

Dear Reviewer,

We appreciate your careful consideration of our manuscript. We have carefully responded to all of your [point-by-point](#) comments and issues and have revised the manuscript accordingly. These revisions are described in detail below.

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**Reviewer 1#**

The dominant formation pathway of sulfate aerosols under haze conditions is still under debate. Liu et al. investigated the formation mechanism of particulate sulfate based on a statistical analysis of long-term observations in Shijiazhuang and Beijing supported with flow tube experiments. They found that the uptake of SO<sub>2</sub> is the rate-determining step of sulfate formation. Ammonium nitrate plays an important role in sulfate formation by impacting the aerosol liquid water content and the phase state of particles. Overall, the paper is well written and the findings have important implications for understanding the sulfate chemistry under haze conditions and improving the air quality in urban environments. However, I have some concerns regarding methods and data analysis that must be addressed before the paper can be considered for publication.

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**Response:** Thank you for your positive comments.

**General comments:**

Uptake kinetics of SO<sub>2</sub>: the authors stated that the RDS of sulfate formation should be the uptake of SO<sub>2</sub> because the dependence of sulfate formation rate on RH is opposite to the dependences of SOR and  $\gamma_{\text{SO}_2}$  on RH. Did the uptake of SO<sub>2</sub> refer to the mass transfer of SO<sub>2</sub> to aerosol particles? If yes, the rates of mass transfer of SO<sub>2</sub> and aqueous oxidation of S(IV) can be calculated using a resistance model (Cheng et al. 2016). According to Cheng et al. (2016), the mass transfer of SO<sub>2</sub> is not the rate-determining step.

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**Response:** Thank you for your good comments and suggestion. Yes, the uptake of SO<sub>2</sub> refers to the mass transfer of SO<sub>2</sub> to aerosol particles. Fig. R1 shows the probability weighted production rate of sulfate through mass transfer (uptake) and aqueous phase oxidation of SO<sub>2</sub> in Shijiazhuang. The mass transfer of SO<sub>2</sub> to aerosol particles is the

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RDS, in particular, when RH is lower than 70%. We added the production rate of sulfate through mass transfer in Fig. 4 in the revised manuscript.

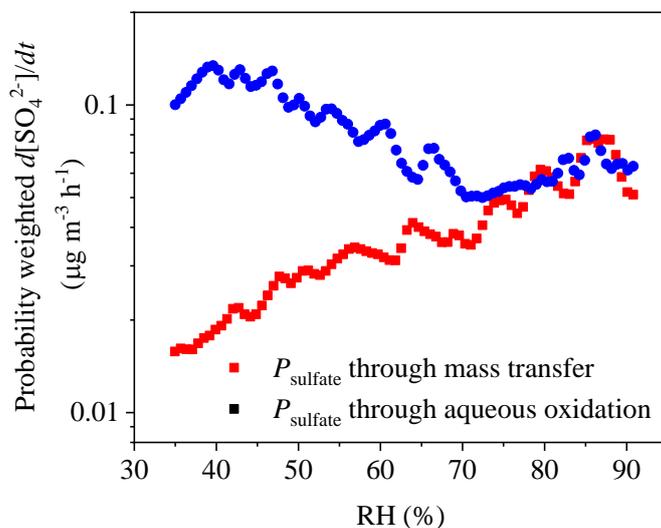


Fig. R1. Dependence of the probability weighted production rate of sulfate through mass transfer and aqueous oxidation of S(IV) in Shijiazhuang.

