

South America 2020 Regional Smoke Plume: Intercomparison with previous years, impact on solar radiation and the role of Pantanal biomass burning season

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Abstract. The 2020 biomass burning season in Brazil was marked by an atypical amount of fire across the Pantanal biome, which led to high levels of smoke within the biome and downwind areas. The present study analyzes fire counts and smoke over Pantanal in 2020, comparing this particular year's data with those from the previous seventeen years (2003-2019). Taking as references the most polluted years in this period, the regional smoke plume and its impact on surface solar radiation were also evaluated. In 2020, the regional smoke plume core covered an area of ~2.6 million km² at the peak of the burning season, an area well above that of the previous six years, but smaller than areas observed in a more remote past, as in 2007 and 2010 (> 5.0 million km²). The smoke loading was lower (mean AOD_{550 nm} ~0.7) than that of 2007 and 2010 (mean AOD_{550 nm} ~1.0). The plume radiation absorption efficiency, when compared with the previous year's plumes, did not present significant differences. Regarding the Pantanal burning season, it revealed some atypical features. Fire counts were up to 3.0 times higher than for the years from 2003 to 2019. Smoke loading over Pantanal, which is typically a fraction of that over Amazonia, in 2020 was higher than over Amazonia, an indication that local smoke surpassed smoke advection from upwind regions. The observed intraseasonal variability of smoke over Pantanal revealed to be largely driven by the nature of the burned areas in the biome. From September on, there was a significant increase of fire count in conservation and indigenous areas, where higher biomass density is present, which would explain the larger smoke plumes over Pantanal, even during October when the fire count was reduced. In October, the biome was covered by a thick smoke layer, which resulted in a mean deficit of surface solar radiation up to 200 Wm⁻². Despite Pantanal biomes' massive burning in 2020, the regional smoke plume was not far from its climatological features. Nevertheless, Pantanal 2020 burning season represents the worst combination of a climate extreme applied to a fire-prone environment, coupled with inadequately enforced environmental regulations, from which there is much to be learned.

32 **1 Introduction**

34 In South America, the Regional Smoke Plume (RSP) produced during the biomass burning season is the most important
signature of anthropogenic activities from the point of view of injection of pollutants into the atmosphere at continental scale
(Prins and Menzel, 1992, Artaxo et al., 1998, Freitas et al., 2004, Longo et al., 2007). Its geographical dimension and pollutants
36 loading, along with its climate and air quality effects, have been comprehensively studied during the last decades (Freitas et
al., 2004, Ignotti et al., 2010, Artaxo et al., 2013, Rosario et al., 2013, Chen et al., 2013, Sena et al., 2015, Moreira et al. 2017,
38 Thornhill et al., 2018). The RSP affects the regional climate reducing the availability of solar energy at the surface and
perturbing the cloud microphysics and atmospheric thermodynamics and chemistry (Procópio et al., 2001, Schafer et al., 2002,
40 Yamasoe et al., 2006, Rosario et al., 2013, Moreira et al. 2017). In the context of air quality, especially near biomass burning
areas, RSP impact on air pollution levels surpasses the most polluted urban areas of the continent (Ignotti et al., 2010, Sena et
42 al., 2013, Rosario et al., 2013). Systematically, along the years, the southern portion of the Amazon rainforest and the eastern
part of the Cerrado ecosystem have been the major sources of pollutants to the RSP. Consequently, in general, these are the
44 most affected areas by the smoke pollution (Artaxo et al., 2013, Pereira et al., 2016), in particular, when compared with the
remaining Brazilian biomes, Pantanal, Caatinga, Mata Atlantica and Pampas (Figure 1). From the mid-2000s to the earliest
46 years of the 2010 decade, following the trends observed in deforestation and fire counts in Amazonia and Cerrado, the RSP
loading and dimension were significantly reduced (Reddington et al. 2015). However, in recent years, due to the increase in
48 deforestation rates and occurrence of major biomass burning events, such as those occurred during 2020 in Pantanal and related
to the massive regional smoke plume in 2019, that contribute to darkened cities in the southeastern region of Brazil, there is a
50 need to perform a historical contextualization of these recent events. Therefore, one can realize the advance achieved in the
task of bringing down the RSP loading and also to highlight the risks of jeopardizing such achievement with the current
52 environmental governance scenario in Brazil. The 2020 biomass burning in Brazil, amid the COVID-19 pandemic, has been
claimed to be one of the worst in recent years, an evaluation largely driven by the extension of fire count and burned area
54 across Pantanal biome (WWF, 2020, FSP, 2020, BBC, 2020, NYT, 2020, Le Monde, 2020, FSP, 2021). Despite a significant
number of news and several literatures produced focusing on the 2020 biomass burning season (Libonati et al., 2020, Marengo
56 et al., 2021, Pletsch et al., 2021), there is still a lack of clarity about what was effectively exceptional during this Brazil
particular biomass burning season. From an integrated perspective of fire counts and RSP loading, and focusing on the biomes
58 most affected by the RSP, Amazonia, Cerrado and Pantanal, this manuscript sets out to analyze how the 2020 biomass burning
season compares with previous biomass burning seasons. To what extent fire count in Amazonia, Cerrado and Pantanal biomes
60 during the 2020 biomass burning season and the produced RSP surpassed the previous years? Did Pantanal biome's exceptional
2020 biomass burning season resemble an exceptional RSP? These are some of the questions that the present analysis sought
62 to address in order to contribute to a comprehensive contextualization of Amazonia, Cerrado and Pantanal 2020 biomass
burning season. Additionally, we also explore the 2020 RSP loading and its impact on surface solar radiation when compared
64 with the RSPs from previous highly polluted years. From here on, the article is organized as follows: Section 2 presents a brief

description of the study region, a summary of the data used and the methods adopted; Section 3 presents the results divided in
66 2 sub-sections: Sub-section 3.1 analyses the intraseasonal and interannual variability of fire counts, aerosol optical depth
(AOD) at 550 nm for the three biomes, Amazon tropical rainforest, Cerrado ecosystem, and Pantanal, for the period from 2003
68 to 2020. Sub-section 3.2 focuses on September and October of the most polluted years of the considered period, providing an
interannual analysis of the spatial distribution of smoke loading, wind field and the anomaly of downward solar radiation at
70 the surface under cloudless conditions. In Section 4, the main conclusions are summarized.

2 Study Region, Data and Methods

72 This study focused on the three Brazilian biomes most affected by the regional smoke plume: Amazon tropical forest (or
Amazonia), Cerrado ecosystem and Pantanal wetlands (Figure 1). The Pampa grasslands biome, in the south of the country,
74 and the rainforest of Mata Atlantica, distributed along the Brazilian coast, from the northeast to the southern states, are less
exposed to the RSP. They are eventually affected by the occurrence of transport of smoke toward southeast and southern
76 Brazil. Caatinga biome, located in the northeast of Brazil and composed of shrubland and dry forests, is in general upwind of
Amazonia and Cerrado, therefore it is less affected by the RSP. According to Pivello (2011), the Amazon tropical forest is a
78 fire-sensitive ecosystem and cannot tolerate fire, with consequent mortality of trees after repeated fires. Although Pantanal and
Cerrado are considered fire-dependent biomes and, therefore, are more adapted to fire occurrence, human-driven fire frequency
80 increased significantly in Brazil, which can lead to land degradation and biodiversity loss (Pivello, 2011). Despite its smaller
domain, when compared to Amazonia and Cerrado, Pantanal consists of one of the world's largest freshwater wetland
82 ecosystems, characterized by highly complex hydrological regime that support two major river systems, the Cuiabá and the
Paraguay rivers, and an abundance of vegetation and animal life (Alho, 2008, Alho et al., 2019). Large scale biomass burning
84 within its domain and the RSP affect the local and regional climate equilibrium and also pose a direct threat to the health of
Pantanal's singular biodiversity and population. MODIS (Moderate Resolution Imaging Spectroradiometer) Enhanced
86 Vegetation Index (EVI, Kidan 2021) was applied in land use characterization across Pantanal during the 2020 burning season,
specially to highlight conservation and indigenous areas, which tend to present higher biomass density.

