



The impact of peripheral circulation characteristics of typhoon on

2 sustained ozone episodes over the Pearl River Delta region, China

- 3 Ying Li^{1,2}, Xiangjun Zhao^{1,2,3*}, Xuejiao Deng^{4*}, Jinhui Gao^{1,2,5}
- 4 ¹Department of Ocean Sciences and Engineering, Southern University of Science and Technology, Shenzhen, China
- 5 ² Southern University of Science and Technology, Shenzhen, China
- 6 ³ School of Mathematics and Finance, Chuzhou University, Anhui 239000, China
- 7 ⁴ Institute of Tropical and Marine Meteorology/Guangdong Provincial Key Laboratory of Regional Numerical Weather
- 8 Prediction, China Meteorological Administration, Guangzhou, China
- 9 ⁵ Plateau Atmosphere and Environment Key Laboratory of Sichuan Province, School of Atmospheric Sciences, Chengdu
- 10 University of Information Technology, Chengdu, China.
- 11 * Corresponding author e-mail address: (iamzxj841025@163.com) and (dxj@gd121.cn)

12 Abstract. It is widely reported that the peripheral circulation of typhoon favors for the formation of sustained ozone 13 episodes. However, the process how it impact on the day-to-day ozone pollution levels during the episodes have not been 14 clearly studies, which is crucial for better prediction of the daily ozone variation. In this study, the analysis of ground 15 observation, wind profile data, and model simulation are integrated. By analyzing the wind profile radar observations, we 16 found a weak winds deepening (WWD; vertical depth of the weak winds increased), which is more correlated to the 17 ground-level ozone variation than surface weak wind. Long-term statistical analyses show that the WWD is a common 18 weather phenomenon that occurs in the peripheral subsidence region of typhoons and was generally accompanied by ozone 19 pollution episodes. WRF-Chem with process analysis simulation show that under the impact of the peripheral subsidence 20 chemical formation (CHEM) and vertical mixing (VMIX) effects are two major contributors to the enhancement of ozone 21 levels, while the advection (ADV) are always negative values. But regarding the daily variability of the daytime ozone 22 levels during the episode, it do not determined by the daily variation of daytime CHEM and VMIX, but that of the ADV 23 term. A detail day-to-day analysis show that weak subsidence associated with typhoon periphery provide the premise for the 24 clear sky and warmer air, which is conducive for the ozone photolysis formation (CHEM) above the ground in planetary 25 boundary layer (PBL) and compensate the ozone through the positive VMIX effects on the ground. The WWD induced by 26 the peripheral circulation of typhoon system provide the premise for the day-to-day positive contribution of ADV term to 27 ozone enhancement throughout the whole planetary boundary layer (PBL), which play an important role in determining the 28 day-to-day daytime ozone variation. These results indicate the important role of the WWD in the lower troposphere for the 29 formation of sustained ozone episodes due to the peripheral circulation of the typhoon, which helps to better predict the daily 30 changes of daytime ozone levels.





32 **1. Introduction**

33	The Pearl River Delta (PRD) located in the coastal region of South China, which is often affected by typhoon
34	systems, has experienced major economic development and urbanization in the past two decades, and has been
35	accompanied by large increases in air pollution and decreases in visibility (Wang et al., 1998, 2001; Lai and Sequeira,
36	2001). Ozone pollution is the most important air pollution issue in this region; ozone has been the 'primary pollutant'
37	since 2014 (Ministry of Ecology and Environment of China, 2016). Ozone is harmful to human health and has adverse
38	effects on vegetation and crops, among others(Aunan et al., 2000; Felzer et al., 2007; Feng et al., 2015). Ozone
39	concentrations are determined by the photochemical reactions of its precursors and by the local meteorological
40	conditions. However, ozone pollution episodes are mainly triggered by weather conditions rather than by sudden
41	increases from emission sources(Ziomas et al., 1995; Giorgi and Meleux, 2007; Lin et al., 2019).
42	The Guangdong Haze Weather Bulletin(Wang, 2017) has classified the weather patterns affecting regional pollution
43	events into cold fronts, cold high-pressure systems moving towards the sea, uniform pressure fields, Western Pacific
44	subtropical high (WPSH), tropical cyclone (TC) peripheries, and weak cold high-pressure ridges. By using observational
45	data, several studies have reported the impacts of TC activity on meteorological factors that are favourable for air
46	pollution over the PRD region(Feng et al., 2007; Chen et al., 2008; Wu et al., 2013). TCs are typical weather systems that
47	are responsible for both high ozone and PM _{2.5} pollution over the PRD(Chen et al., 2008; Deng et al., 2019).
48	Many studies in the PRD region and other coastal regions of China have shown the significant impact of TCs on
49	forming ozone (TCs-Ozone) episodes in recent years(Zhang et al., 2012; Li et al., 2013, 2014; Zhang et al., 2013; Jiang et
50	al., 2015; Huang et al., 2015; Shu et al., 2016, 2019; Tan et al., 2018; Chen et al., 2018; Han et al., 2019). TCs-Ozone
51	episodes generally occur when weather conditions include high temperatures, radiation flux, low relative humidity, and
52	weak winds(Cheng et al., 2016; Liu et al., 2017). There were large amount of observational-based studies reporting the
53	TCs-Ozone episodes are weak wind related, however it is very few about the study of the influence mechanism of weak
54	wind on ozone in TCs-Ozone episodes. In addition, previous integrated process rate(IPR) analysis based on numerical
55	modelling simulations have reported that the chemical (CHEM) and vertical mixing (VMIX) effects are two major



56



57 Wang et al., 2009). The inconsistencies between observational and simulated results of wind contributions to ozone 58 episodes are poorly understood, which may be due to the lack of studies of influence mechanism of weak wind on ozone 59 concentration enhancement. 60 In addition, for the air quality forecast and prevention, it is more important to understand the mechanism leading to the 61 day-to-day variation of the daytime ozone levels, since the ozone levels always reach its peak values in the daytime due 62 to photo-chemistry and ozone converted to NO2 temporarily in the absence of light incidents at nights. However, though 63 the TCs-Ozone episodes have been widely reported, the studies of mechanism on the daily daytime variation of during 64 sustained TCs-Ozone episodes are quite limited. 65 Thus, the objective of this study is to the impact processes of typhoon circulation characteristics on the day-to-day 66 variation of daytime ozone concentration in TCs-ozone episode. The analysis of ground observation, wind profile data,

contributors to ozone episodes, whereas advective transport (ADV) is generally a consumptive process(Shu et al., 2016;

- 67 and WRF-Chem model simulation with process analysis are integrated. Detailed data and model description are
- 68 provided in Sect. 2, followed by the results and discussion in Sect. 3. The last section summaries the main conclusions.

69 2. Data and model

70 **2.1 Data**

In this study, hourly surface ozone concentrations from 2016 over mainland China were obtained from the Ministry of Environmental Protection of China. The 3D wind profiler data, automatic weather station data, cloud data, and solar radiation measurements were provided by the China Meteorological Administration and were used for the meteorological analyses of Typhoon Nepartak. The Final (FNL) Operational Global Analysis data that were used to describe the circulation of Typhoon Nepartak have a horizontal resolution of 1° x 1° with 27 vertical levels and were obtained from the National Centers for Environmental Prediction(NCEP), USA .

