



Formation characteristics of aerosol triplet state and coupling effect between the separated components with different polarity

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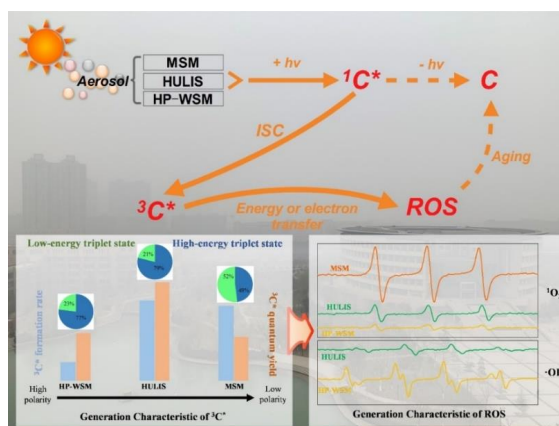
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1 **Abstract:** Atmospheric aerosols contain organic matter that can form triplet state ($^3\text{C}^*$)
2 excited by sunlight, which plays a critical role in the aging process of aerosols. In order to
3 understand the triplet state reaction mechanism of complex aerosol components, the
4 formation characteristics of $^3\text{C}^*$ in the aerosol components with different polarity, i.e., the
5 highly polar water-soluble matter (HP-WSM), humic-like substances (HULIS) and
6 methanol-soluble matter (MSM) were investigated. The coupling effect of generation of $^3\text{C}^*$
7 and reactive oxygen species (ROS) between different aerosol components was also
8 examined. The results show that the $^3\text{C}^*$ generation characteristics is strongly dependent on
9 the polarity of these components. HULIS has the strongest generation ability of $^3\text{C}^*$, and the
10 MSM contribute the most to the total generation of $^3\text{C}^*$. It is found that the high-energy
11 triplet states ($E_T \geq 250 \text{ kJ mol}^{-1}$) of HULIS and HP-WSM components account for up to 80%.
12 These $^3\text{C}^*$ has an important contribution to the photochemically generation of ROS, and the
13 generated ROS of different components are also different, which is determined by the
14 chromophore composition of complex organic matter. Tyrosine-like chromophore is the
15 main substance leading to the formation of water-soluble $^3\text{C}^*$, while the highly oxidized
16 HULIS chromophore plays a leading role in the water-insoluble component. This study also
17 found that there is a coupling effect between HP-WSM and HULIS on $^3\text{C}^*$ and ROS
18 generation. The $^3\text{C}^*$ generation rate increases by about 40% after mixing, but the generation
19 of $^1\text{O}_2$ is severely reduced. Overall, this study provides deep insights into the generation
20 characteristics of the triplet state of atmospheric aerosols.

21 **Key words:** atmospheric aerosols; photochemistry; triplet state; reactive oxygen species;
22 coupling effect

23 **TOC:**



24



25 1. Introduction

26 Atmospheric photochemical processes play an important role in the formation and
27 aging of aerosols (Atkinson, 2000; Derwent et al., 1998; Lim et al., 2005; Ervens et al., 2011;
28 Blando and Turpin, 2000). For example, the photo-oxidation process of organic matter and
29 nitrogen oxides affect the cycle of carbon and nitrogen elements (Rollins et al., 2012;
30 Goldstein and Galbally, 2007), and the volatile organic compounds undergo photochemical
31 reactions to generate secondary organic aerosols (SOA) (Griffith et al., 2013; Arce et al.,
32 2008). Part of the photochemical reaction of aerosol is direct photolysis, and in most cases, it
33 is driven by photochemically generated reactive intermediates, including triplet organics
34 ($^3\text{C}^*$) and reactive oxygen species (ROS) (Warneck, 1999; Herrmann et al., 2015; Rincon et
35 al., 2009; Lee et al., 2014).

36 Environmental $^3\text{C}^*$ is discovered earlier in natural water, and it has been proven to be
37 an important environmental photochemical reaction intermediate (Vione et al., 2014; Xu et
38 al., 2011). After absorbing solar radiation, this kind of substance will transition from the
39 ground state to the excited singlet state ($^1\text{C}^*$), and then rapidly transition to $^3\text{C}^*$ through the
40 intersystem. $^3\text{C}^*$ directly participates in the degradation of pollutants through the process of
41 energy transfer or electron transfer (Canonica et al., 2006; Parker et al., 2013). It can also
42 indirectly degrade pollutants by generating ROS ($^1\text{O}_2$, $\cdot\text{OH}$, O_2^- , $\cdot\text{HO}_2$, etc.) (Wenk et al.,
43 2011; Garg et al., 2011; Glover et al., 2013). Because the lifetime of $^1\text{C}^*$ is much shorter
44 than that of $^3\text{C}^*$, the concentration of $^3\text{C}^*$ in the environment is higher than that of $^1\text{C}^*$, and
45 the impact of $^3\text{C}^*$ on the environment is also greater (McNeill and Canonica, 2016).
46 Compared with the water environment, there are fewer studies on the triplet state in the
47 atmospheric environment. Kaur and Anastasio (2018) used probe technology to conduct
48 photochemical experiments on atmospheric fog droplets and confirmed the existence of $^3\text{C}^*$
49 in the atmospheric environment for the first time. Atmospheric chromophores are organic
50 substances with light-absorbing properties in the atmospheric environment, which are the
51 precursors of $^3\text{C}^*$ in aerosols. For example, imidazoles and pyrazines that are widely present
52 in the atmosphere can absorb light radiation to form $^3\text{C}^*$ (De-Haan et al., 2010; Hawkins et
53 al., 2018; Laskin et al., 2015). However, because the atmospheric chromophore is not a



54 single substance, but a complex organic mixture, this also increases the difficulty of
55 studying the generation mechanism of $^3\text{C}^*$ in actual atmospheric environment.

56 The environmental organic matter components are composed of complex compounds
57 with different polarities, which are expected to have different optical properties and
58 photochemical reactivity. A study explored the differences in the optical properties of
59 aerosol components with different polarities, and found that substances with lower polarities
60 have a stronger ability to absorb light radiation (Chen et al., 2017). Zhou et al. (2017)
61 separated three organic components of hydrophobic, hydrophilic and transitional hydrophilic
62 (medium hydrophilic) from polluted wastewater, and found a significant difference between
63 $^3\text{C}^*$ quantum yield coefficient (f_{TMP}) and singlet oxygen quantum yield ($\Phi_{1\text{O}_2}$) of each
64 component. $^3\text{C}^*$ converts O_2 into $^1\text{O}_2$ by means of energy transfer, so this process can be
65 used to determine the energy (E_T) distribution of $^3\text{C}^*$, which can be divided into low-energy
66 triplet states ($94 \text{ kJ mol}^{-1} \leq E_T < 250 \text{ kJ mol}^{-1}$) and high-energy triplet states ($E_T \geq 250 \text{ kJ}$
67 mol^{-1}) (Wilkinson et al., 1993; Kellogg and Simpson, 1965). Zhou et al. (2019) found that
68 high-energy $^3\text{C}^*$ accounted for more of the organic matter in water, while soil organic matter
69 had more low-energy $^3\text{C}^*$. Although there are some researches on the characteristics of $^3\text{C}^*$
70 formation in water environments, there are very few studies on the characteristics of $^3\text{C}^*$
71 types in the atmospheric environment. Due to the large differences in the sources and
72 chemical processes of organic matter in the atmospheric environment and the water
73 environment, the research conclusions on the $^3\text{C}^*$ generation characteristics of water bodies
74 may not be suitable for atmospheric aerosols.

75 Three-dimensional matrix excitation-emission (EEM) fluorescence spectroscopy
76 technology has been widely used to explore the composition of complex atmospheric
77 chromogenic organic matter (Chow et al., 2004). Chen et al. (2020) used EEM technology to
78 establish the relationship between the atmospheric chromophores and the types and sources
79 of substances, and expanded the application of EEM in the field of atmosphere. The cause of
80 the formation of $^3\text{C}^*$ is directly related to the chromophore. Therefore, the EEM method
81 may be conducive to revealing the chemical mechanism of aerosol triplet states. ROS is an
82 important driving factor in the aerosol aging process. Many studies have shown that $^3\text{C}^*$ can
83 drive the generation of ROS, and the generation characteristics and mechanisms of ROS by



84 different $^3\text{C}^*$ species are not yet known (Vione et al., 2006; Perri et al., 2009;). In summary,
85 in order to explore the formation characteristics and mechanism of triplet in atmospheric
86 aerosols, this study separated complex aerosol components according to polarity, and
87 obtained highly polar water soluble (HP-WSM), humic-like substances (HULIS) and
88 methanol soluble components (MSM, representative of water-insoluble organic matter). Use
89 chemical probe methods (including 2,4,6-trimethylphenol, furfuryl alcohol, and sorbic
90 alcohol) to characterize the generation characteristics of triplet states of different
91 components, including $^3\text{C}^*$ generation rate, quantum yield coefficient, and energy
92 distribution (Zhou et al., 2017; Zhou et al., 2019). This study also explored the
93 structure-activity relationship between the type of chromophore and the generation of $^3\text{C}^*$,
94 and demonstrated that the generation characteristics and mechanism of ROS depends on the
95 generation of $^3\text{C}^*$. Finally, this study confirmed the coupling effect of generation of $^3\text{C}^*$ and
96 ROS between different aerosol components (HP-WSM and HULIS).

