~~On the understanding of tropospheric fast photochemistry: airborne~~ Airborne observations of peroxy radicals during the EMeRGe-campaign in Europe ~~campaign~~

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Abstract. In this study, airborne measurements of the sum of hydroperoxyl (HO_2) and organic peroxy (RO_2) radicals that react with nitrogen monoxide ~~nitrogen monoxide,~~ (NO)~~,~~ ~~to produce nitrogen dioxide, ~~nitrogen dio~~(NO_2)~~, i.e. RO_2^*~~, coupled with actinometry and other key trace gases measurements, have been used to test the current understanding of the fast photochemistry in the outflow of major population centres ~~(MPCs)~~. The~~All~~ measurements were made during the airborne campaign of the EMeRGe (**E**ffect of **M**egacities on the transport and transformation of pollutants on the **R**egional to **G**lobal scal**e**s) project in Europe on-board the **H**igh **A**ltitude **Lo**ng range research aircraft (HALO). The ~~on-board~~ measurements of RO_2^* on HALO were made using the in-situ instrument **P**eroxy **R**adical **C**hemical **E**nhancement and **A**bsorption **S**pectrometer (PeRCEAS). RO_2^* is to a good approximation the sum of peroxy radicals reacting with NO to produce NO_2. RO_2^* mixing ratios up to 120 pptv were observed in air masses of different origins and composition under different local actinometrical conditions during seven HALO research flights in July 2017 over Europe.

~~The range and variability of the RO_2^* measurements agree reasonably well with R~~radical production rates were estimated using knowledge of the photolysis frequencies and the RO_2^* precursor concentrations measured on-board, as well as the relevant rate coefficients. Generally, high RO_2^* were measured in air masses with high production rates. In the airmasses investigated RO_2^* is primarily produced ~~following~~ by the ~~photolysis of ozone (O_3)~~reaction of O^1D with water vapour, and the photolysis of nitrous acid (HONO~~,~~) and of the ox~~y~~igen~~t~~ated volatile organic compounds~~,~~ (OVOC ~~(~~, e.g. formaldehyde (HCHO), and glyoxal (CHOCHO))~~,~~ ~~and nitrous acid (HONO) in the airmasses investigated~~. Due to their short lifetime in most environments, the RO_2^* concentrations are expected to be in a photostationary steady state (PSS), i.e., it is assumed a balance between production and loss rates. The RO_2^* production and loss rates and the suitability of ~~photostationary steady state (~~PSS~~)~~ assumptions to estimate the RO_2^* mixing ratios and ~~the~~ variability ~~of RO_2^*~~ during the airborne observations ~~is~~ are ~~investigated~~discussed. The PSS assumption f~~r~~or RO_2^* is considered robust enough to calculate RO_2^* mixing ratio~~n~~s for most conditions encountered in the air masses measured. The similarities and discrepancies between measured and PSS calculated RO_2^* mixing ratios are discussed~~analysed stepwise~~. ~~The parameters, which predominantly control the RO_2^* mixing ratios under different chemical and physical regimes, are identified during the analysis.~~ The dominant ~~removal~~ terminating processes for~~of~~ RO_2^* in the ~~airmasses~~ pollution plumes measured up to 2000 m are the formation of nitrites and nitrates from radical reactions ~~loss of OH and RO through the reaction~~ with NO_x ~~during the radical interconversion~~. Above 2000 m, $HO_2 – HO_2$ and $HO_2 – RO_2$ reactions dominate the RO_2^* ~~loss reactions~~removal. RO_2^*

calculations ~~underestimated (< 20 %) the measurements~~ by the PSS analytical expression inside the pollution plumes probed often underestimated the measurements. The underestimation is attributed to the limitations of the PSS ~~assumption that RO$_2$* is in PSSnalysis~~ equation used for the analysis. In particular, ~~T~~this expression ~~to take~~ does not~~into~~ account ~~-~~for the ~~production~~ yields of RO$_2$* from the ~~through~~ oxidation and photolysis of volatile organic compounds, VOCs, and ~~the~~ OVOCs ~~not~~ other than measured during the EMeRGe ~~in Europe~~ research flights in Europe. In air masses with NO mixing ratios ≤ 50 pptv and low VOC/NO ratios, the overestimation of the RO$_2$* ~~measured~~ observed may be caused by the formation of H$_2$O and O$_2$ from OH and HO$_2$, being about 4 times faster than the rate of the OH oxidation reaction of the dominant OVOCs considered.
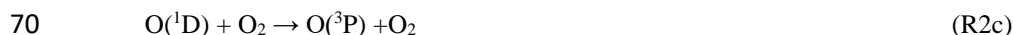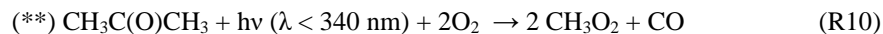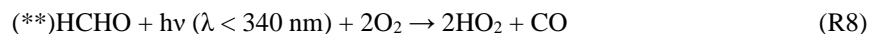
1. Introduction

Hydroperoxyl (HO$_2$) and organic peroxy (RO$_2$, where R stands for any organic group) radicals are reactive species that play a key role in the chemistry of the troposphere. In combination with the hydroxyl (OH) radical, HO$_2$ and RO$_2$ take part in rapid chemical processes that control the lifetime of many key trace constituents in the troposphere. Examples of key tropospheric processes involving HO$_2$ and RO$_2$ are as follows:

- the catalytic cycles~~,~~ which produce and destroy ozone (O$_3$)
- the generation of ~~key~~ inorganic acids, which are precursors of aerosol (e.g. sulphuric acid, H$_2$SO$_4$) and important chemical constituents (e.g. nitric acid, HNO$_3$) in both summer and winter smog
- the generation of organic acids; the production of hygroscopic hydrogen peroxide (H$_2$O$_2$) and organic peroxides (ROOH), which enter aerosol and cloud droplets
- the generation of organic peroxy nitrates (RO$_2$NO$_2$), peroxyacetyl nitrate (CH$_3$COO$_2$NO$_2$, PAN) and other summer smog constituents.

The abundance of HO$_2$ and RO$_2$ in the free troposphere has a non-linear and complex dependency on photochemistry, initiated by solar actinic radiation, and on the concentration of the precursors, such as carbon monoxide (CO), **v**olatile **o**rganic **c**ompounds (VOCs), and peroxides. It also strongly depends on the amounts of nitrogen monoxide (NO) and nitrogen dioxide (NO$_2$) due to the gas-phase reactions of NO and NO$_2$ with the OH and organic oxy (RO) radicals formed during the radical interconversion. The main production and loss processes of HO$_2$ and RO$_2$ in the troposphere are summarised as follows:

a) Production processes of HO$_2$ and RO$_2$ ~~through photolysis and oxidation by OH formed through photolysis~~

$$O_3 + h\nu \ (\lambda < 320 \text{ nm}) \rightarrow O(^1D) + O_2 \tag{R1}$$

$$O(^1D) + H_2O \rightarrow 2OH \tag{R2a}$$

$$O(^1D) + N_2 \rightarrow O(^3P) + N_2 \tag{R2b}$$

$$O(^1D) + O_2 \rightarrow O(^3P) + O_2 \tag{R2c}$$

$$HONO + h\nu \ (\lambda \leq 400 \text{ nm}) \rightarrow OH + NO \tag{R3}$$

$$H_2O_2 + h\nu \rightarrow 2OH \tag{R4}$$

$$OH + O_3 \rightarrow HO_2 + O_2 \tag{R5}$$

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$$OH + CO + O_2 \rightarrow HO_2 + CO_2 \quad\quad\quad (R6)$$

$$(*)OH + CH_4 + O_2 \rightarrow CH_3O_2 + H_2O \quad\quad\quad (R7)$$

$$(**)HCHO + h\nu\ (\lambda < 340\ nm) + 2O_2 \rightarrow 2HO_2 + CO \quad\quad\quad (R8)$$

$$(*)(**)CH_3CHO + h\nu\ (\lambda < 340\ nm) + 2O_2 \rightarrow CH_3O_2 + HO_2 + CO \quad\quad\quad (R9)$$

$$(**)\ CH_3C(O)CH_3 + h\nu\ (\lambda < 340\ nm) + 2O_2 \rightarrow 2\ CH_3O_2 + CO \quad\quad\quad (R10)$$

$$(**)CHOCHO + h\nu + 2O_2 \overset{M}{\rightarrow} 2HO_2 + 2CO \quad\quad\quad (R11)$$

(*) The CH$_3$ produced from the oxidation of CH$_4$ or the photolysis of VOCs further reacts with O$_2$ to form CH$_3$O$_2$. The net or overall reaction is used~~written,~~ because ~~since~~ the formation of CH$_3$O$_2$ is much faster than the CH$_3$ formation due to the high amount of O$_2$ present in the atmosphere.
(**) H and CHO formed through the VOC photolysis further react with O$_2$ to form HO$_2$. The net reaction is used,~~written~~ because ~~since~~ the formation of HO$_2$ is much faster than the H and CHO formation due to the high amount of O$_2$ present in the atmosphere.

$$VOCs + OH \rightarrow OH + HO_2 + RO_2\ \text{and other oxidation products} \quad\quad\quad (R12)$$

$$alkenes + O_3 \rightarrow OH + RO_2 + \text{other oxidation products} \quad\quad\quad (R13)$$

b) Loss processes of HO$_2$ and RO$_2$

$$HO_2 + HO_2 \rightarrow H_2O_2 + O_2 \quad\quad\quad (R14)$$

$$HO_2 + RO_2 \rightarrow ROOH + O_2 \quad\quad\quad (R15)$$

$$RO_2 + RO_2 \rightarrow ROOH + R_{C-1}CHO + O_2 \quad\quad\quad (R16a)$$

$$OH + HO_2 \rightarrow H_2O + O_2 \quad\quad\quad (R17)$$

$$OH + OH \overset{M}{\rightarrow} H_2O_2 \quad\quad\quad (R18)$$

$$OH + NO \overset{M}{\rightarrow} HONO \quad\quad\quad (R19)$$

$$OH + NO_2 \overset{M}{\rightarrow} HNO_3 \quad\quad\quad (R20)$$

$$OH + HONO \rightarrow H_2O + NO_2 \quad\quad\quad (R21)$$
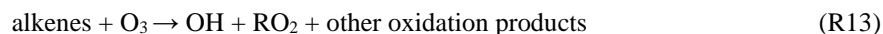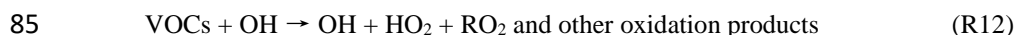
$$RO + NO \overset{M}{\rightarrow} RONO \quad\quad\quad (R22)$$

In addition, HO$_2$ and RO$_2$ ~~undergo~~ participate in radical interconversion ~~processes~~ reactions ~~through the following reactions~~such as:

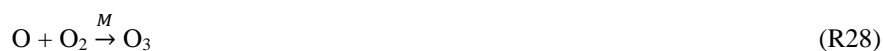$$RO_2 + RO_2 \rightarrow RO + RO + O_2 \quad\quad\quad (R16b)$$

$$HO_2 + NO \rightarrow OH + NO_2 \quad\quad\quad (R23)$$

$$HO_2 + O_3 \rightarrow OH + 2O_2 \quad\quad\quad (R24)$$

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$$RO_2 + NO \rightarrow RO + NO_2 \qquad\qquad\qquad\qquad (R25)$$

$$RO + O_2 \rightarrow R_{H-1}O + HO_2 \qquad\qquad\qquad\qquad (R26)$$

R23 and R25 ~~is~~ are ~~one~~two ~~of the most important reactions in the troposphere as ~~it~~they lead~~s~~ to O_3 formation via the reactions ~~through~~ R27 and R28. The rate of R22 in the atmosphere compared to that of R26 is negligible.

~~Provided that there is sufficient insolation to ensure rapid photochemical processing and all key species involved are measured~~known, ~~T~~the sum of HO_2 and RO_2 that react with NO to produce NO_2 can be estimated by assuming that the interconversion of NO to NO_2 reaches ~~from~~ a ~~photochemical~~ photostationary steady-state (PSS)~~. assumption~~ in which production and loss are to a good approxiamation equal~~mechanisms are equally important~~. ~~The $HO_2 + RO_2$ concentrations and mixing ratios are~~can be estimated using the PSS assumption for NO_2 ~~assuming the following~~ by considering the following ~~chemical mechanism~~reactions:

The PSS assumption for $[NO_2]$ in the following mechanism (R23 to R29) leads to Eq. 1

$$HO_2 + NO \rightarrow OH + NO_2 \qquad\qquad (R23)$$
$$RO_2 + NO + O_2 \rightarrow R_{H-1}O + NO_2 + HO_2 \qquad\qquad (R25 + R26)$$
$$NO_2 + h\nu \,(\lambda < 400 \text{ nm}) \rightarrow NO + O \qquad\qquad (R27)$$
$$O + O_2 \xrightarrow{M} O_3 \qquad\qquad (R28)$$
$$NO + O_3 \rightarrow NO_2 + O_2 \qquad\qquad (R29)$$

~~Assuming a PSS for [NO₂,]. this then leads to Eq. 1~~

$$[HO_2 + RO_2]_{PSS} = \frac{k_{NO+O_3}}{k_{NO+(HO_2+RO_2)}}\left(\frac{j_{NO_2}[NO_2]}{k_{NO+O_3}[NO]} - [O_3]\right) \qquad\qquad (Eq.1)$$

where ~~-~~j_{NO_2} is the photolysis frequency of NO_2 ; k_{NO+O_3} (1.9×10^{-14} cm^3 molecules^{-1} s^{-1} at 298K and 1 atm.) is the rate coefficient of the reaction of NO with O_3 and ~~-~~$k_{NO+(HO_2+RO_2)}$ is usually estimated for the most abundant peroxy radicals HO_2 and CH_3O_2 by assuming a 1:1 HO_2 to CH_3O_2 ratio and averaging the k_{NO+HO_2} (8.2×10^{-12} cm^3 molecules^{-1} s^{-1} at 298K and 1 atm.) and $k_{NO+CH_3O_2}$ (7.7×10^{-12} cm^3 molecules^{-1} s^{-1} at 298K and 1 atm.) rate coefficients ~~the weighted average rate coefficient assumed~~ for the reaction~~s~~ ~~of peroxy radicals~~ with NO. As noted by Parrish et al. (1986), the PSS assumption requires conditions with sufficient and stable solar irradiation, ensuring NO_2 stable photolysis rates (j_{NO_2}).

The PSS radical calculation made on the assumption of the NO_2 steady state is very sensitive to the accuracy of the NO_2 to NO ratio and the O_3 measurements. The comparison of $[HO_2 + RO_2]_{PSS}$ calculated using ~~-~~Eq.1 with ground-based (e.g. Ridley et al., 1992; Cantrell et al., 1997; Carpenter et al., 1998; Volz-Thomas et al., 2003), and airborne measurements, has shown in the past different degree~~s~~ of agreement. The underestimations and overestimations found in air masses with different chemical compositions are not well understood. For~~In~~ the case of airborne measurements, the PSS calculation generally overestimates that measured peroxy radicals (Cantrell et al., 2003a, 2003b). The differences observed could not be attributed to systematic changes in NO, altitude, water vapour and temperature, although these variables are often correlated.

Ground-based (Mihelcic et al., 2003; Kanaya et al., 2007, 2012; Elshorbany et al., 2012; Lu et al., 2012, 2013; Tan et al., 2017, 2018; Whalley et al., 2018, 2021; Lew et al., 2020) and airborne (Crawford et al., 1999; Tan et al., 2001; Cantrell et al., 2003b) measurements have also been compared with model simulations of HO_2 and RO_2. The discrepancies encountered depend upon the chemical composition of the air mass and the chemical mechanisms and constraints used in the model simulations. Tan et al., 2019 and Whalley et al., 2021 reported experimental radical budget calculations using~~based on~~ the published reaction rate coefficients of the ~~fundamental~~ reactions (R1 to R26), which control~~ling~~ OH, HO_2 and RO_2 in the lower troposphere, and the~~and~~ ground-based measurements of all relevant reactants and photolysis frequencies.

In this study, a similar approach has been used to calculate the amount of peroxy radicals in the air masses measured on-board of the **H**igh **A**ltitude **Lo**ng range (HALO) research aircraft over Europe during the first campaign of the EMeRGe (**E**ffect of **Me**gacities on the transport and transformation of pollutants on the Regional to Global scal**e**s) project. The available on-board measurements of RO_2^* are defined as the total sum of OH, RO and peroxy radicals ~~reacting with NO to produce NO₂~~ (i.e., $RO_2^* =$ OH $+ \sum RO + HO_2 + \sum RO_2$, where RO_2 are the organic peroxy radicals ~~reacting~~ producing NO_2 in their reaction with NO ~~to produce NO₂~~). As~~Since~~ the amount of OH and RO is much smaller, RO_2^* to a good approximation is the sum of HO_2 and those RO_2 radicals that react with NO to produce NO_2. For the calculation, RO_2^* is assumed to be in PSS, and an analytical expression is developed with a manageable degree of complexity to estimate the concentration and mixing ratios of RO_2^*. The simultaneous on-board measurements of trace gases and photolysis frequencies are used to constrain the estimate of the RO_2^* concentration.

In contrast to other experimental deployments, the concentrations and/or mixing ratios of the majority of the key species involved in reactions R1 to R26 were continuously measured on-board HALO during the EMeRGe campaign. This enables the use of a large number of measurements to constrain the ~~minimises the number of assumptions required for~~ the PSS calculation~~s~~ of RO_2^*. Consequently, this data set provides an excellent opportunity to gain ~~a~~ deeper insight into the source and sink reactions of RO_2^* and the applicability of the PSS assumption for the different pollution regimes and related weather conditions in the free troposphere.

2.   EMeRGe field campaign in Europe

The overarching objective of the EMeRGe project is to test and improve the current understanding of the photochemical and heterogeneous processing of pollution outflows from **m**ajor **p**opulation **c**entres (MPCs) and their impact on the atmosphere. Two **i**ntensive **o**bservational **p**eriods (IOP) were carried out to investigate selected European and Asian MPC outflows. The European IOP took place from 10 to 28 July 2017 (http://www.iup.uni-bremen.de/emerge/home/home.html). An extensive set of in-situ and remote-sensing airborne measurements of trace gases and aerosol particles were made on-board the HALO aircraft (see www.halo-spp.de) along flight tracks in the lower layers of the troposphere from northwest Europe to the Mediterranean region.