- 77 The observations of a typical ozone episode occurred in the PRD region during 7–10 July 2016 (local standard time; LST)
- 78 before Typhoon Nepartak made landfall was collected and deeply analyzed . Typhoon Nepartak intensified into a super





- typhoon at 20:00 on 5 July, then gradually moved northwest due to the forcing of the WPSH over its northeastern side (Fig. S2). At 05:50 on 8 July, the typhoon made landfall in Taitung County, Taiwan, with a maximum wind speed of 60 m s⁻¹, and again in Shishi City, Fujian at 14:00 on 9 July, with a maximum wind speed of 23 m s⁻¹. At 03:00 on 10 July, the typhoon weakened into a tropical depression.
- 83

84 **2.2 Model descriptions**

85 WRF-Chem is a widely used and fully coupled online 3D Eulerian chemical transport model (https://ruc.noaa.gov/wrf/wrf-chem/) that considers both chemical and physical processes(Zhang et al., 2010; Forkel et 86 87 al., 2012); version 3.9.1.1 was applied in this study. Detailed descriptions of the meteorological and chemical aspects of 88 the WRF-Chem model can be found in Grell et al.(2005) and Skamarock et al.(2008). For the simulation, two nested 89 domains (Fig. S1) were set up with horizontal resolutions of 27 and 9 km and grids of 283×184 and 223×163 for the 90 parent domain (D1) and nested domain (D2), respectively. D1 was centred at (28.5°N, 114.0°E) covering most of China, 91 the surrounding countries, and the ocean. Corresponding simulations provided meteorological and chemical boundary 92 conditions for D2, which covered most of southern China. 93 There were 39 vertical layers that extended from the surface up to a pressure maximum of 50 hPa, 12 of which were 94 located in the lowest 2 km to fully describe the vertical structure of the PBL. Carbon Bond Mechanism Z (CBM-Z), 95 which includes 133 chemical reactions for 53 species and extends the model framework to function for a longer time 96 period and at a larger spatial scale than its predecessor, was used as the gas-phase chemical mechanism(Zaveri and Peters, 97 1999). The corresponding aerosol chemical mechanism was the Model for Simulating Aerosol Interactions and Chemistry (MOSAIC) with eight bins(Zaveri et al., 2008), which is extremely efficient and does not compromise 98 99 accuracy of the aerosol model calculations. Other major model configuration settings are listed in Table 1. 100 Table 1. Major model configuration options used in the simulations.

ITEM	Selection	
Long wave radiation Shortwave radiation Microphysics scheme	RRTMG RRTMG Lin scheme	



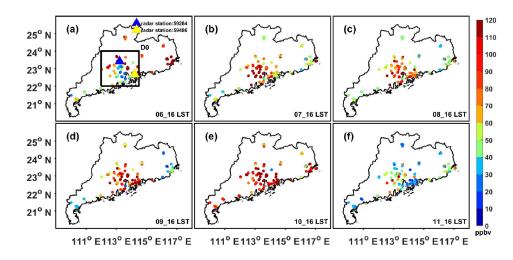


Boundary layer scheme	Yonsei University (YSU) scheme
Land surface option	Noah land surface model
Photolysis scheme	Fast-J photolysis
Dry deposition	Wesely scheme

101 **3. Results and discussion**

102 **3.1 Episodic data analysis**

103 The ozone pollution level and the meteorological conditions of the typhoon Nepartak case was first analyzed. As shown 104 in Fig. 1, Guangdong province experienced a sever ozone pollution during the period. 28% (7 July) to 57% (10 July) of the air quality stations in Guangdong Province exceeded the national air quality standard level-II for ozone (200 µg m⁻³) 105 106 at the daily peaks (16:00 LST). To show the vertical motion of the typhoon centre and peripheral region, we constructed 107 a cross section through the typhoon system (points A and B; Fig. 2a-d) and plotted the corresponding vertical velocities 108 (Fig. 2e-h) using the NCEP data. As shown in Fig. 2e-f, the western subsiding branches of vertical typhoon circulation 109 were located over the PRD during 7-8 July, when ozone concentrations increased significantly compared to those of 6 110 July. After Typhoon Nepartak made landfall at Shishi City on 9 July, the peripheral subsidence had moved to the western 111 area of the PRD region (Fig. 2g-h) and the PRD region was influenced by weak vertical motion and a weak horizontal 112 wind field. Peak ozone levels exceeded 100 ppb at most of the monitoring stations in the PRD at this time. On 11 July, 113 Typhoon Nepartak dissipated and the surface ozone concentrations began to decrease (Fig. 1f).

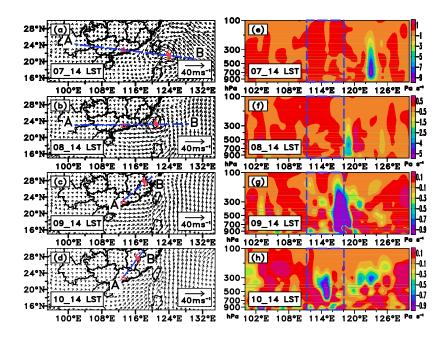


115 Figure 1. The horizontal distribution of surface ozone concentration over PRD at 16:00 from (a) 6 July 2016 to (f) 11 July 2016. The yellow





- 116 and blue triangles in (a) denote the positions of wind profiler station 59486 and 59284. The black box D0 indicates the area where the severe
- 117 ozone pollution event occurred.



118

119 Figure 2. (a)-(d) 1000 hPa wind vectors of NCEP-FNL data from 14:00 7 July to 14:00 10 July with red triangle and typhoon signs

120 representing PRD center and Nepartak locations, respectively. (e)-(h) vertical cross sections of vertical velocity along the four straight lines

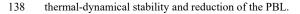
linking PRD and the centers of Typhoon Nepartak in (a)-(d) from 14:00 7 July to 14:00 10 July 2016. The four blue dashed boxes denote the
 longitude range of PRD in (e)-(h).

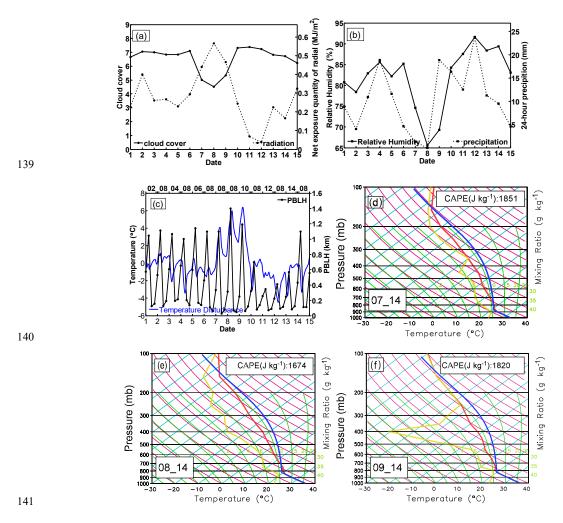
123	The weather over the PRD region was characterized by clear sky, strong solar radiation (Fig. 3a), low relative humidity
124	(Fig. 3b), and high temperatures (Fig. 3c), when the subsiding branches of vertical typhoon circulation were located over
125	the PRD during 7-8 July (Fig. 2e-f). The variations in these surface meteorological variables exhibited favorable
126	conditions for increasing ozone concentrations(Cheng et al., 2016; Liu et al., 2017). However, the height of the PBL
127	increased significantly on 8 and 9 July (Fig. 3c), and the atmosphere was under unstable conditions, which was indicated
128	by the overlapping temperature soundings and the parcel traces below 800 hPa (Fig. 3d-f). This instability is also shown
129	by the large values of convective available potential energy (CAPE; Fig. 3d-f), which is another criterion used to
130	determine the stability of atmosphere. In general, when the CAPE is $\sim 1000 \text{ J kg}^{-1}$, the atmosphere is unstable, which is
131	favorable for convection. These results illustrate that, under the control of typhoon periphery, the PBL height can be
132	increased in unstable atmospheric conditions, which is opposite from the observations in some TCs-haze events (Wu et





al., 2005 and Feng et al., 2007). For example, the research of Wu et al.(2005) reported that the TC produces a strong descending motions in the lower troposphere, a weak surface wind speeds, and a lower PBL. As a result, the strong peripheral subsidence of TC causes descending air motions to force the aerosol particles into a very shallow layer, and the weak horizontal winds keep the pollutant aerosols inside the source region, resulting in very high concentrations. Our observational results indicate that the TCs-Ozone episodes are not dependent on the enhancement of atmospheric





