

Effects of marine fuel sulfur restrictions on particle number concentrations and size distributions in ship plumes at the Baltic Sea

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Abstract. Exhaust emissions from shipping are a major contributor to particle concentrations in coastal and marine areas. Previously, marine fuel sulfur content (FSC) was restricted globally to 4.5 m/m% but the limit was changed to 3.5 m/m% at the beginning of 2012 and further down to 0.5 m/m% in January 2020. In sulfur emission control areas (SECA), the limits are stricter; FSC restriction was originally 1.50 m/m% but it decreased first to 1.00 m/m% in July 2010 and again to 0.10 m/m% in January 2015. In this work, the effects of the FSC restrictions on particle number concentrations (PNC) and size distributions (NSD) are studied at the Baltic Sea SECA. Measurements were made on a small island (Utö, Finland; 59° 46'50N, 21° 22'23E) between 2007 and 2016. Ship plumes were extracted from the particle number size distribution data, and the effects of the FSC restrictions on the observed plumes as well as on the ~~total~~ ambient concentrations were investigated.

Altogether 42 322 analyzable plumes were identified during the 10-year measurement period. The results showed that both changes in the FSC restrictions reduced the PNC of the plumes. The latter restriction (to 0.10 m/m% in January 2015) decreased also the ~~total~~ ambient particle number concentrations, as a significant portion of particles in the area originated from ship plumes that were diluted beyond the plume detection limits. The overall change in the PNC of the plumes and ambient air was 27 and 32 %, respectively, for the total FSC change from 1.50 to 0.10 m/m%. The decrease in plume particle number concentration was caused mostly by a decrease in the concentration of particle sizes of ~ 335 - 1434 nm. The latter restriction also reduced the [count-mediangeometric mean](#) diameter of the particles, which was probably caused by the fuel type change from residual oil to distillates during the latter restriction. The PNC was larger for the plumes measured in daytime compared to those measured in nighttime likely because of the photochemical aging of particles due to UV-light. The difference decreased with the reducing FSC indicating that lower FSC has also an impact on the atmospheric processing of ship plumes.

1 Introduction

Particulate matter (PM) from shipping contributes to a significant fraction of PM₁₀, PM_{2.5}, and PM₁ in European coastal areas (1-7, 1-14, and ≥11 %, respectively; (Viana et al., 2014). PM from shipping emissions can also be transported hundreds of kilometers inland (Lv et al., 2018), and have been linked to increased cardiovascular mortality and morbidity (Lin et al., 2018; [Partanen et al., 2013](#)). The reduction of PM_{2.5} emissions from shipping has been [shown](#) [estimated](#) to reduce the negative health

effects (Barregard et al., 2019; [Chen et al., 2019](#); Partanen et al., 2013; Sofiev et al., 2018). In multiple studies, shipping has also been reported to be a significant contributor to atmospheric particle number concentrations (PNC) both in the sea and coastal areas, especially in the ultrafine size range (Ausmeel et al., 2019; Gobbi et al., 2020; Karl et al., 2020; Kivekäs et al., 2014; Kukkonen et al., 2016; Zanatta et al., 2020). Atmospheric PNC has been considered to cause significant negative health effects such as airway inflammation, impaired lung function, and cardiovascular mortality (Hennig et al., 2018; Strak et al., 2012).

Sulfur dioxide (SO_2) is a gaseous air pollutant that is harmful to human health and the environment. Fossil fuels commonly contain sulfur that forms sulfur oxides (SO_x) in combustion (Arnold et al., 2006; Hoang and Pham, 2018). Most of the sulfur in diesel fuels are oxidized to SO_2 (Kozak and Merksiz, 2005; Walker, 2004), however, some SO_2 is further oxidized to SO_3 (3-5 %) (Cordtz et al., 2013). Already in the high-temperature exhaust, SO_3 can react with water molecules resulting in the formation of gaseous sulfuric acid (GSA) (Arnold et al., 2012; Rönkkö et al., 2013). When the exhaust is cooled immediately after its emission into the atmosphere, the saturation ratio of GSA increases quickly leading to the homogenous nucleation of sulfuric acid and water (Arnold et al., 2006; Kozak and Merksiz, 2005) possibly supported by the exhaust of partly oxidized organic compounds (Arnold et al., 2012; Pirjola et al., 2015). After leaving the exhaust pipe, sulfuric acid contributes also to particle growth by condensing on existing particles formed in an engine, entrained aerosol particles, or newly formed nanoparticles (Arnold et al., 2006; Riipinen et al., 2012). It should be noted that the oxidation of emitted SO_2 and the formation of sulfuric acid continues in the atmosphere leading also to the secondary formation of sulfuric acid and particulate matter (see e.g. Mylläri et al., 2016).

Fuel sulfur content (FSC) has a large effect on the formed particles and it has been reported to have a strong correlation with particle mass and number concentrations in the exhaust emissions of shipping (Alföldy et al., 2013; Diesch et al., 2013). The effect is especially clear in the nanoparticle size range (Ushakov et al., 2013). Using fuels with lower FSC has been shown to reduce PM emissions (Lehtoranta et al., 2019) but FSC has also an influence on the chemical composition of exhaust particles. Typically, ship plume particles consist of organic matter (OM), black carbon (BC), and an inorganic fraction that is almost entirely composed of sulfuric acid (Kumar et al., 2013). At least sulfate and OM concentrations have been reported to be dependent on FSC (Lack et al., 2009). Also, the emission factors (EF) of SO_2 and organic carbon (OC) have been observed to decrease when fuels with lower FSC were used (Wu et al., 2019). In the study of Celik et al. (2019), OM was twice as large for fuels with FSC of 2.5 m/m% (mass per mass percentage) compared to FSC of 0.5 m/m%.

To reduce the harmful effects of ship emissions, restrictions for marine FSC have been set for Emission control areas (ECAs) and global marine transportation (IMO, 2008). Previously, the limit of FSC was globally 4.5 m/m% but it was changed to 3.5 m/m% at the beginning of 2012 and further down to 0.5 m/m% in January 2020. In sulfur emission control areas (SECAs) specifically, the limit of FSC is lower and was changed from 1.50 m/m% to 1.00 m/m% in July 2010 and further to 0.10 m/m% in January 2015. In addition to these limits, the maximum FSC is limited to 0.10 m/m% (from the beginning of January 2010) for the berthed ships in EU ports (European Parliament and the Council of the European Union, 2005). The FSC limits and emissions control areas are described in more detail in (IMO Marpol Annex VI, Regulation 14). The decline of the real FSC

65 of the marine vessels has been shown to follow the FSC restrictions well (Kattner et al., 2015; Pirjola et al., 2014), and therefore
the FSC limits can be expected to resemble the real maximum FSC of used fuels. However, Pirjola et al. (2014) reported that
during the FSC restriction period of 1.50 m/m% (before July 2010), some vessels already used fuels with less than 1.00 m/m%
FSC. This implies that the real FSC change has followed a gradual – rather than a step-wise – behavior at least for the July
2010 restriction enforcement.

70 The long-term effects of shipping emissions on atmospheric particle characteristics have been studied only in a few studies in
SECAs (Ausmeel et al. 2019; Kivekäs et al., 2014). In this study, a 10-year time series of particle number size distribution
(NSD) data (from 2007 to 2016), measured near several busy shipping lanes on the island of Utö in the Baltic Sea SECA, was
investigated. The effects of two changes in the FSC restrictions (from 1.50 to 1.00 and from 1.00 to 0.10 m/m%) on the PNC
and NSD of ship exhaust plumes and ambient aerosol are discussed respectively. Additionally, the impacts of photochemical
75 aging on ship exhaust plumes are investigated.

2 Experimental

2.1 Utö measurement site

The PNC and NSD data were measured at the atmospheric measurement station of the Finnish Meteorological Institute (FMI)
at Utö between January 11th, 2007, and December 31st, 2016. Utö is a small island (0.81 km²) at the southern edge of the
80 Finnish archipelago in the Baltic Sea (59° 46'50N, 21° 22'23E) and it has a year-round population of 20 to 30 inhabitants. The
exact location of the measurement station on the island is shown in Fig. 1. The measurement station and location have been
described in detail previously (Engler et al., (2007); and Laakso et al., 2018).

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2.2 Ship routes and vessel types near Utö

Utö island is within the vicinity of busy shipping lanes. The closest shipping lane is on the western side of the Utö island, about
85 500 m away from the shore, and another busy shipping lane is 6 km south from Utö. Besides, Utö has a pilot station and harbor
with a regular passenger ferry connection to the mainland. The shipping routes near Utö are shown in Fig. 1. In Fig. 1, the data
(white points) represents the Automatic Identification System (AIS) data from the Baltic Marine Environment Protection
Commission (HELCOM). The AIS is a system that automatically produces and transmits information about vessels to other
vessels and coastal authorities (IMO AIS transponders). This information includes identity, type, position, course, speed, and
90 navigational status of the vessel, and other safety-related information. Only a fraction of the AIS signals were plotted as the
total number of signals was excessive. In addition to estimating the presence and movement of vessels around Utö, the fractions
of different types of vessels were also attained from the AIS data for this work. The AIS data were analyzed only for the vessels
with International Maritime Organization (IMO) numbers as these are larger vessels, and smaller vessels without IMO numbers
are not likely to contribute significantly to identified plumes, as shown by Kivekäs et al. (2014).

95 The sectors of the nearby shipping lane, distant shipping lanes, the harbor of Utö, and aging (aging discussed in Sect. 3.6) are identified by colors in Fig. 1. While there is no clear upper limit to the distance from which the plumes may be arriving in the measurement station, ~~an~~ limit radius (blue circle) was set at 25 km to count the number of AIS signals. This limit was chosen so that the harbors of other islands in northwest and northeast directions would not be included in the data as they would have increased the number of AIS signals from berthed ferries extensively.

100 The AIS markings from ships within the 25 km radius from Utö were divided into 10°-sectors (Fig. 2) and the activity of different types of vessels was estimated in each sector. As HELCOM AIS signals are recorded every 6 minutes, the residence time (RT) of vessels occupying different sectors could be estimated by giving every AIS signal a 6 min RT. These RTs were plotted for the three FSC restriction periods and are presented in Fig. 2. Based on the HELCOM AIS data, the vessels were classified into six different categories best representing their specific type: cargo ships, large passenger ships, medium-sized
105 passenger vessels, large work vessels, small vessels, and others.

The majority of the vessels around Utö were cargo and large passenger ships. While the amount of passenger vessel traffic and the distribution of traffic between the sectors remain almost the same between the different periods, the total amount of shipping, estimated from total RT of vessels, reduced from 71.1 (FSC < 1.50 m/m%) to 64.2 (FSC < 1.00 m/m%) and to 48.7 h d⁻¹ (FSC < 0.10 m/m%). The reduction was mostly caused by reduced cargo shipping activity. Notable is also that there were
110 almost no vessels detected in the eastern directions (70-130°). The RT of vessels on the western side of the island was low because the ships on the nearby shipping lane pass through the sectors quickly. The high number of passenger vessels at the direction of 330-340° was because of the signals from berthed ferries at the Utö ferry harbor.

2.3 Particle measurement instrumentation

The aerosol was sampled with a PM_{2.5} inlet, which was positioned on the roof of the measurement container approximately
115 3 m above the ground. NSD and PNC were measured with a differential mobility particle sizer (DMPS). Before entering the DMPS, the sample aerosol was dried with a Nafion dryer. The DMPS consisted of a 28 cm long Vienna-type differential mobility analyzer (DMA) and a continuous flow condensation particle counter₋(CPC). [The model of the CPC was TSI model 3010.](#) The theory and the response functions of DMA have been first described in Hoppel, (1978). The particle size range of the DMPS was 7-53700 nm, and the size spectra were given in 30 size bins. [DMA losses were assumed to be 15% independent of the particle size based on the laboratory tests.](#)
120 [The time resolution of the instrument was 5 minutes 20 seconds. During the measurements, parts of the DMPS including the CPC the DMA and CPC were changed a few times due to malfunctions.](#)

2.4 Data processing method for plume identification

First, the concentration peaks caused by the local sources (e.g. passing by of a tractor) were removed from the DMPS data. Identification of the peaks caused by the local sources was done by examining the time series of [ambient_{total}](#) number size
125 distributions (NSD_{tot}) and [ambient_{total}](#) particle number concentrations (PNC_{tot}) visually as the concentration peaks from tractors, for example, were expected to last only tens of seconds whereas the durations of the ship plumes are measured in

minutes. Secondly, the days with only four or fewer measurement cycles were removed from further analysis to ensure the representativeness of the data. The remaining data was characterized as the cleaned data. The coverage of the cleaned data varied between ~26 - 96% depending on the year. The figure presenting the data coverage is presented in the supplemental data (Fig. S1).

Plumes emitted from the ships were separated from the cleaned DMPS data by using a modified version of the plume detection method developed and tested by Kivekäs et al. (2014). The method was originally developed for [a](#)-scanning mobility particle sizer (SMPS) data, but it is applicable also for the DMPS data. Since the method is described in detail in Kivekäs et al. (2014), only a short description of the method, the modifications made to it, and otherwise relevant information are given below.

