

Retrieving CH₄ Emissions from Coal sampled with UAV-based Aircore system by using GA-IPPF model

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Abstract

Quantifying CH₄ emissions from coal mines has large uncertainty owing to the lack of effective monitoring methods. In this study, we developed a genetic algorithm–interior point penalty function (GA-IPPF) model to calculate the emission rate of large point sources of CH₄. This model can provide the detailed optimized dispersion parameters and has self-calibration characteristics that reduce the accuracy requirements for auxiliary data. We evaluate the influence of different parameters on the retrieved CH₄-emission rate by the GA-IPPF, including the uncertainty of CH₄ concentration measurements, number of CH₄ measurements, and meteorological data. Furthermore, based on the atmospheric CH₄ concentration measurements downwind of a Pniówek coal-mine ventilation shafts from a UAV-based AirCore system and the GA-IPPF model, we retrieve the CH₄-emission rates from the Pniówek coal (Silesia coal mining region mine, Poland) ventilation shaft. Results show that the CH₄-emission rates are variable even in a single day, ranging from 5.6±0.2 kt/year to 12.4±0.6 kt/year on August 18, 2017 and from 3.0±0.3 kt/year to 12.7±0.5 kt/year on August 21, 2017. The combination of the flexible UAV-based AirCore CH₄ measurements and the robust GA-IPPF model provides a effective means of quantifying CH₄ emissions from coal mines.

1.Introduction

The release of CH₄ into the atmosphere during coal mining is very concerning because it contributes to increased atmospheric concentration of CH₄, one of the most important greenhouse gases and waste resources(Cardoso-Saldana and Allen 2020; Zhang et al. 2020). However, CH₄ emissions during coal mining are not always stable owing to collection or manufacturing processes, weather fluctuations, terrain effects(Nathan et al. 2015b). Bottom-up inventories could provide us with approximate information on CH₄-emission rates from strong sources. However, inventories cannot serve as a standard for formulating policies to reduce anthropic CH₄ emission because of their low temporal resolution and uncertainty. The temporal resolution and accuracy of bottom-up inventory are too low to obtain emission

43 information instantaneously(Liu et al. 2020; Pan et al. 2021). Based on the measurements of CH₄
44 concentration around the emission source, developing a fast retrieval model to obtain emission intensity
45 is an acceptable method. With the development of remote-sensing technologies, the CH₄ emission rate
46 has become possible to quantify based on CH₄ concentration samples or measurement.

47 Greenhouse gases observing satellite and TROPOspheric Monitoring Instrument could obtain the
48 column concentration of CH₄ (XCH₄, ppb) with a spatial resolution of 10 km×10 km and 5 km×7.5 km.
49 The regional CH₄ flux could be retrieved by assimilating the measured XCH₄ into an atmospheric
50 dispersion model(Feng et al. 2016; Tu et al. 2022). PRISMA hyperspectral imaging satellite and GHGsat
51 could detect increased CH₄ caused by strong emission sources with high spatial resolution, and the
52 comprehensive CH₄ emission could be quantified by integrated mass enhancement or cross-sectional
53 flux method(Guanter et al. 2021; Varon et al. 2020). However, CH₄ emission from coal is not constant
54 even in a short time, and the spatial and temporal resolutions of satellites are not allowed to repeat the

55 quantification of CH₄ emission from coals in the same day(Schneising et al. 2020; Varon et al. 2019). An
56 airborne platform could fly in low altitude to improve the acquisition of CH₄ concentration(Elder et al.
57 2020; Wolff et al. 2021a) and estimate CH₄ emission from strong sources by cross-sectional flux method
58 or Gaussian dispersion method. However, it has strict requirements for the flight track (downwind
59 direction) and amount of measured CH₄ concentration data. Most ground-based sensors have the
60 advantage to sample the concentration around the source continuously, but it could collect data only near
61 the surface or could only measure column concentration(Caulton et al. 2017; Robertson et al. 2017; Zhou
62 et al. 2021), which are insufficient to generate the distribution characteristic of the emission source.
63 Ground-based different-absorption LIDAR could obtain the CH₄ profile concentration in different
64 altitudes, which is suitable as the input of the emission-retrieval model(Shi et al. 2020a), but it has high
65 requirements in terms of performance of hardware and system stability(Shi et al. 2020b). An unmanned
66 aerial vehicle (UAV) could reach any location rapidly around the CH₄ sources, which could sample CH₄
67 concentration with location information(Iwaszenko et al. 2021; Nathan et al. 2015b), when equipped
68 with an in-situ gas sensor. It could also acquire the characteristic of distribution with adequate data, which
69 is beneficial to retrieving the emission rate.

70 Mass-balance method has been applied in determining CH₄ emissions based on UAV-based
71 samples(Allen et al. 2018). Emission rates calculated by this method contain large uncertainty because
72 the main kernel is Kriging interpolation(Nathan et al. 2015a), which causes obvious uncertainty in
73 representing the actual feature of diffusion. The Gaussian dispersion model has also been applied in
74 retrieving gas emission from strong sources(Shah et al. 2019)³⁰, and is used to quantify CH₄ emission
75 from a coal-mine ventilation shaft in this study. However, existing emission-retrieval models need priori
76 information on diffusion parameters to retrieve the emission rate. Moreover, the accuracy of
77 measurements of auxiliary meteorological data influences the result of CH₄-emission calculation.
78 Therefore, we develop herein a model to overcome these shortcomings. Our model could calculate the
79 diffusion parameters without prior information and reduce the impact of meteorological data on the
80 calculated CH₄-emission rate.

81 In the present study, we collected CH₄ concentration ^{data} around a coal-mine ventilation shaft by using
82 UAV-based active AirCore system with high accuracy(Andersen et al. 2018) for a total of seven Flights.
83 Then, a CH₄ emission-retrieval model based on genetic algorithm (GA) combined with interior point-
84 penalty function (IPPF) is presented. GA-IPPF could help us obtain detailed information on emission
85 intensity and diffusion parameters. Finally, the performance of GA-IPPF model is compared with other
86 quantification methods for CH₄ emission.

87 2.Data and methods

88 2.1.Active AirCore System

89 The active AirCore system comprises a ~50 m coiled stainless-steel tube that works in conjunction
90 with a micropump and a small pinhole orifice (45 μm) to sample air along the trajectory of a drone. As
91 long as the pressure downstream of the orifice is more than half of that of the upstream (ambient) pressure,
92 a critical flow through the orifice is obtained. This finding means that the flow rate depends only on two
93 variables, namely, the air temperature and the upstream (ambient) pressure, both of which are monitored
94 during the flight. After obtaining the air sample, the sample is analyzed on a cavity ring down
95 Spectrometer model G2401-m for CO₂, CH₄, and CO. For CH₄, the accuracy of samples is ± 0.02 parts
96 per million (ppm). The active AirCore system is controlled using an Arduino-built data logger, which
97 records the temperature inside the carbon fiber housing. It also records the ambient temperature, ambient
98 pressure, relative humidity, and pressure downstream of the pinhole orifice to ensure that critical flow is
99 achieved. The datalogger also logs the GPS coordinates. The weight of the active AirCore system is ~1
100 kg. The active AirCore system is attached to a DJI Inspire Pro 1, which is capable of providing flights of
101 ~12 min.