During EMeRGe in Europe, HALO made ~~carried out~~ a total of 53 flight hours distributed over seven flights to investigate the chemical composition of the outflows from the target MPCs: London, Paris, Benelux/ Ruhr metropolitan area, Po Valley, and urban agglomerations such as Rome, Madrid, and Barcelona. The flight tracks are shown in Fig. 1. All HALO flights started from the HALO base at the DLR in Oberpfaffenhofen, southwest of Munich, Germany. To achieve the scientific goals, 60 % of the flights flew at altitudes~~were carried out~~ below 3000 m. Vertical profiles of trace constituents were typically made by keeping the HALO altitude constant at different at ~~in three stable~~ flight levels upwind and downwind of the target MPCs. The flights are named E-EU-FN, where E stands for EMeRGe, EU for Europe, and FN is the two-digit flight number. More details about the EMeRGe

170 IOP in Europe and the set of instruments deployed on-board the HALO aircraft are described elsewhere (Andrés Hernández et al., ~~2021~~2022).
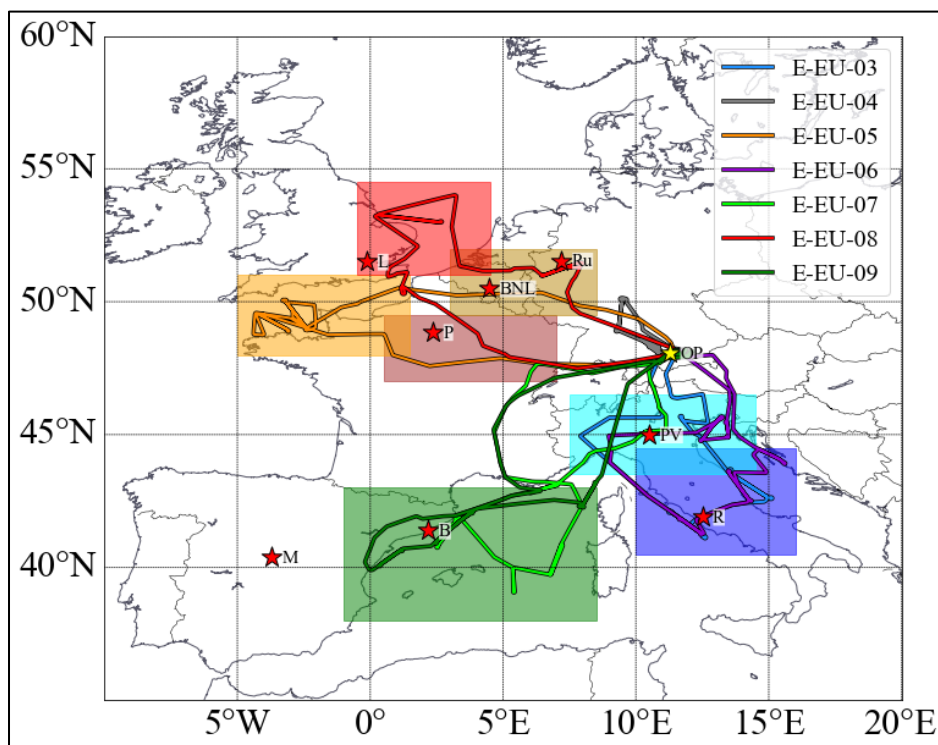
Figure 1: The research flight tracks made by HALO during the EMeRGe-Europe campaign on 11, 13, 17, 20, 24, 26 and 28 July 2017 (E-EU-03 to E-EU-09, respectively, colour coded). MPC target areas are colour coded by shading, and the targeted

175 locations/regions are marked with red stars, M: Madrid, B: Barcelona, P: Paris, L: London; BNL: BeNeLux; Ru: Ruhr area; PV: Po Valley, R: Rome. The location of the HALO base at the DLR in Oberpfaffenhofen, Germany (OP) is indicated by a yellow star.

3. PeRCEAS and other instruments on-board HALO during EMeRGe

The RO_2^* measurements on-board the HALO research aircraft during EMeRGe were made using the **Pe**roxy **R**adical **C**hemical **E**nhancement and **A**bsorption **S**pectrometer (PeRCEAS). PeRCEAS combines the **Pe**roxy **R**adical **C**hemical **A**mplification

180 (PeRCA) and **C**avity **R**ing-**D**own **S**pectroscopy (CRDS) techniques in a dual-channel instrument. Each channel has a separate chemical reactor and detector, which operate alternatively in both background and amplification modes ~~, i.e. without and with the addition of CO,~~ to account for the rapid background variations during airborne measurements. In both modes NO is continuously added to the air sampled at the reactor, while CO is only added in the amplification mode to initiate the chain conversion of RO_2^* into NO_2. In the amplification mode, the sum of the NO_2 produced from ambient RO_2^* through the chain reaction, the ambient

185 NO_2, the NO_2 produced from the ambient O_3 – NO reagent gas reaction and the NO_2 produced in the inlet from any other sources (e.g. thermal decomposition of PAN) is measured. In the background mode, the sum of the ambient NO_2, the NO_2 produced from the ambient O_3 – NO reagent gas reaction and NO_2 produced in the inlet from any other sources is measured. The RO_2^* is retrieved by dividing the difference in NO_2 concentration (ΔNO_2) between amplification and background mode by the conversion efficiency of RO_2^* to NO_2, which is referred to as eCL (**e**ffective **c**hain **l**ength). The PeRCEAS instrument and its specifications have been

190 described in detail elsewhere (Horstjann et al., 2014, George et al., 2020).

The two chemical reactors for sampling the ambient air are part of the **DU**al channel **A**irborne peroxy radical Chemica**l** Amplifi**er** (DUALER) inlet installed inside a pylon located on the outside of the HALO fuselage. During the EMeRGe campaign in Europe, a reagent gas mixing ratio of 30 ppmv NO ([NO] = 1.46 $\times10^{14}$ molecules cm^{-3} at 296 K, 200 mbar) and of 9 % CO ([CO] = 4.4

$\times 10^{17}$ molecules cm^{-3} at 296 K, 200 mbar) were added to the sample flow for the chemical conversion of RO$_2^*$ to NO$_2$. The DUALER inlet was operated at an internal pressure of 200 mbar to achieve stable chemical conversion. The HO$_2$ and RO$_2$ detection sensitivity depends on the rates of loss of HO$_2$ and RO$_2$ by the R19 and R22 reactions. The latter depend on the concentration of the reagent gas NO added ~~concentration due to~~ and the reactions rate coefficients, where k_{22} ~~of reaction R22 being~~is larger than ~~that for~~ kR_{19}. The average eCL for a 1:1 HO$_2$ to CH$_3$O$_2$ mixture under ~~these~~ the DUALER conditions during the campaign in Europe was determined to be 50 ± 8 from laboratory calibrations, where the error is the $\pm 1\sigma$ standard deviation estimated from the reproducibility of the experimental determinations. ~~For the measurement conditions used during the IOP in Europe, t~~Likewise, the ratio $\alpha = \text{eCL}_{CH_3O_2}/\text{eCL}_{HO_2}$ ~~is~~was determined to be 65% for the measurement conditions (George et al., 2020). The values obtained from calibrations before and after the campaign agreed within their experimental errors.

Although the DUALER pressure is kept constant below the ambient pressure, variations in dynamical pressure > 10 mbar during the flight ~~can~~change the residence time and induce turbulences inside the inlet (Kartal et al., 2010; George et al., 2020). These may lead to different physical losses of radicals before amplification and affect the eCL. In the measurements presented in this study, variations in dynamical pressure of this magnitude were only encountered during flight level changes of the aircraft. When used during the analysis, these data sets are either excluded or flagged (P_flag). The effect of the ambient air humidity on eCL (Mihele and Hastie, 1998; Mihele et al., 1999; Reichert et al., 2003) has been accounted for by a calibration procedure reported in George et al. (2020). The [H$_2$O] in the DUALER inlet was lower than 1×10^{17} molecules cm^{-3} for 60 % of measurements during EMeRGe in Europe, for which the eCL$_{wet}$ = 76 % of eCL$_{dry}$. At the highest humidity observed during the campaign, i.e., [H$_2$O]$_{inlet}$ = 2×10^{17} molecules cm^{-3}, the eCL$_{wet}$ is 55 % of eCL$_{dry}$ (see Fig. S1 in the supplementary information).

In addition to the measurement of RO$_2^*$ from PeRCEAS, other in-situ and remote-sensing measurements and basic aircraft data from HALO are used in this study. Details of the corresponding instruments are summarised in Table 1. The remote sensing instruments used on HALO during EMeRGe were the mini Differential Optical Absorption (minDOAS) and the Heidelberg Airborne Imaging DOAS Instrument (HAIDI). ~~Concerning the data obtained by the remote sensing instruments,~~ The miniDOAS observes the atmosphere using six telescopes: two being optimised for the ultraviolet, two for the visible, and two for the near infrared. Three telescopes observe in nadir viewing and three in limb viewing. The three limb scanning telescopes point to the starboard side perpendicular to the aircraft fuselage axis. They are rotated to compensate for roll relative to the horizon. A variant of the DOAS retrieval technique uses least square fitting of the measured and radiative transfer modelled absorption along the line of sight to retrieve ~~the miniDOAS retrieves~~ the differential Slant Column Density (dSCD) of the target gas and a scaling reference gas. The latter is the dimer of molecular oxygen (O$_4$). As the vertical profile of the concentrations of O$_2$ and thus O$_4$ are known then the mixing ratios of the target gas at the flight altitude is obtained from the target gas and O$_4$ dSCDs ~~towards the horizon at the flight altitude. From this, mixing ratios of the targeted gas within the line of sight is estimated using RT modelling~~ (for more details see Stutz et al., 2017; Hüneke et al., 2017; Kluge et al., 2020; Rotermund et al., 2021). The HAIDI ~~instrument~~nadir observations are used to ~~retrieves~~ dSCDs below the aircraft. The dSCDs from HAIDI are then converted to mixing ratios using knowledge of the aircraft altitude and the corresponding geometric Air Mass Factor (AMF), calculated by a radiative transfer model under a well-mixed NO$_2$ layer assumption. As a result of this assumption, the calculated mixing ratios for HAIDI -target gases are lower limits and ~~close~~ similar to the actual values while flying within and close to a well-mixed boundary layer. In spite of ~~Despite~~ the differences in sampling volume and temporal and spatial resolution ~~in the~~between the in-situ and remote sensing measurement techniques, the concentration of the gas HCHO measured by both techniques were in good agreement and the concentrations of the NO$_2$ (remote sensing) and NO$_y$ (in situ) were consistent~~common and related species obtained are in reasonable agreement~~ (for more details see Schumann, 2020).

Table 1: List of the airborne measurements and instrumentation used in this study. PeRCA: Peroxy Radical Chemical Amplification; CRDS: Cavity Ring-Down Spectroscopy; PTR-MS: Proton-Transfer-Reaction Mass Spectrometer; AT-BS: Adsorption Tube and Bag air Sampler; TD-GC-MS: Thermal Desorption Gas Chromatography and Mass Spectrometry; DOAS: Differential Optical Absorption Spectrometry; Univ: University; KIT: Karlsruher Institut für Technologie; DLR: Deutsches Zentrum für Luft- und Raumfahrt; IPA: Institut für Physik der Atmosphäre; FZ: Forschungszentrum; FX: Flugexperimente.

Trace gas-in situ measurements				
Species/parameters	Acronym	Institution	Technique/Instrument	Reference
$RO_2^* = HO_2 + \sum RO_2$	PeRCEAS	Univ. Bremen	PeRCA + CRDS	George et al., 2020
OVOC	HKMS	KIT Karlsruhe	PTR-MS	Brito and Zahn, 2011
O_3	FAIRO	KIT Karlsruhe	UV-Photometry/ Chemiluminescence	Zahn et al., 2012
O_3, CO	AMTEX	DLR-IPA	UV-Photometry/ VUV-Fluorimetry	Gerbig et al., 1996
NO, NO_y	AENEAS	DLR-IPA	Chemiluminescence/ Gold converter	Ziereis et al., 2004
CO_2, CH_4	CATS	DLR-IPA	CRDS	Chen et al., 2010
Trace gas- remote sensing measurements				
Species/parameters	Acronym	Institution	Technique/Instrument	Reference
NO_2, HONO, CH_2O, $C_2H_2O_2$, $C_3H_4O_2$	miniDOAS	Univ. Heidelberg	DOAS / UV-nIR; 2D optical spectrometer	Hüneke et al., 2017
NO_2	HAIDI	Univ. Heidelberg	DOAS / 3x2D-imaging spectrometers	General et al., 2014
Other parameters				
Species/parameters	Acronym	Institution	Technique/Instrument	Reference
Spectral actinic flux density (up/down) Photolysis frequencies	HALO-SR	FZ Jülich	CCD spectro- radiometry	Bohn and Lohse, 2017
Basic aircraft data	BAHAMAS	DLR-FX	various	Mallaun et al., 2015

4. Results and discussion

4.1. Airborne RO_2^* measurements during EMeRGe in Europe

RO_2^* mixing ratios up to 120 pptv were measured during the campaign, as shown in Fig. 2. Typically, the highest RO_2^* mixing ratios were observed below 3000 m over Southern Europe. ~~This is attributed to the higher insolation and temperatures favouring the rapid production of RO_2^* from the photochemical oxidations of CO and VOCs.~~

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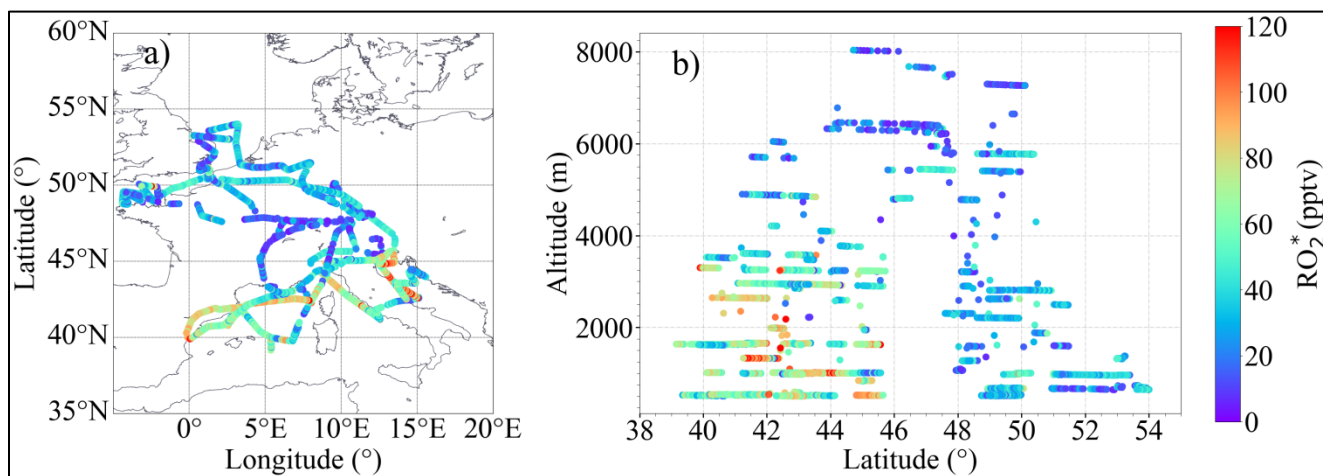
Figure 2: RO_2^* measured during EMeRGe-Europe: a) as a function of longitude and latitude, b) as a function of latitude and altitude.

The origin and thus the composition of the air sampled during the seven flights over Europe were different and heterogeneous. Typically, the air masses measured were influenced by emissions from MPCs and their surroundings, and sometimes by biomass burning transported over short or long distances. The concentration and mixing ratio of RO_2^* rather depends on the insolation and the chemical composition of the air ~~masses~~ probed, particularly on the abundance of RO_2^* precursors, than on the origin of the air masses. ~~Provided that insolation conditions and a sufficient number of key participating precursors are comparable, the air mass origin is irrelevant for calculating RO_2^* concentrations and mixing ratios. This is because the~~ Since RO_2^* ~~concentration is~~ are controlled by fast chemical and photochemical processes, the air mass origin and trajectory are not used in the calculation of RO_2^* concentrations and mixing ratios but are of interest as the source of . RO_2^* precursors. Thus, the RO_2^* variability and its production rates provide valuable insight into the photochemical activity of the air masses probed.

Changes in RO_2^* as a function of latitude and altitude, as shown in Fig. 2, confirm the heterogeneity of the photochemical activity in the air masses probed. Figure 3 shows the RO_2^* vertical profiles averaged for the EMeRGe flights over Europe in 500 m altitude bins. The error bars are standard errors (i.e. ± 1σ standard deviation of each bin). The vertical profiles may be biased as the higher altitudes have fewer measurements than those below 3000 m, as mentioned in section 2. The vertical profiles are a composite from averaging ~~different~~ flights with legs carried out at different longitude and latitudes, and are only shown to summarise the variability in the composition of the air masses measured during the campaign.
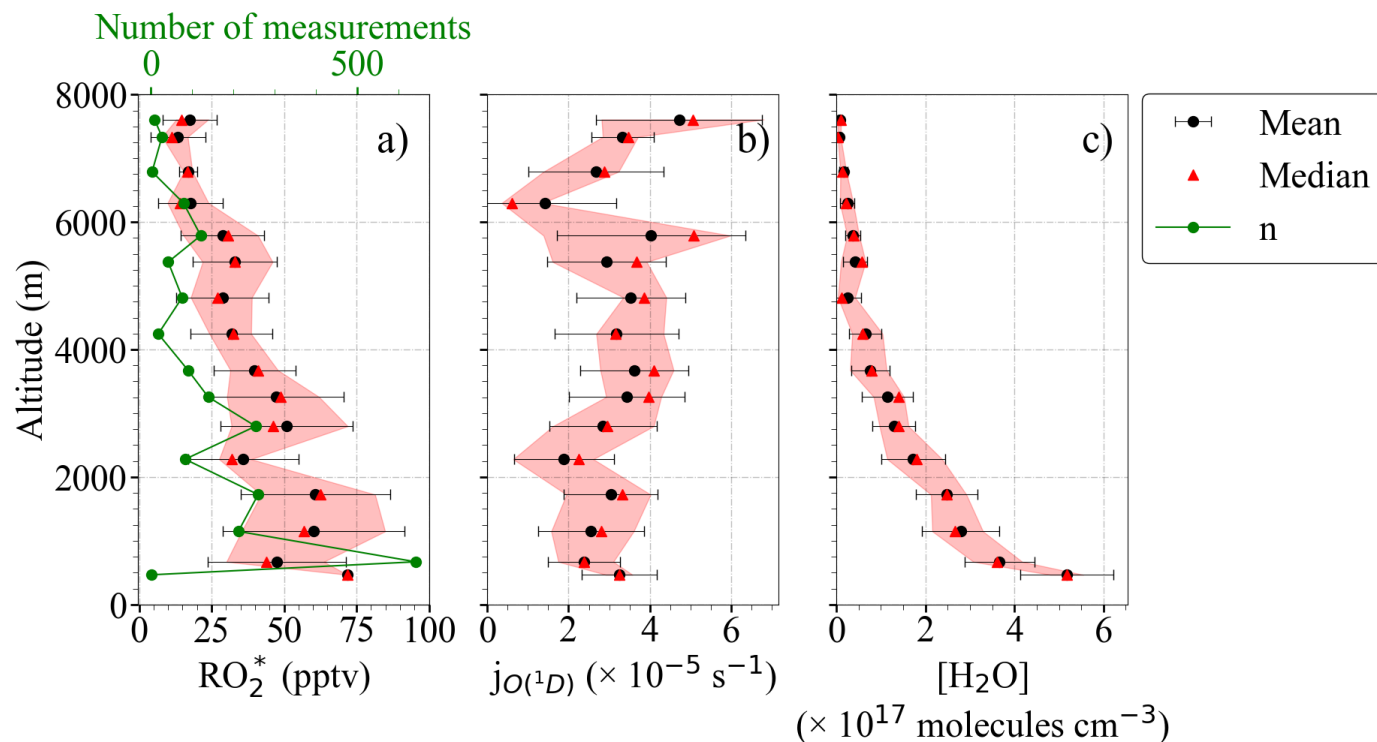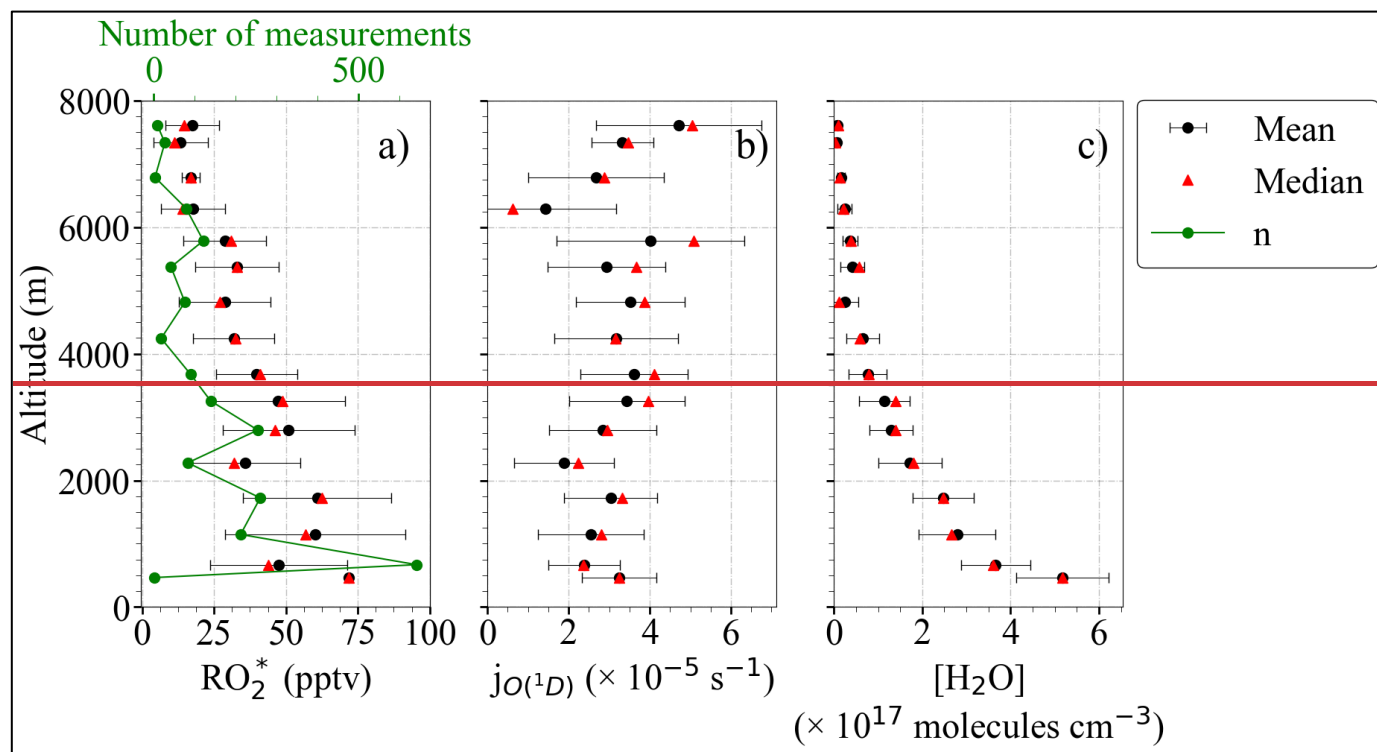
Figure 3: Composite average vertical profiles of a) RO_2^*, b) $j_{O(^1D)}$ and c) $[H_2O]$ observations. The measurements are binned over 500 m altitude. The error bars are the ~~standard errors (i.e.~~ $\pm 1\sigma$ standard deviation of each bin~~)~~. Median values (red triangles) the interquartile 25-75% range (red-shaded area) and the number of individual measurements, n, for each bin (in green) are additionally plotted.

