1	Projected increases in wildfires may challenge regulatory curtailment of PM2.5 over the
2	eastern US by 2050
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Abstract

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Anthropogenic contribution to the overall fine particulate matter (PM_{2.5}) concentrations has 22 been declining sharply in North America. In contrast, a steep rise in wildfire-induced air 23 24 pollution events with recent warming is evident in the region. Here, based on coupled fireclimate-ecosystem model simulations, summertime wildfire-induced PM_{2.5} concentrations are 25 projected to nearly double in North America by the mid-21st century compared to the 26 27 present. More strikingly, the projected enhancement in fire-induced PM_{2.5} ($\sim 1-2 \,\mu \text{g/m}^3$) and its contribution (~15-20%) to the total PM_{2.5} are distinctively significant in the eastern US. 28 This can be attributed to downwind transport of smoke from future enhancement of wildfires 29 in North America to the eastern US and associated positive climatic feedback on PM_{2.5} i.e. 30 perturbations in circulation, atmospheric stability and precipitation. Therefore, the anticipated 31 reductions in PM_{2.5} from regulatory controls on anthropogenic emissions could be 32 significantly compromised in the future in the densely populated eastern US. 33

34 **Key points:**

- 1) Wildfire-PM_{2.5} associations studied based on unprecedented two-way coupled fireclimate-ecosystem model simulations
 - 2) A steep rise in wildfire-induced air pollution events with recent warming is evident in the region
 - 3) The transported smoke from enhanced wildfires in North America can severely affect air quality over Eastern US
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- 42 **Keywords:** wildfire emissions, climate change, air quality, smoke transport, wildfire-climate-
- 43 ecosystem interactions

1. Introduction

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Wildfires are widespread burning events in forests, shrub lands, and grazing lands. In North America (mainly Canada and the US), particulate matter emissions from wildfires are a significant source of regional air pollution (Shi et al., 2019; McClure and Jaffe, 2018; Van Der Werf et al., 2010; Jaffe et al., 2008). Since the 1980s, the number of large wildfires and the length of wildfire season have been increasing, and the trends are projected to continue in the future over the western US, Alaska and Canada (Kitzberger et al., 2017; Kirchmeier-Young et al., 2017; Abatzoglou and Williams, 2016; Partain et al., 2016; Jolly et al., 2015; Westerling et al., 2006; Gillett et al., 2004). Accordingly, particulate emissions from wildfires are also anticipated to increase in North America in the 21st century (Knorr et al., 2017; Liu et al., 2016; Val Martin et al., 2015). Human exposure to high concentrations of wildfireemitted airborne particulate matter of diameter ≤2.5 µm (PM_{2.5}) is known to have substantial adverse effects on pulmonary and cardiovascular functioning (Anjali et al., 2019; Black et al., 2017), which contribute significantly to global and regional all-cause mortality (Zhang et al., 2020; Hong et al., 2019; Yang et al., 2019; Ford et al., 2018; Johnston. et al., 2012). Therefore, a better understanding of the future changes in wildfire-induced PM_{2.5} and its contribution to the total surface PM_{2.5} is essential.

In the last two decades, ambient air quality in the US has substantially improved due to a decline in $PM_{2.5}$ by ~ 40 % (US EPA, 2018). The decrease in $PM_{2.5}$ is primarily due to curtailment of anthropogenic emissions resulting from US-based efforts to meet regulations such as the Clean Air Act (US EPA, 2009), Cross-State Air Pollution Rule, Regional Haze Rule, and the motor vehicles emissions standards. Consequently, air quality over the contiguous US (CONUS) and Canada has improved steadily such that it is predicted to achieve the targeted National Ambient Air Quality Standards in the future (Nolte et al.,

2018). Under this promising scenario, the influence of wildfire-emissions on the total PM_{2.5} becomes even more crucial. Depending on the competition between climate-induced increase in wildfires and the regulatory control on anthropogenic emissions, future enhancement in wildfire-induced PM_{2.5} may compromise the reduction in anthropogenic PM_{2.5} concentrations in certain regions. In agreement, recent studies have highlighted the potential for future enhancement in wildfire-induced pollution to diminish the reducing trend in PM_{2.5}, primarily over the western US (O'Dell et al., 2019; Ford et al., 2018; Val Martin et al., 2015; Yue et al., 2013).

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While the fractional wildfire burnt area and fire intensities are the greatest over the western US and Canadian regions within North America, anthropogenic emissions dominate the ambient PM_{2.5} concentration over the eastern US. The inherent geographical separation between the regions with large wildfire emissions and anthropogenic emissions leads to a pertinent question: will future enhancement in wildfires over the western US and Canada have significant effects on PM_{2.5} over the eastern US? Addressing this question is crucial because the declining trend in PM_{2.5} over the eastern US is the major contributor to the observed 40% decrease in PM_{2.5} over the US in the last two decades (US EPA, 2018). Eastward advection of wildfire smoke from Canada and the western US has been found to severely hamper the surface air quality of the central and eastern US under the influence of the prevailing westerlies during the summer months (Brey et al., 2018; Wu et al., 2018; Gunsch et al., 2018; Kaulfus et al., 2017; Dempsey, 2013). The transported wildfire smoke can influence the meteorology and climate via the radiative impact of carbonaceous emissions, changes in land albedo and cloud system perturbations (Ward et al., 2012; Liu et al., 2014). These fire-weather interactions can have positive feedback on the locally-emitted PM_{2.5} in the eastern US by surface cooling and boundary layer suppression(Guan et al., 2020). At the same time, fire-triggered ecosystem changes can induce negative feedback on

PM2.5 by reducing the future wildfires over North America (Zou et al., 2020). Thus, two-way interactions between fires and climate that are important for predicting future changes in wildfire locations, intensities, and durations (Harris et al., 2016) as well as associated particulate emissions is essential. However, past studies have mostly employed simple statistical models based on statistical regressions of present-day fire burnt area on the meteorological fields (Liu et al., 2016; Spracklen et al., 2009; Yue et al., 2013; Val Martin et al., 2015), and more recently, one-way coupled modelling (Ford et al., 2018; O'Dell et al., 2019).

Here, based on new two-way coupled fire-climate-ecosystem simulations, we demonstrate the significance of wildfire-induced contributions to ambient PM_{2.5} over the eastern US due to enhanced wildfire smoke transportation and smoke-induced changes in weather in eastern US. This enhancement in wildfire-induced PM_{2.5} may potentially challenge the targeted policy-driven reduction of PM_{2.5} in the eastern US. Next, our model setup, experiments and methodology are explained in Section 2, followed by results and discussion in Section 3. The study is summarized in Section 4.

2. Materials and Methods

2.1. RESFire-CESM Model description

We employ the open-source REgion-Specific ecosystem feedback fire (RESFire) model coupled with the Community Land Model version 4.5 and the Community Atmosphere Model version 5 (CAM5) of the Community Earth System Model (CESM) version 1 (Zou et al., 2019; Neale et al., 2013) to perform two-way coupled simulations. RESFire provides state-of-the-art capabilities to simulate the complex fire-climate-ecosystem interactions globally for fires occurring over wildland, cropland, and peatland. Although wildfires dominate in the North American region, RESFire simulates both wildfires and

prescribed fires. Moreover, this integrated setup includes climatic feedback from fire-induced aerosol direct and indirect radiative effects and associated weather changes. It also includes feedback from fire-induced vegetation distribution changes and associated biophysical processes such as evapotranspiration and surface albedo. Sofiev et al. (2012) described the fire plume rise parameterization. Other features in CLM4.5 and CAM5, such as the photosynthesis scheme (Sun et al., 2012), the MAM3 aerosol module (Liu et al., 2012), and the cloud macrophysics scheme (Park et al., 2014), allow for more comprehensive assessments of the climate effects of fires through their interactions with vegetation and clouds. Fire-ecosystem interactions are modelled by simulating fire-induced vegetation mortality and regrowth (and associated land cover change) in RESFire. This approach has been introduced in Zou et al. (2019) and the simulated ecological and climatic effects of wildfires have been evaluated in two sets of sensitivity experiments in Zou et al. (2020). Although fire-climate-ecosystem interactions are considered in this study, our focus is on the fire-induced changes in PM_{2.5} over Canada and the US, so the two vegetation-focused sensitivity experiments reported in Zou et al. (2020) are not included in this paper. Please refer to Zou et al. (2019) and Zou et al. (2020) for more details about the simulation of fireecosystem interactions.

2.2 Numerical Experiment and Methodology

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We designed two sets of simulations for the present day and future scenarios to quantify the impacts of fire-climate-ecosystem interactions (Table 1). The spatial resolution is 0.9° (lat) \times 1.25° (lon) with a time step of 30 min. In each set of simulations, we conducted a default all emission included control run (X_{ALL}, where x=2000 or 2050 indicates the present day or future, respectively) and a sensitivity run with no wildfire emissions to the atmosphere (X_{WEF}, where X is the same as for the control runs). The ALL runs are designed to simulate fully interactive fire disturbances such as fire emissions with plume rise and fire induced land

cover changes of the present day (representative of the 2000s, 2000_{ALL}) and a moderate future emission scenario (representative of the 2050s, 2050_{ALL}) via the RCP4.5. The only difference between the ALL and WEF scenario is that wildfire emissions are absent in the WEF scenario. Specifically, in the WEF runs, the online simulated fire emissions are not passed to the CAM5 atmosphere model so that the difference between the ALL and WEF runs can be used to isolate the atmospheric impacts of fire-climate interactions.

