

1 **Functionality-Based Formation of Secondary Organic Aerosol from *m*-Xylene**

2 **Photooxidation**

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11 **ABSTRACT.** Photooxidation of volatile organic compounds (VOCs) produces condensable
12 oxidized organics (COOs) to yield secondary organic aerosol (SOA), but the fundamental chemical
13 mechanism for gas-to-particle conversion remains uncertain. Here we elucidate the production of
14 COOs and their roles in SOA and brown carbon (BrC) formation from *m*-xylene oxidation by
15 simultaneous monitoring the evolutions of gas-phase products and aerosol properties in an
16 environmental chamber. Four COO types with the distinct functionalities of dicarbonyls,
17 carboxylic acids, polyhydroxy aromatics/quinones, and nitrophenols are identified from early-
18 generation oxidation, with the yields of 25%, 37%, 5%, and 3%, respectively. SOA formation
19 occurs via several heterogeneous processes, including interfacial interaction, ionic
20 dissociation/acid-base reaction, and oligomerization, with the yields of (20 ± 4)% and (32 ± 7)%
21 at 10% and 70% relative humidity (RH), respectively. Chemical speciation shows the dominant
22 presence of oligomers, nitrogen-containing organics, and carboxylates at high RH and
23 carboxylates at low RH. The identified BrC includes N-heterocycles/N-heterochains and

24 nitrophenols, as evident from reduced single scattering albedo. The measured uptake coefficient
25 (γ) for COOs is dependent on the functionality, ranging from 3.7×10^{-4} to 1.3×10^{-2} . A
26 functionality-based kinetic framework is developed to predict SOA production from the
27 measurements of the concentrations and uptake coefficients for COOs, which well reproduces
28 SOA formation from *m*-xylene oxidation. Our results reveal that photochemical oxidation of *m*-
29 xylene represents a major source for SOA and BrC formation under urban environments, because
30 of its large abundance, high reactivity with OH, and high yields for COOs.

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33 **1. Introduction**

34 Photooxidation of anthropogenic and biogenic volatile organic compounds (VOCs)
35 produces tropospheric ozone, secondary organic aerosol (SOA), and brown carbon (BrC), with
36 profound implications for air quality, human health, and climate (Pope et al., 2002; Li et al., 2007;
37 IPCC, 2013; NASEM, 2016; Molina, 2021; Zhang et al., 2021). For example, SOA contributes to
38 the Earth energy budget, directly by scattering solar radiation and indirectly by serving as cloud
39 condensation nuclei to influence cloud formation and precipitation (IPCC, 2013; Wang et al., 2014;
40 Zhu et al., 2017). Also, light absorbing BrC interferes with solar radiation transfer, contributing to
41 positive radiative forcing (Wang et al. 2013; NASEM, 2016). VOC oxidation is initiated by various
42 oxidants (e.g., OH, O₃, NO₃, etc.) and proceeds via multiple pathways and stages (Atkinson, 2000;
43 Suh et al., 2001; Zhang et al., 2002; Zhao et al., 2004; Wennberg et al., 2018), yielding condensable
44 oxidized organics (COOs) to form SOA and BrC via gas-to-particle conversion (Finlayson-Pitts
45 and Pitts, 2000; Moise et al., 2015; Seinfeld and Pandis, 2016). Currently, the enormous chemical
46 complexity for VOC oxidation and gas-to-particle conversion represents one of the greatest
47 challenges in atmospheric chemistry research (Ravishankara, 1997; Zhang et al., 2015; NASEM,

48 2016). Aromatic hydrocarbons (e.g., benzene, toluene, xylenes, and trimethylbenzene) account for
49 20-30% of the total VOCs and are the major anthropogenic SOA precursors in the urban
50 atmosphere (Calvert et al., 2002; Ng et al., 2007; Song et al., 2007; Guo et al., 2014; Seinfeld and
51 Pandis, 2016). For example, *m*-xylene or $C_6H_4(CH_3)_2$ represents an important type of aromatic
52 hydrocarbons, which is emitted primarily from industrial and traffic sources. The concentration of
53 *m*-xylene ranges in a level from sub part per billion (ppb) up to several tens of ppb under urban
54 environments (Calvert et al., 2002; Fortner et al., 2009). Photooxidation of *m*-xylene is primarily
55 initiated by OH with a rate constant of $2.4 \times 10^{-11} \text{ cm}^3 \text{ molecule}^{-1} \text{ s}^{-1}$, which is nearly four times
56 higher than that of toluene (Fan and Zhang, 2008; Ji et al., 2017). A laboratory study identified *m*-
57 tolualdehyde, *m*-dimethylphenols, and dicarbonyls (i.e., glyoxal, methylglyoxal, unsaturated
58 dicarbonyls, and epoxy carbonyls) as the main gas-phase products from OH-initiated oxidation of
59 *m*-xylene (Zhao et al., 2005).

60 Several types of COOs with distinct functionality and volatility are produced from *m*-
61 xylene photooxidation, contributing to aerosol nucleation and grow (Zhang et al., 2004; Zhang et
62 al., 2015; Guo et al., 2020). SOA formation is conventionally considered to be mainly resulted
63 from equilibrium gas-particle partitioning of semi-, intermediate-, or low-volatile products
64 (Shrivastava et al., 2017). Aromatic oxidation by OH yields low-volatile polyhydroxy
65 aromatics/quinones (denoted as PAQ hereafter), which condense to the aerosol-phase (Schwantes
66 et al., 2017). Also, chain autoxidation reactions of RO_2 from aromatics photooxidation likely yield
67 highly oxygenated molecules (HOMs) with low volatility (Molteni et al., 2018; Garmash et al.,
68 2020; Wang et al., 2020), although the reported yield of HOMs from *m*-xylene oxidation is small
69 (1.0-1.7%) (Molteni et al., 2018). Several key oxidation products of aromatic oxidation have been
70 shown to readily engage in multiphase reactions and contribute to SOA formation (Ji et al., 2020,

71 Li et al., 2021a). A recent experimental study demonstrated that SOA formation from toluene
72 photooxidation is mainly contributed by volatile dicarbonyls and organic acids (Li et al., 2021b).
73 Moreover, the aqueous reactions between small α -dicarbonyls and base species (e.g., ammonia
74 and amines) produce light-absorbing brown carbon (BrC) (De Haan et al., 2011, 2017; Marrero-
75 Ortiz et al., 2019; Li et al., 2021a,b). Previous experimental studies identified the presence of
76 hydroxy (C-OH), carbonyl (C=O), and acetal (C-O-C) functional groups in aromatic-derived SOA
77 formed via aqueous reactions, implicating a role of oxygenated organics in SOA formation (Jia
78 and Xu, 2014; Jia and Xu, 2018; Zhang et al., 2019). Also, laboratory experiments showed that
79 coating of *m*-xylene SOA on black carbon particles significantly enhances scattering, absorption,
80 and single scattering albedo (SSA) (Guo et al., 2016).

81 Current atmospheric models mainly parameterize SOA formation based on equilibrium
82 partitioning for semi- to low-volatile COOs (Shrivastava et al., 2017). However, gas-to-particle
83 conversion corresponds to decreasing entropy (i.e., $\Delta S \ll 0$), representing non-equilibrium
84 chemical processes (Peng et al., 2021). Moreover, the occurrence of particle-phase reactions
85 significantly alters the physiochemical properties (including volatility, hygroscopicity, and optical
86 properties) for COOs (Tan et al., 2012; Faust et al., 2017; Ji et al., 2020; Li et al., 2021a,b; Liu et
87 al., 2021). Notably, the volatility-based approach consistently under-predicts SOA formation
88 (Heald et al., 2005; Zhang et al., 2015; Hodzic et al., 2016), particularly during haze formation
89 under polluted conditions (Guo et al., 2014; Peng et al., 2021). Here we investigate the production
90 of COOs with different functionalities and assess their roles in SOA formation from *m*-xylene
91 oxidation. A primary objective of this work is to establish a functionality-based framework to
92 predict SOA formation from VOC photooxidation.

93 2. Experimental Methodology

94 2.1 Chamber experiments

95 The production of COOs and their roles in SOA formation from *m*-xylene oxidation were
96 investigated using a 1 m³ chamber (Teflon® PFA, Fig. S1 in the Supplementary Information or
97 SI), similar to our previous studies (Li et al., 2021a,b). The environmental chamber was equipped
98 with eighteen black light lamps (18 × 30W, F30T8/350BL, Sylvania). A water bubbler at a
99 temperature of 30°C was used to humidify the chamber to 10%, 30%, 50%, or 70% relative
100 humidity (RH), and all experiments were performed at 298 K. Three types of seed particles were
101 produced using a constant output atomizer (Model 3076, TSI) to represent various chemical
102 compositions, i.e., NH₄HSO₄ - ammonium bisulfate (ABS), (NH₄)₂SO₄ - ammonium sulfate (AS)
103 in the presence and absence of NH₃, and NaCl - sodium chloride. Seed particles were dried to RH
104 of ~2% by a Nafion drier (PD-070–18T-12SS, Perma Pure) and size-selected for an initial size of
105 100 nm by a differential mobility analyzer (DMA, Model 3081, TSI). Size-selected particles were
106 injected into the chamber with an initial particle concentration of 1.5 × 10⁴ cm⁻³ measured by a
107 condensation particle counter (CPC, Model 3760A, TSI). The acidity of seed particles was
108 estimated using a thermodynamic model (Fountoukis and Nenes, 2007; Wang et al., 2018), with
109 the pH value of 3~5 in the presence of gaseous NH₃ and 0~1 in the absence of gaseous NH₃ for
110 (NH₄)₂SO₄, -0.5 for NH₄HSO₄, and 7 for NaCl. To investigate the effects of NH₃ and NO_x,
111 commercially available gaseous NH₃ (2000 ppm NH₃ in N₂) and NO₂ (500 ppm NO₂ in air) was
112 injected into the chamber with a flow of 100 sccm. 600 μL H₂O₂ (35 wt%, Sigma-Aldrich) and 10
113 μL *m*-xylene (99.5%, Sigma-Aldrich) were injected into a glass reservoir, which was then flushed
114 into the chamber by a 10 slpm flow from a pure air generator (Aadco 737-11, Aadco Inc.) for 10
115 min. The gases were then mixed by a fan inside the chamber. Once desired concentrations were
116 established, the black light lamps were turned on to initiate H₂O₂ photolysis to generate OH

117 radicals. The initial concentration for *m*-xylene and the steady-state concentration for OH were
118 estimated to be 1.9 ppm and 2.1×10^6 molecules cm^{-3} , respectively (Fig. 1a). The experimental
119 conditions are summarized in Table 1.