Using a resistance model, Cheng et al (2016) proposed out that the mass transfer of SO<sub>2</sub> is not the RDS. It should be noted that a large mass accommodation coefficient of SO<sub>2</sub> ( $\alpha = 0.11$ ) was used in their calculations (Cheng et al., 2016). According to the relationship between the mass accommodation coefficient ( $\alpha$ ) and the uptake coefficient ( $\gamma$ ) of SO<sub>2</sub> (Kulmala and Wagner, 2001), the  $\alpha_{SO_2}$  on particles is on the same order of the  $\gamma_{SO_2}$  (Fig. R2). This means that the mass accommodation of SO<sub>2</sub> on particles should be much lower than the value used by Cheng et al (2016) according to the measured uptake coefficient of SO<sub>2</sub> in this work. Therefore, the mass transfer rate coefficient ( $k_{MT}$ ) is much lower than their value. We added a paragraph “We further calculated the production rate of sulfate through uptake of SO<sub>2</sub> (mass transfer to aerosol particles) according to,

$$\frac{d[SO_4^{2-}]}{dt} = 3600 \cdot \frac{96}{64} \cdot \frac{\gamma_{SO_2} A_s \omega c_{SO_2}}{4} \quad (8)$$

where,  $A_s$  is the surface area concentration of PM<sub>2.5</sub>,  $\omega$  is the mean molecular velocity of SO<sub>2</sub> and  $c_{SO_2}$  is the mass concentration of SO<sub>2</sub>. As shown in Fig. 4C, the probability weighted production rate of sulfate through uptake of SO<sub>2</sub> (the grey line) is lower than

that through aqueous oxidation of S(IV), in particular, when RH is lower than 70%. It should be noted the mass transfer of SO<sub>2</sub> was not thought as the RDS using a large mass accommodation coefficient of SO<sub>2</sub> ( $\alpha = 0.11$ ) (Cheng et al., 2016). According to the relationship between the mass accommodation coefficient ( $\alpha$ ) and the uptake coefficient ( $\gamma$ ) of SO<sub>2</sub> (Kulmala and Wagner, 2001), the  $\alpha_{\text{SO}_2}$  on particles is on the same order of the  $\gamma_{\text{SO}_2}$ . This means that mass transfer rate might be greatly overestimated by Cheng et al. (2016)” in lines 432-444 in the revised manuscript.

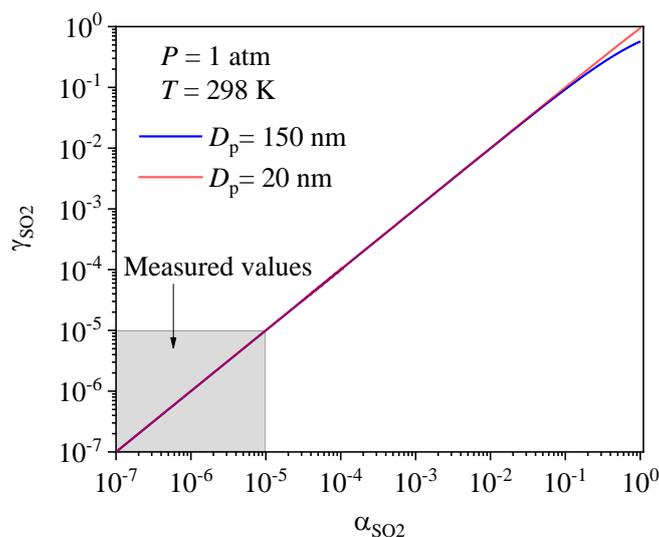


Fig. R2. The relationship between the mass accommodation coefficient ( $\alpha$ ) and the uptake coefficient ( $\gamma$ ) of SO<sub>2</sub> for 20 nm and 150 nm particles at 1 atm and 298 K.

Flow tube experiments: (1) The wall loss of SO<sub>2</sub> on the inner surface of the outside tube and the outside surface of the sample holder was subtracted. However, the wall loss of SO<sub>2</sub> in the presence of NH<sub>3</sub> and/or NO<sub>2</sub> would be larger even in the absence of seed aerosols (Ge et al., 2019), which may lead to an overestimation of  $\gamma_{\text{SO}_2}$  in the presence of NH<sub>3</sub> and/or NO<sub>2</sub>. Did the authors measure the  $\gamma_{\text{SO}_2}$  in the presence of NH<sub>3</sub> and/or NO<sub>2</sub> without the presence of seed aerosols?

