



1 **Projected increases in wildfires may challenge regulatory curtailment of PM_{2.5} over the**
2 **eastern US by 2050**

3

4 Chandan Sarangi^{1,2*}, Yun Qian^{2*}, L. Ruby Leung², Yang Zhang³, Yufei Zou^{4,2}, Yuhang
5 Wang⁴

6

7 ¹ Indian Institute of Technology Madras, Chennai, India

8 ² Pacific Northwest National Laboratory, Richland, WA, USA

9 ³ Department of Civil and Environmental Engineering, Northeastern University, Boston, MA

10 ⁴ School of Earth and Atmospheric Sciences, Georgia Institute of Technology, Atlanta, GA,
11 USA.

12

13

14 *Corresponding Authors:

15 chandansarangi@iitm.ac.in

16 yun.qian@pnnl.gov

17

18

19

20



21 **Abstract**

22 Anthropogenic contribution to the overall fine particulate matter (PM_{2.5}) concentrations has
23 been declining sharply in North America. In contrast, a steep rise in wildfire-induced air
24 pollution events with recent warming is evident in the region. Here, based on coupled fire-
25 climate-ecosystem model simulations, summertime wildfire-induced PM_{2.5} concentrations are
26 projected to nearly double in North America by the mid-21st century compared to the
27 present. More strikingly, the projected enhancement in fire-induced PM_{2.5} (~ 1-2 μg/m³) and
28 its contribution (~15-20%) to the total PM_{2.5} are distinctively significant in the eastern US.
29 This can be attributed to downwind transport of smoke from future enhancement of wildfires
30 in North America to the eastern US and associated positive climatic feedback on PM_{2.5} i.e.
31 increased atmospheric stability and reduced precipitation. Therefore, the anticipated
32 reductions in PM_{2.5} from regulatory controls on anthropogenic emissions could be
33 significantly compromised in the future in the densely populated eastern US.

34 **Key points:**

- 35 1) Wildfire-PM_{2.5} associations studied based on unprecedented two-way coupled fire-
36 climate-ecosystem model simulations
- 37 2) A steep rise in wildfire-induced air pollution events with recent warming is evident in
38 the region
- 39 3) The transported smoke from enhanced wildfires in North America can severely affect
40 air quality over Eastern US

41

42 **Keywords:** wildfire emissions, climate change, air quality, smoke transport, wildfire-climate-
43 ecosystem interactions



44 1. Introduction

45 Wildfires are widespread burning events in forests, shrub lands, and grazing lands. In
46 North America (mainly Canada and the US), particulate matter emissions from wildfires are a
47 significant source of regional air pollution (Shi et al., 2019; McClure and Jaffe, 2018; Van
48 Der Werf et al., 2010; Jaffe et al., 2008). Since the 1980s, the number of large wildfires and
49 the length of wildfire season have been increasing, and the trends are projected to continue in
50 the future over the western US, Alaska and Canada (Kitzberger et al., 2017; Kirchmeier-
51 Young et al., 2017; Abatzoglou and Williams, 2016; Partain et al., 2016; Jolly et al., 2015;
52 Westerling et al., 2006; Gillett et al., 2004). Accordingly, particulate emissions from wildfires
53 are also anticipated to increase in North America in the 21st century (Knorr et al., 2017; Liu
54 et al., 2016; Val Martin et al., 2015). Human exposure to high concentrations of wildfire-
55 emitted airborne particulate matter of diameter ≤ 2.5 μm ($\text{PM}_{2.5}$) is known to have substantial
56 adverse effects on pulmonary and cardiovascular functioning (Anjali et al., 2019; Black et al.,
57 2017), which contribute significantly to global and regional all-cause mortality (Zhang et al.,
58 2020; Hong et al., 2019; Yang et al., 2019; Ford et al., 2018; Johnston. et al., 2012).
59 Therefore, a better understanding of the future changes in wildfire-induced $\text{PM}_{2.5}$ and its
60 contribution to the total surface $\text{PM}_{2.5}$ is essential.

61 In the last two decades, ambient air quality in the US has substantially improved due
62 to a decline in $\text{PM}_{2.5}$ by ~ 40 % (US EPA, 2018). The decrease in $\text{PM}_{2.5}$ is primarily due to
63 curtailment of anthropogenic emissions resulting from US-based efforts to meet regulations
64 such as the Clean Air Act (US EPA, 2009), Cross-State Air Pollution Rule, Regional Haze
65 Rule, and the motor vehicles emissions standards. Consequently, air quality over the
66 contiguous US (CONUS) and Canada has improved steadily such that it is predicted to
67 achieve the targeted National Ambient Air Quality Standards in the future (Nolte et al.,



68 2018). Under this promising scenario, the influence of wildfire-emissions on the total $PM_{2.5}$
69 becomes even more crucial. Depending on the competition between climate-induced increase
70 in wildfires and the regulatory control on anthropogenic emissions, future enhancement in
71 wildfire-induced $PM_{2.5}$ may compromise the reduction in anthropogenic $PM_{2.5}$ concentrations
72 in certain regions. In agreement, recent studies have highlighted the potential for future
73 enhancement in wildfire-induced pollution to diminish the reducing trend in $PM_{2.5}$, primarily
74 over the western US (O'Dell et al., 2019; Ford et al., 2018; Val Martin et al., 2015; Yue et al.,
75 2013).

76 While the fractional wildfire burnt area and fire intensities are the greatest over the
77 western US and Canadian regions within North America, anthropogenic emissions dominate
78 the ambient $PM_{2.5}$ concentration over the eastern US. The inherent geographical separation
79 between the regions with large wildfire emissions and anthropogenic emissions leads to a
80 pertinent question: will future enhancement in wildfires over the western US and Canada
81 have significant effects on $PM_{2.5}$ over the eastern US? Addressing this question is crucial
82 because the declining trend in $PM_{2.5}$ over the eastern US is the major contributor to the
83 observed 40% decrease in $PM_{2.5}$ over the US in the last two decades (US EPA, 2018).
84 Eastward advection of wildfire smoke from Canada and the western US has been found to
85 severely hamper the surface air quality of the central and eastern US under the influence of
86 the prevailing westerlies during the summer months (Brey et al., 2018; Wu et al., 2018;
87 Gunsch et al., 2018; Kaulfus et al., 2017; Dempsey, 2013). The transported wildfire smoke
88 can influence the meteorology and climate via the radiative impact of carbonaceous
89 emissions, changes in land albedo and cloud system perturbations (Ward et al., 2012; Liu et
90 al., 2014). These fire-weather interactions can have positive feedback on the locally-emitted
91 $PM_{2.5}$ in the eastern US by surface cooling and boundary layer suppression (Guan et al.,
92 2020). At the same time, fire-triggered ecosystem changes can induce negative feedback on



93 PM_{2.5} by reducing the future wildfires over North America (Zou et al., 2020). Thus, two-
94 way interactions between fires and climate that are important for predicting future changes in
95 wildfire locations, intensities, and durations (Harris et al., 2016) as well as associated
96 particulate emissions is essential. However, past studies have mostly employed simple
97 statistical models based on statistical regressions of present-day fire burnt area on the
98 meteorological fields (Liu et al., 2016; Spracklen et al., 2009; Yue et al., 2013; Val Martin et
99 al., 2015), and more recently, one-way coupled modelling (Ford et al., 2018; O’Dell et al.,
100 2019).

101 Here, based on new two-way coupled fire-climate-ecosystem simulations, we
102 demonstrate the significance of wildfire-induced contributions to ambient PM_{2.5} over the
103 eastern US due to enhanced wildfire smoke transportation and smoke-induced changes in
104 weather in eastern US. This enhancement in wildfire-induced PM_{2.5} may potentially challenge
105 the targeted policy-driven reduction of PM_{2.5} in the eastern US. Next, our model setup,
106 experiments and methodology are explained in Section 2, followed by results and discussion
107 in Section 3. The study is summarized in Section 4.

108 **2. Materials and Methods**

109 **2.1. RESFire-CESM Model description**

110 We employ the open-source REgion-Specific ecosystem feedback fire (RESFire)
111 model coupled with the Community Land Model version 4.5 and the Community
112 Atmosphere Model version 5 (CAM5) of the Community Earth System Model (CESM)
113 version 1 (Zou et al., 2019; Neale et al., 2013) to perform two-way coupled simulations.
114 RESFire provides state-of-the-art capabilities to simulate the complex fire-climate-ecosystem
115 interactions globally for fires occurring over wildland, cropland, and peatland. Although
116 wildfires dominate in the North American region, RESFire simulates both wildfires and



117 prescribed fires. Moreover, this integrated setup includes climatic feedback from fire-induced
118 aerosol direct and indirect radiative effects and associated weather changes. It also includes
119 feedback from fire-induced vegetation distribution changes and associated biophysical
120 processes such as evapotranspiration and surface albedo. Sofiev et al. (2012) described the
121 fire plume rise parameterization. Other features in CLM4.5 and CAM5, such as the
122 photosynthesis scheme (Sun et al., 2012), the MAM3 aerosol module (Liu et al., 2012), and
123 the cloud macrophysics scheme (Park et al., 2014), allow for more comprehensive
124 assessments of the climate effects of fires through their interactions with vegetation and
125 clouds. Fire-ecosystem interactions are modelled by simulating fire-induced vegetation
126 mortality and regrowth (and associated land cover change) in RESFire. This approach has
127 been introduced in Zou et al. (2019) and the simulated ecological and climatic effects of
128 wildfires have been evaluated in two sets of sensitivity experiments in Zou et al. (2020).
129 Although fire-climate-ecosystem interactions are considered in this study, our focus is on the
130 fire-induced changes in PM_{2.5} over Canada and the US, so the two vegetation-focused
131 sensitivity experiments reported in Zou et al. (2020) are not included in this paper. Please
132 refer to Zou et al. (2019) and Zou et al. (2020) for more details about the simulation of fire-
133 ecosystem interactions.