97 **2. Experimental Section**

98 *2.1 Experimental materials*

99 A Mn^{2+} standard in ZnS and Cr^{3+} standard in MgO were purchased from Freiberg
100 Instruments Inc., Delfter, Germany. Glucose ($\geq 99.5\%$),
101 4-Hydroxy-2,2,6,6-tetramethylpiperidine (TEMP, $\geq 98\%$), L-Histidine ($\geq 99\%$), Phenol
102 ($\geq 99.5\%$), 2,4,6-Trimethylphenol (TMP, $\geq 98\%$), Furfuryl alcohol (FFA, 98%) and
103 5,5-Dimethyl-1-pyrroline-N-oxide (DMPO, 97%) were purchased from Aladdin Reagent
104 Company (Shanghai, China). Trans, trans-2,4-hexadien-1-ol (Sorbic alcohol, 97%) was
105 purchased from Sigma-Aldrich. 4-Nitroanisole (PNA, $\geq 99\%$) was purchased from
106 Thermo-Fisher. Pyridine (pyr, $\geq 99\%$) was purchased from Alfa-Aesar. C18 cartridge (500
107 mg/6 mL) was obtained from Agela Technologies. All chemicals were used as received.

108 *2.2 Sample collection and preparation*

109 Atmospheric $\text{PM}_{2.5}$ samples were collected in Xi'an city, China during the winter from
110 2019.12.21 to 2020.2.1 (Sample list is shown in **Table S1**). The sample collection point was
111 located on the top of Shaanxi University of Science and Technology, about 40 m above the
112 ground. A large flow sampler (XT-1025, Shanghai Xintuo, China) was used to collect



113 atmospheric particulate samples for 23.5 hours. The samples were collected on a 20×25 cm
114 quartz membrane pre-fired in a muffle furnace (2500 QAT-UP, Pallflex Products Co., US),
115 then the collected sample film is stored at -20 °C until use. The simulated combustion PM
116 samples include wheat straw, rice straw, poplar wood, bituminous coal, and lignitous coal.
117 The combustion samples were prepared through a self-built combustion-gas-gathering
118 device. The final particles were collected on a 47 mm diameter quartz membrane and stored
119 at -20 °C (**Text S1** for specific sample collection process and sample information).

120 The samples were continuously extracted with water and methanol ultrasonically for 20
121 minutes, and then filtered with a 0.45 μm PTFE filter to obtain the water-soluble extract
122 (WSM) and methanol extract (MSM), respectively. Note that the MSM here does not
123 actually contain water-soluble substances, thus it represents water-insoluble organic matter.
124 Then, according to the previous method, the solid-phase extraction technology was used
125 extract HP-WSM and HULIS from WSM (Lin et al., 2012; Chen et al., 2016), as shown in
126 **Text S1** (the separation process) and **Table S3** (sample component information). The final
127 concentration of all solution samples after dilution or concentration is controlled to 20 mg-C
128 L⁻¹, and stored at 4 °C for later use.

129 *2.3 Organic carbon (OC)/elemental carbon (EC) analysis*

130 Both original sample films and the prepared sample solution were used for the analysis
131 of organic carbon (OC)/elemental carbon (EC), and an OC/EC analyzer (Model 4, Sunset,
132 America) with IMPROVE_A temperature protocol was employed for quantitative analysis.
133 The original sample was directly measured with a 6 mm diameter filter membrane, and the
134 sample solution was subjected to OC/EC analysis by loading 100 μL of the solution onto the
135 17 mm diameter membrane, and then drying the solvent with nitrogen blowing. The
136 background filter sample was also processed and analyzed with the same procedure, and the
137 background signal interference was subtracted from the final result.

138 *2.4 Spectroscopic characterization*

139 The sample's absorption spectrum and fluorescence spectrum were obtained at the same
140 time in the "fluorescence + absorbance" mode by an Aqualog EEM analyzer (Horiba
141 Scientific, USA). In order to reduce the influence of internal filtration effect on EEM
142 measurement, the solution sample used for optical analysis was diluted before analysis. The



143 concentration of the solution during analysis is shown in **Table S4**. The instrument analysis
144 parameters are set as follows: excitation wavelength of 200-600 nm, emission wavelength of
145 250-800 nm, scanning interval of 2 nm and exposure time of 1 s. The background sample
146 was also analyzed in the same way and subtracted from the actual sample spectrum. A total
147 of 80 sample EEM spectra were analyzed through the PARAFAC model to identify the
148 chromophore types. According to the change trend of the minimum residual error of the 2-7
149 component PARAFAC model (**Fig.S10**) and the interpretability of the chromophores to the
150 actual meaning, finally 4 component model was used (**Fig.S11**).

151 *2.5 Photochemical experiments*

152 The sunlight is simulated by a 300 W xenon light source (PLS-SXE 300, Perfectlight,
153 China) and equipped with a VISREF filter. By combining a cold water circulation machine
154 and a magnetic stirrer, the experiment temperature is kept at about 25 °C. The sample
155 solution (120 µL) added with the chemical probe is placed in a special quartz reaction dish
156 (diameter 12 mm, thickness 2 mm) for light reaction, setting a series of light time and taking
157 samples at specific time intervals to perform HPLC analysis (**Text S2** for lighting
158 equipment and specific steps). A blank control was performed during the experiment, and 3
159 parallel experiments were set up in each group.

160 *2.6 ³OM* and ¹O₂ measurements and calculation of quantum yields*

161 In order to determine the production of ³C* and ¹O₂, TMP (20 µM) and FFA (20 µM)
162 were used as capture agents for ³C* and ¹O₂, respectively (Halladja et al., 2007; Dalrymple
163 et al., 2010). The two capture agent solutions have almost no loss under the direct irradiation
164 of the light source (**Fig.S7**). The probe concentration was analyzed by a high performance
165 liquid chromatography (LC-100, WuFeng, Shanghai) equipped with a C-18 column
166 (4.6×250 mm, 5 µm, Xuanmei). The acetonitrile (ACN) and ultrapure water were used as the
167 mobile phase, and the flow rate is 1 mL min⁻¹. The detection condition of TMP is mobile
168 phase ACN/Water = 50/50, and the detection wavelength is 210 nm; while for FFA, the
169 related terms are 30/70 and 219 nm, respectively.

170 The sorbic alcohol (1 mM) is used as a high-energy quencher to quench high-energy
171 ³C* (Zhou et al., 2017a), and combine the Φ_{1O_2} to quantify the energy distribution of
172 different ³C*. The reaction consumption rate of TMP (k_{TMP}) and the calculated quantum



173 yield coefficient (f_{TMP}) of the triplet state are used to reflect the $^3C^*$ generation rate and
174 ability. According to previous articles, the Φ_{1O_2} and f_{TMP} are calculated as follows
175 (Dalrymple et al., 2010; Bodhipaksha et al., 2015; Mostafa and Rosario-Ortiz, 2013).

$$176 \quad \phi_{^1O_2} = \frac{R_{^1O_2}}{R_a} \quad (1)$$

$$177 \quad f_{TMP} = \frac{k_{TMP}}{R_a} \quad (2)$$

$$178 \quad R_a = \sum \lambda \frac{I_\lambda(1-10^{-l a_\lambda})}{l} \quad (3)$$

$$179 \quad R_{^1O_2} = \frac{k_{FFA} \times k_d}{k'_{FFA}} \quad (4)$$

180 $R_{^1O_2}$ is the rate of generation of singlet oxygen ($M s^{-1}$); R_a is the light absorption rate
181 ($einsteins\ cm^{-3}\ s^{-1}$); k_{TMP} and k_{FFA} are the pseudo-first order reaction rate constant consumed
182 by TMP and FFA (s^{-1}); I_λ is the photon flux at λ wavelength ($Einsteins\ cm^{-2}\ s^{-1}\ nm^{-1}$), as
183 shown in S3 for specific calculation; a_λ represents the unit absorbance of the solution at the
184 wavelength of λ (cm^{-1}); l is the thickness of the solution (0.1 cm); k'_{FFA} is the rate constant of
185 1O_2 and FFA reaction ($1.2 \times 10^8\ M^{-1}\ s^{-1}$); k_d is the rate constant of water quenching 1O_2
186 ($2.4 \times 10^5\ s^{-1}$). The wavelength range of light radiation for calculation in this study is from
187 320 nm to 600 nm.

188 2.7 Reactive oxygen detection (1O_2 , $\cdot OH$)

189 In order to explore the characteristics and mechanism of ROS generated by $^3C^*$, the
190 combination of free radical capture technology and electron paramagnetic resonance
191 spectrometer (EPR, MS5000, Freiberg) was used to detect two ROS including 1O_2 and $\cdot OH$.
192 The concentration of the experimental sample solution is controlled to 100 mg-C/L, using
193 TEMP as 1O_2 capture agent, DMPO as a capture agent for $\cdot OH$. The concentration of the
194 capture agent after mixing is 10 mM. When capturing 1O_2 , the illumination time is 60 min,
195 and when capturing $\cdot OH$, the illumination time is 10 min. At the same time, the 0 point of
196 illumination is set up and the non-illumination control group is carried out. In this study,
197 Sorbic alcohol and histidine (both 50 mM after mixing) were used as quenchers of $^3C^*$ and
198 1O_2 , respectively, to explore the generation mechanism of 1O_2 and $\cdot OH$.



