Measurement report: Contrasting elevation-dependent changes in light absorption of <u>by</u> black and brown carbon: lessons from *in-_situ* measurements from highly polluted Sichuan Basin to pristine Tibetan Plateau

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Abstract. The scientific knowledge on light absorption of by aerosols is extremely limited at the eastern slope of the Tibetan Plateau (ESTP). We conducted the first aerosol field experiment at six sites

- 20 (Chengdu, Sanbacun, Wenchuan, Lixian, Maerkang, Hongyuan) along the ESTP ranging in elevation extending elevation from 500 m to 3500 m. The fraction of light absorption by brown carbon (BrC) to total carbon increases from 20% The light absorption of brown carbon (BrC) accounting for that of total carbon increases from 20% to 50% with altitude, and the mass absorption efficiency (MAE) of BrC over the TP is 2–3 times higher than that inside the SiChuan-Sichuan Basin (SCB), especially in winter.
- 25 In contrast, the MAE of elemental carbon (EC) in winter decreases with altitudeContrary to BrCaerosols, winter EC (elemental carbon) mass absorption efficiency declines with altitude. The contrasting variation of EC and BrC MAE with altitude is mainly attributed to source difference between the TP and SCB. Emissions from theThe more urban sources (motor vehicles, industries, etc.) inside the SCB fail to be transported to the TP due to the stable air in winter inside to winter stable air-
- 30 inside the basin, which also is also favorable for aerosol ageing to enhance absorption efficiency. The

radiative forcing of BrC relative to EC varies from 0.10 to 0.42 as altitude increases with the higher OC/EC ratio over the TP than SCB₇. Thus, the reason of the enhanced relative BrC to EC radiative forcing from polluted SCB to pristine TP is that the BrC concentration decreases more slowly than the EC concentration with altitude thus the enhanced radiative forcing of BrC relative to EC from-

polluted SCB to pristine TP is because the concentration of OC decreases more slowly with altitude than does EC. This study contributes to the understanding of the difference in light absorption by EC
 and BrC with altitude, from polluted lower-altitude basins to the pristine TP, and provides a data set for
 regional climate model validation. This study will deepen the understanding of EC or BrC light absorption difference between the highly polluted basins and clean TP and provide a basic data set for
 optimization of regional climate modeling.

1 Introduction

Some *in__situ* observations, available satellite data_ and model simulations indicated that <u>a greater</u> surface warming trend over time occurs at higher altitudes <u>infor the</u>-mountainous regions <u>worldwideall-over the world</u> (Gao et al., 2018; Guo et al., 2019; Mountain Research Initiative EDW Working Group,

- 2015; Palazzi et al., 2017; Pepin et al., 2019; Rangwala and Miller, 2012; You et al., 2020). Rangwala and Miller (2012) reviewed elevation-dependent warming (EDW) and its possible causes over four high mountain regions, i.e.; the Swiss Alps, the Colorado Rocky Mountains, the Tibetan Plateau (TP), and the Tropical Andes. Their examinations found that the available observations indicate that some mountain regions show show much greater warming rates at seasonal scales than others. The
 mechanisms that can produce enhanced warming rates at higher altitudes may be related to the
- differential sensitivities of surface warming to changes in the climate drivers, at different elevations, such as snow-ice cover, clouds, atmospheric water vapour, aerosols, land use, and vegetation, at <u>different elevations</u> (Rangwala and Miller, 2012; You et al., 2020).
- 25 <u>The Tibetan Plateau (TP, hereafter)</u>, known as <u>the "third pole", '</u> is an ideal <u>placelocation</u> to examine EDW and its mechanism (Guo et al., 2021). The warming rates (rising temperature per 10 years) over <u>the TP weare</u> found to be the most notable in winter and autumn (Liu and Chen, 2000), especially for the central and eastern <u>pPlateaus</u> (Duan and Wu, 2006), which may be partly associated with human activities, such as more anthropogenic emissions <u>inat</u> the sub-regions (Lu et al., 2010). The effect of

carbonaceous aerosols on regional and even global climate is more uncertain <u>becausedue ofto their</u> short<u>er</u> life than <u>the long-lived aerosolsones</u>, such as carbon dioxide and methane (Chung et al., 2012; Ramanathan and Carmichael, 2008). <u>AThe absorbing aerosols</u> (black carbon and dust) from local emissions or long-range transport heat the atmosphere in two ways (Tian et al., 2018). They absorb

- 5 radiation and decrease the surface albedo when deposited on snow and ice (Kang et al., 2019; Lau et al., 2010; Xu et al., 2009). Ramanathan and Carmichael (2008) suggested that black carbon (BC) in the Himalayas, arising from anthropogenic activities <u>inst the Indo-Gangetic Plain (IGP)</u>, could account for half of the local warming during the past several decades. In addition to the well-known BC, the recent work by Wu et al. (2018) suggested that the light absorption efficiency (LAE) of brown carbon (BrC, a.
- 10 certain type of organic aerosols) in winter is 2–_3 times higher than that in summer for the central Tibetan-Plateau. However, the scientific knowledge ofn the optical properties of carbonaceous aerosols (elemental carbon (EC)EC, BrC) is extremely limited over the eastern TP is extremely limited, and *in_situ* aerosol measurements at varying altitudes from the the heavily polluted Basins to the relatively clean TP awere crucial important for better understanding their light absorption.

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<u>PThe previous in -situ</u> measurements have mainly primarily focussed on the southern and northern slopes (Cong et al., 2015; Huang et al., 2007; Kang et al., 2020), whereas ile fewer observations have beenwere conducted onat the eastern slope of the TP (ESTP). The SiG huan Basin (SCB), a highly 20 eastern side of the TP (Zhao et al., 2018). The BrC LAE was strong inside the Basin (Peng et al., 2020a), especially infor the rural areas, because of due to increased more biomass and coal burning impacts (Zhao et al., 2021). Our previous studiesworks indicated that aerosols from the SCB are transported upslope along the ESTP and reach the eastern part of the TP by gradient in-situ observations at the ESTP (Yin et al., 2020). The recent study by S. Y. Zhao et al. (2020) suggested the 25 strong light-absorbing BrC from biomass and coal burning inside the Basin can be transported to the main part of the TP by the enhanced "heat pump". in response to rapid warming over the TP. The aerosols over the TP from local emissions and long-range transport from the surrounding highly polluted areas affected its weather, climate, and water cycle (C. F. Zhao et al., 2020). CThe clouds and radiation are particularly sensitive to aerosols over pristine regions (Garrett and Zhao, 2006; Zhang et 30 al., 2021). However, it is unclearfuzzy whether that the light absorption and radiative forcing of

carbonaceous aerosols change from the highly polluted SCB to the cleaner TP.

In this <u>studywork</u>, we investigated the changes in <u>the</u> light absorption of carbonaceous aerosols (EC, BrC) and calculated the relative radiative forcing of BrC to EC aerosols from the SCB to TP in the four

5 seasons. The sources and origins were also-were determined by using some statistical methods and the hybrid single-particle Lagrangian integrated trajectory (HYSPLIT) back_trajectory model. Our goals-areWe aimed to understand the difference in EC or BrC light absorption-difference between the highly polluted Basins and clean TP₁-and to reveal the corresponding mechanismscauses of the difference, and to provide-generate a basic data set for the optimiszation of regional climate modelling.

