Measurement report: Three years of size-resolved eddy-covariance particle number flux measurements in an urban environment

Agnes Straaten, Stephan Weber

Institute of Geoecology, Technische Universität Braunschweig, Braunschweig, 38106, Germany

5   *Correspondence to*: Agnes Straaten (agnes.straaten@tu-braunschweig.de)

Abstract. Size-resolved particle number fluxes in the size range 10 nm < particle diameter (D_p) < 200 nm were measured over a 3-year period (April 2017 - March 2020) using the eddy covariance technique at an urban site in Berlin, Germany. The observations indicated the site as a net source of particles with a median total particle number flux of F_{TNC} = 0.86 x 10^8 m^{-2} s^{-1}. The turbulent surface-atmosphere exchange of particles was clearly dominated by ultrafine particles (D_p <

10   100 nm) with a share of 96 % of total particle number flux (F_{UFP} = 0.83 x 10^8 m^{-2} s^{-1}).

Annual estimates of median F_{TNC} and F_{UFP} slightly decreased by -9.6 % (-8.9 % for F_{UFP}) from the first to the second observation year and a further -5.9 % (-6.1 % for F_{UFP}) from the second to the third year. The annual variation might be due to different reasons such as variation of flux footprints in the individual years, a slight reduction of traffic intensity in the third year or a progressive transition of the vehicle fleet towards a higher share of low-emission standards or electric drive. Size-

15   resolved measurements illustrated events of bidirectional fluxes, i.e. simultaneous emission and deposition fluxes within the size spectrum, which occurred more often in spring, late summer and autumn than in winter. Multi-year observations of size-resolved particle fluxes proved to be important for deeper understanding of particle exchange processes with the urban surface and the pronounced influence of traffic at this urban site.

1 Introduction

20   Airborne particles, especially ultrafine particles (UFP) with diameters D_p < 100 nm, may enter the circulatory system and are associated to acute and chronic effects on human health (Oberdörster et al., 1995; Nemmar et al., 2002; HEI Review Panel on Ultrafine Particles, 2013; Schraufnagel, 2020). UFP are abundant in cities due to a number of different anthropogenic emission sources such as private and public traffic or commercial and industrial combustion (Tsang et al., 2008; Tie et al. 2009; Bäfver et al. 2011; Weber et al., 2013; Kumar et al., 2014; Gerling et al., 2020). In order to understand the spatio-temporal variation

25   of population exposure to ultrafine particles, information on the strength and location of sources and sinks, and the turbulent exchange between surface and atmosphere are essential (Buzorius et al., 2000; Longley et al., 2003; Mårtensson et al., 2006; Weber et al., 2013).

The turbulent surface-atmosphere-exchange of pollutants may be quantified using the micrometeorological eddy covariance (EC) method (e.g., Baldocchi 2003; Burba and Anderson, 2005; Aubinet et al., 2012). During the past decades EC was applied

30   over non-urban and urban terrain to quantify gaseous pollutant and particle fluxes (e.g. Buzorius et al., 1998; Suni et al., 2003; Damay et al., 2009; Ripamonti et al., 2013; Gioli et al., 2013; Deventer et al., 2018).

Urban particle number flux measurements using EC with varying lower cut-off diameters of the particle counting instruments were conducted in a couple of cities across Europe (cf. Table 1). These studies generally report the city to be a net source of particles, i.e. a positive average particle number flux, which by convention indicates an upward directed emission flux to the

35   atmosphere in contrast to a negative (downward directed) deposition flux. A significant contribution to the particle number flux is attributed to vehicle traffic with peak mode diameters of fresh traffic emission in the size range $D_p < 30$ nm (Wiedensohler et al., 2002; Zhu et al., 2007; Rönkkö et al., 2017). On the diurnal cycle, particle number fluxes are relatively small at night but increase in the early morning hours due to anthropogenic activity and enhanced turbulence, resulting in peak upward fluxes at around noon or in the early afternoon (Dorsey et al., 2002; Mårtensson et al., 2006; Deventer et al., 2013;

40   Conte et al., 2018). Variation in particle number fluxes is shown to be related to different land use within the flux source area with lower emission fluxes from vegetated surfaces or residential areas and higher fluxes from urban land-use with intense traffic (e.g. Järvi et al. 2009, Ripamonti et al. 2013).

Although aerosol processes and dynamics such as particle deposition or the influence on human health are strongly related to the particle diameter, information on size-resolved urban particle number fluxes, especially in the ultrafine size range, remains

45   scarce. Size-resolved fluxes, however, help to deepen understanding on particle sources and dynamics in the urban boundary layer. Additionally, the flux data is essential to parameterise dry deposition velocities and to validate urban air quality models (Saylor et al., 2019; Farmer et al., 2021). Size-resolved flux observations may point to situations of bidirectional particle fluxes, i.e. simultaneously occurring upward and downward fluxes in different ranges of the particle size spectrum (Schmidt and Klemm, 2008; Deventer et al., 2015). Usually upward fluxes tend to occur in the smaller size ranges and may point to emission

50   sources such as vehicle traffic whereas downward fluxes mainly occur in larger size ranges, i.e. accumulation mode particles. Particle flux studies usually comprise observation periods of a few weeks or months within different seasons of the year. Only few studies report observations for longer periods such as one-year or multi-annual periods (e.g. Ripamonti et al., 2013; Deventer et al., 2015, cf. Table 1). This is especially true for size-resolved particle number fluxes from urban areas for which multi-annual data has not been available, yet. These longer term studies, however, are important to analyse temporal variation

55   of urban particle number fluxes and their driving mechanisms.

Here, we report on three years of size-resolved urban particle number fluxes in the diameter size range 10 nm $< D_p <$ 200 nm. The measurements were carried out in central Berlin, Germany, in the framework of the urban climate research project [UC]²-3DO (Scherer et al., 2019). We analyse the diurnal, seasonal and annual variation of ultrafine and size-resolved particle number fluxes and study the contribution from surface drivers such as land-use on temporal variation of particle fluxes.

60

Table 1: Overview of total particle number fluxes of measurement campaigns from ten different cities. Right column: The information in brackets specifies the data basis of the individual studies for estimating the observed particle fluxes.

City	Study	Duration/ Season	Size-resolved	Lower cut-off (nm)	Particle number fluxes (10^8 m^{-2}s^{-1})
Münster	Schmidt & Klemm, 2008	~ 3 months, summer	yes	30	-0.43 – 1.44
	Deventer et al., 2013	1-2 months, spring	yes	55	0 – 0.1 (mean diurnal cycle)
	Deventer et al., 2015	1 year	yes	55	0.02 – 0.1 (seasonal fluctuations)
Innsbruck	Deventer et al., 2018	~1 months, summer	yes	6	0.5 – 3.0 [a]
	Heyden et al., 2018	~1 months, summer	no	10	-1.9 – 2.8 (10-90 percentile)
Lecce	Contini et al., 2012	1-2 month, spring	no	9	2.0 – 10.8 [b] (IQR)
	Conte et al., 2018	1-2 months, spring	yes (3 size bins)	8	2.82 [b]
Helsinki	Järvi et al., 2009	1 year	no	6	0.44 – 8.4 [a]
	Ripamonti et al., 2013	3 years	no	6	0.05 – 11 (daily variation)
	Kurppa et al., 2015	2 years	no	6	0.1 – 8.2 (5-95 percentile)
London	Martin et al., 2009	~ 1 month, autumn	no	10	0.5 – 6.5 (mean diurnal cycle)
	Harrison et al., 2012	~ 1 month, autumn	yes / no	50 / 10	up to 3 / 0.5 – 7
Edinburgh	Nemitz et al., 2000	3 weeks each, spring, autumn	no	100	0.1 – 0.6
	Dorsey et al., 2002	2 weeks each, spring, 2x autumn	no	11	0.9 – 9 (mean diurnal cycle)
	Martin et al., 2009	summer, autumn	no	11	5 – 11 (mean diurnal cycle)
Stockholm	Mårtensson et al., 2006	2-3 months, spring	no	11	0.1 – 3
Manchester	Longley et al., 2004	2 weeks, autumn	yes (3 size bins)	100	0.11 – 0.37 (mean diurnal cycle)
	Martin et al., 2009	1-1.5 months, summer, winter	no	3	3 – 20 (mean diurnal cycle)

Gothenburg	Martin et al., 2009	~1 month, winter	no	5	1 – 7 (mean diurnal cycle)
Berlin	this study	3 years	yes	10	0.2 – 2.0 (mean diurnal cycle)

[a] average daytime emission, [b] measurement height 14 m (roughness sub-layer)

2 Materials and Methods

2.1 Measurement site

Size-resolved particle number fluxes were measured from 01 April 2017 to 31 March 2020 near "Ernst-Reuter-Platz" in central Berlin, Germany, atop the main building of Technische Universität Berlin. The site surroundings comprise residential built-up surfaces, traffic and green areas (Figure 1). To the north of the flux site, the busy main road "Straße des 17. Juni" with an average daily traffic intensity of about 37,300 vehicles day^{-1} is located (Umweltatlas Berlin, 2017). To represent traffic intensity in the flux footprint, data from two traffic counting stations at "Hardenbergstraße" (HS, south-west of measurement site) and "Straße des 17. Juni" (S17J, east of the site) were available (cf. Figure 1). Data was provided by the traffic information centre Berlin (VMZ Berlin Betreibergesellschaft mbH) at hourly resolution.

2.2 Instrumentation

The particle flux instrumentation consisted of an electric mobility particle sizer (Engine Exhaust Particle Sizer Spectrometer, EEPS 3090, TSI Inc., Minnesota, USA) and a 3D ultrasonic anemometer (USA-1, Metek GmbH, Elmshorn, Germany). Both instruments synchronously sampled at a frequency of 10 Hz. The EEPS measures the particle number size distribution (PNSD) over the size range 5.6 nm $< D_p <$ 560 nm in 32 size channels. Through a stainless steel tube of 0.01 m diameter, the air was sampled at a flow rate of 10 L min^{-1} resulting in a laminar sampling flow (Reynolds number $\approx$ 1300; Hinds, 1999). The steel tube was attached to a 10 m meteorological rooftop mast with the sample inlet located next to the sonic anemometer at a height of 57 m above ground level. We sampled dry aerosol using a Nafion dryer MD-700 (Perma Pure LLC, length 0.9 m) to keep relative humidity in sample air < 40 %.

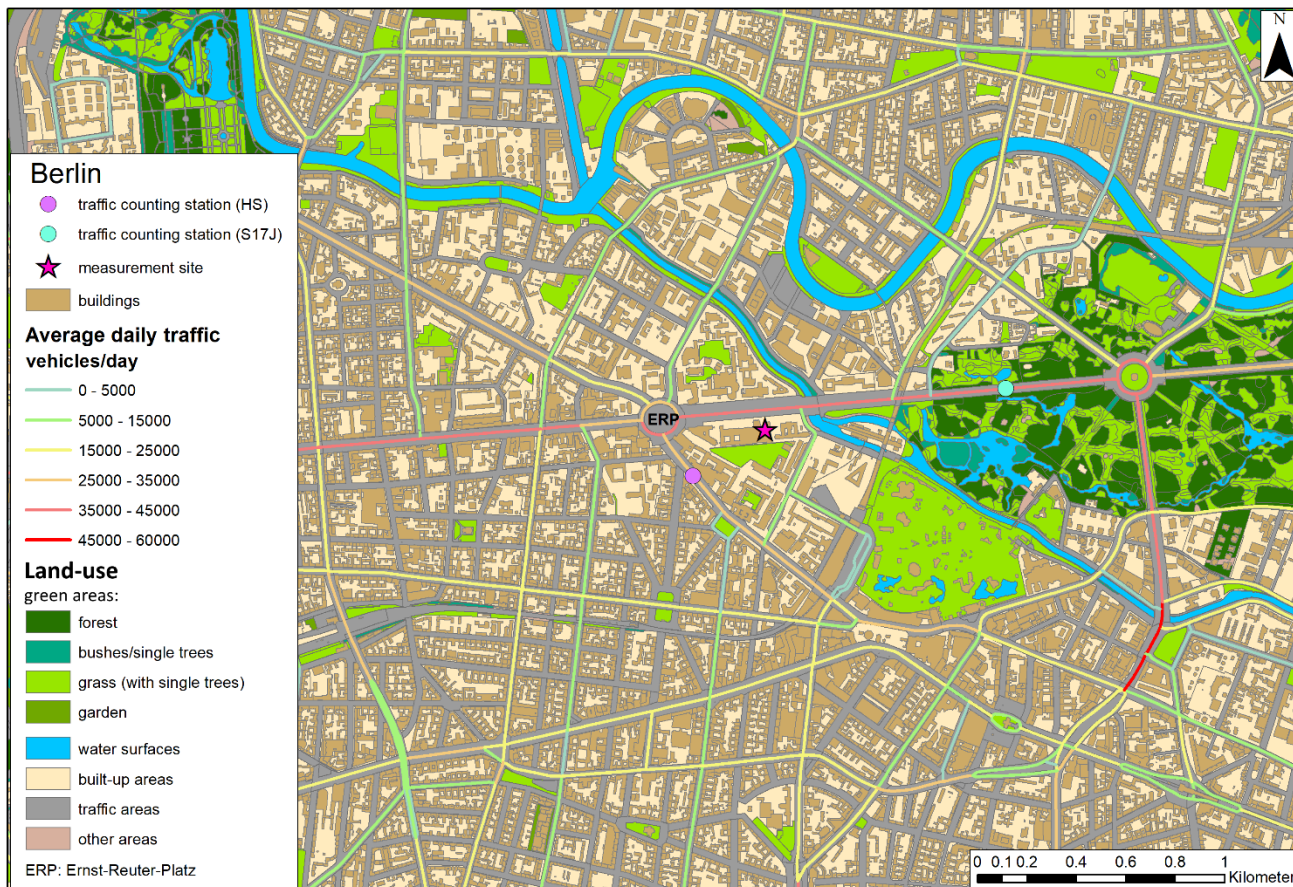
The particle spectrometer EEPS 3090 classifies particles based on their differential electrical mobility. Charged particles entering the electric field are reflected outwards where the charge is delivered to 22 electrodes and converted into particle number concentrations (TSI Inc., 2015). To limit baseline noise in the electrometer readings as can be caused by dirt, the instrument electrometers were zero-checked every second to third day using the instrument firmware. The instrument was primarily developed for measurements of engine exhaust emissions that show higher particle concentrations than usually observed in the urban background. Hence, EEPS readings of specific size channels may fall below the analyser's minimum threshold concentration resulting in non-valid concentration readings. To avoid data gaps within the PNSD, a cubic natural spline interpolation was used as gap-filling strategy to obtain complete PNSDs (cf. section 2.3).

