

1
2
3
4 **Molecular Insights into New Particle Formation**
5 **in Barcelona, Spain**
6

7 **James Brean¹, David C.S. Beddows¹, Zongbo Shi¹,**
8 **Brice Temime-Roussel², Nicolas Marchand², Xavier Querol³,**
9 **Andrés Alastuey³, María Cruz Minguillón³, and**
10 **Roy M. Harrison^{1a*}**
11

12 **¹Division of Environmental Health and Risk Management**
13 **School of Geography, Earth and Environmental Sciences**
14 **University of Birmingham, Edgbaston, Birmingham B15 2TT**
15 **United Kingdom**
16

17 **²Aix Marseille Univ, CNRS, LCE**
18 **Marseille, 13003, France**
19

20 **³Institute of Environmental Assessment and**
21 **Water Research (IDAEA-CSIC), Barcelona, 08034 Spain**
22

23 **^aAlso at: Department of Environmental Sciences / Center of**
24 **Excellence in Environmental Studies, King Abdulaziz University, PO**
25 **Box 80203, Jeddah, 21589, Saudi Arabia**
26

* To whom correspondence should be addressed (Email: r.m.harrison@bham.ac.uk)

27 **ABSTRACT**

28 Atmospheric aerosols contribute some of the greatest uncertainties to estimates of global radiative
29 forcing, and have significant effects on human health. New particle formation (NPF) is the process
30 by which new aerosols of sub-2 nm diameter form from gas-phase precursors and contributes
31 significantly to particle numbers in the atmosphere, accounting for approximately 50% of cloud
32 condensation nuclei globally. Here, we study summertime NPF in urban Barcelona in NE Spain
33 utilising particle counting instruments down to 1.9 nm and a Nitrate CI-API-ToF. The rate of
34 formation of new particles is seen to increase linearly with sulphuric acid concentration, although
35 particle formation rates fall short of chamber studies of H₂SO₄-DMA-H₂O, while exceeding those
36 of H₂SO₄-BioOxOrg-H₂O nucleation, although a role of highly oxygenated molecules (HOMs)
37 cannot be ruled out. The sulphuric acid dimer:monomer ratio is significantly lower than that seen in
38 experiments involving sulphuric acid and DMA in chambers, indicating that stabilization of
39 sulphuric acid clusters by bases is weaker in this dataset than in chambers, either due to rapid
40 evaporation due to high summertime temperatures, or limited pools of stabilising amines. Such a
41 mechanism cannot be verified in this data, as no higher-order H₂SO₄-amine clusters, nor H₂SO₄-
42 HOM clusters were measured. The high concentrations of HOMs arise from isoprene, alkylbenzene,
43 monoterpene and PAH oxidation, with alkylbenzenes providing greater concentrations of HOMs
44 due to significant local sources. The concentration of these HOMs shows a dependence on
45 temperature. The organic compounds measured primarily fall into the SVOC volatility class arising
46 from alkylbenzene and isoprene oxidation. LVOC largely arise from oxidation of alkylbenzenes,
47 PAHs and monoterpenes, whereas ELVOC arise from primarily PAH and monoterpene oxidation.
48 New particle formation without growth past 10 nm is also observed, and on these days oxygenated
49 organic concentrations are lower than on days with growth by a factor of 1.6, and thus high
50 concentrations of low volatility oxygenated organics which primarily derive from traffic-emitted
51 VOCs appear to be a necessary condition for the growth of newly formed particles in Barcelona.
52 These results are consistent with prior observations of new particle formation from sulphuric acid-

53 amine reactions in both chambers and the real atmosphere, and are likely representative of the urban
54 background of many European Mediterranean cities. A role for HOMs in the nucleation process
55 cannot be confirmed or ruled out, and there is strong circumstantial evidence for the participation of
56 HOMs across multiple volatility classes in particle growth.

