Effects of *a priori* profile shape assumptions on comparisons between satellite NO₂ columns and model simulations

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13 Abstract

A critical step in satellite retrievals of trace gas columns is the calculation of the air mass factor 14 15 (AMF) used to convert observed slant columns to vertical columns. This calculation requires a priori information on the shape of the vertical profile. As a result, comparisons between satellite-16 17 retrieved and model-simulated column abundances are influenced by the *a priori* profile shape. We examine how differences between the shape of the simulated and *a priori* profile can impact 18 19 the interpretation of satellite retrievals by performing an adjoint-based 4D-Var assimilation of synthetic NO_2 observations for constraining NO_x emissions. We use the GEOS-Chem Adjoint 20 21 model to perform assimilations using a variety of AMFs to examine how *a posteriori* emission 22 estimates are affected if the AMF is calculated using an *a priori* shape factor that is inconsistent 23 with the simulated profile. In these tests, an inconsistent *a priori* shape factor increased root mean square errors in *a posteriori* emission estimates by up to 30% for realistic conditions over 24 polluted regions. As the difference between the simulated profile shape and the *a priori* profile 25 shape increases, so do the corresponding assimilated emission errors. This reveals the importance 26 27 of using simulated profile information for AMF calculations when comparing that simulated output to satellite retrieved columns. 28

29

30 **1. Introduction**

Satellite observations provide a wealth of information on the abundance of trace gases in
the troposphere (Fishman et al., 2008). The next generation of satellite instruments, including the

upcoming geostationary constellation of TEMPO (Chance et al., 2013; Zoogman et al., 2017),
Sentinal-4 (Bazalgette Courrèges-Lacoste et al., 2011; Ingmann et al., 2012), and GEMS (Bak et
al., 2013; Kim, 2012), will provide information on NO₂ and other air quality relevant pollutants
on unprecedented spatial and temporal scales. Insight into processes that affect atmospheric
composition, including emissions (Streets et al., 2013), lifetimes (Fioletov et al., 2015; de Foy et
al., 2015; Laughner and Cohen, 2019), and deposition (Geddes and Martin, 2017; Kharol et al.,
2018) can be gained by interpreting this information with atmospheric chemistry models.

40 There are three main stages in retrieving trace gas abundances from ultraviolet and visible solar backscatter radiance measurements: calculating a light-path "slant column" by 41 fitting observed spectra to known spectral signatures of trace gases, removing the stratospheric 42 portion of the column, and converting the slant column to a vertical column density using an air 43 44 mass factor (AMF). AMFs are calculated using a radiative transfer model and are a function of viewing geometry, surface reflectance, clouds, and radiative transfer properties of the 45 46 atmosphere. AMF calculations also require an *a priori* estimate of the trace gas vertical profile and are sensitive to the profile shape (Eskes and Boersma, 2003; Palmer et al., 2001). 47 48 Uncertainties in AMF calculations are the dominant source of uncertainty in satellite NO2 49 retrievals over polluted regions (Boersma et al., 2007; Martin et al., 2002) largely due to 50 sensitivity to surface reflectance, clouds, aerosols, and a priori profile information (Lorente et al., 2017). 51

Boersma et al. (2016) highlighted the issue of representativeness errors in comparing model simulated values with UV-Vis satellite-retrieved columns. Vertical representativeness errors arise from the satellite's altitude-dependent sensitivity due to atmospheric scattering and can degrade the quality of model-measurement comparisons beyond errors that arise from either modeling or measurements alone. A consistent accounting of the altitude-dependent sensitivity is necessary to limit these errors.

58 Two common methods are used to account for vertical representativeness. In one method, 59 observed slant columns are converted to vertical columns using an air mass factor calculated 60 with scattering weights to represent instrument vertical sensitivity and shape factors to represent 61 the vertical profile (Palmer et al., 2001). Another commonly used method employs an AMF 62 provided with the retrieval to convert slant columns to vertical columns, and then applies an 63 averaging kernel to the simulated profile to resample the simulated profile in a manner that

mimics the satellite vertical sensitivity (Eskes and Boersma, 2003). In this method both the
averaging kernel and the retrieval AMF are calculated using an *a priori* NO₂ profile that may
have a different shape than the simulated profile, which may introduce errors in the observation-

67 simulation comparison (Zhu et al., 2016).