Table 1: Summary of the sensitivity simulations performed

Scenario	Present-day		Future	
Experiment Name	2000 _{ALL}	2000 _{WEF}	2050 _{ALL}	2050 _{WEF}
Simulation years	2001-2010	2001-2010	2051-2060	2051-2060
Atmosphere	CAM5	CAM5	CAM5	CAM5
Land	CLM4.5	CLM4.5	CLM4.5	CLM4.5
Ocean	Climatology	Climatology	RCP4.5	RCP4.5
Sea ice	Climatology	Climatology	RCP4.5	RCP4.5
Non-fire emissions	ACCMIP	ACCMIP	RCP4.5	RCP4.5
Fire emissions	Online fire aerosols with plume rise	-	Online fire aerosols with plume rise	1
Land cover	Fire disturbances on present-day condition	Fire disturbances on present-day condition	Fire disturbances on RCP4.5 condition	Fire disturbances on RCP4.5 condition

For the present-day experiments, we used the spun-up states from Zou et al. (2019) as initial conditions for both meteorological and chemical variables. Sea surface temperature (SST) for the present day was obtained from the Met Office Hadley Centre (HadISST). Present-day non-fire emissions from anthropogenic and other sources were based on ACCMIP (Lamarque et al., 2010) for the year 2000. We replaced the prescribed GFED2 fire emissions (van der Werf et al., 2006) in the default setting of CESM with the online-coupled fire emissions generated by the RESFire model. Zou et al. (2019) provided more details of the physics parameterizations and modeling experiment settings used in these simulations.

Land use and land cover data for 2000 and 2050 from the Land-Use History A product (Hurtt et al., 2006) are used to initialize the 2000_{ALL}/2000_{WEF} and 2050_{ALL}/2050_{WEF} simulations, respectively. Following the above setup, the future scenario 2050_{ALL} experiment accounts for both fuel load changes associated with the projected land use and land cover change (LULCC) in the 2050s and fire weather changes driven by the SST and sea ice forcing from a coupled CESM simulation following the greenhouse gas (GHG) forcing of the RCP4.5 scenario. The global mean GHG mixing ratios in the CAM5 atmosphere model were fixed at the year 2000 levels in all the present-day experiments and they were replaced by those of the RCP4.5 scenario with the well-mixed assumption and monthly variations. However, the future population and socioeconomic conditions were identical to those of the present day so there was no explicit impact of human-induced mitigation/enhancement effects on wildfires in the future projection in all the future experiments. Future human impacts were considered implicitly in LULCC-induced fuel load changes in the RCP4.5 scenario.

The net projected changes by 2050s in emissions, meteorology and air quality during summer (JJA: June, July, August) months are estimated by comparing decadal-mean values simulated by 2000_{ALL} with 2050_{ALL}. Wildfire-induced enhancement in PM_{2.5} concentration in the present day and mid-21st century is estimated by comparing 2000_{ALL} with 2000_{WEF} and 2050_{ALL} with 2050_{WEF}, respectively. Further, the projected increase in wildfire-induced PM_{2.5} in the future is calculated by comparing the simulated wildfire effect of the 2050s (2050_{ALL}-2050_{WEF}) with that of the 2000s (2000_{ALL}-2000_{WEF}). With large spatiotemporal variability, the projected changes in transported fire-emissions from the western US and Canada to the eastern US by the 2050s and the corresponding impacts are summarized using probability distribution functions. The latter provide information not only for the mean but also variability and extreme values to quantify the simulated changes for the three subregions.

3. Results and Discussion

3.1 Model Evaluation

Zou et al. (2019) performed comprehensive evaluation of the RESFire simulated wildfire burnt area distribution, associated carbon emissions and terrestrial carbon balance to demonstrate reasonable model skill. Zou et al. (2020) compares global fire simulations by CESM-RESFire with modeling results reported in the literature to show better agreement with the GFED4.1s benchmark data and predicts more prominent changes in the future than those predicted by Kloster et al. (2010, 2012). These differences might come from differences in the climate sensitivities of the fire models and scenarios and other input data used to make future projections.

Here, we evaluate the simulated surface PM_{2.5} against satellite-estimates (Figure 1) over North America. The PM_{2.5} concentration is calculated as the sum of sulfate, nitrate, fine sea salt (first 2 size bins), fine dust (first size bin), black carbon (BC), and organic aerosol (OC) at the surface-level of model. OC is the sum of primary organic matter (POM) and secondary organic aerosol (SOA), and SOA is the sum of secondary species formed from toluene, monoterpenes, isoprene, benzene, and xylene. Figure 1 compares the observed and simulated mean annual PM_{2.5} averaged over 2001-2010. The 10-year average satellite AOD-derived annual mean surface PM_{2.5} concentrations (Van Donkelaar et al., 2018) are regridded to the model grid (Figure 1A) and then compared with the RESFire simulations in the 2000_{ALL} present-day run (Figure 1B). The spatial distribution of annual surface PM_{2.5} is reasonably well simulated but also have some biases. To quantify the biases, we also estimated the correlation coefficient as well as normalized mean biases (NMB) of the simulated values compared against the satellite retrieved values over two subregions.

Quantitatively, the NMB values over the western US (WUS) and eastern US (EUS) are 18% and 7%, respectively (Figure 1C-D). In addition, the spatial variability of the 2001-2010

averaged annual AOD distribution (Supplementary Figure 1) is also well represented in our simulation, although the model underestimates high AOD values. Similar spatial variability and biases in AOD and PM_{2.5} were also found when a comparison was performed for only summer months (June through August; JJA). The simulated PM_{2.5} has also been evaluated against the ground-based Interagency Monitoring of Protected Visual Environments (IMPROVE) data, showing similar spatial pattern and biases (10-25%) (Supplementary Figure 2). The biases are smaller over Eastern US and Southwestern US region. The simulated PM_{2.5} values over California matches quite well with the observed annual mean values. However, the biases over Northwestern US region are ~30-40%, a portion of which could be attributed to possible biases in model's meteorology in northwestern US region. Nonetheless, both satellite and in situ evaluation indicate that our simulation biases are largely within the uncertainty range among the various satellite and ground-based datasets, which have normalized mean biases ranging from -3.3% to 33.3% when benchmarked against the ground-based IMPROVE data over the contiguous US (Diao et al., 2019; Val Martin et al. (2015)).

Discrepancies between the simulated and observed PM_{2.5} values may be attributed to several potential reasons. First, the satellite-derived data has a non-zero lower bound of PM_{2.5} concentrations, so the ambient background concentrations for relatively cleaner regions such as the western US may be overestimated (Figure 1C), also the sampling frequency between these datasets are different. Second, year 2000-based constant non-fire emissions were used in the RESFire simulation, which may result in overestimation of the PM_{2.5} concentrations from non-fire sources during 2001-2010 when anthropogenic emissions and PM_{2.5} concentrations continue to decrease (US EPA, 2018). This overestimation is prominent in regions dominated by non-fire sources such as the southeastern US. Third, large uncertainties in fuel consumption and emission factors preclude an accurate estimation of the primary fire

emissions in the model, especially for the eastern US where large fractions of low-intensity prescribed fires consume only under-canopy fuels such as litter and duff layers. The fire model may fail to capture the subtle distinctions between low-intensity prescribed fires and forest fires, so more fuels are consumed and result in higher emissions. Lastly, comparison of a coarsely resolved simulation against in-situ observations also contributes to uncertainty. Differences in the degree to which fire-climate interactions and other physical processes and feedbacks are represented by the models can explain the slights differences in estimating the present day mean wildfire-induced change in PM2.5 over local and downwind regions between our simulations and previous studies. Nonetheless, reasonable simulation of the spatial distribution of wildfire burnt area, AOD, and near surface particulate concentration (mean bias of ~10-20 %) instills confidence about the fidelity of our model setup in particulate pollution simulation, which is the focus of this study.