88 To analyze the RSP spatial dimension and loading, monthly mean aerosol optical depth at 550 nm, from 2003 to 2020, taken
from MODIS atmosphere Level-3 products from Aqua satellite (MYD08, Platnick et al., 2017) was analyzed. The RSP border
90 definition was somehow arbitrary, since the dispersion of the smoke is a continuous process and it is impossible to track the
effective ending of its influence. To provide a perspective on the plume spatial dimension and loading at its core, isolines of
92 AOD at 550 nm of 0.5 were used as reference to evaluate the plume dimension. The value of 0.5 was found convenient since
it is above typical monthly average values of AOD at 550 nm observed over major urban areas of South America, which
94 prevents a misidentification related to these urban plumes. Based on this threshold to define the main area of the plume, the
RSP loading reference was taken as the mean AOD at 550 nm within the delimited area. The mean AOD at 550 nm for each
96 year was also calculated for the domains of the biomes Amazonia, Cerrado and Pantanal. AERONET (Holben et al., 1998)

aerosol optical depth at 550 nm, estimated using Angström exponent at channels 440 nm and 675 nm and AOD at 440 nm, single scattering albedo at 440 nm (Dubovik and King, 2000) and aerosol direct radiative forcing at the surface (Garcia et al., 2012), level 1.5, from three sites located in different parts of Brazil and one in Bolivia, were included in the analysis. Such data were used to help identify differences in the aerosol intrinsic optical properties from different years and locations within the RSP influence area. Since no level 2.0 data are available for the most recent period, we compared AERONET data from levels 1.5 and 2.0 from previous years in Figure S.1 (Supplement), to demonstrate the good quality of level 1.5 data as well.

To better contextualize the 2020 biomass burning season, we also analyzed the interannual variability (2003 to 2020) of the fire counts in each of the considered biomes. Table 1 presents the variables used and their respective applications, data sources and references. Fire counts were obtained from the Brazilian Space Agency (Instituto Nacional de Pesquisas Espaciais - INPE). These are important data to explore the spatial dynamic of smoke emission sources and, to some extent, to analyze the distribution of illegal biomass burning, which has been the main driver of the regional smoke plume in Brazilian biomes. The fire counts used are based on Moderate Resolution Spectroradiometer (MODIS) data and two algorithms. In summary, according to Morisette et al. (2005), the algorithms use empirically derived thresholds based on digital numbers (DNs) at channels 20 (at around 3.7 μm) and 9 (around 440 nm). In the daytime algorithm, pixels are classified as “fire” if two conditions are satisfied: DNs higher than 3000 at channel 20 and lower than 3300 at channel 9. The nighttime algorithm requires only one condition: DNs higher than 3000 at channel 20.

To estimate the direct effect of the aerosol plume on the reduction of surface solar radiation (SSR), instantaneous retrievals of CERES (Cloud and the Earth’s Radiant Energy System) Single Scanner Footprint - Level 3, during Aqua overpasses were used. This variable is estimated by a combination of CERES SW upward irradiance at the top of the atmosphere retrievals, ancillary meteorological data, surface, aerosol, gases and cloud properties. These inputs are used in the Langley parameterized shortwave algorithm (LPSA) (Kratz et al., 2020), to estimate shortwave downward irradiance at the surface from 0.2 to 5.0 μm . The output is gridded in 1° by 1° latitude/longitude resolution. Cloud-free scenes, when cloud fraction is less than 0.1% in a given footprint (NASA, 2018), in the presence of aerosols were selected to account only for the aerosol direct effect on radiation. The aerosol model used in CERES surface irradiance calculations is very briefly described in Kratz et al. (2020). It uses near-real time daily AODs from the Model for Atmospheric Transport and Chemistry (MATCH) (Collins et al., 2001, Rasch et al., 1997). The Optical Properties of Aerosols and Clouds (OPAC) Global Aerosol Data Set (GADS) was used to obtain the intrinsic aerosol properties, such as single scattering albedo and asymmetry parameter (Hess et al., 1998).

Wind circulation at 850 hPa plays a determinant role on the transport of smoke over South America (Freitas et al. 2004) and, therefore, on the regional smoke plume structure, especially for areas downwind of the main biomass burning areas, namely, south of Amazon rainforest and the western portion of Cerrado ecosystem. Due to its location, Pantanal biome is subject to smoke transport from Amazon Forest and Cerrado. The wind dataset to perform this analysis was taken from MERRA-2 (Modern-Era Retrospective analysis for Research and Applications, version 2) reanalysis (Gelaro, et al., 2017). The interannual variability of the fire counts and AOD were analyzed, based on monthly mean values (July, August, September and October), and from the perspective of the selected biomes. Subsequently, the interannual variability of single scattering albedo and mean

radiative forcing were analyzed. Those variables were also based on monthly mean values from AERONET stations located
132 in different positions of the regional plume. September and October were selected to carry out a more comprehensive
geographical analysis of the features and impact on SSR of the 2020 regional smoke plume. Targeting a comparison analysis,
134 RSPs of the most polluted years in the period analyzed were included in this analysis. The radiative impact of these plumes,
based on the anomaly in downward SSR, was also compared.

136 **3 Results**

138 **3.1 Fire count and smoke aerosol loading over Amazon, Cerrado and Pantanal: intraseasonal and interannual variability**

This section addresses the interannual and the intraseasonal variability of monthly fire counts and mean smoke aerosol loading
140 for the three selected biomes (Amazon, Cerrado and Pantanal) during the biomass burning season of the last 18 years, from
2003 to 2020. Figure 2 shows, for each biome and multiple years, the intraseasonal evolution (July, August, September and
142 October) of fire counts and smoke aerosol loading. The typical intra seasonality of the biomass burning season is observed for
most of all the 18 years as follows: from the transition month of July, the number of fire counts increased significantly in the
144 following months, in general, reaching its peak in September. After that, the number of fire counts decreased, but still remained
high, in October, until the next wet season. The burned area variability, not shown here, closely follows the variability of fire
146 counts. For the Amazonia biome, in terms of fire count and mean smoke aerosol loading, 2020 biomass burning season closely
followed the general features of the last nine years (2011 – 2020). Regarding the peak of smoke aerosols over Amazonia, in
148 the last 10 years, September 2017 overpowered all Septembers, including that of 2020. Looking further back in time, the
Amazonia biomass burning seasons of the years of 2004, 2005, 2007 and 2010 were the most fire active and polluted years,
150 with mean AOD at 550 nm over the biome reaching a value close to one, in some case almost twice as that observed in 2020.
A similar feature can be stated for fire count. Regarding the intraseasonal behavior, for all highly polluted biomass burning
152 seasons, September stood out as the most polluted month in Amazonia. For the Cerrado ecosystem, while fire count during the
2020 biomass burning season was not significantly different from the previous nine years (2011 – 2020) biomass burning
154 seasons, smoke aerosol loading was exceptionally high in October, but with precedent when one looks further in the past (ex.
2007). Similar to Amazonia, during the mid-2000's Cerrado presented its more fire active and polluted years (2004, 2005,
156 2007) in the period here analyzed, and including 2010. However, when one looks at the Pantanal biome biomass burning
seasons time series, 2020 was indeed an exceptional season in terms of fire count, with almost three and four times the fire
158 counts typically observed in August and September of the last decades, respectively. Those figures surpass the fire counts of
the years of 2005 and 2007, two of the most fire active biomass burning seasons in the last two decades in Pantanal. The 2020
160 high fire counts scenario for Pantanal is also seen in aerosol mean loading over the biome during September, with twice the
values observed in recent years (2011 - 2020), but not unprecedented in the history of Pantanal. For the years of 2004, 2005,
162 2007 and 2010, Pantanal smoke aerosol loading was as high as, or even higher than, that observed in the 2020 peak. In 2007,