**Response:** Thank you. We agree with you that the wall loss of SO<sub>2</sub> in the presence of NH<sub>3</sub> and/or NO<sub>2</sub> would be larger in the absence of seed aerosols. The  $\gamma_{\text{SO}_2}$  was  $2 \times 10^{-7}$  in the presence of NH<sub>3</sub> and NO<sub>2</sub> and in the absence of seed particles. This value is significantly lower than that in the presence of particles ( $\sim 1 \times 10^{-5}$ ). On the other hand,

the  $c_0$  and  $c$  of  $\text{SO}_2$  are required when we calculating the uptake coefficient. Figure R3A shows the configuration of the flow tube reactor for measuring the  $c_0$ , and Figure R3B shows that for measuring the  $c$ . Because all the inner surface of the sample tube is covered by particles, the contribution of the wall loss of  $\text{SO}_2$  to the drop of the  $c$  is limited (less than 3%). We added a short paragraph “It should be noted that the wall loss of  $\text{SO}_2$  in the presence of  $\text{NH}_3$  and/or  $\text{NO}_2$  would be larger in the absence of seed aerosols. Additional control experiments in the presence of  $\text{NO}_2$  and  $\text{NH}_3$  demonstrate that the contribution of wall loss of  $\text{SO}_2$  should be less than 3 % to the measured  $\gamma$ ” in lines 211-214 in the revised manuscript.

