

1 **Formation of condensable organic vapors from anthropogenic and** 2 **biogenic VOCs is strongly perturbed by NO_x in eastern China**

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24 25 **Abstract**

26
27 Oxygenated organic molecules (OOMs) are the crucial intermediates linking volatile
28 organic compounds (VOCs) to secondary organic aerosol (SOA) in the atmosphere, but
29 **comprehensive** understandings of the characteristics of OOMs and their formations
30 from VOCs are still **missing**. Ambient observations of OOMs using recently developed
31 mass spectrometry techniques are still limited, especially in polluted urban atmosphere
32 where VOCs and oxidants are extremely variable and complex. Here, we investigate
33 OOMs, measured by a nitrate-ion-based chemical ionization mass spectrometer at
34 Nanjing in eastern China, through performing positive matrix factorization on binned
35 mass spectra (binPMF). The binPMF analysis reveals three factors about anthropogenic
36 VOCs (AVOCs) daytime chemistry, three isoprene-related factors, three factors about
37 biogenic VOCs (BVOCs) nighttime chemistry, and three factors about nitrated phenols.
38 All factors are influenced by NO_x in different ways and to different extents. Over 1000
39 non-nitro molecules have been identified and then reconstructed from the selected

40 solution of binPMF, and about 72% of the total signals are contributed by nitrogen-
41 containing OOMs, mostly regarded as organic nitrates formed through peroxy radicals
42 terminated by nitric oxide or nitrate-radical-initiated oxidations. Moreover, multi-
43 nitrates account for about 24% of the total signals, indicating the significant presence
44 of multiple generations, especially for isoprene (e.g., $C_5H_{10}O_8N_2$ and $C_5H_9O_{10}N_3$).
45 Additionally, the distribution of OOMs concentration on carbon number confirm their
46 precursors driven by AVOCs mixed with enhanced BVOCs during summer. Our results
47 highlight the decisive role of NO_x on OOMs formation in densely populated areas, and
48 encourage more studies on the dramatic interactions between anthropogenic and
49 biogenic emissions.

51 **1 Introduction**

52
53 Secondary organic aerosol (SOA), as an important and complex component of
54 submicron particles (Zhang et al., 2007; Jimenez et al., 2009; Huang et al., 2014), is fully
55 involved in affecting climate (Intergovernmental Panel on Climate, 2014) and causing
56 health risks (Nel, 2005; Lim et al., 2012). Volatile organic compounds (VOCs) are
57 ubiquitous in the atmosphere and are recognized as main precursors of SOA (Hallquist
58 et al., 2009; Ziemann and Atkinson, 2012). However, the missing intermediate processes
59 from VOCs to SOA are yet to be elucidated (Hallquist et al., 2009; Ehn et al., 2014).

60
61 Benefitting from the state-of-the-art measurement techniques (Bertram et al.,
62 2011; Jokinen et al., 2012; Lee et al., 2014), many previously unreported oxygenated
63 organic molecules (OOMs), as intermediates from VOCs to SOA (Ziemann and
64 Atkinson, 2012), have been discovered. Among OOMs, highly oxygenated organic
65 molecules (HOMs), first observed in the gas phase at a boreal forest site (Ehn et al.,
66 2010; Ehn et al., 2012) and have been reviewed by Bianchi et al. (2019), are so
67 functionalized and low volatile that they can participate at the beginning of new particle
68 formation (NPF) by stabilizing sulfuric acid (Kulmala et al., 2013; Riccobono et al.,
69 2014) or through clustering alone (Kirkby et al., 2016; Bianchi et al., 2016), and
70 condense on existing particles to be responsible for a large fraction of SOA (Ehn et al.,
71 2014). In addition to conventional VOCs-to-OOMs mechanisms summarized in the
72 Master Chemical Mechanism (MCM) (<http://mcm.york.ac.uk/>, last access: 09 February
73 2021), recent studies have proposed new pathways, such as autoxidation (Crouse et
74 al., 2013; Jokinen et al., 2014) and multigenerational oxidation (Rollins et al.,
75 2012; Wang et al., 2020b), to form condensable vapors by adding oxygen atoms
76 efficiently. The productions of OOMs, especially HOMs, from precursors such as
77 monoterpenes (Ehn et al., 2014; Jokinen et al., 2015; Kirkby et al., 2016; Berndt et al.,
78 2016), sesquiterpenes (Richters et al., 2016), isoprene (Jokinen et al., 2015; Zhao et al.,
79 2020), aromatics (Wang et al., 2017; Molteni et al., 2018; Garmash et al., 2020), and
80 alkanes (Wang et al., 2021) have been investigated in laboratories by using the chemical
81 ionization atmospheric pressure interface time-of-flight mass spectrometer with nitrate
82 reagent ions (nitrate CI-API-TOF).

84 New insights and a general understanding about OOMs have been attained, yet many
85 critical details about OOMs formation and properties need to be addressed. First, the
86 current kinetic descriptions of OOMs obtained from experiments are still limited, such
87 as the lack of individual H-shift rates for autoxidation and of reaction rates of multi-
88 generational products with oxidants. Furthermore, the complexity of the real
89 atmosphere makes it more difficult to apply experimental results to ambient
90 environments. The precursors compete for oxidants and vice versa, and their products
91 will interact mechanistically in mixtures of atmospheric vapors (McFiggans et al.,
92 2019;Heinritzi et al., 2020). However, in the laboratory we usually study simple
93 systems with a single precursor and a single oxidant. Moreover, most experiments are
94 carried out for environments dominated by biogenic VOCs (BVOCs), while
95 anthropogenic emissions receive less attention. In addition to classic anthropogenic
96 VOCs (AVOCs), large amounts of primary emissions of oxygenated VOCs are also
97 present in urban areas (Karl et al., 2018). The effect of NO_x on OOMs is another key
98 issue. NO_x can terminate peroxy radicals (RO₂), outcompeting autoxidation
99 propagation reactions and other bimolecular reactions (RO₂ + RO₂, RO₂ + HO₂), and
100 change the products distribution, and consequently, size-dependently modulate the
101 growth rates of organic aerosol particles (Yan et al., 2020). Additionally, NO_x
102 contributes non-linearly to atmospheric oxidants, which also influence the productions
103 of OOMs (Pye et al., 2019). It is anticipated that NO_x plays a varied role in the
104 formations of OOMs as well as SOA in different environments.

105
106 Therefore, more extensive OOMs observations are needed to validate the atmospheric
107 implications of experiments, to couple with the global or regional model, and finally to
108 comprehensively understand the fate of OOMs in the atmosphere. Until now, only a
109 few ambient observations of OOMs using nitrate CI-API-TOF have been reported
110 (Bianchi et al., 2019), and almost all of them focus on rural or forested or remote
111 atmospheres (Yan et al., 2016;Massoli et al., 2018;Zhang et al., 2020;Beck et al., 2021).

112
113 The Yangtze River delta (YRD) is one of the most developed regions in eastern China.
114 Fine particulate matter, with an aerodynamic diameter smaller than 2.5 μm (PM_{2.5}), has
115 been significantly reduced in eastern China after the implementation of “Action Plan
116 for Air Pollution Prevention and Control” since 2013 (Ding et al., 2019), while
117 (secondary) organic aerosol are still much more abundant than in clean areas (Zhang et
118 al., 2017;Sun et al., 2020). Here we investigated condensable oxygenated organic
119 vapors observed by nitrate CI-API-TOF in August-September 2019 at the Station for
120 Observing Regional Processes and the Earth System (SORPES) in the western part of
121 the YRD, an anthropogenic-emissions-dominated environment (Fu et al., 2013;Xu et
122 al., 2017) mixed with enhanced biogenic emissions during summer (Wang et al.,
123 2020a;Xu et al., 2021). A variety of oxidants (Liu et al., 2019;Li et al., 2020;Xia et al.,
124 2020) with numerous precursors (VOCs) suggest very complicated atmospheric
125 oxidation processes and thousands of products (OOMs). Thereby, positive matrix
126 factorization (PMF) (Paatero and Tapper, 1994) was applied to time-resolved mass
127 spectra which had been pre-divided into small bins (binPMF, Zhang et al., 2019), to

128 separate various sources or processes of OOMs. Combined with summarizing the
129 ensemble chemical characteristics of OOMs, some interesting inspirations about the
130 conversion of VOCs to OOMs were obtained.

132 2 Methodology

134 2.1 Study site

136 The SORPES station (32°07'14" N, 118°57'10" E; 62 m a.s.l.) is located at Nanjing in
137 the western part of YRD, one of the most developed regions in eastern China. Due to
138 its unique location, this site can be influenced by air masses from different source
139 regions of anthropogenic emissions, biomass burning, dust and biogenic emissions
140 (Ding et al., 2013; Ding et al., 2016). Detailed descriptions for the station can be found
141 in previous studies (Nie et al., 2015; Xie et al., 2015; Xu et al., 2018; Wang et al.,
142 2018a; Sun et al., 2018; Shen et al., 2018).

144 2.2 Instrumentation

146 The nitrate CI-APi-TOF (Aerodyne Research Inc. and ToFwerk AG), combining a
147 chemical ionization source (CI) and an atmospheric pressure interface time-of-flight
148 mass spectrometer (APi-TOF) equipped with a long-TOF model (LTOF) with mass
149 resolution of 8000-12000 Th/Th, was deployed to detect the ambient sulfuric acid and
150 OOMs. The ambient air was pulled into a laminar flow reactor, where the sample flow
151 (10 L min⁻¹) is surrounded by a purified airflow serving as the sheath flow (25 L min⁻¹)
152 through a stainless-steel tube (a 100 cm long, 3/4 in. diameter). Nitrate reagent ions
153 were generated in the sheath flow by exposing air-containing nitric acid to a
154 PhotoIonizer X-Ray (Model L9491, Hamamatsu, Japan). Detailed description of the
155 instrument has been described elsewhere (Junninen et al., 2010; Jokinen et al., 2012).
156 The data were acquired at 1 Hz time resolution and analyzed with a tofTools package
157 (version 6.11) based on MATLAB (Math Works Inc.). **Due to the diversity and unknown
158 molecular structures of oxygenated organic compounds, standards for OOMs
159 measurable by the nitrate CI-APi-TOF are still lacking. Like other studies have done
160 (Kirkby et al., 2016; Trostl et al., 2016; Stolzenburg et al., 2018), an empirical method
161 was used to quantify the concentrations of OOMs based on the ionization kinetics
162 (pseudo first order reaction approximation) in the reaction tube of CI (Eq. 1) (Heinritzi
163 et al., 2016).**

$$165 \quad [\text{OOM}_i] = \ln \left(1 + \frac{\sum_{n=0}^1 [\text{OOM}_i \cdot (\text{HNO}_3)_n \cdot \text{NO}_3^- + (\text{OOM}_i - \text{H})^-]}{\sum_{n=0}^2 [(\text{HNO}_3)_n \cdot \text{NO}_3^-]} \right) \times C \times T_i \quad (1)$$

166 Here [OOM_i] is the concentration (molecules cm⁻³) of one OOM. On the right side of
167 the equation, the numerator in the parenthesis is the detected total signals (ions/s) of
168 one OOM **charged by nitrate ions in forming-adduct or deprotonated ways**, the
169 denominator is the sum of all reagent ion signals (ions/s). First, a H₂SO₄-based

171 calibration factor C, with a value of 4.2×10^9 molecules cm^{-3} , was obtained from a
172 calibration using H_2SO_4 (Kuerten et al., 2012) proceeding taking into account the
173 diffusion loss in the sampling line by assuming that all detected OOMs have the same
174 ionization efficiency as H_2SO_4 . The collision frequency of HOMs with nitrate clusters
175 is comparable to that of sulfuric acid with nitrate clusters (Ehn et al., 2014; Hyttinen et
176 al., 2015), yet the collision frequency of some moderately oxygenated molecules with
177 nitrate clusters is relatively slower. Therefore, calibration by this method leads to a
178 lower limit estimate of OOMs concentrations (Ehn et al., 2014; Trostl et al., 2016), but
179 the accurate quantification of OOMs is not the main concern of this study and the errors
180 in the quantification of OOMs do not change our conclusions. Second, a mass
181 dependent transmission efficiency T_i of APi-TOF was inferred in a separate experiment
182 by depleting the reagent ions with several perfluorinated acids (Heinritzi et al., 2016).

183

184 VOCs precursors were measured by a proton transfer reaction time-of-flight mass
185 spectrometer (PTR-ToF-MS, Ionicon Analytik, Innsbruck, Austria, TOF 1000 ultra).
186 $\text{PM}_{2.5}$ was measured with a combined technique of light scattering photometry and beta
187 radiation attenuation (Thermo Scientific SHARP Monitor Model 5030). The chemical
188 compositions of $\text{PM}_{2.5}$ was determined on-line using time-of-flight aerosol chemical
189 speciation monitor (TOF-ACSM, Aerodyne Research Inc.). PMF analysis was further
190 used to separate the organic aerosol (OA) to primary and secondary organic aerosols
191 (POA and SOA). The number concentrations of particles were measured by the
192 scanning mobility particle sizer (SMPS) with nano DMA (4.0 to 63.8 nm) and long
193 DMA (41.4 to 495.8 nm) and the aerodynamic particle sizer (APS) (0.5 to 18.0 μm).
194 NO and NO_2 were measured using a chemiluminescence analyzer equipped with a blue-
195 light converter (TEI, Model 42I-TL); O_3 , SO_2 , and CO were measured using the
196 ultraviolet photometry, pulsed-UV fluorescence, and IR (infrared) photometry
197 techniques (TEI, Model 49I, 43C, and 48C), respectively. Zero and span calibrations
198 for trace gases were performed weekly during the campaign. Meteorological
199 measurements including relative humidity (RH), wind speed, wind direction, and air
200 temperature were recorded by Automatic Weather Station (CAMPEEL co., AG1000).
201 $\text{J}(\text{O}^1\text{D})$ was measured by ultra-fast CCD-detector spectrometer, UVB enhanced
202 (Meteorologieconsult GmbH, Germany).

