

1 Free amino acids quantification in cloud water at the puy de Dôme 2 station (France)

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12 **Abstract.**

13 Eighteen free amino acids (FAAs) were quantified in cloud water sampled at the puy de Dôme station (PUY - France)
14 during 13 cloud events. This quantification has been performed without concentration neither derivatization, using LC-
15 MS and the standard addition method to correct for matrix effects. Total concentrations of FAAs (TCAAs) vary from 1.2
16 μM to 7.7 μM , Ser (Serine) being the most abundant AA (23.7 % in average) but with elevated standard deviation,
17 followed by Glycine (Gly) (20.5 %), Alanine (Ala) (11.9 %), Asparagine (Asn) (8.7 %), and Leucine/Isoleucine (Leu/I)
18 (6.4 %). The distribution of AAs among the cloud events reveals high variability. TCAA constitutes between 0.5 and 4.4 %
19 of the dissolved organic carbon measured in the cloud samples. AAs quantification in cloud water is scarce but the results
20 agree with the few studies that investigated AAs in this aqueous medium. The environmental variability is assessed
21 through a statistical analysis. This work shows that AAs are correlated with the time spent by the air masses within the
22 boundary layer, especially over the sea surface before reaching the PUY. The cloud microphysical properties fluctuation
23 does not explain the AAs variability in our samples confirming previous studies at PUY. We finally assessed the sources
24 and the atmospheric processes that potentially explain the prevailing presence of certain AAs in the cloud samples. The
25 initial relative distribution of AAs in biological matrices (proteins extracted from bacterial cells or mammalian cells, for
26 example) could explain the dominance of Ala, Gly and Leu/I. AA composition of aquatic organisms (*i.e.*, diatoms species)
27 could also explain the high concentrations of Ser in our samples. The analysis of the AAs hygroscopicity also indicates a
28 higher contribution of AAs (80 % in average) that are hydrophilic or neutral revealing the fact that other AAs
29 (hydrophobic) are less favorably incorporated into cloud droplets. Finally, the atmospheric aging of AAs has been
30 evaluated by calculating atmospheric lifetimes considering their potential transformation in the cloud medium by biotic
31 or abiotic (mainly oxidation) processes. The most concentrated AAs encountered in our samples present the longest
32 atmospheric lifetimes and the less dominant are clearly efficiently transformed in the atmosphere, potentially explaining
33 their low concentrations. However, this cannot fully explain the relative contribution of several AAs in the cloud samples.
34 This reveals the high complexity of the bio-physico-chemical processes occurring in the multiphase atmospheric
35 environment.

36 1. Introduction

37 Free or combined amino acids (AAs) that make up proteins and cell walls in living organisms are ubiquitous chemical
38 compounds found in various environments. In the atmosphere, they are commonly detected in the condensed phases due
39 to their low vapor pressures. They have been studied and characterized in atmospheric particles (Barbaro et al., 2020;
40 Matos et al., 2016), rain water (Mace et al., 2003a; Mace et al., 2003b; Xu et al., 2019; Yan et al., 2015), fog water (Zhang
41 and Anastasio, 2003b) and more recently in cloud water (Bianco et al., 2016b; Triesch et al., 2021). Many efforts have
42 been made in the past to assess their sources, their role in the atmospheric chemical and physical processes and their fate
43 (Cape et al., 2011). However, despite those investigations, their exact role in the atmosphere is still poorly understood.
44 They have been studied for their hygroscopic properties since they can modify the ability of the particle to act as cloud
45 condensation nuclei (CCN) (Chan et al., 2005; Kristensson et al., 2010; Li et al., 2013) or ice nuclei (IN) (Pummer et al.,
46 2015; Szyrmer and Zawadzki, 1997). More recently, the role of AAs in new particle formation has been also discussed
47 (Ge et al., 2018). This raises the question of their role in aerosol and cloud formation and hence in the radiative forcing
48 of the Earth's surface. In atmospheric aqueous phases, some AAs have been found to potentially influence atmospheric
49 chemistry by reacting with atmospheric oxidants (Bianco et al., 2016b; McGregor and Anastasio, 2001; Zhang and
50 Anastasio, 2003a); the study from De Haan et al. even showed that AAs can react with glyoxal to form secondary aerosol
51 mass (De Haan et al., 2011). AAs are part of the proteinaceous fraction of aerosol particle that significantly contribute to
52 the organic carbon and organic nitrogen fraction of aerosol particles. Their presence in aerosol particles can modify their
53 chemical properties such as acidity/basicity and buffering ability (Cape et al., 2011; Zhang and Anastasio, 2003b). Finally,
54 AAs are also transferred by atmospheric deposition to other ecosystems such as aquatic surfaces where they act as
55 nutrients since they are particularly bioavailable (Wedyan and Preston, 2008). Atmospheric AAs can therefore contribute
56 to the nutrient cycling at global scale as well as the global carbon and nitrogen cycles.

57 AAs have been detected in the atmosphere under various contrasted environmental scenarios such as urban area (Barbaro
58 et al., 2011; Di Filippo et al., 2014; Ren et al., 2018; Zhu et al., 2020), background/rural sites (Bianco et al., 2016b; Helin
59 et al., 2017; Samy et al., 2011; Song et al., 2017), marine environment (Mandalakis et al., 2011; Matsumoto and Uematsu,
60 2005; Triesch et al., 2021; Violaki and Mihalopoulos, 2010) and polar regions (Barbaro et al., 2015; Feltracco et al., 2019;
61 Mashayekhy Rad et al., 2019; Scalabrin et al., 2012). The quantity and type of AAs detected in all the compartments
62 (aerosol particles, cloud water, rainwater) vary over a wide range. Indeed, their emissions, residence times and spatial and
63 temporal distributions are driven by complex bio-physico-chemical processes occurring in the atmosphere (transport,
64 chemical and biological transformations, deposition, *etc.*). Proteinaceous materials detected in the atmosphere are in
65 majority linked with emissions of primary biological aerosol particles that notably include viruses, bacteria, fungi, algae,
66 spores, pollens and fragments of plants and insects (Després et al., 2012; Fröhlich-Nowoisky et al., 2016). The main
67 source is consequently from biogenic origin, but several anthropogenic sources can also contribute (industry, agricultural
68 practices, wastewater treatment). It is suggested that AAs are directly emitted into the atmosphere or result from the
69 transformations of proteins by enzymatic activity, decomposition by the temperature or the photochemistry (Mopper and
70 Zika, 1987). There are some studies highlighting other possible sources such as emissions by volcanoes (Scalabrin et al.,
71 2012), biomass burning emissions (Chan et al., 2005) and marine emissions by sea bubble bursting (Barbaro et al., 2015;
72 Matsumoto and Uematsu, 2005). Due to the wide variety of AAs sources in the atmosphere, it is rather difficult to correlate
73 AA concentration and speciation with specific sources: Abe et al. (2016) recently proposed to use AAs as markers for
74 biological sources in urban aerosols (Abe et al., 2016); Matsumoto and Uematsu (2005) suggested that the major source

75 of free amino acids (FAAs) in aerosols over the remote North Pacific are related to long-range transport from continental
76 areas; Scalabrin et al. (2012) used AAs ratio to evaluate aerosol aging in the atmosphere.

77 The analysis of AAs in the atmosphere is essential and has been widely conducted to document the concentrations of
78 aerosol particles, their environmental variability, and their effects on atmospheric physico-chemical processes. AAs can
79 also be transferred into the atmospheric aqueous media after activation of aerosol particles into cloud droplets. They
80 consequently contribute to the complex dissolved organic matter measured in clouds that is composed by a significant
81 fraction of biological-derived material (lipids, peptides, carbohydrates...) (Bianco et al., 2018; Cook et al., 2017; Zhao et
82 al., 2013). However, only few studies focus on the detection of AAs in cloud water (Bianco et al., 2016b; Triesch et al.,
83 2021) mainly because of the inherent difficulty to sample clouds. AA concentration in cloud water results from the
84 dissolution of the soluble fraction of the aerosol particles acting as CCN and IN; some very recent studies also argue that
85 AAs could be processed in the cloud medium by the biological activity (Bianco et al., 2019). For instance, the
86 biodegradation of AAs was demonstrated to occur in rainwater (Xu et al., 2020) and in microcosms mimicking the cloud
87 environment (Jaber et al., 2021). The presence of transcripts of genes coding for AAs and proteins biosynthesis and
88 biodegradation has been also shown directly in cloud water samples (Amato et al., 2019). AAs can also be photo-
89 transformed by abiotic processes mainly implying oxidants (Jaber et al., 2021). They can produce other compounds such
90 as carboxylic acids, nitrate, and ammonia (Berger et al., 1999; Berto et al., 2016; Bianco et al., 2016a; Marion et al., 2018;
91 Pattison et al., 2012), thus potentially contributing to the formation in aqueous phase of secondary organic aerosol
92 (“aqSOA”). It is therefore crucial to document AA concentration levels and speciation in clouds.

93 This aim of this work is devoted to the quantification of FAAs in cloud waters. This is quite a challenge due to the
94 chemical complexity of the cloud medium and the low concentration of FAAs ($\approx \mu\text{M}$). In atmospheric waters, namely
95 fog (Zhang and Anastasio, 2003b), rain (Gorzelska et al., 1992; Mopper and Zika, 1987; Xu et al., 2019; Yan et al., 2015)
96 and clouds (Bianco et al., 2016b), the main technique that has been commonly used is liquid chromatography coupled
97 with fluorescence detection. This approach is based on pre- or post-column derivatization of the AAs to increase the
98 sensitivity and simplify the separation by chromatography, but it is time consuming. More recently, Triesch et al. (2021)
99 have used liquid chromatography hyphenated to mass spectrometry (LC-MS) to detect derivatized AAs after concentration
100 of cloud water samples. The use of LC-MS represents a significant improvement as it allows a unique identification. We
101 propose here to go further using LC-MS without pre-concentration and derivatization of the sample. In addition, to
102 overcome matrix effect, we propose to quantify the AAs by the standard addition method (Hewavitharana et al., 2018).
103 Cloud sampling is performed at the puy de Dôme station (PUY) in France offering possibility to collect 13 samples for
104 various environmental conditions. Variability of cloud AA concentrations together with cloud bio-physico-chemical
105 properties and air mass history is thus discussed in this work.

106 **2. Methods/Materials**

107 **2.1 Site and cloud sampling**

108 13 clouds were sampled from 2014 to 2020 at the puy de Dôme station (PUY) in France (45.77 °N, 2.96 °E; 1465m a.s.l.).
109 This mountain observatory is part of the multi-site platform CO-PDD combining *in situ* and remote sensing observations
110 at different altitudes (Baray et al., 2020). PUY belongs to international atmospheric survey networks: ACTRIS (Aerosols,
111 Clouds, and Trace Gases Research Infrastructure), EMEP (the European Monitoring and Evaluation Program) and GAW
112 (Global Atmosphere Watch) as example. Meteorological parameters, atmospheric gases, aerosols, and clouds are

113 monitored over long-term period, to investigate the bio-physico-chemical processes linking those elements and to evaluate
114 the anthropogenic forcing on climate.