102 2.2. Meteorological measurements

103 A radiosonde (Sparv Embedded AB, Sweden, model S1H2-R) measures ambient temperature, ambient
104 pressure, ambient relative humidity, wind speed, and wind direction. The detection range of the
105 temperature sensor is $-40\text{ }^{\circ}\text{C}$ to $+80\text{ }^{\circ}\text{C}$, with an accuracy of $0.3\text{ }^{\circ}\text{C}$. The pressure sensor has a detection
106 range of 300–1100 mbar, with an accuracy of 1 mbar. The relative humidity sensor measures a range of
107 0%–100%, with an accuracy of approximately 2%. Owing to the good connection between the
108 radiosonde and satellites, we assume that the uncertainty in the wind direction is low. The wind speed
109 can be estimated within a range of 0–150 m/s, with an uncertainty of approximately 5%. If the wind-
110 speed reading is less than 4 m/s, a minimum uncertainty of 0.2 m/s is given. The radiosonde is lifted by
111 a ~30 L helium-filled balloon and is tethered onto a fishing line for easier retrieval after making a vertical
112 profile.

113 2.3 Emission retrieve model

114 2.3.1.Gaussian dispersion model

115 The Gaussian dispersion model is used to analyze the CH₄ fugitive from the coal mine in this work.
116 The location of emission source is regarded as the coordinate origin; X-axis is the direction of the
117 downwind, Y-axis is cross-wind direction, and Z-axis is the altitude above the ground. Based on the
118 established coordinate system, the Gaussian plume could be modeled by Equation 1:

$$119 C(x, y, z) = \frac{q}{2\pi u \sigma_y \sigma_z} \exp\left(\frac{-(y)^2}{2\sigma_y^2}\right) \left\{ \exp\left(\frac{-(z-H)^2}{2\sigma_z^2}\right) + \alpha \cdot \exp\left(\frac{-(z+H)^2}{2\sigma_z^2}\right) \right\} + B \quad (1)$$

$$120 \sigma_y = a \cdot x^b \quad (2)$$

$$121 \sigma_z = c \cdot x^d \quad \text{methane is being emitted not coal} \quad (3)$$

122 C is the concentration (g/m^3), q (g/s) is the emission rate of coal from stack, u is the mean wind speed
123 around the stack (m/s), H is the effective stack height, σ_y is the dispersion coefficient in the horizontal
124 direction, σ_z is the dispersion coefficient in the vertical direction, u is the wind speed (m/s), and B is the
125 background concentration of CH₄. Moreover, α is the reflection index of this phenomenon; and x, y, and
126 z are the positions of the samples in the determined coordinate system.

127 2.3.2.GA-IPPF model.

128 First, the genetic algorithm (GA) kernel calculates Q and other dispersion parameters as first guess(Liu

129 and Michalski 2016). It guarantees the retrieved unknown parameters through the global optimum
 130 solution, including Q and diffusion parameters, as shown in Figure1. Then, the results calculated by GA
 131 serve as **oral input parameters** and constraints in the IPPF model, and actual values of the concerned
 132 parameters are retrieved by IPPF.

133 Based on the Gaussian dispersion model, auxiliary meteorological data, location information, and CH₄
 134 samples, we determine the unknown parameters in equations 1 to 3 by using GA, including q, H, a, b, c,
 135 d, and α . First, the locations and concentration of CH₄ samples and wind serve as an initial input of
 136 equation 1. Then, the fitness value evaluates the applicability of the calculated parameters in each step.
 137 We define the fitness value as

$$138 \quad F = \sum_{i=1}^n (C_m^i - C_s^i)^2 \quad (4)$$

$$139 \quad C_s^i(x, y, z) = \sum_{i=1}^n \frac{q'}{2\pi u' \sigma_y' \sigma_z'} \exp\left(-\frac{(y)^2}{(\sigma_y')^2}\right) \left\{ \exp\left(-\frac{(z-H')^2}{(\sigma_z')^2}\right) + \alpha' \cdot \exp\left(-\frac{(z+H')^2}{(\sigma_z')^2}\right) \right\} + B' \quad (5)$$

140 F is the fitness value; C_m^i is the sample CH₄ concentration; i is the number of samples; C_s^i is the
 141 simulated concentration of CH₄ in the location of samples calculated by formula 5; and q' , u' , σ_y' , σ_z' , H' ,
 142 α' , and B' are the calculated CH₄-emission rate, wind speed, diffusion parameters, emission height, reflect
 143 index, and background CH₄ concentration, respectively, acquired from the “Mutation” in Figure1. When
 144 f is less than the threshold value (1×10^{-5}) of the fitness value, the corresponding parameters are treated
 145 as the results of output.

146 IPPF rebuilds the inequality constraint conditions to the unconstrained solution process. It forces the
 147 start point to satisfy the constraints, as shown in formula 6.