during September, the highest aerosol loading was observed within the period analyzed. Unlike 2020, in the polluted years of 2000s and the year 2010, large smoke loading was also observed upwind of Pantanal, in the southern of Amazon Forest and western part of Cerrado, as can be seen in the Hovmoller diagram of AOD at 550 nm (Figure 3). This indicates the relevant contribution of smoke advection from these biomes to the high aerosol loading observed over Pantanal, especially during 2010 when local fire counts were relatively low. From 2011 to 2019, both fire counts and smoke levels in Pantanal were well below the observed values during the previous polluted years, even when high levels of smoke were observed in the upwind regions, for instance in 2017 (Figure 3). In that year, the core of the transport toward the south was over Paraguay. Regarding the high levels of smoke in the northern part of the Amazon basin at the end of the year 2015 (Figure 3), and which did not have a significant influence on Pantanal region, it is worth mentioning that it occurred during an El Nino event, when this part of the Amazon typically experiences drought scenarios and higher incidence of fire counts. Indeed, according to Marengo et al. (2017), the Amazon onset of the rainy season in 2015 occurred later than normal, and the region was characterized by drought in 2016. As in 2015 (Marengo et al., 2017) and 2020 (Marengo et al., 2021), drought conditions also were present, either across central or north region of Brazil, during the polluted years of 2005, 2007 and 2010 (Jimenez et al., 2018, Libonati et al., 2021), corroborating the critical role of climate variability combined with human factors to the exacerbation of biomass burning severity in Brazil. Regarding the years characterized by less polluted scenarios in Pantanal (2011-2019), they were consistently associated with relatively low local fire counts. This scenario changed during the 2020 biomass burning season, when the total fire counts at the peak of the biomass burning season in Pantanal surpassed all the previous analyzed years (Figure 2). According to Marengo et al. (2021), since the beginning of the fire activity monitoring in Brazil, in 1998, the largest number of fires over Pantanal was detected in 2005 and, in 2020, it was 76% higher. As mentioned, the level of aerosol loading over the biome at the peak of the 2020 burning season was similar to the early polluted years (Figure 2), in the middle of 2000s. However, opposite to the former years, smoke loading in Pantanal in 2020 peaked in October, when the smoke loading upwind of Pantanal, mainly in the Amazon biome, was atypically much lower than that over Pantanal (Figure 3). This suggests that Pantanal itself was the main source of smoke in its domain, and with advection from outer regions playing a secondary role. October followed also a highly polluted September, when there was an explosion in fire counts across Pantanal. The increase in smoke over Pantanal, in mid-September, occurred simultaneously with the peak of smoke in Amazonia, suggesting that at this time the smoke plume over Pantanal had also received contribution from biomass burning emission from Amazonia (Figure 3). While the aerosol loading observed during the 2020 biomass burning over Pantanal was not unique from historical perspective, there are relevant differences between 2020 and previous polluted years. Regarding the intraseasonal variability of smoke aerosol loading, October 2020 in Pantanal can be evaluated as exceptional when compared with October from the previous years of the time series. An interesting aspect worth emphasizing for October 2020 is that Pantanal presented much lower fire counts compared to August and September (Figure 3). However, August's higher fire count in Pantanal did not translate into a high level of smoke over the wetland biome. Further analysis was done to clarify these aspects, focusing on the correlation between smoke loading over Pantanal and local fire count (Figure 4a) and on the correlation between smoke over Pantanal and over Amazonia (Figure 4b). The scatter plot of smoke loading over Amazonia versus smoke over Pantanal shows

that the smoke loading over Pantanal has a stronger relationship with smoke over Amazon than with fire counts within the
198 biome. This suggests that, in general, smoke over Pantanal is more affected by advection than by local fires. However, 2020
stands out as an outlier. Typically, mean AOD at 550 nm over Pantanal domain is similar or a fraction of that over Amazonia
200 domain. That was not the case for September and October of 2020, when mean AOD at 550 nm over Pantanal was much higher
than over Amazonia, an indication that locally produced smoke played a major role on the amount of smoke over the Pantanal
202 atmosphere column.

To analyze why in August 2020 the higher fire count did not translate in higher smoke amount, compared with September and
204 October 2020, the monthly distribution of fire count across Pantanal (Figure 5a) was analyzed as a function of land used
(conservation units and indigenous areas) and Enhanced Vegetation Index (EVI). The analysis (Figures 5) shows that, despite
206 the large fire count occurrence, August presented a reduced number of fires within conservation areas (where higher biomass
density is present). However, during September and October, there was a significant increase of fire number within those areas,
208 which could explain the larger aerosol emissions and, consequently, higher AOD values over Pantanal compared to August.

A possible shift in the mean composition of the regional smoke aerosol plume optical properties (mainly light absorption
210 efficiency) during the 2020 biomass burning season was also explored via retrievals of single scattering albedo (SSA) from
AERONET stations. Since there are no AERONET stations operating in Pantanal, the analysis of SSA was carried out from
212 stations in the neighbourhood of Pantanal (Cuiaba) and somewhere else across Amazonia and at Santa Cruz, in Bolivia. The
results obtained did not show significant change between mean SSA for 2020 biomass burning season and from previous years
214 (Figure 6). Referring to Cuiaba, the closest AERONET site to Pantanal, the mean SSA on 2020 was within the site typical
variability, that also was the case of Rio Branco, Alta Floresta and Santa Cruz sites. Multiyear instantaneous aerosol radiative
216 forcing versus AOD at 550 nm from these AERONET stations, and as function of SSA, were also analyzed (Figure 7) by
comparing 2020 with the previous years (2003 to 2019). It is possible to observe that not only AOD but also SSA affects the
218 downward solar irradiance at the surface. However, no exceptional difference can be noticed in 2020 data when compared
with the entire dataset. It is worth mentioning that the AERONET data from quality level 2.0 for 2020 is not yet available, so
220 level 1.5 was used. Therefore, future and further analysis on this matter is highly recommended in order to evaluate the posed
hypotheses, although as observed in Figure A.1, there is very good agreement between levels 1.5 and 2.0 SSA retrievals.