10 2 Data and methods

2.1 Observation sites and aerosol sampling

Compared <u>with-to that of the coarser fraction of particulate matter (PM)</u>, the size of strong lightabsorbing carbonaceous particles <u>wasare mainlyprimarily located</u> in the submicron range. Therefore, <u>samples PM₁-(of particulate matter with an aerodynamic diameter smaller than 1 µm (PM₁) samples</u>

- 15 were collected at six sites (Chengdu, Sanbacun, Wenchuan, Lixian, Maerkang_{*} and Hongyuan) from the western SCB to the eastern part of the TP at elevations varying with varying elevation from 500 m to 3500 m (Figure 1, Table 1). Each sampling site wais selected to represent the background level at the local scale as completely as possible_{*} without local emission impacts. A total of The 1024 PM₁ samplesin total were collected from 21 December-21, 2018 to 18 December-18, 2019 on a day/night day / night
- 20 pattern <u>using anby</u> aerosol sampler (LY-2034, Laoying Instrument Co., Ltd., China) at <u>athe</u> flow rate of 100 L min⁻¹. The samples were stored frozen in pre_baked glass jars until further analysis (Kawamura et al., 2010). <u>MThe meteorological variables</u> (temperature, relative humidity, wind speed, and direction) were downloaded <u>fromby the</u> China Meteorological Data Service Center (http://data.cma.cn/). PM₁ samples were collected near the meteorological observation sites; and thus,
- 25 the meteorological variables <u>couldean</u> represent the situation <u>inat</u> the study region. <u>The-MODIS</u> active fire data (https://earthdata.nasa.gov/active-fire-data) <u>also</u>-were used in this study.

2.2 Chemical analysis

A quarter of each filter was used to analysze water-soluble inorganic ions (Na⁺, NH₄⁺, K⁺, Ca²⁺, Mg²⁺,

F⁻, Cl⁻, SO₄²⁻, and NO₃⁻), and the ions were extracted and filtered <u>usingby</u> ultrapure water and a 0.45 μ m pore syringe filter. The concentrations of the cations and anions were measured <u>usingby</u> ion chromatography (DX-600 & ICS-2500, Dionex, USA). <u>CThe carbonaceous aerosols, that isi.e.</u>, organic carbon (OC) and <u>ECelemental carbon (EC)</u>, were <u>determined analysed usingby</u> a seven-wavelength

- 5 carbon analyszer (Model-2015, DRI, USA). The carbon analyszer measured the OC and EC concentrations using the thermal/optical reflectance (TOR) method (Chow et al., 2007). Briefly, the OC/ECOC/EC was determined by progressively heating the sub-filter. The OC fractions were determined by heating at 120 °C (OC1), 250 °C (OC2), 450 °C (OC3)_a and 550 °C (OC4) in a pure He atmosphere_a; Thesubsequently, EC fractions were measured at 550 °C (EC1), 700 °C (EC2)_a and 800 °C
- 10 (EC3) in an oxidiszing atmosphere of 2% O₂ and 98% He._-The <u>carbon</u> involved-<u>carbon</u> is oxidiszed to CO₂ and then reduced to CH₄ for detection by a flame ioniszation detector. The pyrolyszed organic carbon (OPC) was monitored when the reflected laser signal returned to its initial value after <u>the</u> <u>introduction introducing of O₂ into the analysis atmosphere</u>. The OC was defined as the sum of OC1, OC2, OC3, OC4, and OPC, whereas ile EC was defined as EC1 + EC2 + EC3 OPC. The EC and BrC
- 15 were derived from the light absorption coefficient (b_{abs}) depending on the transmittance attenuation. For the seven-wavelength carbon analyszer, the filter transmittance (FR_{λ}, fraction of light transmitted through the filter) uncertainties ranged from 5% to 18%, with the best precision <u>observedshown</u> at 450 nm and 808 nm (Chen et al., 2015). Thise uncertainty is attributed to the quality of the laser and the sensitivity of the photodiode detector for different wavelengths.

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The coefficient of variation $(CV)_{2}$ in conjunction with correlation coefficients $(r)_{2}$ can be used to characterisze the intra-location variability of chemical species (Zhao et al., 2021). The CV wais calculated usingby the folbelowing equation:

$$CV_{jk} = \sqrt{\frac{1}{p} \sum_{1}^{p} \left(\frac{x_{ij} - x_{ik}}{x_{ij} + x_{ik}}\right)^{2}},$$
(1)

25 where x_{ij} and x_{ik} are the average concentrations <u>offer a</u>-chemical component *i* at sites *j* and *k*, respectively; and *p* is the number of samples. The CV values of zero and approaching one <u>indicatemean</u> no difference and absolute heterogeneity between the two sites for the specific chemical component, respectively. <u>A</u>The CV lower than 0.2 is usually considered to represent <u>a</u> relatively similarity of spatial pattern (Wang et al., 2018).

2.3 Calculation of light absorption parameters

The BrC light absorption increases sharply as <u>the wavelength decreases</u> decreased wavelength, and thus, it can be separated from <u>the EC</u> (Peng et al., 2020a). The light absorption induced by carbonaceous aerosols (<u>the sum of EC and BrC</u>) on a quartz filter was estimated <u>usingby</u> an algorithm of transmittance attenuation (*ATN*):

$$ATN_{\lambda} = \ln\left(\frac{FT_{\lambda,a}}{FT_{\lambda,b}}\right),\tag{2}$$

where, $FT_{\lambda,a}$ and $FT_{\lambda,b}$ in the right hand-represent the filter transmittance after and before thermal analysis for the specific wavelength (λ), respectively. Referring to the work by Chen et al. (2015), the relationship betweenof ATN and withe absorption optical depth (τ_a) can be given as follows:

$$\tau_{a,\lambda} = a_{\lambda} \times ATN_{\lambda}^{2} + c_{\lambda} \times ATN_{\lambda}$$
(3)-

This study use<u>s</u> the two coefficients (a_{λ} and c_{λ}) reported by Chen et al. (2015). The light <u>absorption</u>absorption coefficients (b_{abs}) <u>werecan be</u> calculated <u>using the following with the equation</u>:

$$b_{abs,\lambda} = \tau_{a,\lambda} \times \left(\frac{A}{V}\right),\tag{4}$$

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where, *A* and *V* are <u>the</u> filter area and <u>the</u> sampling volume, respectively. The total b_{abs} can be separated into EC and BrC <u>using</u> a simplified two-component model (Chen et al., 2015):

$$b_{abs,\lambda} = b_{abs,\lambda,EC} + b_{abs,\lambda,BrC} = K_1 \times \lambda^{-AAE_{EC}} + K_2 \times \lambda^{-AAE_{BrC}},$$
(5)

where, K_1 and K_2 are fitting coefficients. AAE_{EC} and AAE_{BrC} represent the EC and BrC absorption Ångström exponents (AAE), respectively. They do not change as with wavelength. AAE_{EC} was