115 The sampling is performed using aluminum cloud water collectors under non-precipitating and non-freezing conditions
116 as described in Deguillaume et al. (2014). Cloud droplets are collected by impaction onto a rectangular plate which then
117 flows directly into a sterilized bottle going through a funnel. The impactor has an estimated cut-off diameter of 7 μm .
118 Before cloud collection, cloud impactors are cleaned using milliQ water and sterilized by autoclaving. Immediately after
119 sampling, a fraction of the aqueous volume is filtered using a 0.2 μm nylon filter (Fisherbrand™) to eliminate
120 microorganisms. The samples are then stored in the dark and frozen at -20 °C (ATP, ion chromatography, total organic
121 carbon, and amino acids). For cell counts, samples are stored at 4 °C after adding a fixative. The analyzes are performed
122 shortly thereafter.

123 **2.2 Physical, chemical, and microbiological characterization of clouds**

124 A systematic characterization is performed on cloud samples allowing to document the available PUYCLOUD database
125 (http://opgc.fr/vobs/so_interface.php?so=puycloud) of the cloud water chemical and biological composition (Renard et
126 al., 2020). These data are reported in Table S1 for the studied cloud events.

127 Chemical composition analyses are performed on cloud samples: pH, total organic carbon (TOC) concentration, and
128 concentrations of main inorganic ionic species. TOC analyses are performed with a TOC analyzer (Shimadzu, TOC-
129 5050A). The spectrofluorimetric method based on the reactivity of p-hydroxyphenilacetic acid with horseradish
130 peroxidase is used to measure the concentration of H₂O₂ in cloud water (Wirgot et al., 2017). Ionic inorganic species
131 (Ca²⁺, K⁺, Mg²⁺, Na⁺, NH₄⁺, Cl⁻, SO₄²⁻ and NO₃⁻) are measured by ion chromatography (Deguillaume et al., 2014).

132 Cloud microphysical properties are determined with the Gerber particle volume monitor-100 (PVM-100) providing liquid
133 water content (LWC) and effective droplet radius (r_e) parameters (Gerber, 1991).

134 The biology of cloud water is also assessed by quantification of bacteria density (CFU mL⁻¹) at 17°C (Vaïtilingom et al.,
135 2012) and adenosine triphosphate (ATP) concentration is measured using the BioThema[®] ATP Biomass kit HS (Koutny
136 et al., 2006).

137 **2.3 Quantification of Amino Acids (AAs)**

138 **2.3.1 Sample preparation**

139 Before analysis by LC-MS, in order to apply the standard addition method to quantify AAs (Hewavitharana et al., 2018),
140 standard solutions are used to spike cloud water samples. Standard solutions are prepared in ultra-pure water and
141 contained alanine (Ala, SIGMA-ALDRICH), arginine (Arg, SIMAFEX), asparagine (Asn, SIGMA-ALDRICH),
142 aspartate (Asp, SIGMA-ALDRICH), glutamine (Gln, SIGMA-ALDRICH) glutamic acid (Glu, SIGMA-ALDRICH),
143 glycine (Gly, MERCK), histidine (His, SIGMA-ALDRICH), leucine/isoleucine (Leu/I, SIGMA-ALDRICH), lysine (Lys,
144 SIGMA-ALDRICH), methionine (Met, SIGMA-ALDRICH), phenylalanine (Phe, ACROS organics), proline (Pro,
145 SIGMA-ALDRICH), serine (Ser, SIGMA-ALDRICH), threonine (Thr, SIGMA-ALDRICH), tryptophan (Trp, SIGMA-
146 ALDRICH), tyrosine (Tyr, SIGMA-ALDRICH), valine (Val, SIGMA-ALDRICH), cysteine (Cys, SIGMA-ALDRICH).
147 The ratio between the sample volume and the standard solution volume is respectively (9:1). The mixture is then vortex
148 mixed for 1 min.

149 Ten samples ready for LC-MS analysis are prepared, from approximately 1 mL of cloud water, containing the original
150 cloud water added with 20 AAs at final concentrations set to 1.0, 5.0, 10, 25, 50, 100, 150, 500 $\mu\text{g L}^{-1}$. This range of
151 concentrations is appropriate considering previous quantification of AAs in cloud waters sampled at PUY (Bianco et al.,
152 2016b). This also allows to cover large range of AA concentrations that can be highly variable depending on the cloud
153 events. L-Lys isotope ($^{13}\text{C}_6$, 99 %; $^{15}\text{N}_2$, 99 %) is also added to each sample at the concentration of 15 $\mu\text{g L}^{-1}$ for mass
154 calibration ($m/z = 155.1273$).

155 2.3.2 LC-MS analysis

156 LC-MS analyzes are performed using an UltiMate™ 3000 (Thermo Scientific™) LC equipped with a Q-Exactive™
157 Hybrid Quadrupole-Orbitrap™ Mass Spectrometer (Thermo Scientific™) ionization chamber. Chromatographic
158 separation of the analytes is performed on BEH Amide/HILIC (1.7 μm , 100 mm \times 2.1 mm) column with column
159 temperature of 30°C. The mobile phases consist of 0.1 % formic acid and water (A) and 0.1% formic acid and acetonitrile
160 (B) with a 0.4 mL min^{-1} flow rate. A four-step linear gradient is applied during the analysis: 10 % A and 90 % B in 8 min,
161 42 % A and 58% B in 0.1 min, 50 % A and 50 % B for 0.9 min and 10 % A and 90 % B for 3 min.

162 The Q-Exactive ion source is equipped with electrospray ionization (ESI) and the Q-Orbitrap™. The volume of injection
163 is 5 μL , and the flow injection analyses are performed for individual AA solutions to obtain the mass spectral data, from
164 which ions are carefully chosen for analysis in the selected ion monitoring (SIM) mode, using the above-mentioned
165 parameter conditions. The mass resolution is set to 35000 FWHM (full width at half maximum), and the instrument is
166 tuned for maximum ion throughput. AGC (automatic gain control) target or the number of ions to fill C-Trap is set to 10^5
167 with injection time of 100 ms. Tests with standard solution and cloud water samples show a better sensitivity in positive
168 mode of ionization for all AAs and with a preference for $[\text{M}+\text{H}]^+$ ionization (ESI+). Other Q-Exactive™ generic
169 parameters are: N_2 flow rate set at 13 a.u, sheath gas (N_2) flow rate set at 50 a.u, sweep gas flow rate set at 2 a.u, spray
170 voltage set at 3.2 kV in positive mode, capillary temperature set at 320°C, and heater temperature set at 425°C.

171 Analysis and visualization of the data set are performed using Xcalibur™ 2.2 software; it allows controlling and
172 processing data from Thermo Scientific™ LC-MS systems and associated instruments. Examples of chromatograms and
173 MS spectra for three AAs (Ser, Val and Trp) are presented in Figures S1 (a, b, c). For quality control, one cloud sample
174 has been analyzed in MS^2 to check the presence of isobaric molecules. The peak with a retention time of 2.89 min and
175 m/z of 118,0867 $[\text{M}+\text{H}]^+$ has been found to correspond to the mixture of 2 isobaric molecules: valine and betaine (Figure
176 S2). Therefore, Val cannot be quantified. Leu and Ile could not be also distinguished as there are isobaric with the same
177 retention time (hereafter, Leu/I). Cys is not quantifiable as it forms S-S bonds. Consequently, 18 AAs can be quantified
178 in this study: Ala, Arg, Asn, Asp, Gln, Glu, Gly, His, Leu/I, Lys, Met, Phe, Pro, Ser, Thr, Trp, Tyr. The retention times
179 and exact masses measured by LC-MS of all the AAs are summarized in Table S2.

180 2.3.3 Standard addition

181 Cloud water is a complex mixture, conducive to disturbance in the LC-MS analytical signal. To restrain this matrix effect,
182 the AAs quantification is performed with the method of the standard addition, which consists of the addition of a series
183 of small volumes of concentrated standard to an existing unknown. For each AA, this method provides a calibration curve.
184 Figure S3 presents, as an example, how the concentration of Gly is measured for a particular cloud event (11-Mar-20
185 cloud) using the standard addition method. The magnitude of the intercept on the x-axis of the trendline is the original
186 concentration of Gly.

187 Table S3 displays calibration curve data measured for the 13 different cloud samples for each AA. The linearity of the
188 calibration curves is attested by the high R^2 values (> 0.95). The AA concentrations and their standard deviation (STD)
189 are calculated according to the equation from Bader (1980). More details can be found in SI (Figure S3 and attached
190 explanations of the calculations).

191 **2.4 Evaluation of air mass history**

192 The CAT model (Computing Atmospheric Trajectory Tool) is a three-dimensional (3D) forward/backward kinematic
193 trajectory code which has been recently developed and used to characterize the atmospheric transport of air masses
194 reaching PUY station (Renard et al., 2020). Backward trajectories clusters have been calculated for all clouds included in
195 the PUYCLOUD database. The temporal resolution of the backtrajectories is 15 min and the total duration is 72 h. The
196 model is initialized with wind fields from the ECMWF ERA-5 reanalyzed with a horizontal resolution of 0.5° and 23
197 vertical pressure levels between 200 and 1000 hPa. On the basis of the atmospheric boundary layer height (ABLH) and
198 altitude of topography interpolated for each trajectory point, this numerical tool allows to calculate the percentage of
199 points above the sea and the continental surfaces (sea surface *vs* continental surface), hereafter named “zone”. A “zone
200 matrix” is thus constructed from CAT model outputs and used for a statistical classification of each cloud event. All the
201 data relative to the 13 clouds of this study are reported in Table S1.

202 According to the classification proposed by Renard et al. (2020), cloud samples are classified into four categories,
203 “marine”, “highly marine”, “continental” and “polluted” by means of an agglomerative hierarchical clustering (AHC),
204 based solely on their chemical concentrations (Cl^- , Mg^{2+} , Na^+ , NH_4^+ , SO_4^{2-} and NO_3^-). This allows clustering clouds in
205 four categories: “highly marine”, “marine”, “continental” and “polluted”. The “marine” clouds have the lowest ion
206 concentrations and most of them come predominantly from western sectors. Marine category is predominant and the most
207 “homogeneous” in terms of concentrations. The “highly marine” category with a similar air mass history, gathers clouds
208 with the highest sea-salts concentrations (Cl^- , Mg^{2+} and Na^+). The continental category corresponds mainly to air masses
209 with high concentrations of potentially anthropogenic ions (NH_4^+ , NO_3^- and SO_4^{2-}) arriving predominantly from the
210 northeastern sector. Finally, the “polluted” category gathers cloud samples with the highest anthropogenic ion
211 concentrations. All the data relative to the clouds studied in the present work are reported in Table S1 and come from the
212 PUYCLOUD dataset.

213 **2.5 Statistical analysis**

214 With the objective to categorize cloud samples, we performed agglomerative hierarchical clustering (AHC), an iterative
215 classification, based on AA concentrations. The AHC dendrogram shows the progressive grouping of the data. The
216 dissimilarity between samples is calculated with the Ward’s agglomeration method using Euclidean distance. The number
217 of categories to retain is automatically defined on the base of the entropy (Addinsoft, 2020).