222

3.2 Regional Smoke Plume and Surface Solar Radiation anomaly: Spatial and interannual analysis

224 Climatologically, September is the peak of the biomass burning season over a large area of South America, and of Brazil in
particular. Therefore, September was selected to explore the spatial distribution, loading and impact on the downward solar
226 radiation at the surface of the regional smoke plume for 2020 and the most polluted years in the time series analyzed, namely
2004, 2005, 2007, 2010, 2017. Considering the particularity of smoke over Pantanal in October of 2020, that month was also
228 added to this analysis of the RSP feature and impact on SSR. Figure 8 presents the interannual variability of the geographical
distribution of the RSP for September and October for the most polluted years. Although it was one of the most polluted in the

230 last five recent years, the 2020 regional plume could not be considered as an exceptional plume when compared, for instance,
with the historical RSP of the years 2007, 2010 and 2005. Among the years analyzed, 2007 stands as the top polluted, for both
232 September and October. However, during October 2020, it is possible to visualize that the peak of the smoke over the continent
was centered in Pantanal (Figure 8b), corroborating the previous discussion that 2020 biomass burning exceptionality was
234 restricted to the Pantanal biome. The area of the RSP core, considered as the extension of the smoke plume delimited by the
isoline of AOD at 550 nm of 0.5 (area) and its respective mean intensity (AOD at 550 nm level) are also identified. The 2007
236 RSP presented the largest (5740811 km²) and most polluted domain (mean AOD at 550 nm > 1.0), a characteristic that extended
to the month of October. The area and AOD at 550 nm figures during September for the 2020 RSP were, respectively, 2450641
238 km² and 0.68.

Due to the wind circulation pattern in the region, in general, the smoke from fires in the Amazon Forest and Cerrado ecosystem
240 are transported westward reaching other South-American countries until it reaches the Andes Mountain range barrier (Freitas
et al., 2004), where the wind circulation becomes predominantly meridional and southwards with a variable zonal component
242 that can fluctuate the transport of smoke between the southeast and the southern portions of Brazil. Going southwards, the
plume can reach the Pantanal region, among other locations, but the smoke usually gets diluted during this transport process.
244 Related to the flow of the regional smoke plume towards the population centers of the southeastern coast, a stronger feature of
2020 regional smoke plume, the analysis of previous years evidenced that this has also been seen in the past. For example, in
246 2004 and 2005 the monthly flow patterns were also towards the highly populated centers in the southeast of Brazil.

Figure 9 also shows that the anomaly of solar radiation in cloud-free conditions occurred over the regions more severely
248 affected by the aerosol plume, consistently with the spatial distribution of AOD, as expected. Again, in 2020, the solar energy
reduction was surpassed by previous top polluted years, specially 2007, when SSR anomaly over a large portion of the
250 continent, including the western portion of Pantanal, reached values up to -300 Wm⁻². It is also possible to see that in 2005,
as occurred in 2020, the southeast portion of Brazil experienced a larger SSR reduction as a result of the persistent transport
252 of smoke toward that region. As suggested by the AOD field, available solar energy reaching the ground in October 2020 over
Pantanal was significantly reduced (up to 200 Wm⁻²).

254 **4 Conclusion**

The 2020 biomass burning season in Brazil attracted unprecedented attention from national and international media as well as
256 the general society. Pantanal biome was a hotspot in this entire discussion. The wetland biome's role in the Regional Smoke
Plume (RSP) has been marginal throughout the years. However, with the explosion of fire counts across Pantanal in 2020,
258 there was a question about the role of Pantanal's smoke emissions to the RSP during the 2020 biomass burning season. In this
study we analyzed to what extent the RSP produced during the 2020 biomass burning season and its impact on Solar Surface
260 Radiation under cloudless scenarios differs from previous years, from 2003 to 2019, in particular the highly polluted former
RSPs. Additionally, we analyzed the interannual and intraseasonal variability of fire counts and aerosol loading with emphasis

262 on the biomes Amazonia tropical rainforest, Cerrado and Pantanal. In the last eighteen years, from the point of view of the
Amazon Forest and Cerrado ecosystem, 2020 can be considered as an ordinary biomass burning season, far from the most fire
264 active and polluted years of 2010 and 2007. For these years, 2010 and 2007, Amazonia and the western portion of Cerrado
experienced their largest fire counts and aerosol loading within the analyzed period. However, for Pantanal, under certain
266 aspects, 2020 was a very particular year. Not exactly due to the smoke loading over the biome, since in September of 2007 the
biome experienced a higher smoke loading, but due to the relative contribution of the local fire when compared to advection
268 from upwind areas and the amount of smoke observed in October. The analysis revealed that, typically, Pantanal smoke loading
has a stronger relationship with smoke over Amazon than with fire counts within the biome, suggesting that AOD over Pantanal
270 is more affected by advection than by local fires. However, in 2020 that was not the case, local fire dominated. Usually, smoke
loading over Pantanal is a fraction of that over Amazonia domain, but for September and October of 2020, mean smoke loading
272 over Pantanal was much higher than over Amazonia, an indication that smoke produced locally played an atypical role to the
smoke level over Pantanal. In the 2020 biomass burning season, fire counts in Pantanal were 3.4 times higher than the mean
274 value from 2003 to 2020. As important as the amount of fire, the nature of the areas being burned within Pantanal revealed to
determine the amount of smoke produced during September and October. From September on, there was a significant increase
276 of fire counts within conservation and indigenous areas, where higher biomass density is present, which would explain the
large smoke plumes over Pantanal, even during October when the number of fires were significantly reduced. The entire biome
278 was continuously covered by a thick layer of smoke from west to east and from south to north for almost one month and a half,
which resulted in a monthly (October) mean deficit of solar radiation at the surface up to 200 Wm^{-2} . The impact of this
280 reduction of incoming solar radiation on biological and surface-atmosphere interactions processes and are yet to be evaluated.
Additionally, considering the plume transport towards the highly productive central and southeast regions of Brazil, the
282 country's capacity of production of renewable energy based on solar radiation is also affected. In a period when cloud cover
is less frequent and the generation of hydro-electric power, the main energy source in Brazil, decreases due to the lack of
284 precipitation, revealing a need for a comprehensive governance.

In conclusion, from RSP loading and area perspective and when compared with previous years (2003-2019), 2020 biomass
286 burning season could not be identified as an exceptional year. However, when focusing on the Pantanal domain, there are three
aspects that differ from typical features, the fire count observed in September, the smoke loading over the Pantanal biome
288 during October and the relative contribution of local fire to the local smoke loading when compared to the advection of smoke
from Amazonia domain. Pantanal biomass burning has received less attention throughout the years, when compared with
290 Amazonia and Cerrado biomes. However, Pantanal is a critical piece in the regional hydrological cycle, with strong connection
with the two larger biomes. Pantanal surface-atmosphere interaction processes' role in the support of local and regional
292 biodiversity is indisputable. Although it is too early to infer a possible pattern change in Pantanal biomass burning features,
studies focusing on the 2020 atypical biomass burning season may shed light on potential scenarios that the region may
294 experience in the future. Because of the likelihood of an increase in drought frequency and intensity, policy makers will need
to take measures to mitigate future scenarios similar to the one shown here in 2020. Current knowledge on the Pantanal biome

296 fire emission dynamics and smoke plume radiative properties and impacts is far limited compared to the Amazon Forest and
Cerrado ecosystem.

298 Although the 2020 regional biomass burning season was not among the most polluted in the history of biomass burning in
Brazil, the Pantanal 2020 biomass burning season represents the worst combination of a climate extreme applied to a fire-
300 prone environment, coupled with inadequately enforced environmental regulations (Marengo et al. 2021, Libonati et al., 2020,
Vale et al., 2021). According to Marengo et al. (2021), the years of 2019 and 2020 were characterized by the worst drought in
302 50 years in Pantanal. Vale et al. (2021) pointed out a large reduction in environmental fines during the pandemic. This was
caused by a decrease in fine enforcement, not due to a decrease in environmental violations, as an increase in Amazonian
304 deforestation was observed. The slashing of funding for environmental protection and climate actions during the pandemic
years has compounded the environmental harm imposed by legislative acts aimed at degrading environmental protection. There
306 are still open questions about the specific behavior of mankind intervention in Pantanal in 2020, a lack of studies of human
causes and responses to fires in the Pantanal has been recognized as a challenge to a full comprehension of what happened
308 (Libonati et al., 2020). One also has to point out that there is a need to advance scientific knowledge on the Pantanal climate
process in the context of both local environmental degradation and global climate change scenarios. This evaluation has as
310 basis the notable lack of regular monitoring and comprehensive experimental campaigns focusing on the biome and its complex
interdependence relationship with the neighborhood biomes, Amazonia and Cerrado and downwind regions.