$$148 \quad \min F(x, r_k) = f(x) + r_k B(x) \quad (6)$$

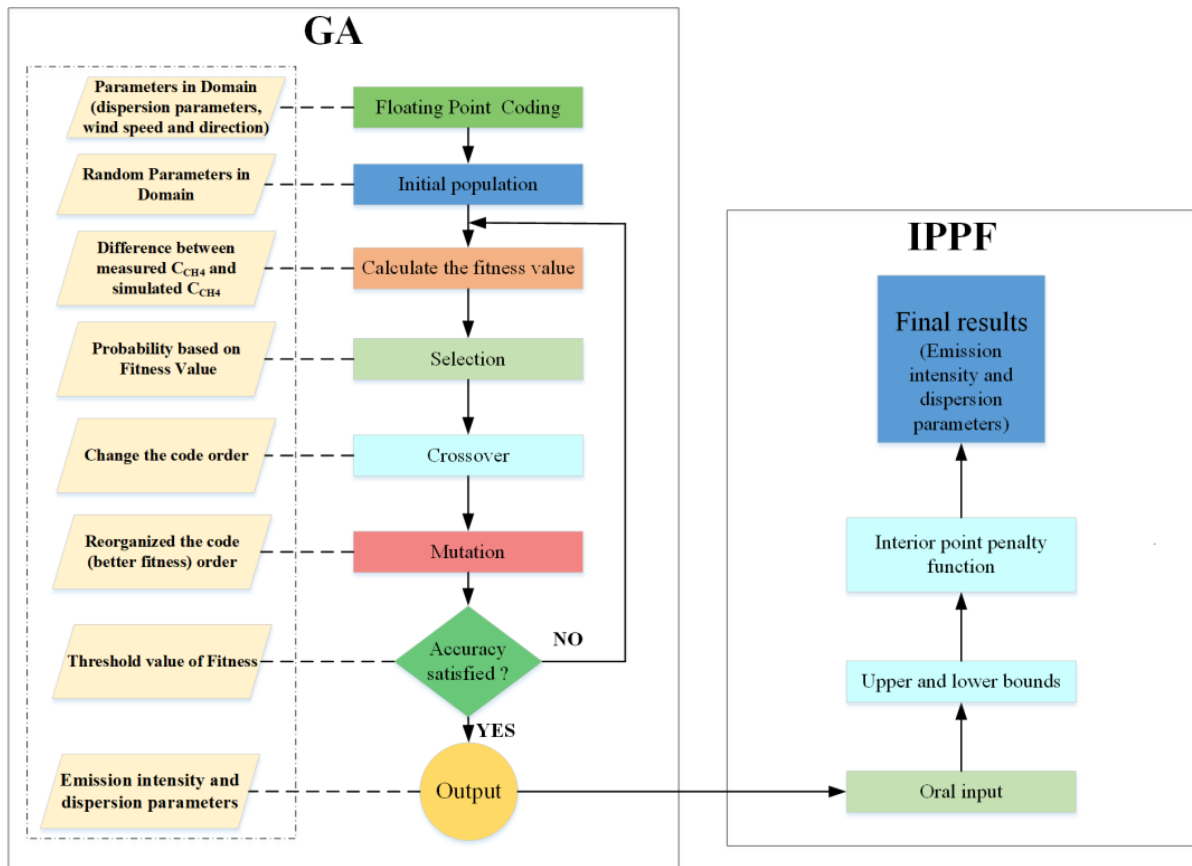
149 $f(x)$ is the unconstrained equation, and r_k is the coefficient of the constrained equation $B(x)$. When the
 150 solution parameters are out of the constraints, $r_k B(x)$ is large, thereby ensuring that the final solution is
 151 feasible under the inequality constraint conditions.

152 To obtain the inequality constraints, GA is repeated 1000000 times, and the mean values of the
 153 calculated wind speed, wind direction, H, a, b, c, d, and α are treated as the oral input of IPPF model.
 154 The domains of H, a, b, c, d, and α are determined by two times the standard deviation of the
 155 corresponding results in GA. The constraint values of wind speed and direction are set according to the
 156 precision of actual measurements, $m \pm \sigma$, whereas m is the measured value of wind speed or wind direction,
 157 and σ is their precision. Actual B values are considered to range within 1800–2500 ppb. Then, the Pearson
 158 correlation coefficient (R) values of the actual samples and simulated one work as judgment in the
 159 solution process of formula 7.

$$160 \quad R = \frac{\sum_{i=1}^n (C_s^i - \bar{C}_s)(C_m^i - \bar{C}_m)}{\sqrt{\sum_{i=1}^n (C_s^i - \bar{C}_s)^2} \sqrt{\sum_{i=1}^n (C_m^i - \bar{C}_m)^2}} \quad (7)$$

161 The results are treated as the final retrieved values of the concerned parameters when the R reaches
 162 the maximum. We use the “fmincon (interior point)” toolbox in MATLAB 2020a to implement the IPPF
 163 model.

164



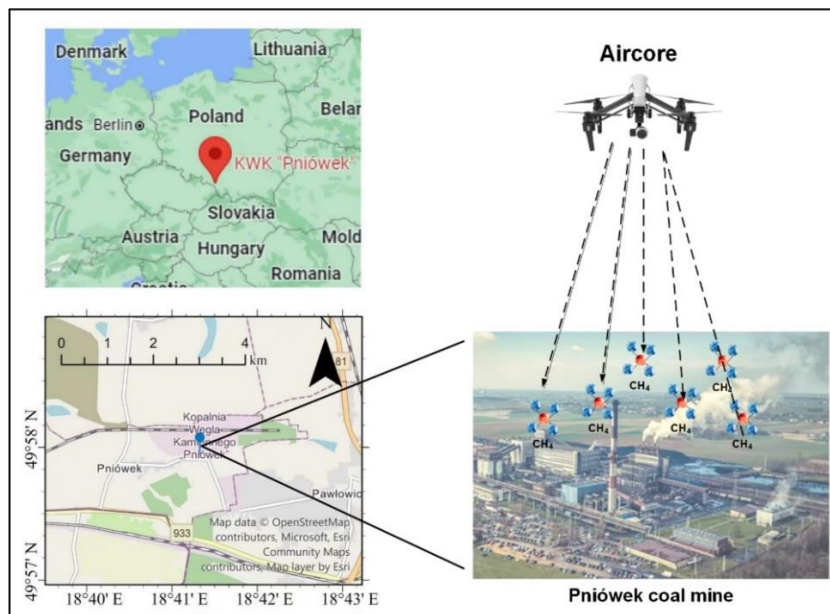
165

166

Figure1. Flow chart of Genetic Algorithm and IPPF model.

167 *This section needs to go before the model section as the location is mentioned. You should also make it clear what type of source this is as*
 168 *2.4. Measurement Site*
 169 *this will make a big difference to the model. Is a large open coal pit, or emission from a deep mine shaft or vent, or is it a combination of both.*

168 The Pniówek coal mine (49.975° N, 18.735° E) is a large mine in south Poland in Pniówek, Silesian
 169 Voivodeship, which is 190 km southwest of the capital Warsaw, see Figure2. It has a large coal reserve
 170 estimated to be about 101.3 million tons. Its coal production is about 5.16 million tons per year.



171

172

Figure2. The Pniówek coal mine

173

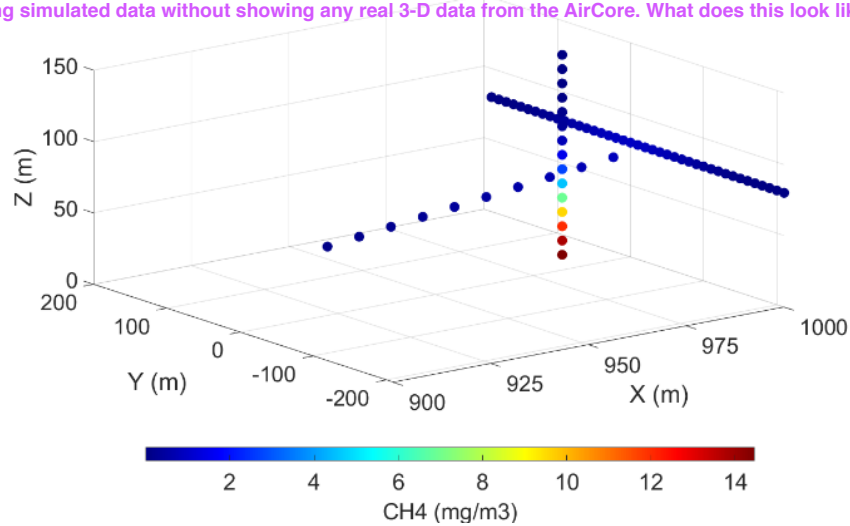
3. Results

Caption needs more explanation. What actually is the zoomed map showing? Where is the mine? Is pink the mine buildings? Is it deep mine with shaft or open pit?

