

A comprehensive study on hygroscopic behaviour and nitrate depletion of NaNO_3 and dicarboxylic acid mixtures: Implications for nitrate depletion in tropospheric aerosols

Shuaishuai Ma^a, Qiong Li^b, Yunhong Zhang^c

5 ^a College of Chemical and Material Engineering, Quzhou University, Quzhou 324000, PR China

^b Shanghai Key Laboratory of Atmospheric Particle Pollution and Prevention, Department of Environmental Science & Engineering, Institute of Atmospheric Sciences, Fudan University, Shanghai 200433, PR China

^c The Institute of Chemical Physics, School of Chemistry and Chemical Engineering, Beijing Institute of Technology, Beijing 100081, PR China

10 Correspondence to: Shuaishuai Ma (mass@qzc.edu.cn) and Yunhong Zhang (yhz@bit.edu.cn)

Abstract. The nitrate depletion and HNO_3 release in internally mixed nitrate and dicarboxylic acids (DCAs) particles have been widely detected in field and laboratory studies. Nevertheless, considerable discrepancies are still present among these measurements, and the influence factors for this acid-displacement reaction have not yet been elucidated. In this work, the hygroscopic growth and chemical composition evolution of mixtures of NaNO_3 and DCAs, i.e., oxalic acid (OA), malonic acid (MA), and glutaric acid (GA), were measured using attenuated total reflectance Fourier transform infrared spectroscopy (ATR-FTIR) and vacuum FTIR techniques. The nitrate depletion from NaNO_3 /OA mixtures was observed in both the measurements, owing to the relatively higher acidity of OA compared to MA and GA. At the same time, the NaNO_3 phase state was found to act as a key regulator of nitrate depletion. Amorphous NaNO_3 solids at relative humidity (RH) < 5% were inert to liquid OA. With increasing RH, the mixtures experienced three interesting stages of phase changes showing different HNO_3 release rates, e.g., at around 15% RH, the slow HNO_3 release was detected by the vacuum IR spectra, potentially indicating the transformation of amorphous solids to semisolid NaNO_3 ; in the second stage (sudden RH increase from ~ 15% to 61%), the HNO_3 release rate was increased by about an order of magnitude; when NaNO_3 deliquescence occurred in the third stage, this displacement reaction proceeded due to more available NO_3^- ions formation. Compared to OA, MA and GA reacted with nitrate only in vacuum FTIR measurement, while in ATR-FTIR measurement, the mixtures tended to be effloresced completely without nitrate depletion. Further, the influences of ambient pressure, chemical composition, and water activity on HNO_3 release rates were estimated via Maxwell steady-state diffusive mass transfer equation. The results showed that weaker acidity of MA and GA as well as relatively lower HNO_3 diffusion rate in ambient gas phase mainly contributed to the unobserved nitrate depletion in ATR-FTIR measurement. Our findings reveal that chemical component, phase state, and water activity of particles, as well as HNO_3 gas phase diffusion play crucial roles on HNO_3 release from

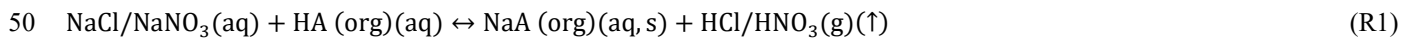
30 nitrate and DCAs mixtures. This work may provide a new perspective on nitrate depletion in the aging processes during transport of tropospheric aerosols.

1 Introduction

Aerosol particles in the atmosphere can play crucial roles in determining Earth's climate, air quality, and human health (Pöschl, 2006; Stevens and Feingold, 2009; Carslaw et al., 2013; Ramanathan et al., 2001; Brown et al., 2006), depending on
35 their various physicochemical properties, e.g., chemical composition, phase state, volatility, reactivity, hygroscopicity, the ability to absorb and scatter solar light as well as act as cloud condensation nuclei (CCN) (Pöschl, 2006; McFiggans et al., 2006; Haywood and Boucher, 2000; Farmer et al., 2015; Shiraiwa et al., 2017; Freney et al., 2018; Kuwata and Martin, 2012).

Sea salt and mineral dust aerosols can provide highly reactive surfaces for nitrates production through heterogeneous
40 reactions of gaseous nitrogen oxides such as HNO_3 , N_2O_5 , NO_2 , and NO_3 (Gibson et al., 2006; Song and Carmichael, 2001; Finlayson-Pitts and Hemminger, 2000). Furthermore, heterogeneous and aqueous oxidation of dicarboxylic acid precursors and gas-particle partitioning of dicarboxylic acids (DCAs) in the atmosphere will cause the internal mixing of DCAs with sea salt and mineral dust aerosols (Tervahattu et al., 2002; Sullivan and Prather, 2007; Laskin et al., 2012; Wang et al., 2010), greatly influencing the hygroscopic behaviour and surface tension of mixed particles (Facchini et al., 1999; Jing et al., 2018; Peng et al., 2001; Ovadnevaite et al., 2017; Hodas et al., 2015).

It is well known that the displacement of strong acids, i.e., HCl or HNO_3 , by weak organic acids, e.g., DCAs, is not thermodynamically favoured in bulk solutions (Laskin et al., 2012). Whereas, the nitrate and chloride depletion in mixed nitrate/chloride and organic acid particles has been widely detected in field and laboratory measurements, which could be expressed as (Laskin et al., 2012; Wang and Laskin, 2014; Ma et al., 2019b; Ghorai et al., 2014; Shao et al., 2018)



The driving force for this displacement reaction is mainly governed by the acidity difference and volatility difference between the organic acids and HCl/HNO_3 (Laskin et al., 2012; Wang and Laskin, 2014; Chen et al., 2021). The acidity difference tends to shift the reaction equilibrium to the left, demonstrating the more dominating driving force for the substitution of strong acids by weak organic acids is the volatility difference (Laskin et al., 2012; Chen et al., 2021). For
55 instance, dissociated HCl concentration is $\sim 10^{10}$ times higher than dissociated citric acid within mixed NaCl /citric acid (1/1 molar ratio) droplets, while the equilibrium gas phase concentration of HCl far exceeds that of citric acid with a factor of $\sim 10^{19}$, suggesting the HCl partition into the gas phase would manage the direction of acid-displacement reaction (Laskin et al., 2012). In general, the less acidity difference, i.e., higher acid dissociation constants (K_{a1}) of organic acids, and larger volatility difference, i.e., higher Henry's law constants (K_{H}) of organic acids, are favourable for gaseous HCl/HNO_3
60 liberation.

As indicated by previous studies, reactions between nitrate or chloride and oxalic acid (OA), which had a log ionization constant ($pK_{a1} = -\log_{10}K_{a1}$) of 1.23 (Haynes and Lide, 2011), always occurred in mixed aerosols, causing the formation of less hygroscopic metal oxalates (Ma et al., 2013, 2019a, b; Ma and He, 2012). Nevertheless, there were considerable discrepancies among earlier observations for internally mixed NaCl and malonic acid (MA) ($pK_{a1} = 2.83$). Laskin et al. (2012), Ghorai et al. (2014), Laskina et al. (2015), and Li et al. (2017) supported the occurrence of HCl substitution by MA, while Choi and Chan (2002), Pope et al. (2010), and Ma et al. (2013) did not observe this displacement reaction. Likewise, succinic acid (SA) ($pK_{a1} = 4.20$) was suggested not to react with internally and externally mixed NaCl by Ma et al. (2013), while Ghorai et al. (2014) found that glutaric acid (GA) ($pK_1 = 4.32$) was reactive to NaCl. In another publication, nitrates were proved to be reactive to MA and GA leading to nitrate depletion (Wang and Laskin, 2014). Indeed, the differences in particle sizes, substrate materials, acidity and volatility of DCAs may be responsible for these controversial results, but it is true that the interactions between organic acids and nitrate/chloride are still not clear.

To the best of our knowledge, the influence of particle-phase state on nitrate depletion has not been considered yet. As Wang et al. (2017) indicated, the OA droplets deposited on polytetrafluoroethylene (PTFE) substrate would crystallize to form OA dihydrate at $\sim 71\%$ relative humidity (RH) during dehydration, which further lost crystalline water to form anhydrous OA at $\sim 5\%$ RH; meanwhile, no deliquescence was observed upon hydration. Pure NaNO_3 droplets might not be effloresced but convert into highly concentrated droplets at low RH upon drying (Liu et al., 2008), or effloresced at certain RH values, e.g., levitated NaNO_3 droplets effloresced at RH near 40% (Lamb et al., 1996), or deposited NaNO_3 particles on hydrophobic and ZnSe substrates effloresced at 34% RH and 62.5% RH, respectively (Zhang et al., 2014). Furthermore, atmospheric aerosols can exist in highly viscous (i.e., semisolid or glassy), solid, liquid, or solid-liquid mixing states (Krieger et al., 2012; Mikhailov et al., 2009; Virtanen et al., 2010; Koop et al., 2011), depending on varying ambient RH, temperature, heterogeneous inclusions and so on (Ma et al., 2021b). Additionally, the particle-phase state has been proved to play a critical role in determining the reactivity of secondary organic material (SOM) upon ammonia exposure (Kuwata and Martin, 2012). These scenarios pose a key issue concerning how the particle-phase state would affect nitrate depletion in sea salt and mineral dust particles.

In this work, two measurement techniques, i.e., attenuated total reflectance Fourier transform infrared spectroscopy (ATR-FTIR) and vacuum FTIR, were carried out to measure the hygroscopic behaviour and nitrate depletion of mixed NaNO_3 and DCAs particles. The effects of HNO_3 gas phase diffusion, as well as chemical component, phase state, and water activity of mixed particles on nitrate depletion were further explored. This work does enhance our understanding of interaction mechanisms of DCAs and nitrate.

2.1 Sample preparation

The 0.1 mol L⁻¹ solutions of NaNO₃/OA, NaNO₃/MA, and NaNO₃/GA with molar ratio of 1:1 (or 3:1) were prepared by dissolving nitrate and DCAs into ultrapure water (18.2 MΩ cm resistivity). The bulk solutions were nebulized ultrasonically to produce aerosol droplets deposited on two CaF₂ windows in vacuum FTIR measurement and the Ge substrate in ATR-FTIR measurement.

2.2 Vacuum FTIR measurement

The vacuum FTIR technique was composed of a vacuum FTIR spectrometer and a RH controlling system. The experimental apparatus and method have been described in detail by our previous studies (Leng et al., 2015; Zhang et al., 2017; Ma et al., 2019c). Briefly, the vacuum FTIR spectrometer (Bruker VERTEX 80v) consisted of a vacuum optics bench, a sample compartment and a vacuum pump. The RH controlling system was composed of a high purity water reservoir, a sample chamber, and another vacuum pump. Water vapor from the water reservoir was fed into the sample chamber, and was pumped out by the vacuum pump. Two solenoid valves were installed in the pipeline of water vapor to instantaneously switch water vapor inlet and outlet. Meanwhile, two needle valves were configured to respectively adjust the flow rates of water vapor inlet and outlet.

The radius of deposited droplets on CaF₂ windows was in the range of ~ 1-3 μm (Zhang et al., 2017). After the nebulization, the CaF₂ windows were installed onto the sample chamber to seal it. The air in optics bench, sample compartment and sample chamber was pumped out to remove water vapor and CO₂. The baseline pressure in optics bench and sample compartment was pumped to ~ 0.21 kPa and the sample chamber arrived to ~ 0.01 kPa to remove water vapor and CO₂. A differential pressure transmitter (Rosemount 3051, accuracy > 0.5%) was used to measure water vapor pressure in the sample chamber, which could be used to calibrate the ambient RH (Zhang et al., 2017). The water content of deposited particles could be estimated by the integrated absorbance of stretching vibration band of liquid water molecules at 3400 cm⁻¹ (Ma et al., 2019c). The resolution of collected IR spectra was 4 cm⁻¹. All measurements were made at 23-26 °C.

