



1 **Detection of Stratospheric Air Intrusion Events From Ground-based**
2 **High-resolution $^{10}\text{Be}/^7\text{Be}$ by Accelerator Mass Spectrometry**

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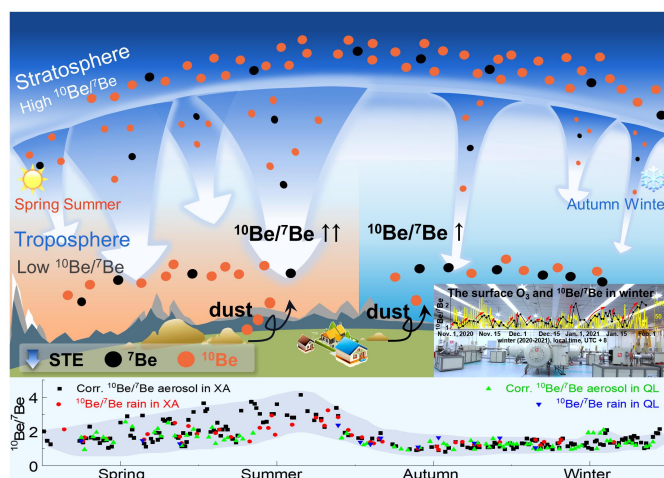
15 **Abstract**

16 Locally rapid stratospheric air intrusions facilitate the transport of stratospheric
17 material to the troposphere. Long-term continuous monitoring of such events by
18 traditional techniques, such as sounding technology, is challenging. Beryllium-7 (^7Be)
19 and beryllium-10 (^{10}Be) offer an alternative. These isotopes are formed by cosmic
20 rays and are mainly produced in the lower stratosphere and upper troposphere. Due to
21 their similar geochemical properties and substantial difference in half-lives favor
22 relatively high $^{10}\text{Be}/^7\text{Be}$ ratios in the stratosphere, as compared to the troposphere.
23 Monitoring surface $^{10}\text{Be}/^7\text{Be}$ ratios affords a potential means to identify stratospheric
24 air intrusions. However, high temporal resolution $^{10}\text{Be}/^7\text{Be}$ observational records must
25 be taken and corrected for dust-borne ^{10}Be to identify stratospheric air intrusions. In



26 this study, we use Accelerator Mass Spectrometry to measure both ^7Be and ^{10}Be in
27 rain and aerosol (down to ~ 200 cubic meters air) with an error of $\sim 1.5\%$. We correct
28 for dust-borne ^{10}Be using soil Al. This method provides precise measurements with
29 daily resolution. We present annual beryllium isotopes (^7Be , ^{10}Be , and $^{10}\text{Be}/^7\text{Be}$ ratio)
30 record for the Chinese Loess Plateau that includes several regional sites. We show that
31 for the city of Xi'an, the proportion of dust-borne (resuspended) ^{10}Be was $\sim 24\%$ in
32 2020/21. Our results confirm that stratospheric air intrusion events in the Loess
33 Plateau are frequent and rapid throughout the year and are strongest in the spring
34 (March-July), when $^{10}\text{Be}/^7\text{Be}$ values were observed to increase about a factor of 3.
35 Even in winter, weaker stratospheric air intrusion events can be detected. Calculated
36 $\Delta(^{10}\text{Be}/^7\text{Be})$ values in winter suggest stratospheric ozone transport can lead to an
37 $\sim 25\%$ cumulative increase the surface ozone.

38



39

40

Abstract graph

41 **Key words:** Cosmogenic nuclides, Stratospheric air intrusion, $^{10}\text{Be}/^7\text{Be}$ ratio,



42 Accelerator mass spectrometry, Ozone transport, Loess Plateau

43

44 1. Introduction

45 Stratospheric air intrusions are a vital process in stratosphere-troposphere
46 exchange (STE), providing a transport mechanism for stratospheric materials to the
47 ground (Holton et al., 1995;Randel et al., 2010). Local stratospheric air intrusions are
48 difficult to quantify and may have hard-to-capture frequent intrusion events
49 (Hodnebrog et al., 2016;Sudo et al., 2003). Although these local stratospheric air
50 intrusions are relatively weak, they can still facilitate the transport of natural and
51 anthropogenic trace chemicals and significantly affects the atmospheric chemistry,
52 atmospheric heat budget, and radiation properties. For example, stratospheric air
53 intrusions can facilitate ozone (O₃) transport to the troposphere (Baylon et al., 2016).
54 Tropospheric O₃ affects human health and ecological environments. Due to the
55 scarcity of observational data in the lower stratosphere and upper troposphere, there
56 are few observational studies on this process. Traditional observation methods of
57 stratospheric air intrusion mainly rely on sounding technology with multi-element
58 sensors (Fischer et al., 2008). However, the observation process is complicated, and it
59 is challenging to observe continuously due to the influence of weather, let alone
60 accurate capture of frequent intrusion events in local areas.

61 Meteoric ⁷Be and ¹⁰Be are produced by cosmic rays interacting with oxygen and
62 nitrogen, mainly in the lower stratosphere and upper troposphere (Brown et al., 1989).
63 After their formation, these beryllium isotopes are rapidly absorbed and transported



64 by submicron aerosols (Lal et al., 1958), and it is generally believed that
65 ground-based ^7Be analyses offer a means of identifying and tracing stratospheric air
66 intrusions (Bhandari et al., 1966; Hernandez-Ceballos et al., 2017). Using ^7Be alone is
67 complicated by tropospheric weather conditions and the latitudinal dependence of ^7Be
68 production at different latitudes (Masarik and Beer, 1999). Using $^{10}\text{Be}/^7\text{Be}$ ratios can
69 avoid these effects, making it more accurate and sensitive than ^7Be concentrations
70 alone to trace stratospheric air intrusions (Jordan et al., 2003). The transport and
71 deposition processes of ^7Be and ^{10}Be are same, and their production rates ratio at each
72 latitude is consistent (Masarik and Beer, 1999). The two isotopes have different
73 residence times in the stratosphere and troposphere and the very different half-lives:
74 ^7Be ($T_{1/2} = 53.29$ d) and ^{10}Be ($T_{1/2} = 1.36$ Ma). These factors make stratospheric
75 $^{10}\text{Be}/^7\text{Be}$ ratios much larger than $^{10}\text{Be}/^7\text{Be}$ ratios observed near the surface air (Jordan
76 et al., 2003). When stratospheric air intrudes, the $^{10}\text{Be}/^7\text{Be}$ ratio in the atmosphere at
77 the earth's surface increases (Brown et al., 1989).