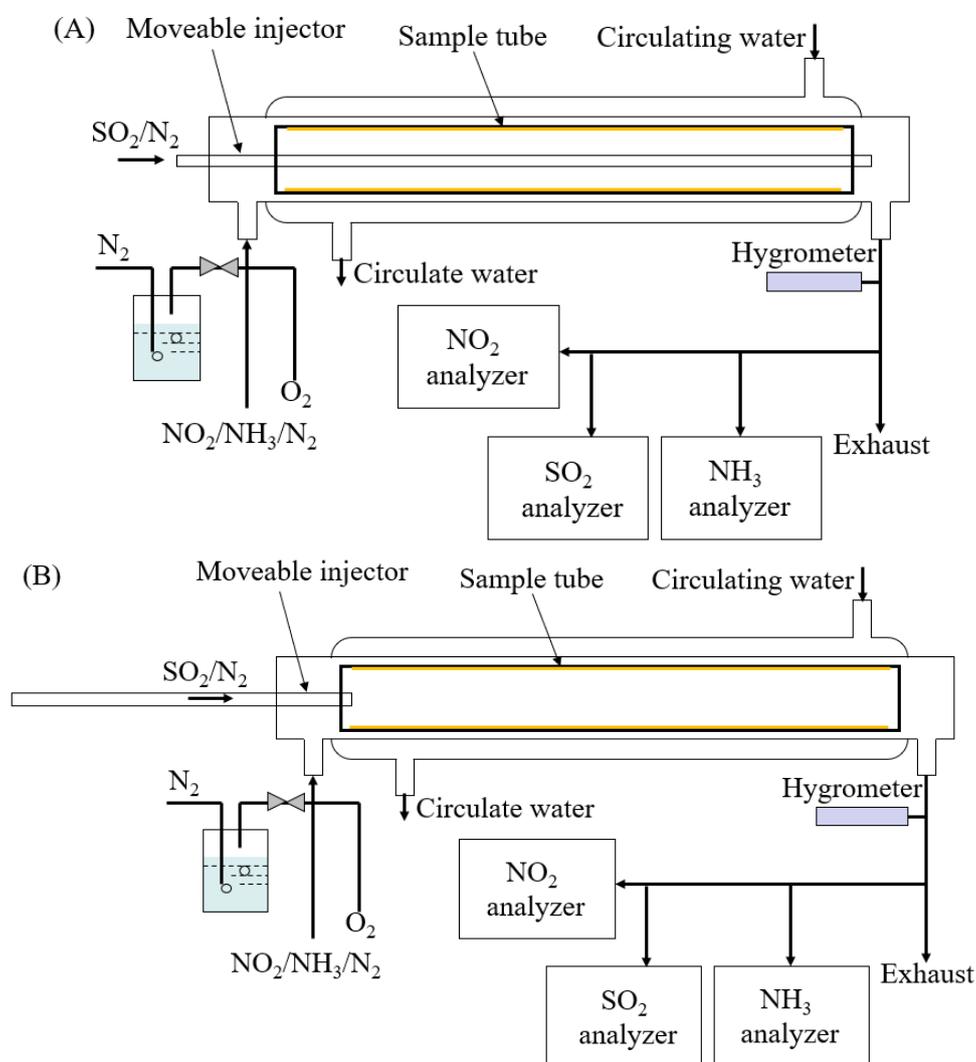


Fig. R3. The configuration of the flow tube for measuring the  $c_0$  and  $c$  of  $\text{SO}_2$ .

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(2) Can 100 ppb of  $\text{NO}_2$  oxidize 190 ppb of  $\text{SO}_2$  at a detectable rate in around 1 min?

The comparable  $\gamma_{\text{SO}_2}$  in the absence and presence of  $\text{NO}_2$  may not demonstrate that  $\text{NO}_2$  is not an important oxidant of  $\text{SO}_2$  if the enhanced uptake of  $\text{SO}_2$  in the presence of  $\text{NO}_2$  is too low under the experimental conditions of the paper.

90 **Response:** Thank you so much for your good comment. The  $\gamma_{\text{SO}_2}$  at 80% RH was  $1.7 \pm 0.3 \times 10^{-6}$  on the mixture of dust and  $\text{NH}_4\text{NO}_3$  in the absence of  $\text{NH}_3$  and  $\text{NO}_2$  (Fig. R4 and Fig. 3). It increased to  $3.7 \pm 0.2 \times 10^{-6}$  in the presence of  $\text{NO}_2$  at the same RH. This means that  $\text{NO}_2$  can promote  $\text{SO}_2$  uptake at high RH in the absence of  $\text{NH}_3$ . Thus, the short reaction time should not be a problem. However, in the presence of  $\text{NH}_3$ , the  $\gamma_{\text{SO}_2}$  showed no difference between with and without  $\text{NO}_2$ . This means that the promotion effect of  $\text{NO}_2$  on  $\text{SO}_2$  uptake might be too low to be detected in the presence of  $\text{NH}_3$ . Because  $\text{NH}_3$  is abundant in North China, we think the promotion effect of  $\text{NO}_2$  alone to  $\text{SO}_2$  uptake should be limited in the atmosphere. In the revised manuscript, we added a paragraph “It should be pointed out that the  $\gamma_{\text{SO}_2}$  at 80% RH was  $1.7 \pm 0.3 \times 10^{-6}$  on the mixture of dust and  $\text{NH}_4\text{NO}_3$  in the absence of  $\text{NH}_3$  and  $\text{NO}_2$  (Fig. 3). It increased to  $3.7 \pm 0.2 \times 10^{-6}$  in the presence of  $\text{NO}_2$ . This is consistent with the promotion effect of  $\text{NO}_2$  for converting  $\text{SO}_2$  to sulfate in the absence of  $\text{NH}_3$  as observed in both a smog chamber (Wang et al., 2016) and a bubbling reactor (Chen et al., 2019d). However, the enhanced uptake of  $\text{SO}_2$  induced by  $\text{NO}_2$  might be too low to be measured in the presence of  $\text{NH}_3$ . Therefore, the weak promotion effect by  $\text{NO}_2$  alone cannot explain the negative correlation between the SOR and the concentration of  $\text{NO}_2$  in Fig. 6F” in lines 607-615 in the revised manuscript.

