

Aqueous chemical bleaching of 4-nitrophenol brown carbon by hydroxyl radicals; products, mechanism and light absorption

Bartłomiej Witkowski^{1*}, Priyanka Jain¹, Tomasz Gierczak¹

¹University of Warsaw, Faculty of Chemistry, al. Żwirki i Wigury 101, 02-089 Warsaw, Poland

5 *Correspondence to:* Bartłomiej Witkowski (bwitk@chem.uw.edu.pl)

Abstract. The reaction of hydroxyl radicals (OH) with 4-nitrophenol (4NP) in an aqueous solution was investigated at pH=2 and 9. The molar yield of the phenolic products quantified was 0.21 ± 0.01 at pH=2 and 0.31 ± 0.02 at pH=9. The yield of 4-nitrocatechol (4NC) was higher at pH=9. At the same time, a lower number of phenolic products was observed at pH=9 due to irreversible reactions of some phenols formed when the pH>7. Mineralization investigated with a total organic carbon (TOC) analyzer showed that after 4NP was completely consumed, approximately 85% of the organic carbon remained in the aqueous solution. Moreover, as inferred from the TOC measurements and the molar yields of the phenols formed, 64% of the organic carbon that remained in the aqueous solution was attributed to the non-aromatic products. The light absorption of the reaction solution between 250 and 600 nm decreased as a result of the OH reaction with 4NP. However, the 4NP solution showed a noticeable resistance to the chemical bleaching reaction investigated due to the formation of light-absorbing by-products. This phenomenon effectively prolongs the time scales of the chemical bleaching of 4NP by OH by a factor of 3–1.5 at pH 2 and 9, respectively. The experimental data acquired indicated that both photolysis and the reaction with OH can be important processes for the removal of light-absorbing organic compounds from cloud water particles containing 4NP.

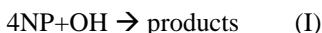
1 Introduction

Atmospheric brown carbon (BrC) is a subfraction of organic aerosols (OA) characterized by strong, wavelength-dependent absorption of electromagnetic irradiation in the near ultraviolet (UV) and visible (Vis) regions (Laskin et al., 2015; Yan et al., 2018). BrC is primarily produced by biomass burning (BB) and has a negative impact on local air quality and human health (Laskin et al., 2015; Hems et al., 2021). Due to the high UV-Vis absorption, BrC greatly contributes (up to 50%) to the radiative forcing of OA (Feng et al., 2013; Wang et al., 2014; Lu et al., 2015; Cordell et al., 2016; Zhang et al., 2017; Yan et al., 2018). Numerous organic compounds contribute to atmospheric BrC (Laskin et al., 2015; Li et al., 2020; Fleming et al., 2020; Vidović et al., 2020; Hettiyadura et al., 2021). At the same time, a significant fraction of BrC chromophores remain poorly characterized (Laskin et al., 2015; Bluvshtein et al., 2017).

Nitrophenols are widespread nitroaromatic compounds that have been identified among the major chromophores of atmospheric BrC (Harrison et al., 2005b; Kitanovski et al., 2012; Claeys et al., 2012; Laskin et al., 2015; Frka et al., 2016; Bluvshtein et al., 2017). 4-Nitrophenol (4NP) is one of the most atmospherically abundant and environmentally widespread nitrophenols (Harrison et al., 2005b; Laskin et al., 2015). 4NP also exhibits a strong absorption of electromagnetic radiation in the UV-Vis region (Jacobson, 1999). For these reasons, 4NP was found to contribute significantly to the light absorption of ambient BrC aerosols (Mohr et al., 2013; Bluvshtein et al., 2017; Kitanovski et al., 2020). 4NP was detected in the air, rain, surface waters, and snow as well as in atmospheric particulate matter (PM) (Jaber et al., 2007; Kitanovski et al., 2012; Claeys et al., 2012; Kahnt et al., 2013; Mohr et al., 2013; Balasubramanian et al., 2019; Liang et al., 2020; Kitanovski et al., 2020). Large quantities of 4NP are produced by the combustion of fossil fuels and biomass (Desyaterik et al., 2013; Mohr et al., 2013; Inomata et al., 2015; Xie et al., 2019). Moreover, 4NP is introduced into the environment by industrial processes (Majewska et al., 2021). Several studies have confirmed that 4NP has an adverse impact on human health (Majewska et al., 2021), is a threat to aquatic organisms (Tenbrook et al., 2003), and contributes to the decline of forests (Natangelo et al., 1999).

The formation, chemical processing and removal (bleaching) of BrC can occur in air as well as in atmospheric aqueous particles, and can involve direct photolysis and reactions with hydroxyl radicals (OH) (Laskin et al., 2015; Forrister et al., 2015; Moise et al., 2015; Hems and Abbatt, 2018; Hems et al., 2020; Li et al., 2020; Jiang et al., 2021). Due to the high value of Henry's law constant (Sander, 2015), and its high solubility in water, 4NP can readily partition into atmospheric aqueous particles (Harrison et al., 2005b; Vione et al., 2009). The chemical and photochemical reactions in the atmospheric aqueous phase contribute to the formation (Vione et al., 2003; Harrison et al., 2005a; Heal et al., 2007), transformation (Vione et al., 2005; Vione et al., 2009) and removal (Harrison et al., 2005a; Braman et al., 2020) of 4NP. The chemical and photochemical processing (aging) in the aqueous phase result in a change in the light absorption of aqueous particles containing 4NP (Zhang et al., 2003; Zhao et al., 2015; Braman et al., 2020). However, the connection between the light absorbance and chemical composition of the aqueous 4NP solution that has been subjected to photolysis and oxidation by OH is poorly characterized (Zhang et al., 2003; Zhao et al., 2015).

50 The aqueous reaction of 4NP with OH (reaction I) is known to produce aromatic products, including hydroquinone (HH), 1,2,4-trihydroxylbenzene (1,2,4-THB), 4-nitrocatechol (4NC) and 4-nitropyrogallol (4NPG) (Tauber et al., 2000; Oturan et al., 2000; Zhang et al., 2003; Daneshvar et al., 2007; Biswal et al., 2013).



55 However, the yields of the substituted phenols from the reaction (I) remain ambiguous (Tauber et al., 2000; Oturan et al., 2000; Zhang et al., 2003; Daneshvar et al., 2007; Ding et al., 2016). Reaction (I) was previously investigated at a molecular level but almost exclusively in the context of wastewater treatment via advanced oxidation processes (AOP) (Tauber et al., 2000; Oturan et al., 2000; Zhang et al., 2003; Ding et al., 2016) where the reaction conditions cannot be considered as atmospherically-relevant (Tauber et al., 2000; Zhang et al., 2003; Daneshvar et al., 2007; Ding et al., 2016). Consequently, it is currently difficult to evaluate if reaction (I) is a relevant source of atmospheric BrC (Oturan et al., 2000; Tauber et al., 2000; Kavitha 60 and Palanivelu, 2005; Biswal et al., 2013; Xiong et al., 2015).

In some clouds and fogs (Herrmann et al., 2015), 4NP ($\text{pK}_a \approx 7.2$) (Rived et al., 1998) can exist in both protonated and deprotonated forms (Fig. S1). At the same time, little information is available about the impact of pH on the distribution of the products of the reaction (I) (Tauber et al., 2000; Oturan et al., 2000). Furthermore, little information exists regarding the pH dependence of the light absorbance of a 4NP solution that has been subjected to OH oxidation (Biswal et al., 2013; Zhao et 65 al., 2015). It should also be noted that the UV-Vis absorption of 4NP and its aromatic oxidation products are strongly pH-dependent (Biswal et al., 2013; Braman et al., 2020).

The objective of this study was to investigate the mechanism of OH reactions with 4NP in the aqueous phase in the context of atmospheric BrC formation and processing. Therefore, reaction (I) was investigated at 298 K in an aqueous solution under acidic ($\text{pH}=2$) and basic ($\text{pH}=9$) conditions using the photoreactor developed in the host laboratory (Witkowski et al., 2019).

70 Additionally, the phenolic products of the reaction (I) were analyzed together with the changes in the UV-Vis absorption of the reaction solution. The phenols under investigation were quantified using gas chromatography coupled to mass spectrometry (GC/MS). The possible mineralization of 4NP and the formation of volatile products were monitored with a total organic carbon (TOC) analyzer. The UV-Vis absorption of the reaction solution, as well as the molar absorption (ϵ , $\text{mol}^{-1} \times \text{L} \times \text{cm}^{-3}$) of the phenols under investigation, were measured between pH 2 and 9.

75 2. Experimental section

The materials and reagents used are listed in section S1 of the electronic supplementary information (SI). All solutions were prepared using deionized (DI) water ($18 \text{ M}\Omega \times \text{cm}^{-1}$).

2.1 Aqueous phase photoreactor

The aqueous photoreactor was described previously (Witkowski et al., 2019); more details are provided in section S4.1. The 80 reaction vessel was a quartz jacketed flask with an internal volume of 100 ml, surrounded by eight 4W lamps. All experiments

were carried out at 298 K; the temperature of the reaction solution was maintained by a circulating water bath (SC100-A10, Thermo Fisher Scientific). Two 4W lamps (TUV, peak emission 254 nm) and six 4W lamps (33-640, emission >400 nm, Philips) were used to irradiate the reaction solution.

2.2 Experimental procedure

85 The reaction mixture was a 100–250 μ M aqueous solution of 4NP; total volume 100 ml. The pH of this solution was not adjusted (unbuffered, no acids or buffers added) or it was adjusted to pH 2 or 9 with HCl, HClO₄ or Na₂HPO₄ (50 mM) to investigate the reaction of fully protonated and deprotonated forms of 4NP (Fig. S1). Hydrogen peroxide (H₂O₂, concentration 5 mM) was photolyzed with UV irradiation (254 nm) to generate OH with an estimated steady-state concentration = 1.4×10^{-9} M (section S3) (Tan et al., 2009). Under these conditions, 4NP was almost completely consumed within 1h. Aliquots of the
90 reaction mixture were sampled every 5 min and analyzed with GC/MS, a UV-Vis spectrophotometer and, a TOC analyzer. The experimental procedure is described in detail in section S4.1.