78 ^7Be and ^{10}Be are typically measured by gamma spectrometry and accelerator
79 mass spectrometry (AMS), respectively. Due to the low counting efficiency of the
80 gamma spectrometry, large samples are required ($> 10^3$ m³ air or > 10 L rain), as well
81 as lengthy measurements (> 1 day) to obtain optimal measurement uncertainties
82 (Tiessen et al., 2019; Yamagata et al., 2019). It is even more challenging to obtain ^7Be
83 data with low uncertainty at a daily resolution based on gamma spectrometry,
84 especially in short-period field sampling and quickly stratospheric air intrusion event
85 research. Measuring ^7Be by AMS significantly reduces sample size requirements and



86 improves the detection limit by approximately 10-100 times (Raisbeck and Yiou,
87 1988). Hence, we measure both isotopes with AMS.

88 Due to the high abundance of ^{10}Be in topsoil, resuspended dust-borne ^{10}Be must
89 be accounted (Czymzik et al., 2018; Yamagata et al., 2010). In particular, corrections
90 are necessary for STE studies in high dust areas such as the Chinese Loess Plateau
91 (one of the leading dust source areas in Asia (Zhang et al., 1997)), that reflect
92 $^{10}\text{Be}/^7\text{Be}$ from dry deposition (Heikkila et al., 2013). We have devised a means of
93 making these corrections based in part on our previous studies of ^{10}Be in loess (Beck
94 et al., 2018). In addition, high-precision ^7Be and ^{10}Be data can contribute to our
95 understanding of atmospheric material transport and circulation (Chham et al.,
96 2019; Liu et al., 2022), solar variability (Adolphi et al., 2014), and paleoclimate (Beck
97 et al., 2018).

98 In this work, we use ground-based $^{10}\text{Be}/^7\text{Be}$ measurements as a means of
99 documenting stratospheric air intrusion events through time. In order to achieve this,
100 we first developed an AMS measurement method to quantify ^7Be and ^{10}Be together
101 (named $^7\text{Be}\text{-}^{10}\text{Be}\text{-AMS}$). Next, we identified an analytical means to remove the
102 resuspended dust-borne ^{10}Be component, ultimately enabling the acquisition of
103 $^{10}\text{Be}/^7\text{Be}$ records with low uncertainty and high temporal resolution in various
104 environments. We made numerous analyses of aerosol and rain samples ($n = 398$)
105 from the Loess Plateau, and an annual $^{10}\text{Be}/^7\text{Be}$ record with the daily resolution was
106 obtained for the first time. We show a relationship between these stratospheric air
107 intrusions in winter with surface O_3 .



108 **2. Methods**

109 **2.1 Sample collection**

110 From May 2020 to June 2021, aerosol and rainwater samples from multiple
111 locations were collected. These sampling points include (Fig. S1): Xi'an (XA,
112 34°13'20"N, 109°00'18"E) and Qinling (QL, 34°3'43"N, 108°20'48"E) for long-term
113 monitoring. Zhongwei (ZW, 37°30'57"N, 105°11'8"E), Taiyuan (TY, 37°48'26"N,
114 112°34'55"E), Chengdu (CD, 30°56'44"N, 103°40'41"E), Nanning (NN, 22°50'18"N,
115 108°16'51"E), and Lianyungang (LYG, 34°46'4"N, 119°26'55"E) for short-term
116 intermittent monitoring. Aerosols and rainwater were collected at XA and QL, while
117 only aerosols were collected at the remaining sampling sites.

118 Total suspended particulate (TSP) samples passed through a large-flow particle
119 collector (TH-1000CII, Wuhan Tianhong Instrument. Co., Ltd.) and a portable
120 small-flow particle collector (TH-150H, Wuhan Tianhong Instrument. Co., Ltd.), prior
121 to collection on polypropylene filter membranes. Each collection period was 24 h and
122 sampled approximately 200-500 m³ air. For rain sample collection, the initial rainfall
123 in the first half-hour was removed to avoid interference from suspended dust.
124 Rainwater samples were collected and stored separately in acidified polyethylene
125 containers to prevent beryllium in the rainwater from adsorbing on the container walls.
126 Detailed rain sample collection information is described elsewhere (Zhang and Fu,
127 2017). Among multiple sampling points, the meteorological data collector
128 (HOBO-U30 Station) established at the XA sampling site was used to accurately
129 monitor meteorological data such as wind speed, precipitation, and solar radiation



130 intensity. Surface O₃ concentration data at the XA site was downloaded from the
131 government website <https://www.aqistudy.cn/>. The sampling times provided in this
132 article are based on local time (UTC+8).

133 **2.2 Extraction of Be and Al in the sample and preparation of the BeO target**

134 ⁷Be and ¹⁰Be targets were prepared according to established experimental
135 procedures for loess ¹⁰Be (Zhou et al., 2007), and rainwater ⁷Be and ¹⁰Be (Zhang and
136 Fu, 2017). The existing experimental process for aerosol samples followed 3 steps
137 (Fig. S2): 1) ⁷Be and ¹⁰Be extraction; 2) ion-exchange separation and purification; and
138 3) BeO preparation. The first step removes organic at 600 °C, and the remaining
139 aerosol component is dissolved in acid. Next, Be(OH)₂ is obtained by reaction with a
140 weak base (NH₃·H₂O), and the precipitate is oxidized to BeO at 900 °C. The Al
141 content of aerosols extracted by acid immersion was measured by ICP-AES
142 (ULTIMA-2, HORIBA Jobin Yvon, France) according to the method proposed by
143 Yamagata et al. (2010).

144 **2.3 AMS analysis of ⁷Be and ¹⁰Be**

145 ⁷Be and ¹⁰Be were analyzed in the same target by the 3 MV multi-nuclide AMS
146 at the Xi'an Accelerator Mass Spectrometry Center, Institute of Earth Environment,
147 Chinese Academy of Sciences. The analysis method follows the approach we
148 established in 2017 (Zhang and Fu, 2017), with some subsequent refinements. In
149 particular, the transmission efficiency after the second stripping foil was substantially
150 improved, reaching approximately 24%, which greatly improved the total
151 transmission efficiency and further improved the analysis precision. ¹⁰Be⁴⁺ or ⁷Be⁴⁺



152 were analyzed in a gas ionization detector (energy spectrum shown in Fig. S3).
153 Measurement details are given in the Methods section of the Supplementary
154 Information.

155 Data quality was assessed considering chemical preparation, measurement
156 uncertainties, and parallel sample monitoring results. A threshold sample recovery rate
157 for the chemical treatment process is maintained at > 80%. Each measurement batch
158 included a standard sample and a blank sample. Standard samples were used for
159 calibration, and blank samples were used for monitoring procedures. The results for
160 the blank samples were well below measured sample values ($\sim 10^3$ times). The
161 measurement results of parallel samples are consistent within 1σ (Table S1).

