



1 A revised mineral dust emission scheme in GEOS-Chem: 2 improvements in dust simulations over China

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10 **Abstract:**

11 Mineral dust plays a significant role in climate change and air quality, but large uncertainties remain in terms of dust
12 emission prediction. In this study, we improved the treatments of dust emission process in a Global 3-D Chemical Transport
13 model (GEOS-Chem) v12.6.0, by incorporating the geographical variation of aerodynamic roughness length (Z_0), smooth
14 roughness length (Z_{0s}), soil texture, introducing Owen effect and Lu and Shao (1999) formulation of sandblasting efficiency
15 α . To investigate the impact of the modifications incorporated in the model, several sensitivity simulations were performed
16 for a severe dust storm during March 27, 2015 to April 2, 2015 over northern China. Results show that simulated threshold
17 friction velocity is very sensitive to the updated Z_0 and Z_{0s} field, with the relative difference ranging from 10% to 60%
18 compared to the original model with uniform value. An inclusion of Owen effect leads to an increase in surface friction
19 velocity, which mainly occurs in the arid and semi-arid regions of northwest China. The substitution of fixed value of α
20 assumed in original scheme with one varying with friction velocity and soil texture based on observations reduces α by 50%
21 on average, especially over regions with sand texture. Comparisons of sensitivity simulations and measurements show that
22 the revised scheme with the implement of updates provides more realistic threshold friction velocities and PM_{10} mass
23 concentrations. The performance of the improved model has been evaluated against surface PM_{10} observations as well as
24 MODIS aerosol optical depth (AOD) values, showing that the spatial and temporal variation of mineral dust are better
25 captured by the revised scheme. Due to the inclusion of the improvement, average PM_{10} concentrations at observational
26 sites are more comparable to the observations, and the average mean bias (MB) and normalized mean bias (NMB) values
27 are reduced from $-196.29\mu g m^{-3}$ and -52.79% to $-47.72\mu g m^{-3}$ and -22.46% respectively. Our study suggests that the
28 erodibility factor, sandblasting efficiency and soil-related properties which are simply assumed in the empirical scheme
29 may lack physical mechanism and spatial-temporal representative. Further study and measurements should be conducted
30 to obtain more realistic and detailed map of these parameters in order to improve dust representation in the model.

31 **1 Introduction**

32 Mineral dust is typically produced by wind erosion from regions with arid and semi-arid surfaces in the world and
33 exerts significant impacts on the atmospheric radiation balance (Tegen et al., 1996; DeMott et al., 2010; Kumar et al., 2014;
34 Saidou Chaibou et al., 2020a), climate (DeMott et al., 2003; Mahowald and Kiehl, 2003; Zhao et al., 2012; Chen et al.,
35 2014; Chin et al., 2014), air quality (Giannadaki et al., 2014; Tian et al., 2019) and human health (Goudie, 2014; Tong et
36 al., 2017). Dust emission process has been recognized as a leading contributor to dust aerosol loading. Global mineral dust
37 particles are mainly emitted from North Africa, the Arabian Peninsula, Central Asia, East Asia, Australia and North



38 America, with East Asia (including the Gobi and Taklimakan deserts) accounting for ~20% of the global dust emission
39 (Ginoux et al., 2004; Nagashima et al., 2016).

40 In order to properly reproduce dust emission process, many dust emission schemes have been developed and
41 implemented in both global and regional chemical transport models (CTMs) (e.g., Marticorena and Bergametti, 1995; Lu
42 and Shao, 1999; Alfaro and Gomes, 2001; Shao, 2001, 2004; Shao et al., 2011; Zender et al., 2003; Kok, 2011a, 2011b).
43 Nevertheless, some intercomparison studies demonstrated that there are large discrepancies among different dust emission
44 models (Uno et al., 2006; Todd et al., 2008; Huneus et al., 2011; Su and Fung, 2015; Ridley et al., 2016; Chen et al., 2017;
45 Chen et al., 2019; Ma et al., 2019; Wu et al., 2019; Saidou Chaibou et al., 2020b; Zhao et al., 2020). Ma et al. (2019)
46 quantitatively evaluated the performance of three dust schemes in WRF-Chem, two schemes in both CHIMERE and
47 CMAQ, and one scheme in CAMx during a dust episode over northern China. Large differences between observed surface
48 PM₁₀ concentrations and modelling results of each model were found. Among schemes in WRF-Chem, AFWA and
49 UOC_Shao2004 are better correlated with observations compared to GOCART but tend to overestimate dust
50 concentrations. Kang et al. (2011) compared the performance of three dust emission schemes in WRF-Chem over East
51 Asia, showing that the difference of dust emission fluxes between three schemes ranges from an order of 10¹ to 10². Ridley
52 et al. (2016) showed that the estimated global dust AOD vary by over a factor of 5 among four global models (including
53 GEOS-Chem, WRF-Chem, CESM and MERRAero), and dust emissions across North Africa are overestimated while
54 emissions from Asia and the Middle East are underestimated overall. An intercomparison of 14 CTMs as part of the Model
55 Inter-Comparison Study for Asia (MICS-Asia) phase III project (Chen et al., 2019) showed that nearly all participant
56 models underestimate PM₁₀ levels and current CTMs have difficulty producing similar dust emissions when adopting
57 different dust schemes.

58 The uncertainties in dust emission models can be attributed to a number of issues, such as threshold friction velocity,
59 surface wind speed, soil texture, particle size distribution, other soil/surface parameters (e.g., soil moisture, vegetation
60 cover, aerodynamic roughness length) and different physical mechanisms (Tegen, 2003; Zhao et al., 2013; Liu et al., 2018;
61 Chen et al., 2019). Darnenova et al. (2009) conducted a detailed comparison between two schemes developed by
62 Marticorena and Bergametti (1995) and Shao et al. (1996), indicating that wind friction velocity is a significant factor in
63 simulating dust emission while the aerodynamic roughness length as well as vegetation cover may play an important role
64 at higher wind speed. Many sensitivity experiments have been conducted and shown that the modeled threshold friction
65 velocity can be modified by soil moisture (Cheng et al., 2008; Mokhtari et al., 2012; Gherboudj et al., 2015; Ju et al., 2018),
66 soil texture (Menut et al., 2013; Gherboudj et al., 2015; Perlwitz et al., 2015a, 2015b; Kontos et al., 2018) and surface
67 roughness (Cheng et al., 2008; Astitha et al., 2012; Menut et al., 2013), which in turn affects the predicted dust emission.
68 In addition, a more accurate value of sandblasting mass efficiency (α) has been reported to be a crucial factor for a better
69 performance of dust emission flux (Mokhtari et al., 2012; Klingmüller et al., 2018; Kontos et al., 2018; Ma et al., 2019).

70 Based on the above studies, it is necessary to take these key parameters, including soil-related properties and empirical
71 input parameters, into fully consideration in a dust emission parameterization. Unfortunately, due to limited observations,
72 many of these parameters are not well included in the model. For example, most dust models simply assume a constant
73 values of aerodynamic roughness length and soil clay fraction (Ginoux et al., 2001; Tegen et al., 2002; Zender et al., 2003),
74 ignoring the temporal and spatial variability of them, which may cause uncertainties to the estimated surface friction
75 velocity and threshold friction velocity. During recent decades, with the development of observation technology, the
76 detailed information on the surface characteristics appropriate for global and regional models have been provided (Laurent
77 et al., 2005, 2008; Prigent et al., 2005, 2012; Shangguan et al., 2014; Perlwitz et al., 2015a, 2015b). Therefore, adopting
78 more accurate and detailed soil datasets is expected to improve the dust model performance.

79 In this study, we present an improvement of the dust emission scheme in GEOS-Chem model by incorporating the
80 updated soil texture and aerodynamic roughness length with spatial variability, Owen effect, drag partition correction factor
81 as well as the updated formulation of sandblasting efficiency, which together significantly improve the prediction of dust



82 emission flux and concentrations over China. The objective is to obtain more realistic surface friction velocity (u_*) and
83 threshold friction velocity (u_{*t}) by considering the effect of the soil moisture, surface roughness and soil texture, thus
84 improving the representation of dust emission in the model.