A common application of comparisons between satellite observed columns and model simulations is to constrain NO_x emissions (e.g. Ding et al., 2018; Ghude et al., 2013; Lamsal et al., 2011; Martin et al., 2003; Vinken et al., 2014). One such approach is the use of fourdimensional variational (4D-Var) data assimilation, which seeks to minimize a cost function that accounts for the difference between simulated and retrieved values. As the cost function is a difference between observed and simulated NO₂ columns, it is susceptible to vertical

representativeness errors resulting from inconsistent *a priori* vertical profile information.

In this work we examine how *a priori* profile assumptions impact satellite-model comparisons and use the GEOS-Chem adjoint as a case study to assess how this impact can affect the interpretation of satellite observations. Section 2 provides the mathematical framework for AMF calculations and satellite-model comparisons. Section 3 describes the adjoint model and synthetic observations for the case study. Section 4 discusses the results.

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81 **2.** Mathematical frameworks

82 2.1 AMFs and averaging kernels

The air mass factor translates the line-of-sight slant column abundances (Ω_s) retrieved from satellite observed radiances into vertical column abundances (Ω_v). An air mass factor is the ratio of Ω_s to Ω_v and depends on the atmospheric path as determined by geometry, NO₂ vertical profile (*n*), surface reflectance, and radiative transfer properties of the atmosphere. Here we use M(n) to represent an air mass factor derived using the vertical number density profile *n*:

$$M(\boldsymbol{n}) = \frac{\Omega_s}{\Omega_v} \tag{1}$$

88

89 In the method described by Palmer et al. (2001), a radiative transfer model is used calculate

90 scattering weights w(z) (also known as box air mass factors) which characterize the sensitivity of

91 backscattered radiance I_B to the abundance of a trace gas at altitude *z*:

$$w(z) = -\frac{1}{M_g} \frac{\alpha_{a,z}}{\alpha_{eff}} \frac{\partial \ln (I_B)}{\partial \tau}$$
(2)

93 where $\alpha_{a,z}$ is the temperature-dependent absorption cross section (m² molec⁻¹), α_{eff} is the effective 94 (weighted average) absorption cross section (m² molec⁻¹) and $\partial \tau$ is the incremental trace gas 95 optical depth. M_G represents a geometric path correction accounting for the satellite viewing 96 geometry:

$$M_G = \sec \theta_0 + \sec \theta \tag{3}$$

97 where θ is the solar zenith angle and θ_o is the satellite viewing angle. This information is then 98 combined with an *a priori* NO₂ shape factor (i.e. normalized vertical profile)

$$\boldsymbol{S}(z) = \frac{\boldsymbol{n}(z)}{\Omega_{v}} \tag{4}$$

99 typically calculated with an atmospheric chemistry model to provide an air mass factor via:100

$$M(\boldsymbol{n}) = \int_0^{tropopause} \boldsymbol{w}(z) \boldsymbol{S}(z) dz$$
(5)

where S(z) is calculated using vertical profile n(z). An attribute of the formulation of Palmer et al. (2001) is the independence of atmospheric radiative transfer properties w(z) and the vertical trace gas profile S(z). The AMF definition in Equation (1) combined with Eq. (4) indicates that a slant column can be calculated from a known vertical profile via:

$$\Omega_s = \int_0^{tropopause} \boldsymbol{w}(z)\boldsymbol{n}(z)dz \tag{6}$$

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106 In an alternative formulation, the air mass factor is represented as part of an averaging 107 kernel. As formulated by Rodgers and Connor (2003), the averaging kernel (A) provides the 108 information needed to relate the retrieved quantity \hat{n} to the true atmospheric profile n:

109

$$\widehat{\boldsymbol{n}} - \boldsymbol{n}_a = \boldsymbol{A}(\boldsymbol{n} - \boldsymbol{n}_a) \tag{7}$$

110 where n_a is an assumed *a priori* profile of number density. The elements of the column

averaging kernel are related to the scattering weights by:

$$A(z) = \frac{w(z)}{M(n_a)}$$
(8)

112 where $M(n_a)$ is an air mass factor calculated using *a priori* vertical profile information. It is

important to note that unlike scattering weights, averaging kernels depend on the *a priori*assumed vertical profile shape.

115 A lexicon is given in Table 1 as notation used to describe these treatments has varied 116 across the literature. We choose M for air mass factor as a single letter is clearer in equations, w117 for scattering weights to maintain the original formulation of Palmer et al. (2001), n for number 118 density following IUPAC recommendations, and Ω for column densities as is common in 119 radiative transfer literature.