2.3 Quantification of the phenolic products with gas chromatography coupled to mass spectrometry

Analyses were carried out using a GC-MS-QP2010Ultra gas chromatograph coupled with a single quadrupole mass spectrometer and equipped with an AOC-5000 autosampler (Shimadzu). Analytes were separated using a capillary column
95 ZB-5MSPlus (Phenomenex). The mass spectrometer was equipped with an electron ionization source (EI, 70 eV) and was operating in the selected ion monitoring (SIM) mode. The instrument was calibrated with the standard solutions of 4NP, HH, 1,2,4-THB, and 4NC that were identified as products of the reaction (I). 2-Nitrophloroglucinol was used as a surrogate standard for the quantification of 4-nitropyrogallol (4NPG) and 5-nitropyrogallol (5NPG), both identified among the products of the reaction (I). Because phloroglucinol was not identified as the product of reaction (I) (Zhao et al., 2013; Xiong et al., 2015), it
100 was used as an internal standard (IS). The phenols were derivatized with acetic anhydride (AA) and analyzed with GC/MS (Regueiro et al., 2009). A detailed description of the analytical procedure is provided in section S4.2.

The yields of the phenols formed from reaction (I) were derived using eq. (I).

$$[Product]_t = Yield \times \Delta[4NP]_t \quad (I)$$

In eq. (I), [Product]_t is the concentration (or the sum of the concentrations) of the products at a given time (t) from the onset
105 of the reaction (mM). [4NP] is the amount of the precursor that reacted at a given time (t) from the onset of the reaction (mM). The yield (molar) is derived as the slope of the linear portion of the plot obtained using eq. (I) (Gierczak et al., 2021).

2.4 UV-Vis spectrophotometry

UV-Vis measurements were performed with an i8 dual-beam spectrophotometer (Envisense) in 4 ml cuvettes with a 1 cm absorption pathlength; the spectral range was 230 to 600 nm. The absorbance of the aliquots of the reaction solution was
110 initially measured at pH=2 or 9, which was the pH of the reaction mixture (section 2.2). Subsequently, a small amount of

NaOH or H₃PO₃ was added to each sample taken from the reactor, the pH was adjusted by 1 unit and the absorbance was measured again. This procedure was repeated until the absorbance of each aliquot of the reaction solution was recorded between pH 2 and 9 (section S4.3). Separately, the wavelength-dependent absorption cross-sections, ϵ , (mol⁻¹ × L × cm⁻¹) were measured (Fig. S5) for 4NP, HH, 1,2,4-BT, 4NC, and 2-NPG using the commercially available standards and are listed in
115 Appendix 1.

2.5 Total organic carbon analyses

Non-purgeable organic carbon (NPOC) was quantified with a TOC-5050A analyzer (Shimadzu) equipped with an ASI-5000A autosampler (Shimadzu). The 1.5 ml of the reaction solution was diluted with the same volume of water, filtered through a 0.22 μ m PTFE membrane, and placed in the TOC autosampler vial. Then, 50 μ l of 2M HCl was added by the autosampler and
120 each sample was purged with O₂ for 2 min before being injected to remove the CO₂ and the sparingly soluble, volatile organic compounds. The injection volume was ca. 20 μ l and each sample was injected into the instrument three times. The TOC analyzer was calibrated with the standard solutions of 4NP (section S4.4).

2.6 Light absorption and atmospheric lifetimes

The production of light-absorbing compounds following reaction (I) was evaluated using eq. (II).

$$\frac{\left(\int_{250nm}^{600nm} A_{10}^{R,mix}(pH) \right)_t d\lambda}{\left(\int_{250nm}^{600nm} A_{10}^{4NP}(pH) \right)_t d\lambda} = \left(\frac{[4-NP]_0}{[4-NP]_t} \right)^{K_{abs}} \quad (II)$$

In eq. (II), $A_{10}^{R,mix}$ and A_{10}^{4NP} are integrated absorbance peak areas between 250 and 600 nm (d λ) for the reaction mixture measured between pH 2 and 9 at different time intervals (t), [4NP]₀ and [4NP]_t are the initial (0) and intermediate (t) concentrations of 4NP measured with GC/MS, and the absorbance (A_{4NP}) of 4NP was calculated by the Beer-Lambert law using the ϵ measured in this study between pH 2 and 9. The expression described by eq. (II) followed function $y = A \times x^K$
130 (section S8 and Fig. S10).

The atmospheric lifetime of BrC was evaluated by calculating the empirical $k_{bleaching}$ rate constant (M⁻¹s⁻¹) – eq. (III).

$$k_{bleaching} = k_{OH}(4 - NP) \times \frac{k_{A,rmix}}{k_{A,4-NP}} \quad (III)$$

In eq. (III), k_{OH} is the second-order rate constant (M⁻¹s⁻¹) for the reaction of 4NP or 4-nitrophenolate (4NPT) with OH (Garcia Einschlag et al., 2003; Biswal et al., 2013), k_A and $k_{A,rmix}$ are the first-order disappearance rate constants (min⁻¹) of the
135 integrated absorbance peak for the 4NP and the reaction mixture, respectively (Fig. S9). The $k_{A,rmix}$ values derived showed little dependence on the pH at which the absorbance was measured, thus average values were used.

The TOC-normalized mass absorption coefficients (MAC_{TOC}) of the reaction mixture were calculated using eq. (IV) (Laskin et al., 2015; Bluvshtein et al., 2017; Jiang et al., 2021).

$$MAC_{TOC}(\text{cm}^2 \times g_{TOC}^{-1}) = \frac{\ln(10) \left(\int_{250nm}^{600nm} A_{10}^{R,mix} \times l_{solution}^{-1} \right) d\lambda}{NPOC} \times 10^{-6} \text{ (IV)}$$

140 In eq. (IV), l is the optical path length (cm^{-1}), NPOC is the concentration of non-purgeable organic carbon ($\text{mg} \times \text{L}^{-1}$). The TOC-normalized rate of sunlight absorption (R_{abs}) by the reaction solution was estimated using eq. (V) (Jiang et al., 2021).

$$R_{abs}(\text{photons} \times s^{-1} \times mg_{TOC}^{-1}) = \left(\frac{\ln(10) \times \left(\int_{250nm}^{600nm} A_{10}^{R,mix} \times l_{solution}^{-1} \times I_\lambda \right) d\lambda}{NPOC} \right) \times 10^{-3} \text{ (V)}$$

In eq. (V), I_λ is the actinic flux ($\text{photons} \times s^{-1} \times \text{cm}^{-2} \times \text{nm}^{-1}$) estimated by the TUV calculator for zenith angles 0-50° (Ncar, 2016).

145 **2.7 Control experiments and uncertainty**

The stability of the phenols under investigation in the presence of H_2O_2 or UV-Vis irradiation alone was studied in control experiments (section S6). Also, for the experiments at $\text{pH}=2$, HCl or HClO_4 was used to confirm that the buffering agent used did not affect the distribution of the detected products. The control experiments revealed that all the phenols under investigation were stable at $\text{pH} \leq 7$, within the time-scale of the experiments, but 1,2,4-THB, 4NPG, 5NPG, and HH underwent irreversible

150 dark reactions at $\text{pH} > 7$.

Experimental uncertainties are reported as 2σ from triplicate measurements, other uncertainties were calculated with the exact differential method unless otherwise noted.

3. Results and discussion

3.1 Products and reaction mechanism

155 The products of the reaction (I) formed under acidic pH conditions are shown in Fig. 1.

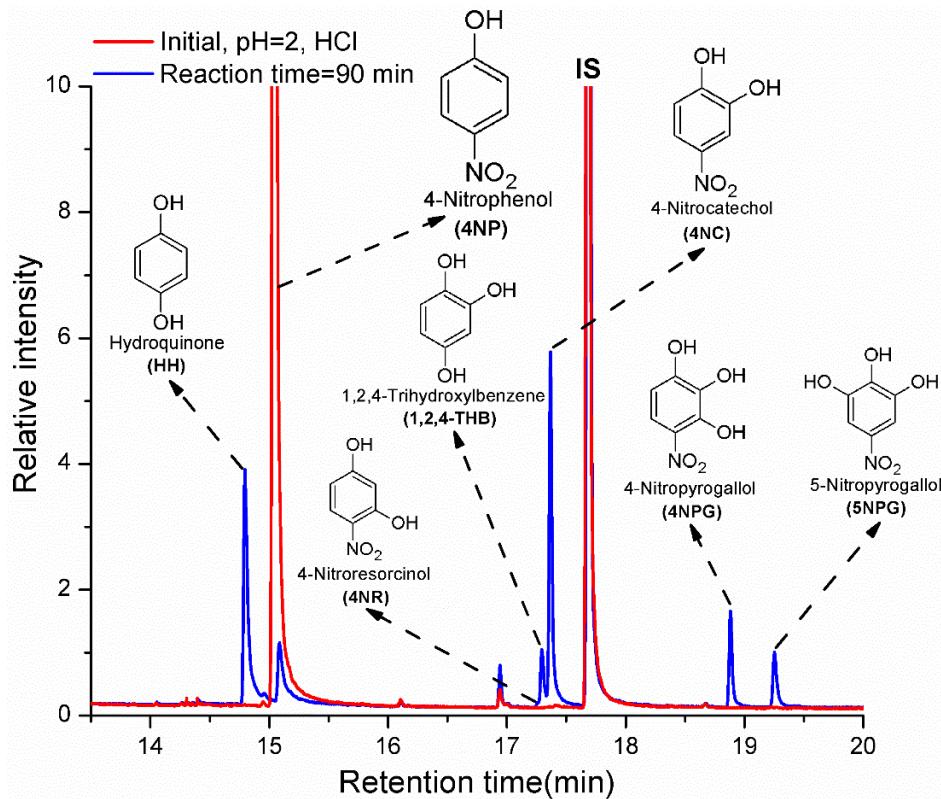
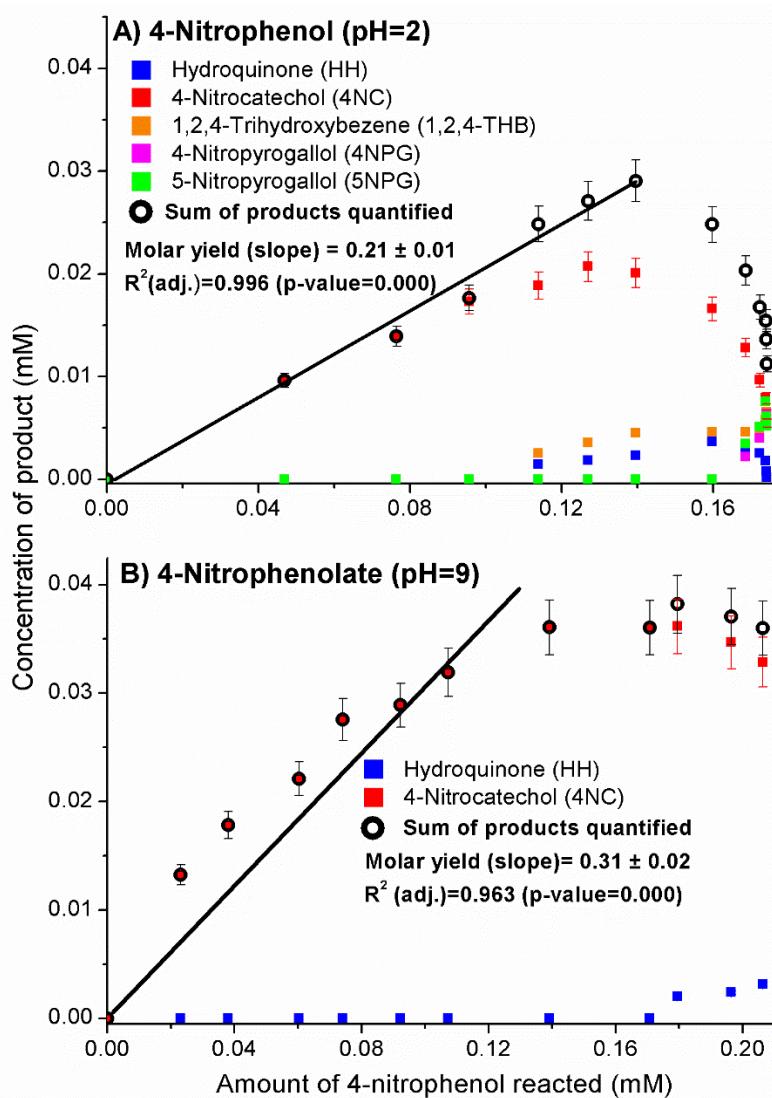


