

## Response to Comments of Editor

(Comments in *italics*)

**Manuscript number:** acp-2021-119

**Title:** Impacts of aerosol-photolysis interaction and aerosol-radiation feedback on surface-layer ozone in North China during a multi-pollutant air pollution episode

*Please respond to the comments of the reviewer. In particular, it seems that the paper is still not communicating clearly why these 3 high pollution episodes are relevant or important to study, and consequently, how to interpret the conclusions to make broader inferences about other time periods.*

*To me, analyzing 3 short episodes in one location is only sufficient for a paper in ACP if the authors motivate well why those episodes are important to study, putting them in a broader context.*

### Response:

Thanks for the Editor's helpful comments to this manuscript. We have carefully addressed the Reviewer's comments.

In this study, we aim to quantify the impacts of aerosol-radiation interactions on O<sub>3</sub> through two processes (aerosol-photolysis interaction (API) and aerosol-radiation feedback (ARF)) during the complex air pollution events with high concentrations of both PM<sub>2.5</sub> and O<sub>3</sub>. These events were selected to obtain the strongest signals of API and ARF. These events were selected also considering that the measurements of J[NO<sub>2</sub>] during 2014 to 2015 from Peking University site (Wang et al., 2019) can help to constrain the simulated photolysis rates of NO<sub>2</sub>. The new findings from our work are that API and ARF both have an effect of reducing O<sub>3</sub>, but the role of API is much larger than that of ARF during all the simulated episodes, which are important for understanding summertime O<sub>3</sub> pollution in China.

### Reference:

Wang, W., Li, X., Shao, M., Hu, M., Zeng, L., Wu, Y., and Tan, T.: The impact of aerosols on photolysis frequencies and ozone production in Beijing during the 4-year period 2012–2015, Atmos. Chem. Phys., 19, 9413–9429, <https://doi.org/10.5194/acp19-9413-2019>, 2019.

## Response to Comments of Reviewer #2

(Comments in *italics*)

**Manuscript number:** acp-2021-119

**Title:** Impacts of aerosol-photolysis interaction and aerosol-radiation feedback on surface-layer ozone in North China during a multi-pollutant air pollution episode

*The revised reversion of Yang et al. does address several of the manuscripts original limitations. The manuscript is still clear and well written and the scientific problem addressed is important to the scientific community. However, a few issues remain that leave questions as to whether the findings in the manuscript are robust and meaningful. My major and minor comments on the revised manuscript are as follows:*

**Response:**

Thanks to the reviewer for the valuable comments and suggestions which are helpful for us to improve our manuscript. We have revised the manuscript carefully, as described in our point-to-point responses to the comments.

**Major Comments:**

- 1. The authors have included two additional episodes to address the issue of scientific robustness. With these two additional episodes, it appears that the impact of aerosol radiation interactions, usually via API, are similar in all episodes despite variability in the magnitude and spatial extent of the CAPAs. With these finding we can reasonably conclude that these values are indeed representative of CAPAs in this region during the period of 2014-2016. However, this did not address any issues with changes in time/emissions (i.e, 2001-2005 or more currently 2018- 2020). If the authors are not going to do any additional episodes, they need to convincingly justify why the period of 2014-2016 is representative of /or important for current/future conditions.*

**Response:**

There might be some misunderstanding about the objectives of this work. As we described in the Introduction section, we aim to quantify the impacts of aerosol-radiation interactions on O<sub>3</sub> through two processes (aerosol-photolysis interaction (API) and aerosol-radiation feedback (ARF)) during the complex air pollution events with high concentrations of both PM<sub>2.5</sub> and O<sub>3</sub>. These events were selected to obtain the strongest signals of API and ARF. These events were selected also considering that the measurements of J[NO<sub>2</sub>] during 2014 to 2015 from Peking University site (Wang et al., 2019) can help to constrain the simulated photolysis rates of NO<sub>2</sub>. The new findings from our work are that API and ARF both have an effect of reducing O<sub>3</sub>, but the role of API is much larger than that of ARF during all the simulated episodes, which are important for understanding summertime O<sub>3</sub> pollution events in China.

We agree with the reviewer that the approach used in this work can be used for different timescales (for example, from 2001-2005 to 2018-2020) to examine how the aerosol-radiation interactions (API and/or ARF) contributed to the decadal changes in seasonal mean O<sub>3</sub> in eastern

China. We actually have planned to examine how API and/or ARF contributed to the increases in summertime  $O_3$  over 2013-2019 in addition to the roles of changes in emissions and heterogeneous reactions. However, these are beyond the scope of this work and are the subjects of our next paper.

2. *The authors have added caveats to the conclusion to address the issue of lacking SOA formation pathways in their simulations. These listed caveats are important, but the authors have overlooked the possibility that increased  $O_3$  from  $PM_{2.5}$  reductions will generate more SOA via increased oxidation. This feedback could partially compensate the increased  $O_3$  formation the authors predict will happen.*

**Response:**

Thanks for pointing this out. The discussions about the impacts of SOA have been revised as follows: “There are some limitations to this work. The lack of secondary organic aerosol (SOA) may result in uncertainties in our model results. First, the absence of SOA would lead to underestimation of API and ARF because of the low bias in AOD, so the reductions in  $O_3$  by API and/or ARF could have been underestimated. Second, the lack of SOA would lead to weaker heterogeneous reactions, leading to higher  $O_3$  concentrations (Li et al., 2019). The net effect of these two processes needs further studies.” (Page 15, Line 405-411)

**Minor Comments:**

1. The response to the previous Reviewer2 Minor comments 7-10 should be included in the manuscript if not already done to facilitate ease of understanding.

**Response:**

The previous Reviewer 2’s Minor Comments 7-10 have been added into Section 4.4 of the revised manuscript as follows:

- (1) Since VMIX increases the surface  $O_3$  concentrations by transporting  $O_3$  from aloft (where  $O_3$  concentrations are high) to the surface layer (Tang et al., 2017; Xing et al., 2017; Gao et al., 2018). (Page 11, Line 283-285)
- (2) The positive change in VMIX due to API may be associated with the different vertical gradient of  $O_3$  between BASE and NOAPI cases (Gao et al., 2020), as shown in Fig. 8a. (Page 11, Line 298-300)
- (3) The positive VMIX\_DIF is related to the evolution in boundary layer during the daytime. The VOCs/ $NO_x$  ratio is calculated to classify sensitivity regimes and to indicate the possible  $O_3$  responses to changes in VOCs and/or  $NO_x$  concentrations.  $O_3$  production is VOC-limited if the ratio is less than 4, and is  $NO_x$ -limited if the ratio is larger than 15 (Edson et al., 2017; Li et al., 2017). The ratio of VOCs/ $NO_x$  ranging around 4-15 indicates a transitional regime, where ozone is nearly equally sensitive to both species (Sillman, 1999). As shown in Fig. S7, (a-f),  $O_3$  is mainly formed under the VOC-limited and the transition regimes in CAPAs. As shown in Figs. S7(g-i) and S7(j-l), both the surface concentrations of VOCs and  $NO_x$  are increased when the impacts of ARF are considered. Thus, the contribution of CHEM in NOAPI is larger than that in NOALL. (Page 12, Line 315-325)
- (4) The positive variation in VMIX due to API may be associated with the different vertical gradient of  $O_3$  between BASE and NOAPI again. (Page 13, Line 340-341)

2. The color bars for Figures 5 and 6 needs to be the same for all episodes to facilitate easy comparison.

**Response:**

Following your suggestion, we now use the same color bars in Figures 5 and 6 in the revised manuscript. **(Pages 30 and 31)**

3. In Figures 2 and 3 the y-axis need to be consistent for all episodes to facilitate ease of comparison

**Response:**

We now use the same y-axis in Figures 2 and 3. **(Pages 27 and 28)**

**Reference:**

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- Gao, J., Li, Y., Zhu, B., Hu, B., Wang, L., and Bao, F.: What have we missed when studying the impact of aerosols on surface ozone via changing photolysis rates?, *Atmos. Chem. Phys.*, 20, 10831–10844, <https://doi.org/10.5194/acp-20-10831-2020>, 2020.
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- Wang, W., Li, X., Shao, M., Hu, M., Zeng, L., Wu, Y., and Tan, T.: The impact of aerosols on photolysis frequencies and ozone production in Beijing during the 4-year period 2012–2015, *Atmos. Chem. Phys.*, 19, 9413–9429, <https://doi.org/10.5194/acp19-9413-2019>, 2019.
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1   **Impacts of aerosol-photolysis interaction and aerosol-radiation**  
2   **feedback on surface-layer ozone in North China during multi-**  
3   **pollutant air pollution episodes**

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## Abstract

We examined the impacts of aerosol-radiation interactions, including the effects of aerosol-photolysis interaction (API) and aerosol-radiation feedback (ARF), on surface-layer ozone ( $O_3$ ) concentrations during three multi-pollutant air pollution episodes characterized by high  $O_3$  and  $PM_{2.5}$  levels during 28 July to 3 August 2014 (Episode1), 8-13 July 2015 (Episode2) and 5-11 June 2016 (Episode3) in North China, by using the Weather Research and Forecasting with Chemistry (WRF-Chem) model embedded with an integrated process analysis scheme. Our results show that aerosol-radiation interactions decreased the daytime shortwave radiation at surface by  $92.4\sim100.3\text{ W m}^{-2}$  averaged over the complex air pollution areas in these three episodes. The dimming effect reduced the near-surface photolysis rates of  $J[NO_2]$  and  $J[O^1D]$  by  $1.8 \times 10^{-3}\sim2.0 \times 10^{-3}\text{ s}^{-1}$  and  $5.7 \times 10^{-6}\sim6.3 \times 10^{-6}\text{ s}^{-1}$ , respectively. However, the daytime shortwave radiation in the atmosphere was increased by  $72.8\sim85.2\text{ W m}^{-2}$ , which made the atmosphere more stable. The stabilized atmosphere decreased the planetary boundary layer height and 10 m wind speed by  $129.0\sim249.0\text{ m}$  and  $0.05\sim0.12\text{ m s}^{-1}$ , respectively. The weakened photolysis rates and changed meteorological conditions reduced daytime surface-layer  $O_3$  concentrations by up to  $9.3\sim11.4\text{ ppb}$ , with API and ARF contributing  $74.6\%\sim90.0\%$  and  $10.0\%\sim25.4\%$  of the  $O_3$  decrease in these three episodes, respectively. Process analysis indicated that the weakened  $O_3$  chemical production made the greatest contribution to API effect while the reduced vertical mixing was the key process for ARF effect. This study implies that future  $PM_{2.5}$  reductions will lead to  $O_3$  increases due to weakened aerosol-radiation interactions. Therefore, tighter controls of  $O_3$  precursors are needed to offset  $O_3$  increases caused by weakened aerosol-radiation interactions in the future.

## 1 Introduction

China has been experiencing severe air pollution in recent years, characterized by high loads of PM<sub>2.5</sub> (particulate matter with an aerodynamic equivalent diameter of 2.5 micrometers or less) and high levels of ozone (O<sub>3</sub>). Observational studies exhibited positive correlations and synchronous occurrence of PM<sub>2.5</sub> and O<sub>3</sub> pollution in North China during summer (Zhao et al., 2018; Zhu et al., 2019), indicating that complex air pollution is becoming a major challenge for North China.