Figure 1 shows examples of typical shape factor, scattering weight, and averaging kernel 120 profiles for a range of atmospheric conditions. NO₂ shape factors have significant variability; 121 Shape factors peak near the surface in urban regions due to local pollution sources, but peak in 122 the upper troposphere in more remote regions due to lightning. The shape of a scattering weight 123 124 profile depends strongly on surface reflectance and cloud conditions. Sensitivity in the lower troposphere increases over reflective surfaces. Clouds increase sensitivity above due to their 125 126 reflectance but shield the satellite from observing the atmosphere below. Averaging kernels have similarities with scattering weights but depend on both the shape of the prior and the satellite 127 128 sensitivity. As AMF calculations are a convolution of the shape factor and the scattering weight profiles, these shapes affect NO₂ retrievals. For these examples, the AMF for a clear sky 129 130 observation with surface reflectance of 0.01 can range from 0.7 in an urban region to 1.7 in a 131 remote region. This large difference demonstrates the importance of the assumed profile shape to 132 the retrieval process.

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134 **2.2** Comparing satellite observations to simulated values

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The following section expresses mathematically how satellite-model comparisons are madeusing various *a priori* profiles.

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139 **2.2.1 Using scattering weights**

Following Palmer et al. (2001), a retrieved vertical column $(\widehat{\Omega}_{v,o})$ is estimated using an observed slant column $\Omega_{s,o}$ and a simulation-based air mass factor $M(n_m)$, which can be calculated with Eq. (5) using the model-simulated NO₂ profile (n_m) :

$$\widehat{\Omega}_{\nu,o} = \frac{\Omega_{s,o}}{M(\boldsymbol{n_m})} \tag{9}$$

- 145 The difference Δ_m between the estimated retrieved column and the model-simulated vertical
- 146 column ($\Omega_{v,m}$) is:

$$\Delta_m = \Omega_{\nu,m} - \widehat{\Omega}_{\nu,o} \tag{10}$$

$$\Delta_m = \left(\sum_{0}^{tropopause} n_m\right) - \frac{\Omega_{s,o}}{M(n_m)} \tag{11}$$

Equation (11) describes how this comparison is used in practice. However, we can rearrange this expression in terms of model ($\Omega_{s,m}$) and observed ($\Omega_{s,o}$) slant columns using the definition of air mass factor:

$$\Delta_m = \frac{\Omega_{s,m}}{M(\boldsymbol{n_m})} - \frac{\Omega_{s,o}}{M(\boldsymbol{n_m})}$$
(12)

$$\Delta_m = \frac{1}{M(\boldsymbol{n_m})} \left(\Omega_{s,m} - \Omega_{s,o} \right) \tag{13}$$

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151 **2.2.2 Using averaging kernels**

152

153 Comparison of simulated and retrieved columns using the averaging kernel is described 154 by Eskes and Boersma (2003) and in the retrieval documentation in Boersma et al. (2011). The 155 averaging kernel is applied to the simulated profile in order to sample the simulated column in a 156 manner that reflects the retrieval sensitivity:

$$\widehat{\Omega}_{\nu,m} = \sum_{0}^{tropopause} A \boldsymbol{n_m}$$
⁽¹⁴⁾

157 The resampled simulated column is then compared to the retrieved vertical column ($\Omega_{v,o}$) using 158 the *a priori*-based air mass factor $M(n_a)$ supplied with the retrieval dataset:

159

$$\Delta_a = \widehat{\Omega}_{\nu,m} - \Omega_{\nu,o} \tag{15}$$

$$\Delta_a = \left(\sum_{i=0}^{tropopause} A_i n_{m,i}\right) - \frac{\Omega_{s,o}}{M(n_a)}$$
(16)

Equation (16) describes how this method is used in practice. To facilitate the comparison with
Eq. (13), Eq. (16) can be rewritten using an alternative formulation relating averaging kernels to
scattering weights:

$$\Delta_{a} = \left(\sum_{i=0}^{tropopause} \frac{\boldsymbol{w}_{i} \boldsymbol{n}_{m,i}}{M(\boldsymbol{n}_{a})}\right) - \frac{\Omega_{s,o}}{M(\boldsymbol{n}_{a})}$$
(17)

$$\Delta_a = \frac{1}{M(\boldsymbol{n}_a)} \left(\Omega_{s,m} - \Omega_{s,o} \right) \tag{18}$$