85 Section 2 gives a detailed description of GEOS-Chem model and the modifications of the improved scheme as well
86 as numerical experiments and data description. Sensitivity results are compared in Section 3.1 to examine the impacts of
87 the modifications. Section 3.2 presents the comparisons of the improved scheme and original version with observations, to
88 evaluate the performance of the improved scheme. Uncertainties, limitations, and future improvements of the emission
89 scheme are discussed in Section 3.3, followed by a summary in Section 4.

90 2 Model and measurements

91 2.1 Model description

92 The GEOS-Chem model is a global three-dimensional chemical transport model driven by assimilated meteorology.
93 In this work, we use v12.6.0 of GEOS-Chem driven by GEOS-FP assimilated meteorological field with a spatial resolution
94 of $0.25^\circ \times 0.3125^\circ$ and 72 vertical levels for China region ($15\text{-}55^\circ\text{N}$, $70\text{-}140^\circ\text{E}$) during the period of March 27, 2015 to
95 April 2, 2015. The lateral boundary conditions is provided by a global simulation of $2^\circ \times 2.5^\circ$.

96 GEOS-Chem includes detailed atmospheric chemical mechanism and online aerosol calculations. In this work, we
97 simulate the dust emission with a combination of the dust entrainment and deposition (DEAD) mobilization scheme
98 (Zender et al., 2003) and Global Ozone Chemistry Aerosol Radiation and Transport (GOCART) source function. Dry
99 deposition velocities for dust aerosols are computed with the gravitational settling scheme of Fairlie et al. (2007) and
100 aerosol deposition scheme from Zhang et al. (2001). Wet deposition scheme, which includes scavenging in convective
101 updrafts, as well as rainout and washout of soluble tracers, is described in Liu et al. (2001). Aerosol optical depth is derived
102 online from aerosol concentrations with externally-mixed assumption using RH-dependent aerosol optical properties from
103 Latimer and Martin (2019). Dust optics are from Ridley et al. (2012).

104 2.2 Improvement on the dust emission scheme in GEOS-Chem

105 The standard dust emission scheme in GEOS-Chem is based on a semi-empirical formulation developed by Zender et
106 al. (2003) and is combined with GOCART source function (Ginoux et al., 2001). In this scheme, the vertical dust flux (F)
107 is proportional to the horizontal saltation flux (Q_s), which is the function of surface friction velocity (u_*) and threshold
108 friction velocity (u_{*t}):

$$109 \quad F = (1 - A_s)S\alpha Q_s \quad (1)$$

$$110 \quad Q_s = C_z \frac{\rho_{air}}{g} u_*^3 \left(1 - \frac{u_{*t}}{u_*}\right) \left(1 + \frac{u_{*t}}{u_*}\right)^2 \quad u_* > u_{*t} \quad (2)$$

111 where α is the vertical-to-horizontal flux ratio or sandblasting efficiency, based on the soil clay content (Marticorena and
112 Bergametti, 1995). S is based on GOCART source function (see Fig. S1), also named as the soil erodibility factor,
113 representing the grid cell fraction of the bare land suitable for mobilization. A_s is the fraction of snow-covered surface. C_z
114 is the saltation constant ($C_z=2.61$).

115 According to the equation, u_{*t} , u_* as well as α are the key input parameters in the accurate prediction of dust
116 emission flux. u_{*t} is used to describe the characteristics of soil and land surface condition, representing the resistance of
117 surface to wind erosion. In the standard dust scheme, u_{*t} is calculated using a semi-empirical formulation as a function
118 of air density and soil particle density (Iversen and White, 1982). Furthermore, two correction terms, including soil
119 moisture correction (Fécan et al., 1999) and drag partition correction (Marticorena and Bergametti, 1995), are also applied
120 to modify u_{*t} . It should be noted that in the original scheme, the drag partition correction term is eliminated.

121 u_* is a description of surface wind speed, which mainly depends on 10m wind speed taken from meteorological field



122 assuming neutral stability (Bonan, 1996). Owen effect, which represents a positive feedback between saltation process and
123 friction speed (Owen, 1964), is often adopted in models to modify u_* . However, Owen effect is eliminated in the original
124 scheme.

125 Sandblasting efficiency α is parameterized according to the empirical relation described by Marticorena and
126 Bergametti (1995) (MB95), which depends on the soil clay content (M_{clay}) and is restricted to $M_{\text{clay}} < 20\%$:

$$127 \quad \alpha = 100e^{(134M_{\text{clay}}-6)\ln 10} \quad (3)$$

128 However, in the global model, α tends to be overly sensitive to M_{clay} . Due to this reason, a globally fixed value of
129 $M_{\text{clay}} = 20\%$ is assumed in current model (Zender et al., 2003).

130 It should be noted that, some input parameters, data or formulations are quite simplified and need to be improved
131 based on the original dust scheme described above. For example, the aerodynamic roughness length (Z_0), the smooth
132 roughness length (Z_{0s}) as well as the mass fraction of clay in the soil (M_{clay}) are assumed as a constant uniformly, despite
133 the fact that it may vary with time and location. As a result, the simulation of related processes, such as drag partition effect
134 or soil moisture effect, may lack spatial representation. Therefore further modifications on these variables should be made
135 in order to obtain more realistic dust emission. Fig. S2 presents the schematic diagram of the dust emission schemes in the
136 standard model and the modifications incorporated in this study. The details of the parameterization options and required
137 input parameters are presented in following sections.

138 2.2.1 Soil Type and Soil Texture Data

139 In the model, M_{clay} can have an impact on u_{*t} through modifying soil moisture correction term, thus influencing the
140 modeled dust emission flux. The soil moisture correction term, defined as f_w , is parameterized according to Fécan et al.
141 (1999), which accounts for the increase of u_{*t} with soil moisture content.

$$142 \quad f_w = \begin{cases} 1 & w < w' \\ [1 + 1.21(w - w')^{0.68}] & w \geq w' \end{cases} \quad (4)$$

$$143 \quad w'(\%) = a(0.0014M_{\text{clay}}^2 + 0.17M_{\text{clay}}) \quad (5)$$

144 where w is gravimetric soil moisture and w' is soil residual moisture.

145 With the increase of soil moisture, soil cohesion can be enhanced, particularly over regions with high clay content,
146 thus inhibiting sandblasting process to some extent. However, as stated above, M_{clay} is assumed as a constant equal to 20%
147 in the original scheme, which can cause uncertainty in dust prediction. In the improved scheme, we employ the gridded
148 data of clay content from the Global Soil Dataset for use in Earth System Models (GSDE) (Shangguan et al., 2014), which
149 is based on the Soil Map of the World and various regional and national soil databases. Fig. 1 shows the updated M_{clay} from
150 Shangguan et al. (2014) with the horizontal resolution of $2^\circ \times 2.5^\circ$ at the global scale. Compared to the original fixed value
151 of 20%, the updated M_{clay} is generally lower in most of the dust source areas over East Asia.

152 2.2.2 Surface roughness length

153 The drag partition is used to describe the impact of roughness elements (such as rocks, pebbles, vegetation, etc.) on
154 u_{*t} . According to Marticorena and Bergametti (1995), the roughness correction term, f_d , is a function of the aerodynamic
155 roughness length Z_0 and the smooth roughness length (Z_{0s}):

$$156 \quad f_d = 1 - \frac{\ln(\frac{Z_0}{Z_{0s}})}{\ln\left[0.7\left(\frac{12255cm}{Z_{0s}}\right)^{0.8}\right]} \quad (6)$$

157 where the required roughness lengths are set as the constant values of $Z_0 = 0.01$ cm and $Z_{0s} = 0.0033$ cm globally.

158 Z_0 represents the roughness length of the overlying non-erodible elements (solid obstacles, such as rocks), which
159 transfers part of the wind momentum from the atmosphere to the surface, dissipating the shear force for particle saltation.
160 Prigent et al. (2005) derived global aerodynamic roughness length in arid and semi-arid areas which are retrieved from the
161 ERS-1 satellite measurements with a horizontal resolution of $0.25^\circ \times 0.25^\circ$. Here we apply the global monthly mean Z_0



162 fields provided by Prigent et al. (2005) and then re-grid the map to $2^\circ \times 2.5^\circ$ horizontal resolution for the incorporation
163 into GEOS-Chem. As Fig. 2 shows, compared to the fixed constant assumed in the original version, the updated global Z_0
164 is generally higher. Over northern China, the Z_0 value ranges from approximately 0.01cm over desert regions to 0.07cm.

