

Secondary Organic Aerosol Formation from the Oxidation of Decamethylcyclopentasiloxane at Atmospherically Relevant OH Concentrations

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Abstract. Decamethylcyclopentasiloxane (D5, C₁₀H₃₀O₅Si₅) is measured at ppt levels outdoors and ppb levels indoors. Primarily used in personal care products, its outdoor concentration is correlated to population density. Since understanding the aerosol formation potential of volatile chemical products is critical to understanding particulate matter in urban areas, the secondary organic aerosol yield of D5 was studied under a wide range of OH concentrations, OH exposures, NO_x concentrations, and temperatures. ~~The and, correspondingly, OH exposures using both batch-mode chamber and continuously run flow tube experiments. These results were comprehensively analyzed and compared to two other secondary organic aerosol (SOA) yield from the datasets from literature. It was found that the SOA yield from the~~ oxidation of D5 is extremely dependent on either the OH concentration, and differing measurements of the SOA yield from the literature are resolved in this study. Here, we compare experimental results from environmental chambers and flow tube reactors. Generally, there are high SOA yields (> 68%) at OH mixing ratios of 5 × 10⁹ or exposure. For OH concentrations of ≤ 10⁷ molec cm⁻³. At atmospherically relevant OH concentrations or OH exposures of ≤ 2 × 10¹¹ molec s cm⁻³, the SOA yield is largely <5% and usually ~1%. This is significantly lower than SOA yields ~~used in emission and particulate matter inventories and demonstrates the necessity of OH concentrations similar to the ambient environment when extrapolating SOA yield data to the outdoor atmosphere. previously reported. Using a two-product absorptive partitioning model for the upper-bound SOA yields, the stoichiometric mass fraction and absorptive partitioning coefficients are, for the first product, α₁ = 0.056 and K_{OM,1} = 0.022 m³ μg⁻¹; for the second product, they are α₂ = 7.7 and K_{OM,2} = 4.3 × 10⁻⁵ m³ μg⁻¹.~~ Generally, there are high SOA yields (> 90%) at OH mixing ratios of 5 × 10⁹ molec cm⁻³ or OH exposures above 10¹² molec s cm⁻³.

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1 Introduction

20 Present in outdoor mixing ratios as high as ~40 ppt, decamethylcyclopentasiloxane (D5, $C_{10}H_{30}O_5Si_5$) has been observed in cities, rural areas, and the Arctic (Buser et al., 2013, 2014; Ahrens et al., 2014; McLachlan et al., 2010; Xu et al., 2019). D5 is used in personal care products, as well as for industrial purposes (Mackay et al., 2015); in 2004, over 17000 tons were used in the then European Union (Safron et al., 2015). Outdoor observations of D5 are population-dependent (Janecek et al., 2017; Gkatzelis et al., 2021) and this dependence is sufficiently reliable ~~to be used as a tracer to differentiate the effects of population~~
25 ~~from that of motor vehicles~~ that it can be used to tease out personal care product emission patterns from other common urban emissions (Coggon et al., 2018). The impact of D5 does not stop at population centers; its long atmospheric lifetime means that it is even found in areas with low population densities.

Likely more than 90% of the D5 used is emitted into the atmosphere (Balducci et al., 2012; Hughes et al., 2012), though much of this may be first emitted indoors and only later exchanged to the outdoors: in an engineering classroom in the U.S. in
30 2014, ~30% by mass of the total volatile organic compounds (VOCs) were D5 (Tang et al., 2015). In an athletic center in the morning, D5 mixing ratios exceeded 6 ppb and emissions were attributed to the humans in the room (Finewax et al., 2020). Even the international space station ~~was found to contain~~ contains trace amounts of D5 in the air (Carter et al., 2015).

Given the abundance of D5 in the ambient atmosphere, it is important to understand its fate. The major loss source of D5 is reaction with the hydroxyl radical; losses by reaction with NO_3 , ~~O_3 , and Cl~~ are all and O_3 are negligible (Atkinson,
35 1991; Alton and Browne, 2020). Global losses by reaction with Cl are less than 5%, though can be higher in polluted areas (Alton and Browne, 2020). The half-life of D5 outdoors is between 3.5 and 7 days, depending on the assumed global average OH concentration and the exact method used to ~~calculate~~ determine the reaction rate of OH with D5 (Safron et al., 2015; Xiao et al., 2015; Alton and Browne, 2020). Outside, both wet and dry deposition of D5 are negligible and methylsiloxanes do not photolyze in the actinic region (~~Hobson et al., 1997). Previous studies by Janecek et al. (2019) and Wu and Johnston (2017)~~
40 ~~measured~~ (Janecek et al., 2017; Hobson et al., 1997).

Chandramouli and Kamens (2001) detected a single product from OH oxidation with D5, 1-hydroxynonamethylcyclopentasiloxane. A study of 1-hydroxynonamethylcyclopentasiloxane partitioning found a large temperature, aerosol seed type, and humidity dependence (Latimer et al., 1998). Though, an investigation by Wu and Johnston (2017) found both dimer products and ring-opened species from the OH oxidation of D5. SOA morphology from all D5 oxidation products was investigated by Janecek et al. (2019).

45 Previous studies investigating the secondary organic aerosol (SOA) yields ~~of D5 oxidation with OH found values between 8 and 50% (Janecek et al., 2019; Wu and Johnston, 2017). SOA yields are~~ the ratios of the mass of organic aerosol formed to the mass of the precursor reacted, ~~between 8 and 50%~~. This is a wide range, and the conditions for these experiments were performed at OH concentrations much higher than those in the ambient atmosphere. By measuring the SOA formation potential
50 of D5, we can better understand the contribution of volatile chemical products (VCPs) to aerosol levels in urban areas.

VCPs are a major (and perhaps majority) source of secondary organic aerosol in ~~cities in the some~~ U.S. ~~, even urban areas that are not megacities~~ cities (McDonald et al., 2018; Gkatzelis et al., 2021). Resolving uncertainties in the mass of SOA

formed from VCPs is critical for refining SOA estimates and for creating policy to reduce SOA non-compliance in urban areas (Burkholder et al., 2017).

55 ~~Researchers use both flow reactors~~For many years, researchers have used flow reactors, usually run at steady-state, and atmospheric chambersto measure, usually run in batch mode, to understand SOA formation and the SOA yields of various compounds (Bruns et al., 2015). Batch mode is where all reactants are added before oxidation and, during each experiment, the evolution of the reactor's contents are tracked in time. While many results agree between the two methods of analysis, different reactors have varying benefits and operating conditions (e.g., OH concentrations, experiment length, precursor concentrations, humidity values). One must account for the particular attributes of the different reactors when extrapolating to the atmosphere.

The chamber experiments conducted in this study were performed for multiple hours at OH concentrations representative of what is found in the atmosphere. Since these experiments are time-limited, and D5's outdoor half-life is multiple days, the chamber experiments tend to have OH exposures slightly less than what is representative of ambient conditions. OH exposure is the quantity of OH concentration over time and is a measure of atmospheric aging (Renbaum and Smith, 2011).

60 To understand SOA yields at higher OH concentrations, a flow reactor was run at steady-state. The residence time of this reactor was short (on the order of minutes), but the OH mixing ratios were higher than used in the chamber experiments.

It is well established that OH concentration and exposure are not necessarily interchangeable (Renbaum and Smith 2011; Liu et al. 2011; McNeill et al. 2008). Additionally, there is precedent for studying an overlapping range of OH exposure using both environmental chambers and flow reactors. For example, Lambe et al. (2015) showed that oxidation experiments over a range of OH exposures can be comparable between both types of reactors.

We start with a discussion of results from these two reactors, which show agreement when either the OH concentrations or exposures overlap. Then, we provide two-product absorptive partitioning parameters and fits for the data collected here. We close by comparing these results to other SOA yield studies in the literature: that of Wu and Johnston (2017) and Janecek et al. (2019).

75 **2 Methods**

Chamber experiments (C1–8) were performed in a temperature-controlled 19 m³ FEP Teflon Environmental Chamber run in batch mode. The ~~chamber is~~uncertainty associated with the reported chamber temperature is < 0.5°C. The chamber is hung in an enclosure, to reduce charge on the surface of the chamber, and is surrounded by ultraviolet lights centered at ~350 nm. Since the walls of the chamber are not rigid and data were collected continuously, the chamber decreased slightly in volume throughout the experiment, but never by more than 15%.