162 AMS ^7Be measurements were cross-checked by gamma spectrometry. Large
163 samples (approximately 2000 m³ air) were collected, and polypropylene membranes
164 with high aerosol concentrations were selected for comparison. About 1/4 of them
165 were analyzed by ^7Be -AMS. The ^7Be in the remaining 3/4 filter membranes were
166 measured using a high purity germanium (HpGe) detector. The results show that the
167 AMS results were consistent with the measurement HpGe detector results (Table S2).
168 Furthermore, for the same samples, the uncertainties for samples measured by AMS
169 were uniformly lower than those measured by HpGe detector (detailed measurement
170 information of the HpGe detector is included in the support information).

171 **2.4 Quantification of the resuspended dust contribution to ^{10}Be**

172 The resuspended dust-borne ^{10}Be contribution was corrected as:
173 $[^{10}\text{Be}]_{\text{corr}} = [^{10}\text{Be}]_{\text{bulk}} - [^{10}\text{Be}]_{\text{dust}}$. The dust proportion was estimated from the



174 aluminum content $P = \frac{[Al]_{aerosol}}{Eff \cdot [Al]_{soil}}$, and this value was used to calculate the dust

175 component from the soil ^{10}Be : $[^{10}\text{Be}]_{dust} = P \cdot [^{10}\text{Be}]_{soil}$ (Yamagata et al., 2010):

176
$$[^{10}\text{Be}]_{corr} = [^{10}\text{Be}]_{bulk} - \frac{[Al]_{aerosol}}{Eff \cdot [Al]_{soil}} \cdot [^{10}\text{Be}]_{soil} \quad (1)$$

177 where $[Al]_{aerosol}$ is the aerosol Al concentration extracted by acid dissolution
178 ($\text{g} \cdot \text{m}^{-3}$). Eff is the acid leaching efficiency of Al in aerosols, which is 51% (Yamagata
179 et al., 2010), as determined by comparisons between leached samples analyzed by
180 ICP-AES and a large number of samples analyzed by NAA. $[Al]_{soil}$ is the Al content
181 of the topsoil (wt%); $[^{10}\text{Be}]_{soil}$ is the ^{10}Be concentration of the topsoil ($\text{atoms} \cdot \text{g}^{-1}$).

182 For $[^{10}\text{Be}]_{soil}$ in equation (1), XA, QL, and TY belong to the Loess Plateau.
183 According to our previous results, this value is $2.13 \cdot 10^8 \text{ atoms} \cdot \text{g}^{-1}$ (Zhou et al., 2007).
184 According to survey results of ^{10}Be in topsoil (Yi et al., (2019b) (Fig. S1), the
185 $[^{10}\text{Be}]_{soil}$ values of ZW, LYG, CD, and NN are $11.70 \cdot 10^8 \text{ atoms} \cdot \text{g}^{-1}$, $3.75 \cdot 10^8 \text{ atoms} \cdot \text{g}^{-1}$,
186 $4.45 \cdot 10^8 \text{ atoms} \cdot \text{g}^{-1}$, and $2.50 \cdot 10^8 \text{ atoms} \cdot \text{g}^{-1}$, respectively. The $[Al]_{soil}$ contents are 7.41
187 wt% (Xiong et al., 2010), 7.57 wt% (Qiu et al., 2014), 8.32 wt% (Tan et al., 2013),
188 and 10.50 wt% (Qiu et al., 2014) on the Loess Plateau, LYG, CD, and NN,
189 respectively.

190 Researchers have shown that removing the initial precipitation (the first
191 half-hour) for rain samples reduces the resuspended dust-borne ^{10}Be (Graham et al.,
192 2003) and allows for a straightforward estimation of ^{10}Be from wet deposition.
193 However, in this approach, the particles associated with the STE source are discarded,
194 obscuring the relationship with aerosol ^{10}Be . Therefore, to verify our dust-borne ^{10}Be



195 corrections, we compare $^{10}\text{Be}/^{7}\text{Be}$ of aerosols and precipitation.

196 **3. Results and discussion**

197 **3.1 Observations of atmospheric ^{7}Be and ^{10}Be deposited on the Chinese Loess** 198 **Plateau**

199 Measured atmospheric and rainwater ^{7}Be and ^{10}Be concentrations, and $^{10}\text{Be}/^{7}\text{Be}$
200 ratios from the XA and QL sites (May 2020 to June 2021) are presented in Fig. 1. At
201 the XA site, the average annual aerosol ^{7}Be concentration was $(3.80 \pm 0.06) \cdot 10^4$
202 $\text{atoms} \cdot \text{m}^{-3}$, and the average annual aerosol ^{10}Be concentration was $(8.09 \pm 0.13) \cdot 10^4$
203 $\text{atoms} \cdot \text{m}^{-3}$. XA rainwater ^{7}Be values averaged $(4.00 \pm 0.16) \cdot 10^4 \text{ atoms} \cdot \text{g}^{-1}$, and ^{10}Be
204 values averaged $(6.42 \pm 0.26) \cdot 10^4 \text{ atoms} \cdot \text{g}^{-1}$. At the QL site (Fig. 1b), the average
205 annual aerosol ^{7}Be concentration was $(4.08 \pm 0.07) \cdot 10^4 \text{ atoms} \cdot \text{m}^{-3}$, and the average
206 annual aerosol ^{10}Be concentration was $(6.76 \pm 0.11) \cdot 10^4 \text{ atoms} \cdot \text{m}^{-3}$. QL rainwater ^{7}Be
207 values averaged $(4.86 \pm 0.19) \cdot 10^4 \text{ atoms} \cdot \text{g}^{-1}$, and ^{10}Be values averaged $(6.35 \pm$
208 $0.24) \cdot 10^4 \text{ atoms} \cdot \text{g}^{-1}$. The average annual $^{10}\text{Be}/^{7}\text{Be}$ ratio for XA was 2.22 ± 0.12 , and
209 the average for the ZQL site was 1.62 ± 0.08 (Fig. 1c). The complete dataset is given
210 in Tables S3 and S4.