90

Figure 1: Eddy covariance measurement site in central Berlin near Ernst-Reuter-Platz (ERP; data sources: Geoportal Berlin, 2014 (modified), 2021; Umweltatlas Berlin, 2017).

To quality-check EEPS measurements for concentrations and fluxes, an on-site inter-comparison to a water-based condensational particle counter (WCPC 3787, TSI Inc., Minnesota, USA) was carried out.

95 **2.3 Data handling**

The post-processing of particle data and fluxes comprised the following procedures:

Size-dependent diffusional particle losses within the sampling line and the Nafion dryer were corrected according to Hinds (1999). The Nafion dryer MD-700 was assumed as stainless steel tube of 0.9 m length as no specific correction factors were available for the dryer.

100 Subsequently, gaps in the PNSD (cf. section 2.2) were filled using a natural spline interpolation following Meyer-Kornblum et al. (2019). As gap-filling caused larger uncertainties in the boundary regions of the PNSD, the size spectrum was limited to 21 size channels in the size range 10 nm $< D_p <$ 200 nm. According to Meyer-Kornblum et al. (2019), PNSDs were gap-filled as long as less than 9 gaps in total (minimum of 12 remaining data points within the PNSD) or 5 contiguous gaps (5

neighbouring size channels) occurred in that size range. PNSDs not fulfilling these requirements were discarded. We measured

105   a total of 8.43 x 10^8 PNSDs during the entire study period, of which 85.4 % were gap-filled, 14.0 % were rejected and 0.6 % were without gaps and did not need to be gap-filled. The total particle number concentration (TNC, 10 nm < D_p < 200 nm), ultrafine particles (UFP, 10 nm < D_p < 100 nm) and three modal concentrations, i.e. nucleation mode (NUC, 10 nm < D_p < 30 nm), Aitken mode (AIT, 30 nm < D_p < 100 nm), and accumulation mode (ACC, 100 nm < D_p < 200 nm) were calculated from gap-filled PNSDs.

110   For particle flux calculation, the missing samples allowance for each half-hourly block average was set to 20 %. Wind vectors and particle number concentrations were checked for plausibility concerning a realistic range of absolute values and spikes were removed following Vickers and Mahrt (1997). Additionally, double coordinate rotation for tilt correction and spectral corrections of high-pass (Moncrieff et al., 2004) and low-pass (Moncrieff et al., 1997) filtering effects were applied. For time lag compensation between particle and sonic data, we used covariance maximization with a specified time lag window.

115   Following other urban particle flux studies, we applied linear detrending (e.g. Mårtensson et al., 2006; Vogt et al., 2011a; Deventer et al., 2013; Deventer et al., 2018; Heyden et al., 2018). The particle flux calculation was performed using the software EddyPro® v6.2.2. To calculate particle number fluxes for WCPC data (cf. section 2.2), the same procedure as described above was adopted. By definition, positive flux values indicate upward (emission) fluxes to the atmosphere whereas negative values denote downward (deposition) fluxes to the surface.

120   The response time of the EEPS for fast changes in concentration was reported as 0.5 s (Johnson et al., 2003, 2004). This response time is similar to other CPCs or size spectrometers used in particle flux measurements (e.g. Buzorius, 2001; Dorsey et al., 2002; Vogt et al., 2011b; Deventer et al., 2015). Limited sensor response is one reason for underestimation of turbulent particle fluxes. We corrected sensor response according to Horst (1997) resulting in an average flux correction of between +2.4 and +3.0 % in the 21 size channels. Finally, particle number fluxes with quality flags > 6 according to Foken et al. (2004)

125   were discarded from subsequent analysis.

    For the flux footprint prediction, we used the two-dimensional parameterisation of Kljun et al. (2015). To analyse the impact of varying amounts of major and minor roads in the annual flux footprints, we calculated a footprint-weighted average daily traffic intensity (ADT). For this, the ADT shape file (data source: Umweltatlas, 2017) was converted into a raster grid (4 km x 4 km with 4 m spatial resolution, comparable to the footprint climatology raster) and weighted with the footprint climatology

130   to calculate ADT.

2.4 Data availability

    For the entire three-year period, data availability was 59.2 % for total particle number flux (F_{TNC}) (59.3 % for UFP flux, F_{UFP}). For the individual observation years, data availability of F_{TNC} decreased from 62.9 % (F_{UFP}: 63.0 %) in the first year to 61.1 %

135   (F_{UFP}: 61.2 %) in the second and 53.6 % (F_{UFP}: 53.6 %) in the third year. The data availability varied due to different reasons such as quality checks, PNSDs that not fulfil the requirements for gap-filling, and off-times due to maintenance of the EEPS.

The lower data availability in the third year was due to a 2-month period (22 May – 24 July 2019) during which the EEPS was not available at the site due to inter-comparison measurements. Data availability was generally lower during night (higher rejection of EC data due to quality flags as is common in EC measurements due to reduced turbulence and stronger stability) and showed highest diurnal average data availability for F_{UFP} in autumn (69 %), followed by winter (58 %), spring (56 %), and summer (54 %).

2.5 Quantification of measurement uncertainty

Atmospheric observations are prone to measurement uncertainty, e.g. sampling errors as well as sensor and gap-filling uncertainty. In the following, we briefly outline the approach to quantify the potential sources of uncertainty.

The gap-filling of size channels below the minimum threshold concentration of the EEPS causes uncertainty due to the natural spline interpolation method (Meyer-Kornblum et al., 2019). Hence, percentage errors resulting from the interpolation process are dependent on the number of gaps within a PNSD and the particle diameter. The highest average percentage errors of 13.2 up to 31.6 % due to gap-filling occur in the size range of 10 – 20 nm. The corresponding uncertainty estimates for the size ranges of 20 – 50 nm, 50 – 100 nm, and 100 – 200 nm were between 5.6 – 17.2 %, 2.0 – 3.7 %, and 2.2 – 6.7 %, respectively (Meyer-Kornblum et al., 2019).

The random error for eddy covariance flux measurements was estimated according to Finkelstein and Sims (2001). The median random flux error varies between 27 and 40 % for the different size channels (25 % for F_{UFP} and F_{TNC}) and generally increases towards the boundary size regions of the PNSD. This coincides with random flux errors of between 28 and 39 % as reported by Deventer et al. (2018).

2.6 Land-use regression analysis

A multiple linear land use regression (LUR) analysis was carried out to estimate the relationship of F_{TNC} and F_{UFP} with land-use in the flux footprint. The land-use types were determined based on a biotope type mapping provided by city authorities (Geoportal Berlin, 2014). We binned data into 16 wind sectors, which consisted of three land-use types, namely "built-up areas", "traffic areas", and "green areas". Green areas (cf. Figure 1) were not classified into more specific subtypes since the surface fractions of forest, bushes/single trees, and garden areas in the flux footprints were small (cf. Figure 5). Additionally, the available data on specific subtypes of green areas was limited and could not differentiate, for instance, between grass areas with or without single trees. The land-use types "water surfaces" and "other areas" were eliminated to avoid collinearity due to high correlation (Pearson's $r \geq 0.6$ or $r \leq -0.6$) with other land-use types and the small amount of these surface types within the footprint. Since the model is thought to provide an estimate of the variation of particle emission and deposition with land-use, non-significant variables were not eliminated.

3 Results

3.1 Meteorological characterisation of the study period

For an overview of the meteorological situation in the measurement period, data from the German weather service station "Berlin Tempelhof" (7 km SE from the flux site) was used. The average annual air temperatures during the study period were 10.3°C, 12.5°C, and 12.0°C (average annual temperature 1981 – 2010: 9.9°C). In terms of annual precipitation, the first observation year (04/2017 – 03/2018) with an annual precipitation sum of 798 mm was significantly wetter than the long-term average of 577 mm (1981 – 2010). However, precipitation in the second year (04/2018 – 03/2019) and the third year (04/2019 – 03/2020) amounted to only 384 mm and 433 mm with the year 2018 characterised by a summer drought and hot spells. Wind at the "Tempelhof" site blows dominantly from SW (average wind direction in 1981 – 2010: 243°). The annual average wind direction only slightly varied with 255°, 272°, and 248° over the three years, respectively (Figure 2, Figure A1).

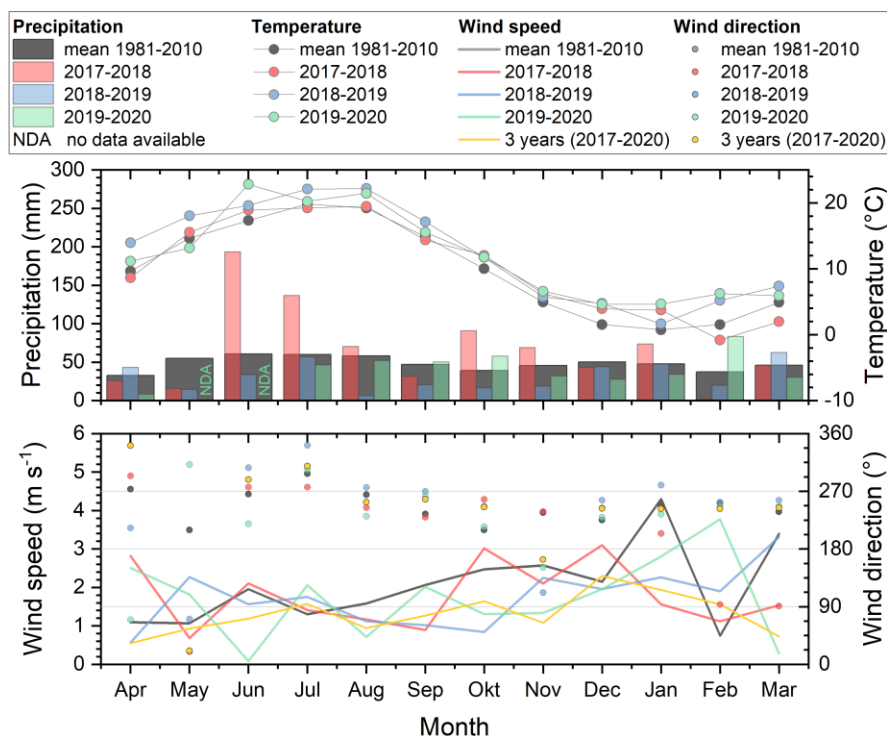



Figure 2: Meteorological data such as temperature, precipitation, wind speed, and wind direction from the German weather service station "Berlin Tempelhof" (German weather service site ID: 433) located 7 km SE from the flux site.

3.2 Data quality assurance and quality control

3.2.1 Quality check of EEPS

The on-site inter-comparison of particle number concentration between EEPS and WCPC (cf. section 2.2) indicated good
185   agreement with an average relative deviation of 10.6 % (slope 0.94, $R^2 = 0.84$, Figure A2a). It has to be noted that we compare
the WCPC data (lower cut-off of 5 nm) with the gap-filled EEPS data (10 nm < D_p < 200 nm). The deviation in particle
concentrations is plausible as it is caused by different lower and upper cut-off diameters.

The time series of WCPC and EEPS fluxes were strongly correlated (r = 0.91), although the EEPS underestimated fluxes
measured by WCPC (slope 0.66, $R^2 = 0.84$, average relative deviation of 26 %, Figure A2b). We argue that the underestimation
190   is due to the lower WCPC cut-off (5 nm) and the dominance of nucleation mode fluxes at the present site (cf. section 3.5).

3.2.2 Spectral analysis

Spectral analysis is used in EC applications to study the frequency response of the measurement setup. Either the measured
cospectra were compared to ideal cospectra (e.g. Kaimal et al., 1972) or to the sonic sensible heat flux cospectrum which often
resembles an ideal cospectrum (Aubinet et al., 2012).
195   We calculated normalised daytime cospectra for F_{UFP} for all wind directions under neutral stratification. Most of the cospectra
values were positive (blue dots) indicating the frequent occurrence of emission fluxes (Figure 3). The particle flux cospectrum
agreed well with sensible heat and the ideal Kaimal spectrum. Furthermore, the cospectrum follows the theoretical -4/3 slope
in the inertial subrange indicating reasonable sensor frequency response.

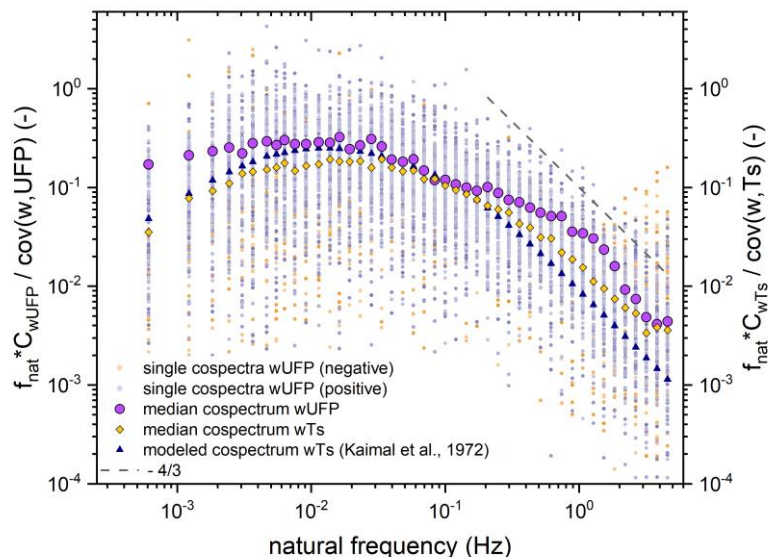



Figure 3: Normalised (median) cospectra of F_{UFP} (n = 160), sonic sensible heat flux, and ideal cospectrum after Kaimal et al. (1972). Additionally, the -4/3 slope in the inertial subrange is indicated. Blue dots represent positive values of the cospectrum, yellow dots negative values with inverse sign.