164

By comparing Eq. (13) to Eq. (18), it is evident that the underlying difference between the two approaches is the choice of *a priori* profile information used to calculate the AMF, as the averaging kernel method is not independent of *a priori* profile assumptions. This bias could be addressed by replacing the *a priori* -based AMF in Eq. (18) with a simulation-based AMF using the following relationship (Boersma et al., 2016; Lamsal et al., 2010):

$$M(\boldsymbol{n}_m) = M(\boldsymbol{n}_a) \frac{\sum \boldsymbol{A} \boldsymbol{n}_m}{\sum \boldsymbol{n}_m}$$
(19)

170 It should be noted that both the averaging kernel and scattering weight methods are 171 equivalent for comparisons that examine ratios of retrieved and modeled columns:

$$r_{m} = \frac{\widehat{\Omega_{v,o}}}{\widehat{\Omega_{v,m}}} = \frac{\frac{\Omega_{s,o}}{M(n_{m})}}{\sum n_{m}} = \frac{\frac{\Omega_{s,o}}{\sum n_{m}} \sum n_{m}}{\sum wn_{m}} = \frac{\Omega_{s,o}}{\sum wn_{m}}}$$

$$r_{a} = \frac{\frac{\Omega_{v,o}}{\widehat{\Omega_{v,m}}}}{\sum An_{m}} = \frac{\frac{\Omega_{s,o}}{M(n_{a})}}{\sum wn_{m}/M(n_{a})} = \frac{\frac{\Omega_{s,o}}{\sum wn_{m}}}{\sum wn_{m}}$$
(20)
(21)

172

For ratios, both methods are dependent on geophysical assumptions used to calculate scattering
weights but are independent of *a priori* profile information. Lastly, some studies (e.g., Buscela et
al., 2013; Qu et al., 2017) may directly assimilate slant column densities rather than vertical
column densities using

$$\Delta_{s,a} = \widehat{\Omega}_{s,m} - \Omega_{s,o} \tag{22}$$

$$= \left(\sum_{i=0}^{tropopause} \boldsymbol{w}_{i}\boldsymbol{n}_{m,i}\right) - \Omega_{s,o}$$
⁽²³⁾

This approach is also still dependent upon the scattering weights but not upon external *a priori*profile information. Overall, the choice of approach may be influenced by whether or not

scattering weights are available from either the NO₂ retrieval product or radiative transfer
calculations applied to the model. In contrast, use of Eq. (11) or (16) are applicable when these

181 are not explicitly available or provided.

182

183 **3. Tools and Methodology**

184 **3.1 GEOS-Chem and its adjoint**

The GEOS-Chem chemical transport model (www.geos-chem.org) is used to create 185 186 synthetic NO₂ observations and for their analysis. The GEOS-Chem version used here is version 35j of the GEOS-Chem Adjoint model. GEOS-Chem includes a detailed oxidant-aerosol 187 chemical mechanism (Bey et al., 2001; Park et al., 2004) and uses assimilated meteorological 188 fields from the Goddard Earth Observation System (GEOS-5), with 47 vertical levels up to 0.01 189 hPa and a horizontal resolution of 4°x5°. Global anthropogenic NO_x emissions are provided by 190 the Emission Database for Global Atmospheric Research (EDGAR) inventory (Olivier et al., 191 192 2005) with regional overwrites over North America (EPA/NEI99), Europe (EMEP), Canada (CAC), Mexico (BRAVO, (Kuhns et al., 2005)), and East Asia (Streets et al., 2006). Other NO_x 193 194 sources include biomass burning (GFED2 (Van der Werf et al., 2010)), lightning (Murray et al., 2012), and soils (Wang et al., 1998). This model has been used previously to constrain NO_x 195 196 emissions (Cooper et al., 2017; Henze et al., 2009; Qu et al., 2017, 2019; Xu et al., 2013; Zhang et al., 2016). 197

The GEOS-Chem adjoint (Henze et al., 2007, 2009) is used here to perform a 4D-Var data assimilation. The adjoint seeks to iteratively minimize a cost function generally defined by the difference between satellite retrieved and simulated columns (Δ , from either Eq. (11) if using a simulation-based air mass factor or Eq. (16) if using the retrieval *a priori*-based air mass factor):

$$J = \frac{1}{2} \boldsymbol{\Delta}^{T} \boldsymbol{S}_{o}^{-1} \boldsymbol{\Delta} + \frac{1}{2} \gamma_{R} (\boldsymbol{E} - \boldsymbol{E}_{a})^{T} \boldsymbol{S}_{E}^{-1} (\boldsymbol{E} - \boldsymbol{E}_{a})$$
(24)

where *E* and *E_a* are the *a posteriori* and *a priori* emissions, S_o and S_E are the retrieval and *a* priori emission error covariance matrices, and γ_R is a regularization parameter that allows for weighting the cost function towards the retrieved columns or *a priori* emissions. Tests performed here required 20-30 iterations to minimize the cost function.