142 Figure 3. Time series of diurnal mean (a) cloud cover, radiation at 59287 observation station, (b) relative humidity, 24-h precipitation

143 and averaged (c) PBLH and temperature anomaly of region D0 from 1 to 15 July; The SkewT/LogP at 14:00 on 7 July (d), 8 July (e) and 9

144 July (f); the solid thick red, blue and yellow lines in d, e and f denote the temperature sounding, the parcel path from surface upward and the





- 145 dewpoint sounding, respectively.
- 146 The evolution of the vertical profile of horizontal winds at representative station 59284 is shown in Fig. 4a. Before 5 147 July, the wind speed was increasing with the vertical atmospheric layers. There were relatively larger wind speeds above 148 the PBL and relatively weaker wind speeds below ~700 m, with relatively low surface ozone concentrations (<40 ppbv). 149 On 5 July, the daily ozone concentration started to increase (>70 ppbv) as the depth of WWD increased simultaneously. 150 The depth of WWD was ~3 km during 7-9 July with a sustained increasing ozone peak. On the night of 11 July, the 151 horizontal wind speed above ~1 km increased significantly and the ozone concentration decreased sharply. Variations in 152 the wind profile and surface ozone at another representative station are also shown in Fig. 4b. At this station, the depth of 153 WWD started to increase on 7 July, with a gradually increasing ozone peak value. Co-variations of the ozone 154 concentration and WWD at other radar stations were also observed (Fig. S3-5). This co-variation is not a local effect, but
 - 20 18 16 14 12 10 8 6 5310 (a) E 4110 **Altitude** 2910 1710 700 $m s^{-1}$ 100 07/03 07/05 07/07 07/09 07/11 07/13 Date 20 18 16 14 12 10 2839 160 (b) Altitnde 1297 922 561 140 120 100 (qdd) 80 80 __0 m s^{−1} 139 07/03 07/05 07/07 07/09 07/11 07/13
- 155 is instead a regional phenomenon.





¹⁵⁸ 59284 and (b) 59486 wind profile radar station.

159 By analyzing the wind profile data (Fig. 4), we noticed that the vertical depth of the horizontal weak wind generally 160 increased from the surface up to the lower troposphere ($\sim 2-3$ km) and the surface ozone concentration changed with the 161 change of vertical depth of the horizontal weak wind. To further illustrate the different impact of the surface weak wind 162 and the WWD on surface ozone concentrations, the correlation coefficients between the surface ozone concentrations



168



- 163 and the average wind speeds from surface to different altitudes(up to 6 km) at different radar stations were calculated
- 164 (Fig. 5). The correlation coefficients show an increasing trend with altitude, reach their maximum values between 2–3
- 165 km and remain stable above ~ 2.5 km. The average correlation coefficient at the surface was 0.57 (0.41–0.67) and the
- average correlation coefficient above 2000 km was ~0.75 (0.69-0.83) for seven radar stations. This indicates the
- 167 potential impact of WWD on the ozone pollution episode induced by Typhoon Nepartak.

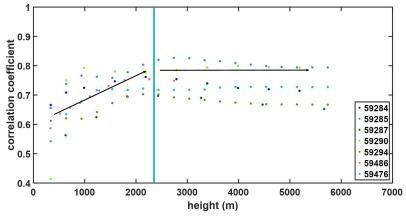


Figure 5. Correlation coefficient between the evolution of average wind speed and the evolution of ground ozone concentration in different
 altitude ranges of each wind profile radar station.

171 **3.2** Long-term statistical analysis of the relationship between WWD and the O3 episode

172 The above observational analysis shows that there was no stable atmospheric stratification and a decrease in the height 173 of the boundary layer in this ozone pollution episode. The analysis of wind profile radar data and the correlation 174 coefficients between the surface ozone concentrations and the average wind speeds between the surface and the altitude 175 of each vertical layer (up to 6 km) indicate that in this episode of ozone pollution, WWD may have played an important 176 role in the increasing of ozone pollution at the surface. Guangdong Province is located on the western coast of the Pacific 177 Ocean and is frequently affected by typhoons annually. To investigate whether the relationship between WWD and 178 ground-level O₃ only occurred in this case study or is a common phenomenon, a long-term statistical analysis of 179 historical data was conducted. A statistical analysis of tropical cyclone wind fields in the Northwestern Pacific Ocean 180 from 2014 to 2018 (based on Guangdong wind profiler data) was conducted. As not all the radar stations in Guangdong 181 province are available during a typhoon, the available statistics number of each radar station for the 38 typhoons were





182	recorded as M. The number of WWD instances at each radar station was recorded as n. Ozone concentrations above 100
183	μ g m ⁻³ are harmful to human health(Organization, 2005). The PRD regional background ozone concentration is
184	generally less than 80–100 μg m $^{-3}$ and the ozone concentrations at most stations can exceed 160 μg m $^{-3}$ (national AQ
185	standard Level-I) during a regional ozone pollution event. Therefore, ozone concentrations of 100–160 μ g m ⁻³ and above
186	160 μ g m ⁻³ were used to denote regional light and heavy ozone pollution in the statistics. The numbers of regional light
187	and heavy ozone pollution events at each radar station were recorded as n1 and n2, respectively. As shown in Table 2, the
188	number of WWD occurrences (n) accounts for 87–97% of the available number(M) of radar stations in the 38 typhoon
189	statistics for the seven radar stations. The average value of n/M for the seven radar stations is 93%. This indicates that,
190	when there is a tropical cyclone in the Northwestern Pacific Ocean, WWD will occur in whole or part of Guangdong
191	province. The number of ozone pollution occurrences (n1+n2) accounts for 78%-100% of the number of WWD
192	occurrences(n). The average value of $(n1+n2)/n$ for the seven radar stations is 94%. The above statistical results show
193	that WWD may be a common phenomenon on the periphery of typhoons and is often accompanied by significant
194	increases in ozone concentrations.

195 Table 2. The statistical results of the peripheral weak wind of 38 tropical cyclones for 7 radar stations in Guangdong

196

Province and ozone concentration from 2014 to 2018.

Radar station number	n/M	(n1 + n2)/n
59294	33/38 (87%)	(21+11)/33 (97%)
59486	32/33 (97%)	(18+12)/32 (94%)
59476	29/30 (97%)	(22+5)/29 (93%)
59285	33/36 (92%)	(21+12)/33 (100%)
59287	35/38 (92%)	(23+12)/35 (100%)
59284	24/25 (96%)	(19+5)/24 (100%)
59290	28/30 (93%)	(13+9)/28 (78%)
Ave.	93% (87%-97%)	94%(78%-100%)





198	The above correlation coefficients and statistical analysis indicate that WWD may be a common weather phenomenon
199	in the periphery of typhoon and could play an important impact on the ground-level ozone concentration. Therefore, the
200	following attempts to give the influence mechanism of WWD on ground-level ozone pollution and the impact of typhoon
201	peripheral circulation on sustained ozone enhancement during Typhoon Nepartak through a WRF-chem numerical
202	simulation.