174 3.1. Validation of performance of GA-IPPF model through simulations

175 First, the dispersion of CH₄ emission from **a coal** is simulated by equation 1, and the dispersion
 176 parameters are shown in Table 1. To make the simulations close to the actual measurement scenarios,
 177 random errors are added to the CH₄ concentration samples (5%), wind speed (± 0.3 m/s), and wind
 178 direction ($\pm 20^\circ$). The spatial resolution of the supposed samples is 10 m, and 70 samples are selected
 179 from the simulated dispersion to represent the data acquired by the UAV-based AirCore. Then, the
 180 concerned parameters are retrieved by the GA-IPPF method. The input parameters include hypothetical
 181 wind speed, wind direction, and 70 samples, as shown in Figure3. Simulations are repeated 10 000 times,
 182 and the average values of the corresponding parameters are treated as the “Retrieved” results in Table 1.

Showing simulated data without showing any real 3-D data from the AirCore. What does this look like? What are you trying to simulate?



183
184
185

Figure3. The determined 70 CH₄ samples in simulations

Table 1. Set parameters in dispersion simulation and the retrieved results

Parameters	Actual	Retrieved
Emission intensity (g/s)	300	300.5±0.01
Wind speed (m/s)	3	3±0.01
Wind direction (°)	90	90±0.01
a	0.11	0.13±0.02
B	0.9	0.9±0.02
c	0.1	0.12±0.01
d	0.82	0.8±0.01
B (ppb)	1900	1900±3
Emission height (m)	20	19.7±1.2
α	0.9	0.89±0.03

186 “Actual” means the set values of parameters, and “Retrieved” means the average values of parameters
 187 retrieved by GA-IPPF model through 10 000 times of simulation.

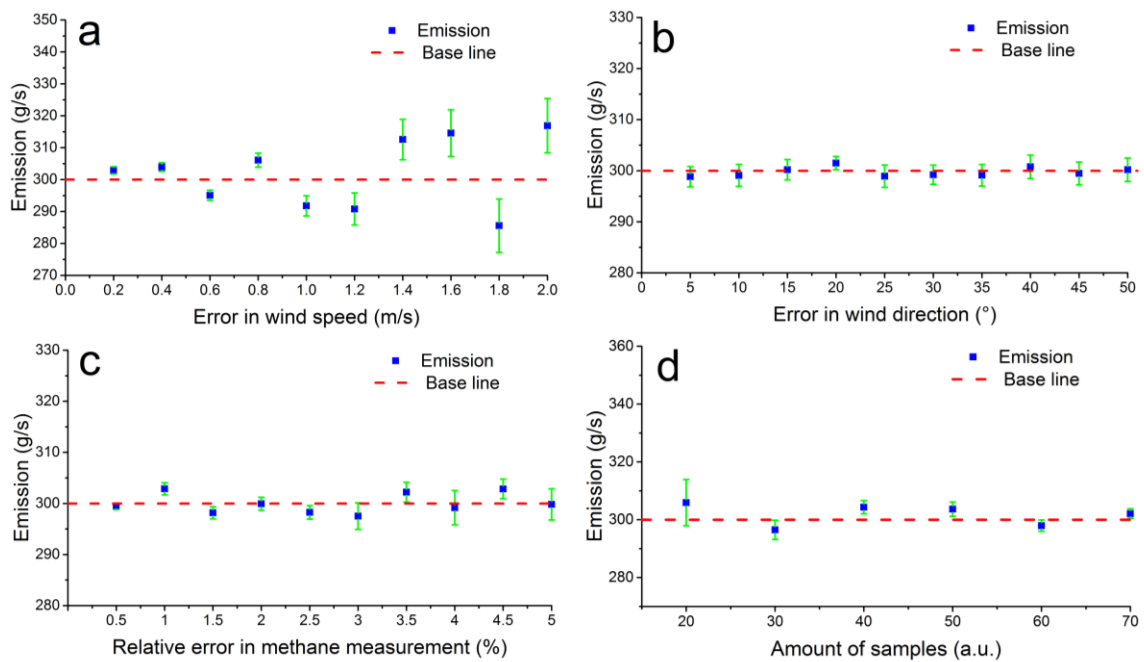
188 **As shown in Table 1, q retrieved by GA-IPPF has only 0.17% bias compared with the set values.**

189 Emission height only has 0.3 m bias to set one, and uncertainty is only 0.6% to 20 m. Other retrieved
 190 parameters also show high consistency with the original settings.

191 3.2. Stability analyses

192 The necessary input parameters in GA-IPPF contain meteorological data (wind speed and direction),
 193 accuracy of CH₄ samples, and amount of CH₄ samples. In formula 1, wind speed nearly has a linear
 194 relationship with the emission estimation. Wind speed is also an important factor that determines

195 atmospheric stability according to the Pasquill–Gifford method(Venkatram 1996) as it affects the
 196 diffusion parameters of σ_y and σ_z . The coordinate is built according to the wind direction, **which is defined**
 197 **the plane coordinates of CH₄ samples**. According to formulas 2 to 3, errors in wind-direction
 198 measurement lead to wrong σ_y and σ_z on each position of samples. CH₄ samples are the most important
 199 factors to determine the Gaussian diffusion. The accuracy of samples influences the judgment of “fitness”
 200 in the GA process. More samples collected in different positions help rebuild the spatial-distribution
 201 characteristics of the plume because it provides larger possibility for fitting process in IPPF and helps
 202 determine the optimum solution. To evaluate the influence of errors in the measurements of necessary
 203 parameters on the final retrieved results, the same settings in Table 1 are used as actual results. The
 204 performance of the GA-IPPF model with additional random errors in each parameter are simulated
 205 10 000 times, as shown in Figure4.



206
 207 **Figure4.** Influence of accuracy of parameters on final results. Baseline represents the set emission rate
 208 of CH₄, 300 g/s: (a) wind speed, with additional error ranging within 0.2–2 m/s and an interval of 0.1
 209 m/s, (b) wind direction, with additional error ranging within 5°–40° and an interval of 5 °, (c) accuracy
 210 of CH₄ samples, with additional error ranging within 0.5%–5.0% and an interval of 0.5%, and (d) amount
 211 of CH₄ samples, randomly selected as 20–70 among the defined 70 samples.

212 In Figure 4 (a), the mean value of q retrieved by GA-IPPF has nearly the same with the baseline if the
 213 error in wind speed is less than 0.4 m/s and the maximum bias to the baseline is 16.3 g/s. Fluctuation of
 214 q occurs obviously if error in wind speed exceeds 0.4 m/s. The standard errors of q are positively
 215 correlated with the values of errors in wind speed, indicating that the accuracy of wind-speed
 216 measurements largely influences the stability of the GA-IPPF model. This model has a self-adjustment
 217 function for wind speed; for example, when the oral wind speed is 3 m/s, the maximum standard error of
 218 q is only 8.5 g/s (3.5% to the 300 g/s) when the additional error of wind is 2 m/s (66.7% to 3 m/s).