165 Z_{0s} characterizes the roughness length of the uncovered, bare erodible surface. Instead of setting Z_{0s} to a fixed value,
166 some studies suggested that Z_{0s} can be estimated as 1/30 of the coarse-mode mass median diameter (MMD) of soil particles,
167 which will provide more realistic representation of soil texture distribution (Marticorena and Bergametti, 1995; Laurent et
168 al., 2006; Mokhtari et al., 2012). In the improved version, we adopt this empirical formula, based on updated soil texture
169 classification (Mokhtari et al., 2012; Xi et al., 2015), to estimate Z_{0s} :

$$170 \quad z_{0s} = MMD/30 \quad (7)$$

171 where MMD is the median diameter of the coarsest mode for various soil textures shown in Table 1. The corresponding
172 Z_{0s} for different soil types are listed in Table 1 and its global distribution is shown in Fig. 3. The soil texture map is obtained
173 based on the Harmonized World Soil Database (HWSD; [http://www.iiasa.ac.at/Research/LUC/External-World-soil-](http://www.iiasa.ac.at/Research/LUC/External-World-soil-database/HTML)
174 [database/HTML](http://www.iiasa.ac.at/Research/LUC/External-World-soil-database/HTML)), which provides global sand, silt, and clay contents at 30 arc-second resolution. The soil texture dataset
175 is re-gridded to $2^\circ \times 2.5^\circ$ resolution, and then is applied to identify the global soil texture by using the United States
176 Department of Agriculture (USDA) soil texture triangle (based on the amount of sand, clay, and silt contents;
177 <http://soils.usda.gov/technical/aids/investigations/texture/>). There are 12 classes of soil defined by USDA. It can be seen
178 from Fig. 4 that loam, sandy loam and clay loam, are the dominant soil types over China. Among them, sandy loam and
179 loam occupy the major part of northwest China.

180 2.2.3 Sandblasting efficiency α

181 Sandblasting efficiency α is important in the dust emission calculation as it is used to convert the horizontal saltation
182 mass flux to a vertical dust mass flux. In the original scheme, α is simply expressed as a function of M_{clay} , which is a fixed
183 value of 20%. The assumption in the original scheme might cause uncertainty in modeled flux and make the spatial
184 variation less representative (Mokhtari et al., 2012).

185 In order to reduce this uncertainty, a more physically-based function from Lu and Shao (1999) (LS99) is adopted in
186 our study. Based on wind tunnel experiments carried out by Rice et al. (1996a, b), Lu and Shao (1999) derived the
187 expression of α through theoretical calculation and some simplifications:

$$188 \quad \alpha = \frac{C_\alpha g f \rho_b}{2p} (0.24 + C_\beta u_* \sqrt{\frac{\rho_p}{p}}) \quad (8)$$

190 where f is the fine particles content in the soil volume, p is soil plastic pressure, which represents the magnitude of the
191 surface resistance (N m^{-2}), ρ_b and ρ_p are the bulk soil density and particle density, respectively, g is the gravitational
192 acceleration (m s^{-2}), u_* is friction velocity (m s^{-1}), and C_α and C_β are empirical constants. Among these parameters, the
193 values of ρ_b and p depend upon different soil textures. Some studies (Kang et al., 2011; Foroutan et al., 2017; Ma et al.,
194 2019) have implemented this formulation in the model and proposed the proper range of these parameters over different
195 soil types.

196 Many measurements from laboratory experiments and field observations have demonstrated the close relationship
197 between α and u_* (Gillette et al., 1997; Gomes et al., 2003; Rajot et al., 2003; Roney and White, 2006; Macpherson et al.,
198 2008; Panebianco et al., 2016; Zhang et al., 2016). To improve the original scheme, we extract α from these measurements
199 over different soil types, based on the expression of LS99, as depicted in Fig. 5 and Table 2.

200 2.3 Experiments design

201 Several sensitivity experiments (Table 3) are conducted to assess the model performance of the modifications in the
202 improved scheme. Control is the control run with the dust emission scheme originally implemented by Fairlie et al. (2007).
203 Sen_mclay, Sen_owen, Sen_ratio, Sen_drag and Sen_ Z_0Z_{0s} are the same as the control run but including the modification



204 of M_{clay} , Owen effect, sandblasting efficiency, drag partition effect and updated surface roughness length (Z_0 and Z_{0s})
205 respectively. Sen_all represents the simulation with the improved scheme which accounts for all the modification described
206 above.

207 2.4 Measurements

208 The data used in this study includes the Moderate Resolution Imaging Spectrometer (MODIS) Level 3 AOD data,
209 hourly observational data of surface PM_{10} concentration, and meteorological field taken from the Meteorological
210 Information Comprehensive Analysis and Process System (MICAPS). The data used in this study is for the period of March
211 27, 2015 to April 2, 2015.

212 MODIS aerosol products are used to evaluate model results of AOD. MODIS AOD at 550 nm is obtained from the
213 daily level-3 product from Aqua satellites (MYD08_D3, $1^\circ \times 1^\circ$ gridded data) and is combined with Deep Blue retrievals
214 which can provide AOD over bright surfaces (i.e., desert regions).

215 Hourly surface observed PM_{10} concentration data, collected from about 1000 environmental monitoring stations
216 maintained by Chinese Ministry of Environmental Protection (MEP; <http://datacenter.mep.gov.cn>), is used to validate the
217 model performance of surface dust concentrations.

218 Meteorological fields of wind speed taken from the Meteorological Information Combine Analysis and Process system
219 (Micaps) developed by the Chinese National Meteorological Center (NMC) are used for evaluation of wind field in the
220 model. Fig. S3-S4 show that the 10m wind field used in the model scheme generally agree well with the Micaps
221 observations over most sites. However, comparisons of averaged surface wind field between the model input and
222 observations (Fig. S5) show that although the circulation patterns in the model are identical with the observations, surface
223 wind speed in the model tend to be larger than observations, especially over western and northeastern Inner Mongolia.

224 3 Results and discussion

225 3.1 Sensitivity study

226 In order to assess the sensitivity of the dust emission to the modified input parameters or physical processes, several
227 numerical experiments are conducted and compared. Fig. 6a presents the relative difference (%) of averaged u_{*t} during
228 study period between these sensitivity simulations and the control run. The u_{*t} simulated by Control run are generally
229 small, with values less than 0.3m/s (not shown). Wu et al. (2013) indicated that u_{*t} over source regions in northern China
230 calculated by Zender et al. (2003) are generally lower (with values ranging from 0.2 to 0.25 m/s) than the measurement
231 (with values ranging from 0.34 to 0.69 m/s) and the values calculated by Shao (2004), which is closer to the observations.
232 The sensitivity simulations show that the update of M_{clay} in Sen_mclay can lead to higher u_{*t} over northern China and
233 lower u_{*t} over southern China than the control simulation, which overestimates M_{clay} over northern China while
234 underestimates it over southern China by setting M_{clay} to 20%. In northern China, particularly in arid and semi-arid regions,
235 the updated M_{clay} will decrease the soil moisture threshold w' and increase soil moisture term f_w , thus leading to a slight
236 increase in u_{*t} (with magnitude <10%). The inhibition of dust emission by surface roughness elements is not taken into
237 account in the original scheme, i.e., $f_d=1$. Some studies (Darmenova et al., 2009; Menut et al., 2013) have demonstrated
238 f_d as a function of Z_0 and Z_{0s} , implying that f_d increase with Z_0 and decrease with Z_{0s} . Compared to the fixed values used
239 in the original scheme, updated Z_0 field used in Sen_ Z_0Z_{0s} are generally larger and updated Z_{0s} field are smaller. Therefore,
240 f_d are increased significantly, particularly over the regions with non-erodible elements (larger Z_0). Result shows that u_{*t}
241 is increased when considering the drag partition effect (increased by 10% in Sen_drag with constant Z_0 field), particularly
242 with the updated Z_0 and Z_{0s} field (increased by 10%~60% in Sen_ Z_0Z_{0s}). In general, due to the inclusion of Z_0 , Z_{0s} and
243 M_{clay} , f_d and f_w are modified, which results in significant alteration in u_{*t} (ranging from -8%~72% in Sen_all) over
244 China. It can be found that the modification of f_d due to updated Z_0 and Z_{0s} makes more contribution to the increase in



245 u_{*t} .