203

204 **2.3 Hydroxyl radical (OH) estimate**

205 The OH concentration was calculated by applying the Eq. (2), based on the assumption
206 that gaseous sulfuric acid is mostly produced from the oxidation of SO_2 by OH and
207 primarily loss by condensing onto particles:

208

$$209 \quad [\text{OH}] = \frac{[\text{H}_2\text{SO}_4] \cdot \text{CS}}{k_{\text{OH}+\text{SO}_2} \cdot [\text{SO}_2]} \quad (2)$$

210

211 Where $k_{\text{OH}+\text{SO}_2}$ is a termolecular reaction constant for the rate-limiting step of the
212 formation pathway of H_2SO_4 in the atmosphere (Finlayson-Pitts and Pitts, 2000), and

213 condensation sink (CS) is the loss rate of H₂SO₄ by condensation to aerosol surface.
214 The value of $k_{\text{OH}+\text{SO}_2}$ is inferred from the IUPAC Task Group on Atmospheric
215 Chemical Kinetic Data Evaluation (<https://iupac-aeris.ipsl.fr/>, last access: 09 August
216 2021). The value of CS was calculated following Eq. (3) (Kulmala et al., 2012):

217

$$218 \quad \text{CS} = 2\pi D \sum_i \beta_{m_i} d_{p_i} N_i \quad (3)$$

219

220 Where D is the diffusion coefficient of gaseous sulfuric acid, β_m is a transition-regime
221 correction factor dependent on the Knudsen number (Fuchs and Sutugin, 1971), and
222 d_{p_i} and N_i are the diameter and number concentration of particles in size bin i .

223

224 It has been proved that $\frac{k_{\text{OH}+\text{SO}_2} \cdot [\text{SO}_2] \cdot [\text{OH}]}{\text{CS}}$ is a very reliable proxy for H₂SO₄ during the
225 day (Lu et al., 2019). The ozonolysis of alkenes can form stabilized Criegee
226 intermediates (SCIs) in addition to OH, and SCIs can also oxidize SO₂ to form H₂SO₄
227 (Mauldin Iii et al., 2012; Guo et al., 2021). A previous study on H₂SO₄ proxy in this site
228 has revealed that the reactions of SO₂ with products from the ozonolysis of alkenes
229 generate a moderate amount of nighttime sulfuric acid, with little effect on daytime
230 sulfuric acid (Yang et al., 2021). Thus, OH may be overestimated during nighttime. In
231 this study, OH was used to calculate the production rates of RO₂ during daytime (Fig.
232 4), the error of OH do not change the relative distribution of RO₂ from different
233 precursors.

234

235 **2.4 binPMF**

236

237 binPMF has been used to analyze the measured HR mass spectrometry data. Briefly,
238 the raw spectra were divided into narrow bins with a width of 0.006 Th after mass
239 calibration. The data matrix and error matrix were prepared according to the methods
240 described by Zhang et al. (2019) for the PMF model inputs (Section S2 in the
241 supplement). Different from the traditional PMF such as using unit mass resolution
242 (UMR) or HR data as input, binPMF still retains HR information as much as possible,
243 avoids the uncertainty of HR peak fitting influencing the results of PMF, and separate
244 the complex overlapping peaks for fitting. The PMF analysis in this work uses the
245 IGOR based analyzing interface SoFi (solution finder, version 6.8) and ME-2 as
246 described in (Canonaco et al., 2013). After select the PMF solution, we fitted the HR
247 peaks in each factor through toftool.

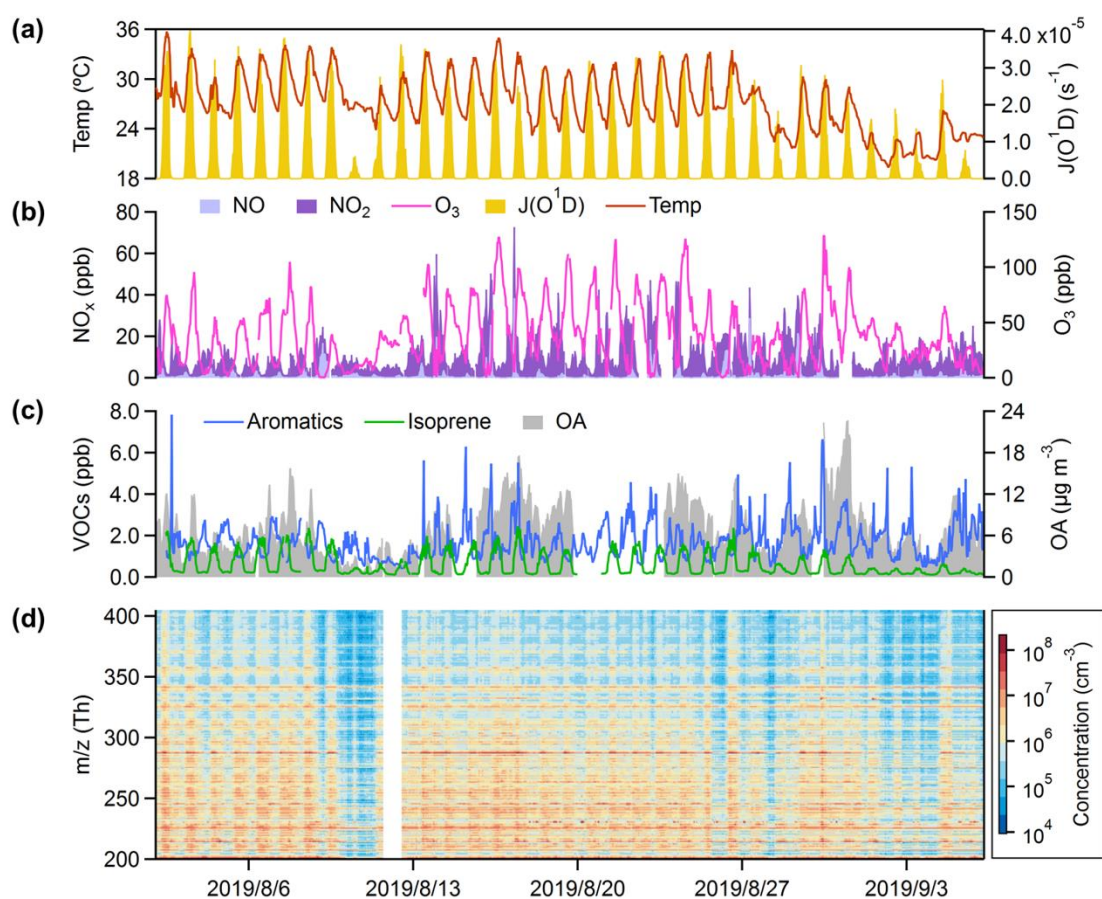
248

249 **3 Results and Discussions**

250

251 Figure 1 shows temporal variation of OOMs and related parameters at the SORPES

252 station in the northeastern suburb of Nanjing from August 02 to September 06, 2019.
253 During the observation period, 22 of 35 days had maximum hourly temperatures above
254 30 degrees Celsius, and 29 days had maximum hourly $J(O^1D)$ above $2 \times 10^{-5} s^{-1}$. High
255 temperature and solar radiation indicate strong photochemistry, producing a large
256 amount of ozone, with concentration often exceeding 80 ppb. Even at night, the
257 concentration of ozone is rarely lower than 10 ppb, resulting from the weak titration of
258 low NO. At the same time, the reaction between ozone and high concentration of NO₂
259 can provide sufficient NO₃ radicals, dominating nocturnal degradation of certain
260 volatile organic compounds (VOCs) (Wayne et al., 1991). The elevated mixing level of
261 total aromatic hydrocarbons is one of the main characteristics of the atmosphere in
262 densely populated areas, in addition to which there should be many alkanes and alkenes
263 which cannot be observed by PTR-ToF-MS (Fu et al., 2013; Xu et al., 2017). In the
264 daytime with strong photochemical reaction ($J(O^1D) > 1 \times 10^{-5} s^{-1}$), we instead observed
265 higher concentrations of isoprene than total aromatics (Fig.S1(a)). The complex
266 mixtures of anthropogenic and biogenic VOCs can be oxidized through a variety of
267 pathways to produce OOMs, of which some low volatile components will condense
268 into particles, forming organic aerosol. The concentrations of OOMs with mass-to-
269 charge ratio (m/z) below 360 Th are usually higher than 10^6 molecules cm^{-3} , and some
270 can even reach up to 10^7 - 10^8 molecules cm^{-3} . Clustered peaks on the spectra of OOMs
271 and their clear daily variations imply a lot of chemical and physical dynamics
272 information (Fig. 1(d) and Fig.S1(b)), which is the main aspect we want to explore in
273 this work.
274



275

276 Figure 1. Overview of the observation. Time series of (a) temperature (Temp) and the
 277 photolysis frequency of O₃ (JO¹D), (b) O₃ and NO_x (NO+NO₂), (c) total aromatics
 278 (benzene + toluene + C₈ aromatics + C₉ aromatics + C₁₀ aromatics + styrene), isoprene,
 279 and OA, and (d) mass spectra of nitrate CI-APi-TOF with m/z in the range of 202-404
 280 Th.

281

282 The binPMF analysis was performed to characterize the sources or processes of OOMs.
 283 A 14-factor solution was selected to interpret the data set, including 3 factors about
 284 AVOCs daytime chemistry, 3 isoprene-related factors, 3 factors about BVOCs
 285 nighttime chemistry, 3 factors about nitrated phenols (NP), and 2 factors excluded from
 286 the following discussion. One of these two disregarded factors is mainly composed of
 287 fluorinated contaminations (F-contaminations), and the other is mainly a mixture of
 288 nitrated phenols and fluorinated contaminations (Mixed contaminations). When
 289 naming these factors, we prioritize the description of dominated species or their
 290 precursors, but if the precursors are complex mixtures, our naming highlights the
 291 characteristics of the chemical processes that drive certain factors. Although this may
 292 not be the optimal PMF solution, it still separates a lot of useful information. We also
 293 stress that the urban OOMs mix is unlikely to be a perfect combination of independent,
 294 unchanging factors, which is an underlying assumption in the PMF algorithm. As such,
 295 there will be no solution which is complete and perfect, but we chose a solution from
 296 which we were able to provide us with interesting insights. Details of the PMF
 297 diagnostics is provided in section S2 in the supplement (Fig. S2-S6). For the

298 convenience of discussions, we have grouped these factors based on shared
 299 characteristics of certain factors.

300

301 Table 1. Summary of molecular characteristics of 9 discussed non-nitrated-phenols
 302 factors. The calculation of the relevant parameters is given in section S3 in
 303 the supplement. Major peaks of each factor are summarized in section S4 in
 304 the supplement.

Factor	Average concentration (cm ⁻³)	Effective formulae	MW (g mol ⁻¹)	O:Sc	O:C	N:C	DBE	log ₁₀ (C* (μg m ⁻³)) in 300K
Aro-OOMs	1.86E+07	C _{9.1} H _{14.3} O _{6.1} N _{0.6}	230.2	-0.52	0.73	0.08	2.6	-1.7
Temp-related	4.50E+07	C _{6.8} H _{10.2} O _{6.0} N _{0.5}	195.8	-0.02	0.95	0.08	2.5	-1.4
Aliph-OOMs	2.11E+07	C _{7.5} H _{12.2} O _{6.7} N _{1.2}	225.7	-0.55	0.96	0.17	1.9	0.0
Photo-related	4.77E+07	C _{6.9} H _{11.0} O _{7.4} N _{1.2}	228.3	-0.28	1.18	0.20	1.8	-1.1
O _x & SOA-related	2.59E+07	C _{6.6} H _{9.8} O _{6.8} N _{1.1}	214.2	-0.24	1.11	0.19	2.2	-0.3
Isop-OOMs	2.83E+07	C _{5.5} H _{9.6} O _{6.9} N _{1.4}	205.8	-0.51	1.34	0.28	0.9	1.2
BVOCs-OOMs I	1.68E+07	C _{7.2} H _{11.5} O _{7.0} N _{1.0}	224.1	-0.26	1.06	0.16	2.0	-1.4
BVOCs-OOMs II	9.05E+06	C _{9.2} H _{14.6} O _{7.1} N _{0.9}	251.3	-0.45	0.83	0.11	2.5	-2.8
BVOCs-OOMs III	1.57E+07	C _{8.6} H _{13.7} O _{6.9} N _{1.2}	243.3	-0.64	0.87	0.16	2.1	-0.7

305 Note: MW is the molecular weight, OSc is the carbon oxidation state, O:C is the oxygen
 306 to carbon ratio, N:C is the nitrogen to carbon ratio, DBE is the double bond equivalent,
 307 C* the saturation concentration and log₁₀(C*) is the volatility.

308

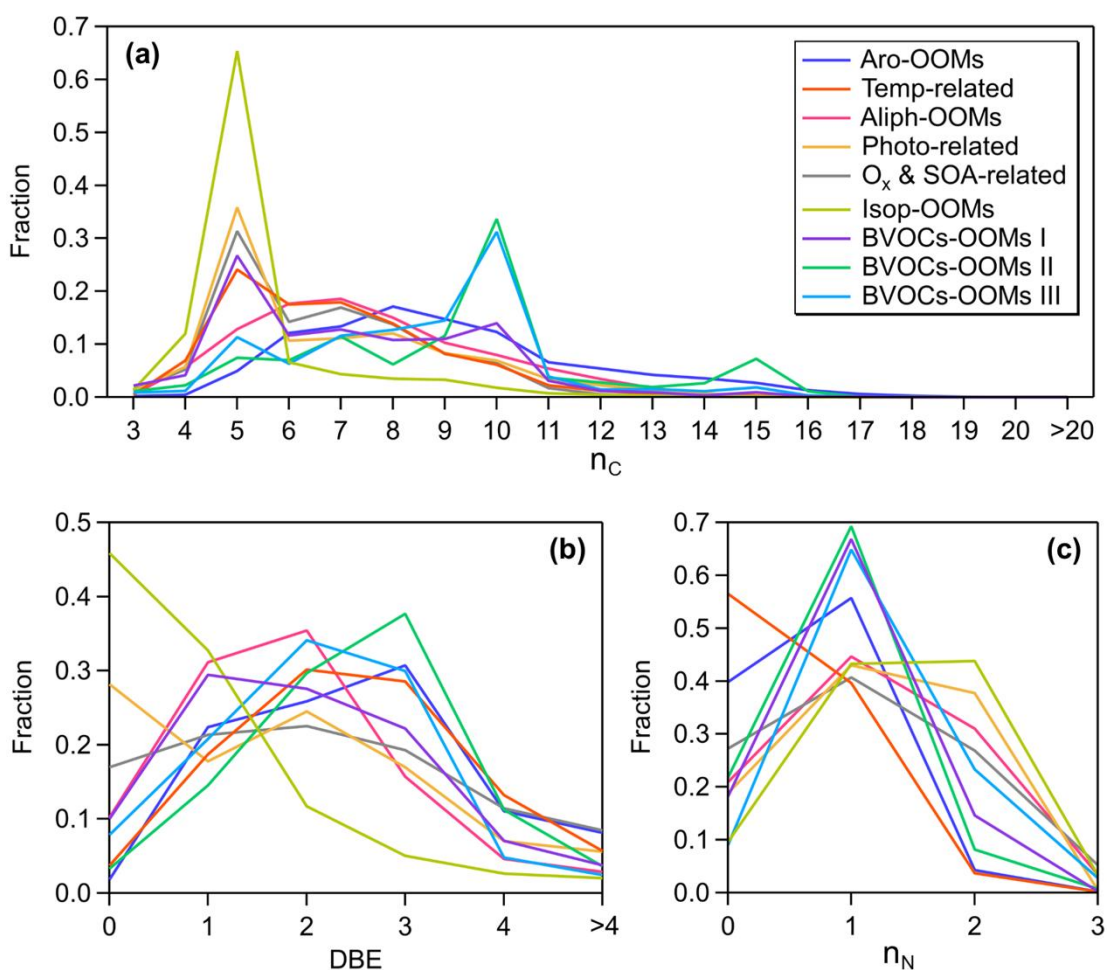
309

310 3.1 AVOCs daytime chemistry

311

312 The following daytime factors are characterized by C₆-C₉ OOMs (Fig. 2(a)), considered
 313 to be derived from the oxidation of anthropogenic VOCs in this urban atmosphere,
 314 while we cannot completely exclude the present of BVOCs-derived OOMs, such as C₅
 315 and C₁₀ OOMs.