- 20 assumed to beas 1 (Bond, 2001), and the other three parameters in Eq. (5) were obtained for AAE_{BrC} values between 2 and 8 with anthe increment of 0.1, by least-square linear regression. T, and thhe AAE_{BrC} that led to the overall best fit in terms of R² was selected as the effective BrC AAE. The mass absorption efficiency (MAE) was obtained fromby the ratio of light absorption coefficients ($b_{abs,\lambda,EC}$ or $b_{abs,\lambda,BrC}$) to the corresponding EC or OC mass concentrations (Olson et al., 2015). The estimated
- 25 MAE_{BrC} was much lower than the true value by replacing BrC with OC_a <u>becausedue to</u> BrC account<u>sing</u> for only a small fraction of OC. The main shortcoming of the separation of <u>the</u> total aerosol absorption into EC and BrC (Eq. 5) <u>does not consideris lack of considering the</u> mineral dust

impacts. According to <u>athe</u> recent study <u>byof</u> Zhang et al. (2021), mineral dust may be <u>aimportant n-</u> important <u>atmospheric aerosol speciesspecies of the atmospheric aerosols</u> over the Tibetan Plateau. However, the study region is located <u>onat</u> the eastern slope of <u>the</u> TP-<u>during our campaign</u>, which is more easily affected by anthropogenic sources from <u>the</u> heavily polluted Sichuan-CBasin than natural

- 5 sources such as mineral dust (Yin et al., 2020) as compared to the north<u>ern</u> areas close to <u>the</u>. Taklimakan and Gobi Deserts. <u>TheOne</u> main aim of this study <u>wais</u> to reveal the gradient distributions of aerosol optical properties from the pollution <u>of the Sichuan-SCBasin</u> to <u>the</u> eastern TP_{is}, and thus, the impact of thise shortcoming may be negligible when studying the spatial heterogeneity of aerosol optical properties at <u>a</u> relatively small spatial scale. <u>In aA</u>dditionally, <u>the</u> AAE of EC <u>wais</u> assumed <u>to</u>.
- 10 <u>beas</u> 1, and the ageinging of EC <u>was not considered</u> did not take when separating the total aerosol absorption into EC and BrC (Eq. 5) in our study.

The <u>light</u> absorbed-<u>light</u> by <u>the</u> carbonaceous component can be estimated as follows (Huang et al., 2018):

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$$\frac{I_0 - I}{I_0} (\lambda, EC) = 1 - e^{-\left(MAE_{\lambda 0, EC} \times \left[\frac{\lambda_0}{\lambda}\right]^{AAE_{EC}} \times C_{EC} \times PBLH\right)}$$
(6)
$$\frac{I_0 - I}{I_0} (\lambda, BrC) = 1 - e^{-\left(MAE_{\lambda 0, BrC} \times \left[\frac{\lambda_0}{\lambda}\right]^{AAE_{BrC}} \times C_{OC} \times PBLH\right)},$$
(7)

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where, 405 nm is <u>thedetermined as</u> reference wavelength λ_0 , and C_{EC} and C_{OC} represent <u>the EC</u> and OC concentrations, respectively. The planetary boundary layer (<u>PBL</u>) height (<u>PBLH</u>) was obtained from the HYSPLIT model, and <u>we assumed no vertical gradients were assumed</u> within the PBL. Thise assumption might overestimate the radiative forcing of aerosols, while <u>it-that</u> has <u>a</u> small effect on the radiative forcing of BrC relative to EC (*f*), which can be estimated <u>usingby</u> the <u>folbelowing</u> equation (Zhao et al., 2019):

$$f = \frac{\int I_0(\lambda) \left[\frac{I_0 - I}{I_0}(\lambda, BrC) \right] d\lambda}{\int I_0(\lambda) \left[\frac{I_0 - I}{I_0}(\lambda, EC) \right] d\lambda},$$
(8)

where $I_0(\lambda)$ is <u>the</u> wavelength-dependent solar emission flux, which is <u>the</u> clear sky <u>aAir mMass 1</u> <u>gGlobal <u>hH</u>orizontal solar irradiance (Levinson et al., 2010). <u>LThe-light</u> absorption by BrC at 405 nm and 445 nm is much stronger than that at <u>the-longer wavelengths</u> inside the SCB (Zhao et al., 2021).</u>

The 405 nm <u>wavelength wasis</u> the lower limit of detection <u>ofby</u> the <u>DRI-2015</u> instrument of <u>DRI 2015</u>. Therefore, the fraction (*f*) wais obtained by numerical integration of the above formula <u>inst</u> the wavelength ranges of 405–980 nm and 405–445 nm for each sample, <u>respectively</u>. <u>NThe nighttime</u> samples were excluded when calculating the radiative forcing of <u>the BrC</u> relative to <u>the EC</u>.

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The exponential function was selected to fit the relationships between BrC MAE and altitude (AT). The equation is given as follows:

$$MAE_{\lambda,BrC} = a_{\lambda} \cdot e^{b \times AT}, \qquad (9)$$

where, a_{λ} and b are the fitted coefficients. The EC MAE can be parameteriszed with <u>the</u> altitude by replacing the subscript of *BrC* with *EC* in Eq. (9).

2.4 HYSPLIT backward trajectory model

<u>The hyHYbrid sSingle pParticle Lagrangian iIntegrated tTrajectory (HYSPLIT)</u> model developed by the National Oceanic and Atmospheric Administration's (NOAA) is a complete system for computing simple air parcel trajectories (Draxler et al., 2009). It <u>HYSPLIT iscontinues to be</u> one of the most extensively used atmospheric transport and dispersion models. A common application is a back_ trajectory trajectory analysis to determine the origin of air masses and establish source-receptor relationships. In this study, the HYSPLIT model was used to determine the potential source regions of air pollutants during the four seasons at the six sites. The 96-h backward trajectories arriving at 500 m above ground level (AGL) and initializing at each hour of the day were calculated with the 0.25°×0.25°

Global Data Assimilation System (GDAS) data from <u>the National Centers</u> for Environmental Prediction (NCEP). The gridded back_-trajectory frequencies were calculated <u>using the with</u> Openair package <u>inof</u> Rplot.

25 **2.5 PMF receptor model**

<u>The</u> EPA PMF receptor model (version÷5.0) is a mathematical approach for quantifying the contribution of sources to samples based on the<u>ir</u> composition or fingerprints of the sources. A specialisted data set can be viewed as a data matrix X of $i \times by j$ dimensions, in which *i* number of samples and *j* chemical species were measured, with uncertainties *u*. The goal of the PMF model wais

to <u>solve-determine</u> the chemical mass balance between <u>the</u> measured species concentrations and source profiles, as shown in <u>the below Eq.</u> (10), with <u>the</u> number of factors p, <u>the</u> species profile f of each source, and <u>the</u>-amount of mass g contributed by each factor to each sample:

$$x_{ij} = \sum_{k=1}^{p} g_{ik} f_{kj} + e_{ij}, \qquad (10)$$

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where e_{ij} is the residual <u>of each sample or species</u> for each <u>sample/species</u>. In this study, the uncertainties of the chemical species concentrations were estimated <u>by-using the</u> Eq. (11):

$$Unc = \sqrt{\left(0.1 \times concentration\right)^2 + \left(0.5 \times MDL\right)^2}, \qquad (11)$$

where MDL is <u>the</u> species-specific method detection limit. The water-soluble ions and carbonaceous aerosols in the four seasons at the six sites were used as input variables to run<u>the</u> PMF model. The MDL of the species can refer to Cui et al. (2019).