219 The retrieved q shows less sensitivity on errors in wind direction, see Figure4(b). When errors in the
 220 wind direction are 5° to 40°, all biases of q are within 1.1 g/s and the standard errors are around 2.3 g/s.
 221 Wind direction determines the spatial location of the sampling point, and **wrong location information**
 222 **leads to distinct errors in emission estimation**. GA-IPPF shows highly accurate ability in wind direction

This implies that there could be big errors in q if the source is not a point source, but unequally distributed over an area, which could be the case with a farm or landfill site emission.

223 to obtain the global optimum solution.

224 Sampling accuracy has small impact on the retrieved q within different settings in CH_4 samples' accuracy, see Figure4(c). Standard deviation is positively correlated with errors in CH_4 measurements.

225 accuracy, see Figure4(c). Standard deviation is positively correlated with errors in CH_4 measurements.
226 The standard deviation is 3.1 g/s when the measurement error reaches 5%. Notably, the uncertainty of
227 CH_4 samples measured by UAV-based AirCore system is far less than 5%. The AirCore system could

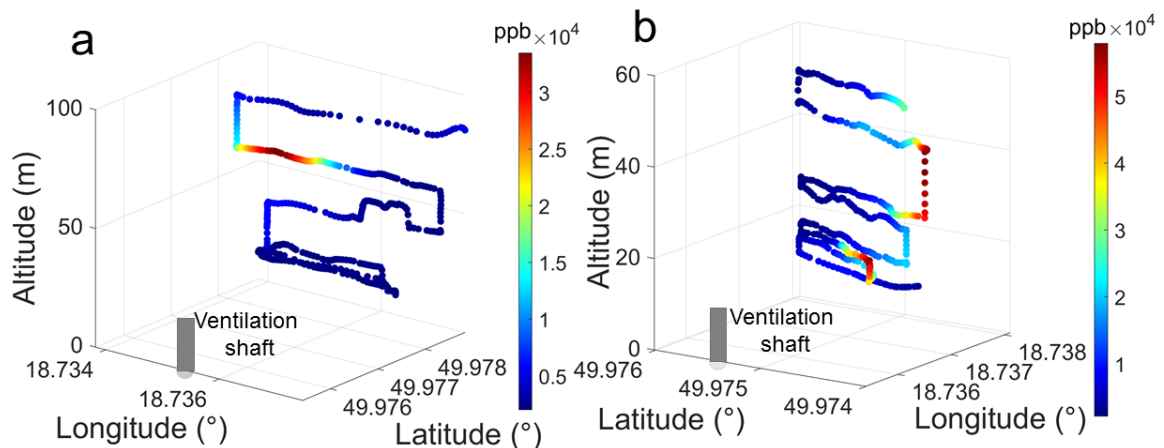
228 acquire more than 70 CH_4 samples in actual feasible measurements, thereby guaranteeing the accuracy
229 of the retrieved CH_4 emission by coal to exceed 99.2%. Not clear what this means. Number of air Core samples that can be collected
230 and analysed within a given time period within the same source emission plume

231 The number of measurement points obviously influences the final accuracy of q by the GA-IPPF model,
232 see Figure4(d). It has 5.9 g/s bias to 300 g/s when n is 20. The accuracy of q and the standard error are
233 negatively correlated with n , which provides the number of judging criteria for the fitting process in the
234 retrieval model. Hence, n directly influences the retrieved results. The AirCore system has the advantage
235 of continuous sampling during flight, which integrates the atmospheric signals along the flight path and
236 helps reduce the uncertainty of the retrieved q . On the other hand, the smoothing of the atmospheric
237 signal also reduces the spatial resolution of the measurements, which needs to be considered during the
238 optimization³⁰.

239 IPPF can suitably solve the problem of inequality constraints, and the calculated solution guarantees
240 the calculated parameters to be within the feasible region. In this section, GA-IPPF model performance
241 and its adjustments on the concerned four input parameters were discussed.

241 3.3. Actual experiments

242 Fifteen active AirCore flights around Pniówek coal mine are collected successfully on August 18, 2017
243 and August 21, 2017. The sample data in Flight 8 (18/8/2017) and Flight 15 (21/8/2017) are used to
244 evaluate the GA-IPPF model, as shown in Figure5.



245 **Figure5.** Samples of CH_4 in two Flights: (a) Flight 6 and (b) Flight 15.
246

247 In Flight 5, the AirCore system collects CH_4 around the coal spirally from 0 m to 98 m, for a total of
248 376 samples, and the measurement period is 7 min, ranging from 1980.1 ppb to 49 113.9 ppb. In Flight
249 15, the AirCore system collects a total of 400 samples, and the measurement period is 9 min, ranging
250 from 2131.7 ppb to 57 265.3 ppb. Both Flights show high spatial variability in CH_4 exhaust from coal
251 mine. Subsequently, we input the wind speed, wind direction, location information, and CH_4 samples
252 collected from Flights into the GA-IPPF model. To express the final retrieved emission (Q) in g/s, the
253 dry-air mixing ratio of CH_4 C (ppb) is transformed into mass concentration m (g/m^3) as follows:

254
$$m = C \cdot \frac{M_{\text{CH}_4}}{M_{\text{Air}}} \quad (8)$$

Surely an AirCore is one sample that has enough air to allow 400 measurements when attached to the Picarro?

255 M_{CH_4} is the molar mass of CH_4 , M_{air} is the molar mass of air.

256 The retrieved results are shown in Table 2. Notably, the emission height in Flight 15 is larger than that
257 of Flight 6, which may be caused by the difference in thermal energy and vertical wind speed of the two
258 flights. The background concentrations of CH_4 are 2001.3 and 2002.1 ppb in Flights 6 and 15,
259 respectively, which show little difference. The dates of the two Flights are very close, so the background
260 concentration of CH_4 in two days have nearly the same seasonal characteristic. The exhaust gases of coal
261 mine are emitted through the stack with effective emission heights of 5.7 and 3.64 m, respectively.

262 To evaluate the rationality of the retrieved results, these parameters are used to simulate CH_4 diffusion
263 from the Pniówek coal mine according to equation 1. The comparison between simulated CH_4
264 concentration and actual samples in the same locations is shown in Figure6.