316



317

318 Figure 2. The distributions of observed non-nitro OOMs grouped by (a) the number of
 319 carbon atoms (n_C), (b) DBE, and (c) the number of nitrogen atoms (n_N) in 9 factors.
 320 Since the signals of RO₂ are very weak, RO₂ from BVOCs OOMs I and BVOCs OOMs
 321 II are excluded in (b) to keep the integer value of DBE.

322

323

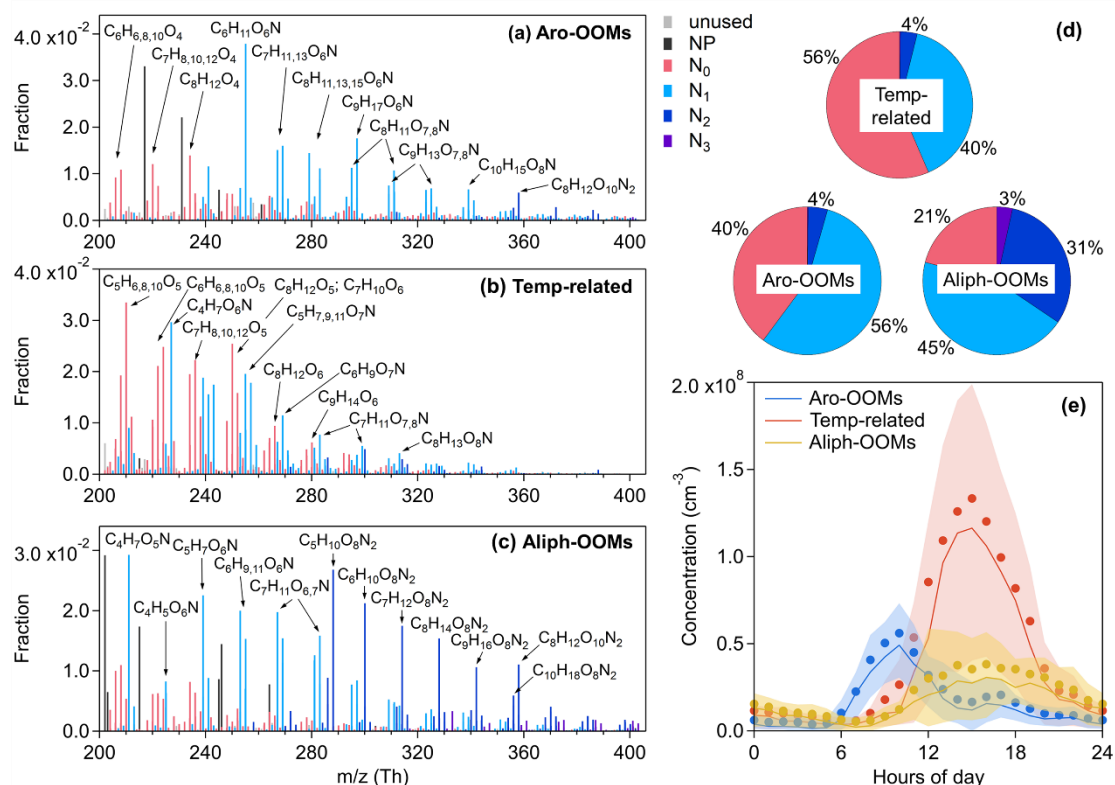
324 Aro-OOMs factor

325

326 The effective DBE of this factor is the largest among all factors (Table 1), with main
 327 signals come from compounds with DBE > 2 (Fig. 2(b)) and consistent with the nature
 328 of the oxidation products of aromatics (Fig. 3(a)). Combined with the correlation with
 329 the production rates of OH-initiated primary peroxy radicals (RO₂) from aromatics
 330 calculated by Eq. (4) (Fig. 4), this factor is supposedly dominated by aromatics-derived
 331 OOMs (Aro-OOMs). The Aro-OOMs factor increases from 5:00 LT with a maximum
 332 at 10:00 LT and a sub peak around 16:00 LT (Fig. 3(e)), following the diurnal variations
 333 of P_{RO₂} of C₇-C₁₀ aromatics (Fig. 4(b-d)) but poorly correlated with P_{RO₂} of benzene
 334 (Fig. 4(a)). Furthermore, OOMs with 8 carbon atoms have the highest signal in this
 335 factor (Fig. 2(a)), derived from the most abundant C₈-aromatics + styrene RO₂ (Fig.
 336 4(f)). Both of these can be explained by the fact that substituted aromatics have higher
 337 OH reactivity (Bloss et al., 2005) and higher HOM yields (Wang et al., 2017; Molteni

338 et al., 2018) than their homologues with less carbon atoms. In terms of molecular
 339 formula, the aromatics-derived OOMs have an overlap with monoterpenes-derived
 340 OOMs (Mehra et al., 2020). Monoterpenes can contribute more C₁₀ OOMs than
 341 aromatics ($P_{\text{MT-RO}_2} > P_{\text{C}_{10} \text{ Aro-RO}_2}$), but aromatics play a more important role in total
 342 in this factor since they provide more RO₂ in the urban atmosphere (Fig. 3(f)).

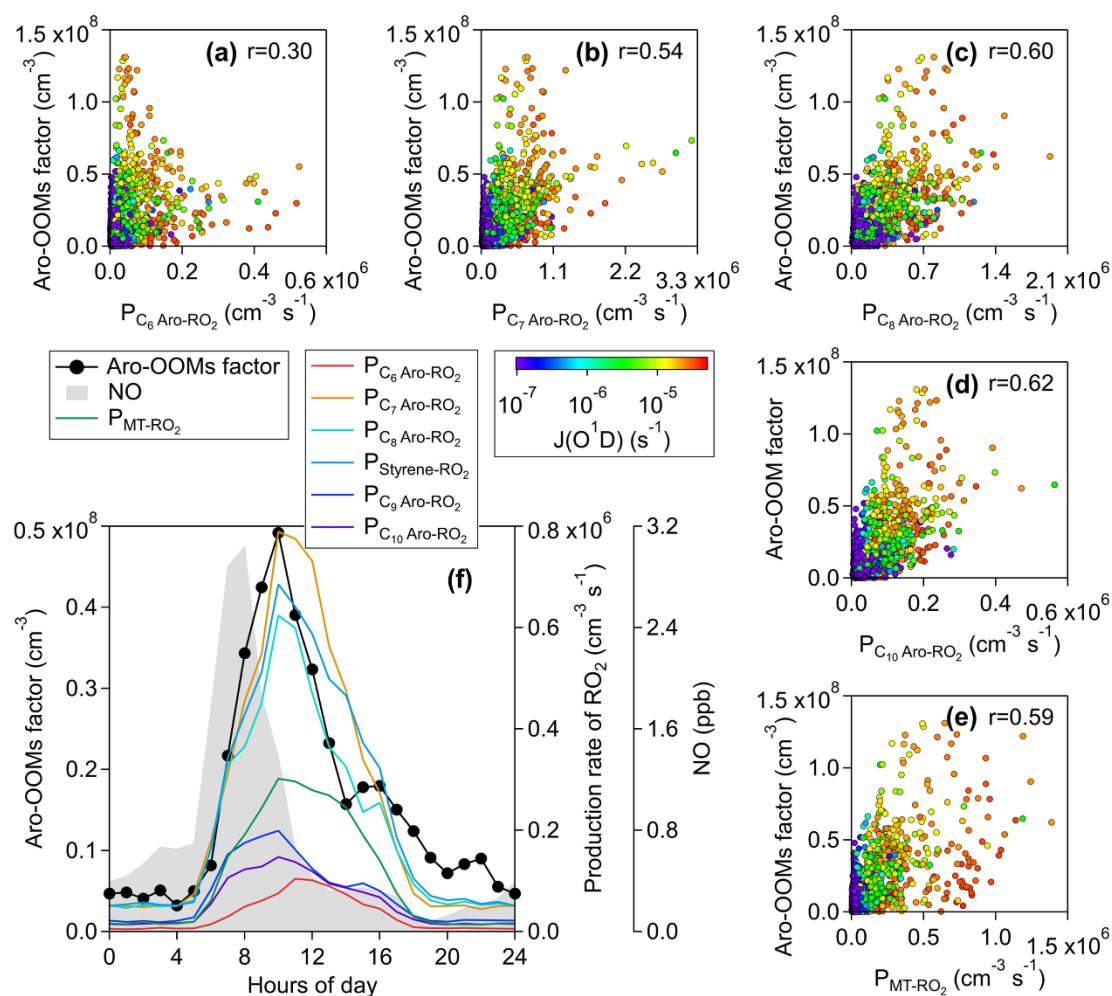
$$P_{\text{RO}_2} = k_{\text{OH+VOC}} \cdot [\text{OH}] \cdot [\text{VOC}] \quad (4)$$



346
 347 Figure 3. Mass spectra of (a) the Aro-OOMs factor, (b) the Temp-related factor, (c) the
 348 Aliph-OOMs factor, and the elemental formulas of major peaks are labeled above them.
 349 Peaks are color-coded by n_N as indicated at the top right of the figure, and the fractions
 350 of peaks grouped by n_N are reported in (d) the pie chart. The gray sticks are fluorinated
 351 contaminations, or non-identified compounds. The nitrated phenols are drawn
 352 separately with black peaks in (a), (b) and (c), and were not included in (d). So n_N can
 353 more reliably represent the number of nitrate groups in each molecule. Diurnal patterns
 354 (Beijing time) of these three factors are shown in (e), the bold solid lines are the median
 355 values, shaded areas represent percentiles of 75 % and 25 % and solid circles represent
 356 mean values.

357
 358 The main molecules of the Aro-OOMs factor are summarized in Table S2. The C_xH_{2x-5}O₆N
 359 (x=6-12, of which C₈H₁₁O₆N is the most intense) series can be produced by the
 360 reaction (R1a) of NO with the bicyclic peroxy radicals (HO-Ar-(O₂)₂), the key
 361 intermediates for aromatics oxidation proposed in the MCM (Bloss et al., 2005; Birdsall
 362 and Elrod, 2011). And here dihydroxy nitro-BTEX (C_xH_{2x-7}O₄N, x=6-8) can be treated

363 as indicators of aromatics oxidation. In addition to the conventional products, $C_9H_{13}O_7$ -
 364 $_9N$ from the $C_xH_{2x-5}O_{7-9}N$ ($x=7-13$) series are also significant in the OH-initiated and
 365 NO_x -influenced oxidation experiments of 1,2,4-trimethylbenzene (Zaytsev et al.,
 366 2019) and of 1,3,5-trimethylbenzene (Tsiligiannis et al., 2019). More oxygenated
 367 compounds may come from auto-oxidation and multigenerational OH attacks. However,
 368 the effective OSc of this factor (Table 1) is lower than that of oxidation products of
 369 aromatics in recent laboratories (Zaytsev et al., 2019;Tsiligiannis et al., 2019;Garmash
 370 et al., 2020;Wang et al., 2020c). We speculate that the abundances of NO_x relative to
 371 oxidants and precursors in these experiments are not sufficient to reproduce the
 372 atmospheric conditions during our observation, or that HOMs are more concentrated in
 373 aerosols due to the large condensation sink on this site (Qi et al., 2015). Although
 374 species with DBE < 3 (Fig. 2(b)) in this factor are most likely produced from multiple
 375 OH attacks in aromatics oxidation, we can't rule out the contribution of alkanes co-
 376 mitted with aromatics, such as the series $C_xH_{2x-1}O_6N$ ($x=5-14$).
 377



378
 379 Figure 4. Correlations of the Aro-OOMs dominated factor with production rate of RO_2
 380 from OH-initiated oxidation of (a) benzene ($P_{C_6 \text{ Aro-}RO_2}$), (b) toluene ($P_{C_7 \text{ Aro-}RO_2}$), (c)
 381 C_8 aromatics ($P_{C_8 \text{ Aro-}RO_2}$), (d) C_{10} aromatics ($P_{C_{10} \text{ Aro-}RO_2}$), and (e) Monoterpenes

382 (P_{MT-RO_2}). All the scatters are colored by $J(O^1D)$, to show the difference between day
383 and night. The median diurnal patterns of this factor and related parameters are plotted
384 in (f).

385

386 **Temp-related factor**

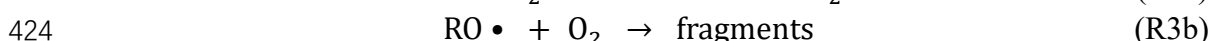
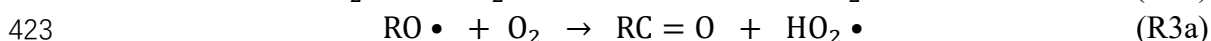
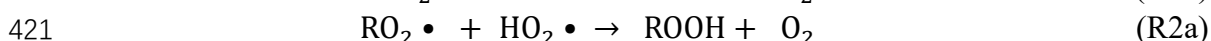
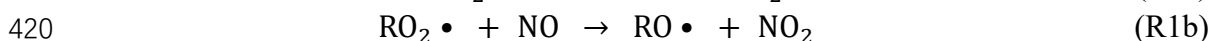
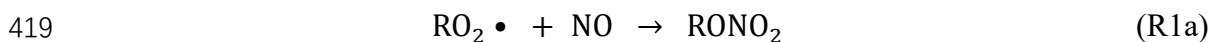
387

388 This factor is named due to good correlation with temperature (Fig. 5), and shows
389 maximum intensity in the afternoon around 15:00 (Fig. 3(e)). The Temp-related factor
390 is the only one dominated by non-nitrogenous organics (Fig. 3(b) and (d)), and has the
391 highest effective OSc (Table 1) among all the factors. The $C_xH_{2x-4}O_5$ ($x=5-11$,
392 summarized in Table S3), $C_xH_{2x-2}O_5$ ($x=5-10$), $C_xH_{2x-6}O_5$ ($x=5-11$), and $C_xH_{2x-4}O_6$
393 ($x=5-10$) series are possibly products from RO_2 terminated by HO_2 (R2a), or closed-
394 shell products from RO in reactions R3a or R3b. Temperature starts to rise at 6:00 LT
395 (Fig. 12(b)), but this factor does not accumulate significantly until after about 10:00 LT
396 (Fig. 3(e)), when the mixed level of NO is reduced to 1 ppb (Fig. 4(f)). This
397 phenomenon suggests a probability of HO_2 -driven chemistry of this factor under low
398 NO conditions, since that NO can consume HO_2 and compete with HO_2 for RO_2 . Such
399 low-NO atmospheric oxidation pathways have been suggested to be non-negligible in
400 the afternoon in central Beijing (Newland et al., 2021).