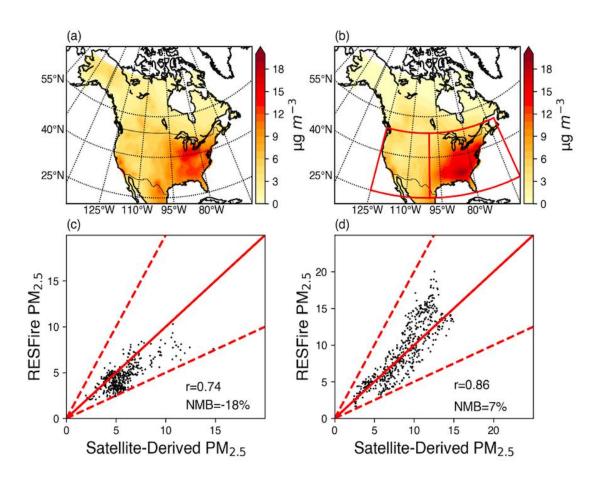


Figure 1: Comparison of the 10-year (2001-2010) averaged annual mean surface PM_{2.5} concentration between observations and RESFire simulations. (a) Satellite-derived surface PM_{2.5} concentrations (with dust and sea-salt removed) estimated by Donkelaar et al., 2018 (available at https://sedac.ciesin.columbia.edu/data/set/sdei-global-annual-gwr-pm2-5-modis-misr-seawifs-aod; last access: 5 November, 2021); (b) 2000_{ALL} Simulated surface PM_{2.5} concentrations (with dust and sea-salt removed) averaged over 2001-2010; The red boxes denote the two subregions (EUS and WUS) shown in Fig. 2 in the main text. (c) comparison of simulated and satellite based gridded surface PM_{2.5} concentrations in the WUS subregion; Number of samples is equal to the number of land grids ~450 (d) same as (c) but in the EUS subregion. Number of samples is equal to the number of land grids ~375 The red solid and dashed lines denote the 1:1 ratio line and ±100% biases, respectively. The correlation coefficients and NMB values are shown at the lower-right corner of each subplot.

3.2 Fire-induced changes in burnt area and PM_{2.5}

The decadal-mean annual fire burnt area simulated for the present day shows widespread wildfires over the entire North America (Figure 2A). Specifically, Canada and the forested areas of the northwestern (> 36 N latitude) and southeastern (< 36 N latitude) US are most intensely affected by wildfires in the present day. By the mid-21st century, a striking increase of 2-5 times in fire burnt area is projected over Canada, Alaska, the Pacific Northwest and portions of the western US by the 2050s (Figure 2B). A distinct positive shift in the probability density function (PDF) of annual fire burnt area is evident in the future, with the decadal-mean difference statistically significant at the 99% confidence level (Zou et al., 2020). A small and statistically insignificant change in interannual variability (~ 0.4 Mha yr⁻¹) of fire burnt areas is also simulated between the present and future. Specifically, our model predicts more than a doubling of burnt area in boreal regions of Canada in the future, in line with a previous projection for Canada (Wotton et al., 2017). Future enhancement in fire burnt area is ~ 20-50% in most fire grids over the western coast of US, which is higher than that over the eastern US (Figures 2A and 2C). The increase over the western US is closer to the lower bound of that derived from statistical model ensemble projections for the western US in the mid-21st century (Yue et al., 2013). The statistics-based projections of future burnt

area over North America were likely too high because fire-induced land cover change, fuel load reduction and factors could induce a negative fire feedback, which was not considered in previous fire projection studies (Zou et al. 2020).

Annual fire burnt area in the southeastern US shows a decline in the future (Figure 2C), as precipitation is projected to increase in that region (discussed later). Note that all future fire changes between 2050_{ALL} and 2000_{ALL} are primarily associated with climate warming in response to the increase in greenhouse gas (GHG) concentrations in the RCP4.5 scenario. No direct impacts of population and socioeconomic changes on wildfires are included in our simulations, although these factors contribute to changes in GHG emissions (via the RCP scenario) that influence the climate simulated in 2000_{ALL} and 2050_{ALL}. As about 80% of the projected fire changes in the future is restricted to the summer season (June through August; JJA) (Figure 2D), we focus on analysis of the summer-mean wildfire-induced PM_{2.5} and its projected future changes over North America.

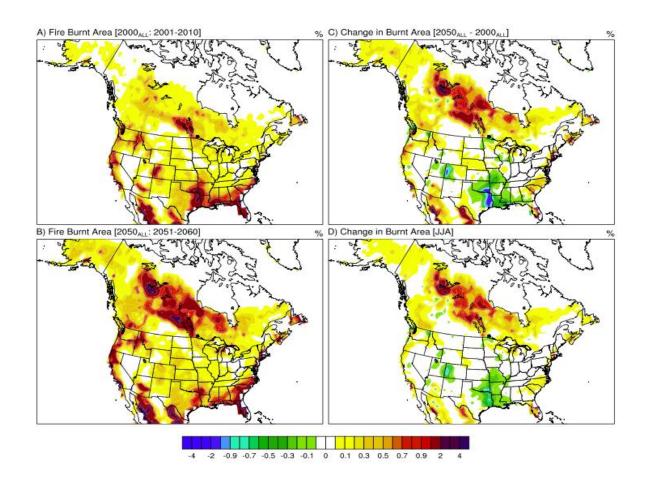


Figure 2: **Spatial distribution of fire burnt area. A-D**, Spatial distribution of simulated decadal-mean annual burnt area (as percentage) over North America for present day (A), mid-21st century (B) and the net change between the 2050s and the 2000s (C). **D**, same as (C), but for wildfire burnt area during summer only (June through August; JJA). The colorbar illustrate grid fraction of area burnt.

The simulated 10-year averaged summer-mean wildfire-induced PM_{2.5} values in 2000_{ALL} are more than $0.5~\mu g/m^3$ over a large part of North America in the present day, with noticeably larger values (> 1 $\mu g/m^3$) in Canada and the northwestern, central, and southeastern US (Figure 3A). Interestingly, the spatial distribution of wildfire-induced PM_{2.5} > 1 $\mu g/m^3$ resembles an inverted horse-shoe shape. The inverted horse-shoe shaped spatial distribution is also consistent with the wildfire-smoke climatology derived from the satellite-guided operational smoke product of the Hazard Mapping System (HMS) during 2005-2015 (Brey et al., 2018; Kaulfus et al., 2017). By the mid-21st century, the spatial extent of the horse-shoe shape for areas with wildfire-induced PM_{2.5} enhancement > 1 $\mu g/m^3$

expands significantly to span most regions of North America, with the most pronounced enhancement occurring over Canada (Figure 3B). The PDFs of the spatial distribution for the three regions can be seen in Figure 3C-E. Specifically, wildfire induced PM_{2.5} in the 2000s over Canada, WUS and EUS during summer is ~ 1 -3 μ g/m³, 1-3 μ g/m³ and 0.6-1.2 μ g/m³, respectively. Maximum values within the WUS region are found over the Pacific Northwest, with most areas having wildfire induced PM_{2.5} values of ~ 2 -3 μ g/m³. Similarly, the southern states have relatively high wildfire induced PM_{2.5} concentrations of ~ 2 -4 μ g/m³ within the EUS in the present-day simulation.

Compared to the 2000s, the wildfire induced JJA-averaged PM_{2.5} values are almost doubled to $\sim 3\text{-}6~\mu\text{g/m}^3$ over Canada in the 2050s (Figure 3B and Figure 3C). Consistently, the values of wildfire induced PM_{2.5} over WUS (mainly coastal) also doubled in 2050s compared to 2000s, with modal values of $\sim 2\text{-}2.5~\mu\text{g/m}^3$ (Figure 3D). Most interestingly, the enhancement in wildfire-induced summer-mean PM_{2.5} over the northern EUS is also significant by the 2050s (Figures 3B). Largely, the summer-mean wildfire-induced PM_{2.5} concentration over EUS increases from ~ 0.8 to $\sim 2~\mu\text{g/m}^3$ in the mid-century to values of 1.2-3.0 $\mu\text{g/m}^3$ (Figure 3E). The summer-mean wildfire-induced PM_{2.5} is thus projected to double in North America by the 2050s compared to the 2000s, with a substantial coverage over the EUS. An important finding from these PDFs appears to be that there are fewer grids with $< 1~\mu\text{g/m}^3$ wildfire induced PM_{2.5}, or alternatively, that more regions are being influenced by PM_{2.5}, and many areas that were already seeing wildfire impacts are seeing enhanced impacts. Such enhancement is found not only at the surface but also in an elevated atmospheric layer over EUS between 900 and 700 hPa. This is nonintuitive given the fact that the increase in fire-burnt area by mid-century over the EUS is not substantial.