2.3 ATR-FTIR measurement

Detailed description of ATR-FTIR measurement has been reported elsewhere (Zhang et al., 2014; Ren et al., 2016). Briefly, the IR spectra of deposited particles on a horizontal ATR (Spectra-Tech Inc. USA) accessory with the Ge substrate were measured by a Nicolet Magna-IR model 560 FTIR spectrometer equipped with a liquid-nitrogen cooled mercury-cadmium-telluride (MCT) detector. The ambient RH in the sample cell was controlled by adjusting the flow rate ratio of gas streams of dry and humidified nitrogen with a total flow rate of 800 mL·min⁻¹. The ambient RH and temperature were recorded by a hygrometer (Centertek Center 310, accuracy of ± 2.5%) in the outlet of sample cell. The diameter of deposited droplets was 1-5 μm with a median diameter of ~ 3 μm (Zhang et al., 2014). The IR spectra were collected between 4000 and 600 cm⁻¹

with a resolution of 4 cm⁻¹. Similar to vacuum FTIR measurement, the water content in aerosol particles was determined by the integrated absorbance of the stretching vibration band of liquid water.

3 Results and Discussions

3.1 Hygroscopic growth and chemical composition evolution of NaNO₃/OA mixtures

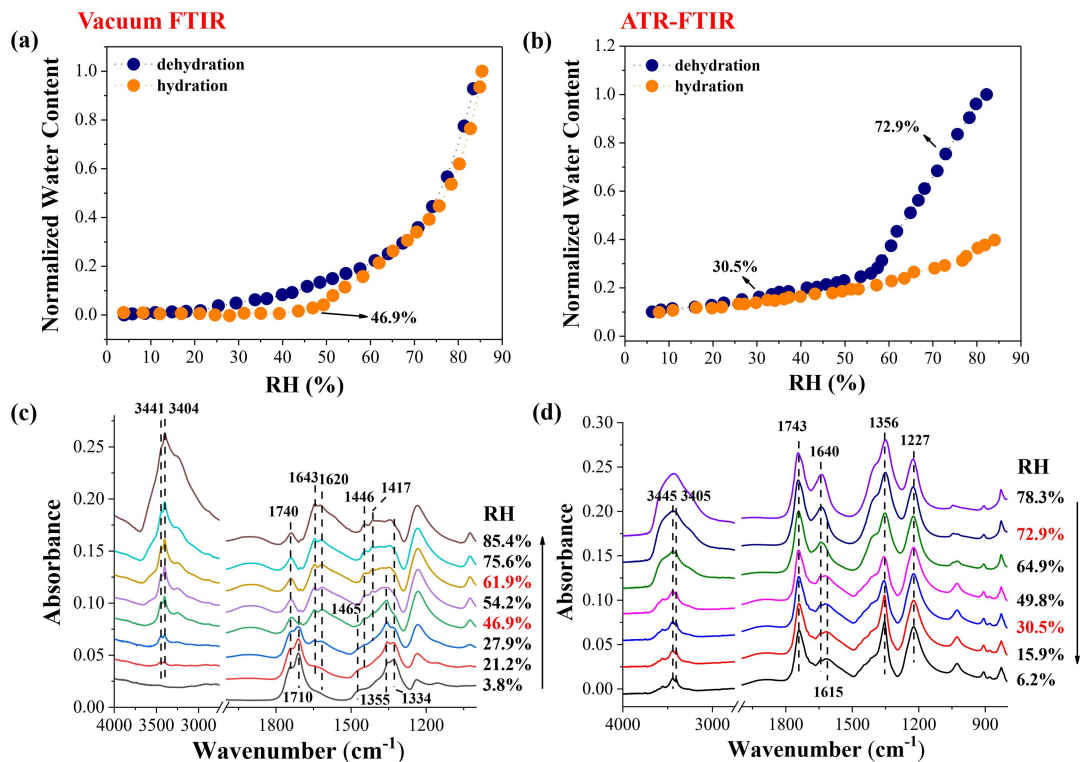
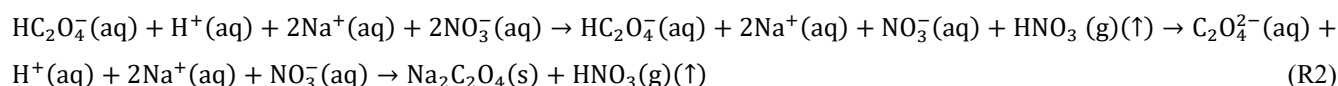


Figure 1. Hygroscopic growth curves of 1:1 mixed NaNO₃/OA particles measured by vacuum FTIR (a) and ATR-FTIR (b), as well as corresponding IR spectra during the humidification in vacuum FTIR measurement (c) and during the dehumidification in ATR-FTIR measurement (d).

The hygroscopic behaviour and IR features of individual components, i.e., NaNO₃, OA, MA, and GA, are shown in the Supplement. Fig. 1 displays the hygroscopic growth and chemical composition evolution of 1:1 NaNO₃/OA mixtures during a RH cycle. In vacuum FTIR measurement, the deposited droplets on CaF₂ windows are first dried in vacuum, and then undergo a humidification-dehumidification cycle. As shown in Fig. 1c, there are two feature bands assigned to the stretching mode of COOH functional groups ($\nu(\text{COOH})$) at ~ 3.8% RH, i.e., the 1740 cm⁻¹ band assigned to liquid OA and 1710 cm⁻¹ band assigned to crystalline anhydrous OA (Wang et al., 2019), indicating the coexistence of liquid OA and anhydrous OA.

135 The 1355 cm⁻¹ band is attributed to NO₃⁻ asymmetric stretching vibration ($\nu_3(\text{NO}_3^-)$) of amorphous NaNO₃ solids, as discussed in the Supplement. The weak absorption at 1620 cm⁻¹ indicates small amounts of crystalline oxalate formation (Hind et al., 1998; Wang et al., 2019), judged from the IR features of Na₂C₂O₄ solids shown in Fig. S4. In addition, the peak at 1465 cm⁻¹ is attributed to O-H bending mode of HC₂O₄⁻ ions (Villepin and Novak, 1971), indicative of the dissociated HC₂O₄⁻ formation. Likewise, Wang et al. (2017) observed the formation of NH₄HC₂O₄ in mixed (NH₄)₂SO₄/OA droplets
 140 upon drying. As already indicated, the release of HNO₃ and associated organic acid salts formation have been detected in several previous studies, thus herein, these observations can demonstrate the HNO₃ release and nitrate depletion in NaNO₃/OA mixtures, which is expected to occur in the vacuuming process. As RH increases to 21.2%, the feature bands at 3441 and 3404 cm⁻¹ appear, indicating the transformation of anhydrous OA to OA dihydrate. After that, the stronger 1620 cm⁻¹ and 1417 cm⁻¹ bands, assigned to C₂O₄²⁻ ion vibrating (Wang et al., 2019), and weaker 1465 cm⁻¹ band indicate the
 145 conversion of aqueous NaHC₂O₄ to crystalline Na₂C₂O₄. Thus, the acid-displacement reaction for NaNO₃/OA mixtures can be expressed as



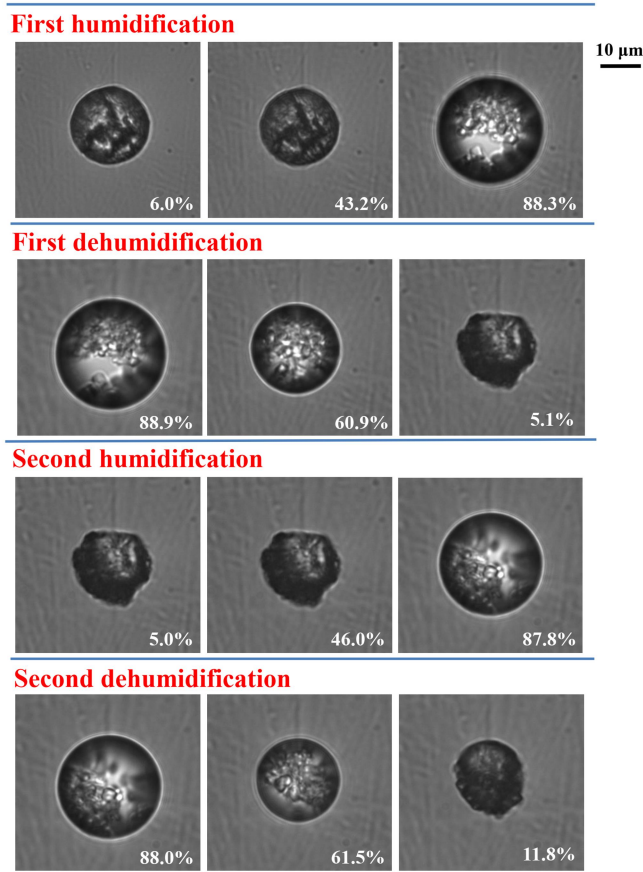
In previous studies, the reaction between NaCl or NaNO₃ and OA was found to produce disodium oxalate (Ma et al., 2013,
 150 2019b). To our knowledge, the formation of intermediate product of sodium hydrogen oxalate is first observed here.

As shown in Fig. 1a and 1c, we can infer that the NaNO₃ deliquescence proceeds at 46.9%-61.9% RH in NaNO₃/OA mixtures, significantly lower than the DRH of pure NaNO₃ particles (seen Fig. S2). Indeed, the aerosol DRH can be significantly reduced by the mixing of organic acids and inorganic salts, according to experimental measurements and thermodynamic model predictions (Bouzidi et al., 2020; Hodas et al., 2015; Marcolli et al., 2004). Furthermore, the OA
 155 dihydrate and crystalline Na₂C₂O₄ cannot be fully deliquesced due to their very high DRHs (Ma et al., 2013; Wang et al., 2019). Note that a solid-liquid equilibrium may be established, i.e., small amounts of OA dihydrate and Na₂C₂O₄ dissolve into the aqueous phase (Hodas et al., 2016). The IR spectra during the dehumidification are shown in Fig. S5a. As seen, the 1357 cm⁻¹ band becomes sharper at ~ 13.6% RH, indicating the remaining NaNO₃ efflorescence. Moreover, the residue of reactants indicates the incomplete reaction between OA and NaNO₃, consistent with the observations by Ma et al. (2019b)
 160 for Ca(NO₃)₂/OA, NaNO₃/OA and Zn(NO₃)₂/OA mixed systems.

Figure 1b and 1d display the hygroscopic growth and IR spectra of 1:1 NaNO₃/OA mixtures measured by ATR-FTIR. Mixed particles undergo a dehumidification-humidification cycle. Upon dehydration, the water content of particles gradually decreases with decreasing RH. When the RH attains ~ 72.9%, the shoulder bands at 3445 and 3405 cm⁻¹ assigned to OA dihydrate appear, as shown in Fig. 1d. Meanwhile, a new peak located at 1615 cm⁻¹ is observed, suggesting the formation of
 165 disodium oxalate. As RH decreases to 30.5%, the 1356 cm⁻¹ band assigned to $\nu_3(\text{NO}_3^-)$ becomes sharper, indicating the NaNO₃ efflorescence. Upon hydration, the water content above ~ 60% RH (orange solid circles in Fig. 1b) is far below that upon dehydration (blue solid circles), owing to the persistence of crystalline OA and Na₂C₂O₄ (Ma et al., 2013; Wang et al., 2017; Peng and Chan, 2001; Wu et al., 2011). Only NaNO₃ solids are deliquesced, identified by the broader NO₃⁻ feature

band at 1356 cm^{-1} (seen Fig. S5b). For a better illustration of phase state changes of mixed particles, the optical images of
170 1:1 NaNO_3/OA mixture during two RH cycles are determined by an optical microscope operated with a video camera, which
has been described in detail elsewhere (Ma et al., 2021a). For comparison with our vacuum FTIR measurement, the sample
droplet first undergoes the rapid drying process and then two humidification-dehumidification cycles. As shown in Fig. 2, at
6.0% RH in the first humidification process, the particle shape is roughly round and smooth, indicative of the amorphous
structure of NaNO_3 component (Hoffman et al., 2004; Laskin et al., 2006; Tang and Fung, 1997). As RH increases to ~
175 43.2%, the particle morphology becomes darker, potentially indicating the uptake of trace amounts of water. After NaNO_3
deliquescence, nonhygroscopic OA dihydrate and $\text{Na}_2\text{C}_2\text{O}_4$ still remain in crystalline states. The mixed particle exists in
solid-liquid mixing state involving liquid OA and NaNO_3 , crystalline $\text{Na}_2\text{C}_2\text{O}_4$ and OA dihydrate. When the RH decreases to
5.1%, the particle surface becomes irregular, differing from the initial particle morphology, which may be attributed to the
higher fraction of crystalline $\text{Na}_2\text{C}_2\text{O}_4$ as a reaction product. During the second RH cycle, the particle absorbs and releases
180 water routinely, indicative of the presence of residual nitrate.