246 Relative difference of u_* with respect to the control run are also compared in Fig. 6b. Considering Owen effect in
247 Sen_owen leads to an increase in u_* by 0%~39%, especially over northwest China where surface wind is strong. Modeled
248 u_* is obtained from u_{10m} and Z_0 under neutral conditions (Bonan, 1996). It can be seen that updated Z_0 in Sen_ Z_0Z_{0s} can
249 modify u_* by influencing the boundary-layer exchange properties. u_* over northern China is generally increased by
250 10%~22% with higher values of Z_0 in Sen_ Z_0Z_{0s} , while it is slightly decreased over Taklimakan and Gobi deserts. In
251 Sen_all, modeled u_* is increased by 5%~50% over most parts of China.

252 Fig. 6c presents the percentage difference in terms of sandblasting efficiency α . In the original version, α is set as a
253 uniformly constant value (around 0.04) due to the assumption of a fixed M_{clay} . In Sen_ratio and Sen_all, u_* -dependent-
254 ratio following LS99, which varies with different soil texture according to observations, is adopted. On average, α is
255 decreased by 50% with the modification in Sen_ratio and Sen_all. The largest reduction occur over regions with sand
256 texture such as over Taklimakan and Gobi Desert.

257 As seen from Fig. 6d, the simulated dust emission flux (F) vary significantly among different experiments. Due to the
258 inclusion of updated M_{clay} , soil moisture term increases in Sen_mclay, which leads to higher u_{*t} and lower F over most
259 regions. Accounting for Owen effect in Sen_owen results in an increase in F of 0%~314%, particularly over northern part
260 of Gansu Province and northwestern Inner Mongolia. A significant reduction in arid and semi-arid regions of northern
261 China is caused by updated α (Sen_ratio). In Sen_drag and Sen_ Z_0Z_{0s} , F are influenced by -100% ~-4% and -100%~50%
262 respectively as a result of the inclusion of f_d with constant Z_0 and updated Z_0 , Z_{0s} respectively. Due to the combined
263 effects of the modifications, F simulated by Sen_all is generally reduced over northern China, except
264 in some regions of northwest China, where Owen effect plays a dominant role.

265 Four sites closer to dust source area or significantly influenced by dust-storms (Beijing, Huhehaote, Jiuquan and Kuele,
266 locations shown in Fig. S1) are selected to evaluate the performance of control and sensitivity simulations. Comparisons
267 of the modeled u_{*t} (Fig. 7) show that in all sites, modeled u_{*t} are increased in Sen_mclay, Sen_drag, Sen_ Z_0Z_{0s} and
268 Sen_all, compared with the original model, with the highest u_{*t} simulated by Sen_all. Modeled u_{*t} increase from
269 0.22~0.25m/s in Control to 0.32~0.37m/s in Sen_all. The reported u_{*t} values over arid and semi-arid regions of China
270 are around 0.3~0.5m/s (Wang et al., 2009). Wu et al. (2013) summarized that u_{*t} range from 0.34~0.69m/s over East Asia
271 and indicated that u_{*t} calculated by Zender et al. (2003) are relatively lower, ranging from 0.2m/s to 0.25m/s. It is apparent
272 that modeled u_{*t} are greatly increased in the revised simulation, which is much closer to the observed values. This
273 improvement is mainly attributed to the update of Z_0 and Z_{0s} . Comparisons between the modeled averaged PM_{10}
274 concentrations and the observational values in four sites show that PM_{10} levels simulated by Sen_all are closer to
275 observations than many other cases. In summary, Sen_all shows the better agreement with the observations in terms of
276 u_{*t} and PM_{10} concentrations.

277 3.2 Comparison between the improved scheme and the original scheme with observations

278 In order to validate the model performance of the improved scheme, time series of the observed surface PM_{10}
279 concentrations are compared against the modeled values from Control (the original scheme) and Sen_all (the improved
280 version) during a dust episode from 27 March to 2 April of 2015 over northern China. The intensity and evolution of this
281 dust event has been described by Wang et al. (2017), illustrating that dust particles were mainly emitted from Mongolia
282 and Inner Mongolia province of China and a dust backflow event took place over North China on March 29. Fig. 8 compares
283 the hourly modeled PM_{10} concentrations and observed values for nine selected sites (locations shown in Fig. S4), which
284 are closer to the dust sources or severely affected by the dust event. It shows that the dust concentrations are generally
285 underestimated in Control run, particularly when dust concentrations are quite high, indicating that the original scheme has
286 difficulty in accurately reproducing the dust emission process. Sen_all generally reproduce the PM_{10} levels better than
287 Control run. Both the magnitude and the temporal evolution of PM_{10} concentrations are captured in Sen_all quite well,



288 with peak values much closer to the observations. Among these sites, Sen_all shows better performance over North China,
289 e.g., Beijing, Tianjin and Huhehaote. But both Control run and Sen_all fail to capture the peak values from 29 March to 30
290 March. During this period, dust particles, mixed with anthropogenic pollutants, flew back due to the south wind over North
291 China (Wang et al., 2017). Uncertainties in the meteorological field and dust heterogeneous reactions in the model may
292 cause the model bias.

293 For specific periods, however, modeled peak values of some sites occur earlier (several hours) than the observations
294 at some sites (e.g., Beijing and Tianjin in 28 March), which could be considered as a result from the uncertainty in the wind
295 field used in the model. It shows that the surface wind is stronger in the model than the observations (Fig. S3), which may
296 lead to stronger transport of the dust from source regions to downwind areas such as Beijing, Tianjin and Kuele. Instead,
297 modeled and observed peak values of some sites in the source regions (e.g., Huhehaote, Xilinguole and Hami) almost
298 simultaneously occur.

299 In order to quantify the performance of the model result, some statistical parameters, including the mean values,
300 correlation coefficient (R), mean bias (MB), normalized mean bias (NMB), are calculated and listed in this paper. The
301 statistical performance for the modeled surface PM₁₀ concentrations from Control run and Sen_all against observations
302 are presented in Table 4. It shows that dust concentrations at all selected sites are significantly underestimated in Control
303 run, especially over northwest regions, with the MB and NMB values ranging from -163.5 μg m⁻³ to -503.61 μg m⁻³ and
304 -64.61% to -68.48% respectively. It is obvious that Sen_all with updated modification greatly improves the dust
305 concentration prediction, with mean values more comparable to the observations, and the average MB and NMB values
306 reduce from -196.29 μg m⁻³ and -52.79% in Control run to -47.72 μg m⁻³ and -22.46% respectively. The largest
307 improvement occurs at northwest stations (e.g. Hami, Akesu and Kuele), which are located close to Taklimakan desert.
308 Over other regions, such as North China (e.g., Beijing, Tianjin, Huhehaote and Xilinguole), the model performance of
309 Sen_all are slightly better than Control run.

310 Although the MB and NMB values of most stations are generally lower and the mean values are much closer to
311 observation for Sen_all simulation, i.e., modifications included improve the underestimation in Control run to some extent,
312 the dust concentrations are still generally underestimated. For stations closer to Gobi desert, such as Xilinguole, Jiuquan
313 and Baiyin, dust concentrations are greatly underestimated with NMB < -30%, which is likely due to the uncertainty in the
314 erodibility factor over Gobi desert used in our study (Ginoux et al., 2001). Similarly, Su and Fung (2015) evaluated the
315 performance of dust emission schemes in WRF-Chem over East Asia, pointing out that the erodibility factor from Ginoux
316 et al. (2001) over the Gobi desert is significantly underestimated, which may result in the underestimation of the dust
317 emission over the Gobi desert. Given that simulated dust emission flux is directly scaled by erodibility factor, we suggest
318 that the erodibility factor used in our model needs to be updated or improved.