218 A large variability of the AA concentrations and relative proportions in the 13 cloud samples from PUY is observed. In
219 order to better understand this variability, a partial least square (PLS) regression is performed to analyze the correlations
220 between the explanatory (X) and dependent (Y) variables. The Xs variables gather the biological (ATP and bacteria
221 density), physical (temperature and pH) and chemical (TOC, Ca^{2+} , K^+ , Mg^{2+} , Na^+ , NH_4^+ , Cl^- , NO_3^- and SO_4^{2-} concentrations)
222 parameters, the “zone” matrix (sea/continental surface \neq ABLH), as well as the seasons. The Ys variables are the 18
223 AA concentrations (Ala, Arg, Asn, Asp, Gln, Glu, Gly, His, Leu/I, Lys, Met, Phe, Pro, Ser, Thr, Trp, Tyr). The Mann–

224 Whitney nonparametric tests are carried out to validate significant differences (p-value < 0.05) between two groups
225 (Renard et al., 2020).

226 **3. Results**

227 **3.1 Evaluation of LC-MS technique for a direct measurement of AAs in cloud**

228 The analytical method used in this study allows assaying AAs directly in cloud samples. MS coupled to LC allows the
229 analysis of the underivatized and non-concentrated analytes, avoiding potential biases and time-consuming processes.
230 The standard addition method also restrains matrix effects which are very commonly encountered with environmental
231 matrices (Hewavitharana et al., 2018). 18 AAs in cloud water sampled at PUY have been identified and their
232 concentrations quantified (Table S1). Concentrations values obtained for all AAs and cloud samples – as well as the
233 standard deviation of the measurements (STD_M) (*i.e.*, the precision of the measurements of AA concentrations) – are
234 reported in Table S3 and detailed in Figure S3. The median STD_M is 12 nM (ranging from 6 nM for Trp to 44 nM for
235 Ser). The relative standard deviation (RSD) ranges from 8 % for Ala to 119 % for Arg (median: 23 %).

236 The STD_M values, as calculated in this work in the context of a standard addition (equation detailed in Figure S3), could
237 be compared to the limit of quantification (LOQ) established in works using internal standard method (Bader, 1980). Both
238 equations are similar and provide comparable results. However, the precision depends on the number of standard points
239 added in the method, and not on the number of replicates. The values in this work are globally low and consistent with
240 those reported in previous works on cloud waters and aerosol particles (Table S4). A recent study performed by Triesch
241 et al. (2021) was able to quantify Val in cloud water samples, but they could not measure Arg, Asn, His, Lys, Cys and
242 Tyr concentrations. Triesch et al. (2021) is also based on LC-MS (Orbitrap™), but with samples concentrated (factor 44)
243 and derivatized with a pre-column. They reported LOQ values ranging from 0.2 to 1.0 µg L⁻¹, vs STD_M from 1.1 to 4.6
244 µg L⁻¹ in this study. STD_M values are also within the same range of magnitude than those reported on aerosol particles by
245 Helin et al. (2017) using direct injection of extracted AAs in LC-MS (triple-quadrupole technology), with values varying
246 from 4 to 160 nM, vs STD_M values from 8 to 44 nM in this work.

247 Looking more carefully at the median of the AA concentrations RSDs (calculated from data displayed in Table S3), it
248 appears that some AAs (Ala, Gly, Leu/I, Pro, Ser and Thr) have low RSDs (from 8 to 13 %) while others (Tyr, Lys, Trp,
249 Gln, Met and Arg) present higher RSDs (from 44 to 119 %). The RSD values obtained in this work are within the same
250 range of order than those reported by Helin et al. (2017). To conclude, these uncertainties do not change the final range
251 of magnitude of the AA concentrations.

252 **3.2 Cloud physico-chemical characteristics**

253 Table S1 presents data characterizing properties of cloud samples (chemical composition, microphysical properties, air
254 mass history). Among the 13 studied clouds, 12 clouds are classified as “marine” according to their ion concentrations
255 (Renard et al. 2020). 17-Jul-20 cloud from the North-East is classified as “continental” due to its NH₄⁺, NO₃⁻ and SO₄²⁻
256 concentrations that are significantly higher than the other studied clouds (Table S1).

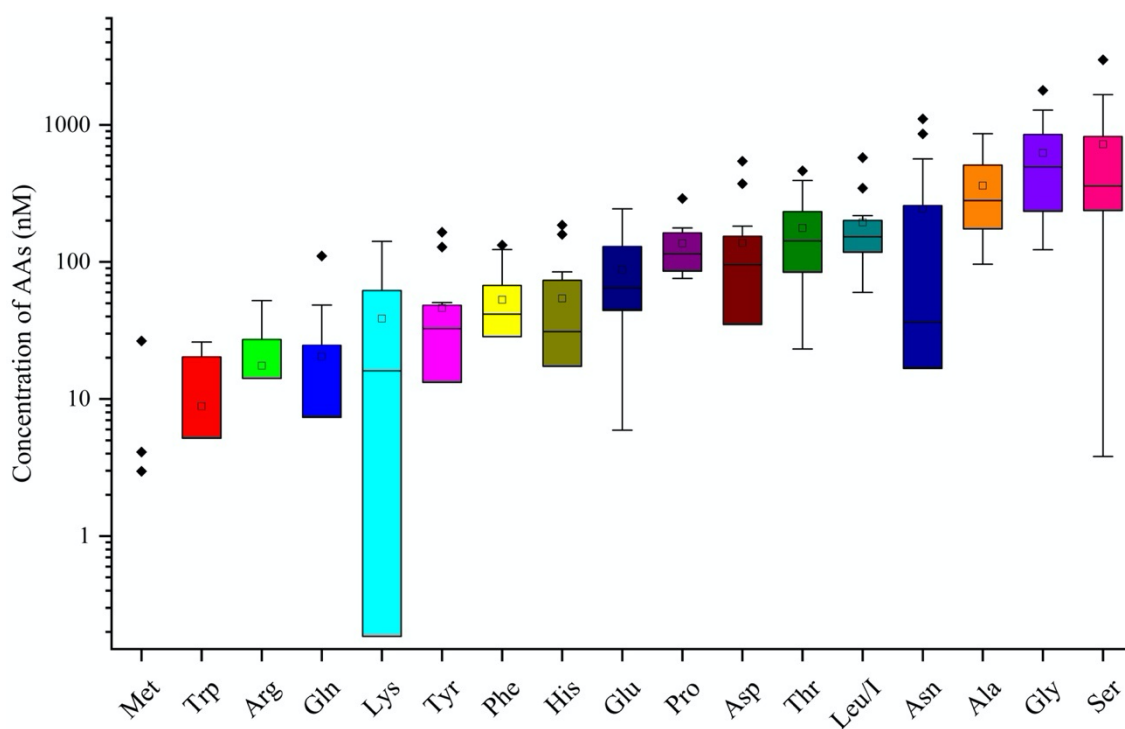
257 Similarly, to the work performed by Renard et al. (2020), the CAT model is used to characterize the air mass history of
258 the cloud samples. Figure 1 represents the mean backtrajectories calculated over the sampling period of the 13 cloud
259 samples; Figure S4 presents the backtrajectory calculations, every hour, for individual cloud event over the sampling
260 period. The CAT model provides a “zone matrix” (Table S1) gathering the percentage of time spent by the air masses

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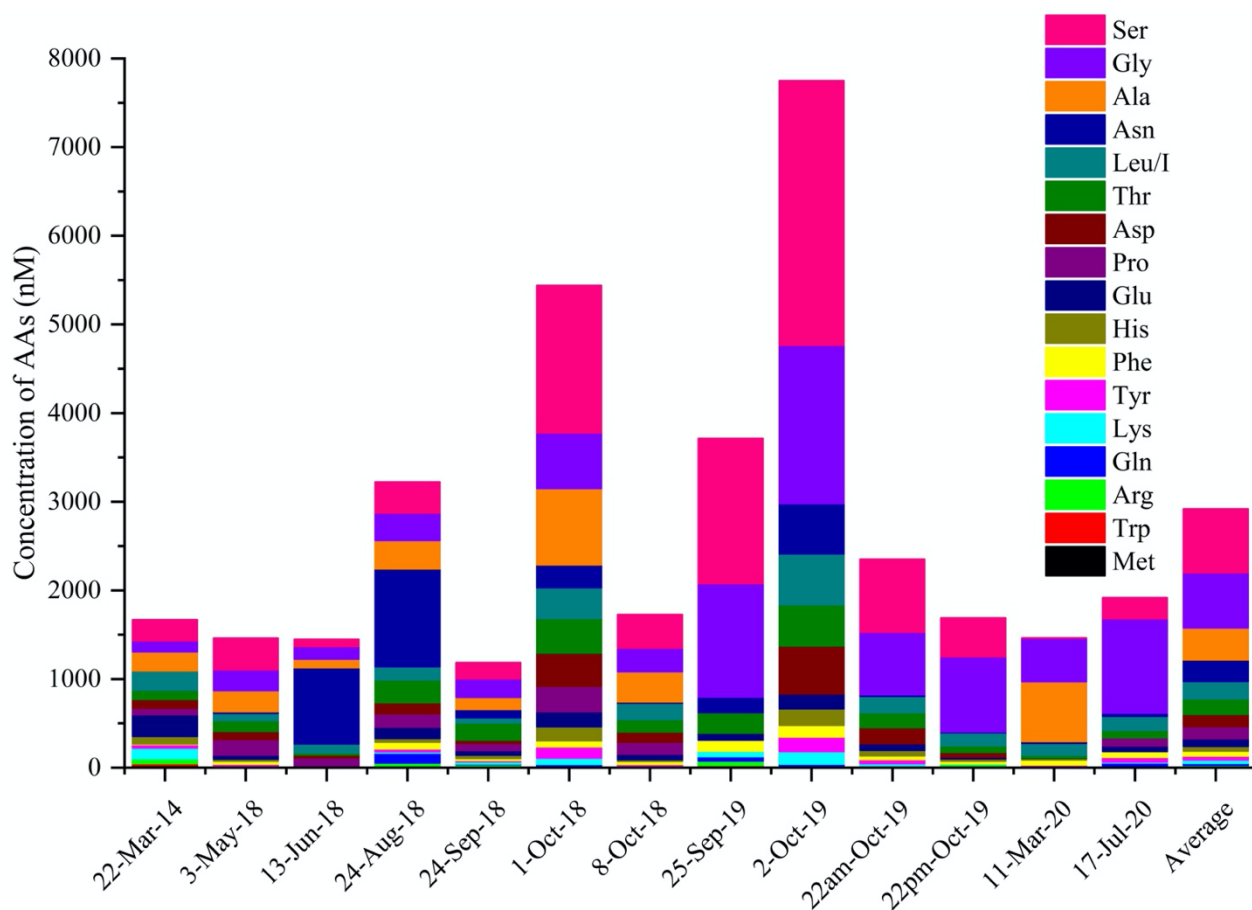
Table 1. Distribution of AA concentrations measured in the 13 clouds sampled at PUY: minimum, maximum, mean, standard deviation (STD), and relative standard deviation (RSD).