3 Results and discussion

3.1 Light absorption of EC and BrC

Table_2 summarizes the seasonally mean OC and EC concentrations, light absorption coefficients, and efficiencies (b_{abs} , MAE) of EC and BrC at 405 nm, and the meteorological variables at the six sites

- 15 during the campaign. The average winter EC concentration ranges from 2.2 μg m⁻³ at Maerkang to 7.9 μg m⁻³ at Sanbacun, which is <u>a bout</u> 2–6 times higher than th<u>atose</u> in the other seasons in response to more primary emissions in winter with similar wind speeds (Table_2). The much higher OC/EC ratios at the Plateau sites than those that at the Basin sites suggests that more secondary OC is formed by chemical reactions over the TP, which has been corroborated by can be supported by the works of Wu et
- 20 al. (2018). <u>HThe higher OC/EC ratios with increasing the altitude can also result from stronger EC emissions at lower altitudes. Combined with those obtained from the previous studies, the winter OC concentration iwas found to vary from 15.0 to 20.1 μg m⁻³, whereas ile-EC is ranged between 4.3 and 4.7 μg m⁻³ at urban areas inside the SCB, which wais substantially significantly lower than that at the Indo-Gangetic-Plain (Table S1). However, OC and EC concentrations inat the eastern TP weare much</u>
- 25 more abundant than th<u>oseat inat the</u> western and southern TP sites <u>becausedue ofto themore</u> denser population and <u>widespread</u> industrial activitiesy (Table S1). BrieflyThus, carbonaceous aerosol pollution is much more severe inside the Basin than that over the TP, indicating that the large amounts

of air pollutants areas trapped inside the deep Basin because of due to calm and stable air.

Figure 2 compares <u>the</u> spectral total and separated EC and BrC b_{abs} in spring and winter at <u>the</u>-six sites along the ESTP. The measured (green hollow points) and calculated b_{abs} (yellow dashed lines) for total

- 5 carbon ((TC₂₇-sum of EC and BrC) wereis comparable, and the difference wais within 5%. For Sanbacun, a rural site inside the Basin, the b_{abs} is much higher than the other sites, especially for theshorter wavelengths, because of due to more BrC emissions from coal and biomass burning for cooking and heating inst rural areas inside the SCB (Zhao et al., 2021). The light absorption of EC aerosols decreaseds with altitude, primarily because of the decreaslined EC concentration (see Table 2). Thise
- phenomenon may be partly <u>due to caused by the stable air inside the deep Basin (Feng et al., 2020)</u>;⁵
 <u>but however, itthat would also apply to BrC asin so</u> far as EC and BrC share sources, and vertical mixing is primarily due to fair weather convection rather than deep convective storms (Zhang et al., 2017). However, the light absorption by BrC does not monotonically change <u>withas</u> altitude <u>becausedue ofto the more complicated sources and origins of BrC. The 405 nm b_{abs} of BrC accounting for that of TC increaseds from 20% at Chengdu to ~ 50% at Hongyuan, whereas ile-the proportion
 </u>
- significantly reduce<u>ds</u> with increas<u>inged</u> wavelength (Figure S1), suggesting that light absorption of BrC aerosols is much stronger at high altitudes than that at lowlands.

Compared towith b_{abs}, MAE can better reflect the LAElight absorption efficiency of aerosols. The
average winter MAE_{EC} is 6.0±1.0 m² g⁻¹ among all sites, which is within the range of 3.9–11.9 m² g⁻¹ over the TP and the the surrounding basins (Tables 2 and S1). Except our result in the rural site, the mean winter MAE_{BrC} of 0.7–0.8 m² g⁻¹ inside the SCB is approximately bout half of that at the Indo-Gangetic Plain (IGP) probably because of due to the differences in BrC emissions, PM size distribution and chemical composition between the SCB and IGP (Choudhary et al., 2018). Figures 3 and S2 show
box plots of spectral MAE_{BrC} and MAE_{EC} in the four seasons from the Basin to Plateau sites, extending elevation from 0.5 to 3.5 km. In contrast toDifferent from EC, MAE_{BrC} at 405 nm over the TP wais 2–3 times higher than that inside the SCB with strong elevation-dependent light absorptionlight absorbing, and the only clear dependence wais in winter. Wu et al. (2018) found that winter MAE_{BrC} wais 4.5 m² g⁻¹ for a pristine environment over the TP (Nam Co, 4730 m asl), which is significantly higher than that

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at Hongyuan (3500 asl) infor our study. The average winter OC/EC ratio of 14.1 at Nam Co wais

<u>significantlylargely</u> higher than that at our sampling sites. Therefore, the clearly increased winter MAE_{BrC} with altitude may be related to BrC composition seasonally, <u>whereaswhile</u> winter MAE_{EC} decreases with altitude, possibly <u>becausedue ofto</u> the difference in source composition and ageinging aerosols inside the deep Basin (Liu et al., 2020). Th<u>ise</u> mechanism <u>iswill be</u> discussed in the following

5 sections.

Figure 4 shows MAE_{BrC} and MAE_{EC} variations as altitude in spring and winter during the campaign. The relationships between average MAE and altitude of the measurement sites were fitted by exponential function, and coefficients of determination (R^2) were given in the figure. R^2 reflects the

strength of the relationships between two parameters. The contrasting MAE variation as altitude between BrC and EC in winter (R² of 0.89 for MAE_{BrC} and 0.86 for MAE_{EC}) is more significant than that ose-in spring (R² of 0.45 for MAE_{BrC} and 0.06 for MAE_{EC}). The better relationships in winter may be because more urban and aged aerosols are trapped inside the deep Basin in response to strong winter temperature inversion (Feng et al., 2020). The relation of MAE_{BrC} or MAE_{EC} at 405 nm with altitude
can be parameteriszed with exponential function (Eq. 9). The spring and winter MAE_{BrC} can be

parameterized with altitude (AT) as follows:

$$MAE_{405,BrC,spr} = 1.33 \cdot e^{0.18 \cdot AT}$$
(12)

$$MAE_{405,BrC,win} = 0.82 \cdot e^{0.33 \cdot AT}$$
(13)

Similarly, the winter MAE_{EC} can be parameteriszed by altitude (AT) as follows:

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$$MAE_{405,EC,win} = 11.35 \cdot e^{-0.18 \cdot AT}$$
 (14).