134 **2.2 Numerical Experiment and Methodology**

135 We designed two sets of simulations for the present day and future scenarios to
136 quantify the impacts of fire-climate-ecosystem interactions (Table 1). The spatial resolution is
137 0.9° (lat) \times 1.25° (lon) with a time step of 30 min. In each set of simulations, we conducted a
138 default all emission included control run (X_{ALL} , where $x=2000$ or 2050 indicates the present
139 day or future, respectively) and a sensitivity run with no wildfire emissions to the atmosphere
140 (X_{WEF} , where X is the same as for the control runs). The ALL runs are designed to simulate
141 fully interactive fire disturbances such as fire emissions with plume rise and fire induced land



142 cover changes of the present day (representative of the 2000s, 2000_{ALL}) and a moderate future
143 emission scenario (representative of the 2050s, 2050_{ALL}) via the RCP4.5. The only difference
144 between the ALL and WEF scenario is that wildfire emissions are absent in the WEF
145 scenario. Specifically, in the WEF runs, the online simulated fire emissions are not passed to
146 the CAM5 atmosphere model so that the difference between the ALL and WEF runs can be
147 used to isolate the atmospheric impacts of fire-climate interactions.

148 Table 1: Summary of the sensitivity simulations performed

Scenario	Present-day		Future	
Experiment Name	2000 _{ALL}	2000 _{WEF}	2050 _{ALL}	2050 _{WEF}
Simulation years	2001-2010	2001-2010	2051-2060	2051-2060
Atmosphere	CAM5	CAM5	CAM5	CAM5
Land	CLM4.5	CLM4.5	CLM4.5	CLM4.5
Ocean	Climatology	Climatology	RCP4.5	RCP4.5
Sea ice	Climatology	Climatology	RCP4.5	RCP4.5
Non-fire emissions	ACCMIP	ACCMIP	RCP4.5	RCP4.5
Fire emissions	Online fire aerosols with plume rise	—	Online fire aerosols with plume rise	—
Land cover	Fire disturbances on present-day condition	Fire disturbances on present-day condition	Fire disturbances on RCP4.5 condition	Fire disturbances on RCP4.5 condition

149

150 For the present-day experiments, we used the spun-up states from Zou et al. (2019) as
151 initial conditions for both meteorological and chemical variables. Sea surface temperature
152 (SST) for the present day was obtained from the Met Office Hadley Centre (HadISST).
153 Present-day non-fire emissions from anthropogenic and other sources were based on
154 ACCMIP (Lamarque et al., 2010) for the year 2000. We replaced the prescribed GFED2 fire
155 emissions (van der Werf et al., 2006) in the default setting of CESM with the online-coupled
156 fire emissions generated by the RESFire model. Zou et al. (2019) provided more details of
157 the physics parameterizations and modeling experiment settings used in these simulations.



158 Land use and land cover data for 2000 and 2050 from the Land-Use History A product (Hurtt
159 et al., 2006) are used to initialize the 2000_{ALL}/2000_{WEF} and 2050_{ALL}/2050_{WEF} simulations,
160 respectively. Following the above setup, the future scenario 2050_{ALL} experiment accounts for
161 both fuel load changes associated with the projected land use and land cover change
162 (LULCC) in the 2050s and fire weather changes driven by the SST and sea ice forcing from a
163 coupled CESM simulation following the greenhouse gas (GHG) forcing of the RCP4.5
164 scenario. The global mean GHG mixing ratios in the CAM5 atmosphere model were fixed at
165 the year 2000 levels in all the present-day experiments and they were replaced by those of the
166 RCP4.5 scenario with the well-mixed assumption and monthly variations. However, the
167 future population and socioeconomic conditions were identical to those of the present day so
168 there was no explicit impact of human-induced mitigation/enhancement effects on wildfires
169 in the future projection in all the future experiments. Future human impacts were considered
170 implicitly in LULCC-induced fuel load changes in the RCP4.5 scenario.

171 The net projected changes by 2050s in emissions, meteorology and air quality during
172 summer (JJA) months are estimated by comparing decadal-mean values simulated by
173 2000_{ALL} with 2050_{ALL}. Wildfire-induced enhancement in PM_{2.5} concentration in the present
174 day and mid-21st century is estimated by comparing 2000_{ALL} with 2000_{WEF} and 2050_{ALL} with
175 2050_{WEF}, respectively. Further, the projected increase in wildfire-induced PM_{2.5} in the future
176 is calculated by comparing the simulated wildfire effect of the 2050s (2050_{ALL}-2050_{WEF}) with
177 that of the 2000s (2000_{ALL}-2000_{WEF}). With large spatiotemporal variability, the projected
178 changes in transported fire-emissions from the western US and Canada to the eastern US by
179 the 2050s and the corresponding impacts are summarized using probability distribution
180 functions. The latter provide information not only for the mean but also variability and
181 extreme values to quantify the simulated changes for the three subregions.

182



183 **3. Results and Discussion**

184 **3.1 Model Evaluation**

185 Zou et al. (2019) performed comprehensive evaluation of the RESFire simulated
186 wildfire burnt area distribution, associated carbon emissions and terrestrial carbon balance to
187 demonstrate reasonable model skill. Zou et al. (2020) compares global fire simulations by
188 CESM-RESFire with modeling results reported in the literature to show better agreement
189 with the GFED4.1s benchmark data and predicts more prominent changes in the future than
190 those predicted by Kloster et al. (2010, 2012). These differences might come from differences
191 in the climate sensitivities of the fire models and scenarios and other input data used to make
192 future projections.

193 Here, we evaluate the simulated surface $PM_{2.5}$ against satellite-estimates (Figures 1)
194 over North America. The $PM_{2.5}$ concentration is calculated as the sum of sulfate, nitrate, fine
195 sea salt (first 2 size bins), fine dust (first size bin), black carbon (BC), and organic aerosol
196 (OC) at the surface-level of model. OC is the sum of primary organic matter (POM) and
197 secondary organic aerosol (SOA), and SOA is the sum of secondary species formed from
198 toluene, monoterpenes, isoprene, benzene, and xylene. Figure 1 compares the observed and
199 simulated mean annual $PM_{2.5}$ averaged over 2001-2010. The 10-year average satellite AOD-
200 derived annual mean surface $PM_{2.5}$ concentrations (Van Donkelaar et al., 2018) are regridded
201 to the model grid (Figure 1A) and then compared with the RESFire simulations in the
202 2000_{ALL} present-day run (Figure 1B). The spatial distribution of annual surface $PM_{2.5}$ is
203 reasonably well simulated but also have some biases. To quantify the biases, we also
204 estimated the correlation coefficient as well as normalized mean biases (NMB) of the
205 simulated values compared against the satellite retrieved values over two subregions.
206 Quantitatively, the NMB values over the western US (WUS) and eastern US (EUS) are 18%
207 and 7%, respectively (Figure 1C-D). In addition, the spatial variability of the 2001-2010

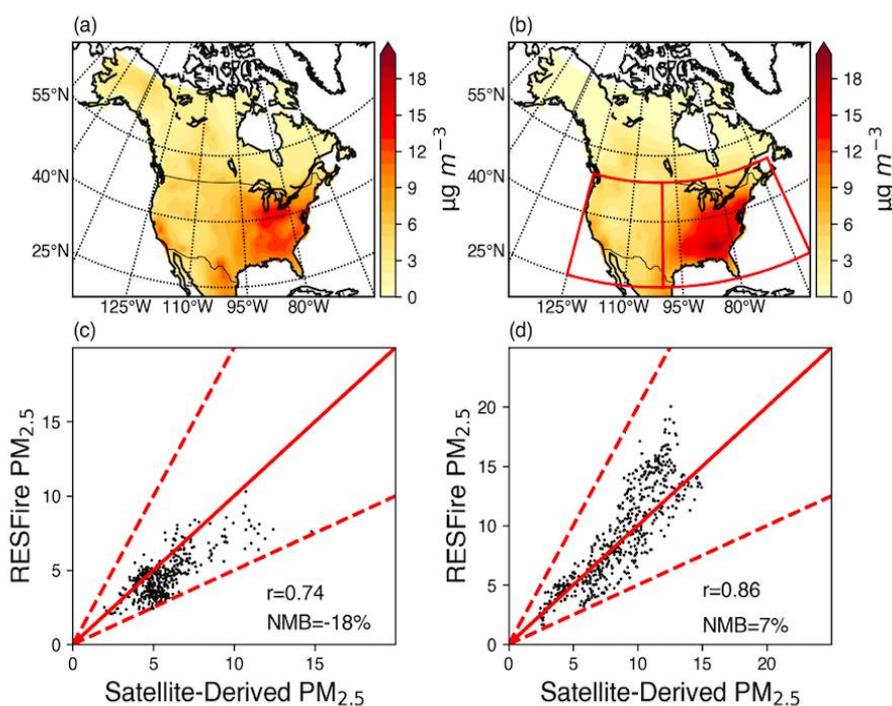


208 averaged annual AOD distribution (Supplementary Figure 1) is also well represented in our
209 simulation, although the model underestimates high AOD values. Similar spatial variability
210 and biases in AOD and PM_{2.5} were also found when a comparison was performed for only
211 summer months (JJA). Our simulation biases are within the uncertainty range among the
212 various satellite and ground-based datasets, which have normalized mean biases ranging from
213 -3.3% to 33.3% when benchmarked against the ground-based IMPROVE data over the
214 contiguous US (Diao et al., 2019). Val Martin et al. (2015) also reported a similar range of
215 biases of PM_{2.5} in their simulation over the US.

216 Discrepancies between the simulated and observed PM_{2.5} values may be attributed to
217 several potential reasons. First, the satellite-derived data has a non-zero lower bound of PM_{2.5}
218 concentrations, so the ambient background concentrations for pristine regions such as the
219 remote areas in Alaska, Canada, and the western US may be overestimated. Second, year
220 2000-based constant non-fire emissions were used in the RESFire simulation, which may
221 result in overestimation of the PM_{2.5} concentrations from non-fire sources during 2001-2010
222 when anthropogenic emissions and PM_{2.5} concentrations continue to decrease (US EPA,
223 2018). This overestimation is prominent in regions dominated by non-fire sources such as the
224 southeastern US. Third, large uncertainties in fuel consumption and emission factors preclude
225 an accurate estimation of the primary fire emissions in the model, especially for the eastern
226 US where large fractions of low-intensity prescribed fires consume only under-canopy fuels
227 such as litter and duff layers. The fire model may fail to capture the subtle distinctions
228 between low-intensity prescribed fires and forest fires, so more fuels are consumed and result
229 in higher emissions. Lastly, comparison of a coarsely resolved simulation against in-situ
230 observations also contributes to uncertainty. Differences in the degree to which fire-climate
231 interactions and other physical processes and feedbacks are represented by the models can
232 explain the slight differences in estimating the present day mean wildfire-induced change in



233 PM_{2.5} over local and downwind regions between our simulations and previous studies.
234 Nonetheless, reasonable simulation of the spatial distribution of wildfire burnt area, AOD,
235 and near surface particulate concentration (mean bias of ~10-20 %) instills confidence about
236 the fidelity of our model setup in particulate pollution simulation, which is the focus of this
237 study.