The identification of plumes assumes that ship plumes entail significantly higher particle concentration than the natural background particle number concentration (PNC_{bg}). The PNC_{bg} was defined as a sliding 25th percentile of 40 consecutive measurements and the excess particle number concentration (PNC_e) was calculated by subtracting the background concentration from the PNC_{tot} . Maxima of the PNC_e were examined if they fulfilled at least one of the two criteria set for the plume detection:

- 1) The PNC_e had to be at least 500 cm^{-3} or
- 2) the ratio of the PNC_{tot} to PNC_{bg} (R_e) had to be at least 1.5.

When the criteria was fulfilled, the PNC_e was considered to be caused by a ship plume and is hereafter later in this study referred to as particle number concentration in the plume (PNC_{pl}). In Fig. 3, a typical day with the plumes exceeding both limits is presented. Periods, when the red dotted line exceeds the black vertical line, are considered ship plumes. If two consecutive peaks exceed the detection limit and are separated by lower concentration, they are considered as two separate plumes. In terms of number size distributions, the studied NSDs are excess number size distributions (NSD_e), which are calculated by subtracting the background number size distributions (NSD_{bg}) from the ~~ambient~~^{total} number size distributions (NSD_{tot}).

As shown by Kivekäs et al. (2014), fast and large changes in the PNC_{bg} can cause an error in the detected plumes during these changes. These errors occurred as the sliding 25th percentiles used for defining PNC_{bg} reacted roughly 10 measurement points too early to decreasing concentrations and 10 measurement points too late to increasing concentrations. This could have caused distortions to the plumes measured during the fast PNC_{bg} change, or the fast changes could themselves have been counted as plumes. Therefore, the fast PNC_{bg} changes were excluded from the data was divided into valid and invalid data periods by counting absolute and the relative change rates of PNC_{bg} and calculating the smoothed sliding average of six consecutive measurement points. When the absolute or relative change of the smoothed value was greater than $\pm 53\text{ cm}^{-3}$ or $\pm 5\%$ respectively, the data was marked as invalid. These values were set so that they corresponded for 67 % of what was needed to define a plume. This validation procedure is described in detail in Kivekäs et al. (2014). Another factor that could have caused uncertainty to the detected plumes is the relatively long measurement cycle of the DMPS (5 min 20 s). However, the large data set renders the effect of these uncertainties to be small. The errors related to the long measurement cycle are discussed in more detail in the supplemental material.

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3 Results and discussions

3.1 Detected exhaust plumes

During the 10-year measurement period, a total of 71 811 ship exhaust plumes were observed, and, based on the data validation procedure described in the experimental section, 42 322 (~ 59 %) of these were considered valid and representative. The number and the frequency of the valid plumes are shown in Fig. 4 as a function of wind direction (in 10° sectors). The two directions with the highest numbers of plumes (Fig. 4A) were 220-250° and 330-340°. The peak observed in the 330-340° direction corresponds to the highest number of AIS markings and it is the direction where both the Utö ferry harbor and pilot station are located. The second-largest number of plumes came from 220-250° direction, where only a few AIS signals were observed within a 25 km radius. Similarly, a noticeable number of plumes were detected coming from a direction of 70-130°, even though there were very few AIS signals in that direction. This indicates that some of the detected plumes had a distant source, which for these directions, is likely to be the shipping lanes situated further away. However, all the emissions from these distant shipping lanes are not expected to be detected as plumes as they [are](#) probably diluted to be unidentifiable by the detection method. Therefore, it is likely that the plumes detected from these directions have originally been the plumes with high initial concentrations. When the numbers of plumes are normalized with the times of wind blowing from each direction (Fig. 1C), the plume frequencies from different wind directions are more even (Fig. 4B). Exceptions are the direction of 50 - 110° with the longest distance to the shipping lane, and the direction of the Utö ferry harbor (330-340°), where the residence time of the berthed vessels is high (Fig. 2).

3.2 Particle number concentrations of exhaust plumes

Particle number concentrations of the plumes (PNC_{pi}) were averaged annually, and the annual data were averaged over the FSC restriction periods (Fig. 5). The results showed that the restrictions of FSC were effective; the average PNC_{pi} decreased from 1379 ± 30 (1.50 m/m%) to 1290 ± 170 (1.00 m/m%) and further down to 1010 ± 90 cm^{-3} (0.10 m/m%). Moreover, the two lowest annual average PNC_{pi} (940 ± 990 and 1070 ± 920 cm^{-3}) were observed during the 0.10 m/m% FSC restriction period and three out of the four highest annual averages (1430 ± 2080 , 1370 ± 1780 , and 1355 ± 2120 cm^{-3}) were measured during the 1.50 m/m% restriction period. Notable is also that the standard deviations are significantly lower for the annual averages measured with the tighter FSC restriction periods indicating that especially the fraction of plumes with very high concentrations have decreased. However, the highest annual average (1570 ± 2100 cm^{-3}), as well as the largest variability in the PNC_{pi} (the widest 25th and 75th percentile range), were observed in 2014, unexpectedly. Numerous intense nucleation events occurred in 2014, however, they are likely to have been identified as invalid data and therefore do not affect the detected plumes.

PNC_{pi} attained in this work are similar to the concentrations of plume observed in similar studies; [An-The](#) median PNC of 7050 and 1470 cm^{-3} were measured during winter and summertime, respectively, in southern Sweden (Ausmeel et al. 2019). A noticeably higher PNC concentration (2100 cm^{-3}) of ship plumes was measured with the airborne measurements right over

the shipping lane at the southern Baltic Sea by Zanatta et al. (2020) during the FSC restriction of 0.10 m/m%. Concentrations observed by Zanatta et al. (2020) were higher likely because of the fresher plumes (plume age ~200-800 s). Also, the small sample size (only 9 continuous periods of single or multiple exhaust plumes each lasting a couple of minutes at maximum) might cause statistical uncertainty to the results. In addition to previous the local meteorology, dilution, and aging process of the plume, as well as the measured particle size range, may influence the measured concentrations ~~variation of the results~~ at least to some degree.

In direct emission measurements, PNC of raw ship exhaust emission from idling have been reported being $432 \pm 1.83 \cdot 10^8$ for idling and $24.6 \pm 0.26 \cdot 10^7 \text{ cm}^{-3}$ for HFO and MDO, respectively (Anderson, et al., 2015b) and in the range of 10^6 - 10^7 cm^{-3} after the dilution by a ratio of 8 (Karl et al., 2020). Therefore, fresh exhaust emission PNC can be expected to be around 10^8 cm^{-3} . Similar concentrations of 0.9 - $1.94 \cdot 10^8 \text{ cm}^{-3}$, depending on fuel and engine load, were found by Zhou et al. (2019). As the concentrations measured from the plumes are in the range of 10^3 cm^{-3} in this study, the dilution ratio for the measured plumes is estimated to be approximately $1:10^5$ to $1:10^7$.

3.3 Particle number size distribution of the exhaust plumes

Average number size distributions observed for the plumes (NSD_{pl}) are shown as a function of wind direction in Fig. 6. The NSD_{pl} were first divided into 10° -degree sectors and then averaged. These averaged NSD_{pl} were interpolated to form a continuous distribution. If the plumes had been plotted individually from every angle, the number of plumes from certain angles would have been very low (even zero), and the individual plumes would have had a disproportional effect on the distribution.

The plumes with the highest PNC_{pl} arrived from the sector of 190 - 240° and the plumes with the lowest concentrations from the sector of 50 - 90° (Fig. 6). Near Utö, the direction of 190 - 240° was the direction with the highest shipping activity excluding the berthed ferries at Utö harbor (Fig. 2). Especially the cargo vessel activity in that direction was high. In contrast, in the direction of 50 - 90° the shipping activity was almost nonexistent near Utö. PNC_{pl} from the direction of the ferry harbor of Utö (310 - 340°) was generally lower than PNC_{pl} from the plumes arriving from 190 - 240° , despite many of the plumes that were expected to originate from the berthed ferries at the harbor (Fig. 1D). This might be due to the small size of the berthed ferries, which produced smaller amounts of particulate matter. The small ferries also operated on distillate fuels during all the FSC restrictions. Distillate fuels have been shown to produce lower particulate emissions on lower engine loads (Zhou et al., 2019). Stricter FSC limits were seen as a decreased PNC_{pl} in the direction of nearby ship traffic (Fig. 6). The effect was visible after both restriction changes, but the decrease was significantly larger after the change of FSC restriction from 1.00 to 0.10 m/m%. Besides PNC_{pl} , the change of FSC limit from 1.00 to 0.10 m/m% was observed to reduce the particle size. To the knowledge of the authors, no previous studies are reporting the effects of FSC on particle size and concentrations from atmospheric exhaust plume measurements. Furthermore, the results from direct emission measurements are divided in literature; Zetterdahl et al. (2016) reported that the stricter FSC limits reduce the particle size, but not the produced PNC_{pl} ,

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225 whereas, for example, Streibel et al. (2017) found lower emission factors for PNC_{pl} and larger particle size when using distillate fuels instead of the HFO.

230 ~~From angles 190-210°, the PNC_{pl} s were significantly higher than from angles 260-280°. The reason for this was investigated by calculating the fractions of specific ship types observed at these angles. These fractions are presented in the supplemental material (Fig. S2). No significant difference was observed except the higher fraction of Roll-on/roll-off passenger vessels in angles 190-210°. Therefore, the difference in the concentrations is not likely to be related to different ship fleets but it can be caused by the meteorology of the area. The lower concentrations of the plumes in angles 260-280° might be caused by higher vertical dilution or the plumes might be transported higher up in the marine boundary layer. However, the effect of meteorology could not be validated.~~

235 In terms of plume number size distributions, it is notable that most of the plume particles were smaller than 100 nm in diameter and case of the FSC restriction of 0.10 m/m% even smaller than 50 nm (Fig. 6). Therefore, the contribution of ship emissions to particle mass is likely to be small, and the characterization of ship plumes should be based on number concentrations rather than mass concentrations in plumes. This has also been proposed by Viana et al. (2014).

240 ~~Count-median-diameters~~Geometric mean diameters (GMD_{GMD}) of particles in plumes were calculated and are shown as a function of wind direction in Fig. 7. The GMD_{GMD} remained almost unchanged during the change from the FSC restrictions of 1.50 to 1.00 m/m%. However, the FSC restriction change from 1.00 to 0.10 m/m% significantly reduced the GMD_{GMD} of particles arriving from all wind directions. Overall, the largest GMD_{GMD} of the plumes was observed from eastern and southeastern directions (60-170°), where the distance to the shipping lanes is large, and the smallest GMD_{GMD} for the plumes from north to northwestern directions (300-330°), which is the direction of Utö ferry harbor. This variation of the GMD_{GMD} from a minimum of 36, 39, and 31 nm to the maximum of 58, 55, and 51 nm for the FSC restrictions of 1.50, 1.00, and 0.10 m/m%, respectively, implies that particles either grow significantly or small particles evaporate when they are transported in the atmosphere. Zanatta et al. (2020) found similar behavior of increasing particle size as the age of the plumes increased. ~~They proposed this to be a result of coagulation and dilution-controlled processes.~~ The atmospheric processing of plume particles is discussed more in Sect. 3.6.

250 In direct emission measurements, the FSC has been related especially to produced nanoparticles, and the effect on the larger particles has been observed to be smaller (Ushakov et al., 2013). Therefore, in this work, particles were divided into two size ranges: 7-1434 nm and 1455-537499 nm, and their concentration changes were studied separately. The two size ranges were calculated from the DMPS bins of 1-21 (midpoints of the size bins were 7-134 nm) and 22-30 (midpoints of the size bins were 155-499 nm). ~~The size ranges were chosen based on testing to find a limit for the effects of the FSC restrictions on PNC_{pl} s.~~ Average NSD_{pl} s for all the plumes during different sulfur restrictions are presented in the supplemental material (Fig. S3).

255 ~~From these distributions, it is visible that the particle size of 144 nm acts as a limit for the effects of the FSC restrictions.~~ Based on the two size ranges, FSC restrictions were more effective in 7-1434 nm particles. The PNC_{pl} of the smaller particles decreased clearly, while that of the larger particles remained almost unchanged when the FSC limit became tighter (Fig. 8). PNC_{pl} were also calculated using slightly different size ranges and a figure with ranges, 7-125416 nm, and 12534-537499 nm

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is presented in the supplemental material (Fig. S2S4). The figure shows that if the lower limit for the larger particle size is smaller than 1455 nm, the change in the FSC restriction is also seen in the larger size class as reduced PNC_{pl} after the FSC change from 1.00 to 0.10 m/m%. Therefore, the effect of sulfur restrictions seems to be mostly limited to the PNC_{pl} in the particle sizes smaller than 1434 nm.

The PNC_{pl} in the smaller size range (7-1434 nm) was observed to be highest when the wind was blowing from the direction of 190-340° (Fig. 8), where there is plenty of shipping activity nearby e.g. nearby shipping lane and Utö harbor (see Fig. 1D). In contrast, the PNC_{pl} of the larger size range (1455-537499 nm) was highest when the wind was blowing from the direction of 60-210°, where there is a significant amount of ship traffic at longer distances from the measurement site (see Fig. 1A). This supports the assumption that particles grow in the atmosphere to larger particle sizes through atmospheric processing (e.g. coagulation or condensation) and the increased GMDCMD is not only due to the evaporation of small particles. For the larger particle size range (1455-537499 nm), the similarity of the shapes of the PNC_{pl} patterns in different FSC restriction periods indicated also that the ship fleet was responsible for the plumes and the shipping routes have stayed relatively constant between different restriction periods. This strengthens the assumption that the differences observed in this study were caused by the FSC restriction changes and not the change in the ship fleet.