57

58

59 1. INTRODUCTION

60 Atmospheric aerosols, defined as liquid or solid droplets suspended in a gas, affect the climate both
61 directly by scattering and absorbing radiation, and indirectly by acting as cloud condensation nuclei
62 (CCN) (Penner et al., 2011), providing great uncertainties in estimates of global radiative forcing
63 (IPCC, 2014). Further, fine ambient aerosols (defined as those with diameter below 2.5 μm) are the
64 fifth greatest global mortality risk factor, resulting in 103.1 million disability-adjusted life year loss
65 in 2015 (Cohen et al., 2017). The number concentration of the ultrafine fraction of these (aerosols
66 with diameter below 0.1 μm , referred to as ultrafine particles or UFP) pose potentially significant
67 health risks also, due to their high concentration and surface area. The more diffuse, gas-like
68 behaviour of UFP allows them to penetrate into the deep lung and enter the bloodstream (Miller et
69 al., 2017). Ultrafine particles occur in the urban environment either as primary emissions (e.g., from
70 car exhaust (Harrison et al., 2018)) or secondarily as the product of new particle formation (NPF)
71 (Brines et al., 2015; Guo et al., 2014; Kulmala et al., 2017; Lee et al., 2019)

72

73 NPF is the formation of aerosol particles from gas-phase precursors. NPF can be considered a two-
74 step process involving initial formation of a cluster of gas molecules at the critical diameter at
75 around 1.5 nm - the diameter at which a free-energy barrier must be overcome to allow the
76 spontaneous phase transition from gas to liquid or solid (Zhang et al., 2012), and the subsequent
77 growth of this droplet to a larger aerosol particle. The first step of this process is dependent upon
78 the stability and abundance of the clustering molecules. Sulphuric acid, water, and dimethylamine
79 (DMA), for example, efficiently form particles as the strong hydrogen bonding between the acid
80 base pair produces near negligible evaporation, much lower than the evaporation rate seen for the
81 more weakly bound sulphuric acid-ammonia-water system. Nucleation of sulphuric acid, DMA and
82 water proceeds at, or near to the kinetic limit in a chamber at 278 K when DMA mixing ratios are
83 sufficient (Almeida et al., 2013; Kürten et al., 2014). Once past this 1.5 nm diameter, condensation
84 and coagulation will drive particle growth. Both the abundance of condensable gases and their

85 vapour pressures limit condensational growth. Vapour pressures are especially important for the
86 initial growth stages, as the Kelvin effect barrier impairs condensation of more volatile species, with
87 this condition of low vapour pressures becoming less significant as the diameter of the particle
88 increases (Tröstl et al., 2016). Once sufficiently large (>50 nm), the loss processes of coagulation
89 and evaporation of these particles become inefficient, resulting in a significant atmospheric lifetime.
90 It is from these these diameters onwards the climate forcing effects of these particles become most
91 pronounced.

92

93 NPF processes happen globally, across a diverse range of environments from pristine polar regions,
94 to polluted urban megacities (Kerminen et al., 2018), and represent a significant source of CCN,
95 with 10-60% of NPF events shown to produce CCN and enhancement factors to CCN count ranging
96 from 0.5 – 11 (Lee et al., 2019 and references within). Strong NPF events are observed across a
97 range of urban environments, despite high condensation sinks $>10^{-2} \text{ s}^{-1}$ (Bousiotis et al., 2019; Yu et
98 al., 2016), and can act as a precursor to strong haze events (Guo et al., 2014). The occurrence of
99 urban NPF has only been partially explained by growing understanding from recent in-depth studies
100 (Yao et al., 2018). Recent advances in instrumentation allow for the measurement of particles down
101 to the critical diameter with instruments such as the particle size magnifier (PSM), and (Neutral) Air
102 Ion Spectrometer (NAIS/AIS) (Lee et al., 2019), and mass spectral techniques for measuring the
103 abundance and composition of neutral (Jokinen et al., 2012) and charged (Junninen et al., 2010)
104 clusters. Elucidated mechanisms with these techniques involve sulphuric acid and ammonia in
105 remote environments (Jokinen et al., 2018; Yan, 2018), monoterpene derived highly oxygenated
106 molecules (HOM) in remote environments (Rose et al., 2018), iodic acid in coastal environments
107 (Sipilä et al., 2016), and sulphuric acid and DMA in polluted urban environments (Yao et al., 2018).