312

Data availability. All the dataset (AERONET, MODIS, CERES, MERRA-2, INPE fire counts) used in this study are publicly
314 available and can be downloaded from their respective sites provided in Table 1 of this manuscript.

316 Supplement

The supplement related to this article is available online at: <xxxxxxxxxxxxxx>.

318

Author contributions. N. E. R., E. S. T and M. A. Y. designed and performed the research, analyzed the data, and wrote the
320 paper.

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322

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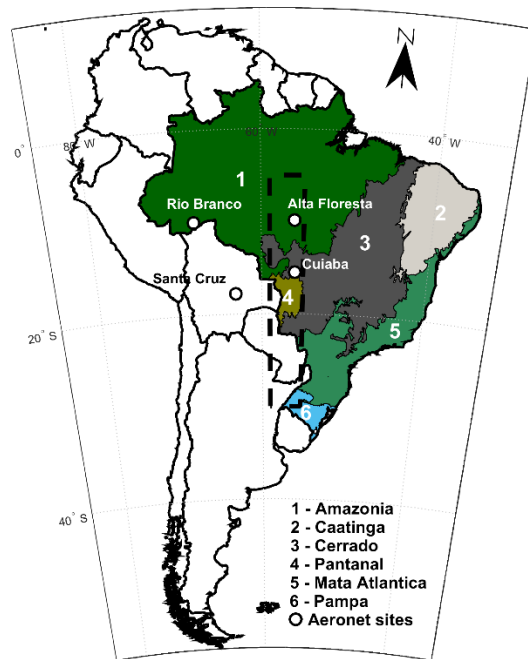
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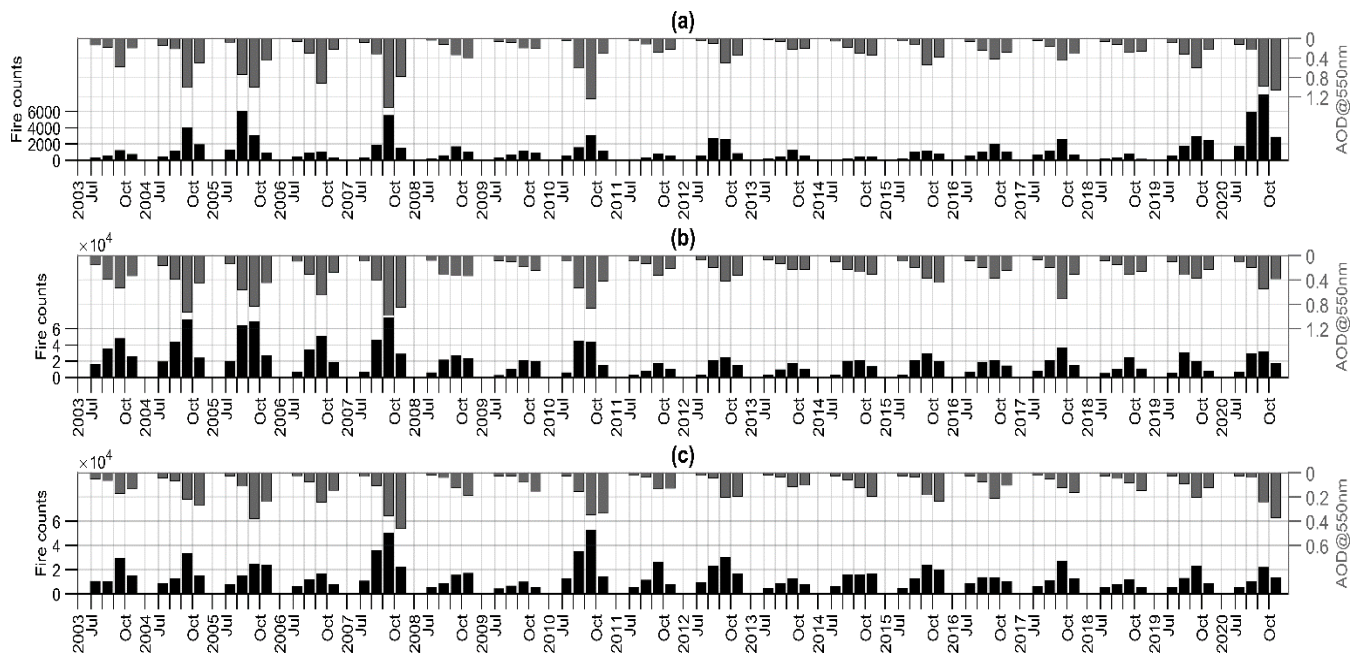


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486 **Figure 1: Spatial distribution of the Brazilian biomes of Amazon tropical forest, Caatinga, Cerrado, Pantanal, Mata Atlântica and**
 488 **Pampa. Locations of the AERONET stations considered in this study are also depicted. The area confined by the dashed red box**
represents a transect defined to study north-south smoke variability and transport taking Pantanal west and east borders as
reference.

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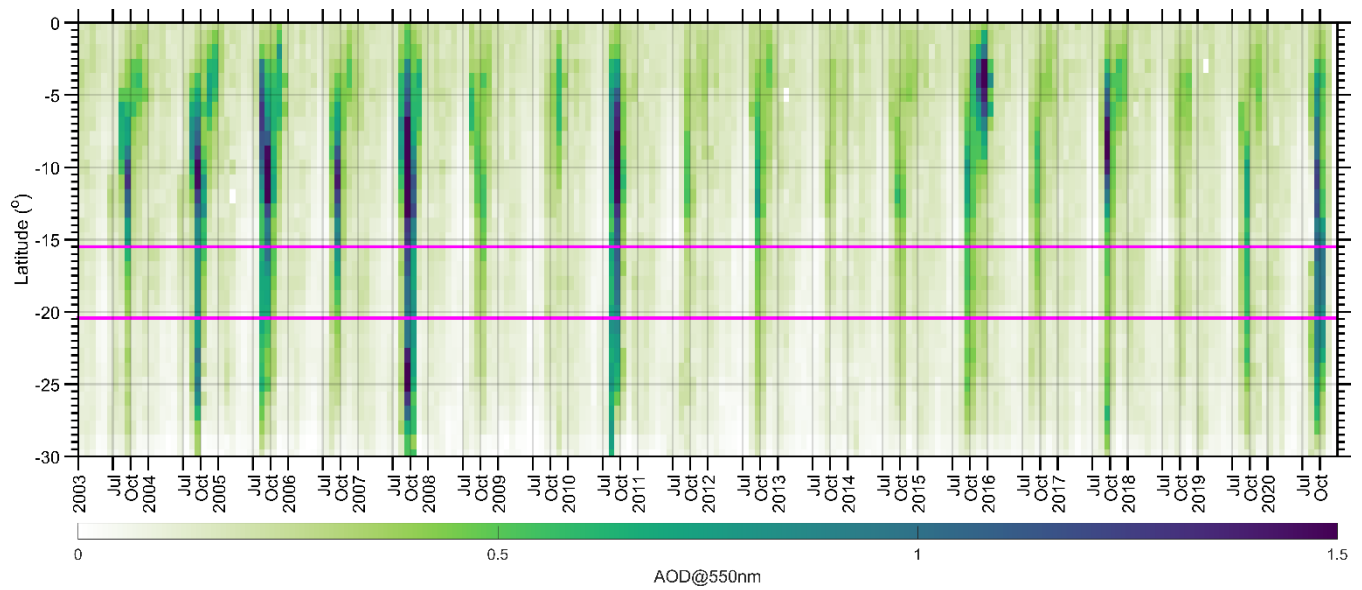
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Figure 2: Monthly (from July to October) interannual evolution of fire counts (black), aerosol optical depth at 550 nm (grey) for the years from 2003 to 2020 for (a) Pantanal, (b) Amazonia, and (c) Cerrado biomes.