3.4 Plume number size distributions for shipping lane sectors

To investigate the number size distributions of plumes (NSD_{pl}) in more detail, two sectors, distant shipping lanes (70-130°) and nearby shipping lanes (280-300°), were chosen for a separate analysis. The sectors were selected so that the distant shipping lane represented plumes that were transported longer distances whereas nearby shipping lanes represented fresher plumes. These assumptions of plume ages are based on the locations of the shipping lanes (Fig. 1), the GMDCMDs of the plumes (Fig. 7), and the dependency of particle concentrations on wind directions (Fig. 8). Even though the GMDCMDs for the plumes from the harbor of Utö (320-350°) were the smallest, they were not included as a large number of these plumes were expected to be originated from the berthed ferries operating on fuel with FSC < 0.10 m/m% during all three FSC restriction periods. In the case of distant shipping lanes (70-130°), it was ensured from Fig. 2 that there were no additional vessels between the measurement station and the shipping lanes (a distance of ~25 km). In the nearby shipping lanes sector, almost all of the AIS markings were from cargo vessels from the nearby shipping lane (a distance of ~1-2 km).

For both the nearby and distant shipping lane sectors, the shapes of the NSD_{pl} remain nearly unchanged after the change from the FSC restriction of 1.50 to 1.00 m/m% (Fig. 9). In the distant shipping lanes sector, the GMDCMD shifted from 51 to 49 nm, and in the nearby shipping lanes sector from 40 to 41 nm, when the FSC restriction changed from 1.50 to 1.00 m/m%. After the FSC restriction changed from 1.00 to 0.10 m/m%, the shape of the NSD_{pl} altered more clearly as the GMDCMD of the particles shifts to significantly smaller particle sizes; from 49 to 44 nm in the distant shipping lanes sector (70-130°) and from 41 to 32 nm in the nearby shipping lane sector (280-300°). The reduction of the particle size with decreasing FSC has also been reported by Zetterdahl et al. (2016), who observed the decrease in the particle size mode from 34 to 19 nm in the raw exhaust gas for the ships using fuels with FSC of 0.48 and 0.092 m/m%, respectively.

The NSD_{pi} from the distant shipping lanes sector (70-130°) were observed to be bimodal consisting of a nucleation mode with larger particle number concentrations at ~30-60 nm and a soot mode with smaller number concentrations at ~100-200 nm. The mode in the particle sizes larger than 100 nm in ship emission exhaust has been identified to be a soot mode (Ntziachristos et al., 2016). The similar shape of the NSD has been attained for ship diesel exhaust emissions from direct emission measurements by Corbin et al. (2020) having similar proportions between the nucleation and soot modes and with the nucleation mode particle size being slightly smaller ~35 nm. This might be because the plumes measured in this study had time to grow in the atmosphere by coagulation, as indicated in Fig. 7. [Increasing particle size via coagulation has been also reported by \(Petzold et al., 2008\).](#) In the literature, NSDs of ship plumes have been typically observed to be unimodal (Celik et al., 2019) or bimodal (Petzold et al., 2008), whereas emissions measured directly from the engine exhaust are often bimodal (Anderson, Salo, & Fridell, 2015a; Ushakov et al., 2013). In addition to the former, also trimodal NSDs measured directly from the marine engine exhaust have been reported by (Tian et al., 2014), who found a distinct volatile mode at 15 nm and two combustion-related modes at 38 and 155 nm.

3.5 ~~Total P~~particle concentrations and size distributions in ambient air

The main focus of this work was to investigate the effects of the FSC restrictions on ship plumes, but in addition to this, shipping emissions are expected to contribute also to the ambient ~~total~~ particle number concentrations (PNC_{tot} , including background concentrations, PNC_{pi} , and the part of PNC_e that was not considered as valid plumes) at Utö. The annual average PNC_{tot} was calculated for each year (2007-2016) and averaged for the different FSC restriction periods (Figure 10). The average PNC_{tot} was almost the same during the FSC restriction of 1.50 (2910 ± 110) and 1.00 m/m% ($2960 \pm 1140 \text{ cm}^{-3}$) However, after the FSC restriction was set to 0.10 m/m%, the average PNC_{tot} decreased significantly, to $1970 \pm 200 \text{ cm}^{-3}$, probably because of the switch between heavy residual fuels and lighter distillates. The distillate fuels with lower FSC have been shown to produce lower particle emissions (Zhou et al., 2019).

Despite the more stringent FSC restriction, the annual average PNC_{tot} was as high as $4840 \pm 5270 \text{ cm}^{-3}$ in 2014. This was attributed, at least partially, to the numerous intense nucleation events occurring in that year (see [an](#) example of nucleation event in supplemental material Fig. [S3S5](#)). These nucleation events increased the concentrations of small particles by several times, and the concentration levels stayed elevated for extended periods. These events fulfilled the criteria for the nucleation event set by Dal Maso et al. (2005): appearing of a new mode to the size distribution in the nucleation mode size range, this mode has to prevail over several hours and show signs of growth. These criteria differentiate from the ship plumes as the ship plumes last much shorter periods and do not show significant growth. Moreover, the plumes had even the peak of NSD_{pi} at larger particle sizes (~30-60 nm) compared to the definition of nucleation mode (3-25 nm) by Dal Maso et al., (2005). The existence of the nucleation events is also indicated by the much larger variability in the PNC_{tot} values in 2014 than for the other years (Fig. 10).

To further investigate the effects of FSC restriction changes on the ambient aerosol at Utö, the average ~~ambienttotal~~ number size distributions (NSD_{tot}) were plotted as a function of wind direction for the three FSC restriction periods (Fig. 11). The

325 highest PNC_{tot} was observed during the FSC restrictions of 1.50 and 1.00 m/m% with the wind directions between 90-120°
and 170-230°, with some elevated PNC_{tot} in and 270-290°. During the FSC restriction of 0.10 m/m%, the PNC_{tot} was
elevated also in the same directions, however, the average PNC_{tot} were significantly lower. The highest PNC_{tot} was observed
in the size range of 40 to 100 nm, which is a typical size range for ship emission particles (Alanen et al., 2020; Celik et al.,
2019; Jonsson et al., 2011; Petzold et al., 2008; Westerlund et al., 2015). All these sectors with elevated PNC_{tot} correlated well
330 with the distant shipping lanes (directions of distant shipping lanes estimated from figure 1A; 80-110°, 170-250°, 270-290°),
indicating that a large fraction of the measured background particle concentrations in the size range of 7-500-537 nm originated
from the distant, diluted ship plumes. A large fraction of PNC_{tot} being from diluted ship plumes is also supported by the fact
that the elevations of PNC_{tot} in directions of the distant shipping lanes decrease with the more stringent FSC restriction.
However, the effect of FSC restriction changes on the PNC_{tot} might be slightly exaggerated, as also the shipping activity in the
335 immediate area of Utö has decreased (Fig. 2). The change of ship traffic on the distant shipping lanes, however, was not directly
studied and might differ from the change seen in Fig. 2. In a previous study of by Kecorius et al. (2016), it was found that
ambient PNC_{tot} increased 26-53 % when the air mass was transported over the Baltic Sea, and most of the increase was related
to shipping traffic.

The ship exhaust plumes arriving from long distances are not observed as individual plumes as they do not cause detectable
340 peaks in the data due to excessive dilution. Excessive dilution causing ship plumes to become undetectable is presented also
in other studies: ship plumes were undetectable when plumes arrived from at least 100 km (Kecorius et al., 2016). Petzold et
al. (2008) estimated that in 24 hours, the ship exhaust plumes get entirely mixed with the marine boundary layer. Chen et al.
(2005) found that due to the dispersion, the lifetime of a plume was 2.5 ± 0.6 hours when the detection criteria were the same
as the second criteria used in this study ($R_e > 1.5$). As the FSC restrictions seem to affect both PNC_{pl} and PNC_{tot}, they have a
345 larger effect on the air quality than only the direct effects measured from nearby ship exhaust plumes would assume.

From the NSD_{pl}s and NSD_{tot}s, the masses of PM_{0.144} and PM_{0.537} (PM of particles smaller than 144 and 537 nm) were calculated.
Both PM_{0.144} and PM_{0.537} were calculated for the valid plumes and all measured plumes. Also, the contributions of both to
ambient PM_{0.144} and PM_{0.537} were calculated by taking into account the duration and number of the plumes (Table 1). Both
only the valid and all plumes were calculated to give approximate lower and upper limits for the contributions as valid plumes
350 almost certainly did not include all the plumes and all the measured plumes likely included also other phenomena in addition
to plumes. A steady decrease of the contributions of the ship plumes to PM_{0.144} and PM_{0.537} can be seen after the FSC restriction
changes with the contributions reducing overall from 5.5 - 14.0 % ($31.6 - 80.8 \text{ ng m}^{-3}$) to 3.9 - 8.9 % ($12.9 - 29.8 \text{ ng m}^{-3}$) for
PM_{0.144} and from 2.8 - 7.4 % ($106.3 - 283.3 \text{ ng m}^{-3}$) to 2.4 - 5.5 % ($60.0 - 136.2 \text{ ng m}^{-3}$) for PM_{0.537}. Similar values of 34 ± 19
ng m⁻³ (summer) and $18 \pm 8 \text{ ng m}^{-3}$ (winter) for PM_{0.15} during the FSC restriction of 0.10 m/m% have been reported by Ausmeel
355 et al., (2020). Contribution to PM_{0.5} reported by Ausmeel et al., (2020) is slightly lower compared to this study, $37 \pm 20 \text{ ng m}^{-3}$
(summer) and $29 \pm 13 \text{ ng m}^{-3}$ (winter). An important thing to note is that these contributions are only the direct contributions
of the detected plumes to ambient PM_{0.144} and PM_{0.537}. The real contributions are likely to be higher because of the contribution
from ship plumes diluted beyond detection. In calculating the masses of plume particles, a density of 1.23 g cm^{-3} was used.

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corresponding to the density of ship exhaust particles calculated based on the particle chemistry (Petzold et al., 2008). For ambient aerosol, a density of 1.10 g cm^{-3} was used. This is the average of the effective density values reported by Geller et al. (2006) for coastal aerosols of different sizes. They reported density values of 1.19, 1.14, 0.99, and 1.06 g cm^{-3} for particle sizes of 50, 118, 146, and 202 nm, respectively.

3.6 Photochemical aging of the plumes

Photochemical aging changes the physical and chemical properties of the ship plume particles. Sulfur content in fuel affects SO_2 emitted from marine engines, but Zetterdahl et al. (2016) measured also reduced volatile organic compound (VOC) emissions related to the smaller FSC. However, in contrast to decreased secondary-organic aerosol (SOA) precursor observations (SO_2 and VOC), Wu et al. (2019) observed an increase in SOA formation due to changing from residual fuel oils to distillate fuels. At Utö, the SO_2 rich ship emissions from the Utö harbor area have been previously linked to being a possible cause of local nucleation events (from the secondary aerosol formation) especially in the size range of 10-30 nm (Hyvärinen et al., 2008).

In this study, the effect of photochemical aging on the exhaust plumes was examined by comparing the number size distributions (NSD_{pl}) and volume size distributions (VSD_{pl}) measured in nighttime (total radiation intensity $I_{\text{tot}} < 0 \text{ W m}^{-2}$) and in the daytime when $I_{\text{tot}} > 200 \text{ W m}^{-2}$. The NSD_{pl} and VSD_{pl} were calculated as averages in 10° -sectors and the number of plumes exceeding the limit values per angle are presented in the supplemental material (Fig. S4S6). Only the sectors of 220° - 260° and 320° - 350° had enough data for both nighttime and daytime plume analysis by exceeding 50-40 plumes per 10° . With fewer plumes per sector, noise in the data was high (Fig. S5S7). Moreover, the sector of 320° - 350° was the direction of the ferry harbor of Utö and was therefore excluded from further analysis.

Night- and daytime NSD_{pl} and VSD_{pl} differed from each other during all the three FSC restriction periods (Fig. 12). A large fraction of the change between day and night might be related to the different meteorological conditions and ship traffic during day and night. Assuming that the wind speed, humidity, temperature, and other meteorological parameters, as well as the composition of the ship fleet, stayed relatively constant between the restriction periods, the remaining variable is FSC, and the differences between the restriction periods are caused by the changes in the FSC limits. These assumptions cause large uncertainty in this result and the results are therefore only considered qualitatively.

The most striking difference between the day- and nighttime NSD_{pl} was that the size distribution maximum shifted to the larger particle size in the daytime (Fig. 12). The shift was largest for the FSC limit of 0.10 m/m% and smallest for the FSC limit of 1.50 m/m%. Moreover, during the FSC limit of 0.10 m/m%, the maximum of the NSD_{pl} was clearly at the smaller particle size than the maxima at 1.00 and 1.50 m/m% at nighttime whereas in the daytime the maximum was almost independent of the FSC limit.