3.2 Experiment Outline

In this study we perform 4D-Var data assimilation experiments to infer surface NO_x 209 210 emissions using synthetic NO₂ observations. We use synthetic observations built from known emission inventories to provide a "truth" that can be used to evaluate the inversion results. To 211 demonstrate how a priori profile information can propagate in an assimilation, we use either the 212 model profile (Δ_m , Eq. (11)) or an *a priori* profile (Δ_a , Eq. (16)) in the cost function. A one-week 213 spin-up window at the start of each adjoint iteration is used to allow NO_x to reach steady state. 214 Observation error covariances S_{a} are described as a relative error of 30% of the slant column 215 density, plus an absolute error of 10¹⁵ molecules cm⁻², which is representative of typical satellite 216 retrieved NO₂ column uncertainties (Boersma et al., 2007; Martin et al., 2002). We omit the a 217 *priori* emissions constraint in the cost function (i.e. set $\gamma_R=0$) to isolate the impact of the 218 219 observations.

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221 **3.2.1 Synthetic observations**

222 Synthetic observations (*Obs*₅) are created using a GEOS-Chem simulation where random 223 Gaussian noise with a standard deviation of 5% is added to the anthropogenic NO_x emissions. 224 Additional tests using observations where noise with a standard deviation of 30% is added 225 (*Obs*₃₀) are also used. No additional noise is added to the individual observations to isolate the 226 impact of AMF errors against additional sources of uncertainty. Figure 2 shows the standard (*a* 227 *priori*) anthropogenic NO_x emissions and the changes used to create the "true" emissions for the 228 synthetic observations.

229 For these tests, we use one observation per hour per 4°x5° grid box for a period of two weeks in July 2010. Observations consist of synthetic slant columns ($\Omega_{s,o}$) created by applying 230 231 scattering weights to the synthetic vertical profiles using Eq. (6). Scattering weights are 232 calculated using the LIDORT radiative transfer model (Spurr, 2002) by providing LIDORT with the observation conditions of OMI observations during July 2010, which are used to represent 233 234 typical viewing conditions of low earth orbit satellite observations, and aerosol profiles from the GEOS-Chem base simulation. To represent typical conditions, these representative scattering 235 236 weight profiles for each grid box are used to produce the synthetic slant columns. Tests performed for all 4°x5° grid boxes used here indicate that the mean relative difference between 237 an air mass factor calculated using an average scattering weight profile and the average of air 238

239	mass factors using observation-specific scattering weight profiles is less than 4%.
240	
241	3.2.2 Shape Factors
242	To test the impact of a priori profile information, seven different tests are performed
243	using seven different NO ₂ profile shapes for AMF calculations:
244	• Case SF_M : The GEOS-Chem model simulated profile (n_m) , updated at each iteration
245	of the adjoint run
246	• Case <i>SF</i> _{prior} : The <i>a priori</i> GEOS-Chem simulated profile, without updating.
247	• Case SF_{n30} : An <i>a priori</i> profile created by a GEOS-Chem simulation where global
248	anthropogenic NO _x emissions were perturbed with random Gaussian noise with a
249	standard deviation of 30%. In cases where this results in negative emissions, a value
250	of zero is used.
251	• Case <i>SF_{diffem}</i> : An <i>a priori</i> profile created by a GEOS-Chem simulation where regional
252	emission overwrites are turned off.
253	• Case <i>SF_{finer}</i> : An <i>a priori</i> profile created by a GEOS-Chem simulation run at finer
254	$(2^{\circ}x2.5^{\circ})$ resolution.
255	• Case SF_{trop} : An <i>a priori</i> profile that assumes the NO ₂ profile shape is uniform from
256	the surface to the tropopause (~200 hPa).
257	• Case SF_{BL} : An <i>a priori</i> profile that assumes the NO ₂ profile shape is uniform from the
258	surface to the boundary layer (~800 hPa).
259	
260	An advantage of using scattering weights and the simulated shape factor in a 4D-Var framework
261	is that it allows for the shape factor, and thus the AMF, to be updated at each iteration. When a
262	priori profiles from an external source are used it is not possible for them to update during the
263	inversion. The SF_M and SF_{prior} cases test the impact that iterative updates to the AMF have on a
264	posteriori estimates. The additional cases test for the impact of using an averaging kernel based
265	on a priori profile assumptions that are inconsistent with the model. In practice, averaging
266	kernels and <i>a priori</i> profiles included in retrieval data sets are generally derived from chemical