238
239

240 Figure 1: Comparison of the 10-year (2001-2010) averaged annual mean surface PM_{2.5}
241 concentration between observations and RESFire simulations. (a) Satellite-derived surface
242 PM_{2.5} concentrations (with dust and sea-salt removed) estimated by Donkelaar et al., 2018
243 (available at [https://sedac.ciesin.columbia.edu/data/set/sdei-global-annual-gwr-pm2-5-modis-](https://sedac.ciesin.columbia.edu/data/set/sdei-global-annual-gwr-pm2-5-modis-misr-seawifs-aod)
244 [misr-seawifs-aod](https://sedac.ciesin.columbia.edu/data/set/sdei-global-annual-gwr-pm2-5-modis-misr-seawifs-aod); last access: 5 November, 2021); (b) 2000_{ALL} Simulated surface PM_{2.5}
245 concentrations (with dust and sea-salt removed) averaged over 2001-2010; The red boxes
246 denote the two subregions (EUS and WUS) shown in Fig. 2 in the main text. (c) comparison
247 of simulated and satellite based gridded surface PM_{2.5} concentrations in the WUS subregion;
248 (d) same as (c) but in the EUS subregion. The red solid and dashed lines denote the 1:1 ratio
249 line and $\pm 100\%$ biases, respectively. The correlation coefficients and NMB values are shown
250 at the lower-right corner of each subplot.

251



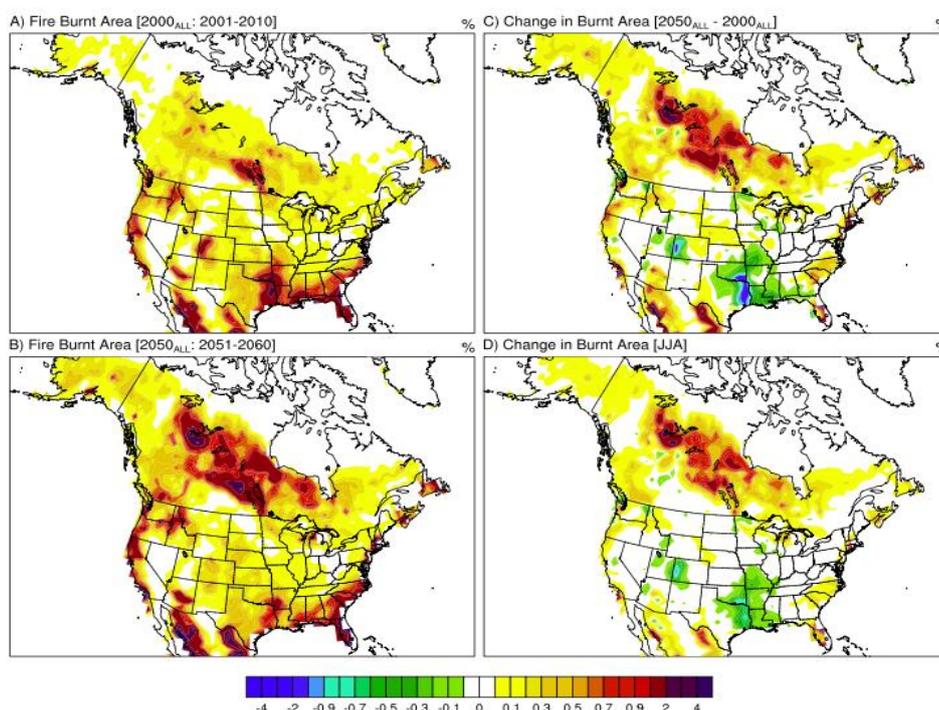
252 **3.2 Fire-induced changes in burnt area and PM_{2.5}**

253 The decadal-mean annual fire burnt area simulated for the present day shows
254 widespread wildfires over the entire North America (Figure 2A). Specifically, Canada and the
255 forested areas of the northwestern (> 36 N latitude) and southeastern (< 36 N latitude) US are
256 most intensely affected by wildfires in the present day. By the mid-21st century, a striking
257 increase of 2-5 times in fire burnt area is projected over Canada, Alaska, the Pacific
258 Northwest and portions of the western US by the 2050s (Figure 2B). A distinct positive shift
259 in the probability density function (PDF) of annual fire burnt area is evident in the future,
260 with the decadal-mean difference statistically significant at the 99% confidence level (Zou et
261 al., 2020). A small and statistically insignificant change in interannual variability (~ 0.4 Mha
262 yr⁻¹) of fire burnt areas is also simulated between the present and future. Specifically, our
263 model predicts more than a doubling of burnt area in boreal regions of Canada in the future,
264 in line with a previous projection for Canada (Wotton et al., 2017). Future enhancement in
265 fire burnt area is ~ 20-50% in most fire grids over the western coast of US, which is higher
266 than that over the eastern US (Figures 2A and 2C). The increase over the western US is closer
267 to the lower bound of that derived from statistical model ensemble projections for the western
268 US in the mid-21st century (Yue et al., 2013). The statistics-based projections of future burnt
269 area over North America were likely too high because fire-induced land cover change, fuel
270 load reduction and factors could induce a negative fire feedback, which was not considered in
271 previous fire projection studies (Zou et al. 2020).

272 Annual fire burnt area in the southeastern US shows a decline in the future (Figure
273 2C), as precipitation is projected to increase in that region (discussed later). Note that all
274 future fire changes between 2050_{ALL} and 2000_{ALL} are primarily associated with climate
275 warming in response to the increase in greenhouse gas (GHG) concentrations in the RCP4.5
276 scenario. No direct impacts of population and socioeconomic changes on wildfires are



277 included in our simulations, although these factors contribute to changes in GHG emissions
278 (via the RCP scenario) that influence the climate simulated in 2000_{ALL} and 2050_{ALL}. As about
279 80% of the projected fire changes in the future is restricted to the summer season (June
280 through August; JJA) (Figure 2D), we focus on analysis of the summer-mean wildfire-
281 induced PM_{2.5} and its projected future changes over North America.



282

283 **Figure 2: Spatial distribution of fire burnt area.** A-D, Spatial distribution of simulated
284 decadal-mean annual burnt area (as percentage) over North America for present day (A),
285 mid-21st century (B) and the net change between the 2050s and the 2000s (C). D, same as
286 (C), but for wildfire burnt area during summer only (June through August; JJA). The colorbar
287 illustrate grid fraction of area burnt.

288

289

The simulated 10-year averaged summer-mean wildfire-induced PM_{2.5} values in
290 2000_{ALL} are more than 0.5 $\mu\text{g}/\text{m}^3$ over a large part of North America in the present day, with
291 noticeably larger values ($> 1 \mu\text{g}/\text{m}^3$) in Canada and the northwestern, central, and
292 southeastern US (Figure 3A). Interestingly, the spatial distribution of wildfire-induced

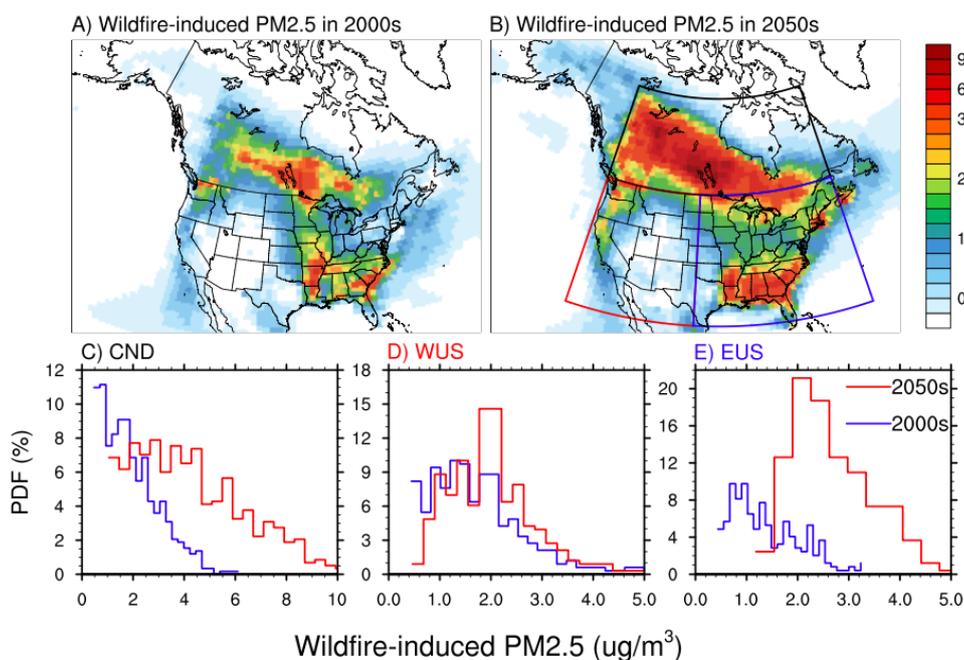


293 $PM_{2.5} > 1 \mu g/m^3$ resembles an inverted horse-shoe shape. The inverted horse-shoe shaped
294 spatial distribution is also consistent with the wildfire-smoke climatology derived from the
295 satellite-guided operational smoke product of the Hazard Mapping System (HMS) during
296 2005-2015 (Brey et al., 2018; Kaulfus et al., 2017). By the mid-21st century, the spatial extent
297 of the horse-shoe shape for areas with wildfire-induced $PM_{2.5}$ enhancement $> 1 \mu g/m^3$
298 expands significantly to span most regions of North America, with the most pronounced
299 enhancement occurring over Canada (Figure 3B). The PDFs of the spatial distribution for the
300 three regions can be seen in Figure 3C-E. Specifically, wildfire induced $PM_{2.5}$ in the 2000s
301 over Canada, WUS and EUS during summer is $\sim 1-3 \mu g/m^3$, $1-3 \mu g/m^3$ and $0.6-1.2 \mu g/m^3$,
302 respectively. Maximum values within the WUS region are found over the Pacific Northwest,
303 with most areas having wildfire induced $PM_{2.5}$ values of $\sim 2-3 \mu g/m^3$. Similarly, the southern
304 states have relatively high wildfire induced $PM_{2.5}$ concentrations of $\sim 2-4 \mu g/m^3$ within the
305 EUS in the present-day simulation.