Besides, as shown in Fig. 1c, the absorbance of 1620 cm^{-1} band assigned to oxalate shows a slight increase at RH as
low as 21.2%, implying the nitrate depletion proceeds at relatively low RH. Therefore, the effect of nitrate phase state on
nitrate depletion needs to be further explored.



185 **Figure 2.** Optical images of 1:1 mixed NaNO_3/OA particle during two RH cycles. The corresponding ambient RH is marked at the bottom right of the images.

3.2 Effect of nitrate phase state on nitrate depletion for NaNO_3/OA mixtures

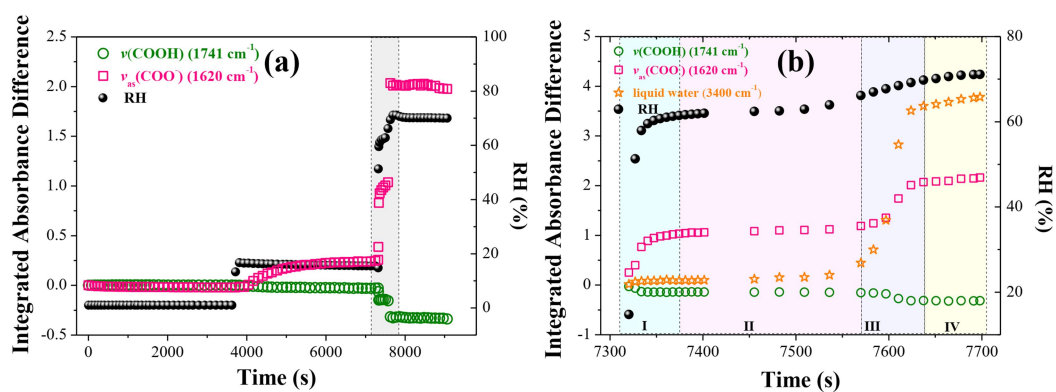


Figure 3. (a) Temporal changes in ΔA values of $\nu(\text{COOH})$ and $\nu_{\text{as}}(\text{COO}^-)$ bands with stepwise increasing RH for 3:1 NaNO_3/OA mixed particles. (b) Temporal changes in ΔA values of $\nu(\text{COOH})$, $\nu_{\text{as}}(\text{COO}^-)$ and liquid water bands with increasing RH corresponding to the shaded period in (a).

To better understand the phase state effect on this acid-displacement reaction, 3:1 NaNO_3/OA mixtures are dried in vacuum and then exposed to stepwise increasing RH. The integrated absorbance difference (ΔA), derived from the deducted spectra of IR spectra at a certain time to that at the initial time, of 1741 cm^{-1} band (assigned to COOH stretching mode, $\nu(\text{COOH})$) and 1620 cm^{-1} band (assigned to COO^- asymmetric stretching vibration, $\nu_{\text{as}}(\text{COO}^-)$) as a function of time is determined and shown in Fig. 3a. First, the ΔA remains almost unchanged at $\text{RH} < 5\%$. As RH increases to $\sim 15\%$, the ΔA of $\nu_{\text{as}}(\text{COO}^-)$ band exhibits a considerable increase, indicative of the occurrence of nitrate depletion. At the same time, the ΔA value of ~ 0.3 when the reaction equilibrium is reached implies that the nitrate depletion is limited at low RH. Note that the IR spectra changes with time at constant $\sim 15\%$ RH are supplied in Fig. S6. Almost no 1620 cm^{-1} band is observed at the initial time, suggesting almost no $\text{Na}_2\text{C}_2\text{O}_4$ formation in the vacuuming process, potentially owing to the minimization of HNO_3 release arising from rapid water evaporation (Ma et al., 2013). Then, the 1620 cm^{-1} band appears and becomes stronger with time, suggesting $\text{Na}_2\text{C}_2\text{O}_4$ can be continuously produced at constant $\sim 15\%$ RH. Considering the constant droplet water activity at constant 15% RH, these scenarios may provide more direct and powerful evidences for oxalate formation and associated nitrate depletion. Besides, coupled with the particle morphology changes shown in Fig. 2, we can infer the conversion of amorphous NaNO_3 solids to highly viscous NaNO_3 semisolids due to the uptake of trace amounts of moisture. As RH continues to increase, the ΔA values of $\nu_{\text{as}}(\text{COO}^-)$ and $\nu(\text{COOH})$ bands respectively increase or decrease. Finally, the ΔA values of the two bands remain constant with time.

Further, the variation in ΔA and water content as RH increases from $\sim 14.8\%$ to $\sim 69.2\%$ (shaded period in Fig. 3a) is processed and depicted in Fig. 3b. First, four stages are classed. In stage I, as RH increases from $\sim 14.8\%$ to $\sim 61.0\%$, the liquid water content (yellow pentagram) shows a very slight increase, while the ΔA of $\nu_{\text{as}}(\text{COO}^-)$ band (pink pane) and $\nu(\text{COOH})$ band (green circle) increases or decreases significantly. This implies that the absorbed moisture with increasing RH is favourable for nitrate depletion. In stage II, the RH increases slightly, meanwhile, the water content and the ΔA values remain almost unchanged. As the RH increases to $\sim 66.2\%$, the deliquescence of remaining NaNO_3 occurs, causing a sharp increase in liquid water content (stage III); meanwhile, the ΔA values of $\nu_{\text{as}}(\text{COO}^-)$ and $\nu(\text{COOH})$ bands respectively increase or decrease, indicating further nitrate depletion due to more available NO_3^- ions formation. In stage IV, the reaction equilibrium is reached and ΔA values remain unchanged. In addition, corresponding IR spectra changes of 3:1 mixtures during the shaded period are supplied in Fig. S7. At $\sim 69.2\%$ RH, the presence of 1741 cm^{-1} band indicates the excess of liquid OA, suggesting this displacement reaction tends to reach equilibrium with comparable final concentrations of “reactants” and “products” (Wang and Laskin, 2014).

For the quantitative evaluation of phase state effect, the kinetics of this displacement reaction should be further explored. For a given second-order reaction: $\text{A} + \text{B} \rightarrow \text{P}$ (P = products), the reaction rate can be determined as

$$\frac{d[\text{A}]}{dt} = -k[\text{A}][\text{B}] \quad (1)$$

where k refers to the second-order rate constant ($\text{cm}^3 \text{ molecule}^{-1} \text{ s}^{-1}$); $[A]$ and $[B]$ refer to the concentration of reactants A and B, respectively. Herein, during stepwise RH increasing, the concentration of reactant A, i.e., aqueous NO_3^- , is limited and the concentration of reactant B, i.e., liquid OA, is in excess, and hence, this displacement reaction can be treated as a pseudo first-order reaction (He et al., 2017; Gao et al., 2018). The rate equation (1) can be rewritten as

$$\frac{d[A]}{dt} = -k_{\text{app}}[A] \quad (2)$$

And then,

$$[A] = [A_0]e^{-k_{\text{app}}t} \quad (3)$$

where k_{app} is the pseudo first-order rate constant (s^{-1}) and equals to $k[B]$. $[A_0]$ refers to the initial concentration of reactant A. Based on this, the concentrations of reactant A and products P will change exponentially. As well, the integrated absorbance of 1620 cm^{-1} ($\nu_{\text{as}}(\text{COO}^-)$) band can be used to describe the product concentration, thus the pseudo-first-order rate constant k_{app} can be determined by the exponential changes in absorbance difference of $\nu_{\text{as}}(\text{COO}^-)$ band, $\Delta\bar{A}$, as a function of reaction time (Hung and Ariya, 2007; Hung et al., 2005; Gao et al., 2018). Namely, $\Delta\bar{A} = A_{\infty}e^{-k_{\text{app}}t}$, where $\Delta\bar{A} = A_t - A_{\infty}$, A_t and A_{∞} represent the integrated absorbance at time t and time infinite, respectively.

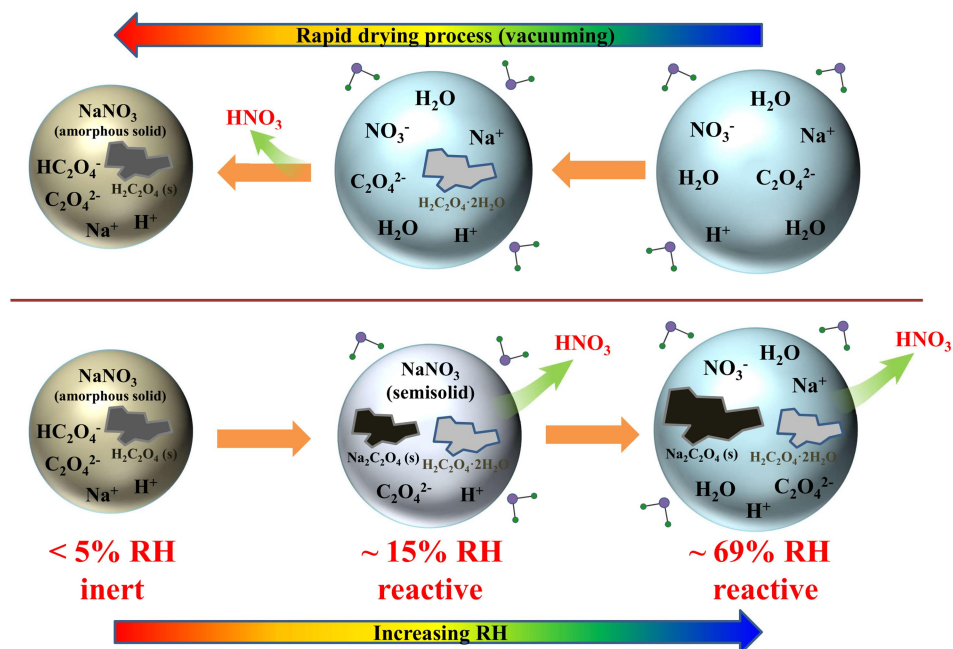
As shown in Fig. 3, the HNO_3 release process can be divided into three stages, corresponding to three RH ranges, i.e., constant $\sim 15\%$, sudden RH increase from $\sim 14.8\%$ to $\sim 61.0\%$, and $66.2\text{-}69.2\%$. The measured k_{app} values are shown in Table 1 and the $\Delta\bar{A}$ changes with initialized reaction time are depicted in Fig. S8. As seen, the HNO_3 release is relatively slow with a k_{app} value of $\sim 1.70 \times 10^{-3} \text{ s}^{-1}$ at constant $\sim 15\%$ RH, possibly owing to limited NO_3^- concentration and mass transfer limitation within the particle phase; the humidification process at $14.8\text{-}61.0\%$ RH exhibits the fastest HNO_3 release with the k_{app} value of $\sim 7.45 \times 10^{-2} \text{ s}^{-1}$; during NaNO_3 deliquescence, the HNO_3 release rate slightly decreases but is still about one order of magnitude higher than that at constant $\sim 15\%$ RH. It is noteworthy that the R^2 of exponential fitting of $\Delta\bar{A}$ as a function of initialized reaction time at $66.2\text{-}69.2\%$ RH is only 0.812, implying this displacement reaction may no longer obey to pseudo first-order reaction after NaNO_3 deliquescence due to the formation of large amounts of available NO_3^- ions, in other words, the concentration of aqueous NO_3^- is no longer limited. Furthermore, the fractions of reacting liquid OA at the end of three RH ranges are measured by the ratio of integrated absorbance of 1741 cm^{-1} ($\nu(\text{COOH})$) band at a certain time to that before the reaction, which show a value of $\sim 32.3\%$ before NaNO_3 deliquescence and reach $\sim 71.0\%$ after deliquescence. These results further confirm that aqueous NaNO_3 tends to be more reactive to liquid OA than amorphous and semisolid NaNO_3 due to the presence of large amounts of available NO_3^- ions.

Table 1 The pseudo-first-order rate constant k_{app} and fractions of reacting liquid OA corresponding to different time periods and RH ranges.