- transport models that have different physical processes, emissions, or spatial resolutions. The 267
- SF_{n30} and SF_{diffem} tests are representative of inversions that use *a priori* profile information from 268
- a different chemical transport model with similar resolution but different emissions. The SF_{finer} 269

- test represents an inversion that uses *a priori* profiles from a chemical transport model with a
- different horizontal resolution. The SF_{BL} and SF_{trop} tests do not represent any modern retrieval
- algorithms, but are used as extreme examples of an *a priori* that assumes no spatial variability.
- 273 The SF_{BL} profile is representative of polluted regions as indicated by the typical urban profile in
- Fig. 1, while the *SF*_{trop} profile is representative of a typical rural profile. Table 2 provides global
- 275 mean AMFs for these test cases, which range from 1.3-2.1, and the resulting global mean
- observed vertical columns, which range from $0.9-1.5 \ge 10^{15}$ molec/cm². Global mean 'observed'
- vertical columns are 33% higher for SF_{2x25} than for SF_M , and up to 66% higher for SF_{BL} . Global
- 278 mean 'observed' vertical columns for SF_{n30} and SF_{diffem} are similar to SF_M , although individual
- observations may differ by up to 18% for SF_{n30} and 28% for SF_{diffem} .
- 280

281 **4. Results**

Figure 3 shows root mean square errors (RMSE) for the *a posteriori* emissions estimated by the 4D-Var assimilations of *Obs5* synthetic observations. All tests successfully reduce the *a priori* emission error by an order of magnitude or more. The SF_M has the lowest RMSE indicating that it can best estimate the "true" emissions. The next lowest RMSE is for the SF_{prior} test, which uses the same initial model shape factor but does not update during the adjoint iterations, followed by the SF_{finer} , SF_{diffem} , SF_{n30} , SF_{trop} , and SF_{BL} tests.

- Figure 4 shows maps of the difference in RMSE between the SF_M test and the other tests for Obs_5 observations. The SF_M test has a lower RMSE than the other tests in 65-72% grid boxes where the difference is nonzero. Again, the SF_{prior} test is closest to the SF_M test with a root mean square difference of 2.9x10⁷ molec/cm²/s, followed by SF_{finer} (3.6x10⁷ molec/cm²/s), SF_{n30} (3.8x10⁷ molec/cm²/s), SF_{diffem} (4.0x10⁷ molec/cm²/s), SF_{trop} , (7.8x10⁷ molec/cm²/s), and SF_{BL} (9.0x10⁷ molec/cm²/s).
- Table 3 summarizes additional error statistics focused on grid boxes with significant emission sources. Errors in *a posteriori* emission estimates are correlated with the "true" emissions in the SF_{trop} and SF_{n30} tests, and weakly correlated in the SF_{BL} , SF_{prior} , and SF_{diffem} tests, indicating that these tests are not well constraining the emissions. Differences between tests are more significant over polluted regions where AMF errors are more influential; For example, in the regions with the highest NO_x emissions, RMSE values indicate SF_M outperforms SF_{n30} by 30% and SF_{trop} by >80%. Another sign of adjoint inversion quality is a low variance in errors.

While the posterior error is reduced relative to the *a priori* error in all tests, error standard deviations are 30% higher for SF_{n30} and 90% higher for SF_{trop} compared to SF_M . The global maximum error for the SF_{trop} test is 30% higher than for the SF_M test. All metrics indicate that the SF_M test best represents the "true" emissions.

Tests using Obs_{30} observations and the SF_M and SF_{trop} shape factors were also performed. Despite the difference between *a priori* observed vertical columns using these shape factors as indicated by Table 2, these assimilations produced similar *a posteriori* results, with RMSE of 2.9×10^8 molec/cm²/s for SF_M and 2.8×10^8 molec/cm²/s for SF_{trop} .