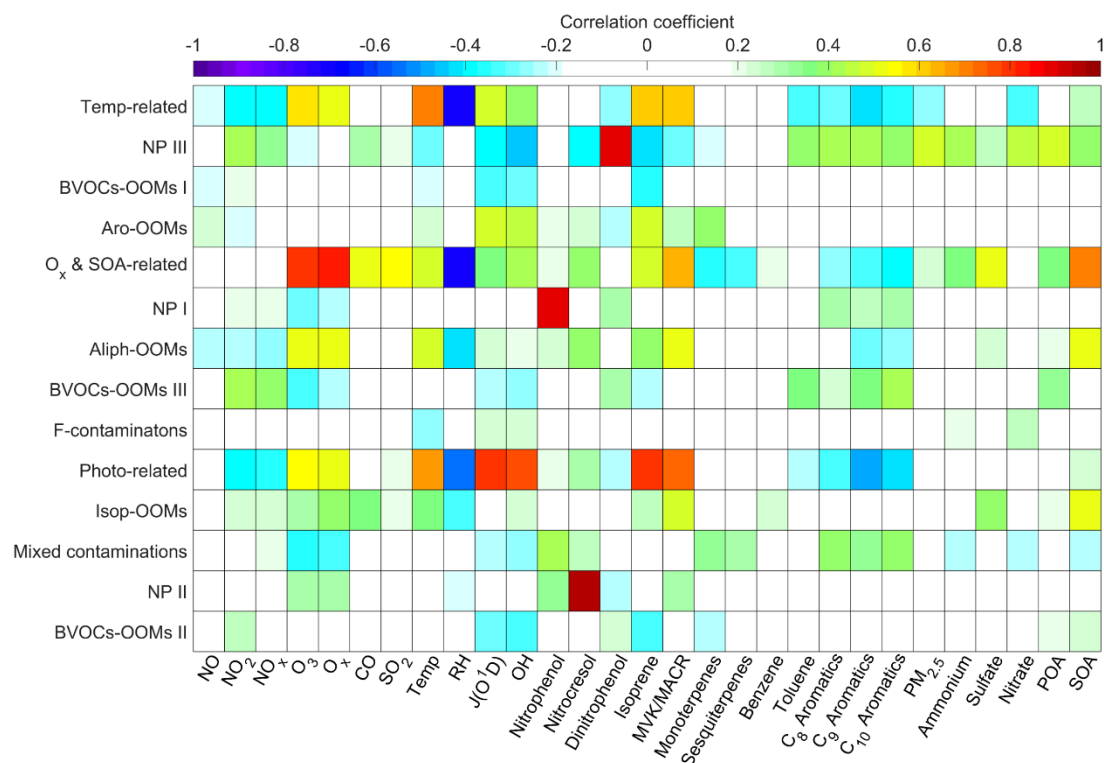
401

402 A factor caused by similar chemical processes called isoprene afternoon was discovered
403 in the nitrate CI-API-TOF data collected at a forest site in Centreville, Alabama, USA
404 (Massoli et al., 2018), correlated well with HO_2 , O_3 , and temperature. We also observed
405 a number of isoprene oxidation products in the Temp-related factor ($nC = 4, 5$ in Fig.
406 2(a)). Many of the $C_xH_{2x-1}O_6N$ ($x=3-7$) and $C_xH_{2x-3}O_6N$ ($x=4-9$) series were also present
407 in the light HOMs factor which was supposed to be fragments from the oxidation of
408 larger VOCs (e.g., monoterpene) in Hyytiälä Finland (Yan et al., 2016). While at the
409 SORPES station, the C_6 - C_9 ions should mainly come from the oxidation of
410 anthropogenic VOCs. At lower temperatures, the propensity of condensable organic
411 molecules to condense into aerosol makes the concentration measured using nitrate CI-
412 API-TOF lower. Thus, the total concentration of the Temp-related factor in the gas and
413 aerosol phases was calculated based on gas-particle equilibrium (section S5 in
414 the supplement), and was found to be still temperature dependent (Fig. S7), illustrating
415 the temperature-influenced chemical process controlling the factor. For instance,
416 Unimolecular reaction rates like RO_2 H-shifts increase qualitatively with temperature
417 (Bianchi et al., 2019; Frege et al., 2018).

418



425



427

428 Figure 5. Correlations of PMF factors with external gas-phase and particulate tracers
 429 from other instruments deployed at the SOPRES station, with the color representing the
 430 Pearson correlation coefficients. From left to right, the tracers are gas-phase species
 431 (NO , NO_2 , NO_x , O_3 , CO , SO_2), meteorological data (temperature (Temp), relative
 432 humidity (RH), photolysis constants ($J(\text{O}^1\text{D})$)), nitrate CI-APi-TOF data (OH,
 433 nitrophenol, nitrocresol, dinitrophenol), PTR-ToF-MS data (isoprene, methyl vinyl
 434 ketone/methacrolein (MVK/MACR)), monoterpenes, sesquiterpenes, benzene, toluene,
 435 C_8 aromatics, C_9 aromatics, C_{10} aromatics), $\text{PM}_{2.5}$, and ACSM data (ammonium, sulfate,
 436 nitrate, POA, SOA).

437

438 Aliph-OOMs factor

439

440 This factor is dominated by organic nitrates (Fig. 3(c) and (d)), and contains the bulk
 441 of anthropogenic di-nitrates and tri-nitrates. The $\text{C}_x\text{H}_{2x-2}\text{O}_8\text{N}_2$ ($x=4-13$, summarized in
 442 Table S4) and $\text{C}_x\text{H}_{2x}\text{O}_8\text{N}_2$ ($x=4-9$) series have not been reported in aromatics oxidation
 443 experiments under high NO_x conditions (Tsiligiannis et al., 2019; Wang et al., 2020c),
 444 and nor in the forest or rural environments (Yan et al., 2016; Massoli et al., 2018). A
 445 reasonable assumption is that these saturated or nearly saturated compounds are the
 446 products of aliphatics (including alkanes, alkenes, aliphatic alcohol, etc.) during their
 447 oxidation affected intensively by NO_x in the urban atmosphere. The Aliph-OOMs factor
 448 has a broad afternoon peak lasting from 14:00 to 19:00 LT (Fig. 3(e)), suggesting that
 449 the formation of multi-nitrate requires enough OH exposure time.

450

451 Considering a simple scenario of alkane photo-oxidation under high NO_x conditions:

452 the RO₂ generated from OH attack is completely terminated by NO (Fig. 6(a)). The
453 chain-retaining products are C_nH_{2n}O (one more carbonyl group than the precursor) and
454 C_nH_{2n+1}O₃N (one more nitrate group than the precursor), and the re-oxidation of these
455 products is a repetition of the above process which is defined as the basic reaction
456 scheme. The multiple (1st to 3rd) generation products of alkanes summarized in Fig.
457 6(b) are regarded as reference compounds, which we compare OOMs with to
458 investigate other mechanisms that differ from those shown in Fig. 6(a). Specifically,
459 this comparison is performed between the reference molecule and OOMs with the same
460 numbers of carbon, hydrogen and nitrogen atoms, but different numbers of oxygen
461 atoms. The number of extra oxygen ($n_{\text{O}_{\text{extra}}}$) from each aliphatic OOM over its
462 corresponding reference molecule was calculated by Eq. (5), that is, subtracting
463 carbonyl and nitrate oxygens from the molecule. Thus, the $n_{\text{O}_{\text{extra}}}$ can represent the
464 additional oxygenated moieties such as hydroxyl group (-OH), peroxy group (-OOH),
465 and possibly ether group. These functional groups may come from RO isomerization
466 (Orlando et al., 2003), the addition of OH to alkenes, or pre-existing moieties in the
467 precursor, RO₂ autoxidation or specific RO₂ bimolecular termination reactions
468 (RO₂+RO₂, RO₂+HO₂).

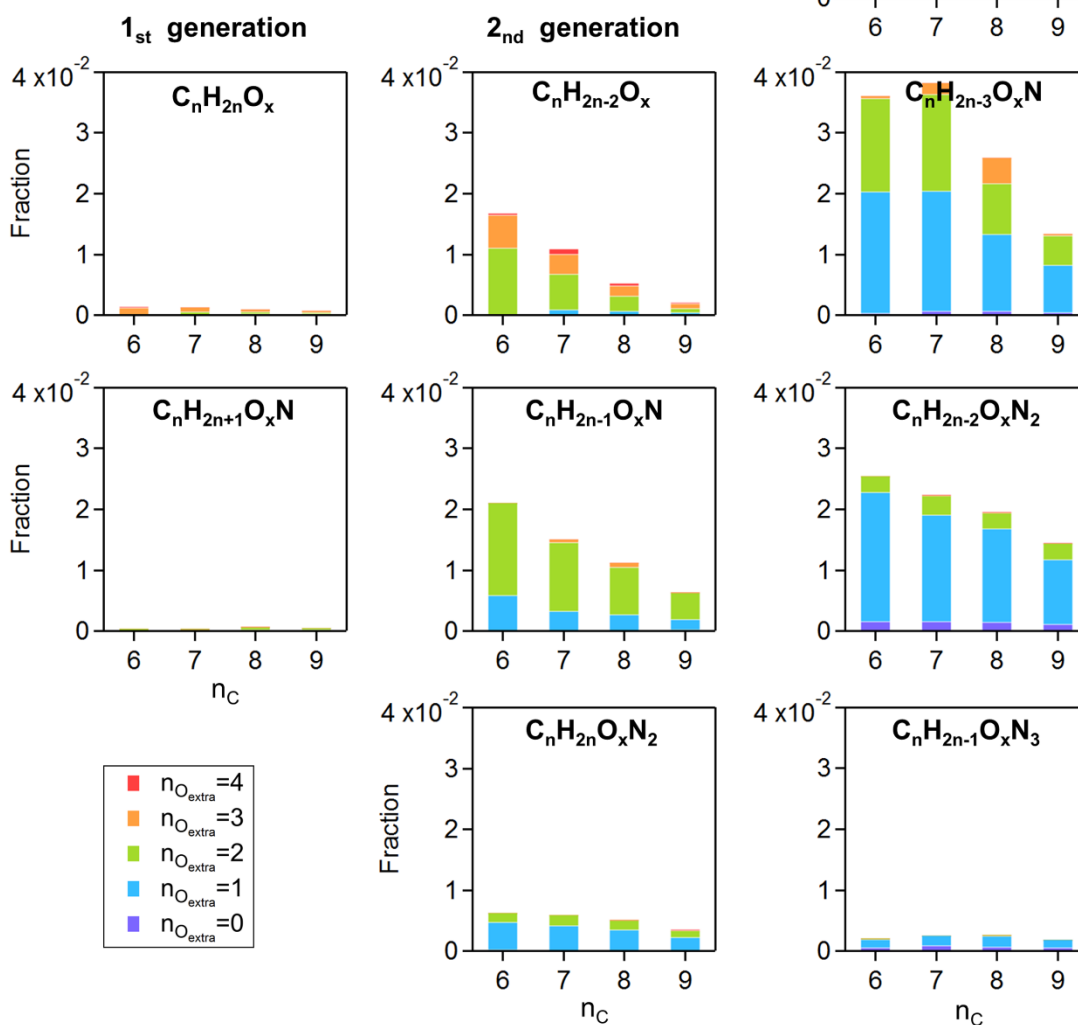
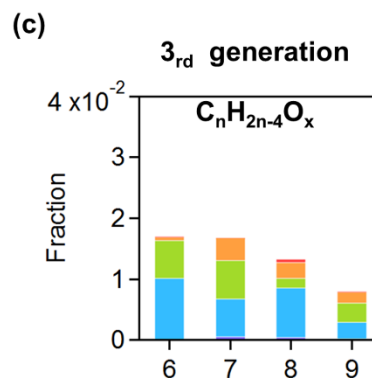
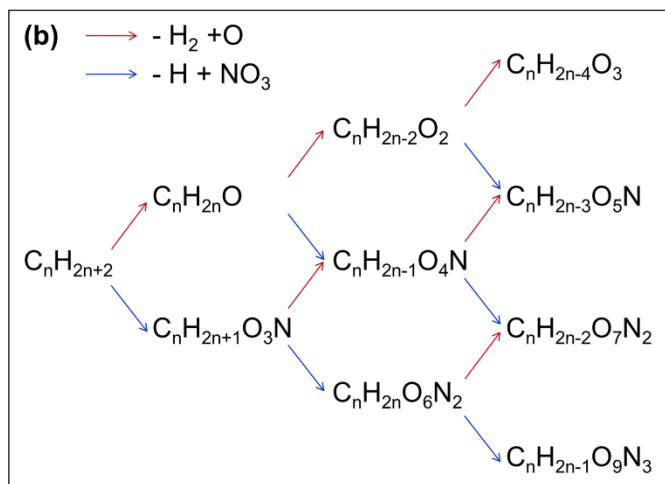
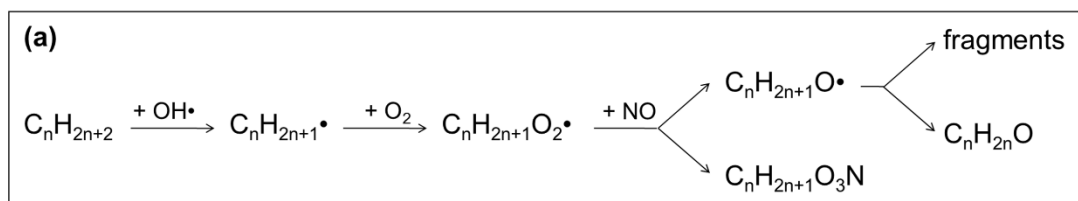
469

$$470 \quad n_{\text{O}_{\text{extra}}} = n_{\text{O}} - \text{DBE} - 3 \times n_{\text{N}} \quad (5)$$

471

472 As showed in Fig. 6(c), aliphatic OOMs in this factor are mainly the third-generation
473 products followed by the second-generation products, and both of which have one or
474 two oxygen-containing functional groups in addition to the carbonyls and nitrates. It
475 should be noted that the first-generation (Fig. 6(a)) and basic products (Fig. 6(b)) here
476 are underestimated due to the low sensitivity of nitrate CI-API-TOF to these compounds.
477 The multifunctional products of aliphatics are condensable to form SOA (correlation
478 coefficients with SOA showed in Fig. 5). Recent work has showed that autoxidation is
479 more common than previously thought (Wang et al., 2021), and more studies are needed
480 to explore the oxidation mechanisms of anthropogenic aliphatics and to evaluate their
481 contribution to SOA.