203 **3.3 Model simulation and validation**

204 To investigate the impact of typhoon periphery and WWD on forming the sustained ozone episode, the numerical model 205 with the process analysis was applied in this study. Before applying the model to carry out any analysis, the model performance was validated by using the available observations. Figure S6a-d presents the measured and simulated data 206 207 for temperatures, wind speeds, wind directions, and ozone concentrations at Guangzhou from 00:00 on 3 July to 07:00 on 208 15 July 2016. With regards to the meteorological variables, there was good agreement between the measured and 209 modelled results, especially the shifting wind features, implying that the model successfully captured the synoptic 210 features. However, ozone concentrations (Fig. S6d) overestimated low values or underestimated high values some times. 211 But the simulated results and observed data reasonably agreed with each other and captured the ozone episode in the 212 region.

213 Statistical metrics including the index of agreement (IOA), mean bias (MB), root mean square error (RMSE), and 214 normalised mean bias (NMB) were used to further examine the model performance (Table 3). The IOA of the wind 215 direction was determined according to Kwok et al.(2010), and the IOA values for the other variables were calculated 216 following the approach of Lu et al.(1997). Generally, our simulation of the time series of ozone concentrations and 217 meteorological variables was reasonable. All the meteorological parameters were close to the corresponding simulation 218 results in the PRD region(Wang et al., 2006; Li et al., 2007; Hu et al., 2016). IOAs for temperature and wind speed (0.89 and 0.66, respectively) reached the criteria (as presented in the brackets of Table 3). The model performed well at 219 220 capturing the wind directions, with a small MB of 7.72°. MBs and NMBs for temperature and wind speed exceeded the





- 221 benchmarks; however, they are comparable to the findings of Li et al.(2013) with a slight overestimation, which is
- 222 probably due to the incomplete resolution of the urban morphology impact in the model(Chan et al., 2013). Moreover,
- ozone concentrations are generally well simulated, with an IOA of 0.84 and an NMB of 4.83. Time series comparisons of
- ozone concentrations and meteorological factors at Shenzhen, Zhongshan and Zhuhai are presented in Figs. S6a1-d1,
- a2-d2 and a3-d3. The overall results suggests that the model has the capability to reproduce ozone concentrations and
- 226 capture the transport features in southern China during this period.

Table 3. Statistical comparison between the observed and simulated variables. The benchmarks are based on Emery et al.(2007) and EPA (Doll, 1991). Values that did not reach the criteria are marked in grey.

Variable ^a	IOA ^b	MB ^b	RMSE ^b	NMB ^b (%)
Temp (°C)	0.89 (≥0.8)	0.75 (≤±0.5)	1.90	2.68
Wspd. (m s ⁻¹)	0.66 (≥0.6)	0.65 (≤±0.5)	1.45 (≤±2.0)	37.81
Wdir. (°)	0.77	7.72 (≤±10)	85.88	4.24
Ozone (ppbv)	0.84	9.53	37.15	4.83 (≤15)

229 a Temp. = temperature; Wspd. = wind speed; Wdir. = wind direction.

230 ^b IOA is the index of agreement; MB is the mean bias; RMSE is the root mean square error; NMB is the normalized mean bias.

3.4 IPR of the impact of typhoon peripheral circulation on sustained ozone enhancement and

232 influence mechanism of WWD on ground-level ozone

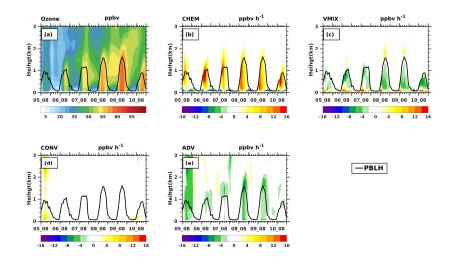
233 As variations in ozone concentration are directly caused by physical and chemical processes(Zhu et al., 2015), the fact 234 that peripheral circulation of a typhoon affects ozone concentration can be discussed using an IPR analysis. The 235 following processes were taken into account in this analysis: (1) advective transport (ADV), which is strongly related to 236 wind and ozone concentration gradients from upwind areas to downwind areas; (2) vertical mixing (VMIX) caused by 237 atmospheric turbulence and vertical gradients of ozone concentrations, which are related to variations in the PBL(Zhang 238 and Rao, 1999; Gao et al., 2017); (3) chemistry (CHEM), which is the result of chemical calculations that include ozone 239 chemical production and consumption; and (4) convective processes (CONV), i.e. the ozone contribution due to convective movements. Complete details on the analytical process of the WRF-Chem model can be found in previous 240 241 studies(J. Gao et al., 2016; H. Zhang et al., 2014) and in the WRF-Chem user guide.

Figure 6a shows the profile evolution of the average ozone concentrations in region D0 (black box D0 in Fig. 1) from





243 08:00 on 5 July to 20:00 on 10 July. The ozone concentrations gradually increased from 6 to 9 July throughout the PBL, 244 with an increase in PBL height of up to ~1.5 km. On 10 July, the PBL height decreased to less than 1 km, and the ozone 245 concentrations above 1 km decreased with the PBL; however, the regional average surface ozone concentrations were 246 still high but lower than that on 9 July. Figure 6b-e show the vertical distributions of the processes that contribute to the 247 ozone concentrations. 248 It can be seen from Fig. 6b-e: during the period from 08:00 to 20:00 on July 5-10, the contributions of CONV in PBL 249 were basically zero; CHEM on the ground were strong negative contributions, and VMIX on the ground were strong 250 positive contributions; ADV in PBL were weak negative contributions during 6 to 7 July, and the negative contributions 251 of ADV in PBL were strengthened on July 8 and 9. Therefore, the contributions of ground VMIX and CHEM played a 252 major role in the change of the PBL ozone concentrations, which is consistent with previous studies. At the same time, 253 the changes in the strength of ADV contributions in PBL might also have a certain impact on the changes in the ozone 254 concentrations on the ground.



255

256 Figure 6. The profile evolution of averaged (a) ozone concentration and (b)-(e) CHEM, VMIX, CONV, and ADV of region D0 from 08:00 5

257 July to 20:00 10 July. The black lines denote the planetary boundary layer height(PBLH).

258 In order to investigate the cause of the continued increase of the daytime ozone concentration during the sustained

259 ozone episode, the numerical relationship between the daytime average ozone concentration difference of two adjacent