Label	Minimum (nM)	Maximum (nM)	Mean (nM)	σ (nM)	RSD
Ser	4	2983	721	866	120 %
Gly	123	1787	622	507	81 %
Ala	96	862	360	270	75 %
Asn	8	1105	264	375	142 %
Leu/I	60	577	194	141	72 %
Thr	23	462	176	133	75 %
Asp	33	543	165	166	100 %
Pro	76	290	137	72	53 %
Glu	6	244	87	70	81 %
His	16	185	65	61	93 %
Phe	6	133	57	39	68 %
Tyr	13	165	55	50	91 %
Lys	0	141	50	48	96 %
Gln	2	111	33	36	108 %
Arg	4	52	25	17	69 %
Trp	3	26	14	9	66 %
Met	3	27	11	13	119 %
TCAA	1187	7749	2696	1936	72 %

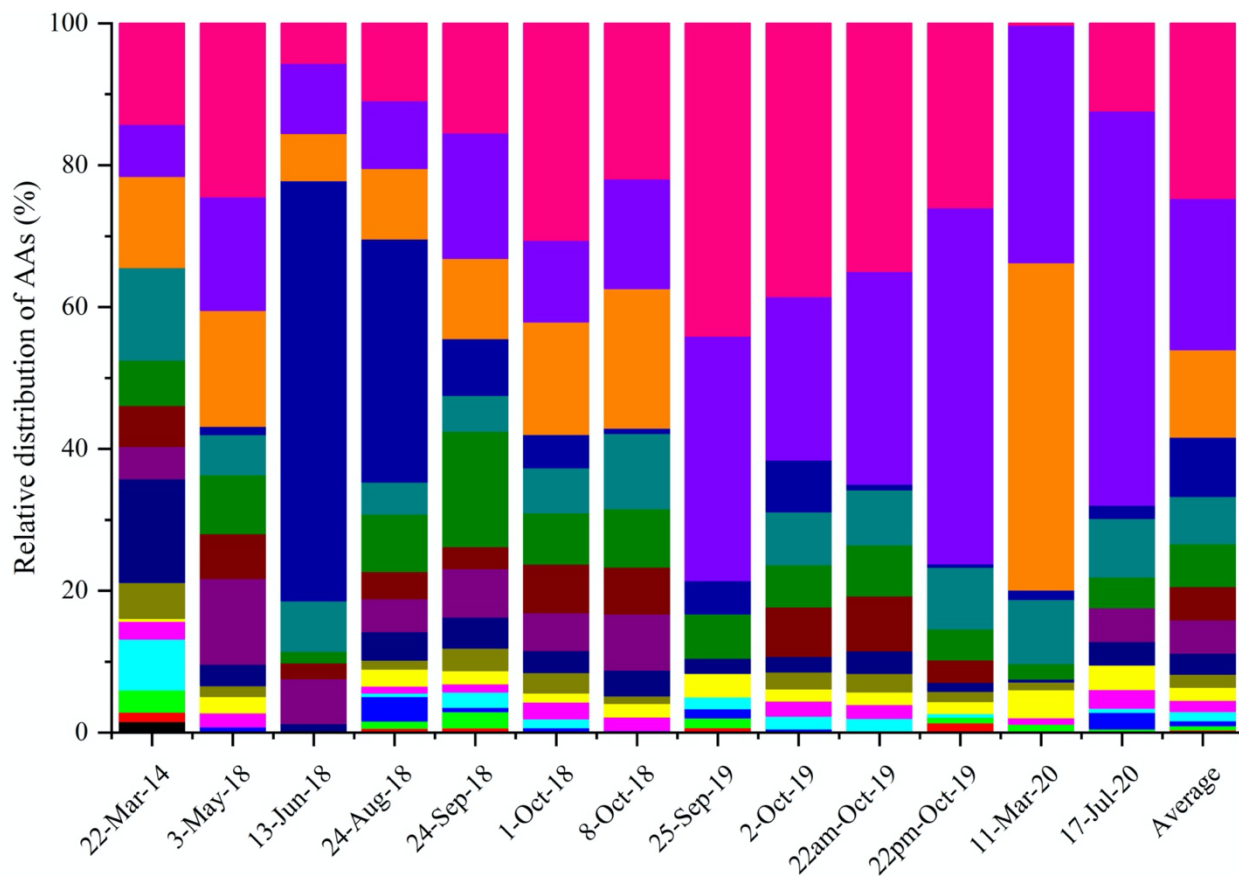
284 The total concentrations of free amino acids (TCAAs) vary significantly between cloud samples: the lowest concentration
 285 is 1.2 μM (24-Sep-18 cloud), and the highest one is 7.7 μM (2-Oct-19 cloud) while the mean value is equal to 2.7 μM
 286 (Table 1 and Figure 3a). In detail, Ser is the most abundant AA in the 13 cloud samples, with the highest STD (from 4 to
 287 2983 nM), followed by Gly (from 123 to 1787 nM), Ala (from 96 to 862 nM), and Asn (from 8 to 1105 nM) (Figure 2).
 288 This ranking seems common and ubiquitous, from polar to urban sites, in clouds as in rainwaters or aerosols (Table S4).
 289 Ser is also preponderant in marine clouds at Cape Verde Islands (Triesh et al., 2021) and rural fogs in Northern California
 290 (Zhang and Anastasio, 2003b).



291
 292 **Figure 2. Distribution of each AA for the 13 cloud samples. AA concentrations are in logarithmic scale. The bottom and top**
 293 **lines of the box correspond to the 25th and 75th percentiles, respectively. The middle line represents the median value and the**
 294 **square the mean value. The ends of the whiskers are the 10th and 90th percentiles, and the filled diamonds are outliers**
 295 **(concentrations above the 90th percentile).**



296



297

298

299

Figure 3. a. Distribution(nM) and b. relative contributions (% nM) of AAs molar concentrations in each cloud event sampled at PUY.

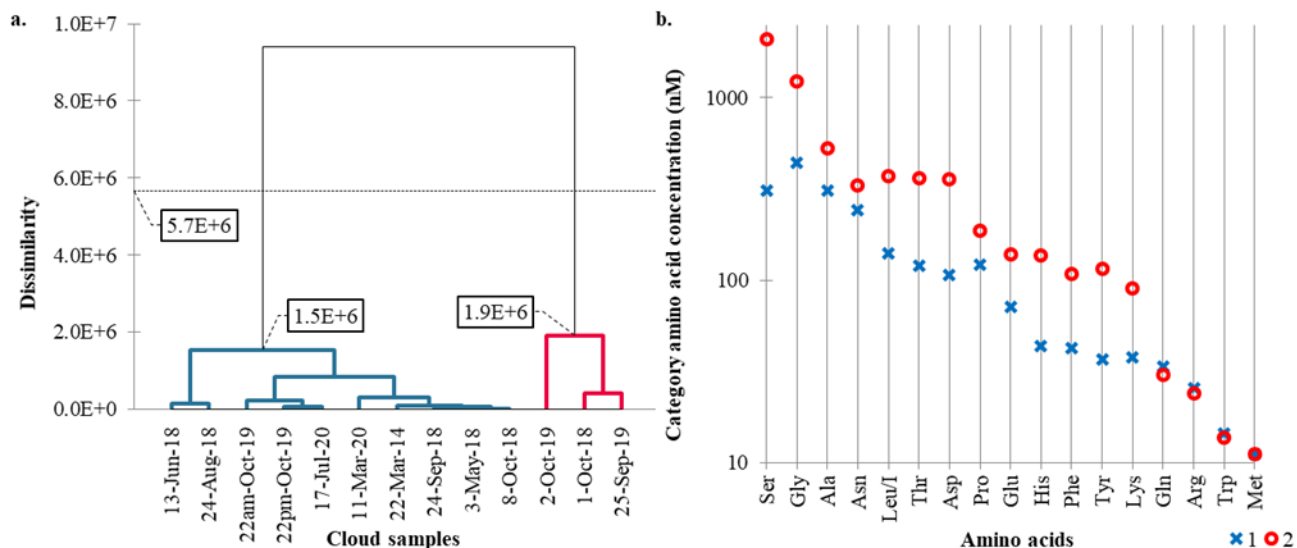
300 Figure 3 illustrates for each cloud event the relative and absolute molar concentrations of AAs. As discussed above, the
 301 TCAAs strongly vary between the different cloud events (Figure 3a). Their relative concentrations (Figure 3b) also vary
 302 among the cloud samples. For example, Ser contribution exceeds 50 % in 25-Sep-19 cloud, while Ser is almost absent in
 303 11-Mar-20 cloud sample, and vice versa for Ala. Asn prevails in 13-Jun-18 and 24-Aug-18 clouds. Nevertheless, the
 304 relative concentrations are quite similar, and the highest TCAAs do not seem to be explained by the mere presence, in
 305 excess, of a single AA.

306 Agglomerative hierarchical clustering (AHC), used to categorize cloud samples based on the AA concentration,
 307 successfully groups the 13 observations, with a satisfactory cophenetic correlation (correlation coefficient between the
 308 dissimilarity and the Euclidean distance matrices) of 0.79/1 (Figure 4a). The dotted line in Figure 4a represents the degree
 309 of truncation (dissimilarity = $5.7 \cdot 10^6$) of the dendrogram used for creating categories. This truncation is automatically
 310 chosen, based on the entropy level. The AHC profile plot (Figure 4b) details the average composition of these two
 311 categories determined from the 18 AAs.

312 AHC establish two different categories which reflect the variability of AAs in the 13 cloud samples. In detail, the blue
 313 category gathers 10 cloud samples with lower AA concentrations. This blue category is the most homogeneous (within-
 314 class variance = $3.7 \cdot 10^5$), compared to the red category (within-class variance = $1.2 \cdot 10^6$). Conversely, the red one, more
 315 heterogeneous, gathers 3 cloud samples with higher AA concentrations except for Met (absent in most of the 13 samples).

316 AHC reveals two categories significantly different, which are not explained by a punctual excess of certain AAs such as
 317 Ser or Gly. This cannot be concluded by only analyzing Figure 3 and confirms the gain of using AHC. AHC allows to
 318 perform a non-parametric test (Mann-Whitney test, not shown). Because the computed p-value is lower than the
 319 significance level $\alpha = 0.05$, the distribution of 9 AAs (Asp, Gly, His, Leu/I, Lys, Phe, Ser, Thr and Tyr) concentrations
 320 can be accepted as significantly different between both AHC categories.

321



322 **Figure 4. a. Dendrogram representing the agglomerative hierarchical clustering (AHC) based on dissimilarities using the**
 323 **Ward's method on concentrations of the 18 AAs. The 13 cloud samples are assigned to one of two established categories by**
 324 **entropy (i.e., dissimilarity < $5.7 \cdot 10^6$). b. Profile plot established by the AHC from the 18 main AAs. The Y axis, in logarithmic**
 325 **scale, displays the average AA concentrations of the category.**

326 Note that the 13-Jun-18 and 24-Aug-18 cloud samples are isolated in the AHC blue category due to their high Asn
 327 concentration.