VSD_{pl} was bimodal during all three restriction periods. The mode at the particle size of $\sim 90 \text{ nm}$ was smaller with relation to the mode at $\sim 300 \text{ nm}$, the difference between the modes being especially noticeable during the FSC restriction of 0.10 m/m%.

Similar to the NSD_{pl} , the mode at ~90 nm shifted to the larger particle size in [the](#) daytime compared to that in the nighttime. Again, the shift was largest for the FSC of 0.10 m/m% and smallest for the FSC of 1.50 m/m%.

When the effects caused by dilution to the PNC_{pl} are excluded, the remaining factors in the atmosphere influencing PNC_{pl} are coagulation and new particle formation (Celik et al., 2019). Also, condensation and evaporation can change NSD_{pl} , and in the

case of evaporation, it can affect the PNC_{pl} by evaporating particles completely or reducing particle size below the detection limit of the measurement instrument. Celik et al. (2019) suspected that during nighttime, coagulation reduced the particle number emission factors, but they did not find any correlation between plume age and plume particle number emission factor (E_n) during the daytime. It was proposed that this could be because new particle formation during daytime overrides the effects of coagulation. Shiraiwa et al. (2013) have presented that in photochemical aging, semi-volatile organic compounds (SVOC) equilibrate with the particle population and the majority of SVOC condense on the largest particles because of their greater absorption capacity and reduced equilibrium vapor pressure on the surface of the particle (the Kelvin effect). They found that the maxima of the NSD got narrower and shifted to a larger particle size at the diameter of 200 nm in 20 h of photochemical aging and depleted basically all particles below 100 nm. This kind of behavior is an extreme case but could be another plausible explanation for the observed behavior of NSD_{pl} seen in this work.

Table 24 represents total PNC_{pl} and volume concentration of the plume particles (VC_{pl}) for the three different FSC restrictions for the day- and nighttime. Overall, the PNC_{pl} was higher during nighttime plumes than during the daytime. The difference was largest for the FSC restriction of 1.50 m/m% and the smallest for the 0.10 m/m% restriction. Each change in the FSC restrictions was observed to reduce the PNC_{pl} both in the day- and night time, however, the dependency of the PNC_{pl} on the FSC was greater in the nighttime, and therefore the night-day difference in the PNC_{pl} decreased with decreasing FSC. In contrast to PNC_{pl} , the VC_{pl} was higher during the daytime plumes but similar to the PNC_{pl} , the VC_{pl} reduced with the stricter FSC limit. For the VC_{pl} , the difference between the day- and nighttime concentrations were less dependent on the FSC limit than for the PNC_{pl} .

Higher total VC_{pl} during daytime than during nighttime indicates the condensation of gaseous matter on particles. Meanwhile, the PNC_{pl} was lower in the daytime, which might indicate ~~either coagulation or~~ evaporation of small particles. The decrease of PNC_{pl} and increase of VC_{pl} in daytime diminishes as the FSC reduces implying that the higher FSC increases ~~either the coagulation or~~ evaporation of small particles and the condensation ~~on~~ larger particles. ~~Coagulation cannot explain the~~ higher number of smaller particles observed in Fig. 12, ~~can be explained by~~ but the evaporation of liquid matter from the surface of medium-sized particles exposing the small solid core particles ~~could explain this~~. The evaporated material from these particles, and the wholly evaporated particles together with other VOCs in plume condensate over larger particles.

4 Conclusions

This study investigated the effects of the fuel sulfur content (FSC) restrictions on the number size distributions (NSD_{pl}) and particle number concentrations (PNC_{pl}) at the Utö island in the Baltic Sea during the 10-year time-period (2007-2016). The

focus was both on the changes in the ship exhaust plumes and the ambient aerosol. In general, FSC restrictions were observed to have a clear impact on particle number concentrations and size distributions in the marine environment next to the shipping lanes. Besides, the FSC limitations affected the size distributions and the growth of particles during the plume aging.

The overall change in FSC from 1.50 to 0.10 m/m% decreased the PNC_{pl} and ~~ambient~~~~total~~ particle number concentration (PNC_{tot}) by 27 and 32 %, respectively. In both cases, the largest decrease was observed after the latter FSC restriction change from 1.00 to 0.10 m/m%, likely because the fuel type was changed from residual oil to distillates.

Particle size distribution in the plumes also changed after the FSC restriction change from 1.00 to 0.10 m/m%. The FSC restriction affected especially the concentrations of the small particles by reducing the particle number concentration in particle sizes $\sim 335-1434$ nm, whereas particles with $D_p \geq 1441-55$ nm remained mostly unaffected. Due to the small size of particles in ship exhaust plume emissions (~~CDM-GDM~~ typically 30-60 nm), the contribution of ship emissions to $PM_{2.5}$ is likely to be small and the effect of ship emissions should be characterized based on PNC rather than PM.

In addition to the isolated ship plumes, there was a decrease in the ~~total~~-ambient particle number concentrations (PNC_{tot}) with the stricter FSC limits. When the wind was blowing over distant shipping lanes, the PNC_{tot} was significantly elevated because of large numbers of diluted ship plumes arriving ~~at~~~~in~~ Utö. This observation indicates that shipping is a major source of particles in the marine environment, even though the distance to the shipping lanes can be substantial. Jonson et al. (2019) proposed that the contribution of shipping related particle emissions to $PM_{2.5}$ would be negligible after the FSC restriction of 0.10 m/m%. However, based on the results of this study, this does not seem to be the case, as the reduction of PNC_{pl} and particle size were moderate and the concentrations in the largest particle sizes remained unchanged. However, the largest particle sizes can be affected by other sources than shipping in Utö, for example, long-range transported particles from other combustion sources in the mainland.

~~Count median particle size~~Geometric mean diameter (~~GMD~~~~CMD~~) was found to increase with the increasing residence time of ship exhaust plumes in the atmosphere. ~~GMD~~~~CMD~~ was larger for the particles in plumes that traveled long distances.

Furthermore, particle volume concentration was larger for the plumes measured in daytime compared to the plumes measured in nighttime likely due to the photochemical aging of particles due to UV-light. This difference between day and nighttime decreased with the reducing FSC. This indicates that lower FSC might reduce the condensation of gaseous matter on larger particles and ~~either~~ diminishes ~~coagulation or~~ evaporation of smaller particles. This result includes several uncertainties and therefore needs further investigation.

We note that during a decade long measurement period, also the ship traffic somewhat decreased in the vicinity of Utö and it is possible that the decrease in activity, in addition to FSC limits, has affected the observed ambient particle number concentrations. However, especially the change from 1.00 to 0.10 m/m% FSC was observed to cause a clear change also in ambient particle concentrations. Also, the decrease in ship traffic is assumed to be gradual and is not expected to cause this kind of step-change. Also, the amount of ship traffic does not affect the measured concentrations and number size distributions of plumes as these average values do not depend on the number of measured plumes.

As conclusions, the FSC restrictions can be expected to cause impacts on the environment and human health due to the lower particle number concentrations in the ultrafine particle size range. Concentration reduction of ultrafine particles is especially effective in reducing negative health effects as the respirable fraction of ultrafine particles is high (Heyder, 2004) and the lower fraction of the respiratory tract is more sensitive to pathogen numbers compared to [the](#) upper parts of the respiratory tract (Thomas, 2013). [The reduction of particle concentration in the size range of ~ 33-144 nm observed in this study implies significant effects on radiative properties of marine clouds. Particle diameter range from 40 to 120 nm has been identified to be crucial for cloud condensation nuclei, as larger particles are almost always activated to form cloud droplets, while particles smaller than 40 nm are not activated with realistic supersaturations \(Dusek et al., 2006\). Therefore, ships running with low-sulfur fuel produce fewer cloud condensation nuclei in this critical size range. This can lead to a substantial inverse Twomey effect, increasing the albedo of marine clouds, and thus, having an indirect warming effect on the climate \(Twomey, 1977\). Reduction of particle concentrations in this study in the size range of ~35-134 nm has significant contributions to cloud formation as 40-120 nm \(electrical mobility diameter\) has been identified as the most crucial size for cloud condensation nuclei as larger particles are almost always activated to form cloud droplets and the particles smaller than 40 nm are not activated with realistic supersaturations \(Dusek et al., 2006\). Therefore, ships produce fewer cloud condensation nuclei, and this might have a large warming effect on the climate as marine clouds usually have a cooling effect on the atmosphere.](#) Increasing average cloud droplet concentrations from 100 to 375 cm⁻³ has been presented to have more than double of a cooling effect on the climate than the warming caused by the doubling of CO₂ (Latham et al., 2008). Therefore, even a slight decrease of cloud condensation nuclei over sea areas might have a strong cooling effect on the climate.

Data availability: Data are available upon request from the corresponding author Sami Seppälä (sami.seppala@fmi.fi).

Author contributions. Sami Seppälä handled the formal analysis, software, visualization, and writing- original-draft, Joel Kuula, Antti-Pekka Hyvärinen, Sanna Saarikoski, Topi Rönkkö, Jorma Keskinen, Jukka-Pekka Jalkanen contributed to ideas, review and editing of the article, Hilikka Timonen acted as a supervisor and contributed to reviewing and editing of the article.

Competing interests. The authors declare that they have no conflict of interest

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References

- Alanen, J., Isotalo, M., Kuittinen, N., Simonen, P., Martikainen, S., Kuuluvainen, H., Honkanen, M., Lehtoranta, K., Nyyssö, S., Vesala, H., Timonen, H., Aurela, M., Keskinen, J., and Ro, T.: Physical Characteristics of Particle Emissions from a Medium Speed Ship Engine Fueled with Natural Gas and Low-Sulfur Liquid Fuels, *Environ. Sci. Tech.*, 54, 5376–5384, <https://doi.org/10.1021/acs.est.9b06460>, 2020.
- Alföldy, B., Lööv, J. B., Lagler, F., Mellqvist, J., Berg, N., Beecken, J., Weststrate, H., Duyzer, J., Bencs, L., Horemans, B., Cavalli, F., Putaud, J. P., Janssens-Maenhout, G., Csordás, A. P., Van Grieken, R., Borowiak, A., and Hjorth, J.: Measurements of air pollution emission factors for marine transportation in SECA, *Atmos. Meas. Tech.*, 6(7), 1777–1791, <https://doi.org/10.5194/amt-6-1777-2013>, 2013.
- Anderson, M., Salo, K., and Fridell, E.: Particle- and Gaseous Emissions from an LNG Powered Ship, *Environ. Sci. Tech.*, 49(20), 12568–12575, <https://doi.org/10.1021/acs.est.5b02678>, 2015a.
- Anderson, M., Salo, K., Hallquist, Å. M., and Fridell, E., Characterization of particles from a marine engine operating at low loads, *Atmos. Environ.*, 101(2015), 65–71. <https://doi.org/10.1016/j.atmosenv.2014.11.009>, 2015b.
- Arnold, F., Pirjola, L., Aufmhoff, H., Schuck, T., Lähde, T., and Hämeri, K.: First gaseous sulfuric acid measurements in automobile exhaust: Implications for volatile nanoparticle formation, *Atmos. Environ.*, 40(37), 7097–7105, <https://doi.org/10.1016/j.atmosenv.2006.06.038>, 2006.
- Arnold, F., Pirjola, L., Rönkkö, T., Reichl, U., Schlager, H., Lähde, T., Heikkilä, J., and Keskinen, J.: First online measurements of sulfuric acid gas in modern heavy-duty diesel engine exhaust: Implications for nanoparticle formation, *Environ. Sci. Tech.*, 46(20), 11227–11234, <https://doi.org/10.1021/es302432s>, 2012.
- Ausmeel, S., Eriksson, A., Ahlberg, E., and Kristensson, A.: Methods for identifying aged ship plumes and estimating contribution to aerosol exposure downwind of shipping lanes, *Atmos. Meas. Tech.*, 12(8), 4479–4493, <https://doi.org/10.5194/amt-12-4479-2019>, 2019.
- Barregard, L., Molnr, P., Jonson, J. E., and Stockfelt, L.: Impact on population health of baltic shipping emissions, *Int. J. Environ. Res.*, 16(11), <https://doi.org/10.3390/ijerph16111954>, 2019.
- Celik, S., Drewnick, F., Fachinger, F., Brooks, J., Darbyshire, E., Coe, H., Paris, J.-D., Eger, P., Schuladen, J., Tadic, I., Friedrich, N., Dienhart, D., Hottmann, B., Fischer, H., Crowley, J., Harder, H., and Borrmann, S.: Influence of vessel characteristics and atmospheric processes on the gas and particle phase of ship emission plumes: In-situ measurements in the Mediterranean Sea and around the Arabian Peninsula, *Atmos. Chem. Phys. Discuss.*, 1–36, <https://doi.org/10.5194/acp-2019-859>, 2019.
- Chen, C., Saikawa, E., Comer, B., Mao, X., and Rutherford, D.: Ship Emission Impacts on Air Quality and Human Health in the Pearl River Delta (PRD) Region, China, in 2015, With Projections to 2030, *GeoHealth*, 3(9), 284–306, <https://doi.org/10.1029/2019GH000183>, 2019.