3.2 Sources impacting on-the light absorption of EC and BrC

<u>The OC/EC</u> ratio can be used to <u>approximaterough</u>ly identify sources of carbonaceous aerosols, and the ratio of aerosols from fossil <u>fuel</u> combustion is generally lower than that <u>fromof</u> biomass burning

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(Bond et al., 2004). Figure S3 shows the relationship between OC and EC concentrations inside the SCB and that over the TP during the campaign, and the OC/EC ratio was obtained by fitting the relationships with univariate linear regression. The significantly simultaneous change between OC and EC ($R^2 = 0.80$ for SCB₇ and $R^2 = 0.75$ for TP) indicated that the sources may be similar. The OC/EC ratios of 2.14 for western SCB and 2.06 for eastern TP are significantly lower than those at at Nam Co

(13.8–14.1, Wu et al., 2018) representing a pristine environment over central TP (Cong et al., 2009), wh<u>ereas ile</u> the ratios are much higher than th<u>oseat at Lhasa</u>, the largest city o<u>f the ver</u>-TP (1.46, Li et al., 2016). The OC/EC ratios for in our study <u>are is</u> slightly lower than th<u>ose at at for</u> urban areas in eastern China and Helsinki in Finland (Han et al., 2014; Viidanoja et al., 2002), indicating that carbonaceous aerosols <u>inet</u> western SCB and eastern TP may be significantly affected by primary sources.

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In addition to Besides primary sources, secondary formation largely contributes to OC aerosols₂₅ and thus, secondary organic carbon (SOC) was calculated <u>using the with EC-tracer method</u> (Turpin and Lim, 2001). To better understand <u>the</u> light absorption of primary <u>organic carbon (POC)</u> and <u>secondary</u> <u>SOC</u>, Figures S4 and 5 show sample-to-sample and average MAE_{BrC} variations as SOC and POC concentrations for each site in spring and winter during the campaign, respectively. The <u>LAE light</u>absorption efficiency of BrC significantly declineds as the <u>increased</u>-OC composition <u>increased</u> with the better relationships for POC at each site (Figure S4). The average winter MAE_{BrC} decreased by

approximatelyabout 70% as POC increasesd from 3.0 μg m⁻³ at Hongyuan to more higher than 20 μg m⁻³ at Chengdu (Figure 5). SOC accounting for OC significantly increaseds from the western SCB to the eastern TP, and it wasis higher than 50% at Maerkang and Hongyuan because of due to relatively fewer primary sources over the TP. The large winter MAE_{BrC} increment inas theas the SOC/POC ratio indicates that the more SOC and the fewer POC are is a favourable condition for BrC light absorption
enhancement (Figure 5). Therefore, the strong elevation-dependent MAE_{BrC} in winter (Figure 4) may

be induced by SOC/POC ratio variations from the western SCB to the eastern TP.

The EC LAElight absorption efficiency largely decreased as the reduces as EC concentrations increased at each site (Figure S5). However, the average winter MAE_{EC} inside the highly polluted SCB **25 is**-was much higher than that over the clean TP, wh<u>ereasile</u> for similar EC concentrations among the Plateau sites, <u>the MAE_{EC} at Wenchuan is-was approximatelyabout two</u>² times higher than that at Hongyuan, with <u>a</u> strong dependence on elevation. Therefore, winter aerosol ag<u>einging</u> inside the deep Basin and large source differences may induce light absorption reduction from the-western SCB to <u>the</u> eastern TP. The increase <u>inof MAE_{EC}</u> as the ratios of water_-soluble ions (K⁺, Cl⁻, SO₄²⁻, and NO₃⁻) to

30 EC concentrations <u>aton</u> different levels suggests that EC light absorption <u>was affected is certainly</u>

impacted by many anthropogenic sources at the six sites (Figure S6). A specific inorganic component can be considered as an the indicator of a the specific emission sources. K⁺ and Cl⁻ ions are usually used to characterise for characterizing biomass burning (BB) and coal combustion (CC), respectively (Tao et al., 2016). NO₃⁻ and SO₄²⁻ can reflect motor vehicle and industry source impacts, respectively.

- 5 Therefore, to further <u>identify the find-key</u> sources impacting MAE_{EC} , we <u>checked-verified</u> the spring and winter mean MAE_{EC} variations as the concentrations of K⁺, Cl⁻, NO_{3⁻2} and SO_{4²⁻} ions at the six sites (Figure 6). Compared <u>with-to</u> the spring value, the winter MAE_{EC} <u>was is the lower owingdue</u> to high EC concentrations and <u>was is</u>-more sensitive to <u>the-the</u> chemical species from anthropogenic emissions. Furthermore, <u>the NO₃⁻</u> difference among the sites (Figure 6a) <u>wais</u> much larger than <u>that of</u>
- 10 K^+ , Cl^-_3 and SO_4^{2-} because of due to the combustion of fossil fuelsmany fossil fuel combustion at the Chengyu City Clusters inside the Basin. The spatial heterogeneity in the $(NO_3^- + SO_4^{2-})$ -/- $(K^+ + Cl^- + NO_3^- + SO_4^{2-})$ ratio in winter wais more significant than that in spring, and winter MAE_{EC} evident obvious ly increaseds as the ratio from the TP to the Basin sites. Therefore, the emissions from fossil fuel combustion may be a key source-influence on ing-winter MAE_{EC}.

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The above paragraphs separately <u>analyseanalyzed the LAE light absorption efficiency</u> of BrC and EC and their variations as chemical species, <u>while and the change in radiative forcing of BrC relative to EC</u> (*f*) from Chengdu to Hongyuan is show<u>ned in Figure 7a to reveal the mechanism. <u>PThe parameter (*f*)</u> reflects <u>the light absorption strength of BrC at the shorter wavelengths as compared to that of EC</u> aerosols at <u>all the whole</u>-wavelengths. The much higher *f* values indicated that <u>the radiative forcing of</u> BrC aerosols is much stronger for <u>the similar EC radiate forcing₂₇ and thus</u> this parameter can be used to better understand the radiative forcing of secondary aerosols relative to primary aerosols at a specific location. The altitude (*AT*) increased by 3 km, <u>andwhile</u> the median *f* increase<u>d</u>s from <u>approximatelyabout</u> 0.10 inside the Basin to 0.42 over <u>the eastern TP</u>. The relationship between *f* and altitude can be parameteriszed as follows<u>the below equation</u>:</u>

$$f = 0.077 \cdot e^{0.480 \cdot AT} \tag{15}$$

Some <u>A few</u> studies <u>have</u> found that the direct radiative forcing of BrC-/-(BrC+EC) increases with altitude <u>simply becausedue to the fact that</u> the concentration of BrC decreases more slowly with altitude <u>than-than that of does-</u>EC (Liu et al., 2014, 2015; Zeng et al., 2020; Zhang et al., 2017).

Therefore, we also checked the median OC/EC ratio variations from the Basin to <u>the-</u>Plateau sites during the campaign (Figure 7b). The OC/EC ratio change<u>d</u>s within the range of 2–4, and the 75th percentiles of the ratio increase<u>d</u>s more significantly than the median values from the Basin to Plateau sites. Therefore, the increased *f* from <u>the-</u>western SCB to <u>the-</u>eastern TP may be closely related to more secondary formation and fewer primary emissions over the TP than <u>over the SCB</u> (also see Figure 5c).