Aerosols can absorb and scatter solar radiation to affect Earth's energy balance. They can also act as cloud condensation nuclei and ice nuclei, and further modify the microphysical characteristics of clouds (Albrecht et al., 1989; Haywood et al., 2000; Lohmann et al., 2005). Both ways perturb meteorological variables, e.g., temperature, planetary boundary layer height (PBLH), and precipitation, and eventually influence air pollutants (Petäjä et al., 2016; Miao et al., 2018; Zhang et al., 2018). Many studies are focused on the feedback between aerosol and meteorology (Gao et al., 2015; Gao et al., 2016a; Qiu et al., 2017; Chen et al., 2019; Zhu et al., 2021). Gao et al. (2015) used the WRF-Chem model to investigate the feedbacks between aerosols and meteorological variables over the North China Plain in January 2013, and pointed out that aerosols could cause a decrease in surface temperature by 0.8-2.8 °C but an increase of 0.1-0.5 °C around 925 hPa. The more stable atmosphere caused by surface cooling and higher layer heating led to the decreases of surface wind speed and PBLH by 0.3 m s<sup>-1</sup> and 40-200 m, respectively, which further resulted in overall PM<sub>2.5</sub> increases by 10-50 µg m<sup>-3</sup> (2-30%). By using the same WRF-Chem model, Qiu et al. (2017) reported that the surface downward shortwave radiation and PBLH were reduced by 54.6 W m<sup>-2</sup> and 111.4 m due to aerosol radiative forcing during 21 and 27 February 2014 in the North China Plain. As a result, the surface PM<sub>2.5</sub> concentration averaged over the North China Plain was increased by 34.9 µg m<sup>-3</sup> (20.4%).

Aerosols can also influence O<sub>3</sub> through aerosol-radiation interactions, including aerosol-photolysis interaction and aerosol-radiation feedback. Aerosols can scatter and absorb UV radiation, and therefore directly affect O<sub>3</sub> photochemistry reactions,

71 which is called aerosol-photolysis interaction (API) (Dickerson et al., 1997; Liao et al.,  
72 1999; Li et al., 2011; Lou et al., 2014). The changed meteorological variables due to  
73 aerosol radiative forcing can indirectly affect O<sub>3</sub> concentrations, which is called  
74 aerosol-radiation feedback (ARF) (Hansen et al., 1997; Gao et al., 2018; Liu et al.,  
75 2020). Although the effects of API or ARF on O<sub>3</sub> have been examined by previous  
76 studies (Xing et al., 2017; Gao et al., 2018; Gao et al., 2020), the combined effects of  
77 API and ARF on O<sub>3</sub>, especially under the conditions of synchronous occurrence of  
78 high PM<sub>2.5</sub> and O<sub>3</sub> concentrations, remain largely elusive.

79 The present study aims to (1) quantify the respective/combined contributions of  
80 API and ARF on surface O<sub>3</sub> concentrations by using the WRF-Chem model; (2)  
81 explore the prominent physical and/or chemical processes responsible for API and  
82 ARF effects by using an integrated process rate (IPR) analysis embedded in WRF-  
83 Chem model. In order to draw the general conclusions, three multi-pollutant air  
84 pollution episodes characterized by high O<sub>3</sub> and PM<sub>2.5</sub> levels during 28 July to 3  
85 August 2014 (Episode1), 8-13 July 2015 (Episode2) and 5-11 June 2016 (Episode3)  
86 in North China are analyzed in this study. The model configuration, numerical  
87 experiments, observational data, and the integrated process rate analysis are described  
88 in section 2. Section 3 shows the model evaluation. The presentation and discussion  
89 of the model results are exhibited in section 4, and the conclusions and discussions are  
90 provided in section 5.

## 91 **2 Methods**

### 92 **2.1 Model configuration**

93 The version 3.7.1 of the online-coupled Weather Research and Forecasting with  
94 Chemistry (WRF-Chem) model (Grell et al., 2005; Skamarock et al., 2008) is used in  
95 this study to explore the impacts of aerosol-radiation interactions on surface-layer O<sub>3</sub>  
96 in North China. WRF-Chem can simulate gas phase species and aerosols coupled with  
97 meteorological fields, and has been widely used to investigate air pollution over North  
98 China (Gao et al., 2016a; Gao et al., 2020; Wu et al., 2020). As shown in Fig. 1, we  
99 design two nested model domains with the number of grid points of 57 (west-east) ×



41 (south–north) and 37 (west–east)  $\times$  43 (south–north) at 27 and 9 km horizontal resolutions, respectively. The parent domain centers at (39 °N, 117 °E). The model contains 29 vertical levels from the surface to 50 hPa, with 14 levels below 2 km for the fully description of the vertical structure of planetary boundary layer (PBL).

The Carbon Bond Mechanism Z (CBM-Z) is selected as the gas-phase chemical mechanism (Zaveri and Peters, 1999), and the full 8-bin MOSAIC (Model for Simulating Aerosol Interactions and Chemistry) aerosol module with aqueous chemistry is used to simulate aerosol evolution (Zaveri et al., 2008). The photolysis rates are calculated by the Fast-J scheme (Wild et al., 2000). Other major physical parameterizations used in this study are listed in Table 1.

The initial and boundary meteorological conditions are provided by the National Centers for Environmental Prediction (NCEP) Final Analysis data with a spatial resolution of  $1^\circ \times 1^\circ$ . In order to limit the model bias of simulated meteorological fields, the four-dimensional data assimilation (FDDA) is used with the nudging coefficient of  $3.0 \times 10^{-4}$  for wind, temperature and humidity (no analysis nudging is applied for the inner domain) (Lo et al., 2008; Otte, 2008). Chemical initial and boundary conditions are obtained from the Model for Ozone and Related chemical Tracers, version 4 (MOZART-4) forecasts (Emmons et al., 2010).

Anthropogenic emissions in Episode1 are taken from the 2010 MIX Asian emission inventory, and the Multi-resolution Emission Inventory for China (MEIC) is used in Episode2 and Episode3 (<http://www.meicmodel.org/>) (Li et al., 2017a). These emission inventories provide emissions of sulfur dioxide (SO<sub>2</sub>), nitrogen oxides (NO<sub>x</sub>), carbon monoxide (CO), non-methane volatile organic compounds (NMVOCs), carbon dioxide (CO<sub>2</sub>), ammonia (NH<sub>3</sub>), black carbon (BC), organic carbon (OC), PM<sub>10</sub> (particulate matter with aerodynamic diameter is 10  $\mu$ m and less) and PM<sub>2.5</sub>. Emissions are aggregated from four sectors, including power generation, industry, residential, and transportation, with  $0.25^\circ \times 0.25^\circ$  spatial resolution. Biogenic emissions are calculated online by the Model of Emissions of Gases and Aerosols from Nature (MEGAN) (Guenther et al., 2006).

## 2.2 Numerical experiments

To quantify the impacts of API and ARF on O<sub>3</sub>, three case simulations have been conducted: (1) BASE – the base simulation coupled with the interactions between aerosol and radiation, which includes both impacts of API and ARF; (2) NOAPI – the same as the BASE case, but the impact of API is turned off ( aerosol optical properties are set to zero in the photolysis module), following Wu et al. (2020); (3) NOALL – both the impacts of API and ARF are turned off (removing the mass of aerosol species when calculating aerosol optical properties in the optical module), following Qiu et al. (2017). The differences between BASE and NOAPI (i.e., BASE minus NOAPI) represent the impacts of API. The contributions from ARF can be obtained by comparing NOAPI and NOALL (i.e., NOAPI minus NOALL). The combined effects of API and ARF on O<sub>3</sub> concentrations can be quantitatively evaluated by the differences between BASE and NOALL (i.e., BASE minus NOALL).

All the experiments in Episode1, Episode2 and Episode3 are conducted from 26 July to 3 August 2014, 6-13 July 2015 and 3-11 June 2016, respectively, with the first 40 hours as the model spin-up in each case. Simulation results from the BASE cases of the three episodes are used to evaluate the model performance.

## 2.3 Observational data

Simulation results are compared with meteorological and chemical measurements. The surface-layer meteorological data (2 m temperature (T<sub>2</sub>), 2 m relative humidity (RH<sub>2</sub>), and 10 m wind speed (WS<sub>10</sub>)) with the temporal resolution of 3 h at ten stations (Table S1) are obtained from NOAA's National Climatic Data Center (<https://gis.ncdc.noaa.gov/maps/ncei/cdo/hourly>). The radiosonde data of temperature at 08:00 and 20:00 LST in Beijing (39.93 °N, 116.28 °E) are provided by the University of Wyoming (<http://weather.uwyo.edu/>). Observed hourly concentrations of PM<sub>2.5</sub> and O<sub>3</sub> at thirty-two sites (Table S2) in North China are collected from the China National Environmental Monitoring Center (CNEMC). The photolysis rate of nitrogen dioxide (J[NO<sub>2</sub>]) measured at the Peking University site (39.99 °N, 116.31 °E) is also used to evaluate the model performance. More details

about the measurement technique of  $J[\text{NO}_2]$  can be found in Wang et al. (2019). The aerosol optical depth (AOD) at Beijing site (39.98°N, 116.38°E) is provided by AERONET (level 2.0, <http://aeronet.gsfc.nasa.gov/>). The AOD at 675 nm and 440 nm are used to derive the AOD at 550 nm to compare with the simulated ones.

## 2.4 Integrated process rate analysis

Integrated process rate (IPR) analysis has been widely used to quantify the contributions of different processes to  $\text{O}_3$  variations (Goncalves et al., 2009; Gao et al., 2016b; Tang et al., 2017; Gao et al., 2018). In this study, four physical/chemical processes are considered, including vertical mixing (VMIX), net chemical production (CHEM), horizontal advection (ADVH), and vertical advection (ADVZ). VMIX is initiated by turbulent process and closely related to PBL development, which influences  $\text{O}_3$  vertical gradients. CHEM represents the net  $\text{O}_3$  chemical production (chemical production minus chemical consumption). ADVH and ADVZ represent transport by winds (Gao et al., 2016b). In this study, we define ADV as the sum of ADVH and ADVZ.

## 3 Model evaluation

Reasonable representation of observed meteorological and chemical variables by the WRF-Chem model can provide foundation for evaluating the impacts of aerosols on surface-layer ozone concentrations. The model results presented in this section are taken from the BASE cases in the three episodes. The concentrations of air pollutants are averaged over the thirty-two observation sites in Beijing, Tianjin and Baoding. To ensure the data quality, the mean value for each time is calculated only when concentrations are available at more than sixteen sites, as did in Li et al. (2019a).

### 3.1 Chemical simulations

Figure 2 shows the temporal variations of observed and simulated  $\text{PM}_{2.5}$  and  $\text{O}_3$  concentrations over North China for the three episodes. As shown in Fig. 2, the temporal variations of observed  $\text{PM}_{2.5}$  can be well performed by the model with correlation coefficients (R) of 0.66, 0.56 and 0.73 and normalized mean bias (NMB)

of -19.2%, -3.9% and 30.4% during Episode1, Episode2 and Episode3, respectively. The model also tracks well the diurnal variation of  $O_3$  over the North China, with R of 0.86, 0.91 and 0.86 and NMB of -12.0%, 0.4% and 1.6% for Episode1, Episode2 and Episode3, respectively.

Figure S1 shows the correlation between observed and simulated AOD at 550 nm in Beijing. In the WRF-Chem model, the AOD at 550 nm are calculated by using the values at 400 and 600 nm according to the Angstrom exponent. Analyzing Fig. S1, the model can reproduce the observed AOD with R of 0.7 and NMB of 7.9%.