211 The XA sampling site is a typical high-dust locality on the Loess Plateau, and the
212 QL sampling site is relatively humid with lower dust content. Both sites experience
213 distinct similar seasonal changes. ^{7}Be and ^{10}Be aerosol concentrations show large
214 fluctuations associated with precipitation (Fig. 1a, 2b). The ^{10}Be concentrations and
215 $^{10}\text{Be}/^{7}\text{Be}$ ratio from dry deposition at the XA site are significantly higher than those at
216 the QL site, as well as coastal areas near the same latitude during the same period

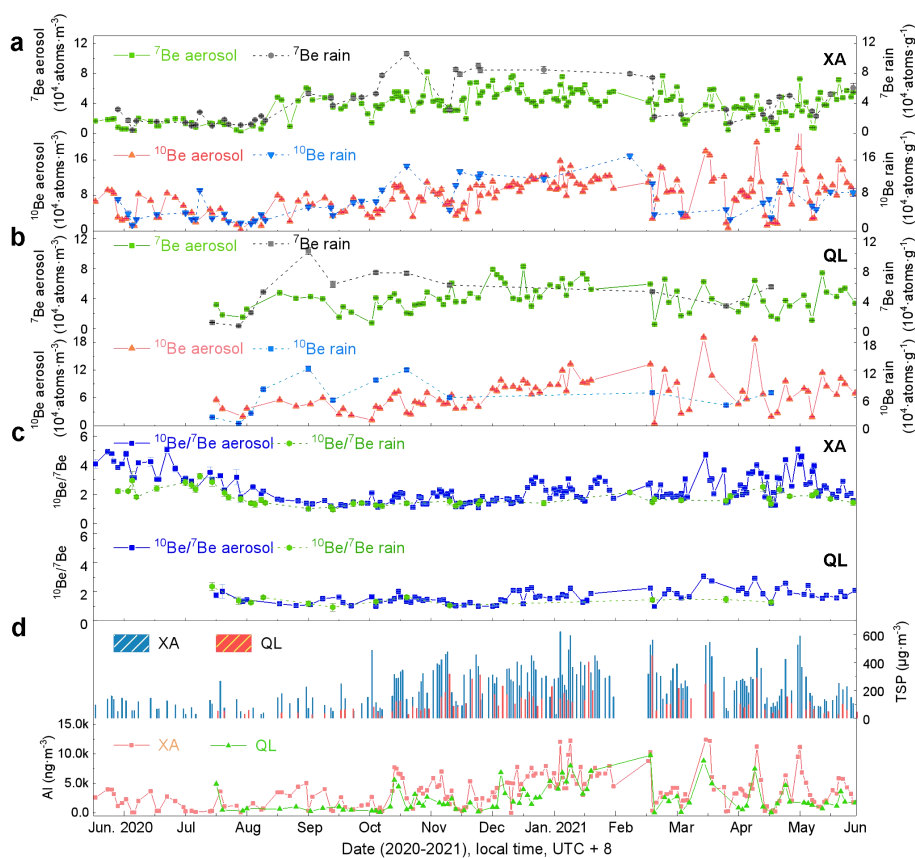


217 (such as Dazaifu and Tokyo in Japan (Yamagata et al., 2019)), especially in the dry
218 season of the Loess Plateau, such as spring. QL serves as a control site with relatively
219 higher precipitation and is located in the Qinling National Nature Reserve with a high
220 normalized vegetation index (NDVI) (He et al., 2019). QL is only 60 km away from
221 XA. As shown in Fig. 1c, the $^{10}\text{Be}/^7\text{Be}$ ratio of dry and wet deposition at the two
222 sampling sites during wet periods is consistent and reflects the amount of deposition.
223 Aerosol $^{10}\text{Be}/^7\text{Be}$ ratios at the two sites during dry periods are higher than the
224 rainwater $^{10}\text{Be}/^7\text{Be}$ from the same period. This phenomenon is more obvious at dusty
225 XA than at QL. Fig. 1d shows that the mean TSP concentration at XA ($234.20 \mu\text{g}\cdot\text{m}^{-3}$)
226 was higher than the mean at QL ($119.94 \mu\text{g}\cdot\text{m}^{-3}$). The average aerosol Al content at
227 XA ($3.6 \mu\text{g}\cdot\text{m}^{-3}$) is significantly higher than at QL ($1.5 \mu\text{g}\cdot\text{m}^{-3}$) and in coastal areas
228 (the mean in Japan is approximately $2.31 \mu\text{g}\cdot\text{m}^{-3}$ (Yamagata et al., 2010)), which
229 confirms that dust from the XA site has a higher probability of re-suspension.

230 XA is located in a warm temperate area with a semi-humid and semi-arid
231 continental monsoon climate. Its rainy seasons are unevenly distributed, and the
232 annual average precipitation is 500-750 mm (Chen et al., 2020b). The East Asian
233 winter monsoon and high-altitude westerly jets that carry dust from the western and
234 northern deserts profoundly impact the supply of aeolian materials on the Loess
235 Plateau (Shen et al., 2010). Dust-borne ^{10}Be from resuspended soil dust contributes
236 significantly to the surface aerosol ^{10}Be , and obscures atmospheric mass movement
237 information such as the STE (Fig. 1a-c). As expressed in equation (1), to obtain the
238 actual deposition flux and associated ^{10}Be and $^{10}\text{Be}/^7\text{Be}$, it is necessary to remove



239 dust-borne interference, as discussed below.



240

241 **Fig. 1** ^7Be and ^{10}Be concentrations and $^{10}\text{Be}/^7\text{Be}$ ratios in the atmosphere above the
242 Loess Plateau. Daily variation in ^7Be and ^{10}Be concentrations at XA (a) and at QL (b).
243 (c). Daily $^{10}\text{Be}/^7\text{Be}$ ratios (wet and dry deposition) of the Loess Plateau. (d). Daily
244 atmospheric TSP content and aerosol Al content at XA and QL.

245

246 3.2 Correction for dust-borne ^{10}Be on aerosol $^{10}\text{Be}/^7\text{Be}$ values

247 Due to its long half-life (1.36×10^6 years), ^{10}Be accumulates after being deposited
248 on the ground, and ^{10}Be abundances per gram of soil are approximately 10^4 times



249 higher than corresponding ^{10}Be abundances in the near-ground atmosphere. In
250 contrast, ^7Be can not accumulate in the soil due to its relatively short half-life (53.3
251 days) (Yamagata et al., 2010). Therefore, resuspended dust-borne ^{10}Be can increase
252 the atmospheric ^{10}Be concentration and $^{10}\text{Be}/^7\text{Be}$ ratios. Dust-borne ^{10}Be can be
253 corrected by considering all of the sources of ^{10}Be in the bulk measurements:

$$254 \quad [^{10}\text{Be}]_{\text{bulk}} = [^{10}\text{Be}]_{\text{bg}} + [^{10}\text{Be}]_{\text{STE}} + [^{10}\text{Be}]_{\text{dust}} \quad (2)$$

$$255 \quad [^{10}\text{Be}]_{\text{corr}} = [^{10}\text{Be}]_{\text{bg}} + [^{10}\text{Be}]_{\text{STE}} \quad (3)$$

256 The measured aerosol ^{10}Be concentration, $[^{10}\text{Be}]_{\text{bulk}}$ contains three distinct ^{10}Be
257 components (equation 2). The background value $[^{10}\text{Be}]_{\text{bg}}$ refers to the ^{10}Be
258 concentration in the surface-atmosphere at equilibrium with no STE component.
259 $[^{10}\text{Be}]_{\text{STE}}$ is the ^{10}Be carried by the STE, and $[^{10}\text{Be}]_{\text{dust}}$ is resuspended dust-borne ^{10}Be
260 (all concentrations in $\text{atoms}\cdot\text{m}^{-3}$). The dust-borne ^{10}Be must be removed to calculate
261 the corrected value $[^{10}\text{Be}]_{\text{corr}}$ (equation 3).