3.3 Particle number concentrations

Over the course of the three-year study period, we observed median TNC concentrations of 7,300 cm^{-3} and 6,450 cm^{-3} for UFP, respectively (cf. Table 2). The half-hourly average TNC varied between a minimum of 3,099 cm^{-3} and a maximum of 53,879 cm^{-3}. Ultrafine particles dominated rooftop concentrations as indicated by an average UFP/TNC ratio of 0.9. The majority of particles in UFP size range occurred in the NUC mode.

Table 2: Statistical quantities of TNC, UFP as well as particle number concentrations of the three modes NUC, AIT, and ACC.

Particle number concentration (cm^{-3})	TNC	UFP	NUC	AIT	ACC
Minimum	3,099	2,828	1,732	801	77
2-percentile	3,794	3,488	1,986	1,126	177
Median	7,300	6,447	3,136	3,027	681
Mean	8,337	7,522	3,995	3,528	814
98-percentile	19,947	18,734	12,725	9,500	2,287
Maximum	53,879	48,145	37,442	30,820	16,741

3.4 Footprint analysis

The particle fluxes measured at our rooftop site represented a surface area of about 5.3 km² (with respect to the 80 % contour line). The flux peak contribution, however, was from an area situated to the W/SW of the site at a distance of about 50 to

350 m (Figure 4). The peak contribution source area slightly shifted northwards in the second year and southwards in the third year. However, the source areas were very similar in the individual years. A detailed analysis of source-area weighted contribution of land-use indicates 60 % of built-up areas, 26 % traffic areas, 11 % green areas, and 2 - 3 % water surfaces in the flux footprint (Figure 5). Green areas were mainly characterised by grass with or without single trees whereas forests, bushes/single trees, and gardens contribute little to the flux footprint. The land-use fractions were similar in the individual study years. However, it has to be taken into account that streets vary in terms of traffic intensity, so that the strength of traffic-related sources might differ between individual years.

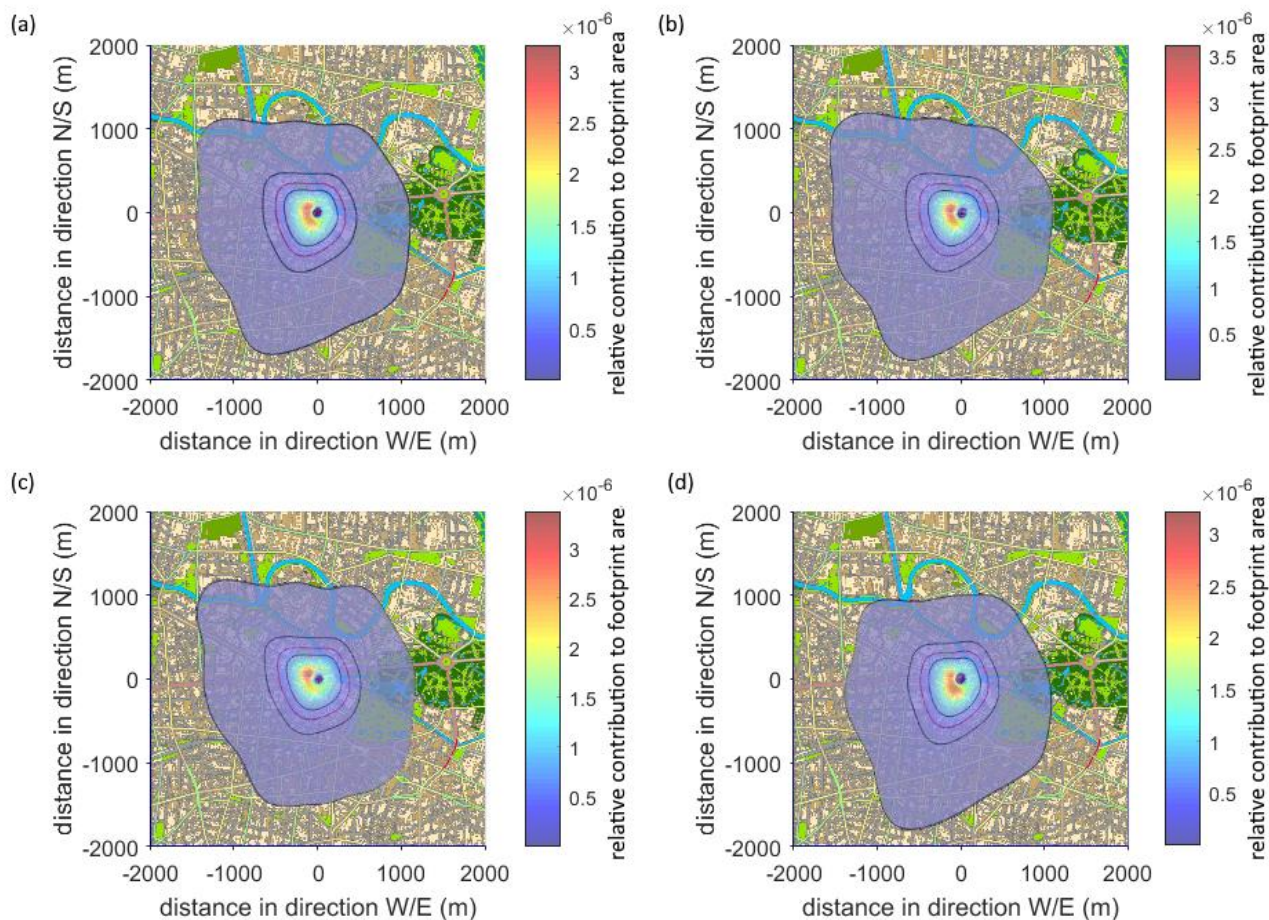



Figure 4: Footprint climatology of (a) the entire three-year period, (b) the first year, (c) the second year, and (d) the third year. Flux footprints were calculated for a 4 km x 4 km area (4 m spatial resolution). For reasons of clarity only the 40 %, 50 % (purple), 60 %, and 80 % contour lines are shown (footprint model: Kljun et al.,2015; map data sources: Geoportal Berlin, 2014 (modified), 2021; Umweltatlas Berlin, 2017).

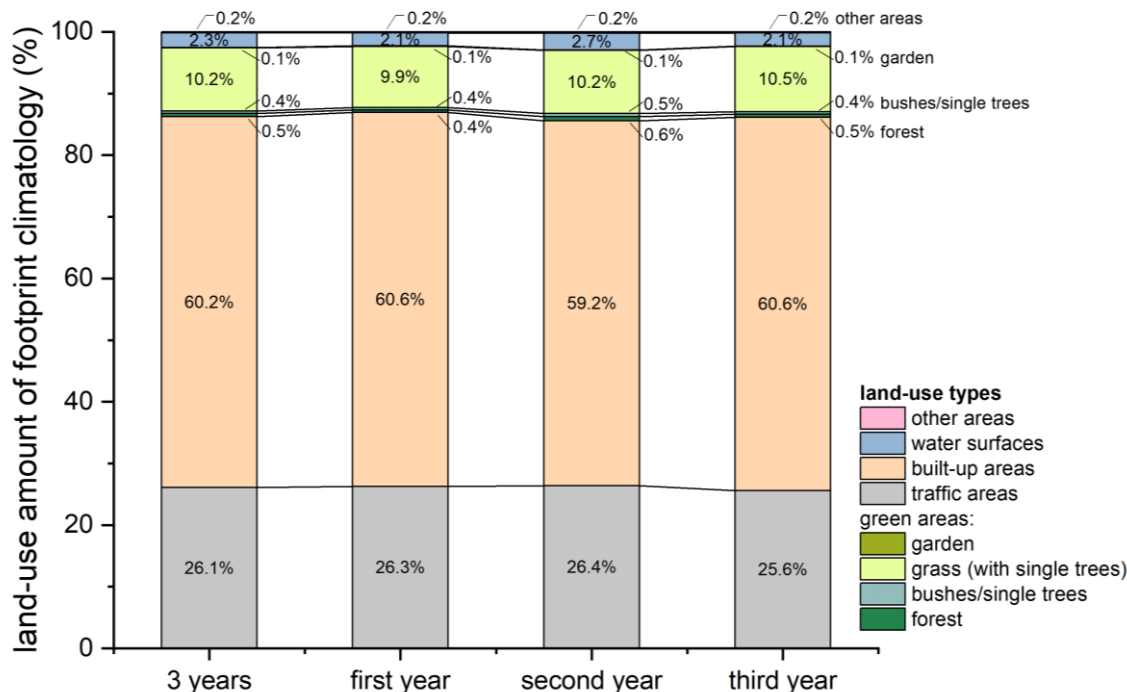



Figure 5: Amount of land-use contributing to the flux footprint climatologies separated according to the different time periods of measurement. For data analysis, the footprint information in a 4 km x 4 km grid centred on the measurement site was used (4 m spatial resolution). This footprint area completely includes the 80 % contour line.

230

3.5 Total and ultrafine particle number fluxes

The measured particle number fluxes indicated the study area in central Berlin as a net source of particles with a three year median F_{TNC} of 0.86×10^8 m^{-2} s^{-1} and 0.83×10^8 m^{-2} s^{-1} for F_{UFP} (cf. Table 3). The majority of 91.4 % of fluxes were upward (emission) fluxes, whereas 8.6 % of data showed downward (deposition) fluxes. Annual estimates of median F_{TNC} and F_{UFP}

235 showed a decrease by -9.6 % (-8.9 % for F_{UFP}) from the first to the second year and a further -5.9 % (-6.1 % for F_{UFP}) from the second to the third year (Figure 6a, Table 3). We observed a limited number of frequency of higher F_{UFP} (e.g. $> 1.2 \times 10^8$ m^{-2} s^{-1}) both in the second and third year (Figure 6b). The ratio of F_{UFP} to F_{TNC} is 0.96 whereas the average nucleation mode particle flux (F_{NUC}) accounted for 63 % of F_{TNC} (data not shown here). As the ultrafine size range clearly dominated F_{TNC}, we will focus on F_{UFP} in subsequent analysis.

240

Table 3: Mean, median, minimum, and maximum F_{TNC} and F_{UFP} for the three measurement years and the four meteorological seasons spring (MAM), summer (JJA), autumn (SON), and winter (DJF).

F_{TNC} (x 10^8 m^{-2} s^{-1})	3 years	First year	Second year	Third year	MAM	JJA	SON	DJF
Mean	1.19	1.32	1.15	1.07	1.21	1.12	1.07	1.36
Median	0.86	0.94	0.85	0.80	0.86	0.82	0.81	0.97
Minimum	-16.43	-16.43	-14.77	-10.69	-11.51	-16.3	-8.54	-14.30
Maximum	26.78	22.48	24.88	26.78	26.78	24.88	18.38	22.8
F_{UFP} (x 10^8 m^{-2} s^{-1})	3 years	First year	Second year	Third year	MAM	JJA	SON	DJF
Mean	1.14	1.27	1.10	1.04	1.17	1.08	1.03	1.31
Median	0.83	0.90	0.82	0.77	0.83	0.79	0.78	0.94
Minimum	-16.51	-16.51	-14.74	-10.73	-11.53	-16.51	-8.57	-14.34
Maximum	26.72	22.37	24.76	26.72	26.72	24.76	16.95	22.37

F_{UFP} was characterised by distinct seasonal variation, i.e. largest average fluxes occurred in winter whereas fluxes were lower in autumn and summer (Figure 7, Table 3). Generally, a lower frequency of deposition events and higher emission fluxes prevailed during winter. However, due to consistent positive monthly median values the site footprint was a net particle emission source in every single month.

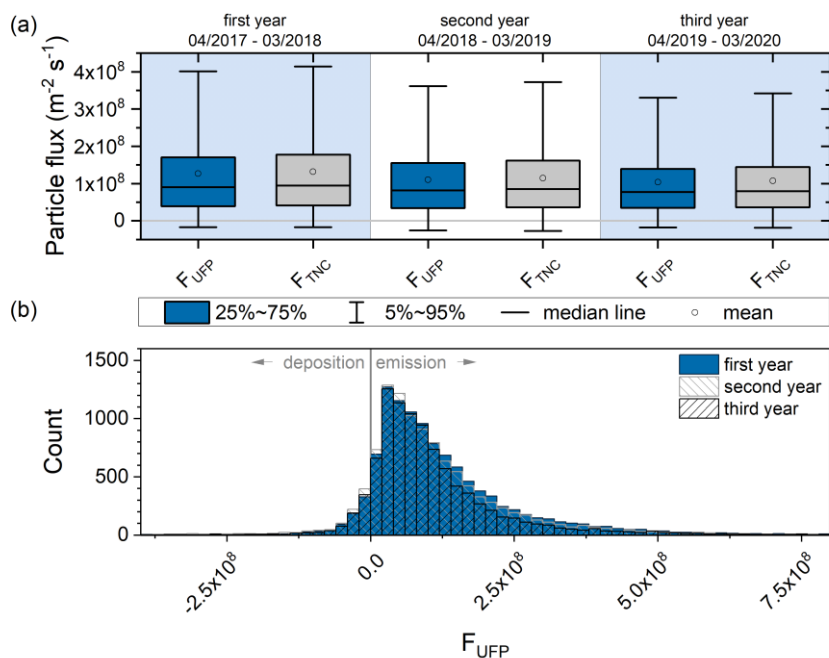



Figure 6: (a) Range of F_{TNC} and F_{UFP} and (b) the frequency distributions of F_{UFP} for each of the three years of measurement.