120 2.2 Analytical instruments

121 The particle size growth, density, light scattering/absorption were simultaneously and
122 continuously monitored, after monodispersed seed particles were exposed to the products of *m*-
123 xylene-OH oxidation in the reaction chamber. A scanning mobility particle sizer (SMPS)
124 consisting of a DMA and a CPC was used to continuously measure the particle size growth factor,
125 $GF = D_p/D_0$, where D_p is the diameter after exposure and $D_0 = 100$ nm is the initial diameter. Note
126 both D_p and D_0 correspond to the dry particle diameter measured after passing through a Nafion
127 dryer ($\sim 2\%$ RH). An aerosol particle mass analyzer (APM, Model 3600, Kanomax) was used to
128 derive the particle density throughout the experiments (Li et al., 2021b). The density for the SOA
129 materials (ρ_{SOA}) is calculated by,

$$\rho_{SOA} = \frac{\rho \times GF^3 - \rho_0}{GF^3 - 1} \quad (1)$$

130 where ρ and GF are measured particle density (g cm^{-3}) and growth factor at t (min), respectively.
131 ρ_0 is the initial density of seed particles. The uncertainty in the density measurements was
132 estimated to be ± 0.03 g cm^{-3} .

133 To measure the optical properties of exposed particles, a 1-slp flow from the chamber
134 was diluted by 4-slp N_2 and introduced into a commercial integrating nephelometer (TSI 3563)
135 and a home-built cavity ring-down spectrometer (CRDS) throughout the experiment. Light
136 scattering (b_{sca}) and extinction (b_{ext}) at 532 nm were determined by the nephelometer and CRDS,
137 respectively. The nephelometer was calibrated by comparing measured b_{sca} with b_{ext} from CRDS
138 when measuring ammonium sulfate particles with diameters of 100 nm, 150 nm, 200 nm, 250 nm,

139 and 300 nm. The absorption coefficient (b_{abs}) was determined from ($b_{\text{ext}} - b_{\text{sca}}$), and SSA was
140 calculated from $b_{\text{sca}}/b_{\text{ext}}$.

141 The ID-CIMS using the hydronium reagent ion (H_3O^+) was employed to analyze gaseous
142 concentrations with a sample flow rate of 0.5 slpm from the chamber. The sample line was heated
143 to 70°C to reduce wall loss. Gas-phase concentrations of *m*-xylene and its oxidation products were
144 quantified by ID-CIMS. Briefly, the concentration of species A from the proton transfer reaction
145 ($\text{H}_3\text{O}^+ + \text{A} \rightarrow \text{H}_2\text{O} + \text{HA}^+$) is determined by,

$$[\text{A}] = \frac{S_{\text{A}}}{k_{\text{PTR}} \times \Delta t \times S_{\text{RI}}} \quad (2)$$

146 where S_{A} and S_{RI} are the mass spectrum intensities for A and reagent ions, respectively. k_{PTR} is the
147 proton-transfer rate constant between A and H_3O^+ (Zhao and Zhang, 2004). $\Delta t = \frac{l}{U_i}$ is the retention
148 time of ions, which is determined by the length of the drift tube $l = 8$ cm and the ion drift velocity
149 U_i . The ion drift velocity U_i is determined from:

$$U_i = u_0 \left(\frac{760 \text{ torr}}{p} \right) \left(\frac{T}{273 \text{ K}} \right) E \quad (3)$$

150 where $u_0 = 2 \text{ cm}^2 \text{ V}^{-1} \text{ s}^{-1}$ is the reduced ionic mobility (Dotan et al., 1976), $p = 3.0$ torr, $T = 298$ K,
151 and $E = 133 \text{ V cm}^{-1}$ is the voltage gradient in the drift tube. Alternatively, the concentrations of *m*-
152 xylene, dicarbonyls, and organic acids were also calibrated by preparing the vapors of *m*-xylene,
153 methylglyoxal with known mixing ratios in N_2 , and using a permeation tube of propanoic acid,
154 respectively (Li et al., 2021a). The detection limit (defined as 3 times of the ratio of signal to noise)
155 for the oxidation products from *m*-xylene-OH reactions was estimated to be 50 ppt by the ID-
156 CIMS.

157 The steady-state OH concentration was determined by exponential fitting ($y = ke^{-bx}$) of
158 the decay of *m*-xylene concentration (Fig. 1a),

$$[\text{OH}] = \frac{b}{k_1} \quad (4)$$

159 where $k_1 = 2.4 \times 10^{-11} \text{ cm}^3 \text{ molecule}^{-1} \text{ s}^{-1}$ is the second-order rate constant for *m*-xylene-OH
 160 oxidation (Fan and Zhang, 2008). The estimated concentration of OH was $2.1 \times 10^6 \text{ molecules cm}^{-3}$
 161 ³ in our experiments.

162 We also estimated loss of condensable vapors to the chamber wall using the first-order
 163 wall-loss coefficient, $k_w \text{ (s}^{-1}\text{)}$, by considering gas-phase transport within the chamber (Zhang et
 164 al., 2014)

$$k_w = \left(\frac{A}{V}\right) \frac{\alpha_w \bar{c}}{4 + \frac{\pi}{2} \left(\frac{\alpha_w \bar{c}}{\sqrt{k_e D_g}}\right)} \quad (5)$$

165 where A/V is the surface to volume ratio of the chamber (equal to $6/L$ for a square chamber, where
 166 $L = 1 \text{ m}$ is its dimension), α_w is the mass accommodation coefficient of vapors onto Teflon chamber
 167 walls, \bar{c} is the mean thermal speed of the molecules, k_e is the coefficient of eddy diffusion, which
 168 is estimated to be 0.5 based on the loss rate of particles (McMurry and Rader, 1985),

$$k_e = \frac{\left(\frac{\pi L \beta_0}{12}\right)^2}{D_b} \quad (6)$$

169 where β_0 is the fractional loss rate of particles,

$$\beta_0 = \frac{\Delta N}{N \times \Delta t} = \frac{15000 \text{ cm}^{-3} - 7500 \text{ cm}^{-3}}{15000 \text{ cm}^{-3} \times 7200 \text{ s}} = 1.4 \times 10^{-5} \text{ s}^{-1} \quad (7)$$

170 where N is the particle number concentration, $D_b = 6.75 \times 10^{-10} \text{ m}^2 \text{ s}^{-1}$ is the Brownian diffusion
 171 coefficient for particles. D_g is the gas-phase diffusion coefficient. For NH_3 , the mass
 172 accommodation coefficient of vapors onto the chamber wall is 0.05 (Zhang et al., 2014). Gas-
 173 phase diffusion constant D_g is $1.98 \times 10^{-5} \text{ m}^2 \text{ s}^{-1}$. The mean thermal speed \bar{c} is 603 m s^{-1} . The first-
 174 order wall-loss coefficient of NH_3 was calculated to be $1.2 \times 10^{-2} \text{ s}^{-1}$.

175 The average concentration of NH₃ or NO_x (noted as X) is estimated by,

$$\overline{[X]} = \frac{(1 - e^{-k_w \times \Delta t_i})}{k_w \times \Delta t_i} [X]_0 \quad (8)$$

176 For initial injection of 400 ppb NH₃, the average concentration of NH₃ was estimated to be 19 ppb.

177 The first-order wall-loss rate of NO_x was measured to be $2.5 \times 10^{-3} \text{ h}^{-1}$, leading to negligible wall-

178 loss (Qi et al., 2020).

179 The particle-phase chemical composition was analyzed by thermal desorption (TD) - ID -
180 CIMS (TD-ID-CIMS) (Wang et al., 2010). Seed particles after 20 min of exposure to
181 photooxidation were collected for 2 hours by a platinum filament (with a collection voltage of
182 about 3000 V) in a 2.5 slpm flow from the reaction chamber, and the analytes were evaporated by
183 heating the filament to 350°C for 2 s and detected by ID-CIMS using H₃O⁺ as the reagent ions.
184 The mass resolution of TD-ID-CIMS was about 0.5 amu. The desorption signal was represented
185 by the relative intensity (RI) of the integrated peak area during heating. The uncertainties of TD-
186 ID-CIMS measurements arose from the flow rate, voltage, collection time, evaporation voltage,
187 and mass spectrometer ionization/detection efficiencies during particle collection and were
188 represented by the standard deviation of three repeated measurements. In both ID-CIMS and TD-
189 ID-CIMS configurations, an ion drift-tube was used. An electric field of E/N = 138 Td was applied
190 for ID-CIMS at high RH (i.e., 70%), and E/N = 110 Td was used for TD-ID-CIMS at low RH (<
191 1%). Formation of cluster ions was effectively suppressed, since a uniform electrical field in the
192 drift tube decomposed weakly-bonded cluster ions. Fragmentation using ID-CIMS or TD-ID-
193 CIMS was minimal, and all gaseous and particle-phase products were detected at their respective
194 parent-peaks, indicating little dissociation effect from either thermal desorption or ionization by
195 H₃O⁺ (Yuan et al., 2017). For example, plausible fragment-ion peaks for oligomers and organic
196 acids are the monomeric ion-peaks and [M-17]⁺/[M-45]⁺ from α-cleavage, respectively. From our

197 detected particle-phase products (Fig. 2c), non-fragmental oligomers/NCO and organic acids
198 represented the dominant peaks, while negligible monomer (e.g., $m/z = 73$) and fragments for
199 organic acids (e.g., [M-17] or [M-45]) were detected, confirming little fragmentation by TD-ID-
200 CIMS measurement. Also, the major acetal/hemiacetal oligomers detected by TD-ID-CIMS at
201 high RH from *m*-xylene oxidation were consistent with a previous study showing intact oligomers
202 with thermal desorption (Claflin and Ziemann, 2019).

