
741 site in India, *Atmos. Environ.*, 36, 603-610, [https://doi.org/10.1016/S1352-2310\(01\)00527-1](https://doi.org/10.1016/S1352-2310(01)00527-1), 2002.

742 Nakanishi, M. and Niino, H.: An improved Mellor–Yamada Level-3 model: Its numerical stability and
743 application to a regional prediction of advection fog, *Boundary-Layer Meteorol.*, 119, 397-407,
744 <https://doi.org/10.1007/s10546-005-9030-8>, 2006.

745 Olauson, J.: ERA5: The new champion of wind power modelling?, *Renew. Energ.*, 126, 322-331,
746 <https://doi.org/10.1016/j.renene.2018.03.056>, 2018.

747 Nikolic, J., Zhong, S., Pei, L., Bian, X., Heilman, W. E., & Charney, J. J.: Sensitivity of low-level jets to
748 land-use and land-cover change over the continental US, *Atmosphere*, 10(4), 174,
749 <https://doi.org/10.3390/atmos10040174>, 2019.

750 Prtenjak, M. T., Jeričević, A., Klaić, Z. B., Alebić-Juretić, A., and Bulić, I. H.: Atmospheric dynamics
751 and elevated ozone concentrations in the northern Adriatic, *Meteorol. Appl.*, 20, 482-496,
752 <https://doi.org/10.1002/met.1312>, 2013.

753 Ploeger, F., Diallo, M., Charlesworth, E., Konopka, P., Legras, B., Laube, J. C., ... & Riese, M.: The
754 stratospheric Brewer–Dobson circulation inferred from age of air in the ERA5 reanalysis, *Atmos. Chem.*
755 *Phys.*, 21(11), 8393-8412, <https://doi.org/10.5194/acp-21-8393-2021>, 2021.

756 Ravishankara, A. R.: Are chlorine atoms significant tropospheric free radicals?, *Proc. Natl. Acad. Sci. U.*
757 *S. A.*, 106, 13639-13640, <https://doi.org/10.1073/pnas.0907089106>, 2009.

758 Salmond, J. A. and McKendry, I. G.: Secondary ozone maxima in a very stable nocturnal boundary layer:
759 Observations from the Lower Fraser Valley, BC, *Atmos. Environ.*, 36, 5771-5782,
760 [https://doi.org/10.1016/S1352-2310\(02\)00698-2](https://doi.org/10.1016/S1352-2310(02)00698-2), 2002.

761 Seibert, P., Feldmann, H., Neiningner, B., Baumle, M., and Trickl, T.: South foehn and ozone in the
762 Eastern Alps - case study and climatological aspects, *Atmos. Environ.*, 34, 1379-1394,
763 [https://doi.org/10.1016/S1352-2310\(99\)00439-2](https://doi.org/10.1016/S1352-2310(99)00439-2), 2000.

764 Sen, P. K.: Estimates of the regression coefficient based on Kendall's tau, *J. Am. Stat. Assoc.*, 63(324),
765 1379-1389, <https://doi.org/10.2307/2285891>, 1968.

766 Seto, K. C., Guneralp, B., and Hutyra, L. R.: Global forecasts of urban expansion to 2030 and direct
767 impacts on biodiversity and carbon pools, *Proc. Natl. Acad. Sci. U. S. A.*, 109, 16083-16088,
768 <https://doi.org/10.1073/pnas.1211658109>, 2012.

769 Shen, J., Zhang, Y., Wang, X., Li, J., Chen, H., Liu, R., Zhong, L., Jiang, M., Yue, D., Chen, D., and Lv,
770 W.: An ozone episode over the Pearl River Delta in October 2008, *Atmos. Environ.*, 122, 852-863,
771 <https://doi.org/10.1016/j.atmosenv.2015.03.036>, 2015.

772 Shith, S., Awang, N. R., Latif, M. T., and Ramli, N. A.: Fluctuations in nighttime ground-level ozone
773 concentrations during haze events in Malaysia, *Air Qual. Atmos. Hlth.*, 14, 19-26,
774 <https://doi.org/10.1007/s11869-020-00908-5>, 2021.

775 Sousa, S. I. V., Alvim-Ferraz, M. C. M., and Martins, F. G.: Identification and origin of nocturnal ozone
776 maxima at urban and rural areas of northern Portugal – Influence of horizontal transport, *Atmos. Environ.*,
777 45, 942-956, <https://doi.org/10.1016/j.atmosenv.2010.11.008>, 2011.