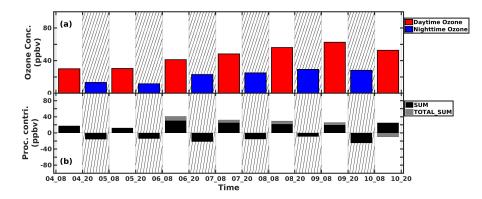


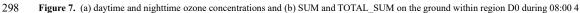
260 days and the various physical and chemical processes is need to be presented. In the numerical IPR analysis, the ozone 261 concentration at any location at time t+1 follows Eq. (1): $C_{t+1} = C_t + SUM_{t+1}$ 262 (1)where C_{t+1} and C_t are the ozone concentrations at time t+1 and time t, respectively. SUM_{t+1} is the net change in 263 contributions from all of the physical and chemical processes from time t to time t + 1, and is shown in Eq. (2): 264 $SUM_{t+1} = ADV_{t+1} + CHEM_{t+1} + VMIX_{t+1} + CONV_{t+1}$. (2) 265 266 As specified in Eqs. (1) and (2), ozone concentration is a cumulative amount. Then, according to Eq. (1), we obtain: $C_{t+24} - C_t = \sum_{i=1}^{j=24} SUM_{t+j}, (t = 08:00,09:00,...,20:00),$ (3) 267 268 where C_t and C_{t+24} are the ozone concentrations at the corresponding time on two adjacent days. For example, if C_t is the ozone concentration at 8:00 in the morning on a certain day, C_{t+24} represents the ozone concentration at 8:00 in the 269 270 next morning. SUM_{t+i} is the sum of the contributions from all of the physical and chemical processes at the 271 corresponding time over the time slots. For example, when t is 08:00, SUM_{08+1} indicates the SUM at 9:00 in the 272 morning, and SUM_{08+24} indicates the SUM at 8:00 in the next morning. To give the daytime average ozone 273 concentration difference of two adjacent days, we use 08:00 and 20:00 as the daytime and nighttime boundaries to reprocess the hourly data into a half-day average. If the daytime average ozone concentrations for two adjacent days are 274 denoted as C_{d1} and C_{d2} , the difference between the daytime average ozone concentrations on two adjacent days can be 275 276 further expressed by three continuous contribution terms from 09:00 on the first day (d1) to 20:00 on the second day (d2): $C_{d2} - C_{d1} = \frac{1}{N} \sum_{t_1=0}^{t_1=20} (t_1 - 8) \cdot SUM_{t_1} + \sum_{t_2=21}^{t_2=08} SUM_{t_2} + \frac{1}{N} \sum_{t_3=09}^{t_3=20} (21 - t_3) \cdot SUM_{t_3},$ (4) 277 278 where C_{d2} and C_{d1} are the daytime average ozone concentrations on two adjacent days. N is the total number of time slots for the daytime period. Due to the daytime period is between 08:00-20:00, N is 13. When the right side of Eq. (4)>0, 279 280 the daytime average ozone concentration will increase compared to the daytime average concentration from the previous 281 day, and vice versa. The three terms on the right side of Eq. (4) are referred to as $SUM_{d,d1}$, $SUM_{n,d1}$, and $SUM_{d,d2}$, 282 respectively. $SUM_{d,d1}$ and $SUM_{d,d2}$ reflect the daytime contributions on two adjacent days. $SUM_{n,d1}$ reflects the





283 nighttime contribution to difference in daytime average concentration (DDAC) between the two adjacent days. The sum 284 of these three terms is referred to as TOTAL SUM. It can be seen from Eq. (4): TOTAL SUM is consistent with the 285 evolution of daytime average ozone concentration, that is, when TOTAL SUM>0, daytime average ozone concentration 286 increases; when TOTAL SUM<0, daytime average ozone concentration decreases. It worthy note that the ozone chemistry between the daytime and nighttime is totally different. The SUM value during daytime is always positive 287 288 while the SUM of the nighttime is always negative. In terms of the daily daytime variation, the separated three terms 289 illiterate that the daily variation of daytime ozone level not only determined by the daytime increase but also influence by 290 the nighttime ozone variation between the two adjacent day. For example, the nighttime accumulation of ozone (as well 291 as precursors) could also contribute to the daytime ozone increase of the next day. It can be seen from Fig. 7, during the 292 daytime of 6-9 July, TOTAL SUM was positive, and the corresponding daytime average ozone concentrations gradually 293 increased; On the 10 July, TOTAL SUM was negative, and daytime average ozone concentration began to decrease; However, the daytime SUM on 10 July was still positive. The above analyses indicate that TOTAL_SUM can well 294 295 reflect the changing trend of daytime average ozone concentrations, so the cause of the sustained increase in ozone 296 concentrations can be analyzed according to Eq. (4).





299 July to 20:00 10 July

297

300 According to Eq. (2), Eq. (4) can be further decomposed into the following form:

301
$$C_{d2} - C_{d1} = \frac{1}{N} \sum_{t_1=09}^{t_1=20} (t_1 - 8) \cdot CHEM_{t1} + \sum_{t_2=21}^{t_2=08} CHEM_{t2} + \frac{1}{N} \sum_{t_3=09}^{t_3=20} (21 - t_3) \cdot CHEM_{t3}$$





302
$$+\frac{1}{N}\sum_{t1=09}^{t1=20}(t1-8)\cdot VMIX_{t1} + \sum_{t2=21}^{t2=08}VMIX_{t2} + \frac{1}{N}\sum_{t3=09}^{t3=20}(21-t3)\cdot VMIX_{t3}$$

$$3 + \frac{1}{N} \sum_{t_{1}=09}^{t_{1}=20} (t_{1}-8) \cdot \text{CONV}_{t_{1}} + \sum_{t_{2}=21}^{t_{2}=08} \text{CONV}_{t_{2}} + \frac{1}{N} \sum_{t_{3}=09}^{t_{3}=20} (21-t_{3}) \cdot \text{CONV}_{t_{3}}$$

304
$$+ \frac{1}{N} \sum_{t_{1}=0.9}^{t_{1}=20} (t_{1}-8) \cdot ADV_{t_{1}} + \sum_{t_{2}=21}^{t_{2}=0.8} ADV_{t_{2}} + \frac{1}{N} \sum_{t_{3}=0.9}^{t_{3}=20} (21-t_{3}) \cdot ADV_{t_{3}} .$$
 (5)

305 The decomposed items are respectively denoted as:

306 TOTAL_SUM_CHEM =
$$\frac{1}{N} \sum_{t_1=09}^{t_1=20} (t_1 - 8) \cdot CHEM_{t_1} + \sum_{t_2=21}^{t_2=08} CHEM_{t_2} + \frac{1}{N} \sum_{t_3=09}^{t_3=20} (21 - t_3) \cdot CHEM_{t_3}$$
,

307 TOTAL_SUM_VMIX =
$$\frac{1}{N} \sum_{t_1=09}^{t_1=20} (t_1 - 8) \cdot VMIX_{t_1} + \sum_{t_2=21}^{t_2=08} VMIX_{t_2} + \frac{1}{N} \sum_{t_3=09}^{t_3=20} (21 - t_3) \cdot VMIX_{t_3}$$

308 TOTAL_SUM_CONV =
$$\frac{1}{N} \sum_{t_1=09}^{t_1=20} (t_1 - 8) \cdot \text{CONV}_{t_1} + \sum_{t_2=21}^{t_2=08} \text{CONV}_{t_2} + \frac{1}{N} \sum_{t_3=09}^{t_3=20} (21 - t_3) \cdot \text{CONV}_{t_3}$$

309 TOTAL_SUM_ADV =
$$\frac{1}{N} \sum_{t_1=0.9}^{t_1=20} (t_1 - 8) \cdot ADV_{t_1} + \sum_{t_2=2.1}^{t_2=0.8} ADV_{t_2} + \frac{1}{N} \sum_{t_3=0.9}^{t_3=20} (21 - t_3) \cdot ADV_{t_3}$$
.

310 Equation (5) shows that the daytime average ozone concentration difference of two adjacent days is determined by

311 TOTAL_SUM_CHEM, TOTAL_SUM_VMIX, TOTAL_SUM_CONV and TOTAL_SUM_ADV.





313 Table 4. TOTAL SUM CHEM, TOTAL SUM VMIX, TOTAL SUM CONV, and TOTAL SUM ADV on the

314 ground.