The seasonal pattern is also evident on the mean diurnal cycle (Figure 8a). Strongest emission fluxes occurred in winter with a maximum F_{UFP} of about 2.0×10^8 m^{-2} s^{-1}, whereas lowest fluxes were observed in autumn and summer. F_{UFP} increased after sunrise and showed two local maxima at around 9:30 LT and in the early afternoon (12:30 to 14:00 LT). The moderate decrease at around noon was related to reduced traffic intensity (cf. Figure 8d). Average F_{UFP} varied between 0.2×10^7 m^{-2} s^{-1} at night and 2.0×10^8 m^{-2} s^{-1} during day. The morning increase in F_{UFP} coincided with an increase in atmospheric turbulence and a more unstable boundary layer. In the evening, F_{UFP} decreased as atmospheric stratification changed from unstable to neutral conditions (Figure 8c).

The daily amplitude of F_{UFP} clearly differed between working days and weekends (Figure 8b). Whereas working days showed a strong morning increase of F_{UFP} and a large amplitude over the diurnal course, the weekend was characterised by a reduced morning increase and only half the amplitude in comparison to working days. This difference was in accordance with the variation of traffic intensity between working days and weekends (Figure 8d).

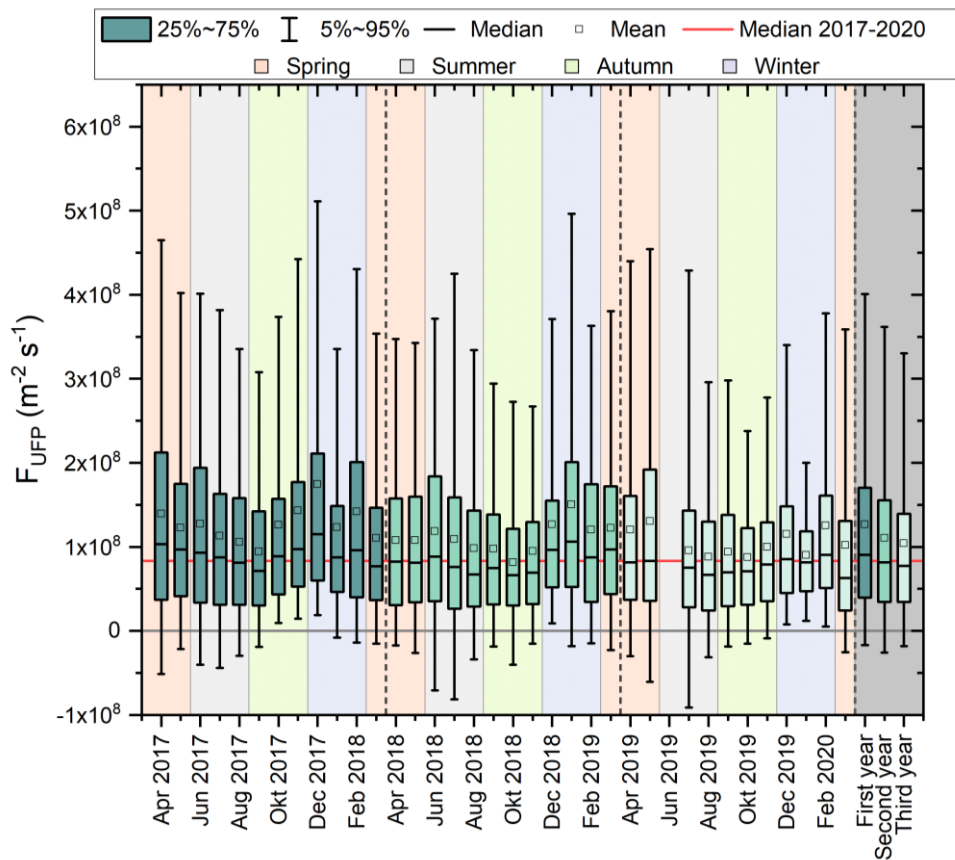



Figure 7: Variation of monthly and annual F_{UFP} during the study period.

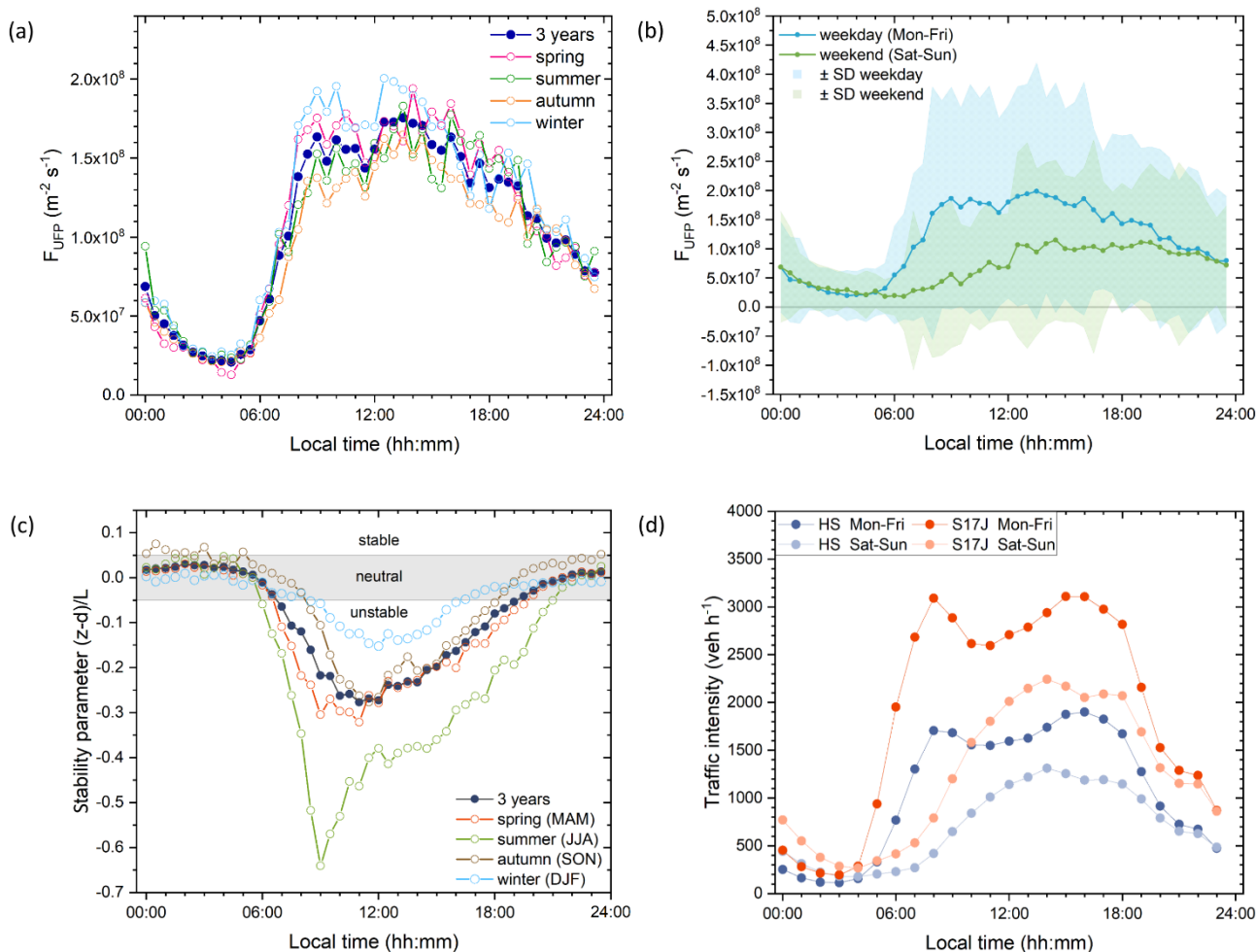



Figure 8: Diurnal cycles of (a) average and seasonal F_{UFP}, (b) average F_{UFP} for weekdays and weekends with standard deviation (SD), (c) median stability parameter (z-d)/L for each season and the whole measurement period with z = measurement height, d = displacement height, and L = Monin-Obukhov length, and (d) average traffic intensity at HS and S17J.

3.6 Size-resolved particle number fluxes

Size-resolved particle number fluxes were characterised by distinct temporal variation on the diurnal cycle (Figure 9a). Strongest emission prevailed in the smallest size bin (D_p < 12 nm) with a maximum in the early afternoon. Coarser particles (D_p > 100 nm), however, showed little variation regardless of the day of week. Although deposition fluxes occurred (Figure 9b), positive average fluxes were prevailing in all size ranges emphasising the dominance of net particle emission from the flux footprint. The frequency of particle deposition events peaked in the ACC mode (Figure 9b). On working days, the deposition frequency of UFP was highest during night, especially in the second half of the night. On weekends, UFP deposition

more often occurred in the morning to pre-noon hours. The timing of deposition fluxes was associated with the daily minimum in F_{UFP}, low traffic intensity, and a neutral or stable atmosphere (cf. Figure 8). Generally, NUC mode particles showed higher deposition probabilities than AIT mode particles. This was evident during the entire day and for every day of the week.

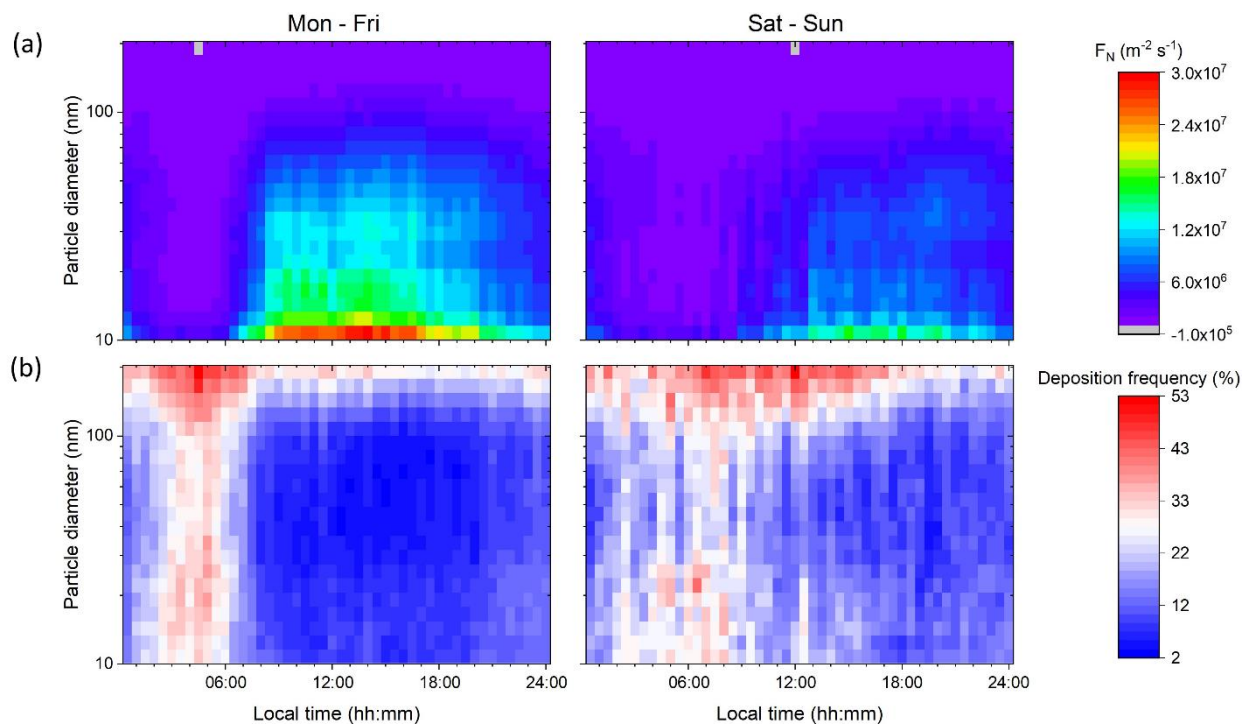



Figure 9: Diurnal cycles for a weekday (Mon – Fri) and a day of weekend (Sat – Sun) of (a) size-resolved average particle number fluxes (F_N) and (b) the frequency of particle deposition. In (a) colours indicate the strength of emission fluxes while grey colour symbolises a slight particle deposition.

On the seasonal cycle, highest emission fluxes predominantly occurred in the smallest size bins ($D_p < 30$ nm) during winter (average F_{TNC}: 1.44×10^8 m^{-2} s^{-1} and F_{UFP}: 1.38×10^8 m^{-2} s^{-1} in December, Figure 10a, b). The lowest average particle number fluxes were evident in late summer and autumn. Additionally, the frequency of particle deposition events was higher in summer/autumn with a maximum in August and September (Figure 10c). The variation in the frequency of deposition fluxes over the size spectrum pointed to the occurrence of bidirectional fluxes.

The difference between the maximum and minimum of deposition frequencies in a monthly size spectrum (max. Δ deposition frequency, Figure 10c) might be interpreted as a measure to quantify the occurrence of bidirectional fluxes. Thus, bidirectionality in particle fluxes occurred more frequently in spring, late summer and autumn than in winter. Bidirectionality was often associated with simultaneous emission fluxes in the ultrafine and deposition fluxes in the coarser particle size range. The months of May, June, and July showed less particle deposition events but higher F_{TNC} and F_{UFP} than the other months in spring and summer. This was probably due to the variation of wind direction and, consequently, flux source area. The average

three-year wind directions were from NNE in May, WNW in June, and NW in July. In those directions strong particle sources such as the busy road "Straße des 17. Juni" and the traffic circle "Ernst-Reuter-Platz" were located (cf. Figure 1).