### 3.2 Meteorological simulations

Figure 3 shows the time series of observed and simulated  $T_2$ ,  $RH_2$ ,  $WS_{10}$  and  $J[NO_2]$  during the three episodes. The observed  $T_2$ ,  $RH_2$ ,  $WS_{10}$  are averaged over the ten meteorological observation stations, and the  $J[NO_2]$  are measured at Peking University. Most of the monitored  $J[NO_2]$  in Episode3 are unavailable, so the comparison of  $J[NO_2]$  in Episode3 is not shown. Generally, the model can depict the temporal variations of  $T_2$  fairly well with R of 0.98 and the mean bias (MB) of -1.9~0.9 °C. For  $RH_2$ , the R and MB are 0.91~0.97 and -4.0%~1.9%, respectively. Although WRF-Chem model overestimates  $WS_{10}$  with the MB of 0.6~0.9 m s<sup>-1</sup>, the R for  $WS_{10}$  is 0.70~0.89 and the root-mean-square error (RMSE) is 0.9~1.5 m s<sup>-1</sup>, which is smaller than the threshold of model performance criteria (2 m s<sup>-1</sup>) proposed by Emery et al. (2001). The positive bias in wind speed can also be reproduced in other studies (Zhang et al., 2010; Gao et al., 2015; Liao et al., 2015; Qiu et al., 2017). The predicted  $J[NO_2]$  agrees well with the observations with R of 0.97~0.98 and NMB of 6.8%~6.9%. We also conduct comparisons of observed and simulated temperature profiles at 08:00 and 20:00 LST in Beijing during the three episodes (Fig. S2). The vertical profiles of observed temperature can be well captured by the model in these three complex air pollution episodes. Generally, the WRF-Chem model can reasonably reproduce the temporal variations of observed meteorological parameters.

## 4 Results

It is known that co-occurrence of  $PM_{2.5}$  and  $O_3$  pollution is frequently observed

nowadays over China (Dai et al., 2021). The complex air pollution characterized by high  $\text{PM}_{2.5}$  and  $\text{O}_3$  levels has already received widespread attention from both scientists and policy-makers. Therefore, we examine the impacts of aerosol-radiation interactions on  $\text{O}_3$  concentrations with a special focus on the complex air pollution areas (CAPAs, Fig. S3) in the three episodes, where the mean simulated daily  $\text{PM}_{2.5}$  and MDA8 (maximum daily 8-h average)  $\text{O}_3$  concentrations are larger than  $75 \mu\text{g m}^{-3}$  and 80 ppb, respectively, based on the National Ambient Air Quality Standards (<http://www.mee.gov.cn>).

#### 4.1 Impacts of aerosol-radiation interactions on meteorology

Figure 4 shows the impacts of aerosol-radiation interactions on shortwave radiation at the surface (BOT\_SW), shortwave radiation in the atmosphere (ATM\_SW), PBLH, and  $\text{WS}_{10}$  during the daytime (08:00-17:00 LST) from Episode1 to Episode3. Analyzing the results of the interactions between aerosol and radiation (the combined impacts of API and ARF), BOT\_SW is decreased over the entire simulated domain in the three episodes with the decreases of  $93.2 \text{ W m}^{-2}$  (20.5%),  $100.3 \text{ W m}^{-2}$  (19.5%) and  $92.4 \text{ W m}^{-2}$  (19.2%) over CAPAs, respectively. Contrary to the changes in BOT\_SW, ATM\_SW is increased significantly in the three episodes with the increases of  $72.8 \text{ W m}^{-2}$  (25.3%),  $85.2 \text{ W m}^{-2}$  (29.0%) and  $73.7 \text{ W m}^{-2}$  (26.4%) over CAPAs, respectively. The decreased BOT\_SW perturbs the near-surface energy flux, which weakens convection and suppresses the development of PBL (Li et al., 2017b). The mean PBLHs over CAPAs are decreased by 129.0 m (13.0%), 249.0 m (20.9%) and 224.6 m (19.0%), respectively.  $\text{WS}_{10}$  exhibits overall reductions over CAPAs and is calculated to decrease by  $0.12 \text{ m s}^{-1}$  (3.6%),  $0.05 \text{ m s}^{-1}$  (1.6%), and  $0.12 \text{ m s}^{-1}$  (3.0%) for the three episodes, respectively. We also examine the changed meteorological variables caused by API and ARF respectively. As shown in Fig. S4 and S5, API has little impact on meteorological variables; which means the major contributor to the meteorology variability is ARF.

#### 4.2 Impacts of aerosol-radiation interactions on photolysis

Figure 5 shows the spatial distributions of mean daytime surface-layer  $\text{PM}_{2.5}$

concentrations simulated by BASE cases and the changes in  $J[\text{NO}_2]$  and  $J[\text{O}^1\text{D}]$  due to aerosol-radiation interactions from Episode1 to Episode3. When the combined impacts (API and ARF) are considered,  $J[\text{NO}_2]$  and  $J[\text{O}^1\text{D}]$  are decreased over the entire domain in the three episodes, and the spatial patterns of changed  $J[\text{NO}_2]$  and  $J[\text{O}^1\text{D}]$  are similar to that of simulated  $\text{PM}_{2.5}$ . Analyzing the three simulated episodes, the surface  $J[\text{NO}_2]$  averaged over CAPAs are decreased by  $1.8 \times 10^{-3} \text{ s}^{-1}$  (40.5%),  $2.0 \times 10^{-3} \text{ s}^{-1}$  (36.8%) and  $1.8 \times 10^{-3} \text{ s}^{-1}$  (36.0%), respectively. The decreased surface  $J[\text{O}^1\text{D}]$  over CAPAs are  $6.1 \times 10^{-6} \text{ s}^{-1}$  (48.8%),  $6.3 \times 10^{-6} \text{ s}^{-1}$  (41.4%) and  $5.7 \times 10^{-6} \text{ s}^{-1}$  (44.6%), respectively. Figure S6 exhibits the impacts of API and ARF on surface  $J[\text{NO}_2]$  and  $J[\text{O}^1\text{D}]$ . Conclusions can be summarized that  $J[\text{NO}_2]$  and  $J[\text{O}^1\text{D}]$  are significantly modified by API and little affected by ARF.

#### 4.3 Impacts of aerosol-radiation interactions on $\text{O}_3$

Figure 6 shows the changes in surface-layer  $\text{O}_3$  due to API, ARF, and the combined effects (denoted as ALL) from Episode1 to Episode3. As shown in Fig. 6(a1-a3), API alone leads to overall surface  $\text{O}_3$  decreases over the entire domain with average reductions of 8.5 ppb (10.1%), 9.0 ppb (10.6%) and 8.3 ppb (10.4%) over CAPAs in the three episodes, respectively. The changes can be explained by the substantially diminished UV radiation due to aerosol loading, which significantly weakens the efficiency of photochemical reactions and restrains  $\text{O}_3$  formation. However, the decreased surface  $\text{O}_3$  concentrations due to ARF are only 2.9 ppb (3.1%, Fig. 6(b1)), 1.0 ppb (1.2%, Fig. 6(b2)) and 1.0 ppb (1.1%, Fig. 6(b3)) for the three episodes, which indicates that API is the dominant way for  $\text{O}_3$  reduction related to aerosol-radiation interactions. Fig. 6(c1-c3) presents the combined effects of API and ARF. Generally, aerosol-radiation interactions decrease the surface  $\text{O}_3$  concentrations by 11.4 ppb (13.5%), 10.0 ppb (11.9%) and 9.3 ppb (11.6%) averaged over CAPAs in the three episodes, respectively.

#### 4.4 Influencing mechanism of aerosol-radiation interactions on $\text{O}_3$

Figure 7a shows mean results of the three episodes (Episode1, Episode2 and Episode3) in diurnal variations of simulated daytime surface-layer  $\text{O}_3$  concentrations

from BASE, NOAPI and NOALL cases averaged over CAPAs. All the experiments (BASE, NOAPI and NOALL) present O<sub>3</sub> increases from 08:00 LST. It is shown that the simulated O<sub>3</sub> concentrations in BASE case increase more slowly than that in NOAPI and NOALL cases. To explain the underlying mechanisms of API and ARF impacts on O<sub>3</sub>, we quantify the variations in contributions of different processes (ADV, CHEM, and VMIX) to O<sub>3</sub> by using the IPR analysis.

Figure 7b shows hourly surface O<sub>3</sub> changes induced by each physical/chemical process (i.e., ADV, CHEM, and VMIX) in BASE case averaged from Episode1 to Episode3. The significant positive contribution to the hourly variation in O<sub>3</sub> is contributed by VMIX, and the contribution reaches the maximum at about 09:00 LST. Since VMIX increases the surface O<sub>3</sub> concentrations by transporting O<sub>3</sub> from aloft (where O<sub>3</sub> concentrations are high) to the surface layer (Tang et al., 2017; Xing et al., 2017; Gao et al., 2018). The CHEM process makes negative contributions at around 09:00 and 16:00 LST, which means that the chemical consumption of O<sub>3</sub> is stronger than the chemical production. At noon, the net chemical contribution turns to be positive due to stronger solar UV radiation. The contribution from all the processes (NET, the sum of VMIX, CHEM, and ADV) to O<sub>3</sub> variation is peaked at the noon and then becomes weakened. After sunset (17:00 LST), the NET contribution turns to be negative over CAPAs, leading to O<sub>3</sub> decrease.

Figure 7c shows the changes in hourly process contributions caused by API averaged from Episode1 to Episode3. The chemical production of O<sub>3</sub> is suppressed significantly due to aerosol impacts on photolysis rates. The weakened O<sub>3</sub> chemical production decreases the contribution from CHEM, and results in a negative value of CHEM\_DIF (-3.2 ppb h<sup>-1</sup>). In contrast to CHEM\_DIF, the contribution from changed VMIX (VMIX\_DIF) to O<sub>3</sub> concentration due to API is always positive, and the mean value is +3.0 ppb h<sup>-1</sup>. The positive change in VMIX due to API may be associated with the different vertical gradient of O<sub>3</sub> between BASE and NOAPI cases (Gao et al., 2020), as shown in Fig. 8a. The impact of API on ADV process is relatively small (-0.26 ppb h<sup>-1</sup>). NET\_DIF, namely the sum of VMIX\_DIF, CHEM\_DIF and ADV\_DIF, indicates the differences in hourly O<sub>3</sub> changes caused by API. As shown in Fig. 7c,

NET\_DIF is almost negative during the daytime over CAPAs with the mean value of -0.46 ppb h<sup>-1</sup>. This is because the decreases in CHEM and ADV are larger than the increases in VMIX caused by API; the O<sub>3</sub> decrease is mainly attributed to the significantly decreased contribution from CHEM. The maximum difference in O<sub>3</sub> between BASE and NOAPI appears at 11:00 LST with a value of -11.1 ppb (Fig. 7a).