203 While the gaseous concentrations of the oxidation products, GF, SSA, and density were
204 temporally resolved in our experiments, the particle chemical composition reflected an overall
205 aerosol chemical makeup during an experiment. To assess the effects of different seed particles,
206 RH, and NH₃ on SOA and BrC formation, we compared the measured GF and SSA values after
207 120-min exposure to the *m*-xylene-OH oxidation.

208 2.3 Uptake coefficient (γ) determination

209 The uptake coefficient (γ) for COOs is calculated by,

$$\gamma = \frac{4 \times FR \times kp}{\bar{c} \times S} \quad (9)$$

210 where *FR* is the percent contributions of the COO species from *m*-xylene-OH oxidation to the
211 particle-phase mass intensity measured by TD-ID-CIMS. *kp* is measured first-order particle uptake
212 rate constant for COOs (i.e., glyoxal, methylglyoxal, γ -dicarbonyls, organic acids, PAQ, and
213 nitrophenols), which is calculated from the measured growth rate of seed particles,

$$kp = \frac{\pi \times (D_p^3 - D_0^3) \times \rho \times \bar{N} \times N_A}{6 \times MW \times [A] \times \Delta t} \quad (10)$$

214 where D_p and D_0 are the final and initial diameters of the particles during the time-period of 120
215 min, respectively, ρ is the density of total SOA, $\bar{N} = 11250 \text{ cm}^{-3}$ is the average particle number
216 concentration during 120 min, $\Delta t = 120 \text{ min}$ is the exposure time, N_A is Avogadro constant, MW

217 is the molar weight of species A, and $[A]$ is the average gas phase concentration of species A. \bar{c} is
 218 the mean thermal speed of A. S is the average surface area of aerosols ($\text{cm}^2 \text{ surface}/\text{cm}^3 \text{ air}$) during
 219 the period of $\Delta t = 120 \text{ min}$,

$$S = \frac{\pi \int_{t=0}^{120 \text{ min}} D^2 \times HGf^2 \times \bar{N} dt}{\Delta t} \quad (11)$$

220 where D is the time-dependent particle diameter, HGf is the hygroscopic growth factor of particles
 221 (Seinfeld and Pandis, 2016). The error bars of reactive coefficients are derived from the
 222 propagation of uncertainties for the parameters in eqs. 9-11 (i.e., 1σ of at least three repeated
 223 measurements for each parameter).

224 **2.4 Simulation of gaseous oxidation products**

225 The gas-phase concentration of *m*-xylene-OH oxidation products were simulated by a tri-
 226 exponential kinetic model to account for the gas-phase oxidation, chamber wall loss, and particle
 227 uptake. The precursor A with an initial concentration of $[A]_0$ undergoes multiple steps of oxidation
 228 to yield multi-generation products, P_n (where $n = 1, 2, 3, \dots$). Once formed, each product engages
 229 in further oxidation with a pseudo first-order rate constant of $k'_{n+1} = k_{n+1}[\text{OH}]_{\text{ss}}$, particle uptake
 230 with a rate constant of kp_n , and wall loss with a rate of kw_n . The forward rate constants, k_{n+1} , were
 231 taken from MCM 3.3.1 (Jenkin et al., 2003). The rate constant of uptake kp_n is expressed by,

$$kp_n = \frac{1}{4} \gamma_n \bar{c} S \quad (12)$$

232 and the first-order wall loss rate constants were estimated by eq. 5 with the mass accommodation
 233 coefficients α_w constrained by the measured time-dependent concentration. The gas-phase
 234 concentrations of P_1 to P_3 are calculated by,

$$[P_1] = \frac{k'_1 [A]_0}{k'_2 + kw_1 + kp_1 - k'_1} (e^{-k'_1 t} - e^{-(k'_2 + kw_1 + kp_1)t}) \quad (13)$$

$$[P_2] = C_1 e^{-(k'_3 + kw_2 + kp_2)t} + C_2 e^{-k'_1 t} - C_3 e^{-(k'_2 + kw_1 + kp_1)t} \quad (14)$$

$$[P_3] = C_4 e^{-(k'_4 + kw_3 + kp_3)t} + C_5 e^{-(k'_3 + kw_2 + kp_2)t} + C_6 e^{-k'_1 t} - C_7 e^{-(k'_2 + kw_1 + kp_1)t} \quad (15)$$

235 where C_1 to C_7 are constants, and k'_n denotes the pseudo-first order rate constant,

$$k'_n = k_n[\text{OH}] \quad (16)$$

236 The rate constants and values for C_1 to C_7 are summarized in Table S1.

237 2.5 COO yield

238 The COO yield (Y_{COO}) from *m*-xylene oxidation is determined from,

$$239 Y_{\text{COO}} = [\text{COO}_i] / \Delta[\text{C}_6\text{H}_4(\text{CH}_3)_2] \quad (17)$$

240 where i represents the type of COO species, and $\Delta[\text{C}_6\text{H}_4(\text{CH}_3)_2]$ denotes the concentration of *m*-
 241 xylene consumed due to OH oxidation. There were four types of COOs identified from our
 242 experiments, i.e., dicarbonyls for $i = 1$, carboxylic acids for $i = 2$, PAQ for $i = 3$, and nitrophenols
 243 for $i = 4$. The measured gaseous concentration for each COO is corrected for wall-loss and
 244 unreacted earlier generation products,

$$[\text{COO}_i] = \left\{ [\text{COO}_{i,f}] + \left(\frac{1}{4} \gamma_i \bar{c} S + kw_i \right) \overline{[\text{COO}_i]} \Delta t \right\} / \left(1 - \frac{[A_f]}{[A_t]} \right) \quad (18)$$

245 where $[\text{COO}_{i,f}]$ and $\overline{[\text{COO}_i]}$ are the final and time-averaged gas-phase concentrations of COO_i at
 246 120 min, respectively. $[A_f]$ and $[A_t]$ are the unreacted and the total production of the earlier
 247 generation species A, respectively.

248 2.6 SOA mass concentration and yield

249 The measured SOA mass concentration is calculated based on the aerosol size growth and
 250 density,

$$M_{\text{SOA,meas.}} = \frac{1}{6} \pi (D_p^3 \times \rho_t - D_0^3 \times \rho_0) \times \bar{N} \quad (19)$$

251 where ρ_t and ρ_0 are the final and initial particle density during 120 min, respectively.
 252 Alternatively, by categorizing COOs in accordance with the functional groups, we established a
 253 framework to predict SOA formation from the measured production and reactivity. The SOA mass
 254 concentration (M_{SOA}) is predicted from the measured uptake coefficients and average gas-phase
 255 concentrations of COOs,

$$M_{SOA} = \sum_i^4 \frac{1}{4} \gamma_i \bar{c}_i S [A_i] MW_i \quad (20)$$

256 where MW_i is the molar weight of species i . γ_i is averaged over all identified species for each type
 257 (i) of COOs, and $[A_i]$ is the sum of the measured concentrations of all identified species for type i
 258 (time-averaged). The SOA mass yield is expressed by,

$$Y_{SOA} = M_{SOA} / \{ \Delta [C_6H_4(CH_3)_2] \times MW_m \} \quad (21)$$

260 where MW_m is the molar weight of m -xylene.

261 To compare the measured and predicted SOA formation at RH = 10% and 70%, we
 262 corrected the increased vapor wall loss and reduced particle surface area at RH = 70% compared
 263 to 10% RH. A correction factor is determined from the average measured gas concentration ratios
 264 (R_c) between 10% and 70% RH for each type of COOs. Additionally, a hygroscopic volume
 265 enhancement factor (HVE) of 1.2 between 10% and 70% RH is adopted. The RH-corrected SOA
 266 mass for each COO is estimated by,

$$M_{i,corr} = R_c \times M_{SOA} \times \frac{S_1}{S_2} \times HVE \quad (22)$$

267 where S_1 and S_2 are the particle surface areas at 10% and 70% RH, respectively.