482



483

484 Figure 6. (a) Simplified oxidation mechanism for alkanes attacked by OH once under

485 NO_x-controlled conditions. (b) summarizes the changes in molecular formula of the 1st
486 to 3rd generation products of alkanes, based on the basic reaction scheme in (a). (c)
487 shows the fraction of potential alkanes-derived compounds in the Aliph-OOMs factor.
488 The compounds listed in (c) are grouped according to the molecular formulas in (b),
489 i.e., the same number of carbon, hydrogen and nitrogen atoms, but different numbers
490 of oxygen atoms. The bars in (c) are colored with $n_{O_{\text{extra}}}$. Please see text for details

491 about $n_{O_{\text{extra}}}$.

492

493 **3.2 Isoprene-related chemistry**

494

495 The following factors are characterized by C₅ OOMs (Fig. 2(a)), of which an isoprene
496 dihydroxyl dinitrate C₅H₁₀O₈N₂ (charged by NO₃⁻ at m/z 288 Th) is the fingerprint
497 molecule (Fig. 7). Apart from isoprene-derived compounds, OOMs formed from other
498 precursors undergoing the similar chemical processes are also allocated to these three
499 factors.

500

501 **Photo-related factor**

502

503 This factor is defined based on its correlation with J(O¹D) (Fig. 5), having an apparent
504 diurnal cycle with a peak at 12:00 LT (Fig. 7(e)). The major peak of the Photo-related
505 factor is C₅H₁₀O₈N₂ (Fig. 7(a)), most probably generating from double OH attack
506 proceed with double RO₂+NO termination (Jenkin et al., 2015). C₅H₁₀O₈N₂ can be also
507 produced in NO₃+ isoprene system (Ng et al., 2008;Zhao et al., 2020), whereas in this
508 study, the nocturnal C₅H₁₀O₈N₂ is principally from the Isop-OOMs factor (Fig. 8(b))
509 which will be discussed later. Other peaks with nC≤5, like C₅H₇O₇N, C₄H₇O₆N,
510 C₅H₉O₆N, are also likely to be the isoprene products. The total signal of compounds
511 with nC > 5 is not low, although their respective proportions are not as prominent as C₅
512 species (Fig. 7(d)), implying the contribution of other precursors together with isoprene.
513 In addition, the relationship of this factor with isoprene and J(O¹D) together (Fig. 5)
514 reveals the effect of light-dependent emission of isoprene on it.

515

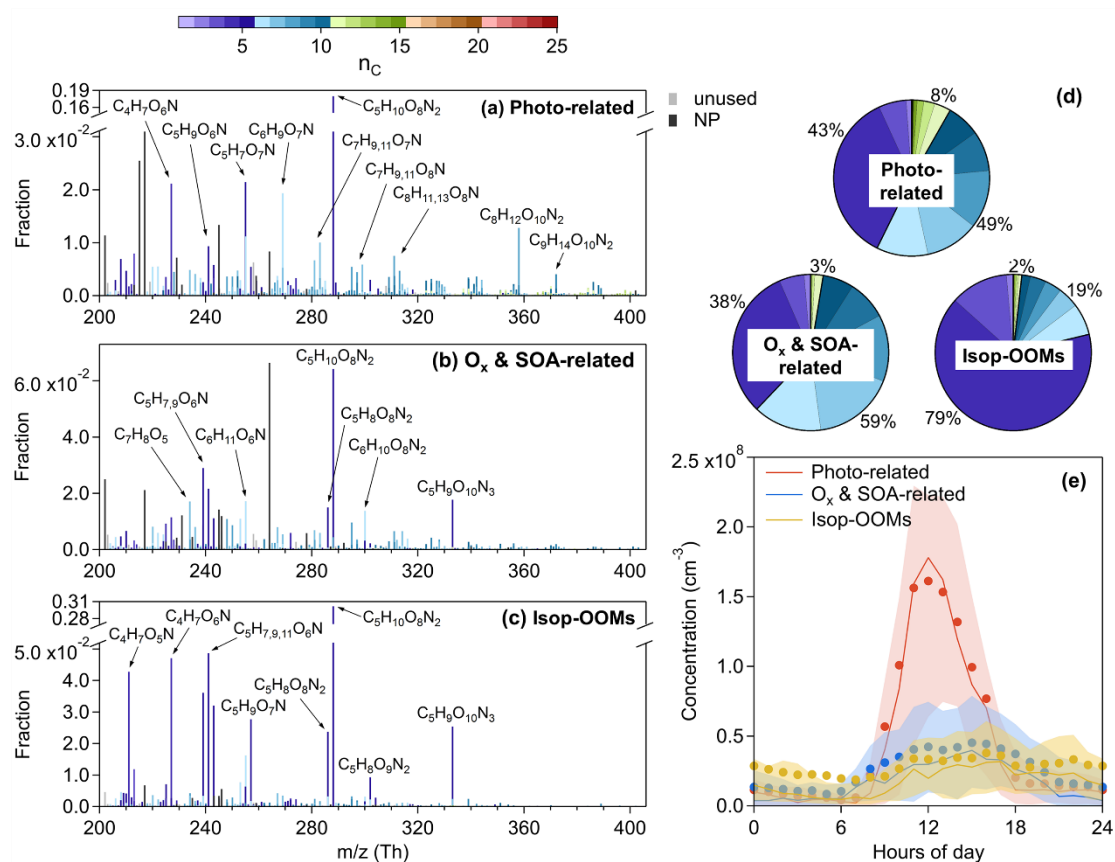
516 **O_x & SOA-related factor**

517

518 The atmospheric oxidation of VOCs produces low-volatile compounds, forming SOA
519 through gas-particle partitioning, and concurrently promotes ozone formation
520 (Atkinson, 2000). Both SOA and O_x have long lifetimes (>12 h), and their correlations
521 have been extensively investigated (Herndon et al., 2008;Wood et al., 2010;Hu et al.,
522 2016). The OOMs factor related to ozone and SOA together (Fig. 5), having slightly
523 elevated concentrations during daytime (Fig. 7(e)), is considered to be generated from
524 this photochemical aging process. Apart from C₅H₁₀O₈N₂, other isoprene multi-nitrates
525 are also present in this factor. C₅H₉O₁₀N₃, an isoprene hydroxyl trinitrate requiring at
526 least two steps of oxidation found in the experimental study on isoprene oxidation by

527 NO₃ (Zhao et al., 2020), naturally does not appear in the photo-related factor at all, but
 528 is mostly apportioned into the O_x & SOA-related factor and the Isop-OOMs factor (Fig.
 529 8(c) and 8(d)). Like the photo-related factor, isoprene is a significant but not the only
 530 precursor of this factor (Fig. 2 and 7). The biggest peak of the O_x & SOA-related factor
 531 is an ion at m/z 264 with formula C₆H₅O₃N (HNO₃NO₃⁻), identified as an adduct of
 532 nitrophenol (C₆H₅O₃N) with nitrate dimer (HNO₃NO₃⁻). The time variation of
 533 C₆H₅O₃N (HNO₃NO₃⁻) is influenced by the reagent ions in addition to the atmospheric
 534 nitrophenol. So far, we don't know why this compound share the same processes with
 535 others, but we did a test that removing the bins with unit m/z = 264 from the input
 536 matrix and still got this factor from PMF model.

537



538

539 Figure 7. Mass spectra of (a) the Photo-related factor, (b) the O_x & SOA-related factor,
 540 (c) the Isop-OOMs factor, and the elemental formulas of major peaks are labeled above
 541 them. Peaks are color-coded by n_C as indicated at the top of the figure, and the fractions
 542 of peaks grouped by n_C are reported in (d) the pie chart. The gray sticks are fluorinated
 543 contaminations, or non-identified compounds. The nitrated phenols are drawn
 544 separately with black peaks in (a), (b) and (c). The molecules represented by the gray
 545 and black sticks were not included in (d). Diurnal patterns of the three factors are shown
 546 in (e), the bold solid lines are the median values, shaded areas represent percentiles of
 547 75 % and 25 % and solid circles represent mean values.

548

549 **Isop-OOMs factor**

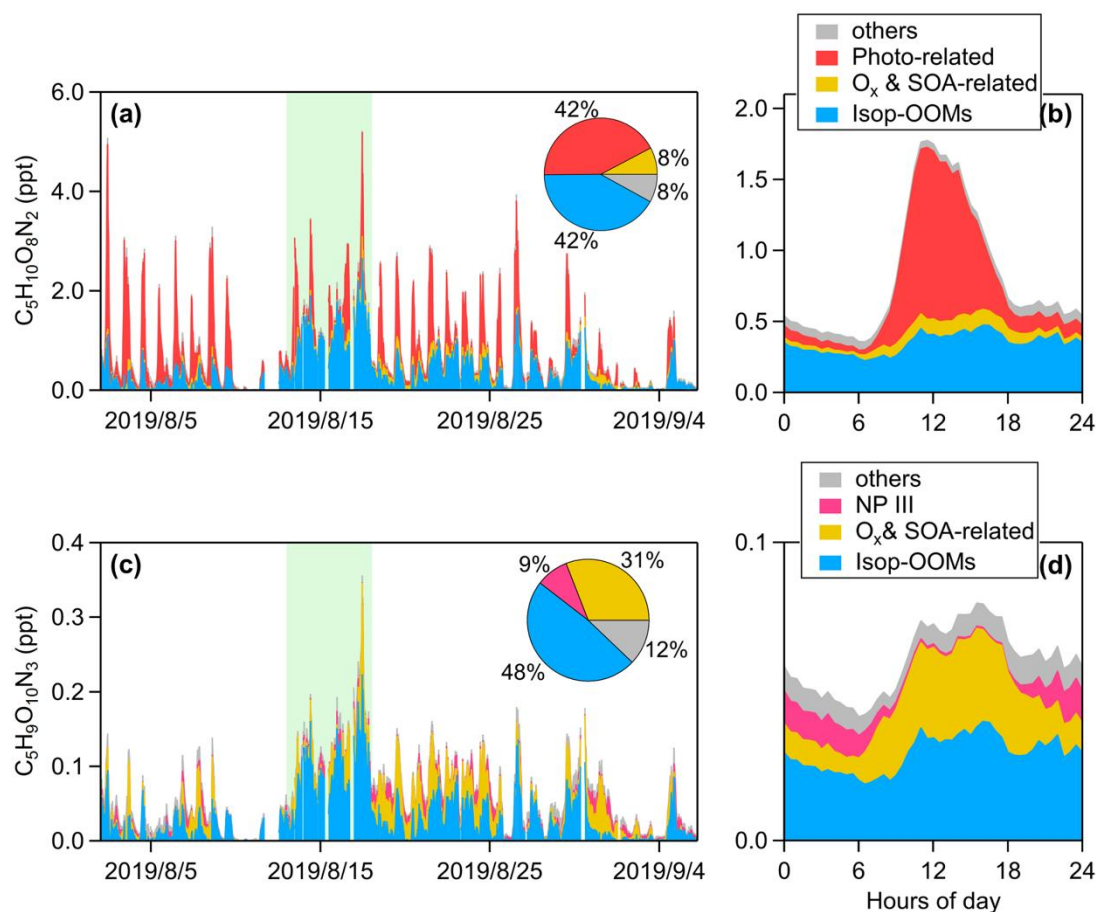
550

551 The mass spectra of the Isop-OOMs factor, as its name implies, is exclusively
552 contributed by isoprene-derived compounds (Fig. 7(c)). $C_5H_{10}N_2O_8$ contributes about
553 30% of the intensity of this factor, and the dominance of $C_5H_{10}N_2O_8$ was also found in
554 the isoprene nitrates type I factor in Centreville (Massoli et al., 2018). In addition to
555 multi-nitrates ($C_5H_{10}O_{7-8}N_2$, $C_5H_8O_{6-9}N_2$, and $C_5H_9O_{10}N_3$ summarized in Table. S6),
556 several mononitrate series ($C_4H_7O_{5-7}N$, $C_5H_9O_{4-9}N$, $C_5H_7O_{5-8}N$, and $C_5H_{11}O_{5-6}N$) of
557 this factor are also abundant in the isoprene nitrates type II factor in Centreville
558 (Massoli et al., 2018). Many of isoprene nitrates here have been specially investigated
559 in our previous observations in the YRD (Xu et al., 2021), and have been discovered in
560 other filed measurements (Lee et al., 2016; Massoli et al., 2018) and in many
561 laboratories (Ng et al., 2008; Lambe et al., 2017). Generally, these compounds are
562 second- and third-generation OH oxidation products of isoprene under high- NO_x
563 conditions (Wennberg et al., 2018).

564

565 The diurnal pattern of the Isop-OOMs factor is relatively unclear (Fig. 7(e)), with
566 obvious differences between mean and median values usually caused by plume events.
567 This indicates that isoprene chemistry, usually varying evidently from day (OH-
568 initiated) to night (NO_3 -initiated), is not the driver of this factor. This factor correlates
569 positively with MVK / MACR and SOA ($r > 0.50$, showed in Fig. 5), but not with
570 isoprene and OH. It seems that these isoprene OOMs are produced elsewhere and then
571 transported due to their longer lifetime determined by their relatively high volatility
572 (Table 1). The Isop-OOMs factor in the continental air masses are more intensive than
573 those in the coastal and YRD air masses (Fig. S8), consistent with the spatial
574 distribution of isoprene emissions (Sindelarova et al., 2014). An archetypal episode
575 affected by continental air masses (August 13 to August 17, 2019) is showed in Fig. 8.
576 During this period, $C_5H_9O_{10}N_3$ was almost entirely transported, while $C_5H_{10}O_8N_2$ has
577 strong in situ photochemical generation, in addition to the source of transport.

578



579

580 Figure 8. Stacked (a) time series and (b) mean diurnal pattern of isoprene dihydroxyl
 581 dinitrate ($C_5H_{10}O_8N_2$). Stacked (c) time series and (d) mean diurnal pattern of isoprene
 582 hydroxyl trinitrate ($C_5H_9O_{10}N_3$). The contribution ratios of each PMF factor to these
 583 two compounds are reported in the pie chart respectively. The light green shaded area
 584 represents a typical episode influenced by transported continental air masses (August
 585 13 to August 17, 2019).