Prior to each chamber experiment, the contents of the chamber were flushed with air stripped of ozone, nitrogen oxides, water vapor, and organic carbon for > 24 h. ~~H₂O₂~~Hydrogen peroxide (H₂O₂), when used as an OH source (C1–7), was injected by flowing air at 5 Lpm over liquid H₂O₂ in a ~42°C water bath to obtain an [H₂O₂] ≈ 2 ppm. For ~~the experiment that used experiment C8, an evacuated glass bulb was filled with~~ methyl nitrite (CH₃ONO) ~~, a glass bulb was evacuated and then filled to the desired pressure to obtain~~ and diluted with nitrogen. The bulb was then flushed into the chamber with nitrogen to obtain

~~a mixing ratio of ~600 ppb in the chamber. After bringing the bulb up to atmospheric pressure with nitrogen, it was flushed into the chamber with nitrogen.~~ CH₃ONO forms OH as described in Schwantes et al. (2019).

D5 (99%, TCI America) was injected into the chamber for ~~Experiments 1–8~~ experiments C1–8 at room temperature by flowing nitrogen through a glass bulb at 5 Lpm for > 60 min. To obtain the desired initial surface area concentration, a sonicated, 90 0.06 M (0.15 M for Experiment 2C2) (NH₄)₂SO₄ solution was atomized to create aerosol that was then dried, passed through a TSI Model 3088 soft x-ray neutralizer, and injected into the chamber. For Experiment 7C7, no aerosol was injected. For Experiments 5–7 experiments C5–7, NO (506.9 ppm ± 2%, Airgas Specialty Gases, Certified Standard) was injected prior to the beginning of the experiment to achieve initial NO mixing ratios between 80 and 100 ppb. ~~During Experiments~~ The estimated uncertainty of [NO₀] is 5 and 6 ppb. For Experiments 5–7 During C5 and C6, 1 ppb ~~/min~~ min⁻¹ of NO was injected 95 from the inception of radiation to the end of the experiment. All experiments began with [NO₂]₀ = 0 ppb.

Experiments at higher OH mixing ratios were conducted in the Caltech Photooxidation Flow Tube (CPOT, Huang et al., 2017) at steady-state and a constant flow rate of 4.88 Lpm and 23.0±0.1°C. The mean residence time of the CPOT was 671 ± 15 s and the diffusivity was 15 ± 2 cm² s⁻¹, as calculated with a step injection of SO₂ using Equation 4 in Huang and Seinfeld (2019). For Experiments 9–15 experiments F9–15, clean air flowed through an ozone generator (UVP, 97-0067-01); 100 for Experiments 16–19 F16–19, O₂ flowed through the same generator to create higher concentrations of O₃. The 254 nm lights photolyze O₃ to form O(¹D), which reacts with H₂O to form 2OH. After conditions were changed in the CPOT, no results were collected for at least 2 h. Data were averaged over between 1 and 11 h. D5 was injected through a syringe pump (Harvard Apparatus).

For all experiments, the concentration of D5 was measured with an HP 6890N gas chromatograph with a flame ionization 105 detector (GC-FID) and a DB-5 column. Prior to the beginning of oxidation for the chamber experiments, all contents of the reactor were left to sit for 4 h (2.8 h for Experiment 7C7) and the initial concentration of D5 was taken as the mean concentration during this time. For the CPOT experiments, the initial concentration of D5 was calculated by measuring the outlet flow with lights off, no water source, and the absence of O₃. For Experiment 9F9, the change in D5 was sufficiently small that it was within the uncertainty. For calculating the SOA yield for this experiment, we used the OH exposure calculated from the change 110 in SO₂ concentration to find the change in D5 (7 ppb).

To calibrate the GC-FID, a small Teflon bag was filled with 35 ppm of D5 and later diluted to 9 ppm. This bag was sampled using the GC-FID, and the concentration was verified with a Fourier transform infrared absorption (FT-IR) spectrometer with a 19 cm path length and absorption cross sections from the Pacific Northwest National Laboratory (PNNL) database. To minimize vapor-wall-loss to the FT-IR enclosure, multiple samples were taken until a consistent spectrum was achieved.

115 Gas-phase oxidation products were evaluated with a CF₃O⁻ chemical ionization mass spectrometer (CIMS) equipped with a Varian 1200 triple quadrupole mass analyzer. Concentrations of NO and NO₂ were measured with a Teledyne Nitrogen Oxide Analyzer (Model T200) and O₃ was found with a Horiba Ambient Monitor. Temperature and humidity were determined using a Vaisala HMM211 probe.

Aerosol volume was measured by a custom-built scanning mobility particle sizer (SMPS) with a 3081 TSI Differential 120 Mobility Analyzer (DMA) and a TSI 3010 butanol condensation particle counter (CPC). The sheath flow rate was 2.64 Lpm

and the aerosol flow rate from the chamber was 0.515 Lpm. A voltage scan from 15 to 9875 V was performed in 240 s every 330 s. Aerosol from the chamber flowed through an x-ray source to provide a known charge distribution, and the size distributions were determined using the data inversion method described by Mai et al. (2018). Experiment ~~2-C2~~ required a logarithmic fit to the largest particles present, as described in Charan et al. (2020), which is the source of the higher SOA yield uncertainty than in the other experiments (see Table 1). Conversions to mass concentration were performed by assuming that the aerosol density was ~~1.1-1.52 ± 0.04 g cm⁻³. This was estimated from data in Wu and Johnston (2017), which measured D5 secondary organic aerosol particles (using information in their Fig. S1a and Table S1).~~, which was the density calculated at [OH] ≈ 9.4 × 10⁹ molec cm⁻³ in a flow reactor (Xu and Collins, 2021) using an Aerosol Particle Mass Analyzer and a SMPS system as described in Malloy et al. (2009).

Uncertainty estimates for all the instruments used in this study were determined as described in Charan et al. (2020). For the chamber experiments, particle-wall-deposition corrections were performed by calculating a diameter-independent first-order exponential fit ($\beta = 1-7 \times 10^{-4} \text{ min}^{-1}$) to the particle volume concentration during the 3 h prior to the onset of oxidation and applying that correction to the rest of the experiment. This method was chosen because it aligns with a diameter-dependent fit as determined using the method in Charan et al. (2018) but is simpler and because, for the chamber experiments, minimal organic aerosol formed and ~~so~~, so, the particle diameters changed insignificantly throughout the duration of the experiment. For ~~Experiment 7~~experiment C7, in which no initial aerosol was present, no aerosol was generated throughout the experiment and so no correction was necessary to determine an apparent SOA yield of 0.

For the CPOT experiments, an upper estimate of the wall-deposition-corrected SOA mass was calculated with the inverse of the particle-size-dependent penetration efficiency of the flow-tube component of the reactor (data from Fig. 9d in Huang et al. 2017). Since particles nucleated in the CPOT, the penetration efficiency of only the flow-tube component (and not the static mixer prior to the region of reaction) was used. The penetration efficiency, however, is based on the entire flow tube and nucleated particles may not form immediately at the beginning of the flow-tube component; thus, the wall-deposition correction performed here is an upper bound of the correction. Note that this correction also neglects particle growth throughout the reactor and any particle-particle coagulation. Diagrams of the CPOT and its static mixer are in Huang et al. (2017).

SOA yield (Y) is defined by $Y = \frac{\Delta \text{SOA}_{\text{corr}}}{\Delta \text{D5}}$, where $\Delta \text{SOA}_{\text{corr}}$ is the wall-deposition-corrected change in the aerosol mass concentration and ΔD5 is the mass concentration of reacted D5. Calculations were performed as described by Charan et al. (2020) and with the assumption that a particle, once deposited on the reactor wall, no longer acts as a condensation sink (Trump et al., 2016). Note that since so little aerosol was formed during the chamber experiments, this assumption had a negligible effect on the chamber results. For the CPOT experiments, any deviation from this assumption would have prevented the ~~data~~ reactor from reaching steady-state.

Table 1 gives the upper estimate of the wall-deposition-corrected SOA yield for the CPOT experiments. The lower bound of the SOA yields of the CPOT experiments is the reported uncertainty subtracted from the uncorrected Y and the upper bound is the sum of the corrected Y and the reported uncertainty.