Chen, G., Huey, L. G., Trainer, M., Nicks, D., Corbett, J., Ryerson, T., Parrish, D., Neuman, J. A., Nowak, J., Tanner, D.,
520 Holloway, J., Brock, C., Crawford, J., Olson, J. R., Sullivan, A., Weber, R., Schauffler, S., Donnelly, S., Atlas, E., and
Fehsenfeld, F.: An investigation of the chemistry of ship emission plumes during ITCT 2002, *J. Geophys. Res. Atmos.*,
110(10), 1–15, <https://doi.org/10.1029/2004JD005236>, 2005.

Corbin, J. C., Peng, W., Yang, J., Sommer, D. E., Trivanovic, U., Kirchen, P., Miller, J. W., Rogak, S., Cocker, D. R.,
Smallwood, G. J., Lobo, P., and Gagné, S.: Characterization of particulate matter emitted by a marine engine operated with
525 liquefied natural gas and diesel fuels, *Atmos. Environ.*, 220(October 2019), <https://doi.org/10.1016/j.atmosenv.2019.117030>,
2020.

Cordtz, R., Schramm, J., Andreasen, A., Eskildsen, S. S., and Mayer, S.: Modeling the distribution of sulfur compounds in a
large two stroke diesel engine, *Energy & Fuels*, 27(3), 1652–1660. <https://doi.org/10.1021/ef301793a>, 2013.

Dal Maso, M., Kulmala, M., Riipinen, I., Wagner, R., Hussein, T., Aalto, P. P., and Lehtinen, K. E. J.: Formation and growth
530 of fresh atmospheric aerosols: Eight years of aerosol size distribution data from SMEAR II, Hyytiälä, Finland, *Boreal Environ.*
Res., 10(5), 323–336, 2005.

Diesch, J. M., Drewnick, F., Klimach, T., and Borrmann, S.: Investigation of gaseous and particulate emissions from various
marine vessel types measured on the banks of the Elbe in Northern Germany, *Atmos. Chem. Phys.*, 13(7), 3603–3618,
<https://doi.org/10.5194/acp-13-3603-2013>, 2013.

535 Jonson, J. E., Gauss, M., Jalkanen, J. P., and Johansson, L.: Effects of strengthening the Baltic Sea ECA regulations, *Atmos.*
Chem. Phys., 19(21), 13469–13487, <https://doi.org/10.5194/acp-19-13469-2019>, 2019.

Dusek, U., Frank, G. P., Hildebrandt, L., Curtius, J., Schneider, J., Walter, S., Chand, D., Drewnick, F., Hings, S., Jung, D.,
Borrmann, S., and Andreae, M. O.: Size matters more than chemistry for cloud-nucleating ability of aerosol particles, *Science*,
312(5778), 1375–1378, <https://doi.org/10.1126/science.1125261>, 2006.

540 The European Parliament and the Council of the European Union: Directive 2005/33/EC of the European Parliament and the
Council of 6 July 2005 amending Directive 1999/32/EC, available: <https://eur-lex.europa.eu/LexUriServ/LexUriServ.do?uri=OJ:L:2005:191:0059:0069:EN:PDF>, (last access: 28 August 2020), 2005.

Engler, C., Lihavainen, H., Komppula, M., Kerminen, V. M., Kulmala, M., and Viisanen, Y.: Continuous measurements of
545 aerosol properties at the Baltic Sea, *Tellus B Chem. Phys. Meteorol.*, 59(4), 728–741, <https://doi.org/10.1111/j.1600-0889.2007.00285.x>, 2007.

[Geller, M., Biswas, S., and Sioutas, C.: Determination of Particle Effective Density in Urban Environments with a Differential
Mobility Analyzer and Aerosol Particle Mass Analyzer, *Aerosol Sci. Technol.*, 40:9, 709-723,
<https://doi.org/10.1080/02786820600803925>, 2006.](https://doi.org/10.1080/02786820600803925)

550 Gobbi, G. P., Di Liberto, L., and Barnaba, F.: Impact of port emissions on EU-regulated and non-regulated air quality
indicators: The case of Civitavecchia (Italy), *Sci. Tot. Environ.*, 719, 134984, <https://doi.org/10.1016/j.scitotenv.2019.134984>,
2020.

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Hennig, F., Quass, U., Hellack, B., Küpper, M., Kuhlbusch, T. A. J., Stafoggia, M., and Hoffmann, B.: Ultrafine and fine particle number and surface area concentrations and daily cause-specific mortality in the Ruhr area, Germany, 2009–2014, *Environ. Health Perspectives*, 126(2), 2009–2014, <https://doi.org/10.1289/EHP2054>, 2018.

Heyder, J.: Deposition of inhaled particles in the human respiratory tract and consequences for regional targeting in respiratory drug delivery. *Proceed. Am. Thorac. Soc.*, 1(4), 315–320, <https://doi.org/10.1513/pats.200409-046TA>, 2004.

Hoang, A. T., and Pham, V. V.: A review on fuels used for marine diesel engines, *J. Mech. Eng. Res. Dev.*, 41(4), 22–32, <https://doi.org/10.26480/jmerd.04.2018.22.32>, 2018.

Hoppel, W.A.: Determination of aerosol size distribution from the mobility distribution of the charged fraction of aerosols, *J. Aerosol. Sci.*, 9, 41–54, [https://doi.org/10.1016/0021-8502\(78\)90062-9](https://doi.org/10.1016/0021-8502(78)90062-9), 1978.

Hyvärinen, A. P., Komppula, M., Engler, C., Kivekäs, N., Kerminen, V. M., Dal Maso, M., Viisanen, Y., and Lihavainen, H.: Atmospheric new particle formation at Utö, Baltic Sea 2003–2005, *Tellus B Chem. Phys. Meteorol.*, 60 B(3), 345–352, <https://doi.org/10.1111/j.1600-0889.2008.00343.x>, 2008.

IMO: AIS transponders, IMO (International Maritime Organization), Available: <http://www.imo.org/en/OurWork/Safety/Navigation/Pages/AIS.aspx> (last access: 28 August 2020)

IMO: Amendments to the Annex of the Protocol of 1997 to amend the International Convention for Prevention of Pollution from Ships 1973, as modified by the Protocol of 1978, relating thereto, Revised MARPOL Annex VI, IMO (International Maritime Organization), available at: [http://www.imo.org/en/KnowledgeCentre/IndexofIMOResolutions/Marine-Environment-Protection-Committee-\(MEPC\)/Documents/MEPC.176\(58\).pdf](http://www.imo.org/en/KnowledgeCentre/IndexofIMOResolutions/Marine-Environment-Protection-Committee-(MEPC)/Documents/MEPC.176(58).pdf) (last access: 28 August 2020), 2008.

IMO: MARPOL 73/78 Annex VI - Regulations for the prevention of Air Pollution from ships – Technical and Operational implications, Det Norske Veritas, Available: http://hulpinnood.nl/wp-content/uploads/2015/03/BIJLAGE3_Marpol-annex-VI.pdf, (last access: 28 August 2020).

Jonsson, Å. M., Westerlund, J., and Hallquist, M.: Size-resolved particle emission factors for individual ships, *Geophys. Res. Lett.*, 38(13), 1–5, <https://doi.org/10.1029/2011GL047672>, 2011.

Karl, M., Pirjola, L., Karppinen, A., Jalkanen, J. P., Ramacher, M. O. P., and Kukkonen, J.: Modeling of the concentrations of ultrafine particles in the plumes of ships in the vicinity of major harbors, *Int. J. Environ. Res.*, 17(3), 1–24, <https://doi.org/10.3390/ijerph17030777>, 2020.

Kattner, L., Mathieu-Öffing, B., Burrows, J. P., Richter, A., Schmolke, S., Seyler, A., and Wittrock, F.: Monitoring compliance with sulfur content regulations of shipping fuel by in situ measurements of ship emissions, *Atmos. Chem. Phys.*, 15(17), 10087–10092, <https://doi.org/10.5194/acp-15-10087-2015>, 2015.

Kecorius, S., Kivekäs, N., Kristensson, A., Tuch, T., Covert, D. S., Birmili, W., Lihavainen, H., Hyvärinen, A. P., Martinsson, J., Sporre, M. K., Swietlicki, E., Wiedensohler, A., and Ulevicius, V.: Significant increase of aerosol number concentrations in air masses crossing a densely trafficked sea area, *Oceanography*, 58(1), 1–12., <https://doi.org/10.1016/j.oceano.2015.08.001>, 2016.

Kivekäs, N., Massling, A., Grythe, H., Lange, R., Rusnak, V., Carreno, S., Skov, H., Swietlicki, E., Nguyen, Q. T., Glasius, M., and Kristensson, A.: Contribution of ship traffic to aerosol particle concentrations downwind of a major shipping lane, *Atmos. Chem. Phys.*, 14(16), 8255–8267, <https://doi.org/10.5194/acp-14-8255-2014>, 2014.

Kozak, M. and Merksiz, J.: the Mechanics of Fuel Sulphur Influence on Exhaust Emissions From Diesel Engines, TEKA Kom.

590 Mot. Energ. Roln, 5, 96–106, <http://www.pan-ol.lublin.pl/wydawnictwa/TMot5/Kozak.pdf>, 2005.

Kukkonen, J., Karl, M., Keuken, M. P., Denier Van Der Gon, H. A. C., Denby, B. R., Singh, V., Douros, J., Manders, A., Samaras, Z., Moussiopoulos, N., Jonkers, S., Aarnio, M., Karppinen, A., Kangas, L., Lützenkirchen, S., Petäjä, T., Vouitsis, I., and Sokhi, R. S.: Modelling the dispersion of particle numbers in five European cities, *Geosci. Model Dev.*, 9,9(2), 451–478, <https://doi.org/10.5194/gmd-9-451-2016>, 2016.

595 Kumar, P., Pirjola, L., Ketzel, M., and Harrison, R. M.: Nanoparticle emissions from 11 non-vehicle exhaust sources - A review, *Atmos. Environ.*, 67, 252–277, <https://doi.org/10.1016/j.atmosenv.2012.11.011>, 2013.

Laakso, L., Mikkonen, S., Drebs, A., Karjalainen, A., Pirinen, P., and Alenius, P.: 100 Years of atmospheric and marine observations at the Finnish Utö Island in the Baltic Sea, *Ocean Sci.*, 14(4), 617–632, <https://doi.org/10.5194/os-14-617-2018>, 2018.

600 Lack, D. A., Corbett, J. J., Onasch, T., Lerner, B., Massoli, P., Quinn, P. K., Bates, T. S., Covert, D. S., Coffman, D., Sierau, B., Herndon, S., Allan, J., Baynard, T., Lovejoy, E., Ravishankara, A. R., and Williams, E.: Particulate emissions from commercial shipping: Chemical, physical, and optical properties, *J. Geophys. Res. Atmos.*, 114(4), 1–16, <https://doi.org/10.1029/2008JD011300>, 2009.

Latham, J., Rasch, P., Chen, C. C., Kettles, L., Gadian, A., Gettelman, A., Morrison, H., Bower, K., and Choulaton, T.: Global temperature stabilization via controlled albedo enhancement of low-level maritime clouds, *Philos. Trans. R. Soc. A*, 366(1882), 3969–3987, <https://doi.org/10.1098/rsta.2008.0137>, 2008.

Lehtoranta, K., Aakko-Saksa, P., Murtonen, T., Vesala, H., Ntziachristos, L., Rönkkö, T., Karjalainen, P., Kuittinen, N., and Timonen, H.: Particulate Mass and Nonvolatile Particle Number Emissions from Marine Engines Using Low-Sulfur Fuels, Natural Gas, or Scrubbers, *Environ. Sci. Tech.*, 53(6), 3315–3322, <https://doi.org/10.1021/acs.est.8b05555>, 2019.

610 Lin, H., Tao, J., Qian, Z. (Min), Ruan, Z., Xu, Y., Hang, J., Xu, X., Liu, T., Guo, Y., Zeng, W., Xiao, J., Guo, L., Li, X., and Ma, W.: Shipping pollution emission associated with increased cardiovascular mortality: A time series study in Guangzhou, China, *Environ. Pollut.*, 241(June), 862–868, <https://doi.org/10.1016/j.envpol.2018.06.027>, 2018.

Lv, Z., Liu, H., Ying, Q., Fu, M., Meng, Z., Wang, Y., Wei, W., Gong, H., and He, K.: Impacts of shipping emissions on PM_{2.5} pollution in China, *Atmos. Chem. Phys.*, 18(21), 15811–15824, <https://doi.org/10.5194/acp-18-15811-2018>, 2018.

615 Mylläri, F., Asmi, E., Anttila, T., Saukko, E., Vakkari, V., Pirjola, L., Hillamo, R., Laurila, T., Häyrynen, A., Rautiainen, J., Lihavainen, H., O'Connor, E., Niemelä, V., Keskinen, J., Dal Maso, M., and Rönkkö, T.: New particle formation in the fresh flue-gas plume from a coal-fired power plant: Effect of flue-gas cleaning, *Atmos. Chem. Phys.*, 16(11), 7485–7496, <https://doi.org/10.5194/acp-16-7485-2016>, 2016.

Field Code Changed

Ntziachristos, L., Saukko, E., Lehtoranta, K., Rönkkö, T., Timonen, H., Simonen, P., Karjalainen, P., and Keskinen, J.: Particle emissions characterization from a medium-speed marine diesel engine with two fuels at different sampling conditions, *Fuel*, 186, 456–465, <https://doi.org/10.1016/j.fuel.2016.08.091>, 2016.

