Long-range transport of anthropogenic air pollutants into the marine air: Insight into fine particle transport and chloride depletion on sea salts

Liang Xu¹, Xiaohuan Liu², Huiwang Gao², Xiaohong Yao², Daizhou Zhang¹, Lei Bi¹, Lei Liu¹,
Jian Zhang¹, Yinxiao Zhang¹, Yuanyuan Wang¹, Qi Yuan², Weijun Li¹,*

¹Department of Atmospheric Sciences, School of Earth Sciences, Zhejiang University, Hangzhou 310027, China
²Key Laboratory of Marine Environment and Ecology, Ministry of Education, Ocean University of China, Qingdao 266100, China
³Faculty of Environmental and Symbiotic Sciences, Prefectural University of Kumamoto, Kumamoto 862-8502, Japan

*Corresponding author: W. Li (liweijun@zju.edu.cn)

Abstract

Long-range transport of anthropogenic air pollutants from East Asia can affect the downwind marine air quality during spring and winter. Long-range transport of continental air pollutants and their interaction with sea salt aerosols (SSA) significantly modify the radiative forcing of marine aerosols and influence ocean biogeochemical cycling. Previous studies poorly characterize variations of aerosol particles along with air mass transport from the continental edge to the remote ocean. Here, the research ship R/V Dongfanghong 2 traveled from the eastern China seas
(ECS) to the northwestern Pacific Ocean (NWPO) to understand what and how air pollutants were transported from the highly polluted continental air to clean marine air in spring. A transmission electron microscope (TEM) was used to find the long-range transported anthropogenic particles and the possible Cl-depletion phenomenon of SSA in marine air. Primary and secondary anthropogenic aerosols were identified and dramatically declined from 87% to 8% by number from the ECS to remote NWPO. For the SSA aging, 86% of SSA particles in the ECS were identified as fully aged, while the proportion of fully aged SSA particles in the NWPO decreased to 31%. The result highlights that anthropogenic acidic gases in the troposphere (e.g., SO$_2$, NO$_x$, and volatile organic compounds) were transported longer distances compared to the anthropogenic aerosol and could exert a significant impact on marine aerosols in the NWPO. These results show that anthropogenic particles and gases from East Asia significantly perturb aerosol chemistry in marine air. The optical properties and cloud condensation nucleation of the modified SSA particles should be incorporated into the more accurately modeling of clouds in the ECS and NWPO in spring and winter.
1. Introduction

Marine aerosols play an important role in the global aerosol emission budget and greatly impact the Earth’s radiative forcing and biogeochemical cycling (ODowd Colin and de Leeuw, 2007). Sea salt aerosol (SSA) is one crucial component of marine aerosols, especially in the remote marine atmosphere (Lewis and Schwartz, 2004). SSA is composed of Na and Cl with minor amounts of Mg, Ca, K, and S (Li et al., 2016a). Fresh SSA usually has the form of a cubic NaCl core associated with MgCl$_2$ and CaSO$_4$ coating (Chi et al., 2015). SSA in polluted air can serve as reactive surfaces through the uptake of acidic gaseous SO$_2$, NO$_x$, and organic acids, releasing gaseous reactive chlorine compounds (chloride depletion) (Laskin et al., 2012; Yao and Zhang, 2012). The chloride depletion further releases Cl- and I-containing compounds into the air which could drive marine new particle formation (He et al., 2021; Yu et al., 2019). Moreover, the SSA aging processes can transform fresh SSA into partially aged SSA and finally into fully aged SSA that mainly contain NaNO$_3$, Na$_2$SO$_4$, and organic sodium salts (Chi et al., 2015; Laskin et al., 2012). These products from the chemical aging processes also modify hygroscopic properties of individual SSA (Cravigan et al., 2020; Ghorai et al., 2014). Some studies found that the aged SSA could alter global climate directly by scattering incoming solar radiation or indirectly by acting as cloud condensation nuclei (CCN) or ice nuclei (IN) in marine air (Hu et al., 2005; Murphy et al., 1998; Pierce and Adams, 2006).