While the vapor-wall-deposition lifetime of D5 to the chamber walls was estimated to be on the order of weeks, the propensity of vapor-wall-deposition of the oxidation products is not extensively investigated in this study. Even at high initial seed

Table 1. Experimental conditions. All experiments began with $[\text{NO}_2]_0 = 0$ ppb. The estimated uncertainty of $[\text{NO}]_0$ is 5 ppb. Experiments in which NO was continuously injected maintained a rate of 1 ppb min^{-1} , starting at the inception of radiation. For Experiment 8, due to a pulse of OH when the CH_3NO_2 initially photolyzes followed by a steadier concentration, two OH mixing ratios are given: $[\text{OH}] = 2 \times 10^8 \text{ molec cm}^{-3}$ at the beginning of the experiment and $[\text{OH}] = 1 \times 10^6 \text{ molec cm}^{-3}$ at its end. The uncertainties associated with the chamber and CPOT temperature are $< 0.5^\circ\text{C}$ and 0.1°C , respectively. The uncertainty in the OH concentrations for the chamber and CPOT experiments are $5 \times 10^5 \text{ molec cm}^{-3}$ and $1 \times 10^6 \text{ molec cm}^{-3}$, respectively. Uncertainties in the OH exposure for the CPOT experiments are $5 \times 10^{10} \text{ molec s cm}^{-3}$. For the chamber experiments, the uncertainties in the negative direction are $4 \times 10^{10} \text{ molec cm}^{-3}$ and in the positive direction are the value of OH exposure reported for the experiment. For the CPOT experiments, the upper estimate of the wall-deposition-corrected SOA yield is shown in parentheses. The negative uncertainty in the SOA yields of the CPOT experiments is the reported uncertainty subtracted from the uncorrected Y and the positive uncertainty is the sum of the corrected Y and the reported uncertainty.

Label	Reactor	$[\text{NO}]_0$ (ppb)	Contin. NO Injection?	OH source	$[\text{OH}]$ (molec cm^{-3})	OH exposure (molec s cm^{-3})	T ($^\circ\text{C}$)	RH (%)	$[\text{D5}]_0$ (ppb)	$1 - \frac{[\text{D5}]_{\text{end}}}{[\text{D5}]_0}$	$[\text{Surf Area}]_0$ ($10^3 \mu\text{m}^2 \text{ cm}^{-3}$)	SOA Yield
<u>C1</u>	Chamber	0	No	H_2O_2	4.5×10^6	9×10^{10}	26.6	<u><5</u>	497 ± 5	0.18	3.6 ± 0.3	<u>$1.5 \pm 1.5\%$</u>
<u>C2</u>	Chamber	0	No	H_2O_2	3.8×10^6	8×10^{10}	26.5	<u><5</u>	298 ± 3	0.17	5.1 ± 0.3	<u>$5.7 \pm 8.0\%$</u>
<u>C3</u>	Chamber	0	No	H_2O_2	2.2×10^6	6×10^{10}	27.6	<u><5</u>	30 ± 1	0.09	0.8 ± 0.1	<u>$0 \pm 0.3\%$</u>
<u>C4</u>	Chamber	0	No	H_2O_2	1.6×10^6	3×10^{10}	17.7	<u><5</u>	580 ± 5	0.08	2.4 ± 0.2	<u>$2.6 \pm 4.0\%$</u>
<u>C5</u>	Chamber	82	Yes	H_2O_2	5.0×10^6	1.3×10^{11}	26.6	<u><5</u>	696 ± 9	0.22	3.9 ± 0.3	<u>$0.7 \pm 0.7\%$</u>
<u>C6</u>	Chamber	86	Yes	H_2O_2	4.3×10^6	9×10^{10}	26.6	<u><5</u>	650 ± 6	0.18	1.1 ± 0.1	<u>$0.2 \pm 0.3\%$</u>
<u>C7</u>	Chamber	76	No	H_2O_2	5.5×10^6	1.3×10^{11}	23.6	<u><5</u>	591 ± 2	0.24	0	<u>$0 \pm 0.1\%$</u>
<u>C8</u>	Chamber	84	No	CH_3NO_2	$10^6 - 10^{8.3}$	2.3×10^{11}	26.6	<u><5</u>	603 ± 5	0.38	0.6 ± 0.1	<u>$0 \pm 0.1\%$</u>
<u>F9</u>	CPOT	0	No	O_3	2.0×10^7	1.4×10^{10}	23.0	<u>2</u>	262 ± 10	0.03	0	<u>$1.1(1.9) \pm 1.1\%$</u>
<u>F10</u>	CPOT	0	No	O_3	2.3×10^8	1.5×10^{11}	23.0	<u>3</u>	262 ± 10	0.26	0	<u>$1.8(2.9) \pm 0.2\%$</u>
<u>F11</u>	CPOT	0	No	O_3	5.0×10^8	3.3×10^{11}	23.0	<u>4</u>	262 ± 10	0.28	0	<u>$6.0(9.2) \pm 0.6\%$</u>
<u>F12</u>	CPOT	0	No	O_3	2.3×10^8	1.5×10^{11}	23.0	<u>3</u>	262 ± 10	0.38	0	<u>$4.6(6.7) \pm 0.4\%$</u>
<u>F13</u>	CPOT	0	No	O_3	1.2×10^9	7.8×10^{11}	23.0	<u>10</u>	262 ± 10	0.60	0	<u>$14(19) \pm 1\%$</u>
<u>F14</u>	CPOT	0	No	O_3	1.5×10^9	1.0×10^{12}	23.0	<u>16</u>	262 ± 10	0.67	0	<u>$24(32) \pm 2\%$</u>
<u>F15</u>	CPOT	0	No	O_3	1.6×10^9	1.1×10^{12}	23.0	<u>33</u>	262 ± 10	0.71	0	<u>$35(49) \pm 2\%$</u>
<u>F16</u>	CPOT	0	No	O_3	4.7×10^9	3.2×10^{12}	23.0	<u>25</u>	246 ± 3	0.97	0	<u>$109(157) \pm 7\%$</u>
<u>F17</u>	CPOT	0	No	O_3	4.8×10^9	3.2×10^{12}	23.0	<u>30</u>	246 ± 3	0.98	0	<u>$110(158) \pm 7\%$</u>
<u>F18</u>	CPOT	0	No	O_3	4.7×10^9	3.1×10^{12}	23.0	<u>23</u>	82 ± 3	1.00	0	<u>$102(138) \pm 6\%$</u>
<u>F19</u>	CPOT	0	No	O_3	4.9×10^9	3.3×10^{12}	23.0	<u>33</u>	82 ± 3	1.00	0	<u>$94(128) \pm 6\%$</u>

surface area concentrations, the SOA yield is still quite small (see Fig. S1). Alton and Browne (2020) estimated that, for their unseeded $\sim 1 \text{ m}^3$ FEP Teflon chamber, 5% of the ester product of D5 oxidation might partition to the chamber walls during the reaction. The volume of the chamber used in this study is 19 m^3 and seed aerosol is introduced prior to the experiment (except for Experiment 7, which was performed in the absence of seed aerosol). Even though other products may have higher wall-loss rates, if 5% of the oxidation products were lost to the chamber walls in C1–8, the SOA yields would still be within the reported uncertainty and sufficiently small so as not to affect any conclusions, smaller than expected previously. The CPOT reactor is operated at steady-state (see Fig. S2) and, therefore, any oxidation products that are sufficiently high volatility and in equilibrium with the bulk flow (i.e., not lost permanently to the quartz walls) do not need a vapor-wall-deposition correction. While

165 very low-volatility compounds may be lost to the reactor's walls, as was seen by Krechmer et al. (2020) in a continuously run Teflon reactor, we do not expect there to be significant irreversible gas-phase wall loss of siloxanes or their oxidation products in this reactor. Note that Wu and Johnston (2017) did see higher SOA yields in seeded experiments in their steady-state PFA Teflon reactor than in unseeded ones, indicating that there may be some irreversible loss, even when run continuously.

For chamber experiments that employed H₂O₂, the OH concentration was calculated by fitting the gas-phase D5 concentration to a first-order exponential, fixing the initial point of the fit as the initial D5 concentration (fits had R² > 0.75), and using
170 the value for the reaction rate of OH with D5, $k = 2.1 \pm 0.1 \times 10^{-12} \text{ cm}^3 \text{ molec}^{-1} \text{ s}^{-1}$, which was measured using the relative rate method at $297 \pm 3 \text{ K}$ (Alton and Browne, 2020). Note that other experimental evaluations of the reaction rate of OH with D5 that use the relative rates method vary by less than a factor of 2 (the reasons for this difference are not known), which would not affect the order of magnitude of the OH concentration estimate (Kim and Xu, 2017; Safron et al., 2015; Xiao et al., 2015). OH is the major loss source in the atmosphere and, we expect, in these experiments: ~~losses to O₃, NO₃, and Cl are all~~
175 ~~negligible~~; note that there was no Cl in these experiments (Atkinson, 1991; Alton and Browne, 2020). The ozone concentration did not affect the SOA yield results: ~~Experiments 7 and 9 C7 and F9~~, which were performed at substantially different O₃ concentrations, still gave similar results for the SOA yield ($0 \pm 0.1\%$ and $0.8 \pm 0.8\%$ with an upper wall-deposition-corrected bound of 1.4%, respectively). For ~~Experiment 8, experiments performed with NO_x (Appendix A), there was no observed NO_x dependence.~~

180 For C8, in which CH₃NO₂ served as the OH source, the sharp decrease in the D5 mixing ratio immediately after the commencement of radiation, followed by a gradual decrease in its concentration, indicates that two OH concentrations ~~are~~ were relevant for this experiment ([OH]= $2 \times 10^8 \text{ molec cm}^{-3}$ at the beginning of the experiment and [OH]= $1 \times 10^6 \text{ molec cm}^{-3}$ at its end). Since the D5 concentration ~~is was~~ measured every ~21 min, and the pulse with high OH concentrations ~~occurs~~ occurred within the first 30 min of oxidation, the initial OH concentration is estimated with a two-point first-order exponential
185 fit to the initial concentration and the first data point (12.3 min into radiation). The second OH concentration is estimated with a first-order exponential fit of the second point (33.3 min into radiation) to the end of the experiment.