Partanen, A. I., Laakso, A., Schmidt, A., Kokkola, H., Kuokkanen, T., Pietikäinen, J. P., Kerminen, V. M., Lehtinen, K. E. J., Laakso, L., and Korhonen, H.: Climate and air quality trade-offs in altering ship fuel sulfur content, *Atmos. Chem. Phys.*, 13(23), 12059–12071, <https://doi.org/10.5194/acp-13-12059-2013>, 2013.

Petzold, A., Hasselbach, J., Lauer, P., Baumann, R., Franke, K., Gurk, C., Schlager, H., and Weingartner, E.: Experimental studies on particle emissions from cruising ship, their characteristic properties, transformation and atmospheric lifetime in the marine boundary layer, *Atmos. Chem. Phys.*, 8(9), 2387–2403, <https://doi.org/10.5194/acp-8-2387-2008>, 2008.

Pirjola, L., Karl, M., Rönkkö, T., and Arnold, F.: Model studies of volatile diesel exhaust particle formation: Are organic vapours involved in nucleation and growth?, *Atmos. Chem. Phys.*, 15(18), 10435–10452, <https://doi.org/10.5194/acp-15-10435-2015>, 2015.

Pirjola, L., Pajunaja, A., Walden, J., Jalkanen, J. P., Rönkkö, T., Kousa, A., and Koskentalo, T.: Mobile measurements of ship emissions in two harbour areas in Finland, *Atmos. Meas. Tech.*, 7(1), 149–161, <https://doi.org/10.5194/amt-7-149-2014>, 2014.

Riipinen, I., Yli-Juuti, T., Pierce, J. R., Petäjä, T., Worsnop, D. R., Kulmala, M., and Donahue, N. M.: The contribution of organics to atmospheric nanoparticle growth, *Nat. Geosci.*, 5(7), 453–458. <https://doi.org/10.1038/ngeo1499>, 2012.

Rönkkö, T., Lähde, T., Heikkilä, J., Pirjola, L., Bauschke, U., Arnold, F., Schlager, H., Rothe, D., Yli-Ojanperä, J., and Keskinen, J.: Effects of gaseous sulphuric acid on diesel exhaust nanoparticle formation and characteristics, *Environ. Sci. Tech.*, 47(20), 11882–11889, <https://doi.org/10.1021/es402354y>, 2013.

Shiraiwa, M., Yee, L. D., Schilling, K. A., Loza, C. L., Craven, J. S., Zuend, A., Ziemann, P. J., and Seinfeld, J. H.: Size distribution dynamics reveal particle-phase chemistry in organic aerosol formation, *Proc. Natl. Acad. Sci. U.S.A.*, 110(29), 11746–11750, <https://doi.org/10.1073/pnas.1307501110>, 2013.

Sofiev, M., Winebrake, J. J., Johansson, L., Carr, E. W., Prank, M., Soares, J., Vira, J., Kouznetsov, R., Jalkanen, J. P., and Corbett, J. J.: Cleaner fuels for ships provide public health benefits with climate tradeoffs., *Nature Communications*, 9(1), 1–12. <https://doi.org/10.1038/s41467-017-02774-9>, 2018.

Strak, M., Janssen, N. A. H., Godri, K. J., Gosens, I., Mudway, I. S., Cassee, F. R., Lebre, E., Kelly, F. J., Harrison, R. M., Brunekreef, B., Steenhof, M., and Hoek, G.: Respiratory health effects of airborne particulate matter: The role of particle size, composition, and oxidative potential-the RAPTES project, *Environ. Health Perspectives*, 120(8), 1183–1189, <https://doi.org/10.1289/ehp.1104389>, 2012.

Streibel, T., Schnelle-Kreis, J., Czech, H., Harndorf, H., Jakobi, G., Jokiniemi, J., Karg, E., Lintelmann, J., Matuschek, G., Michalke, B., Müller, L., Orasche, J., Passig, J., Radischat, C., Rabe, R., Reda, A., Rüger, C., Schwemer, T., Sippula, O., and Zimmermann, R.: Aerosol emissions of a ship diesel engine operated with diesel fuel or heavy fuel oil, *Environ. Sci. Pollut. Res.*, 24(12), 10976–10991, <https://doi.org/10.1007/s11356-016-6724-z>, 2017.

Thomas, R. J.: Particle size and pathogenicity in the respiratory tract, *Virulence*, 4(8), 847–858, <https://doi.org/10.4161/viru.27172>, 2013.

Tian, J., Riemer, N., West, M., Pfaffenberger, L., Schlager, H., and Petzold, A.: Modeling the evolution of aerosol particles in a ship plume using PartMC-MOSAIC, *Atmos. Chem. Phys.*, 14(11), 5327–5347, <https://doi.org/10.5194/acp-14-5327-2014>, 2014.

Twomey, S.: The influence of Pollution on the Shortwave Albedo of Clouds, *J. Atmos. Sci.*, 34, 1149-1152, [https://doi.org/10.1175/1520-0469\(1977\)034<1149:TIOPOT>2.0.CO;2](https://doi.org/10.1175/1520-0469(1977)034<1149:TIOPOT>2.0.CO;2), 1977.

Ushakov, S., Valland, H., Nielsen, J. B., and Hennie, E.: Effects of high sulphur content in marine fuels on particulate matter emission characteristics, *J. Mar. Eng. Technol.*, 12(3), 30–39, <https://doi.org/10.1080/20464177.2013.11020283>, 2013.

Viana, M., Hammingh, P., Colette, A., Querol, X., Degraeuwe, B., Vlioger, I. de, and van Aardenne, J.: Impact of maritime transport emissions on coastal air quality in Europe, *Atmos. Environ.*, 90, 96–105, <https://doi.org/10.1016/j.atmosenv.2014.03.046>, 2014.

Walker, A. P.: Controlling particulate emissions from diesel vehicles. *Top. Catal.*, 28(1–4), 165–170, <https://doi.org/10.1023/B:TOCA.0000024346.29600.0e>, 2004.

Westerlund, J., Hallquist, M., and Hallquist, Å. M.: Characterization of fleet emissions from ships through multi-individual determination of size-resolved particle emissions in a coastal area, *Atmos. Environ.*, 112(X), 159–166, <https://doi.org/10.1016/j.atmosenv.2015.04.018>, 2015.

Wu, Z., Zhang, Y., He, J., Chen, H., Huang, X., Wang, Y., Yu, X., Yang, W., Zhang, R., Zhu, M., Li, S., Fang, H., Zhang, Z., and Wang, X.: Dramatic increase of reactive VOC emission from ships at berth after implementing the fuel switch policy in the Pearl River Delta Emission Control Area, *Atmos. Chem. Phys. Discuss.*, 1–25. <https://doi.org/10.5194/acp-2019-897>, 2019.

Zanatta, M., Bozem, H., Köllner, F., Schneider, J., Kunkel, D., Hoor, P., de Faria, J., Petzold, A., Bundke, U., Hayden, K., Staebler, R. M., Schulz, H., and Herber, A. B.: Airborne survey of trace gases and aerosols over the Southern Baltic Sea: from clean marine boundary layer to shipping corridor effect, *Tellus B. Chem. Phys. Meteorol.*, 72(1), 1–24, <https://doi.org/10.1080/16000889.2019.1695349>, 2020.

Zetterdahl, M., Moldanová, J., Pei, X., Pathak, R. K., and Demirdjian, B.: Impact of the 0.1% fuel sulfur content limit in SECA on particle and gaseous emissions from marine vessels. *Atmos. Environ.*, 145, 338–345, <https://doi.org/10.1016/j.atmosenv.2016.09.022>, 2016.

Zhou, S., Zhou, J., and Zhu, Y.: Chemical composition and size distribution of particulate matters from marine diesel engines with different fuel oils, *Fuel*, 235(August 2018), 972–983, <https://doi.org/10.1016/j.fuel.2018.08.080>, 2019.

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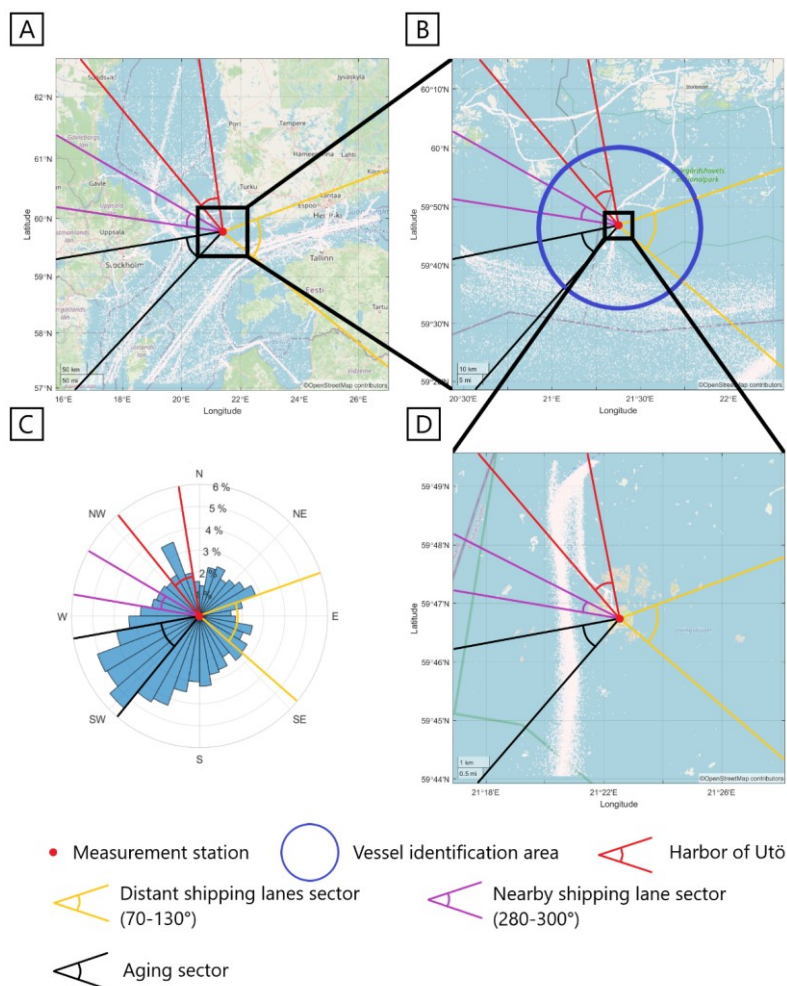


Figure 1: A) The location of the Utö island and location of the vessels (every 10 000th shown) with IMO numbers in 360 000 km² area around Utö. B) Location of the vessels (every 500th shown) with IMO numbers in 10 000 km² area around Utö. C) Percentages of the time that wind is blowing from different directions during valid measurement periods divided into 10°-sectors. D) Location of the vessels (every 10th shown) with IMO numbers in 100 km² area around Utö. White dots in pictures A, B, and D, mark the locations of vessels (0.01, 0.2, and 10 % of the locations plotted for figures A, B, and C respectively). Distant shipping lanes sector, nearby shipping lane sector, and harbor of Utö are marked with colored lines. Base map and data from OpenStreetMap and OpenStreetMap Foundation (OSMF). Data from © OpenStreetMap contributors is licensed under the Open Data Commons Open Database License (ODbL) by the OSMF.

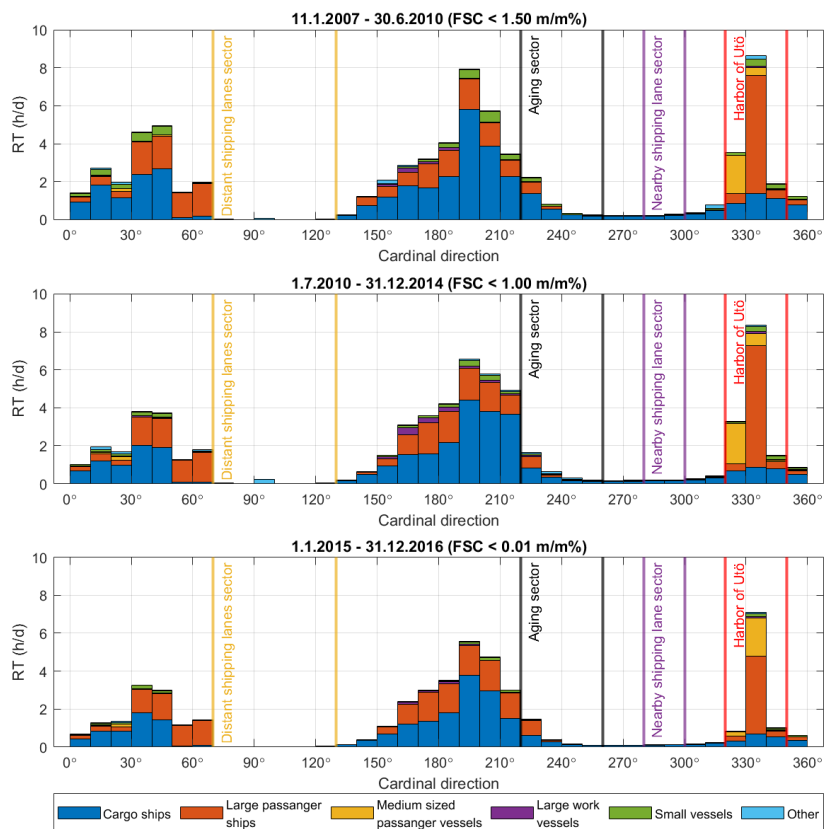


Figure 2: Ship types around Utö during the different FSC restriction periods A) FSC < 1.50 m/m%, (11.1.2007-30.6.2010), B) FSC < 1.00 m/m%, (1.7.2010-31.12.2014), C) FSC < 0.10 m/m%, (1.1.2015-31.12.2016). Ship types were classified as cargo ships, large passenger ships, medium-sized passenger vessels, large work vessels, small vessels, and others. On the y-axis, RT is the average residence time in hours per day of ships in the category on each of the 10°-sectors given on the x-axis. Distant shipping lanes sector, aging sector, nearby shipping lane sector, and harbor of Utö are marked with vertical line pairs.