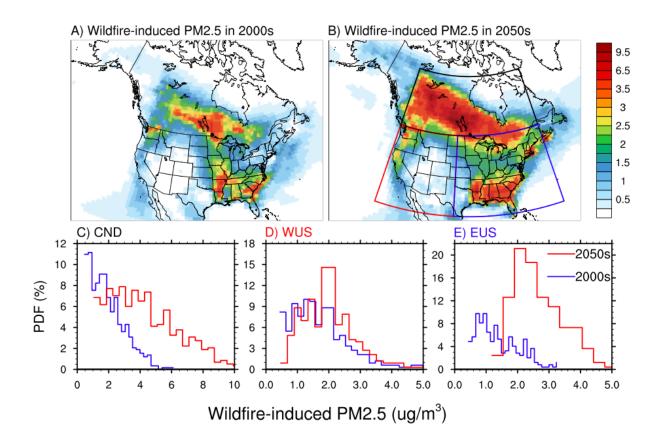


Figure 3: **Spatial distribution of** $PM_{2.5}$ **concentrations.** Spatial distribution of decadal-mean wildfire-induced enhancement in summer (June through August; JJA) $PM_{2.5}$ concentration over North America for present day (A, 2000_{ALL} - 2000_{WEF}) and future (B, 2050_{ALL} - 2050_{WEF}). **C-E**, Probability density functions (PDFs) of wildfire contribution within the three regions shown in Figure 2B for Canada (CND: black box) (C), WUS (red box) (D), and EUS (blue box) (E), respectively, for 2000s (blue) and 2050s (red). Only grids over land in North America are used to generate the PDFs. The y-axis indicates the probability of occurrence of different $PM_{2.5}$ values shown in the x-axis. The colorbar illustrates $PM_{2.5}$ in ug/m³.

As anthropogenic- and wildfire-induced PM_{2.5} concentrations may change differently with time across North America, next, we investigate the relative contribution of wildfire-induced PM_{2.5} to the total PM_{2.5} in the future. Prominent enhancement of the wildfire contribution is apparent in the entire domain by the 2050s (Figures 4A-B). Largely, during the 2000s, the simulated fractional contribution of wildfires to PM_{2.5} is ~15-50 % in Canada (Figure 4A). Specifically, a bi-modal distribution is simulated over Canada with modal values around 18% and 30% (Figure 4C). Over WUS, the present day simulated percentage

contributions of wildfire-induced values are 5-25% (Figure 4A), with modal values between 10-20% (Figure 4D). Note that many areas located in the Pacific Northwest have higher values of ~30-40% (Figure 4A). At the same time, the fractional contribution by wildfire-induced PM_{2.5} is ~5-10% in most areas of EUS in present day (Figure 4F). Nevertheless, some areas in the central US also have higher values of ~10-25% (Figure 4A).

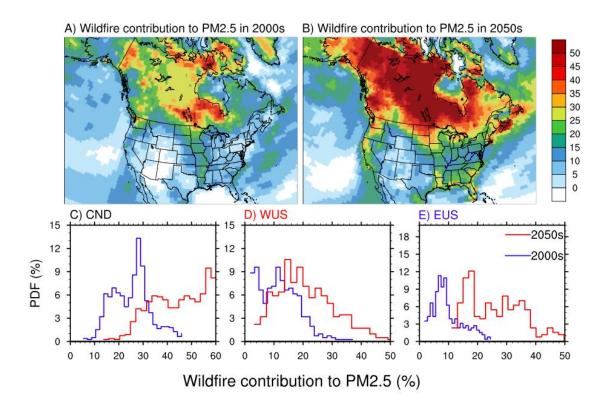


Figure 4: **Spatial distribution and probability density function of the percentage contribution of wildfire emissions. A-B**, Spatial distribution of the percentage contribution of wildfire emissions to decadal-averaged summer (June through August; JJA) mean PM_{2.5} concentrations over North America during present day (A) and future (B). The percentage contribution of wildfire-induced PM_{2.5} to the total PM_{2.5} concentrations is calculated as ([2000_{ALL}-2000_{WEF}]/2000_{ALL}) and ([2050_{ALL}-2050_{WEF}]/2050_{ALL}) for the present and future, respectively. **C-E**, Probability density functions (PDFs) of the percentage wildfire contribution within the three regions shown in Figure 2D for Canada (CND: black box) (C), WUS (red box) (D), and EUS (blue box) (E), respectively, for the 2000s (blue) and the 2050s (red). Only grids over land in North America are used to generate the PDFs. The y-axis indicates the probability of occurrence of different PM_{2.5} values shown in the x-axis.

The wildfire contributions in the 2050s show a clear shift towards higher values in all sub-regions compared to the 2000s (Figure 4B). Over Canada, the values shifted from 15-30 % in the 2000s to ~30-60% in the 2050s, a nearly two-fold increase in the fractional contribution of wildfire emissions to the total PM_{2.5} concentration is simulated (Figure 4B and corresponding PDF in Figure 4C). Similarly, the contribution values increased to ~ 10-35 % in the 2050s, compared to 10-20% in the 2000s over WUS (Figure 4B), thereby featuring a broadening of the bi-modal distribution of wildfire contribution (Figure 4D). The shift in the percentage contribution is most prominent for the higher values, corresponding to some areas located in the Pacific Northwest and west coast of the US (Figure 4B). Consistent with Figure 3B, the shift in the contribution values over EUS is also very distinct, revealing an increase in the mode values from 6-10% in the 2000s to ~16-20 % by the 2050s (Figure 4B and Figure 4E). Thus, not only in absolute values, but our results also underscore a large increase in the contribution of wildfire emissions over EUS in the future.

3.3. Mechanistic understanding of the underlying processes

The larger enhancement in the relative contribution of wildfire emissions to the total surface PM_{2.5} in EUS in the 2050s can be explained by three mechanisms. First, due to the increase in Canadian and western US wildfires, downwind transport of wildfire smoke plumes to EUS will be enhanced by the 2050s. This long-range transport to the atmospheric column of EUS can happen within a few days of the fire occurrence (Supplementary Figures 3A and 3B). Using Hazard Mapping System (HMS)-detected smoke plumes, recent studies identified a strong positive association between the transported smoke plumes in the atmospheric column and collocated surface PM_{2.5} enhancement in EUS (Brey et al., 2018; Wu et al., 2018; Gunsch et al., 2018; Kaulfus et al., 2017; Larsen et al., 2017; Dempsey, 2013). Hazard Mapping System (HMS) is an operational smoke detection product over North

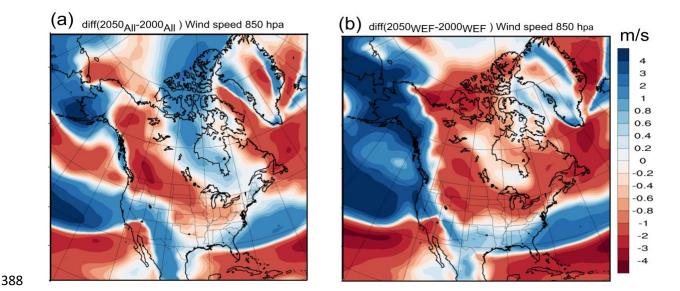


Figure 5a: Spatial distribution of decadal mean summer (June through August; JJA) wildfire-induced future changes $[(2050_{ALL}-2050_{WEF}) - (2000_{ALL}-2000_{WEF})]$. A) Wind speed at 850 hpa for $[(2050_{ALL}-2000_{WEF})]$.

America known as developed by the National Oceanic and Atmospheric Administration (NOAA) and operated by National Environmental Satellite, Data, and Information Service (NESDIS), available at http://satepsanone.nesdis.noaa.gov/FIRE/fire.html. Specifically, these studies found that the smoke plumes transported from Canada are located at an altitude of ~ 1-3 km over EUS (Colarco et al., 2004; Wu et al., 2018). Due to mixing by the daytime boundary layer and deposition, the smoke plumes enhance the surface PM_{2.5} concentration over EUS (Wu et al., 2018; Colarco et al., 2004; Rogers et al., 2020; Dreessen et al., 2015). Hence HMS smoky days may be a useful proxy for wildfire-induced surface PM2.5 over North America. In agreement, Brey et al. (2018) showed that the HMS-based smoke plumes observed over EUS is significantly aged, suggestive of their long-range transport origin. Consistent with the observed temporal change in HMS pattern, Xue et al. (2021) estimated using the mid-visible Multi Angle Implementation of Atmospheric Correction (MAIAC) satellite-derived Aerosol Optical Depth (AOD) that Canadian and western US fires have caused an increase in the daily PM_{2.5} over Montana, North Dakota, South Dakota and Minnesota by 18.3, 12.8, 10.4 and 10.1 μg m⁻³, respectively, between August 2011 (a low

fire month) and August 2018 (a high fire month). In summary, the visually apparent satellite-based signatures of wildfire-smoke across Canada and EUS provide a necessary, though not sufficient, support for the influence of Canadian smoke plumes on EUS air quality. Although, the change in burnt area over northeastern EUS is negligible compared to the western US and Canadian regions, however, there are some enhancements seen over east coast of US, which can also contribute to enhanced fire emissions.

In future, the simulated speed of the westerly jet flows over Canada wildfire regions is reduced in both the scenarios (Figure 5 A-B). It indicates that the westerly-induced transported wildfire emissions from Canada boreal forests to the eastern half of Northern America and EUS will be slower in future compared to that in present era. On the one hand, it implies that the advection of smoke plumes will be slightly reduced in future. But on the other hand, this phenomenon can also contribute to the enhanced PM2.5 values at surface as these transporting plumes will be subject to relatively more boundary layer mixing over the EUS and dry deposition/settling enhances. At the same time, the westerly winds over western US below 40 °N is strengthened in future (Figure 5 A-B) compared to present day which indicates more advection flux of wildfire emissions to EUS. Thus, the net effect is more removal of wildfire-emitted PM2.5 from WUS and more influx of wildfire-emitted PM2.5 in EUS.