OH exposure was calculated, for chamber experiments (~~Experiments 1–8 C1–8~~) and experiments from Wu and Johnston (2017), as

$$[\text{OH}] * t = \frac{-1}{k_{\text{OH}+\text{D5}}} \ln \left(\frac{[\text{D5}]_{\text{end}}}{[\text{D5}]_0} \right), \quad (1)$$

190 where $k_{\text{OH}+\text{D5}} = 2.1 \times 10^{-12} \text{ cm}^3 \text{ molec}^{-1} \text{ s}^{-1}$ (Alton and Browne, 2020). For the CPOT experiments (~~Experiments 9–19 F9–19~~), OH exposure was calculated as

$$[\text{OH}] * t = \frac{-1}{k_{\text{OH}+\text{SO}_2}} \ln \left(\frac{[\text{SO}_2]_{\text{end}}}{[\text{SO}_2]_0} \right), \quad (2)$$

where $k_{\text{OH}+\text{SO}_2} = 9 \times 10^{-13} \text{ cm}^3 \text{ molec}^{-1} \text{ s}^{-1}$ for ~~an identical a CPOT~~ an identical a CPOT setup with SO₂ instead of D5 (~~Janecek et al., 2019~~) and otherwise identical flows and conditions, as in the method described by Janecek et al. (2019). Uncertainties in the OH exposure for the CPOT experiments are $5 \times 10^{10} \text{ molec s cm}^{-3}$. The correlation between [H₂O] and OH exposure used to find
195 the OH exposure for ~~Experiments 9–19 experiments F9–19~~ Experiments 9–19 experiments F9–19 is plotted in Fig. ~~S2S3~~. Since the O₃ concentrations differed in

Experiments 9–15 and 16–19 (F9–15 and F16–19), the correlation between the [H₂O] and the OH exposure is also different. Note that, for Experiments 9–19 (F9–19), the OH exposure calculated using Equation 2 is about twice that calculated using Equation 1. This effect may be because of a lower OH concentration in the D5 system due to the regeneration of OH during the oxidation of reaction of OH with the later D5 or the absorption of OH into the aerosol particles, oxidation products or the aerosol surfaces. Ideally, we seek to report the OH exposure excluding any regeneration. So, for Experiments 9–19 (F9–19), the OH exposure is calculated with Equation 2. Experiments 1–8 (C1–8) and those from Wu and Johnston (2017) use Equation 1, but have a positive uncertainty equal to their calculated OH exposure. They have a negative uncertainty of 4×10^{10} molec s cm⁻³. This does not include the potential additional uncertainty in the value of k_{OH+D5} as calculated using similar methods in different laboratories. Note that, for Experiments 1–8 (C1–8), OH concentration is calculated independently of the OH exposure. For Experiments 9–19 (F9–19), OH concentration is the ratio of the OH exposure to the residence time of the reactor. The uncertainty in the OH concentrations for the chamber and CPOT experiments are 5×10^5 molec cm⁻³ and 1×10^6 molec cm⁻³, respectively.

In order to determine the absorptive partitioning coefficients, C1-3, C5-8, and F9-17 were fit to a two-product model (Odum et al., 1996):

$$Y = M \left(\frac{\alpha_1 K_{OM,1}}{1 + M K_{OM,1}} + \frac{\alpha_2 K_{OM,2}}{1 + M K_{OM,2}} \right), \quad (3)$$

where Y is the SOA yield; M is the organic aerosol mass concentration; α_1 and α_2 are the stoichiometric mass fractions of products 1 and 2, respectively; and $K_{OM,1}$ and $K_{OM,2}$ are the absorptive partitioning coefficients for products 1 and 2, respectively. Fits were performed for starting points that varied from 10^{-4} to 10^4 and from 10^{-10} m³ μg⁻¹ to 10^{10} m³ μg⁻¹ for the α s and K_{OM} s, respectively. The fits with the highest R² were chosen.

3 Results

SOA yields and experimental conditions are given in Table 1 with estimated uncertainties. These SOA yields vary from 0 to 79% (114% at the upper bound of the wall-deposition-corrected value), an even wider range than that reported by the literature of 8–50% (Janecek et al., 2019; Wu and Johnston, 2017).

Between the experiments performed here and those in the literature, the OH concentrations and OH exposures vary widely. Determining which of these is the relevant parameter is critical to extrapolating the SOA yield data to the atmosphere: environmentally relevant OH concentrations are on the order of 10^6 molec cm⁻³, but since D5 is primarily lost to OH and has a half life of 3.5–7 days, OH exposures on the order of 10^{12} molec s cm⁻³ are also relevant. Due to experimental limitations, in particular an inability to perform experiments for multiple days without diluting the sample and otherwise changing the conditions, these two variables are often correlated.

Nonetheless, differentiating the effects of these two variables is possible. If a chemical process occurs in which the reaction of D5 and OH forms an intermediate or a second-generation product that then either reacts with OH or fragments, then the competition between the two outcomes is moderated by the relative time required for self-reaction or reaction with OH. This

means that, as the OH concentration increases, the OH reaction product will predominate. If this is the chemistry that D5
230 undergoes, then we would expect the SOA yield to depend solely on the OH concentration and not on the OH exposure.

Figure 1a shows the relationship between the measured OH concentrations and SOA yields for the experiments performed
here as well as those from the literature. There is very good agreement between the chamber and CPOT experiments for similar
OH concentration (shown in purple and orange, respectively). Moreover, the sharp increase in measured Y starting at $[\text{OH}]$
 $\approx 10^9 \text{ molec cm}^{-3}$ matches the hypothesis that there is a competitive process moderated by OH concentration.

235 3.1 Agreement between the CPOT and chamber experiments

The SOA yields from the CPOT and chamber experiments vary from 0 to 110% (158% at the upper bound of the wall-deposition-corrected
value), an even wider range than that reported by the literature of 8–50% (Table 1 shows all experimental conditions and
SOA yields). However, the measured SOA yields for these experiments (and those in the literature) correlate with the OH
concentrations and the OH exposures (Fig. 1).

240 All chamber experiments (C1–8) give small SOA yields. The only experiment of these with $Y > 3\%$ is C2, which had a large
uncertainty due to aerosol sizes outside the measurement range. It seems likely, then, that at OH concentrations $\lesssim 10^7 \text{ molec}$
 cm^{-3} , and OH exposures $\lesssim 2 \times 10^{11} \text{ molec s cm}^{-3}$, the SOA yield of D5 oxidation is $< 3\%$ and close to 1% or 0%.

At similar OH concentrations and OH exposures, the chamber and flow tube data agree. While F9–F19 had uniformly
higher OH mixing ratios than C1–7, experiment C8 had an $[\text{OH}]$ between F9 and F10–19 at its very beginning (though a
245 lower $[\text{OH}]$ after $\lesssim 30$). Since F9 and C1–8 have similarly low SOA yields, there appears to be agreement between the SOA
yields when viewed as a function of OH concentration. In Fig. 1a, which shows the relationship between the measured OH
concentrations and SOA yields for the experiments performed here as well as those from the literature, the chamber and CPOT
experiments coincide (shown in purple circles and orange triangles, respectively). The SOA yield increases above $\sim 5\%$ only at
 $[\text{OH}] > 5 \times 10^8 \text{ molec cm}^{-3}$.

250 There is also agreement between the chamber and the CPOT experiments when SOA yield is viewed as a function of OH
exposure, as is shown in Fig. 1b. Experiment F9 has the lowest OH exposure of any other experiments performed in this study,
and its low SOA yield matches that of C1–8. F10 and F12 have lower OH exposures than C8 and their measured SOA yields
match those of the chamber experiments. Only starting at OH exposures of $\geq 3 \times 10^{11} \text{ molec s cm}^{-3}$ does the SOA yield begin
increasing significantly.