Figure 7d shows the impacts of ARF on each physical/chemical process contribution to the hourly O<sub>3</sub> variation averaged from Episode1 to Episode3. At 08:00 LST, the change in VMIX due to ARF is large with a value of -3.5 ppb h<sup>-1</sup>, resulting in a net negative variation with all processes considered. The decrease in O<sub>3</sub> reaches the maximum with the value of 5.2 ppb at around 08:00 LST over CAPAs (Fig. 7a). During 09:00 to 16:00 LST, the positive VMIX\_DIF (mean value of +0.20 ppb h<sup>-1</sup>) or the positive CHEM\_DIF (mean value of +0.55 ppb h<sup>-1</sup>) is the major process to positive NET\_DIF. The positive VMIX DIF is related to the evolution in boundary layer during the daytime. The VOCs/NO<sub>x</sub> ratio is calculated to classify sensitivity regimes and to indicate the possible O<sub>3</sub> responses to changes in VOCs and/or NO<sub>x</sub> concentrations. O<sub>3</sub> production is VOC-limited if the ratio is less than 4, and is NO<sub>x</sub>-limited if the ratio is larger than 15 (Edson et al., 2017; Li et al., 2017c). The ratio of VOCs/NO<sub>x</sub> ranging around 4-15 indicates a transitional regime, where ozone is nearly equally sensitive to both species (Sillman, 1999). As shown in Fig. S7, (a-f), O<sub>3</sub> is mainly formed under the VOC-limited and the transition regimes in CAPAs. As shown in Figs. S7(g-i) and S7(j-l), both the surface concentrations of VOCs and NO<sub>x</sub> are increased when the impacts of ARF are considered. Thus, the contribution of CHEM in NOAPI is larger than that in NOALL.

When both impacts of API and ARF are considered, the variation pattern of the difference in hourly process contribution shown in Fig. 7e is similar to that in Fig. 7c, which indicates that API is the dominant factor to surface-layer O<sub>3</sub> reduction.

Figure 8 presents the vertical profiles of simulated daytime O<sub>3</sub> concentrations in three cases (BASE, NOAPI, and NOALL), and the differences in contributions from each physical/chemical process to hourly O<sub>3</sub> variations caused by API, ARF and the combined effects averaged over CAPAs from Episode1 to Episode3. As shown in Fig.



8a, the O<sub>3</sub> concentration is lower in BASE than that in other two scenarios (NOAPI and NOALL), especially at the lower 12 levels (below 863.0 m), owing to the impacts of aerosols (API and/or ARF).

The changes in each process contribution caused by API are presented in Fig. 8b. The contribution from CHEM\_DIF is -2.0 ppb h<sup>-1</sup> for the first seven layers (from 27.6 to 342.8 m). Conversely, the contribution from VMIX\_DIF shows a positive value under the 342.8 m (between the first layer to the seventh layer) with the mean value of +1.7 ppb h<sup>-1</sup>. The positive variation in VMIX due to API may be associated with the different vertical gradient of O<sub>3</sub> between BASE and NOAPI again. The contributions of changed advections (ADVH\_DIF and ADVZ\_DIF) are relatively small, with mean values of +0.07 and -0.21 ppb h<sup>-1</sup> below the first seven layers, which may result from small impact of API on wind field (Fig. S5(a4-c4)). The net difference is a negative value (-0.44 ppb h<sup>-1</sup>); API leads to O<sub>3</sub> reduction not only nearly surface but also aloft.

Figure 8c shows the differences in O<sub>3</sub> budget due to ARF. When the ARF is considered, the vertical turbulence is weakened and the development of PBL is inhibited, which makes VMIX\_DIF negative at the lower seven layers (below the 342.8 m) with a mean value of -0.64 ppb h<sup>-1</sup>, but the variation in CHEM caused by ARF is positive with a mean value of +0.72 ppb h<sup>-1</sup>. The enhanced O<sub>3</sub> precursors due to ARF can promote the chemical production of O<sub>3</sub> (Tie et al., 2009; Gao et al., 2018). The changes of ADVZ and ADVH (ADVZ\_DIF and ADVH\_DIF) caused by ARF are associated with the variations in wind field. When ARF is considered, the horizontal wind speed is decreased (Fig. S7S8(a)), which makes ADVH\_DIF positive at the lower twelve layers with a mean value of +0.25 ppb h<sup>-1</sup>. However, ADVZ\_DIF is negative at these layers with a mean value of -0.27 ppb h<sup>-1</sup> because aerosol radiative effects decrease the transport of O<sub>3</sub> from the upper to lower layers (Fig. S7S8(b)).

In Fig. 8d, the pattern and magnitude of the differences in process contributions between BASE and NOALL are similar to those caused by API, indicating the dominate contributor of API on O<sub>3</sub> changes. The impacts of API on O<sub>3</sub> both near the surface and aloft are greater than those of ARF.

Figure S8-S9 and S9-S10 detailed show the influencing mechanism of aerosol-

radiation interactions on O<sub>3</sub> in each episode. Similar variation characteristics can be found among the three episodes as the mean situation discussed above, with the larger impacts of API on O<sub>3</sub> both near the surface and aloft than those of ARF, indicating the major contributor of API on O<sub>3</sub> reduction related with aerosol-radiation interactions.

## 5 Conclusions and Discussions

In this study, the fully coupled regional chemistry transport model WRF-Chem is applied to investigate the impacts of aerosol-radiation interactions, including the impacts of aerosol-photolysis interaction (API) and the impacts of aerosol-radiation feedback (ARF), on O<sub>3</sub> during summertime complex air pollution episodes during 28 July to 3 August 2014 (Episode1), 8-13 July 2015 (Episode2) and 5-11 June 2016 (Episode3). Three sensitivity experiments are designed to quantify the respective and combined impacts from API and ARF. Generally, the spatiotemporal distributions of observed pollutant concentrations and meteorological parameters can be captured fairly well by the model with correlation coefficients of 0.56~0.91 for pollutant concentrations and 0.70~0.98 for meteorological parameters.

Sensitivity experiments show that aerosol-radiation interactions decrease BOT\_SW, WS<sub>10</sub>, PBLH, J[NO<sub>2</sub>], and J[O<sup>1</sup>D] by 92.4~100.3 W m<sup>-2</sup>, 0.05~0.12 m s<sup>-1</sup>, 129.0~249.0 m,  $1.8 \times 10^{-3}$ ~ $2.0 \times 10^{-3}$  s<sup>-1</sup>, and  $5.7 \times 10^{-6}$ ~ $6.3 \times 10^{-6}$  s<sup>-1</sup> over CAPAs, and increase ATM\_SW by 72.8~85.2 W m<sup>-2</sup>. The changed meteorological variables and weakened photochemistry reaction further reduce surface-layer O<sub>3</sub> concentrations by up to 9.3~11.4 ppb, with API and ARF contributing 74.6%~90.0% and 10.0%~25.4%, respectively.

We further examine the influencing mechanism of aerosol-radiation interactions on O<sub>3</sub> by using integrated process rate analysis. API can directly affect O<sub>3</sub> by reducing the photochemistry reactions within the lower several hundred meters and therefore amplify the O<sub>3</sub> vertical gradient, which promotes the vertical mixing of O<sub>3</sub>. The reduced photochemistry reactions of O<sub>3</sub> weaken the chemical contribution and reduce surface O<sub>3</sub> concentrations, even though the enhanced vertical mixing can partly counteract the reduction. ARF affects O<sub>3</sub> concentrations indirectly through the

changed meteorological variables, e.g., the decreased PBLH. The suppressed PBL can weaken the vertical mixing of O<sub>3</sub> by turbulence. Generally, the impacts of API on O<sub>3</sub> both near the surface and aloft are greater than those of ARF, indicating the dominant role of API on O<sub>3</sub> reduction related with aerosol-radiation interactions.

This study provides a detailed understanding of aerosol impacts on O<sub>3</sub> through aerosol-radiation interactions (including both API and ARF). The results imply that future PM<sub>2.5</sub> reductions will lead to O<sub>3</sub> increases due to weakened aerosol-radiation interactions. A recent study emphasized the need for controlling VOCs emissions to mitigate O<sub>3</sub> pollution (Li et al., 2019b). Therefore, tighter controls of O<sub>3</sub> precursors (especially VOCs emissions) are needed to counteract future O<sub>3</sub> increases caused by weakened aerosol-radiation interactions, and the contributions of different mitigation strategies with the impacts of aerosol-radiation interactions to O<sub>3</sub> air quality will be discussed detailedly in our future work.

There are some limitations to this work. The lack of secondary organic aerosol (SOA) may result in uncertainties in our model results. First, the absence of SOA would lead to underestimation of API and ARF because of the low bias in AOD, so the reductions in O<sub>3</sub> by API and/or ARF could have been underestimated. Second, the lack of SOA would lead to weaker heterogeneous reactions, leading to higher O<sub>3</sub> concentrations (Li et al., 2019c). The net effect of these two processes needs further studies.~~The uncertainty of the lack of secondary organic aerosols (SOA), and the missing mechanisms of some heterogeneous reactions may result in large uncertainties in the final simulation results. Gao et al. (2017) added some SOA formation mechanisms into the MOSAIC module by using the volatility basis set (VBS) in WRF-Chem and found that the surface PM<sub>2.5</sub> concentrations in urban Beijing were reduced by 1.9 μg m<sup>-3</sup> due to the weakened ARF effect during Asia-Pacific Economic Cooperation (APEC). Similar magnitude can also be found in Zhou et al. (2019) (1.8 μg m<sup>-3</sup>) who did not consider the impacts of SOA in WRF-Chem when analyzing the impacts of weakened ARF on PM<sub>2.5</sub> during APEC. Therefore, more work should be conducted to explore the impacts of ARF on PM<sub>2.5</sub> and O<sub>3</sub> concentrations under consideration of SOA in future.~~

## **Data availability**

The observed hourly surface concentrations of air pollutants are derived from the China National Environmental Monitoring Center (<http://www.cnemc.cn>). The observed surface meteorological data are obtained from NOAA's National Climatic Data Center (<https://gis.ncdc.noaa.gov/maps/ncei/cdo/hourly>). The radiosonde data are provided by the University of Wyoming (<http://weather.uwyo.edu/>). The photolysis rates of nitrogen dioxide in Beijing are provided by Xin Li ([li\\_xin@pku.edu.cn](mailto:li_xin@pku.edu.cn)). The aerosol optical depth in Beijing is obtained from the AERONET level 2.0 data collection (<http://aeronet.gsfc.nasa.gov/>). The simulation results can be accessed by contacting Lei Chen ([chenlei@nuist.edu.cn](mailto:chenlei@nuist.edu.cn)) and Hong Liao ([hongliao@nuist.edu.cn](mailto:hongliao@nuist.edu.cn)).

## **Author contributions**

HY, LC, and HL conceived the study and designed the experiments. HY and LC performed the simulations and carried out the data analysis. JZ, WW, and XL provided useful comments on the paper. HY prepared the paper with contributions from all co-authors.

## **Competing interests**

The authors declare that they have no competing interests.