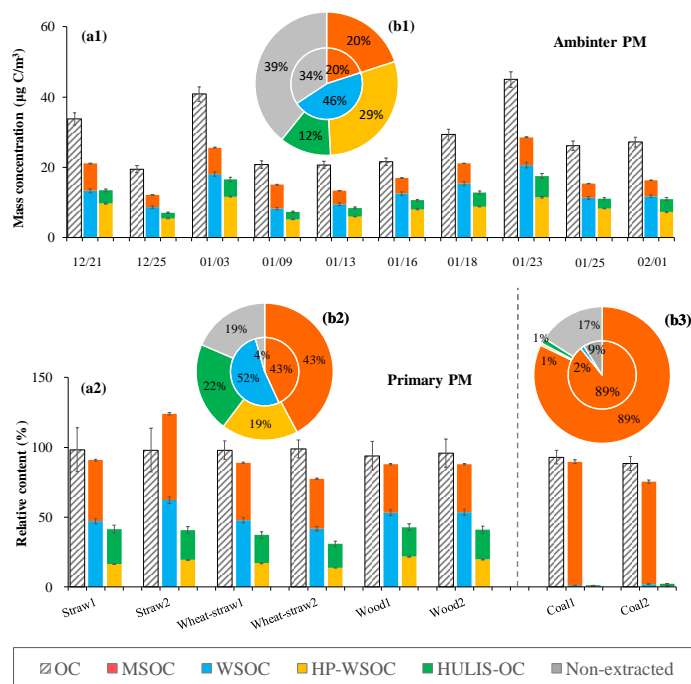
199 The EPR detection parameters for the $^1\text{O}_2$ and $\cdot\text{OH}$ are the same: magnetic field
200 strength, 330-342 mT; detection time, 180 s; modulation amplitude, 0.2 mT; number of
201 detections, 1; and microwave intensity, 8.0 mW.

202 3. Results and Discussion

203 3.1 Carbon composition of different polar aerosols

204 Organic carbon is the main precursor for the formation of triplet states in aerosols, and
205 the composition of organic carbon is different between components of different polarities.
206 The results of carbon analysis are shown in **Fig.1**. The OC values of environmental $\text{PM}_{2.5}$
207 samples in winter is in the range of (19.51-44.97) $\mu\text{g-C}/\text{m}^3$, of which the WSOC component
208 accounts for about half of the OC, mainly the HP-WSOC component (25%-37%). The
209 average concentration of HULIS-C in this study is (3.39 \pm 1.38) $\mu\text{g-C}/\text{m}^3$, and the
210 concentration is between clean areas and areas severely affected by biomass burning
211 (Nguyen et al., 2014; Wang et al., 2017). Affected by source and environmental conditions,
212 the composition and concentration of HULIS vary greatly, generally accounting for 8-74%
213 of WSOC (Feczko et al., 2007). In this study, HULIS accounts for 26% of WSOC, which is
214 at a moderate level. The water-insoluble organic matter MSOC accounts for about 20% of
215 the OC of the $\text{PM}_{2.5}$ sample. This result indicates that there is a considerable part of the
216 water-insoluble organic carbon in the particulate sample. It is worth noting that the content
217 of water-insoluble organic carbon may be higher, considering that 34% of the organic carbon
218 has not been extracted (**Fig.1 b1**).

219 The MSOC in the primary combustion sample is generally higher than that in the actual
220 PM sample. In the biomass sample, 35%-64% of the components in OC are MSOC, while
221 the HP-WSOC and HULIS in OC are basically the same at 19%-22%. There are very few
222 water-soluble components in coal combustion samples. Nearly 90% of OC components are
223 MSOC. MSOC should be macromolecular organics, especially tar-like substances formed
224 during coal combustion. Whether it is coal combustion or biomass combustion, the content
225 of coking organic carbon in thermo-optical analysis is significantly higher than that of actual
226 atmospheric samples (**Fig.S11**).



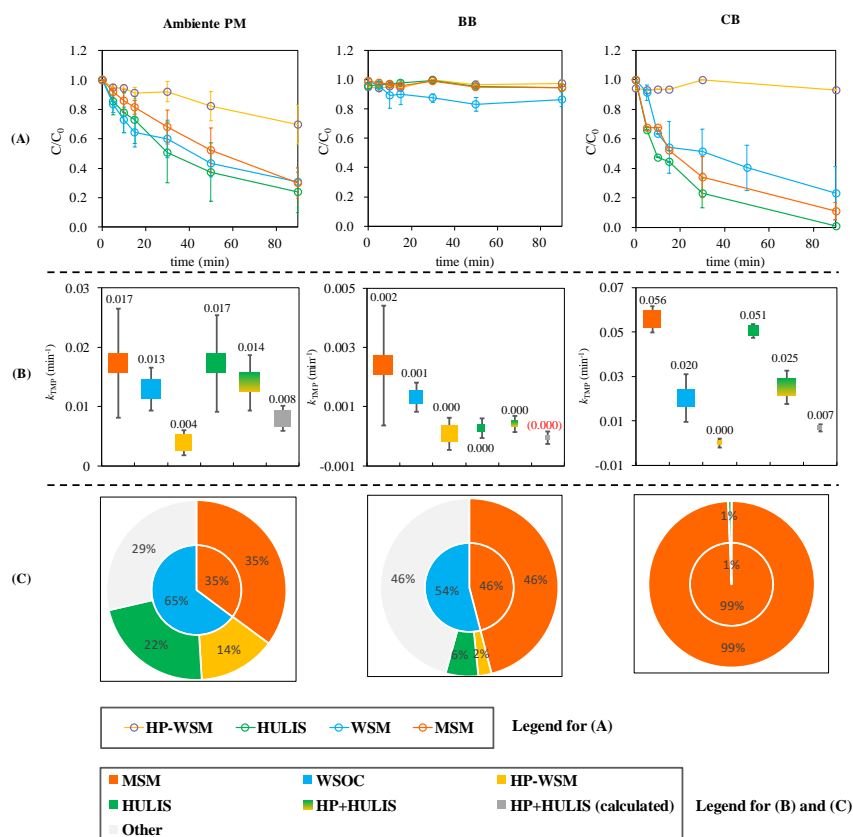
227
 228 **Fig.1** The content of organic carbon in different polar components of atmospheric aerosol. (a1) and (a2) represent
 229 the organic carbon content in actual atmospheric aerosol samples and primary combustion source samples,
 230 respectively; (b1), (b2) and (b3) represent the average relative content of organic carbon of different polarity
 231 groups in actual atmospheric aerosols, biomass combustion aerosols, and coal combustion aerosol samples,
 232 respectively. The error bars represent the relative standard deviations obtained for a set of three parallel samples.

233 3.2 Triplet generation ability of aerosols with different polarities

234 The characteristics of triplet generation of components with different polar in aerosol
 235 are different. The pseudo-first order reaction kinetic rate constant k_{TMP} of the triplet state and
 236 TMP is usually used to characterize the rate of formation of $^3C^*$, and the triplet quantum
 237 yield coefficient (f_{TMP}) reflects the ability of the triplet state to be generated (Ervens et al.,
 238 2011; Zhou et L., 2017b). **Fig.2(A)(B)** shows the average attenuation curve and average k_{TMP}
 239 value of the reaction of different components with TMP under illumination conditions. The
 240 k_{TMP} of each sample is shown in **Table S5**, and the average TMP attenuation fitting curve is
 241 shown in **Fig.S16**. The results show that the aerosol components with different polarities
 242 exhibit different generation rates of $^3C^*$. The k_{TMP} of the components with different polar in
 243 the actual PM sample is between 0.004-0.017 min^{-1} , and the k_{TMP} of the HP-WSM
 244 component is the lowest. In the biomass combustion samples, all polar components did not



245 significantly consume TMP; although HP-WSM in coal samples did not consume TMP,
 246 HULIS and MSM components were consumed significantly, with an average k_{TMP} value of
 247 0.054 min^{-1} . **Table 1** lists the f_{TMP} values of the components with different polar of each
 248 sample. The f_{TMP} value of the HULIS component of the atmospheric PM sample is between
 249 $(35\text{-}180) \text{ M}^{-1}$, which is higher than the HP-WSM component $(5\text{-}90) \text{ M}^{-1}$ and MSM
 250 component $(14\text{-}70) \text{ M}^{-1}$, indicating that the HULIS has the strongest ability to generate
 251 triplet states.



252
 253 **Fig.2** (A) The average attenuation curve of TMP consumption by aerosol samples of different components; (B)
 254 the triplet generation rate constant; (C) the average relative contribution to the overall triplet generation rate of
 255 the aerosol. The error bar represents one standard deviation of the experimental values of all samples.
 256

257 **Table 1.** Formation rate constant (k_{TMP}), quantum yield coefficient (f_{TMP}) and singlet oxygen quantum yield (Φ_{1O_2})
 258 of $^3C^*$ in the different polar components of each sample.