255 If OH concentration or exposure were the strict determinant of the SOA yield, F16–17 should give the same SOA yields as
F18–19. These experiments do have similar SOA yields, and F16–18 are all within error. If there is a difference, it might be
attributed to a dependence on the organic aerosol mass concentration at high OH concentrations or exposures. Viewing SOA
yield as a function of organic aerosol mass concentration also gives agreement between the chamber and CPOT experiments,
as shown in Fig. 2.

260 Assuming an absorptive partitioning model (Odum et al., 1996), a two-product fit was performed for experiments C1–3,
C5–8, and F9–17 using Equation 3. This model assumes the same temperature for all experiments, which is why C4 was
excluded (conducted at 17.7°C). While the chamber and CPOT experiments also varied in temperature, all the rest were within

5°C of one another. The absorptive partitioning model also assumes that the parent compound is still present; so F18 and F19, where the precursor is completely consumed, were excluded from the fit. A single-product version of Equation 3 did not provide a good fit.

The two-product fit that included the non-wall-loss-corrected values for the CPOT experiments gave parameters: $\alpha_1 = 0.044$, $K_{OM,1} = 0.027 \text{ m}^3 \mu\text{g}^{-1}$, $\alpha_2 = 5.5$, and $K_{OM,2} = 6.0 \times 10^{-5} \text{ m}^3 \mu\text{g}^{-1}$. For the upper bound, which is a fit with the wall-loss-corrected values for experiments F9–17, $\alpha_1 = 0.056$, $K_{OM,1} = 0.022 \text{ m}^3 \mu\text{g}^{-1}$, $\alpha_2 = 7.7$, and $K_{OM,2} = 4.3 \times 10^{-5} \text{ m}^3 \mu\text{g}^{-1}$. Both of these fits are shown in Fig. 2; panel b shows the full range and panel c zooms in on the region with $M < 120 \mu\text{g m}^{-3}$.

Note that, as shown in Fig. 2a, the experiments with larger M similarly had larger OH exposures (and OH concentrations), since these were the experiments with higher SOA yields. Also, when all the D5 is consumed, the absorptive partitioning model no longer applies; in Fig. 2a, experiment F15 (triangle with $M=991 \mu\text{g m}^{-3}$) has a similar M as F18 and F19 (triangles with M of 1267 and 1175 $\mu\text{g m}^{-3}$, respectively), but they have very different SOA yields.

The major difference in F16–17 and F18–19 is the percent of D5 that reacted by the end of the experiment: 97% for F16, 98% for F17, and 100% for F18–19. Figure 3 shows the fraction reacted compared to the SOA yield for experiments performed in this study and those in the literature. The color axis in Fig. 3 shows the OH exposure corresponding to the fraction of D5 reacted. This fit could indicate that there are later generation oxidation products that form large amounts of aerosol and that the gas-phase reaction rate to form the low-volatility later-generation oxidation product is slower than the gas-phase reaction rate to form the first-generation product (Kroll and Seinfeld, 2008).

We do not expect that either relative humidity or temperature affect the SOA yield sufficiently that these would account for the vastly different measured SOA yields under different OH concentration or exposures. Experiments C1–8 were performed at relative humidity (RH) levels between 2 and 6%, experiments F9–19 were between 0 and 30% RH, those by Wu and Johnston (2017) were performed at 27°C and a RH of 8–10%, and the experiments from Janecek et al. (2019) were run at 24°C and an RH of 25% or 45%. At similar values of relative humidity but different OH concentrations (e.g., F9–12, which all have RH $\leq 6\%$), the OH concentration or exposure matters for determining the SOA yield. For Experiments C3 and C4, the lowest and highest temperatures studied here (17.7 and 27.6°C, respectively), the measured SOA Y varies by $< 2\%$ $< 3\%$, which is within the uncertainty.

The NO_x concentrations also do not seem to affect the SOA yield for OH concentrations $\sim 5 \times 10^6 \text{ molec cm}^{-3}$, as discussed in Appendix A. While the D5 oxidation chemistry may depend on the NO mixing ratio (but not on the NO_2 mixing ratio), this has no effect on the measured SOA yield for the chamber experiments.

3.2 Agreement with literature

Also plotted in Fig. 1a is the correlation between the OH and Y for Figs. 1–3 are the experiments performed by Wu and Johnston (2017). These results and Janecek et al. (2019). The results from Wu and Johnston (2017), which were used by McDonald et al. (2018) for evaluating the contributions of D5 to aerosol levels in the Los Angeles Basin, were performed in a 50 L PFA photooxidation chamber with reported OH concentrations of $\sim 10^8 \text{ molec cm}^{-3}$ (the error of which “was difficult to assess”) at 27°C and a relative humidity (RH) of 8–10%. Their data are neither vapor- nor wall-deposition corrected. For similar

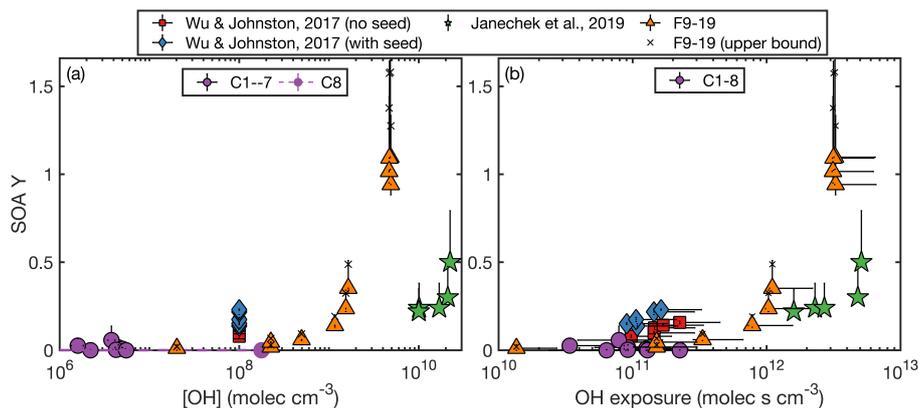


Figure 1. Measured SOA yield as a function of the (a) OH concentration and (b) OH exposure ~~normalized to the amount of reacted D5~~ for the experiments performed here and by Janecek et al. (2019) and Wu and Johnston (2017). In panel (a), ~~Experiment 8C8~~, which was performed with methyl nitrite, is shown in purple and not outlined in black and the initial and final OH concentrations are connected with a purple dashed line. The vast majority of this experiment was performed under the lower OH concentration. The same OH concentration is used for all experiments from Wu and Johnston (2017).

initial D5 concentrations (and, hence, for similar organic aerosol concentrations), the measured SOA yields were uniformly higher in experiments that were initiated with ammonium sulfate seed than those that were not (see Fig. 2b1). We, therefore, show the seeded and unseeded experiments in Figs. 1–3 as diamonds and squares, respectively. In Fig. 1a, ~~these points are, respectively,~~ blue and red, ~~respectively.~~

OH concentration in the experiments reported by Wu and Johnston (2017) were calculated by replacing the precursor with SO₂, measuring the formation of aerosol, and assuming that all the SO₂ ~~reacts-reacted~~ with OH to form H₂SO₄ and all the sulfuric acid ~~forms-formed~~ aerosol with minimal wall loss (Hall et al., 2013). Because of the uncertainties present for each step of this measurement, it seems reasonable that this [OH] estimate could be too low by at least a factor of 2. If the experiments from Wu and Johnston (2017) had OH concentrations more than twice as large, points in Fig. 1a would roughly agree.

Other instrumental and analysis uncertainties might close the gap between the OH concentrations measured by Wu and Johnston (2017) and the OH concentrations found in the experiments performed here. For example, the CPOT experiments and the Wu and Johnston (2017) experiments calculate the total OH exposure experienced in the flow reactor and then find the OH concentration by taking the ratio of this exposure and the residence time. Since the reactor used by Wu and Johnston (2017) is a rectangular bag, regions will exist with differing OH concentrations. If this reactor has slightly higher concentrations in some points or its residence time is overestimated or if the residence time for CPOT is a slight underestimate (we calculated an uncertainty of ~2%), this could account for the remaining disagreement between the data from the two experimental setups.

~~Furthermore, differences~~ We calculated the OH exposure in the experiments by Wu and Johnston (2017) in the same manner as for experiments C1–8, and there appears to be fairly good agreement between the SOA yields measured as a function of OH exposure (see Fig. 1b).