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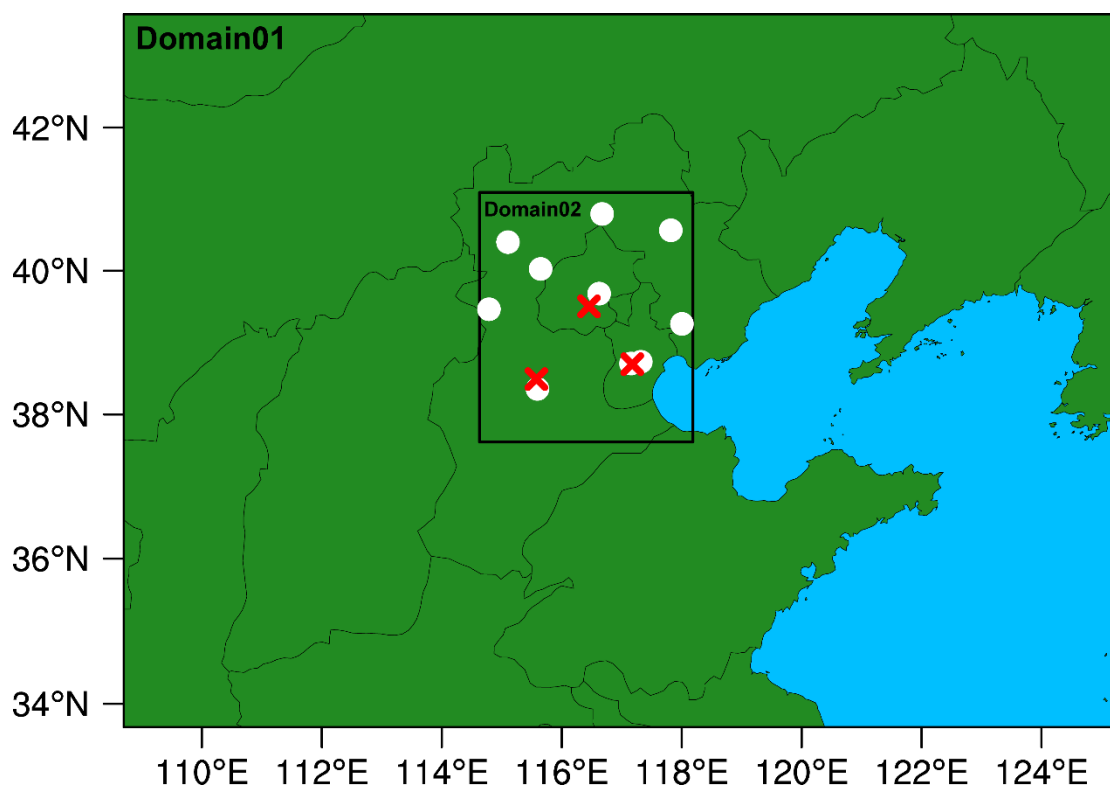
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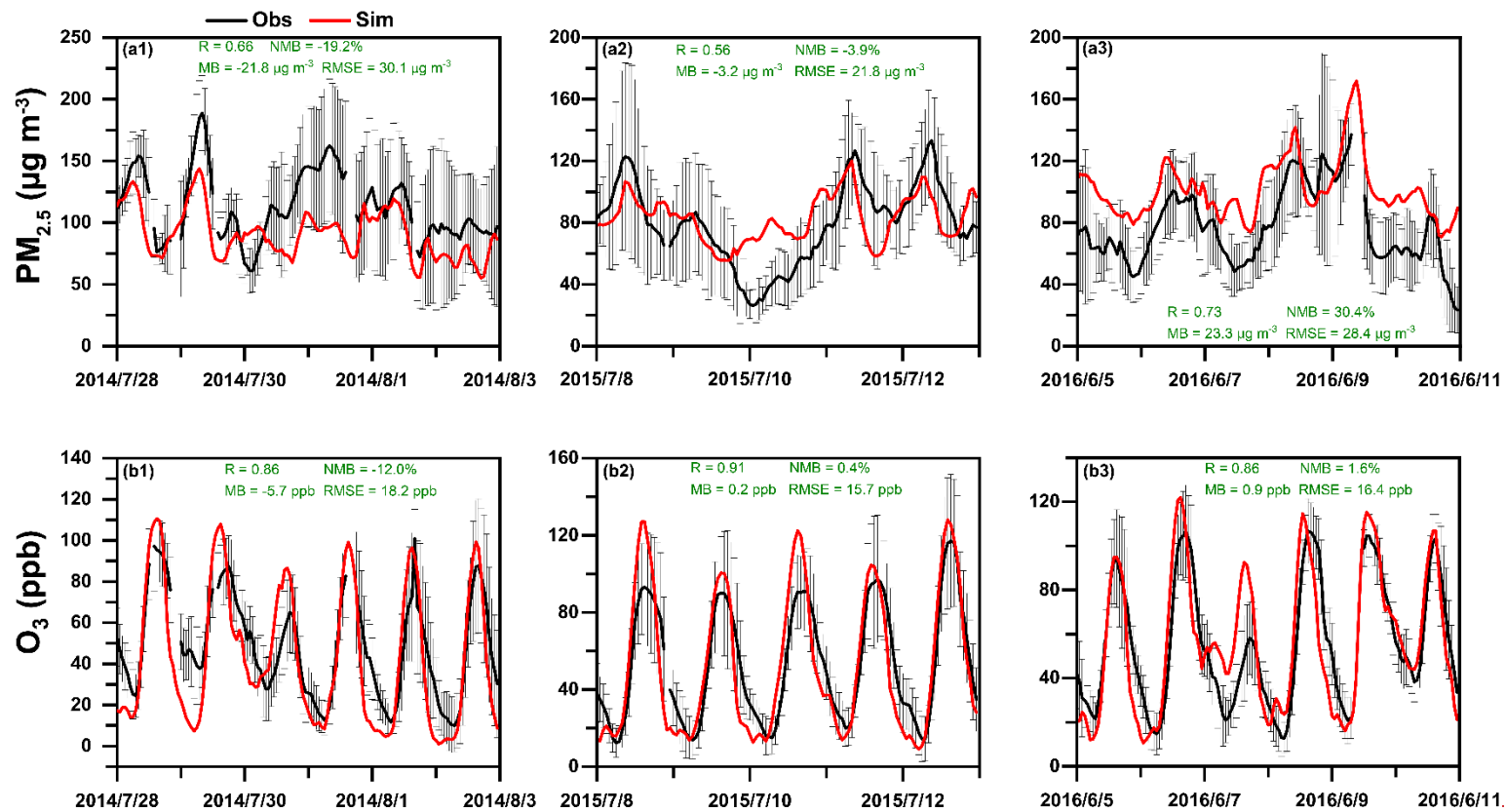
1 **Table 1.** Physical parameterization options used in the simulation.

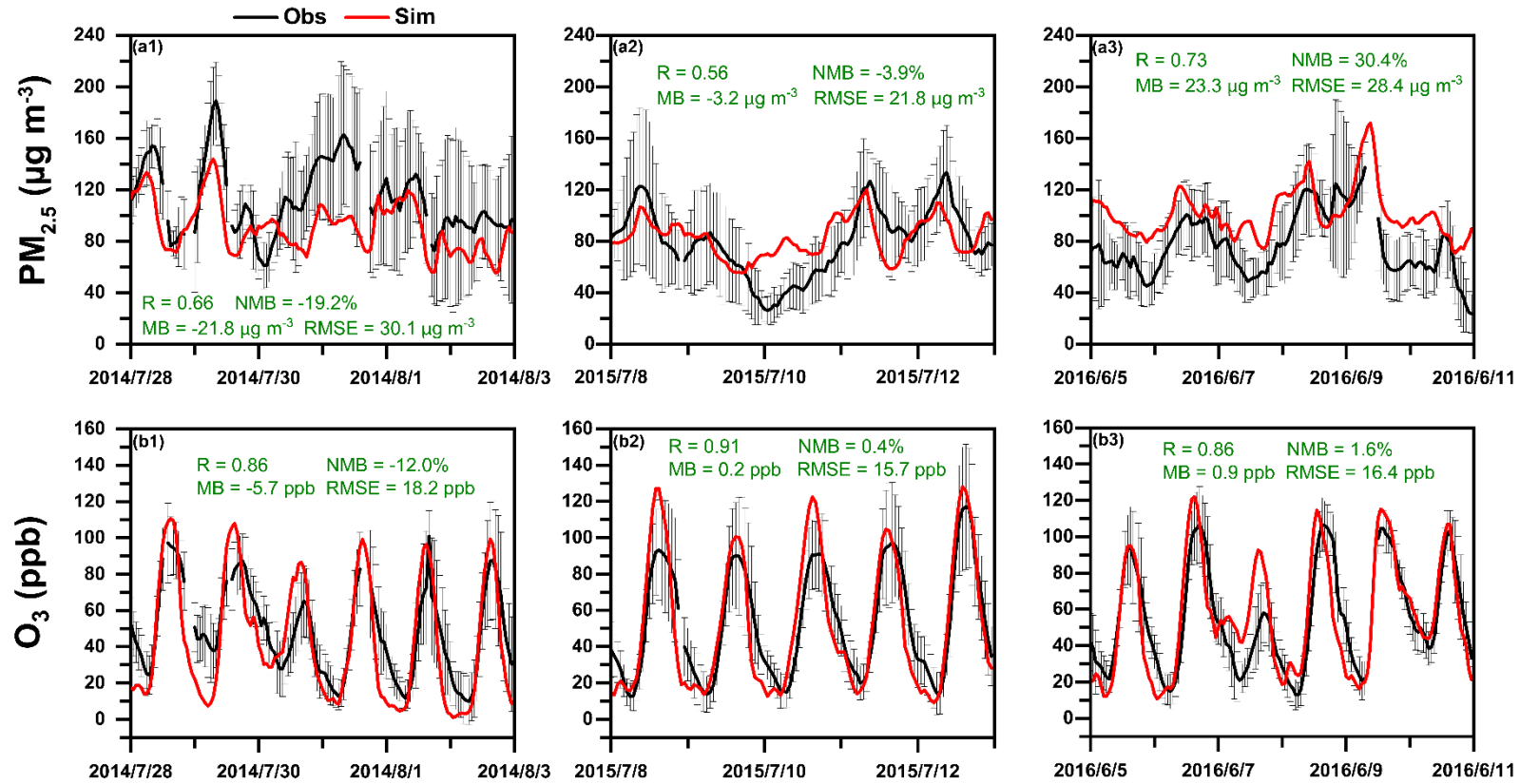
| Options                    | Schemes   |
|----------------------------|---|
| Microphysics scheme        | Lin (Purdue) scheme (Lin et al., 1983)                  |
| Cumulus scheme             | Grell 3D ensemble scheme                                |
| Boundary layer scheme      | Yonsei University PBL scheme (Hong et al., 2006)        |
| Surface layer scheme       | Monin-Obukhov surface scheme (Foken, 2006)              |
| Land-surface scheme        | Unified Noah land-surface model (Chen and Dudhia, 2001) |
| Longwave radiation scheme  | RRTMG (Iacono et al., 2008)                             |
| Shortwave radiation scheme | RRTMG (Iacono et al., 2008)                             |