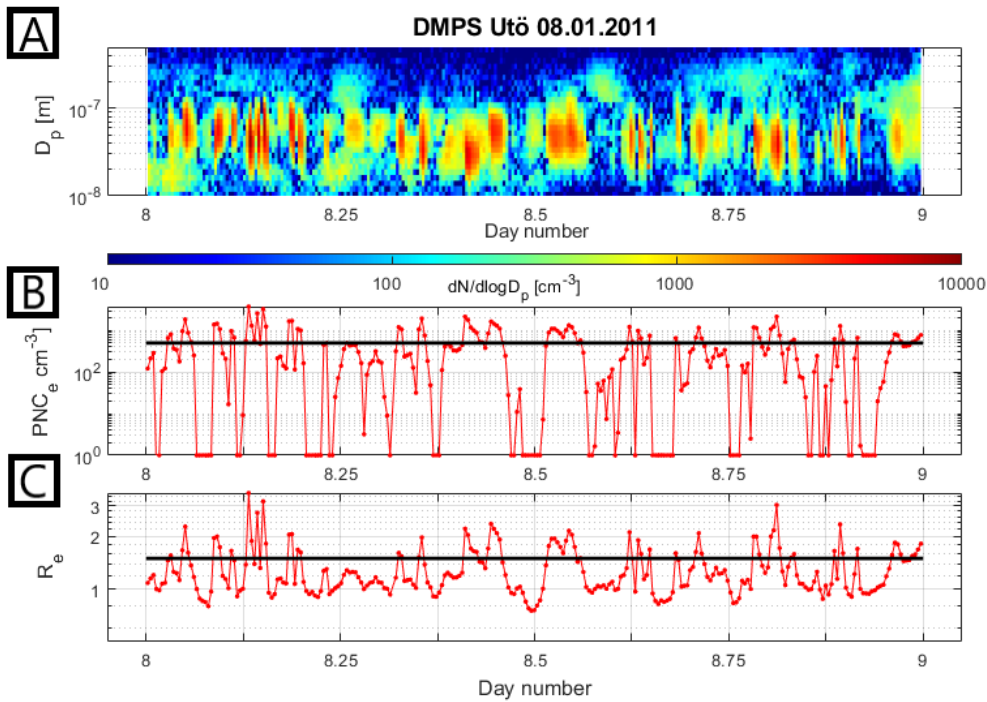


Figure 3: An example of the time series of A) number size distribution of excess particles (NSD_e), B) particle number concentration of excess particles (PNC_e) and C) the ratio of [ambient](#) to total particle number concentration to the background particle number concentration (R_e) for a single day. In B) – C) the plume detection limits have been marked with black lines.

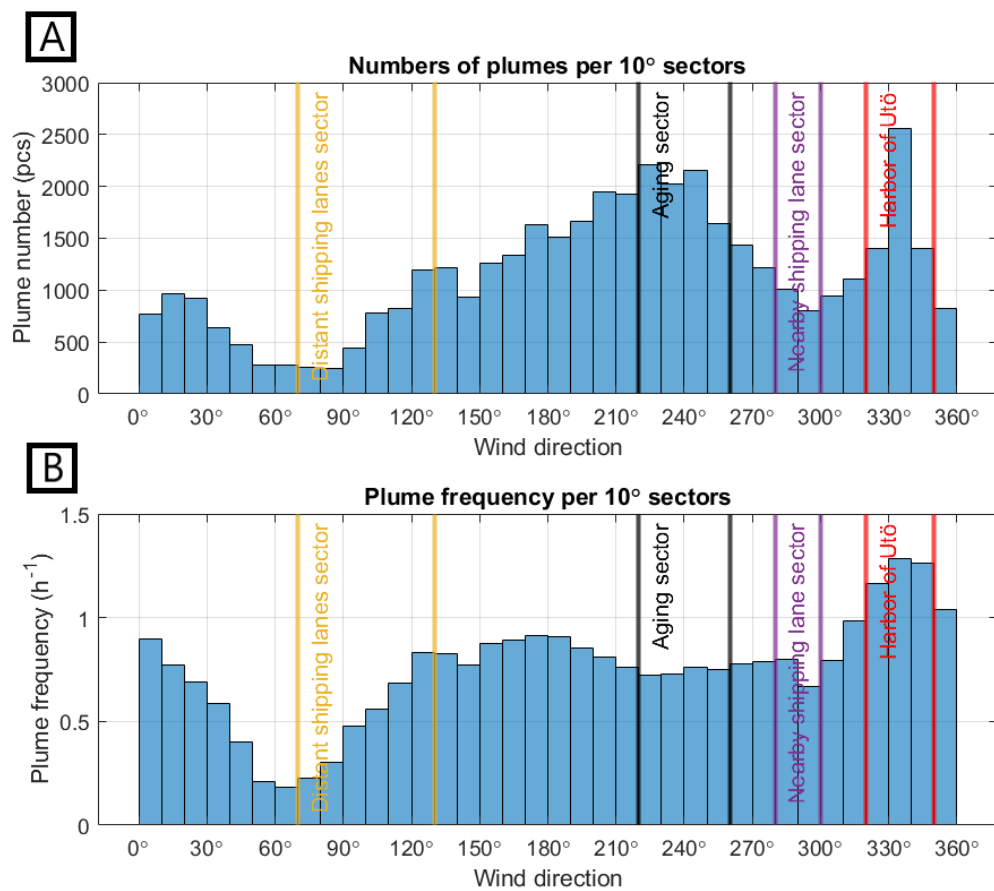


Figure 4: A) Number of valid plumes and B) the frequencies of the plumes per hour from different wind directions during valid time periods. Distant shipping lanes sector, aging sector, nearby shipping lane sector, and harbor of Utö are marked with vertical line pairs.

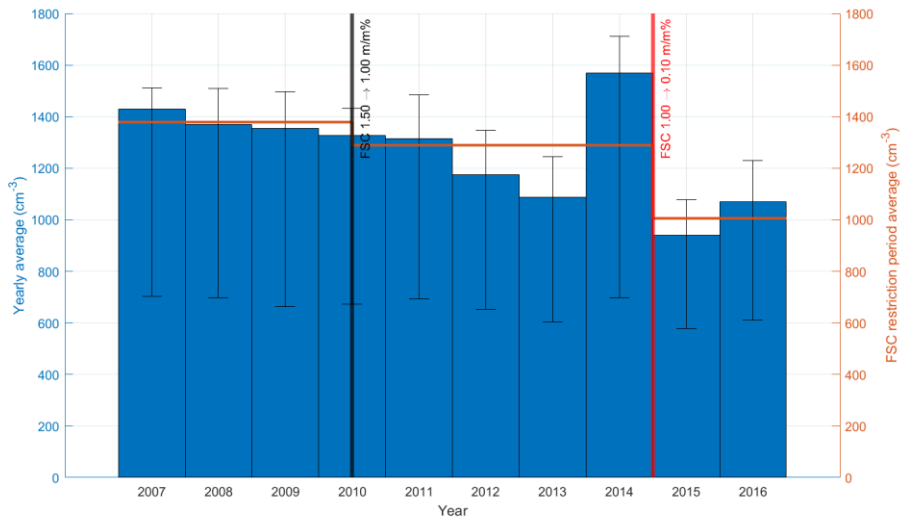
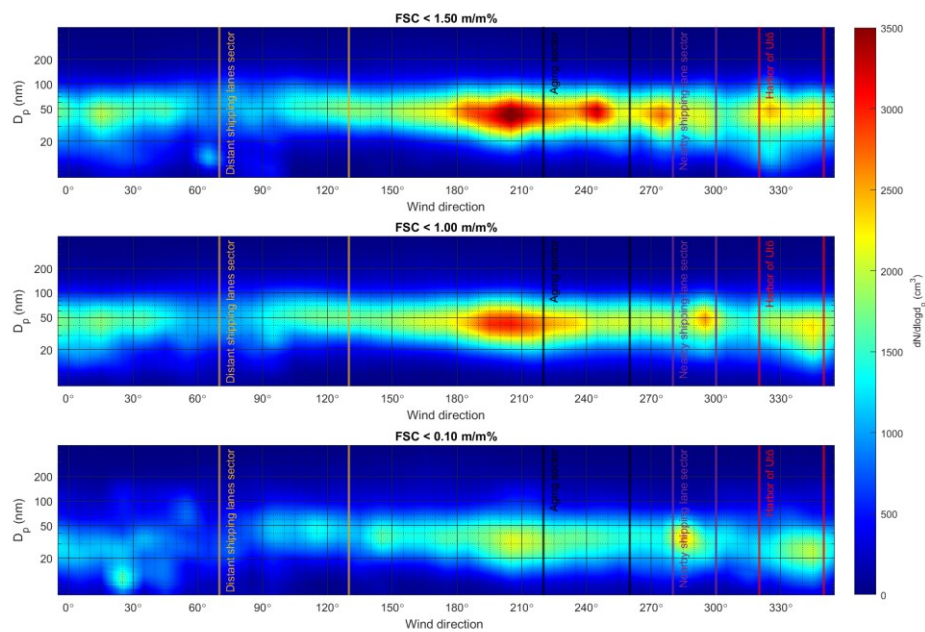


Figure 5: Annual average particle number concentrations of plumes (PNC_{pl}, blue bars) and the average PNC_{pl} during different FSC restriction periods (orange lines). Changes in the FSC restrictions are marked with the black and red vertical lines. The 25th and 75th percentiles are marked with the black error bars.



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Figure 6: sectors) for different FSC restrictions. Wind directions are shown on the x-axis, particle mobility diameters (D_p) on the y-axis, and normalized particle number concentration ($dN/d\log D_p$) with colors. Distant shipping lanes sector, aging sector, nearby shipping lane sector, and harbor of Utö are marked with vertical line pairs.

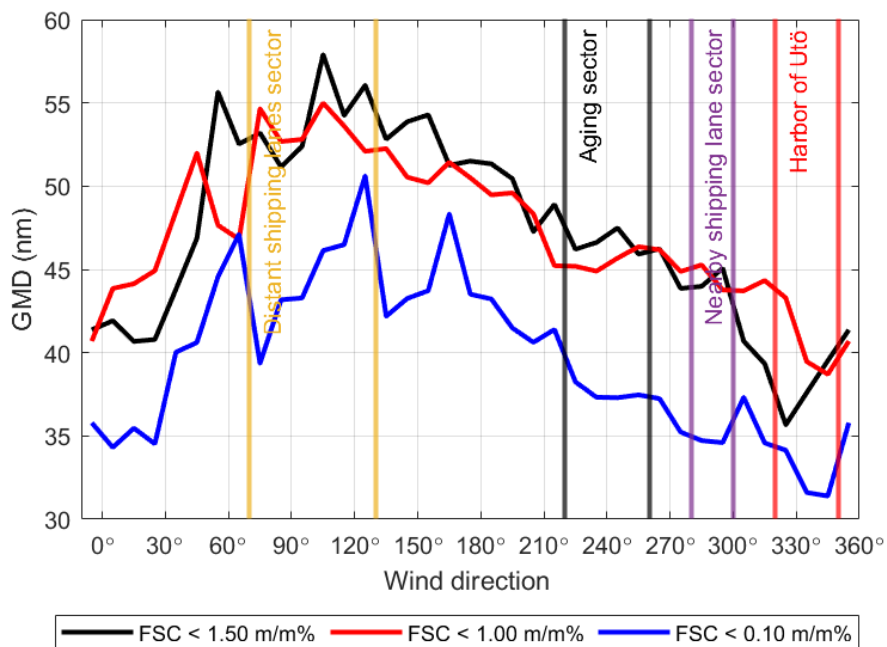


Figure 7: Average count-median-geometric mean diameters (GMD) of particles in plumes originating from different directions during the different fuel sulfur content (FSC) restrictions. Distant shipping lanes sector, aging sector, nearby shipping lane sector, and harbor of Utö are marked with vertical line pairs.