306 Compared to the 2000s, the wildfire induced JJA averaged $PM_{2.5}$ values are almost
307 doubled to $\sim 3-6 \mu g/m^3$ over Canada in the 2050s (Figure 3B and Figure 3C). Consistently,
308 the values of wildfire induced $PM_{2.5}$ over WUS (mainly coastal) also doubled in 2050s
309 compared to 2000s, with modal values of $\sim 2-2.5 \mu g/m^3$ (Figure 3D). Most interestingly, the
310 enhancement in wildfire-induced summer-mean $PM_{2.5}$ over the northern EUS is also
311 significant by the 2050s (Figures 3B). Largely, the summer-mean wildfire-induced $PM_{2.5}$
312 concentration over EUS increases from ~ 0.8 to $\sim 2 \mu g/m^3$ in the mid-century to values of $1.2-$
313 $3.0 \mu g/m^3$ (Figure 3E). The summer-mean wildfire-induced $PM_{2.5}$ is thus projected to double
314 in North America by the 2050s compared to the 2000s, with a substantial coverage over the
315 EUS. Also, the spatial variability is enhanced which will enhance the exposure over
316 previously clean grids. An important finding from these PDFs appears to be that there are
317 fewer grids with $< 1 \mu g/m^3$ wildfire induced $PM_{2.5}$, or alternatively, that more regions are



318 being influenced by $\text{PM}_{2.5}$, and many areas that were already seeing wildfire impacts are
319 seeing enhanced impacts. Such enhancement is found not only at the surface but also in an
320 elevated atmospheric layer over EUS between 900 and 700 hPa. This is nonintuitive given
321 the fact that the increase in fire-burnt area by mid-century over the EUS is not substantial.



322

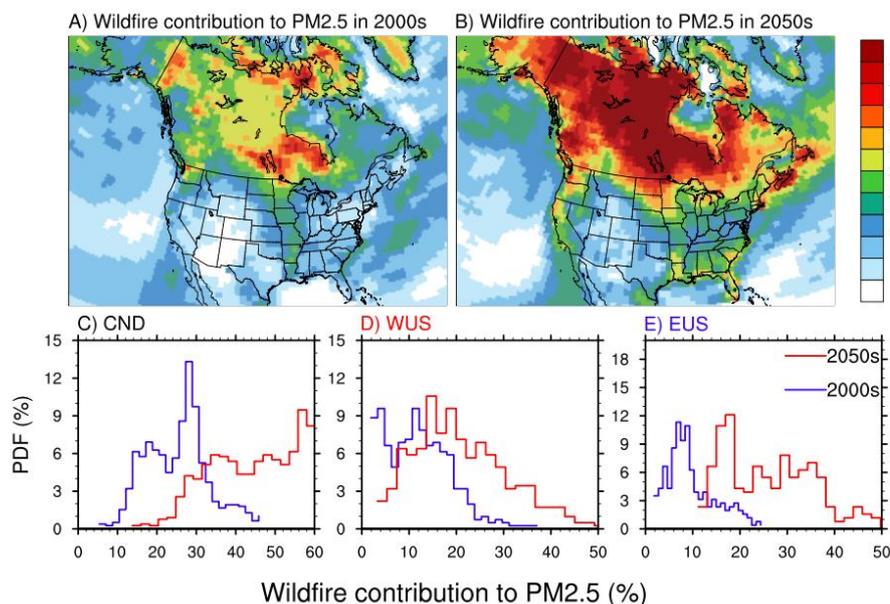
323 **Figure 3: Spatial distribution of $\text{PM}_{2.5}$ concentrations.** Spatial distribution of decadal-mean
324 wildfire-induced enhancement in summer (JJA) $\text{PM}_{2.5}$ concentration over North America for
325 present day (A, 2000_{ALL}-2000_{WEF}) and future (B, 2050_{ALL}-2050_{WEF}). C-E, Probability density
326 functions (PDFs) of wildfire contribution within the three regions shown in Figure 2B for
327 Canada (CND: black box) (C), WUS (red box) (D), and EUS (blue box) (E), respectively, for
328 2000s (blue) and 2050s (red). Only grids over land in North America are used to generate
329 the PDFs. The y-axis indicates the probability of occurrence of different $\text{PM}_{2.5}$ values shown in
330 the x-axis. The colorbar illustrates $\text{PM}_{2.5}$ in $\mu\text{g}/\text{m}^3$.

331

332 As anthropogenic- and wildfire-induced $\text{PM}_{2.5}$ concentrations may change differently
333 with time across North America, next, we investigate the relative contribution of wildfire-
334 induced $\text{PM}_{2.5}$ to the total $\text{PM}_{2.5}$ in the future. Prominent enhancement of the wildfire



335 contribution is apparent in the entire domain by the 2050s (Figures 4A-B). Largely, during
336 the 2000s, the simulated fractional contribution of wildfires to $PM_{2.5}$ is ~15-50 % in Canada
337 (Figure 4A). Specifically, a bi-modal distribution is simulated over Canada with modal values
338 around 18% and 30% (Figure 4C). Over WUS, the present day simulated percentage
339 contributions of wildfire-induced values are 5-25% (Figure 4A), with modal values between
340 10-20% (Figure 4D). Note that many areas located in the Pacific Northwest have higher
341 values of ~30-40% (Figure 4A). At the same time, the fractional contribution by wildfire-
342 induced $PM_{2.5}$ is ~5-10% in most areas of EUS in present day (Figure 4F). Nevertheless,
343 some areas in the central US also have higher values of ~10-25% (Figure 4A).



344

345 **Figure 4: Spatial distribution and probability density function of the percentage**
346 **contribution of wildfire emissions. A-B**, Spatial distribution of the percentage contribution
347 of wildfire emissions to decadal-averaged summer (JJA) mean $PM_{2.5}$ concentrations over
348 North America during present day (A) and future (B). The percentage contribution of
349 wildfire-induced $PM_{2.5}$ to the total $PM_{2.5}$ concentrations is calculated as $([2000_{ALL}-$
350 $2000_{WEF}]/2000_{ALL})$ and $([2050_{ALL}-2050_{WEF}]/2050_{ALL})$ for the present and future, respectively.
351 **C-E**, Probability density functions (PDFs) of the percentage wildfire contribution within the



352 three regions shown in Figure 2D for Canada (CND: black box) (C), WUS (red box) (D), and
353 EUS (blue box) (E), respectively, for the 2000s (blue) and the 2050s (red). Only grids over
354 land in North America are used to generate the PDFs. The y-axis indicates the probability of
355 occurrence of different $PM_{2.5}$ values shown in the x-axis.

356

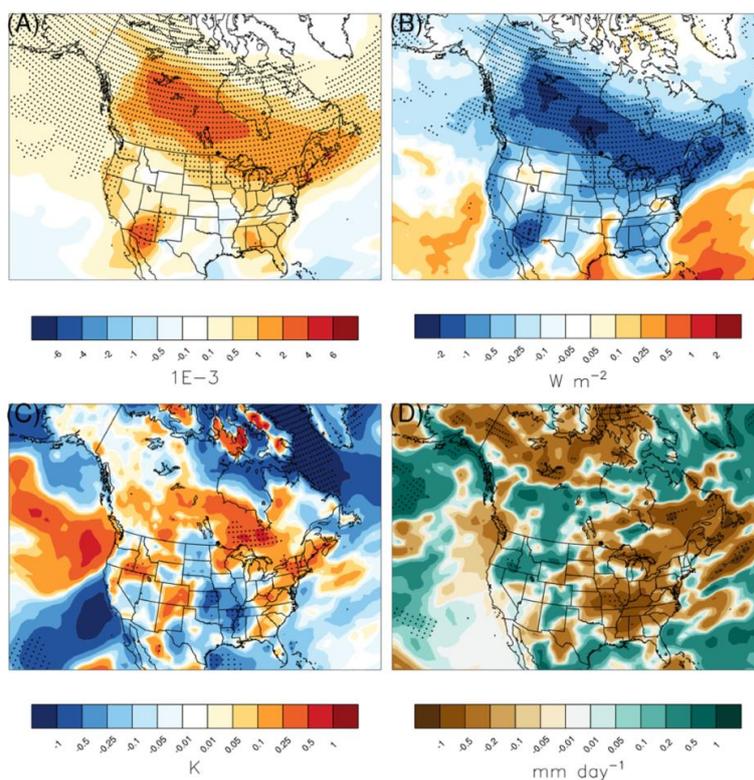
357 The wildfire contributions in the 2050s show a clear shift towards higher values in all
358 sub-regions compared to the 2000s (Figure 4B). Over Canada, the values shifted from 15-30
359 % in the 2000s to ~30-60% in the 2050s, a nearly two-fold increase in the fractional
360 contribution of wildfire emissions to the total $PM_{2.5}$ concentration is simulated (Figure 4B
361 and corresponding PDF in Figure 4C). Similarly, the contribution values increased to ~ 10-35
362 % in the 2050s, compared to 10-20% in the 2000s over WUS (Figure 4B), thereby featuring a
363 broadening of the bi-modal distribution of wildfire contribution (Figure 4D). The shift in the
364 percentage contribution is most prominent for the higher values, corresponding to some areas
365 located in the Pacific Northwest and west coast of the US (Figure 4B). Consistent with Figure
366 3B, the shift in the contribution values over EUS is also very distinct, revealing an increase in
367 the mode values from 6-10% in the 2000s to ~16-20 % by the 2050s (Figure 4B and Figure
368 4E). Thus, not only in absolute values, but our results also underscore a large increase in the
369 contribution of wildfire emissions over EUS in the future.