2



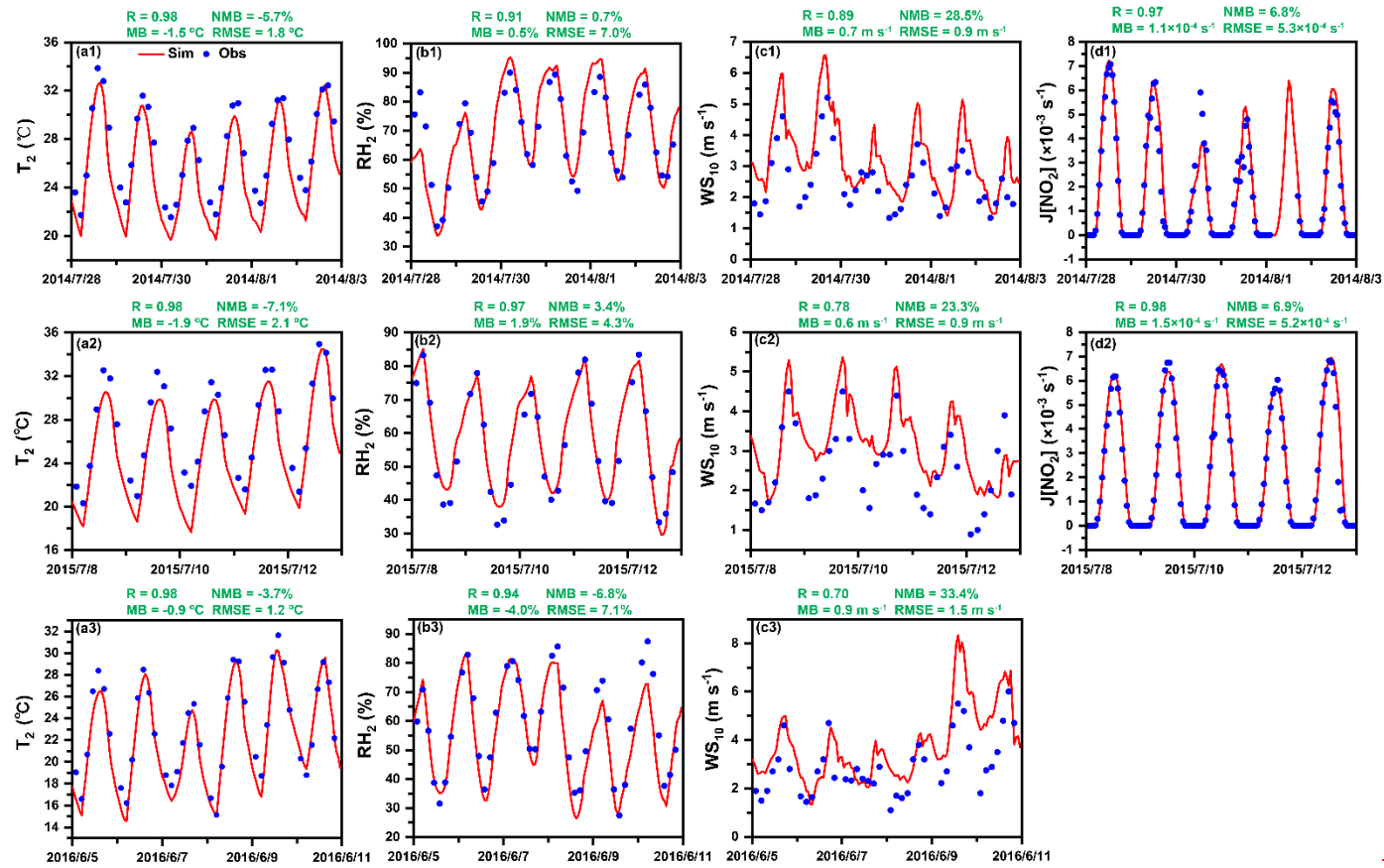
**Figure 1.** Map of the two WRF-Chem modeling domains with the locations of meteorological (white dots) and environmental (red crosses) observation sites used for model evaluation.

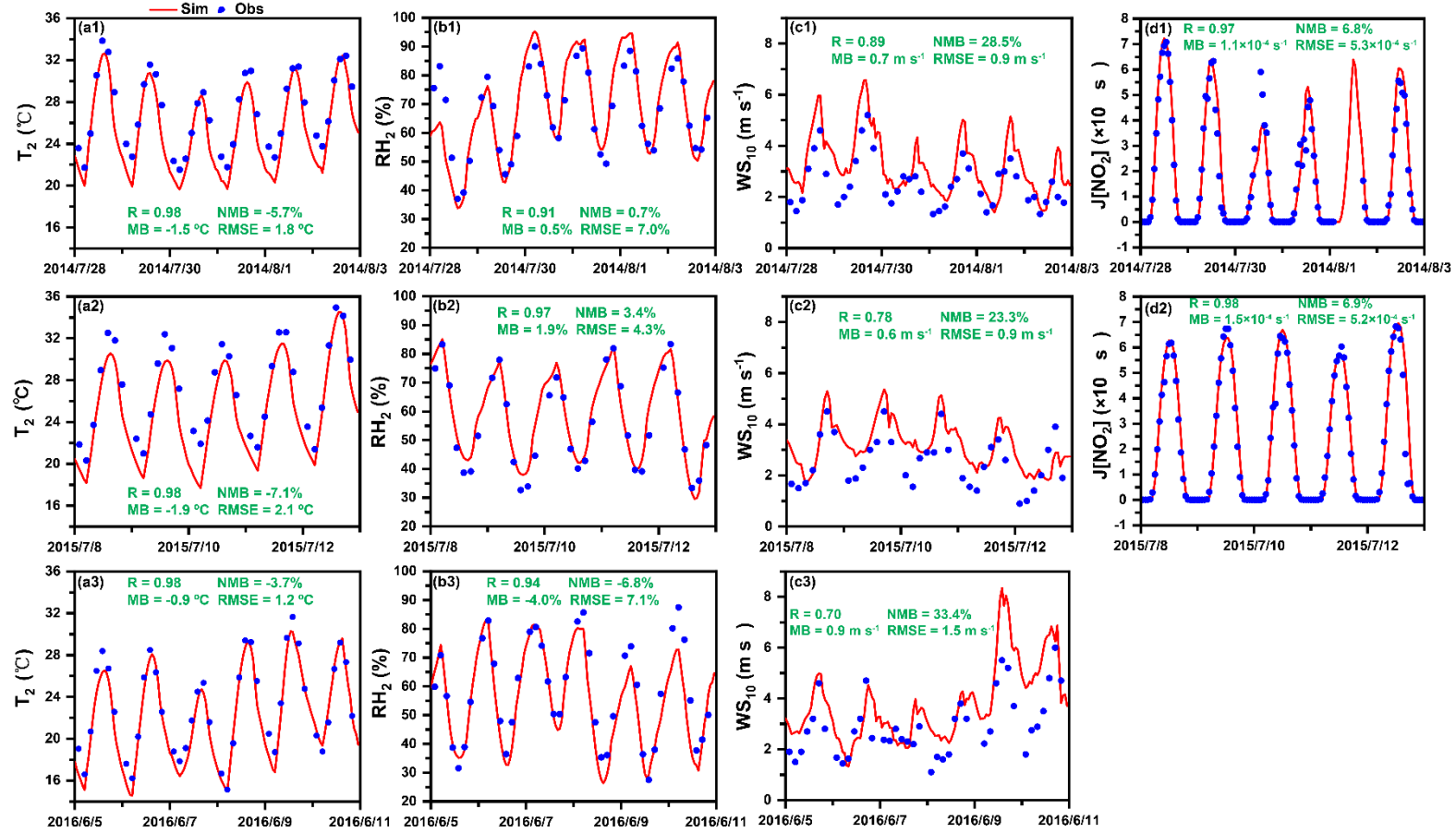




**Figure 2.** Time series of observed (black) and simulated (red) hourly surface (a)  $PM_{2.5}$  and (b)  $O_3$  concentrations averaged over the thirty-two observation sites in Beijing, Tianjin, and Baoding during 28 July to 3 August 2014 (Episode1, a1-b1), 8-13 July 2015 (Episode2, a2-b2) and 5-11 June 2016 (Episode3, a3-b3). The error bars represent the standard deviations. The calculated correlation coefficient (R), mean bias (MB), normalized mean bias (NMB) and root-mean-square error (RMSE) are also shown.