It is well known that the interaction of the continental-marine air masses not only releases some active substances into marine air but also supplies many nutrients (e.g.,
Fe, N, and P) for biological growth (e.g., plankton) to the ocean’s surface (Li et al., 2017; Shi et al., 2012). Some studies have reported that the continental anthropogenic and natural pollutants can be carried to remote marine air through long-range atmospheric transport (Guo et al., 2014; Li et al., 2017; Moffet et al., 2012; Uematsu et al., 2010). The continental pollutants that are deposited into the ocean increase the nutrient input to the seawater, and finally alter the primary productivity in the open sea (Fu et al., 2018; Luo et al., 2016; Mahowald et al., 2018; Shi et al., 2012). Moreover, previous studies found that large amounts of light-absorbing aerosols (e.g., black carbon and brown carbon) from continental polluted air can be transported into the open ocean air and significantly influence the radiative balance of the marine boundary layer (Kang et al., 2018; Kondo et al., 2016; Ueda et al., 2018; Zhang et al., 2014). Therefore, it is significant to understand the physicochemical properties of continental anthropogenic aerosol particles in marine air.

The eastern China seas (ECS: the Yellow Sea and the East China Sea) and the northwestern Pacific Ocean (NWPO) can be affected by the Asian continental air masses under the prevailing westerly winds in winter and spring (Uematsu et al., 2010; Uno et al., 2009). At present, there have been many in-depth studies on the physicochemical properties of aerosols in air masses before they leave the Asian continent. For example, Li et al. (2014) collected aerosol particles at a background site in the Yellow River Delta and determined their physicochemical properties before leaving the Asian continent. Feng et al. (2012) studied the sources and formation pathways of PM$_{2.5}$ at Changdao Island, a resort island in Bohai Sea/Yellow
Sea, which is in the transport path of the Asian continental outflow to the Pacific Ocean. Shi et al. (2019) investigated aerosol particles from Asian continental outflow in Qingdao and found the solubility of phosphorus was related to the sources and atmospheric acidification processes. However, these atmospheric field observations were limited to some isolated continental sites.

Shipboard observations are an effective way to study marine aerosol properties in remote areas. Using the single particle analysis method (i.e., electron microscope), previous shipboard atmospheric studies have observed chloride depletion and sulfur enrichment in SSA particles from the marine boundary layer (Bondy et al., 2017; McInnes et al., 1994; Mouri and Okada, 1993). At a coastal city in southwestern Japan, Zhang et al. (2003) found that dust particles from Asian continent could mix with SSA particles in the marine atmosphere and further restrained chloride depletion from the sea-salt component in the particles. However, these studies did not examine how anthropogenic pollutants influence the physicochemical properties of SSA from the margin sea air to the remote marine air. Furthermore, they cannot continuously trace the changes of anthropogenic aerosol particles along the pollutants’ transport path to the remote NWPO. These information is critical to comprehensively understand the influence of continental anthropogenic air pollutants on the marine air.

To achieve this aim, the research ship R/V Dongfanghong 2 was designed to cruise from the ECS to the NWPO so that we could understand what and how air pollutants are transported from the highly polluted continental air to clean marine air in spring. After the cruise, a transmission electron microscope was used to obtain
composition, size, morphology, and mixing states of marine aerosol particles. Based on this information, we compared aerosol characteristics over the ECS and the remote NWPO. Furthermore, we also discussed how the continental air masses influence marine aerosols in the ECS and the NWPO air.

2. Experiments

2.1 Aerosol sampling and analyses

Aerosol samples were collected on board the R/V Dongfanghong 2 during the cruise from 17 March to 22 April 2014. The cruise path crossed the ECS and the NWPO (Figure 1). Aerosol particles were sampled onto copper TEM grids (carbon type-B, 300-mesh copper; Tianld Co., China) using a DKL-2 sampler (Genstar Electronic Technology, China). The sampler was equipped with a single-stage impactor with a 0.5 mm diameter jet nozzle at an airflow of 1.0 L/min. If the particle density is 2 g cm\(^{-3}\), the collection efficiency of the sampler is 50% for particles with a 260 nm aerodynamic diameter. All the samples were collected at the ship’s bow to avoid contamination from the exhaust. During the same cruise, the short-period contribution from contamination of particulate matter was less than 3% (Luo et al., 2016). The sampling duration varied from 2 to 3 min to avoid individual particles overlap on the substrate. After the collection, all the samples were sealed in dry plastic capsules and stored in a desiccator at 25 °C and 20 ± 3% relative humidity (RH) for further analysis.