268 3. Results and Discussions

269 3.1 Evolution of oxidation products

270 To elucidate the production of COOs, we measured time-dependent gaseous concentrations
271 of *m*-xylene oxidation products. The OH-*m*-xylene reactions occur via dominantly OH-addition to
272 the aromatic ring to yield *m*-xylene-OH adducts and minorly H-abstraction from the methyl group
273 to form methylbenzyl radicals (about 4%) (Fan and Zhang, 2008). The *m*-xylene-OH adducts then
274 react with O₂ to form dimethylphenol via H-extraction or OH-*m*-xylene-O₂ peroxy radicals (RO₂)
275 via O₂ addition (Fig. S2a). While the *m*-xylene mixing ratio exhibits an exponential decay
276 throughout the experiments (Fig. 1a), the formation of the gas-phase oxidation products follows
277 the rate-determining steps involving successive OH-oxidation with three major generations (Fig.
278 1b-e). The first-generation products (P₁) include dimethylphenol ($m/z = 123$) and
279 methylbenzaldehyde ($m/z = 121$), corresponding to OH addition to the aromatic ring (~96%) and
280 hydrogen extraction (~4%) from the methyl group, respectively (Fan and Zhang, 2008; Li et al.,
281 2021b). The second-generation products (P₂) mainly consist of methylglyoxal ($m/z = 73$),
282 methylbutenedial ($m/z = 99$), toluic acid ($m/z = 137$), and dimethylresorcinol (Fig. 1f and Fig. S2b),
283 which are produced from the P₁ reactions with OH/HO₂. For example, methylglyoxal and
284 methylbutenedial are formed from OH oxidation of dimethylphenol and subsequent ring-
285 opening or from the primary peroxy radical. The latter undergoes cyclization to form the bicyclic
286 radical, which then reacts with O₂ to form the secondary peroxy radical, followed by reactions
287 with HO₂ and subsequent ring-cleavage (Fan and Zhang, 2008). The third-generation products (P₃)
288 contain mainly multi-functional organic acids, including pyruvic acid ($m/z = 89$), 4-oxo-2-
289 pentenoic acid ($m/z = 115$), and 3-methyl-4-oxo-2-pentenoic acid ($m/z = 129$) produced from the
290 subsequent reactions of dicarbonyls with OH. Note that P_n in Fig. 1b denotes the lumped product
291 of the n^{th} generation, which is related the sequence of OH addition for *m*-xylene oxidation and
292 accounts for the various species detected by ID-CIMS (Fig. S2b).

293 The P₁ concentrations rise immediately upon initiating photooxidation and reach the peak
294 mixing ratios of about 20 ppb for dimethylphenol and 9 ppb for methylbenzaldehyde. There is a
295 slight decline in P₁ concentrations after 70 to 100 mins, reflecting their consumption from further
296 OH oxidation (Fig. 1c). The P₂ concentrations follow those of P₁ (Fig. 1d and Fig. S3a) and reach
297 the peak mixing ratios ranging from 0.4 to 2.4 ppb, dependent on their subsequent oxidation by
298 OH as well as loss to the chamber wall and aerosols. Note that the concentration of toluic acid
299 rises later than those of the dicarbonyls, attributable to the slower rates for H-abstraction by OH
300 from the methyl group than OH-addition to the aromatic ring (Fan and Zhang, 2008). The P₃
301 concentrations increase monotonically after a delay of 10-20 mins and reach a mixing ratio from
302 sub-ppb to 2.5 ppb at 120 min (Fig. 1e and Fig. S3b). The initial concentration of *m*-xylene (e.g.,
303 1940 ppb) in our experiment is higher than that in the atmosphere, potentially inducing self- and
304 cross-reactions of RO₂ to form alkoxy radicals (RO) or dialkyl peroxides (ROOR') leading to
305 HOMs. However, negligible products relevant to HOMs were detected in our experiments,
306 indicating a minor importance for the self- and cross-reactions of RO₂ compared to the competing
307 reactions between RO₂ and HO₂/NO/RO₂ to form ring-opening products. Our results for
308 insignificant contribution of HOMs to SOA formation from *m*-xylene oxidation are consistent with
309 a small yield of HOMs reported in a previous study (Molteni et al., 2018).

310 We simulated the time-dependent evolution of the gaseous products by employing a tri-
311 exponential kinetic model (Fig. 1b and Table S1). Multi-generation oxidation of *m*-xylene occurs
312 via consecutive reaction steps to produce a multi-generation product, P_{*n*} (where *n* denotes the
313 sequence of OH oxidation), which subsequently engages in additional OH oxidation, wall loss,
314 and aerosol loss. Fig. 1b-d indicates that the simulated concentrations well reproduce those

315 measured for P₁ to P₃ by adopting our measured γ (to be discussed below) and estimated wall loss
316 rates (see Methods and Table S1 in SI).

317 **3.2 Particle size growth, SSA, and chemical speciation**

318 To evaluate SOA and BrC formation from COOs, we measured the time-dependent GF,
319 SSA, and density, after (NH₄)₂SO₄ seed particles are exposed to the *m*-xylene oxidation products
320 in the presence of 19 ppb NH₃ and at 70% RH (Fig. 2a-c, Exp. 1). The GF increases monotonically
321 and reaches a value of (2.41 ± 0.03) at 120 min (Fig. 2a), while the SSA declines steadily
322 throughout the exposure and reaches the value of (0.91 ± 0.01) at 120 min (Fig. 2b). The latter is
323 indicative of the production of light-absorbing BrC. The measured SSA for *m*-xylene is lower than
324 that previously reported for toluene under comparable experimental conditions (Li et al., 2021b).
325 The difference in SSA/BrC formation between toluene and *m*-xylene is explained by a higher yield
326 of methylglyoxal from *m*-xylene than from toluene (Jenkin et al., 2003; Nishino et al., 2010), since
327 methylglyoxal produces BrC more efficiently via aqueous reaction than glyoxal (Li et al., 2021a).
328 Analysis of particle chemical composition using TD-ID-CIMS reveals high abundances of
329 oligomers, nitrogen-containing organics (NCO) consisting of N-heterocycles/N-heterochains,
330 carboxylates, along with a small amount of dimethylresorcinol/benzoquinone (Fig. 2c). The
331 identified mass peaks are summarized in Tables S2 to S4. Assuming similar sensitivity to proton-
332 transfer reactions for the aerosol-phase organics, i.e., with the proton transfer reaction rate
333 constants of $(2 \text{ to } 4) \times 10^{-9} \text{ cm}^3 \text{ molecule}^{-1} \text{ s}^{-1}$ (Zhao and Zhang, 2004), the contributions of
334 oligomers, NCO, carboxylates, and dimethylresorcinol/benzoquinone to the total SOA formation
335 are estimated to be 22%, 23%, 47%, and 8%, respectively, at 70% RH (Table S5). Note that the
336 TD-ID-CIMS method preserves the identify for all organic species without fragmentation,
337 providing unambiguous chemical speciation for the aerosol-phase products.

338 3.3 Effects of seed particles, NH₃, RH, and NO_x

339 We assessed the dependence of SOA/BrC formation from *m*-xylene-OH oxidation on
340 different seed particles (Exp. 1-4), NH₃ concentration (Exp. 1-12), RH (Exp. 1,13-15), and NO_x
341 concentration (Exp. 1,16-18). Fig. 3a shows that (NH₄)₂SO₄ in the presence of 19 ppb NH₃ exhibits
342 the largest GF and lowest SSA at 120 min, which are explained by the chemical mechanisms
343 leading to the formation of the observed aerosol-phase products. While dicarbonyls undergo
344 aqueous phase reactions to form oligomers (Fig. S4) (Ji et al., 2020; Li et al., 2021a), organic acids
345 engage in ionic dissociation and acid-base reaction to yield carboxylates (Fig. 3b). In addition, the
346 reaction of dicarbonyls with NH₃ produces N-heterocycles/N-heterochains (Fig. S5), which are
347 light-absorbing (Marrero-Ortiz et al., 2019; Li et al., 2021a). Oligomerization of dicarbonyls
348 involves protonation, hydration, and deprotonation, which are mediated by carbenium ions via
349 nucleophilic addition (Ji et al., 2020; Li et al., 2021a). While protonation of dicarbonyls to yield
350 carbenium ions is promoted by hydrogen ion activity, high acidity hinders hydration and
351 deprotonation, explaining the largest GF and the smallest SSA on weakly acidic (NH₄)₂SO₄
352 particles in the presence NH₃ (Fig. 3a). In addition, uptake of dicarbonyls, organic acids, and PAQ
353 is likely facilitated on aqueous (NH₄)₂SO₄ particles, because of surface propensity of charge-
354 separation at the interface (Hua et al., 2011; Shi et al., 2020). Specifically, surface-abundant NH₄⁺
355 cations arising from interfacial charge separation of (NH₄)₂SO₄ likely exert electrostatic attraction
356 to gaseous oxygenated species (with a negative charge character) to enhance uptake for
357 dicarbonyls, organic acids, and PAQ (Li et al., 2021a,b). The most efficient formation of SOA and
358 BrC on (NH₄)₂SO₄ particles in the presence of NH₃ are also consistent with the measured highest
359 intensities for oligomers and NCO (Fig. 3b). On the other hand, ionic dissociation and acid-base
360 reaction to yield carboxylates occur efficiently in the presence of NH₃/NaCl, as evident from the

361 increasing carboxylate intensity from NH_4HSO_4 to NaCl (Fig. 3b). The GF and SSA increase and
362 decrease, respectively, with NH_3 from 0 to 27 ppb for all three types of seed particles (Fig. 4a,b),
363 relevant to the reactions of NH_3 with dicarbonyls to form NCO and with organic acids to form
364 ammonium carboxylates. The analysis of the particle composition confirms that the intensities of
365 oligomers, NCO, and carboxylates increase with the NH_3 concentration (Fig. 4c-e).