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504 **Figure 3: Hovmoller diagram of mean AOD at 550 nm from Amazon Forest latitudes to Pantanal downwind regions considering the**
 506 **average between Pantanal borders at west (58.5°W) and east (54.96°W) longitudes (see Figure 1). The magenta line represents the**
 508 **north and southern limits of Pantanal. Each year's label is defined at its beginning, January first, that is also the case of the months**
 510 **of July and October, which are used to highlight the biomass burning season period.**

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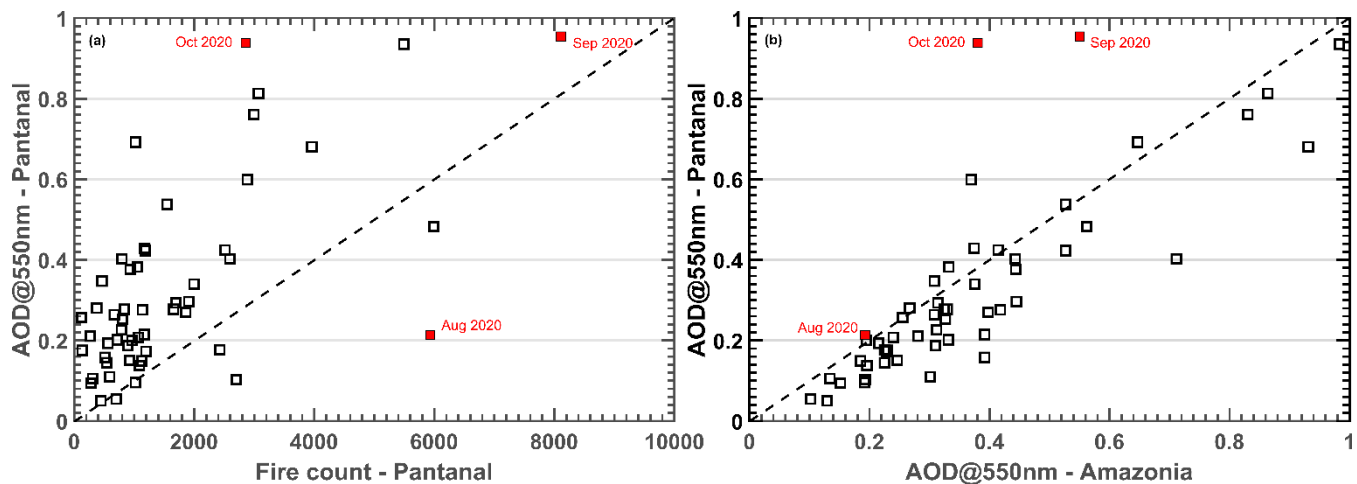
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522 **Figure 4: Multiyear analysis (2003-2020) of (a) monthly mean aerosol optical Depth at 550 nm (AOD@550nm) as function of fire**
 524 **count for Pantanal biome; b) monthly mean aerosol optical Depth at 550 nm (AOD@550nm) over Pantanal as function of monthly**
 526 **mean AOD@550nm over Amazonia. August, September and October of the year 2020 are red-highlighted.**

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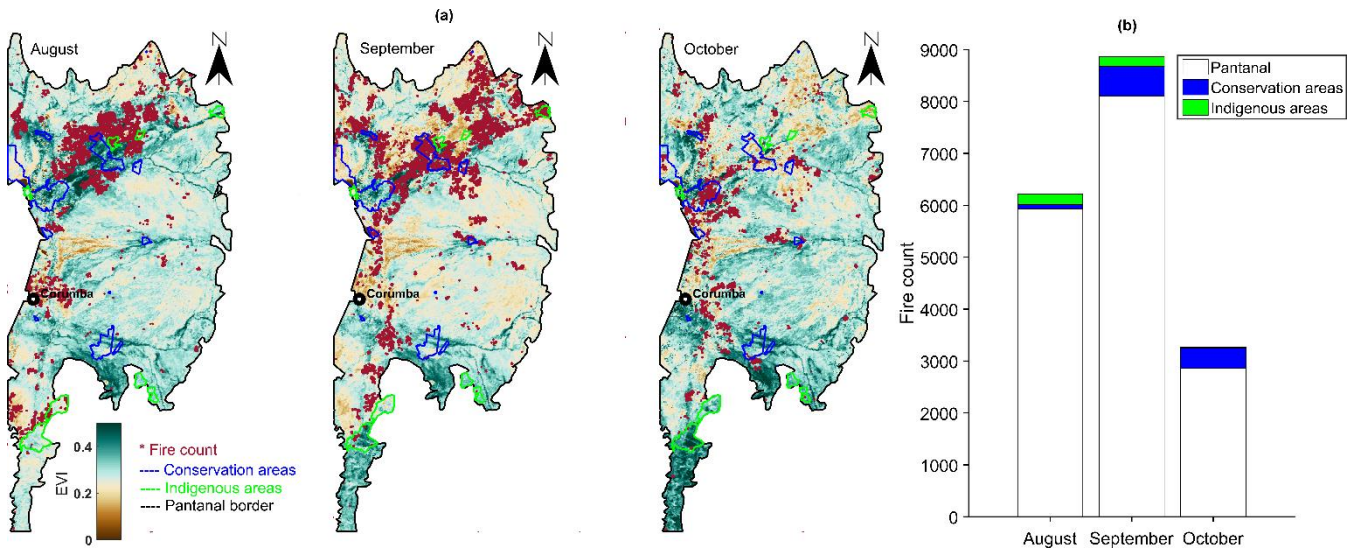
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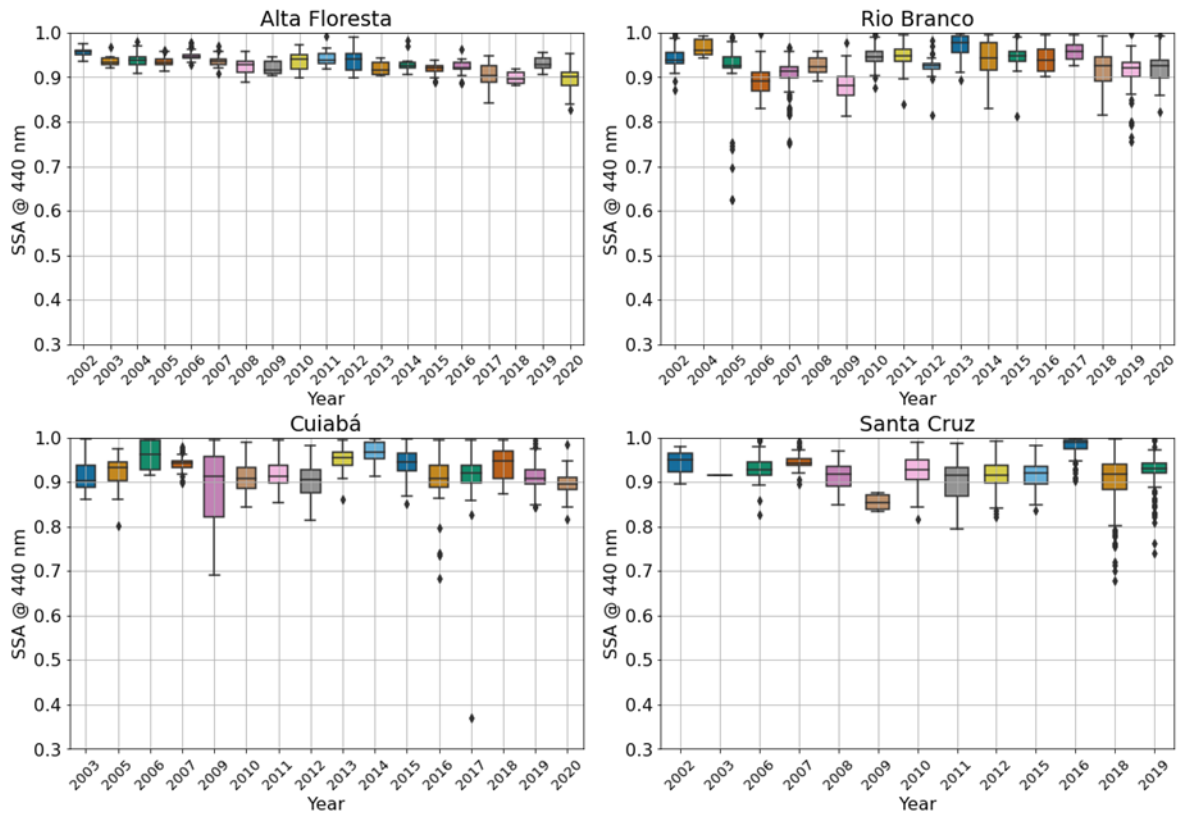