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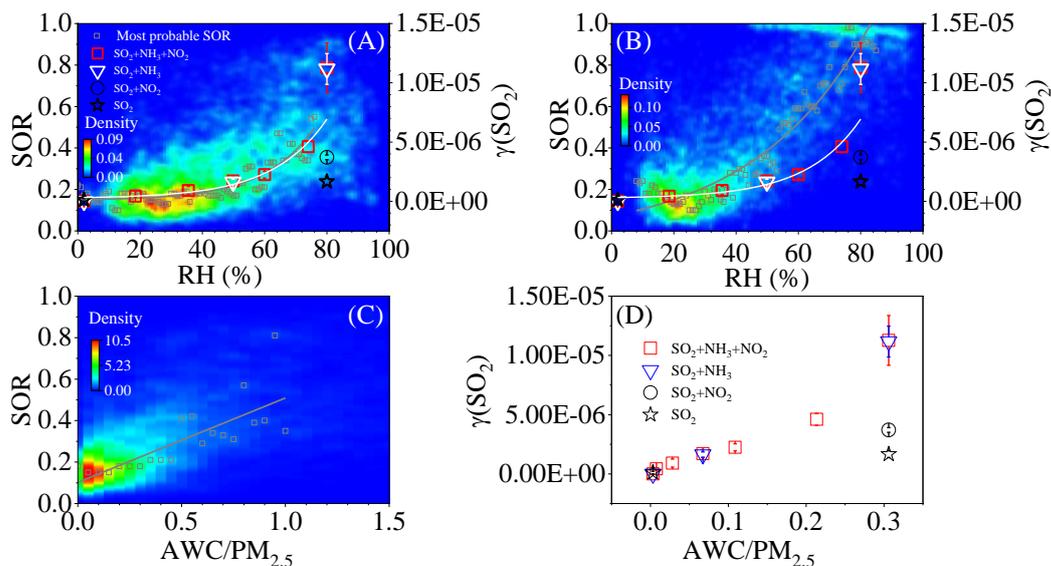


Fig. R4. Relationship between SOR and  $\gamma_{\text{SO}_2,\text{BET}}$  on dust internally mixed with  $\text{NH}_4\text{NO}_3$  (2:1) and RH in (A) Shijiazhuang and (B) Beijing, and the correlation of (C) SOR in Shijiazhuang and (D)  $\gamma_{\text{SO}_2,\text{BET}}$  with  $\text{AWC}/\text{PM}_{2.5}$ . The initial concentrations of  $\text{SO}_2$ ,  $\text{NO}_2$  and/or  $\text{NH}_3$  in the flow tube reactor were  $190 \pm 2.5$ ,  $100 \pm 2.5$  and/or  $50 \pm 2.5$  ppb, respectively. The grey lines are the fitting curves for the most probable SOR and the white lines are the fitting curves for the  $\gamma_{\text{SO}_2,\text{BET}}$ .

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Specific comments:

Lines 191-192: Did the control experiments run in the presence of  $\text{NH}_3$  and  $\text{NO}_2$ ?

**Response:** Thank you. Yes, it has been done. We added a short paragraph “It should be noted that the wall loss of  $\text{SO}_2$  in the presence of  $\text{NH}_3$  and/or  $\text{NO}_2$  would be larger in the absence of seed aerosols. Additional control experiments in the presence of  $\text{NO}_2$  and  $\text{NH}_3$  demonstrate that the contribution of wall loss of  $\text{SO}_2$  should be less than 3 % to the measured  $\gamma$ ” in lines 211-214 in the revised manuscript.

Lines 368-372: The oxidation of  $\text{SO}_2$  by  $\text{O}_2$  on the aqueous microdroplets has been found to occur under acidic conditions ( $\text{pH} < 3$ ). What is the aerosol pH of the mixture of ammonium nitrate and dust?