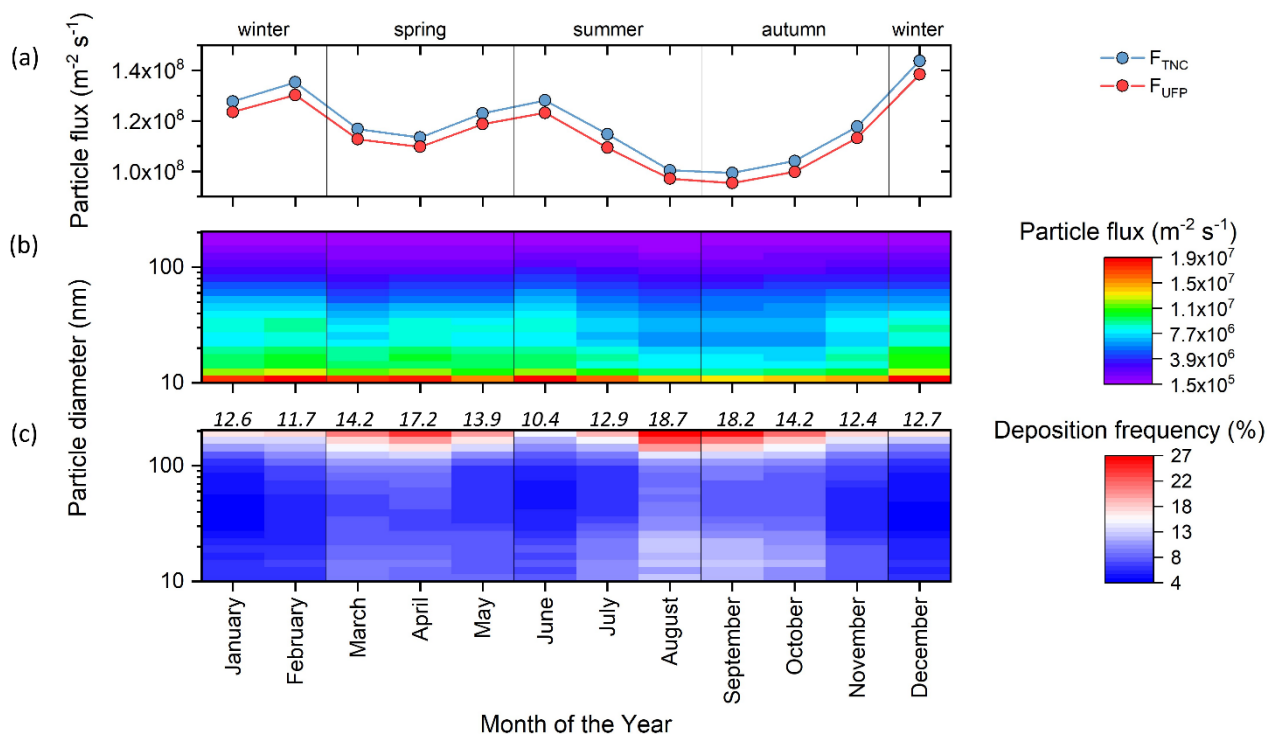



Figure 10: Monthly variation of (a) average F_{TNC} and F_{UFP}, (b) size-resolved average particle number fluxes, and (c) size-resolved frequency of particle deposition. In (c) the values given in italics quantify the maximum difference in deposition frequency within each monthly size spectrum (max. Δ deposition frequency, given in percentage points).

3.7 Wind sector analysis of modal particle number fluxes

The wind sector-based analysis indicated net emission of particles from every wind sector (Figure 11a). The largest average F_{TNC} were observed for westerly and northerly wind directions (W-N-NE), i.e. the direction in which "Straße des 17. Juni" and the traffic circle "Ernst-Reuter-Platz" are located (cf. Figure 1). In contrast, particle fluxes from easterly and southerly wind directions were smaller. Vegetated greenspaces, built-up areas but also the main road "Hardenbergstraße" were located in this direction. For all wind directions highest particle number fluxes prevailed in NUC mode (57-72 %), followed by particles in AIT (26-38 %) and ACC modes (2-5 %, Figure 11b).

When net fluxes were distinguished into emission and deposition fluxes, the emission fluxes indicated similar behaviour such as the average F_{TNC} due to the high frequency of emission events (cf. chapter 3.5, Figure 12a). Strongest emission fluxes occurred for wind from W to NE whereas the most frequent emission fluxes were evident for wind from 180 to 315° (Figure 12c). Strongest deposition fluxes prevailed under north-westerly wind whereas lower deposition was evident for the remaining directions (Figure 12b). However, the most frequent deposition fluxes were observed under winds from southerly and north-

westerly directions (Figure 12d). Since particle deposition occurred most often during night (cf. Figure 9b), the wind direction related frequency distribution of deposition fluxes (Figure 12d) was similar to the night time wind rose (not shown here) which was characterised by most frequent winds from southerly and north-westerly directions. The wind direction related frequency distribution of emission fluxes, however, was similar to the daytime wind rose (not shown here). Generally, ACC particles were more frequently characterised by deposition fluxes than other particle modes.

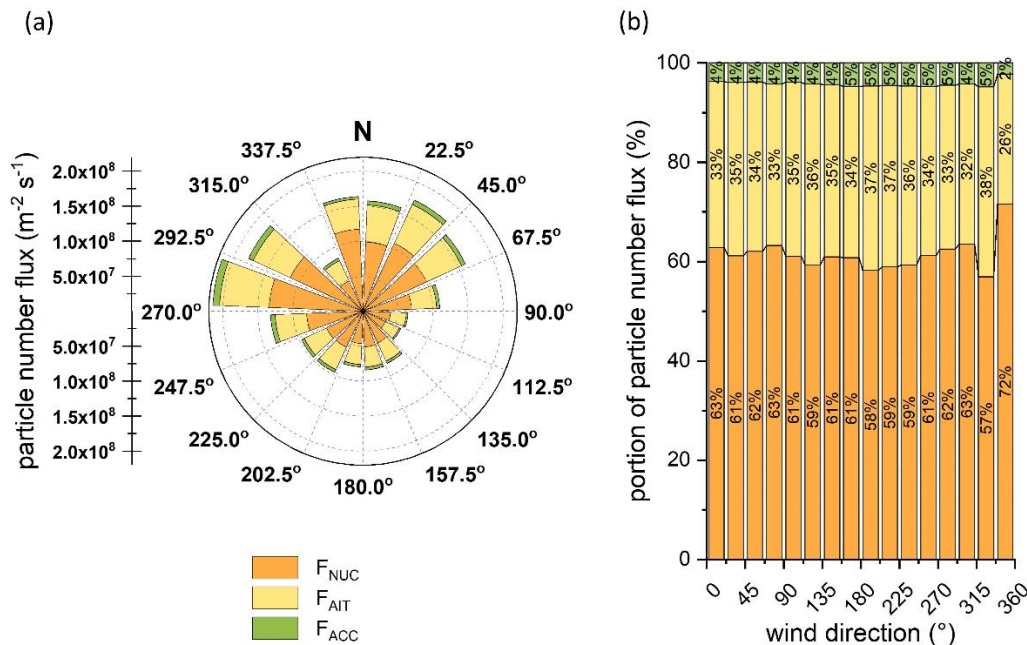



Figure 11: (a) Average particle number fluxes and (b) fraction of specific particle mode fluxes binned into 22.5° sectors for F_{NUC}, F_{AIT}, and F_{ACC}.

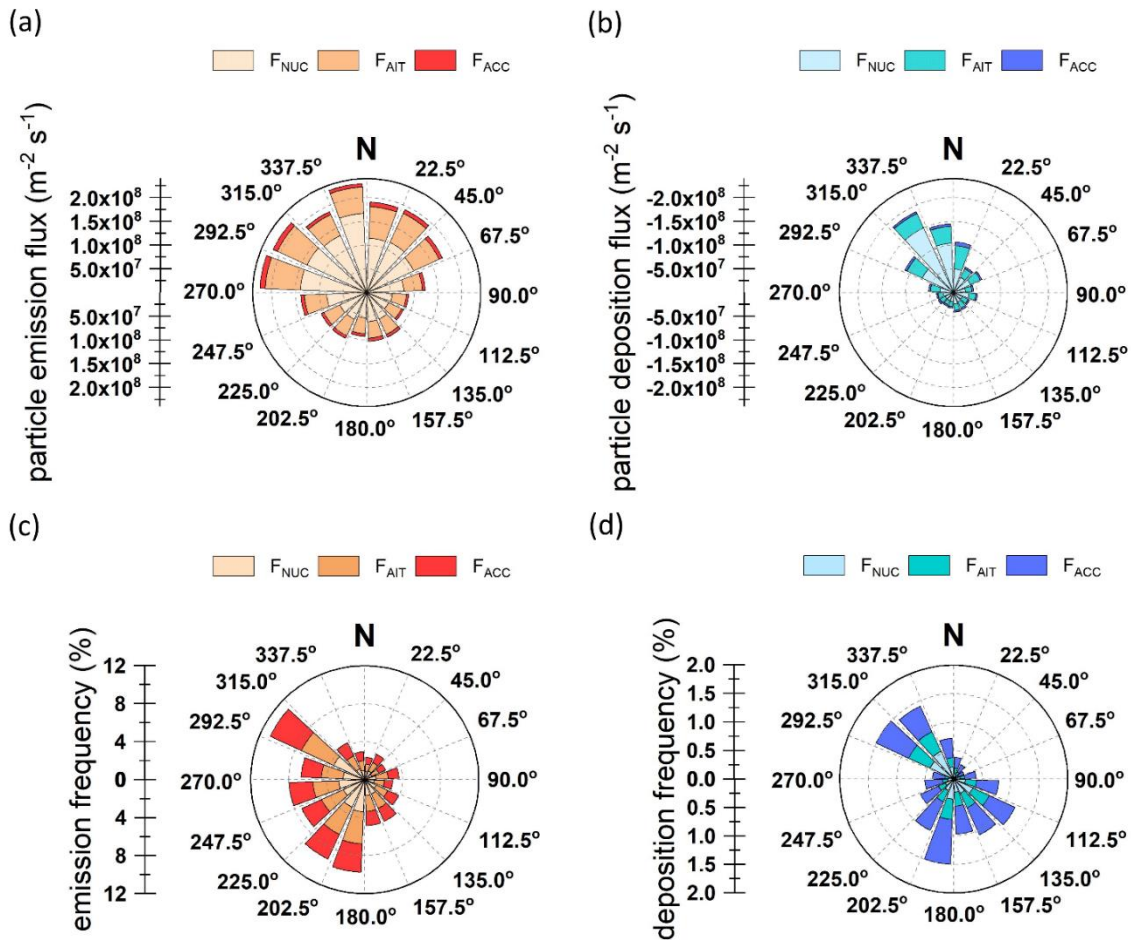



Figure 12: Average (a) emission and (b) deposition particle number fluxes as well as the frequencies of (c) emission and (d) deposition events per wind direction sector for F_{NUC}, F_{AIT}, and F_{ACC}.

3.8 Land-use regression analysis

The three-year data set was analysed with regard to variation of land-use across the flux footprint (cf. Figure 13). We observed the sectors with the highest fraction of traffic areas to match with the highest particle fluxes, e.g. the sector with the highest traffic fraction of 53 % resulted in the largest particle emission flux with 2.05×10^8 m^{-2} s^{-1}.

To further analyse this relationship, a LUR analysis was carried out for F_{TNC} and F_{UFP} in the 16 wind sectors. At a confidence level of 95 %, only the land-use type "traffic areas" was highly significant with a p-value < 0.001 (Table 4).

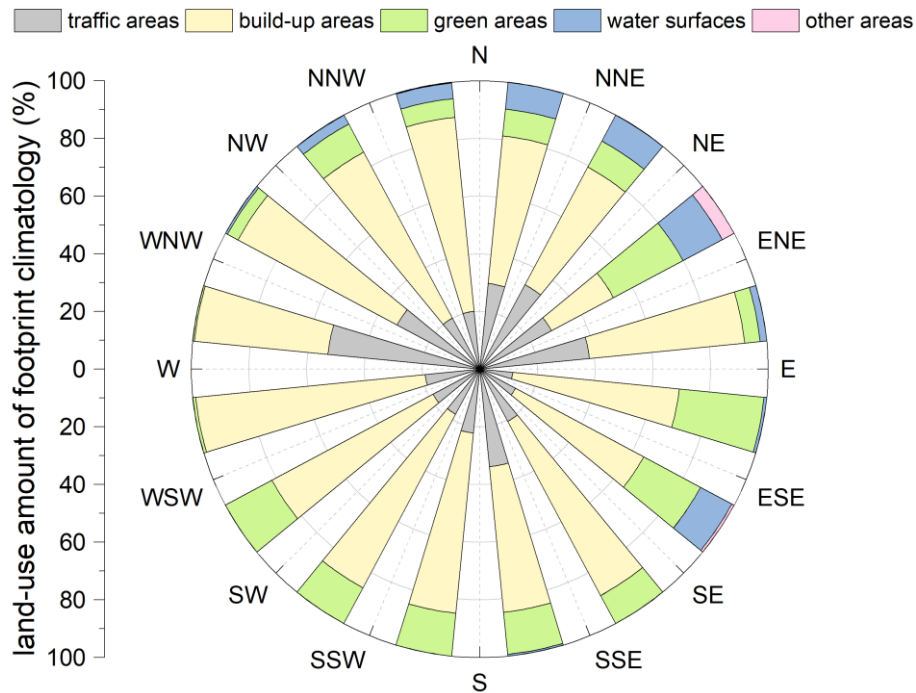



Figure 13: Fraction of land-use per wind direction sector contributing to the three-year flux footprint. For data analysis the flux footprint was defined as a circle centred on the flux site with a radius of 2 km.

The LUR analysis based on the three-year data set gives further evidence for the important role of traffic as UFP emission source. Traffic areas emitted $F_{UFP} = 3.44 \times 10^8$ m^{-2} s^{-1}, followed by built-up areas with 4.87×10^7 m^{-2} s^{-1} (not significant, Table 4). Green areas seem to be a sink of particles with $F_{UFP} = -1.46 \times 10^7$ m^{-2} s^{-1} although the relationship was statistically not significant. However, the analysis does not take temporal variation of traffic intensity into account. A comparison of hourly traffic data from counting stations with particle number fluxes showed a distinct increase in F_{UFP} with intensifying traffic (Figure 14). The strongest increase was obvious for F_{NUC}, followed by F_{AIT}. F_{ACC} only slightly increased with traffic intensity.