328 3.4 Comparison with previous studies on clouds, fogs, and rain

329 To our knowledge, only two studies refer to the AA characterization in cloud water (Table 2). A first one has been
330 performed at PUY on 25 cloud samples; 16 AAs have been quantified by a different analytical procedure, using high-
331 performance liquid chromatography connected to a fluorescence detection after derivatization of the AAs (Bianco et al.,
332 2016b). They report a mean TCAA concentration of 2.67 μM with values ranging from 1.30 to 6.25 μM . These reported
333 concentrations are within the same range of magnitude those of the present study (from 1.2 to 7.7 μM). However, the
334 main difference between the present study and Bianco's study lies in the relative concentrations of the various AAs. Trp,
335 Leu/I, Phe and Ser were the four most concentrated AAs (mean concentrations of 563, 548, 337 and 281 nM, respectively)
336 while we found Ser, Gly Ala, and Asn as the most abundant AAs (mean concentrations of 721, 622, 360, and 264 nM,
337 respectively). This discrepancy could result from sampling characteristics, *i.e.*, cloud waters in the Bianco's study have
338 been sampled during two short periods (March/April and November 2014), whereas in the present work, cloud waters
339 have been collected over 6 years and covering different seasons.

340 The second one is a recent study reporting the characterization of AAs in cloud waters sampled at a marine site, the Cape
341 Verde Islands (Triesch et al., 2021). Results also indicate variability of AA concentrations in cloud samples with values
342 varying from 11.2 to 489.9 ng m^{-3} . These TCAAs are within the same range of magnitude as observed in this study (from
343 39 to 244 ng m^{-3}). In both studies (Cape Verde Islands and PUY), Ser, Ala and Gly are amongst the major AAs, but Asp
344 is found to be highly concentrated in the Cape Verde's study. They also find that the relative distributions of these four
345 AAs greatly change during the campaign period. Gly and Ser are found to be the dominant AAs in the first seven cloud
346 samples, while Ala and Asp are also highly present together with Gly and Ser during the last part of the campaign (3
347 samples). They conclude that these differences are due to the different types of clouds sampled during this campaign.
348 Triesch et al. (2021) show that some clouds present low TCAAs (less than 65 ng m^{-3}) with a dominance of Gly and Ser
349 and a second group with elevated TCAAs (more than 250 ng m^{-3}) and Ser as major AA, followed by Ala and Gly. This
350 enrichment of cloud waters in AAs could be due to oceanic sources or may be the result of *in situ* formation of AAs in
351 cloud water by for example enzymatic degradation of proteins, as reported by the authors. These hypotheses are also
352 supported by elevated concentrations of Asp at the end of the campaign that is a biologically produced AA. Globally, the
353 concentrations and major groups of AAs reported by Triesch et al. (2021) agree with the present work. This can be
354 explained by the remoteness of both locations and also the relevant marine influence encountered at PUY (Renard et al.,
355 2020).

356 In fog waters, at Davis in Northern California, Zhang et al. (2003b) measured elevated concentration of TCAAs with a
357 mean concentration of 20 μM . This is probably due to the proximity of the sampling site collection with local emission
358 of aerosol particles in this rural environment; however, dominant AAs are the same (Ser, Gly, Ala, Asn and Leu/I). Two
359 other studies in rainwater display similar AA concentrations and concentration ranking (Yan et al., 2015; Xu et al., 2019).
360 The study in Korea measured lower AA concentrations (Free and combined AAs) at Seoul (an inland urban area) than
361 those at Uljin (a coastal rural area) attributed to differences in contributing sources (Yan et al., 2015). A similar work has
362 been performed at a suburban site in Guiyang (China) over one year and have shown a seasonal effect with a maximum
363 level of AAs (Free and combined AAs) at spring and a minimal one at winter (Xue et al., 2019).

364 To conclude, few studies presented above report concentrations of AAs in cloud and fog waters. This is a challenging
365 issue to compare those 3 studies that have been performed for contrasted environmental conditions and for a limited
366 number of samples.

Table 2. FAA concentrations in atmospheric aqueous samples: cloud, fog, and rain (n is relative to the number of the samples).

Localization	Environment/ medium	Period / Samples	Separation/ Detection Method	Concentrations of FAAs (Range and mean values)	Distribution Major FAAs	Reference
Puy de Dôme Mountain, France (1465 m)	Rural + marine influence (Cloud)	03/2014 05-10/2018 09-10/2019 03-07/2020 13 samples	HPLC-MS/MS Standard addition	Range: 39 - 244 ng m ⁻³	Ser > Gly > Ala > Asn > Leu/I	(This work)
Puy de Dôme Mountain, France (1465 m)	Rural + marine influence (Cloud)	03-04/2014 (spring) 11/2014 (winter) 25 samples	HPLC-Fluorescence OPA-Derivatization	Mean: 118.6 ± 97.6 ng m ⁻³	Trp > Leu/I > Phe > Ser	(Bianco et al., 2016b)
Cape Verde Islands (744 m)	Marine (Cloud)	09-10/2017 (winter) 10 samples	HPLC-MS Waters AccQ-Tag-Derivatization	Range: 11.2 - 489.9 ng m ⁻³	Ser > Asp > Ala > Gly > Thr	(Triesch et al., 2021)
Northern California Davis, US	Rural (Fog)	1997 - 1999 (winter) 11 samples	HPLC-Fluorescence OPA-Derivatization	Mean: 40.8 ± 38.0 ng m ⁻³ (FAAs, protein type)	Ser > Gly > Leu > Ala > Val	(Zhang and Anastasio, 2003b)
Atlantic Ocean, Gulf Mexico (Cruise)	Marine (Rain)	09-10/1985 (n = 3) 02, 06, 09/1986 (n = 4) 7 samples	HPLC OPA/NAC-Derivatization	Mean: 604 ± 585 µg L ⁻¹	Gly > Ser > Ala > acidic AAs	(Mopper and Zika, 1987)
Seoul, South Korea (17 m)	Urban (Rain)	03/2012 - 04/2014 36 samples	HPLC	Mean: 21.0 ± 17.9 µg L ⁻¹	THAA:	(Yan et al., 2015)
Ulsan, South Korea (30 m)	Marine (Rain)	02/2011 - 01/2012 31 samples	OPA-Derivatization	Mean: 100.9 ± 110.2 µg L ⁻¹	Gly > Glu, Ala, Asp, Ser	
Guiyang, China (1300 m)	Suburban (Rain)	05/2017 – 04/2018 Summer (n = 29) Autumn (n = 9) Winter (n = 14) Spring (n = 13) 65 samples	HPLC OPA-Derivatization	Total range: 1.1 - 10.1 µM Mean: 3.7 µM Summer range: 1.3 - 6.6 µM Mean Summer: 2.9 µM Autumn range: 1.1 - 8.8 µM Mean Autumn: 4.4 µM Winter range: 1.5 - 9.9 µM Mean Winter: 3.4 µM Spring range: 2.6 - 10.1 µM Mean Spring: 5.2 µM	THAA: Glu + Gln, Gly, Pro > Asp, Ala	(Xu et al., 2019)

- 368 • HPLC: High performance liquid chromatography
369 • OPA: Ortho-phthalaldehyde
370 • THAA: Total hydrolyzable amino acids
371

372 3.5 Analysis of the environmental variability

373 A large variability of the AA concentrations and relative proportions in the 13 cloud samples from PUY is observed
374 (Table S1). To better understand this variability, data are analyzed in parallel with various environmental factors such as
375 the air masses history and quantitative physical, chemical, and biological measurements. During their atmospheric

376 transports, the air masses received chemical species under various forms and from various sources, and could also undergo
 377 multiphase chemical transformations, as well as deposition. This section is devoted to the correlation between the AA
 378 concentrations and the air mass history. To this end, PLS regressions are performed, and the results are validated with
 379 nonparametric tests (Mann–Whitney tests).

380 The PLS matrix of the explanatory variables (the “Xs”) is composed of the “zone matrix” (sea/continental surface
 381 </> ABLH) from the CAT model, to which are added the temperature, the pH, the inorganic ion concentrations, the
 382 bacteria density, the ATP concentration, and the seasons (Table S1). The matrix of the dependent variables (the “Ys”) is
 383 composed of the AA concentrations.

384 **Table 3. PLS correlation matrix between AA concentrations and “zone matrix” (sea/continental surface </> ABLH) from the**
 385 **CAT model, and temperature, pH, cation and anion concentrations, TOC and H₂O₂ concentrations, bacteria density (CFU mL⁻¹)**
 386 **and ATP concentration, and the seasons (Fall/Winter and Spring/Summer) determined from 13 cloud sampled at PUY. Highest**
 387 **correlations are displayed in dark red and highest anti-correlations in dark blue. R > 0.5 (or R < -0.5) with p-values < 0.1**
 388 **are underlined.**

Variables	Ser	Gly	Ala	Asn	Leu/I	Thr	Asp	Pro	Glu	His	Phe	Tyr	Lys	Gln	Arg	Trp	Met
Sea surface (< ABLH)	0.88	0.68	0.18	0.38	0.76	0.82	0.74	0.31	0.53	0.70	0.84	0.71	0.58	0.21	0.20	0.08	0.08
Sea surface (> ABLH)	-0.57	-0.45	0.03	-0.14	-0.50	-0.52	-0.42	-0.28	-0.04	-0.35	-0.37	-0.58	-0.11	0.16	0.31	0.38	0.25
Continental surface (< ABLH)	0.54	0.33	0.45	0.26	0.57	0.63	0.55	0.63	0.33	0.56	0.22	0.68	0.18	0.17	-0.48	-0.06	0.00
Continental surface (> ABLH)	-0.31	-0.21	-0.26	-0.23	-0.27	-0.32	-0.33	-0.11	-0.48	-0.36	-0.40	-0.18	-0.43	-0.36	-0.38	-0.41	-0.31
Temperature (°C)	-0.08	0.29	-0.14	0.29	-0.10	-0.15	-0.07	-0.25	-0.41	-0.28	0.40	-0.13	-0.36	0.51	-0.13	0.09	-0.20
pH	-0.10	0.00	-0.13	-0.19	0.02	-0.25	-0.11	-0.23	0.40	0.06	-0.32	-0.03	0.22	0.09	0.34	0.61	0.67
Na ⁺ (μM)	0.77	0.93	0.17	-0.07	0.62	0.44	0.65	0.03	0.07	0.46	0.80	0.54	0.43	0.02	0.03	0.12	-0.01
NH ₄ ⁺ (μM)	-0.38	-0.03	-0.13	-0.10	-0.21	-0.43	-0.17	-0.36	-0.13	-0.10	-0.11	-0.16	-0.21	0.03	-0.22	-0.05	0.13
Mg ₂ ⁺ (μM)	0.06	-0.11	0.09	0.10	0.17	0.11	0.15	0.23	0.46	0.41	-0.02	0.31	0.18	-0.11	-0.08	-0.08	0.30
K ⁺ (μM)	0.75	0.85	0.18	0.08	0.72	0.55	0.75	0.08	0.08	0.48	0.73	0.61	0.40	0.00	-0.24	-0.16	-0.16
Ca ²⁺ (μM)	-0.08	-0.34	0.04	0.30	-0.08	0.20	0.00	0.38	-0.05	0.09	-0.04	0.11	-0.28	-0.08	-0.39	-0.53	-0.41
SO ₄ ²⁻ (μM)	-0.11	0.12	0.37	-0.29	0.13	-0.11	0.12	-0.01	0.41	0.18	-0.15	0.08	0.13	0.10	-0.13	0.21	0.58
NO ₃ ⁻ (μM)	-0.17	0.00	-0.05	0.02	-0.01	-0.03	0.02	-0.12	0.02	-0.07	0.01	-0.05	-0.21	0.07	-0.26	-0.39	-0.05
Cl ⁻ (μM)	0.67	0.76	0.01	-0.05	0.60	0.34	0.50	-0.18	0.34	0.44	0.63	0.44	0.65	0.03	0.40	0.28	0.38
TOC (mgC L ⁻¹)	0.03	0.09	-0.09	0.25	0.00	-0.10	-0.10	0.15	-0.35	-0.09	0.16	0.19	-0.26	0.04	-0.27	0.06	-0.16
H ₂ O ₂ (μM)	-0.19	-0.01	-0.44	-0.10	-0.26	-0.46	-0.40	-0.32	-0.26	-0.25	-0.24	-0.15	-0.16	0.00	0.21	0.43	0.20
Bacteria (17°C; CFU/mL)	0.28	0.07	0.50	0.20	0.41	0.51	0.41	0.33	0.43	0.44	0.21	0.36	0.36	0.11	-0.19	-0.39	0.19
ATP (nM)	0.13	0.58	0.00	-0.14	0.19	-0.04	0.22	-0.32	0.05	0.19	0.33	0.17	0.01	0.19	-0.16	0.14	0.16
Fall / Winter	0.44	0.40	0.46	-0.44	0.32	0.36	0.36	0.29	-0.15	0.15	0.21	0.25	0.21	-0.39	-0.23	-0.31	-0.25
Spring / Summer	-0.44	-0.40	-0.46	0.44	-0.32	-0.36	-0.36	-0.29	0.15	-0.15	-0.21	-0.25	-0.21	0.39	0.23	0.31	0.25