Along with this dynamical changes, other climatic feedbacks simulated can also contribute to enhancement of EUS pollution. Specifically, the enhancement of wildfire-induced smoke aerosols increases solar absorption and scattering in the future (Figure 6A). This reduces the incoming solar radiation reaching at the surface (Figure 6B) and induces surface cooling. With atmospheric warming and surface cooling, lower-tropospheric stability is enhanced by wildfire aerosols in the future (Figure 6C). The smoke plumes which reaches eastern US are

at an elevated altitude due to self-lofting property of absorbing aerosols as they travel downwind but the smoke over western US is at near surface elevation as it is at its source region. This can explain the more significant atmospheric stability simulated over the eastern US compared to the source regions in western US and boreal forests of Canada. Relatively stronger atmospheric stability over eastern US impose a stronger thermal capping that traps more anthropogenic aerosols and particulate matter near the surface over EUS (already an emission hotspot). At the same time, future increase in wildfire emissions also leads to greater reduction of monthly rainfall (Figure 6D) over EUS, which may additionally strengthen the positive feedback to surface PM2.5 over EUS by reducing wet scavenging of transported wildfire smoke to EUS. Thus, wildfire-emitted aerosols induce positive feedback on the surface PM2.5 concentration over EUS through fire-climate interactions that vary on a regional scale. Moreover, the above discussed dynamical changes in future can also feedback these simulated thermodynamical and precipitation changes, exaggerating the enhancement in PM2.5 values over EUS in future. However, due to computational constrains, no direct quantification of the magnitude of these feedback (with aerosol-radiation and aerosol-cloud interactions turned off) on PM2.5 is performed and would be taken up in future studies.

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Lastly, the reason of why the contribution of wildfire emissions to the total surface PM_{2.5} in EUS is so substantial in the 2050s is the drastic reduction of anthropogenic contribution to the surface PM_{2.5} over EUS in the future primarily due to policy-driven reduction in anthropogenic emissions under the RCP4.5 scenario. Specifically, the simulated ambient summer mean PM_{2.5} concentration exhibits widespread declines in the future (Supplementary Figure 4), with reduction in PM_{2.5} concentration over eastern US in the range of 4-15 µg/m³, which is greatest within North America. Thus, large reduction in anthropogenic contribution combined with increased downwind advection of Canadian

smoke to EUS and the associated positive feedbacks can explain the projected dominance of wildfire emissions over EUS in future.

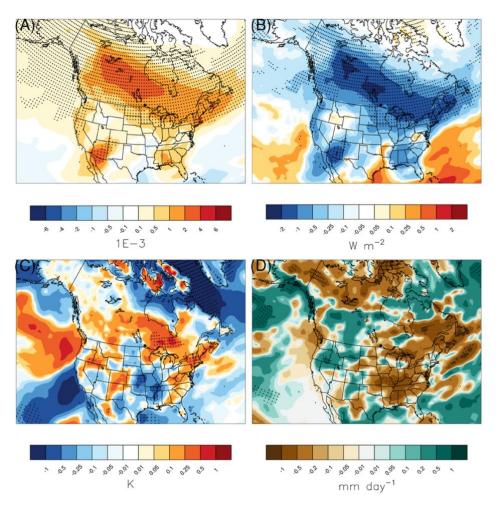


Figure 6: Spatial distribution of decadal mean summer (June through August; JJA) wildfire-induced future changes $[(2050_{ALL}-2050_{WEF}) - (2000_{ALL}-2000_{WEF})]$. A) aerosol absorption optical depth at 550 nm, B) aerosol direct radiative forcing at surface, C) lower-tropospheric stability calculated as the difference between the potential temperature at 900 hPa and 1000 hPa, D) summer averaged precipitation rates, over North America. Areas marked with black dots indicate grids where changes are significant at the 95% confidence level.

3.4 Future Implications and uncertainties

However, is the simulated future enhancement in wildfire contribution over EUS substantial enough to affect the surface $PM_{2.5}$ values over EUS in future? The World Health Organization (WHO) air quality guidelines for annual and daily $PM_{2.5}$ concentration are 10 $\mu g m^{-3}$ and 25 $\mu g m^{-3}$, respectively. As no specific guideline for seasonal-mean $PM_{2.5}$ in the

summer is available, we use the annual guideline value as a reference to understand the implication of wildfire emissions on ambient PM_{2.5} concentration in the future. Interestingly, the mean summertime PM_{2.5} concentration in the wildfire emission free (WEF) scenario is projected to remain within 10 μ g m⁻³ over most of North America, except for the southeastern US (~15% of the domain) (Figure 7A). However, the ALL-scenario projects an increase in the exposure concentration level such that values > 10 μ g/m³ are common in Canada and EUS in the future (Figure 7B). Quantitatively, over Canada, the entire PDF of PM_{2.5} concentration shifts towards higher values by ~5-6 μ g/m³. Specifically, the modal value shifts from ~ 6 μ g/m³ in 2050_{WEF} to 11-12 μ g/m³ in 2050_{ALL} (Figure 7C), so PM_{2.5} concentration is projected to surpass the WHO guidelines over a large fraction of Canada in the future. Similarly, the entire PDF of PM_{2.5} concentration shifts towards higher values by ~2-3 μ g/m³ over EUS, with the mode of the PDF increasing from ~ 7-8 μ g/m³ in 2050_{WEF} to ~ 10-11 μ g/m³ in 2050_{ALL} (Figure 7E). The modal value of summer mean PM_{2.5} over WUS increases from ~ 6 μ g/m³ in 2050_{WEF} to ~ 7-8 μ g/m³ in 2050_{ALL} (Figure 7D), although a few grid cells show PM_{2.5} values greater than 10 μ g/m³ (Figure 7B).

Clearly, the climate-induced enhancement in fires and its influence via the advected wildfire smoke to EUS can have significant implications for air quality management in the future. The PM2.5 enhancement in future over the southern states within EUS is large (Figure 7A-B), which is consistent with Figure 3 and 4 results. However, the future change in burnt area over the same region is negligible or mostly reducing (Figure 1C-D). Thus, it can be argued that the simulated enhancement is mostly related with the dynamic perturbations and thermodynamical feedbacks due to wildfire emissions (Figure 6). As the rate of anthropogenic emissions is also regionally highest over the Southeastern states, the impact of these wildfire-induced climatic feedbacks on local air quality is distinctly seen over the EUS.

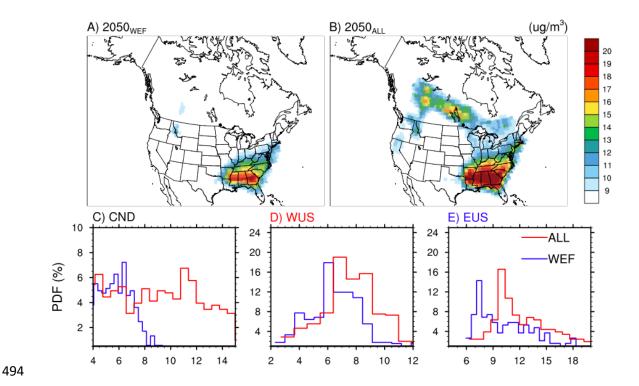


Figure 7: **Spatial distribution and probability density function of** PM_{2.5} **concentration in 2050s. A-B**, Spatial distribution of decadal-average summer (June through August; JJA) mean PM_{2.5} concentration over North America in mid-21st century from 2050_{WEF} (wildfire emission-free) (A) and 2050_{ALL} (wildfire emission-inclusive) (B). **C-E**, Probability density functions (PDFs) of the same within the three regions shown in Figure 2B for Canada; CND (C), western US; WUS (D), and eastern US; EUS (E), respectively, for the 2050_{WEF} (blue) and 2050_{ALL} (red) runs. The y-axis indicates the probability of occurrence of different PM_{2.5} values shown in the x-axis. Only grids over land in North America are used to generate the PDFs. Note the different ranges of values shown in the y- and x-axis in **C-E**. The colorbar and the x-axis for Panel C-E indicates PM_{2.5} values.