Sample number	HP-WSM			HULIS			MSM		
	k_{TMP} (min^{-1})	f_{TMP} (M^{-1})	Φ_{102} (%)	k_{TMP} (min^{-1})	f_{TMP} (M^{-1})	Φ_{102} (%)	k_{TMP} (min^{-1})	f_{TMP} (M^{-1})	Φ_{102} (%)
2019/12/21	0.005	51.9	6.63	0.020	89.9	4.57	0.007	0.007	7.43
2019/12/25	0.006	56.9	5.78	0.025	158.3	7.60	0.014	0.014	11.91
2020/1/3	0.005	32.0	2.78	0.013	35.6	3.43	0.018	0.018	5.40
2020/1/9	0.006	71.5	9.23	0.021	82.7	6.45	0.033	0.033	14.37
2020/1/13	0.001	5.2	5.15	0.037	167.6	4.68	0.032	0.032	7.08
2020/1/16	0.002	24.5	6.53	0.025	140.9	6.82	0.022	0.022	8.44
2020/1/18	0.006	60.4	4.39	0.013	75.8	4.60	0.007	0.007	6.76
2020/1/23	0.008	91.8	11.92	0.010	43.7	3.57	0.015	0.015	7.03
2020/1/25	0.002	39.5	3.59	0.018	102.9	5.88	0.011	0.011	6.42
2020/2/1	0.002	22.6	6.79	0.009	40.3	5.37	0.028	0.028	8.29
Wheat-straw1	0.000	0.00	3.31	0.000	4.6	1.70	0.000	0.000	4.89
Wheat-straw2	0.000	0.00	2.18	0.000	5.4	1.49	0.000	0.000	6.34
Rice-straw1	0.000	0.00	5.97	0.001	0.0	3.09	0.005	0.005	7.10
Rice-straw2	0.000	0.00	3.66	0.001	0.0	4.66	0.004	0.004	4.75
Wood1	0.001	16.3	2.45	0.000	2.0	1.95	0.001	0.001	8.80
Wood2	0.001	10.3	2.40	0.000	1.5	3.02	0.003	0.003	7.61
coal 1	0.000	3.2	1.49	0.047	270.8	38.85	0.050	0.050	25.92
coal 2	0.000	0.0	1.47	0.054	417.2	34.18	0.062	0.062	48.77
Ambient samples	0.004	45.64	6.28	0.02	93.78	5.30	0.019	0.019	8.31
(mean±SD)	±0.002	±24.54	±2.56	±0.01	±45.97	±1.31	±0.009	±0.009	±2.61
Primary samples	0-0.001	0-16.3	1.5-6.0	0-0.05	0-417	1.5-38.9	0-0.06	0-0.06	4.8-48.8
(range(mean))	(0.0002)	(3.7)	(2.9)	(0.01)	(88)	(11.1)	(0.02)	(0.02)	(14.3)

259

260 The ability to generate triplet states should be related to the chemical composition of
 261 components with different polar. The HP-WSM component has a low ability to generate
 262 triplet states, which may be related to the fact that the component contains more carboxylic
 263 acids, alcohols and sugars and other small molecules with high polarity substances, which
 264 cannot form triplet states (Bodhipaksha et al., 2015; Zhang et al., 2014). At the same time,
 265 these small molecules are also easy to quench the triplet state. The combustion sample
 266 showed different results from the actual sample (Fig.S13), although the particulate matter of
 267 biomass combustion has a strong light absorption capacity (Lin et al., 2017). However, in
 268 this study, the f_{TMP} all the components with different polar in the biomass combustion
 269 samples is small, and the f_{TMP} of some components is even close to zero, indicating that the
 270 light-absorbing organic matter of the biomass combustion source is not very capable of
 271 generating $^3\text{C}^*$. The HP-WSM component of the coal combustion sample has a very low



272 ability to generate $^3\text{C}^*$, while the HULIS and MSM components have higher f_{TMP} values.
273 Especially the HULIS component, the content of the coal sample is very small, but it has the
274 highest generation ability of $^3\text{C}^*$ among all samples. From the results, the k_{TMP} and f_{TMP}
275 trends of different polar components are basically the same. The difference is that the total
276 consumption rate of TMP k_{TMP} for the MSM component is larger than that of the HULIS
277 component, but f_{TMP} is smaller than that of HULIS, although MSM has stronger light
278 absorption capacity than HULIS (Fig.S13).

279 Colored organics and metal ions can change their original light-absorbing
280 characteristics through chelation (Wan et al., 2019; Kikuchi et al., 2017). Therefore, it is
281 expected that there will be an interaction between the HP-WSM and HULIS components
282 containing metal ions to affect the generation of triplet states. The difference between the
283 reaction k_{TMP} value and the theoretical calculation value (the sum of the half of the respective
284 k_{TMP} values) of the two components under the equal concentration and equal volume mixing
285 is comparatively studied. The result is shown in Fig.2(B). The mixed sample consumes more
286 TMP, and the actual environmental sample increases the generation rate of triplet by 1.8
287 times. This experiment demonstrates the existence of the interaction effect of photochemical
288 reactions between different components.

289 According to the proportion of different polar components in the sample OC, the
290 average contribution of different components to the total generation of $^3\text{C}^*$ is calculated, as
291 shown in Fig.2(C). The contribution of the WSM component to the total k_{TMP} of PM is 65%.
292 After separating it into single components of HP-WSM and HULIS, they contribute 14%
293 and 22% of the triplet generation rate, respectively. Obviously, 29% has not been explained.
294 As the result described in the previous paragraph, this is partly due to the interaction
295 between HP-WSM and HULIS that promotes the generation of triplet states. The main
296 possible mechanism is that the metal elements in the system chelate with certain functional
297 groups of HULIS, forming more chromophores that can generate $^3\text{C}^*$, or reducing the
298 energy required for the original electronic transition, thereby increasing and enhancing the
299 generation rate of $^3\text{C}^*$. Note that this study cannot completely rule out the possibility that
300 inorganic salt components consume TMP, such as $\cdot\text{SO}_4^-$ (Fang et al., 2013). But the
301 experiments we added show that nitrate and sulfate do not consume TMP significantly under



302 light conditions. There is no relevant research on the formation of $^3\text{C}^*$ from water-insoluble
303 aerosol components. This study found that the contribution of MSM components to the
304 formation of $^3\text{C}^*$ can reach 35% in actual samples, and even more than 99% in coal-burning
305 samples, indicating that the contribution of water-insoluble organic carbon to aerosol triplet
306 photochemistry is significant.

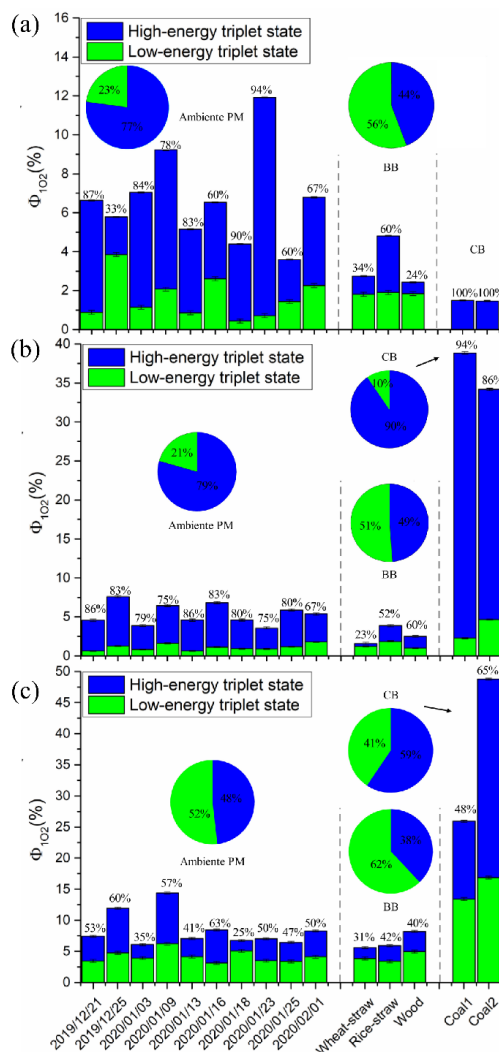
307 *3.3 The energy distribution of triplet states of aerosols with different polarities*

308 Different organic matter in the aerosol can form the triplet state with different energy.
309 $^3\text{C}^*$ and O_2 in the environment can generate $^1\text{O}_2$ through energy transfer. Therefore, this
310 study uses the $\Phi_{1\text{O}_2}$ to explore the energy distribution characteristics of $^3\text{C}^*$. The experiment
311 uses FFA as the quencher of $^1\text{O}_2$, combined with sorbic alcohol as the quencher of
312 high-energy $^3\text{C}^*$. After adding sorbic alcohol, the calculated $\Phi_{1\text{O}_2}$ is considered to be the
313 contribution of low-energy $^3\text{C}^*$, which is different from the contribution of $\Phi_{1\text{O}_2}$ when no
314 quencher is added. The difference is considered to be the contribution of high-energy $^3\text{C}^*$,
315 and the difference in reaction ability of $^3\text{C}^*$ with different energies is obtained (Zhou et al.,
316 2019).

317 **Fig.3** shows the $^3\text{C}^*$ energy distribution of different polar components. According to the
318 results of $\Phi_{1\text{O}_2}$, the high-energy triplet states in environmental samples accounted for
319 3.59%-11.92%, 3.57%-7.6%, 6.08%-14.37% in HP-WSM, HULIS and MSM components,
320 respectively. Among water-soluble components, high-energy $^3\text{C}^*$ accounts for an average of
321 close to 80%, which is significantly higher than that of natural water (33%) and sewage
322 (65%) samples (Zhou et al., 2019). This result is as stated in the introduction: atmospheric
323 samples and water environment samples are quite different in terms of source and
324 composition, and the conclusions of previous studies on the triplet state of water
325 environment samples are not necessarily suitable for the atmospheric PM samples. In MSM,
326 the proportion of low-energy $^3\text{C}^*$ increased, reaching an average of 52%, which should be
327 related to the organic matter containing polycyclic conjugated electrons in the
328 water-insoluble component. Especially for biomass combustion samples, the low-energy $^3\text{C}^*$
329 of different polar components has reached more than 50%. The $\Phi_{1\text{O}_2}$ yield of the HP-WSM
330 component of coal samples is very low, while the $\Phi_{1\text{O}_2}$ yield of the HULIS and MSM
331 components is the highest among all samples, reaching an average of 36% and 37%,



332 respectively. Unlike other samples, most of the coal combustion samples are contributed by
 333 high-energy $^3\text{C}^*$.