319 As stated above, although the model can capture the overall temporal variations of surface dust concentrations, the
320 occurrence of modeled peak values show earlier (about six hours) than the observations over several stations, which may
321 be attributed to strong transport due to stronger surface wind used in the model. It should be noted here that this model bias
322 contributes a lot to the simulation error, leading to smaller R and greater MB, NMB values. R values will be greatly
323 improved if this bias is eliminated, implying that the input assimilated meteorological field is important for dust emission
324 simulation and needs to be further evaluated and adjusted.

325 The averaged modeled surface PM₁₀ concentrations with and without the modifications (Sen_all and Control run
326 respectively) and observational values at ~1400 stations over China during the study period are compared in Fig. 9. It
327 shows that dust concentrations are generally underestimated in Control run (NMB=-16%, regression slope=0.4), which
328 could be attributed to the crude representation of soil properties, roughness length and other related elements. Incorporating
329 improvements in scheme makes the modeling result much closer to the observations, with R values increasing from 0.6 to
330 0.7, NMB values changing from -16% to -11%, regression slope ranging from 0.4 to 0.6. However, the improved model
331 still tends to underestimate the dust concentrations. Unrealistic soil properties (e.g., soil texture; roughness length) and



332 insufficiently accurate potential source map (the erodibility factor) used to scale dust emission flux could be the possible
333 causes.

334 To further investigate the model performance, spatial distributions of averaged simulated surface PM_{10} concentrations
335 from Control run and Sen_all and their comparisons against observations are presented in Fig. 10. Results show that both
336 Control run and Sen_all can reproduce the pattern of dust concentrations in the study region, with high values located over
337 northwest China, North China and some areas of northeast China, indicating that GEOS-Chem can represent the main
338 features of dust emission and transport during the dust storm. It is found that for most sites in Control run, the simulated
339 magnitude are close to the observational values, but are underestimated over northwest China (where Gobi and Taklimakan
340 deserts are located) and North China plain. The simulated values from Sen_all are generally larger than Control run, and
341 are more consistent with measurements both in magnitude and in area extent, especially
342 over the desert region of northwest China. However, dust concentrations are still underestimated over North China plain,
343 possibly due to outdated source map or some potential dust source regions over Inner Mongolia are not well included. In
344 addition, missing mechanism of secondary aerosol source in the model such as heterogeneous reactions could also cause
345 the model bias (Zheng et al., 2015; Cheng et al., 2016).

346 Fig. 11 shows the spatial distribution of simulated averaged AOD from Control run and Sen_all as well as MODIS
347 AOD for the study period. For better comparison, simulated AOD at 13:00 local time (Aqua passage time) are extracted.
348 Result shows that Control run reproduces the major regions with high AOD values, e.g., eastern China, but with lower
349 magnitude. Control run also fails to capture the high-AOD area over the Taklimakan desert, while Sen_all could capture it.
350 Compared with Control run, Sen_all generally reproduce the spatial coverage and magnitude of the observed AOD.

351 3.3 Discussion

352 In our study, we point out that the erodibility factor (S) in the model may introduce uncertainty in modeled dust
353 concentrations, especially over Gobi desert. Several studies indicated that S from Ginoux et al.(2001) over the Gobi Desert
354 has been significantly underestimated and needs to be improved (Su and Fung, 2015; Zeng et al., 2020). Wu and Lin (2014)
355 have demonstrated that the potential source regions in the southeast of Mongolia and the middle-east of Inner Mongolia
356 are not well characterized by the S from GOCART scheme, which results in the underestimation of dust concentration in
357 this area and its downwind regions. In addition, the source function may not provide precise enough information about the
358 recent expansion of dust source areas over northern China, with the desertification and deforestation (Ku and Park, 2013).
359 Studies have demonstrated that implementing a physically based parameterization instead of an empirical dust source
360 function which is usually time-invariant and lacks physical treatment (Kok et al., 2014a, 2014b), or adopting the dynamic
361 dust source function (Xi et al., 2015), could improve the representation of dust emission. Therefore, the dust source function
362 should be precisely established with new updates and higher resolution using various measurements.

363 In terms of sandblasting efficiency α , many modeling studies as well as observational analysis have investigated its
364 magnitude and expression, but the results may vary greatly (Kang et al., 2011; Ma et al., 2019). The formulation for α used
365 in our improved scheme is based on LS99, which establishes the relationship between α and u_* , along with other soil-
366 related parameters dependent on soil textures. In this study, we derived α for different soil types based on the reported
367 values, but uncertainties still remain due to limited available measurements. In addition to the expression from MB95 and
368 LS99, there are other α formulations proposed by Shao et al. (1996) (Shao96) and Shao04. Different from the empirical
369 function, expressions of Shao96 and Shao04 are more sophisticated, which is the function of u_{*t} and u_* respectively,
370 along with some size information of soil particles. Comparisons of different formulations of α for different soil types (Kang
371 et al., 2011; Ma et al., 2019) have shown that the variation in α can reach up to several orders of magnitude, and few
372 equations could reproduce the measured positive correlation between α and u_* , suggesting that α for different soil texture
373 should be further investigated and observed to improve the model accuracy.

374 In this study, surface conditions including erodibility factor, soil texture, clay content and surface roughness length



375 play a significant role in improving the model performance of u_* , u_{*t} and F . We conclude that substituting globally fixed
376 values of these properties with more realistic and physical-based ones could reduce the model uncertainty and improve the
377 understanding of dust emission mechanism. In physically-based scheme, the importance of accurate input surface
378 properties, including soil particle size distribution (Darmenova et al., 2009; Kok, 2011a, 2011b), soil texture (Shao et al.,
379 2011; Foroutan et al., 2017), surface roughness length (Darmenova et al., 2009; Kontos et al., 2018) and soil plastic pressure
380 (Lu and Shao, 1999; Kang et al., 2011), have also been highlighted by many studies. Therefore, accurate and abundant
381 observation data of soil-related properties are urgently needed, particularly over dust source region. Moreover, various and
382 comprehensive observation methods (e.g., experimental data, field and satellite observations) are recommended in order
383 to correct and update the input data.

384 4 Summary and Conclusion

385 In this study, we revised the treatments of dust emission processes by considering the effect of soil moisture, surface
386 roughness, soil texture, as well as Owen effect and more physically-based formulation of sandblasting efficiency in GEOS-
387 Chem version 12.6.0, in order to improve dust simulation over China. Several sensitivity simulations were conducted
388 during a severe dust storm between March 27, 2015 to April 2, 2015 over northern China to analyze the effects of these
389 modifications on u_* , u_{*t} and emission flux.

390 In the improved scheme, we substituted global constant value of Z_0 , assumed M_{clay} in the original version with
391 geographical variation map obtained from the measurement provided by Prigent et al. (2005) and Shangguan et al. (2014)
392 respectively. Z_0 s and sandblasting efficiency were calculated with formulations based on soil texture data from FAO dataset,
393 which is more physically-based than the original version. In addition, Owen effect and drag partition correction factor were
394 considered in the improved version.

395 Sensitivity result showed that the modified f_d and f_w by inclusion of the updated Z_0 , Z_0 s and M_{clay} resulted in
396 significant alteration in u_{*t} (ranging from -8%~72%) over China. u_{*t} was increased when including the drag partition
397 effect, particularly with the updated Z_0 and Z_0 s field (increased by 10%~60%), which induced the modeled u_{*t} much
398 closer to the measurements. Considering Owen effect increased modeled u_* by 0%~39%, especially over northwest China
399 where surface wind is strong. In general, modeled u_* was increased by 5%~50% over most parts of China due to the
400 inclusion of Owen effect and updated Z_0 . In terms of sandblasting efficiency, it was decreased by 50% on average with
401 the updated u_* -dependent-ratio following LS99, with the largest reduction occurring over regions with sand texture. Due
402 to the combined effect of updated treatments, emission flux simulated by improved scheme was generally decreased over
403 northern China, except in some regions of northwest China, where Owen effect played a dominant role. Better agreement
404 between the improved model results and observational values was achieved in terms of the u_{*t} and surface PM_{10}
405 concentrations in selected typical sites over northern China.