Note that our simulated present-day estimates of wildfire induced PM_{2.5} values as well as the percentage contribution of wildfire emissions are within the range of reported values in previous studies over the domain, which augment the fidelity our future projections. Specifically, our simulated present-day estimates of wildfire induced PM_{2.5} values are also within the range of reported values in previous studies over the domain. Reported values of wildfire-induced PM_{2.5} over WUS during summertime vary from ~1 μ g/m³ (Jaffe et al., 2008) to ~2 μ g/m³ (Park et al., 2007) and ~3 μ g/m³ (Ford et al., 2018), with the highest values documented over the Pacific Northwest and west coast regions (~1-4 μ g/m³) (O'Dell et al., 2019). The wildfire-induced PM_{2.5} over EUS during summertime varies from ~1 μ g/m³ (Park

et al., 2007) to $\sim 2.5 \,\mu\text{g/m}^3$ ($\sim 3 \,\mu\text{g/m}^3$ in the southeastern US) (Ford et al., 2018). Consistently, our simulated present-day estimates of wildfire contribution values are also within the range of reported values in previous studies. For example, Meng et al. (2019) found that wildfires can be the largest sectoral contributor (~18-59%) to the populationweighted PM_{2.5} in various subregions of Canada. Over WUS, the present-day percentage contribution of wildfire induced PM_{2.5} to the total PM_{2.5} is reported to be $\sim 12\%$ (Liu et al., 2017), ~ 15% (Park et al., 2007) and ~30% (Ford et al., 2018), with higher values of ~40% in the Pacific Northwest (O'Dell et al., 2019). Over EUS our simulated values are also within the range of previously reported values of ~ 5% (Park et al., 2007) and ~15-18% (Ford et al., 2018). However, our two-way coupled simulations illustrate that future enhancement in the wildfire associated PM_{2.5} over the EUS could be greater compared to the western US, which is not emphasized explicitly in any of the previous studies (although Ford et al., 2018 illustrated increase in PM2.5 over mid and central US from Canadian fires). These could be since inclusion of the wildfire-induced climatic feedbacks in our simulation is an unprecedented exercise. Please also note that our study is focused on JJA period and the wildfires in western US mainly occurs during August-September months, so the results should be compared consciously.

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Nonetheless, inherent limitations in our simulations may introduce uncertainties in the projected future changes. For example, our reported changes in PM_{2.5} concentrations based on relatively coarse resolution simulations and decadal averages likely represent a low-end estimate compared to changes at regional and daily/weekly scales. Moreover, our experiments do not consider the direct human influences such as population change and socioeconomic development on wildfires, which may aggravate the increase in PM_{2.5} concentrations over the densely populated EUS in the future. Common sources of uncertainty in modeling burnt area and fire emission and fire aerosol and smoke are also present in our

model. Fire smoke, in particular, is extremely hard to measure and evaluate. Lastly, inherent uncertainties in the physics parameterizations used in the model, sensitivity of climate to GHGs emissions, and the RCP scenarios should also be noted. Thus, ensemble modeling considering different emissions scenarios, population and future time periods, and the use of a finer spatial resolution may provide a more robust and better quantification of the wildfire-induced impact on policy regulated improvements in PM_{2.5} over EUS.

4. Conclusion

In summary, online coupled fire-climate-ecosystem simulations project a nearly twofold increase in wildfire-induced summer-mean surface PM_{2.5} concentration by the mid-21st century over the entire North America. In a wildfire-emission free future, a large portion of North America will have PM_{2.5} values below the WHO guidelines. But in a future with wildfire emissions, the improvements from policy-driven reductions in anthropogenic PM_{2.5} will be compromised by the projected doubling of PM_{2.5} from wildfires. More strikingly, wildfire-induced enhancement in surface PM_{2.5} values and percentage contribution of the wildfire emissions over EUS could be substantial by mid-century. This is mainly because of the large enhancement in fires over Northern America by 2050s and associated increase in amount of downwind transport of smoke to EUS. In addition, enhancement of smoke transport induces a positive climate feedback to PM_{2.5} concentrations over EUS by increasing the lower-tropospheric stability and reducing wet scavenging rates. Despite the inherent limitations, this study highlights the natural versus anthropogenic contributions and the non-local nature of air pollution that can complicate regulatory strategies aimed at improving air quality over the eastern US in a warmer future.

Data availability statements

- The HMS data used in this paper are available free through the link
- https://www.ospo.noaa.gov/Products/land/hms.html. The model simulations data are
- available at https://portal.nersc.gov/project/m1660/yang560/wildfire

Code availability statements

- The model code and scripts are available at
- 567 https://portal.nersc.gov/project/m1660/yang560/wildfire

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Authors Contribution

- YQ, CS and RL designed this study. CS did the model and satellite analysis and wrote the
- first draft of the manuscript. YZou performed the simulations. All authors provided inputs
- throughout the study and helped in the drafting and submission process.

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References

- Abatzoglou, J. T. and Williams, A. P.: Impact of anthropogenic climate change on wildfire
- 587 across western US forests, Proc. Natl. Acad. Sci., 113(42), 11770 LP 11775,
- 588 doi:10.1073/pnas.1607171113, 2016.
- Andela, N., Morton, D. C., Giglio, L., Chen, Y., van der Werf, G. R., Kasibhatla, P. S.,
- DeFries, R. S., Collatz, G. J., Hantson, S., Kloster, S., Bachelet, D., Forrest, M., Lasslop, G.,
- Li, F., Mangeon, S., Melton, J. R., Yue, C. and Randerson, J. T.: A human-driven decline in
- 592 global burned area, Science (80-.)., 356(6345), 1356 LP 1362,
- 593 doi:10.1126/science.aal4108, 2017.
- 594 Anjali, H., Muhammad, A., Anthony, D. M., Karen, S., R., S. M., Mick, M., M., T. A., J., A.
- M. and Martine, D.: Impact of Fine Particulate Matter (PM_{2.5}) Exposure During Wildfires on
- 596 Cardiovascular Health Outcomes, J. Am. Heart Assoc., 4(7), e001653,
- 597 doi:10.1161/JAHA.114.001653, 2019.
- 598 Black, C., Tesfaigzi, Y., Bassein, J. A. and Miller, L. A.: Wildfire smoke exposure and
- 599 human health: Significant gaps in research for a growing public health issue, Environ.
- Toxicol. Pharmacol., 55, 186–195, doi:https://doi.org/10.1016/j.etap.2017.08.022, 2017.
- Brey, S. J., Ruminski, M., Atwood, S. A. and Fischer, E. V: Connecting smoke plumes to

- sources using Hazard Mapping System (HMS) smoke and fire location data over North
- 603 America, Atmos. Chem. Phys., 18(3), 1745–1761, doi:10.5194/acp-18-1745-2018, 2018.
- Dempsey, F.: Forest Fire Effects on Air Quality in Ontario: Evaluation of Several Recent
- 605 Examples, Bull. Am. Meteorol. Soc., 94(7), 1059–1064, doi:10.1175/BAMS-D-11-00202.1,
- 606 2013.
- Dominici, F., Peng, R. D., Bell, M. L., Pham, L., McDermott, A., Zeger, S. L. and Samet, J.
- 608 M.: Fine Particulate Air Pollution and Hospital Admission for Cardiovascular and
- Respiratory Diseases, JAMA, 295(10), 1127–1134, doi:10.1001/jama.295.10.1127, 2006.
- Diao, M., Holloway, T., Choi, S., O'Neill, S.M., Al-Hamdan, M.Z., Van Donkelaar, A.,
- Martin, R.V., Jin, X., Fiore, A.M., Henze, D.K. and Lacey, F., 2019. Methods, availability,
- and applications of PM2. 5 exposure estimates derived from ground measurements, satellite,
- and atmospheric models. Journal of the Air & Waste Management Association, 69(12),
- 614 pp.1391-1414.
- Ford, B., Val Martin, M., Zelasky, S. E., Fischer, E. V, Anenberg, S. C., Heald, C. L. and
- Pierce, J. R.: Future Fire Impacts on Smoke Concentrations, Visibility, and Health in the
- 617 Contiguous United States, GeoHealth, 2(8), 229–247, doi:10.1029/2018GH000144, 2018.
- 618 Gillett, N. P., Weaver, A. J., Zwiers, F. W. and Flannigan, M. D.: Detecting the effect of
- climate change on Canadian forest fires, Geophys. Res. Lett., 31(18),
- doi:10.1029/2004GL020876, 2004.
- 621 Gunsch, M. J., May, N. W., Wen, M., Bottenus, C. L. H., Gardner, D. J., VanReken, T. M.,
- Bertman, S. B., Hopke, P. K., Ault, A. P. and Pratt, K. A.: Ubiquitous influence of wildfire
- 623 emissions and secondary organic aerosol on summertime atmospheric aerosol in the forested
- 624 Great Lakes region, Atmos. Chem. Phys., 18(5), 3701–3715, doi:10.5194/acp-18-3701-2018,
- 625 2018.
- 626 Guan S, Wong DC, Gao Y, Zhang T, Pouliot G. Impact of wildfire on particulate matter in
- the southeastern United States in November 2016. Sci Total Environ. 2020;724:138354.
- 628 doi:10.1016/j.scitotenv.2020.138354.
- Johnston F. H., Henderson. S. B., Yang, C., T., R. J., Miriam, M., S., D. R., Patrick, K.,
- 630 M.J.S., B. D. and Michael, B.: Estimated Global Mortality Attributable to Smoke from
- 631 Landscape Fires, Environ. Health Perspect., 120(5), 695–701, doi:10.1289/ehp.1104422,
- 632 2012.
- Harris, R. M. B., Remenyi, T. A., Williamson, G. J., Bindoff, N. L., and Bowman, D. M. J.
- 634 S.: Climate-vegetation fire interactions and feedbacks: trivial detail or major barrier to
- projecting the future of the Earth system?, Wires Clim. Change, 7, 910-931,
- 636 10.1002/wcc.428, 2016.
- Hantson, S., Arneth, A., Harrison, S. P., Kelley, D. I., Prentice, I. C., Rabin, S. S., et al.
- 638 (2016). The status and challenge of global fire modelling. Biogeosciences, 13, 3359–3375.
- 639 https://doi.org/10.5194/bg-13-3359-2016
- Hu, X., Yu, C., Tian, D., Ruminski, M., Robertson, K., Waller, L. A. and Liu, Y.:
- Comparison of the Hazard Mapping System (HMS) fire product to ground-based fire records
- in Georgia, USA, J. Geophys. Res. Atmos., 121(6), 2901–2910, doi:10.1002/2015JD024448,
- 643 2016.