~~Differences between mean and median values indicate less RO_2^* variability in the air masses probed above 3000 m.~~ Most of the EMeRGe measurements below 2000 m were carried out in the outflow of MPCs, which are expected to contain significant amounts of RO_2^* precursors. HALO flew at the lowest altitudes during flight legs over the English Channel, the Mediterranean and the

North Sea. The H_2O concentration in the air masses decreases steadily with altitude as expected. The higher relative variability in H_2O observed at 3000 m and the increase at 5000 m is associated with measurements under stormy conditions, often over the Alps.

4.2. RO_2^* production rates

275 The ~~total~~ rate of production ~~rate~~ of ~~OH and~~ RO_2^* ~~($P_{OH+HO_2+RO_2}$) can be estimated~~ from the reactions R1 to R13 ~~as follows~~is given by:

$$P_{RO_2^-} = 2jO_D^1[O_3]\frac{k_{O_D^1 + H_2O}\,[H_2O]}{k_{O_D^1 + H_2O}\,[H_2O] + k_{O_D^1 + O_2}\,[O_2] + k_{O_D^1 + N_2}\,[N_2]} + j_{HONO}\,[HONO] + 2j_{H_2O_2}\,[H_2O_2] + 2\sum_i j_i\,[OVOC_i] +$$

$$\sum_j k_{OH+VOC_j}[OH][VOC_j] + \sum k_{O_3+alkenes_k}[O_3][alkenes_k] \qquad\qquad\qquad\qquad \text{(Eq. 2)}$$

where OVOC stands for oxygenated volatile organic compounds.

280 In this ~~work~~study, Eq. 2 has been applied to the measurements taken within the EMeRGe campaign in Europe. There were no H_2O_2 measurements available for EMeRGe. However, ~~from~~ the results reported by Tan et al. (2001), indicate that the rate of OH production from the H_2O_2 photolysis is not ~~become~~ significant except when ~~at low~~ NO_x ~~conditions~~is low. To be more precise, for conditions having NO < 50 ppt, the partitioning of HO_x is strongly shifted to HO_2. HO_2 then predominantly reacts with itself or RO_2 to form peroxides, which can in turn photolyse. For conditions with NO > 50 pptv the rates of reactions of HO_x with NO_x are 285 faster than those of HO_2 with HO_2 and RO_2. As the NO mixing ratio was higher than 50 pptv in 75 % of the air masses probed in Europe~~Since the $[NO_x] > 8 \times 10^{12}$ molecules cm^{-3} for 60 % of the measurements during the IOP, as a first approximation~~, the ~~production~~ rate ~~of OH from H_2O_2~~of the photolysis of H_2O_2 ~~is~~was as a first approximation assumed not to be ~~negligible~~ significant source of OH for the EMeRGe dataset considered in this study.

Formaldehyde (HCHO), acetaldehyde (CH_3CHO), acetone, ($CH_3C(O)CH_3$), and glyoxal (CHOCHO) were the OVOCs measured 290 in EMeRGe forming directly radicals through photolysis. They are produced in the photolysis and oxidation of VOCs and are likely the ~~Similarly, the VOC photolysis was assumed to dominate the RO_2^* production over the oxidation by OH and ozonolysis of alkenes. The~~ most abundant and reactive ~~oxygenated volatile organic compounds~~ (OVOCs) present. ~~measured have been taken~~In this study they were assumed to be the dominant ~~as a surrogate for the sum of~~ VOCs. in the air masses probed.

There were no measurements of alkenes provided in EMeRGe. Consequently the ozonolyis term in Eq. 2 was not included in the 295 analysis.

The above~~e~~ assumptions lead to Eq. 3, which ~~estimates~~calculates the RO_2^* production rate ($P_{RO_2^*}$) for the EMeRGe measurements as follows:

$$P_{RO_2^*} = 2j_{O(^1D)}[O_3]\frac{k_{O_D^1 + H_2O}\,[H_2O]}{k_{O_D^1 + H_2O}\,[H_2O] + k_{O_D^1 + O_2}\,[O_2] + k_{O_D^1 + N_2}\,[N_2]} + j_{HONO}\,[HONO] + 2j_{HCHO}\,[HCHO] + 2j_{CH_3CHO}\,[CH_3CHO] +$$

$$2j_{CH_3C(O)CH_3}[\,CH_3C(O)CH_3] + 2j_{CHOCHO}\,[CHOCHO] \qquad\qquad\qquad\qquad \text{(Eq.3)}$$

300 ~~Eq. 3 yields the rate of production of RO_2^* molecules.~~ The production rate of RO_2^* molecules can be expressed in units of mixing ratio of RO_2^* by dividing with the air concentration at each altitude, ~~estimated~~ calculated from the pressure and temperature measurements (for the vertical profile and the latitudinal distribution of $P_{RO_2^*}$ see Fig. S2 and S3 in the supplementary information ~~.~~).

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Figure 4 shows the composite averaged vertical profile of all measured RO_2^* mixing ratios colour-coded with the ~~estimated~~ calculated $P_{RO_2^*}$. ~~Small circles show the 1-minute measurements binned for $P_{RO_2^*}$ up to 0.8 pptv s^{-1} in 0.1 pptv s^{-1} intervals. The production rates above 0.8 pptv s^{-1} are binned to the 0.8 pptv s^{-1} bin. Larger circles in the figure result from further binning the small circles over 500 m altitude steps. The error bars are the standard errors for each altitude bin.~~ For the sake of representativeness and comparability, the number of measurements in each altitude bin is shown in Fig. 4b. ~~The~~ Hhigher RO_2^* mixing ratios observed below 4000 m are typically associated with $P_{RO_2^*} \geq 0.4$ pptv s^{-1}. Above 4000 m both $P_{RO_2^*}$ and RO_2^* start to decrease with altitude, as expected. This is related to the decrease in H_2O and other radical precursor concentrations with altitude, as detailed in Fig. 5 and Fig. 6. In previous airborne campaigns at various parts of the world, RO_2^* vertical distributions showed a local maximum between 1500 and 4000 m, as reported by Tan et al. (2001), Cantrell et al. (2003a, 2003b), and Andrés-Hernández et al. (2009). In the present work, this local maximum is more evident for measurements with $P_{RO_2^*} \geq 0.5$ pptv s^{-1}.


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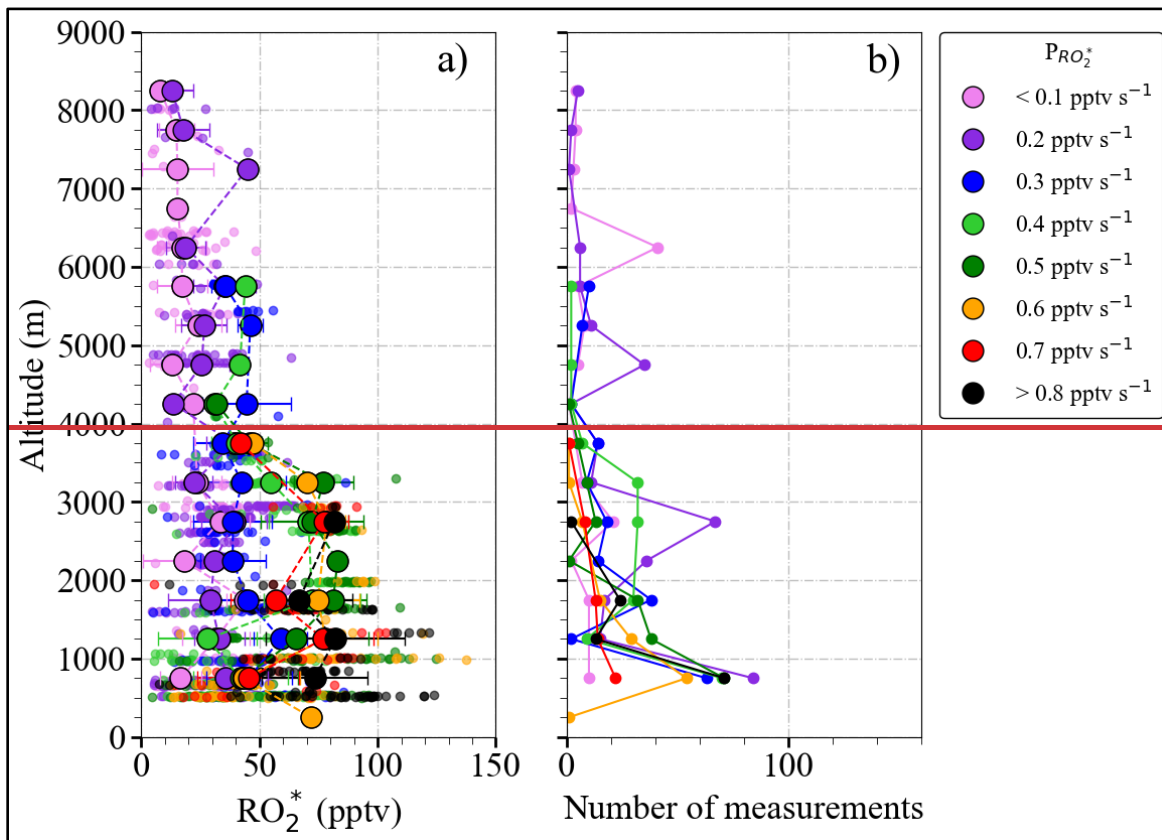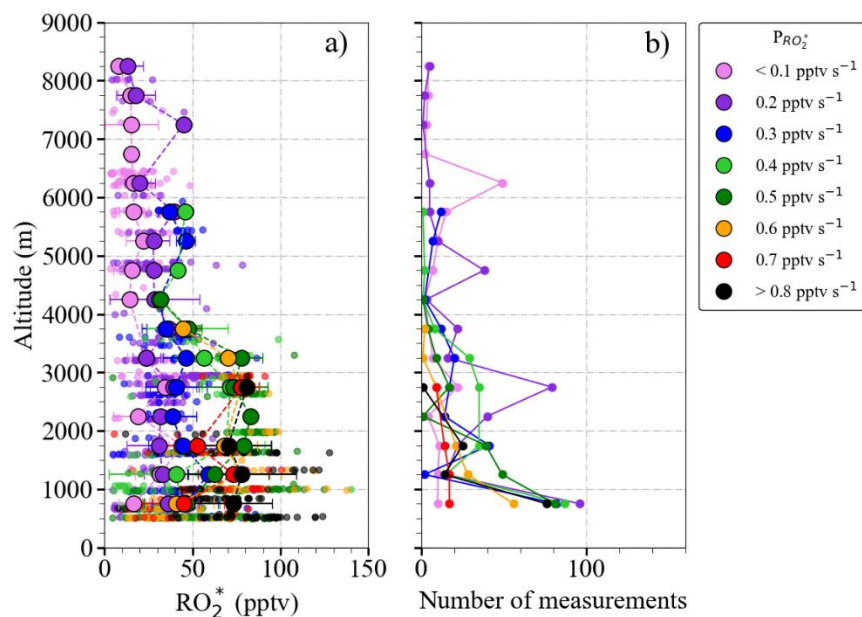
Figure 4: a) Composite averaged vertical distribution of measured RO_2^* colour-coded according to the value of $P_{RO_2^*}$, b) the number of measurements in each altitude bin. Small circles are 1-minute individual measurements binned with $P_{RO_2^*}$ values in 0.1 pptv s^{-1} intervals. Larger circles result from a further binning over 500 m altitude steps. All the production rates below 0.1 pptv s^{-1} and above 0.8 pptv s^{-1} are binned to 0.1 pptv s^{-1} and 0.8 pptv s^{-1}, respectively. The error bars are the standard deviation for each altitude bin.

Figure 5 shows the fractional contribution of the production rate from each radical precursor reaction included in Eq. 3 as a function of altitude. The data are classified into three groups according to the rate of change of ~~production~~ of the RO_2^* mixing ratio $P_{RO_2^*}$ < 0.07 pptv s^{-1} (5a), 0.07< $P_{RO_2^*}$ < 0.8 pptv s^{-1} (5b), and $P_{RO_2^*}$ > 0.8 pptv s^{-1} (5c) to show the lowest, most common, and highest ranges, respectively, encountered during the ~~IOP~~campaign. For 89 % of the measurements, $0.07 < P_{RO_2^*} < 0.8$ pptv s^{-1} applies, while the rest of the data are equally distributed in the other two $P_{RO_2^*}$ ranges. The data in each group are always binned over 500 m when available.

Typically, the high amount of H_2O in the air masses probed ~~results in the O_3 photolysis and subsequent~~ leads to the reaction of O^1D with H_2O (R1-R2a) ~~and is~~being the highest RO_2^* radical production rate ($\geqslant$ 50 %) below 4000 m. As the amount of H_2O reduces with altitude, the relative contribution from O_3 photolysis decreases. Above 4000 m, HCHO, HONO, and CHOCHO photolysis contributions range between 20 % to 40 %, 2.5 % to 30 %, and 5 % to 25 %, respectively. The HCHO contribution increases up to 80% during measurements above 6000 m. The contributions of CH_3CHO and $CH_3C(O)CH_3$ photolysis are, in contrast, practically negligible (< 5 %).
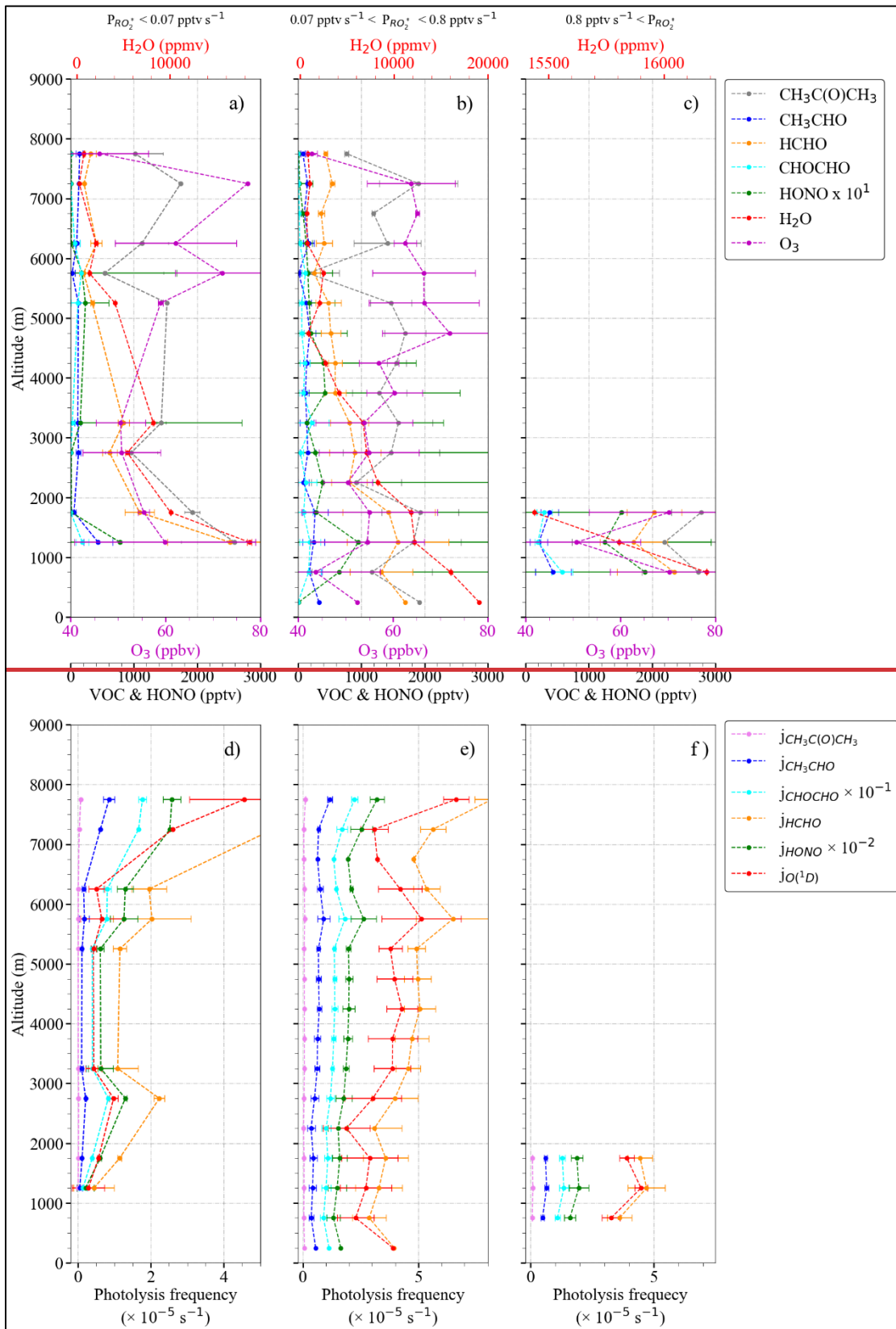
~~-~~The vertical changes of the precursor mixing ratios and photolysis frequencies used to calculate $P_{RO_2^*}$ in Fig. 5 are shown in Fig. 6a to 6f. $P_{RO_2^*}$ < 0.07 pptv s^{-1} is associated with measurements under cloudy conditions, towards sunset where the photolysis frequencies are low, or at altitudes above 5000 m in air masses with a low amount of RO_2^* precursors. $P_{RO_2^*}$ > 0.8 pptv s^{-1} are found for air masses, measured below 2000 m in the outflow of MPCs over the sea, for conditions having sufficient insolation ($j_{O(^1D)}$ > 3×10^{-5} S^{-1}) and a high content of RO_2^* precursors (HCHO > 1000 pptv and HONO >100 pptv). The increase in the photolysis frequencies as a function of altitude is concurrent with decreases in precursor concentrations. As a result, the $P_{RO_2^*}$ do not significantly vary with altitude in the air masses investigated.