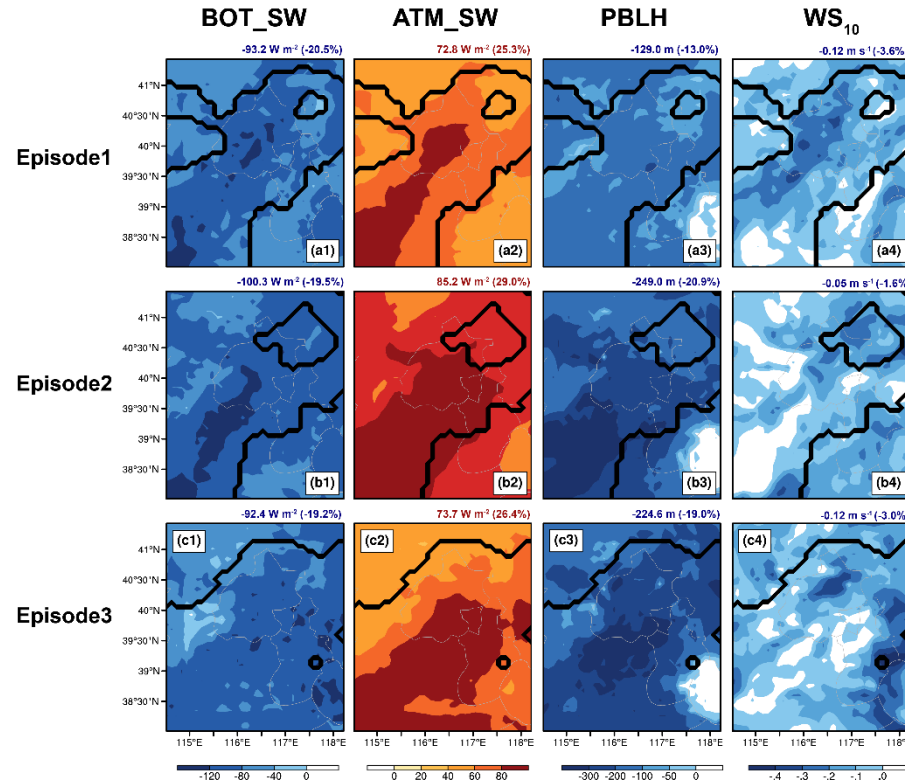




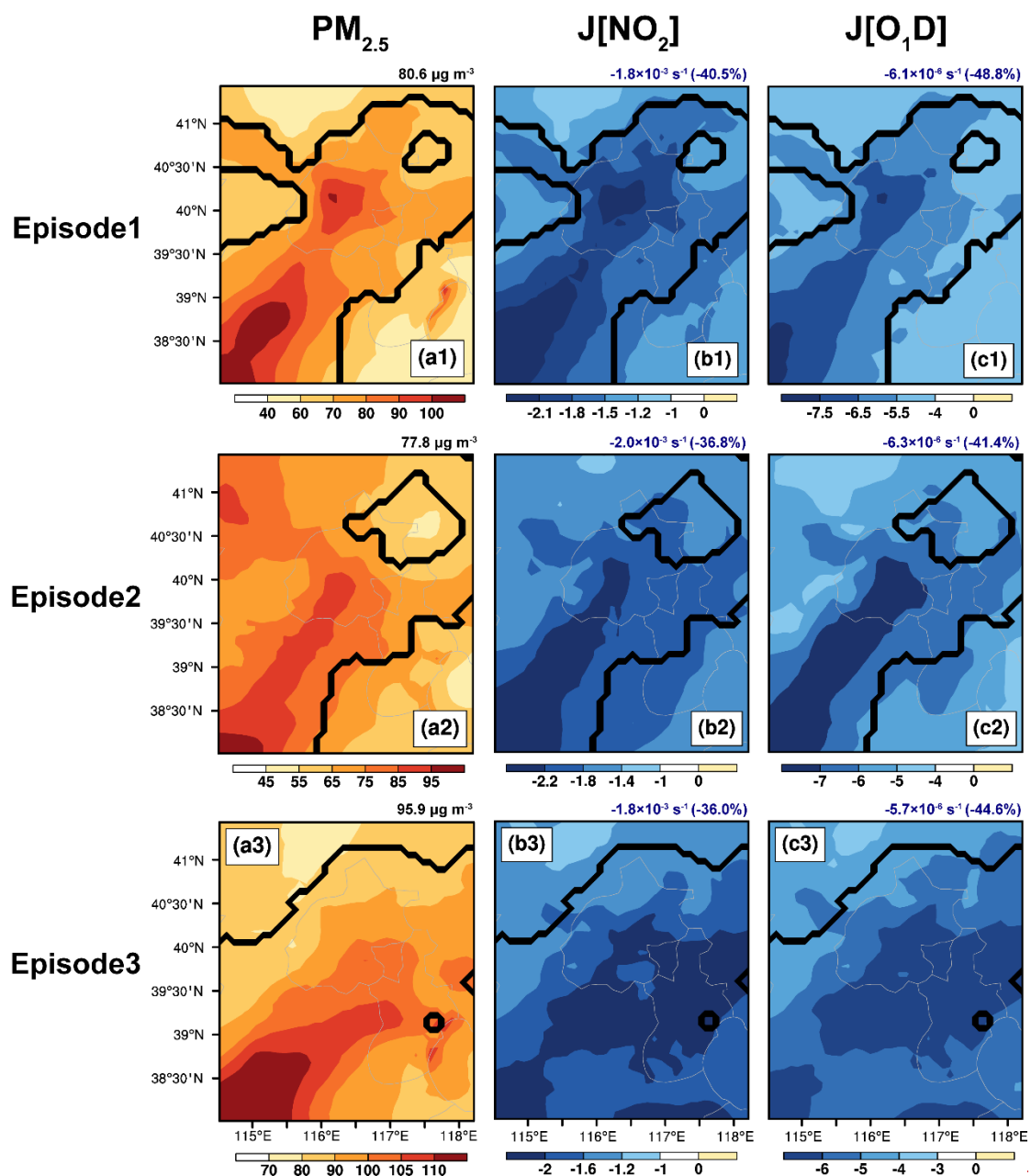


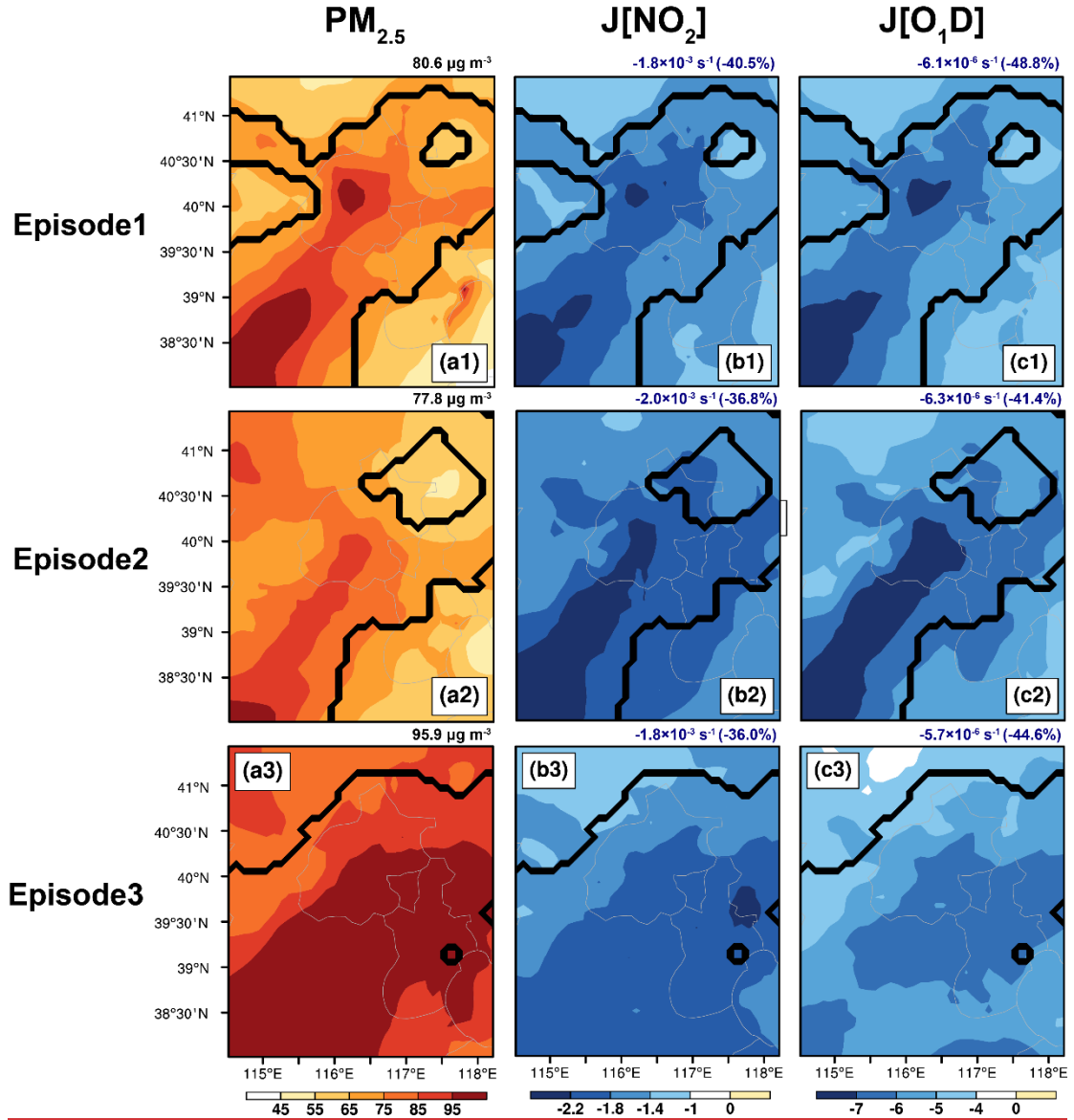
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2 **Figure 3.** Time series of 3-hourly observed (blue dots) and hourly simulated (red lines) (a) 2-m temperature ( $T_2$ ), (b) 2-m relative humidity  
3 ( $RH_2$ ), (c) wind speed at 10 m ( $WS_{10}$ ) averaged over ten meteorological observation stations, and (d) surface photolysis rate of  $NO_2$  ( $J[NO_2]$ )  
4 during 28 July to 3 August 2014 (Episode1, a1-d1), 8-13 July 2015 (Episode2, a2-d2) and 5-11 June 2016 (Episode3, a3-c3). The calculated  
5 correlation coefficient (R), mean bias (MB), normalized mean bias (NMB) and root-mean-square error (RMSE) are also shown.