334
 335 **Fig.3** $^3\text{C}^*$ energy distribution in different polar sample components. (A) Sample HP-WSM component (b)
 336 Sample HULIS component (c) Sample MSM component; pie chart represents the average proportion of high and
 337 low energy triplet states; the percentage represents the $\Phi_{1\text{O}_2}$ caused by the high energy triplet in the total $\Phi_{1\text{O}_2}$.
 338 The error bar represents one standard deviation of the experimental values of all samples.

339 High-energy $^3\text{C}^*$ has higher energy and is more reactive than low-energy $^3\text{C}^*$, and it is
 340 easier to effectively collide with other substances at the molecular level. $^3\text{C}^*$ can not only
 341 generate $^1\text{O}_2$, but also react with other organic substances to generate active intermediates,



342 while the energy of low-energy $^3\text{C}^*$ only allows the conversion of O_2 to $^1\text{O}_2$. According to
343 previous reports, among the substances that can form triplet states, aromatic ketones and
344 other carbonyl compounds are likely to contribute a lot to high-energy $^3\text{C}^*$, while polycyclic
345 aromatic hydrocarbons and quinones are unlikely to be the main sources of high-energy $^3\text{C}^*$
346 (McNeill and Canonica, 2016; Kuznetsova and Kaliya, 1992).

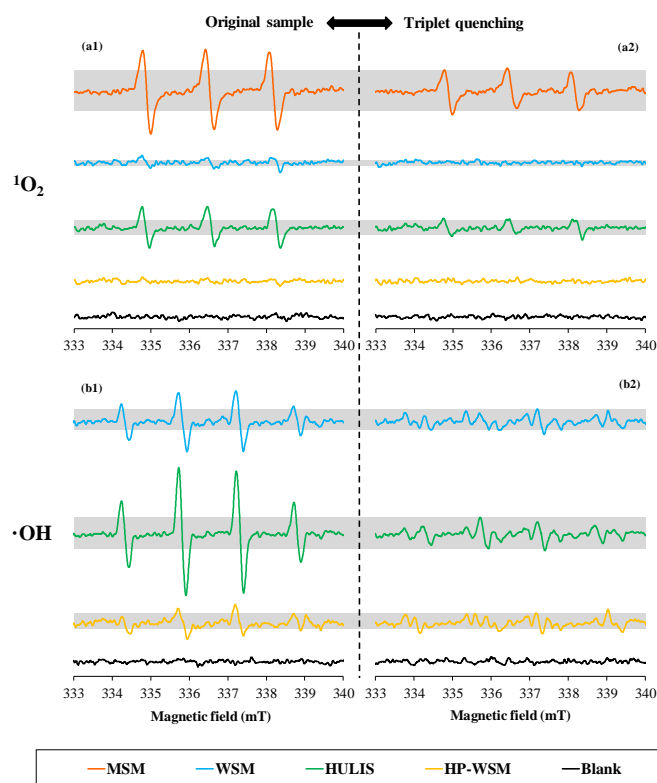
347 3.4. ROS generation caused by $^3\text{C}^*$

348 $^3\text{C}^*$ can induce the generation of ROS, and the ability of triplet substances with
349 different polar components to generate ROS is different. This study uses EPR technology to
350 identify $^1\text{O}_2$ and $\cdot\text{OH}$ produced by different polar components (Blough and Zepp, 1995; Shi
351 et al., 2003; Cote et al., 2018). The triplet quencher is used to quantitatively judge the
352 contribution of the triplet process to the production of $^1\text{O}_2$. **Fig.4** shows the EPR spectra of
353 the photochemical reaction products of different components of atmospheric aerosols. It can
354 be found that the output of $^1\text{O}_2$ is $\text{MSM} > \text{HULIS} > \text{WSM} > \text{HP-WSM}$, and HP-WSM is almost
355 not detected. When the triplet quencher sorbic alcohol is added, the signal of $^1\text{O}_2$ decreases
356 by 44%, 43% and 56%, respectively, indicating $^3\text{C}^*$ is an important precursor for the
357 photochemical generation of $^1\text{O}_2$. This study further explored the characteristics of $\cdot\text{OH}$
358 generated by water-soluble components under light conditions. The results showed that the
359 EPR spectrum of the HULIS sample showed a strong $\cdot\text{OH}$ signal, and the HP-WSM
360 component also showed a certain signal. When the triplet quencher is added, the $\cdot\text{OH}$ signal
361 is significantly reduced, and the signals of each component are reduced by 59% (WSM), 75%
362 (HULIS) and 26% (HP-WSM), respectively. This result demonstrates that $^3\text{C}^*$ is also an
363 important precursor of $\cdot\text{OH}$. Because the methanol in the MSM sample quenched the $\cdot\text{OH}$,
364 MSM was not tested.

365 The experimental results show that $^3\text{C}^*$ plays a leading role in the production of $^1\text{O}_2$
366 and $\cdot\text{OH}$. When there are triplet substances in the environment, the triplet state transfers
367 energy to O_2 molecules to form $^1\text{O}_2$, and $^1\text{O}_2$ may undergo complex reactions to
368 generate $\cdot\text{OH}$ (Chen et al., 2020; Mu et al., 2020). In this study, the generation mechanism
369 of $\cdot\text{OH}$ varies with the components. For the HP-WSM component, it contains a lot of
370 inorganic substances such as metal ions, so it can form a Fenton-like system. For example,
371 in the presence of H_2O_2 , Fe^{2+} , Cu^{2+} will convert it into $\cdot\text{OH}$ (Shi et al., 2018). For the organic



372 component HULIS, the production of $\cdot\text{OH}$ is mainly the effect of organic matter. HULIS
373 organic matter is excited to produce $^3\text{C}^*$ and then generates $^1\text{O}_2$, and finally, $^1\text{O}_2$ leads to the
374 formation of $\cdot\text{OH}$ (Fig.S17). It should be noted that the possibility of triplet substances
375 directly producing $\cdot\text{OH}$ cannot be ruled out. This potential mechanism is worth exploring in
376 the future.



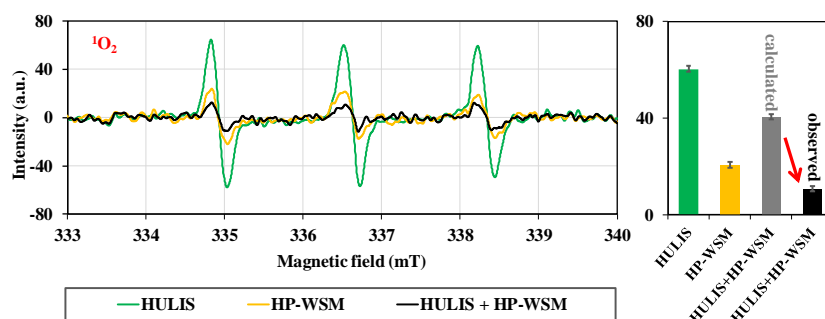
377

378 **Fig.4** Different aerosol components produce (a1) singlet oxygen and (b1) hydroxyl radicals under illumination
379 conditions, and the addition of triplet quencher (sorbic alcohol) produces (a2) singlet oxygen and (b2) hydroxyl
380 radicals feature.

381 The interaction between different components also affects the photochemical
382 generation of ROS. **Fig.5** is the EPR signal spectrum of $^1\text{O}_2$ produced by mixing HP-WSM
383 and HULIS (All actual atmospheric samples mixed in equal amounts). The above results
384 have shown that the appearance of $^3\text{C}^*$ is promoted due to the interaction effect (**Fig.2b**).
385 Furthermore, the theoretical production of $^1\text{O}_2$ should be more than half of the sum of the
386 production of the two separate components, but **Fig.5** shows a significant inhibitory effect



387 that the production of $^1\text{O}_2$ is greatly reduced after mixing. This is not the expected result, and
388 there may be a different mechanism than expected: the complex formed by metal ions and
389 HULIS increases the total rate of $^3\text{C}^*$ formation, but it is not sensitive to the reaction
390 pathway of energy transfer. It is easier to undergo chemical reactions through electron
391 transfer, resulting in a decrease in the production of $^1\text{O}_2$ (McNeill and Canonica, 2016; Wan
392 et al., 2019; Kikuchi et al., 2017; Blough and Zepp, 1995).