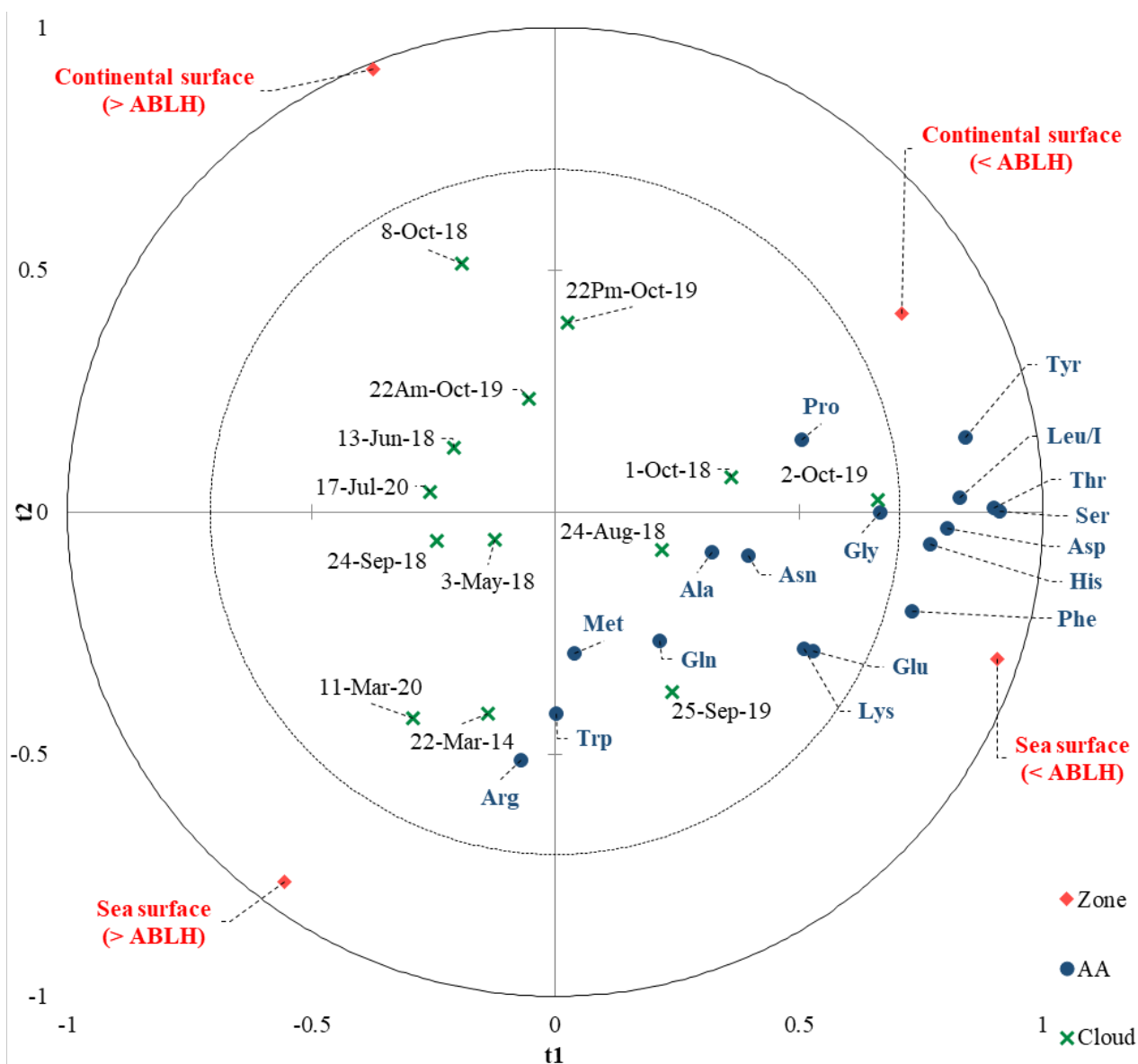
390 The correlation matrix of this PLS (Table 3) displays significant (anti-) correlations. First, 9 of the 18 AAs (Gly, His, Tyr,
 391 Asp, Leu/I, Thr, Phe and Ser) are robustly correlated with sea surface below the atmospheric boundary layer height
 392 (< ABLH), with correlation coefficients (r) ranging from 0.68 to 0.88. These 9 AAs are also significantly anticorrelated
 393 with sea surface in free atmosphere (> ABLH) (r ranging from -0.35 to -0.58), confirming direct influences from the
 394 boundary layer. These 9 AAs coherently correlate with Na⁺, Cl⁻ and K⁺ concentrations, confirming a marine influence for
 395 those AAs, similar to the observations of Triesch et al. (2021).

396 To a lesser extent, the same tendency (correlation / anticorrelation) is observed with continental surface
 397 (< ABLH / > ABLH). PUY is a remoteness site and the presence of anthropic ions, such as NO₃⁻ and NH₄⁺, are correlated
 398 with continental surface (> ABLH) (Renard et al., 2020). Thereby, the AAs and, in particular, the 9 aforementioned AAs
 399 are slightly anticorrelated with these anthropic ions.

400 No correlation appears between TOC concentration and the most abundant AAs, confirming the boundary layer influence,
 401 as well as the variability of AA proportion in organic carbon. These 9 AAs are also slightly anti-correlated with H₂O₂
 402 concentration, suggesting a potential influence of the photochemistry on AA concentrations (Lundeen et al., 2014). The
 403 biological parameters, in particular the bacteria density, are overall correlated with the AA concentrations. The 9 AAs
 404 most correlated with the sea surface (< ABLH) are, to a lesser extent, also correlated with Fall/Winter.

405 The PLS regression is a powerful statistical tool adapted for particular data conditions such as small sample sizes or data
 406 with non-normal distribution (Chin and Newsted, 1999). However, with only 13 samples, all the results in this work
 407 should be considered as “trends” that need to be investigated.

408 To go further in modelling the environmental variability of the AA concentrations in our cloud samples, we performed a
 409 simplified PLS restricting the Xs to the parameters of the CAT model (*i.e.*, the zone matrix). The predictive quality index
 410 of the models obtained with the PLS ($Q^2 = 0.19$ with one component) is satisfactory given the complexity of the cloud
 411 composition. In details, Figure 5 displays PLS correlation chart with t component on axes t1 and t2. The main axis (t1) is
 412 linked to the ABLH and most of the AAs are correlated with “< ABLH”. The t2 axis is linked to the zone (sea / continental
 413 surfaces) and it reveals a preponderance of marine influence, which is consistent with the dominant western oceanic air
 414 masses at PUY.



415

416

417 **Figure 5. Partial least squares (PLS) chart with t component on axes t1 and t2. The correlations map superimposes the**
 418 **dependent variables from the chemical matrix (blue circles), the explanatory variables (red diamonds) and the cloud**
 419 **events (green cross).**

420 Cloud water is a complex matrix resulting from the interaction of many factors; cloud samples more influenced by
421 continental zones (Northeast) could modify this model, and the predictive model provided by this PLS needs further
422 investigations to be validated. However, it appears that the air mass history remains the prevailing parameter, as observed
423 in Renard et al. (2020), after considering more cloud events. The CAT model could be used to estimate the AA
424 concentration, and this work helps to propose scientific plausible reasons explaining the environmental variability of AA
425 composition.

426 The following section is devoted to the analysis of the processes occurring in the atmosphere that could potentially explain
427 the AA levels and distributions in the clouds sampled at PUY. These processes are linked to their sources and to their
428 potential biotic and/or abiotic transformations in the atmosphere.

429 **4. Discussion**

430 Results reveal high variability in the relative concentrations of FAAs among cloud samples; however, some major FAAs
431 could be detected following this relative concentration ranking: Ser > Gly > Ala > Asn > Leu/I. On the contrary Trp and
432 Met present very low concentrations.

433 **4.1 Potential influence of the initial AAs distribution in biological matrices**

434 As free AAs are mostly from biological origin, we first compared the AA composition of various biological
435 macromolecules (proteins, peptidoglycans, *etc.*) that can be the source of AAs after hydrolysis with the relative
436 concentrations of AAs measured in the studied cloud samples.

437 Studies report the relative distributions of AAs in proteins extracted from different taxa (archaea, bacteria and eucaryotes)
438 (Bogatyreva et al., 2006; Gaur, 2014; Jordan et al., 2005). Although there are some differences between mammalian,
439 invertebrate, plant, protozoa, fungi and bacterial protein composition, some AAs (Ala, Gly, Leu/I and Val) are clearly
440 dominant while others are in low amounts (Cyst, Trp, His and Met). Globally, this relative abundance of AAs initially
441 constituting proteins presents similarities with the relative concentrations present in our samples. In particular, Gly, Ala,
442 and Leu/I are the most abundant in our samples as in the proteins, while Trp, and Met whose concentrations are the lowest
443 in our cloud samples are also minor components of proteins. Ser that is the major AA in our samples is present in proteins
444 in average and is not dominant.