In previous airborne campaigns, Tan et al. (2001) and Cantrell et al. (2003b) reported a reduction of the fractional contribution of

340 the reaction of $O(^1D)$ with H_2O as the $P_{RO_2^*}$ value decreases. At very low $P_{RO_2^*}$ values (< 0.03 pptv s^{-1}), the sum of all other production terms exceeded the fraction from the $O(^1D) + H_2O$ term. For these conditions, H_2O_2 and VOCs photolysis dominated the $P_{RO_2^*}$. For~~In the case of~~ the EMeRGe data set in Europe, only 6 % of~~-~~ $P_{RO_2^*}$ are below 0.06 pptv s^{-1}.
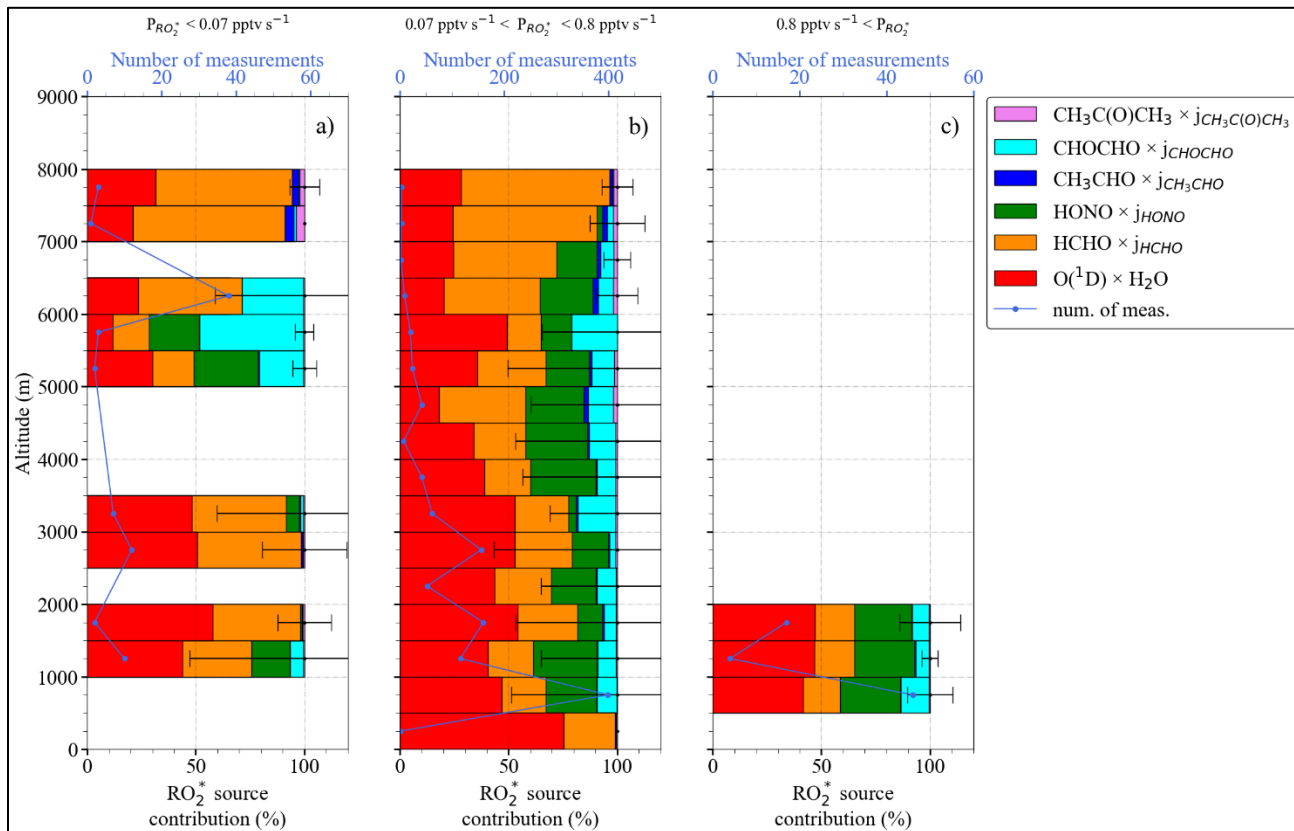
Figure 5: Total $P_{RO_2^*}$ and fractional precursor contributions estimated using Eq. 5 as a function of altitude, for:~~-~~ a) $P_{RO_2^*}$ ~~-~~< 0.07

345 pptv s^{-1}, b) 0.07 pptv s^{-1} < $P_{RO_2^*}$ < 0.8 pptv s^{-1}, and c) $P_{RO_2^*}$ > 0.8 pptv s^{-1}. Note the different scales in the number of measurements.
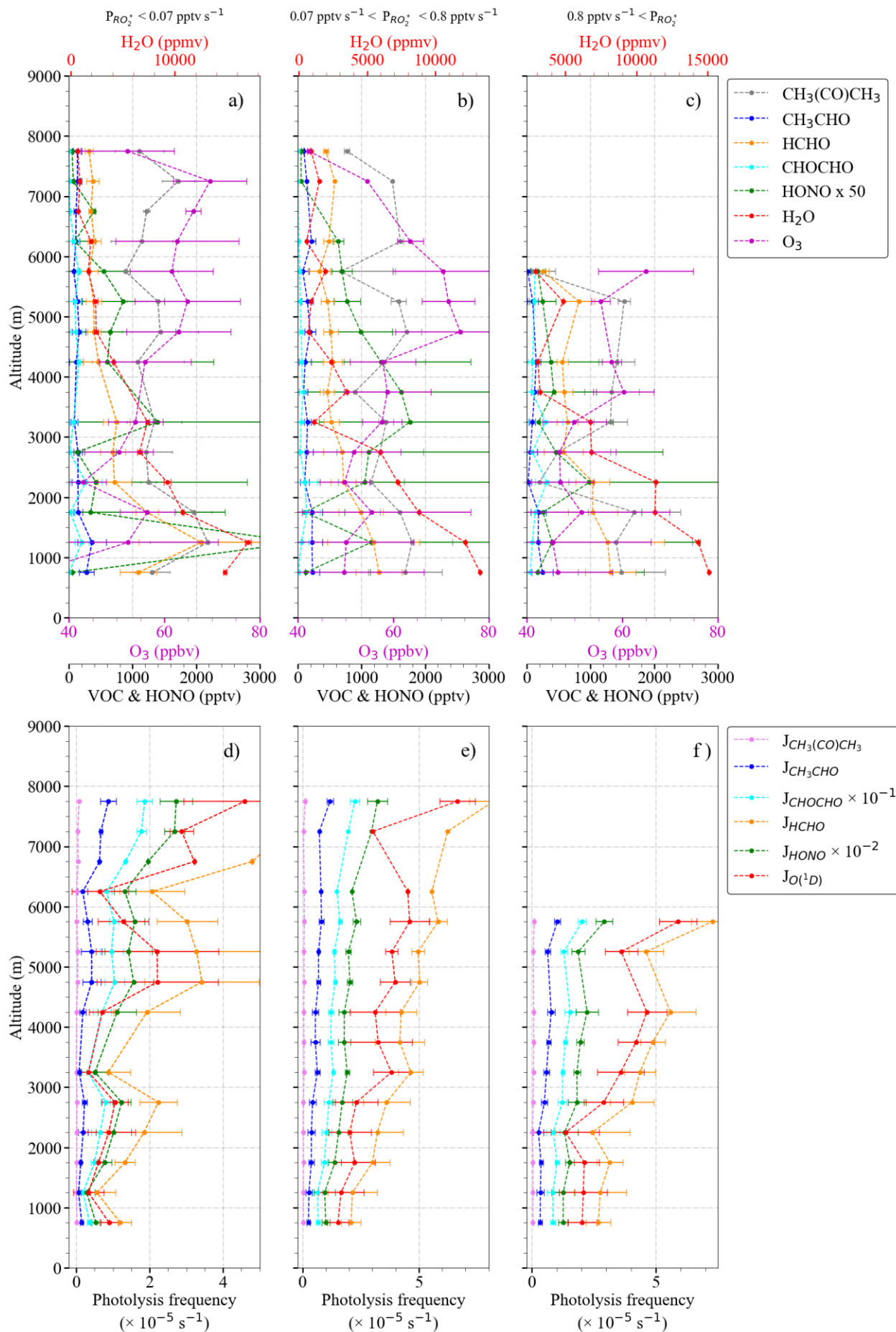
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4.3. PSS estimation of ~~the~~ RO_2^* mixing ratios

Under most ambient conditions in the troposphere, the RO_2^*, ~~which to a first approximation is the sum of HO_2 and RO_2, which react with NO to produce an RO_2,~~ are short-lived, and the chemical lifetime of RO_2^* is much shorter than the chemical transport time into and out of an air mass being probed. Consequently, pseudo-steady-state conditions prevail, and the radical production
355 and ~~loss~~ destruction rates are balanced:

$$P_{HO_2+RO_2^*} = LD_{HO_2+RO_2^*} \qquad\qquad (Eq.4)$$

~~If the interconversion reactions between OH, RO, HO_2 and RO_2~~ The (R5 to R7, R12, R16b, and R23 to R26) are interconversion reactions between OH, RO, HO_2 and RO_2 and do consequently occur without radical losses~~,~~. ~~then the radical number~~
360 ~~densities~~concentrations are calculated by solving Eq. 4. Solving Eq. 4 leads to Eq. 5 i~~If~~ ~~the~~ RO_2^* – RO_2^* reactions are assumed to be the dominant radical ~~loss~~terminating processes~~, Eq. 4 leads to Eq. 5~~.

$$2j_{O(^1D)}[O_3]\frac{k_{O_D^1+H_2O}[H_2O]}{k_{O_D^1+H_2O}[H_2O]+k_{O_D^1+O_2}[O_2]+k_{O_D^1+N_2}[N_2]} + j_{HONO}[HONO] + 2j_{HCHO}[HCHO] + 2j_{CH_3CHO}[CH_3CHO]$$
$$+ 2j_{CH_3C(O)CH_3}[CH_3C(O)CH_3] + 2j_{CHOCHO}[CHOCHO] = k_{RO_2^*+RO_2^*}[RO_2^*]^2$$

$$(Eq.\ 5)$$

365 where j_{HCHO}, j_{CH_3CHO}, $j_{CH_3C(O)CH_3}$ j_{CHOCHO} are respectively j_8, j_9, $j_{10a,b}$ and j_{11}, as in Table 1 in the supplementary information. and $k_{RO_2^*+RO_2^*}$ represents ~~is~~ ~~the~~ an effective RO_2^* self-reaction rate coefficient, ~~which is defined as the weighted average rate~~comprising ~~coefficient between~~ HO_2 – HO_2, HO_2 – RO_2 and RO_2 – RO_2 reaction rates.

Consequently, the RO_2^* concentrations are expected to correlate with the square root of the $P_{RO_2^*}$.

Figure 7 shows the relationship between the measured $[RO_2^*]$ and the ~~estimated~~ calculated $\sqrt[2]{P_{RO_2^*}}$. Generally, both $[RO_2^*]$ and
370 $\sqrt[2]{P_{RO_2^*}}$ increase with the photolysis frequency of O_3 ($j_{O(^1D)}$). ~~The~~ Measurements in which $[RO_2^*]$ ~~<~~were less than 0.5×10^{12} molecules cm^{-3} ~~,~~ ~~and~~ $\sqrt[2]{P_{RO_2^*}}$ ~~<~~less than 1000 and with $j_{O(^1D)} > 5 \times 10^{-5}$ ~~belong to the measurements~~were made above 6000 m, where the amount of RO_2^* precursors is low. The relatively weak correlation observed between $[RO_2^*]$ and $\sqrt[2]{P_{RO_2^*}}$ indicates the ~~presence~~ necessity of other radical ~~loss~~terminating processes and/or missing ~~production~~radical formation terms in the $P_{RO_2^*}$ calculation. Apart from this, the spread in the diagram confirms that the effective RO_2^* self-reaction rate $k_{RO_2^*+RO_2^*}[RO_2^*]^2$ varies
375 widely in the air masses probed likely due to the effect of changes in HO_2 and $\sum RO_2$ concentrations in the individual loss reaction rate coefficients. ~~As mentioned in section 4.1,~~ ~~P~~photochemical processing ~~was~~is expected to be enhanced over Southern Europe due to the prevailing conditions of high insolation and temperatures during the ~~measurements~~EMeRGe flights~~.~~, which might lead to the rapid production of RO_2^* from the photochemical oxidations of CO and VOCs. This is also reflected in the higher $P_{RO_2^*}$ and $[RO_2^*]$ observed in Southern Europe as compared to those in Northern Europe (Fig. 7b).
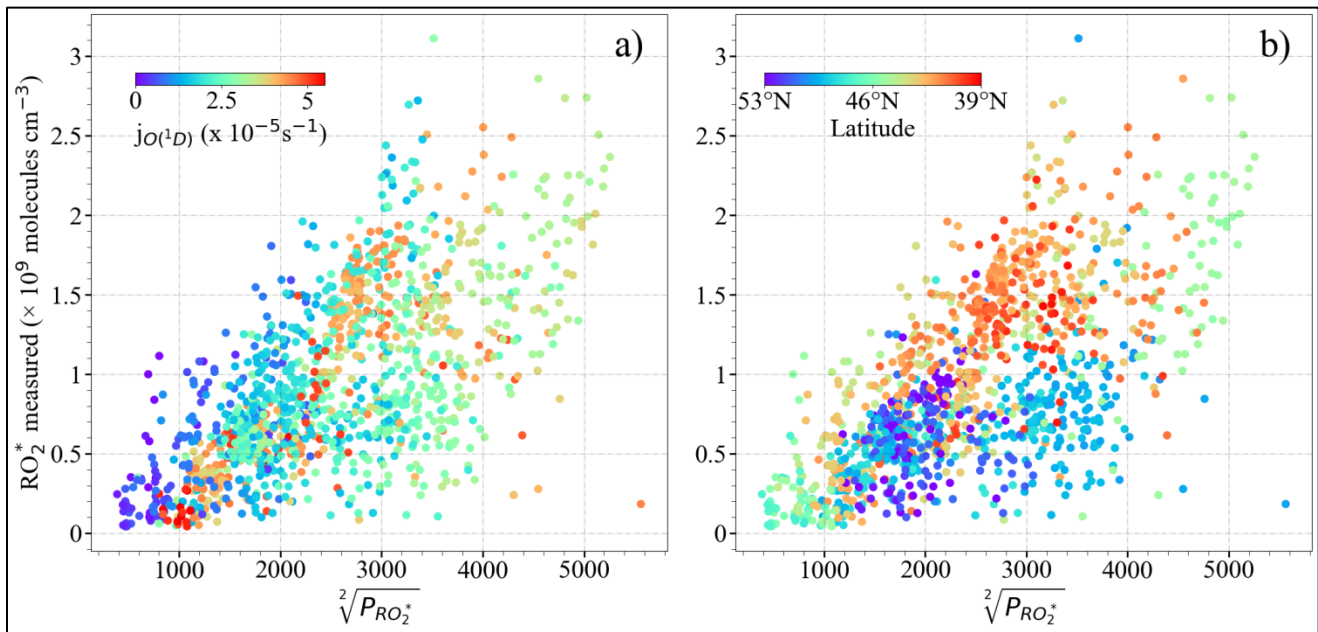
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Figure 7: ~~Plot of the M~~measured [RO$_2^*$] versus ~~estimated~~ calculated $\sqrt[2]{P_{RO_2^*}}$ colour-coded for values of a)- j$_{O(^1D)}$ and b) latitude.