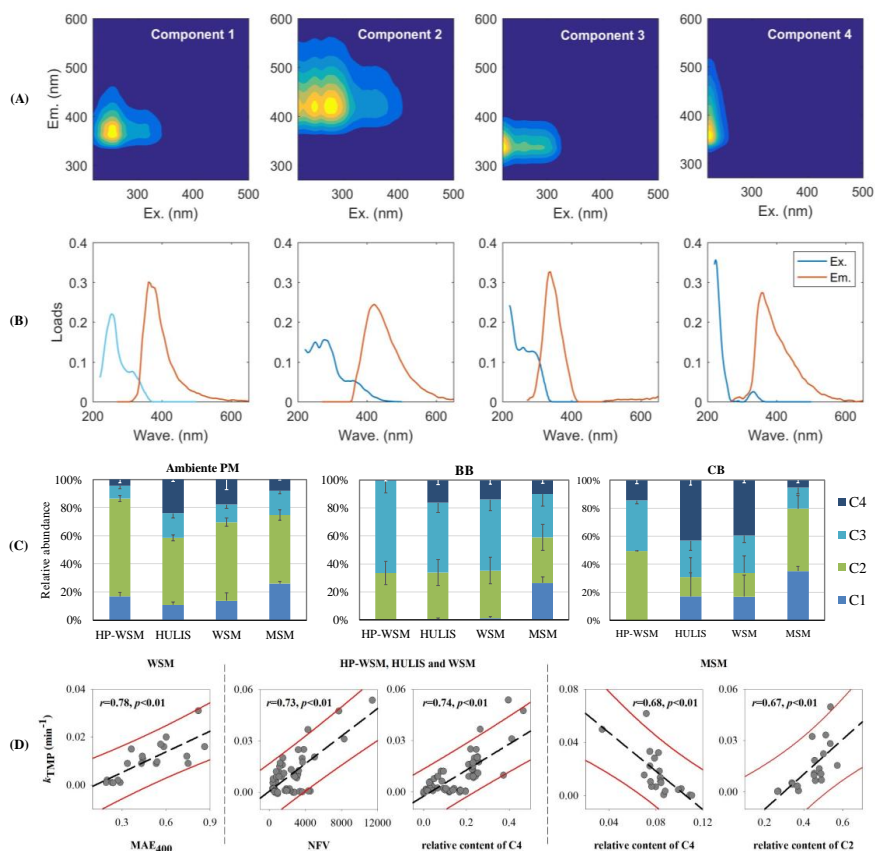


393

394 **Fig.5** The mixed sample of HP-WSM and HULIS produces the EPR spectrum of $^1\text{O}_2$. The Figure shows the
395 result of the actual atmospheric sample extract sample after 60 minutes of illumination (the signal of the control
396 sample placed in the dark for 60 minutes has been subtracted). Reaction conditions: the mixed sample is
397 illuminated for 60 minutes.

398 3.5 The relationship between optical properties and $^3\text{C}^*$ generation

399 The generation of $^3\text{C}^*$ should be related to optical characteristics. The study gives the
400 absorbance and fluorescence parameters of each sample (As shown in **Table S5** and
401 **Fig.S15**). The results are similar to previous research reports. The non-polar components
402 have greater light absorption and fluorescence capabilities (Chen et al., 2017). Fluorescent
403 substances are the direct precursors for the formation of triplet states, which can be found by
404 the correlation between k_{TMP} and standard fluorescence volume (NFV) (**Fig.6**). The sample
405 with stronger fluorescence intensity produces $^3\text{C}^*$ at a higher rate. However, the fluorescent
406 substances are still complex and diverse. In order to explore which chromophore is the most
407 important triplet precursor, this study further explored the structure-activity relationship
408 between chromophore types and triplet generation.



409
410 **Fig.6** (a and b) The four chromophores identified and (c) the relative content in different samples and
411 components, and (d) the correlation between MAE_{400} , NFV, C4, C2 and k_{TMP} . The error bar in the Figure (c)
412 represents one standard deviation of all sample values; the black dashed line in the Figure (d) represents the
413 linear fitting result, and the red line represents the 95% prediction interval.

414 This study identified four different chromophores C1-C4 from the samples using the
415 EEM-PARAFAC method (**Fig.6**). Refer to previous reports that the profile of C1 is similar
416 to low-oxidation HULIS; C2 may be highly oxidized HULIS; C3 may be Tryptophan-like
417 OM; and C4 is similar to Tyrosine-like chromophore (Chen et al., 2021; Murphy et al., 2013;
418 Chen et al., 2016; Chen et al., 2019). **Fig.6**(b) shows the proportions of the four
419 chromophores in each component sample. The results show that the HULIS chromophore is
420 the main light-absorbing organic matter in the water-soluble components, and compared
421 with the combustion source sample, the actual aerosol contains more highly oxidized HULIS,



422 which may be different from the actual atmospheric aerosol in the actual atmospheric
423 environment. That is probably related to the atmospheric aging process. Tryptophan-like OM
424 accounts for a relatively large proportion of combustion source samples. Recent studies have
425 shown that similar chromophores may be phenolic substances (Chen et al., 2020), and
426 biomass and coal combustion can emit a large number of phenolic substances.

427 The correlation between the different polar components k_{TMP} and the composition of
428 chromophores were explored (Fig.6), and the correlation analysis between the quantum
429 yield f_{TMP} and different types of chromophores was also carried out, and the results were
430 similar to k_{TMP} Trend (Fig.S18). As shown in Fig.6, there is a positive correlation between
431 Tyrosine-like chromophores and k_{TMP} in water-soluble components, and a negative
432 correlation in water-insoluble components. On the contrary, among the water-insoluble
433 components, the highly oxidized HULIS chromophore has an important role in promoting
434 the $^3C^*$ production of the MSM component. Analogous to the related literature, this
435 chromophore is likely to be a class of aromatic hydrocarbons containing oxygen functional
436 groups (Zhou et al., 2017; Zhou et al., 2019), structural substances are beneficial to the
437 formation of triplet states (Fig.S19). It is found that both high-energy $^3C^*$ and low-energy
438 $^3C^*$ in the MSM component have a positive correlation with the low-oxidation HULIS
439 chromophore, while the water-soluble component has no obvious correlation with the four
440 chromophores, which may be due to that the triplet reaction of the chromophore is different
441 (electron transfer and energy transfer). The low-oxidation HULIS chromophore may
442 generate $^3C^*$ through energy transfer, while the high-oxidation HULIS chromophore is
443 through electron transfer. For the MSM component, the triplet reaction is mainly an energy
444 transfer type reaction. HULIS chromophores play an important role. For example, quinones
445 may form low-energy $^3C^*$, while aromatic ketones may contribute a lot to high-energy $^3C^*$
446 (McNeill and Canonica, 2016; Kuznetsova and Kaliya, 1992).

447 4. Environmental Implications

448 $^3C^*$ plays an important role in the formation and aging process of atmospheric aerosols.
449 On the one hand, $^3C^*$ itself is reactive and can directly react with other substances. On the
450 other hand, it can produce 1O_2 and $\cdot OH$ and other ROS substances, which indirectly



451 participate in the generation reaction of aerosol components. In this study, the complex
452 aerosol samples were divided into three components including HP-WSM, HULIS and MSM
453 according to polarity, and their optical and photochemical reaction characteristics were
454 discussed respectively. This study demonstrated that the ${}^3\text{C}^*$ generation characteristics of
455 different polar components in atmospheric particulate matter samples are different. The
456 low-polarity components have strong light absorption and fluorescence capabilities, and the
457 related k_{TMP} is also enhanced. Among them, the f_{TMP} of HULIS component is the strongest,
458 and the water-insoluble components contribute the most to the total ${}^3\text{C}^*$ generation. This
459 result means that the triplet photochemical reaction can enhance the heterogeneous aerosol
460 reaction. This study also demonstrated the complexity of the triplet types. The distribution of
461 different energy ${}^3\text{C}^*$ was indirectly investigated through $\Phi_{1\text{O}_2}$, and it was found that
462 high-energy ${}^3\text{C}^*$ is the main form of ${}^3\text{C}^*$ (80%) in the water-soluble components of aerosols.
463 The high-energy ${}^3\text{C}^*$ in the insoluble components accounts for 50% or even lower, and the
464 high-energy ${}^3\text{C}^*$ plays a more important role in the production of ${}^1\text{O}_2$ and $\cdot\text{OH}$. Note that
465 there are only few reports on the energy distribution of the triplet state of aerosols so far.
466 Using EEM-PARAFAC technology, the structure-activity relationship between chromophore
467 composition and triplet generation was explored, and it was found that in the water-soluble
468 components, the chromophore previously defined as Tyrosine-like chromophore has a strong
469 correlation with the triplet generation rate. The highly oxidized HULIS chromophore in the
470 water-insoluble component plays an important role in promoting the production of ${}^3\text{C}^*$. This
471 result means that the photochemical reaction characteristics of different chromophores are
472 different, which ultimately determine the overall photochemical reactivity of the sample.
473 Finally, the obtained results also proved that there is a coupling effect of photochemical
474 reaction between HP-WSM and HULIS. After mixing, the production of ${}^3\text{C}^*$ is enhanced,
475 but the production of singlet oxygen is weakened. This result means that the photochemistry
476 of aerosol ${}^3\text{C}^*$ is not reacts alone, but is affected by the aerosol composition. In particular,
477 metal ions are most likely to undergo chelation with chromophore substances, thereby
478 changing the original optical and photochemical reaction characteristics of the chromophore.
479 It is necessary to explore this aspect in the future.



480 **Data availability.** All data that support the findings of this study are available in this article
481 and its Supplement or from the corresponding author on request.

482 **Supporting information.** Additional information as noted in the text, including three texts
483 (sampling information, photochemical experiment and calculation of quantum yields), six
484 tables and nineteen **Figures** (experimental detailed data).

485 **Author contributions.** DJ and QC designed the experiments and data analysis. DJ, JW and
486 LX performed sample collection. HL performed the EPR analysis. Other experiments are
487 performed by DJ, YQ, XF and TC polished the article. QC prepared the paper with the
488 contributions from all co-authors.

489 **Competing interests.** The authors declare that they have no conflict of interest.