536 **Figure 5: a) Fire count distribution over Pantanal for August, September and October of 2020 using as background the Enhanced**
 538 **Vegetation Index (EVI); b) Monthly (August, September and October) distribution of fire count over Pantanal and over conservation**
and indigenous areas located within the biome.

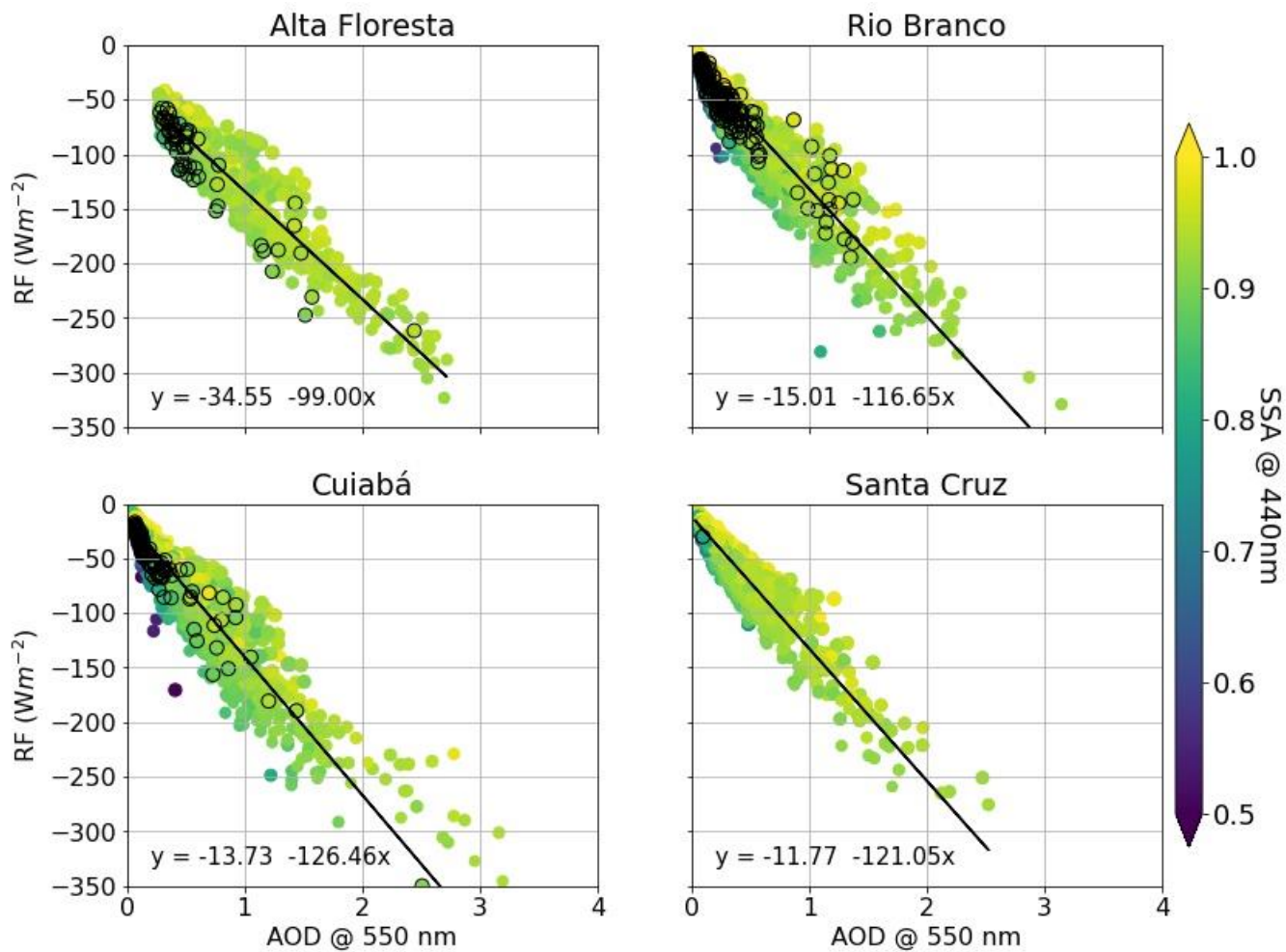
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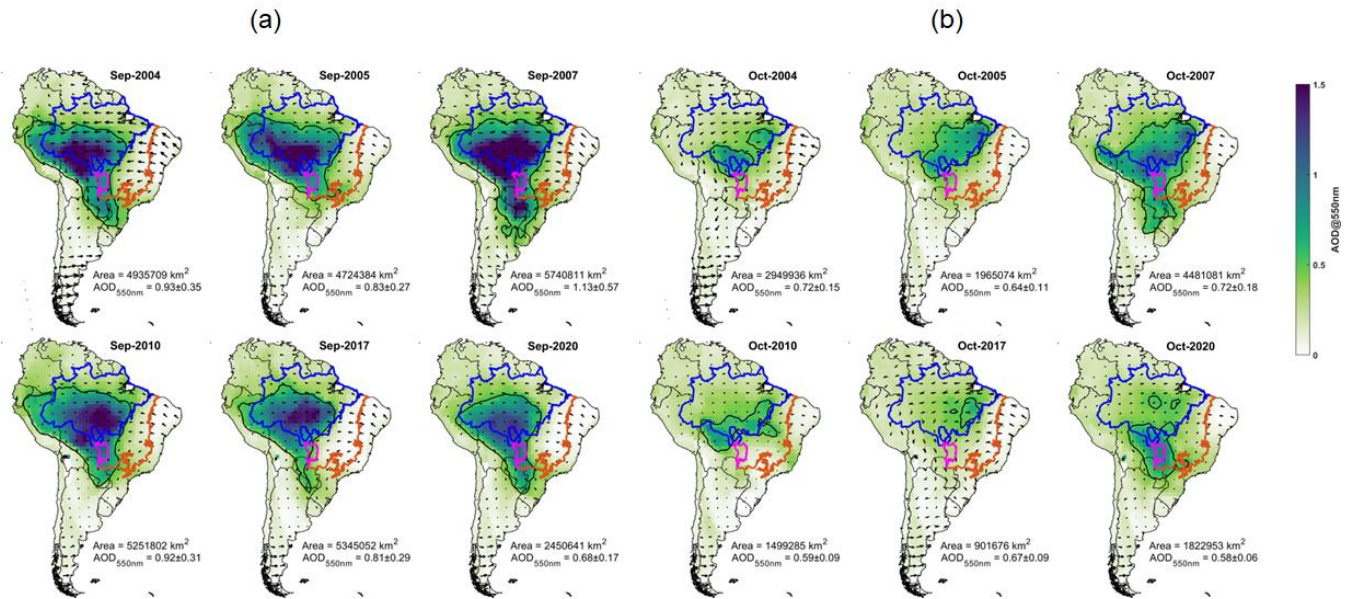
546 **Figure 6: Multiyear box plots of the monthly single scattering albedo at 440 nm (SSA@440nm) from 2003 to 2020 for four**
 548 **AERONET sites: Alta Floresta, Cuiabá, Rio Branco and Santa Cruz. In each box, the black line in the center represents the median,**
and the lower and upper limits are the first and the third quartiles, respectively. The vertical lines extending from the box represent
the spread of instantaneous SSA@440 nm with the length being 1.5 times the interquartile range.