108

109 Urban Barcelona sees frequent, strong summer-time NPF events occurring on 28% of days. These
110 events are associated with high insolation and elevated ozone ($\sim 60 \mu\text{g m}^{-3}$) when considering the

111 whole year (Brines et al., 2014, 2015). Ground-level observations report NPF events starting
112 typically at midday, and either occurring in urban Barcelona and the surrounding regional
113 background simultaneously, or isolated to either urban Barcelona or just the regional background
114 (Dall'Osto et al., 2013). Vertical profiles over urban Barcelona reveal that NPF occurs at higher
115 altitudes, and starts earlier in the day, as at a given altitude these events are not suppressed by early
116 traffic peaks contributing to particle load (Minguillón et al., 2015). Here, we examine gas phase
117 mass spectral evidence and particle formation rates at the critical diameter from sulphuric acid in
118 Barcelona, with possible contribution from strong bases and highly oxygenated organic molecules
119 (HOMs), as well as factors influencing subsequent particle growth.

120

121 **2. METHODS**

122 **2.1 Sampling Site**

123 The Palau Reial site in Barcelona (41°23'15" N, 2°6'53.64" E) is representative of the urban
124 background of Barcelona, located at the Institute of Environmental Assessment and Water Research
125 (IDAEA-CSIC) in the north-west of the city. Sampling was performed from a container 20 m from
126 a low traffic road, and 200 m from the nearest main road (Avinguda Diagonal). Data were taken
127 from 2018/06/28 through 2018/07/18.

128

129 **2.2 Chemical Ionisation Atmospheric Pressure Interface Time of Flight Mass**

130 **Spectrometry**

131 The Aerodyne Nitrate Chemical Ionisation Atmospheric Pressure interface Time of Flight Mass
132 Spectrometer (CI-APi-ToF) was used to make measurements of neutral oxygenated organic
133 compounds, organic and inorganic acids, bases, and their molecular clusters at high time resolution
134 with high resolving power. The ionization system charges molecules by adduct formation, such as
135 in the case of organic compounds with two or more hydrogen bond donor groups (Hytinen et al.,
136 2015), or proton transfer in the case of strong acids like sulphuric acid (Jokinen et al., 2012).

137 Hydroxyl or hydroperoxyl functionalities are both common hydrogen bond donating groups, with
138 hydroperoxyl being the more efficient hydrogen bond donor (Møller et al., 2017). This instrument
139 has been explained in great detail elsewhere (Jokinen et al., 2012; Junninen et al., 2010), but briefly,
140 the front end consists of a chemical ionisation system where a 10 L min^{-1} sample flow is drawn in
141 through the 1 m length 1" OD stainless steel tubing opening. A secondary flow was run parallel and
142 concentric to this sample flow, rendering the reaction chamber effectively wall-less. A $3 \text{ cm}^3 \text{ min}^{-1}$
143 flow of a carrier gas (N_2) is passed over a reservoir of liquid HNO_3 , entraining vapour which is
144 subsequently ionised to NO_3^- via an X-ray source. Ions are then guided into the sample flow. The
145 nitrate ions will then charge molecules either by clustering or proton transfer. The mixed flows
146 travelling at 10 L min^{-1} enter the critical orifice at the front end of the instrument at 0.8 L min^{-1} and
147 are guided through a series of differentially pumped chambers before reaching the ToF analyser. All
148 data analysis was carried out in the Tofware package in Igor Pro 7 (Tofwerk AG, Switzerland).
149 Signals except for those of amines and ammonia are divided by the sum of reagent ion signals and
150 multiplied by a calibration coefficient to produce a concentration. A calibration coefficient of 3×10^9
151 cm^{-3} was established based upon comparison with a sulphuric acid proxy (Mikkonen et al., 2011)
152 and is in line with a prior calibration with our instrument (Brean et al., 2019). Uniform sensitivity
153 between H_2SO_4 and all other species measured by CI-APi-ToF bar amines and ammonia was
154 assumed in this work. This introduces some uncertainties, as it relies upon both collision rates and
155 charging efficiencies to be the same within the ionisation source for all species. Amine and
156 ammonia signals are normalised to the nitrate trimer signal (Simon et al., 2016). Prior reports of
157 ammonia and amines as measured by CI-APi-ToF employed corona discharge systems, which
158 utilise higher concentrations of nitric acid, thus we report normalised signals. We present
159 correlations of each of these bases clustered with the nitrate dimer plotted against measurements
160 with the nitrate trimer, as well as their intercorrelations and example peak fits across Figure S1. C_2
161 amines, C_4 amines and ammonia were the only molecules of this kind found in our mass spectra.
162 Systematic uncertainties of +100% / -50 % arising from this method are assumed.