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493 **References**

- 494 Arce, R., Pino, E. F., Valle, C., Agreda, J.: Photophysics and photochemistry of 1-nitropyrene, *J Phys Chem A.*,
495 112, 10294-10304, <https://doi.org/10294-10304>, 10.1021/jp803051x, 2008.
- 496 Atkinson, R.: Atmospheric chemistry of VOCs and NO_x, *Atmos. Environ.*, 34, 2063-2101,
497 [https://doi.org/10.1016/S1352-2310\(99\)00460-4](https://doi.org/10.1016/S1352-2310(99)00460-4), 2000.
- 498 Blando, J. D., Turpin, B. J.: Secondary organic aerosol formation in cloud and fog droplets: a literature
499 evaluation of plausibility, *Atmos. Environ.*, 34, 1623-1632, [https://doi.org/10.1016/S1352-2310\(99\)00392-1](https://doi.org/10.1016/S1352-2310(99)00392-1),
500 2000.
- 501 Blough, N. V., Zepp, R. G.: Reactive oxygen species in natural waters, *Active Oxygen in Chemistry.*, 2, 280-333,
502 https://doi.org/10.1007/978-94-007-0874-7_8, 1995.
- 503 Bodhipaksha, L. C., Sharpless, C. M., Chin, Y. P.: Triplet photochemistry of effluent and natural organic matter in
504 whole water and isolates from effluentreceiving rivers, *Environ. Sci. Technol.*, 49, 3453-3463,
505 <https://doi.org/10.1021/es505081w>, 2015.
- 506 Canonica, S., Hellrung, B., Müller, P., Wirz, J.: Aqueous oxidation of phenylurea herbicides by triplet aromatic
507 ketones, *Environ. Sci. Technol.*, 40, 6636-6641, <https://doi.org/10.1021/es0611238>, 2006.
- 508 Chen, Q. C., Fumikazu, I., Hayato, H., Daichi, A., Michihiro, M.: Chemical structural characteristics of HULIS
509 and other fractionated organic matter in urban aerosols: results from mass spectral and FT-IR analysis,
510 *Environ. Sci. Technol.*, 50, 1721-1730, <https://doi.org/10.1021/acs.est.5b05277>, 2016.



- 511 Chen, Q. C., Ikemori, F., Nakamura, Y., Vodicka, P., Kawamura K., Mochida M.: Structural and light-absorption
512 characteristics of complex water-insoluble organic mixtures in urban submicrometer aerosols, *Environ. Sci.*
513 *Technol.*, 51, 8293-8303, <https://doi.org/10.1021/acs.est.7b01630>, 2017.
- 514 Chen, Q. C., Li, J. W., Hua, X. Y., Jiang, X. T., Mu, Z.: Identification of species and sources of atmospheric
515 chromophores by fluorescence excitation-emission matrix with parallel factor analysis, *Sci. Total. Environ.*,
516 718, 137322-137332, <https://doi.org/10.1016/j.scitotenv.2020.137322>, 2020.
- 517 Chen, Q. C., Miyazaki, Y., Kawamura, K.: Characterization of chromophoric water-soluble organic matter in
518 urban, forest, and marine aerosols by HR-ToF-AMS analysis and excitation-emission matrix spectroscopy,
519 *Environ. Sci. Technol.*, 50, 10351-10360, <https://doi.org/10.1021/acs.est.6b01643>, 2016.
- 520 Chen, Q. C., Mu, Z., Song, W.: Size-resolved characterization of the chromophores in atmospheric particulate
521 matter from a typical coal-burning city in China, *J. Geophys. Res. Atmos.*, 124, 10546-10563,
522 <https://doi.org/10.1029/2019JD031149>, 2019.
- 523 Chen, Q. C., Mu, Z., Xu, Li.: Triplet-state organic matter in atmospheric aerosols: Formation characteristics and
524 potential effects on aerosol aging, *Atmos. Environ.*, 252, 118343,
525 <https://doi.org/10.1016/j.atmosenv.2021.118343>, 2021.
- 526 Chow, J. C., Watson, J. G., Kuhns, H.: Source profiles for industrial, mobile, and area sources in the big bend
527 regional aerosol visibility and observational study, *Chemosphere.*, 54, 185-208,
528 <https://doi.org/10.1016/j.chemosphere.2003.07.004>, 2004.
- 529 Cote, C. D., Schneider, S. R., Lyu, M., Gao, S., Gan, L., Holod, A. J., Chou, T. H. H., Styler, S. A.:
530 Photochemical production of singlet oxygen by urban road dust, *Environ. Sci. Technol. Lett.*, 5, 92-97,
531 <https://doi.org/10.1021/acs.estlett.7b00>, 2018.
- 532 Dalrymple, R. M., Carfagno, R. K., Sharpless, R. M.: Correlations between dissolved organic matter optical
533 properties and quantum yields of singlet oxygen and hydrogen peroxide, *Environ. Sci. Technol.*, 44,
534 5824-5829, <https://doi.org/10.1021/es101005u>, 2010.
- 535 De-Haan, D. O., Hawkins, L. N., Kononenko, J. A., Turley, J. J., Corrigan, A. L., Tolbert, M. A., Jimenez, J. L.:
536 Formation of nitrogen-containing oligomers by methylglyoxal and amines in simulated evaporating cloud
537 droplets, *Environ. Sci. Technol.*, 45, 984-991, <https://doi.org/10.1021/es102933x>, 2010.
- 538 Derwent, R. G., Jenkin, M. E., Saunders, S. M., Pilling, M. J.: Photochemical ozone creation potentials for
539 organic compounds in Northwest Europe calculated with a master chemical mechanism, *Atmos. Environ.*,
540 32, 2429-2441, [https://doi.org/10.1016/S1352-2310\(98\)00053-3](https://doi.org/10.1016/S1352-2310(98)00053-3), 1998.
- 541 Ervens, B., Turpin, B. J., Weber, R. J.: Secondary organic aerosol formation in cloud droplets and aqueous
542 particles (aqSOA): A review of laboratory, field and model studies, *Atmos. Chem. Phys.*, 11, 11069-11102,
543 <https://doi.org/10.5194/acp-11-11069-2011>, 2011.
- 544 Fang, G.D., Gao, J., Dionysiou, D. D., Liu, C., Zhou, D.: Activation of persulfate by quinones: Free radical
545 reactions and implication for the degradation of PCBs, *Environ. Sci. Technol.*, 47, 4605-4611,
546 <https://doi.org/10.1021/es400262n>, 2013.
- 547 Feczko, T., Puxbaum, H., Kasper-Giebl, A., Handler, M., Limbeck, A., Gelencser, A.: Determination of water and
548 alkaline extractable atmospheric humic-like substances with the TU Vienna HULIS analyzer in samples
549 from six background sites in Europe, *J. Geophys. Res. Atmos.*, 112, D23S10,
550 <https://doi.org/10.1029/2006JD008331>, 2007.
- 551 Garg, S., Rose, A. L., Waite, T. D.: Photochemical production of superoxide and hydrogen peroxide from natural
552 organic matter, *Geochimica Et Cosmochimica Acta.*, 75, 4310-4320,
553 <https://doi.org/10.1016/j.gca.2011.05.014>, 2011.