Table 4: Coefficients of the LUR analysis concerning F_{TNC} and F_{UFP} in the different wind sectors. Significant variables (significance level p < 0.001) are highlighted (*).

	F_{TNC} (m^{-2} s^{-1})	Standard error F_{TNC} (m^{-2} s^{-1})	P-value	Lower 95 % (m^{-2} s^{-1})	Upper 95 % (m^{-2} s^{-1})
Intersect	0	-	-	-	-
Green areas	-1.58×10^7	9.69×10^7	0.87	-2.25×10^8	1.94×10^8
Built-up areas	5.07×10^7	3.73×10^7	0.20	-2.98×10^7	1.31×10^8
Traffic area	3.59×10^8	6.87×10^7	0.16×10^{-3} *	2.10×10^8	5.08×10^8
Adjusted R²					0.85

330

340

	F_{UFP} (m⁻² s⁻¹)	Standard error F_{UFP} (m⁻² s⁻¹)	P-value	Lower 95 % (m⁻² s⁻¹)	Upper 95 % (m⁻² s⁻¹)
Intersect	0	-	-	-	-
Green areas	-1.46×10^7	9.44×10^7	0.88	-2.18×10^8	1.89×10^8
Built-up areas	4.87×10^7	3.63×10^7	0.20	-2.97×10^7	1.27×10^8
Traffic area	3.44×10^8	6.69×10^7	0.19×10^{-3} *	1.99×10^8	4.88×10^8
Adjusted R²					0.84

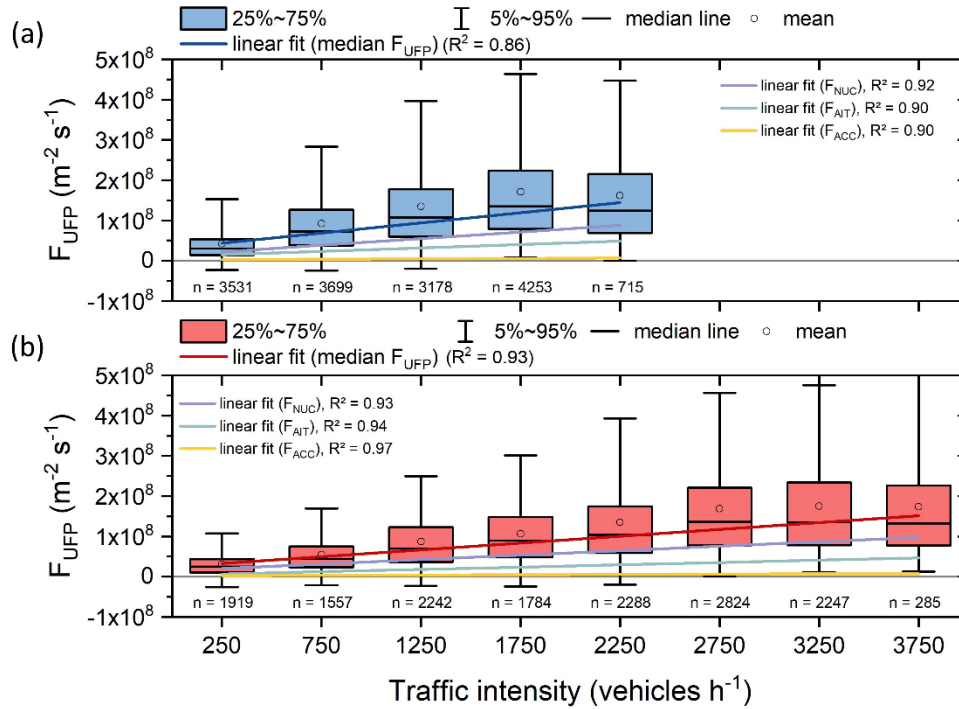



Figure 14: Relationship between F_{UFP} and traffic intensity at (a) "Hardenbergstraße" (HS) and (b) "Straße des 17. Juni" (S17J). Linear regression lines for the particle mode fluxes F_{NUC}, F_{AIT}, and F_{ACC} are given.

4. Discussion

The particle number fluxes observed at the Berlin site show dominance of particle emission with a three-year median F_{TNC} of 0.86×10^8 m⁻² s⁻¹ and F_{UFP} of 0.83×10^8 m⁻² s⁻¹ which makes the site a net emission source of ultrafine particles to the urban boundary layer. Urban particle number flux measurements from different locations are, however, not easy to compare given the different diameter cut-offs of particle counting instruments. Due to the dominating impact of UFP on the number flux, the lower cut-off is of particular importance. However, our results coincide with other studies from cities across Europe using a similar lower cut-off in the particle counting instruments (cf. Table 1). Average total number fluxes in Stockholm (Mårtensson et al., 2006), Innsbruck (Deventer et al., 2018; Heyden et al., 2018) and Lecce (Conte et al., 2018) are in the same order of magnitude varying between 0.5 and 3.0×10^8 m⁻² s⁻¹.

355   Although the three observation years point to constant emission of ultrafine particles from the urban surface, annual estimates show a median F_{UFP} decrease of -8.9 % from the first to the second and -6.1 % from the second to the third observation year. The observed decrease might be due to several reasons. Analysing the temporal variation of traffic count data, we found a slight reduction in the third year (Figure A3b). This could be one reason for the lower median F_{UFP} in the third observation year. In addition, the decrease might be associated to the slight spatial shift of the peak contribution location in the flux footprint

360   towards southerly directions. These source areas are characterised by a lower fraction of traffic areas and reduced traffic intensity in comparison to northerly directions. In the second year, the average wind direction and also the peak contribution location in the flux footprint shifted towards north-westerly directions. These were related to stronger particle deposition fluxes (cf. Figure 12) that tend to reduce F_{UFP}. As traffic areas in the flux footprint have different traffic intensity (i.e. major and minor roads), the annual variation of surface fractions of traffic areas in the flux footprints might not be sufficient to explain

365   particle number flux differences. However, the variation of traffic impact can be estimated using a footprint-weighted ADT, which estimates annual variation of traffic influence by quantifying the amount of major and minor roads contributing to the annual flux footprint (Figure A3a). This indicates a relationship between particle fluxes and footprint-weighted traffic contribution, which we assume to be the major reason for the observed variation in annual average particle fluxes. Additionally, the transition of the vehicle fleet towards a higher share of low-emission standards or electric drive might play a role. According

370   to official car registration data, the proportion of electric vehicles in Berlin increased by a factor of three in the time period from 2017 to 2020 (Kraftfahrt-Bundesamt, 2021). In the third observation year lower traffic intensity due to the COVID-19 lockdown restrictions starting in mid-March 2020 might have influenced our measurements and may explain the local minimum fluxes in March 2020 (cf. Figure 7). However, an in-depth analysis of effects related to reduced traffic intensity during the COVID-19 lockdown is beyond the scope of this study but will be addressed in future work.

375   The analysis of size-resolved particle number fluxes points to the dominance of emission fluxes over the diurnal cycle. The strongest emission fluxes occur during daytime with a maximum of smallest particles in the early afternoon (e.g. Schmidt and Klemm, 2008; Deventer et al., 2018). The diurnal courses of particle fluxes coincided with traffic intensity demonstrating strong influence of traffic on F_{NUC}. The delayed increase and reduced amplitude of traffic intensity weekends compared to a working day was clearly evident in particle flux data. At this time of low traffic intensity, particle deposition occurred much

380   more frequently. While coarse particles ($D_p > 180$ nm) were deposited with a frequency > 20 % at any time throughout the day, UFP deposition preferably occurred in the second half of the night or other low-emission periods, e.g. weekend mornings. Simultaneous occurrence of upward and downward directed fluxes, i.e. bidirectional fluxes, was recently reported from other cities (Schmidt and Klemm, 2008; Deventer et al., 2013, 2015, 2018). A tipping point, that defines the particle diameter separating average emission from average deposition fluxes, was reported for $160 < D_p < 190$ nm (Deventer et al., 2013, 2015).

385   On the mean diurnal cycle no such tipping point can be identified at the present site, since our data does not show simultaneously occurring average upward and downward fluxes across the size spectrum.

   We found a dominating contribution of F_{UFP} to the total flux (96 %) which corresponds to findings from Innsbruck (99 %; size range: 6 nm < D_p < 637 nm; Deventer et al., 2018) and London (Harrison et al., 2012). In addition, the diurnal course of F_{UFP}

showing a nocturnal minimum and an early afternoon maximum that is driven by traffic intensity and atmospheric stability is comparable to other findings (Dorsey et al., 2002; Deventer et al., 2013; Conte et al., 2018). A typical double rush-hour pattern of traffic intensity that is often not occurring in particle flux data (Dorsey et al., 2002; Järvi et al., 2009; Martin et al., 2009; Deventer et al., 2015) is slightly evident in the diurnal cycle of average F_{UFP} in Berlin. Also the differences in the diurnal cycles of F_{UFP} between working and weekend days point to dominating influence of traffic on the measured fluxes at our site. This was confirmed by wind-dependent analysis of flux variation, the regression of fluxes against land-use in the flux source area (LUR analysis) and the temporal variation with traffic data from counting stations. Similar findings of higher fluxes from traffic areas and lower fluxes from areas with higher amounts of green surfaces and buildings were reported by Järvi et al. (2009), Ripamonti et al. (2013), and Mårtensson et al. (2006).

In each wind direction sector, we found the highest particle number fluxes in nucleation mode ($F_{NUC} \geq 57$ % of average F_{TNC}, cf. Figure 11b). Road traffic is the main source of particles in the size range below 50 nm (Morawska et al., 2008). Especially rush-hour periods are characterised by a maximum of traffic-related particle emission in the size range between 20 and 30 nm (Wiedensohler et al., 2002). Thus, the large contribution of F_{NUC} confirms the high influence of traffic intensity on fluxes at this site.

The deposition frequency seems to be related to the strength of local sources as indicated by the increase of deposition frequency with decreasing F_{UFP} and F_{TNC} (cf. Figure 10). This may also be the reason for the higher frequency of particle deposition in ACC than UFP, since ACC emission fluxes are significantly lower than AIT and NUC mode fluxes (cf. Figure 9a). The occurrence of bidirectional fluxes in the annual course also seems to depend on the strength of the surrounding sources with a more frequent appearance in seasons with lower emissions. Thus, they occur more often in summer and autumn than in winter (cf. Figure 10c). Deventer et al. (2015) also reported differences in particle deposition between summer and winter and found an increase of deposition fluxes for $D_p > 170$ nm in spring and summer.

The highest particle number fluxes were observed in winter (cf. Figure 10a) due to a larger number of emission sources (e.g. domestic heating) and reduced atmospheric dilution. However, the LUR analysis illustrates that the relationship between sources from built-up areas and particle number fluxes is statistically not significant and that emission fluxes from built-up areas are one order of magnitude smaller than traffic area fluxes (cf. Table 4). Thus, other local emission sources such as domestic heating in winter are of limited influence for particle number fluxes at this site in contrast to traffic related sources.

5 Conclusions

We report on the first multi-annual data set of urban size-resolved particle number fluxes that were measured over the size range $10 < D_p < 200$ nm from April 2017 to March 2020 in an urban area of Berlin, Germany, using the eddy-covariance technique. The Berlin site was a net source of particles with a majority of ultrafine particle emission as indicated by a F_{UFP}/F_{TNC} ratio of 0.96. The magnitude and temporal variation of F_{TNC} and F_{UFP} were rather similar in the individual observation years, however, a reduction of median F_{UFP} of -8.9 % and -6.1 % for the second and third years was evident. This might be related to

variation in flux source area, a modification in the vehicle fleet, and an effect of reduced traffic intensity in the third measurement year.

The results clearly point to traffic as the dominant influence on ultrafine particle number fluxes at the Berlin site. Due to the health effects associated with UFP, measures that result in reduced traffic intensity, e.g. by promoting alternative urban mobility concepts, or a transition of the traffic fleet towards a higher share of electric or low-emission vehicles may help to reduce personal exposure towards UFP. Long-term observations of size-resolved particle number fluxes from different urban environments would be an important tool to monitor and asses the success of different mobility strategies and reduction measures.