370 **3.3. Mechanistic understanding of the underlying processes**

371 The larger enhancement in the relative contribution of wildfire emissions to the total
372 surface $PM_{2.5}$ in EUS in the 2050s can be explained by three mechanisms. First, due to the
373 increase in Canadian and western US wildfires, downwind transport of wildfire smoke
374 plumes to EUS will be enhanced by the 2050s. This long-range transport to the atmospheric
375 column of EUS can happen within a few days of the fire occurrence (Supplementary Figures
376 2A and 2B). Using Hazard Mapping System (HMS)-detected smoke plumes, recent studies
377 identified a strong positive association between the transported smoke plumes in the



378 atmospheric column and collocated surface PM_{2.5} enhancement in EUS (Brey et al., 2018;
379 Wu et al., 2018; Gunsch et al., 2018; Kaulfus et al., 2017; Larsen et al., 2017; Dempsey,
380 2013). Hazard Mapping System (HMS) is an operational smoke detection product over North
381 America known as developed by the National Oceanic and Atmospheric Administration
382 (NOAA) and operated by National Environmental Satellite, Data, and Information Service
383 (NESDIS), available at <http://satepsanone.nesdis.noaa.gov/FIRE/fire.html>. Specifically, these
384 studies found that the smoke plumes transported from Canada are located at an altitude of ~
385 1-3 km over EUS (Colarco et al., 2004; Wu et al., 2018). Due to mixing by the daytime
386 boundary layer and deposition, the smoke plumes enhance the surface PM_{2.5} concentration
387 over EUS (Wu et al., 2018; Colarco et al., 2004; Rogers et al., 2020; Dreessen et al., 2015).
388 Hence HMS smoky days may be a useful proxy for wildfire-induced surface PM_{2.5} over
389 North America. In agreement, Brey et al. (2018) showed that the HMS-based smoke plumes
390 observed over EUS is significantly aged, suggestive of their long-range transport origin.
391 Consistent with the observed temporal change in HMS pattern, Xue et al. (2021) estimated
392 using the mid-visible Multi Angle Implementation of Atmospheric Correction (MAIAC)
393 satellite-derived Aerosol Optical Depth (AOD) that Canadian and western US fires have
394 caused an increase in the daily PM_{2.5} over Montana, North Dakota, South Dakota and
395 Minnesota by 18.3, 12.8, 10.4 and 10.1 $\mu\text{g m}^{-3}$, respectively, between August 2011 (a low
396 fire month) and August 2018 (a high fire month). In summary, the visually apparent satellite-
397 based signatures of wildfire-smoke across Canada and EUS provide a necessary, though not
398 sufficient, support for the influence of Canadian smoke plumes on EUS air quality. Although,
399 the change in burnt area over northeastern EUS is negligible compared to the western US and
400 Canadian regions, however, there are some enhancements seen over east coast of US, which
401 can also contribute to enhanced fire emissions.



402

403 Figure 5: Spatial distribution of decadal mean summer (JJA) wildfire-induced future changes
404 [(2050_{ALL}-2050_{WEF}) - (2000_{ALL}-2000_{WEF})]. A) aerosol absorption optical depth at 550 nm, B)
405 aerosol direct radiative forcing at surface, C) lower-tropospheric stability calculated as the
406 difference between the potential temperature at 900 hPa and 1000 hPa, D) summer averaged
407 precipitation rates, over North America. Areas marked with black dots indicate grids where
408 changes are significant at the 95% confidence level.

409

Secondly, distinct climatic feedbacks are also simulated which contributes to
410 enhancement of EUS pollution. Specifically, the enhancement of wildfire-induced smoke
411 aerosols increases solar absorption and scattering in the future (Figure 5A). This reduces the
412 incoming solar radiation at the surface (Figure 5B) and induces surface cooling. With
413 atmospheric warming and surface cooling, lower-tropospheric stability is enhanced by
414 wildfire aerosols in the future (Figure 5C). These changes impose a thermal capping that
415 traps more anthropogenic aerosols and particulate matter near the surface over EUS. At the
416 same time, future increase in wildfire emissions also leads to greater reduction of monthly



417 rainfall (Figure 5D) over EUS, which may subsequently strengthen the positive feedback to
418 surface PM_{2.5} over EUS by reducing wet scavenging of transported wildfire smoke to EUS.
419 Thus, wildfire-emitted aerosols induce positive feedback on the surface PM_{2.5} concentration
420 over EUS through fire-climate interactions that vary on a regional scale. However, due to
421 computational constrains, no direct quantification of the magnitude of these feedback (with
422 aerosol-radiation and aerosol-cloud interactions turned off) on PM_{2.5} is performed and would
423 be taken up in future studies.

424 Lastly, the reason of why the contribution of wildfire emissions to the total surface
425 PM_{2.5} in EUS is so substantial in the 2050s is the drastic reduction of anthropogenic
426 contribution to the surface PM_{2.5} over EUS in the future primarily due to policy-driven
427 reduction in anthropogenic emissions under the RCP4.5 scenario. Specifically, the simulated
428 ambient summer mean PM_{2.5} concentration exhibits widespread declines in the future
429 (Supplementary Figure 3), with reduction in PM_{2.5} concentration over eastern US in the range
430 of 4-15 μg/m³, which is greatest within North America. Thus, large reduction in
431 anthropogenic contribution combined with increased downwind advection of Canadian
432 smoke to EUS and the associated positive feedbacks can explain the projected dominance of
433 wildfire emissions over EUS in future.

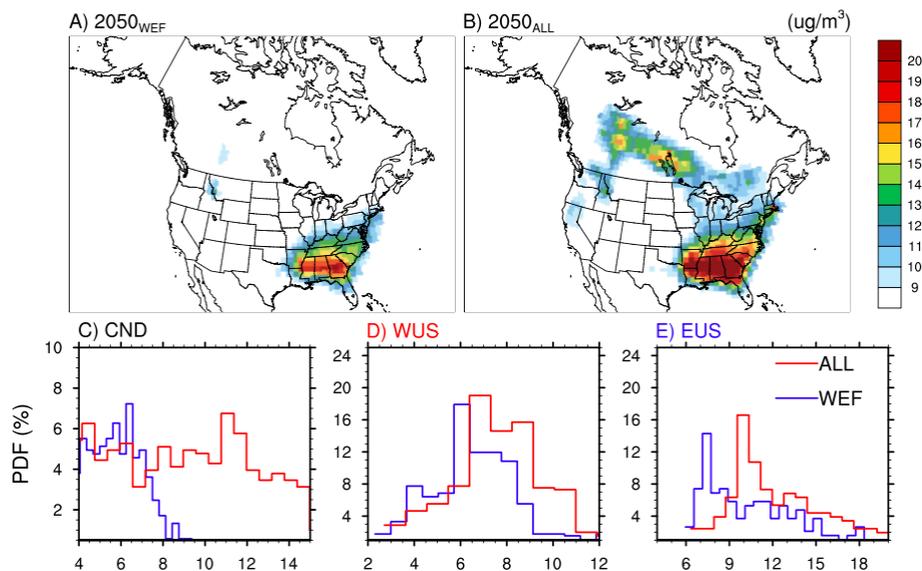
434 **3.4 Future Implications and uncertainties**

435 However, is the simulated future enhancement in wildfire contribution over EUS
436 substantial enough to affect the surface PM_{2.5} values over EUS in future? The World Health
437 Organization (WHO) air quality guidelines for annual and daily PM_{2.5} concentration are 10
438 μg m⁻³ and 25 μg m⁻³, respectively. As no specific guideline for seasonal-mean PM_{2.5} in the
439 summer is available, we use the annual guideline value as a reference to understand the
440 implication of wildfire emissions on ambient PM_{2.5} concentration in the future. Interestingly,



441 the mean summertime PM_{2.5} concentration in the wildfire emission free (WEF) scenario is
442 projected to remain within 10 µg m⁻³ over most of North America, except for the
443 southeastern US (Figure 6A). However, the ALL-scenario projects an increase in the
444 exposure concentration level such that values > 10 µg/m³ are common in Canada and EUS in
445 the future (Figure 6B). Quantitatively, over Canada, the entire PDF of PM_{2.5} concentration
446 shifts towards higher values by ~5-6 µg/m³. Specifically, the modal value shifts from ~ 6
447 µg/m³ in 2050_{WEF} to 11-12 µg/m³ in 2050_{ALL} (Figure 6C), so PM_{2.5} concentration is projected
448 to surpass the WHO guidelines over a large fraction of Canada in the future. Similarly, the
449 entire PDF of PM_{2.5} concentration shifts towards higher values by ~2-3 µg/m³ over EUS, with
450 the mode of the PDF increasing from ~ 7-8 µg/m³ in 2050_{WEF} to ~ 10-11 µg/m³ in 2050_{ALL}
451 (Figure 6E). The modal value of summer mean PM_{2.5} over WUS increases from ~ 6 µg/m³ in
452 2050_{WEF} to ~ 7-8 µg/m³ in 2050_{ALL} (Figure 6D), although a few grid cells show PM_{2.5} values
453 greater than 10 µg/m³ (Figure 6B).

454 Clearly, the climate-induced enhancement in fires and its influence via the advected
455 wildfire smoke to EUS can have significant implications for air quality management in the
456 future. The PM_{2.5} enhancement in future over the southern states within EUS is large (Figure
457 6A-B), which is consistent with Figure 3 and 4 results. However, the future change in burnt
458 area over the same region is negligible or mostly reducing (Figure 1C-D). Thus, it can be
459 argued that the simulated enhancement is mostly related with the positive climatic feedbacks
460 due to wildfire emissions i.e. enhanced lower atmospheric stability and reduced surface
461 precipitation (Figure 5). As the rate of anthropogenic emissions is also regionally highest
462 over the Southeastern states, the impact of these wildfire-induced climatic feedbacks on local
463 air quality is distinctly seen over the EUS.