The correlation between [RO$_2^*$] and $\sqrt[2]{P_{RO_2^*}}$ improves when the measurements south and north of 47°N are separately analysed (Fig. 8). For a given [RO$_2^*$], the P$_{RO_2^*}$ calculated is higher for the measurements north of 47°N than south of 47°N. The lowest [RO$_2^*$] to $\sqrt[2]{P_{RO_2^*}}$ ratios are associated with higher NO$_x$ (NO + NO$_2$), especially north of 47°N, indicating the urban character and higher amounts of the ~~content in~~ RO$_2^*$ precursors of the air probed (Fig. 8d). Please n~~N~~ote that these results are only valid for the data set acquired over Europe during EMeRGe flights and do not yield a relationship between [RO$_2^*$] and $\sqrt[2]{P_{RO_2^*}}$, which is generally applicable under all conditions for these two latitude windows.
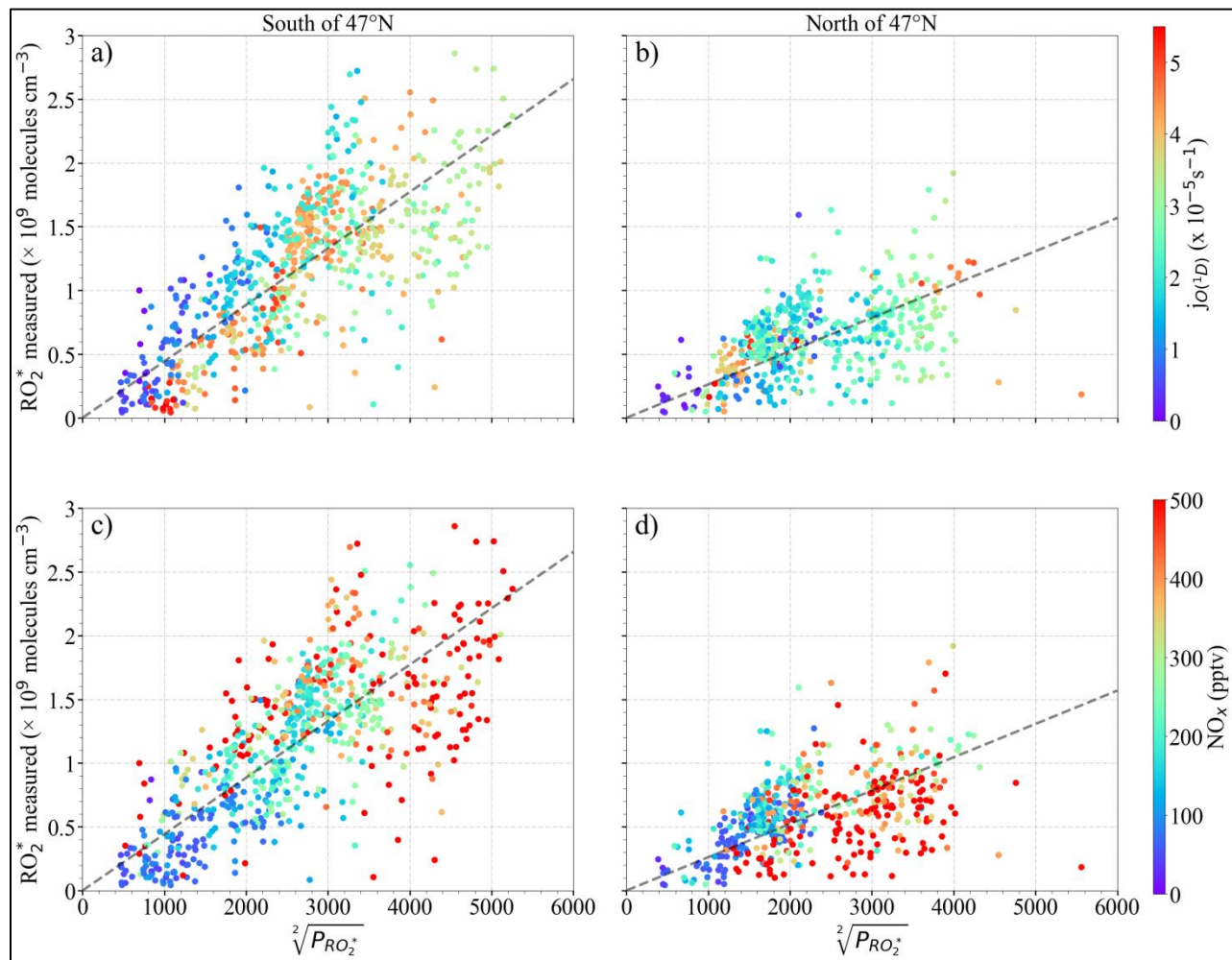
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Figure 8: ~~Plots of the~~ Mmeasured $[RO_2^*]$ vs $\sqrt[2]{P_{RO_2^*}}$ for the following latitudes: a) and c) south of 47°N; b) and d) north of 47°N. Note that a) and b) are colour-coded with $j_{O(^1D)}$; c) and d) are colour-coded by NO_x mixing ratio. The dashed lines indicate the linear fit for visual support.

The relationship between RO_2^* and $P_{RO_2^*}$ is further investigated to identify the dominant RO_2^* loss process in the air masses considered in this study. As stated in section 3, HO_2 and RO_2 are not speciated but retrieved as RO_2^* by the PeRCEAS instrument. Because not all peroxy radicals are detected equally by the instrument, the comparison of measured and calculated RO_2^* values is complicated. To investigate this ~~So, the effect of~~ changes in the HO_2 to the total RO_2^* ratios~~, represented~~ have been taken into consideration by δ, i.e., $[HO_2] = δ[RO_2^*]$ and $[CH_3O_2] = (1- δ) [RO_2^*]$, ~~is investigated~~in the analysis. As a first approach, RO_2 is assumed to consist only of CH_3O_2 ~~reactions are taken as a surrogate for all RO_2 reactions~~ to reduce the complexity of the calculations by considering only CH_3O_2 reaction rate constants. Moreover, in a previous study the ratio $α = eCL_{CH_3O_2} /eCL_{HO_2}$ was determined to be 65% for the measurement conditions (George et al., 2020).

~~Furthermore,~~The Eq. 5 is additionally extended to include RO_2^* ~~production~~effective yields from VOC oxidation~~, radical conversion reactions between oxy and peroxy radicals~~ and ~~OH and RO~~radical losses through HONO, HNO_3~~, and organic nitrate formation.~~: ~~Consequently, Eq. 5 is accordingly modified:~~

$$(2j_1[O_3]\beta + j_3[HONO])(1 - \rho) + 2j_8[HCHO] + 2j_9[CH_3CHO] + 2(j_{10a} + j_{10b})[CH_3C(O)CH_3] + 2j_{11}[CHOCHO] =$$

$$\delta[RO_2^*](k_{23}[NO] + k_{24}[O_3])\rho + \left(2k_{16b}\left((1-\delta)[RO_2^*]\right)^2 + k_{25}(1-\delta)[RO_2^*][NO]\right)\left(\frac{k_{22}[NO]}{(k_{22}[NO]+k_{26}[O_2])}\right) + 2k_{15}\delta(1 -$$

405 $$\delta)[RO_2^*]^2 + 2k_{16a}\left((1-\delta)[RO_2^*]\right)^2 + 2k_{14}(\delta[RO_2^*])^2 \qquad \text{(Eq. 6)}$$

where β is the effective yield of OH ~~production~~ in the reaction of $O(^1D)$ with H_2O given by:

$$\beta = \left(\frac{k_{2a}[H_2O]}{k_{2a}[H_2O] + k_{2b}[O_2] + k_{2c}[N_2]}\right),$$

On the left hand side of Eq. 6, 1-ρ accounts for the effective yield of HO_2+RO_2 through the radical initiation reactions R2a and R3 and reactions R5 to R7 and R12. As the calculation is constrained with on-board measurements, only the reactions of

410 measured VOCs were considered in R12. Similarly, ρ accounts for the effective yield of HONO, HNO_3 and H_2O formation through reactions R19 to R21 and the $HO_2 + NO$ and $HO_2 + O_3$ reactions (R23 and R24 respectively) on the right hand side of Eq. 6.

Consequently, ρ is given by:

415 $$\rho = \frac{(k_{19}[NO] + k_{20}[NO_2] + k_{21}[HONO])}{(k_5[O_3] + k_6[CO] + k_7[CH_4] + k_{12a}[HCHO] + k_{12b}[CH_3CHO] + k_{12c}[CH_3C(O)CH_3] + k_{12d}[CH_3OH] + k_{12e}[CHOCHO] + k_{17}[HO_2] + k_{19}[NO] + k_{20}[NO_2] + k_{21}[HONO])}$$

Measurements of CH_4, HCHO, CH_3CHO, CHOCHO, CH_3OH, and $CH_3C(O)CH_3$ on-board HALO are available and implemented in Eq. 6. These comprise the most abundant and reactive OVOCs and are considered to be a representative ~~the~~ surrogate for the

420 VOCs, ~~which~~ that act as RO_2^* precursors through oxidation and photolysis. During the EMeRGe ~~IOP~~ campaign in Europe, $k_{12a} \times$ HCHO and $k_{12b} \times CH_3CHO$ have the highest contribution to the $1 - \rho$ from all the OVOC measured. ~~HCHO and CH₃CHO are the dominant radical precursors from the OVOC oxidations.~~ Their impact on ~~the rate of pridtcion of RO₂* and~~ the RO_2^* budget is found to be similar because their respective concentrations compensate the difference in the rate coefficients of their reactions with OH ($k_{OH12a+HCHO} = 8.5 \times 10^{-12}$ cm^3 molecule^{-1} s^{-1} and $k_{OH+CH_3CHO12b} = 1.5 \times 10^{-11}$ cm^3 molecule^{-1} s^{-1} at 298K and 1 atm.). Despite

425 its high mixing ratios measured, $CH_3C(O)CH_3$ is ~~a~~ less important ~~source of RO₂*~~ in the $1 - \rho$ term. This is because the rate coefficient $k(T)_{12c}$ is significantly slower than $k_{OH+HCHO12a}$ and $k_{OH+CH_3CHO12b}$ (see Table S1 in the supplement). Similarly, the contribution of CHOCHO and CH_3OH is an order of magnitude lower than that of HCHO and CH_3CHO.

Concerning the term $\delta[RO_2^*](k_{23}[NO] + k_{24}[O_3])\rho$ on the right hand side of Eq.6, the HO_2 reaction with O_3 has a negligible effect ~~As~~ as $k_{HO_2+O_324}$ is almost four orders of magnitude smaller than k_{HO_2+NO23} and the NO concentrations remained about three

430 orders of magnitude smaller than the O_3 measured during the campaign, ~~the HO₂ reaction with O₃ has a negligible effect in Eq. 6~~.

The impact of the methylglyoxal ($CH_3C(O)C(O)H$) photolysis was also investigated by using the $CH_3C(O)C(O)H^*$ measurements provided by the miniDOAS instrument. The $CH_3C(O)C(O)H^*$ measured is the sum of $CH_3C(O)C(O)H$, and a fraction of other substituted dicarbonyls (mainly 2,3-butanedione, $C_3H_6O_2$), with similar visible absorption spectra. For the calculation, $CH_3C(O)C(O)H$ was assumed to be half of $CH_3C(O)C(O)H^*$ as recommended by Zarzana et al. (2017) and Kluge et al. (2020).

435 The RO_2^* calculated by including $CH_3C(O)C(O)H$ photolysis systematically overestimated the measurements. As the adequacy of the recommended factor of 0.5 varies with the actual air mass composition, $CH_3C(O)C(O)H$ was not included in the calculations.

Figure 9 shows the fractional contribution of the destruction rate ($D_{RO_2^*}$) calculated for a 1:1 mixture of HO_2 and CH_3O_2 using the reactions included in Eq. 6 as a function of altitude. The data are classified into three groups according to the rate of destruction of RO_2^* mixing ratio $D_{RO_2^*} < 0.01$ pptv s^{-1} (a), $0.01 < D_{RO_2^*} < 0.9$ pptv s^{-1} (b), and $D_{RO_2^*} > 0.9$ pptv s^{-1} (c) to show the lowest, most common, and highest ranges, respectively, encountered during the EMeRGe campaign. For 90 % of the measurements, $0.01 < D_{RO_2^*} < 0.9$ pptv s^{-1} applies, while the rest of the data are equally distributed in the other two $D_{RO_2^*}$ ranges. The data in each group are always binned over 500 m when available.

As can be seen in Fig. 9, the ±1σ standard deviation of the obtained bins is very high. In spite of this, the $HO_2 - CH_3O_2$ and $HO_2 - HO_2$ reactions seem to dominate the radical destruction processes in the air masses probed. Their combined contribution is > 70 % in all the cases except in the 1000 m bin of $D_{RO_2^*} > 0.9$ pptv s^{-1}. Other significant radical losses occur through the HONO and HNO_3 formation. The contribution of the $CH_3O_2 + CH_3O_2$ reaction to the total RO_2^* destruction rate is < 5 %.".
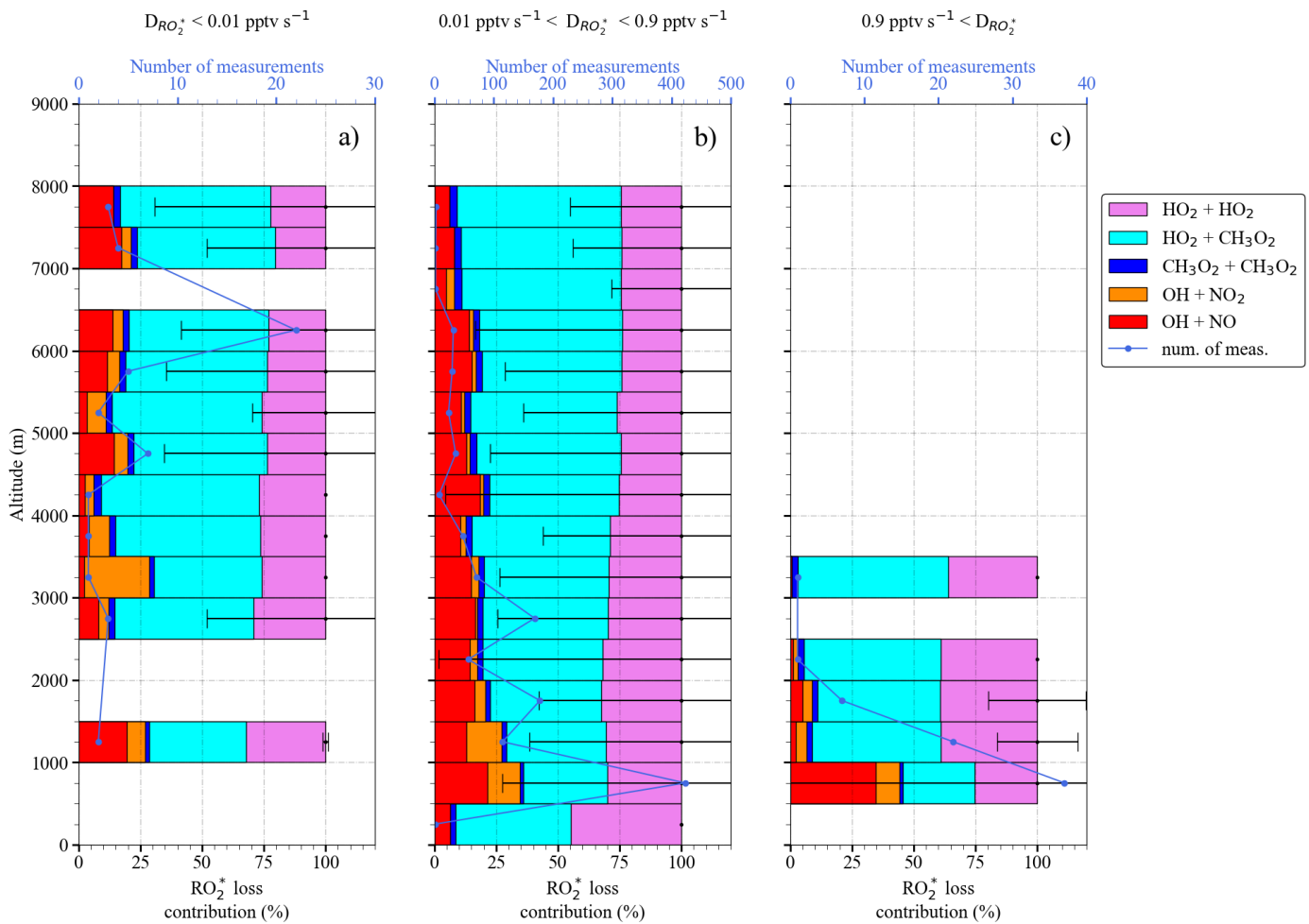

Figure 9: RO_2^* destruction rate $D_{RO_2^*}$ and fractional contributions from loss reactions in Eq.6 as a function of altitude, for: a) $D_{RO_2^*} < 0.01$ pptv s^{-1}, b) $0.01 < D_{RO_2^*} < 0.9$ pptv s^{-1}, and c) $D_{RO_2^*} > 0.9$ pptv s^{-1}. Note the different scales in the number of measurements."

Since Eq. 6, is quadratic in $[RO_2^*]$ it can be solved for ~~can be calculated as~~ $[RO_2^*]_c$ where c stands for calculated, as:

$$[RO_2^*]_c = \frac{-\left(-L_{RO_2^*}\right) - \sqrt{L_{RO_2^*}^2 - 4\left(-2k_{RO_2^*}\right)P_{RO_2^*}}}{2\left(-2k_{RO_2^*}\right)}$$

(Eq. 79)

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where

$$k_{RO_2^*} = \left(\left(k_{16b} \left(\frac{k_{zz}[NO]}{(k_{zz}[NO] + k_{z6}[O_z])} \right) + k_{16a} \right) k_{16a} (1 - \delta)^2 + k_{15}\delta(1 - \delta) + k_{14}\delta^2 \right)$$

$$L_{RO_2^*} = \left(\delta(k_{23}[NO] + k_{24}[O_3])\rho + \left(\frac{k_{zz}[NO]}{(k_{zz}[NO] + k_{z6}[O_z])} \right) k_{25}(1 - \delta)[NO] \right)$$

$$P_{gRO_2^*} = (2j_1[O_3]\beta + j_3[HONO])(1 - \rho) + 2j_8[HCHO] + 2j_9[CH_3CHO] + 2(j_{10a} + j_{10b})[CH_3C(O)CH_3] + 2j_{11}[CHOCHO]$$

where $k_{RO_2^*}$ is a weighed rate coefficient of RO_2^* self reactions for a 1:1 mixture of HO_2 and CH_3O_2, L_{RO2*} comprises the formation of HONO and HNO_3 and Pg_{RO2*} is the gross production of RO_2^*.

The second solution of the quadratic equation gives negative values for $[RO_2^*]_c[RO_2^*]$, therefore is assumed to have~~has~~ no physical meaning. A more detailed derivation of Eq. 6 and Eq. 7 are given in the supplementary information.

Figure 10 shows the measured RO_2^* (hereinafter referred to as $RO_2^*{}_m$) mixing ratio versus the calculated $RO_2^*{}_c$ mixing ratio using Eq. 7. $RO_2^*{}_m$ and $RO_2^*{}_c$ are the measured and calculated RO_2^* respectively for $\delta = 1$, i.e. $RO_2^* = HO_2$ and $\delta = 0.5$, i.e. $HO_2 = RO_2$~~Applying Eq. 9~~Eq. 7 to the measured dataset reduces the overestimation of $RO_2^*{}_m$ by $RO_2^*{}_c$ at NO mixing ratios higher than 250 pptv (Fig. 10~~Fig. 9~~), especially for $RO_2^*{}_m < 40$ pptv~~but does not introduce significant changes in the overall correlations (Table 2).~~ The eCL corresponding to $\delta = 1$ and $\delta = 0.5$ used for the $RO_2^*{}_m$ retrievals were determined in laboratory experiments, as reported by George et al. (2020). The small circles represent 1-minute $RO_2^*{}_m$, whereas the large circles are the mean of the $RO_2^*{}_m$ binned over 10 pptv $RO_2^*{}_c$ intervals.~~As in Fig. 9, plots of the 1-minute $RO_2^*{}_m$ and the average of $RO_2^*{}_m$ binned over 10 pptv $RO_2^*{}_c$ intervals versus $RO_2^*{}_c$ are depicted for $\delta = 1$ and $\delta = 0.5$ in Fig. 10.~~ The RO_2^* data are colour-coded with the on-board NO measurements. The linear regression slopes are around 0.7 ($R^2 = 0.96$), indicating an overall 25 – 30 % underestimation of the $RO_2^*{}_m$.~~ However, the $RO_2^*{}_m$ are mostly overestimated 4000 m under low insolation and underestimated in polluted plumes measured below 2000 m with NO mixing ratios approximately above 50 pptv (see Fig. 11~~Fig. 10 for $\delta = 0.5$). The y-axis intercept is below the instrumental detection limit for most measurement conditions.

Table 2: Linear regression parameters from $RO_2^*{}_m$ versus $RO_2^*{}_c$ using Eq. 7 ~~and Eq. 9~~ from ~~Fig. 9 and~~ Fig.10~~, respectively~~.