Time periods (s)	RH ranges (%)	k_{app} (s^{-1})	R^2	Fractions of reacting liquid OA (%)
4042-5731	~ 15 (constant)	1.70×10^{-3}	0.991	6.9

7320-7374	14.8-61.0	7.45×10^{-2}	0.958	32.3
7570-7651	66.2-69.2	2.77×10^{-2}	0.812	71.0

A summary of phase state changes and chemical compositional evolution of NaNO₃/OA mixed particles is shown in Fig. 4. During the rapid drying (i.e., vacuuming process), mixed droplets release water and then OA dihydrate forms. After that, mixed particles are effloresced to produce amorphous NaNO₃ solids, and OA dihydrate is converted into anhydrous OA. At the same time, the aqueous NaHC₂O₄ is formed accompanied by the release of gaseous HNO₃. For the humidification process, first, amorphous NaNO₃ solids at < 5% RH are inert to liquid OA, due to the unavailable dissociative NO₃⁻ ions for HNO₃ liberation. As RH increases to around 15% (or even lower), amorphous solids are converted into viscous semisolids, which can exhibit chemical reactivity to liquid OA causing nitrate depletion. When the NaNO₃ is deliquesced, more available NO₃⁻ ions are formed, and thus aqueous NaNO₃ is more reactive to liquid OA causing higher nitrate depletion extent. In previous studies, Li et al. (2017) indicated that the reaction between MA and NaCl could occur when aqueous H⁺ from MA and Cl⁻ ions from NaCl were available for HCl liberation, meanwhile, the reaction would slow down or stop once the amount of available H⁺ became small or the particles were effloresced. Besides, Kuwata and Martin (2012) investigated the phase state effect of atmospheric SOM on its reactivity upon ammonia exposure. They found that the semisolid adipic acid and α -pinene SOM could take up small amounts of ammonia even at low RH. At high RH, the particles existed in liquid state, and the absorbed water could act as a plasticizer which decreased the viscosity of particles and increased the diffusion coefficient of ammonia, thereby leading to extensive ammonia uptake. It was noteworthy that the ammonia uptake was also thermodynamic or kinetically limited, showing that aqueous SOM particles were not fully neutralized even for the highest NH₃ concentration.



270 **Figure 4.** Schematic diagram of phase state changes and chemical compositional evolution of NaNO_3/OA mixed particles in the vacuuming and humidification processes.

3.3 Hygroscopic growth and chemical composition evolution of NaNO₃/MA mixtures

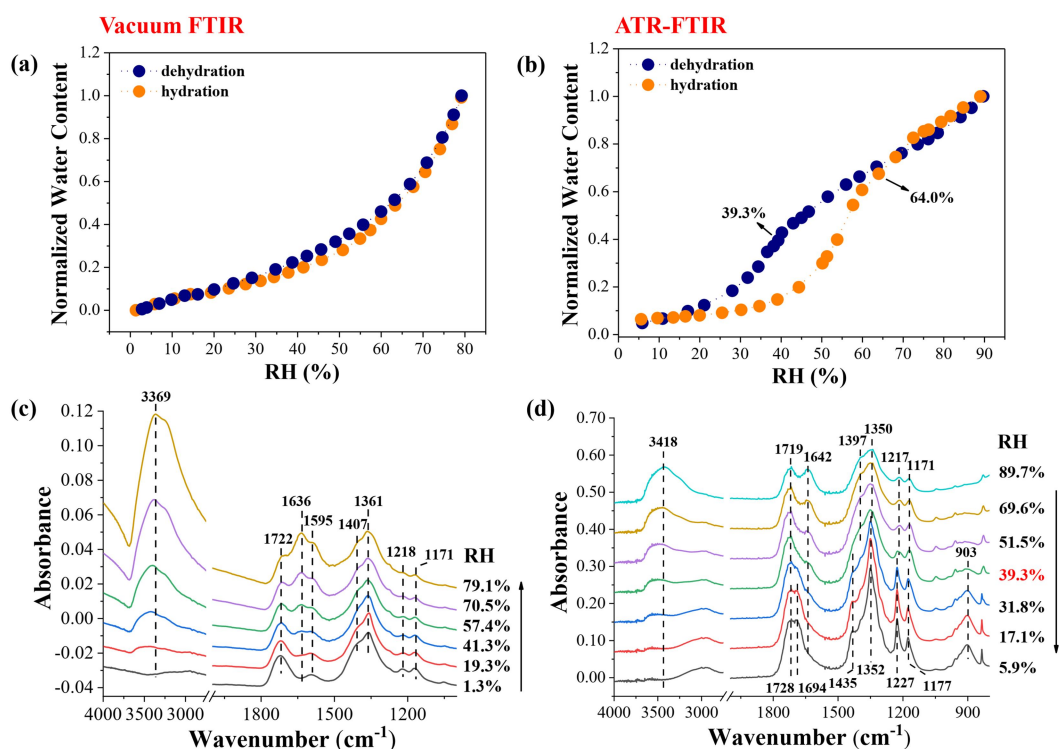
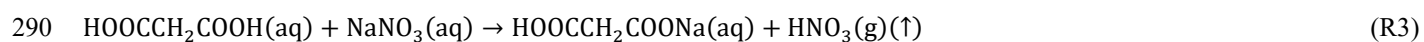


Figure 5. Hygroscopic growth curves of 1:1 mixed NaNO₃/MA particles measured by vacuum FTIR (a) and ATR-FTIR (b), as well as corresponding IR spectra during the humidification in vacuum FTIR measurement (c) and during the dehumidification in ATR-FTIR measurement (d).

Figure 5a and 5b display the hygroscopic behaviour of mixed NaNO₃/MA particles deposited on CaF₂ windows and Ge substrate measured by vacuum FTIR and ATR-FTIR, respectively. Correspondingly, the IR spectra upon hydration and upon dehydration are shown in Fig. 5c and 5d. Note that the IR spectra upon dehydration in vacuum FTIR measurement and upon hydration in ATR-FTIR measurement are shown in Fig. S9a and S9b, respectively. In vacuum FTIR measurement, the water content of mixed particles increases or decreases continuously with changing RH without distinct phase transitions (seen Fig. 5a). In other words, the addition of MA can totally inhibit the NaNO₃ crystallization. Likewise, Braban and Abbatt (2004) and Parsons et al. (2004) did not observe the efflorescence of (NH₄)₂SO₄ or MA in 1:1 mixed (NH₄)₂SO₄/MA particles under dry conditions. Also, no efflorescence was observed by Ghorai et al. (2014) for 1:1 NaCl/MA mixed system. In Fig. 5c, the 1722 cm⁻¹ band is assigned to $\nu(\text{C}=\text{O})$ of aqueous MA, and the 1407, 1218 and 1171 cm⁻¹ bands are also the characteristics of aqueous MA (Shao et al., 2017). The 1361 cm⁻¹ band is assigned to $\nu_3(\text{NO}_3^-)$ of NaNO₃ solution phase. More importantly, a new band located at 1595 cm⁻¹ indicates the formation of monosodium malonate (HOOCCH₂COONa) (Wang et al., 2019;

Shao et al., 2018). Thus, the displacement reaction between NaNO_3 and MA is confirmed by vacuum FTIR measurement as follows:



The production of monosodium malonate has also been observed in mixed NaCl/MA aerosols by Li et al. (2017). They explained that the first acid dissociation constant (K_{a1}) of MA was about 1.5×10^{-3} , which was ~ 3 orders of magnitude larger than the second one ($K_{a2} = 2.0 \times 10^{-6}$), resulting in much more $\text{HOOCCH}_2\text{COO}^-$ dissociated from MA than $\text{CH}_2(\text{COO})_2^{2-}$ (Li et al., 2017). Besides, no disodium salts are observed in the NaNO_3/MA system, differing from the NaNO_3/OA system, which may be due to the higher acidity of OA ($pK_{a2} = 4.19$) than MA ($pK_{a2} = 5.70$). To further validate this displacement reaction, the IR spectra of NaNO_3/MA mixtures before and after the RH cycle are supplied in Fig. S10. It is clear that the 1722 cm^{-1} band assigned to $\nu(\text{COOH})$ of MA becomes weaker, meanwhile, the 1595 cm^{-1} band assigned to $\nu_{\text{as}}(\text{COO}^-)$ of $\text{HOOCCH}_2\text{COO}^-$ ions dissociated from monosodium malonate becomes stronger after the RH cycle.

In ATR-FTIR measurement, deposited particles on the Ge substrate show distinct efflorescence and deliquescence transitions during the RH cycle. The initial efflorescence relative humidity (ERH) is $\sim 39.9\%$ and the RH where the particles are fully deliquesced is $\sim 64.0\%$. As shown in Fig. 5d, the IR features of mixtures at 89.7% RH are comparable to that measured by vacuum FTIR. Whereas, as RH decreases to 39.3% , the 1719 cm^{-1} band, assigned to $\nu(\text{COOH})$ of aqueous MA, is split into two shoulder peaks located at 1728 and 1694 cm^{-1} . Furthermore, the 1397 , 1217 , and 1171 cm^{-1} bands red-shift to 1435 , 1227 , and 1177 cm^{-1} , respectively. A new peak located at 903 cm^{-1} appears. All these scenarios indicate the liquid-solid phase transition of MA (Shao et al., 2017). The 1350 cm^{-1} band assigned to $\nu_3(\text{NO}_3^-)$ of aqueous NO_3^- turns into a sharp peak at 1352 cm^{-1} , suggesting the formation of NaNO_3 solids. In other words, the efflorescence of MA and NaNO_3 occurs synchronously at $\sim 39.3\%$ RH. As compared to vacuum FTIR results, we can infer that the heterogeneous efficacy of Ge substrate is much higher than CaF_2 windows, resulting in heterogeneous nucleation of MA and NaNO_3 upon dehydration (Ma et al., 2021b). As shown in Fig. 5b and S9b, the solid-liquid phase transition of crystalline MA and NaNO_3 ends at $\sim 64.0\%$ RH, which is lower than the deliquescence points of pure components. Specially, no IR features of malonate sodium salts are observed, indicative of no nitrate depletion in ATR-FTIR measurement, which is differed from the vacuum FTIR observation. The causes for this discrepancy will be discussed detailed in Sec. 3.5. Besides, the chlorine depletion in 1:1 mixed NaCl/MA particles with two different RH changing rates is also experimentally detected to further probe the influence factors for acid-displacement reactions, and the detailed discussion is presented in the Supplement. Briefly, the disodium malonate is produced in 1:1 mixed NaCl/MA particles during the dehumidification with the RH changing rate of $0.04\% \text{ RH} \cdot \text{s}^{-1}$, while in the faster dehumidification process ($2.20\% \text{ RH} \cdot \text{s}^{-1}$), no malonate sodium salts are formed and the mixtures are effloresced completely. There are two probable causes for no chloride depletion observed in the fast drying process. One is the minimization of HCl release caused by rapid water evaporation (Ma et al., 2013). Another cause is the efflorescence transition of mixed droplets at relatively high RH arising from lower temperature of droplets caused by rapid water evaporation (Ma et al., 2019c).