464

465 **Figure 6: Spatial distribution and probability density function of PM_{2.5} concentration in**
466 **2050s. A-B**, Spatial distribution of decadal-average summer (JJA) mean PM_{2.5} concentration
467 over North America in mid-21st century from 2050_{WEF} (wildfire emission-free) (A) and
468 2050_{ALL} (wildfire emission-inclusive) (B). **C-E**, Probability density functions (PDFs) of the
469 same within the three regions shown in Figure 2B for Canada; CND (C), western US; WUS
470 (D), and eastern US; EUS (E), respectively, for the 2050_{WEF} (blue) and 2050_{ALL} (red) runs.
471 The y-axis indicates the probability of occurrence of different PM_{2.5} values shown in the x-
472 axis. Only grids over land in North America are used to generate the PDFs. Note the different
473 ranges of values shown in the y- and x-axis in C-E. The colorbar and the x-axis for Panel C-E
474 indicates PM_{2.5} values.

475 Note that our simulated present-day estimates of wildfire induced PM_{2.5} values as well
476 as the percentage contribution of wildfire emissions are within the range of reported values in
477 previous studies over the domain, which augment the fidelity our future projections.
478 Specifically, our simulated present-day estimates of wildfire induced PM_{2.5} values are also
479 within the range of reported values in previous studies over the domain. Reported values of
480 wildfire-induced PM_{2.5} over WUS during summertime vary from $\sim 1 \mu\text{g}/\text{m}^3$ (Jaffe et al., 2008)
481 to $\sim 2 \mu\text{g}/\text{m}^3$ (Park et al., 2007) and $\sim 3 \mu\text{g}/\text{m}^3$ (Ford et al., 2018), with the highest values
482 documented over the Pacific Northwest and west coast regions ($\sim 1-4 \mu\text{g}/\text{m}^3$) (O'Dell et al.,
483 2019). The wildfire-induced PM_{2.5} over EUS during summertime varies from $\sim 1 \mu\text{g}/\text{m}^3$ (Park



484 et al., 2007) to $\sim 2.5 \mu\text{g}/\text{m}^3$ ($\sim 3 \mu\text{g}/\text{m}^3$ in the southeastern US) (Ford et al., 2018).
485 Consistently, our simulated present-day estimates of wildfire contribution values are also
486 within the range of reported values in previous studies. For example, Meng et al. (2019)
487 found that wildfires can be the largest sectoral contributor (~ 18 - 59%) to the population-
488 weighted $\text{PM}_{2.5}$ in various subregions of Canada. Over WUS, the present-day percentage
489 contribution of wildfire induced $\text{PM}_{2.5}$ to the total $\text{PM}_{2.5}$ is reported to be $\sim 12\%$ (Liu et al.,
490 2017), $\sim 15\%$ (Park et al., 2007) and $\sim 30\%$ (Ford et al., 2018), with higher values of $\sim 40\%$ in
491 the Pacific Northwest (O'Dell et al., 2019). Over EUS our simulated values are also within
492 the range of previously reported values of $\sim 5\%$ (Park et al., 2007) and ~ 15 - 18% (Ford et al.,
493 2018). However, our two-way coupled simulations illustrate that future enhancement in the
494 wildfire associated $\text{PM}_{2.5}$ over the EUS could be greater compared to the western US, which
495 is not emphasized explicitly in any of the previous studies (although Ford et al., 2018
496 illustrated increase in $\text{PM}_{2.5}$ over mid and central US from Canadian fires). These could be
497 since inclusion of the wildfire-induced climatic feedbacks in our simulation is an
498 unprecedented exercise. Please also note that our study is focused on JJA period and the
499 wildfires in western US mainly occurs during August-September months, so the results
500 should be compared consciously.

501 Nonetheless, inherent limitations in our simulations may introduce uncertainties in the
502 projected future changes. For example, our reported changes in $\text{PM}_{2.5}$ concentrations based on
503 relatively coarse resolution simulations and decadal averages likely represent a low-end
504 estimate compared to changes at regional and daily/weekly scales. Moreover, our
505 experiments do not consider the direct human influences such as population change and
506 socioeconomic development on wildfires, which may aggravate the increase in $\text{PM}_{2.5}$
507 concentrations over the densely populated EUS in the future. Common sources of uncertainty
508 in modeling burnt area and fire emission and fire aerosol and smoke are also present in our



509 model. Fire smoke, in particular, is extremely hard to measure and evaluate. Lastly, inherent
510 uncertainties in the physics parameterizations used in the model, sensitivity of climate to
511 GHGs emissions, and the RCP scenarios should also be noted. Thus, ensemble modeling
512 considering different emissions scenarios, population and future time periods, and the use of
513 a finer spatial resolution may provide a more robust and better quantification of the wildfire-
514 induced impact on policy regulated improvements in PM_{2.5} over EUS.

515 **4. Conclusion**

516 In summary, online coupled fire-climate-ecosystem simulations project a nearly
517 twofold increase in wildfire-induced summer-mean surface PM_{2.5} concentration by the mid-
518 21st century over the entire North America. In a wildfire-emission free future, a large portion
519 of North America will have PM_{2.5} values below the WHO guidelines. But in a future with
520 wildfire emissions, the improvements from policy-driven reductions in anthropogenic PM_{2.5}
521 will be compromised by the projected doubling of PM_{2.5} from wildfires. More strikingly,
522 wildfire-induced enhancement in surface PM_{2.5} values and percentage contribution of the
523 wildfire emissions over EUS could be substantial by mid-century. This is mainly because of
524 the large enhancement in fires over Northern America by 2050s and associated increase in
525 amount of downwind transport of smoke to EUS. In addition, enhancement of smoke
526 transport induces a positive climate feedback to PM_{2.5} concentrations over EUS by increasing
527 the lower-tropospheric stability and reducing wet scavenging rates. Despite the inherent
528 limitations, this study highlights the natural versus anthropogenic contributions and the non-
529 local nature of air pollution that can complicate regulatory strategies aimed at improving air
530 quality over the eastern US in a warmer future.

531 **Data availability statements**



532 The HMS data used in this paper are available free through the link
533 <https://www.ospo.noaa.gov/Products/land/hms.html>. The model simulations data are
534 available at <https://portal.nersc.gov/project/m1660/yang560/wildfire>

535 **Code availability statements**

536 The model code and scripts are available at
537 <https://portal.nersc.gov/project/m1660/yang560/wildfire>

538

539 **Acknowledgements:**

540 This research was performed at PNNL and funded under Assistance Agreement No.
541 RD835871 by the U.S. Environmental Protection Agency to Yale University through the
542 SEARCH (Solutions for Energy, AiR, Climate, and Health) Center. It has not been formally
543 reviewed by EPA. The views expressed in this document are solely those of the SEARCH
544 Center and do not necessarily reflect those of the Agency. EPA does not endorse any
545 products or commercial services mentioned in this publication. CS is supported by the New
546 Faculty Initiation Grant project number CE/20-21/065/NFIG/008961 from IIT Madras.
547 PNNL is operated by Battelle Memorial Institute for the U.S. Department of Energy under
548 contract DE-AC06-76RLO-1830.

549

550 **Authors Contribution**

551 YQ, CS and RL designed this study. CS did the model and satellite analysis and wrote the
552 first draft of the manuscript. YZou performed the simulations. All authors provided inputs
553 throughout the study and helped in the drafting and submission process.

554

555 **References**

- 556 Abatzoglou, J. T. and Williams, A. P.: Impact of anthropogenic climate change on wildfire
557 across western US forests, *Proc. Natl. Acad. Sci.*, 113(42), 11770 LP – 11775,
558 doi:10.1073/pnas.1607171113, 2016.
- 559 Andela, N., Morton, D. C., Giglio, L., Chen, Y., van der Werf, G. R., Kasibhatla, P. S.,
560 DeFries, R. S., Collatz, G. J., Hantson, S., Kloster, S., Bachelet, D., Forrest, M., Lasslop, G.,
561 Li, F., Mangeon, S., Melton, J. R., Yue, C. and Randerson, J. T.: A human-driven decline in
562 global burned area, *Science* (80-.), 356(6345), 1356 LP – 1362,
563 doi:10.1126/science.aal4108, 2017.
- 564 Anjali, H., Muhammad, A., Anthony, D. M., Karen, S., R., S. M., Mick, M., M., T. A., J., A.
565 M. and Martine, D.: Impact of Fine Particulate Matter (PM_{2.5}) Exposure During Wildfires on
566 Cardiovascular Health Outcomes, *J. Am. Heart Assoc.*, 4(7), e001653,
567 doi:10.1161/JAHA.114.001653, 2019.
- 568 Black, C., Tesfaigzi, Y., Bassein, J. A. and Miller, L. A.: Wildfire smoke exposure and
569 human health: Significant gaps in research for a growing public health issue, *Environ.*
570 *Toxicol. Pharmacol.*, 55, 186–195, doi:https://doi.org/10.1016/j.etap.2017.08.022, 2017.
- 571 Brey, S. J., Ruminski, M., Atwood, S. A. and Fischer, E. V: Connecting smoke plumes to