262 We measured atmospheric ^7Be and ^{10}Be values from different sites (Fig. 2a) and
263 different seasons (Fig. 2b), and used equation (1) to eliminate the influence of soil
264 dust on atmospheric ^{10}Be (Fig. 2c). The resuspension of terrestrial dust is controlled
265 by dryness and wind power (Zhang et al., 1997). Therefore, we chose representative
266 sampling sites with different atmospheric relative humidity and NDVI characteristics.
267 These sampling sites include ZW (h) in an arid and high-dust area, XA (b) and TY (i)
268 in semi-arid and low-humidity dusty regions, and CD (j), NN (k), LYG (l), and QL (g)
269 in humid and low-dust areas. These areas include typical geographic environments
270 such as deserts, coastal regions, humid inland regions, and areas prone to drought.



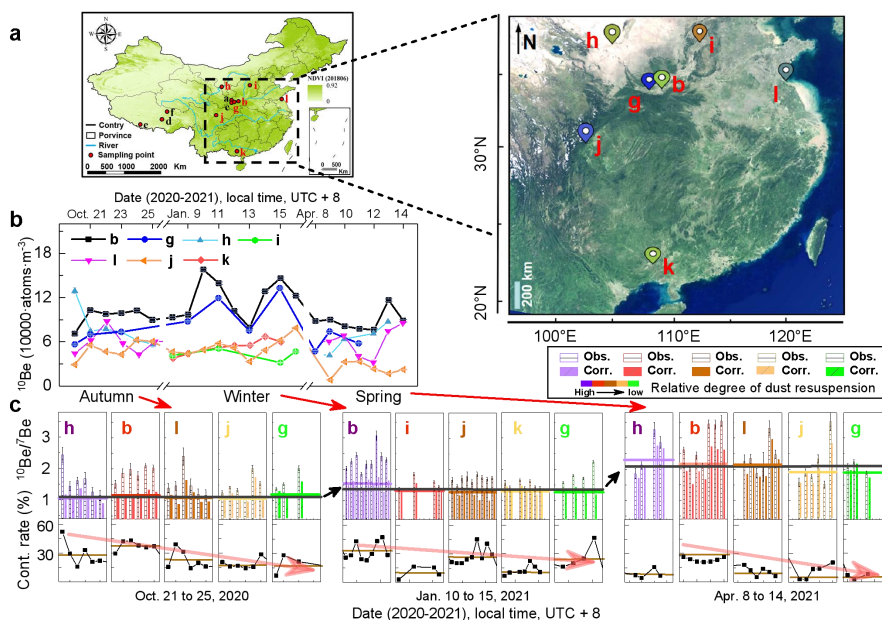
271 The ^7Be concentration in the atmosphere is minimally disturbed by dust
272 (Yamagata et al., 2010). However, as shown in Fig. 2b, in three different seasons
273 (autumn, winter, and spring), ^{10}Be concentration fluctuations caused by the influence
274 of soil dust in different regions in the same season or at different times in the same
275 region are relatively large, which leads to large fluctuations in observed atmospheric
276 $^{10}\text{Be}/^7\text{Be}$ ratios (darker columns in Fig. 2c). These results show that the drier the area
277 is, the greater the short-term fluctuation of $^{10}\text{Be}/^7\text{Be}$ (Fig. 2c). Furthermore, Chen et al.
278 (2020b) pointed out that soil ^{10}Be in areas with less precipitation ($< 1200 \text{ mm}\cdot\text{a}^{-1}$) is
279 more likely to be resuspended and not deposited in the surface soil. This conclusion
280 also shows once again that different amounts of resuspended dust can cause
281 fluctuations in observed atmospheric ^{10}Be concentrations.

282 Based on equation (1), after removing the soil dust contribution of each sampling
283 site (contamination rate shown as the black line, the bottom panel of Fig. 2c), the
284 corrected average $^{10}\text{Be}/^7\text{Be}$ ratios (colored horizontal lines, upper panel of Fig. 2c)
285 from each region are very similar. Among them, the XA $^{10}\text{Be}/^7\text{Be}$ correction ratio (b)
286 in winter (January) is larger than the average value, caused by local STE events. The
287 $^{10}\text{Be}/^7\text{Be}$ correction is relatively large, and the relative average value fluctuates
288 significantly in spring (April), because STE events frequently occur in spring. The
289 autumn-winter-spring trend in the average $^{10}\text{Be}/^7\text{Be}$ correction value is also consistent
290 with the seasonal variation of the $^{10}\text{Be}/^7\text{Be}$ ratio. Detailed ^7Be and ^{10}Be results are
291 given in Table S5.

292 To correct for dust-borne ^{10}Be concentrations in the atmosphere, the soil erosion



293 conditions for each region need to be considered. Early studies made this correction
294 based on Ca/Mg content (Brown et al., 1989; Mann et al., 2011) or U isotope
295 composition (Monaghan et al., 1986). However, these methods overestimate the
296 effects of dust and rely on assumptions that pose weak constraints on the dust
297 composition (Graham et al., 2003). Zhang et al. (1994) pointed out that the Al
298 provides an excellent means to calculate soil dust flux. Through simultaneous
299 observations in different areas, we confirmed that the Al flux method (equation 1)
300 proceeds an effective correction for atmospheric ^{10}Be . The corrected atmospheric ^7Be
301 and ^{10}Be yielded $^{10}\text{Be}/^7\text{Be}$ records that reflected daily subsidence levels as an indicator
302 of surface deposition processes. At the same time, the corrected atmospheric ^{10}Be
303 provides a practical observation method for quantifying the ^{10}Be atmospheric input
304 (Yi et al., 2019a), and to study East Asian summer monsoon rainfall changes and
305 geomagnetic field changes (Kong et al., 2020).