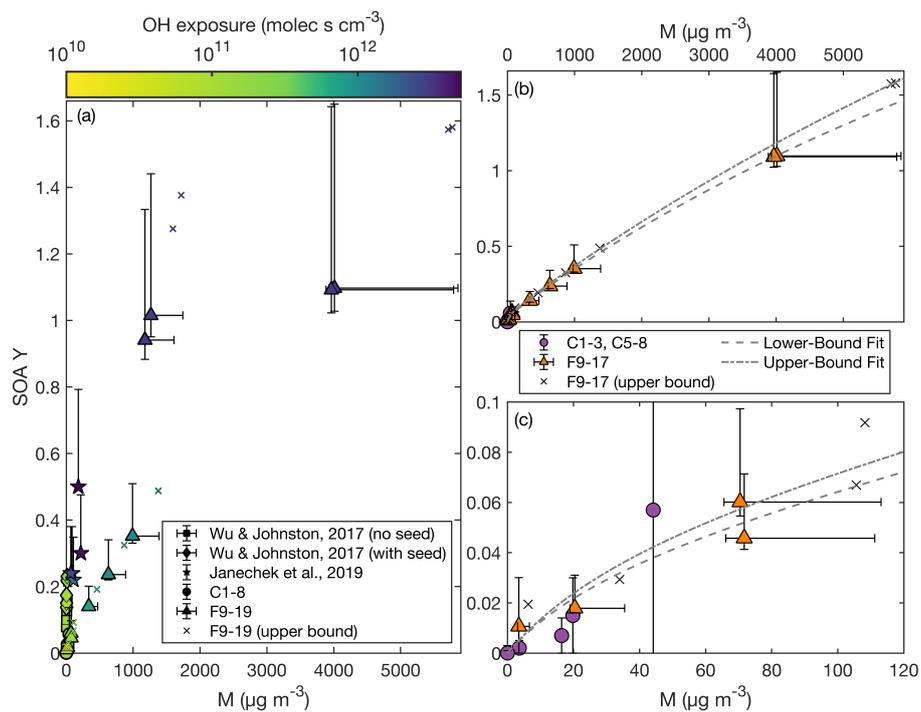


Figure 2. SOA yield as a function of organic aerosol mass concentration (M) as (a) compared to that reported by Wu and Johnston (2017) and Janecek et al. (2019). The color axis is the OH exposure concentration. Panel (b) shows the SOA yield for experiments C1–3, C5–8, the non-wall-deposition-corrected F9–17, and the wall-deposition-corrected F9–17. Also included are the fits of a two-product absorption partitioning model to C1–3, C5–8, and the non-wall-deposition-corrected F9–17 (dashed curve) and to C1–3, C5–8, and the wall-deposition-corrected F9–17 (dash and dotted curve). Panel (b) focuses on experiments with $M < 120 \mu\text{g m}^{-3}$. Panel (a) shows the entire scale and panel (b) focuses on experiments with $M < 34 \mu\text{g m}^{-3}$. Since the experiments reported by Wu and Johnston (2017) showed a dependence on the presence or absence of aerosol seed, those experiments are shown separately. CPOT experiments performed in this study were conducted without seed aerosol.

Differences in the analysis could change the relevant SOA yields calculated. For Experiments 10–19, which could cause better agreement between experiments sets when Y is viewed as a function of both OH concentration and exposure. For F10–19, we measured both the initial and the final D5 concentration and for Experiments 1–8 C1–8 we continuously measured the concentration. Wu and Johnston (2017) measured the initial concentration and calculated the SOA yield by using the [OH] to estimate the amount of reacted D5. If Wu and Johnston (2017) underestimated the [OH], they might have correspondingly overestimated Y because they would have assumed less D5 reacted than in actuality. To achieve agreement to experiments performed here, then, the [OH] concentration could be different by less than a factor of 2 because of due to these confounding variables.

We also assumed that the density of the SOA formed was $1.1\text{--}1.52 \text{ g cm}^{-3}$ and Wu and Johnston (2017) collected the aerosol onto filters and directly measured the mass formed. While we based our density estimate on, crudely, what was found in

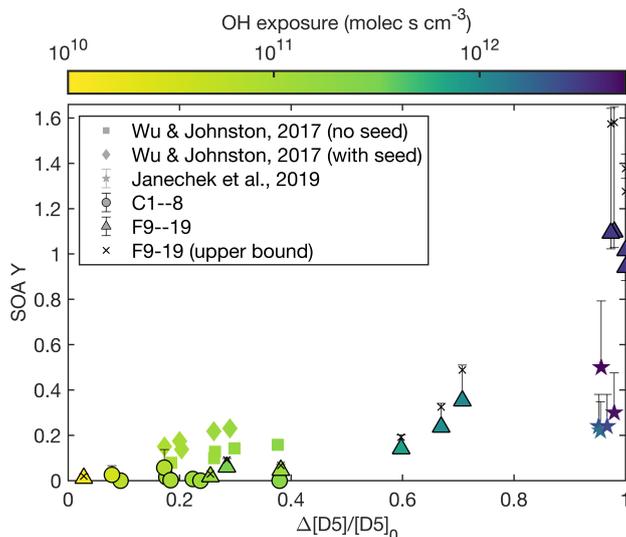


Figure 3. Measured SOA yield as a function of the fraction of D5 reacted at the end of the experiment. The color of each point indicates the OH exposure for the experiment. Experiments performed here are circles and triangles and are outlined in black, those by Janecek et al. (2019) are stars, and those by Wu and Johnston (2017) are squares and diamonds. The wall-deposition-corrected data for Experiments 9–19 are shown as black Xs.

Wu and Johnston (2017) (particularly, their Experiments 1 and 5), the aerosol density in their experiments could have been as much as 1.6 g cm^{-3} . Indeed, much secondary organic aerosol has a density of $1.4\text{--}1.6 \text{ g cm}^{-3}$, which would account for a significant portion (Kostenidou et al., 2007), but deviations from this range could account for some of the discrepancy (Kostenidou et al., 2007). Since between the sets of experiments. Additionally, since the CPOT experiments were unseeded,

330 seeded experiments increased the measured Y, which could also have led to better agreement.

Data from Janecek et al. (2019) show the opposite disagreement: OH concentrations and exposures are a factor of ~ 2 too large to perfectly match the results presented here. Janecek et al. (2019) performed their experiments in a 13.3 L potential aerosol mass oxidation flow reactor (PAM OFR) with OH concentrations on the order of $10^{10} \text{ molec cm}^{-3}$. They reported the total OH exposure and calculated it similarly to the method used for the CPOT (using Equation 2) and we convert this to the [OH] plotted by dividing this OH exposure by the residence time (calculated from the size of the reactor and the reported flow rate) and assuming that OH concentrations throughout the reactor are approximately constant. Just as with the CPOT experiments, their experiments are unseeded, and they measure the initial and final D5 concentration directly. They used an SOA density of 0.959 g cm^{-3} to calculate Y and the positive error bars shown are an adjustment of their SOA yields to the $1.1\text{--}1.52 \text{ g cm}^{-3}$ used in the experiments performed here.

340 While they corrected for particle loss downstream of their reactors, they did not account for those particles lost within their reactor; this could have led to an underestimate of their SOA yields. While the methods to calculate OH concentration and exposure were very similar, the CPOT and the PAM OFR are nevertheless different and, therefore, OH concentrations

could vary locally in dissimilar ways between the reactors. ~~Since the chemical mechanism shift would be based on the local OH concentration and not the average, a~~ factor of 2 disagreement could be within the uncertainty. A comparison between
345 predicted and estimated OH exposures for the PAM OFR indicates agreement only within a factor of 3 (Li et al., 2015; Janecek et al., 2019), so a factor of 2 disagreement in OH-OH concentrations or exposures would seem to be with the uncertainties for the CPOT and the PAM OFR.

~~If the SOA yield depends on the OH exposure, instead of the OH concentration, we would expect that the dependence would actually be on the OH exposure normalized to the amount of reacted D5. That is, the number of OH radicals available per
350 reacted D5 molecule, as is shown in Fig. 1b. This figure shows a factor of 10 disagreement between data from Wu and Johnston (2017) and Janecek et al. (2019) and that from the experiments conducted here. From the CPOT, Experiments 18 (OH exposure/ Δ [D5]=1.5 s, Y=73%) and 19 (OH exposure/ Δ [D5]=1.6 s, Y=68%), which had [D5]₀=82 ppb, differ significantly from Experiments 16 and 17 (OH exposure/ Δ [D5]=0.5 s, Y=79%), which had [D5]₀=246 ppb and otherwise identical experimental conditions. This suggests that OH exposure is not the driving force in determining the SOA yield.~~