Figure 1: GC/MS chromatogram illustrating the formation of phenols from reaction (I) at pH=2.

HH, 1,2,4-THB, 4NC and 5NPG were formed from reaction (I) under acidic pH conditions (Fig.1), which is in good agreement with the previously published results (Oturan et al., 2000; Tauber et al., 2000; Liu et al., 2010; Xiong et al., 2015; Du et al.,

160 2017; Chen et al., 2018). Previously, two unknown isomers of 4NC were detected as products of the reaction (I) (Zhao et al., 2013). Here, 4-nitroresorcinol (4NR), a structural isomer of 4NC, was also tentatively identified among the products for the first time. Moreover, this study is the first to report the formation of 4NPG from reaction (I) (Xiong et al., 2015). Previously, formation of these two products (4NR and 4NPG) might have been difficult to observe due to the lack of standards and the absence of an MS detector (Tauber et al., 2000; Daneshvar et al., 2007; Liu et al., 2010). Furthermore, the insufficient resolving power of the HPLC used to study the products of the reaction (I) likely contributed to the fact that the formation of 4NR and 4NPG was not previously reported (Lipczynska-Kochany, 1991; Oturan et al., 2000; Tauber et al., 2000; Daneshvar et al., 2007; Liu et al., 2010).

165 The phenolic products of the reaction (I) were quantified with GC/MS; the results are presented in Fig. 2.



170 **Figure 2: The formation of phenolic products from the 4-nitrophenol (A, pH=2) and 4-nitrophenolate (B, pH=9) + OH reaction.** Uncertainties of the yields obtained from eq. (I) are standard errors from the linear regression analysis. Adjusted linear coefficients of the determination (R^2) > 0.96 and p-values < 0.05 were obtained, confirming that the linear correlation was statistically significant. The values of the standardized residuals from the linear regression analyses were all < 3 , confirming that none of the data points included in the regression analysis should be classified as an outlier (Table S2 and Fig. S4).

175 The results of the experiments carried out in pure water (no acids or buffers added) are not included in Fig. 2. The pH of the unbuffered reaction solution quickly decreased to ca. 3.5 (Di Paola et al., 2003), likely due to the formation of nitrite (NO_2^-) and nitrate (NO_3^-) (Kotronarou et al., 1991; Lipczynska-Kochany, 1991; Kavitha and Palanivelu, 2005; Liu et al., 2010). For this reason, the distribution of products in the acidic reaction solution and in pure water was the same.

As presented in Fig. 2, 4NC was the major product formed in both the acidic and basic reaction solutions. The other products 180 detected were HH, 1,2,4-THB, 4NPG and, 5NPG. The proposed mechanism of reaction (I) is presented in Fig. 3.

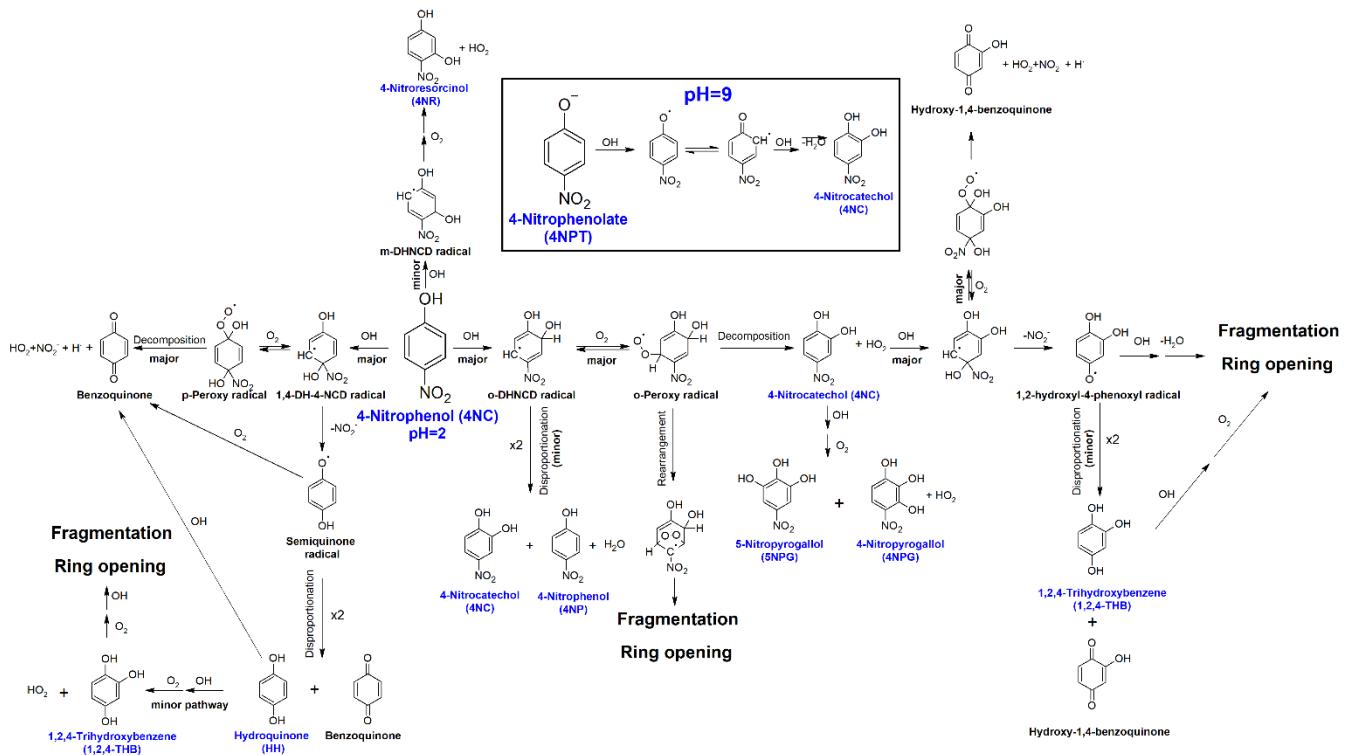


Figure 3: The proposed mechanism for the reaction of OH with 4-nitrophenol in an aqueous solution. The names of the compounds detected with GC/MS are shown in blue. Note that for clarity, all possible fragmentation and reaction pathways of the radical by-products are not shown.

185 Reaction (I) is initiated by the electrophilic addition of OH, primarily in the ortho and para positions of 4NP, yielding dihydroxynitrocyclohexadienyl (DHNCD) radicals – Fig. 3 (Kotronarou et al., 1991; Di Paola et al., 2003; Kavitha and Palanivelu, 2005). Two ortho-DHNCD radicals can then disproportionate to form 4NC and regenerate 4NP (Tauber et al., 2000; Gonzalez et al., 2004; Liu et al., 2010; Zhao et al., 2013). Alternatively, 4NC is formed by reaction of o-DHNCD with O₂ and decomposition of the resulting o-peroxy radical (Oturan et al., 2000; Di Paola et al., 2003; Liu et al., 2010; Xiong et al., 2015; Ding et al., 2016). Due to the lower molar yield of phenols at pH=2 (0.21), the ortho-addition primarily yields ring-opening products. The addition of OH at the meta position of 4NP (minor pathway) leads to the formation of 4-NR (Fig. 1). Likewise, 4NPG and 5NPG (second-generation products) are most likely formed following the 4NC+OH reaction, which also involves the formation of a peroxy-type radical.