306

307 **Fig. 2** Atmospheric $^{10}\text{Be}/^{7}\text{Be}$ observations and the impact of dust-borne ^{10}Be from
 308 each study region. (a). Sample site map with NDVI (modis.gsfc.nasa.gov/), and a
 309 satellite image of the entire study area (from © Google Maps). (b). Short-term
 310 fluctuations in surface aerosol ^{10}Be concentration during 3 different seasons at each
 311 sampling site. (c). Observed aerosol $^{10}\text{Be}/^{7}\text{Be}$ values (darker hollow columns),
 312 corrected $^{10}\text{Be}/^{7}\text{Be}$ values (brighter solid columns), and dust contribution (lower panel)
 313 before and after correction in 3 different seasons and regions. The horizontal lines
 314 represent average $^{10}\text{Be}/^{7}\text{Be}$ values after correction and the average value contribution
 315 rate from dust-borne ^{10}Be . The color of the columns from dark to light indicates that
 316 the relative resuspension amount of dust in various places in the same season
 317 gradually decreases. The arrows indicate trends.

318 3.3 Contribution of soil dust ^{10}Be in aerosols on the Loess Plateau

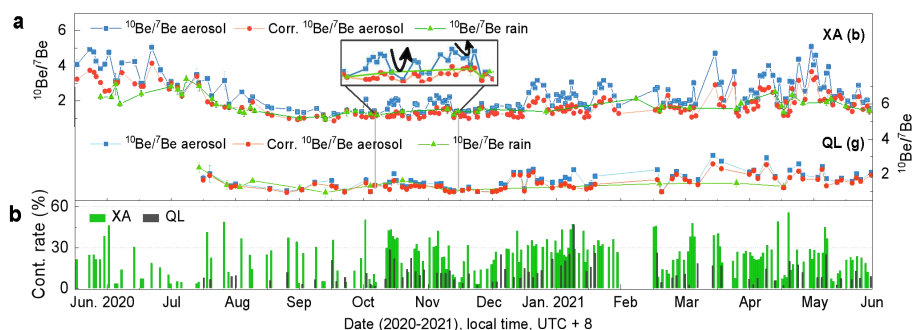
319 The observed aerosol $^{10}\text{Be}/^{7}\text{Be}$ ratios at the XA site during the dry period (blue



320 line in Fig. 3a) deviate significantly from recent rainwater $^{10}\text{Be}/^7\text{Be}$ ratios (green line
321 in Fig. 3a). In contrast, aerosol $^{10}\text{Be}/^7\text{Be}$ observations from the relatively low-dust QL
322 site are consistent with rainwater $^{10}\text{Be}/^7\text{Be}$ (Fig. 3a). This result highlights the
323 importance of correcting for resuspended dust-borne ^{10}Be .

324 The corrected aerosol $^{10}\text{Be}/^7\text{Be}$ ratios (red line in Fig. 3a) follow the rainfall
325 $^{10}\text{Be}/^7\text{Be}$ trend. Corrected aerosol $^{10}\text{Be}/^7\text{Be}$ values from XA and QL are 0.91-3.73 and
326 0.93-2.56, respectively. The average annual contributions of resuspended dust-borne
327 ^{10}Be for XA and QL were ~24% and ~12%, respectively (Fig. 3b; Table S3 and S4).
328 The contribution of soil dust from the XA site is much higher than in low dust areas,
329 such as New Zealand (11%) (Graham et al., 2003), Japan (~15%) (Yamagata et al.,
330 2010), or Seville (10%) (Padilla et al., 2019).

331 The corrected aerosol $^{10}\text{Be}/^7\text{Be}$ ratios (red line in Fig. 3a) remove abrupt transient
332 changes associated with dust-borne ^{10}Be (not STE events. This fluctuation is
333 especially apparent at the XA site and is seen as "V"-shaped changes in the
334 uncorrected record (enlarged view of Fig. 3a). The corrected $^{10}\text{Be}/^7\text{Be}$ ratios seen at
335 the XA and QL sites are very similar, as would be expected once the dust-borne ^{10}Be
336 has been removed. In addition, for atmospheric motion information that cannot be
337 captured by rain samples when there is no precipitation, these corrected dry
338 deposition $^{10}\text{Be}/^7\text{Be}$ ratios are an effective supplement.



339

340 **Fig. 3** Correction of resuspended soil dust in aerosol $^{10}\text{Be}/^{7}\text{Be}$ on the Loess Plateau.

341 (a). Comparison of aerosol $^{10}\text{Be}/^{7}\text{Be}$ observations (blue) and corrected values (red),

342 with rainwater $^{10}\text{Be}/^{7}\text{Be}$ (green). (b). The contribution of soil dust ^{10}Be to aerosol

343 $^{10}\text{Be}/^{7}\text{Be}$ at the XA site (green) and QL site (black).

344

345 3.4 Stratospheric air intrusions on the Loess Plateau and their contribution to 346 surface O_3

347 As shown in Fig. 1a and 1b, high ^7Be and ^{10}Be concentrations in winter are
348 caused by the relatively long-term accumulation of surface aerosols during periods of
349 low precipitation (Yamagata et al., 2019). However, $^{10}\text{Be}/^{7}\text{Be}$ ratio change relatively
350 little in these cases (Fig. 1c). This is because the fixed production rate ratio and the
351 similar geochemical properties make $^{10}\text{Be}/^{7}\text{Be}$ resistant to air mass dilution at
352 different latitudes and rain removal. In addition, the large difference in the half-lives
353 of ^7Be and ^{10}Be and their fixed production rate lead to stratospheric $^{10}\text{Be}/^{7}\text{Be}$ ratios
354 that are significantly larger than in the troposphere. Thus, $^{10}\text{Be}/^{7}\text{Be}$ increases seen near
355 the surface air can be used to detect stratospheric air intrusion events.