~~The major difference in Experiments 16–17 and 18–19 is the percent of D5 that reacted by the end of the experiment: 97% for Experiment 16, 98% for Experiment 17, and 100% for Experiments 18–19. Figure 3 shows the fraction reacted compared to the SOA yield for experiments performed in this study and those in the literature. This fit could indicate that there are later generation oxidation products that form large amounts of aerosol and that the gas-phase reaction rate to form the low-volatility later-generation oxidation product is slower than the gas-phase reaction rate to form the first-generation product
360 (Kroll and Seinfeld, 2008). However, if this were the case, Experiments 18–19 (Y=73% and 68%), in which all of the initial D5 reacted throughout the experiment, should show higher SOA yields than Experiments 16–17 (Y=79%), which they do not. Additionally, if later generation oxidation products produced more aerosol, there should be a correlation between Y and the OH exposure normalized to the amount of reacted D5 (Fig. 1b), which is also not accurate. The color axis in Fig. 3 indicates that across studies the fraction of D5 reacted correlates with the [OH]. So, there is no reason to suspect that it is the later-generation
365 products that matter instead of the OH concentration for determining SOA yield.~~

~~If OH concentration is the strict determinant of the SOA yield, Experiments 16–17 should give the same SOA yields as Experiments 18–19. These experiments do have similar SOA yields, and Experiments 16–18 are all within error. If there is a difference, it might be attributed to a dependence on the organic aerosol mass concentration (M) at high [OH]. This would indicate that at high mass loadings, relatively more low-volatility products partition into the particle phase. This could also
370 explain the disagreement in Fig. 1a between the CPOT experiments and the data from Janecek et al. (2019). Partitioning between the particle and gas phases does not significantly change the dependence of Y on [OH] for the experiments performed in this study: as shown in Fig. 2a, Experiment 15 (triangle with M=717 $\mu\text{g m}^{-3}$) has a similar M as Experiments 18 and 19 (triangles with M of 917 and 851 $\mu\text{g m}^{-3}$, respectively) but due to their differing [OH], they have very different SOA yields. Fig. 2b shows the same for a comparison between the lower [OH] experiments (Experiments 1–10); even at the same M, the
375 OH concentration is what matters for determining Y. Note that the seed-surface-area dependence of experiments performed by Wu and Johnston (2017) is likely a result of the loss of oxidation product to the reactor's walls instead of to condensation onto particles suspended in the bulk of the chamber. The vast majority of the experiments performed under atmospherically relevant~~

~~OH concentrations were also seeded and showed low SOA yields. For all experiments with $[\text{OH}] < 10^8 \text{ molec cm}^{-3}$, the SOA yield is still $< 5\%$ and, in general, is closer to $\sim 1\%$.~~

380 4 Conclusions

The atmospheric aerosol formation potential of D5 was investigated under a range of OH concentrations and exposures. While secondary organic aerosol (SOA) yields can reach 110% (158% at the upper limit) at OH mixing ratios of $\sim 5 \times 10^9 \text{ molec cm}^{-3}$ and OH exposures of $\sim 3 \times 10^{12} \text{ molec s cm}^{-3}$, at lower OH concentrations and exposures ($\lesssim 10^7 \text{ molec cm}^{-3}$ and $\lesssim 2 \times 10^{11} \text{ molec s cm}^{-3}$, respectively), SOA yields do not exceed 6% and are likely $\sim 1\%$.

385 Between the experiments performed here and those in the literature, the SOA yields vary widely; but, these SOA yields are correlated to a similarly large range of OH concentrations and OH exposures. Environmentally relevant OH concentrations are on the order of $10^6 \text{ molec cm}^{-3}$; since D5 is primarily lost to OH and has a half life of 3.5–7 days, OH exposures on the order of $10^{12} \text{ molec s cm}^{-3}$ are also relevant. When viewed as a function of OH concentration or exposure, results here generally agree with those from the literature.

390 Due to experimental limitations, in particular an inability to perform experiments for multiple days without diluting the sample and otherwise changing the conditions, the OH concentration and exposure are often correlated, as was the case for these experiments. The correlation, however, will be different in the atmosphere than in the lab. When extrapolating these laboratory results, atmospheric modelers should be careful about understanding the relevant OH concentrations and exposures because the two variables may have different effects on the chemistry of the system and, correspondingly, the SOA yield. There
395 may also be other variables, not investigated in this work, that affect the chemistry and SOA formation of D5 oxidation.

~~The atmospheric aerosol formation potential of D5 was investigated under a range of OH concentrations and exposures. While secondary organic aerosol (SOA) yields can reach 79% (114% at the upper limit) at OH mixing ratios of $\sim 5 \times 10^9 \text{ molec cm}^{-3}$, at atmospherically relevant OH concentrations ($[\text{OH}] \lesssim 10^{7.5} \text{ molec cm}^{-3}$), SOA yields do not exceed 5% and are likely $\sim 1\%$. It is the OH concentration, and not the OH exposure, that affects the SOA yield.~~

400 ~~This demonstrates the importance of extrapolating to the atmosphere at OH concentrations close to atmospheric levels and of using the appropriate reactor for the chemistry of a precursor to determine the secondary organic aerosol formation: if OH concentration is dominant, environmental chambers may be more useful, but if OH exposure matters, then flow tubes that have high OH mixing ratios may be the best tool.~~

~~Despite the relatively low SOA yields~~ Regardless of the true, ambiently relevant SOA yield of D5 measured here at ambient
405 ~~OH concentrations,~~ silicon has still been observed in ambient aerosol and its concentration is likely somewhat population (and not vehicle) dependent (Bzdek et al., 2014; Pennington et al., 2012). Since D5 is so abundant If a lower SOA yield than previously thought is atmospherically appropriate, it could be possible that the silicon present is from D5 or other volatile methyl siloxanes, just in lower concentrations than expected, since D5 is so abundant. Another possibility is that silicon in the aerosol-phase comes from polydimethylsiloxanes (Weschler, 1988).

410 Since the aerosol formed from volatile chemical products (VCPs) may dominate the high concentrations of particulate matter found in urban areas (McDonald et al., 2018), understanding those VCPs that have high aerosol-formation potential and those which do not is important for formulating policy to reduce human exposure to organic aerosol.

Appendix A: NO_x -Dependence of SOA Yield

415 For [the chamber experiments, which had](#) atmospherically relevant OH concentrations, the SOA yield does not change depending on the NO_x concentration: experiments with no NO_x present are on both the lower and higher end of the SOA yields for the chamber experiments. Those with a continuous injection of NO throughout the experiment, which ensured that the NO/HO_2 ratio remained high even as the NO reacted, had SOA yields similar to both the no NO_x and the initial NO experiments. This indicates that different NO mixing ratios did not have an effect on the measured SOA yield.

420 This does not imply that the chemistry is independent of NO concentration. Indeed, the concentrations of gas-phase fragments detected by the CIMS at m/z 139, 169, 243, and 317, which likely correspond to oxygenated fragments of D5, depend on the NO concentration but not the NO_2 concentration. Figure A1 shows the signal for these fragments normalized to the reagent ion as a function of the NO concentration at any time. Note that, since some of the methyl nitrite is detected as NO, data from [Experiment 8-C8](#) were not included. Figure A2 shows the NO and NO_2 concentrations in each experiment as a function of time.

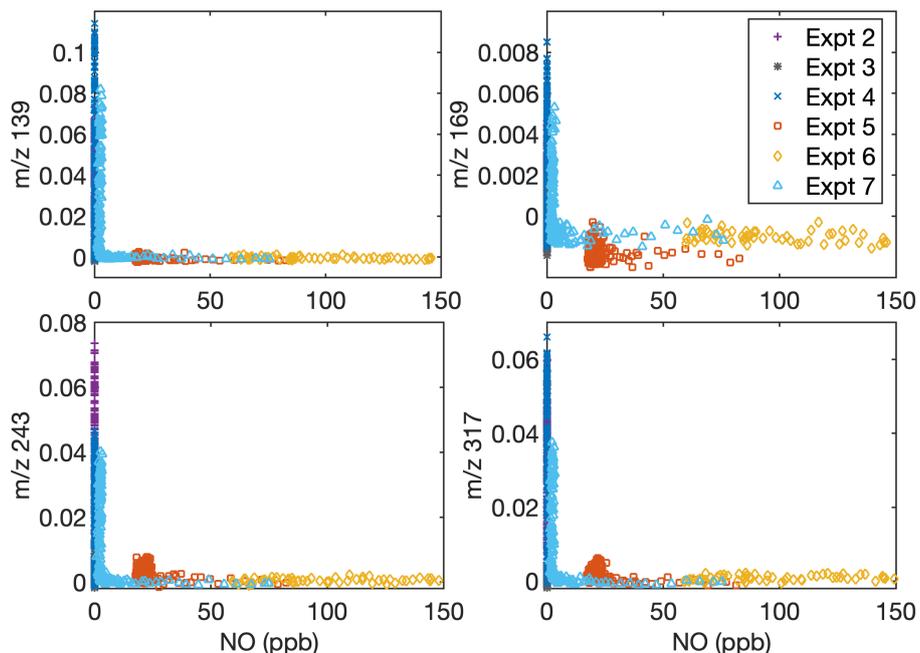


Figure A1. Dependence of gas-phase D5 oxidation products on the NO concentration in the chamber indicates that oxidation chemistry changes depending on NO concentrations. Signals normalized to the reagent concentrations with (a) $m/z=139$, (b) $m/z=169$, (c) $m/z=243$, and (d) $m/z=317$ are shown as a function of NO concentration. Experiment 6-C6 has [NO] extending to >450 ppb, but since the normalized signal remains close to 0, data above [NO]=150 ppb are cut off for clarity. Because of the inaccuracy of NO measurements during oxidation when methyl nitrite is present, Experiment 8-C8 is not included.