The aerosol particles were analyzed by a JEOL JEM-2100 transmission electron
microscope (TEM) operated at 200 kV. The chemical elements (heavier than carbon, $Z \geq 6$) were semi-quantitatively detected by an energy-dispersive X-ray spectrometer (EDS) (Oxford Instruments, UK). The iTEM software (Olympus Soft Imaging Solutions GmbH, Germany) was used to analyze individual particles in the TEM images and obtain their projected area, perimeter, aspect ratio, and equivalent circle diameter (ECD).

A total of 22 samples were analyzed in this study. The location of each sample is shown in Figure 1. The details about sampling dates, times, and meteorological conditions for each sample are listed in Table S1. Due to the influence of the westerlies, the ECS is frequently affected by the air pollutants’ transport from Asia in spring (Shi et al., 2019). The NWPO is less affected by the transport of continental pollutants because of the remote distance (Zhang et al., 2018). According to the sample locations, we defined two sample categories along with the cruise path (Figure 1): 11 samples in the ECS and 11 samples in the NWPO.

2.2 Air mass backward trajectories

The NOAA HYSPLIT (Hybrid Single Particle Lagrangian Integrated Trajectory) trajectory model (Stein et al., 2015) was applied to calculate the backward trajectories for the investigation of air mass sources and transport paths. The total run time was set at 48 hours. We selected an altitude of 500 m as the endpoint in each backward trajectory. In this study, we obtained 21 backward trajectories.

3. Results
3.1 Classification and relative abundance of aerosol particles

The TEM analyses provided the morphology, mixing state, and composition of individual particles. In this study, we analyzed 3,734 particles in 22 samples (2,770 particles collected in the ECS and 964 particles in the NWPO). Aerosol particles were classified into seven types based on their composition and morphology: sulfur-rich (S-rich), organic matter (OM), soot, metal, fly ash, mineral, and sea salt. S-rich particles are considered as secondary inorganic particles (e.g., (NH₄)₂SO₄ and NH₄NO₃), which are formed from their gaseous precursors, such as SO₂, NOₓ, and NH₃. OM particles mainly contain C and certain O and Si. Here we observed two kinds of OM particles: spherical or irregular primary organic matter (POM) particles and secondary organic matter (SOM) particles. The POM is directly emitted from the combustion of fossil fuel and biomass and SOM is formed from volatile organic compounds (VOCs) or the oxidized POM in the atmosphere (Li et al., 2016a; Wang et al., 2021). The SOM is normally mixed with S-rich particles (Li et al., 2016b). Soot particles (i.e., black carbon) are chain-like aggregates of carbonaceous spheres, mainly containing C and minor O. Metal particles mainly contain Fe, Zn, and Pb, and fly ash particles contain Si, Al, and minor Ca and Fe. Metal and fly ash particles both display spherical shapes and are directly emitted from heavy industrial activities such as power plants and steel factories (Li et al., 2017). Mineral particles are composed of Si, Al, Ca, and Fe and present irregular shapes (Figure 2a). Mineral particles originate from arid deserts (e.g., Sahara and Gobi), roads, and construction activities in the continental areas. Sea salt aerosol (SSA, Figure 2b) is from the bursting of air bubbles...
resulting from the waves breaking. SSA is mainly composed of Na and Cl, with minor Mg, Ca, K, and S.

The high-resolution TEM could see through the thin materials on the substrate, so the inner mixing structure of different aerosol components in individual particles can be directly identified (Li et al., 2016b; Riemer et al., 2019). We found that most of the individual non-SSA particles collected in marine air contained two or more different types of anthropogenic aerosols. To elucidate the mixing structure of the non-SSA particles, we further defined six types of non-SSA particles: S-metal (Figure 2c), S-fly ash (Figure 2d), S-soot (Figure 2e), OM coating (Figure 2f), OM-S (Figure 2g), and S-rich (Figure 2h). In the ECS, anthropogenic aerosols accounted for 87% of all particles by number fraction, including S-rich for 42%, S-soot for 21%, S-metal/fly ash for 8%, OM-S for 6%, and OM coating for 10%. Only 8% of the observed particles in the ECS were SSA particles. The remaining 5% was identified as mineral particles. Interestingly, SSA particles became the dominant aerosol at 90% in the NWPO and anthropogenic aerosols only accounted for 8%, suggesting that marine emissions became the primary aerosol source in the NWPO. Therefore, there are large differences between aerosol particles in the ECS and the NWPO (Figure 3).