430

Appendix A

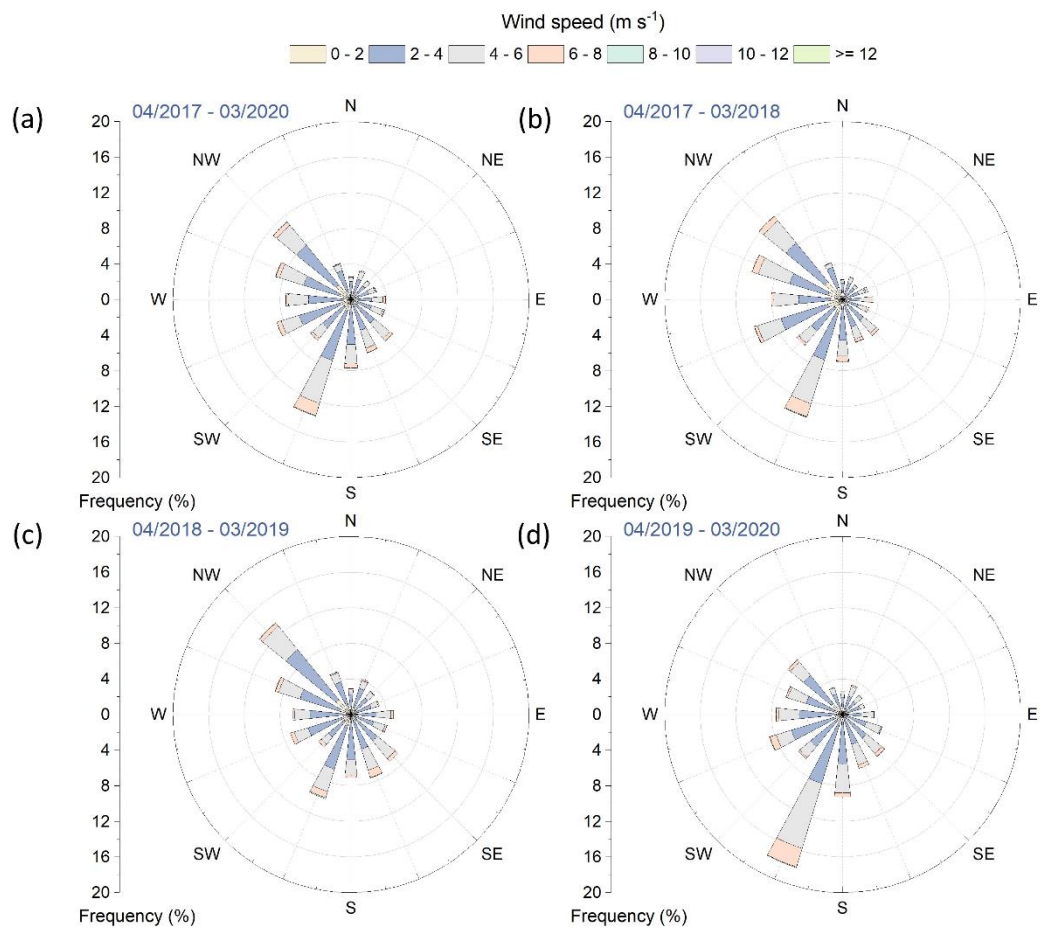
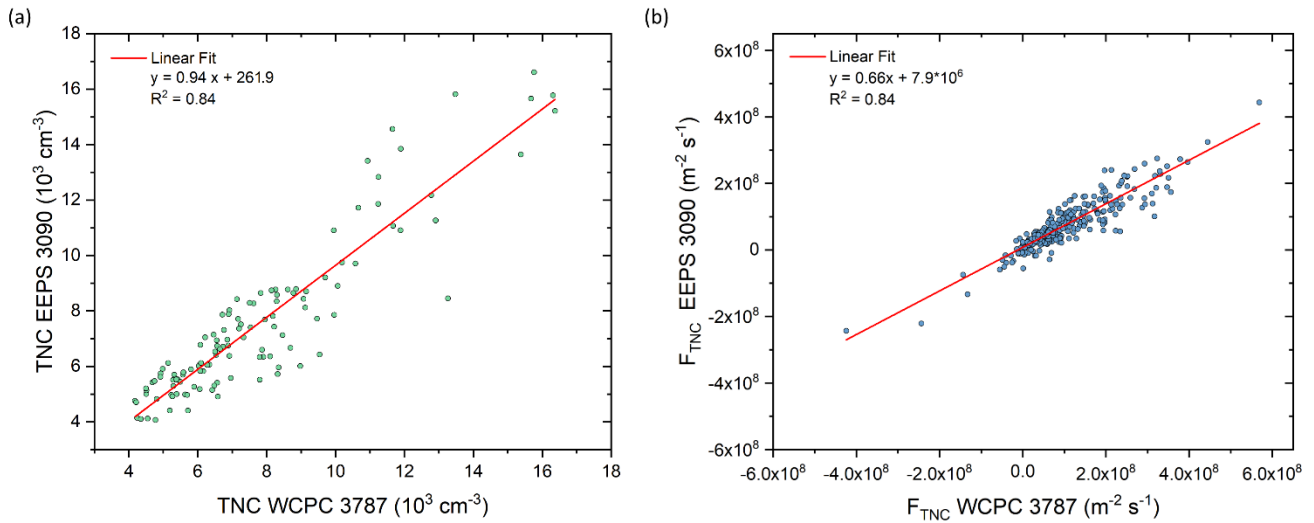
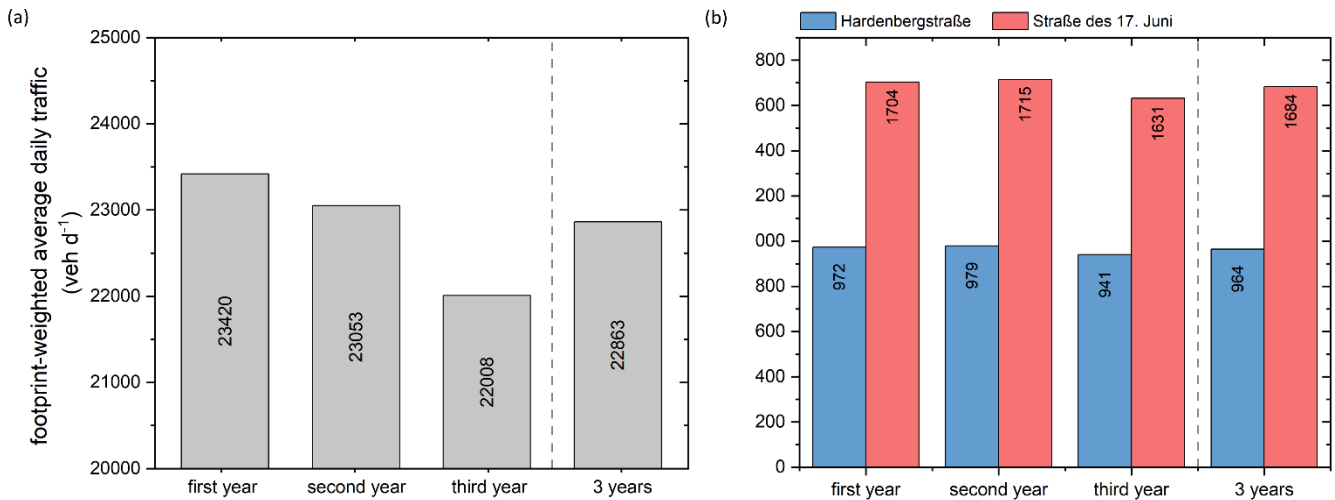



Figure A1: Wind roses for (a) the whole three-year period, (b) the first year, (c) the second year, and (d) the third year.

435 **Figure A2: Measurement comparison of EEPS 3090 with WCPC 3787 at the Berlin site concerning (a) total particle number concentration (TNC, 4 days, July 2020) and (b) total particle number flux (F_{TNC}, 12 days, September 2020). The data of EEPS relate to particle sizes between 10 and 200 nm whereas the WCPC counts particles with $D_p > 5$ nm.**

440 **Figure A3: (a) Footprint-weighted average daily traffic intensity (data source: Umweltatlas, 2017) and (b) average traffic intensity measured at the two traffic counting stations "Hardenbergstraße" (HS) and "Straße des 17. Juni" (S17J) for the entire measurement period and each observation year.**

Data availability

445 The data used in this study are publicly available: "Three years of size-resolved eddy-covariance particle number flux measurements in Berlin, Germany" on https://publikationsserver.tu-braunschweig.de/receive/dbbs_mods_00069675.

Author contribution

SW acted as the supervisor with responsibilities concerning conceptualisation, funding acquisition, project administration as well as reviewing and editing of the paper draft. Together with AS, he designed the experiments. AS took care of data curation,
450 maintaining the flux site, formal data analysis, measurement operation and data collection, methodology, software use and development for data analysis as well as data visualisation. Additionally, AS prepared the original draft of this paper.

Competing interests

The authors declare that they have no conflict of interest.

Acknowledgements

455 This study was supported by the German Federal Ministry of Education and Research (BMBF) under Grants FKZ 01 LP 1602 D (Urban Climate under Change, Module 3DO) and FKZ 01 LP 1912 D (Urban Climate under Change, Phase II, Module 3DO+M). We would like to thank Andreas Müer from VMZ Berlin Betreibergesellschaft mbH for providing traffic data from two counting stations. Furthermore, we would like to thank our project partners (Fred Meier, Dieter Scherer) of the TU Berlin, Institute for Ecology, Chair of Climatology, for providing hourly measurement data of mixing layer height for flux footprint
460 prediction and for being able to install sensors at their site. We thank Hagen Mittendorf, Technische Universität Braunschweig, for his support in setting up and maintaining the flux site.

References

Aubinet, M.; Vesala, T.; Papale, D. (Eds.): Eddy Covariance. A Practical Guide to Measurement and Data Analysis, Springer Netherlands, Dordrecht, doi: 10.1007/978-94-007-2351-1, 2012.
465 Bäfver, L. S.; Leckner, B.; Tullin, C.; Berntsen, M.: Particle emissions from pellets stoves and modern and old-type wood stoves, Biomass Bioenergy, 35 (8), 3648–3655, doi: 10.1016/j.biombioe.2011.05.027, 2011.
Baldocchi, D.: Assessing the eddy covariance technique for evaluating carbon dioxide exchange rates of ecosystems: past, present and future, Glob. Chang. Biol., (9), 479–492, 2003.

Burba, G.; Anderson, D.: A brief practical Guide to Eddy Covariance Flux Measurements. Principles and workflow examples

470   for scientific and industrial applications, LI-COR, Inc., 2005.

Buzorius, G.: Cut-Off Sizes and Time Constants of the CPC TSI 3010 Operating at 1-3 lpm Flow Rates, Aerosol Sci. Technol., 35 (1), 577–585, doi: 10.1080/02786820121505, 2001.

Buzorius, G.; Rannik, Ü.; Mäkelä, J. M.; Vesala, T.; Kulmala, M.: Vertical aerosol particle fluxes measured by eddy covariance technique using condensational particle counter, J. Aerosol Sci., 29 (1-2), 157–171, doi: 10.1016/S0021-8502(97)00458-8,

475   1998.

Buzorius, G.; Rannik, Ü.; Mäkelä, J. M.; Keronen, P.; Vesala, T.; Kulmala, M.: Vertical aerosol fluxes measured by the eddy covariance method and deposition of nucleation mode particles above a Scots pine forest in southern Finland, J. Geophys. Res., 105 (D15), 19905–19916, doi: 10.1029/2000JD900108, 2000.

Conte, M.; Donateo, A.; Contini, D.: Characterisation of particle size distributions and corresponding size-segregated turbulent

480   fluxes simultaneously with CO_2 exchange in an urban area, Sci. Total Environ., 622-623, 1067–1078, doi: 10.1016/j.scitotenv.2017.12.040, 2018.

Contini, D.; Donateo, A.; Elefante, C.; Grasso, F. M.: Analysis of particles and carbon dioxide concentrations and fluxes in an urban area. Correlation with traffic rate and local micrometeorology, Atmospheric Environ., 46, 25–35, doi: 10.1016/j.atmosenv.2011.10.039, 2012.

485   Damay, P. E.; Maro, D.; Coppalle, A.; Lamaud, E.; Connan, O.; Hébert, D.; Talbaut, M.; Irvine, M.: Size-resolved eddy covariance measurements of fine particle vertical fluxes, J. Aerosol Sci., 40 (12), 1050–1058, doi: 10.1016/j.jaerosci.2009.09.010, 2009.

Deventer, M. J.; El-Madany, T.; Griessbaum, F.; Klemm, O.: One-year measurement of size-resolved particle fluxes in an urban area, Tellus B Chem. Phys. Meteorol., 67 (1), 25531, doi: 10.3402/tellusb.v67.25531, 2015.

490   Deventer, M. J.; Griessbaum, F.; Klemm, O.: Size-resolved flux measurement of sub-micrometer particles over an urban area, metz, 22 (6), 729–737, doi: 10.1127/0941-2948/2013/0441, 2013.

Deventer, M. J.; Heyden, L. von der; Lamprecht, C.; Graus, M.; Karl, T.; Held, A.: Aerosol particles during the Innsbruck Air Quality Study (INNAQS). Fluxes of nucleation to accumulation mode particles in relation to selective urban tracers, Atmospheric Environ., 190, 376–388, doi: 10.1016/j.atmosenv.2018.04.043, 2018.

495   Dorsey, J. R.; Nemitz, E.; Gallagher, M. W.; Fowler, D.; Williams, P. I.; Bower, K. N.; Beswick, K. M.: Direct measurements and parameterisation of aerosol flux, concentration and emission velocity above a city, Atmospheric Environ., 36 (5), 791–800, doi: 10.1016/S1352-2310(01)00526-X, 2002.

Farmer, D. K.; Boedicker, E. K.; DeBolt, H. M.: Dry Deposition of Atmospheric Aerosols. Approaches, Observations, and Mechanisms, Annu. Rev. Phys. Chem., 72, 375–397, doi: 10.1146/annurev-physchem-090519-034936, 2021.

500   Finkelstein, P. L.; Sims, P. F.: Sampling error in eddy correlation flux measurements, J. Geophys. Res., 106 (D4), 3503–3509, doi: 10.1029/2000JD900731, 2001.

Foken, T.; Göockede, M.; Mauder, M.; Mahrt, L.; Amiro, B.; Munger, W.: Post-Field Data Quality Control, in: Handbook of Micrometeorology, edited by: Lee, X., Massman, W., Law, B., Atmos. Ocean. Sci. Lib, 29, Springer, Dordrecht, doi: https://doi.org/10.1007/1-4020-2265-4_9, 2004.

505 Geoportal Berlin: Biotoptypen (Umweltatlas), license „dl-de/by-2-0" (www.govdata.de/dl-de/by-2-0), URL: https://fbinter.stadt-berlin.de/fb/berlin/service_intern.jsp?id=s_fb_berlinbtk@senstadt&type=WFS, last access: 20 May 2021, 2014.

Geoportal Berlin: ALKIS Berlin Gebäude, license „dl-de/by-2-0" (www.govdata.de/dl-de/by-2-0), URL: https://fbinter.stadt-berlin.de/fb/berlin/service_intern.jsp?id=s_wfs_alkis_gebaeudeflaechen@senstadt&type=WFS, last access: 20 May 2021,

510 2021.

Gerling, L.; Löschau, G.; Wiedensohler, A.; Weber, S.: Statistical modelling of roadside and urban background ultrafine and accumulation mode particle number concentrations using generalized additive models, Sci. Total Environ., 703, 134570, doi: 10.1016/j.scitotenv.2019.134570, 2020.