366 We carried out additional experiments by varying RH from 10% to 70% in the presence of
367 19 ppb NH_3 (Fig. 5). The GF decreases with increasing RH from (3.10 ± 0.06) at RH = 10% to
368 (2.41 ± 0.03) at RH = 70% (Fig. 5a). The measured SSA at 120 min is close to unity at 10% and
369 30% RH and decreases with increasing RH (Fig. 5a), indicating negligible NCO formation at low
370 RH but significant NCO formation at RH above 50%. Carboxylates represent the dominant
371 constituent throughout the RH range (i.e., 85% at 10% RH to 47% at 70% RH), while the
372 contributions of oligomers and NCO are small at low RH (2-5% at RH < 40%) and become
373 increasingly significant at high RH (15-23% at RH > 50%) (Fig. 5b). The fraction for PAQ (8%)
374 is nearly invariant with RH. For $(\text{NH}_4)_2\text{SO}_4$ particles, the deliquescent and efflorescent points are
375 at 80% and 36% RH, respectively (Li et al., 2021b). At low RH (10% and 30%), aqueous reactions
376 to yield oligomers/NCO are minimal because of small water activity and significantly suppressed
377 protonation and oligomerization. On the other hand, carboxylic acids readily undergo ionic
378 dissociation or acid-base reaction, since organic acids efficiently retain water even at low RH (Xue,
379 2009). The equilibrium vapor pressures for PAQ are much lower than those for organic acids
380 (Table S6), facilitating more efficient condensation. Additionally, wall-loss of COOs is more
381 pronounced at high RH (Li et al., 2021b), explaining the decreased GF with increasing RH.
382 Measurement of gaseous concentrations for COOs shows that wall-loss is 1.5 to 4.6 times more

383 efficient at 70% RH than at 10% RH, while wall loss of *m*-xylene is negligible at both RH levels
384 (Table S7).

385 We determined the total density and the density of the SOA fraction for (NH₄)₂SO₄ seed
386 particles exposed to *m*-xylene oxidation (Fig. 5 c,d). The measured densities are distinct between
387 10 and 70% RH. At 10% RH, the total density decreases monotonically, while the SOA density
388 increases slightly with reaction time, i.e., from 1.27 to 1.39 g cm⁻³ (Fig. 5c), indicating minor
389 oligomers and dominant carboxylic acids at low RH. At 70% RH, the total particle density initially
390 decreases from 1.77 g cm⁻³ to 1.41 g cm⁻³ (at a GF of 1.24) and subsequently increases steadily to
391 1.56 g cm⁻³ at 120 min. The SOA density on (NH₄)₂SO₄ particles increases from 1.26 g cm⁻³ at 10
392 min to 1.55 g cm⁻³ at 120 min (Fig. 5d). The evolution in the densities reflects the variation in the
393 chemical composition. The initial particle growth is dominated by small oligomers, imidazoles
394 from methylglyoxal/methylbutenedial, and early-generation organic acids (e.g., toluic acid), with
395 the densities from 0.98 to 1.27 g cm⁻³ (Table S8). Subsequent particle growth from
396 methylglyoxal/methylbutenedial yields large oligomers, with the densities of 1.71-1.90 g cm⁻³
397 (Table S8).

398 To evaluate the NO_x effects, we performed experiments on (NH₄)₂SO₄ seed particles by
399 varying the initial NO_x concentration from 0 to 500 ppb (Fig. 6). Three major nitrophenols are
400 identified from NO₂ addition to the OH-*m*-xylene adduct, i.e., 4-methyl-2-nitrophenol (*m/z* = 154),
401 dimethyl nitrophenol (*m/z* = 168), and dimethyl-4-nitrocatechol (*m/z* = 184) (Fig. 6a). The
402 production of nitrophenols is much smaller than those for dicarbonyls and organic acids, consistent
403 with a smaller yield (less than 7%) for nitrophenols (Fan and Zhang, 2008). The GF on (NH₄)₂SO₄
404 seed particles with NH₃ decreases from 2.41 ± 0.03 to 2.18 ± 0.03 with 0 to 500 ppb NO_x (Fig. 6b).
405 The SSA decreases significantly from 0.911 ± 0.006 to 0.839 ± 0.003 with increasing NO_x (Fig.

406 6b), because of the formation of light-absorbing nitrophenols (Hems and Abbatt, 2018; Li et al.,
407 2021b). Analysis of particle composition reveals that the decreasing GF with increasing NO_x
408 correlates with simultaneous decreases of carboxylates, oligomers, and NCOs and with increasing
409 nitrophenols in the aerosol-phase (Fig. 6c). Overall, nitrophenols contribute 2% to 4% to the SOA
410 formation. Addition of NO_x not only produces nitrophenols and but also alters the cycling between
411 RO₂ and alkoxy (RO) radicals, leading to re-distributions of COOs. The presence of NO_x decreases
412 dicarbonyls and organic acids, as evident from decreased intensities for oligomers, NCO, and
413 carboxylates in the aerosol-phase with increasing NO_x (Fig. 2h).

414 **3.4 COO and SOA Yields**

415 From the measured GF and COO concentrations, we derived the γ , which is widely
416 employed to represent aerosol formation in atmospheric models (Zhang et al., 2015). The
417 measured γ varies with the functionality of COOs and RH on (NH₄)₂SO₄ seed particles (Fig. 7a).
418 Results of the γ values for COOs are also summarized in Table S9. The γ for dicarbonyls is the
419 highest (from 3×10^{-3} to 1.3×10^{-2}) at 70% RH but is significantly reduced (from 3.7×10^{-4} to 1.0
420 $\times 10^{-3}$) at 10% RH, while the γ 's for organic acids, PAQ, and nitrophenols are slightly higher at
421 70% RH. Among the organic acids, the γ correlates with the acid dissociation constant and
422 solubility, which are the highest for pyruvic acid and the lowest for toluic acid. The standard
423 variation in γ (within one σ) among each COO type is within 50%, 40%, 30%, and 15% for
424 dicarbonyls, organic acids, PAQ, and nitrophenols, respectively. From the measured
425 concentrations of COOs and *m*-xylene, we determined the COO yields (Y_{COO}) using eq. 17, by
426 considering four COO species, i.e., dicarbonyls for $i = 1$, carboxylic acids for $i = 2$, PAQ for $i = 3$,
427 and nitrophenols for $i = 4$ (Table S7). The Y_{COO} values for dicarbonyls, organic acids, PAQ, and

428 nitrophenols are 25%, 37%, 5%, and 3%, respectively (Fig. 7b), indicating significant production
429 of COOs from *m*-xylene oxidation.

430 Our work shows significant increase in particle size (Fig. 2a) and constantly varying
431 particle properties, i.e., SSA (Fig. 2b) and density (Fig. 5c,d), which correlate closely with the
432 gaseous COO production (Fig. 1). These results imply a highly nonequilibrium kinetic process
433 leading to SOA formation from *m*-xylene oxidation, which cannot be described by equilibrium
434 partitioning. The gas-to-particle conversion from *m*-xylene oxidation is dominated by several
435 volatile COO species (i.e., organic acids and dicarbonyls) and with minor contribution from
436 condensation of low-volatility COO (i.e., polyhydroxy aromatics/quinones, and nitrophenols) (Fig.
437 5b). In our study, the vapor pressures of the detected gaseous oxidation products for organic acids
438 and dicarbonyls are too large to explain the measured particle growth via equilibrium partitioning
439 (Table S6). Specifically, the saturation vapor pressures of organic acids detected in the particle-
440 phase range from 1.9×10^{-3} to 6.6×10^{-6} atm, while their gas-phase concentrations range from 0.5
441 to 2.5 ppb. Similarly, the saturation vapor pressures of dicarbonyls detected in the particle-phase
442 range from 1.6×10^{-1} to 3.9×10^{-4} atm, while their gas-phase concentrations range from 0.5 to 2.2
443 ppb. The high saturation vapor pressures and low gas-phase concentrations for dicarbonyls and
444 organic acids render equilibrium partitioning implausible for these COOs in our experiments. Also,
445 the detection of the particle-phase products confirms that organic acids and dicarbonyls directly
446 participate in heterogeneous reactions to yield low-volatile products. Moreover, volatility-based
447 equilibrium partitioning cannot not explain the variations in the particle size growth, SSA, and
448 chemical compositions on different seed particles (Fig. 3). Clearly, the gas-to-particle conversion
449 from *m*-xylene oxidation likely involves several distinct heterogeneous processes, including the

450 interfacial interaction, ionic dissociation/acid-base reaction, and nucleophilic oligomerization (Li
451 et al., 2021a,b).

452 We establish a functionality-based SOA formation to predict the aerosol mass
453 concentration (M_{SOA}) from the gaseous concentrations and γ 's for COOs using eq. 20. The M_{SOA}
454 is calculated from the measurements of the averaged γ (Fig. 7a and Table S10) and the time-
455 averaged concentrations for each COO type (Table S7). While the production of COOs is
456 dependent on the VOC types and photooxidation, the aerosol-phase reactivity of COOs, as
457 represented by γ , is dependent on the functionality for a given aerosol type. Thus, SOA formation
458 from various VOC types can be predicted from the production and reactivity for COOs based on
459 the experimental measurements, by categorizing COOs in accordance with their functional groups.

460 The overall gas uptake onto liquid or solid particles is commonly described by an empirical
461 framework on the basis of an effective γ , which accounts for the gas-phase diffusion, the interfacial
462 process (i.e., accommodation), and the aerosol-phase processes (Zhang et al., 1994; Ravishankara,
463 1997). The aerosol-phase processes include thermodynamically driven surface or volume-limited
464 partitioning (including condensation for low-volatile species or solubility for volatile species) and
465 chemically driven reactivity relevant to the functionality (Gomez et al., 2015). Ultimately, the
466 SOA growth is dependent on the volatility of the aerosol-phase products. Evidently, there are two
467 pathways to produce low-volatile compounds, i.e., gas-phase oxidation to yield low-volatile COOs
468 that subsequently under condensation and multiphase reactions of volatile COOs to form low-
469 volatile products in the aerosol-phase. In our study, the gas-to-particle conversion from *m*-xylene
470 oxidation occurs dominantly via the latter, i.e., reactive uptake for di-carbonyls and organic acids,
471 with a relatively small contribution from the former, i.e., condensation by low-volatile PAQs and
472 nitrophenols. Also, our measured uptake coefficient for each COO type (γ) is derived from the

473 measured particle growth (eqs. 9 and 10), which implicitly accounts for non-continuum diffusion,
474 imperfect accommodation, and evaporation.