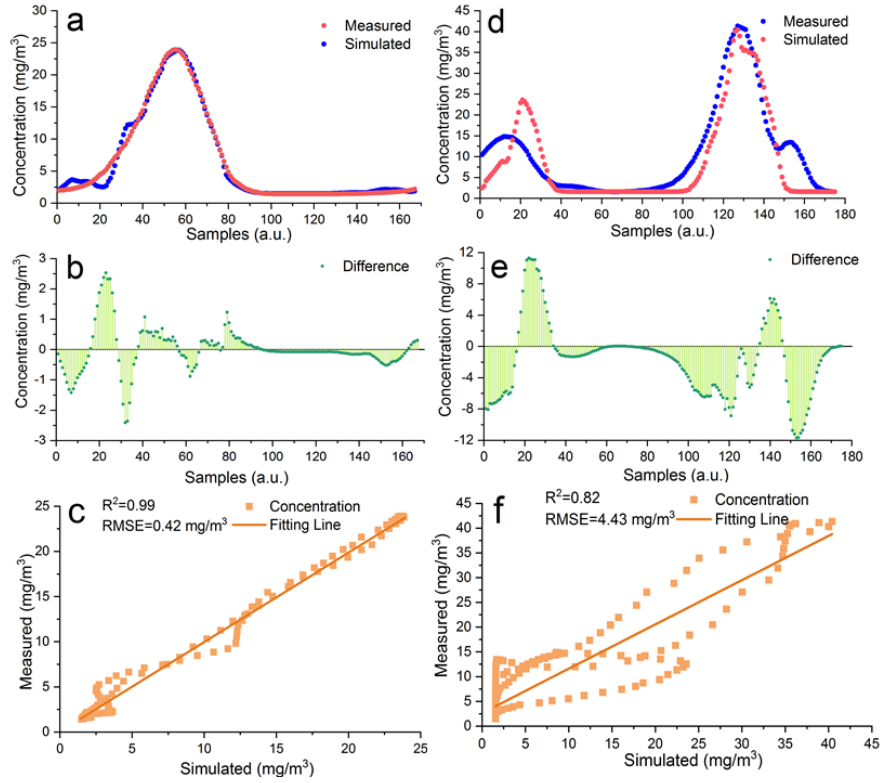
265 Table 2. Results calculated by GA-IPFF model

Parameters	Flight 6	Flight 15
Oral wind speed (m/s)	4.5	4.1
Oral wind direction (°)	310	125.4
Emission intensity (kt/year)	6.1±0.3	8.4±0.5
Wind speed (m/s)	3.25	3.20
Wind direction (°)	349.6°	128.1
a	0.22	0.31
b	0.90	0.90
c	0.006	1.50
d	1.29	0.38
B (mg/m ³)	1.55	1.57
Emission height (m)	59.3	36.3
Reflection index	0.85	1.0

266 Then, we also calculated the difference between the actual measured points and simulated ones as

267
$$D_c = C_s - C_m \quad (9)$$

268 D_c is the difference of CH_4 concentration between actual measured and simulated ones. C_s is simulated
269 CH_4 concentration (mg/m³), C_m is simulated CH_4 concentration (mg/m³).



270

271 **Figure6.** Comparison between the measured samples and the simulated ones based on the parameters in
 272 Table 1: (a). Flight 6 and (d) Flight 15. The difference of simulated CH₄ concentration and actual
 273 measured ones:(b) Flight 6 and (e)Flight 15. Correlation Analysis: (c) Flight 6 and (f)Flight 15.

274 As shown in Figure6(a), the tendency of the simulated concentration data is consistent with the
 275 measured ones in Flight 6. The largest value (NO.55) of the measured CH₄ concentration is 23.9 mg/m³,
 276 whereas simulated one is 23.8 mg/m³ on same location, only 0.42% bias. Dc is ranging from -2.4 to 2.3
 277 mg/m³ in Flight 6 (see Figure6(b)), this little bias indicates the simulated result is reasonable. The R²
 278 of the measured samples and simulated ones is 0.99, root mean square error (RMSE) is 0.42 mg/m³.It
 279 indicates the GA-IPPF model could correctly rebuild the diffusion of CH₄ in Flight 6. Figure6(d) shows
 280 a slight difference between the two items in the first peak and third peak. Since the GA-IPPF method
 281 would adjust more weights to the samples with higher concentration (NO.100 to 150 in Flight 15) to get
 282 the global optimal solution of the relevant parameters, this would lead to the low fitness of the first peak
 283 in Figure6(e). In general, the tendency of the simulated ones remains consistent with that of the actual
 284 samples in Flight 15, especially for the points in the second peak. R² and RMSE of the measured samples
 285 and simulated ones in Flight 15 also show the high applicability of the retrieved parameters.

286 3.4. Uncertainty Analyses

287 The uncertainties in CH₄-emission estimations are derived from the measurements of meteorological
 288 data and CH₄ samples in the discussed Flights. As shown in Figure3, n is the largest source that
 289 contributes to the uncertainty in emission calculation, and the accuracy of AirCore samples, wind speed,
 290 and wind direction should also be of concern. Therefore, the total uncertainty in actual CH₄ emission
 291 retrieved could be calculated as

$$\varepsilon_t = \sqrt{\varepsilon_n^2 + \varepsilon_m^2 + \varepsilon_w^2 + \varepsilon_d^2} \quad (10)$$

292 Where ε_t is the total uncertainty of Q estimation. In this section, ε_n , ε_m , ε_w , and ε_d are the uncertainty
 293 caused by n, accuracy of CH₄ samples, wind speed, and wind direction, respectively. Results in Figure3

294 are used to determine the values of ε_n , ε_m , ε_w , and ε_d in each Flight.

295 **4. Discussion**

296 *4.1. Comparison with other methods*

297 To investigate the difference between our recommended emission model and others, three methods
298 have been applied to estimate CH₄ emission in all Flights, including mass-balance approach, nonlinear
299 least square fit (NLSF), and facility emission.

300 Mass-balance approach quantifies CH₄ emission by calculating the cross-sectional flux perpendicular to
301 the wind direction (Krings et al. 2018). First, a two-dimensional plane is selected according to the **amount**
302 **of CH₄ samples**. Second, the two-dimensional plane is divided into a grid of equal spatial resolution.
303 Third, CH₄ samples are regarded as origin points to interpolate full grids defined by the Kriging
304 interpolation scheme (Mays et al. 2009). Finally, the emission rate of the CH₄ source is calculated by

$$305 \quad F_{(CH_4)} = \iint v \sin(\alpha) \cdot (C_{(x,z)} - C_{bg}) dx dz \quad (11)$$

306 Where v is the wind speed, α is the angle between wind direction and the two-dimensional plane, $C_{(x,z)}$
307 is the density of CH₄ in each grid, and C_{bg} is the background of CH₄ in each grid. The uncertainty analyses
308 of this method are detailed in Nathan et al. (Nathan et al. 2015a).

309 NLSF and the combination of NLSF with Gaussian diffusion model are also extensively used for point-
310 source emission retrieval (Wolff et al. 2021b; Zheng et al. 2020). In this study, NLSF is used to estimate
311 Q in each Flight by fitting the unknown parameters in equation 1, and the uncertainty of the retrieved Q
312 is evaluated with formula 6.