586

587 3.3 BVOCs nighttime chemistry

588

589 The following nighttime factors are characterized by C_{10} OOMs (Fig. 2(a)), which are
 590 identified as the oxidation products of monoterpenes. Except for the BVOCs-OOMs I
 591 factor (Fig. 9(a)), the contribution of isoprene-derived OOMs was much lower in these
 592 factors. Compared to the above isoprene-related factors, $C_5H_{10}O_8N_2$ and $C_5H_9O_{10}N_3$
 593 was no longer significantly present in the following factors.

594

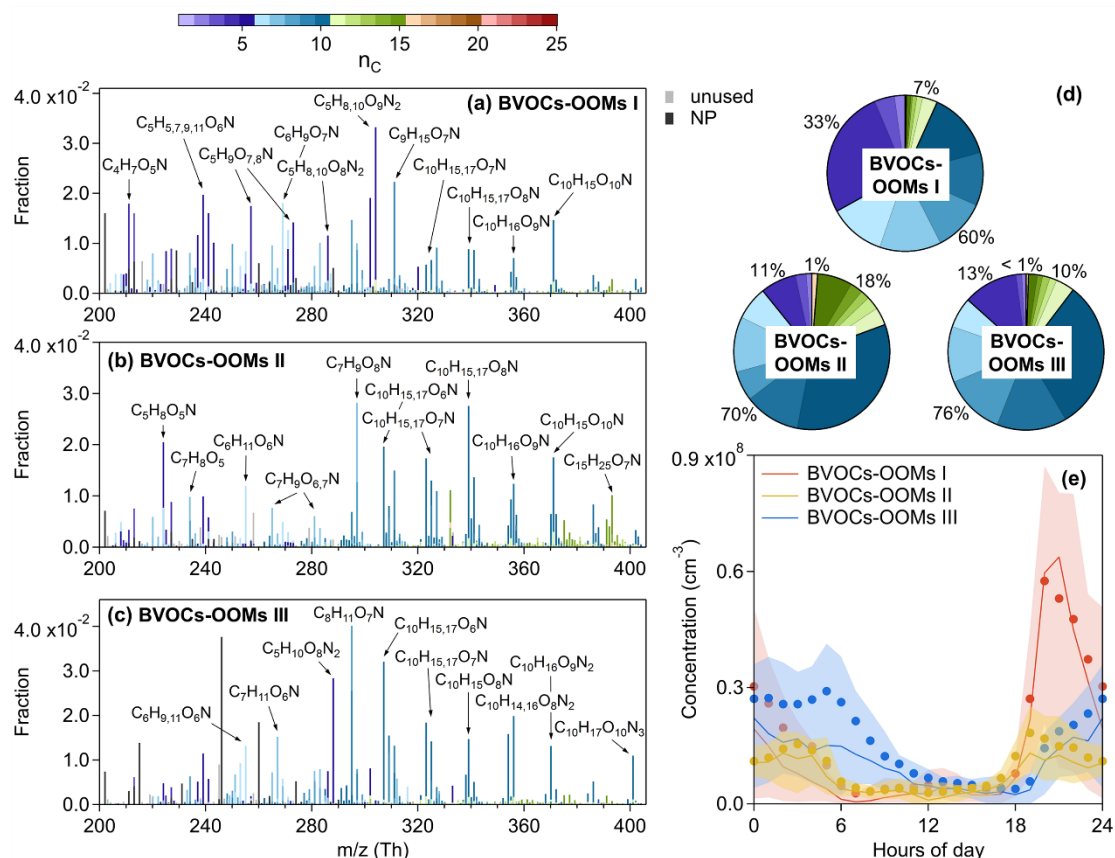
595 BVOCs-OOMs I factor

596

597 The first nighttime factor has its maximum concentration at around 20:00 LT, and
 598 decreases to very low value during the day. It is moderately correlated with the
 599 production rate of NO_3 radical (P_{NO_3} derived from Eq. 6) at night, and reaches high
 600 intensity only under conditions of NO below 1 ppb (Fig. 10(a)), indicating a chemical

601 process of NO₃ radical. The concentration of this factor is mainly from C₅ peaks,
 602 followed by C₆-C₁₀ peaks (Fig. 9(d)), about 80% of which are ONs (Fig. 2(c)),
 603 designating the oxidations of isoprene and monoterpenes by NO₃ (BVOCs-OOMs I).
 604 In the case of isoprene oxidation, the nitrate groups of C₅H₉O₄₋₈N, C₅H₇O₅₋₈N and
 605 C₄H₇O₅₋₆N series (summarized in Table S8) are likely to come from the addition of NO₃.
 606 Next, the C₅H₁₀O₈₋₉N₂ and C₅H₈O₇₋₁₀N₂ series are probably second-generation products.
 607 These compounds derived from isoprene+NO₃ system have been discussed in previous
 608 laboratory (Kwan et al., 2012; Zhao et al., 2020) and ambient data sets (Ayres et al.,
 609 2015; Xiong et al., 2015). Additionally, The C₆-C₁₀ species are potentially the products
 610 of monoterpenes degraded by NO₃.

$$P_{\text{NO}_3} = k_{\text{NO}_2+\text{O}_3} \cdot [\text{NO}_2] \cdot [\text{O}_3] \quad (6)$$



614
 615 Figure 9. Mass spectra of (a) the BVOCs-OOMs I factor, (b) the BVOCs-OOMs II
 616 factor, (c) the BVOCs-OOMs III factor, and the elemental formulas of major peaks are
 617 labeled above them. Peaks are color-coded by n_C as indicated at the top of the figure,
 618 and the fractions of peaks grouped by n_C are reported in (d) the pie chart. The gray
 619 sticks are fluorinated contaminations, or non-identified compounds. The nitrated
 620 phenols are drawn separately with black peaks in (a), (b) and (c). The molecules
 621 represented by the gray and black sticks were not included in (d). Diurnal patterns of
 622 these three factors are shown in (e), the bold solid lines are the median values, shaded
 623 areas represent percentiles of 75 % and 25 % and solid circles represent mean values.

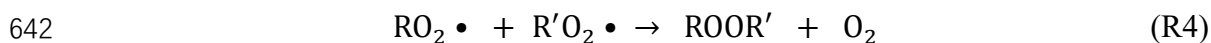
624

625 **BVOCs-OOMs II factor**

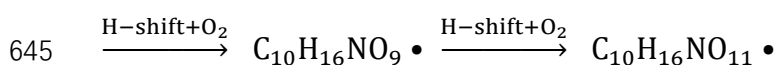
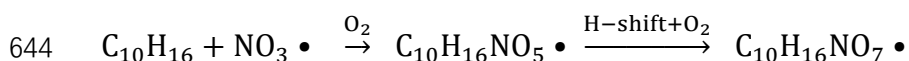
626

627 The second nighttime factor are intense at night and over five times lower during the
628 day. Like the BVOCs-OOMs I factor, this factor has high concentrations when NO is
629 reduced leading to increased NO₃ availability (Fig. 10(b)), and about 80% of
630 compounds in this factor are ONs (Fig. 2 (c)). Accordingly, this may also be a factor
631 strongly influenced by NO₃. It is dominated by C₆-C₁₀ OOMs, among which the highest
632 intensity is at C₁₀ (Fig. 9(d)), coherent with the nature of monoterpene products
633 (BVOCs-OOMs II). This factor has weaker signals at C₁₅ which are plausibly the
634 products of sesquiterpenes but could also be dimmers formed from R4 (monoterpenes
635 + isoprene or monoterpenes + C₅ monoterpene fragments). Compared to the BVOCs
636 OOMs I factor (Fig. 9(d)), this factor has more large mass molecules (C₁₀) and fewer
637 small mass molecules (C₅), resulting in an effective volatility over one order of
638 magnitude lower. A NO₃-initiated factor, called the nighttime type-2 factor, has also
639 been discovered in Hyytiälä Finland (Yan et al., 2016), but the similar factor we found
640 has a higher proportion of organic nitrates, due to the more abundant NO_x here.

641

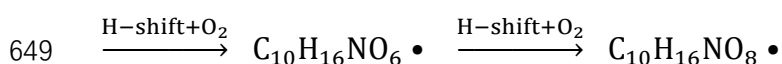
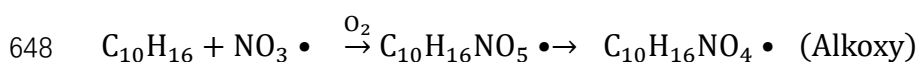


643



646 (R5a)

647



650 (R5b)

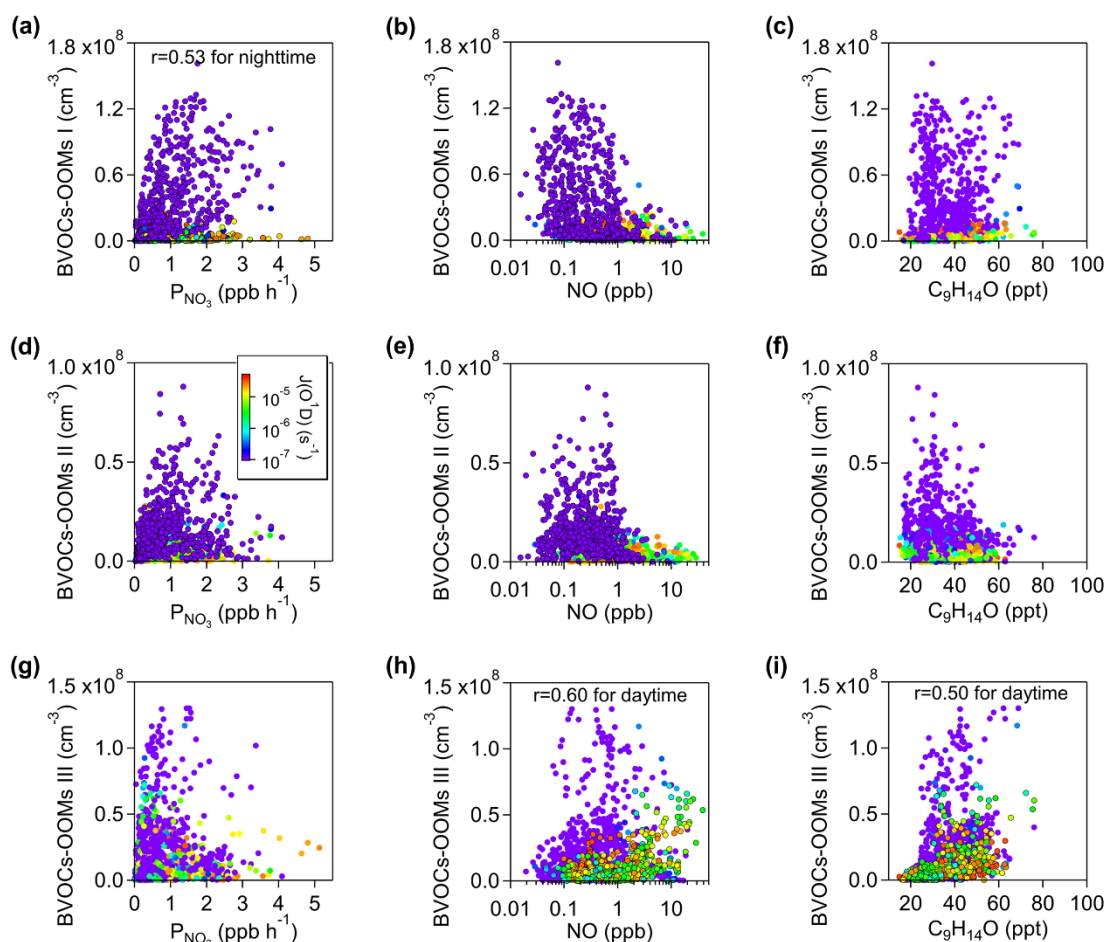
651

652 In terms of fingerprint molecules of this factor (summarized in Table S9), The
653 C₁₀H₁₅O₅₋₁₂N series are carbonyl products from precursor RO₂ or RO terminations,
654 while the C₁₀H₁₇O₅₋₉N series are alcohol or hydroperoxide products from precursor RO₂
655 terminations. The C₇H₉O₆₋₈N, C₉H₁₅O₆₋₉N, C₉H₁₃O₇₋₁₀N, and C₈H₁₃O₇₋₈N series are
656 expected to be fragments. The closed-shell compounds mentioned above have been
657 reported in the experiments of monoterpenes + NO₃ system (Nah et al., 2016; Faxon et
658 al., 2018; Takeuchi and Ng, 2019).

659

660 It is noteworthy that a set of nitrogen-containing radicals, C₁₀H₁₆O₆₋₁₁N (peak fitting
661 are shown in Fig. S9), is present in the BVOCs-OOMs II factor. Starting from a generic
662 monoterpene molecule with the formula C₁₀H₁₆, the NO₃ addition with fast O₂ addition
663 results in a peroxy radical with the formula C₁₀H₁₆O₅N, If the initial intermediate RO₂

664 is capable to proceed via autoxidation by the formal addition of O_2 , we expect radicals,
 665 $C_{10}H_{16}O_{5+2x}N$ (x denotes times of autoxidation performed) with an odd oxygen number,
 666 to be formed (R5a). In addition, peroxy radicals with an even oxygen number,
 667 $C_{10}H_{16}O_{6+2x}N$, are likely produced via reaction chain 5b: (1) RO_2 is propagated to RO
 668 through bimolecular reactions, and (2) RO isomerize to an alcohol by internal H
 669 abstraction forming a carbon-centered radical (Orlando et al., 2003; Orlando and
 670 Tyndall, 2012), (3) the carbon-centered radical can again take up an oxygen molecule
 671 and follow the autoxidation route. The $C_{10}H_{16}O_9N$ radical is also moderately intense in
 672 the BVOCs-OOMs I factor (Fig. 9(a)), testifying the presence of NO_3 chemistry. These
 673 $C_{10}H_{16}O_{6-11}N$ radicals are also reported in the CLOUD chamber (Yan et al., 2020). In
 674 addition to C_{10} radicals, a C_5 radical, $C_5H_8O_5N$ (peak fitting are shown in Fig. S9), is
 675 also found in the BVOCs-OOMs II factor. $C_5H_8O_5N$ are possibly derived from the
 676 oxidation of isoprene initiated by NO_3 , as observed in the laboratory (Zhao et al., 2020).
 677 Another hypothesis is that $C_5H_8O_5N$ is formed from the fragmentation process of
 678 monoterpenes.
 679



680

681 Figure 10. Scatter plots of the BVOCs-OOMs I factor with (a) P_{NO_3} , (b) NO, and (c)

682 nopinone ($C_9H_{14}O$). Scatter plots of the BVOCs-OOMs II factor with (d) P_{NO_3} , (e) NO,

683 and (f) nopinone ($C_9H_{14}O$). Scatter plots of the BVOCs-OOMs III factor with (g) P_{NO_3} ,

684 (h) NO, and (i) nopinone (C₉H₁₄O). All the scatters are colored by J(O¹D), to show the
685 difference between day and night. Pearson correlation coefficient showed in (a) is
686 calculated for nighttime, but the correlation coefficients in (c) are only for daytime.