Gioli, B.; Toscano, P.; Zaldei, A.; Fratini, G.; Miglietta, F.: CO2, CH4 and Particles Flux Measurements in Florence, Italy,

515 Enrgy Proced., 40, 537–544, doi: 10.1016/j.egypro.2013.08.062, 2013.

Harrison, R. M.; Dall'Osto, M.; Beddows, D. C. S.; Thorpe, A. J.; Bloss, W. J.; Allan, J. D.; Coe, H.; Dorsey, J. R.; Gallagher, M.; Martin, C.; Whitehead, J.; Williams, P. I.; Jones, R. L.; Langridge, J. M.; Benton, A. K.; Ball, S. M.; Langford, B.; Hewitt, C. N.; Davison, B.; Martin, D.; Petersson, K. F.; Henshaw, S. J.; White, I. R.; Shallcross, D. E.; Barlow, J. F.; Dunbar, T.; Davies, F.; Nemitz, E.; Phillips, G. J.; Helfter, C.; Di Marco, C. F.; Smith, S.: Atmospheric chemistry and physics in the

520 atmosphere of a developed megacity (London). An overview of the REPARTEE experiment and its conclusions, Atmos. Chem. Phys., 12 (6), 3065–3114, doi: 10.5194/acp-12-3065-2012, 2012.

HEI Review Panel on Ultrafine Particles: Understanding the Health Effects of Ambient Ultrafine Particles. HEI Perspectives 3. Health Effects Institute, Boston, MA, 2013.

Heyden, L. von der; Deventer, M. J.; Graus, M.; Karl, T.; Lamprecht, C.; Held, A.: Aerosol particles during the Innsbruck Air

525 Quality Study (INNAQS). The impact of transient fluxes on total aerosol number exchange, Atmospheric Environ., 190, 389–400, doi: 10.1016/j.atmosenv.2018.07.041, 2018.

Hinds, W. C.: Aerosol technology. Properties, behavior, and measurement of airborne particles, 2. ed., Wiley, New York, 1999.

Horst, T. W.: A simple formula for attenuation of eddy fluxes measured with first-order-response scalar sensors, Boundary

530 Layer Meteorol., 82 (2), 219–233, doi: 10.1023/A:1000229130034, 1997.

Järvi, L.; Rannik, Ü.; Mammarella, I.; Sogachev, A.; Aalto, P. P.; Keronen, P.; Siivola, E.; Kulmala, M.; Vesala, T.: Annual particle flux observations over a heterogeneous urban area, Atmos. Chem. Phys. Discuss., 9 (3), 13407–13437, doi: 10.5194/acpd-9-13407-2009, 2009.

Johnson, T.; Caldow, R.; Pöcher, A.; Mirme, A.; Kittelson, D.: An Engine Exhaust Particle Sizer Spectrometer for Transient
Emission Particle Measurements, Conference: 9th Diesel Engine Emissions Reduction (DEER) Workshop 2003, Newport, RI
(US), 08/24/2003--08/28/2003, 2003.

Johnson, T.; Caldow, R.; Pöcher, A.; Mirme, A.; Kittelson, D.: A New Electrical Mobility Particle Sizer Spectrometer for
Engine Exhaust Particle Measurements, SAE Technical Paper Series 2004-01-1341 (SAE-SP, 1801), 2004.

Kaimal, J. C.; Wyngaard, J. C.; Izumi, Y.; Coté, O. R.: Spectral characteristics of surface-layer turbulence, Q. J. R. Meteorol.
Soc., 98 (417), 563–589, doi: 10.1002/qj.49709841707, 1972.

Kljun, N.; Calanca, P.; Rotach, M. W.; Schmid, H. P.: A simple two-dimensional parameterisation for Flux Footprint
Prediction (FFP), Geosci. Model Dev., 8 (11), 3695–3713, doi: 10.5194/gmd-8-3695-2015, 2015.

Kraftfahrt-Bundesamt: Anzahl der Personenkraftwagen mit Elektroantrieb im Bundesland Berlin im Zeitraum der Jahre 2012
bis 2020, cited from de.statista.com, URL: https://de.statista.com/statistik/daten/studie/709212/umfrage/personenkraftwagen-
mit-elektroantrieb-in-berlin/, last access: 21 July 2021, 2021.

Kumar, P.; Morawska, L.; Birmili, W.; Paasonen, P.; Hu, M.; Kulmala, M.; Harrison, R. M.; Norford, L.; Britter, R.: Ultrafine
particles in cities, Environ. Int., 66, 1–10, doi: 10.1016/j.envint.2014.01.013, 2014.

Kurppa, M.; Nordbo, A.; Haapanala, S. Järvi, L.: Effect of seasonal variability and land use on particle number and CO2
exchange in Helsinki, Finland, Urban Clim., 13, 94–109, doi: 10.1016/j.uclim.2015.07.006, 2015.

Longley, I. D.; Gallagher, M. W.; Dorsey, J. R.; Flynn, M.; Allan, J. D.; Alfarra, M. R.; Inglis, D.: A case study of aerosol
(4.6nm<Dp<10μm) number and mass size distribution measurements in a busy street canyon in Manchester, UK, Atmospheric
Environ., 37 (12), 1563–1571, doi: 10.1016/S1352-2310(03)00010-4, 2003.

Longley, I. D.; Gallagher, M. W.; Dorsey, J. R.; Flynn, M.: A case-study of fine particle concentrations and fluxes measured
in a busy street canyon in Manchester, UK, Atmospheric Environ., 38 (22), 3595–3603, doi: 10.1016/j.atmosenv.2004.03.040,
2004.

Mårtensson, E. M.; Nilsson, E. D.; Buzorius, G.; Johansson, C.: Eddy covariance measurements and parameterisation of traffic
related particle emissions in an urban environment, Atmos. Chem. Phys., 6 (3), 769–785, doi: 10.5194/acp-6-769-2006, 2006.

Martin, C. L.; Longley, I. D.; Dorsey, J. R.; Thomas, R. M.; Gallagher, M. W.; Nemitz, E.: Ultrafine particle fluxes above four
major European cities, Atmospheric Environ., 43 (31), 4714–4721, doi: 10.1016/j.atmosenv.2008.10.009, 2009.

Meyer-Kornblum, A.; Gerling, L.; Weber, S.: Gap-filling Fast Electrical Mobility Spectrometer Measurements of Particle
Number Size Distributions for Eddy Covariance Application, Aerosol Air Qual. Res., 19 (12), 2721–2731, doi:
10.4209/aaqr.2019.06.0291, 2019.

Moncrieff, J. B.; Massheder, J. M.; Bruin, H. de; Elbers, J.; Friborg, T.; Heusinkveld, B.; Kabat, P.; Scott, S.; Soegaard, H.;
Verhoef, A.: A system to measure surface fluxes of momentum, sensible heat, water vapour and carbon dioxide, J. Hydrol.,
188-189, 589–611, 1997.

Moncrieff, J.; Clement, R.; Finnigan, J.; Meyers, T.: Averaging, Detrending, and Filtering of Eddy Covariance Time Series, in: Handbook of Micrometeorology, edited by: Lee, X., Massman, W., Law, B., Atmos. Ocean. Sci. Lib., 29, Springer, Dordrecht, 2004.

570 Morawska, L.; Ristovski, Z.; Jayaratne, E. R.; Keogh, D. U.; Ling, X.: Ambient nano and ultrafine particles from motor vehicle emissions. Characteristics, ambient processing and implications on human exposure, Atmospheric Environ., 42 (35), 8113–8138, doi: 10.1016/j.atmosenv.2008.07.050, 2008.

Nemitz, E.; Fowler, D.; Dorsey, J. R.; Theobald, M. R.; McDonald, A. D.; Bower, K. N.; Beswick, K. M.; Williams, P. I.; Gallagher, M. W.: Direct measurements of size-segregated particle fluxes above a city, J. Aerosol Sci., 31, 116–117, doi: 10.1016/S0021-8502(00)90123-X, 2000.

575 Nemmar, A.; Hoet, P.H.M.; Vanquickenborne, B.; Dinsdale, D.; Thomeer, M.; Hoylaerts, M. F.; Vanbilloen, H.; Mortelmans, L.; Nemery, B.: Passage of Inhaled Particles Into the Blood Circulation in Humans, Circulation, 105 (4),411–414, doi: 10.1161/hc0402.104118, 2002.

Oberdörster, G.; Gelein, R. M.; Ferin, J.; Weiss, B.: Association of particulate air pollution and acute mortality. Involvement of ultrafine particles?, Inhal. Toxicol., 7 (1), 111–124, doi: 10.3109/08958379509014275, 1995.

580 Ripamonti, G.; Järvi, L.; Mølgaard, B.; Hussein, T.; Nordbo, A.; Hämeri, K.: The effect of local sources on aerosol particle number size distribution, concentrations and fluxes in Helsinki, Finland, Tellus B Chem. Phys. Meteorol., 65 (1), 19786, doi: 10.3402/tellusb.v65i0.19786, 2013.

Rönkkö, T.; Kuuluvainen, H.; Karjalainen, P.; Keskinen, J.; Hillamo, R.; Niemi, J. V.; Pirjola, L.; Timonen, H. J.; Saarikoski, S.; Saukko, E.; Järvinen, A.; Silvennoinen, H.; Rostedt, A.; Olin, M.; Yli-Ojanperä, J.; Nousiainen, P.; Kousa, A.; Dal Maso, 585 M.: Traffic is a major source of atmospheric nanocluster aerosol, P. Natl. Acad. Sci. USA, 114 (29), 7549–7554, doi: 10.1073/pnas.1700830114, 2017.

Saylor, R. D.; Baker, B. D.; Lee, P.; Tong, D.; Pan, L.; Hicks, B. B.: The particle dry deposition component of total deposition from air quality models. Right, wrong or uncertain?, Tellus B Chem. Phys. Meteorol., 71 (1), 1–22, doi: 10.1080/16000889.2018.1550324, 2019.

590 Scherer, D.; Ament, F.; Emeis, S.; Fehrenbach, U.; Leitl, B.; Scherber, K.; Schneider, C.; Vogt, U.: Three-Dimensional Observation of Atmospheric Processes in Cities, metz, 28 (2), 121-138, doi: 10.1127/metz/2019/0911, 2019.

Schmidt, A.; Klemm, O.: Direct determination of highly size-resolved turbulent particle fluxes with the disjunct eddy covariance method and a 12 - stage electrical low pressure impactor, Atmos. Chem. Phys., 8 (24), 7405–7417, doi: 10.5194/acp-8-7405-2008, 2008.

595 Schraufnagel, D. E.: The health effects of ultrafine particles, Exp. Mol. Med., 52 (3), 311–317, doi: 10.1038/s12276-020-0403-3, 2020.

Straaten, A.; Weber, S.: Three years of size-resolved eddy-covariance particle number flux measurements in Berlin, Germany, online available: https://publikationsserver.tu-braunschweig.de/receive/dbbs_mods_00069675, doi: 10.24355/dbbs.084-202107221337-0.

600 Suni, T.; Rinne, J.; Reissell, A.; Altimir, N.; Keronen, P.; Rannik, Ü.; Dal Maso, M.; Kulmala, M.; Vesala, T.: Long-term measurements of surface fluxes above a Scots pine forest in Hyytiälä, southern Finland, 1996-2001, Boreal Environ. Res., (8), 287-301, 2003.

Tie, X.; Wu, D.; Brasseur, G.: Lung cancer mortality and exposure to atmospheric aerosol particles in Guangzhou, China, Atmospheric Environ., 43 (14), 2375–2377, doi: 10.1016/j.atmosenv.2009.01.036, 2009.

605 Tsang, H.; Kwok, R.; Miguel, A. H.: Pedestrian Exposure to Ultrafine Particles in Hong Kong Under Heavy Traffic Conditions, Aerosol Air Qual. Res., 8 (1), 19–27, 2008.

TSI, Inc.: Engine Exhaust Particle Sizer Spectrometer Model 3090/3090AK. Operation and Service Manual, 2015.

Umweltatlas Berlin: Verkehrsmengen DTV 2014 (Umweltatlas), license „dl-de/by-2-0" (www.govdata.de/dl-de/by-2-0), URL: https://fbinter.stadt-berlin.de/fb/berlin/service_intern.jsp?id=wfs_07_01verkmeng2014@senstadt&type=WFS, last

610 access: 20 May 2021, 2017.

Vickers, D.; Mahrt, L.: Quality Control and Flux Sampling Problems for Tower and Aircraft Data, J. Atmos. Oceanic Technol., 14 (3), 512–526, doi: 10.1175/1520-0426(1997)014<0512:QCAFSP>2.0.CO;2, 1997.

Vogt, M.; Nilsson, E. D.; Ahlm, L.; Mårtensson, E. M.; Johansson, C.: Seasonal and diurnal cycles of 0.25–2.5 µm aerosol fluxes over urban Stockholm, Sweden, Tellus B Chem. Phys. Meteorol., 63 (5), 935–951, doi: 10.1111/j.1600-

615 0889.2011.00551.x, 2011a.

Vogt, M.; Nilsson, E. D.; Ahlm, L.; Mårtensson, E. M.; Struthers, H.; Johansson, C.: Traffic aerosol emission velocity derived from direct flux measurements over urban Stockholm, Sweden, Atmospheric Environ., 45 (32), 5725–5731, doi: 10.1016/j.atmosenv.2011.07.026, 2011b.

Weber, S.; Kordowski, K.; Kuttler, W.: Variability of particle number concentration and particle size dynamics in an urban

620 street canyon under different meteorological conditions, Sci. Total Environ., 449, 102–114, doi: 10.1016/j.scitotenv.2013.01.044, 2013.

Wiedensohler, A; Wehner, B.; Birmili, W.: Aerosol Number Concentrations and Size Distributions at Mountain-Rural, Urban-Influenced Rural, and Urban-Background Sites in Germany, J. Aerosol Med., 15 (2), 237–243, 2002.

Zhu, Y.; Eiguren-Fernandez, A.; Hinds, W. C.; Miguel, A. H.: In-cabin commuter exposure to ultrafine particles on Los

625 Angeles freeways, Environ. Sci. Technol., 41 (7), 2138–2145, doi: 10.1021/es0618797, 2007.