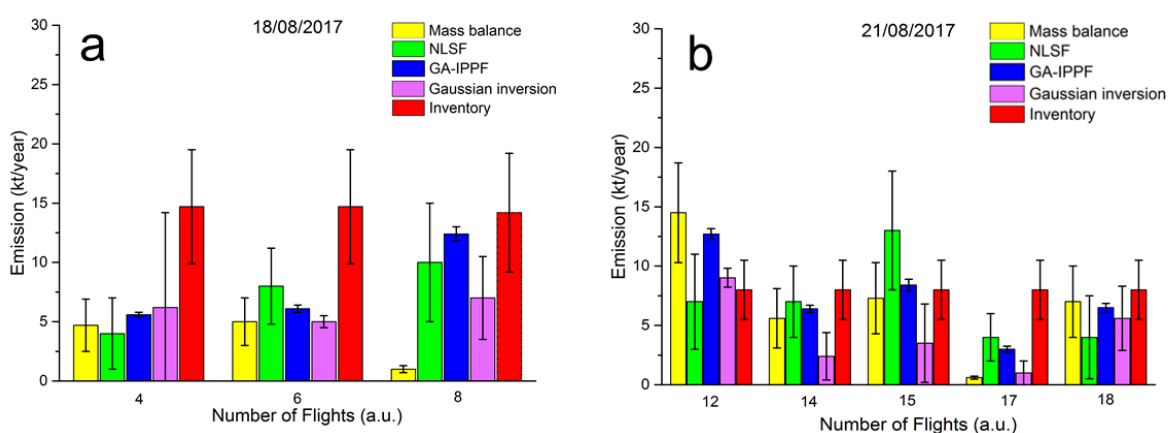
313 Missing date Andersen et al. also developed an inverse Gaussian approach to quantifying CH₄ emissions from coal
314 mine based on the same Flights (Andersen et al. 2021). First, the Gaussian dispersion is built as

$$C(x, y, z) = \frac{q}{2\pi u \sigma_y \sigma_z \cos(\theta)} \exp\left(\frac{-(y)^2}{2\sigma_y^2}\right) \left\{ \exp\left(\frac{-(z-H)^2}{2\sigma_z^2}\right) + \alpha \cdot \exp\left(\frac{-(z+H)^2}{2\sigma_z^2}\right) \right\} \quad (12)$$

315 Where θ is the angle between the wind direction and the perpendicular line of the flight trajectory. This
316 model does not contain the item of background of CH₄. Furthermore, σ_y and σ_z are treated as certain
317 values in equation 8. Then, q and the related parameters are retrieved by “fmincon optimizer” in
318 MATLAB (detailed settings are found in Andersen et al.). The CH₄ emissions in each Flight, as evaluated
319 by Andersen et al., are presented in this section.

320 Facility-emission data and hourly CH₄ emission from shaft are calculated by measuring raw CH₄
321 concentration and air flux through the shafts. The CH₄ emission rates estimated using hourly facility-
322 emission data for 21 August 2017 and 21 August 2017 are 14.4 ± 4.9 and 8.2 ± 2.9 kt/year, respectively,
323 as shown in Figure 7.

Why is the inventory emission lower for flight 8? Seems to be consistent across the second day.



324

325 **Figure7.** Quantified CH₄ emission by different methods based on the collected data: (a) August 18, 2017
 326 and (b) August 21, 2017. The results of CH₄ emission rate calculated by Mass balance and Inverse
 327 Gaussian refer to [Andersen et al³⁰](#). Why does this reference have no date and a number next to it (wrong style)

328 As shown in Figure7, Flights 4, 6, and 8 are measured on 18 August 2017, whereas Flights 12, 14, 15,
 329 17, and 18 are measured on 21 August 2017. Figure7(a) shows that the CH₄-emission rates calculated by
 330 mass balance are smaller than the inventory estimation in all Flights. In Flight 8, q retrieved by mass
 331 balance is extremely lower than those quantified by other methods, whereas q retrieved by GA-IPPF
 332 model (12.4±0.6 kt/year) shows only a slight difference from the inventory. As shown in Figure 7(b),
 333 CH₄ emissions retrieved by mass balance, inverse Gaussian, and GA-IPPF model are overestimated
 334 compared with the inventory in Flight 12. Mass balance and inverse Gaussian method also show
 335 obviously underestimated q in Flight 17. Estimations of retrieved CH₄ emission in Flight 18 show
 336 consistency among methods of mass balance, GA-IPPF, and inverse Gaussian. The CH₄-emission rate of
 337 coal generally has significant variability in each measurement, even on the same day. Mass balance is
 338 very sensitive to the size settings of grids, and different height and length settings can affect the
 339 concentration distribution across the cross-section. NLSF has a high-accuracy requirement for wind
 340 measurements, and errors on these measurements have a linear influence on the final emission estimation.
 341 Notably, the standard errors of q quantified by GA-IPPF always are the least among these methods,
 342 indicating the stability of our developed model.

343 4.2 Application of Reanalysis meteorological database in GA-IPPF model

344 Wind speed and wind direction acquired by the radiosonde or weather station are two main parameters
 345 in GA-IPPF. However, additional sensors are bound to increase the cost and difficulty during actual CH₄-
 346 emission measurements. To explore the possibility of weather reanalysis data instead of actual wind
 347 measurement by sensors, we use 10 m U and V wind components from the ERA5 meteorological
 348 reanalysis database (spatial resolution is 0.1°×0.1°, and temporal resolution is 1 h) developed by the
 349 European Centre for Medium-range Weather Forecast(Hersbach et al. 2020) to evaluate GA-IPPF model.
 350 However, the wind directions from ERA obviously differ from the actual measurements during the
 351 Flights. Hence, we determine the wind direction by using the CH₄ samples, for example, the line between
 352 the shaft and the location of the maximum value of samples in the same heights is treated as the
 353 downwind direction, whose uncertainty is set as 50°. Wind speed from ERA is used for the CH₄-emission
 354 calculation, uncertainty is supposed as 2 m/s. Even oral wind speed and direction obviously differ
 355 between the two sources; however, the GA-IPPF model adjusts them into reasonable ranges. The results
 356 of q during all Flights retrieved by two meteorological data sources have been evaluated, as shown in

357 Table 3.

358

Table 3. Retrieved CH₄ emission by ERA meteorological data

Flights	Actual	ERA
4	5.6±0.2	6.0±0.3
6	6.1±0.3	6.4±0.5
8	12.4±0.6	14.4±0.9
12	12.7±0.5	13.2±0.7
14	6.4±0.3	5.0±0.4
15	8.4±0.5	7.8±0.5
17	3.0±0.3	3.4±0.5
18	6.5±0.4	7.0±0.5

359 Table 3 shows that the values of quantified q between two meteorological sources are within 20% in
360 the same Flight. The standard errors of q retrieved by the ERA5 database are larger than those from actual
361 measurements, which depends on the accuracy of the reanalysis of wind speed and wind direction. Thus,
362 it reduces the necessity of additional equipment except for the AirCore system and the complexity of this
363 program.