406 Compared with both surface PM_{10} observations and MODIS AOD, the revised dust emission scheme produced better
407 performance in both temporal and spatial variation. Result showed that the dust concentrations were generally
408 underestimated at selected sites in the original scheme, particularly when dust concentrations were high. For the improved
409 scheme, both the magnitude and the temporal evolution of PM_{10} concentrations were well captured, with peak values much
410 closer to the observations. According to the statistics, with the implementation of the updates, averaged PM_{10} values at
411 selected sites were more comparable to the observations, and the average MB and NMB values were reduced from -
412 $196.29 \mu\text{g m}^{-3}$ and -52.79% in Control run to $-47.72 \mu\text{g m}^{-3}$ and -22.46% respectively. However, for some sites closer to
413 Gobi desert, dust concentrations were still underestimated, which was likely attributed to the uncertainty in the erodibility
414 factor over Gobi desert. Comparison of the model results and observed averaged PM_{10} concentrations at ~1000 stations
415 showed that the revised scheme improved the model performance, with R values increasing from 0.6 to 0.7, NMB values
416 changing from -16% to -11%. Moreover, the improved scheme demonstrated better performance in reproducing the spatial
417 distribution of AOD than the original scheme, particularly over the desert region of northwest China.



418 In summary, this study indicated that compared to the original scheme, the revised dust emission scheme had an
419 overall better agreement with the measurements. However, more physically-based schemes and more detailed up-to-date
420 input parameters should be further investigated and observed to improve the accuracy of model.

421 *Code and data availability:* Measurements used in this work have been listed in Sect. 2.4 and acknowledgements. For
422 the model outputs and codes can be accessed by contacting Rong Tian (rongtian@nuist.edu.cn).

423 *Competing interests.* The authors declare that they have no conflict of interest.

424 *Author contributions.* RT designed and conducted the model experiments, analyzed the result and wrote the paper. XYM
425 supervised the project, proposed scientific suggestions and revised the paper. JQZ processed the observation data.

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- 644



645 Table 1. Input soil aggregate size distribution parameters dependent on soil texture classification following USDA.

Soil Texture	Mode1			Mode2			Mode3			Z_{0s}
	n	MMD	σ	n	MMD	σ	n	MMD	σ	
Sand	0.9	1000	1.6	0.1	100	1.7	0	10	1.8	33.3
Loamy sand	0.6	690	1.6	0.3	100	1.7	0.1	10	1.8	23
Sandy loam	0.6	520	1.6	0.3	100	1.7	0.1	5	1.8	17.3
Silt loam	0.5	520	1.6	0.35	100	1.7	0.15	5	1.8	17.3
Loam	0.35	520	1.6	0.5	75	1.7	0.15	2.5	1.8	17.3
Sandy clay loam	0.3	210	1.7	0.5	75	1.7	0.2	2.5	1.8	7
Silt clay loam	0.3	210	1.7	0.5	50	1.7	0.2	2.5	1.8	7
Clay loam	0.2	125	1.7	0.5	50	1.7	0.3	1	1.8	4.2
Sandy clay	0.65	100	1.8	0	10	1.8	0.35	1	1.8	3.3
Silty clay	0.6	100	1.8	0	10	1.8	0.4	0.5	1.8	3.3
Clay	0.5	100	1.8	0	10	1.8	0.5	0.5	1.8	3.3
Silt	0.45	520	1.6	0.4	75	1.7	0.15	2.5	1.8	17.3

646 Including three-mode log-normal parameters (mass fraction n (%), mass median diameter MMD (μm), and geometric standard deviation
 647 σ), and smooth aeolian roughness length z_{0s} (μm).

648 Table 2. Input soil-related parameters for different soil texture used in calculation of sandblasting efficiency α .

Soil Texture	$p(N\text{ m}^{-2})$	$f(\%)$	$\rho_b(\text{kg m}^{-3})$	C_α
Sand	5000	6.9	1000	0.01
Loamy sand	5000	18.5	1000	0.008
Sandy loam	10000	22.3	800	0.7
Silt loam	10000	22.3	800	0.7
Loam	10000	22.3	800	0.7
Sandy clay loam	10000	22.3	800	0.9
Silt clay loam	10000	22.3	800	0.7
Clay loam	10000	22.3	800	0.9
Sandy clay	30000	72	700	0.2
Silty clay	30000	72	700	0.2
Clay	30000	72	700	0.2
Silt	10000	22.3	800	0.9

649



650 Table 3. Sensitivity Experiments design and description.

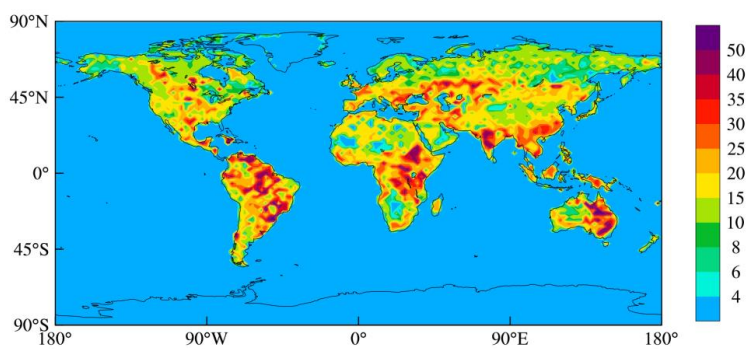
Experiment name	Modifications					Description
	Updated M_{clay}	Owen effect	Updated α	Drag partition correction (Default Z_0 , Z_{0s})	Updated Z_0 , Z_{0s}	
Control	N	N	N	N	N	Original scheme with default configurations. Serves as a control simulation.
Sen_mclay	Y	N	N	N	N	Adopting global M_{clay} from Shangguan et al. (2014).
Sen_owen	N	Y	N	N	N	Considering Owen effect.
Sen_ratio	N	N	Y	N	N	Using updated α from Lu and Shao (1999).
Sen_drag	N	N	N	Y	N	Considering f_d but with $Z_0=0.01$ cm, $Z_{0s}=0.0033$ cm
Sen_ Z_0Z_{0s}	N	N	N	N	Y	Using updated Z_0 from Prigent et al. (2005) and updated Z_{0s} .
Sen_all	Y	Y	Y	Y	Y	Improved scheme including all the modifications described above.

651

652 Table 4. Statistics for observed and simulated (Control and Sen_all) surface PM_{10} concentrations at selected sites.

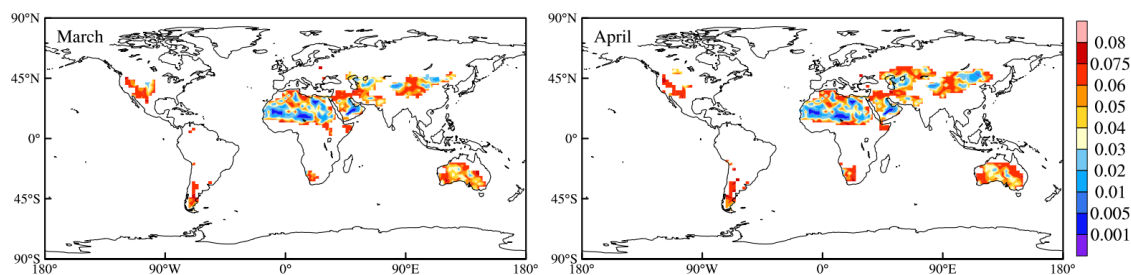
Sites	Obs mean ($\mu g m^{-3}$)	Mod mean ($\mu g m^{-3}$)		R		MB ($\mu g m^{-3}$)		NMB (%)	
		Control	Sen_all	Control	Sen_all	Control	Sen_all	Control	Sen_all
Beijing	232.33	130.54	148.90	0.17	0.15	-87.40	-64.78	-37.62	-27.88
Tianjin	196.68	121.86	135.89	0.01	0.02	-72.87	-52.46	-37.05	-26.67
Huhehaote	148.35	108.76	119.88	0.67	0.66	-39.02	-27.49	-26.30	-18.53
Xilinguole	116.51	48.35	64.11	0.56	0.57	-72.47	-73.17	-62.20	-62.80
Kuele	487.96	163.88	559.67	0.57	0.55	-315.26	123.77	-64.61	25.37
Hami	238.74	146.58	453.88	0.64	0.81	-163.50	-50.57	-68.48	-21.18
Akesu	738.39	236.09	827.03	0.45	0.50	-503.61	79.95	-68.20	10.83
Jiuquan	653.77	320.55	464.01	0.19	0.34	-338.29	-227.17	-51.74	-34.75
Baiyin	295.84	120.45	155.93	0.46	0.69	-174.16	-137.55	-58.87	-46.50
Average	345.40	155.23	325.48	0.41	0.48	-196.29	-47.72	-52.79	-22.46