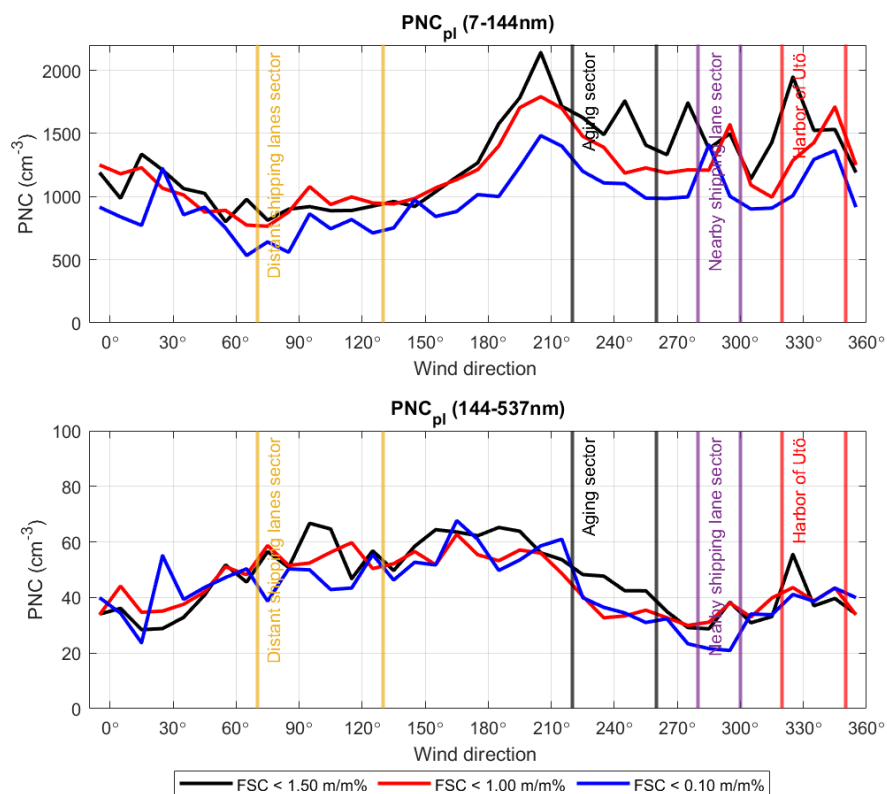


Figure 8: Particle number concentrations of the plume particles (PNC_{pl}) in size range 7-1434 nm and 144-499-537 nm averaged for 10°-sectors. Distant shipping lanes sector, aging sector, nearby shipping lane sector, and harbor of Utö are marked with vertical line pairs.

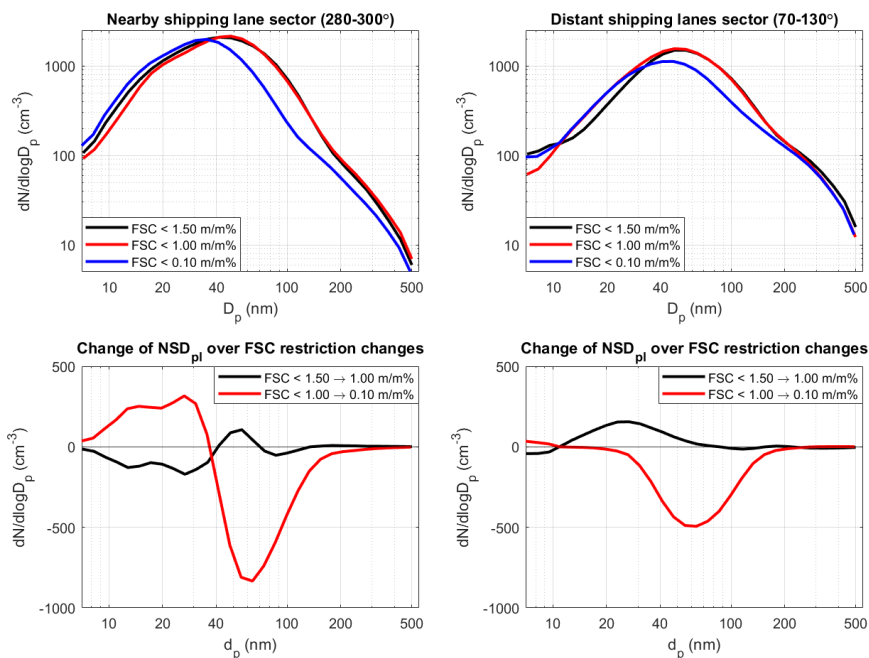


Figure 9: A) Average number size distribution of the plumes (NSD_{pl}) from the nearby shipping lanes sector (280-300°) and B) from the distant shipping lanes sector (70-130°) during the three FSC restriction periods. C) Change of NSD_{pl} in nearby shipping lanes sector (280-300°) and D) in the distant shipping lanes sector (70-130°) caused by the changes of the FSC restriction.

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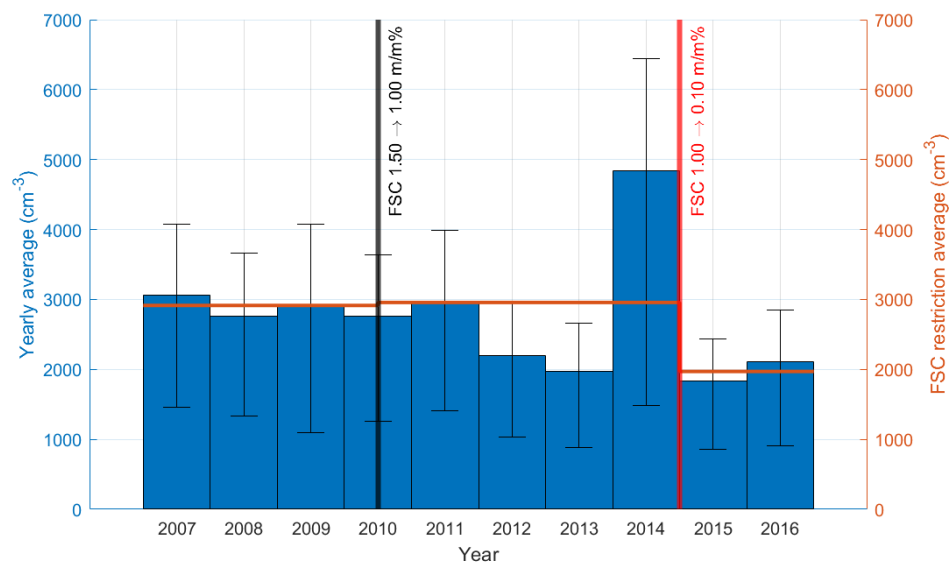


Figure 10: Annual average [ambient total](#) particle number concentrations (PNC_{tot}, blue bars) and the average PNC_{tot} (orange lines) during the different FSC restrictions. Changes of the FSC restrictions are marked with the black and red vertical lines. 25th and 75th percentiles are marked with the black error bars.

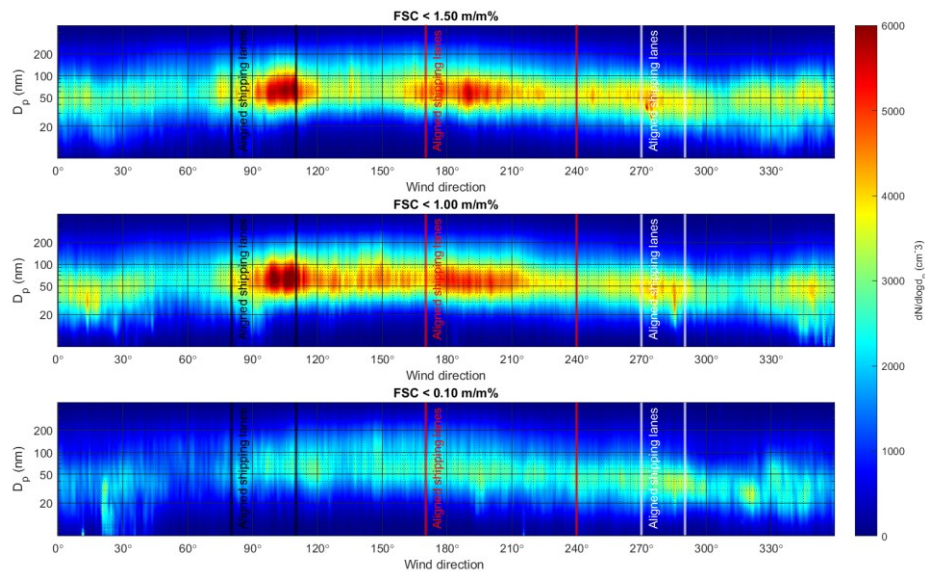


Figure 11: **AmbientTotal** number size distributions (NSD_{tot}) averaged for wind directions. Wind directions are shown on the x-axis, particle mobility diameters on the y-axis, and normalized particle number concentrations ($dN/d\log d_p$) with the colors. Directions, where there are shipping lanes at least partially aligned with the wind directions, are marked with vertical line pairs (black, red, and white). These shipping lanes have been estimated from Figure 1A.

Table 1: Average mass concentrations (MC) of ambient aerosol, valid plumes, and all plumes during the different fuel sulfur content (FSC) restrictions and the contributions of valid and all the plumes to the ambient MC.

Variable	PM _{0.144}			PM _{0.537}		
FSC limit (m/m%)	1.50	1.00	0.10	1.50	1.00	0.10
Average MC of valid plumes (µg cm ⁻³)	0.20	0.18	0.11	0.68	0.64	0.51
Contribution of valid plumes to ambient MC (%)	5.5	4.9	3.9	2.8	2.6	2.4
Average MC of all plumes (µg cm ⁻³)	0.22	0.21	0.12	0.78	0.76	0.55
Contribution of all plumes to ambient MC (%)	14.0	12.1	8.9	7.4	6.7	5.5
Average ambient MC (µg cm ⁻³)	0.58	0.58	0.33	3.81	3.68	2.47

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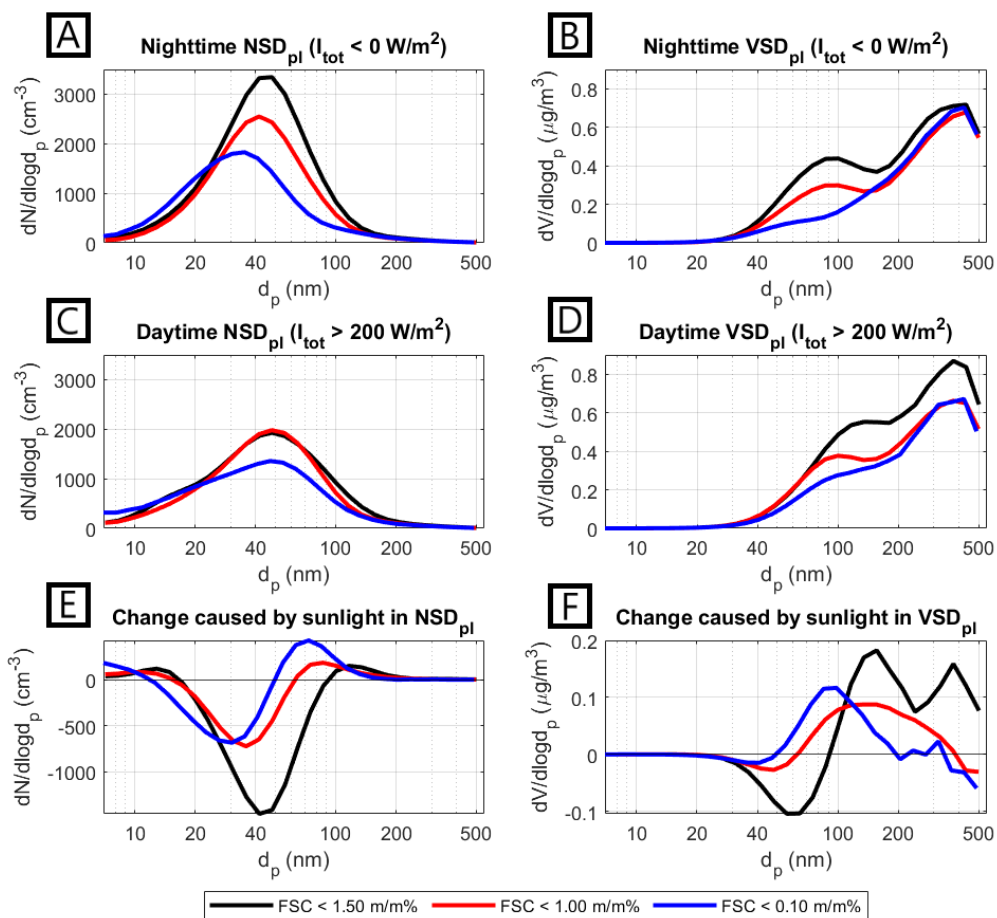


Figure 12: Average number size distributions (NSD_{pl}) and volume size distributions (VSD_{pl}) of the plumes in the sector of 220-260° for three FSC restriction periods. A) nighttime NSD_{pl}, B) nighttime VSD_{pl}, C) daytime NSD_{pl}, D) daytime VSD_{pl}, E) change in NSD_{pl} caused by exposure to sunlight, F) change in VSD_{pl} caused by exposure to sunlight.

Table 24: Particle number (PNC_{pl}) and volume concentrations (VC_{pl}) of plumes during the three different sulfur restriction periods and the change in concentrations between day- and nighttimes.

Variable	PNC _{pl} (cm ⁻³)			VC _{pl} (μm ³ m ⁻³)		
FSC limit (m/m%)	1.50	1.00	0.10	1.50	1.00	0.10
Nighttime	1910	1470	1200	0.527	0.422	0.373
Daytime	1400	1300	1090	0.585	0.458	0.402
Change	-510	-170	-110	0.058	0.036	0.029
Change (%)	-27	-11	-9	+11	+9	+8