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310 5. Discussion & Conclusions

Accounting for the vertical profile dependence of satellite observations is essential to 311 312 accurately interpret those observations. This work examines how the choice of shape factor affects differences between simulated and satellite-retrieved quantities in a data assimilation 313 314 framework. Examination of the mathematical frameworks behind two common methods for comparing simulated and retrieved columns highlights how the method introduced by Palmer et 315 316 al. (2001) facilitates separation of observation sensitivity (scattering weights) from the profile shape (shape factor) enabling the model-retrieval comparison to be independent of *a priori* 317 318 profile assumptions.

In these case studies, vertical representativeness errors were best reduced by using a 319 320 shape factor that was consistent with the model simulation. This was especially true in polluted regions where the AMF errors dominate observation uncertainties, as deviations between the 321 tests were largest in these regions. The further the shape factor deviated from the model state the 322 larger the inversion errors became, as indicated by Fig. 5. The SF_{finer} test indicates that using a 323 324 higher resolution model to generate a priori profiles does not provide an advantage in 325 simulation-observation comparisons, as consistency between the simulation profile and the a *priori* shape factor is of greater importance. Comparing the SF_M and SF_{prior} tests shows that 326 327 allowing for the shape factor to update during the iterative adjoint process further reduces the RMSE by 10%. However, even without allowing for shape factor updates, using a shape factor 328 329 that is consistent with the initial model state produces a more accurate inversion result than using 330 other assumed profile shapes.

331

The case study presented here demonstrates that the shape factor source can have a strong

influence on adjoint inversion results. However, the magnitude of this influence can vary. 332 333 Inversion tests performed using synthetic observations based on random 30% perturbations to 334 emissions were insensitive to the AMF, despite large differences in a priori vertical column densities. In these tests, the cost function was more sensitive to the larger difference between the 335 observed and simulated slant columns (i.e. $\Omega_{s,m}$ - $\Omega_{s,o}$ in Eq. (13) and (19)) than to the AMF. This 336 337 indicates that while the cost function is mathematically dependent on the AMF, the inversion is less sensitive to vertical representativeness errors in cases where emissions are poorly 338 339 constrained, as is the case in recent adjoint inversion studies (e.g. Qu et al., 2017). However, choice of AMF will become increasingly important to adjoint inversions as emission inventories 340 improve. Furthermore, omitting the a priori emissions constraint in the cost function and 341 342 omitting noise in the observations in these tests to isolate the impact of the AMF effectively 343 assumes poorly constrained *a priori* emissions and ideal observations. In practice, cost function sensitivity to AMF choice may be buffered when a priori emissions uncertainties and 344 observational noise are considered. 345

As it is beneficial for a consistent shape factor to be used when comparing satellite retrieved values to model simulated results, it will be useful for data products to provide the information required for this method to the user community. This is most straightforward when scattering weights (rather than averaging kernels) are provided alongside retrieved column data, as scattering weights and shape factors are independently calculated, however simulation-based air mass factors can be calculated using the averaging kernel and *a priori*-based air mass factor via Eq. 19.

In summary, when comparing a model simulation to a satellite retrieved NO₂ column in a data assimilation environment utilizing column differences, calculating the AMF using the simulated shape factor allows for better accuracy in inversion results. This demonstration can provide general guidance for other methods of interpreting satellite observations with models, as using the simulated shape factor assures consistency in the vertical representativeness between model and retrieval.

359

360 6. Author Contributions

MJC and RVM designed the overall study. MJC designed and carried out the case studies and
 their analysis. All co-authors provided guidance in analyzing results. MJC prepared the

- 363 manuscript with contributions from all co-authors.
- 364

365 **7. Competing interests**

- 366 The authors declare that they have no conflict of interest.
- 367

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- 371

372 9. Data Availability

373 The GEOS-Chem chemical transport model and its adjoint are available at www.geos-chem.org

- 374 (last access: 20 August 2017). OMI NO₂ data used in this study is available from the NASA
- Goddard Earth Sciences Data and Information Services Center (https://disc.sci.gsfc.nasa.gov;
- last access: 14 March 2019). AMF code (Spurr, 2002; Martin et al., 2002) used to calculate
- scattering weights and air mass factors is available at http://fizz.phys.dal.ca/~atmos (last access:
- 37819 June 2017).
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546 Figure 1: (Left) Shape factor profiles from a GEOS-Chem simulation for July 2010. Shown are a

- 547 global average, and typical urban (Beijing), rural (Midwest USA), and remote (Tropical Pacific)
- 548 profiles. (Middle) Typical OMI scattering weight profiles for varying surface reflectance and
- 549 cloud height. (Right) Averaging kernels calculated using the same shape factors and scattering
- weights ("Clear Sky" surface reflectance is 0.01, "Cloudy" uses cloud height of 1 km).