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As previously mentioned, <u>the</u> MAE of carbonaceous aerosols largely depends on emission sources. <u>The</u> PMF receptor model is widely used to apportion the sources influencing air pollutants at a specific site based on the fingerprints of the sources $\frac{1}{27}$ for example, K⁺ and Cl⁻ are usually used as tracers for

- 10 <u>BBbiomass burning (BB)</u> and <u>CCeoal combustion (CC)</u>, respectively (Tao et al., 2016). <u>The PMF</u> analysis was conducted in this study for each season. <u>MThe motor vehicles</u>, biomass and coal burning, dust, sea salt, and secondary formations <u>weare</u> found to be the main sources at the six sites. Figure 8 shows mass concentrations of species for each source at each site apportioned <u>using theby</u> PMF model in winter during the campaign. The PMF results for the other seasons are <u>shownillustrated</u> in Figures
- 15 S7–S9. The winter NO_3^- concentrations for secondary nitrate decreased from 3.44 µg m⁻³ inat Sanbacun to 0.07 µg m⁻³ inat Maerkang, which is more heterogeneous than that in summer and fall. As a main source inside the SCB, the winter secondary nitrate is in response to the intensive mixing between themotor vehicle emissions and other primary pollutants trapped inside the Basin by strong capping inversion (Feng et al., 2020). Additionally, high humidity inside the SCB facilities the secondary nitrate
- formation, and the average nitrogen oxidation ratio in Sichuan (average RH = 80%) is 3.1 times of that in winter in Beijing (average RH = 27%) (Wang et al., 2021). EC aerosols from the intensive human activities inside the SCB are easily aged by coating the secondary_formed nitrate in winter, which further causes the enhancement of thes Basin EC light absorption. The latest studies study byof Zhang et al. (2022) found that the light absorption and radiative forcing of BCblack carbon coated withby
 inorganic salts are much stronger than of that inside organic materials. The chemical species (K⁺, Cl⁻) from BBbiomass burning and CCeoal combustion declined from the Basin to Plateau sites; however, but the declining ranges in the warm seasons (summer and fall) were are more significant than those in the cold seasons (spring and winter) because of the usedue to usage of more fuel for heating over the
- TP. Therefore, the primary BrC from <u>BB</u>biomass burning and <u>CC</u>coal combustion for winter heating
 over the TP may partly contribute to the strong TP BrC light absorption.

3.3 Impacts of regional and long-range transport on the light absorption of aerosols

<u>FThe fresh aerosol particles are gradually aged by mixing with other pollutants during the long-range</u> transport, and the <u>rebyn</u> enhancinge their light absorption and radiative forcing. The similarities of <u>the</u>

- 5 major chemical species between <u>the</u> two sites should represent regional air pollution, wh<u>ereas ile</u> the differences should reflect local source impacts. The comparisons between Basin (Chengdu, Sanbacun) and Plateau sites (Wenchuan, Lixian, Maerkang, Hongyuan) <u>related to about</u> the average mass concentrations of water-soluble ions and carbonaceous species in the four seasons are <u>showed shown</u> in Figures 9 and S10–S12. The numerical ranges between the two axes of each subplot <u>weare</u> set to be
- 10 equal to more clearly <u>observe the</u> spatial heterogeneity of the chemical species <u>inst</u> the region. The combination of <u>the</u> CV <u>andwith</u> correlation coefficients can be used to better understand intra-location variability (Wilson et al., 2005). The CV <u>wais</u> between 0 and 1 (see-Eq. 1), <u>whereand athe</u> smaller value representeds athe more uniform particle concentrations. <u>MThe moderate differences and</u>, relatively high CV values (0.22–0.75) <u>were</u>, are observed for the chemical species from anthropogenic
- 15 sources (NH4⁺, K⁺, SO4²⁻, NO3⁻, F⁻, Cl⁻, OC, and EC) in the four seasons. The differences indicated that there weare limited similarities between the Basin and Plateau sites, and the discrepancies were in major anthropogenic sources. The spatial heterogeneity offor K^+ and NO₃⁻ is more obvious evident than that of the other species in the four seasons, which is mainlyprimarily related to more BBbiomassburning and vehicle emissions inside the SCB (Zhao et al., 2021). The weak inter-regional transport 20 between the western SCB and eastern TP suggested that the light absorption of carbonaceous aerosols over the TP is rarely influenced by pollutants from the SCB. Furthermore, the CV values for K⁺, NO₃⁻, and EC in winter weare the lowest among the seasons becausedue ofto increased BBbiomass burning and <u>CCcoal combustion</u> for winter heating over the Tibetan Plateau. Unlike CV, a high correlation coefficient for athe specific chemical component does not necessarily indicate uniformity, which may 25 suggest source similarity between sites. The correlation largely dependeds on the season (Figures 9 and S10–S12). The strong correlations for NH4⁺, K⁺, SO4²⁻, NO3⁻, OC, and EC in the-spring and winter implied vinfer that Basin and Plateau sites shared similar sources for the species, whereas while weak correlations for NO₃⁻, OC, and EC in summer and fall indicated impacts of dissimilar sources impacts between the Sichuan SCBasin and Tibetan Plateau.

Compared towith the species from anthropogenic sources, the lowest CV values for Na⁺, Mg²⁺, and Ca^{2+} among the species indicated that they weare more comparable between the Basin and Plateau sites. Furthermore, changes in Na⁺ values weare more synchronous than those of Mg²⁺ and Ca²⁺ in the summer and fall. Na⁺ concentrations were is found to be high in salt-rich dust from saline soils (Quick and Chadwick, 2011). Dust events frequently occurred in spring and winter over the Tibetan PlateauP and northwest China, where saline and alkaline land and dried salt-lakes are located (Jiang et al.,

2021; Zhang et al., 2009; Zhang et al., 2021), and T thus, the weak correlations for Na⁺, Mg²⁺, and Ca²⁺ values in spring and winter may suggest local and regional dust plume impacts. Therefore, the

lack of considering mineral dust impacts in separation of BrC light absorption from total aerosol

absorption (Eq. 5) might cause some errors. The errors should be much smaller as than those

of compared to the studies onat the north or northeast TP close to the Taklimakan and Gobi Deserts.

MODIS active fire data suggests that <u>BBbiomass burning</u> is <u>mainlyprimarily loconduc</u>ated in South Asia around our study regions, which wais more abundant-frequent in cold than warm seasons during

- 15 our campaign (Figure S13). The PM mass concentrations in conjunction with wind data can be used to identify the local PM origins, and Figure 10 shows that K^+ pollution increasedrose in the four seasons at the six sites. The back--trajectory calculation can providegive PM origins from long-range transport. and Figures 11 and S14-S16 illustrate the gridded back-trajectory frequencies in the four seasons. K⁺ stratification in warm seasons is was more obvious evident than that in cold seasons, which
- 20 implyingsinfers that there were are substantial more BB biomass burning plumes over the Tibetan PlateauP in spring and winter. The change in wind direction is not obvious from the Sichuan SCBasin to the TPibetan Plateau during warm seasons is not evident. However, the predominant wind direction is northwest-southeast in cold seasons for the Basin sites, whereaswhile itthat mainlyprimarily focusses on the southwest for the Plateau sites (Figure 10). The highest frequency of the back trajectory was also
- 25 is in the southwest of the sampling sites in winter (Figure 11). Therefore, the BBbiomass burning emissions originatinged from South Asia are transported to the eastern Tibetan PlateauP by highly frequent southwesterly winds, and thus inducing high K^+ concentrations in spring and winter. The BrC aerosols from the intensive BBbiomass burning in South Asia are gradually aged by internal or external mixing with the other anthropogenic emissions during the long-range transport. The light absorption of the aged BrC aerosols over the TP is enhanced by coating withe inorganic components (Zhang et al.,
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2022), which may partly contribute to the stronger BrC light absorption at the Plateau sites than that at the Basin sites. Unlike the eastern TP, the carbonaceous aerosols in the western SCB are regionally transported from the central and eastern SCB, <u>aswhich</u> can be seen from the increase in pollution rose and back trajectories. The aerosols accumulate and stagnate at the front areas of the mountains <u>because</u> of <u>due to the</u> terrain block_i; and thus, the light absorption of <u>by</u> EC aerosols emitted from motor vehicles is enhanced by the intensive mixing among the air pollutants.