Formula used to calculate RO_2^*	δ	slope	y-intercept (pptv)	R^2
	1.00	0.71	5	0.96
Eq. 7	0.50	0.74	6	0.97
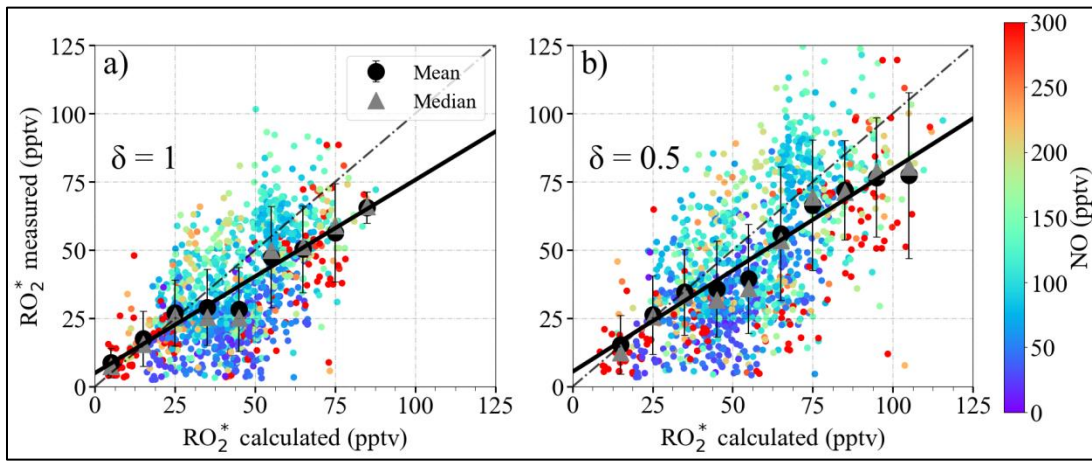
Figure 10: $RO_2^*{}_m$ versus $RO_2^*{}_c$ ~~calculated~~ using ~~Eq. 9~~Eq. 7 for a) $\delta = 1$ and b) $\delta = 0.5$. The data are colour-coded with the measured NO mixing ratios. The 1-minute (small circles), the mean of the binned $RO_2^*{}_m$ over 10 pptv $RO_2^*{}_c$ intervals (large circles), and the median of each bin (grey triangles) are shown. The error bars indicate $\pm 1\sigma$ standard deviation ~~the standard error~~ of each bin. The linear regression for the binned values (solid line) and the 1:1 relation (dashed line) are also depicted for reference.

Figure 11 shows the vertical profiles of $RO_2^*{}_m$ and $RO_2^*{}_c$ mixing ratios calculated for $\delta = 0.5$, averaged for the EMeRGe flights over Europe in 500 m altitude bins. $RO_2^*{}_c$ seems to overestimate $RO_2^*{}_m$ for altitudes above 4000 m. As mentioned in Sect. 4.1, the vertical profiles are a composite from averaging flights with legs carried out at different longitude and latitudes. Therefore, the differences between $RO_2^*{}_m$ and $RO_2^*{}_c$ have been studied in more detail respect to the composition of the individual air masses (see the $RO_2^*{}_m$ and $RO_2^*{}_c$ mixing ratios as a function of latitude and altitude in Fig. S4 in the supplementary information).
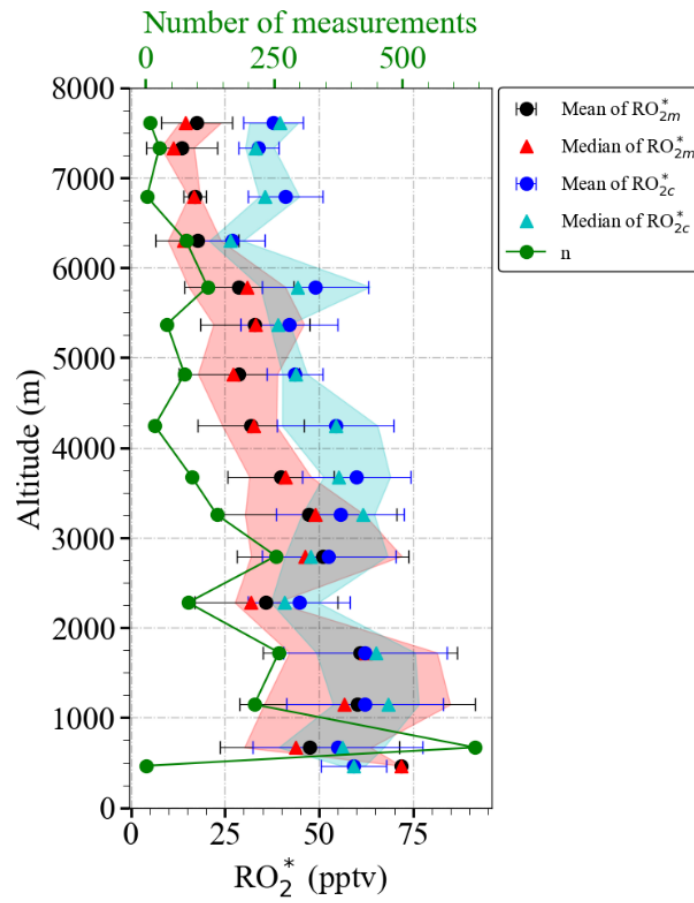
Figure 11: Vertical distribution of the mean $RO_2^*{}_m$ and mean $RO_2^*{}_c$ using Eq. 7 for $\delta = 0.5$ for the EMeRGe data set in Europe. The measurements are binned over 500 m altitude. The error bars are the $\pm 1\sigma$ standard deviation of each bin. Median values (red and green triangles) the interquartile 25-75% range (red and blue shaded areas) and the number of individual measurements, n, for each bin (in green) are additionally plotted.
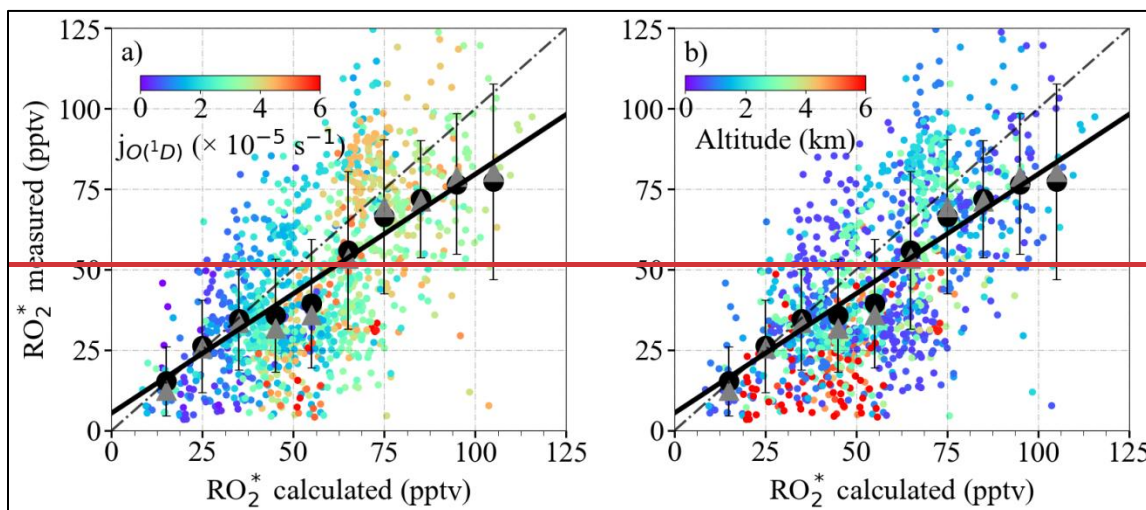
~~Figure 11: $RO_2^*{}_m$ versus $RO_2^*{}_c$ calculated using Eq. 9 for $\delta = 0.5$. The data points are colour-coded for a) photolysis frequency of O_3; b) altitude. The 1-minute (small circles), the mean of the binned $RO_2^*{}_m$ over 10 pptv $RO_2^*{}_c$ intervals (large circles), and the median of each bin (grey triangles) are shown. The error bars indicate the standard error of each bin. The linear regression for the binned values (solid line) and the 1:1 relation (dashed line) are also depicted for reference.~~

Figure 12 shows the data for $\delta = 0.5$ colour-coded with NO, NO_x, the sum of HCHO, CH_3CHO, CHOCHO, CH_3OH, and $CH_3C(O)CH_3$ (from now on referred to as $\Sigma VOCs$), as a surrogate for the amount of OVOCs acting as RO_2^* precursors, and the $\Sigma VOCs$ to NO ratio. The largest differences between $RO_2^*{}_m$ and $RO_2^*{}_c$ ~~is~~ are observed for the bins around 50 pptv. The $RO_2^*{}_c$ overestimate the $RO_2^*{}_m$ mostly for ~~The~~ $RO_2^*{}_m < 25$ pptv observed above $\simeq$4000 m. ~~are overestimated for~~ These air masses are characterised by ~~with low insolation, i.e. $j_{O(^1D)} < 2 \times 10^{-5}$ s^{-1} (Fig. 11~~Fig. 10)~~,~~ NO < 50 pptv, $\Sigma VOCs$ typically below 4 ppbv, ~~and~~ high $\Sigma VOCs$/NO ratios (> 50), and ~~.~~ low insolation conditions , i.e. $j_{O(^1D)} < 2 \times 10^{-5}$ s^{-1} (see Fig. S5 in the supplementary information). Under these insolation conditions, the radical production rate is expected to be low, and the $RO_2^* - RO_2^*$ reactions are expected to dominate the RO_2^* loss processes. As ~~Since~~ OH and H_2O_2 were not measured during the EMeRGe campaign in Europe, ~~Eq. 9~~Eq. 7 does not include the loss reactions R17 and R18, which might be significant under such conditions (Tan et al., 2001) and explain the ~~RO_2^*~~ overestimation of $RO_2^*{}_m$. This is also the case for the overestimations observed below 40 pptv $RO_2^*{}_m$ at other altitudes, where NO < 50 pptv but the $\Sigma VOCs$/NO ratios remain low. The overestimation may therefore be independent of the $\Sigma VOCs$/NO ratios. For NO ≤ 50 pptv, $NO_2 \leq 100$ pptv, $RO_2^* \leq 40$ pptv and HCHO ≤ 1 ppbv, the rate of reaction ~~R18~~R17, which forms H_2O and O_2 from OH and HO_2, is about 4 times faster than the rate of the OH oxidation reaction of the dominant OVOCs (R12) considered in this study or the rate of formation of HONO (R19).

$RO_2^*{}_m$ is ~~systematically~~ both underestimated and overestimated for $\Sigma VOCs$ mixing ratios greater than 7 ppbv. The composition of these air masses is ~~quite~~ very different, as reflected by the $\Sigma VOCs$/NO ratios. This implies that ~~Eq. 9~~Eq. 7 does not capture the peroxy radical ~~production~~ yields adequately from the measured VOCs and OVOC in these cases. The ~~underestimation~~ differences ~~of~~ between $RO_2^*{}_m$ and $RO_2^*{}_c$ may be explained in part by a) changes in OH yields due to ~~recycling through~~ additional VOC

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oxidation processes, which are not in ~~Eq. 9~~Eq. 7 and/or b) RO_2^* production from the photolysis of carbonyls, which were not measured and/or c) RO_2^* production from the ozonolysis of alkenes or unid~~i~~entified biogenic terpene emissions and/or d) overestimation of the loss processes. For VOC < 2ppb and $\sum$VOCs/NO < 20, $RO_2^*{}_m$ is systematically overestimated. This might indicate underestimation in the radical losses through nitrite and nitrate formation.
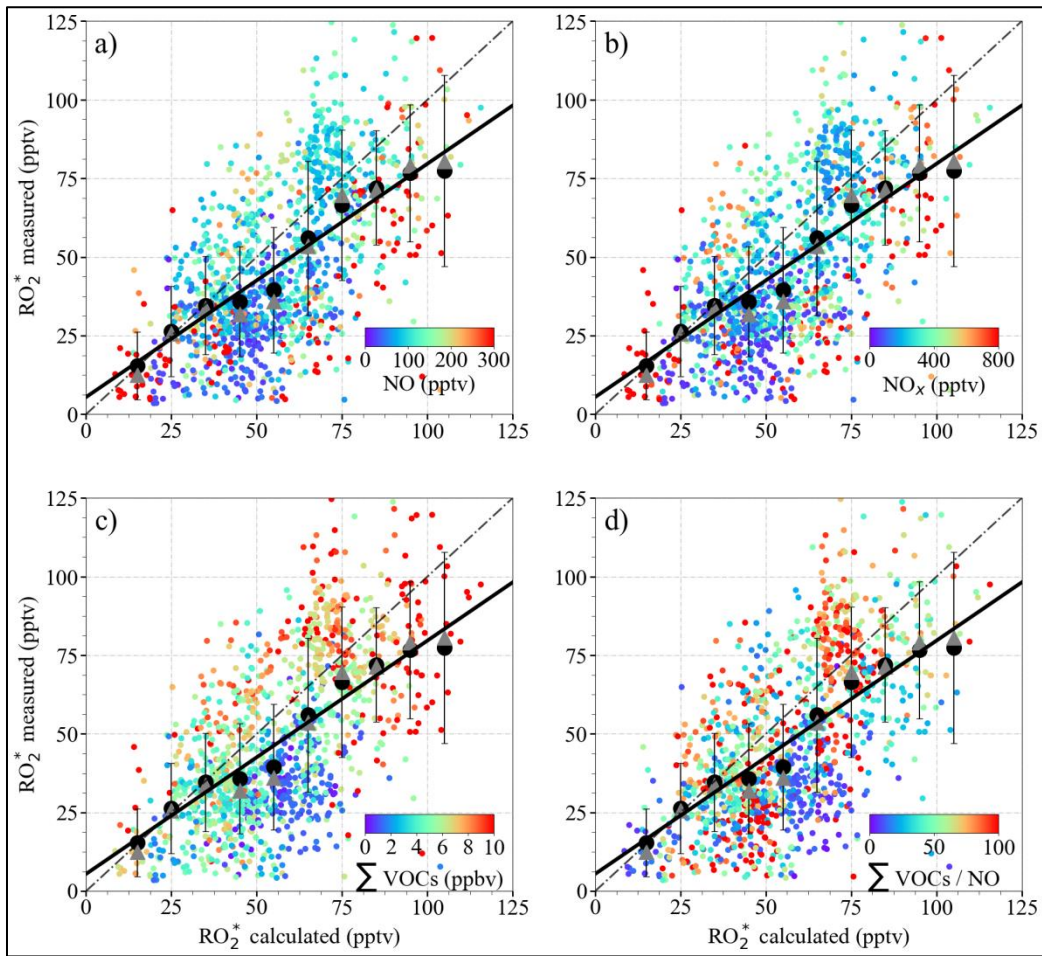

Figure 12: $RO_2^*{}_m$ versus $RO_2^*{}_c$ using ~~Eq. 9~~Eq. 7 for $\delta = 0.5$ colour-coded with the measured a) NO mixing ratio, b) NO_x mixing ratio, c) $\sum$VOCs mixing ratio, where $\sum$VOCs = $HCHO + CH_3CHO + (CHO)_2 + CH_3OH + CH_3C(O)CH_3$, and d) $\sum$VOCs/NO ratio. The 1-minute (small circles), the mean of the binned $RO_2^*{}_m$ over 10 pptv $RO_2^*{}_c$ intervals (large circles), and the median of each bin (triangles) are shown. The error bars represent the $\pm 1\sigma$ standard ~~error~~ deviation of each bin. The linear regression for the binned values (solid line) and the 1:1 relationship (dashed line) are plotted for reference.

Although considered small, the s~~S~~patial and temporal differences in the in-situ measurements of the key trace gases (O_3, NO, H_2O, CO, CH_4, VOCs) as compared to those of the ~~with respect to~~ remote sensing observations (NO_2 and HONO) used in ~~Eq. 9~~Eq. 7 may also contribute to the overall spread observed in Fig. 12. Although the temporal evolution and the amount of the trace gases measured using in-situ and remote sensing instruments agree reasonably well, as shown for HCHO in Fig. 13, the remote sensing instruments have, in general, larger air sampling volumes compared to that of in-situ instruments. This may occasionally lead to significant differences depending on the location of the pollutant layers with respect to HALO. In addition, PTR-MS measurements of HCHO might include interferences from molecular fragments of other compounds in the sample air (Inomata et al., 2008). Further details about the accuracy and comparability of the instrumentation on-board during the campaign can be found elsewhere (Schumann, 2020).
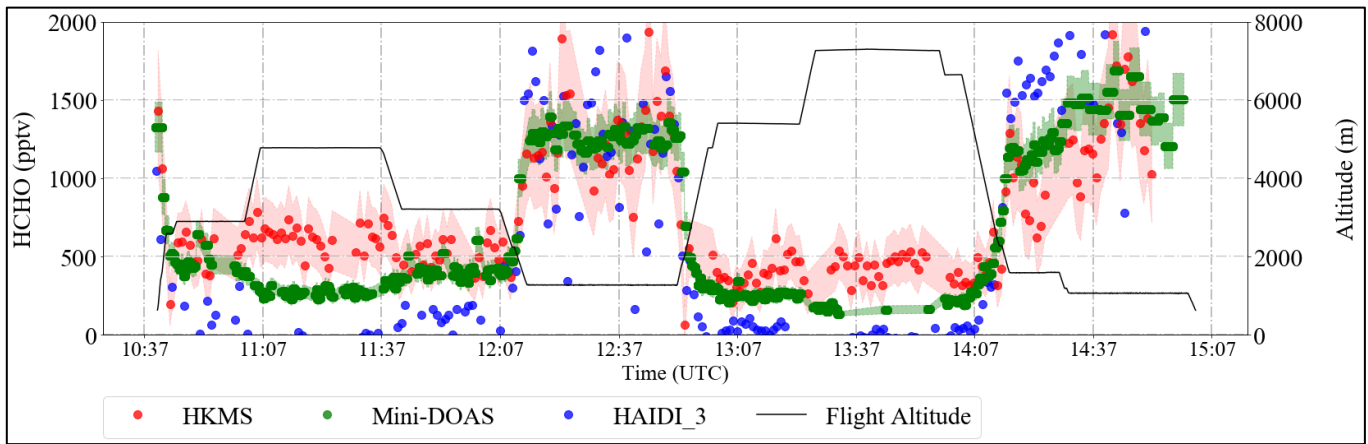
Figure 13: An example of the time series of the measured HCHO mixing ratios retrieved from the remote sensing (HAIDI in blue and miniDOAS in green) and in-situ (HKMS in red) instruments during the E-EU-04 flight on 14.07.2017. The shaded region shows ± 1σ uncertainties of the HKMS and miniDOAS instruments. The flight altitude is depicted in black.

In summary, apart from the inaccuracies in the reaction rate coefficients, the differences between $RO_2^*{}_m$ and $RO_2^*{}_c$ might be caused by a combined effect of the limitations of the analytical expression to simulate complex non-linear chemistry and the measurement uncertainties arising from the spatial heterogeneity of the plume for the remote sensing instruments. Consequently, the quantification of limiting factors in Eq. 7 require the ~~individual~~ analysis of the pollution events encountered along the flights individually ~~is required to quantify limiting factors in Eq. 9~~Eq. 7.