315

Period	4 09 5 20	5 08 (20	(09 7 20	7 09 9 20	0.00.0.20	0.02.10.20
(ppbv)	4_08-5_20	5_08_6_20	6_08-7_20	7_08-8_20	8_08-9_20	9_08-10_20
TOTAL_SUM _CHEM	-138.16	-113.817	-133.376	-96.6765	-75.1189	-133.958
TOTAL_SUM _VMIX	118.845	113.4034	131.0915	88.912	70.3796	105.2292
TOTAL_SUM _CONV	33.7043	13.4999	-1.725	0.8075	-2.7115	12.1332
TOTAL_SUM_ADV	-13.9651	-3.3129	10.9665	15.0615	14.0091	6.9084
TOTAL_SUM_CAC	14.3893	13.0863	-4.0095	-6.9570	-7.4508	-16.5956
TOTAL_SUMs	0.4242	9.7734	6.957	8.1045	6.5583	-9.6872

2	1	4
3	1	C

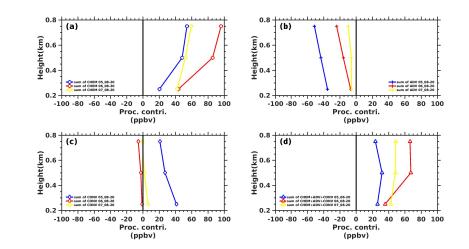
Note: the highlighted column indicate the non-attainment (national-II air quality standard) ozone period. 6 TOTAL SUM CAC is the sum of the TOTAL SUM (CHEM+VMIX+CONV). 317

318 On the ground, regarding to the daily variability of the daytime ozone levels during the episode, the details budget of the TOTAL SUM CHEM, TOTAL SUM VMIX and TOTAL SUM CONV during the episode are presented in Table 319 320 4. More specifically, the CHEM contribution is always negative and VMIX contribution is always positive on the ground which should be the result of the surface NO-titration effect. The CONV contribution is relatively small during the 321 episode (columns highlighted by brown color), while the ADV contribution significantly increased from negative value 322 323 to positive value from 4 July to 10 July. The TOTAL SUMs term is the sum of all the four processes 324 (CHEM+VMIX+CONV+ADV), which show a large daytime ozone enhancement from 5 July to 9 July and a ozone 325 decrease on 10 July. By calculating a sum of CHEM+VMIX+CONV (TOTAL_SUM_CAC in the Table 4) , we found 326 this three processes changed to negative values during the episode period, while the ADV term changed to positive 327 values and determined the sustained increase of daytime ozone. The results indicate that both the VMIX and ADV enhancement contributed to the daily increase of daytime ozone concentration from 6 to 9 July on the ground. 328 329 Because the VMIX contribution to the ground is closely dependent on the vertical gradient of the concentration and 330 the turbulence exchange coefficients (Gao et al. 2020). To confirm why the VMIX contribution to the surface ozone





- 331 reach the maximum (131.0915ppb) from 6 to 7 July, the vertical profiles of accumulative CHEM, ADV, CONV and CAC 332 (CHEM+ADV+CONV) during the time period from 08:00 to 20:00 on 5-7 July are shown in Fig. 8. (For example, the 333 accumulative of CHEM (i.e. the sum of CHEM) from 08:00 to 20:00 on 6 July is denoted as sum of CHEM 06 08-20). As shown, the gradient of vertical profile of accumulative CHEM contribution on 6 July was significantly larger than that 334 335 of vertical profiles of accumulative CHEM contribution on 5 July and 7 July (Fig. 8a). The CHEM increase in PBL 336 should be because of the impact of the periphery of Typhoon (Fig. 12a), which would produce a field of meteorological 337 conditions that was conducive to photochemical reactions. These meteorological conditions also led to an increase in the 338 absolute contribution and gradient of accumulative ADV contribution compared to that of 5 July (Fig. 8b). Therefore, the 339 vertical profile gradient of sum of ALL 06 08-20 was the largest, which contributed to the enhancement of VMIX 340 contribution to the ozone on the ground. In short, both the daytime CHEM and ADV enhancement above the 341 ground throughout the PBL contribute to the increase in VMIX contribution to the ground-level ozone. The 342 CHEM enhancement above the ground should be due to the increase in photochemical formations of precursors 343 within the PBL, while the ADV enhancement above the ground throughout the PBL should be a result of the 344 WWD (weak wind deepening) effect happened in the whole lower troposphere during the episode.
- 345



347 Figure 8. The vertical profiles of accumulative (a) CHEM, (b) ADV, (c) CONV, and (d) ALL (CHEM+ADV+CONV) during the periods





349	In summary, under the impact of the peripheral subsidence of typhoon, the chemical formation (CHEM) and vertical
350	mixing (VMIX) effects are two major contributors to the enhancement of ozone levels, while the ADV and CONV are
351	always negative values. But regarding the daily variability of the daytime ozone levels during the episode, the daily
352	daytime ozone levels do not associated with daily variation of daytime CHEM. By a detail analysis, it is found that the
353	decrease of the negative ADV throughout the PBL could also play an important role. Results show that the weak
354	subsidence associated with typhoon periphery provide the premise for the clear sky and warmer air, which is conducive
355	for the ozone photolysis formation in planetary boundary layer (PBL) above the ground where is dominated by
356	NO-titration effect. The WWD induced by the peripheral circulation of typhoon system provide the premise for the
357	enhanced contribution to ozone levels from daily ADV variation throughout the whole PBL, and the increased
358	contribution to the continue enhancement of ground-level ozone via the VMIX processes.

359 4. Conclusions

360 It is widely reported that the peripheral circulation of typhoon favors for the sustained ozone episodes. However, the 361 process how it impact on the ozone pollution levels during the episodes have not been clearly studies, which is crucial for 362 better prediction of the daily ozone variation during the episode. In this study, the analysis of ground observation, wind 363 profile data, and model simulation are integrated. By analyzing the wind profile radar observations, it was found that not 364 only surface weak winds but also WWD generally appeared in the periphery of Typhoon. The statistics of wind fields and 365 ground-level ozone at 7 wind profile radar stations in PRD during the 38 typhoons in the Northwestern Pacific Ocean 366 from 2014-2018 show that the number of WWD occurrences accounts for 93% (87-97%) of the available number of 367 radar stations for the seven radar stations in average. The number of ozone pollution occurrences accounts for 94% of the 368 number of WWD occurrences in average. The statistical results show that WWD is a common weather phenomenon in 369 the periphery of typhoons associated with pheriphery subsidence of typhoon systerm and is often accompanied by 370 significant increases in ozone concentrations.





371 The WRF-chem model was used to simulate the daily daytime ozone variation in a sustained ozone pollution process 372 in PRD during Typhoon Nepartak in 2016. Validation results show that the model could reasonably reproduce the 373 observed temperature, wind speed, wind direction and O3. Process analysis results show that under the impact of the 374 peripheral subsidence of typhoon, the chemical formation (CHEM) and vertical mixing (VMIX) effects are two major 375 contributors to the enhancement of ozone levels, while the ADV and CONV are always negative or small values. But 376 regarding the daily variability of the daytime ozone levels during the episode, the day-to-day variation of the daytime 377 ozone levels do not determined by the daily variation of daytime CHEM and VMIX, but the ADV term. By a detail 378 day-to-day analysis, it is found that the decrease of the negative ADV on the ground and throughout the PBL play an 379 important role. The integrated effect of the day-to-day variation of the accumulative CHEM above the ground and 380 accumulative ADV contribution throughout the PBL determined the overall day-to-day daytime ozone variation through 381 the VMIX process. The enhanced VMIX contribution associated both to the enhanced CHEM and enhanced ADV in the 382 above PBL.

383 Results show that the weak subsidence associated with typhoon periphery provide the premise for the clear sky and 384 warmer air, which is conducive for the ozone photolysis formation in planetary boundary layer (PBL) above the ground 385 where is dominated by NO-titration effect. The WWD induced by the peripheral circulation of typhoon system provide 386 the premise for the enhanced contribution to ozone levels from daily ADV variation throughout the whole PBL, and the 387 increased contribution to the continue enhancement of ground-level ozone via the VMIX processes. It shows that the 388 peripheral characteristics of approaching typhoon not only form the ozone episode by the enhanced photochemical 389 reactions but also the increase in pollution accumulation throughout the PBL due to the weak wind deepening up to 3~5 390 km (but not a stability condition in thermodynamics). This result explains why daytime ozone continues to increase, 391 although the photochemical contribution began to decrease during the event. It also indicate the important role of the 392 WWD in the lower troposphere for the formation of sustained ozone episodes due to the peripheral circulation of the 393 typhoon, which helps to better predict the daily changes of daytime ozone levels.