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4 Summary and conclusions

Tibetan Plateau (TP)-is surrounded by the three highly polluted regions, i.e., Indo-Gangetic Plain (IGP), Taklimakan and Gobi Deserts (TGDs) and SiChuan Basin (SCB). However, the previous studies have

10 <u>mainly-primarily</u> focussed on the south (IGP) and north slopes (TGDs)₇ and thus t<u>T</u>he first *in_-situ* aerosol measurements were conducted at on the eastern slope of Tibetan Plateau (ESTP) to study the elevation-dependent light absorption of by carbonaceous aerosols from the highly polluted SCB to the pristine TP. The source and origin impacts on light absorption of by aerosols also were also discussed by combining with PMF and HYSPLIT results.

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The EC and BrC light absorptions werewas separated usingby athe simple two-component model. The BrC light absorption coefficient at 405 nm_a accounting for that of total earbon (TC; (sum of EC and BrC), was-is found to increase from ~ 20% inside the SCB to ~ 50% over the TP. The BrC mass-absorption efficiency (MAE) over the eastern TP wais 2–3 times higher than that inside the SCB_a with
strong elevation-dependent absorption. The most significant elevation-dependent winter MAE_{BrC} was is closely related to the high ratio of SOC to POC secondary to primary organic carbon (OC), that isi.e., more OC from secondary formation than those from primary emissions at high altitudes. In contrast. toDifferent from BrC, winter MAE_{EC} declinesd from the highly polluted SCB to the clean TP_x which is because of due to-source differences between the two regions. Substantial More-urban sources-

25 <u>emissions</u> (vehicles, industries, etc.) are-were trapped inside the deep SCB <u>owingdue</u> to poor dispersion and frequent temperature inversion <u>during</u> cold seasons. <u>HThe high</u> primary emissions and weak dispersion conditions <u>were are favourable</u> for mixing and aerosol ageinging to enhance <u>the light</u> absorption inside the Basin. The median radiative forcing of BrC relative to EC increases<u>d</u> from <u>approximatelyabout</u> 0.10 inside the Basin to 0.42 over <u>the eastern TP</u>, which <u>was is</u>-associated with the

OC/EC ratio. Therefore, the enhanced radiative forcing of BrC relative to EC <u>occurredsis</u> because <u>OC</u> the concentration of <u>OC</u>-decreasesd more slowly with altitude than the than does EC concentration.

The first aerosol field experiment was conducted <u>inet athe</u> specific study region<u>; however</u>, <u>but</u> only six sampling sites <u>were used</u> from the deep SCB to <u>the</u>-eastern TP <u>were used</u> in this study. <u>MThe more</u> measurement sites <u>shouldwill</u> be established to better understand the chemical composition and light properties of aerosols <u>inet athe</u> unique region. The light absorption coefficients and efficiencies of BrC <u>could not failed to</u> be separated from th<u>oseat</u> of TC in summer and fall at Maerkang and Hongyuan <u>because of due to</u>-instrument failure, which limited t<u>he revelation o revealof</u> the elevation-dependent

10 light absorption. Furthermore, replacing BrC, OC mass concentration was used to estimate MAE_{BrC} , which may cause <u>significant large</u> uncertainty; , and thus these are expected to be corrected in the future studiesy.

Data availability. Raw data sets (Zhao et al., 2022, DOI: 10.5281/zenodo.6474199) used in this manuscript were available at https://zenodo.org/record/6474199#.YmCn_YtByUk.

Author contributions. Suping Zhao and Ye Yu designed the study. Suping Zhao analyzed the data with help from Ye Yu, Jinbei Chen and Shichang Kang. Daiying Yin and Longxiang Dong collected and analyzed data during the campaign. Shaofeng Qi conducted the field experiment.

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Competing interests. The authors declare that they have no conflict of interest.

Financial support. This work was supported by the National Natural Science Foundation of China (42075185; 41605103), Youth Innovation Promotion Association, CAS (Y2021111), and Gansu Science and Technology Program key projects (20JR10RA037 and 18JR2RA005).

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Figure 1: Geographic location of the six *in_-situ* measurement sites (Chengdu, Sanbacun,
Wenchuan, Lixian, Maerkang, and Hongyuan) along the eastern slope of Tibetan PlateauESTP.
The map is a pure reproduction of Google Maps with added a marks for our study locations.

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Figure 2: Spectral light absorption coefficients (b_{abs}) by EC and BrC in spring and winter at the six sites along the ESTP. The subplots depict the decomposition of total light absorption by EC and BrC with the model given in Eq. 4. Error bars represent uncertainties derived from replicate

5 analyses and lower quantifiable limits.



Figure 3: Box plots of spectral $\frac{BrC}{mass}$ mass absorption efficiency <u>of BrC</u> (MAE_{BrC}) in each season from Chengdu inside the SCB to Hongyuan over the TP <u>extending-ranging in</u> elevation from 500 m to 3500 m. The lines inside the boxes denote the median values, and the two whiskers and the

top and bottom of the boxes denote the 5th and 95th and the 75th and 25th percentiles.

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Figure 4: Variations of (a) MAE_{BrC} and (b) MAE_{EC} at 405 nm as altitude in spring and winter during the <u>field</u> campaign. The solid dots denote the median values, and the two whiskers of the dots denote the 25th and 75th percentiles. The relationships between averaged MAE and altitude of the measurement sites were fitted by exponential function, and the coefficients of determination (R^2) also were given in each subplot. The relationships (R^2 with the superscript of an asterisk) passed the significance level of 0.01.



Figure 5: Variations of spring and winter mean-averaged MAE_{BrC} in spring and winter as (a) SOC, (b) POC, (c) SOC/POC ratio and (d) MAE_{EC} as EC concentrations at the six sites. The hollow and solid dots denote the median values in spring and winter, and the four whiskers of the

5 dots denote the 25th and 75th percentiles of the corresponding two variables. The horizontal axis in each subplot is showed on a logarithmic scale to more clearly see the details.



Figure 6: Variations of spring and winter MAE_{EC} in spring and winter as (a) NO_3^- , (b) SO_4^{2-} , (c) K^+ , and (d) CI^- concentrations, and (e) $(NO_3^- + SO_4^{2-}) / (K^+ + CI^- + NO_3^- + SO_4^{2-})$ ratio at the six sites. The hollow and solid dots denote the median values in spring and winter, and the four whiskers of the dots denote the 25th and 75th percentiles of the corresponding two variables. The axes in each subplot are showed on a logarithmic scale to more clearly see the details.