The ratio of $RO_2^*{}_m$ to $RO_2^*{}_c$ ($RO_2^*{}_m/RO_2^*{}_c$) has been used to assess the applicability of ~~Eq. 9~~Eq. 7 for the calculation of RO_2^* in the air masses probed ~~(Fig. 14)~~. In Fig. 14, the data are colour-coded with respect to $RO_2^*{}_m /RO_2^*{}_c$, H_2O, $\Sigma VOCs$, and NO_x. The air masses probed at altitudes above 2000 m are close to the PSS assumptions used to develop ~~Eq. 9~~Eq. 7, and consequently, the $RO_2^*{}_m/RO_2^*{}_c$ remains ≤ 1. In contrast, $RO_2^*{}_m/RO_2^*{}_c$ is at its highest value below 2000 m, reaching up to 3. At these altitudes, most of the flights in Europe were carried out in pollution plumes, in which both the amount of NO_x and RO_2^* precursors are high. The analytical expression does not capture the RO_2^* variations resulting from fast non-linear photochemistry present in these pollution plumes. This is the case for the measurements made between 42°N and 46°N in the outflow of Po Valley and Rome. $\Sigma VOCs > 7$ ppbv and NO_x mixing ratios > 500 pptv indicate high radical precursor loading and relatively fresh emissions. The $RO_2^*{}_m/RO_2^*{}_c$ is also > 2 in the measurements over the English Channel (between 50°N and 52°N) with $\Sigma VOCs$ and NO_x mixing ratio > 7 ppbv and 1000 pptv, respectively.
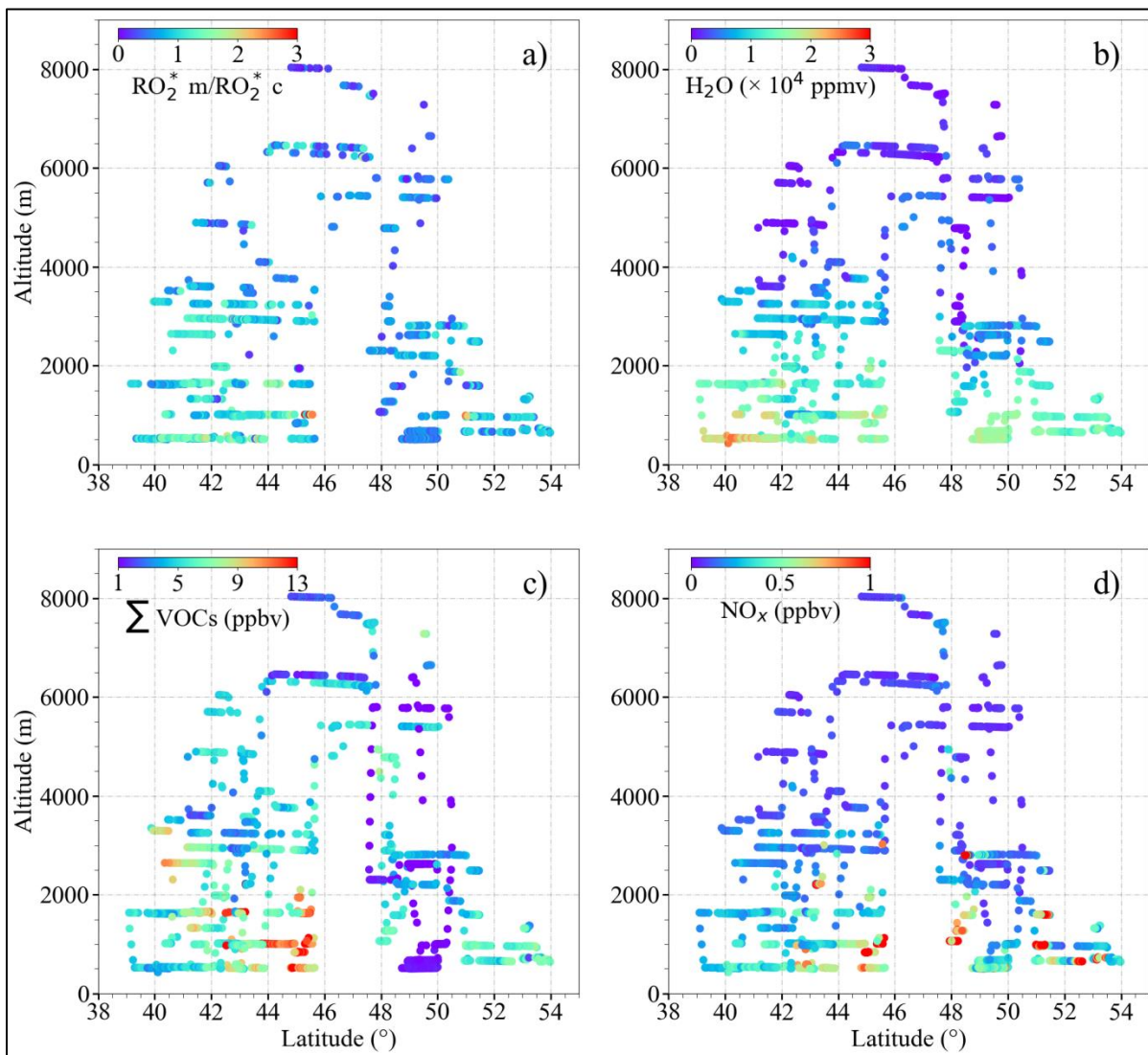
Figure 14: Plots of a) the ratio of $RO_2^*{}_m$ to $RO_2^*{}_c$ ($RO_2^*{}_m/RO_2^*{}_c$) assuming that $\delta = 0.5$; b) H_2O; c) $\sum$VOCs; d) NO_x as a function of latitude and altitude for the EMeRGe measurements in Europe.

The applicability of ~~Eq. 9~~Eq. 7 for calculating the in-flight measurements of RO_2^* along the track of the E-EU-03 flight on 11 July 2017 ~~is shown in Fig. 16Fig. 14~~was studied in more detail. The E-EU-03 flight investigated the outflow of selected MPCs in Italy (i.e., Po Valley and Rome). Consequently, the flight track was routed along the western coast of Italy and included vertical profiling over the Tyrrhenian Sea upwind of Rome (see Fig. ~~15~~S6 in the supplementary information). As indicated by ~~can be seen in~~ $j_{O(^1D)}$, in Fig. 15, cloudless conditions dominated throughout the flight track. The $RO_2^*{}_c$ agree reasonably well with $RO_2^*{}_m$ throughout this period except in the pollution plume measured from 12:05 to 12:25 UTC. In this plume, CO, NO, NO_2, HONO, NO_y, and HCHO were ~~reached~~ 100 ppbv, 180 pptv, 150 pptv, 120 pptv, 1ppbv and 2 ppbv, respectively. The $RO_2^*{}_m$ ~~RO_2^*~~ ~~measurements~~ are approximately 20 % underestimated by $RO_2^*{}_c$ during this period. Backward trajectories calculated using FLEXTRA indicate the transport of pollution through the Mediterranean mixed with dust plumes originating from Tunisia. The NO mixing ratios observed indicate the proximity to emission sources.
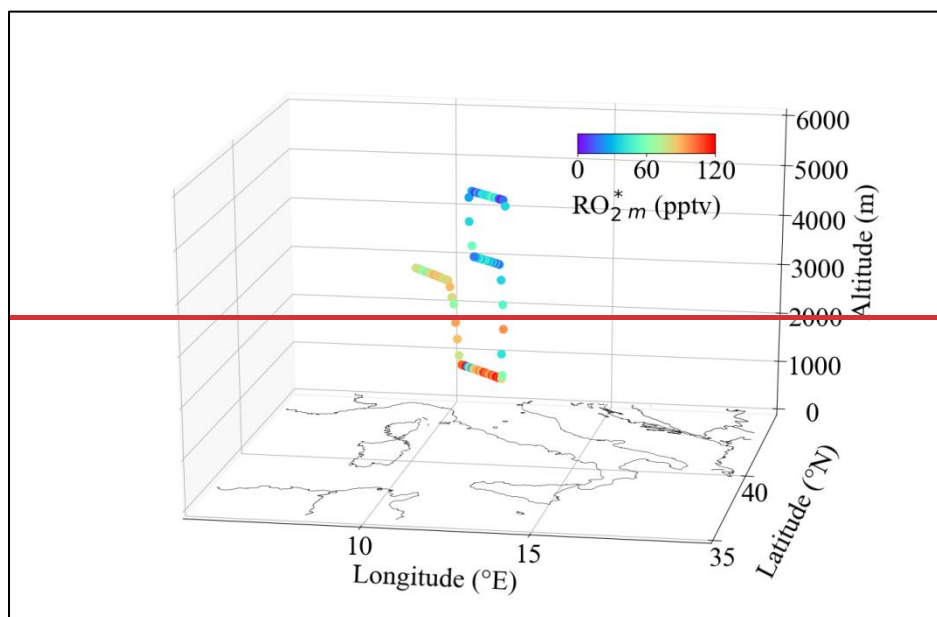
The measurements of VOCs used in ~~Eq. 9~~Eq. 7 may not be representative of the actual complex VOC composition in the plume measured from 12:05 to 12:25 UTC. Consequently, the RO_2 to HO_2 ratio ~~and the RO intermediates involved in the radical interconversion processes~~, the branching ratios and effective rate coefficients for $RO_2^* - RO_2^*$ reactions might not be well represented in ~~Eq. 9~~Eq. 7. Taking CH_3O_2 as a surrogate for all RO_2 might lead to uncertainties in the RO_2^* calculations in the presence of OVOCs with larger organic chains. On the experimental side, changes in the HO_2 to RO_2 ratio affect the accuracy of the PeRCEAS retrieval of the total sum of radicals. As noted in section 3, in this study $RO_2^* = HO_2 + 0.65 \times RO_2$, and the eCL is determined for a 1:1 mixture of $HO_2:CH_3O_2$, i.e. $\delta = 0.5$ is used for the RO_2^* retrieval. However, the HO_2 to CH_3O_2 ratio is not expected to remain constant in all the air masses probed. For a 3:1 ratio of $HO_2:RO_2$, the $RO_{2\,m}^*$ would decrease by 10 %. Similarly, a $HO_2:RO_2$ ratio of 1:3 would lead to an increase of 10 % in the reported $RO_{2\,m}^*$. This uncertainty is well below the in-flight uncertainty of the PeRCEAS instrument indicated by the error bars in Fig. ~~16~~14 (George et al., 2020), and cannot account for the overall ~~20 %~~ underestimation~~s~~. However, it might reduce the differences observed between $RO_{2\,m}^*$ and $RO_{2\,c}^*$ in particular cases. A complete explanation of the variability of RO_2^* in the pollution plumes measured within the ~~IOP~~campaign in Europe is beyond the scope of this analysis and requires an investigation by high-resolution chemical models.
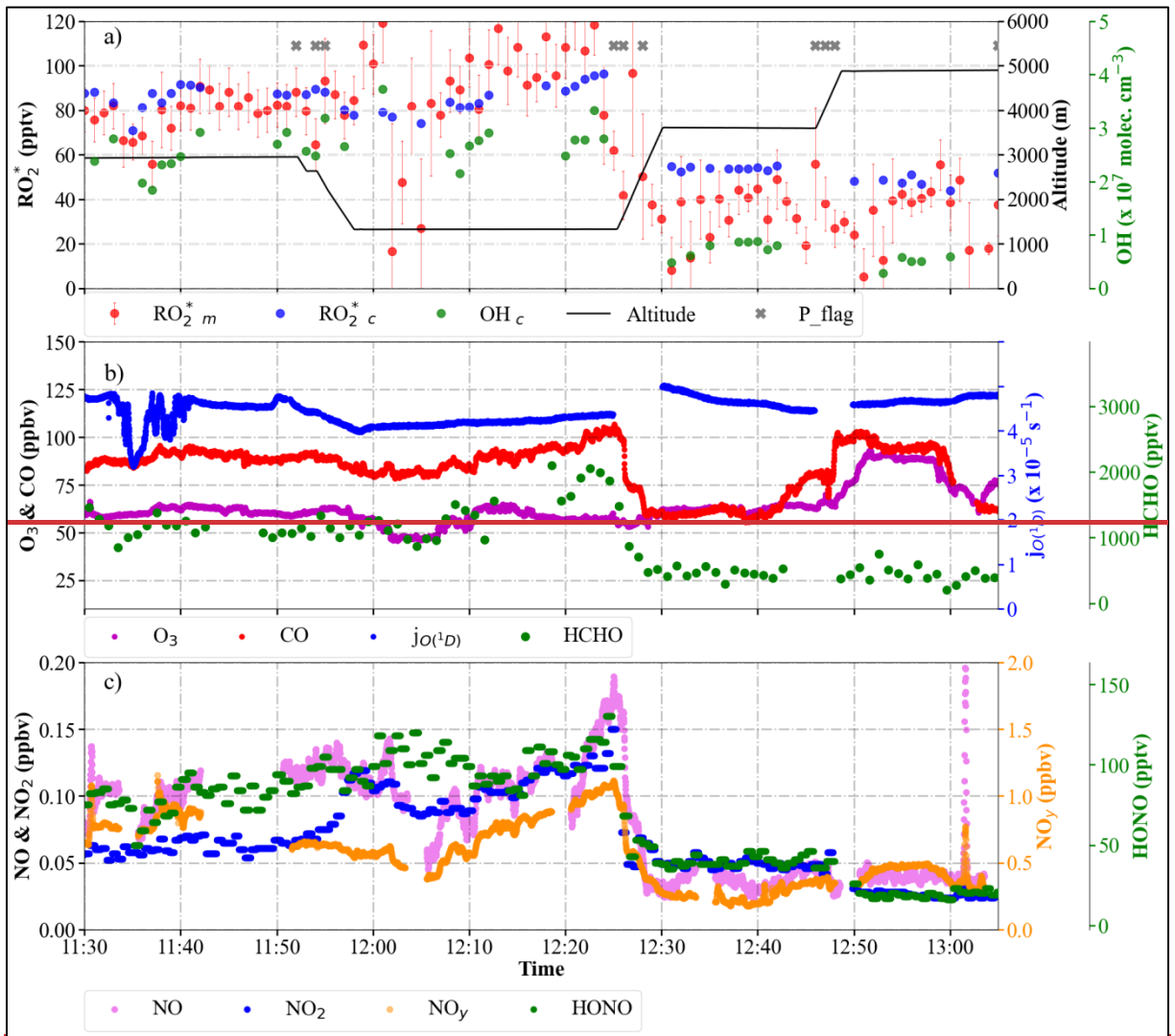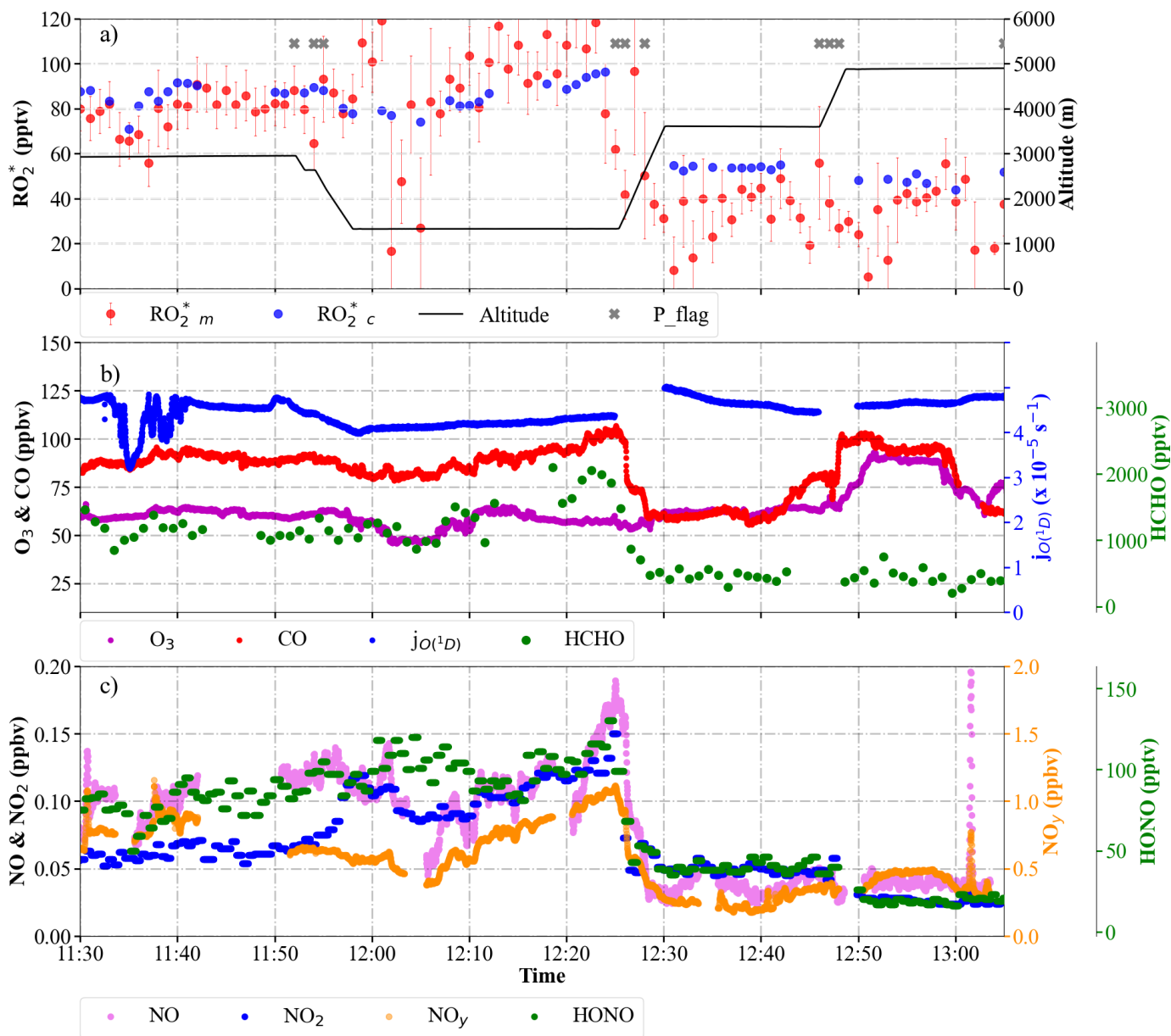
Figure 15: Temporal variation of RO$_2^*$$_m$ and RO$_2^*$$_c$, selected radical precursors and j$_{O(^1D)}$ along the E-EU-03 flight track: a) RO$_2^*$$_m$, RO$_2^*$$_c$ ~~and calculated OH (OH$_c$)~~ mixing ratios. The flight altitude is indicated in black. The P_flag indicates RO$_2^*$ measurements affected by dynamical pressure variation in the inlet; b) O$_3$, CO, HCHO mixing ratios, and j$_{O(^1D)}$; c) NO, NO$_2$, NO$_y$, and HONO mixing ratios.