687

688 **BVOCs-OOMs III factor**

689

690 The third nighttime factor (BVOCs-OOMs III) is dominated by nitrogen-containing
691 species with a contribution ratio about 90%, among which dinitrates account for more
692 than 20% (Fig. 2(c)). When grouped by carbon numbers, C₁₀ OOMs have the strongest
693 signal. Unlike the above two NO₃-related factors, this factor shows no correlation with
694 P_{NO₃}, but has positive correlation with NO, especially during the daytime (Fig. 10(c)).
695 C₉H₁₄O, a typical product of NO-affected monoterpenes oxidation (Calogirou et al.,
696 1999), is found to be correlated with this factor (Fig. 10(c)). It is reasonable to infer that
697 these organic nitrates may come from terminations of monoterpenes-RO₂ by NO. In
698 addition to the elevated intensity during the nighttime, this factor still remains at a
699 relatively high concentration in the morning, which is much higher than that of the two
700 NO₃-related factors (Figure 9(e)). Owing to the suppression of NO to RO₂ autoxidation
701 and the relatively insufficient oxidant in dark environment, the effective OSc of the
702 BVOCs-OOMs III factor is lower than other factors. Apart from the mononitrates
703 summarized in Tabel S10, the C₁₀H₁₆O₇₋₁₀N₂ (dinitrates) and C₁₀H₁₇O₁₀N₃ (a trinitrate
704 charged by NO₃⁻ at m/z 401) are most likely the result of multiple-generation processes
705 involving OH or NO₃ oxidation of monoterpenes proceeding RO₂ + NO terminations.
706 A similar factor, called terpene nitrates, has also been reported in Centreville, USA
707 (Massoli et al., 2018), while in Hyyti ä ä Finland (Yan et al., 2016), it's that the daytime
708 type-1 factor is related to NO.

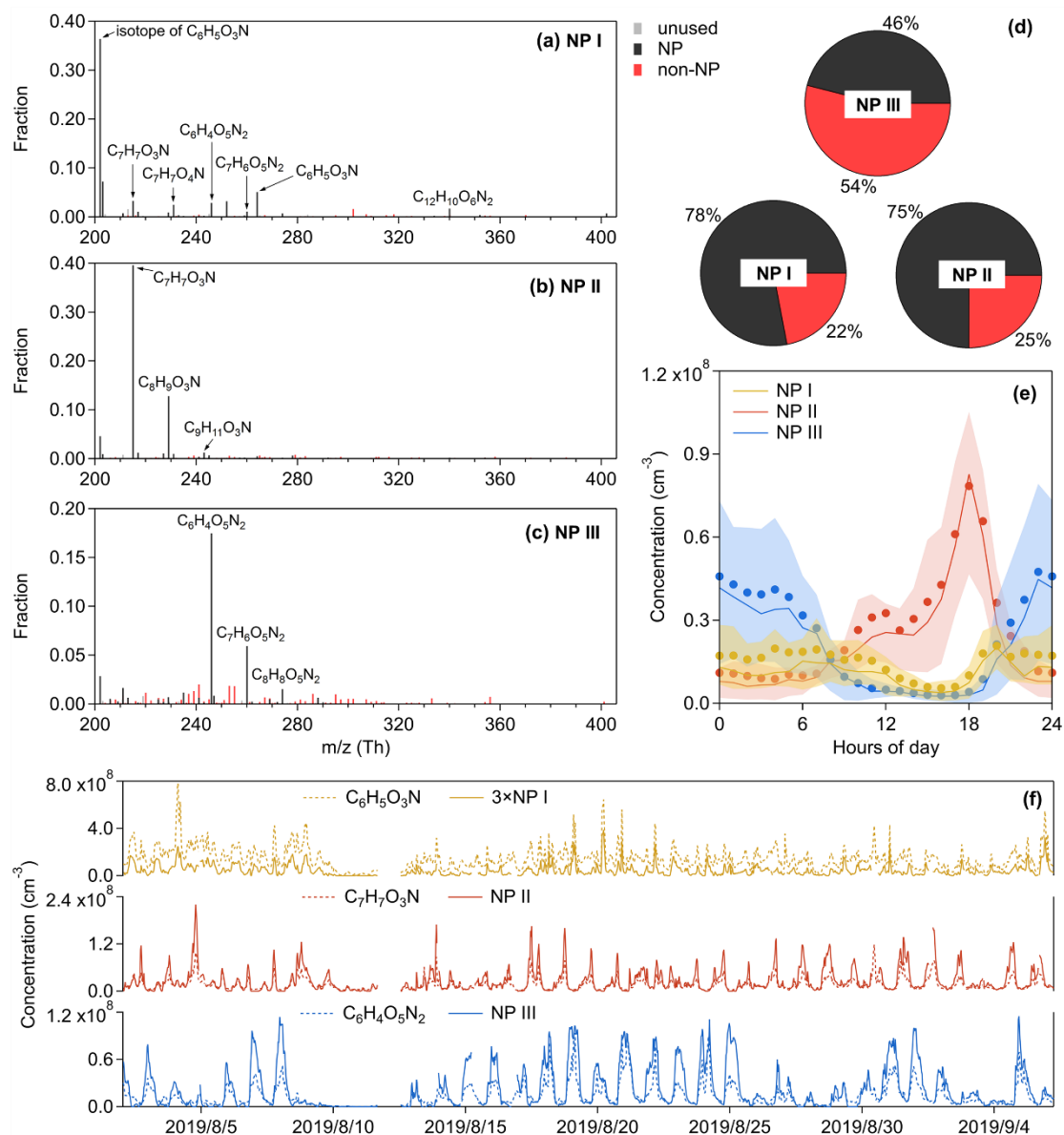
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710 **3.4 Nitrated phenols factors**

711

712 Nitrated phenols are of concern, because of their phytotoxicity (Rippen et al., 1987)
713 and as an important chromophores of brown carbon in aerosol (Desyaterik et al.,
714 2013;Mohr et al., 2013). The sources of these highly volatile compounds are attributed
715 to biomass burning, vehicle exhausts, and secondary gas or aqueous phase production
716 (Harrison et al., 2005). Here we identified three factors about NP, including the NP I
717 factor dominated by nitrophenol, the NP II factor dominated by substituted nitrophenols,
718 and the NP III factor dominated by dinitrophenols. Although the mass spectrum of the
719 NP III factor is less pure than the NP I & II factors (Fig. 11), its time series follows well
720 with C₆H₄O₅N₂ (Fig. 11(f)), implying that this factor is driven by di-nitrated-phenols
721 chemistry. Since nitrated phenols have been broadly investigated and relatively clearly
722 recognized (Harrison et al., 2005;Yuan et al., 2016;Wang et al., 2018b;Cheng et al.,
723 2021), they are not discussed too much here. It seems that the chemistry of nitrated
724 phenols is distinctive to other OOMs.

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Figure 11. Mass spectra of (a) the NP I factor, (b) the NP II factor, and (c) the NP III factor, and the elemental formulas of major peaks are labeled above them. The gray sticks are fluorinated contaminations, or non-identified compounds. The nitrated phenols are drawn separately with black peaks in (a), (b) and (c), while other OOMs are plotted as red peaks. The molecules represented by the gray were not included in (d). Diurnal patterns of these three factors are shown in (e), the bold solid lines are the median values, shaded areas represent percentiles of 75 % and 25 % and solid circles represent mean values. (f) Time series of PMF factors and tracers.

3.5 Ensemble chemical properties

After performing PMF analysis, over 1000 non-nitro molecules have been identified through HR peaks fitting in each factor. The mean concentration of total non-nitro OOMs reconstructed from the selected PMF solution is about 2.1×10^8 molecules cm^{-3} . Ensemble chemical properties of these non-nitro OOMs are summarized in Fig. 12. The

742 number of carbon atoms implies the precursor information of OOMs. C₅ OOMs, which
743 principally consist of isoprene products benefited from the high reactivity and intensive
744 emissions of isoprene in summer, are the most abundant (Fig. 12(c)). While C₆-C₉
745 OOMs are mostly likely formed from the oxidation of AVOCs such as aromatics and
746 aliphatic series in the urban and suburban atmosphere, and as we expected, these
747 AVOCs-derived OOMs account for about 50% of the total signal (Fig. 12(c)). The
748 intensity of OOMs decreases from C₇ to C₉ determined by the concentration distribution
749 of precursors, but becomes a plateau at C₁₀ (Fig. 12(c)), indicating another source of
750 C₁₀ OOMs, such as monoterpenes oxidation. These results underscore the formation of
751 SOA precursors from a mixture of anthropogenic and biogenic emissions, under
752 ongoing forest cover increases (Wang et al., 2020a) in highly urbanized eastern China.

753

754 In addition to the anthropogenic VOCs, another human-induced perturbation on the
755 formation of OOMs is the NO_x-affected chemistry of VOCs, i.e., RO₂ + NO
756 terminations or NO₃-initiated oxidations. As showed in Fig. 12(c), about 70% of OOMs
757 are nitrogen-bearing compounds, regarded as organic nitrates within the allowed range
758 of uncertainty. If isoprene nitrates are not included, organic nitrates peak at C₇ as do the
759 nitrogen-free species, showing the significant production of organic nitrates through the
760 AVOCs + NO_x pathways. The NO_x effect on AVOCs-derived OOMs, typified by the
761 Aro-OOMs factor and the Aliph-OOMs factor, are not showed in previous ambient
762 measurements (Yan et al., 2016; Lee et al., 2016; Massoli et al., 2018).

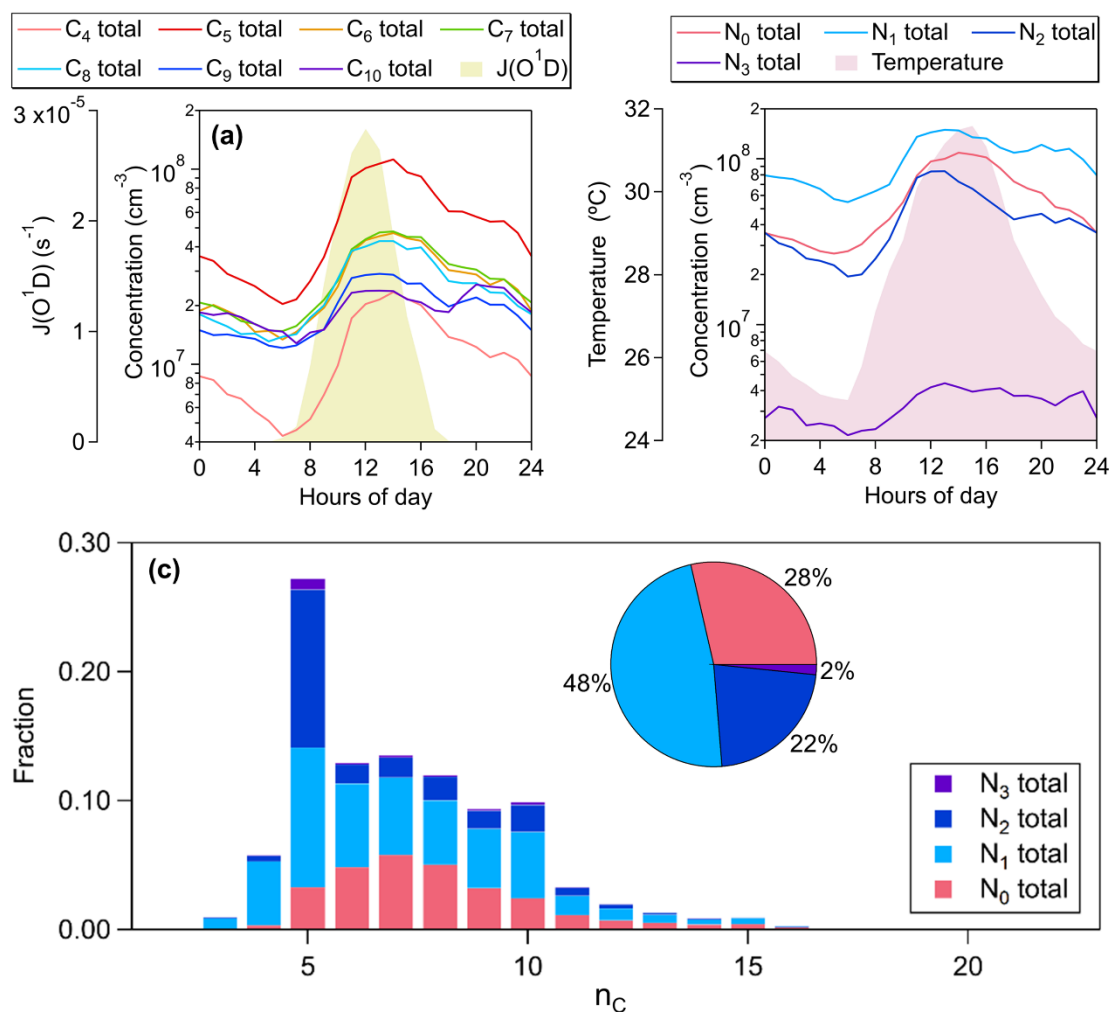
763

764 OOMs grouped by carbon numbers or nitrogen numbers consistently have absolute
765 high concentrations in the daytime (Fig. 12(a) and (b)), revealing the crucial role of
766 photochemical progress, involving RO₂ + NO termination reactions, on OOMs. In
767 addition, The C₅-C₁₀ OOMs are enhanced again during 19:00-22:00 LT, and the
768 nighttime peak of C₁₀ OOMs is even higher than its daytime peak (Fig. 12(a)). The
769 nocturnal C₁₀ OOMs are more intense than C₉ OOMs (Fig. 12(a)), and there are more
770 C₁₀ nitrates than C₉ nitrates (Fig. 12(c)). These results show the fate of VOCs degraded
771 by NO₃ during the nighttime, which are more important to monoterpenes. In contrast to
772 nitrogen-free OOMs, organic nitrates are enriched through the reactions of BVOCs with
773 NO₃ in the early evening (Fig. 12(b)), as indicated by three BVOCs nighttime chemistry
774 factors.