163 Due to the high resolving power of the CI-APi-ToF system (mass resolving power of 3000, and
164 mass accuracy of 20 ppm at 201 m/Q), multiple peaks can be fit at the same unit mass and their
165 molecular formulae assigned. Beyond 500 m/Q, peak fitting and assignment of compositions
166 becomes problematic as signal decreases, mass accuracy decreases, and the total number of possible
167 chemical compositions increases, so peaks above the C₂₀ region have not been assigned (Cubison
168 and Jimenez, 2015), however, signals past this region tended to be extremely low. All ions
169 identified are listed in Table S1. As proton transfer mostly happens with acids, and nearly all HOM
170 molecules will be charged by adduct formation it is possible to infer the uncharged formula;
171 therefore, all HOMs from here onwards will be listed as their uncharged form. The CI-APi-ToF
172 inlet was placed approximately 1.5 m a.g.l. CI-APi-ToF data is only available between the dates
173 2018/07/06 and 2018/07/17.

174

175 **2.2 Particle Size and Number Measurements**

176 Two Scanning Mobility Particle Sizer (SMPS) instruments measured particle size distributions at 5
177 minute time resolution, one Long Column SMPS (TSI 3080 EC, 3082 Long DMA, 3772 CPC, TSI,
178 USA) and one NanoSMPS (3082 EC, 3082 Nano DMA, 3776 CPC, TSI, USA) measuring the
179 ranges 10.9 – 478.3 nm and 4.5 – 65.3 nm respectively. A Particle Size Magnifier (A10, Airmodus,
180 FN) linked to a CPC (3775, TSI, USA) measured the sub-3 nm size fraction. The PSM was run in
181 stepping mode, operating at four different saturator flows to vary the lower size cut of particles that
182 it will grow (defined as the point of 50% efficiency, D₅₀). The instrument provided D₅₀ from 1.4 to
183 2.4 nm. The instrument switched between saturator flows each 2.5 minutes, giving a sub-2.4 nm
184 size distribution every 10 minutes. Aerosol sampling inlets were placed approximately 2 m a.g.l.

185

186 **2.3 Other Measurements**

187 Mixing ratios of non-methane VOC with proton affinity greater than H₃O⁺ were made using the
188 proton transfer reaction time of flight mass spectrometer (PTR-ToF-MS 8000, Ionicon Analytik

189 GmbH, Austria). A detailed description of the instrument is provided by Graus et al., (2010) The
 190 sampling set up, operating conditions, and quantification procedures are similar to those described
 191 in Minguillón et al. (2016). Continual monitoring of composition and mass of submicron aerosol
 192 >75 nm was carried out with an Aerosol Chemical Speciation Monitor (ACSM, Aerodyne, USA)
 193 (Ng et al., 2011). Ozone, NO, and NO₂ were measured by conventional ultraviolet and
 194 chemiluminescence air quality instrumentation. Meteorological data were supplied by the Faculty of
 195 Physics of University of Barcelona, from a nearby (200 m from the measurement site)
 196 meteorological station located at the roof of an 8 floor building.