475 To assess the consistency of our approach, we compared the predicted and measured SOA
476 mass yields (Y_{SOA}), according to eq. 21. The predicted Y_{SOA} is $(11 \pm 3)\%$ and $(6 \pm 1)\%$,
477 respectively, at 10% and 70 % RH, comparable to the measured values of $(14 \pm 0.4)\%$ and $(6.3 \pm$
478 $0.2)\%$ on $(\text{NH}_4)_2\text{SO}_4$ seed particles (Fig. 7c). The slight differences between the predicted and
479 measured Y_{SOA} 's are explainable by the uncertainties related to lumping each COO type (i.e.,
480 averaging the γ values) as well as unaccounted low-abundant COO species. To account for the
481 RH-dependent wall loss for COOs, we corrected Y_{SOA} from the measured gaseous concentrations
482 at varying RH (see Methods and Table S7). Such correction results in an increase of about one to
483 four times for Y_{SOA} at 10% and 70% RH, yielding the values of $(20 \pm 4)\%$ and $(32 \pm 7)\%$,
484 respectively (Fig. 7c). Under atmospheric conditions, SOA/BrC formation from *m*-xylene is
485 expected to be enhanced at high RH, because of significantly enhanced aqueous reaction, larger
486 aerosol surface-area, and higher water activity. Field measurements showed significantly enhanced
487 SOA formation with increasing RH during severe haze evolution (Guo et al., 2014; Peng et al.,
488 2021).

489 **4. Conclusions**

490 In this work we have elucidated the production of COOs and their roles in SOA formation
491 from *m*-xylene oxidation. The major advantages of our analytical methodologies lie in
492 simultaneous monitoring of the evolutions of gaseous oxidation products and aerosol properties
493 when seed particles are exposed to the oxidation products, which allows for assessment of multi-
494 generation production of SOA and BrC. Also, our experimental approach remedies the deficiency
495 of wall loss of reactive gaseous and particles, which represents a main obstacle in experimental

496 investigation using environmental chambers (Zhang et al., 2014; Huang et al., 2018). OH-initiated
497 oxidation of *m*-xylene produces four distinct COO types consisting of dicarbonyls, carboxylic
498 acids, polyhydroxy aromatics/quinones, and nitrophenols from early-generation (P₂ and P₃), with
499 the yields of 25%, 37%, 5%, and 3%, respectively. SOA formation occurs via several
500 heterogeneous processes, including interfacial interaction, ionic dissociation/acid-base reaction,
501 and oligomerization, with the yields of (20 ± 4)% at 10% RH and (32 ± 7)% at 70% RH. The
502 identified aerosol-phase products include dominant oligomers, N-heterocycles/N-heterochains,
503 and carboxylates at high RH and dominant carboxylates at low RH, along with small amounts of
504 PAQ and nitrophenols (in the presence of NO_x). The nitrogen-containing organics consisting of
505 N-heterocycles, N-heterochains, and nitrophenols are light-absorbing, characterized by low SSA.
506 The measured γ for COOs is dependent on the functionality, ranging from 3.7×10^{-4} to 1.3×10^{-2} .
507 A kinetic framework is developed to predict SOA formation from the concentration and γ for
508 COOs. Our results illustrate that SOA formation from *m*-xylene oxidation is accurately quantified
509 from the abundance and reactivity for COOs (i.e., eq. 20). Notably, this kinetic framework
510 accounts for the interfacial process (i.e., mass accommodation) as well as aqueous reactions (i.e.,
511 oligomerization and acid-base reactions) without the assumption of gas-particle equilibrium. We
512 envisage that this functionality-based approach is applicable to predict SOA formation from VOC
513 photooxidation using experiment/field measured or model-simulated COOs, aerosol surface area,
514 and reactivity. We conclude that photochemical oxidation of *m*-xylene represents a major source
515 for SOA and BrC formation under urban conditions, because of its large abundance, high reactivity
516 with OH, and high yields for COOs.

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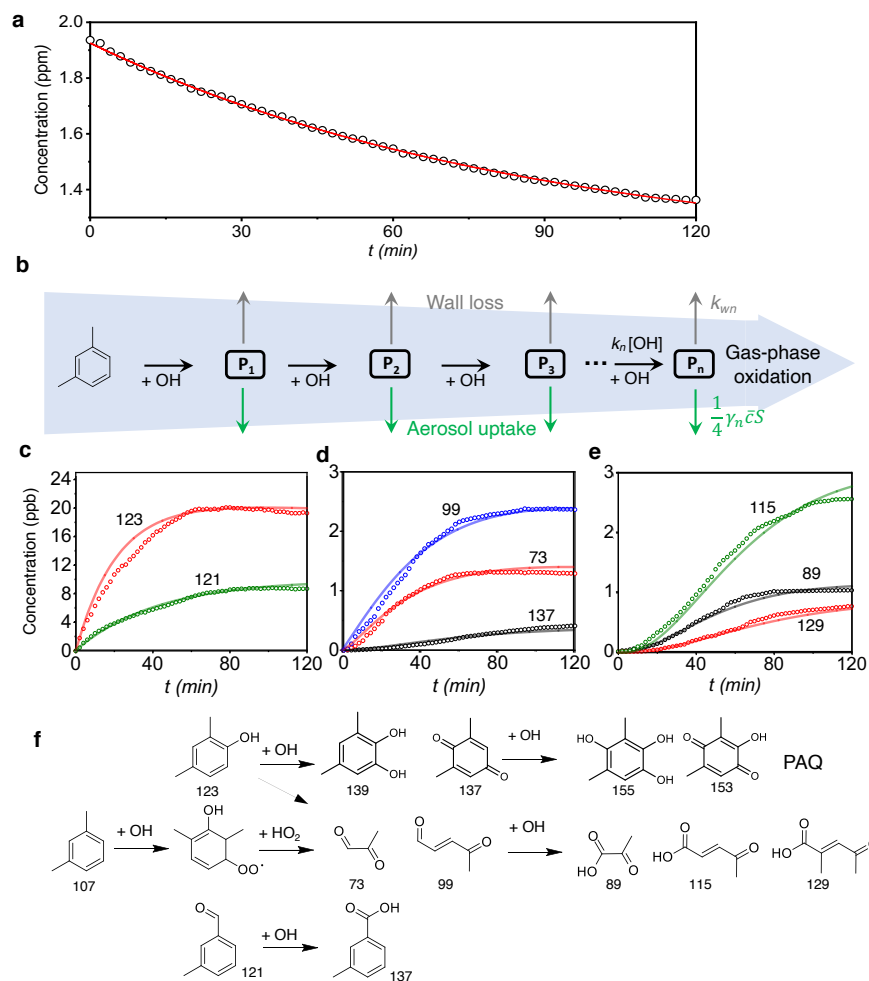
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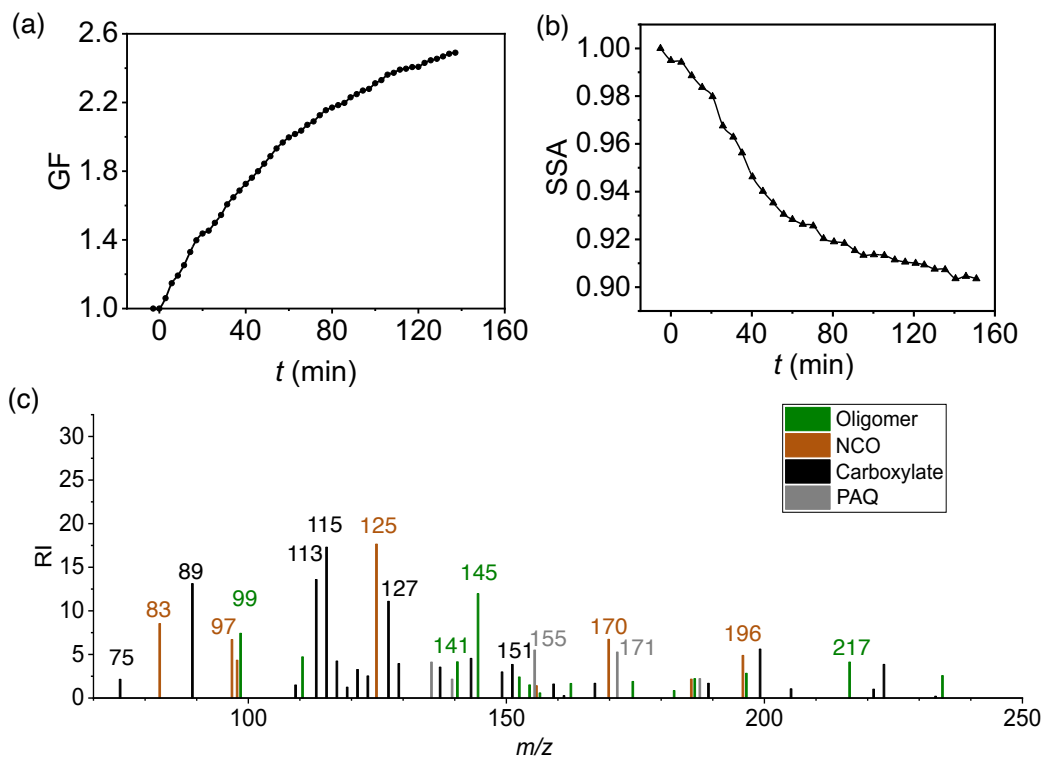
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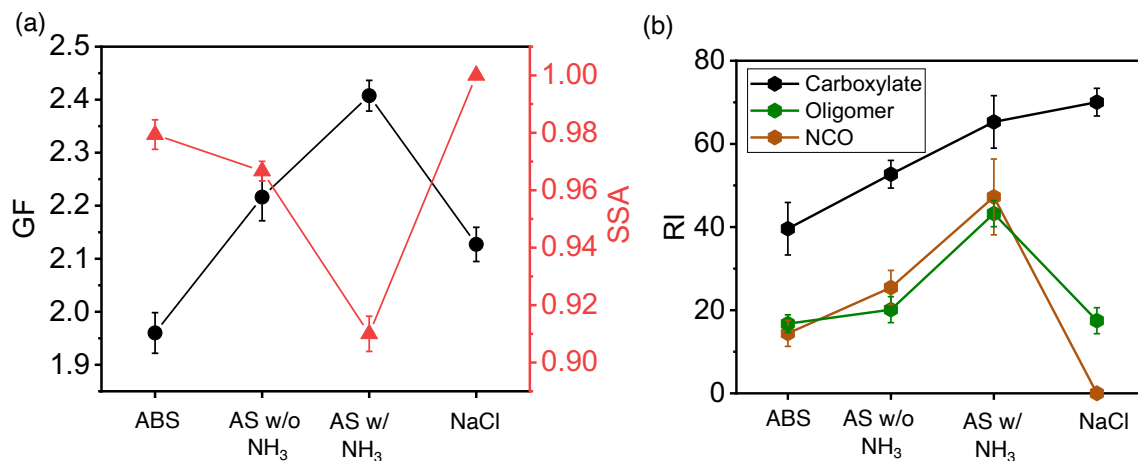
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 748 **Figure 1.** OH-initiated photooxidation of *m*-xylene. (a) Time-dependent concentration of *m*-
 749 xylene. The symbols are from the measurements, and the red solid line corresponds to exponential
 750 fitting of the *m*-xylene concentration. The initial concentration for *m*-xylene and the estimated
 751 steady-state concentration for OH are 1940 ppb and 2.1×10^6 molecules cm^{-3} , respectively, with a
 752 bimolecular rate constant of 2.4×10^{-11} cm^3 molecule $^{-1}$ s $^{-1}$. (b) Schematic representation leading to
 753 the multi-generation product, P_{*n*}, where *n* denotes the sequence of OH-initiated oxidation. The
 754 black, gray, and green arrows denote the gas-phase oxidation, chamber wall loss, and aerosol
 755 uptake, respectively. k_n and k_{wn} are the rate coefficients for the gas-phase oxidation and wall loss,
 756 respectively, γ_n represents the aerosol uptake coefficient, \bar{c} is the thermal velocity, and S is the
 757 aerosol surface area. (c-e) Time-dependent gas-phase concentrations of P₁ (c), P₂ (d), and P₃ (e)
 758 products. The symbols are from measurements, and the solid curves are simulated according to the
 759 tri-exponential kinetics. (f) The gaseous oxidation pathways leading to the detected products. The
 760 top row corresponds to the pathways leading to polyhydroxy aromatics/quinones (PAQ). The
 761 numbers in (c) to (f) represents the mass to charge ratio (*m/z*). All experiments are carried out for
 762 (NH₄)₂SO₄ seed particles with 19 ppb NH₃ at 298 K and RH = 70%. Initiation of photooxidation
 763 by ultraviolet light occurs at $t = 0$.