395

- Author contributions. YL and XZ designed and led the study. JG performed model simulations. XZ and YL analyzed data
 and interpreted results. XZ, YL and XD have discussed the results and commented on the paper. XZ wrote the paper with
 input from all coauthors.
- 401 *Competing interests.* The authors declare that they have no conflict of interest.
- 402
- 403 Acknowledgements. We would like to acknowledge the National Centers for Environmental Prediction (NCEP) for the
- 404 Final Operational Global Analysis data which are freely obtained from the website https://rda.ucar.edu/datasets/ds083.2/.
- 405 The hourly ambient surface O₃ concentration are real-timely released by Ministry of Environmental Protection, China on
- 406 the website http://www.aqistudy.cn/, freely downloaded from http://106.37.208.233:20035/. The meteorological datas,
- 407 such as the wind profiler data, automatic weather station data, cloud data and so on, were provided by the China
- 408 Meteorological Administration and downloaded from http://172.22.1.175. This research was supported by the National
- 409 Natural Science Foundation of China (Grant 41961160728), the Guangdong Province Science and Technology Planning
- 410 Project of China (Grant 2017A050506003), and Shenzhen Peacock Teams Plan (KQTD20180411143441009).
- 411
- 412
- 413





414 **■ References**

- Aunan, K., Berntsen, T. K., and Seip, H. M.: Surface Ozone in China and its Possible Impact on
 Agricultural Crop Yields, AMBIO J. Hum. Environ., 29, 294–301, 2000.
- Chan, A., Fung, J. C. H., and Lau, A. K. H.: Influence of urban morphometric modification on regional
 boundary-layer dynamics, J. Geophys. Res. Atmospheres, 118, 2729–2747, 2013.
- 419 Chen, X. L., Fan, S. J., Jiang-Nan, L. I., Ji, L., Wang, A. Y., and Soi-Kun, F.: typical weather 420 characteristics associated with air pollution in Hong Kong area, J. Trop. Meteorol., 014, 101–104, 2008.
- 421 Chen, Z., Zhuang, Y., Xie, X., Chen, D., Cheng, N., Yang, L., and Li, R.: Understanding long-term
- variations of meteorological influences on ground ozone concentrations in Beijing During 2006-2016.,
- 423 Environ. Pollut., 245, 29–37, 2018.
- Cheng, N. L., Li, Y. T., Zhang, D. W., Chen, T., Wang, X., Huan, N., Chen, C., and Meng, F.:
 Characteristics of Ozone over Standard and Its Relationships with Meteorological Conditions in Beijing
- 426 City in 2014, Environ. Sci., 37, 2016.
- 427 Deng, T., Wang, T., Wang, S., Zou, Y., Yin, C., Li, F., Liu, L., Wang, N., Song, L., and Wu, C. and: Impact
- 428 of typhoon periphery on high ozone and high aerosol pollution in the Pearl River Delta region, Sci. Total
- 429 Environ., 668, 617–630, 2019.
- 430 Doll, D. C.: Guideline for Regulatory Application of the Urban Airshed Model, 1991.
- 431 Emery, C., Tai, E., and Yarwood, G.: Enhanced meteorological modeling and performance evaluation for
- 432 two texas episodes, in: Prepared for the Texas Natural Resource Conservation Commission, by Environ
- 433 International Corp, 2007.
- Felzer, B. S., Cronin, T., Reilly, J. M., Melillo, J. M., and Wang, X.: Impacts of ozone on trees and crops,
 Comptes Rendus Géoscience, 339, 784–798, 2007.
- Feng, Y., Wang, A., Wu, D., and Xu, X.: The influence of tropical cyclone Melor on PM(10)
 concentrations during an aerosol episode over the Pearl River Delta region of China: Numerical modeling
 versus observational analysis, Atmos. Environ., 41, p.4349-4365, 2007.
- Feng, Z., Hu, E., Wang, X., Jiang, L., and Liu, X.: Ground-level O-3 pollution and its impacts on food
 crops in China: A review, Environ. Pollut., 199, 42–48, 2015.
- Forkel, R., Werhahn, J., Hansen, A. B., Mckeen, S., Peckham, S., Grell, G., and Suppan, P.: Effect of
 aerosol-radiation feedback on regional air quality A case study with WRF/Chem, Atmos. Environ., 53,
 202–211, 2012.
- Gao, J., Zhu, B., Xiao, H., Kang, Hou, X., and Shao, P.: A case study of surface ozone source
 apportionment during a high concentration episode, under frequent shifting wind conditions over the
 Yangtze River Delta, China, Sci. Total Environ., 544, 853–863, 2016.
- 447 Gao, J., Zhu, B., Xiao, H., Kang, H., Hou, X., Yin, Y., Zhang, L., and Miao, Q.: Diurnal variations and 448 source apportionment of ozone at the summit of Mount Huang, a rural site in Eastern China, Environ.
- 449 Pollut., 222, 513–522, 2017.





- Gao, J., Li, Y., Zhu, B., Hu, B., Wang, L., and Bao, f.: What have we missed when studying the impact of aerosols on surface ozone via changing photolysis rates?, Atmospheric Chem. Phys., 10831-10844, 2020.
- Giorgi, F. and Meleux, F.: Modelling the regional effects of climate change on air quality, Comptes
 Rendus Geosci., 339, 721–733, 2007.
- 454 Grell, G. A., Peckham, S. E., Schmitz, R., Mckeen, S. A., Frost, G., Skamarock, W. C., and Eder, B.: Fully
- 455 coupled "online" chemistry within the WRF model, 2005.
- 456 Han, H., Liu, J., Shu, L., Wang, T., and Yuan, H.: Local and synoptic meteorological influences on daily
- variability of summertime surface ozone in eastern China, Atmospheric Chem. Phys., 1–51, 2019.
- Hu, J., Chen, J., Ying, Q., and Zhang, H.: One-Year Simulation of Ozone and Particulate Matter in China
 Using WRF/CMAQ Modeling System, Atmospheric Chem. Phys. Discuss., 16, 10333–10350, 2016.
- 460 Huang, J., Liu, H., Crawford, J. H., Chan, C., Considine, D. B., Zhang, Y., Zheng, X., Zhao, C., Thouret,
- V., and Oltmans, S. J.: Origin of springtime ozone enhancements in the lower troposphere over Beijing: in
- situ measurements and model analysis, 15, 5161–5179, 2015.
- 463 Jiang, Y. C., Zhao, T. L., Liu, J., Xu, X. D., Tan, C. H., Cheng, X. H., Bi, X. Y., Gan, J. B., You, J. F., and
- 464 Zhao, S. Z.: Why does surface ozone peak before a typhoon landing in southeast China?,
- 465 ATMOSPHERIC Chem. Phys., 15, 13331–13338, 2015.
- Kwok, R. H. F., Fung, J. C. H., Lau, A. K. H., and Fu, J. S.: Numerical study on seasonal variations of
 gaseous pollutants and particulate matters in Hong Kong and Pearl River Delta Region, J. Geophys. Res.
 Atmospheres, 115, 2010.
- Lai, L. Y. and Sequeira, R.: Visibility degradation across Hong Kong: its components and their relative contributions, Atmos. Environ., 35, 5861–5872, 2001.
- Li, J., Wang, Z., Akimoto, H., Gao, C., Pochanart, P., and Wang, X.: Modeling study of ozone seasonal cycle in lower troposphere over east Asia, J. Geophys. Res. Atmospheres, 112, 2007.
- Li, Y., Lau, A. K. H., Fung, J. C. H., Ma, H., and Tse, Y.: Systematic evaluation of ozone control policies
 using an Ozone Source Apportionment method, Atmos. Environ., 76, 136–146,
 https://doi.org/10.1016/j.atmosenv.2013.02.033, 2013.
- Li, Y., Lau, A., Wong, A., and Fung, J.: Decomposition of the wind and nonwind effects on observed
 year-to-year air quality variation, J. Geophys. Res. Atmospheres, 119, 6207–6220, 2014.
- 478 Lin, X., Yuan, Z., Yang, L., Luo, H., and Li, W.: Impact of Extreme Meteorological Events on Ozone in 479 the Pearl River Delta, China, Aerosol Air Qual. Res., 19, 1307-1324, 480 https://doi.org/10.4209/aagr.2019.01.0027, 2019.
- Liu, J., Wu, D., Fan, S. J., Liao, Z. H., and Deng, T.: Impacts of precursors and meteorological factors on
 ozone pollution in Pearl River Delta, Zhongguo Huanjing Kexuechina Environ. Sci., 37, 813–820, 2017.
- Lu, R., Turco, R. P., and Jacobson, M. Z.: An integrated air pollution modeling system for urban and
 regional scales: 2. Simulations for SCAQS 1987, J. Geophys. Res. Atmospheres, 102, 6081–6098,
 https://doi.org/10.1029/96JD03502, 1997.