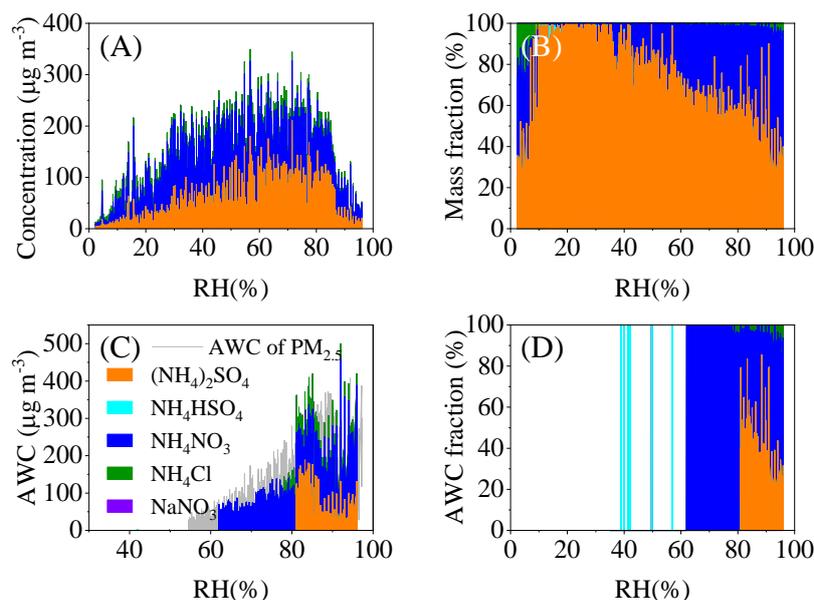
**Response:** Thank you so much for your comment. We cannot calculate or measure the pH of the mixture of  $\text{NH}_4\text{NO}_3$  and dust. The pH of deliquesced  $\text{NH}_4\text{NO}_3$  is around 4.2

calculated using the ISORROPIA II model. This value is close to the literature value  
130 (pH<3.5) (Hung et al., 2018). We revised the paragraph “Another study also observed  
a quick formation of sulfate on the surface of aqueous microdroplets under acidic  
conditions (pH < 3.5) without the addition of other oxidants, which was explained by  
the direct interfacial electron transfer from SO<sub>2</sub> to O<sub>2</sub> on the aqueous microdroplets  
(Hung et al., 2018). The pH of deliquesced NH<sub>4</sub>NO<sub>3</sub> is 4.2 calculated using the  
135 ISORROPIA II model” in lines 398-403 in the revised manuscript.

Fig 5C: The AWC was attributed to individual components using E-AIM model. Are  
the concentrations of the total AWC consistent with the ISORROPIA model? At RH of  
60%-80%, only ammonium nitrate aerosols contributed to the AWC. Does this indicate  
140 that ammonium sulfate aerosols are effloresced and phase-separated with ammonium  
nitrate aerosols? Please explain why ammonium sulfate aerosols and ammonium nitrate  
aerosols are not in the same liquid phase.

**Response:** Thank you for your good comments. The AWC attributed to the individual  
salt cannot be separated from that of PM<sub>2.5</sub> using the ISORROPIA model. Thus, it was  
145 estimated using the reconstructed mass concentration of the salts and the growth factors.  
In Fig. R5 (Fig. 5C in the revised manuscript), we compared the total AWC of PM<sub>2.5</sub>  
calculated using the ISORROPIA model with the sum of the AWC attributed to the  
individual salt using the E-AIM model. Overall, the latter one underestimates around  
13% of the AWC. This should be related to the difference in the mixing state between  
150 these two calculation methods. NH<sub>4</sub>NO<sub>3</sub> dominates the AWC at RH of range 60-80%.  
However, we don't think this means (NH<sub>4</sub>)<sub>2</sub>SO<sub>4</sub> aerosols are effloresced and phase-  
separated with NH<sub>4</sub>NO<sub>3</sub>. As shown in Fig. R5, NH<sub>4</sub>NO<sub>3</sub> explained ~70% of the AWC  
of PM<sub>2.5</sub>. Thus, we think it is reasonable to draw a conclusion that NH<sub>4</sub>NO<sub>3</sub> is the  
dominant contributor to AWC in the RH range of 60-80%. In the revised manuscript,  
155 we added a new paragraph “As shown in Fig. 5C, the sum of the AWC of individual  
salts overall underestimated around 13 % of that calculated using the ISORROPIA II  
model (the gray line) because the mixing state was not considered in the former method.  
However, we can still draw a conclusion that NH<sub>4</sub>NO<sub>3</sub> and (NH<sub>4</sub>)<sub>2</sub>SO<sub>4</sub> are the major

contributors to the AWC” in lines 511-515 in the revised manuscript.