190 Moreover, only trace amounts of 1,2,4-THB were formed from 4NC+OH under acidic conditions – see also section S7 and Fig. S6 (Oturan et al., 2000; Zhang et al., 2003; Xiong et al., 2015; Ding et al., 2016; Du et al., 2017). As previously reported, 4NC was quantitatively converted into 1,2,4-THB in the absence of O₂, which promotes the disproportionation reaction between the two 1,2-hydroxyl-4-phenoxy radicals formed by the ipso addition of OH to 4NC (Di Paola et al., 2003; Gonzalez et al., 2004; Liu et al., 2010). Hence, in the presence of O₂, the formation of 1,2,4-THB from the 4NC+OH reaction is expected

to be a minor process (Liu et al., 2010), which is consistent with the experimental data reported in section S7. Consequently, 200 it is proposed that 1,2,4-THB is likely formed following the ipso addition of OH to 4NP (Zhang et al., 2003; Kavitha and Palanivelu, 2005; Daneshvar et al., 2007). The resulting 1,4-hydroxy-4-nitrocyclohexadienyl (1,4-DH-4-NCD) type radical can then eliminate NO_2^- producing HH and 1,2,4-THB by the disproportionation mechanism (O'neill et al., 1978; Kotronarou et al., 1991). The formation of NO_2^- is also consistent with the observed rapid decrease in the pH of the unbuffered reaction solution.

205 Previously, the formation of 1,2,4-THB from the reaction of OH with HH was observed (Niessen et al., 1988; Barzaghi and Herrmann, 2002; Sobczyński et al., 2004). Therefore, under the experimental conditions used in this work, the low yield of HH is likely due to its rapid oxidation to 1,2,4-THB and BQ (not quantified) (Kotronarou et al., 1991; Oturan et al., 2000; Di Paola et al., 2003; Sobczyński et al., 2004; Gonzalez et al., 2004).

210 Due to the irreversible reactions of some phenols under investigation at pH=9 (Table S3), only 4NC and HH were detected as products of the reaction (I) from 4-NPT – Fig. 2B. At the same time, the total yield of phenols produced (mainly 4NC) increased from 0.21 (pH=2) to 0.31. At pH=9, reaction (I) can proceed by addition as well as by one-electron oxidation mechanisms, as shown in the inset in Fig. 3 (Biswal et al., 2013; Zhao et al., 2013). The reactions of the 4-nitrophenolxy radicals formed following the one-electron oxidation of 4NPT by OH are unclear (Gonzalez et al., 2004; Wojnárovits and Takács, 2008; Liu et al., 2010; Zhao et al., 2013). It was proposed (Niessen et al., 1988; Barzaghi and Herrmann, 2002) that such radicals can 215 exist in two resonance forms; in the case of 4NP, one of these resonance forms may directly react with OH to produce 4NC. This mechanism would explain the higher yield of 4NC from reaction (I) under basic pH conditions.

3.2 Light absorption and the time-evolution of brown-carbon chromophores

220 The absorbance of the reaction solution decreased during the course of the reaction (section S8) for both 4NP and 4NPT. Previously, an initial small increase in the absorbance at 420 nm of the reaction 4NP solution (pH=5) was reported, followed by rapid bleaching (Zhao et al., 2015); these differences can be caused by the slightly different reaction conditions used. Also, in this study, integrated absorbance values were used [eq. (II) and (III)] which may be a more adequate approach due to the shift of the maximum absorbance (A_{max}) of the reaction solution (Fig. S8) (Zhao et al., 2015; Hems and Abbatt, 2018).

The contribution of the light-absorbing products of the reaction (I) to the absorption of the reaction solution was evaluated with eq. (II) – the results are shown in Fig. 4.

225

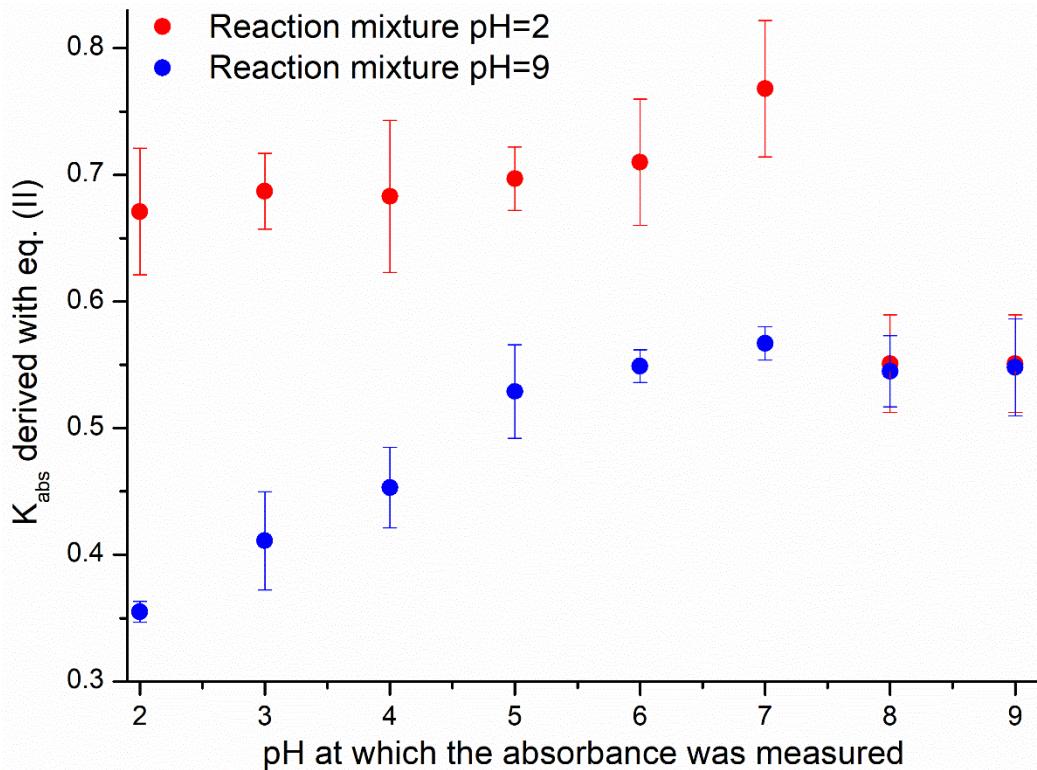
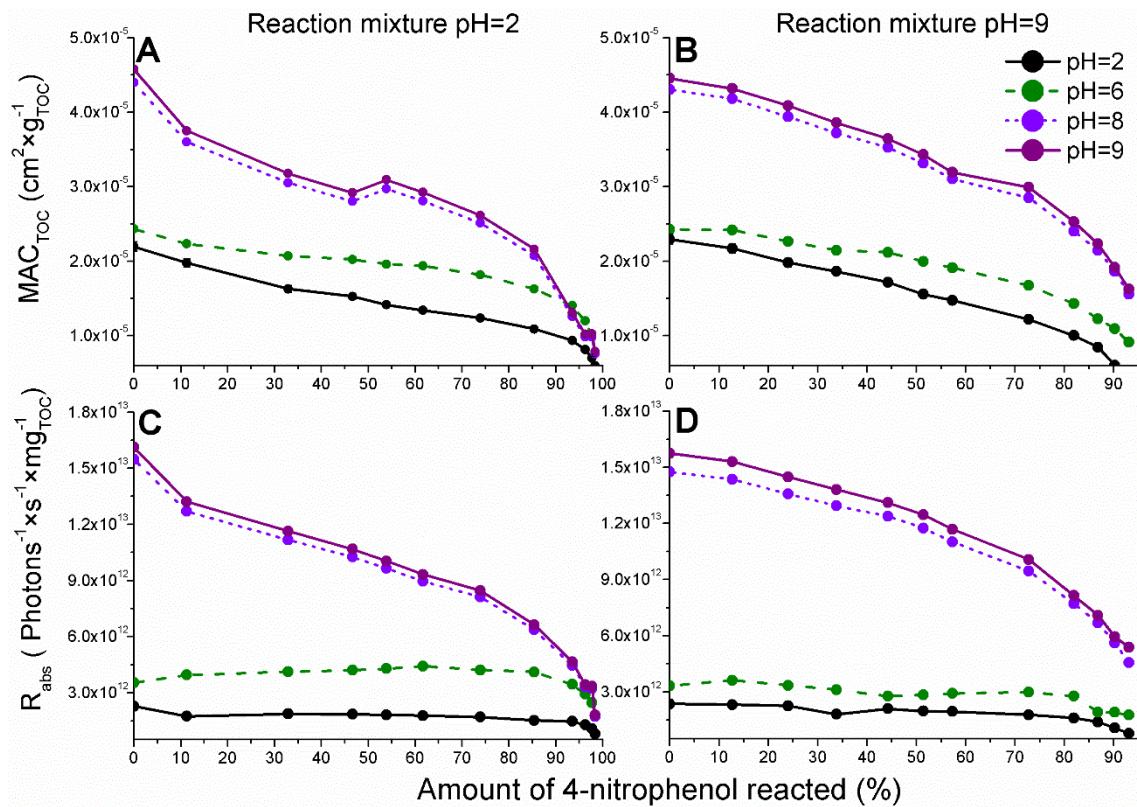


Figure 4: The ability of 4NP to generate products with an absorbance between 250 and 600 nm evaluated with K_{abs} factors derived using eq. (II). Uncertainties are precisions from the regression analysis.

The results presented in Fig.4 indicate that when the reaction was carried out under basic pH conditions the relative absorbance of the products was lower and increased more slowly. The absorbance of the products generated from 4NP (pH=2) decreased sharply when the pH of the solution before the UV-Vis measurement was adjusted to 8 and 9. Such a result indicates that the light-absorbing compounds are not stable at pH>7 which is in good agreement with the results presented in sections 3.1 and S6 (Randolph et al., 2018). The observed increase in the light absorption of the reaction solution (pH=9) between pH=2 and 9 could be due to the pH-dependence of the absorbance of the substituted carboxylic acids; aromatic ring-opening products (Oturan et al., 2000; Zhang et al., 2003; Kavitha and Palanivelu, 2005; Hems and Abbatt, 2018) with pK_a values likely falling between 3 and 6 (Rapf et al., 2017).

The values of MACtoc and R_{abs} calculated using eq. (IV) and (V) are presented in Fig. 5.