775

776 Apart from reflecting the influence of NO_x, multi-nitrates also imply the multiple
777 generations of VOCs oxidation, which is evident in the products of isoprene (e.g.,
778 C₅H₁₀O₈N₂ and C₅H₉O₁₀N₃) due to its two carbon-carbon double bonds. As products of
779 mononitrates, multi-nitrates follow mononitrates in diurnal variation, with double peaks
780 initiated by OH and NO₃ respectively (Fig. 12(b)). Considering that the formation of
781 organic nitrate is only a small branch of RO₂ + NO termination, the contribution of
782 multi-step oxidation should be larger than that shown in Fig. 12(c).

783



784

785 Figure 12. Ensemble chemical properties of non-nitro OOMs reconstructed from the
 786 selected PMF solution. (a) Median diurnal cycles of total compounds with carbon
 787 number of 5-10 respectively. (b) Median diurnal cycles of total compounds with n_N of
 788 0-3 respectively. (c) The distributions of total observed OOMs at different n_C . OOMs
 789 on each carbon number is grouped by nitrogen number, and the total concentration
 790 fractions of each groups are reported in the pie chart. Since we selected peaks in the
 791 m/z range of 202-404 Th, OOMs with $n_C < 5$ or $n_C > 10$ detected by nitrate CI-APi-
 792 TOF are underestimated.

793

794 4 Conclusions

795

796 We have investigated the sources and characteristics of gas-phase OOMs observed
 797 using a nitrate CI-APi-TOF at the SORPES station in the YRD of eastern China, an
 798 environment dominated by anthropogenic emissions with enhanced biogenic emissions
 799 during summer.

800

801 The binPMF analysis, which avoids the uncertainty introduced by high-resolution peak
 802 fitting to the input data matrix, was applied to deconvolve the complexity of the data
 803 set, and it resolved 14 factors, among which 12 factors have been discussed in detail. A

804 morning factor (Aro-OOMs), correlated with the production rates of RO₂ from
805 aromatics, is characterized by unsaturated products of aromatics such as C_xH_{2x-5}O₆₋₉N
806 (x=6-12). An afternoon factor (Aliph-OOMs), containing the bulk of C₆-C₉ dinitrates
807 and trinitrates such as C_xH_{2x-2}O₈N₂ (x=4-13) and C_xH_{2x}O₈N₂ (x=4-8), is assumed to be
808 derived from aliphatics oxidation. A transported factor (Isop-OOMs), correlates with
809 MVK / MACR and SOA, is exclusively dominated by isoprene nitrates (e.g.,
810 C₅H₁₀O₈N₂ and C₅H₉O₁₀N₃). A nighttime factor (BVOCs-OOMs III), related to NO, is
811 dominated by terpenes nitrates such as C₁₀H₁₅O₆N, C₁₀H₁₆O₇₋₁₀N₂ and C₁₀H₁₇O₁₀N₃. In
812 addition to the factors distinguished by precursors, several factors are driven by
813 chemistry. A factor following the J(O¹D) (Photo-related), consisting of isoprene
814 products mixed with others, is thought to be produced by in situ photochemistry. An
815 afternoon factor (Temp-related), having the most abundant nitrogen-free OOMs such
816 as C_xH_{2x-4}O₅₋₆ (x=5-10), C_xH_{2x-2}O₅ (x=5-10), and C_xH_{2x-6}O₄ (x=5-10), is generated
817 involving temperature-influenced chemistry. A daytime factor (O_x & SOA-related),
818 correlated well with O_x and SOA, indicates the photochemical aging process. Two
819 nighttime factors (BVOCs-OOMs I & II), benefiting from NO₃ and suppressed by NO,
820 are considered to be produced from NO₃-initiated oxidation of BVOCs, and both of them
821 have the fingerprint molecule, C₁₀H₁₆O₉N. The remaining three factors are governed
822 by nitrated phenols.

823

824 All of these factors from various precursors are influenced in different ways by NO_x.
825 Over 1000 non-nitro molecules have been identified and then reconstructed from
826 selected solution of binPMF, and about 72% of the total signal are contributed by
827 nitrogen-containing OOMs, almost regarded as organic nitrates formed through RO₂ +
828 NO terminations or NO₃-initiated oxidations. Moreover, multi-nitrates have a
829 contribution ratio of about 23% to total concentration, indicating the significant
830 presence of multiple oxidation generations, especially for isoprene (e.g., C₅H₁₀O₈N₂
831 and C₅H₉O₁₀N₃). The nitrate CI-APi-TOF data set presented here highlight the decisive
832 role of NO_x chemistry on OOMs formation in densely populated areas.

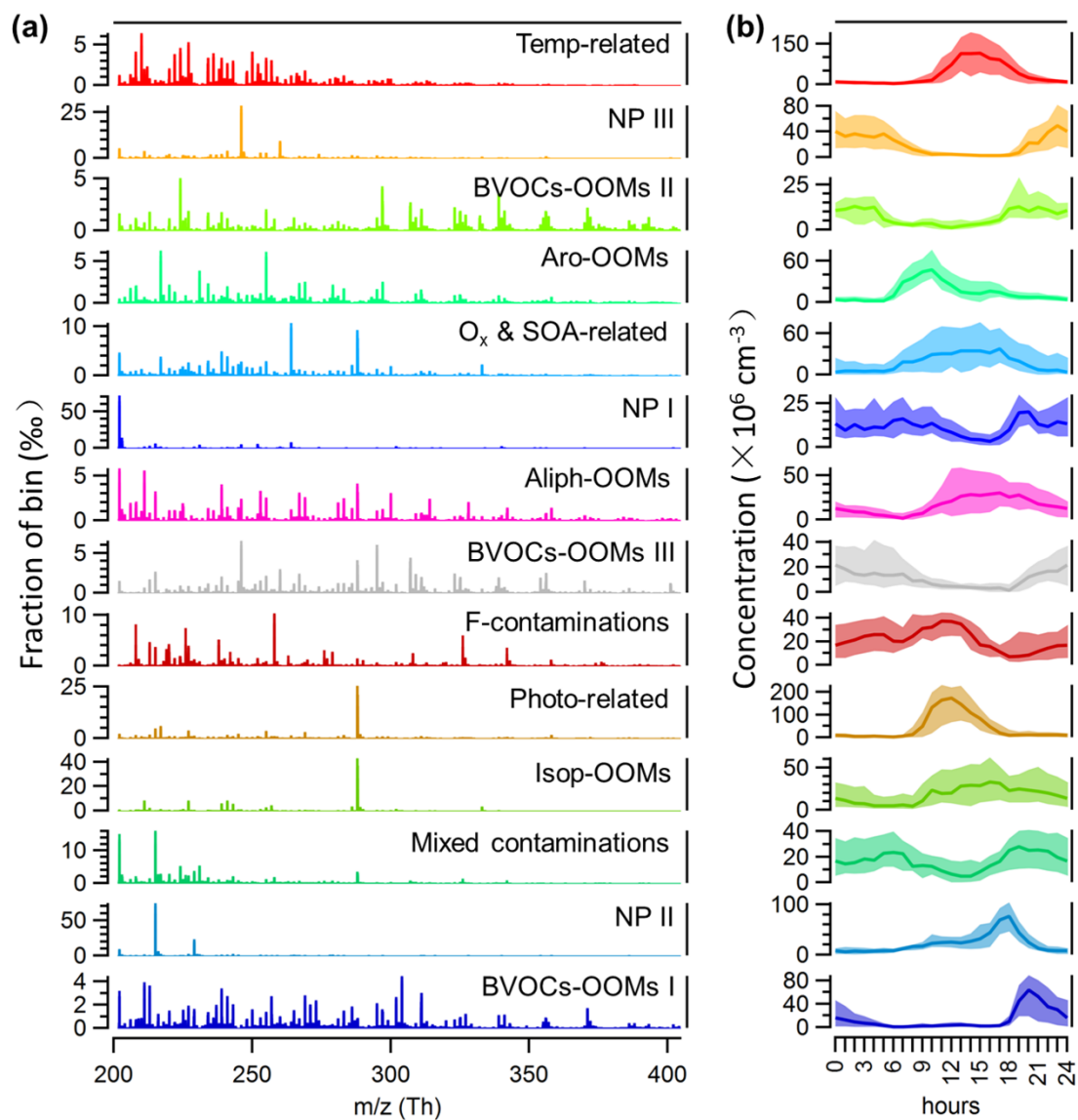
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834 The differences in OOMs observed in different environments are so clear and the
835 underlying causes are well worth thinking. The precursors, oxidants, and formation
836 pathways of OOMs are changing when moving from urbanized areas to pristine regions,
837 as AVOCs and NO_x concentrations decrease, and BVOCs concentrations increase. This
838 process can also occur under the trend of global warming and anthropogenic emissions
839 mitigation, but we still know very little about it. Clarifying the variations of
840 compositions, properties and formation efficiency of OOMs will help to understand the
841 evolution of SOA production during this process. In summary, our findings highlight
842 the dramatic interactions between anthropogenic and biogenic emissions, and
843 encourage more investigations from a mechanistic point of view.

844

845 **Appendix A. The selected solution of binPMF analysis on nitrate CI-APi-TOF data**

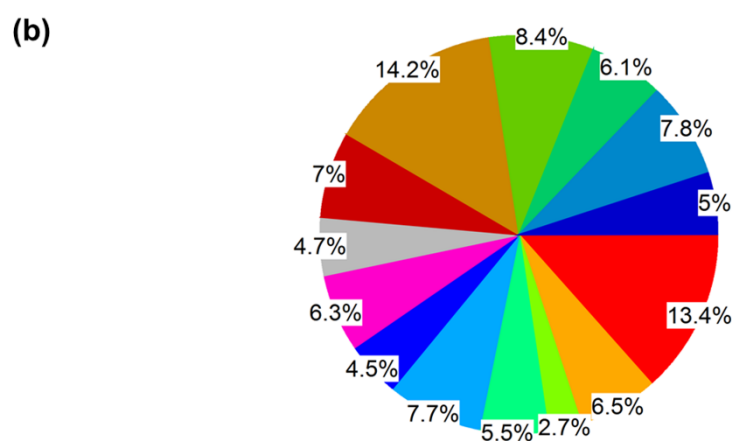
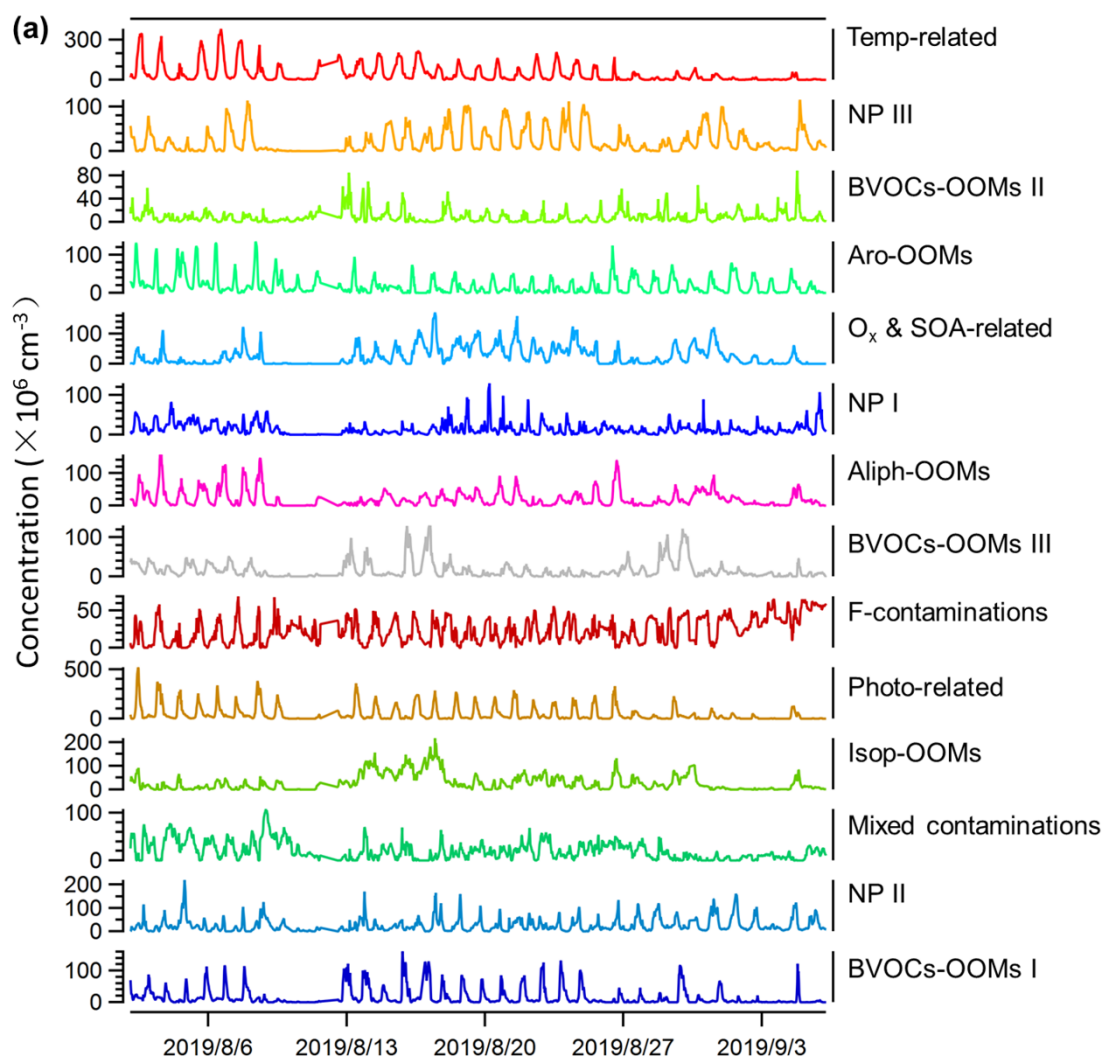
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847

848 Figure A1. The selected solution for binPMF analysis of nitrate CI-APi-TOF data,
 849 showing (a) mass profile and (b) diurnal cycle of different factor.

850



851

852 Figure A2. The selected solution for binPMF analysis of nitrate CI-API-TOF, showing
 853 (a) time series of and (b) contribution to total signal reconstructed by PMF of each
 854 factor.

855

856 **Data availability.** Measurement data at the SORPES station, including OOMs data and
 857 relevant trace gases and aerosol data as well as meteorological data, are available upon

858 request from the corresponding author before the SORPES database is open to the
859 public.

860

861 **Author contributions.** WN and AD designed this research. YL, YL, DG, CL, ZX, LC,
862 TW, LW, PS, XQ, JW and XC conducted the measurements. YL, WN, CY, YZ, DH,
863 ZW, and DW analyzed the data. YL, WN, ND, ME and AD wrote the paper.

864

865 **Competing interests.** The authors declare that they have no conflict of interest.

866

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871

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879 Reference

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