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Fig. R5. Variations of (A) concentrations and (B) fractions of molecular composition of water-soluble ions, and (C) and (D) the corresponding contributions to AWC with RH in Shijiazhuang.

165 Lines 545-551: The authors should rule out the possibility that the enhanced uptake of  $\text{SO}_2$  induced by  $\text{NO}_2$  in the reaction time scale of the flow tube experiments is too low to be measured. Previous smog chamber experiments with longer reaction times have demonstrated that  $\text{NO}_2$  can promote sulfate formation (Wang et al., 2016; Chen et al., 2019).

170 **Response:** Thank you for your suggestion. We added a paragraph as “It should be pointed out that the  $\gamma_{\text{SO}_2}$  at 80% RH was  $1.7 \pm 0.3 \times 10^{-6}$  on the mixture of dust and  $\text{NH}_4\text{NO}_3$  in the absence of  $\text{NH}_3$  and  $\text{NO}_2$  (Fig. 3). It increased to  $3.7 \pm 0.2 \times 10^{-6}$  in the presence of  $\text{NO}_2$ . This is consistent with the promotion effect of  $\text{NO}_2$  for converting  $\text{SO}_2$  to sulfate in the absence of  $\text{NH}_3$  as observed in both a smog chamber (Wang et al., 2016) and a bubbling reactor (Chen et al., 2019d). However, the enhanced uptake of  $\text{SO}_2$  induced by  $\text{NO}_2$  might be too low to be measured in the presence of  $\text{NH}_3$ . Therefore, the weak promotion effect by  $\text{NO}_2$  alone cannot explain the negative correlation between the SOR and the concentration of  $\text{NO}_2$  in Fig. 6F” in lines 607-615 in the revised manuscript.

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Technical comments:

Line 28: Write out “SOR”.

**Response:** Thank you. It has been defined as “sulfur oxidation ratio” in line 28 in the revised manuscript.

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Fig 5: Variations of (A) concentrations...

**Response:** Thank you. We revised the caption as “Variations of (A) the mass concentrations and (B) the mass fractions of molecular composition in PM<sub>2.5</sub>, (C) the estimated AWC attributed to different composition and (D) the corresponding AWC fraction as a function of RH in Shijiazhuang” in the revised manuscript.

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### References:

Cheng, Y., Zheng, G., Wei, C., Mu, Q., Zheng, B., Wang, Z., Gao, M., Zhang, Q., He, K., Carmichael, G., Poschl, U., and Su, H.: Reactive nitrogen chemistry in aerosol water as a source of sulfate during haze events in China, *Sci. Adv.*, 2, <https://doi.org/10.1126/sciadv.1601530>, 2016.

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Hung, H.-M., Hsu, M.-N., and Hoffmann, M. R.: Quantification of SO<sub>2</sub> oxidation on interfacial surfaces of acidic micro-droplets: Implication for ambient sulfate formation, *Environ. Sci. Technol.*, 52, <https://doi.org/9079-9086>, [10.1021/acs.est.8b01391](https://doi.org/10.1021/acs.est.8b01391), 2018.

Kulmala, M., and Wagner, P. E.: Mass accommodation and uptake coefficients — a quantitative comparison, *J. Aerosol Sci.*, 32, 833-841, [https://doi.org/10.1016/S0021-8502\(00\)00116-6](https://doi.org/10.1016/S0021-8502(00)00116-6), 2001.

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