764

765 **Figure 2.** SOA and BrC formation from *m*-xylene photooxidation. Time-dependent growth factor
 766 ($GF = D_p/D_0$, a) and single scattering albedo (SSA, b) of seed particles exposed to *m*-xylene
 767 oxidation products. (c) Chemical speciation of aerosol-phase products by TD-ID-CIMS. Initiation
 768 of photooxidation by ultraviolet light occurred at $t = 0$. RI denotes the relative intensity for the
 769 integrated peak area during thermal desorption. All experiments are carried out for $(NH_4)_2SO_4$ seed
 770 particles with 19 ppb NH_3 at 298 K and $RH = 70\%$.

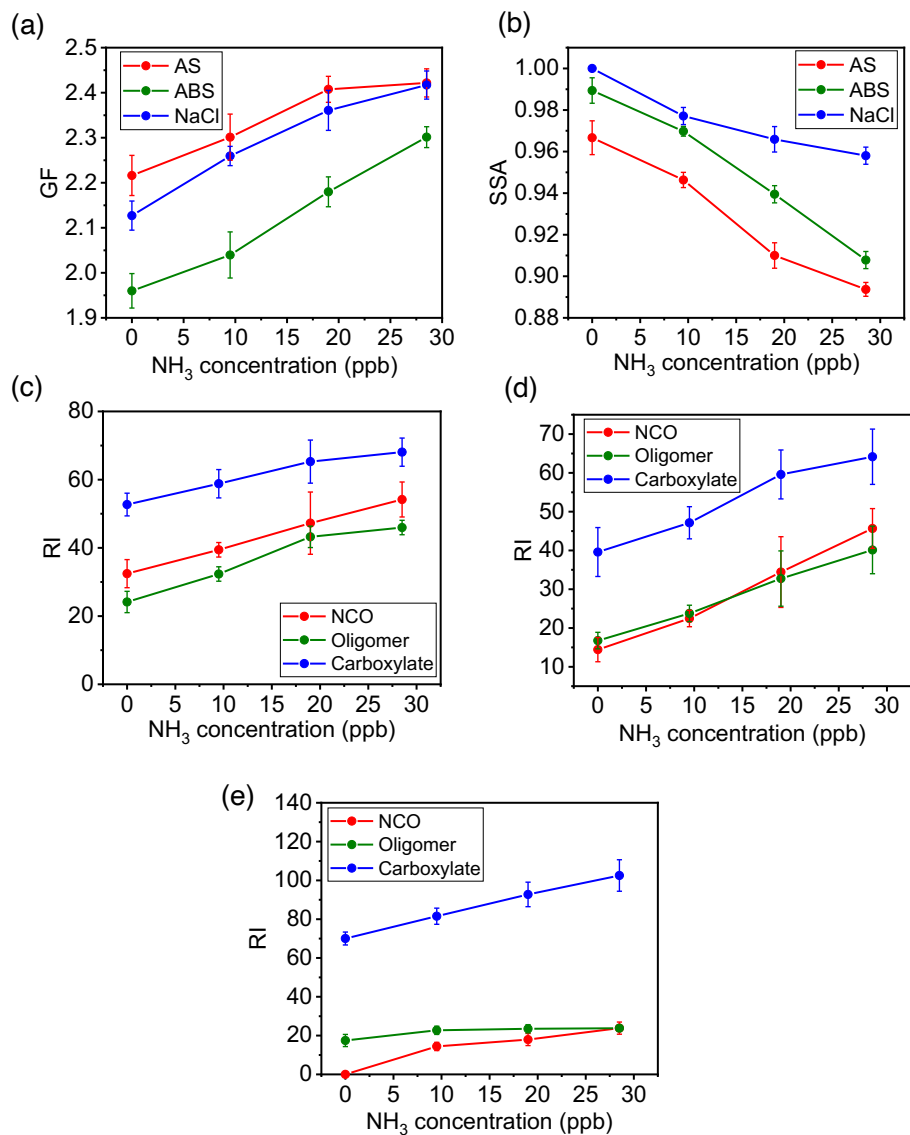
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773 **Figure 3.** Dependence of SOA/BrC formation on seed particles. (a) GF (black) and SSA (red) at
 774 120 mins of exposure for (NH₄)₂SO₄ particles in the presence (AS w/ NH₃) and absence (AS w/o
 775 NH₃) of 19 ppb NH₃ and for NH₄HSO₄ (ABS) and NaCl particles in the absence of NH₃. (b)
 776 Aerosol-phase relative intensity (RI) for carboxylates (black), oligomers (green), and NCO (brown)
 777 on different seed particles. The error bar denotes 1σ of 3 replicated measurements. All particle
 778 properties were measured relevant to dry conditions (less than 5% RH) and were dominantly
 779 contributed by non-volatile aerosol-phase products.

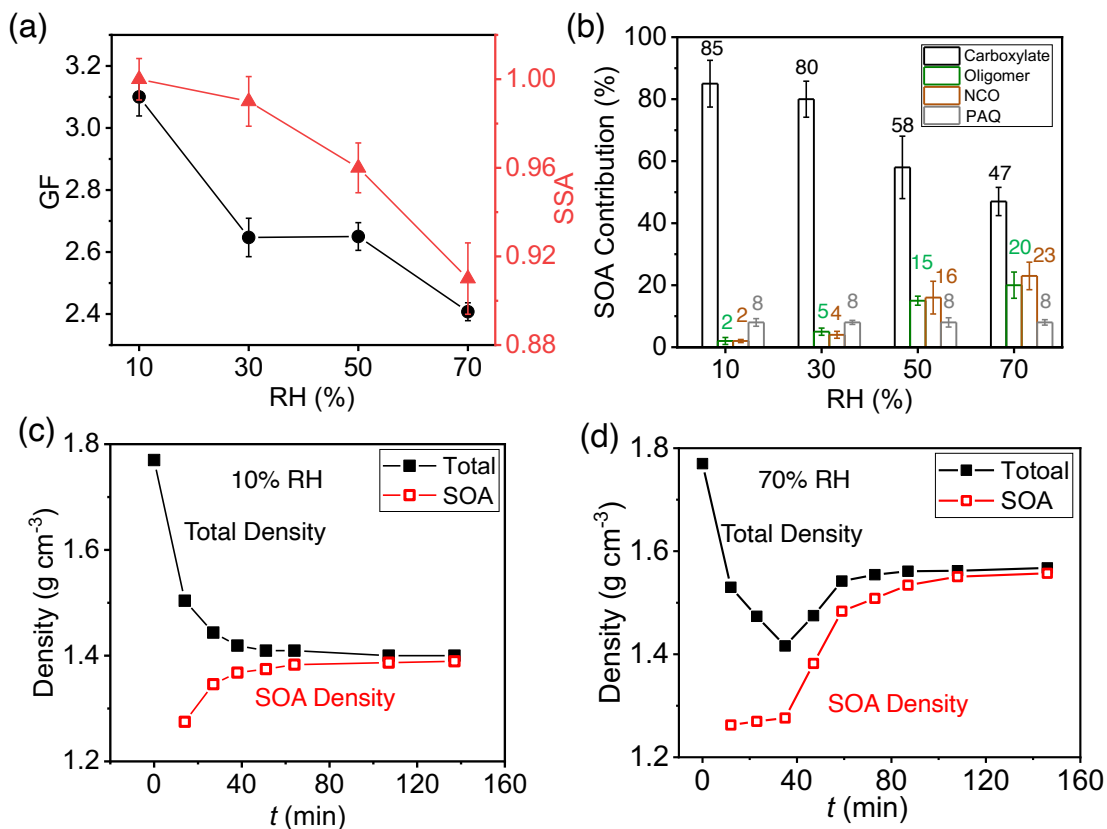
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782 **Figure 4.** Variation of SOA formation with seed particles and NH_3 concentration. (a,b) growth
 783 factor (a) and SSA (b) at $t = 120$ min with varying NH_3 concentration between 0 and 27 ppb for
 784 ammonium sulfate (AS, red), ammonium bisulfate (ABS, green), and sodium chloride (NaCl,
 785 blue) seed particles. (c-e) Aerosol-phase mass relative intensity (RI) for NCO (red), oligomers
 786 (green), and carboxylates (blue) on AS (c), ABS (d), and NaCl (e) seed particles. The error bar
 787 denotes 1σ of 3 replicated measurements.