356 Fig. 4a shows a distinct seasonal pattern in $^{10}\text{Be}/^{7}\text{Be}$ ratios from the XA and QL

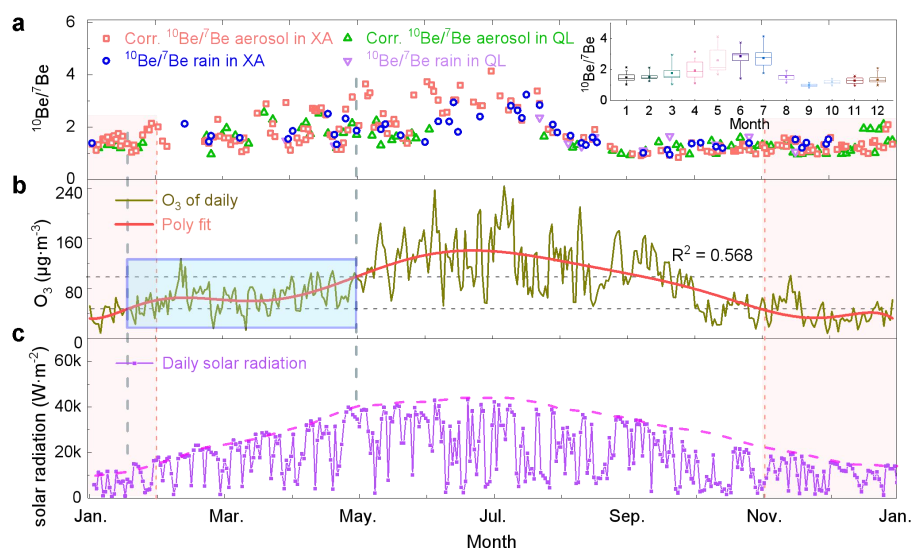


357 sites. This record is consistent with observational records (Jordan et al., 2003). Higher
358 $^{10}\text{Be}/^7\text{Be}$ ratios can indicate stratospheric air intrusion events superimposed on an
359 annual mean value of ~ 1 (Fig. 4). In spring and early summer (March to July), the
360 $^{10}\text{Be}/^7\text{Be}$ values reach a maximum of approximately 3 times higher than the annual
361 average. The increase is mainly affected by the Brewer-Dobson circulation in the
362 northern hemisphere (Butchart, 2014). The monthly $^{10}\text{Be}/^7\text{Be}$ variation box plot in the
363 upper right corner of Fig. 4a shows that the $^{10}\text{Be}/^7\text{Be}$ ratio between March and July is
364 relatively large and that the high-frequency positions are scattered. This indicates that
365 the intensity and frequency of stratospheric air intrusions during the "spring leakage"
366 of the atmosphere are higher. Conversely, the frequency and intensity of $^{10}\text{Be}/^7\text{Be}$
367 changes observed in other months are relatively scattered, indicating fewer but
368 distinct stratospheric air intrusion events during other periods of the year.

369 Sunlight is a natural prerequisite for the photochemical generation of O_3 at the
370 earth's surface (Kondratyev and Varotsos, 1996). The sunlight radiation follows a
371 symmetrical intensity law in the spring and autumn at the XA site (Fig. 4c). However,
372 surface O_3 concentrations at the XA site (Fig. 4b) are inconsistent with sunlight
373 radiation intensity (Fig. 4b blue area). This feature is consistent with the seasonal
374 increase in $^{10}\text{Be}/^7\text{Be}$. We suggest that surface O_3 concentration increases reflect an
375 influx of stratospheric O_3 , indicated by frequent and relatively large changes in the
376 $^{10}\text{Be}/^7\text{Be}$. Similar results have been confirmed during spring in other regions using
377 ground-based lidar observations (Langford et al., 2009), airborne observations
378 (Weigel et al., 2012), and atmospheric models (Zhao et al., 2021). This also explains



379 why surface O_3 concentrations are elevated in spring when the light intensity is
380 comparable to that in autumn. With the arrival of summer, the light intensity gradually
381 increases, and O_3 becomes dominated by the photochemical process at the surface
382 (Fig. 4b).



383

384 **Fig. 4** The $^{10}\text{Be}/^{7}\text{Be}$ ratio response to seasonal stratospheric air intrusions, consistent
385 with seasonal surface O_3 production on the Loess Plateau. (a). Seasonal $^{10}\text{Be}/^{7}\text{Be}$
386 variations. Inset shows a box plot (25%-75%) of monthly atmospheric $^{10}\text{Be}/^{7}\text{Be}$, with
387 the 5%-95% whiskers, and minimum/maximum values indicated by asterisks. (b).
388 Seasonal O_3 variations near the surface air at the XA site. Blue box indicates periods
389 of high surface O_3 concentrations that are higher than expected from corresponding
390 sunlight radiation intensities. (c). Seasonal sunlight radiation intensity at the XA site,
391 with associated trend line (red dashed line).

392



393 **3.5 Evaluation of stratospheric O₃ intrusion with weak STE in winter**

394 In addition to the strong "spring leakage" stratospheric air intrusion events that
395 occur regularly every year, a number of relatively low-intensity stratospheric air
396 intrusions are known to occur in other seasons. Li et al. (2021) pointed out that O₃
397 pollution during the winter haze season on the North China Plain is increasing. Ozone
398 pollution caused by winter stratospheric air intrusions is particularly serious (Chen et
399 al., 2020a). Although low-intensity and rapid stratospheric air intrusions are less
400 significant than in summer, their adverse effects cannot be ignored. Our study shows
401 that processes can be identified using ground-based ¹⁰Be/⁷Be measurements.
402 Traditional gamma spectrometry is less well-suited to obtaining high-precision ⁷Be
403 measurements of low volume air samples. Furthermore, our approach makes it
404 possible to make the daily measurements provide the requisite resolution necessary to
405 understand the rapid and transient chemical reactions in the atmosphere (Zheng et al.,
406 2011).

407 We observe frequent low but statistically significant ozone fluctuations in winter
408 on the Loess Plateau (pink shaded area, Fig. 4). A comparison with our ¹⁰Be/⁷Be data,
409 suggest low-intensity and rapid stratospheric air intrusions (November to January of
410 the following year) at the XA site (Fig. 5). In the absence of stratospheric air
411 intrusions, surface ozone shows a diurnal pattern with nighttime lows. However, when
412 ¹⁰Be/⁷Be increases (blue vertical shaded area in Fig. 5a), surface O₃ concentrations
413 increase (Fig. 5b). During these times, especially at night (shaded area in Fig. 5b),
414 surface O₃ concentrations are not observed to decrease, and diurnal ozone pattern



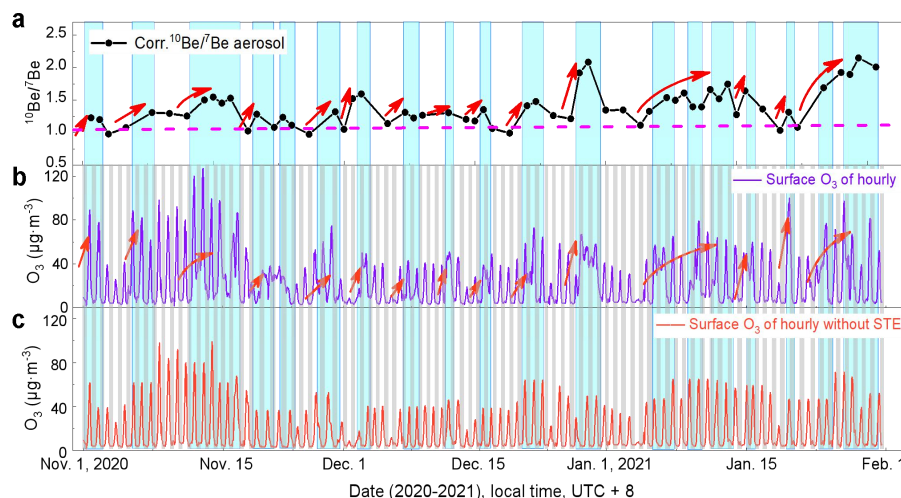
415 disappears. High O₃ concentrations are characteristics of the stratosphere (Kley et al.,
416 2007). Therefore, these phenomena may be related to stratospheric air intrusions.
417 Taken alone, the ozone trends can only be considered as indirect evidence for
418 stratospheric air intrusions. However, considered alongside the ¹⁰Be/⁷Be results at the
419 XA site, we interpret the covariation in winter to be related to rapid (1-3 days)
420 stratospheric air intrusions.