778 Sugimoto, N., Nishizawa, T., Liu, X., Matsui, I., Shimizu, A., Zhang, Y., ... & Liu, J.: Continuous
779 observations of aerosol profiles with a two-wavelength Mie-scattering lidar in Guangzhou in PRD2006,
780 *J. Appl. Meteorol. Clim.*, 48(9), 1822-1830, <https://doi.org/10.1175/2009JAMC2089.1>, 2009.

781 Stull, R. B.: *An introduction to boundary layer meteorology*, Springer Science & Business Media, 1988.

782 Sullivan, J. T., Rabenhorst, S. D., Dreessen, J., McGee, T. J., Delgado, R., Twigg, L., and Sunnicht, G.:
783 Lidar observations revealing transport of O₃ in the presence of a nocturnal low-level jet: Regional
784 implications for “next-day” pollution, *Atmos. Environ.*, 158, 160-171,
785 <https://doi.org/10.1016/j.atmosenv.2017.03.039>, 2017.

786 Tong, N. Y. O. and Leung, D. Y. C.: Ozone diurnal characteristics in areas with different urbanisations,
787 *Int. J. Environ. Pollut.*, 49, 100-124, <https://doi.org/10.1504/Ijep.2012.049771>, 2012.

788 Trier, S. B., Wilson, J. W., Ahijevych, D. A., and Sobash, R. A.: Mesoscale vertical motions near
789 nocturnal convection initiation in PECAN, *Mon. Weather Rev.*, 145, 2919-2941,
790 <https://doi.org/10.1175/mwr-d-17-0005.1>, 2017.

791 Udina, M., Soler, M. R., Olid, M., Jiménez-Estevé, B., and Bech, J.: Pollutant vertical mixing in the
792 nocturnal boundary layer enhanced by density currents and low-level jets: Two representative case
793 studies, *Boundary-Layer Meteorol.*, 174, 203-230, <https://doi.org/10.1007/s10546-019-00483-y>, 2019.

794 Wang, T., Wei, X. L., Ding, A. J., Poon, C. N., Lam, K. S., Li, Y. S., Chan, L. Y., and Anson, M.:
795 Increasing surface ozone concentrations in the background atmosphere of southern China, 1994-2007,
796 *Atmos. Chem. Phys.*, 9, 6217-6227, <https://doi.org/10.5194/acp-9-6217-2009>, 2009.

797 Wang, X., Zhang, Y., Hu, Y., Zhou, W., Lu, K., Zhong, L., Zeng, L., Shao, M., Hu, M., and Russell, A.
798 G.: Process analysis and sensitivity study of regional ozone formation over the Pearl River Delta, China,
799 during the PRIDE-PRD2004 campaign using the Community Multiscale Air Quality modeling system,
800 *Atmos. Chem. Phys.*, 10, 4423-4437, <https://doi.org/10.5194/acp-10-4423-2010>, 2010.

801 Wang, X., Situ, S., Guenther, A., Chen, F. E. I., Wu, Z., Xia, B., and Wang, T.: Spatiotemporal variability
802 of biogenic terpenoid emissions in Pearl River Delta, China, with high-resolution land-cover and
803 meteorological data, *Tellus B Chem. Phys. Meteorol.*, 63, 241-254, <https://doi.org/10.1111/j.1600-0889.2010.00523.x>, 2011.

805 Wang, T., Xue, L., Brimblecombe, P., Lam, Y. F., Li, L., and Zhang, L.: Ozone pollution in China: A
806 review of concentrations, meteorological influences, chemical precursors, and effects, *Sci. Total.
807 Environ.*, 575, 1582-1596, <https://doi.org/10.1016/j.scitotenv.2016.10.081>, 2017.

808 Wyatt Appel, K., Napelenok, S., Hogrefe, C., Pouliot, G., Foley, K. M., Roselle, S. J., Pleim, J. E., Bash,
809 J., Pye, H. O. T., Heath, N., Murphy, B., and Mathur, R.: Overview and evaluation of the Community
810 Multiscale Air Quality (CMAQ) modeling system version 5.2, in: *Air Pollution Modeling and its*

811 Application XXV, Springer Proceedings in Complexity, 69-73, [https://doi.org/10.1007/978-3-319-](https://doi.org/10.1007/978-3-319-57645-9_11)
812 57645-9_11, 2018.

813 Wang, T., Dai, J., Lam, K. S., Nan Poon, C., & Brasseur, G. P.: Twenty-five years of lower tropospheric
814 ozone observations in tropical East Asia: The influence of emissions and weather patterns, *Geophys. Res.*
815 *Let.*, 46(20), 11463-11470, <https://doi.org/10.1029/2019GL084459>, 2019.