240

245

Figure 5: The pH-dependent organic carbon-normalized mass absorption coefficients (MAC_{TOC}) derived using the integrated absorbance peak for the reaction mixture absorption measured for the reaction of 4NP (A, pH=2) and 4NPT (B, pH=9) and the corresponding TOC-normalized rates of sunlight absorption (R_{abs}) for 4NP (C) and 4NPT (D). Only the data for pH =2, 6, 8, 9 are shown for clarity and the complete data are presented in Fig. S11. The colors refer to the pH at which the absorbance was measured (section 2.4). The experimental data are represented by points, with the lines provided to guide the eye.

As presented in Fig. 5, a decrease in the calculated MAC_{TOC} was observed during the course of the reaction (I). Moreover, the experimental data acquired shows an increase of MAC_{TOC} following the increase in pH at which the absorbance was measured (2, 6, 8, or 9). Such a result is likely due to the higher ϵ values for 4NP and 4NC under basic pH conditions (Fig. S5).

250

255

In Fig. 5A and 5B, the disappearance rate constants of MAC_{TOC} are of a similar order. Such a result can be explained by the formation of a higher number of light-absorbing phenols (second-generation products) at pH=2 and also due to the higher formation yield of 4NC, which has high ϵ values at pH=9. Consequently, the disappearance rate constants of the overall light absorption of 4NC following reaction (I) are mostly independent on the pH of the reaction solution and primarily depend on the pH at which the absorbance is measured. The MAC_{TOC} values calculated were slightly higher compared with the values measured for the previously investigated aromatic BrC chromophores (derived from the non-nitrated precursors) (Jiang et al., 2019; Jiang et al., 2021), likely due to the high ϵ values of 4NP and the nitrated phenols generated by the reaction (I).

The R_{abs} values (Fig. 5C and 5D) decrease slower compared with the values of MAC_{TOC} and become stable when the pH at which the absorbance was measured is less than 7 (see also Fig. S11). This is likely due to a red-shift of the A_{max} of the reaction

260 solution, likely connected to the red-shift of the 4NP and 4NCs absorbance following the increase in the pH (Fig. S5). As presented in Fig. S13, the actinic flux exhibits a significant increase between 300 and 400 nm. Consequently, the BrC chromophores – products of the reaction (I) – characterized by a strong absorbance above 400 nm, will contribute to the observed stabilization of the estimated R_{abs} at pH<7.

3.3. Atmospheric implications

265 The average measured Henry's law constant – 5×10^4 (M×atm⁻¹) – indicates that 4NP exists entirely in the aqueous phase in clouds but not in “wet” aerosols (Fig. S12) (Herrmann et al., 2015). Once dissolved in cloud water, 4NP can undergo chemical and photochemical processing; thus, the rates of the bleaching of the 4NP solution due to the reaction with OH were evaluated – Table 1.

Table 1. The pH-dependent second-order rate constants, empirical bleaching rate constants, and the effective quantum yields of photolysis were used to estimate the lifetimes of 4NPs and the corresponding BrC

Reactive species	$K_{A,4NP}$ (min ⁻¹)	$K_{A,rmix}$ (min ⁻¹)	$k_{OH} \times 10^{-9}$ (M ⁻¹ s ⁻¹)	Reference	$k_{bleaching} \times 10^{-9}$ (M ⁻¹ s ⁻¹)	Lifetime of BrC relative to the lifetime of 4NP (this work)
<i>Reaction with OH</i>						
4NP (pH<3)	0.047	0.016	6.2	(Biswal et al., 2013)	4.1	3
			3.8	(García Einschlag et al., 2003)	1.6	
			<u>4.7</u>	<u>Average literature value</u>		
4NPT (pH>8)	0.029	0.019	8.7	(Biswal et al., 2013)	6.0	1.5
<i>Photolysis</i>						
Reactive species			ϕ (molecules × photon ⁻¹) × 10 ⁶		Reference	
4NP (pH<3)			110.0 ^a		(Lemaire et al., 1985; Braman et al., 2020)	

270 **^aThe average quantum yields measured in the presence of photorecalcitrant α -pinene SOA used as a scavenger of the secondary OH formed following the photolysis of HONO formed during the photolysis of 4NP.**

Presented in Table 1, the first-order $k_{\text{bleaching}}$ rate constants were derived with eq. (II) using the experimental data acquired in this study and the lifetimes due to the reaction with OH were estimated with eq. (SIII). The data summarized in Table 1 shows that the lifetimes of BrC chromophores are 3 and 1.5-times longer than the lifetime of 4NP (precursor) under acidic and basic pH conditions, respectively, due to the formation of light-absorbing products.

275 The previously reported average quantum yields (ϕ , molecules \times photon $^{-1}$) (Lemaire et al., 1985; García Einschlag et al., 2003; Biswal et al., 2013; Braman et al., 2020) were used to derive the lifetimes of 4NP due to photolysis using eq. (SIV). The literature ϕ values listed in Table 1 were derived by measuring the decrease in the absorbance of the 4NP solution, hence can be regarded as effective ϕ for the bleaching of chromophores in aqueous solutions of 4NP and 4NPT (Lemaire et al., 1985; Braman et al., 2020). To our knowledge, the wavelength-dependent ϕ values for 4NP and 4NPT are not available. The 280 estimated lifetimes of light-absorbing compounds in the cloud water particles containing 4NP due to direct photolysis and due reaction with the OH were compared (Fig. 6).

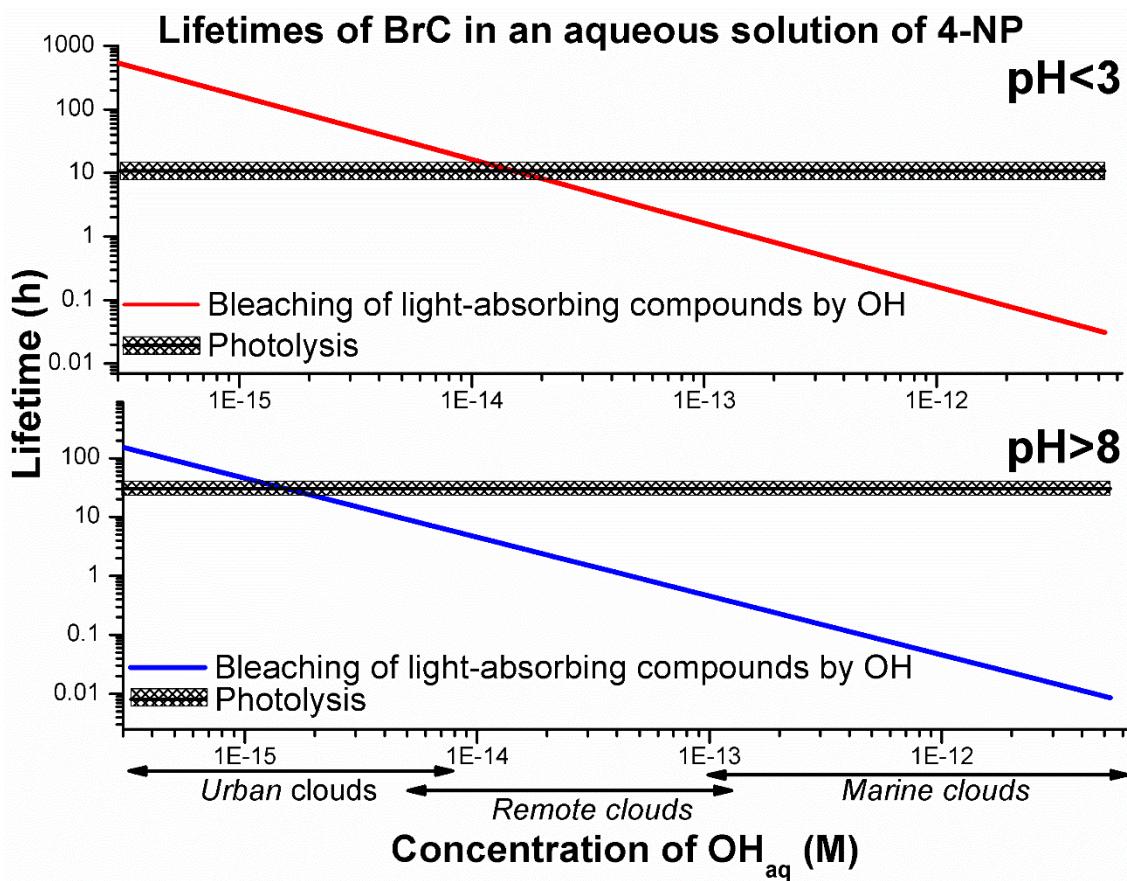


Figure 6: The estimated aqueous-phase lifetimes of light-absorbing compounds in solutions of 4NP and 4NPT. The lifetimes due to the reaction with OH were calculated with $k_{\text{bleaching}}$ rate constants derived with eq. (III) using the experimental data acquired in this 285 study. The average lifetime due to photolysis is shown, shaded areas are 2σ values, representing the range of photolysis lifetimes calculated for the solar zenith angles 0–50°.

As presented in Fig. 6, the photolysis of 4NP BrC may be relevant under realistic atmospheric conditions in urban and remote clouds, with the estimated concentration of OH_{aq} lower than $1 \times 10^{-13} \text{ M}$ (Herrmann et al., 2010). Bleaching by OH is expected to be a more efficient removal mechanism for 4NPT, due to its lower reported quantum yields (ϕ) combined with the higher 290 reactivity of the precursor towards OH at pH>8 (Lemaire et al., 1985; Braman et al., 2020).

A low degree of mineralization of the precursor (ca. 15%) was observed for both 4NP and 4NPT, as presented in Fig. 7.