- 572 sources using Hazard Mapping System (HMS) smoke and fire location data over North
573 America, *Atmos. Chem. Phys.*, 18(3), 1745–1761, doi:10.5194/acp-18-1745-2018, 2018.
- 574 Dempsey, F.: Forest Fire Effects on Air Quality in Ontario: Evaluation of Several Recent
575 Examples, *Bull. Am. Meteorol. Soc.*, 94(7), 1059–1064, doi:10.1175/BAMS-D-11-00202.1,
576 2013.
- 577 Dominici, F., Peng, R. D., Bell, M. L., Pham, L., McDermott, A., Zeger, S. L. and Samet, J.
578 M.: Fine Particulate Air Pollution and Hospital Admission for Cardiovascular and
579 Respiratory Diseases, *JAMA*, 295(10), 1127–1134, doi:10.1001/jama.295.10.1127, 2006.
- 580 Diao, M., Holloway, T., Choi, S., O’Neill, S.M., Al-Hamdan, M.Z., Van Donkelaar, A.,
581 Martin, R.V., Jin, X., Fiore, A.M., Henze, D.K. and Lacey, F., 2019. Methods, availability,
582 and applications of PM_{2.5} exposure estimates derived from ground measurements, satellite,
583 and atmospheric models. *Journal of the Air & Waste Management Association*, 69(12),
584 pp.1391-1414.
- 585 Ford, B., Val Martin, M., Zelasky, S. E., Fischer, E. V., Anenberg, S. C., Heald, C. L. and
586 Pierce, J. R.: Future Fire Impacts on Smoke Concentrations, Visibility, and Health in the
587 Contiguous United States, *GeoHealth*, 2(8), 229–247, doi:10.1029/2018GH000144, 2018.
- 588 Gillett, N. P., Weaver, A. J., Zwiers, F. W. and Flannigan, M. D.: Detecting the effect of
589 climate change on Canadian forest fires, *Geophys. Res. Lett.*, 31(18),
590 doi:10.1029/2004GL020876, 2004.
- 591 Gunsch, M. J., May, N. W., Wen, M., Bottenus, C. L. H., Gardner, D. J., VanReken, T. M.,
592 Bertman, S. B., Hopke, P. K., Ault, A. P. and Pratt, K. A.: Ubiquitous influence of wildfire
593 emissions and secondary organic aerosol on summertime atmospheric aerosol in the forested
594 Great Lakes region, *Atmos. Chem. Phys.*, 18(5), 3701–3715, doi:10.5194/acp-18-3701-2018,
595 2018.
- 596 Guan S, Wong DC, Gao Y, Zhang T, Pouliot G. Impact of wildfire on particulate matter in
597 the southeastern United States in November 2016. *Sci Total Environ.* 2020;724:138354.
598 doi:10.1016/j.scitotenv.2020.138354.
- 599 Johnston F. H., Henderson. S. B., Yang, C., T., R. J., Miriam, M., S., D. R., Patrick, K.,
600 M.J.S., B. D. and Michael, B.: Estimated Global Mortality Attributable to Smoke from
601 Landscape Fires, *Environ. Health Perspect.*, 120(5), 695–701, doi:10.1289/ehp.1104422,
602 2012.
- 603 Harris, R. M. B., Remenyi, T. A., Williamson, G. J., Bindoff, N. L., and Bowman, D. M. J.
604 S.: Climate-vegetation fire interactions and feedbacks: trivial detail or major barrier to
605 projecting the future of the Earth system?, *Wires Clim. Change*, 7, 910-931,
606 10.1002/wcc.428, 2016.
- 607 Hantson, S., Arneth, A., Harrison, S. P., Kelley, D. I., Prentice, I. C., Rabin, S. S., et al.
608 (2016). The status and challenge of global fire modelling. *Biogeosciences*, 13, 3359–3375.
609 <https://doi.org/10.5194/bg-13-3359-2016>
- 610 Hu, X., Yu, C., Tian, D., Ruminski, M., Robertson, K., Waller, L. A. and Liu, Y.:
611 Comparison of the Hazard Mapping System (HMS) fire product to ground-based fire records
612 in Georgia, USA, *J. Geophys. Res. Atmos.*, 121(6), 2901–2910, doi:10.1002/2015JD024448,
613 2016.



- 614 HURTT, G. C., FROLKING, S., FEARON, M. G., MOORE, B., SHEVLIAKOVA, E.,
615 MALYSHEV, S., PACALA, S. W. and HOUGHTON, R. A.: The underpinnings of land-use
616 history: three centuries of global gridded land-use transitions, wood-harvest activity, and
617 resulting secondary lands, *Glob. Chang. Biol.*, 12(7), 1208–1229, doi:10.1111/j.1365-
618 2486.2006.01150.x, 2006.
- 619 Jaffe, D., Hafner, W., Chand, D., Westerling, A. and Spracklen, D.: Interannual Variations in
620 PM_{2.5} due to Wildfires in the Western United States, *Environ. Sci. Technol.*, 42(8), 2812–
621 2818, doi:10.1021/es702755v, 2008.
- 622 Jolly, W. M., Cochrane, M. A., Freeborn, P. H., Holden, Z. A., Brown, T. J., Williamson, G.
623 J. and Bowman, D. M. J. S.: Climate-induced variations in global wildfire danger from 1979
624 to 2013, *Nat. Commun.*, 6, 7537 [online] Available from:
625 <https://doi.org/10.1038/ncomms8537>, 2015.
- 626 Kaulfus, A. S., Nair, U., Jaffe, D., Christopher, S. A. and Goodrick, S.: Biomass Burning
627 Smoke Climatology of the United States: Implications for Particulate Matter Air Quality,
628 *Environ. Sci. Technol.*, 51(20), 11731–11741, doi:10.1021/acs.est.7b03292, 2017.
- 629 Kirchmeier-Young, M. C., Zwiers, F. W., Gillett, N. P. and Cannon, A. J.: Attributing
630 extreme fire risk in Western Canada to human emissions, *Clim. Change*, 144(2), 365–379,
631 doi:10.1007/s10584-017-2030-0, 2017.
- 632 Kitzberger, T., Falk, D. A., Westerling, A. L. and Swetnam, T. W.: Direct and indirect
633 climate controls predict heterogeneous early-mid 21st century wildfire burned area across
634 western and North America, *PLoS One*, 12(12), e0188486 [online] Available from:
635 <https://doi.org/10.1371/journal.pone.0188486>, 2017.
- 636 Knorr, W., Dentener, F., Lamarque, J.-F., Jiang, L. and Arneth, A.: Wildfire air pollution
637 hazard during the 21st century, *Atmos. Chem. Phys.*, 17(14), 9223–9236, doi:10.5194/acp-
638 17-9223-2017, 2017.
- 639 Koplitz, S.N., Nolte, C.G., Pouliot, G.A., Vukovich, J.M. and Beidler, J., 2018. Influence of
640 uncertainties in burned area estimates on modeled wildland fire PM_{2.5} and ozone pollution
641 in the contiguous US. *Atmospheric environment*, 191, pp.328-339.
- 642 Lamarque, J. F., Bond, T. C., Eyring, V., Granier, C., Heil, A., Klimont, Z., Lee, D., Liousse,
643 C., Mieville, A., Owen, 628 B., Schultz, M. G., Shindell, D., Smith, S. J., Stehfest, E., Van
644 Aardenne, J., Cooper, O. R., Kainuma, M., 629 Mahowald, N., McConnell, J. R., Naik, V.,
645 Riahi, K., and van Vuuren, D. P.: Historical (1850-2000) gridded anthropogenic and biomass
646 burning emissions of reactive gases and aerosols: methodology and application, 631 *Atmos.*
647 *Chem. Phys.*, 10, 7017-7039, 10.5194/acp-10-7017-2010, 2010.
- 648 Leibensperger, E. M., Mickley, L. J., Jacob, D. J., Chen, W.-T., Seinfeld, J. H., Nenes, A.,
649 Adams, P. J., Streets, D. G., Kumar, N. and Rind, D.: Climatic effects of 1950–2050
650 changes in US anthropogenic aerosols – Part 2: Climate response, *Atmos. Chem.*
651 *Phys.*, 12(7), 3349–3362, doi:10.5194/acp-12-3349-2012, 2012.
- 652 Li, F., Zeng, X. D., & Levis, S. (2012). A process-based fire parameterization of intermediate
653 complexity in a dynamic global vegetation model. *Biogeosciences*, 9(7), 2761–2780.
654 <https://doi.org/10.5194/bg-9-2761-2012>
- 655 Liu, J. C., Mickley, L. J., Sulprizio, M. P., Dominici, F., Yue, X., Ebisu, K., Anderson, G. B.,
656 Khan, R. F. A., Bravo, M. A. and Bell, M. L.: Particulate Air Pollution from Wildfires in the



- 657 Western US under Climate Change, *Clim. Change*, 138(3), 655–666, doi:10.1007/s10584-
658 016-1762-6, 2016.
- 659 Liu, Y., Goodrick, S. and Heilman, W.: Wildland fire emissions, carbon, and climate:
660 Wildfire–climate interactions, *For. Ecol. Manage.*, 317, 80–96,
661 doi:https://doi.org/10.1016/j.foreco.2013.02.020, 2014.
- 662 Liu, X., Easter, R. C., Ghan, S. J., Zaveri, R., Rasch, P., Shi, X., Lamarque, J. F., Gettelman,
663 A., Morrison, H., Vitt, F., Conley, A., Park, S., Neale, R., Hannay, C., Ekman, A. M. L.,
664 Hess, P., Mahowald, N., Collins, W., Iacono, M. J., Bretherton, C. S., Flanner, M. G., and
665 Mitchell, D.: Toward a minimal representation of aerosols in climate models: description and
666 evaluation in the Community Atmosphere Model CAM5, *Geosci. Model Dev.*, 5, 709– 652
667 739, 10.5194/gmd-5-709-2012, 2012.
- 668 Meng, J., Martin, R.V., Li, C., van Donkelaar, A., Tzompa-Sosa, Z.A., Yue, X., Xu, J.W.,
669 Weagle, C.L. and Burnett, R.T., 2019. Source Contributions to Ambient Fine Particulate
670 Matter for Canada. *Environmental science & technology*, 53(17), pp.10269-10278.
- 671 McClure, C. D. and Jaffe, D. A.: US particulate matter air quality improves except in
672 wildfire-prone areas, *Proc. Natl. Acad. Sci.*, 115(31), 7901 LP – 7906,
673 doi:10.1073/pnas.1804353115, 2018.
- 674 Neale, R. B., Chen, C. C., Gettelman, A., Lauritzen, P. H., Park, S., Williamson, D. L.,
675 Conley, A. J., Garcia, R., Kinnison, D., Lamarque, J. F., Marsh, D., Mills, M., Smith, A. K.,
676 Tilmes, S., Vitt, F., Morrison, H., Cameron671 Smith, P., Collins, W. D., Iacono, M. J.,
677 Easter, R. C., Ghan, S. J., Liu, X. H., Rasch, P. J., and Taylor, M. A.: Description of the
678 NCAR Community Atmosphere Model (CAM 5.0), NCAR 289, 2013
- 679 Nolte, C. G., Spero, T. L., Bowden, J. H., Mallard, M. S. and Dolwick, P. D.: The potential
680 effects of climate change on air quality across the conterminous US at 2030 under three
681 Representative Concentration Pathways, *Atmos. Chem. Phys.*, 18(20), 15471–15489,
682 doi:10.5194/acp-18-15471-2018, 2018.
- 683 Katelyn O’Dell, Bonne Ford, Emily V. Fischer, and Jeffrey R. Pierce *Environmental Science*
684 & *Technology* 2019 53 (4), 1797-1804, DOI: 10.1021/acs.est.8b05430.
685
- 686 Partain, J. L., Alden, S., Strader, H., Bhatt, U. S., Bieniek, P. A., Brettschneider, B. R.,
687 Walsh, J. E., Lader, R. T., Olsson, P. Q., Rupp, T. S., Thoman, R. L., York, A. D. and Ziel,
688 R. H.: An Assessment of the Role of Anthropogenic Climate Change in the Alaska Fire
689 Season of 2015, *Bull. Am. Meteorol. Soc.*, 97(12), S14–S18, doi:10.1175/BAMS-D-16-
690 0149.1, 2016.
- 691 Park, S., Bretherton, C. S., and Rasch, P. J.: Integrating Cloud Processes in the Community
692 Atmosphere Model, Version 5, *J. Climate*, 27, 6821–6856, 10.1175/Jcli-D-14-00087.1, 2014.
- 693 Park, R.J., Jacob, D.J. and Logan, J.A., 2007. Fire and biofuel contributions to annual mean
694 aerosol mass concentrations in the United States. *Atmospheric Environment*, 41(35), pp.7389-
695 7400.
- 696 Pierce, J. R., Val Martin, M., & Heald, C. L. (2017). Estimating the Effects of Changing
697 Climate on Fires and Consequences for U.S. Air Quality, Using a Set of Global and Regional
698 Climate Models (final report no. JFSP-13-1-01-4). Retrieved from
699 https://www.firescience.gov/projects/13-1-01-4/project/13-1-01-4_final_report.pdf