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552 **Figure 7: Aerosol direct instantaneous radiative forcing as function of AOD at 550 nm at AERONET sites highlighting 2020 (black) against historical values (2003 – 2019).**

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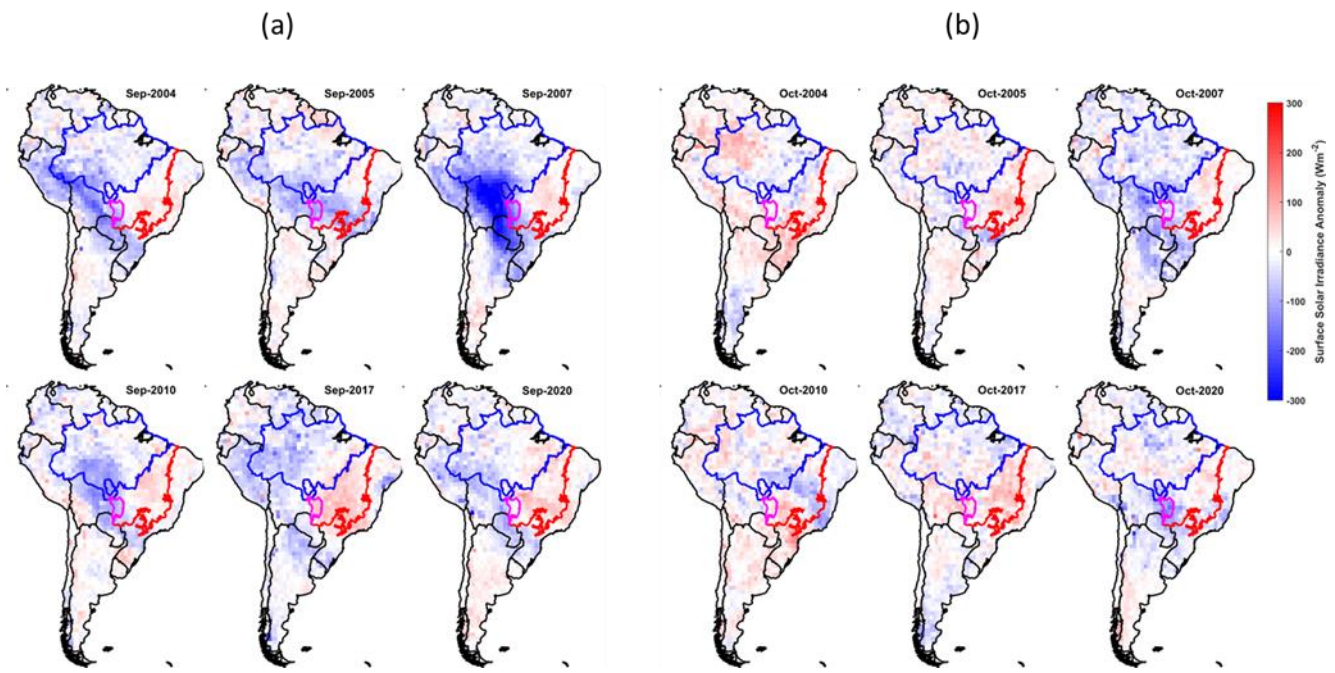


556 Figure 8: Spatial distribution of the mean aerosol optical depth at 550 nm (AOD_{550nm}) and wind pattern at 850 hPa for September
 558 (a) and October (right) of different years. The area limited by the black solid line represents the regions with AOD_{550nm} higher than
 0.5, and its dimension (Area) and mean AOD_{550nm} are also presented. The biomes' borders are represented by the colors blue
 (Amazon Forest), red (Cerrado ecosystem) and green (Pantanal).

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566 **Figure 9: Spatial distribution of the mean surface solar irradiance under cloudless conditions and in the presence of aerosol for the**
 568 **months of September and October of different years. The biome's borders are represented by the colors blue (Amazon Forest), red**
(Cerrado ecosystem) and magenta (Pantanal).

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Table 1: Set of variables used in this study, their respective applications, sources and references.

Variables	Application	Data sources	Reference
Aerosol Optical Depth at 500 nm (AERONET) and at 550 nm (MODIS – Aqua)	Smoke plume loading and dimension	AERONET (Level, 1.5) https://aeronet.gsfc.nasa.gov/ MODIS Atmosphere Level 3 (L3) gridded products (Daily and Monthly) from the Aqua platform (MYD08D 3/M 3) https://giovanni.gsfc.nasa.gov/giovanni/ Accessed on Feb16, 2021.	Holben et al. (1998); King et al., 2013; Levy et al., 2013; Platnick et al., 2017
Aerosol Single Scattering Albedo	Smoke plume absorption efficiency	AERONET (Level 1.5) https://aeronet.gsfc.nasa.gov/	Dubovik and King (2000)
Aerosol radiative forcing	Radiative impact	AERONET (Level 1.5) https://aeronet.gsfc.nasa.gov/	Garcia et al. (2012)
Wind components at 850 hPa	Circulation pattern	MERRA 2- Reanalysis atmospheric variables. https://disc.gsfc.nasa.gov/datasets/M2TMNXSLV_5.12.4 Accessed on November 11, 2021.	Gelaro, et al. (2017); GMAO (2015)
Fire counts	Fire activity	http://queimadas.dgi.inpe.br/queimadas/portal/static/estatisticasestados/ – Accessed on December23, 2020.	Morisette et al. (2005)
Enhanced Vegetation Index (EVI)	Land use	https://doi.org/10.5067/MODIS/MOD13C2.061 Accessed on May, 20, 2022	Didan (2021)

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