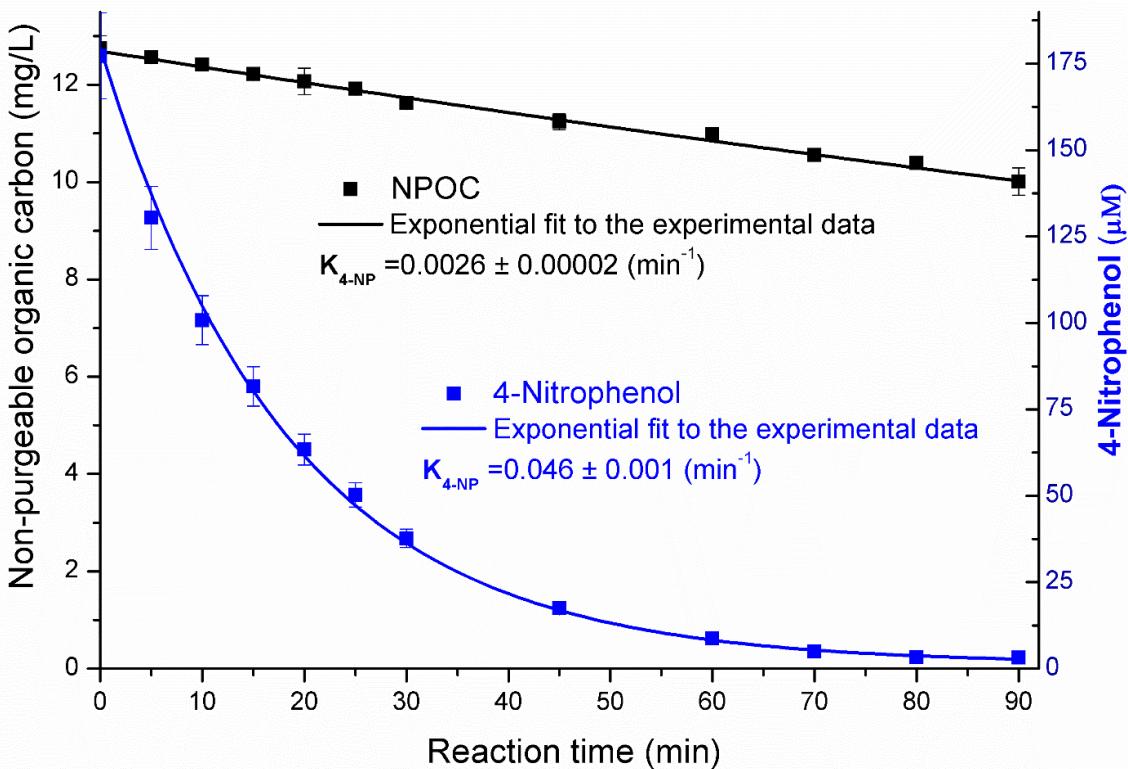


Figure 7: The concentrations of 4NP and non-purgeable organic carbon during the reaction in pure water (the results obtained for pH=2 and 9 reaction solutions were very similar – section S11). The uncertainties for the first-order decay rates (K , min^{-1}) are standard errors from the regression analysis.

295

Due to the low degree of mineralization of 4NP (Fig. 7) combined with the yields of phenolic products calculated as ~0.2 (pH=2) and ~0.3 (pH=9), it is estimated that the yield of non-aromatic products from reaction (I) was between 0.51 and 0.68. These aromatic ring-opening products may include functionalized carboxylic acid, as previously proposed (Oturan et al., 2000; Zhang et al., 2003; Kavitha and Palanivelu, 2005; Hems and Abbatt, 2018).

300

4. Conclusions

305

The results acquired in this study show that the reaction of OH with 4NP in cloud water leads to the bleaching of light-absorbing compounds. Previously, the reaction with OH was concluded to be a major removal mechanism for a number of nitrophenols in the atmospheric aqueous phase (Vione et al., 2009; Albinet et al., 2010; Zhao et al., 2015). As previously reported, the reaction of OH with 5-nitroguaiacol, 4NC, and dinitrophenol initially leads to a small increase in the light absorption followed by rapid bleaching (Zhao et al., 2015; Hems and Abbatt, 2018) but such behavior was not observed for 4NP in this study. By comparing the results obtained in this study with the literature data (Zhao et al., 2015; Hems and Abbatt, 2018), it can also be concluded that more substituted nitrophenols initially yield higher amounts of light-absorbing products compared to 4NP.

Moreover, the atmospheric lifetimes of BrC chromophores due to the reaction with OH are expected to be significantly longer than the lifetimes of the parent nitrophenols (precursors) due to the formation of light-absorbing by-products (Zhao et al., 2015; 310 Hems and Abbatt, 2018). The results obtained here and the data from the literature also indicate that the aqueous reaction of nitrophenols with OH can potentially contribute to cloud water acidity due to the formation of NO_2^- , NO_3^- and organic nitro carboxylic acids (Tilgner et al., 2021). Additionally, the aqueous oxidation of nitrophenols by OH may be an important source of potentially toxic and harmful aqueous SOAs due to the formation of nitrogen-containing organic compounds combined with the low degree of mineralization of such precursors.

315 *Data availability.* The raw data can be obtained by contacting the corresponding author.

Author contributions. BW designed the study, developed the methodology, analyzed the data and wrote the paper. PJ carried out the experiments, optimized the methodology and processed the raw data. TG supervised the experiments, analyzed the data and contributed to the final manuscript. All authors contributed to the interpretation of the results.

Competing interests. The authors declare that they have no conflicts of interest

320 *Acknowledgments.* This work was carried out at the Biological and Chemical Research Centre, University of Warsaw, established within the project co-financed by the European Union from the European Regional Development Fund under the

Operational Programme Innovative Economy, 2007–2013. We thank Dr Bartłomiej Kiersztyn for making the TOC measurements possible. We thank the anonymous reviewers for the very insightful comments and suggestions.

Financial support. This project was funded by the Polish National Science Centre: grant number UMO-325 2018/31/B/ST10/01865. The authors also acknowledge the support from the statutory research funds of the University of Warsaw, decision number BOB-661-199/2020.

References

Albinet, A., Minero, C., and Vione, D.: Phototransformation processes of 2,4-dinitrophenol, relevant to atmospheric water droplets, 80, 753-758, <https://doi.org/10.1016/j.chemosphere.2010.05.016>, 2010.

330 Balasubramanian, P., Balamurugan, T. S. T., Chen, S.-M., and Chen, T.-W.: Simplistic synthesis of ultrafine CoMnO₃ nanosheets: An excellent electrocatalyst for highly sensitive detection of toxic 4-nitrophenol in environmental water samples, 361, 123-133, <https://doi.org/10.1016/j.jhazmat.2018.08.070>, 2019.

Barzaghi, P. and Herrmann, H.: A mechanistic study of the oxidation of phenol by OH/NO₂/NO₃ in aqueous solution, 4, 3669-3675, 10.1039/B201652D, 2002.

335 Biswal, J., Paul, J., Naik, D. B., Sarkar, S. K., and Sabharwal, S.: Radiolytic degradation of 4-nitrophenol in aqueous solutions: Pulse and steady state radiolysis study, 85, 161-166, <https://doi.org/10.1016/j.radphyschem.2013.01.003>, 2013.

Bluvshtein, N., Lin, P., Flores, J. M., Segev, L., Mazar, Y., Tas, E., Snider, G., Weagle, C., Brown, S. S., Laskin, A., and Rudich, Y.: Broadband optical properties of biomass-burning aerosol and identification of brown carbon chromophores, 122, 5441-5456, <https://doi.org/10.1002/2016JD026230>, 2017.

340 Braman, T., Dolvin, L., Thrasher, C., Yu, H., Walhout, E. Q., and O'Brien, R. E.: Fresh versus Photo-recalcitrant Secondary Organic Aerosol: Effects of Organic Mixtures on Aqueous Photodegradation of 4-Nitrophenol, 7, 248-253, 10.1021/acs.estlett.0c00177, 2020.

Chen, C., Han, Y., Guo, J., Zhou, L., and Lan, Y.: Assessing the role of silica gel in the degradation of p-nitrophenol via Zn(0)-activated persulfate, 88, 169-176, <https://doi.org/10.1016/j.jtice.2018.03.053>, 2018.

345 Claeys, M., Vermeylen, R., Yasmeen, F., Gómez-González, Y., Chi, X., Maenhaut, W., Mészáros, T., and Salma, I.: Chemical characterisation of humic-like substances from urban, rural and tropical biomass burning environments using liquid chromatography with UV/vis photodiode array detection and electrospray ionisation mass spectrometry, Environ. Chem., 9, 273-284, <https://doi.org/10.1071/EN11163>, 2012.

Cordell, R. L., Mazet, M., Dechoux, C., Hama, S. M. L., Staelens, J., Hofman, J., Stroobants, C., Roekens, E., Kos, G. P. A., 350 Weijers, E. P., Frumau, K. F. A., Panteliadis, P., Delaunay, T., Wyche, K. P., and Monks, P. S.: Evaluation of biomass burning across North West Europe and its impact on air quality, 141, 276-286, <https://doi.org/10.1016/j.atmosenv.2016.06.065>, 2016.

Daneshvar, N., Behnajady, M. A., and Zorriyeh Asghar, Y.: Photooxidative degradation of 4-nitrophenol (4-NP) in UV/H₂O₂ process: Influence of operational parameters and reaction mechanism, 139, 275-279,

355 <https://doi.org/10.1016/j.jhazmat.2006.06.045>, 2007.

Desyaterik, Y., Sun, Y., Shen, X., Lee, T., Wang, X., Wang, T., and Collett Jr., J. L.: Speciation of “brown” carbon in cloud water impacted by agricultural biomass burning in eastern China, 118, 7389-7399, <https://doi.org/10.1002/jgrd.50561>, 2013.

Di Paola, A., Augugliaro, V., Palmisano, L., Pantaleo, G., and Savinov, E.: Heterogeneous photocatalytic degradation of nitrophenols, 155, 207-214, [https://doi.org/10.1016/S1010-6030\(02\)00390-8](https://doi.org/10.1016/S1010-6030(02)00390-8), 2003.

360 Ding, R., Mao, Z.-Y., and Wang, J.-L.: Synergistic effects of 4-nitrophenol degradation using gamma irradiation combined with a advanced oxidation process, 27, 4, 10.1007/s41365-016-0004-y, 2016.

Du, J., Che, D., Li, X., Guo, W., and Ren, N.: Factors affecting p-nitrophenol removal by microscale zero-valent iron coupling with weak magnetic field (WMF), 7, 18231-18237, 10.1039/C7RA02002C, 2017.