- 700 Pouliot G, Pace TG, Roy B, Pierce T, Mobley D. Development of a biomass burning
701 emissions inventory by combining satellite and ground-based information. *J Appl Remote*
702 *Sens.* 2008;2:021501. doi: 10.1117/1.2939551.
- 703 Randerson, J. T., Chen, Y., van der Werf, G. R., Rogers, B. M., & Morton, D. C. (2012).
704 Global burned area and biomass burning emissions from small fires. *Journal of Geophysical*
705 *Research*, 117, G04012. <https://doi.org/10.1029/2012JG002128>
- 706 Rolph, G. D., Draxler, R. R., Stein, A. F., Taylor, A., Ruminski, M. G., Kondragunta, S.,
707 Zeng, J., Huang, H.-C., Manikin, G., McQueen, J. T. and Davidson, P. M.: Description and
708 Verification of the NOAA Smoke Forecasting System: The 2007 Fire Season, *Weather*
709 *Forecast.*, 24(2), 361–378, doi:10.1175/2008WAF2222165.1, 2009.
- 710 Spracklen, D. V., Mickley, L. J., Logan, J. A., Hudman, R. C., Yevich, R., Flannigan, M. D.,
711 & Westerling, A. L. (2009). Impacts of climate change from 2000 to 2050 on wildfire activity
712 and carbonaceous aerosol concentrations in the western United States. *Journal of Geophysical*
713 *Research*, 114, D20301. <https://doi.org/10.1029/2008JD010966>
- 714 Sun, Y., Gu, L. H., and Dickinson, R. E.: A numerical issue in calculating the coupled carbon
715 and water fluxes in a climate model, *J. Geophys. Res.-Atmos.*, 117,
716 D2210310.1029/2012jd018059, 2012
- 717 Sofiev, M., Ermakova, T., and Vankevich, R.: Evaluation of the smoke-injection height from
718 wild-land fires using remote-sensing data, *Atmos. Chem. Phys.*, 12, 1995-2006, 10.5194/acp-
719 12-1995-2012, 2012
- 720 Shi, H., Jiang, Z., Zhao, B., Li, Z., Chen, Y., Gu, Y., Jiang, J. H., Lee, M., Liou, K.-N., Neu,
721 J. L., Payne, V. H., Su, H., Wang, Y., Witek, M. and Worden, J.: Modeling Study of the Air
722 Quality Impact of Record-Breaking Southern California Wildfires in December 2017, *J.*
723 *Geophys. Res. Atmos.*, 0(0), doi:10.1029/2019JD030472, 2019.
- 724 U.S. EPA (U.S. Environmental Protection Agency). 2009. Integrated Science Assessment
725 (ISA) For Particulate Matter (Final Report). EPA/600/R-08/139F. Washington, DC:U.S. EPA.
- 726 U.S. EPA. 2018. Our Nation's Air. <https://gispub.epa.gov/air/trendsreport/2018/>
- 727 Val Martin, M., Heald, C. L., Lamarque, J.-F., Tilmes, S., Emmons, L. K. and Schichtel, B.
728 A.: How emissions, climate, and land use change will impact mid-century air quality over the
729 United States: a focus on effects at national parks, *Atmos. Chem. Phys.*, 15(5), 2805–2823,
730 doi:10.5194/acp-15-2805-2015, 2015.
- 731 Van der Werf, G. R., Randerson, J. T., Giglio, L., Collatz, G. J., Kasibhatla, P. S., and
732 Arellano, A. F.: Interannual variability in global biomass burning emissions from 1997 to
733 2004, *Atmos. Chem. Phys.*, 6, 3423-3441, DOI 737 10.5194/acp-6-3423-2006, 2006
- 734 Van Der Werf, G. R., Randerson, J. T., Giglio, L., Collatz, G. J., Mu, M., Kasibhatla, P. S.,
735 Morton, D. C., Defries, R. S., Jin, Y. and Van Leeuwen, T. T.: Global fire emissions and the
736 contribution of deforestation, savanna, forest, agricultural, and peat fires (1997-2009), *Atmos.*
737 *Chem. Phys.*, 10(23), 11707–11735, doi:10.5194/acp-10-11707-2010, 2010.
- 738 Van Donkelaar, A., R. V. Martin, M. Brauer, N. C. Hsu, R. A. Kahn, R. C. Levy, A.
739 Lyapustin, A. M. Sayer, and D. M. Winker. 2018. Global Annual PM_{2.5} Grids from MODIS,
740 MISR and SeaWiFS Aerosol Optical Depth (AOD) with GWR, 1998-2016. Palisades NY:



- 741 NASA Socioeconomic Data and Applications Center (SEDAC).
742 <https://doi.org/10.7927/H4ZK5DQS>. Accessed 16 November 2019.
- 743 Ward, D. S., Kloster, S., Mahowald, N. M., Rogers, B. M., Randerson, J. T. and Hess, P. G.:
744 The changing radiative forcing of fires: global model estimates for past, present and future,
745 *Atmos. Chem. Phys.*, 12(22), 10857–10886, doi:10.5194/acp-12-10857-2012, 2012.
- 746 Wotton, B. M., Flannigan, M. D., and Marshall, G. A.: Potential climate change impacts on
747 fire intensity and key wildfire suppression thresholds in Canada, *Environ. Res. Lett.*, 12,
748 095003, <https://doi.org/10.1088/1748-9326/aa7e6e>, 2017.
- 749 Westerling, A. L., Hidalgo, H. G., Cayan, D. R. and Swetnam, T. W.: Warming and Earlier
750 Spring Increase Western U.S. Forest Wildfire Activity, *Science* (80-.), 313(5789), 940 LP –
751 943, doi:10.1126/science.1128834, 2006.
- 752 Wotawa, G. and Trainer, M.: The Influence of Canadian Forest Fires on Pollutant
753 Concentrations in the United States, *Science* (80-.), 288(5464), 324 LP – 328,
754 doi:10.1126/science.288.5464.324, 2000.
- 755 Wu, Y., Arapi, A., Huang, J., Gross, B. and Moshary, F.: Intra-continental wildfire smoke
756 transport and impact on local air quality observed by ground-based and satellite remote
757 sensing in New York City, *Atmos. Environ.*, 187, 266–281,
758 doi:<https://doi.org/10.1016/j.atmosenv.2018.06.006>, 2018.
- 759 Xue, Z., Gupta, P., and Christopher, S.: Satellite-based estimation of the impacts of
760 summertime wildfires on PM_{2.5} concentration in the United States, *Atmos. Chem. Phys.*, 21,
761 11243–11256, <https://doi.org/10.5194/acp-21-11243-2021>, 2021.
- 762 Yang, P.-L., Y. Zhang, K. Wang, P. Doraiswamy, and S.-H. Cho, 2019, Health Impacts and
763 Cost-Benefit Analyses of Surface O₃ and PM_{2.5} over the U.S. under Future Climate and
764 Emission Scenarios, *Environmental Research*, 178, November 2019, 108687,
765 <https://doi.org/10.1016/j.envres.2019.108687>.
- 766 Yue, X., Mickley, L. J., Logan, J. A. and Kaplan, J. O.: Ensemble projections of wildfire
767 activity and carbonaceous aerosol concentrations over the western United States in the mid-
768 21st century, *Atmos. Environ.*, 77, 767–780,
769 doi:<https://doi.org/10.1016/j.atmosenv.2013.06.003>, 2013.
- 770 Zou, Y., Wang, Y., Ke, Z., Tian, H., Yang, J. and Liu, Y.: Development of a REgion-Specific
771 Ecosystem Feedback Fire (RESFire) Model in the Community Earth System Model, *J. Adv.
772 Model. Earth Syst.*, 11(2), 417–445, doi:10.1029/2018MS001368, 2019.
- 773 Zou, Y., Wang, Y., Qian, Y., Tian, H., Yang, J. and Alvarado, E.: Using CESM-RESFire to
774 understand climate–fire–ecosystem interactions and the implications for decadal climate
775 variability, *Atmos. Chem. Phys.*, 20, 995–1020, <https://doi.org/10.5194/acp-20-995-2020>,
776 2020.
- 777 Zhang, Y., P.-L. Yang, Y. Gao, R. L. Leung, and M. Bell, 2020, Health and Economic
778 Impacts of Air Pollution Induced by Climate Extremes over the Continental U.S.,
779 *Environmental International*, 143, 105921, <https://doi.org/10.1016/j.envint.2020.105921>.

780

781