Figure 7: Variation of radiative forcing of BrC relative to EC (f, see Eq. 8) as (a) altitude and (b) OC/EC ratio for each site. The solid dots denote the median values, and the two whiskers of the dots denote the 25th and 75th percentiles of the variables.



Figure 8: Mass concentrations of species for each source at each site apportioned by PMF model in winter during the campaign. The vertical axes are showed on logarithmic scale to better distinguish the concentrations of chemical species among the measurement sites.



Concentrations for basin sites

Figure 9: Relationships of spring PM₁ chemical components concentrations <u>in spring</u> between basin (horizontal axes, including Chengdu and Sanbacun) and plateau sites (vertical axes, including Wenchuan, Lixian, Maerkang and Hongyuan). The correlation coefficients (r) with an

asterisk and two asterisk superscripts passed the significance level of 0.05 and 0.01, respectively.



Figure 10: K⁺ pollution rose in the four seasons at the six sites along the ESTP. Mean K⁺ concentrations and calm frequencies also were given in each subplot.



Figure 11: Gridded back trajectory frequencies with hexagonal binning in winter at the six sites from west Sichuan Basin to Tibetan Plateau. The map is a pure reproduction of Google Maps with added the trajectory frequencies. Copyright © Google Maps.

Name	Latitude (degree)	Longitude (degree)	Altitude (km)
Chengdu	30.67	104.06	0.50
Sanbacun	30.99	103.66	0.65
Wenchuan	31.46	103.61	1.33
Lixian	31.42	103.16	1.89
Maerkang	31.92	102.22	2.62
Hongyuan	32.79	102.55	3.50

Table 1 Summary of the measurement sites (name, location and altitude).

Table2 Seasonally averaged values (mean ± std.) of OC and EC concentrations, light absorption coefficient (b_{abs}), mass absorption efficiency (MAE) and meteorological variables (wind speed (WS), temperature (Tem.), relative humidity (RH)) at the six sites during the measurement campaign. There is no babs or MAE reported for MEK and HY in summer and fall as the used

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DRI instrument does not work at the 2 wavelengths of 405 nm and 445 nm when the samples are measured, and thus separation of EC and BrC cannot be conducted by Eq. (5). Chengdu, Sanbacun, Wenchuan, Lixian, Maerkang and Hongyuan are abbreviated as CD, SBC, WC, LX, MEK and HY, respectively.

Season	Sites	OC	EC	babs (M m ⁻¹))	MAE (m ²	g ⁻¹)	WS	Tem.	RH
		(µg m ⁻³)	(µg m ⁻³)	BrC, 405	EC, 405	BrC, 405	EC, 405	(m s ⁻¹)	(°C)	(%)
Spring	CD	7.9±3.7	2.2±1.2	8.5±2.8	32.4±12.7	1.3±0.6	17.1±4.8	1.6±0.7	17.5±4.3	80.3±19.9
	SBC	6.7±3.0	3.3±1.5	13.7±5.6	44.2±16.9	2.1±0.9	13.7±2.5	1.4±0.6	16.9±4.1	77.6±15.9
	WC	3.2±1.6	1.2±0.8	4.8±2.3	20.2±9.0	1.6±0.9	21.5±11.6	2.4±1.0	15.1±4.4	65.1±17.4
	LX	3.5±1.4	1.0±0.6	$5.4{\pm}2.5$	11.8±5.5	1.7±0.8	13.8±6.9	1.6±0.5	13.3±5.3	61.5±20.4
	MEK	3.0±1.7	0.8±0.6	4.8±2.7	10.2±3.6	1.9±1.2	16.6±9.4	1.1±0.6	10.6±5.5	62.0±26.5
	HY	4.1±1.6	0.9±0.6	11.5±4.9	12.9±6.2	2.8±0.9	17.3±9.8	2.4±1.0	2.4±3.6	70.0±16.6
Summer	CD	5.4±1.2	1.9±0.5	9.0±2.7	29.2±6.9	1.8±0.6	16.4±4.5	1.3±0.4	25.2±2.9	84.6±18.8
	SBC	2.9±1.2	1.5±0.7	21.8±15.0	32.6±7.9	10.1±7.1	29.8±6.5	1.1±0.4	24.1±3.0	82.7±13.9
	WC	2.2±0.8	1.0±0.5	2.2±1.5	18.9±5.5	1.4±1.3	23.5±9.5	1.7±0.7	23.1±3.2	64.5±16.5
	LX	2.7±0.9	0.8 ± 0.5	13.3±5.0	9.5±2.7	5.4±2.5	16.4±11.3	1.4±0.5	20.9±4.0	65.2±18.0
	MEK	2.7±1.5	0.7 ± 0.6	_	_	—	—	1.0±0.4	16.6±4.3	73.3±22.6
	HY	3.0±1.2	0.8±0.6	_	_	—	—	1.8±0.6	10.1±3.3	77.8±11.6
Fall	CD	4.7±1.3	2.3±1.0	5.3±2.5	40.6±16.6	1.1±0.5	18.3±4.0	1.1±0.4	15.6±4.9	88.4±10.8
	SBC	5.3±3.4	3.0±1.8	22.0±13.7	50.8±11.2	6.0±5.6	24.3±9.3	0.9±0.2	14.9±4.4	89.9±11.6
	WC	1.6±0.8	0.8 ± 0.5	3.0±2.0	18.2±7.3	2.3±1.8	27.3±13.9	1.7±0.6	14.1±5.4	72.7±10.0
	LX	2.4±1.0	0.9±0.5	12.7±6.6	10.5±3.4	6.5±3.8	14.7±10.1	1.3±0.3	11.9±5.6	76.8±11.3
	MEK	2.3±1.2	0.9±0.6	—	—	—	—	0.9±0.4	8.8±5.5	78.4±17.0
	HY	3.4±2.2	1.3±1.1	—	_	_	_	1.9±0.7	0.7±5.6	73.9±11.0
Winter	CD	15.0±5.9	4.7±2.0	10.5±4.6	47.6±20.1	0.8±0.5	10.4±2.8	1.2±0.4	6.6±2.7	78.9±16.9
	SBC	18.9±7.6	7.9±3.4	17.1±10.2	74.7±27.9	1.2±1.0	9.9±2.0	1.0±0.3	5.8±2.7	79.2±15.0
	WC	8.2±3.1	2.8±1.3	11.2±3.2	29.7±9.5	1.5±0.5	11.6±4.4	1.9±0.6	3.6±2.4	60.2±9.0
	LX	8.4±2.7	3.0±1.3	17.1±15.4	24.3±9.1	2.2±2.6	8.9±3.9	1.4±0.4	-0.1±2.1	62.4±10.3
	MEK	5.3±2.3	2.2±1.1	13.2±4.0	16.6±6.3	2.5±0.9	8.6±4.4	1.1±0.3	-0.2±3.7	36.1±11.0
	HY	8.4±3.8	3.0±1.6	21.5±11.3	18.9±10.2	2.5±0.7	6.7±4.9	2.1±1.5	-6.5±6.8	42.8±21.8