3.2 Variations of aerosol particles from the ECS to the NWPO

Figure 3 shows variations of aerosol particles along with the cruise pathway from the ECS to the NWPO. Mineral particles can only be transported from continental areas. Figure 3 shows that higher number fractions of mineral particles always
occurred when the sampling sites were close to eastern China. The number fraction of mineral particles rose to 15% in the coastal air during the cruise (Figure 3). The proportion of mineral particles decreased from 15% to 3% for the samples collected in the ECS. When the ship traveled eastward into the NWPO, the proportion of mineral particles dropped to a low level (2%). In contrast, the proportion of mineral particles increased again when the ship returned to the ECS. Altogether, the number fraction of mineral particles was 5% in the ECS, twice as high as that in the NWPO, suggesting that the ECS and NWPO were influenced by westerlies during the sampling period. Indeed, Figure 1 shows that most of 48h backward trajectories of air masses in the ECS were sourced from eastern China and that some of the 48h back trajectories of air masses in the NWPO crossed Japan.

Figure 3 shows that S-metal and S-fly ash (two typical anthropogenic aerosol particles) displayed variation similar to the mineral particles. Number fractions of S-metal and S-fly ash particles in the ECS samples ranged from 17% to 2% with the average value at 8%, but only 0.3% for S-metal and S-fly ash particle was found in the NWPO. In a word, we conclude that aerosol particles from the Asian continent directly exert much greater impacts on the ECS than the NWPO.

### 3.3 Comparison of anthropogenic secondary aerosol particles

Secondary S-rich and OM coating particles are normally considered as arising from the conversion of anthropogenic gaseous pollutants (e.g., SO$_2$, NO$_x$, NH$_3$, and VOCs) (Li et al., 2021). TEM observations clearly identified secondary particles and showed their variation of number fraction in the samples (Figure 3). Our results show
that number fractions of S-rich particles were dominant in the ECS samples with the range of 32%-71%, but their fractions decreased to 5% in the NWPO. The results indicate that secondary particles in the ECS were strongly influenced by anthropogenic pollutants transported from eastern China. Furthermore, we noticed that secondary aerosol particles were frequently mixed with some typical fine primary anthropogenic particles (e.g., soot, fly ash, and metal) and formed S-soot/S-fly ash/S-metal particles (Figure 4). As a result, we conclude that the anthropogenic gases or aerosol pollutants in the continental air masses significantly influence the downwind air quality of the ECS, but they have a minor impact on the NWPO air.

It should be noted that OM coating particles were frequently found in the ECS but barely observed in the NWPO. In other words, S-rich particles in the NWPO had no typical OM coating, although S-rich particles accounted for ~5% in the NWPO samples. Figure 4b shows that these S-rich particles in the NWPO only had one dominant size range smaller than 1 μm, which is different from the larger and broader size distribution of S-rich particles in the ECS. In addition, we noticed the particularly high fraction of S-rich particles in Sample #11 and #12 collected in the NWPO (15% and 24%). These results indicate that these S-rich particles likely formed in the NWPO air. Coincidently, Zhu et al. (2019) observed a new particle formation event in the same area and proposed that the event was likely caused by long-range transported continental gases (e.g., SO₂, NOₓ, and VOCs).

3.4 Aging of sea salt aerosols

As the dominant aerosol particles in marine air, SSA particles accounted for
70%-98% in the NWPO. SSA could serve as reactive surfaces for heterogeneous and multiphase chemical reactions in the atmosphere, and these reactions also could alter the morphology and composition of SSA (Athanasopoulou et al., 2008; Chi et al., 2015; Laskin et al., 2012). Based on the morphology and composition of SSA, we further classified SSA into three categories: fresh SSA, partially aged SSA, and fully aged SSA (Figure 5).

The fresh SSA did not experience any chemical modification in the atmosphere. TEM images of fresh SSA indicate the cubic NaCl core and coating composed of MgCl$_2$ and CaSO$_4$ (Figure 5a). The NaCl core only contains Na and Cl, with the atomic ratio of Na to Cl close to 1:1 (Figure 5d). The major components in the coating are Mg, O, S, Cl, and Ca (Figure 5e), thus, we infer their molecular forms as MgCl$_2$ and CaSO$_4$ (Buseck and Pósfai, 1999; Chi et al., 2015; Geng et al., 2010; Pósfai et al., 1994).