445 We looked at the composition of peptidoglycan that form all the cell wall of Gram positive and Gram negative bacteria
446 that can be an atmospheric source of AAs (Vollmer et al., 2008). Peptidoglycans are complex structures formed by glycan
447 strands (composed of sugars) cross-linked by short penta-peptides. Although some slight variations can exist depending
448 on the bacterial strains, the standard sequence of this peptide is L-Ala-D-Glu-L-Lys-D-Ala-D-Ala. In a few cases Ser and
449 Gly have been also reported in the sequence. In addition, these penta-peptides are connected by inter-peptide bridges
450 varying from 1 to 7 AAs which contain in majority Gly and Ala but also Orn, Lys, Glu or Ser. Peptidoglycans can thus
451 represent a major source of Ala and Gly, that are the major AAs detected in our samples.

452 Finally, we specifically searched in aqueous media for the potential origin of Ser, which is dominant in our sample. Hecky
453 et al. report the AA composition of cell walls from six different diatom species, selected on the basis of taxonomy and
454 habitat diversity (Hecky et al., 1973). Three are from estuarine origin, the other from fresh waters. The protein template
455 of these cell walls is composed of the following AA sequences: Asp-Ser-Ser-Gly-Thr-Ser-Ser-Asp-Ser-Gly. Ser is thus

456 highly abundant in these aquatic organisms and plays an important role in the complexation of the silicon (Si^{4+}). This
457 result confirms previous reported data who showed the prevalence of serine in marine diatoms (Chuecas and Riley, 1969).
458 AAs in sea water during phytoplankton blooms were also investigated (Ittekkot, 1982): Glu concentration is maximum
459 in the early stages of the bloom, while Asp, Gly, Ala and Lys concentrations increase at the end of the bloom. In parallel,
460 Ser was one of the most abundant AA and its concentration remains high all along the bloom period. Ser could come from
461 the cell walls of some phytoplankton species which are diatoms. Hashioka et al. showed that diatoms could contribute up
462 to 80% of the total phytoplankton in the ocean to during bloom events (Hashioka et al., 2013). The high Ser concentration
463 measured in our cloud samples could thus originate from diatoms and could be a marker of their oceanic origin; this has
464 also been proposed by Triesch et al. (2021) who underline the marine origin of Ser present in their samples.

465 In conclusion, combining the composition of proteins, peptidoglycans and diatom cell-walls shows that Ala, Gly, Ser,
466 Leu/I are major AAs, while Trp and Met are minor ones; these ratios are fitting rather well with the concentrations found
467 in our cloud samples.

468 In the following, we aim at discussing more the variability of the AAs distributions and concentrations among the samples
469 looking at the air mass history (*i.e.*, sources) and their atmospheric transformations.

470 **4.2 Potential influence of the air mass origin on the AA concentrations and their relative distribution**

471 Table S4 summarizes the studies that analyze AAs quantity and distribution in various atmospheric media. Interestingly,
472 the systematic presence of Ser, Ala and Gly is observed in the various atmospheric waters including clouds (Triesch et
473 al., 2021), fogs (Zhang and Anastasio, 2003b) and rains (Mopper and Zika, 1987; Yan et al., 2015). These 3 AAs are also
474 significantly present in aerosols over contrasted regions over the world: rural sites (Zhang and Anastasio, 2003b) marine
475 sites (Matsumoto and Uematsu, 2005; Triesch et al., 2021; Violaki and Mihalopoulos, 2010; Wedyan and Preston, 2008),
476 urban or suburban sites (Barbaro et al., 2011; Samy et al., 2013) and polar sites (Scalabrin et al., 2012).

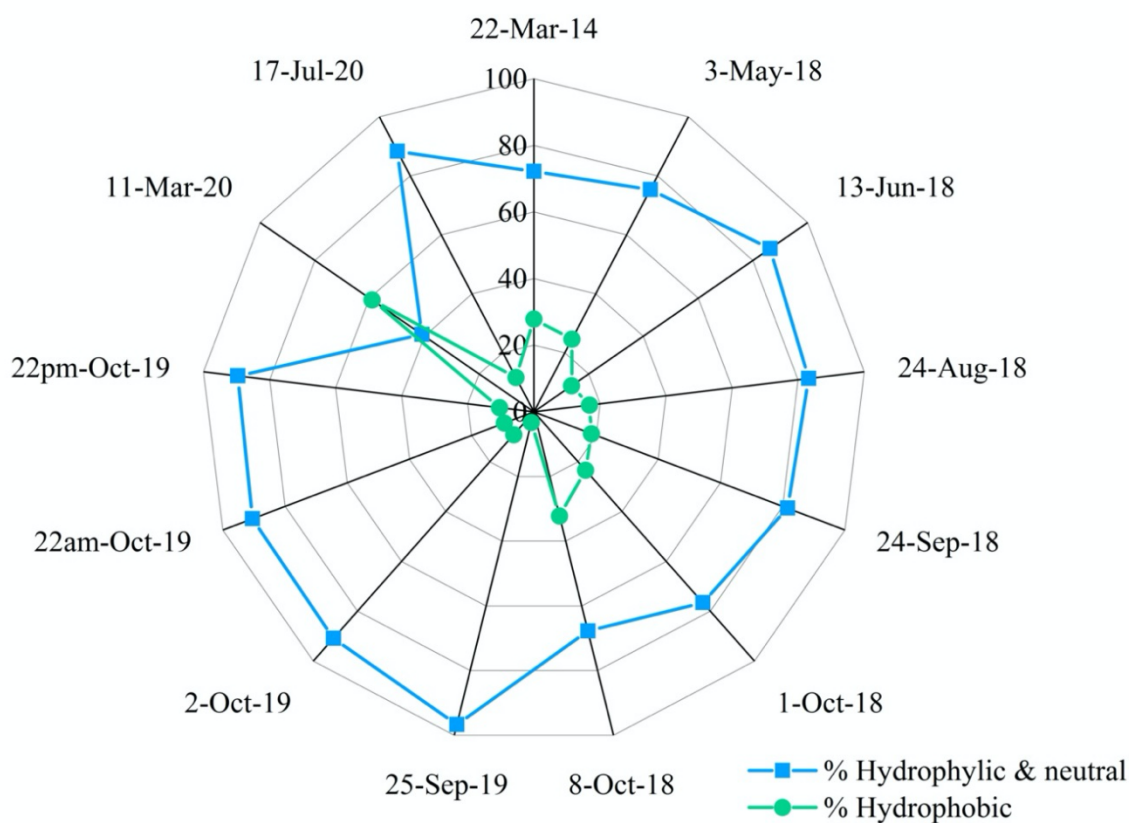
477 Looking more specifically at only two AAs (Gly and Ala), this list of studies can be extended to other works: in rain (Xu
478 et al., 2019), in marine aerosols (Mace et al., 2003b; Mandalakis et al., 2011), in rural aerosols (Ruiz-Jimenez et al., 2021;
479 Samy et al., 2011), in polar and remote sites (Barbaro et al., 2020; Barbaro et al., 2015; Feltracco et al., 2019). We can
480 notice that Gly is globally one of the major FAAs in all the reported studies (see Table S4 and joint explanations). In the
481 present study, we detect significant concentrations of Leu/I in agreement with only 3 other studies (Bianco et al., 2016b;
482 Mashayekhy Rad et al., 2019; Wedyan and Preston, 2008).

483 We overall found the same major groups of AAs that are commonly detected in marine clouds and aerosols. However,
484 one of the main differences is the high concentration of Asn in two of our samples, instead of the more common Asp,
485 suggesting potential conversion Asp/Asn (Jaber et al., 2021), and indicating that the origin of the clouds and aerosols is
486 not the only main driving factor explaining the final observed FAAs relative proportion in the clouds sampled at PUY.
487 Moreover, the presence of similar trends of AA composition in aerosols sampled under different sites (rural, marine,
488 urban, and polar) and in our cloud samples shows various influences from both continental and marine sources.

489 In agreement with the results of the PLS analysis (Figure 5), a significant correlation ($r = 0.78$) is observed between the
490 TCAA and the time spent by the air mass over the sea and below the boundary layer height (sea surface < ABLH). The
491 correlation between the TCAA and sea and the continental surfaces (< ABLH) is even higher ($r = 0.86$) (Figure S5),
492 indicating that the boundary layer influences the total amount of AAs rather than their relative concentration. When the

493 air mass is transported in the free troposphere, the TCAA is lower, possibly because of the remoteness of the direct sources
494 and because of chemical transformations that might be more intense in this upper part of the atmosphere.

495 To go further, Triesch et al. (2021) compared the AA composition of samples collected at Cape Verde Islands (marine
496 environment) in both aerosol and cloud phases. They show that FAAs are partitioned according to their hygroscopic
497 properties. They show that the hydrophobic AAs (Ala, Val, Phe, Leu/I) represent a much lower proportion (about 25 %)
498 of the total AAs present in cloud water, compared to the neutral (Ser, Gly, Thr, Pro, Tyr) plus the hydrophilic AAs (Glu,
499 Asp, Gaba). Figure 6 shows the distribution of the AAs in our samples collected at the PUY station according to their
500 hydrophobic *versus* hydrophobic + neutral properties. Clearly the concentrations of hydrophilic (Glu, Asp, Gln, Asn, His,
501 Lys, Arg) and neutral (Trp, Tyr, Gly, Thr, Ser, Pro) AAs are much higher (average value of 80 %) than that of hydrophobic
502 (Leu/I, Phe, Met, Ala) ones in all the samples, except in the 11Mar-20 sample where the hydrophilic + neutral AAs
503 represent only 40.8 % of the total FAAs. Our results are consistent with those measured in cloud samples at Cape Verde
504 Islands; this suggests that the hydrophobic nature of AAs is less favorable for their incorporation in cloud droplets due to
505 their low solubility.



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507 **Figure 6. Relative composition of AAs grouped by hygroscopicity (hydrophilic + neutral versus hydrophobic AAs) observed in**
508 **each cloud sample.**

509 Although the initial AA composition of the emitted aerosols can greatly impact the type of FAAs, the aging of the samples
510 due to biotic and abiotic processes must be considered to explain the presence of major or minor groups of AAs.

511 4.3 Potential influence of the atmospheric aging of AAs

512 Table 4 reports calculated, and experimental lifetimes of the different AAs targeted in this work considering different
513 biotic and abiotic processes.

514 First, AA theoretical lifetimes are calculated considering the reactivity constants of AAs with HO[•] radicals, O₃ and ¹O₂^{*}.
515 The values issued from the work of Jaber et al. (2021), Triesch et al. (2021) and Mc Gregor and Anastasio (2001) are
516 reported in columns A, B and C, respectively. At first glance it can be noticed that the lifetimes depend on the AAs and
517 can vary from a few hours or even minutes to a few days. Globally, reported values from the three studies are rather
518 consistent although they were calculated using a different set of reactivity constants and different oxidant concentrations
519 (see foot notes of Table 4).

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Table 4. Estimated atmospheric lifetimes of AAs degraded by atmospheric biological and chemical processes. AAs are classified following their mean concentrations measured in the present study. A brief description of the calculations is added below this table. More information can be found in SI for the calculations performed in this study based on the work from Jaber et al. (2021).