653



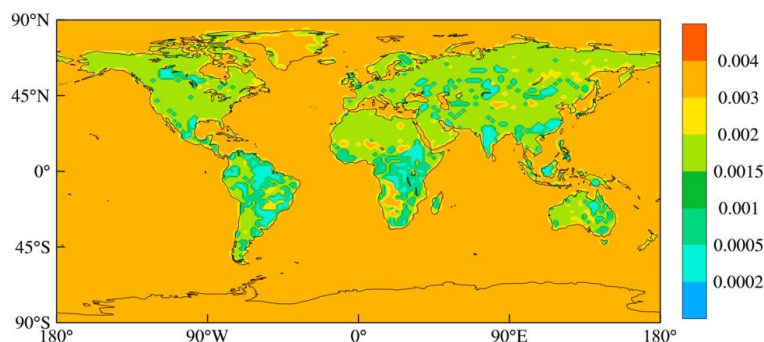
654

655 Fig 1. Updated input data of global M_{clay} (%). Data is derived from Shangguan et al. (2014) and is re-gridded to $2^\circ \times$
656 2.5° horizontal resolution in the model.



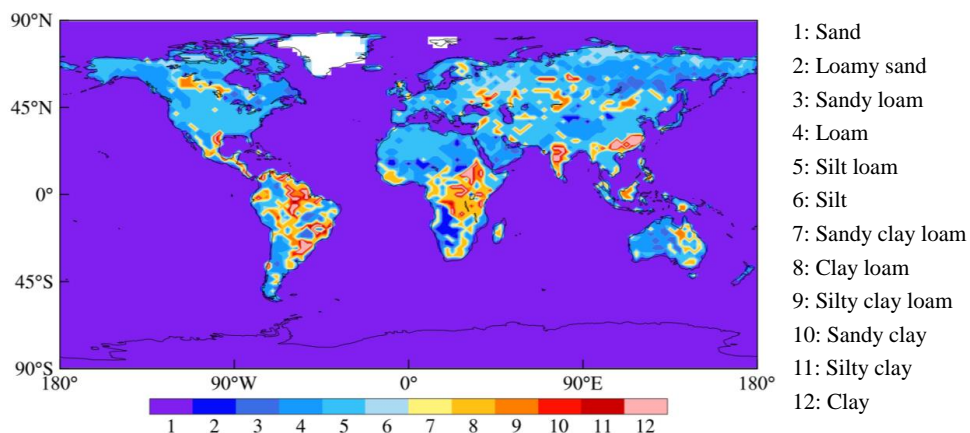
657

658 Fig 2. Monthly updated input data of global aerodynamic roughness length (Z_0) (cm) in March (left) and April (right).
659 Data is derived from Prigent et al. (2005) and re-gridded to $2^\circ \times 2.5^\circ$ horizontal resolution in the model.

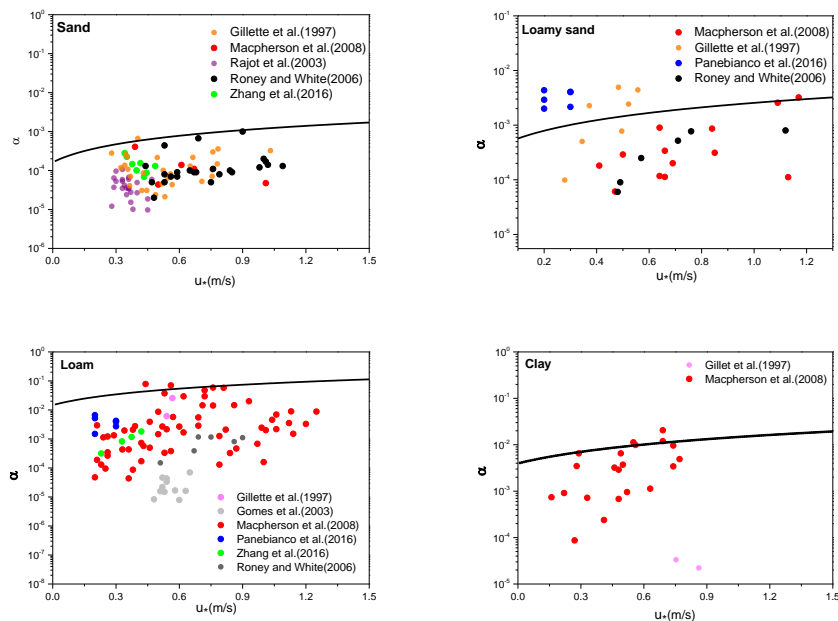


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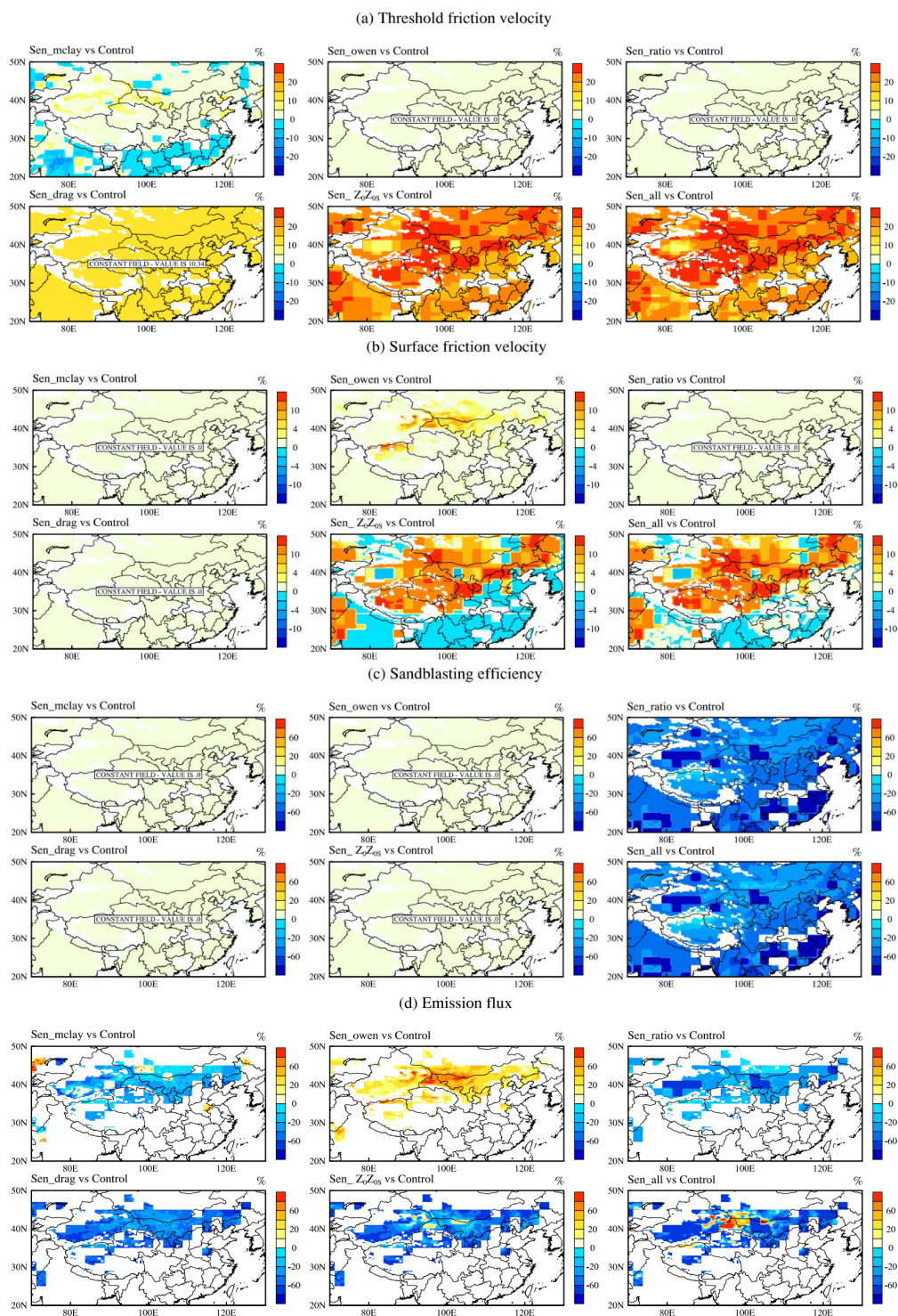
661 Fig 3. Updated global map of smooth roughness length (Z_{0s}) estimated from the empirical relationship with soil
662 texture.