The partially aged SSA represent those SSA particles that undergo chemical modification but still retain part of the NaCl core (Figure 5b). The morphological differences can be observed between the fresh SSA and partially aged SSA. The NaCl core still persists in the partially aged SSA but cannot keep its regular cubic shape. Meanwhile, the coating composition turns into O, Na, Mg, Ca, and S, with decreasing Cl (Figure 5f). The SSA aging is attributed to the Cl-depletion phenomena, which can be expressed as follows (Laskin et al., 2012):

$$\text{NaCl (aq)} + \text{HA} (\text{aq, g}) \rightarrow \text{NaA (aq)} + \text{HCl (aq, g)}$$

where HA represent atmospheric acids (e.g., H$_2$SO$_4$, HNO$_3$, and methanesulfonic...
acid). NaCl in the SSA could react with inorganic (e.g., HNO₃ and H₂SO₄) or organic acid (e.g., methanesulfonic acid), releasing volatile HCl (g) to the atmosphere, leading to depletion in chloride and enrichment in corresponding sodium salts.

We define the fully aged SSA as particles whose NaCl cores have been completely transformed into NaNO₃ and Na₂SO₄. Figure 5c shows that the NaCl cores in the fully aged SSA entirely disappeared, leaving a rounder shape. The Cl element was no longer detected in the fully aged SSA and the major aerosol components were the mixture of NaNO₃ and Na₂SO₄ (Figure 5g).

To evaluate composition differences of SSA, we present triangular diagrams of Na-Cl-S weight ratio based on the EDS. Figure 6a shows that the fresh SSA is around the NaCl, the partially aged SSA is distributed in the center of triangular, and the fully aged SSA is around NaNO₃ and Na₂SO₄. Figure 6b shows that there were large variations of the SSA components in the ECS and NWPO. Our results show that most of the aged SSA in the ECS were the mixture of NaNO₃ and Na₂SO₄, suggesting that SSA in the ECS underwent heterogeneous reactions and became fully aged SSA. SSA in the NWPO is widely distributed between NaCl and NaNO₃/Na₂SO₄ in the triangular diagram, suggesting that these were partially aged SSA particles.

Figure 7 shows the relative abundance of SSA at different sampling sites from the ECS to the NWPO. In the ECS, 86% of SSA particles were fully aged. As the ship traveled into the NWPO, fully aged SSA particles decreased to 31%, meanwhile the proportion of fresh SSA increased to 55%. The size distribution of SSA particles shows that the proportion of fresh SSA increased with the increase of particle size.
from 0.05 to 5 μm (Figure 8). On the contrary, the proportion of fully aged SSA decreased with the increase of particle size from 0.05 to 5 μm. Overall, partially and fully aged SSA accounted for 61% of SSA particles smaller than 3 μm, while fresh SSA dominated (81%) in SSA particles larger than 3 μm. Figure 8 also reveals that 94% of the fully aged SSA particles were smaller than 3 μm.

4. Discussion

Abundant BC, metal, and fly ash particles in the ECS show that long-range transport of anthropogenic aerosol particles from the polluted continental areas constantly influence the ECS air during the spring and winter. Moreover, the existence of OM coating particles (10%, Figure 3) in the ECS indicates that secondary sulfate/nitrate particles underwent aging process and formed the OM coating during their transport (Li et al., 2021). This result suggests that the long-range transported air masses from continental areas brought abundant anthropogenic gases such as NOx, SO2, and VOCs into marine air. Indeed, the HYSPLIT backward trajectories show that the air masses in the ECS were mainly from eastern China (Figure 1). Under the influence of the westerlies, a large number of anthropogenic and natural pollutants are transported to the ECS and further influence its air quality.

Figure 3 shows that both primary and secondary anthropogenic aerosols were relatively low in the NWPO (8%), suggesting a slight influence from continental aerosols. However, we observed severe Cl-depletion in SSA in some samples in the NWPO. The Cl-depletion of SSA is mainly caused by the heterogeneous reactions...
with acidic gaseous pollutants in marine air (Chi et al., 2015; Hsu et al., 2007; Laskin et al., 2012). By comparing the aging degree of SSA and air mass backward trajectories, we found that the air masses with a relatively high proportion of fully aged SSA particles (#13 and #15) mostly formed due to the anthropogenic gaseous pollutants (e.g., SO$_2$ and NO$_x$) from northwest China and Japan (Figure S1). The result suggests that aerosol particles and gases might have different transport distances. Aerosol particles could be significantly removed by dry or wet deposition. However, the anthropogenic gases can be transported further to the NWPO air (Figure 9). Over the western Pacific, Koike et al. (2003) also found that anthropogenic gaseous pollutants (e.g., NO$_x$ and SO$_2$) from the East Asia have higher transport efficiency than aerosols from the same region. Meanwhile, the surface changes of SSA can be considered as a potential indicator for anthropogenic gaseous pollutants in remote marine air.