421 A fixed relationship between O₃ and ⁷Be during stratospheric air intrusions was
422 proposed by Bazhanov and Rodhe (1997). Because ¹⁰Be/⁷Be avoids the interference
423 of tropospheric variability on ⁷Be concentrations, we substitute Δ(¹⁰Be/⁷Be) for the
424 Δ⁷Be to interpret surface ozone increases (ΔO₃). The ΔO₃ and Δ(¹⁰Be/⁷Be) values
425 show a significantly positive correlation (R² = 0.889, p < 0.01. Fig. S4), providing a
426 useful estimate of stratospheric ozone to the surface.

427 A surface O₃ record obtained by the above method is shown in Fig. 5c. We
428 estimate that surface O₃ increased by 25.45% in the XA site during the winter of
429 2020/21. Chen et al. (2020a) used the hemisphere WRF-CMAQ model to find that the
430 contribution of STE to surface O₃ in the winter of 2015 in eastern China was
431 approximately 15%-21%. Stratospheric air intrusion in the Qinghai-Tibet Plateau of
432 China has been shown to increase the O₃ concentrations in the troposphere by
433 approximately 53% (Zhang et al., 2021). The influence of weather on surface O₃ in
434 China varies by region, season, and year. It may be equivalent to or even more
435 significant than the impact of changes in anthropogenic emissions, of which
436 atmospheric motion has a significant contribution (Liu and Wang, 2020). O₃ brought



437 by stratospheric air intrusion seriously affects the chemical balance of the
438 tropospheric atmosphere. For example, increased night-time O₃ (mainly input from
439 the stratosphere under STE) accelerates the conversion of NO_x to nitrate (Tang et al.,
440 2021), which has an important impact on atmospheric environmental pollution
441 process. In addition, this study proved that the relatively weak stratospheric air
442 intrusion process is randomly generated, which is an impact that needs to be
443 considered for the future surface O₃ concentration simulation.



444

445 **Fig. 5** The influence of winter stratospheric air intrusions at the XA site and surface
446 O₃ compared with daily ¹⁰Be/⁷Be measurements. Aerosol ¹⁰Be/⁷Be ratio (a) and
447 surface O₃ concentrations (b) high-frequency variations at the XA site in winter. (c).
448 Surface O₃ concentrations after removing the contribution by stratospheric air
449 intrusions (see text). The purple line segment in Fig. 5a is the baseline of the ¹⁰Be/⁷Be
450 ratio when no STE occurs. The blue shaded areas indicate stratospheric air intrusions



451 determined by $^{10}\text{Be}/^7\text{Be}$. Shaded areas in Fig. 5b and 5c indicate nighttime. Arrows
452 indicate trends.

453

454 **4 Conclusions**

455 STE is an important channel for transmitting atmospheric matters, particularly
456 low-intensity and frequent stratospheric air intrusion events that can quickly transmit
457 chemical matters to affect the environment. This work presents annual ground-based
458 $^{10}\text{Be}/^7\text{Be}$ records documenting stratospheric air intrusions. The main features of these
459 records include the following:

- 460 • High-sensitivity measurements (error $\sim 1.5\%$) of ^7Be and ^{10}Be from single
461 targets prepared from aerosols (down to $\sim 200\text{ m}^3$ air) depending on AMS.
- 462 • Results in different regions confirm that Al content can be used to correct
463 resuspended dust-borne ^{10}Be in aerosol ^{10}Be .
- 464 • By measuring both rainwater and aerosols, $^{10}\text{Be}/^7\text{Be}$ measurements offer a
465 means of identifying stratospheric air intrusions throughout the year, with daily
466 resolution. The first of such records is presented herein, from the Loess Plateau of
467 China. We document $^{10}\text{Be}/^7\text{Be}$ records for 2020-2021 from two sites: 1) XA site -
468 rainwater $^{10}\text{Be}/^7\text{Be}$ was found to be 0.94-3.24, and the corrected aerosol $^{10}\text{Be}/^7\text{Be}$
469 range was found to be 0.91-3.73 after removal of the regional dust-borne component
470 ($\sim 24\%$); 2) QL site - rainwater $^{10}\text{Be}/^7\text{Be}$ was found to be 0.94-2.36, and the corrected
471 aerosol $^{10}\text{Be}/^7\text{Be}$ range was found to be 0.93-2.56 after the removal of the regional
472 dust-borne component ($\sim 12\%$).



473 • Atmospheric $^{10}\text{Be}/^7\text{Be}$ on the Loess Plateau confirms that stratospheric air
474 intrusions occur frequently and rapidly (1-3 days) throughout the year, with the
475 strongest events in spring (March-July). These have a significant influence on surface
476 ozone. It is shown by $\Delta(^{10}\text{Be}/^7\text{Be})$ that even under weak stratospheric air intrusions in
477 winter, the cumulative contribution to surface ozone at the XA site in 2020/21 is
478 ~25%.

479 **Data availability**

480 The author declares that the main data supporting the results of this study can be
481 found in the text and its supplementary materials.

482 **Supplement**

483 Supplementary information is available for this paper.

484 **Author contribution statement**

485 Xuke Liu finished the experiments, wrote the first version manuscript and
486 analysed the data. Yunchong Fu conceived the original idea, performed AMS analysis
487 and revised the initial manuscript. Xuke Liu and Yunchong Fu designed research
488 roadmap. Li Zhang, Yanting Bi, and Guoqing Zhao designed chemical experiments
489 and collected samples. Yunchong Fu assisted in AMS method development and
490 supervised the research project. All the authors discussed the results and commented
491 on the manuscript, and George S. Burr for revising manuscript language.

492 **Competing interests**

493 The authors declare no competing interests.



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