663
 664 Fig 4. Global soil texture map based on the USDA classification.



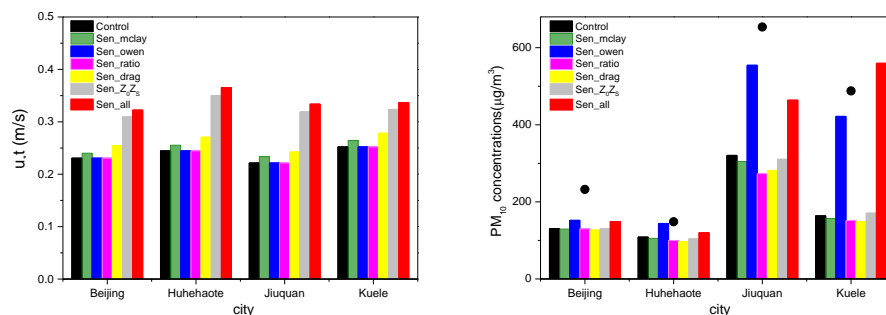
665
 666 Fig 5. Updated sandblasting efficiency α as a function of surface friction velocity u_* , following Lu and Shao (1999)
 667 for sand, loamy loam, loam and clay and observations from the literature.



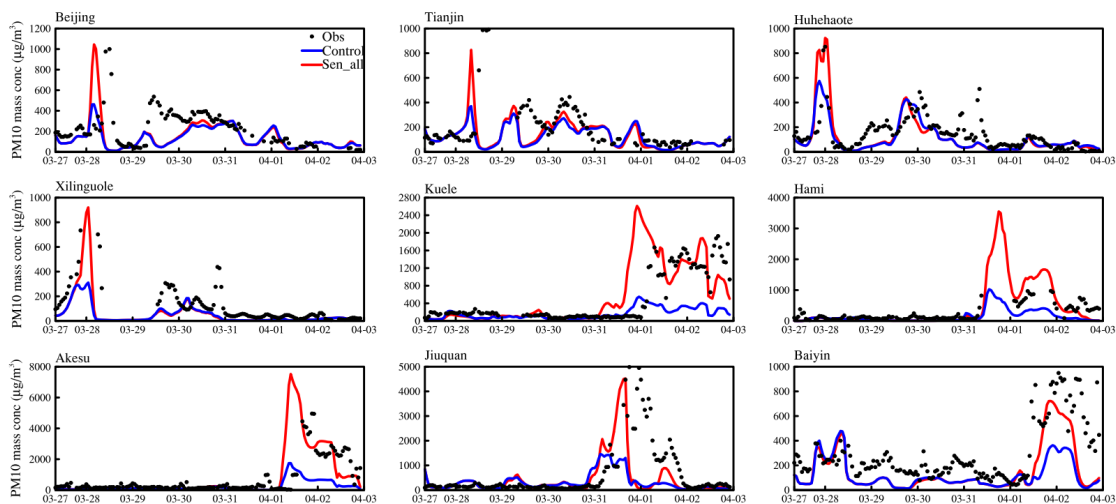
668 Fig 6. Relative difference (%) in simulated averaged threshold friction velocity u_{*t} (a), surface friction velocity u_*



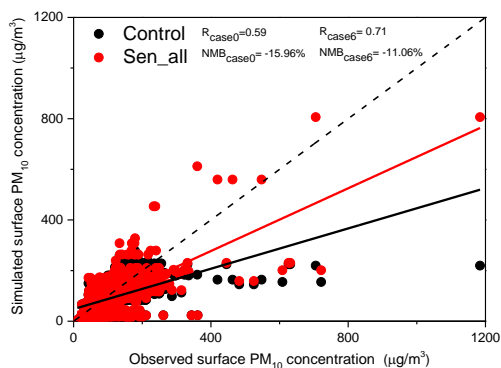
669 (b), sandblasting efficiency α (c) and emission flux (d) between sensitivity simulations and control run during the
670 study period.



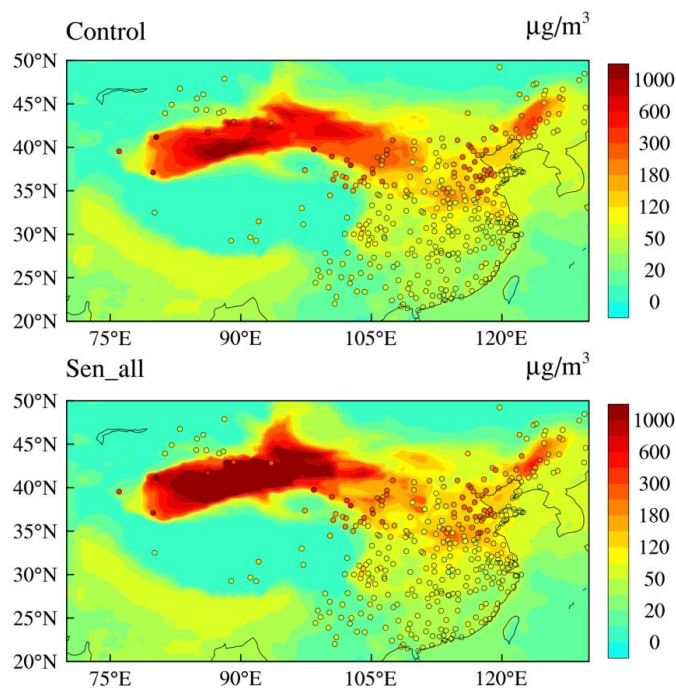
671 Fig 7. Comparisons of simulated averaged threshold friction velocity u_{+t} (left) and PM_{10} concentrations (right) at
672 selected sites. Black dots in right figure indicate the observed averaged PM_{10} concentrations.



673 Fig 8. Temporal variation of hourly PM_{10} concentrations from observations (black dots) and simulations of Control
674 run (blue line) and Sen_{all} (red line) during the study period at nine selected sites.



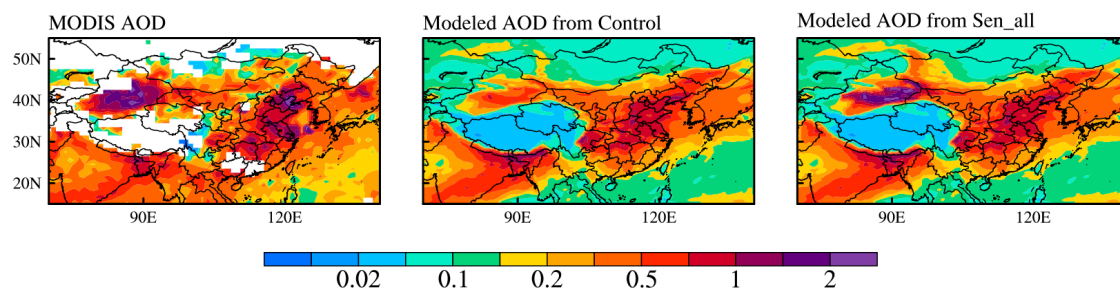
675 Fig 9. Comparison of modelled and measured surface PM₁₀ concentrations at observational sites. The dotted line is
676 the 1:1 line. Model results are taken from Control run and Sen_all respectively.



677
678 Fig 10. Comparison of simulated averaged PM₁₀ surface concentrations from Control run (top) and Sen_all (bottom)
679 with the observed values.



680



681

682 Fig 11. Spatial distribution of MODIS retrieved AOD at 550nm (left column) and simulated AOD at 550 nm from
683 Control run (middle column) and Sen_all (right column). The simulation results are extracted at 13:00 local time to
684 match the MODIS observation time.