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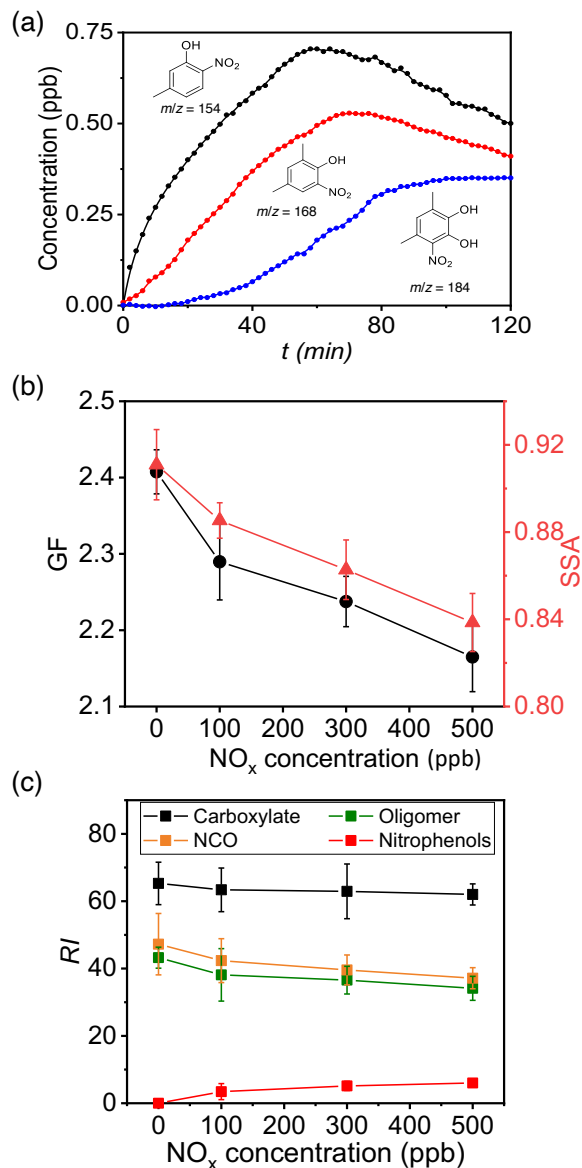


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790 **Figure 5.** Dependence of SOA/BrC formation on RH. (a) GF (black) and SSA (red) at 120 min
 791 with varying RH from 10% to 70%. (b) Contribution of COOs to SOA formation with varying RH
 792 from 10% to 70%: carboxylate (black), oligomers (green), NCO (brown), and PAQ (gray) to SOA
 793 formation. The number denotes the percentage contribution (%) of each type of aerosol-phase
 794 products. (c,d) Time-dependent particle (black) and SOA (red) densities of seed particles exposed
 795 to *m*-xylene oxidation products at RH = 10% (c) and 70% (d), respectively. The error bar denotes
 796 1σ of 3 replicated measurements.

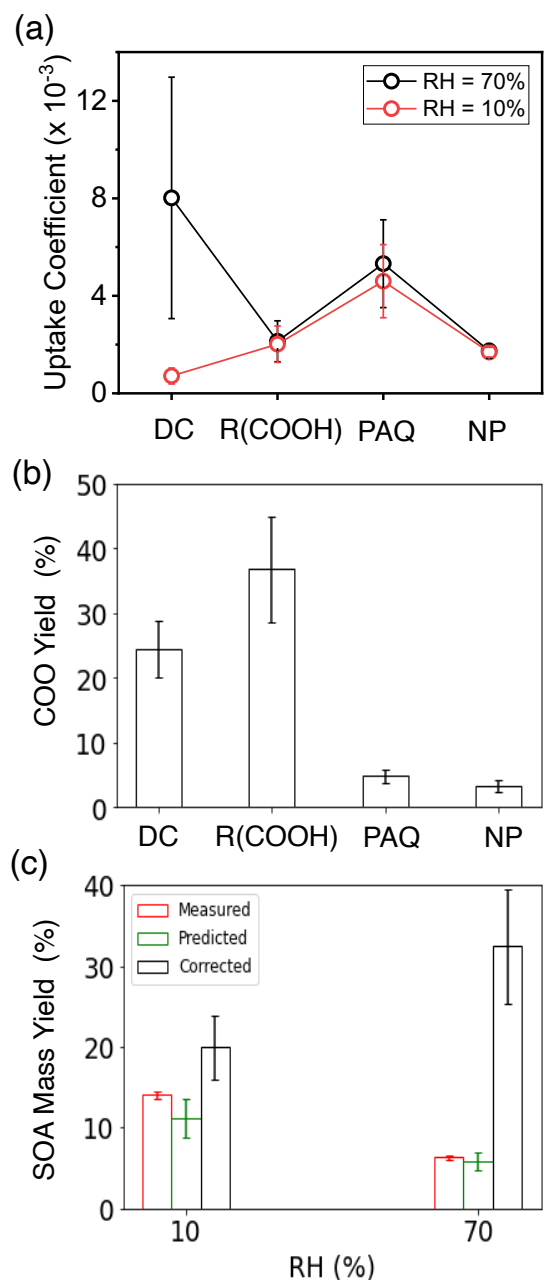
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800 **Figure 6.** The effects of NO_x. (a) Time-dependent gas-phase concentrations of 5-methyl-2-
 801 nitrophenol (black, *m/z* = 154), dimethyl nitrophenol (red, *m/z* = 168), and dihydroxy nitrotoluene
 802 (blue, *m/z* = 184). The numbers denote the mass to charge ratio (*m/z*). Initiation of photooxidation
 803 by ultraviolet light occurred at *t* = 0. (b) GF (black) and SSA (red) at 120 min with varying NO_x
 804 concentration from 0 to 500 ppb. (h) Dependence of aerosol-phase relative mass intensities (RI)
 805 for carboxylates (black), oligomers (green), NCO (brown), and nitrophenols (red) on NO_x
 806 concentration for (NH₄)₂SO₄ seed particles with 19 ppb NH₃ at RH = 70%. The error bar denotes
 807 1σ of 3 replicated measurements.



808

809 **Figure 7.** Uptake coefficient, COO yield, and SOA mass yield. (a) Average uptake coefficients (γ)
 810 for different types of COOs at 10% (red) and 70% (black) RH for $(\text{NH}_4)_2\text{SO}_4$ seed particles with
 811 19 ppb NH_3 at 298 K. (b) COO yields: dicarbonyls, organic acids, polyhydroxy aromatics/quinones,
 812 and nitrophenols are represented by DC, $\text{R}(\text{COOH})_n$, PAQ, and NP, respectively. (c) SOA mass
 813 yields at 10% (left columns) and 70% RH (right). The red, green, and black columns represent the
 814 measured, predicted, and corrected (for wall loss) SOA mass yields according to eq. 19-22. All
 815 experiments are carried out for $14 \mu\text{g m}^{-3}$ $(\text{NH}_4)_2\text{SO}_4$ seed particles with 19 ppb NH_3 at 298 K. The
 816 error bars denote the 1σ of 3 replicated measurements or by accounting for error propagation of
 817 the measured parameters.

818

819 **Table 1. Summary of experimental conditions.**

Experiment #	Seed particle	NH ₃ concentration (ppb)	NO _x concentration (ppb)	RH
Exp. 1	AS	19	0	70%
Exp. 2	AS	0	0	70%
Exp. 3	ABS	0	0	70%
Exp. 4	NaCl	0	0	70%
Exp. 5	AS	9.5	0	70%
Exp. 6	AS	28.5	0	70%
Exp. 7	ABS	9.5	0	70%
Exp. 8	ABS	19	0	70%
Exp. 9	ABS	28.5	0	70%
Exp. 10	NaCl	9.5	0	70%
Exp. 11	NaCl	19	0	70%
Exp. 12	NaCl	28.5	0	70%
Exp. 13	AS	19	0	10%
Exp. 14	AS	19	0	30%
Exp. 15	AS	19	0	50%
Exp. 16	AS	19	100	70%
Exp. 17	AS	19	300	70%
Exp. 18	AS	19	500	70%

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