- 486 Ministry of Ecology and Environment of China: Chinese State of the Environment Bulletin, 1–54, 2016.
- 487 Organization, W. H.: WHO Air quality guidelines for particulate matter, ozone, nitrogen dioxide and488 sulfur dioxide Global update 2005, 2005.
- 489 Shu, L., Xie, M., Wang, T., Gao, D., Chen, P., Han, Y., Li, S., Zhuang, B., and Li, M.: Integrated studies of
- 490 a regional ozone pollution synthetically affected by subtropical high and typhoon system in the Yangtze
- 491 River Delta region, China, Atmospheric Chem. Phys., 16, 15801–15819, 2016.
- 492 Shu, L., Wang, T., Xie, M., Li, M., Zhao, M., Zhang, M., and Zhao, X.: Episode study of fine particle and
- 493 ozone during the CAPUM-YRD over Yangtze River Delta of China: Characteristics and source
- 494 attribution, Atmos. Environ., 203, 87–101, https://doi.org/10.1016/j.atmosenv.2019.01.044, 2019.
- Skamarock, W. C., Klemp, J. B., Dudhia, J., Gill, D. O., Barker, D. M., Duda, M. G., Huang, X.-Y., Wang,
 W., and Powers, J. G.: A Description of the Advanced Research WRF Version 3, 125, n.d.
- 497 Tan, Z., Lu, K., Jiang, M., Su, R., Dong, H., Zeng, L., Xie, S., Tan, Q., and Zhang, Y.: Exploring ozone
- 498 pollution in Chengdu, southwestern China: A case study from radical chemistry to O3 -VOC-NOx
- 499 sensitivity, Sci. Total Environ., 636, 775–786, 2018.
- 500 Wang, N.: Guangdong Haze Weather Bulletin, 21 pp., 2017.
- 501 Wang, T., Lam, K. S., Lee, A. S. Y., Pang, S. W., and Tsui, W. S.: Meteorological and Chemical
- 502 Characteristics of the Photochemical Ozone Episodes Observed at Cape D'Aguilar in Hong Kong, J. Appl.
- 503 Meteorol., 37, 1167–1178, 1998.
- Wang, T., Wu, Y. Y., Cheung, T. F., and Lam, K. S.: A study of surface ozone and the relation to complex
 wind flow in Hong Kong, Atmos. Environ., 35, 3203–3215, 2001.
- 506 Wang, X., Zhang, Y., Hu, Y., Zhou, W., and Russell, A. G.: Process analysis and sensitivity study of
- regional ozone formation over the Pearl River Delta, China, during the PRIDE-PRD2004 campaign using
 the CMAQ model, Atmospheric Chem. Phys. Discuss., 9, 635–645, 2009.
- 509 Wang, Z., Li, J., Wang, X., Pochanart, P., and Akimoto, H.: Modeling of Regional High Ozone Episode
- Observed at Two Mountain Sites (Mt. Tai and Huang) in East China, J. Atmospheric Chem., 55, 253–272, 2006.
- Wu, D., Tie, X., Li, C., Ying, Z., Lau, K. H., Huang, J., Deng, X., and Bi, X.: An extremely low visibility
 event over the Guangzhou region: A case study, Atmos. Environ., 39, p.6568-6577, 2005.
- Wu, M., Wu, D., Fan, Q., Wang, B. M., Li, H. W., and Fan, S. J.: Observational studies of the
 meteorological characteristics associated with poor air quality over the Pearl River Delta in China,
 Atmospheric Chem. Phys., 13, 10755–10766, https://doi.org/10.5194/acp-13-10755-2013, 2013.
- Zaveri, R. A. and Peters, L. K.: A new lumped structure photochemical mechanism for large-scale
 applications, J. Geophys. Res. Atmospheres, 104, 30387–30415, 1999.
- Zaveri, R. A., Easter, R. C., Fast, J. D., and Peters, L. K.: Model for Simulating Aerosol Interactions and
 Chemistry (MOSAIC), J. Geophys. Res. Atmospheres, 113, 2008.
- 521 Zhang, H., DeNero, S. P., Joe, D. K., Lee, H.-H., Chen, S.-H., Michalakes, J., and Kleeman, M. J.:





- 522 Development of a Source Oriented version of the WRF- Chem Model and its Application to the California
 523 Regional PM10/PM2.5 Air Quality Study, 20, 2014.
- 524 Zhang, J. and Rao, S. T.: The Role of Vertical Mixing in the Temporal Evolution of Ground-Level Ozone
- 525 Concentrations, J. Appl. Meteorol., 38, 1674–1691, 1999.
- 526 Zhang, J. P., Zhu, T., Zhang, Q. H., Li, C. C., Shu, H. L., Ying, Y., Dai, Z. P., Wang, X., Liu, X. Y., and
- 527 Liang, A. M.: The impact of circulation patterns on regional transport pathways and air quality over
- 528 Beijing and its surroundings, Atmospheric Chem. Phys., 12, 5031–5053, 2012.
- 529 Zhang, Y., Wen, X. Y., and Jang, C. J.: Simulating chemistry-aerosol-cloud-radiation-climate feedbacks
- 530 over the continental U.S. using the online-coupled Weather Research Forecasting Model with chemistry
- 531 (WRF/Chem), Atmos. Environ., 44, p.3568-3582, 2010.
- Zhang, Y., Mao, H., Ding, A., Zhou, D., and Fu, C.: Impact of synoptic weather patterns on
 spatio-temporal variation in surface {O3} levels in Hong Kong during 1999–2011, Atmos. Environ., 73,
 41–50, 2013.
- 535 Zhu, B., Kang, H., Zhu, T., Su, J., Hou, X., and Gao, J.: Impact of Shanghai urban land surface forcing on
- 536 downstream city ozone chemistry: URBAN LAND-SURFACE FORCING ON OZONE, J. Geophys. Res.
- 537 Atmospheres, 120, 4340–4351, https://doi.org/10.1002/2014JD022859, 2015.
- Ziomas, I. C., Melas, D., Zerefos, C. S., Bais, A. F., and Paliatsos, A. G.: Forecasting peak pollutant levels
- from meteorological variables, Atmos. Environ., 29, 3703–3711, 1995.