AAs	Theo. lifetimes by oxidation (days)	Theo. lifetimes by oxidation (days)	Theo. lifetimes by oxidation (days)	Exp. lifetimes by oxidation processes (days)	Exp. lifetimes by oxidation processes (days)	Exp. lifetimes by biological processes (days)
Reference	This study (adapted from Jaber et al. (2021))	Triesch et al. (2021)	Mc Gregor and Anastasio (2001)	This study (adapted from Jaber et al. (2021))	Mc Gregor and Anastasio (2001)	This study (adapted from Jaber et al. (2021))
Degradation processes	Oxidants: HO [•] , O ₃ and ¹ O ₂ [*]	Oxidant: HO [•]	Oxidants: HO [•] , O ₃ and ¹ O ₂ [*]	Irradiation experiments in artificial cloud medium	Irradiation experiments in fog waters	4 microbial strains in artificial cloud medium
Additional information	(A)	(B)	(C)	(D)	(E)	(F)
Ser	4.47	1.64	/	17.55	> 3.67	0.63 (15.1h)
Gly	41.26	3.09	> 170	\$	> 3.67	4.20
Ala	4.16	6.83	/	22.60	> 3.67	0.31 (7.6h)
Asn	22.05	/	/	\$	> 3.67	0.34 (8.1h)
Leu/I	0.64 (15.4h)	0.29	6.67	43.34	4.2	7.09
Thr	2.21	1.03	/	4.67	/	1.28
Asp	22.47	/	/	\$	2.42	1.55
Pro	1.72	1.70	/	\$	/	0.31 (7.4h)
Glu	5.49	3.29	37.5h	17.64	2.25	0.19 (4.5h)
His	0.10 (2.5h)	/	0.2 (5h)	22.60	1.00-1.83	1.79
Phe	0.17 (4.2h)	0.08	1.75	/	> 3.67	1.80
Tyr	0.08 (2.0h)	0.04	0.05 (1.2h)	3.56	1.25-2.33	0.86 (20.5h)
Lys	3.25	/	/	\$	> 3.67	0.59 (14.3h)
Gln	2.13	0.97	/	\$	/	0.20 (4.8h)
Arg	0.32 (7.7h)	/	/	\$	> 3.67	0.37 (9h)
Trp	0.06 (1.4h)	/	0.01 (0.15h)	10.51	0.11-0.38	6.97
Met	0.01 (0.13h)	0.06	0.01 (0.24h)	6.23	0.07-0.52	2.75

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(A): Theoretical calculations considering kinetic rate constants for the AAs oxidation by HO[•], O₃ and ¹O₂^{*} following Jaber et al. (2021). Aqueous concentrations of HO[•], O₃ and ¹O₂^{*} are respectively equal to 10⁻¹⁴, 5.0 10⁻¹⁰ and 1.0 10⁻¹² M.

(B) Theoretical calculations by Triesch et al. (2021). The mean lifetimes are estimated by considering pH-dependent rate constant of AAs with HO[•]. An HO[•] concentration of 2.2 10⁻¹⁴ M is considered in this study.

(C) Theoretical calculations by Mc Gregor and Anastasio (2001) were done under typical midday, wintertime conditions. Several oxidants were considered: the photoproduction of HO[•] and ¹O₂^{*} in the droplets, the source of HO[•] and O₃ in the aqueous phase by mass transfer.

(D) Experimental irradiation of 19 AAs at a concentration of 1 μM each in an artificial cloud medium were conducted. HO[•] production was performed using Fe-Ethylenediamine-N,N'-disuccinic acid (EDDS) complex solution. HO[•] concentration of 8.3 10⁻¹³ M is estimated. \$: Lifetimes cannot be calculated since a production is observed during the experiment.

(E) Irradiation experiments using simulated sunlight illumination were performed on real fog waters containing AAs.

(F) Biodegradation experiments of 19 AAs were performed by Jaber et al. (2021) using 4 microbial strains (*Rhodococcus enclensis* PDD-23b-28, *Pseudomonas graminis* PDD-13b-3, *Pseudomonas syringae* PDD-32b-74 and *Sphingomonas sp.* PDD-32b-11) in artificial cloud water.

537 These theoretical lifetimes of AAs could explain the very low Met and Trp concentrations measured in our cloud samples
538 which are very reactive, and the high concentrations measured for Gly and Asn, and in some extent for Ser and Ala, which

539 are very slowly transformed. However, they do not fit with the large amounts of Leu/I, except for the values given by Mc
540 Gregor and Anastasio (2001).

541 A second approach is to consider transformation rate measurements to further calculate experimental lifetimes.
542 Experimental investigations were designed to evaluate both abiotic and biotic processes. Photodegradation experiments
543 have been designed to assess oxidations processes, the first one was performed by Jaber et al (2021) in a microcosm
544 mimicking cloud environment using artificial cloud medium (Table 4, Column D), the second one (Mc Gregor and
545 Anastasio, 2001) consisted in irradiating real fog samples (Table 4, column E). In both cases the HO[•] concentration is
546 quantified. Interestingly the obtained experimental lifetimes are globally longer than the theoretical ones, many of them
547 exceeded 3 days and only Trp and Met lifetimes in the fog experiment are less than one hour. Moreover, certain AA
548 lifetimes could not be calculated from transformation rates measured by Jaber et al. (2021) as they observed production
549 and not degradation of some AAs (Gly, Asn, Asp, Pro, Phe, Lys, Arg, Gln). The experimental results displayed Table 4
550 (Columns D, E and F), which are different from theoretical ones (Columns A, B and C), reflect a much higher complexity
551 of the occurring transformations. On the one hand, irradiations are performed on complex media containing a mixture of
552 AAs, as well as other carbon and nitrogen sources. So, the measured transformation rates are net values reflecting both
553 synthesis and degradation processes, and even potential inter-conversion mechanisms. On the other hand, theoretical
554 lifetimes are calculated from reactivity constants measured in pure water containing a single AA without any other C or
555 N components and thus far from the chemical reactivity in real environmental samples. In addition, it is difficult to
556 interpret this data in more detail. Indeed, very few studies have studied the photo-produced compounds during these
557 oxidation processes. Some works report the formation of carboxylic acids, nitrate and ammonia from AAs photo-
558 transformations, or the conversion of AAs in another different AAs (His to Pro, Asp and Asn, Phe to Tyr, Pro to Glu) (see
559 Jaber et al. (2021) for review). More detailed pathways of abiotic transformations are only available for Trp, Tyr and Phe
560 (Bianco et al 2016a; Pattison et al., 2012). In spite of this complex situation, the long lifetimes or net production of Ser,
561 Leu/I, Gly and Asn (Table 4, columns D and E) could explain the relative high concentrations of these compounds in
562 cloud waters collected at PUY. On the contrary the short lifetimes (< 1 day) measured in fogs could explain the low
563 concentrations of Met and Trp. However, the lifetimes reported here cannot fully explain intermediate concentration
564 values measured for most of the other AAs; more work is needed to better understand oxidation pathways in complex
565 atmospheric media and measure additional transformation rates.

566 Potential biological transformation processes have been also evaluated in the lab. Recent works (Xu et al., 2020; Zhu et
567 al., 2021), based on Degradation Index (DI) calculations, suggest that biodegradation of AAs could occur in rain and in
568 aerosols. To calculate biotransformation lifetimes, transformation rates were measured in microcosms with 4 bacterial
569 strains isolated from clouds and representative of this medium and incubated in artificial cloud water (Jaber et al., 2021).
570 As in the previous case of irradiation in the same microcosm, it was shown that some AAs could be degraded but also
571 produced depending on the bacterial strain. The resulting biodegradation rates were thus calculated considering the
572 proportion of each type of cell in real cloud (see Table S4 for more details). From these global reaction rates, lifetimes
573 could be calculated for individual AA (Table 4, Column F). These biological lifetimes are very different from those
574 obtained considering oxidation processes, and globally much shorter. *Per se* they cannot explain the ranking of the larger
575 AAs (Ser, Ala, Leu/I, Asn) and lower AAs (Met, Trp) concentrations in our cloud samples, suggesting they might not be
576 the major contribution to the transformation of these AAs. However, when other compounds are considered with rather
577 low concentrations such as Gln, Arg, Lys, Phe, His, experimental oxidation lifetimes are long while biodegradation
578 lifetimes are much shorter, combination of these two processes could reflect a more realistic situation. Biosynthesis and

579 biodegradation pathways are very complex and interconnected and are fully described in databases (see:
580 <https://www.genome.jp/kegg/pathway.html>). The complexity comes from how different microorganisms use these
581 pathways. Up to now the only biodegradation rates related to atmospheric waters are from Jaber et al. (2021) and might
582 be incomplete; more experimental work should be conducted on real atmospheric samples.

583 **Conclusion**

584 This study reports the quantification of amino acids in cloud waters sampled at the puy de Dôme station using a new
585 approach based on a direct *in situ* analysis of the sample. Concentration of AAs represent in average nearly 2 % of the
586 TOC with a significant variability of TCAAs among the different samples. This heterogeneity is also observed in the AAs
587 distribution between the samples, but certain AAs are more dominant, especially Ser, Gly, Ala, Asp and Leu/I. These AA
588 relative proportions can be explained by the original biological matrices that emit AAs into the atmosphere, by the
589 hygroscopicity of AAs that favours their incorporation in the cloud water and finally by their transformations during their
590 transport into the atmosphere that modulate the total concentration of AAs. At PUY, the residence of the air masses within
591 the boundary layer, especially above the sea, seems also to surely increase the total amount of AAs in cloud water.
592 Conversely, the AA concentrations seem to decrease when the photolysis conditions are more favourable (free
593 troposphere or Spring / Summer period). In other words, the AAs concentration is modulated by the sources (mainly from
594 the boundary layer) and the sinks associated with the photodegradation.

595 However, it is still hard to validate all the formulated hypotheses that have been proposed to explain the differences in
596 the amounts and proportions of the various AAs found in our samples. This variability integrates many factors that are
597 interconnected or decorrelated and that should be investigated in the future. Some future targeted works could be
598 mentioned. First, this study is to our knowledge only the third one performed on cloud samples. More samples should be
599 collected at different seasons and at other sites presenting contrasted environmental conditions. This is crucial to robustly
600 evaluate atmospheric AA variability considering the effect of difference sources and atmospheric transport. Second, a
601 major limitation encountered to interpret the impact of transformation processes on the final distribution of AAs in
602 atmospheric samples lies on the lack of knowledge available in this field. Very few studies report the complex mechanisms
603 of biotic and abiotic transformations of AAs under realistic atmospheric conditions. Photochemists and biologists should
604 develop interdisciplinary work to describe these transformation pathways; this remains a challenging task.

605 **Author contributions.**

606 L. D. & A.-M. D. designed the project. A. B., M. B. & L. D. sampled the clouds at PUY. M. B., S. J. and M. L. conducted
607 the analysis. J.-L. B. used the CAT model to calculate backward trajectories and the “matrix zone”. P. R. and F. R.
608 performed the statistical analysis. P. R., M. B., F. R., A.-M. D. & L. D. wrote the paper.

609 **Competing interests.**

610 The authors declare that they have no conflict of interest.

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