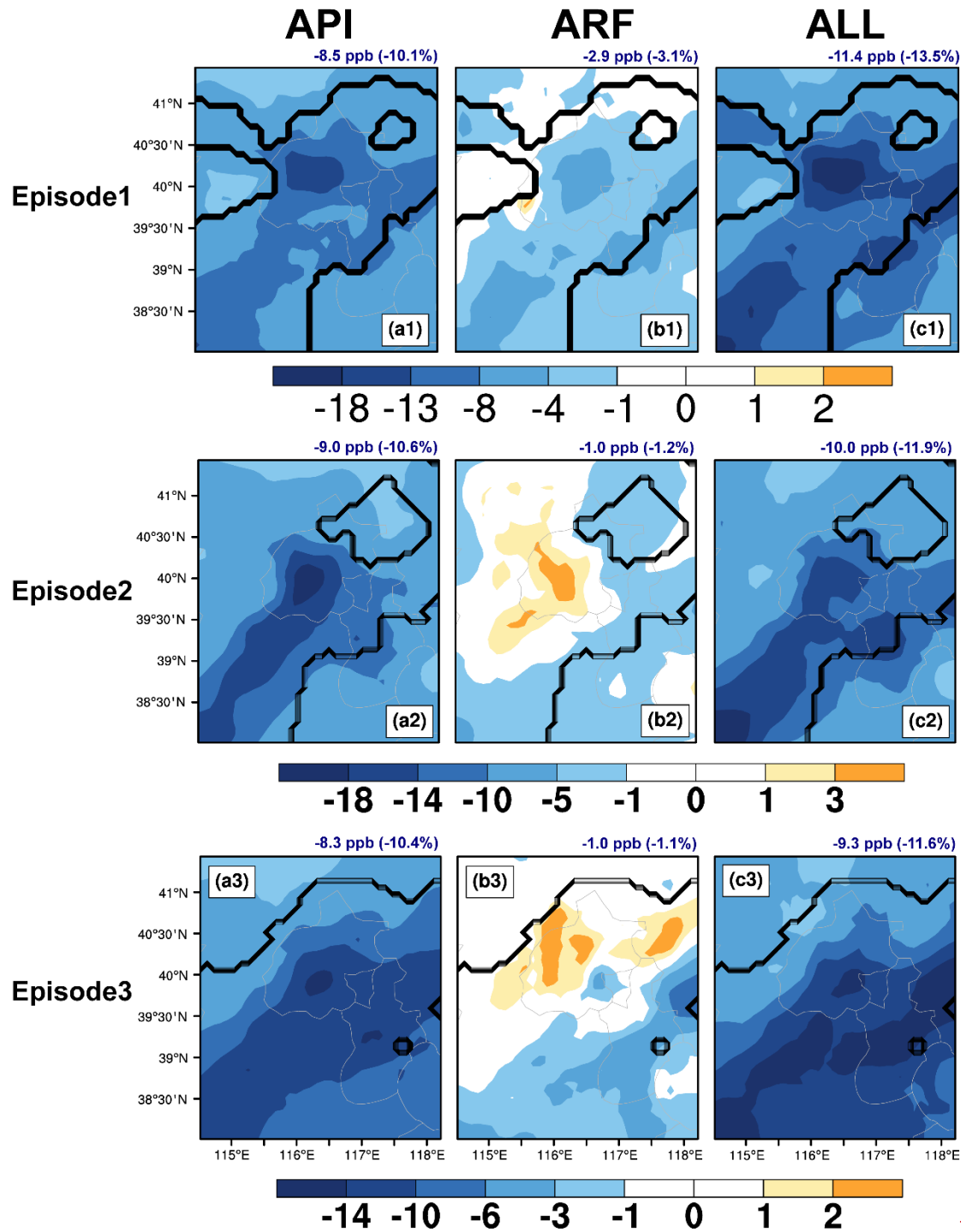


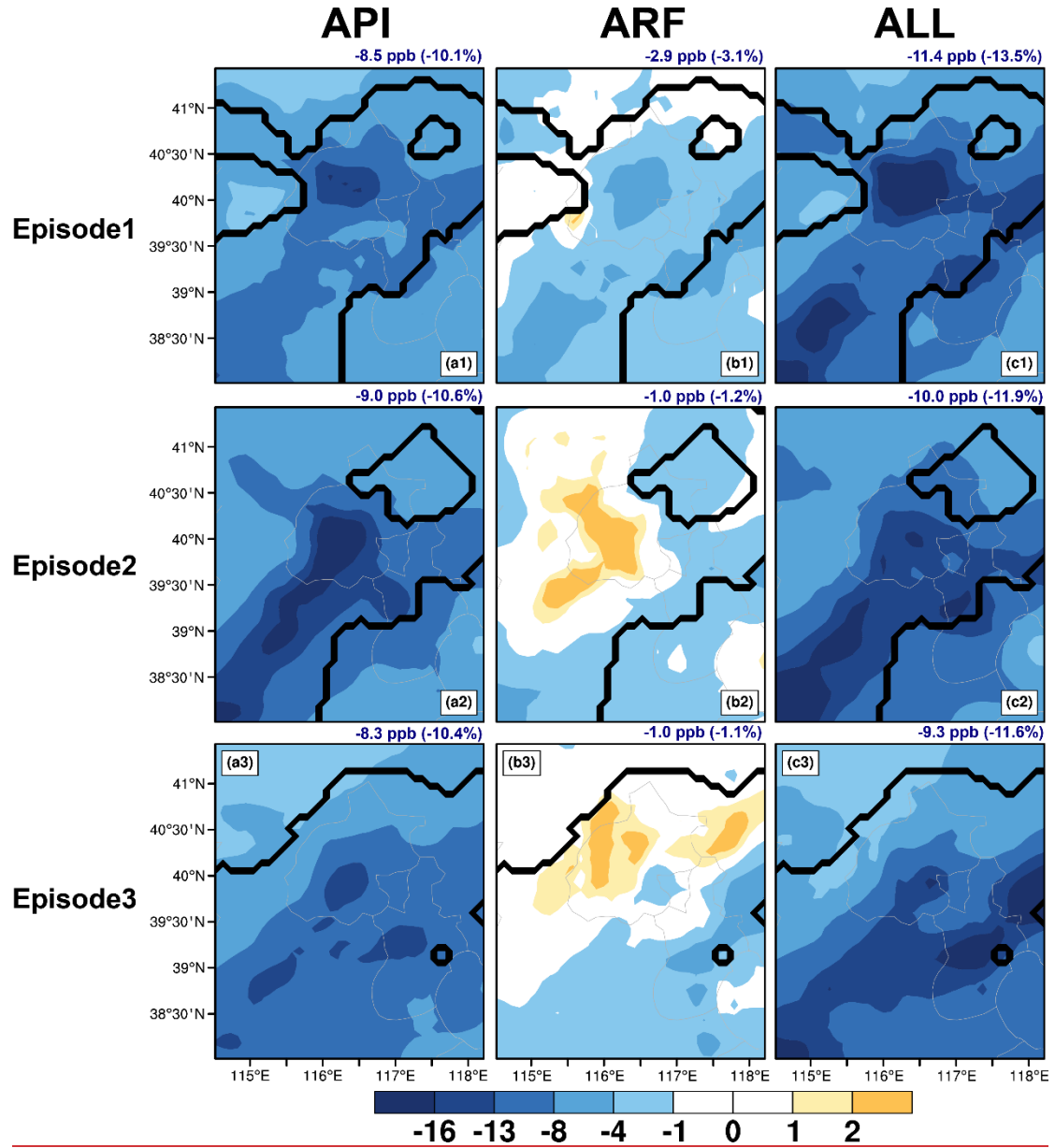
**Figure 4.** The impacts of aerosol-radiation interactions on shortwave radiation at the surface (BOT\_SW), shortwave radiation in the atmosphere (ATM\_SW), PBL height (PBLH), and 10-m wind speed (WS<sub>10</sub>) in the daytime (08:00-17:00 LST) during 28 July to 3 August 2014 (Episode1), 8-13 July 2015 (Episode2) and 5-11 June 2016 (Episode3). The regions sandwiched between two black lines are defined as the complex air pollution areas (CAPAs) where the mean daily PM<sub>2.5</sub> and MDA8 O<sub>3</sub> concentrations in BASE case are larger than 75  $\mu\text{g m}^{-3}$  and 80 ppb. The calculated changes (percentage changes) averaged over CAPAs are also shown at the top of each panel.





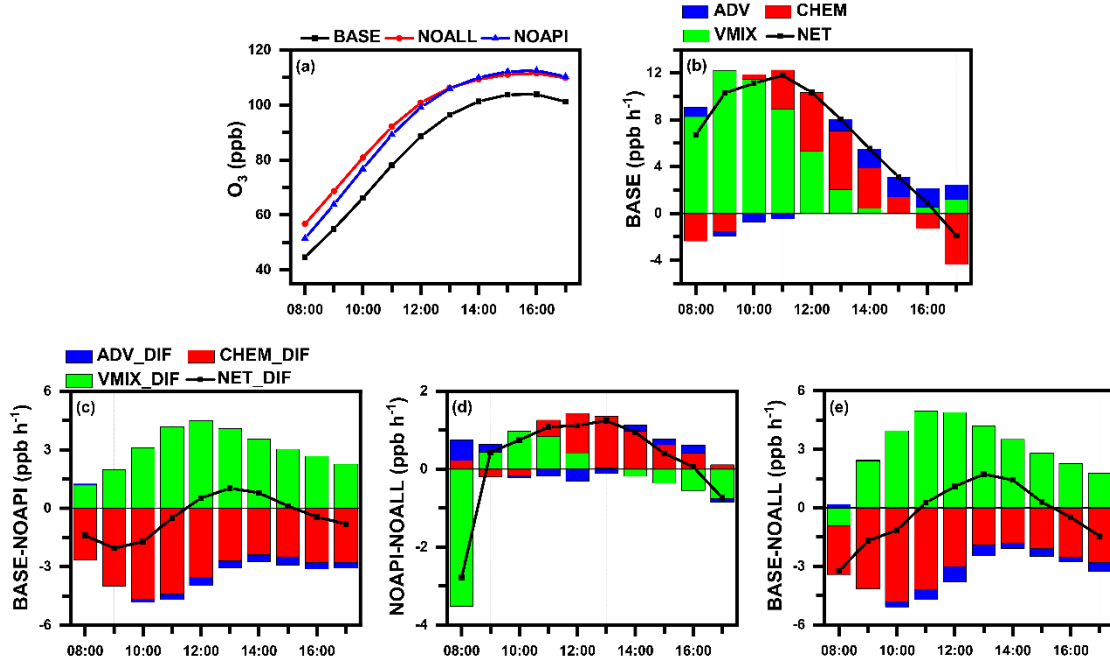
**Figure 5.** Spatial distributions of (a) simulated surface-layer  $PM_{2.5}$  concentrations in BASE cases, and the changes in surface (b)  $J[NO_2]$  and (c)  $J[O^1D]$  due to aerosol-radiation interactions in the daytime (08:00-17:00 LST) during 28 July to 3 August 2014 (Episode1), 8-13 July 2015 (Episode2) and 5-11 June 2016 (Episode3). The calculated values (percentage changes) averaged over CAPAs are also shown at the top of each panel.





**Figure 6.** The changes in surface-layer ozone due to (a) aerosol-photolysis interaction (API), (b) aerosol-radiation feedback (ARF), and (c) the combined effects (ALL, defined as API+ARF) in the daytime (08:00-17:00 LST) during 28 July to 3 August 2014 (Episode1), 8-13 July 2015 (Episode2) and 5-11 June 2016 (Episode3). The calculated mean changes (percentage changes) averaged over CAPAs are also shown at the top of each panel.

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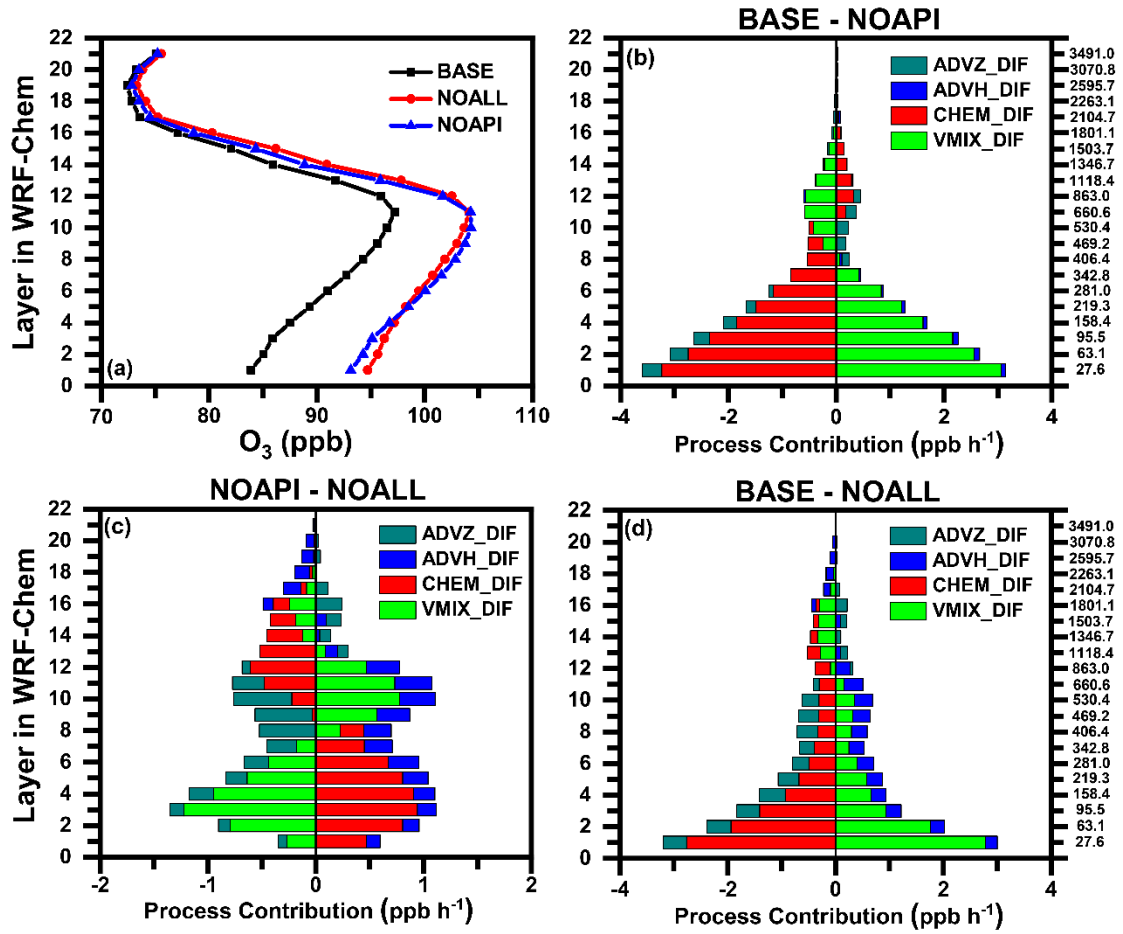


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3 **Figure 7.** Temporal evolution characteristics of aerosol-radiation interactions on O<sub>3</sub>  
4 averaged over the three episodes. (a) Diurnal variations of simulated surface O<sub>3</sub>  
5 concentrations in BASE (black dotted line), NOAPI (blue dotted line), and NOALL  
6 (red dotted line) cases over CAPAs. (b) The hourly surface O<sub>3</sub> changes induced by  
7 each physical/chemical process using the IPR analysis method in BASE case. (c-e)  
8 Changes in hourly surface O<sub>3</sub> process contributions caused by API (BASE minus  
9 NOAPI), ARF (NOAPI minus NOALL), and ALL (BASE minus NOALL) over  
10 CAPAs during the daytime (08:00-17:00 LST). The black lines with squares denote  
11 the net contribution of all processes (NET, defined as VMIX+CHEM+ADV).  
12 Differences of each process contribution are denoted as VMIX\_DIF, CHEM\_DIF,  
13 ADV\_DIF, and NET\_DIF.

14





**Figure 8.** The impacts of aerosol-radiation interactions on vertical O<sub>3</sub> averaged over the three episodes. (a) Vertical profiles of simulated O<sub>3</sub> concentrations in BASE (black dotted line), NOAPI (blue dotted line), and NOALL (red dotted line) cases over CAPAs. (b-d) Changes in O<sub>3</sub> budget due to API, ARF, and ALL over CAPAs during the daytime (08:00-17:00 LST). Differences of each process contribution are denoted by ADVZ\_DIF, ADVH\_DIF, CHEM\_DIF, and VMIX\_DIF.