Figure 6 shows a higher percentage of Cl-free SSA particles (on Cl = 0 line) in the ECS than those in NWPO, suggesting modifications of SSA in the ECS were much more severe than them in the NWPO. Referred to Zhang et al. (2003), we further provide the dash line that Na:Cl:S was changed only by reaction with H$_2$SO$_4$. Thus, particles above the dash line represent that Cl in these SSA particles was replaced by S deposition and other chemical processes (e.g., HNO$_3$ and organic acids). The number of fully aged SSA particles above the dash line was further counted. Surprisingly, we found that Cl-depletion in 70% of the fully aged SSA particles in the ECS was not only caused by S deposition, but the fraction increased to 87% in the
NWPO (Figure 6). The result indicates that a higher proportion of the fully aged SSA particles in the NWPO were reacted with diverse acids besides H$_2$SO$_4$.

It is well known that organic acids could also be a reason of Cl-depletion in SSA (Chi et al., 2015; Laskin et al., 2012). It is well known that the oxidation of dimethyl sulfide (DMS) emitted from phytoplankton contributes to the Cl-depletion in SSA (Sievering et al., 2004). Both anthropogenic acid gases and DMS in the ECS were found to be higher than that in the NWPO (Figures S1 and S2). Compared with the ECS, we found that DMS in our NWPO shipping area had lower mass concentrations and minor fluctuations (3.6 × 10$^{-11}$ kg m$^{-3}$, Fig S2). However, the percentages of aged SSA differed widely (from 23% to 91%) within the NWPO. We deduce that the low concentration of the DMS only slightly modified part of SSA in the NWPO, but it was not enough to influence all the SSA. Moreover, Zhu et al. (2019) reported that secondary sulfate particles in the NWPO were mostly came from the long-range transported acidic gases during the sampling period. The contribution of DMS to non-sea-salt sulfate is less than 6% in the remote ocean of the northern hemisphere (George et al., 2008; Quinn et al., 1990; Savoie et al., 1994). Therefore, a large proportion of aged SSA particles in the NWPO should be mainly attributed to the anthropogenic acidic gases from long-range transport.

Previous studies found that SSA emissions increased with the increase of wind speed (Feng et al., 2017; Pant et al., 2008; Shinozuka et al., 2004). In our study, the proportion of fresh SSA increases with increasing wind speed (Figure 7), consistent with the aforementioned studies. SSA particles of a smaller size have lower dry
deposition velocities and longer lifetimes in the air (Lewis and Schwartz, 2004), which could enhance the Cl-depletion. This could be the reason that partially and fully aged SSA was mostly in the smaller size range (<3 μm) (Figure 8). On the contrary, the newly emitted coarse SSA particles with high dry deposition velocity are more likely to deposit to the ocean, resulting in less reacted SSA. As a result, the fresh SSA from local sea spray was mostly found in the coarse size range (larger than 3 μm) in our samples (Figure 8). This result suggests that it is crucial to study aging process among the size-resolved SSA, especially particles smaller than 3 μm.

In the future, we need to pay more attention to the influence of anthropogenic gaseous pollutants on the SSA aging in remote marine air. On the one hand, the aging processes could modify the hygroscopicity of SSA, determining their morphology and phase state in the humidified marine environment, in the end directly affecting optical properties of SSA (Wang et al., 2019). On the other hand, the hygroscopicity change due to SSA aging could alter CCN activity, and indirectly affect global climate (Hu et al., 2005; Murphy et al., 1998; Pierce and Adams, 2006). The SSA aging processes could release more Cl- and I-containing compounds into air, which might lead to marine new particle formation and growth (He et al., 2021). Meanwhile, the SSA also serves as an important sink for the anthropogenic acidic gases in remote marine areas (Chi et al., 2015; Laskin et al., 2012). Thus, in future research, it would be crucial to quantify the anthropogenic acidic gases scavenged by SSA.