197

198 **2.4 Condensation Sink and Particle Growth Rate**

199 The condensation sink (CS) represents the rate at which a vapour phase molecule will collide with
 200 pre-existing particle surface, and was calculated from the size distribution data as follows (Kulmala
 201 et al., 2012):

202

$$203 \quad CS = 4\pi D \sum_{d_p} \beta_{m,d_p} d_p N_{d_p}, \quad (1)$$

204

205 where D is the diffusion coefficient of the diffusing vapour (assumed sulphuric acid), β_m is a
 206 transition regime correction (Kulmala et al., 2001), d_p is particle diameter, and N_{d_p} is the number of
 207 particles at diameter d_p . The formation rate of new particles at size d_p is calculated as follows:

208

$$209 \quad J_{d_p} = \frac{dN_{d_p}}{dt} + CoagS_{d_p} \cdot N_{d_p} + \frac{GR}{\Delta d_p} \cdot N_{d_p} \quad (2)$$

210

211 where the first term on the right-hand side comprises the rate at which particles enter the size d_p ,
 212 and the latter two terms represent losses from this size by coagulation and growth respectively. J_5
 213 was calculated using the data between 5 - 10 nm, and $J_{1.9}$ was calculated using the measurements
 214 between 1.9 – 4.5 nm. We also calculated $J_{1.9}$ from our NanoSMPS data, employing the equations

215 of Lehtinen et al. (2007). $J_{1.9}$ from both methods showed reasonable agreement ($R^2 = 0.34$).
216 Agreement between J_5 and $J_{1.9}$ for each method was similar ($R^2 = 0.37$ and $R^2 = 0.38$ for $J_{1.9}$
217 calculated from PSM data and from Lehtinen et al. (2019) respectively). $J_{1.9}$ is greater than J_5 as
218 predicted from equation (2) by around a factor of 20. See Kulmala et al. (2001) for more
219 information on calculation of coagulation sinks and formation rates. Growth rates between 4.5 – 20
220 nm were calculated according to the lognormal distribution function method (Kulmala et al., 2012),
221 whereas those between 1.9 and 4.5 nm were calculated from PSM data using a time-delay method
222 between PSM and NanoSMPS data. Systematic uncertainties on our calculated $J_{1.9}$ values include
223 25% method uncertainty (Yli-Juuti et al., 2017), with a further 25% arising from uncertainties in
224 PSM cutoff (± 0.5 nm), as well as a 10% uncertainty in counting errors. A 50% error arising from
225 calculated coagulation sink is also applied (Kurten et al., 2016). The above calculations rely on the
226 assumption of homogeneous air masses, and while air mass advection, as well as primary particle
227 emissions can cause errors in estimations of temporal changes in particle count and diameter, the
228 appearance and persistence of a new mode of particles across a period of several hours is typically
229 indicative of a regional process.

230

231 Growth rates from irreversible condensation of various vapours were calculated as according to
232 Nieminen et al. (2010). At our measured relative humidity, sulphuric acid favours binding to 2 H₂O
233 molecules (Kúrten et al., 2007). As amine concentrations are likely limited, we presume no mass
234 from amines in the condensing species. H₂SO₄ was assigned a density of 1.8 g cm⁻³. For simplicity,
235 the properties of MSA regarding density and hydration were presumed the same as H₂SO₄, and
236 HIO₃ was presumed to have the same hydration as H₂SO₄, with a density of 4.98 g cm⁻³. The
237 density of condensing organic vapours was assumed 1.5 g cm⁻³, and concentration-weighted mean
238 mass (~ 276 g mol⁻¹ for LVOC) and atomic weighted diffusion volumes of organic compounds were
239 used to calculate GRs.

240

241 **2.4 DBE and 2D-VBS**

242 The double bond equivalent (DBE) describes the degree of unsaturation of an organic molecule and
243 is defined simply as:

244

$$245 \quad DBE = N_C - \frac{N_H}{2} - \frac{N_N}{2} + 1 \quad (3)$$

246

247 The saturation vapour pressure at 300 K is defined by the 2D-volatility basis set (2D-VBS) as
248 follows, if all nitrogen functionality is assumed to take the form -ONO₂ (Bianchi 2019; Donahue
249 2011; Schervish and Donahue, 2020):

250

$$251 \quad \text{Log}_{10}(C^*)(300 \text{ K}) = (N_{C0} - N_C)b_C - N_O b_O - 2 \frac{N_O N_C}{N_C + N_O} b_{CO} - N_n b_N \quad (4)$$

252

253 where N_C , N_H , and N_N , are the number of carbon, hydrogen, and nitrogen atoms respectively. N_O is
254 the number of oxygen atoms minus $3N_N$ to account for -ONO₂ groups, N_{C0} is 25 (the carbon
255 number of a 1 $\mu\text{g m}^{-3}$ alkane), b_C , b_O , b_{CO} , and b_N are 0.475, 0.2, 0.9 and 2.5 respectively, and
256 represent interaction and nonideality terms. The final term of equation (4) represents the -ONO₂
257 groups, each reducing the saturation vapour pressure by 2.5 orders of magnitude. C^* values are
258 calculated at 300 K and not corrected for temperature, as 300 K is within 1 K of the campaign
259 average temperature.