816 Wu, X., Yuan, T., Qie, K., & Luo, J.: Geographical distribution of extreme deep and intense convective
817 storms on Earth, *Atmos. Res.*, 235, 104789, <https://doi.org/10.1016/j.atmosres.2019.104789>, 2020.

818 Xue, L., Wang, T., Louie, P. K., Luk, C. W., Blake, D. R., and Xu, Z.: Increasing external effects negate
819 local efforts to control ozone air pollution: A case study of Hong Kong and implications for other Chinese
820 cities, *Environ. Sci. Technol.*, 48, 10769-10775, <https://doi.org/10.1021/es503278g>, 2014.

821 Yang, C., Li, Q., Hu, Z., Chen, J., Shi, T., Ding, K., and Wu, G.: Spatiotemporal evolution of urban
822 agglomerations in four major bay areas of US, China and Japan from 1987 to 2017: Evidence from remote
823 sensing images, *Sci. Total. Environ.*, 671, 232-247, <https://doi.org/10.1016/j.scitotenv.2019.03.154>,
824 2019a.

825 Yang, L., Luo, H., Yuan, Z., Zheng, J., Huang, Z., Li, C., Lin, X., Louie, P. K. K., Chen, D., and Bian,
826 Y.: Quantitative impacts of meteorology and precursor emission changes on the long-term trend of
827 ambient ozone over the Pearl River Delta, China, and implications for ozone control strategy, *Atmos.*
828 *Chem. Phys.*, 19, 12901-12916, <https://doi.org/10.5194/acp-19-12901-2019>, 2019b.

829 Yue, X., Unger, N., Harper, K., Xia, X., Liao, H., Zhu, T., Xiao, J., Feng, Z., and Li, J.: Ozone and haze
830 pollution weakens net primary productivity in China, *Atmos. Chem. Phys.*, 17, 6073-6089,
831 <https://doi.org/10.5194/acp-17-6073-2017>, 2017.

832 Yusoff, M. F., Latif, M. T., Juneng, L., Khan, M. F., Ahamad, F., Chung, J. X., and Mohtar, A. A. A.:
833 Spatio-temporal assessment of nocturnal surface ozone in Malaysia, *Atmos. Environ.*, 207, 105-116,
834 <https://doi.org/10.1016/j.atmosenv.2019.03.023>, 2019.

835 Zhang, R., Lei, W., Tie, X., and Hess, P.: Industrial emissions cause extreme urban ozone diurnal
836 variability, *Proc. Natl. Acad. Sci. U. S. A.*, 101, 6346-6350, <https://doi.org/10.1073/pnas.0401484101>,
837 2004.

838 Zhang, B., Li, J., Wang, M., Duan, P., and Li, C.: Using DMSP/OLS and NPP/VIIRS images to analyze
839 the expansion of 21 urban agglomerations in mainland China, *J. Urban Plan. Dev.*, 147, 04021024,
840 [https://doi.org/10.1061/\(asce\)up.1943-5444.0000690](https://doi.org/10.1061/(asce)up.1943-5444.0000690), 2021.

841 Zheng, B., Tong, D., Li, M., Liu, F., Hong, C., Geng, G., Li, H., Li, X., Peng, L., Qi, J., Yan, L., Zhang,
842 Y., Zhao, H., Zheng, Y., He, K., and Zhang, Q.: Trends in China's anthropogenic emissions since 2010
843 as the consequence of clean air actions, *Atmos. Chem. Phys.*, 18, 14095-14111,
844 <https://doi.org/10.5194/acp-18-14095-2018>, 2018.

845 Zhong, Z., Zheng, J., Zhu, M., Huang, Z., Zhang, Z., Jia, G., Wang, X., Bian, Y., Wang, Y., and Li, N.:
846 Recent developments of anthropogenic air pollutant emission inventories in Guangdong province, China,
847 *Sci. Total Environ.*, 627, 1080-1092, <https://doi.org/10.1016/j.scitotenv.2018.01.268>, 2018.

848 Ziemann, A., Starke, M., & Leiding, T.: Sensitivity of nocturnal low-level jets to land-use parameters
849 and meteorological quantities, *Adv. Sci. Res.*, 16, 85-93, <https://doi.org/10.5194/asr-16-85-2019>, 2019.

850 Zhu, X. W., Ma, Z. Q., Li, Z. M., Wu, J., Guo, H., Yin, X. M., Ma, X. H., and Qiao, L.: Impacts of
851 meteorological conditions on nocturnal surface ozone enhancement during the summertime in Beijing,
852 *Atmos. Environ.*, 225, 117368, <https://doi.org/10.1016/j.atmosenv.2020.117368>, 2020.

853