3.4 Hygroscopic growth and chemical composition evolution of NaNO₃/GA mixtures

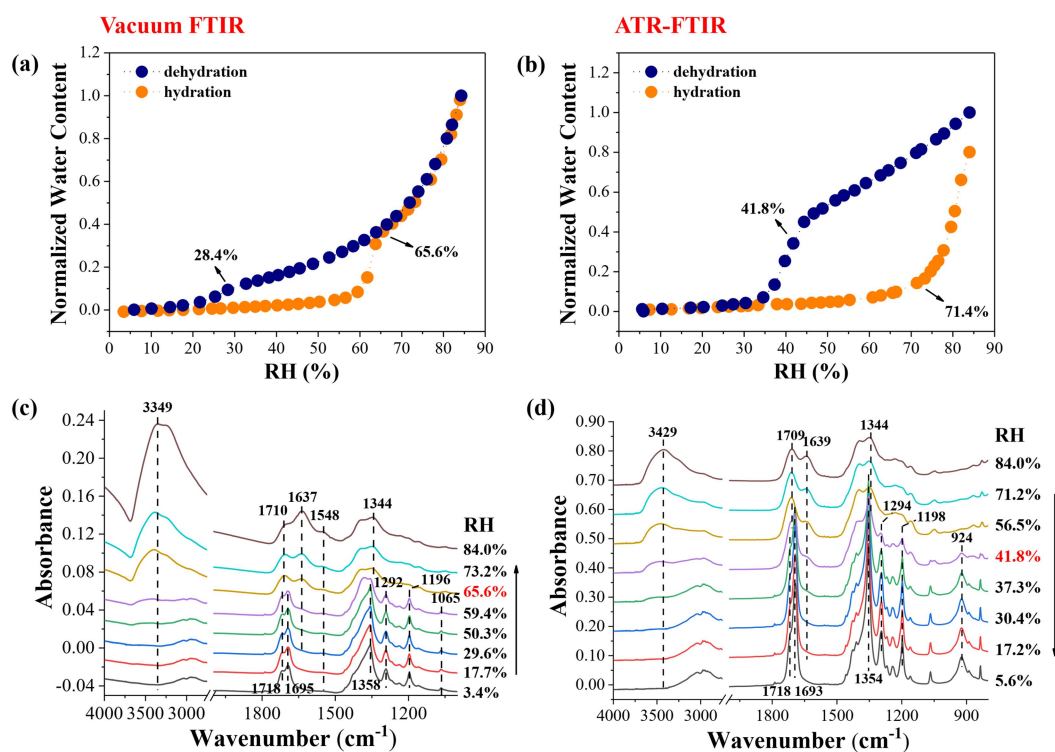


Figure 6. Hygroscopic growth curves of 1:1 mixed NaNO₃/GA particles measured by vacuum FTIR (a) and ATR-FTIR (b), as well as corresponding IR spectra during the humidification in vacuum FTIR measurement (c) and during the dehumidification in ATR-FTIR measurement (d).

The hygroscopic growth curve and corresponding IR spectra of 1:1 mixed NaNO₃/GA particles measured by vacuum FTIR are shown in Fig. 6a and 6c, respectively. Routinely, mixed particles are first dried in vacuum and then undergo a RH cycle. In Fig. 6c, at ~ 3.4% RH, the shoulder peaks located at 1718 and 1695 cm⁻¹ are attributed to $\nu(\text{C=O})$ of crystalline GA, and the $\nu(\text{C-O})$ band at 1292 cm⁻¹ and rocking vibration mode of CH₂ ($\gamma(\text{CH}_2)$) located at 1196 cm⁻¹ are also the characteristics of GA solids (Wu et al., 2019a). The 1358 cm⁻¹ band is attributed to $\nu_3(\text{NO}_3^-)$ of NaNO₃ solids. Indeed, mixed NaNO₃/GA particles exhibit distinctly different efflorescence behaviour compared with NaNO₃/MA mixtures, considerably consistent with the observation by Ghorai et al. (2014) for NaCl/GA(MA) mixed systems. This may be attributed to weaker inhibiting effect of GA than MA on inorganic salt crystallization (Ma et al., 2021b). In addition, a weak peak located at 1548 cm⁻¹ appears upon hydration, indicating the formation of dissociated glutarate sodium salts. The weaker intensity compared with NaNO₃/MA system indicates the substantially weaker chemical reactivity of GA ($pK_{a1} = 4.32$) than MA ($pK_{a1} = 2.83$) (Ghorai et al., 2014). As RH increases to 65.6%, mixed NaNO₃/GA particles are fully deliquesced, and corresponding IR feature changes are consistent with the observation by Wu et al. (2019b). During the dehumidification, aqueous droplets are

effloresced at $\sim 28.4\%$ RH, judged mainly from the IR feature changes of mixtures shown in Fig. S12a. It should be noted that only $\nu_3(\text{NO}_3^-)$ band experiences a red shift from 1344 to 1358 cm^{-1} , implying only NaNO_3 component effloresces and GA cannot be crystallized upon dehydration. The particle morphology of partial crystallization for mixtures of GA and nitrates was also observed by Wang and Laskin (2014). Braban (2004) found that GA would not effloresce upon drying for $(\text{NH}_4)_2\text{SO}_4/\text{GA}$ mixed system, potentially owing to the residual $(\text{NH}_4)_2\text{SO}_4$ in solution phase added extra barrier to the formation of GA crystalline germ (Braban and Abbatt, 2004). Wu et al. (2019b) studied the stepwise efflorescence process of 1:1 mixed NaNO_3/GA particles in the pulsed RH mode with vacuum FTIR method. They found that effloresced particles reversibly absorb and release water with pulsed RH changes when the minimal RH values were below 10%, suggesting there were still small amount of water retained in mixed particles, i.e., the mixed particles were partially crystallized. In addition, the crystallization of GA in the vacuuming process can be explained by the lower temperature of droplets arising from rapid water evaporation, as discussed in the Supplement.

In ATR-FTIR measurement, the efflorescence of mixed NaNO_3/GA particles occurs at $\sim 41.8\%$ RH, showing an abrupt decrease of water content with decreasing RH (seen Fig. 6b). For the humidification process, mixed particles obviously take up water at $\sim 71.4\%$ RH. The water content cannot match with that in the dehumidification process even until $\sim 84.0\%$ RH, suggesting the GA component cannot be deliquesced completely due to its high DRH (Marcolli et al., 2004; Yeung et al., 2010). This is further confirmed by corresponding IR features shown in Fig. S12b. As seen, the feature bands at 1207 and 924 cm^{-1} assigned to GA crystals are still observed at RH as high as 84.0%. Likewise, partial dissolution of GA was also observed for the $(\text{NH}_4)_2\text{SO}_4/\text{GA}$ system by Ling and Chan (2008). More importantly, there are no IR features of glutarate sodium salts observed during the RH cycle, which will be further explained in Sec. 3.5.

3.5 Influence factors for acid-displacement reactions

As already indicated, the driving out of gaseous HCl/HNO_3 from chloride/nitrate and organic acids mixtures is mainly dominated by two factors, i.e., K_{a1} and K_H of organic acids (Wang and Laskin, 2014; Laskin et al., 2012). Note that the acid dissociation constant (pK_{a1} , pK_{a2}), equilibrium constant K_1 , K_2 and K_3 utilized in the pH change simulation, as well as Henry's law constant (K_H) at 298 K for different acids involved in this study are listed in Table 2. In our previous review, the chlorine depletion in mixed $\text{NaCl}/\text{diacids}$ systems was suggested to be related to the acidity and environmental concentration of organic acids, as well as ambient RH and particle size of mixtures (Chen et al., 2021). First, stronger acidity would cause more available H^+ in aqueous phase, favouring the acid-displacement reaction. The acidity of diacids followed the order of $\text{OA} > \text{MA} > \text{SA} > \text{GA}$. Then, the higher acid concentration and lower RH would lead to greater HCl partial pressure, which was favourable for the partitioning of HCl into the gas phase. Finally, the chlorine depletion extent, ξ , was determined to be inversely proportional to r^2 where r was the droplet radius. In other words, the depletion extent ξ would increase greatly with decreasing droplet size.

370 As known, the driving out of volatile species such as HNO_3 from aqueous droplets can be quantified by Maxwell steady-state diffusive mass transfer equation (Chen et al., 2021; Cai et al., 2014; Ray et al., 1979). The mass flux of HNO_3 partitioning from particle to gas phase can be determined as

$$-\frac{dm}{dt} = \frac{4\pi rMD}{RT}(p_\infty - p_r) \quad (4)$$

where m denotes the mass of HNO_3 within the droplets (g); t denotes the evaporation time (s); M and D represent the molecular weight (g mol^{-1}) and diffusion coefficient of HNO_3 in the air ($\text{m}^2 \text{s}^{-1}$), respectively; R is the ideal gas constant ($\text{J mol}^{-1} \text{K}^{-1}$); T is the temperature (K); r is the droplet radius (m); p_∞ and p_r represent the partial pressure of HNO_3 (Pa) at infinite distance and droplet surface, respectively.

When the RH decreases continuously, the evaporation rate of HNO_3 at a certain RH, k_{RH} (g s^{-1}), can be expressed as (assuming $p_\infty = 0$)

$$380 \quad k_{\text{RH}} = \frac{dm}{dt} = \frac{4\pi rMDp_r}{RT} \quad (5)$$

In the ATR-FTIR measurement, the diffusion coefficient of HNO_3 in the air is $(1.18 \pm 0.03) \times 10^{-5} \text{ m}^2 \text{s}^{-1}$ under the condition of $T = 298 \text{ K}$ and $P = 1 \text{ atm}$ (Durham and Stockburger, 1986). Assuming that the droplet radius at 90% RH is $\sim 1.5 \mu\text{m}$, the droplet size r at any RH can be determined according to the size growth factors predicted by the Extended Aerosol Inorganic Model (E-AIM). p_r can also be estimated by the E-AIM model. Note that the E-AIM predictions are performed by UNIFAC model with parameters modified by Peng et al. (2001). In addition, the RH can be converted to water activity, a_w , by Köhler equation to minimize the Kelvin curvature effect, as shown in the Supplement (Jing et al., 2016). Based on these, the k_{RH} for 1:1 mixed NaNO_3/OA , NaNO_3/MA , and NaNO_3/GA systems as a function of a_w in ATR-FTIR measurement can be estimated and shown in Fig. 7a. Note that the solid phase formation is prevented to obtain the simulation data of supersaturated droplets at low RH.

390 Besides, according to Chapman-Enskog method, the diffusion coefficient of HNO_3 in the gas phase can be expressed as (Reid et al., 1987)

$$D = \frac{0.00266}{PM_{\text{AB}}^{1/2} \sigma_{\text{AB}}^2 \Omega_D} \quad (6)$$

where subscripts A and B denote species A and B; P is the ambient pressure (bar); $M_{\text{AB}} = 2[(1/M_A) + (1/M_B)]^{-1}$ (g mol^{-1}); σ_{AB} represents the characteristic length (\AA); Ω_D is the diffusion collision integral. Thus, the HNO_3 diffusion coefficient in vacuum FTIR measurement, D^* , is much higher than that in ATR-FTIR measurement, due to the much lower ambient pressure in vacuum FTIR, P^* , which can be expressed as

$$P^* = P_0 * RH \quad (7)$$

where P_0 denotes the saturated water vapor pressure at 298 K (bar). Based on equations (6) and (7), the D^* value at any RH can be determined. Accordingly, the k_{RH} values in vacuum FTIR measurement can also be calculated for comparison, as shown in Fig. 7a. The k_{RH} values in vacuum FTIR measurement are ~ 35 times higher than those in ATR-FTIR measurement at $a_w = 0.9$, due to much lower ambient pressure in vacuum FTIR. Based on this, the discrepancies among the two

measurements for nitrate depletion in $\text{NaNO}_3/\text{MA}(\text{GA})$ systems can be mainly attributed to the higher HNO_3 release rate arising from faster HNO_3 gas phase diffusion in vacuum FTIR. Previous studies have indicated that numerous atmospheric processes involved aerosol particles were often carried out at high altitudes with significantly lower pressure than the ground level (Zhao et al., 2009; Rosenberger et al., 2018; Schilling and Winterer, 2014), and aerosol properties such as hygroscopicity under reduced pressure should be further characterized (Tang et al., 2019). Furthermore, mixed NaNO_3/OA systems exhibits higher HNO_3 release rate than $\text{NaNO}_3/\text{MA}(\text{GA})$ systems due to the higher acidity of OA. Meanwhile, the HNO_3 release rate for NaNO_3/OA system in ATR-FTIR measurement is comparable to that for NaNO_3/GA system in vacuum FTIR measurement. These indicate that the lower acidity and hence lower reactivity of MA and GA also contribute to the unobserved nitrate depletion in ATR-FTIR measurement. In addition, all the k_{RH} values increase significantly with decreasing droplet water activity. Specifically, the k_{RH} at $a_w = 0.1$ is about two orders of magnitude higher than that at $a_w = 0.9$. It should be noted that in ATR-FTIR measurement, HNO_3 release from $\text{NaNO}_3/\text{MA}(\text{GA})$ aqueous droplets at very low RH would exhibit comparable levels to that for NaNO_3/OA system at relatively higher RH. Therefore, considerable HNO_3 release from $\text{NaNO}_3/\text{MA}(\text{GA})$ mixtures can be expected in case of no complete efflorescence of mixtures in ATR-FTIR measurement. In other words, the crystallization of mixed droplets induced by the Ge substrate, causing the lack of aqueous H^+ and NO_3^- ions available for HNO_3 liberation, tends to be another cause for negligible or even no HNO_3 release from $\text{NaNO}_3/\text{MA}(\text{GA})$ mixtures.