Feng, Y., Ramanathan, V., and Kotamarthi, V. R.: Brown carbon: a significant atmospheric absorber of solar radiation?, 13, 8607-8621, 10.5194/acp-13-8607-2013, 2013.

365 Fleming, L. T., Lin, P., Roberts, J. M., Selimovic, V., Yokelson, R., Laskin, J., Laskin, A., and Nizkorodov, S. A.: Molecular composition and photochemical lifetimes of brown carbon chromophores in biomass burning organic aerosol, 20, 1105-1129, 10.5194/acp-20-1105-2020, 2020.

Forrister, H., Liu, J., Scheuer, E., Dibb, J., Ziembka, L., Thornhill, K. L., Anderson, B., Diskin, G., Perring, A. E., Schwarz, J. P., Campuzano-Jost, P., Day, D. A., Palm, B. B., Jimenez, J. L., Nenes, A., and Weber, R. J.: Evolution of brown carbon in wildfire plumes, 42, 4623-4630, <https://doi.org/10.1002/2015GL063897>, 2015.

Frka, S., Šala, M., Kroflič, A., Huš, M., Čusak, A., and Grgić, I.: Quantum Chemical Calculations Resolved Identification of Methylnitrocatechols in Atmospheric Aerosols, *Environ. Sci. Technol.*, 50, 5526-5535, 10.1021/acs.est.6b00823, 2016.

García Einschlag, F. S., Carlos, L., and Capparelli, A. L.: Competition kinetics using the UV/H₂O₂ process: a structure 375 reactivity correlation for the rate constants of hydroxyl radicals toward nitroaromatic compounds, 53, 1-7, [https://doi.org/10.1016/S0045-6535\(03\)00388-6](https://doi.org/10.1016/S0045-6535(03)00388-6), 2003.

Gierczak, T., Bernard, F., Papanastasiou, D. K., and Burkholder, J. B.: Atmospheric Chemistry of c-C₅HF₇ and c-C₅F₈: Temperature-Dependent OH Reaction Rate Coefficients, Degradation Products, Infrared Spectra, and Global Warming Potentials, 125, 1050-1061, 10.1021/acs.jpca.0c10561, 2021.

380 Gonzalez, M. G., Oliveros, E., Wörner, M., and Braun, A. M.: Vacuum-ultraviolet photolysis of aqueous reaction systems, 5, 225-246, <https://doi.org/10.1016/j.jphotochemrev.2004.10.002>, 2004.

Harrison, M. A. J., Heal, M. R., and Cape, J. N.: Evaluation of the pathways of tropospheric nitrophenol formation from benzene and phenol using a multiphase model, 5, 1679-1695, 10.5194/acp-5-1679-2005, 2005a.

Harrison, M. A. J., Barra, S., Borghesi, D., Vione, D., Arsene, C., and Iulian Olariu, R.: Nitrated phenols in the atmosphere: 385 a review, 39, 231-248, <https://doi.org/10.1016/j.atmosenv.2004.09.044>, 2005b.

Heal, M. R., Harrison, M. A. J., and Neil Cape, J.: Aqueous-phase nitration of phenol by N₂O₅ and ClO₂, 41, 3515-3520, <https://doi.org/10.1016/j.atmosenv.2007.02.003>, 2007.

Hems, R. F. and Abbatt, J. P. D.: Aqueous Phase Photo-oxidation of Brown Carbon Nitrophenols: Reaction Kinetics, Mechanism, and Evolution of Light Absorption, 2, 225-234, 10.1021/acsearthspacechem.7b00123, 2018.

390 Hems, R. F., Schnitzler, E. G., Liu-Kang, C., Cappa, C. D., and Abbatt, J. P. D.: Aging of Atmospheric Brown Carbon Aerosol, *ACS Earth Space Chem.*, 5, 722-748, 10.1021/acsearthspacechem.0c00346, 2021.

Hems, R. F., Schnitzler, E. G., Bastawrous, M., Soong, R., Simpson, A. J., and Abbatt, J. P. D.: Aqueous Photoreactions of Wood Smoke Brown Carbon, 4, 1149-1160, 10.1021/acsearthspacechem.0c00117, 2020.

Herrmann, H., Hoffmann, D., Schaefer, T., Bräuer, P., and Tilgner, A.: Tropospheric Aqueous-Phase Free-Radical 395 Chemistry: Radical Sources, Spectra, Reaction Kinetics and Prediction Tools, 11, 3796-3822, <https://doi.org/10.1002/cphc.201000533>, 2010.

Herrmann, H., Schaefer, T., Tilgner, A., Styler, S. A., Weller, C., Teich, M., and Otto, T.: Tropospheric Aqueous-Phase Chemistry: Kinetics, Mechanisms, and Its Coupling to a Changing Gas Phase, *Chem. Rev.*, 115, 4259-4334, 10.1021/cr500447k, 2015.

400 Hettiyadura, A. P. S., Garcia, V., Li, C., West, C. P., Tomlin, J., He, Q., Rudich, Y., and Laskin, A.: Chemical Composition and Molecular-Specific Optical Properties of Atmospheric Brown Carbon Associated with Biomass Burning, 55, 2511-2521, 10.1021/acs.est.0c05883, 2021.

Inomata, S., Fushimi, A., Sato, K., Fujitani, Y., and Yamada, H.: 4-Nitrophenol, 1-nitropyrene, and 9-nitroanthracene emissions in exhaust particles from diesel vehicles with different exhaust gas treatments, 110, 93-102, <https://doi.org/10.1016/j.atmosenv.2015.03.043>, 2015.

405 Jaber, F., Schummer, C., Al Chami, J., Mirabel, P., and Millet, M.: Solid-phase microextraction and gas chromatography-mass spectrometry for analysis of phenols and nitrophenols in rainwater, as their t-butyldimethylsilyl derivatives, 387, 2527-2535, 10.1007/s00216-006-1115-9, 2007.

410 Jacobson, M. Z.: Isolating nitrated and aromatic aerosols and nitrated aromatic gases as sources of ultraviolet light absorption, 104, 3527-3542, <https://doi.org/10.1029/1998JD100054>, 1999.

415 Jiang, H., Frie, A. L., Lavi, A., Chen, J. Y., Zhang, H., Bahreini, R., and Lin, Y.-H.: Brown Carbon Formation from Nighttime Chemistry of Unsaturated Heterocyclic Volatile Organic Compounds, 6, 184-190, 10.1021/acs.estlett.9b00017, 2019.

420 Jiang, W., Misovich, M. V., Hettiyadura, A. P. S., Laskin, A., McFall, A. S., Anastasio, C., and Zhang, Q.: Photosensitized Reactions of a Phenolic Carbonyl from Wood Combustion in the Aqueous Phase—Chemical Evolution and Light Absorption Properties of AqSOA, 10.1021/acs.estlett.0c07581, 2021.

425 Kahnt, A., Behrouzi, S., Vermeylen, R., Safi Shalamzari, M., Vercauteren, J., Roekens, E., Claeys, M., and Maenhaut, W.: One-year study of nitro-organic compounds and their relation to wood burning in PM10 aerosol from a rural site in Belgium, 81, 561-568, <https://doi.org/10.1016/j.atmosenv.2013.09.041>, 2013.

430 Kavitha, V. and Palanivelu, K.: Degradation of nitrophenols by Fenton and photo-Fenton processes, 170, 83-95, <https://doi.org/10.1016/j.jphotochem.2004.08.003>, 2005.

435 Kitanovski, Z., Grgić, I., Vermeylen, R., Claeys, M., and Maenhaut, W.: Liquid chromatography tandem mass spectrometry method for characterization of monoaromatic nitro-compounds in atmospheric particulate matter, *J. Chromatogr. A*, 1268, 35-43, <https://doi.org/10.1016/j.chroma.2012.10.021>, 2012.

440 Kitanovski, Z., Shahpoury, P., Samara, C., Voliotis, A., and Lammel, G.: Composition and mass size distribution of nitrated and oxygenated aromatic compounds in ambient particulate matter from southern and central Europe – implications for the origin, 20, 2471-2487, 10.5194/acp-20-2471-2020, 2020.

445 Kotronarou, A., Mills, G., and Hoffmann, M. R.: Ultrasonic irradiation of p-nitrophenol in aqueous solution, 95, 3630-3638, 10.1021/j100162a037, 1991.

450 Laskin, A., Laskin, J., and Nizkorodov, S. A.: Chemistry of Atmospheric Brown Carbon, 115, 4335-4382, 10.1021/cr5006167, 2015.

455 Lemaire, J., Guth, J. A., Klais, O., Leahey, J., Merz, W., Philp, J., Wilmes, R., and Wolff, C. J. M.: Ring test of a method for assessing the phototransformation of chemicals in water, 14, 53-77, [https://doi.org/10.1016/0045-6535\(85\)90041-4](https://doi.org/10.1016/0045-6535(85)90041-4), 1985.

460 Li, C., He, Q., Hettiyadura, A. P. S., Käfer, U., Shmul, G., Meidan, D., Zimmermann, R., Brown, S. S., George, C., Laskin, A., and Rudich, Y.: Formation of Secondary Brown Carbon in Biomass Burning Aerosol Proxies through NO₃ Radical Reactions, 54, 1395-1405, 10.1021/acs.est.9b05641, 2020.

465 Liang, Y., Wang, X., Dong, S., Liu, Z., Mu, J., Lu, C., Zhang, J., Li, M., Xue, L., and Wang, W.: Size distributions of nitrated phenols in winter at a coastal site in north China and the impacts from primary sources and secondary formation, 250, 126256, <https://doi.org/10.1016/j.chemosphere.2020.126256>, 2020.

470 Lipczynska-Kochany, E.: Novel method for a photocatalytic degradation of 4-nitrophenol in homogeneous aqueous solution, 12, 87-92, 10.1080/09593339109384985, 1991.

475 Liu, Y., Wang, D., Sun, B., and Zhu, X.: Aqueous 4-nitrophenol decomposition and hydrogen peroxide formation induced by contact glow discharge electrolysis, 181, 1010-1015, <https://doi.org/10.1016/j.jhazmat.2010.05.115>, 2010.