- 554 Glover, C. M., Rosario-Ortiz, F. L.: Impact of halides on the photoproduction of reactive intermediates from
555 organic matter, *Environ. Sci. Technol.*, 47, 13949-13956, <https://doi.org/10.1021/es4026886>, 2013.
- 556 Goldstein, A. H., Galbally, I. E.: Known and unexplored organic constituents in the earth's atmosphere, *Environ.*
557 *Sci. Technol.*, 41, 1514-1521, <https://doi.org/10.1021/es072476p>, 2007.
- 558 Griffith, E. C., Carpenter, B. K., Shoemaker, R. K., Vaida, V.: Photochemistry of aqueous pyruvic acid, *P Natl*
559 *Acad. Sci.*, 110, 11714-11719, <https://doi.org/10.1073/pnas.1303206110>, 2013.
- 560 Hawkins, L. N., Welsh, H. G., Alexander, M. V.: Evidence for pyrazine-based chromophores in cloud water
561 mimics containing methylglyoxal and ammonium sulfate, *Atmos. Chem. Phys.*, 18, 12413-12431,
562 <https://doi.org/10.5194/acp-18-12413-2018>, 2018.
- 563 Halladja, S., Ter-Halle, A., Aguer, J. P.: Inhibition of humic substances mediated photooxygenation of furfuryl
564 alcohol by 2,4,6-trimethylphenol: Evidence for reactivity of the phenol with humic triplet excited states,
565 *Environ. Sci. Technol.*, 41, 6066-6073, <https://doi.org/10.1021/es070656t>, 2007.
- 566 Herrmann, H., Schaefer, T., Tilgner, A., Styler, S. A., Weller, C., Teich, M., Otto, T.: Tropospheric aqueous-phase
567 chemistry: Kinetics, mechanisms, and its coupling to a changing gas phase, *Chem. Rev.*, 115, 4259-4334,
568 <https://doi.org/10.1021/cr500447k>, 2015.
- 569 Kaur, R., Anastasio, C.: First measurements of organic triplet excited states in atmospheric waters. *Environ. Sci.*
570 *Technol.*, 52, 5218-5226, <https://doi.org/10.1021/acs.est.7b06699>, 2018.
- 571 Kellogg, R. E., Simpson, W. T.: Perturbation of singlet-triplet transition energies, *J. Am. Chem. Soc.*, 87,
572 4230-4234, <https://doi.org/10.1021/ja00947a002>, 1965.
- 573 Kikuchi, T., Fujii, M., Terao, K., Jiwei, R., Lee, Y. P., Yoshimura, C.: Correlations between aromaticity of
574 dissolved organic matter and trace metal concentrations in natural and effluent waters: A case study in the
575 Sagami River Basin, Japan, *Sci. Total Environ.*, 576, 36-45, <https://doi.org/10.1016/j.scitotenv.2016.10.068>,
576 2017.
- 577 Kuznetsova, N. A., Kaliya, O. L.: The photochemistry of coumarins, *Chem. Rev.*, 61, 683-696,
578 <https://doi.org/10.1070/RC1992v061n07ABEH000992>, 1992.
- 579 Laskin, A., Laskin, J., Nizkorodov, S.: Chemistry of atmospheric brown carbon, *Chem. Rev.*, 115, 4335-4382,
580 <https://doi.org/10.1021/cr5006167>, 2015.
- 581 Lee, H. J., Aiona, P. K., Laskin, A., Laskin, J., Nizkorodov, S. A.: Effect of solar radiation on the optical
582 properties and molecular composition of laboratory proxies of atmospheric brown carbon, *Environ. Sci.*
583 *Technol.*, 48, 10217-10226, <https://doi.org/10.1021/es502515r>, 2014.
- 584 Lim, H. J., Carlton, A. G., Turpin, B. J.: Isoprene forms secondary organic aerosol through cloud processing:
585 Model simulations. *Environ. Sci. Technol.*, 39, 4441-4446, <https://doi.org/10.1021/es048039h>, 2005.
- 586 Lin, P., Bluvshstein, N., Rudich, Y., Nizkorodov, S. A., Laskin, J., Laskin, A.: Molecular chemistry of atmospheric
587 brown carbon inferred from a nationwide biomass burning event. *Environ. Sci. Technol.*, 51, 11561-11570,
588 <https://doi.org/10.1021/acs.est.7b02276>, 2017.
- 589 Lin, P., Rincon, A. G., Kalberer, M., Yu, J. Z.: Elemental composition of HULIS in the Pearl River Delta Region,
590 China: Results inferred from positive and negative electrospray high resolution mass spectrometric data,
591 *Environ. Sci. Technol.*, 46, 7454-7462, <https://doi.org/10.1021/es300285d>, 2012.
- 592 McNeill, K., Canonica, S.: Triplet state dissolved organic matter in aquatic photochemistry: Reaction
593 mechanisms, substrate scope, and photophysical properties, *Environ. Sci. Proc. Imp.*, 18, 1381-1399,
594 <https://doi.org/10.1039/c6em00408c>, 2016.
- 595 Mostafa, S., Rosario-Ortiz, F. L.: Singlet oxygen formation from wastewater organic matter, *Environ. Sci.*
596 *Technol.*, 47, 8179-8186., <https://doi.org/10.1021/es401814s>, 2013.



- 597 Murphy, K. R., Stedmon, C. A., Graeber, D.: Fluorescence spectroscopy and multi-way techniques-PARAFAC,
598 *Anal. Methods.*, 5, 6557-6566, <https://doi.org/10.1039/c3ay41160e>, 2013.
- 599 Mu, Z., Chen, Q. C., Zhang, L. X., Guan, D. J., Li, H.: Photo-degradation of atmospheric chromophores: Type
600 conversion and changes in photochemical reactivity, *Atmos. Chem. Phys. Discuss.*[preprint],
601 <https://doi.org/10.5194/acp-2020-1223>, in review, 2020.
- 602 Nguyen, Q. T., Kristensen, T. B., Hansen, A. M. K., Skov, H., Bossi, R., Massling, A., Sørensen, L. L., Bilde, M.,
603 Glasius, M., Nøjgaard, J. K.: Characterization of humic-like substances in Arctic aerosols, *J. Geophys. Res.*
604 *Atmos.*, 119, 5011-5027, <https://doi.org/10.1002/2013JD020144>, 2014.
- 605 Parker, K. M., Pignatello, J. J., Mitch, W. A.: Influence of ionic strength on tripletstate natural organic matter loss
606 by energy transfer and electron transfer pathways, *Environ. Sci. Technol.*, 47, 10987-10994,
607 <https://doi.org/10.1021/es401900j>, 2013.
- 608 Perri, M. J.; Seitzinger, S. P.; Turpin, B. J. Secondary organic aerosol production from aqueous photooxidation of
609 glycolaldehyde:Laboratory experiments. *Atmos. Environ.* 2009, 43, 1487-1497.
- 610 Rollins, A. W., Browne, E. C., Min, K. E., Pusede, S. E.: Evidence for NO_x control over nighttime SOA
611 formation, *Science.*, 337, 1210-1212, <https://doi.org/10.1126/science.1221520>, 2012.
- 612 Rincon, A. G., Guzman, M. I., Hoffmann, M. R., Colussi, A. J. J.: Optical absorptivity versus molecular
613 composition of model organic aerosol matter, *Phys. Chem. A.*, 113, 10512-10520,
614 <https://doi.org/10.1021/jp904644n>, 2009.
- 615 Shi, T. M., Schins, R. P. F., Knaapen, A. M.; Kuhlbusch, T., Pitz, M., Heinrich, J., Borm, P. J. A.: Hydroxyl
616 radical generation by electron paramagnetic resonance as a new method to monitor ambient particulate
617 matter composition, *J. Environ. Monit.*, 5, 550-556, <https://doi.org/10.1039/B303928P>, 2003.
- 618 Vione, D., Maurino, V., Minero, C., Pelizzetti, E., Harrison, M. A. J.: Photochemical reactions in the tropospheric
619 aqueous phase and on particulate matter, *Chem. Soc. Rev.*, 35, 441-453, <https://doi.org/10.1039/b510796m>,
620 2006.
- 621 Vione, D., Minella, M., Maurino, V., Minero, C.: Indirect photochemistry in sunlit surface waters: Photoinduced
622 production of reactive transient species, *Chem. Eur. J.*, 20, 10590-10606,
623 <https://doi.org/10.1002/chem.201400413>, 2014.
- 624 Wang, Y. J., Hu, M., Lin, P., Guo, Q. F., Wu, Z. J., Li, M. R., Zeng, L. M., Song, Y., Zeng, L. W., Wu, Y. S., Guo,
625 S., Huang, X. F., He, L. Y.: Molecular characterization of nitrogen-containing organic compounds in
626 humic-like substances emitted from straw residue burning, *Environ. Sci. Technol.*, 51, 5951-5961,
627 <https://doi.org/10.1021/acs.est.7b00248>, 2017.
- 628 Wan, D., Sharma, V. K., Liu, L., Zuo, Y. G., Chen, Y.: Mechanistic insight into the effect of metal ions on
629 photogeneration of reactive species from dissolved organic matter, *Environ. Sci. Technol.*, 53, 5778-5786,
630 <https://doi.org/10.1021/acs.est.9b00538>, 2019.
- 631 Warneck, P.: The relative importance of various pathways for the oxidation of sulfur dioxide and nitrogen dioxide
632 in sunlit continental fair weather clouds, *Phys. Chem. Chem. Phys.*, 1,
633 5471-5483, <https://doi.org/10.1039/A906558J>, 1999.
- 634 Wenk, J., Von, G. U., Canonica, S.: Effect of dissolved organic matter on the transformation of contaminants
635 induced by excited triplet states and the hydroxyl radical, *Environ. Sci. Technol.*, 45, 1334-1340,
636 <https://doi.org/10.1021/es102212t>, 2011.
- 637 Wilkinson, F., Helman, W. P., Ross, A. B.: Quantum yields for the photosensitized formation of the lowest
638 electronically excited singlet state of molecular oxygen in solution, *J. Phys. Chem. Ref. Data.*, 22, 113-262,
639 <https://doi.org/10.1063/1.555934>, 1993.



- 640 Xu, H. M., Cooper, W. J., Jung, J. y., Song, W. H.: Photosensitized degradation of amoxicillin in natural organic
641 matter isolate solutions, *Water Res.*, 45, 632-638, <https://doi.org/10.1016/j.watres.2010.08.024>, 2011.
- 642 Zhang, D. N., Yan, S. W., Song, W. H.: Photochemically induced formation of reactive oxygen species (ROS)
643 from effluent organic matter, *Environ. Sci. Technol.*, 48, 12645-12653, <http://dx.doi.org/10.1021/es5028663>,
644 2014.
- 645 Zhou, H. X., Yan, S. W., Ma, J. Z., Lian, L. S., Song, W. H.: Development of novel chemical probes for
646 examining triplet natural organic matter under solar illumination, *Environ. Sci. Technol.*, 51,
647 <http://dx.doi.org/10.1021/acs.est.7b02828>, 11066-11074, 2017a.
- 648 Zhou, H. X., Lian, L. S., Yan, S.W., Song, W. H.: Insights into the photo-induced formation of reactive
649 intermediates from effluent organic matter: the role of chemical constituents, *Water Res.*, 112, 120-128,
650 <http://dx.doi.org/10.1016/j.watres.2017.01.048>, 2017b.
- 651 Zhou, H. X., Yan, S. W., Lian, L. S., Song, W. H.: Triplet-state photochemistry of dissolved organic matter:
652 triplet-state energy distribution and surface electric charge conditions, *Environ. Sci. Technol.*, 53, 2482-2490,
653 <http://dx.doi.org/10.1021/acs.est.8b06574>, 2019.