~~The OH concentrations in Fig. 16~~Fig. 15 ~~are upper limits calculated by assuming pseudo-steady-state for the OH production (R1-R3, R24 and R26) and loss (R4 to R6, R13 and R15 to R19) reactions as described by Eq. 10~~Eq. 8~~:~~

~~$$2j_1[O_3]\beta + j_4[HONO] + k_{23}[HO_2][NO] + k_{24}[HO_2][O_3] = [OH](k_5[O_3] + k_6[CO] + k_7[CH_4] + k_{12a}[HCHO] + k_{12b}[CH_3CHO] + k_{12c}[CH_3C(O)CH_3] + k_{12d}[CH_3OH] + k_{12e}[CHOCHO] + k_{17}[HO_2] + k_{19}[NO] + k_{20}[NO_2] + k_{21}[HONO]) - 2(k_{18a} + k_{18b})[OH]^2$$~~
(Eq. 10Eq. 8)

~~The OH calculated from Eq. 10~~Eq. 8 ~~assuming δ = 0.5 is much higher than the OH concentration reported in the previous airborne (Crawfor et al., 1999'; Tan et al. 2001) and ground-based measurements (Mihelcic et al., 2003; Kanaya et al., 2007, 2012; Hofzumahaus et al., 2009; Elshorbany et al., 2012; Lu et al., 2012, 2013; Tan et al., 2017, 2018; Whalley et al., 2018, 2021; Michelle et al., 2020) in different urban environments. This indicated that the limited number of OVOCs measurements available~~

4.4. Comparison of results with other studies

4.4.1 RO$_2^*$ production rate

Cantrell et al. (2003b) proposed that the production of RO$_2^*$, P$_{RO_2^*}$, is equal to to be equal to the sum of two terms representing RO$_2^*$ – RO$_2^*$ reactions and the RO$_2^*$ – NO$_x$ reactions in the troposphere. As a result of this assumption, these authors describe the
605 relationship between HO$_2$, RO$_2$, P$_{RO_2^*}$ and NO$_x$ is described byas Eq.11:

$$P_{RO_2^*} = k_{RR} [HO_2 + RO_2]^2 + k_{RN} [HO_2 + RO_2] [NO_x] \qquad \text{(Eq. 11Eq. 9)}$$

where k$_{RR}$ and k$_{RN}$ refer to effective rate coefficients for RO$_2^*$ – RO$_2^*$ and RO$_2^*$ – NO$_x$ reactions, and are calculated as fit parameters. Solving Eq. 11Eq. 9 for [HO$_2$ + RO$_2$]^2 leads to:

$$[HO_2 + RO_2] = \sqrt[2]{A + B^2} - B \qquad \text{(Eq. 12Eq. 10)}$$

610 where A = $\frac{P_{RO_2^*}}{k_{RR}}$ and B = $\frac{k_{RN}[NO_x]}{2\,k_{RR}}$. For low NO$_x$ and/or high P$_{RO_2^*}$, B becomes negligible compared to A. Then [HO$_2$ + RO$_2$] approaches $\sqrt[2]{A}$ and is independent of NO$_x$. For high NO$_x$ and /or low P$_{RO_2^*}$, [HO$_2$ + RO$_2$] approaches zero.

The least-square fitting in Eq. 12 is applied to The RO$_2^*$$_m$ and RO$_2^*$$_c$ forfrom the EMeRGe measurements observations in Europe observations, binned in 0.1 pptv s^{-1} P$_{RO_2^*}$ intervals, arewere fitted according to the procedure of Cantrel et al. (2003b) and the results are as shown in Fig. 1716. The fitobtained fit parameters k$_{RR}$ and k$_{RN}$ were fitted. and for Fig. 17a16a and Fig. 18b16b are k$_{RR}$
615 = 7 × 10^{-5}; k$_{RN}$ = 9 × 10^{-6}. The RO$_2^*$ calculated by Eq. 9Eq. 7 appears to be close to the linear function of the NO$_x$ measured. Similar to the results of the study of Cantrell et al. (2003b), a decrease of RO$_2^*$ with NO$_x$ is identified for NO$_x$ > 1000 pptv, although only for P$_{RO_2^*}$ < 0.7 pptv s^{-1}. In the study of Cantrell et al. (2003b), P$_{RO_2^*}$ only reached values up to 0.275 pptv s^{-1}.

Despite the low agreement of the fitted lines with the RO$_2^*$$_m$, a decrease of the RO$_2^*$$_m$ as a function of NO$_x$ is still observed. The disagreement between the RO$_2^*$$_m$ and the curves estimated using Eq. 12Eq. 10 implies that the simplified Eq. 11Eq. 9 from Cantrell
620 et al.(2003b), is insufficient to adequately describe the chemical and physical processes occurring in the troposphereair masses probed. Part of the disagreement might arise from missing terms in the P$_{RO_2^*}$ calculation or inaccuracies related to the NO to NO$_2$ ratio in the air mass, which are more evident at higher P$_{RO_2^*}$. As expected, the ratio of calculated [RO$_2^*$$_c$] to $\sqrt[2]{P_{RO_2^*}}$ has a negative linear dependency dependence on the measured [NO$_x$] (see Fig. 17c16c). The comparable plot relationship of [RO$_2^*$$_m$] to /$\sqrt[2]{P_{RO_2^*}}$ to [NO$_x$] is not linear for NO$_x$ < 100 pptv approximately (see Fig. 16d). This indicates that the simplified approach of Cantrell et
625 al. (2003b) is not applicable to the more complex non-linear processes are involved in the air masses investigated than those considered in Eq. 11 (see Fig.17d) within EMeRGe.
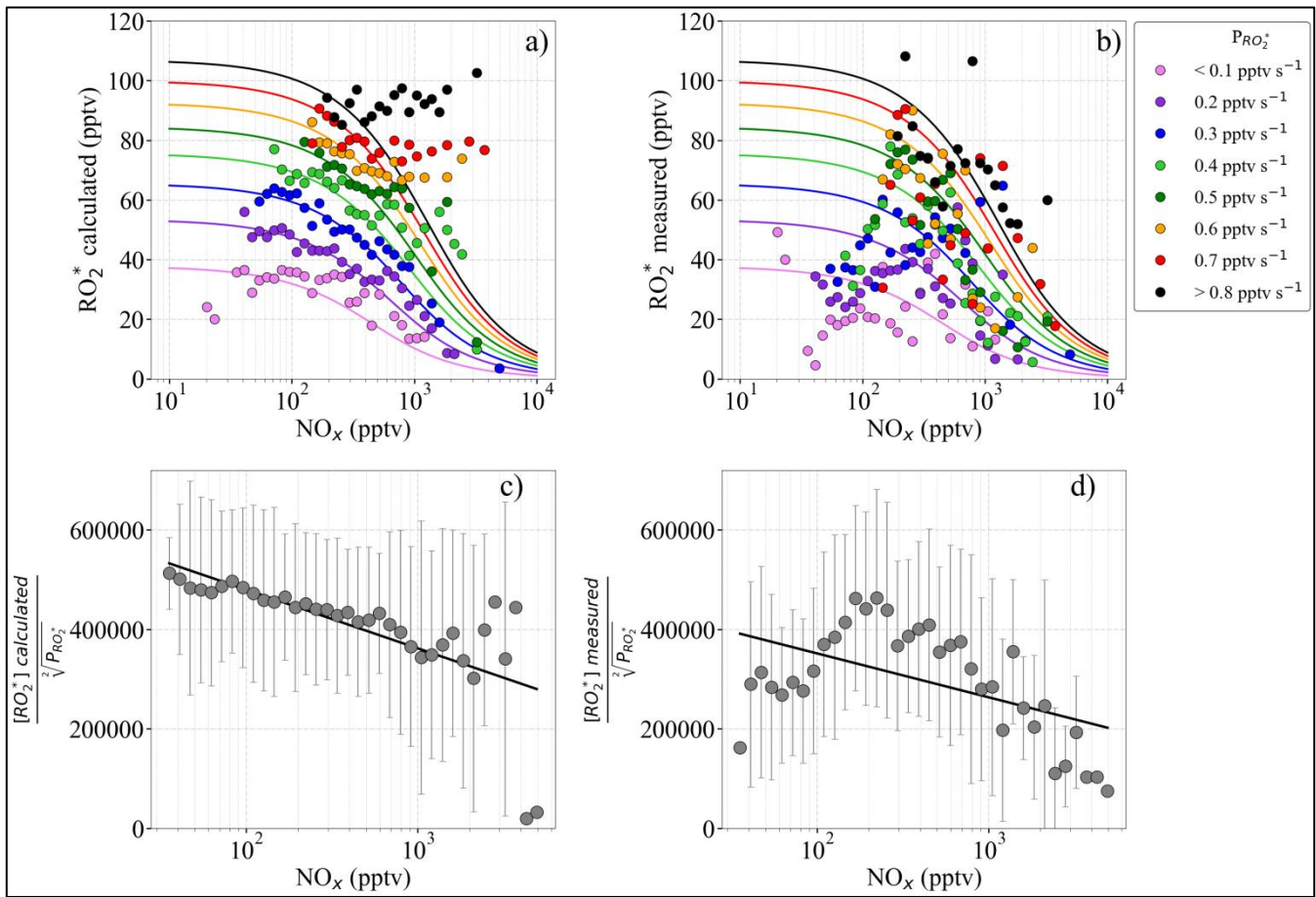
Figure 1716: a) RO$_2$$^*_c$ binned into with P$_{RO_2^*}$ intervals of 0.1 pptv s^{-1} versus binned NO$_x$; b) RO$_2$$^*_m$ binned into P$_{RO_2^*}$ intervals of 0.1 pptv s^{-1} versus binned NO$_x$ binned with P$_{RO_2^*}$; c) ratio of [RO$_2^*$]$_c$ to $\sqrt[2]{P_{RO_2^*}}$ versus NO$_x$ratio; and d) [RO$_2^*$]$_m$ to $\sqrt[2]{P_{RO_2^*}}$ ratio versus NO$_x$. The data in the a) and b) panels are assigned different colours,coloured as a function of thewith radical production rate interval. The NO$_x$ observations are binned into 50 equidistant intervals on a logrithmic scales. The solid lines in paneles a) and b) are the least square fits, obtained using Eq. 11Eq. 9. In panels c) and d) tThe ratio of [RO$_2^*$]$_c$ to $\sqrt[2]{P_{RO_2^*}}$ and [RO$_2^*$]$_m$ to $\sqrt[2]{P_{RO_2^*}}$ RO$_2^*$ forare different P$_{RO_2^*}$ ranges are binned into 50 NOx intervals bins over NO$_x$ equidsistant on in the logarithmic scale from 10 to 10000 pptv. The solid lines are the least square fits obtained using Eq. 11. Error bars indicate the standard error (i.e. ±1σ, where σ is one standard deviation) for the distribution in each binof the bins.

4.4.2 O$_3$ production rate

The O$_3$ production rate (P$_{O_3}$) is calculated from the EMeRGe Europe dataset using the reaction of RO$_2$* with NO in a similar manner to that used in previous studies of photochemical processes in urban environments (e.g. Kleinman et al., 1995; Volz-Thomas et al., 2003; Mihelcic et al., 2003; Cantrell et al., 2003b; and references herein).

$$P_{O_3} = k_{RO_2^*+NO}[RO_2^*][NO] \hspace{3cm} \text{(Eq. 13Eq. 11)}$$

where $k_{RO_2^*+NO}$ is taken as the average of k_{HO_2+NO} and $k_{CH_3O_2+NO}$.

Figure 18-17 shows plots of the mean P$_{O_3}$ calculated using Eq. 13Eq. 11 from the RO$_2$$^*_m$ and RO$_2$$^*_c$ as a function of NO. The measurements are binned into 50 NO mixing ratio bins. The bin size increases with NO to keep the points equidistant onin the logarithmic scale. The calculated P$_{O_3}$ forusing the RO$_2$$^*_m$ and RO$_2$$^*_c$ agree well within the standard deviation of the bins.
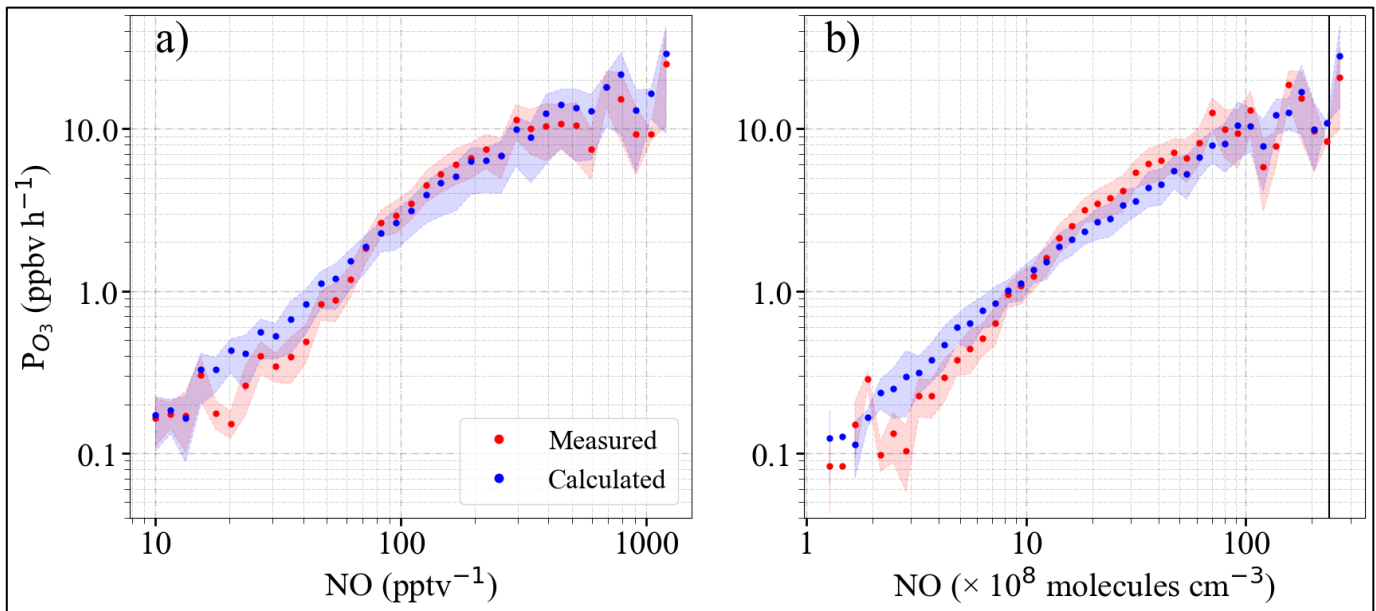
Figure 1817: Calculated O_3 production rate (P_{O_3}) determined using $RO_2^*{}_m$ (red dots) and $RO_2^*{}_c$ (blue dots) as a function of: a) NO mixing ratio; b) NO number density concentration. The 1-minute measurements are binned into 50 bins of ver NO equidistant onin the logarithmic scale for panel a) from 10 to 10000 pptv and for panel b) from and from 5×10^7 to 3.5×10^{10} molecules cm^{-3} in 15a and 15b, respectively. The shaded area shows the standard error(i.e. where σ is one ± 1σ standard deviation of each bin. To facilitate For comparison with ground-based measurements, the black line plotted in panel b)) is the number density concentration corresponding to 1 ppbv NO at 1000 mbar and 25°C has been marked by the solid black line in plot b).

Similar P_{O_3}-values have been reported for ground-based measurements in polluted areas such as Wangdu (Tan et al., 2017) and Beijing (Whalley et al., 2021) and similar ranges of peroxy radicals and NO mixing ratios. In previous work, Whalley et al. (2018) calculated P_{O_3} to be about an order of magnitude lower than that found in this study from observations in central London for about an order of magnitude lower amount of $HO_2 + RO_2$. For NO > 1 ppbv, the P_{O_3} estimated from the measurement of HO_2 and RO_2, or from the assumptions of an HO_2 to RO_2 ratio were underestimated by the models in other studies in the urban atmosphere (e.g. Martinez et al., 2003; Ren et al., 2003; Kanaya et al., 2008; Mao et al., 2010; Kanaya et al., 2012; Ren et al., 2013; Brune et al., 2016; Griffith et al., 2016). This behaviour is generally attributed to an underestimateing of large RO_2 concentrations, which likely undergo multiple bimolecular reactions with NO before forming an HO_2 radical.

During the EMeRGe IOP campaign in Europe, the NO mixing ratios were < 1 ppbv (approximately < 3×10^{10} molecules cm^{-3}). The ozone production rates obtained for both $RO_2^*{}_m$ and $RO_2^*{}_c$ measured and calculated RO_2^* are in reasonable agreement with other modelling studies in urban environments where the mixing ratio of NO is < 1 ppbv. (Tan et al., 2017; Whalley et al., 2021)

5. Summary and conclusions

This study exploits the airborne measurements of various atmospheric constituents on-board the HALO research aircraft over Europe in summer 2017 to investigate radical photochemistry in the probed airmasses. RO_2^* are calculated by assuming a **p**hotostationary **s**teady-**s**tate (PSS) of RO_2^* and compared with the actual measurements. The calculation is constrained by the simultaneous airborne measurements of radical precursors, photolysis frequencies and reactants of RO_2^* such as NO_x and O_3. The calculated radical production rates $P_{RO_2^*}$ do not significantly vary with altitude in the air masses investigated as the increase in the photolysis frequencies as a function of altitude is concurrent with decreases in precursor concentrations.

The significance and the importance of selected ~~production and loss~~initiating and terminating processes in the RO$_2^*$ chemistry are investigated by gradually increasing the complexity of the analytical expression. The agreement of the calculations with the measurements over a wide range of chemical composition and insolation conditions improves when the analytical expression is extended to account for effective ~~oxy peroxy~~ radical yields ~~from~~ VOC oxidation and radical losses through nitrates and nitrites formation. ~~interconversion reactions and subsequent loss of OH and RO during the interconversion.~~ The RO$_2^*$ measured is usually overestimated when NO is < 50 pptv in the air probed. This behavior might be explained by ~~is attributed~~ ~~to~~ RO$_2^*$ loss processes involving reactions with OH ~~, which are not considered in the analytical expression~~(e.g. the reaction of HO$_2$ with OH, but possibly to a lesser extent the three body reaction of OH with itself to make H$_2$O$_2$). These reactions may become significant RO$_2^*$ loss processes at low NO concentrations as measured during the campaign but~~The reactions~~ are excluded from the analytical expression, which is~~to~~ constrained by ~~it with~~ on-board measurements. ~~These reactions may become significant RO$_2^*$ loss processes at low NO concentrations~~.

The RO$_2^*$ calculated under assumption of a photo~~t~~stationary state ~~of~~ ~~results indicate that the~~ ~~steady state calculations~~ mostly underestimated the RO$_2^*$ ~~resulted in estimated which lower than the RO$_2^*$~~ measured~~ments~~ in polluted plumes of urban origin at altitudes below 2000 m. Changes in the HO$_2$ to RO$_2$ ratios in different plumes can ~~partly~~ account for the disagreement in particular cases. In pollution plumes with the sum of the ~~-~~OVOCs measured mixing ratios being ~~>~~ higher than 7 ppbv approximately, the underestimation of the measurements can reach up to 80 %. In these plumes, the oxidation and/or photolysis of ~~non-measured~~ VOCs, which were not measured, ~~-~~and the ozonolysis of alkenes might be significant sources of RO$_2^*$, limiting the accuracy of the analytical expression. ~~The high values for~~ overestimation of the OH concentration calculated ~~using~~ based on the measured reactants also ~~that VOC and~~ indicates missing ~~oxy peroxy radical interconversion reactions~~ are missing in the analytical PSS expression. More information about peroxy radical speciation and VOC partitioning is required to better describe the fast photochemistry in these pollution plumes.

However, t~~T~~he analytical expression developed is robust enough to simulate the radical chemistry in most of the conditions in the free troposphere encountered during EMeRGe ~~IOP~~ in Europe. Speciated radical and VOC measurements in future campaigns would facilitate the estimation of radical loss reactions in air masses having ~~with~~ NO < 50 pptv and improve radical production rates estimations in pollution plumes having ~~with~~ a high amount of VOCs, where non-linear complex chemistry is involved. Comparing RO$_2^*$ measurements with RO$_2^*$ calculations from the analytical expression helps to identify different chemical and physical regimes, which can be used to constrain future model studies.

The calculated O$_3$ production rates for NO < 1 ppbv are in the same order of magnitude as those previously reported for urban environments. This indicates that the selected RO$_2^*$ production and loss processes and observations of the radical precursors on-board are, to a good approximation, adequate for the estimation of the O$_3$ production in the measured airmasses in the free troposphere over Europe.

Acknowledgements

6. References

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