Besides, the pH changes of 1:1 $\text{NaNO}_3/\text{DCAs}$ aqueous droplets with initial concentration of 0.1 mol/L as a function of nitrate depletion fraction are calculated and shown in Fig. 7b. The calculation method is similar to that for ammonium depletion simulation proposed by Wang et al. (2019) (shown in the Supplement). As the NO_3^- in aqueous phase is depleted, the conversion of DCAs to their sodium salts proceeds, leading to the continuous reduction in droplet acidity.

Table 2 The acid dissociation constant (pK_{a1} , pK_{a2}), equilibrium constant K_1 , K_2 and K_3 utilized in the pH change simulation, as well as Henry's law constant (K_{H}) at 298 K for different acids involved in this study.

Species	pK_{a1}^{a}	pK_{a2}^{a}	$K_1 (K_{a1})$	$K_2 (K_{a2})$	$K_3 (1/K_{a1})$	$K_{\text{H}} (\text{M/atm})^{\text{b}}$
HNO_3	-1.27		$> 2 \times 10^1$		5×10^{-2}	$> 2 \times 10^5$
Oxalic acid	1.23	4.19	5.9×10^{-2}	6.5×10^{-5}		$(0.06-6.2) \times 10^8$
Malonic acid	2.83	5.70	1.5×10^{-3}	2.0×10^{-6}		$(0.26-3.3) \times 10^{10}$
Glutaric acid	4.32	5.42	4.8×10^{-5}	3.8×10^{-6}		$(0.35-3.3) \times 10^9$

^a: Data from Haynes and Lide (2011).

^b: Data from Wang and Laskin (2014), Compernelle and Müller (2014), Soonsin et al. (2010), and Bilde et al. (2003).

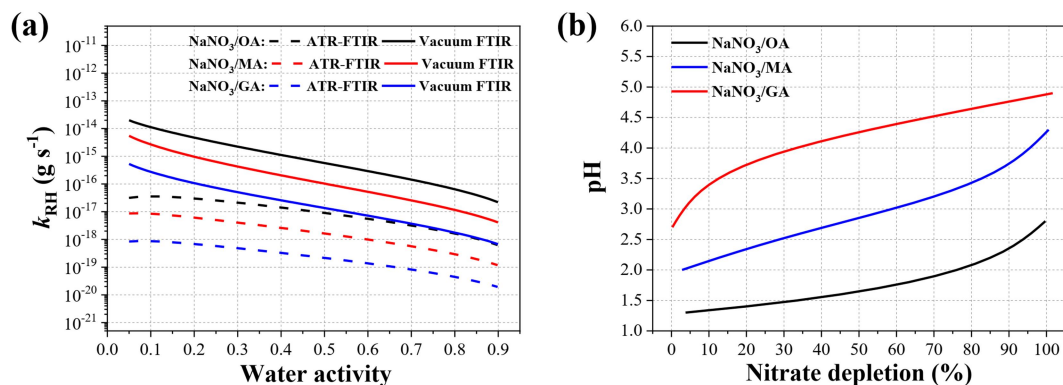


Figure 7. (a) The k_{RH} values for 1:1 mixed droplets of NaNO₃ and DCAs as a function of water activity at 298 K in the two measurements, i.e., ATR-FTIR (dashed lines) and vacuum FTIR (solid lines). Note that the droplet radius is assumed to be $\sim 1.5 \mu\text{m}$ at $a_w = 0.9$. (b) The droplet pH of 1:1 NaNO₃/DCAs mixtures with initial concentration of 0.1 mol/L as a function of nitrate depletion fraction at 298 K.

430 4 Conclusions and atmospheric implications

The conversion of organic acids to organic acid salts and the internal heterogeneity can significantly alter the hygroscopic properties, acidity, optical properties, viscosity, and chemical reactivity of mixed aerosols (Ma et al., 2013, 2019b; Peng et al., 2016; Ghorai et al., 2014). However, the formation and precipitation of organic acid salts have not been considered by the current version of E-AIM. In this work, the hygroscopic behaviour and chemical composition modification of NaNO₃/DCAs mixtures during the RH cycle were investigated by vacuum FTIR and ATR-FTIR techniques. First, OA could react with nitrate causing considerable nitrate depletion in both the measurements. Indeed, substantial metal oxalate complexes, which could not contribute to the CCN activity of aerosols, were frequently detected mixed in tropospheric aerosols (Furukawa and Takahashi, 2011). At the same time, the nitrate phase state would play a critical role in determining the occurrence and extent of nitrate depletion, e.g., amorphous NaNO₃ solids were inert to liquid OA; at around 15% RH, some moisture was absorbed to form semisolid NaNO₃ accompanied by HNO₃ liberation; the nitrate depletion extent and HNO₃ release rate increased with increasing RH; during NaNO₃ deliquescence, the displacement reaction proceeded due to more available NO₃⁻ ions formation. In addition, dried NaNO₃ particles can be treated as a surrogate for a broad class of amorphous or semisolid species existing in tropospheric aerosols, which undergo homogeneous or heterogeneous reactions causing secondary inorganic or organic aerosols formation. Therefore, the phase state effect may also be suitable for relevant aerosol reaction systems.

MA and GA exhibited weaker chemical reactivity to nitrate than OA in vacuum FTIR measurement. More importantly, almost no sodium salts of MA and GA were produced in ATR-FTIR measurements, differing from the vacuum FTIR observations. This discrepancy was confirmed to be mainly dominated by the faster HNO₃ gas phase diffusion arising from much lower ambient pressure in vacuum FTIR. Besides, the weaker acidity of MA and GA, lower HNO₃ release rate at

450 higher RH, as well as the occurrence of efflorescence transition of aqueous droplets also contributed to the unobserved nitrate depletion in ATR-FTIR measurements. These scenarios indicate that organic acids have a potential to deplete nitrate based on the comprehensive consideration of acidity, particle-phase state, droplet water activity, as well as HNO₃ gas phase diffusion (influenced by ambient pressure and so on). Our results reveal that faster HNO₃ gas phase diffusion, higher acidity of organic acids, lower droplet water activity, as well as the absence of efflorescence transition upon drying would be
455 favourable for HNO₃ partition into the gas phase. In atmospheric aerosols, insoluble materials such as mineral dust inclusions frequently induce the heterogeneous nucleation of aerosol droplets at relatively high RH, and thus displacement reactions between MA or GA and nitrates may rarely contribute to the nitrate depletion in aerosols. Indeed, in supermicrometer sea-salt aerosols, the chloride depletion was mainly attributed to sulfate and nitrate, followed by methanesulfonate and oxalate, while malonate and glutarate played a minor role in Cl⁻ loss (Kerminen et al., 1998).

460 Besides, this work will help us understand the discrepancies among previous observations on chloride/nitrate depletion. For instance, in Ghorai's study, the submicrometer NaCl/MA(GA) particles deposited on Si₃N₄ windows and TEM grids tended to stay metastable state under dry conditions, leading to no or partial crystallization of mixtures which facilitated the HCl liberation (Ghorai et al., 2014). Likewise, the micron-sized NaNO₃/MA(GA) particles deposited on carbon-filmed grids and Si₃N₄ windows would exist in amorphous liquid or viscous semisolid state after drying, causing significant HNO₃
465 liberation (Wang and Laskin, 2014). In contrast, Ma et al. (2013) did not observe the chlorine depletion in mixed NaCl/MA(GA) particles placed in aluminum sample holder and deposited on Ge substrate, which might be attributed to the heterogeneous nucleation of mixed droplets in the fast drying process. Also, the formation of Na₂C₂O₄ was observed in internally mixed NaCl/OA particles due to the higher acidity of OA (Ma et al., 2013).

470 *Data availability.* All data are available upon request from the corresponding authors.

Author contributions. SSM and YHZ designed the experimental plan. SSM and QL preformed the measurements. QL helped with data analysis. SSM and YHZ wrote the paper. All authors discussed and contributed to the manuscript.

475 *Competing interests.* The authors declare that they have no conflict of interest.

Acknowledgements. This work was supported by the National Natural Science Foundation of China (Nos. 42127806 and 41875144).

References

480 Bilde, M., Svenningsson, B., Mønster, J., and Rosenørn, T.: Even-odd alternation of evaporation rates and vapor pressures of C3–C9 dicarboxylic acid aerosols, *Environ. Sci. Technol.*, 37, 1371-1378, <https://doi.org/10.1021/es0201810>, 2003.

- Bouzidi, H., Zuend, A., Ondráček, J., Schwarz, J., and Ždimal, V.: Hygroscopic behavior of inorganic–organic aerosol systems including ammonium sulfate, dicarboxylic acids, and oligomer, *Atmos. Environ.*, 229, 117481, <https://doi.org/10.1016/j.atmosenv.2020.117481>, 2020.
- 485 Braban, C. F.: Laboratory studies of model tropospheric aerosol phase transitions, Ph.D. thesis, University of Toronto, Toronto, 2004.
- Braban, C. F., and Abbatt, J. P. D.: A study of the phase transition behavior of internally mixed ammonium sulfate - malonic acid aerosols, *Atmos. Chem. Phys.*, 4, 1451-1459, <https://doi.org/10.5194/acp-4-1451-2004>, 2004.
- Brown, S. S., Ryerson, T. B., Wollny, A. G., Brock, C. A., Peltier, R., Sullivan, A. P., Weber, R. J., Dube, W. P., Trainer, M.,
- 490 Meagher, J. F., Fehsenfeld, F. C., and Ravishankara, A. R.: Variability in nocturnal nitrogen oxide processing and its role in regional air quality, *Science*, 311, 67-70, <https://doi.org/10.1126/science.1120120>, 2006.
- Cai, C., Stewart, D. J., Preston, T. C., Walker, J. S., Zhang, Y. H., and Reid, J. P.: A new approach to determine vapour pressures and hygroscopicities of aqueous aerosols containing semi-volatile organic compounds, *Phys. Chem. Chem. Phys.*, 16, 3162-3172, <https://doi.org/10.1039/c3cp54948h>, 2014.
- 495 Carslaw, K. S., Lee, L. A., Reddington, C. L., Pringle, K. J., Rap, A., Forster, P. M., Mann, G. W., Spracklen, D. V., Woodhouse, M. T., Regayre, L. A., and Pierce, J. R.: Large contribution of natural aerosols to uncertainty in indirect forcing, *Nature*, 503, 67-71, <https://doi.org/10.1038/nature12674>, 2013.
- Chen, Z., Liu, P., Liu, Y., and Zhang, Y. H.: Strong acids or bases displaced by weak acids or bases in aerosols: Reactions driven by the continuous partitioning of volatile products into the gas phase, *Acc. Chem. Res.*, 54, 3667-3678, <https://doi.org/10.1021/acs.accounts.1c00318>, 2021.
- 500 Choi, M. Y., and Chan, C. K.: The effects of organic species on the hygroscopic behaviors of inorganic aerosols, *Environ. Sci. Technol.*, 36, 2422-2428, <https://doi.org/10.1021/es0113293>, 2002.
- Compernelle, S., and Müller, J. F.: Henry's law constants of diacids and hydroxy polyacids: recommended values, *Atmos. Chem. Phys.*, 14, 2699-2712, <https://doi.org/10.5194/acp-14-2699-2014>, 2014.
- 505 Durham, J. L., and Stockburger, L.: Nitric acid-air diffusion coefficient: Experimental determination, *Atmos. Environ.*, 20, 559-563, [https://doi.org/10.1016/0004-6981\(86\)90098-3](https://doi.org/10.1016/0004-6981(86)90098-3), 1986.
- Facchini, M. C., Mircea, M., Fuzzi, S., and Charlson, R. J.: Cloud albedo enhancement by surface-active organic solutes in growing droplets, *Nature*, 401, 257-259, <https://doi.org/10.1038/45758>, 1999.
- Farmer, D. K., Cappa, C. D., and Kreidenweis, S. M.: Atmospheric processes and their controlling influence on cloud
- 510 condensation nuclei activity, *Chem. Rev.*, 115, 4199-4217, <https://doi.org/10.1021/cr5006292>, 2015.
- Finlayson-Pitts, B. J., and Hemminger, J. C.: Physical chemistry of airborne sea salt particles and their components, *J. Phys. Chem. A*, 104, 11463-11477, <https://doi.org/10.1021/jp002968n>, 2000.
- Frenay, E., Sellegri, K., Chrit, M., Adachi, K., Brito, J., Waked, A., Borbon, A., Colomb, A., Dupuy, R., Pichon, J.-M., Bouvier, L., Delon, C., Jambert, C., Durand, P., Bourianne, T., Gaimoz, C., Triquet, S., Féron, A., Beekmann, M., Dulac, F.,
- 515 and Sartelet, K.: Aerosol composition and the contribution of SOA formation over Mediterranean forests, *Atmos. Chem. Phys.*, 18, 7041-7056, <https://doi.org/10.5194/acp-18-7041-2018>, 2018.
- Furukawa, T., and Takahashi, Y.: Oxalate metal complexes in aerosol particles: implications for the hygroscopicity of oxalate-containing particles, *Atmos. Chem. Phys.*, 11, 4289-4301, <https://doi.org/10.5194/acp-11-4289-2011>, 2011.
- Gao, X. Y., Zhang, Y. H., and Liu, Y.: A kinetics study of the heterogeneous reaction of n-butylamine with succinic acid using an ATR-FTIR flow reactor, *Phys. Chem. Chem. Phys.*, 20, 15464-15472, <https://doi.org/10.1039/C8CP01914B>, 2018.
- 520 Ghorai, S., Wang, B. B., Tivanski, A., and Laskin, A.: Hygroscopic properties of internally mixed particles composed of NaCl and water-soluble organic acids, *Environ. Sci. Technol.*, 48, 2234-2241, <https://doi.org/10.1021/es404727u>, 2014.
- Gibson, E. R., Hudson, P. K., and Grassian, V. H.: Physicochemical properties of nitrate aerosols: Implications for the atmosphere, *J. Phys. Chem. A*, 110, 11785-11799, <https://doi.org/10.1021/jp063821k>, 2006.
- 525 Haynes, W. M., and Lide, D. R.: CRC handbook of chemistry and physics, CRC Press, Boca Raton, FL, 2011.
- Haywood, J., and Boucher, O.: Estimates of the direct and indirect radiative forcing due to tropospheric aerosols: A review, *Rev. Geophys.*, 38, 513-543, <https://doi.org/10.1029/1999RG000078>, 2000.
- He, X., Leng, C., Pang, S., and Zhang, Y.: Kinetics study of heterogeneous reactions of ozone with unsaturated fatty acid single droplets using micro-FTIR spectroscopy, *RSC Adv.*, 7, 3204-3213, <https://doi.org/10.1039/c6ra25255a>, 2017.
- 530 Hind, A. R., Bhargava, S. K., Van Bronswijk, W., Grocott, S. C., and Eyer, S. L.: On the aqueous vibrational spectra of alkali metal oxalates, *Appl. Spectrosc.*, 52, 683-691, <https://doi.org/10.1366/0003702981944355>, 1998.