- 644 HURTT, G. C., FROLKING, S., FEARON, M. G., MOORE, B., SHEVLIAKOVA, E.,
- 645 MALYSHEV, S., PACALA, S. W. and HOUGHTON, R. A.: The underpinnings of land-use
- 646 history: three centuries of global gridded land-use transitions, wood-harvest activity, and
- resulting secondary lands, Glob. Chang. Biol., 12(7), 1208–1229, doi:10.1111/j.1365-
- 648 2486.2006.01150.x, 2006.
- Jaffe, D., Hafner, W., Chand, D., Westerling, A. and Spracklen, D.: Interannual Variations in
- PM_{2.5} due to Wildfires in the Western United States, Environ. Sci. Technol., 42(8), 2812–
- 651 2818, doi:10.1021/es702755v, 2008.
- Jolly, W. M., Cochrane, M. A., Freeborn, P. H., Holden, Z. A., Brown, T. J., Williamson, G.
- J. and Bowman, D. M. J. S.: Climate-induced variations in global wildfire danger from 1979
- to 2013, Nat. Commun., 6, 7537 [online] Available from:
- 655 https://doi.org/10.1038/ncomms8537, 2015.
- Kaulfus, A. S., Nair, U., Jaffe, D., Christopher, S. A. and Goodrick, S.: Biomass Burning
- Smoke Climatology of the United States: Implications for Particulate Matter Air Quality,
- 658 Environ. Sci. Technol., 51(20), 11731–11741, doi:10.1021/acs.est.7b03292, 2017.
- 659 Kirchmeier-Young, M. C., Zwiers, F. W., Gillett, N. P. and Cannon, A. J.: Attributing
- extreme fire risk in Western Canada to human emissions, Clim. Change, 144(2), 365–379,
- doi:10.1007/s10584-017-2030-0, 2017.
- Kitzberger, T., Falk, D. A., Westerling, A. L. and Swetnam, T. W.: Direct and indirect
- climate controls predict heterogeneous early-mid 21st century wildfire burned area across
- western and North America, PLoS One, 12(12), e0188486 [online] Available from:
- 665 https://doi.org/10.1371/journal.pone.0188486, 2017.
- Knorr, W., Dentener, F., Lamarque, J.-F., Jiang, L. and Arneth, A.: Wildfire air pollution
- hazard during the 21st century, Atmos. Chem. Phys., 17(14), 9223–9236, doi:10.5194/acp-
- 668 17-9223-2017, 2017.
- Koplitz, S.N., Nolte, C.G., Pouliot, G.A., Vukovich, J.M. and Beidler, J., 2018. Influence of
- uncertainties in burned area estimates on modeled wildland fire PM2. 5 and ozone pollution
- in the contiguous US. *Atmospheric environment*, 191, pp.328-339.
- Lamarque, J. F., Bond, T. C., Eyring, V., Granier, C., Heil, A., Klimont, Z., Lee, D., Liousse,
- 673 C., Mieville, A., Owen, 628 B., Schultz, M. G., Shindell, D., Smith, S. J., Stehfest, E., Van
- Aardenne, J., Cooper, O. R., Kainuma, M., 629 Mahowald, N., McConnell, J. R., Naik, V.,
- Riahi, K., and van Vuuren, D. P.: Historical (1850-2000) gridded anthropogenic and biomass
- burning emissions of reactive gases and aerosols: methodology and application, 631 Atmos.
- 677 Chem. Phys., 10, 7017-7039, 10.5194/acp-10-7017-2010, 2010.
- Leibensperger, E. M., Mickley, L. J., Jacob, D. J., Chen, W.-T., Seinfeld, J. H., Nenes, A.,
- Adams, P. J., Streets, D. G., Kumar, N. and Rind, D.: Climatic effects of 1950–2050
- changes in US anthropogenic aerosols Part 2: Climate response, Atmos. Chem.
- 681 Phys., 12(7), 3349–3362, doi:10.5194/acp-12-3349-2012, 2012.
- 682 Li, F., Zeng, X. D., & Levis, S. (2012). A process-based fire parameterization of intermediate
- complexity in a dynamic global vegetation model. Biogeosciences, 9(7), 2761–2780.
- 684 https://doi.org/10.5194/bg-9-2761-2012
- Liu, J. C., Mickley, L. J., Sulprizio, M. P., Dominici, F., Yue, X., Ebisu, K., Anderson, G. B.,
- Khan, R. F. A., Bravo, M. A. and Bell, M. L.: Particulate Air Pollution from Wildfires in the

- 687 Western US under Climate Change, Clim. Change, 138(3), 655–666, doi:10.1007/s10584-
- 688 016-1762-6, 2016.
- 689 Liu, Y., Goodrick, S. and Heilman, W.: Wildland fire emissions, carbon, and climate:
- 690 Wildfire-climate interactions, For. Ecol. Manage., 317, 80–96,
- 691 doi:https://doi.org/10.1016/j.foreco.2013.02.020, 2014.
- Liu, X., Easter, R. C., Ghan, S. J., Zaveri, R., Rasch, P., Shi, X., Lamarque, J. F., Gettelman,
- 693 A., Morrison, H., Vitt, F., Conley, A., Park, S., Neale, R., Hannay, C., Ekman, A. M. L.,
- Hess, P., Mahowald, N., Collins, W., Iacono, M. J., Bretherton, C. S., Flanner, M. G., and
- Mitchell, D.: Toward a minimal representation of aerosols in climate models: description and
- 696 evaluation in the Community Atmosphere Model CAM5, Geosci. Model Dev., 5, 709-652
- 697 739, 10.5194/gmd-5-709-2012, 2012.
- 698 Meng, J., Martin, R.V., Li, C., van Donkelaar, A., Tzompa-Sosa, Z.A., Yue, X., Xu, J.W.,
- Weagle, C.L. and Burnett, R.T., 2019. Source Contributions to Ambient Fine Particulate
- Matter for Canada. *Environmental science & technology*, 53(17), pp.10269-10278.
- McClure, C. D. and Jaffe, D. A.: US particulate matter air quality improves except in
- 702 wildfire-prone areas, Proc. Natl. Acad. Sci., 115(31), 7901 LP 7906,
- 703 doi:10.1073/pnas.1804353115, 2018.
- Neale, R. B., Chen, C. C., Gettelman, A., Lauritzen, P. H., Park, S., Williamson, D. L.,
- Conley, A. J., Garcia, R., Kinnison, D., Lamarque, J. F., Marsh, D., Mills, M., Smith, A. K.,
- Tilmes, S., Vitt, F., Morrison, H., Cameron671 Smith, P., Collins, W. D., Iacono, M. J.,
- Easter, R. C., Ghan, S. J., Liu, X. H., Rasch, P. J., and Taylor, M. A.: Description of the
- NCAR Community Atmosphere Model (CAM 5.0), NCAR 289, 2013
- Nolte, C. G., Spero, T. L., Bowden, J. H., Mallard, M. S. and Dolwick, P. D.: The potential
- of effects of climate change on air quality across the conterminous US at 2030 under three
- 711 Representative Concentration Pathways, Atmos. Chem. Phys., 18(20), 15471–15489,
- 712 doi:10.5194/acp-18-15471-2018, 2018.
- Katelyn O'Dell, Bonne Ford, Emily V. Fischer, and Jeffrey R. Pierce Environmental Science
- 714 & Technology 2019 53 (4), 1797-1804, DOI: 10.1021/acs.est.8b05430.
- Partain, J. L., Alden, S., Strader, H., Bhatt, U. S., Bieniek, P. A., Brettschneider, B. R.,
- 717 Walsh, J. E., Lader, R. T., Olsson, P. O., Rupp, T. S., Thoman, R. L., York, A. D. and Ziel,
- 718 R. H.: An Assessment of the Role of Anthropogenic Climate Change in the Alaska Fire
- 719 Season of 2015, Bull. Am. Meteorol. Soc., 97(12), S14–S18, doi:10.1175/BAMS-D-16-
- 720 0149.1, 2016.
- Park, S., Bretherton, C. S., and Rasch, P. J.: Integrating Cloud Processes in the Community
- 722 Atmosphere Model, Version 5, J. Climate, 27, 6821-6856, 10.1175/Jcli-D-14-00087.1, 2014.
- Park, R.J., Jacob, D.J. and Logan, J.A., 2007. Fire and biofuel contributions to annual mean
- aerosol mass concentrations in the United States. Atmospheric Environment, 41(35), pp.7389-
- 725 7400.