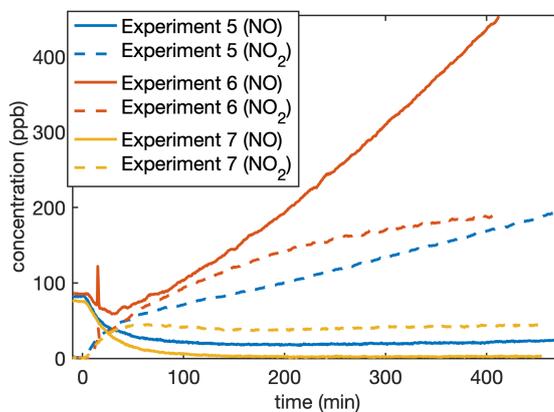


Figure A2. For the experiments that included NO_x , the NO and NO_2 concentrations as a function of the time since the onset of oxidation. Experiment 8-C8 is not included, since methyl nitrite was present. The measurement uncertainty is ~ 5 ppb, but any organonitrates would also be measured as NO_2 .

425 Fu et al. (2020) found that the gas-phase rearrangement of methylsiloxanes is dependent on the NO/HO₂ ratio. A comparison of Figs. A1 and A3 shows that the concentration of some gas-phase fragments is dependent on the NO mixing ratio but not on the NO₂ mixing ratio. This is consistent with gas-phase products depending on the NO/HO₂ ratio. Note that at all NO/HO₂ ratios investigated, aerosol formation is still minimal when [OH] is small.

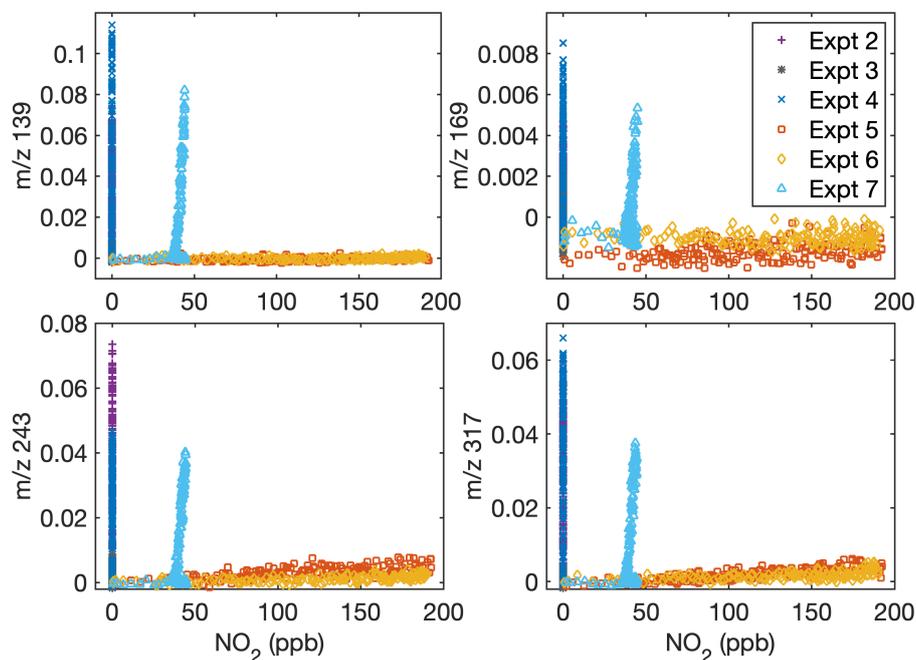


Figure A3. Dependence of gas-phase D5 oxidation products on the NO₂ concentration in the chamber indicates that oxidation chemistry does not depend on NO₂ (but does depend on NO, see Fig. A1). Signals normalized to the reagent concentrations with (a) m/z=139, (b) m/z=169, (c) m/z=243, and (d) m/z=317 are shown as a function of NO₂ concentration. Because of the inaccuracy of NO₂ measurements during oxidation when methyl nitrite is present, Experiment 8-C8 is not included.

430 *Data availability.* Chamber data available upon request and through the Index of Chamber Atmospheric Research in the United States (ICARUS), [experiment sets 220 and 221](https://icarus.ucdavis.edu/experimentset/220), <https://icarus.ucdavis.edu/experimentset/220> and <https://icarus.ucdavis.edu/experimentset/221>.

Author contributions. SMC designed the experiments, carried out the data collection and analysis, and wrote the manuscript. YH assisted with the CPOT experiments. RSB and QL participated in discussions about the study. QL measured the aerosol density. DRC secured funding for the project. JHS supervised the work. All authors reviewed and edited the manuscript.

Competing interests. The authors declare that they have no conflict of interest.

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Supporting Information: Secondary Organic Aerosol Formation from the Oxidation of Decamethylcyclopentasiloxane at Atmospherically Relevant OH Concentrations

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~~Two~~Three Figures (S1–S3), ~~2~~4 pages total

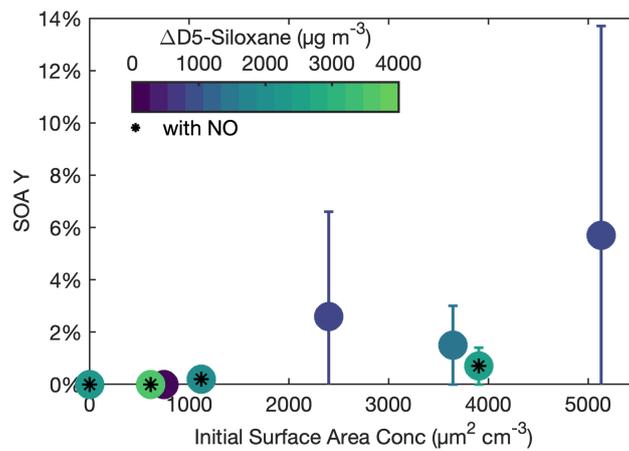


Figure S1. SOA yield at the end of ~~each experiment~~ experiments C1-8 with the associated uncertainty is shown as a function of the initial seed surface area concentration. The color of each point represents the amount of D5 that reacted throughout the entire experiment. Experiments with NO_x present include a black asterisk in their center.

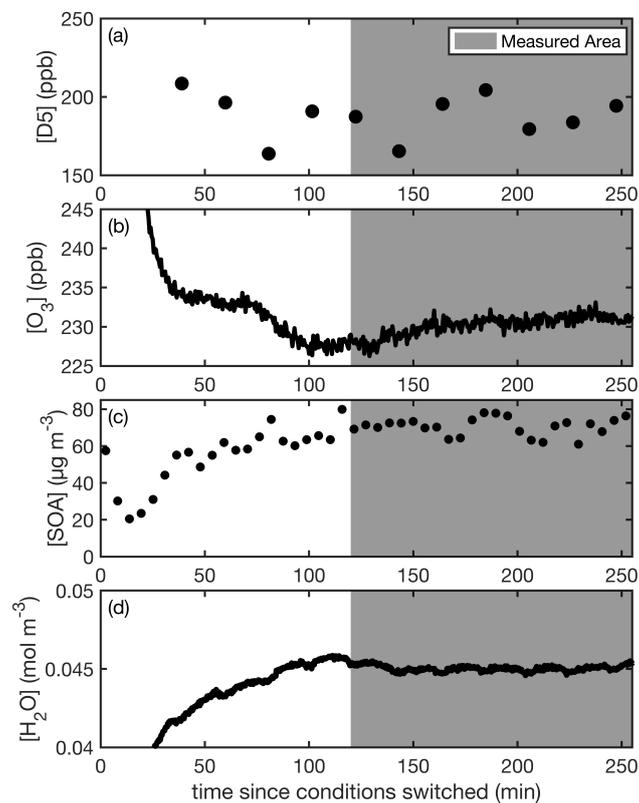


Figure S2. For experiment F11, the stability of the CPOT reactor as shown by measurements at the output of the CPOT of the (a) concentration of D5, (b) concentration of ozone, (c) concentration of SOA, and (d) absolute humidity in the reactor. Data were collected beginning 2 h after conditions were switched. Data prior to 30 min are outside the chosen y-axis limits for panels a, b, and d.