- Hodas, N., Zuend, A., Mui, W., Flagan, R. C., and Seinfeld, J. H.: Influence of particle-phase state on the hygroscopic behavior of mixed organic–inorganic aerosols, *Atmos. Chem. Phys.*, 15, 5027-5045, <https://doi.org/10.5194/acp-15-5027-2015>, 2015.
- 535 Hodas, N., Zuend, A., Schilling, K., Berkemeier, T., Shiraiwa, M., Flagan, R. C., and Seinfeld, J. H.: Discontinuities in hygroscopic growth below and above water saturation for laboratory surrogates of oligomers in organic atmospheric aerosols, *Atmos. Chem. Phys.*, 16, 12767-12792, <https://doi.org/10.5194/acp-16-12767-2016>, 2016.
- Hoffman, R. C., Laskin, A., and Finlayson-Pitts, B. J.: Sodium nitrate particles: physical and chemical properties during hydration and dehydration, and implications for aged sea salt aerosols, *J. Aerosol Sci.*, 35, 869-887, <https://doi.org/10.1016/j.jaerosci.2004.02.003>, 2004.
- 540 Hung, H.-M., Katrib, Y., and Martin, S. T.: Products and mechanisms of the reaction of oleic acid with ozone and nitrate radical, *J. Phys. Chem. A*, 109, 4517-4530, <https://doi.org/10.1021/jp0500900>, 2005.
- Hung, H.-M., and Ariya, P.: Oxidation of oleic acid and oleic acid/sodium chloride(aq) mixture droplets with ozone: changes of hygroscopicity and role of secondary reactions, *J. Phys. Chem. A*, 111, 620-632, <https://doi.org/10.1021/jp0654563>, 2007.
- 545 Jing, B., Tong, S. R., Liu, Q. F., Li, K., Wang, W. G., Zhang, Y. H., and Ge, M. F.: Hygroscopic behavior of multicomponent organic aerosols and their internal mixtures with ammonium sulfate, *Atmos. Chem. Phys.*, 16, 4101-4118, <https://doi.org/10.5194/acp-16-4101-2016>, 2016.
- Jing, B., Wang, Z., Tan, F., Guo, Y. C., Tong, S. R., Wang, W. G., Zhang, Y. H., and Ge, M. F.: Hygroscopic behavior of atmospheric aerosols containing nitrate salts and water-soluble organic acids, *Atmos. Chem. Phys.*, 18, 5115-5127, <https://doi.org/10.5194/acp-18-5115-2018>, 2018.
- 550 Kerminen, V.-M., Teinilä, K., Hillamo, R., and Pakkanen, T.: Substitution of chloride in sea-salt particles by inorganic and organic anions, *J. Aerosol Sci.*, 29, 929-942, [https://doi.org/10.1016/S0021-8502\(98\)00002-0](https://doi.org/10.1016/S0021-8502(98)00002-0), 1998.
- Koop, T., Bookhold, J., Shiraiwa, M., and Poschl, U.: Glass transition and phase state of organic compounds: dependency on molecular properties and implications for secondary organic aerosols in the atmosphere, *Phys. Chem. Chem. Phys.*, 13, 19238-19255, <https://doi.org/10.1039/c1cp22617g>, 2011.
- 555 Krieger, U. K., Marcolli, C., and Reid, J. P.: Exploring the complexity of aerosol particle properties and processes using single particle techniques, *Chem. Soc. Rev.*, 41, 6631-6662, <https://doi.org/10.1039/c2cs35082c>, 2012.
- Kuwata, M., and Martin, S. T.: Phase of atmospheric secondary organic material affects its reactivity, *Proc. Natl. Acad. Sci. U. S. A.*, 109, 17354-17359, <https://doi.org/10.1073/pnas.1209071109>, 2012.
- 560 Lamb, D., Moyle, A. M., and Brune, W. H.: The environmental control of individual aqueous particles in a cubic electrodynamic levitation system, *Aerosol Sci. Technol.*, 24, 263-278, <https://doi.org/10.1080/02786829608965371>, 1996.
- Laskin, A., Cowin, J. P., and Iedema, M. J.: Analysis of individual environmental particles using modern methods of electron microscopy and X-ray microanalysis, *J. Electron Spectrosc. Relat. Phenom.*, 150, 260-274, <https://doi.org/10.1016/j.elspec.2005.06.008>, 2006.
- 565 Laskin, A., Moffet, R. C., Gilles, M. K., Fast, J. D., Zaveri, R. A., Wang, B. B., Nigge, P., and Shutthanandan, J.: Tropospheric chemistry of internally mixed sea salt and organic particles: Surprising reactivity of NaCl with weak organic acids, *J. Geophys. Res.-Atmos.*, 117, D15302, <https://doi.org/10.1029/2012jd017743>, 2012.
- Laskina, O., Morris, H. S., Grandquist, J. R., Qin, Z., Stone, E. A., Tivanski, A. V., and Grassian, V. H.: Size matters in the water uptake and hygroscopic growth of atmospherically relevant multicomponent aerosol particles, *J. Phys. Chem. A*, 119, 4489-4497, <https://doi.org/10.1021/jp510268p>, 2015.
- 570 Leng, C. B., Pang, S. F., Zhang, Y., Cai, C., Liu, Y., and Zhang, Y. H.: Vacuum FTIR observation on the dynamic hygroscopicity of aerosols under pulsed relative humidity, *Environ. Sci. Technol.*, 49, 9107-9115, <https://doi.org/10.1021/acs.est.5b01218>, 2015.
- Li, X., Gupta, D., Lee, J., Park, G., and Ro, C.-U.: Real-time investigation of chemical compositions and hygroscopic properties of aerosols generated from NaCl and malonic acid mixture solutions using in situ Raman microspectrometry, *Environ. Sci. Technol.*, 51, 263-270, <https://doi.org/10.1021/acs.est.6b04356>, 2017.
- 575 Ling, T. Y., and Chan, C. K.: Partial crystallization and deliquescence of particles containing ammonium sulfate and dicarboxylic acids, *J. Geophys. Res.-Atmos.*, 113, D14205, <https://doi.org/10.1029/2008jd009779>, 2008.
- Liu, Y., Yang, Z. W., Desyaterik, Y., Gassman, P. L., Wang, H., and Laskin, A.: Hygroscopic behavior of substrate-deposited particles studied by micro-FT-IR spectroscopy and complementary methods of particle analysis, *Anal. Chem.*, 80, 633-642, <https://doi.org/10.1021/ac701638r>, 2008.
- 580