- 726 Pierce, J. R., Val Martin, M., & Heald, C. L. (2017). Estimating the Effects of Changing
- 727 Climate on Fires and Consequences for U.S. Air Quality, Using a Set of Global and Regional
- 728 Climate Models (final report no. JFSP-13-1-01-4). Retrieved from
- https://www.firescience.gov/projects/13-1-01-4/project/13-1-01-4_final_report.pdf

- Pouliot G, Pace TG, Roy B, Pierce T, Mobley D. Development of a biomass burning
- emissions inventory by combining satellite and ground-based information. J Appl Remote
- 732 Sens. 2008;2:021501. doi: 10.1117/1.2939551.
- 733 Randerson, J. T., Chen, Y., van der Werf, G. R., Rogers, B. M., & Morton, D. C. (2012).
- Global burned area and biomass burning emissions from small fires. Journal of Geophysical
- 735 Research, 117, G04012. https://doi.org/10.1029/2012JG002128
- Rolph, G. D., Draxler, R. R., Stein, A. F., Taylor, A., Ruminski, M. G., Kondragunta, S.,
- Zeng, J., Huang, H.-C., Manikin, G., McQueen, J. T. and Davidson, P. M.: Description and
- Verification of the NOAA Smoke Forecasting System: The 2007 Fire Season, Weather
- 739 Forecast., 24(2), 361–378, doi:10.1175/2008WAF2222165.1, 2009.
- Spracklen, D. V., Mickley, L. J., Logan, J. A., Hudman, R. C., Yevich, R., Flannigan, M. D.,
- Westerling, A. L. (2009). Impacts of climate change from 2000 to 2050 on wildfire activity
- and carbonaceous aerosol concentrations in the western United States. Journal of Geophysical
- 743 Research, 114, D20301. https://doi.org/10.1029/2008JD010966
- Sun, Y., Gu, L. H., and Dickinson, R. E.: A numerical issue in calculating the coupled carbon
- and water fluxes in a climate model, J. Geophys. Res.-Atmos., 117,
- 746 D2210310.1029/2012jd018059, 2012
- Sofiev, M., Ermakova, T., and Vankevich, R.: Evaluation of the smoke-injection height from
- wild-land fires using remote-sensing data, Atmos. Chem. Phys., 12, 1995-2006, 10.5194/acp-
- 749 12-1995-2012, 2012
- 750 Shi, H., Jiang, Z., Zhao, B., Li, Z., Chen, Y., Gu, Y., Jiang, J. H., Lee, M., Liou, K.-N., Neu,
- J. L., Payne, V. H., Su, H., Wang, Y., Witek, M. and Worden, J.: Modeling Study of the Air
- Ouality Impact of Record-Breaking Southern California Wildfires in December 2017, J.
- 753 Geophys. Res. Atmos., 0(0), doi:10.1029/2019JD030472, 2019.
- 754 U.S. EPA (U.S. Environmental Protection Agency). 2009. Integrated Science Assessment
- 755 (ISA) For Particulate Matter (Final Report). EPA/600/R-08/139F. Washington, DC:U.S. EPA.
- 756 U.S. EPA. 2018. Our Nation's Air. https://gispub.epa.gov/air/trendsreport/2018/
- 757 Val Martin, M., Heald, C. L., Lamarque, J.-F., Tilmes, S., Emmons, L. K. and Schichtel, B.
- A.: How emissions, climate, and land use change will impact mid-century air quality over the
- United States: a focus on effects at national parks, Atmos. Chem. Phys., 15(5), 2805–2823,
- 760 doi:10.5194/acp-15-2805-2015, 2015.
- Van der Werf, G. R., Randerson, J. T., Giglio, L., Collatz, G. J., Kasibhatla, P. S., and
- Arellano, A. F.: Interannual variability in global biomass burning emissions from 1997 to
- 763 2004, Atmos. Chem. Phys., 6, 3423-3441, DOI 737 10.5194/acp-6-3423-2006, 2006
- Van Der Werf, G. R., Randerson, J. T., Giglio, L., Collatz, G. J., Mu, M., Kasibhatla, P. S.,
- Morton, D. C., Defries, R. S., Jin, Y. and Van Leeuwen, T. T.: Global fire emissions and the
- contribution of deforestation, savanna, forest, agricultural, and peat fires (1997-2009), Atmos.
- 767 Chem. Phys., 10(23), 11707–11735, doi:10.5194/acp-10-11707-2010, 2010.
- Van Donkelaar, A., R. V. Martin, M. Brauer, N. C. Hsu, R. A. Kahn, R. C. Levy, A.
- Lyapustin, A. M. Sayer, and D. M. Winker. 2018. Global Annual PM_{2.5} Grids from MODIS,
- MISR and SeaWiFS Aerosol Optical Depth (AOD) with GWR, 1998-2016. Palisades NY:

- 771 NASA Socioeconomic Data and Applications Center (SEDAC).
- https://doi.org/10.7927/H4ZK5DQS. Accessed 16 November 2019.
- Ward, D. S., Kloster, S., Mahowald, N. M., Rogers, B. M., Randerson, J. T. and Hess, P. G.:
- The changing radiative forcing of fires: global model estimates for past, present and future,
- 775 Atmos. Chem. Phys., 12(22), 10857–10886, doi:10.5194/acp-12-10857-2012, 2012.
- Wotton, B. M., Flannigan, M. D., and Marshall, G. A.: Potential climate change impacts on
- fire intensity and key wildfire suppression thresholds in Canada, Environ. Res. Lett., 12,
- 778 095003, https://doi.org/10.1088/1748-9326/aa7e6e, 2017.
- Westerling, A. L., Hidalgo, H. G., Cayan, D. R. and Swetnam, T. W.: Warming and Earlier
- 780 Spring Increase Western U.S. Forest Wildfire Activity, Science (80-.)., 313(5789), 940 LP –
- 781 943, doi:10.1126/science.1128834, 2006.
- Wotawa, G. and Trainer, M.: The Influence of Canadian Forest Fires on Pollutant
- 783 Concentrations in the United States, Science (80-.)., 288(5464), 324 LP 328,
- 784 doi:10.1126/science.288.5464.324, 2000.
- 785 Wu, Y., Arapi, A., Huang, J., Gross, B. and Moshary, F.: Intra-continental wildfire smoke
- transport and impact on local air quality observed by ground-based and satellite remote
- sensing in New York City, Atmos. Environ., 187, 266–281,
- 788 doi:https://doi.org/10.1016/j.atmosenv.2018.06.006, 2018.
- Xue, Z., Gupta, P., and Christopher, S.: Satellite-based estimation of the impacts of
- summertime wildfires on PM2.5concentration in the United States, Atmos. Chem. Phys., 21,
- 791 11243–11256, https://doi.org/10.5194/acp-21-11243-2021, 2021.
- Yang, P.-L., Y. Zhang, K. Wang, P. Doraiswamy, and S.-H. Cho, 2019, Health Impacts and
- Cost-Benefit Analyses of Surface O₃ and PM_{2.5} over the U.S. under Future Climate and
- Emission Scenarios, *Environmental Research*, 178, November 2019, 108687,
- 795 <u>https://doi.org/10.1016/j.envres.2019.108687</u>.
- Yue, X., Mickley, L. J., Logan, J. A. and Kaplan, J. O.: Ensemble projections of wildfire
- activity and carbonaceous aerosol concentrations over the western United States in the mid-
- 798 21st century, Atmos. Environ., 77, 767–780,
- 799 doi:https://doi.org/10.1016/j.atmosenv.2013.06.003, 2013.
- Zou, Y., Wang, Y., Ke, Z., Tian, H., Yang, J. and Liu, Y.: Development of a REgion-Specific
- 801 Ecosystem Feedback Fire (RESFire) Model in the Community Earth System Model, J. Adv.
- 802 Model. Earth Syst., 11(2), 417–445, doi:10.1029/2018MS001368, 2019.
- Zou, Y., Wang, Y., Qian, Y., Tian, H., Yang, J. and Alvarado, E.: Using CESM-RESFire to
- understand climate–fire–ecosystem interactions and the implications for decadal climate
- variability, Atmos. Chem. Phys., 20, 995–1020, https://doi.org/10.5194/acp-20-995-2020,
- 806 2020.
- Zhang, Y., P.-L. Yang, Y. Gao, R. L. Leung, and M. Bell, 2020, Health and Economic
- 808 Impacts of Air Pollution Induced by Climate Extremes over the Continental U.S.,
- 809 Environmental International, 143, 105921, https://doi.org/10.1016/j.envint.2020.105921.