Figure 2 (top) Anthropogenic NO_x emissions for July 2010 used in GEOS-Chem. (bottom) Ratio
of "true" emissions used to create *Obs5* synthetic observations to a priori NO_x emissions.



557 Figure 3: Global root mean square error (RMSE) values for 4D-Var estimates of NO_x emissions

558 for tests using various shape factors in AMF calculations.



- 561 Figure 4: Difference between root mean square error (RMSE) of adjoint tests for Obs₅ synthetic
- 562 observations. Root mean square differences between the *a posteriori* emissions estimates
- 563 (molec/cm 2 /s) are inset.





566 Figure 5: Scatterplot of adjoint test results. X-axis represents the deviation of the shape factor

from the model simulated shape factor (root mean square difference). Y-axis represents the *a posteriori* emissions error from the adjoint inversion.

569

Variable		Palmer et al., 2001	Eskes & Boersma, 2003	Boersma et al. 2016	Notation used here	
Air mass	factor	AMF	M	М	Μ	
Slant Col	umn	$\Omega_{\rm S}$	S	Ns	$\Omega_{\rm s}$	
Vertical C	Column	$\tilde{\Omega_V}$	V	Ňv	$\Omega_{\rm v}$	
Scattering	g Weight	w(z)	C_1	m	W	
Shape Fa	ctor	$S_z(z)$			S(z)	
Averagin	g Kernel		А	А	A	
Number of	lensity	n(z)	Х	X1	n(z)	
Geometric AMF		AMF _G			MG	
Test name	Shape factor source	Air Mass Factor	Synthetic observation (Obs_5) vertical column density (x10 ¹⁵ molec/cm ²)			
		(unitless)		•		
		Global Me	ean Global M $(x10^{15} m)$	flean Ma olec/cm ²) dif SF	ximum ference from M(%)	
SF _M	Model	2.1	0.9	-		
SF _{n30}	Model w/ 30% noise	2.1	0.9	19		
SF _{diffem}	Model w/ different emissions	2.1	0.9	28		
SF _{finer}	Model at finer (2°x2.5°) resolution	1.6	1.2	23		
SF_{trop}	Uniform in troposphere	1.8	1.0	57		
SF_{BL}	Uniform in boundary	1.3	1.5	27		

576 Table 2: Global mean air mass factors and synthetic observation vertical column density for

577 shape factors tested here.

578							
Test Name	Shape Factor Source	Correlation (r) of <i>a posteriori</i> error and "true" emissions	<i>a posteriori</i> RMSE (x10 ⁸ molec/cm ² /s)		Error standard deviation (x10 ⁸ molec/cm ² /s)		Maximum error (x10 ⁹ molec/cm ² /s)
		if "true" emissions $> 10^{10}$ molec/cm ² /s	"true" emissions > 10 ¹⁰ molec/cm ² /s	"true" emissions > 10 ¹¹ molec/cm ² /s	"true" emissions > 10 ¹⁰ molec/cm ² /s	"true" emissions > 10 ¹¹ molec/cm ² /s	
SF _M	Model	0.06^{*}	1.8	3.0	1.8	2.9	1.6
$\mathbf{SF}_{\mathrm{prior}}$	a priori	0.11	2.0	3.2	2.0	3.3	1.6
SF _{n30}	Model w/ 30% noise	0.26	2.1	3.9	2.1	3.8	1.8
SF_{diffem}	Model w/ different emissions	0.13	2.0	3.6	2.0	3.7	1.9
SF _{finer}	Model at finer (2°x2.5°) resolution	0.05*	2.1	3.2	2.1	3.2	1.8
SF _{trop}	Uniform in troposphere	0.39	2.8	5.6	2.8	5.5	2.1
SF _{BL}	Uniform in boundary layer	0.17	2.8	4.6	2.8	4.6	1.9

579Table 3: Summary of error statistics for adjoint tests. Values marked * indicate that correlation is580not statistically significant (p>0.05). For comparisons, mean "true" emissions for grid boxes with581emissions>10¹⁰ molec/cm²/s is 4.9x10¹⁰, and mean "true" emissions for boxes with582emissions>10¹¹ molec/cm²/s is 1.6x10¹¹ molec/cm²/s.