- Ma, Q. X., and He, H.: Synergistic effect in the humidifying process of atmospheric relevant calcium nitrate, calcite and oxalic acid mixtures, *Atmos. Environ.*, 50, 97-102, <https://doi.org/10.1016/j.atmosenv.2011.12.057>, 2012.
- Ma, Q. X., Ma, J. Z., Liu, C., Lai, C. Y., and He, H.: Laboratory study on the hygroscopic behavior of external and internal C₂–C₄ dicarboxylic acid–NaCl mixtures, *Environ. Sci. Technol.*, 47, 10381-10388, <https://doi.org/10.1021/es4023267>, 2013.
- Ma, Q. X., Liu, C., Ma, J. Z., Chu, B. W., and He, H.: A laboratory study on the hygroscopic behavior of H₂C₂O₄-containing mixed particles, *Atmos. Environ.*, 200, 34-39, <https://doi.org/10.1016/j.atmosenv.2018.11.056>, 2019a.
- Ma, Q. X., Zhong, C., Liu, C., Liu, J., Ma, J. Z., Wu, L. Y., and He, H.: A comprehensive study about the hygroscopic behavior of mixtures of oxalic acid and nitrate salts: Implication for the occurrence of atmospheric metal oxalate complex, *ACS Earth Space Chem.*, 3, 1216-1225, <https://doi.org/10.1021/acsearthspacechem.9b00077>, 2019b.
- Ma, S. S., Yang, W., Zheng, C. M., Pang, S. F., and Zhang, Y. H.: Subsecond measurements on aerosols: From hygroscopic growth factors to efflorescence kinetics, *Atmos. Environ.*, 210, 177-185, <https://doi.org/10.1016/j.atmosenv.2019.04.049>, 2019c.
- Ma, S. S., Chen, Z., Pang, S. F., and Zhang, Y. H.: Observations on hygroscopic growth and phase transitions of mixed 1, 2, 6-hexanetriol/(NH₄)₂SO₄ particles: investigation of the liquid–liquid phase separation (LLPS) dynamic process and mechanism and secondary LLPS during the dehumidification, *Atmos. Chem. Phys.*, 21, 9705-9717, <https://doi.org/10.5194/acp-21-9705-2021>, 2021a.
- Ma, S. S., Pang, S. F., Li, J., and Zhang, Y. H.: A review of efflorescence kinetics studies on atmospherically relevant particles, *Chemosphere*, 277, 130320, <https://doi.org/10.1016/j.chemosphere.2021.130320>, 2021b.
- Marcolli, C., Luo, B., and Peter, T.: Mixing of the organic aerosol fractions: Liquids as the thermodynamically stable phases, *J. Phys. Chem. A*, 108, 2216-2224, <https://doi.org/10.1021/jp036080l>, 2004.
- Mcfiggans, G., Artaxo, P., Baltensperger, U., Coe, H., Facchini, M. C., Feingold, G., Fuzzi, S., Gysel, M., Laaksonen, A., and Lohmann, U.: The effect of physical and chemical aerosol properties on warm cloud droplet activation, *Atmos. Chem. Phys.*, 6, 2593-2649, <https://doi.org/10.5194/acp-6-2593-2006>, 2006.
- Mikhailov, E., Vlasenko, S., Martin, S. T., Koop, T., and Schl, U. P.: Amorphous and crystalline aerosol particles interacting with water vapor: conceptual framework and experimental evidence for restructuring, phase transitions and kinetic limitations, *Atmos. Chem. Phys.*, 9, 9491-9522, <https://doi.org/10.5194/acp-9-9491-2009>, 2009.
- Ovadnevaite, J., Zuend, A., Laaksonen, A., Sanchez, K. J., Roberts, G., Ceburnis, D., Decesari, S., Rinaldi, M., Hodas, N., Facchini, M. C., Seinfeld, J. H., and O' Dowd, C.: Surface tension prevails over solute effect in organic-influenced cloud droplet activation, *Nature*, 546, 637-641, <https://doi.org/10.1038/nature22806>, 2017.
- Parsons, M. T., Knopf, D. A., and Bertram, A. K.: Deliquescence and crystallization of ammonium sulfate particles internally mixed with water-soluble organic compounds, *J. Phys. Chem. A*, 108, 11600-11608, <https://doi.org/10.1021/jp0462862>, 2004.
- Peng, C. G., Chan, M. N., and Chan, C. K.: The hygroscopic properties of dicarboxylic and multifunctional acids: Measurements and UNIFAC predictions, *Environ. Sci. Technol.*, 35, 4495-4501, <https://doi.org/10.1021/es0107531>, 2001.
- Peng, C., Jing, B., Guo, Y. C., Zhang, Y. H., and Ge, M. F.: Hygroscopic behavior of multicomponent aerosols involving NaCl and dicarboxylic acids, *J. Phys. Chem. A*, 120, 1029-1038, <https://doi.org/10.1021/acs.jpca.5b09373>, 2016.
- Peng, C. G., and Chan, C. K.: The water cycles of water-soluble organic salts of atmospheric importance, *Atmos. Environ.*, 35, 1183-1192, [https://doi.org/10.1016/S1352-2310\(00\)00426-X](https://doi.org/10.1016/S1352-2310(00)00426-X), 2001.
- Pope, F. D., Dennis-Smith, B. J., Griffiths, P. T., Clegg, S. L., and Cox, R. A.: Studies of single aerosol particles containing malonic acid, glutaric acid, and their mixtures with sodium chloride. I. Hygroscopic growth, *J. Phys. Chem. A*, 114, 5335-5341, <https://doi.org/10.1021/jp100059k>, 2010.
- Pöschl, U.: Atmospheric aerosols: composition, transformation, climate and health effects, *Angew. Chem., Int. Ed.*, 44, 7520-7540, <https://doi.org/10.1002/anie.200501122>, 2006.
- Ramanathan, V., Crutzen, P. J., Kiehl, J. T., and Rosenfeld, D.: Aerosols, climate, and the hydrological cycle, *Science*, 294, 2119-2124, <https://doi.org/10.1126/science.1064034>, 2001.
- Ray, A. K., Davis, E. J., and Ravindran, P.: Determination of ultra-low vapor pressures by submicron droplet evaporation, *J. Chem. Phys.*, 71, 582-587, <https://doi.org/10.1063/1.438408>, 1979.
- Reid, R. C., Prausnitz, J. M., and Poling, B. E.: The properties of gases and liquids, McGraw-Hill, New York, 1987.

- 630 Ren, H. M., Cai, C., Leng, C. B., Pang, S. F., and Zhang, Y. H.: Nucleation kinetics in mixed NaNO_3 /glycerol droplets investigated with the FTIR-ATR technique, *J. Phys. Chem. B*, 120, 2913-2920, <https://doi.org/10.1021/acs.jpcc.5b12442>, 2016.
Rosenberger, T., Münzer, A., Kiesler, D., Wiggers, H., and Kruis, F. E.: Ejector-based sampling from low-pressure aerosol reactors, *J. Aerosol Sci.*, 123, 105-115, <https://doi.org/10.1016/j.jaerosci.2018.06.003>, 2018.
- 635 Schilling, C., and Winterer, M.: Preserving particle characteristics at increasing production rate of ZnO nanoparticles by chemical vapor synthesis, *Chem. Vap. Deposition*, 20, 138-145, <https://doi.org/10.1002/cvde.201307094>, 2014.
Shao, X., Zhang, Y., Pang, S. F., and Zhang, Y. H.: Vacuum FTIR observation on hygroscopic properties and phase transition of malonic acid aerosols, *Chem. Phys.*, 483-484, 7-11, <https://doi.org/10.1016/j.chemphys.2016.11.001>, 2017.
Shao, X., Wu, F. M., Yang, H., Pang, S. F., and Zhang, Y. H.: Observing HNO_3 release dependent upon metal complexes in malonic acid/nitrate droplets, *Spectrochim. Acta A*, 201, 399-404, <https://doi.org/10.1016/j.saa.2018.05.026>, 2018.
- 640 Shiraiwa, M., Li, Y., Tsimpidi, A. P., Karydis, V. A., Berkemeier, T., Pandis, S. N., Lelieveld, J., Koop, T., and Poschl, U.: Global distribution of particle phase state in atmospheric secondary organic aerosols, *Nat. Commun.*, 8, 15002, <https://doi.org/10.1038/ncomms15002>, 2017.
Song, C. H., and Carmichael, G. R.: Gas-particle partitioning of nitric acid modulated by alkaline aerosol, *J. Atmos. Chem.*, 40, 1-22, <https://doi.org/10.1023/A:1010657929716>, 2001.
- 645 Soonsin, V., Zardini, A. A., Marcolli, C., Zuend, A., and Krieger, U. K.: The vapor pressures and activities of dicarboxylic acids reconsidered: the impact of the physical state of the aerosol, *Atmos. Chem. Phys.*, 10, 11753-11767, <https://doi.org/10.5194/acp-10-11753-2010>, 2010.
Stevens, B., and Feingold, G.: Untangling aerosol effects on clouds and precipitation in a buffered system, *Nature*, 461, 607-613, <https://doi.org/10.1038/nature08281>, 2009.
- 650 Sullivan, R. C., and Prather, K. A.: Investigations of the diurnal cycle and mixing state of oxalic acid in individual particles in Asian aerosol outflow, *Environ. Sci. Technol.*, 41, 8062-8069, <https://doi.org/10.1021/es071134g>, 2007.
Tang, I. N., and Fung, K. H.: Hydration and Raman scattering studies of levitated microparticles: $\text{Ba}(\text{NO}_3)_2$, $\text{Sr}(\text{NO}_3)_2$, and $\text{Ca}(\text{NO}_3)_2$, *J. Chem. Phys.*, 106, 1653-1660, <https://doi.org/10.1063/1.473318>, 1997.
- 655 Tang, M. J., Chan, C. K., Li, Y. J., Su, H., Ma, Q. X., Wu, Z., Zhang, G., Wang, Z., Ge, M., Hu, M., He, H., and Wang, X.: A review of experimental techniques for aerosol hygroscopicity studies, *Atmos. Chem. Phys.*, 19, 12631-12686, <https://doi.org/10.5194/acp-19-12631-2019>, 2019.
Tervahattu, H., Hartonen, K., Kerminen, V.-M., Kupiainen, K., Aarnio, P., Koskentalo, T., Tuck, A. F., and Vaida, V.: New evidence of an organic layer on marine aerosols, *J. Geophys. Res.-Atmos.*, 107, 4053, <https://doi.org/10.1029/2000JD000282>, 2002.
- 660 Villepin, J. d., and Novak, A.: Vibrational spectra of and isotope effect in hydrogen bonded potassium hydrogen oxalate, *Spectrosc. Lett.*, 4, 1-8, <https://doi.org/10.1080/00387017108078634>, 1971.
Virtanen, A., Joutsensaari, J., Koop, T., Kannosto, J., Yli-Pirila, P., Leskinen, J., Makela, J. M., Holopainen, J. K., Poschl, U., Kulmala, M., Worsnop, D. R., and Laaksonen, A.: An amorphous solid state of biogenic secondary organic aerosol particles, *Nature*, 467, 824-827, <https://doi.org/10.1038/nature09455>, 2010.
- 665 Wang, B. B., and Laskin, A.: Reactions between water-soluble organic acids and nitrates in atmospheric aerosols: Recycling of nitric acid and formation of organic salts, *J. Geophys. Res.-Atmos.*, 119, 3335-3351, <https://doi.org/10.1002/2013jd021169>, 2014.
Wang, G., Xie, M., Hu, S., Gao, S., Tachibana, E., and Kawamura, K.: Dicarboxylic acids, metals and isotopic compositions of C and N in atmospheric aerosols from inland China: implications for dust and coal burning emission and secondary aerosol formation, *Atmos. Chem. Phys.*, 10, 6087-6096, <https://doi.org/10.5194/acp-10-6087-2010>, 2010.
- 670 Wang, N., Jing, B., Wang, P., Wang, Z., Li, J. R., Pang, S. F., Zhang, Y. H., and Ge, M. F.: Hygroscopicity and compositional evolution of atmospheric aerosols containing water-soluble carboxylic acid salts and ammonium sulfate: Influence of ammonium depletion, *Environ. Sci. Technol.*, 53, 6225-6234, <https://doi.org/10.1021/acs.est.8b07052>, 2019.
- 675 Wang, X. W., Jing, B., Tan, F., Ma, J. B., Zhang, Y. H., and Ge, M. F.: Hygroscopic behavior and chemical composition evolution of internally mixed aerosols composed of oxalic acid and ammonium sulfate, *Atmos. Chem. Phys.*, 17, 12797-12812, <https://doi.org/10.5194/acp-17-12797-2017>, 2017.

- Wu, F. M., Wang, N., Pang, S. F., and Zhang, Y. H.: Hygroscopic behavior and fractional crystallization of mixed $(\text{NH}_4)_2\text{SO}_4$ /glutaric acid aerosols by vacuum FTIR, *Spectrochim. Acta A*, 208, 255-261, <https://doi.org/10.1016/j.saa.2018.10.010>, 2019a.
- Wu, F. M., Wang, X. W., Pang, S. F., and Zhang, Y. H.: Measuring hygroscopicity of internally mixed NaNO_3 and glutaric acid particles by vacuum FTIR, *Spectrochim. Acta A*, 219, 104-109, <https://doi.org/10.1016/j.saa.2019.04.034>, 2019b.
- Wu, Z. J., Nowak, A., Poulain, L., Herrmann, H., and Wiedensohler, A.: Hygroscopic behavior of atmospherically relevant water-soluble carboxylic salts and their influence on the water uptake of ammonium sulfate, *Atmos. Chem. Phys.*, 11, 12617-12626, <https://doi.org/10.5194/acp-11-12617-2011>, 2011.
- Yeung, M. C., Ling, T. Y., and Chan, C. K.: Effects of the polymorphic transformation of glutaric acid particles on their deliquescence and hygroscopic properties, *J. Phys. Chem. A*, 114, 898-903, <https://doi.org/10.1021/jp908250v>, 2010.
- Zhang, Q. N., Zhang, Y., Cai, C., Guo, Y. C., Reid, J. P., and Zhang, Y. H.: In situ observation on the dynamic process of evaporation and crystallization of sodium nitrate droplets on a ZnSe substrate by FTIR-ATR, *J. Phys. Chem. A*, 118, 2728-2737, <https://doi.org/10.1021/jp412073c>, 2014.
- Zhang, Y., Cai, C., Pang, S. F., Reid, J. P., and Zhang, Y. H.: A rapid scan vacuum FTIR method for determining diffusion coefficients in viscous and glassy aerosol particles, *Phys. Chem. Chem. Phys.*, 19, 29177-29186, <https://doi.org/10.1039/c7cp04473a>, 2017.
- Zhao, H., Liu, X. F., and Tse, S. D.: Effects of pressure and precursor loading in the flame synthesis of titania nanoparticles, *J. Aerosol Sci.*, 40, 919-937, <https://doi.org/10.1016/j.jaerosci.2009.07.004>, 2009.