This is not a sentence

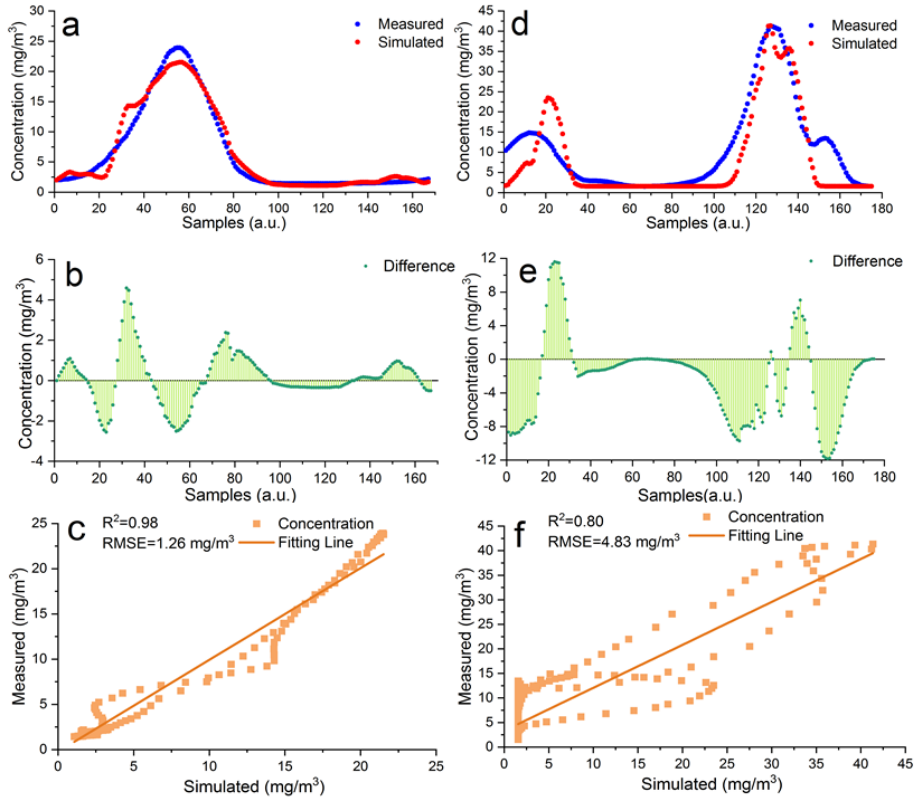
364 To explore the reason that the acceptable difference of calculated methane emission rate by the two
365 sources of meteorological data. The concerned parameters in Flight 6 and Flight 15 calculated based on
366 ERA5 meteorological data were presented in Table 4.

367

Table 4. Results calculated ERA5 meteorological data

Parameters	Flight 6	Flight 15
Oral wind speed (m/s)	2.6	4.1
Oral wind direction (°)	300	120
Emission intensity (kt/year)	6.4±0.5	7.8±0.6
Wind speed (m/s)	2.99	4.52
Wind direction (°)	349.4°	128.1
a	0.28	0.18
b	0.90	0.93
c	0.01	0.13
d	1.26	0.84
B (mg/m ³)	1.56	1.57
Emission height (m)	60.2	36.0
Reflection index	0.80	0.71

368 The oral wind speed and wind direction in Table 4 are obviously different from those in Table 2.
369 However, the calculated wind directions are nearly the same based on the two sources of meteorological
370 data. Diffusion parameters and emission height in Table 8 also show less difference in two Tables (Table
371 2 and Table 4). It is worth nothing that the wind speed and reflection index would be adjusted to reach
372 the global solution by GA-IPPF model, which leads to little bias for the emission rate of CH₄ in Table 3.



373

374 **Figure8.** Comparison between the measured samples and the simulated ones based on the ERA5
 375 meteorological data: (a). Flight 6 and (d) Flight 15. The difference of simulated CH4 concentration and
 376 actual measured ones:(b) Flight 6 and (e)Flight 15. Correlation Analysis: (c) Flight 6 and (f)Flight 15.

377 The simulated concentration of CH₄ in Flight 6 and Flight 15 calculated by parameters in Table
 378 4 are shown in Fig.8. In Fig.8(a), the consistency between actual samples and simulated ones is
 379 slight lower than that in Fig.6(a), D_c is ranging from -2.4 to 4.3 mg/m³, which is an acceptable bias
 380 as only 6 points exceed 2.3 mg/m³. R² (0.98) of measured samples and simulated ones is almost
 381 same to that in Fig.6(c), while RMSE is nearly three times than that in Fig.6(c). In Fig.8(d), the
 382 tendency of simulated CH₄ concentration is similar to Fig.6(d). D_c is ranging from -11.9 to 11.6
 383 mg/m³, which is nearly same as the result in Fig.6 (e), it worth nothing that D_c simulated by ERA
 384 meteorological data is slight larger on samples (NO.1 to 20) compared with that in Fig.6 (e). The R²
 385 and RMSE in Fig. 8(f) indicate that the retrieved results using ERA data are less accurate than that
 386 using actual measured meteorological data. In summary, though we set large uncertainties in ERA
 387 meteorological data, GA-IPPF can still guarantee reasonable and adequate accuracy for the retrieved
 388 emission rate and diffusion parameters.

389 5.Conclusion

390 CH₄ emissions from coal are inconsistent even with short time differences. They usually show a large
 391 difference for different mining volumes and types. Enhancement in CH₄ by the emission source is much
 392 larger than the background concentration, and the distribution of leak gas shows an obvious spatial
 393 difference. Hence, the retrieval time needs to be shortened for each emission measurement. AirCore has
 394 high portability and flexibility to measure CH₄ concentration around emission sources, accompanied by
 395 the GA-IPPF model, which is acceptable to calculate CH₄ emission from coals or other point sources.
 396 This program can help improve the accuracy of estimating CH₄ emission from coals, especially
 397 developed countries that even lack no inventories of gas emission. It can also help governments evaluate

this is a double negative

398 the fugitive CH₄-emission rate during mining and formulate policies to promote the innovation of mining
399 equipment and technology. The recommended program is appropriate for quantifying local sources based
400 on the advantage of hardware and retrieval model. The UAV-based AirCore system helps rebuild the
401 diffusion of CH₄ with flexibility and high sampling rate. The GA model could restrict the calculated
402 emission details within a reasonable range. Therefore, this program has great potential application in the
403 point-source quantitation of other gases, such CO₂, SO₂, etc.

Should include a recommendation of how many Air
Core retrievals from a single point source are
required to minimise the errors on model results

404 Acknowledgements

405 This work was supported by the National Natural Science Foundation of China (Grant No.41971283,
406 41801261, 41827801, 41901274, 41801282, 42171464), the National Key Research and Development
407 Program of China (2017YFC0212600), The Key Research and Development Project of Hubei Province
408 (2021BCA216), the Open Research Fund of National Earth Observation Data
409 Center(NODAOP2021005).

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