5. Conclusions
Individual aerosol particles were collected from 17 March to 22 April, 2014 on board of the ship R/V Dongfanghong 2 from the ECS to the NWPO. We classified aerosol particles based on their composition, morphology, and mixing state: mineral, sea salt, S-metal, S-fly ash, S-soot, OM coating, OM-S, and S-rich. Microscopic analysis showed that anthropogenic aerosols accounted for 87% of the total particle number in the ECS. In particular, higher proportions of secondary particles (i.e., S-rich particles, 42%, and OM coating particles, 10%) were found in the ECS. Meanwhile, primary and secondary anthropogenic aerosols are relatively low in the NWPO (8%).

TEM observations revealed that SSA particles were the most abundant in the NWPO atmosphere, accounting for 90% of all analyzed aerosol particles. The Cl-depletion of sea salt aerosol (SSA) particles caused by the heterogeneous reactions with acidic gaseous pollutants was further observed. Three types of SSA particles, fresh, partially aged, and fully aged were classified. Fully aged SSA particles were the dominant SSA in the ECS (86%), while fully aged SSA particles decreased to 31% in the NWPO. The severe aging of SSA (partially and fully aged, at most 88% of SSA) was still found in the NWPO, despite there being only minor anthropogenic aerosol particles. These results show that aerosol particles from the continent air might be removed by dry and wet deposition, but the air pollutants were transported further to the NWPO. The aging of SSA particles has important effects on their hygroscopic and optical properties, one effect being the promotion of heterogeneous reaction with acidic gases in the NWPO. Our observations show that more attention should be given
to the influence of anthropogenic gaseous pollutants on the Cl-depletion on SSA in remote marine areas.

Supplement

The supplement related to this article is available online at:

Author contributions

LX and WL conceived the study and wrote the article. The sampling during the research cruise was organized by XL, HG, and XY. LX, WL, LL, JZ, YZ, YW, and QY carried out TEM analyses of individual particles. DZ and LB contributed to the improvement of this paper. All authors reviewed and approved the paper.

Competing interests

The authors declare that they have no conflict of interest.

Acknowledgements

We thank Peter Hyde for his editorial comments. We acknowledge the NOAA Air Resources Laboratory for the provision of the HYSPLIT transport and dispersion model and READY website (http://www.ready.noaa.gov) used in this publication.

Financial support

This research was supported by the National Natural Science Foundation of China
(grant nos. 42075096, 91844301, and 41807305), the National Key R&D Program of China (grant no. 2017YFC0212700), Zhejiang Provincial Natural Science Foundation of China (grant no. LZ19D050001), and China Postdoctoral Science Foundation (grant no. 2019M662021).

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Figure 1. (a) Map of the cruise track (black line) and 48 h air mass backward trajectories (red and blue lines) arriving at 500 m above ground level at sampling locations. The interval between two circle symbols is 6 h. The number represents the sample ID in Table S1. (b) Photo of the R/V Dongfanghong 2.
Figure 2. Transmission electron microscope (TEM) images of different types of aerosol particles: (a) mineral; (b) sea salt; (c) metal particles mixed with sulfate; (d) fly ash particles mixed with sulfate; (e) soot particles mixed with sulfate; (f) secondary organic matter (OM) coating on sulfate; (g) primary OM particle mixed with sulfate; (h) S-rich particle.

Figure 3. Relative abundances of eight types of aerosol particles in different samples. The number of analyzed aerosol particles is shown above the column.
Figure 4. Relative abundances of individual particles in different size bins.

Figure 5. Morphology and EDS spectra of the typical fresh, partially aged, and fully aged SSA. The main anionic elements are shown in the square brackets.
Figure 6. Triangular diagram of Na-Cl-S from EDS data (weight percentage) showing the elemental composition of SSA particles. The three stars represent pure NaCl, Na₂SO₄, and NaNO₃, respectively. The dash line indicates that Na:Cl:S is changed only by the postulated reaction of 2NaCl + H₂SO₄ → Na₂SO₄ + 2HCl(g) (Zhang et al., 2003). Particles above the dash line are those which S cannot compensate Cl losses and there should be other processes causing Cl-depletion.
Figure 7. Relative abundances of the fresh, partially aged, and fully aged SSA particles in different samples. Samples with SSA particle less than 30 are excluded due to the small number. The line indicates the wind speed of the corresponding sample.
Figure 8. Relative abundances of three types of SSA particles in different size bins.

Figure 9. Schematic diagram showing the impact of long-range transported anthropogenic air pollutants on marine aerosols.