480 Lu, Z., Streets, D. G., Winikul, E., Yan, F., Chen, Y., Bond, T. C., Feng, Y., Dubey, M. K., Liu, S., Pinto, J. P., and Carmichael, G. R.: Light Absorption Properties and Radiative Effects of Primary Organic Aerosol Emissions, 49, 4868-4877, 10.1021/acs.est.5b00211, 2015.

485 Majewska, M., Khan, F., Pieta, I. S., Wróblewska, A., Szmigelski, R., and Pieta, P.: Toxicity of selected airborne nitrophenols on eukaryotic cell membrane models, 266, 128996, <https://doi.org/10.1016/j.chemosphere.2020.128996>, 2021.

490 Mohr, C., Lopez-Hilfiker, F. D., Zotter, P., Prévôt, A. S. H., Xu, L., Ng, N. L., Herndon, S. C., Williams, L. R., Franklin, J. P., Zahniser, M. S., Worsnop, D. R., Knighton, W. B., Aiken, A. C., Gorkowski, K. J., Dubey, M. K., Allan, J. D., and Thornton, J. A.: Contribution of Nitrated Phenols to Wood Burning Brown Carbon Light Absorption in Detling, United Kingdom during Winter Time, 47, 6316-6324, 10.1021/es400683v, 2013.

495 Moise, T., Flores, J. M., and Rudich, Y.: Optical Properties of Secondary Organic Aerosols and Their Changes by Chemical Processes, 115, 4400-4439, 10.1021/cr5005259, 2015.

500 Natangelo, M., Mangiapan, S., Bagnati, R., Benfenati, E., and Fanelli, R.: Increased concentrations of nitrophenols in leaves from a damaged forestal site, 38, 1495-1503, [https://doi.org/10.1016/S0045-6535\(98\)00370-1](https://doi.org/10.1016/S0045-6535(98)00370-1), 1999.

505 NCAR TUV calculator: https://www.acom.ucar.edu/Models/TUV/Interactive_TUV/, last access: Sep 2021.

Niessen, R., Lenoir, D., and Boule, P.: Phototransformation of phenol induced by excitation of nitrate ions, 17, 1977-1984, [https://doi.org/10.1016/0045-6535\(88\)90009-4](https://doi.org/10.1016/0045-6535(88)90009-4), 1988.

460 O'Neill, P., Steenken, S., van der Linde, H., and Schulte-Frohlinde, D.: Reaction of OH radicals with nitrophenols in aqueous solution, 12, 13-17, [https://doi.org/10.1016/0146-5724\(78\)90070-5](https://doi.org/10.1016/0146-5724(78)90070-5), 1978.

465 Oturan, M. A., Peiroten, J., Chartrin, P., and Acher, A. J.: Complete Destruction of p-Nitrophenol in Aqueous Medium by Electro-Fenton Method, 34, 3474-3479, 10.1021/es990901b, 2000.

Randolph, C., Lahive, C. W., Sami, S., Havenith, R. W. A., Heeres, H. J., and Deuss, P. J.: Biobased Chemicals: 1,2,4-
465 Benzenetriol, Selective Deuteration and Dimerization to Bifunctional Aromatic Compounds, 22, 1663-1671, 10.1021/acs.oprd.8b00303, 2018.

Rapf, R. J., Dooley, M. R., Kappes, K., Perkins, R. J., and Vaida, V.: pH Dependence of the Aqueous Photochemistry of α -Keto Acids, 121, 8368-8379, 10.1021/acs.jpca.7b08192, 2017.

470 Regueiro, J., Becerril, E., Garcia-Jares, C., and Llompart, M.: Trace analysis of parabens, triclosan and related chlorophenols in water by headspace solid-phase microextraction with in situ derivatization and gas chromatography-tandem mass spectrometry, 1216, 4693-4702, <https://doi.org/10.1016/j.chroma.2009.04.025>, 2009.

Rived, F., Rosés, M., and Bosch, E.: Dissociation constants of neutral and charged acids in methyl alcohol. The acid strength resolution, 374, 309-324, [https://doi.org/10.1016/S0003-2670\(98\)00418-8](https://doi.org/10.1016/S0003-2670(98)00418-8), 1998.

475 Sander, R.: Compilation of Henry's law constants (version 4.0) for water as solvent, 15, 4399-4981, 10.5194/acp-15-4399-2015, 2015.

Sobczyński, A., Duczmal, Ł., and Zmudziński, W.: Phenol destruction by photocatalysis on TiO₂: an attempt to solve the reaction mechanism, 213, 225-230, <https://doi.org/10.1016/j.molcata.2003.12.006>, 2004.

Tan, Y., Perri, M. J., Seitzinger, S. P., and Turpin, B. J.: Effects of Precursor Concentration and Acidic Sulfate in Aqueous Glyoxal-OH Radical Oxidation and Implications for Secondary Organic Aerosol, 43, 8105-8112, 10.1021/es901742f, 2009.

480 Tauber, A., Schuchmann, H.-P., and von Sonntag, C.: Sonolysis of aqueous 4-nitrophenol at low and high pH, 7, 45-52, [https://doi.org/10.1016/S1350-4177\(99\)00018-8](https://doi.org/10.1016/S1350-4177(99)00018-8), 2000.

TenBrook, P. L., Kendall, S. M., Viant, M. R., and Tjeerdema, R. S.: Toxicokinetics and biotransformation of p-nitrophenol in red abalone (*Haliotis rufescens*), 62, 329-336, [https://doi.org/10.1016/S0166-445X\(02\)00103-0](https://doi.org/10.1016/S0166-445X(02)00103-0), 2003.

485 Tilgner, A., Schaefer, T., Alexander, B., Barth, M., Collett Jr, J. L., Fahey, K. M., Nenes, A., Pye, H. O. T., Herrmann, H., and McNeill, V. F.: Acidity and the multiphase chemistry of atmospheric aqueous particles and clouds, 21, 13483-13536, 10.5194/acp-21-13483-2021, 2021.

Vidović, K., Kroflič, A., Šala, M., and Grgić, I.: Aqueous-Phase Brown Carbon Formation from Aromatic Precursors under Sunlight Conditions, 11, 131, 2020.

490 Vione, D., Maurino, V., Minero, C., and Pelizzetti, E.: Aqueous Atmospheric Chemistry: Formation of 2,4-Dinitrophenol upon Nitration of 2-Nitrophenol and 4-Nitrophenol in Solution, 39, 7921-7931, 10.1021/es050824m, 2005.

Vione, D., Maurino, V., Minero, C., Borghesi, D., Lucchiari, M., and Pelizzetti, E.: New Processes in the Environmental Chemistry of Nitrite. 2. The Role of Hydrogen Peroxide, 37, 4635-4641, 10.1021/es0300259, 2003.

495 Vione, D., Maurino, V., Minero, C., Duncianu, M., Olariu, R.-I., Arsene, C., Sarakha, M., and Mailhot, G.: Assessing the transformation kinetics of 2- and 4-nitrophenol in the atmospheric aqueous phase. Implications for the distribution of both nitroisomers in the atmosphere, 43, 2321-2327, <https://doi.org/10.1016/j.atmosenv.2009.01.025>, 2009.

Wang, X., Heald, C. L., Ridley, D. A., Schwarz, J. P., Spackman, J. R., Perring, A. E., Coe, H., Liu, D., and Clarke, A. D.: Exploiting simultaneous observational constraints on mass and absorption to estimate the global direct radiative forcing of black carbon and brown carbon, 14, 10989-11010, 10.5194/acp-14-10989-2014, 2014.

500 Witkowski, B., Al-sharafi, M., and Gierczak, T.: Kinetics and products of the aqueous-phase oxidation of β -caryophyllonic acid by hydroxyl radicals, 213, 231-238, <https://doi.org/10.1016/j.atmosenv.2019.06.016>, 2019.

Wojnárovits, L. and Takács, E.: Irradiation treatment of azo dye containing wastewater: An overview, 77, 225-244, <https://doi.org/10.1016/j.radphyschem.2007.05.003>, 2008.

Xie, M., Chen, X., Hays, M. D., and Holder, A. L.: Composition and light absorption of N-containing aromatic compounds in organic aerosols from laboratory biomass burning, 19, 2899-2915, 10.5194/acp-19-2899-2019, 2019.

505 Xiong, X., Sun, Y., Sun, B., Song, W., Sun, J., Gao, N., Qiao, J., and Guan, X.: Enhancement of the advanced Fenton process by weak magnetic field for the degradation of 4-nitrophenol, 5, 13357-13365, 10.1039/C4RA16318D, 2015.

Yan, J., Wang, X., Gong, P., Wang, C., and Cong, Z.: Review of brown carbon aerosols: Recent progress and perspectives, 634, 1475-1485, <https://doi.org/10.1016/j.scitotenv.2018.04.083>, 2018.

510 Zhang, W., Xiao, X., An, T., Song, Z., Fu, J., Sheng, G., and Cui, M.: Kinetics, degradation pathway and reaction mechanism of advanced oxidation of 4-nitrophenol in water by a UV/H₂O₂ process, 78, 788-794, <https://doi.org/10.1002/jctb.864>, 2003.

Zhang, Y., Forrister, H., Liu, J., Dibb, J., Anderson, B., Schwarz, J. P., Perring, A. E., Jimenez, J. L., Campuzano-Jost, P., Wang, Y., Nenes, A., and Weber, R. J.: Top-of-atmosphere radiative forcing affected by brown carbon in the upper troposphere, 10, 486-489, 10.1038/ngeo2960, 2017.

515 Zhao, R., Lee, A. K. Y., Huang, L., Li, X., Yang, F., and Abbatt, J. P. D.: Photochemical processing of aqueous atmospheric brown carbon, 15, 6087-6100, 10.5194/acp-15-6087-2015, 2015.

Zhao, S., Ma, H., Wang, M., Cao, C., and Yao, S.: Study on the role of hydroperoxy radical in degradation of p-nitrophenol attacked by hydroxyl radical using photolytical technique, 259, 17-24, <https://doi.org/10.1016/j.jphotochem.2013.02.012>, 2013.

520