260

261 **3. RESULTS AND DISCUSSION**

262 **3.1 General Conditions of NPF Events**

263 Summer NPF events in the regional background around Barcelona are associated with high
264 insolation, relatively low ozone (high compared with the rest of the year), and lower particulate
265 matter load (Brines et al., 2014; Carnerero et al., 2019). Figure 1 shows an example of a day with no
266 NPF in panel (a), referred to as “non-event” here, where two traffic-associated peaks in particle

267 number are seen during rush hours. Midday traffic peaks are also seen on certain days, but these are
268 easily distinguished from nucleation processes due to the lack of a significant <10 nm mode. Figure
269 1(b) shows a nucleation day with growth to larger sizes >10 nm, termed “full-event”, showing the
270 growth through the course of the day. These fulfil all the criteria of Dal Maso et al. (2005). 4 events
271 of this type were observed with CI-APi-ToF data coverage. Figure 1(c) shows a day with nucleation
272 occurring, but no growth past 10 nm. These days are referred to as “burst-event” days. Here, NPF is
273 seen to occur, but particles fail to grow past the nucleation mode. 2 such events were seen in this
274 data with CI-APi-ToF data coverage, and both are accompanied by a distinct mode appearing
275 beforehand at ~20-40 nm. Condensation sinks were not significantly higher than on full event days,
276 so this failure of particles to grow further cannot be attributed to condensational (or coagulation)
277 losses. $GR_{4.5-20}$ ranged between 2.47 – 7.31 $nm\ h^{-1}$ ($4.69 \pm 2.03\ nm\ h^{-1}$), $GR_{1.9-4.5}$ ranged between
278 3.12 – 5.20 $nm\ h^{-1}$ ($4.36 \pm 1.02\ nm\ h^{-1}$). The survival parameter (P) as suggested by Kulmala et al.
279 (2017) is defined as $CS \cdot 10^{-4} / GR$, and for this data is equal to 82, higher than other European cities.
280 The occurrence of such a high P value should, in theory, inhibit the occurrence of NPF, but we
281 show events happen readily under such conditions, akin to other heavily polluted megacities.

282

283 Figure 2 contains box plots showing condensation sink, temperature and global radiation for all 3
284 NPF types across the entire day (diurnal profiles plotted in Figure S2). Condensation sinks during
285 NPF periods of both types (Figures 1(b) & 1(c)) were not significantly lower than in non-event
286 periods. Condensation sinks were suppressed prior to the beginning of an event for full-events,
287 increasing relative to non-events through the afternoon period. Of the two burst-events, one was
288 similarly characterised by a suppression to condensation sink, whereas the other showed a sharp rise
289 in the midday. Global radiation and temperature were higher for full-events, most significantly for
290 temperature. Figure 3 is as Figure 2 but for sulphuric acid, ammonia and amines, and HOMs as
291 measured by CI-APi-ToF (HOM criteria are discussed in section 3.3.1). Sulphuric acid is elevated
292 during both full-event and burst-event periods. In urban Barcelona, sulphuric acid will primarily

293 arise from oxidation of SO₂ by the OH[•] radical, with anthropogenic emissions such as shipping
294 emissions from the port areas being significant sources of SO₂ (Henschel et al., 2013). Direct traffic
295 emissions have been shown to be a significant primary sulphuric acid source (Olin et al., 2020), but
296 our sulphuric acid data show no traffic peaks. Ammonia and amines show enhancement for full-
297 event periods, but not burst-event periods. Nucleation rates (at typical tropospheric sulphuric acid
298 concentrations) are sensitive to amine concentrations in the range of a few pptv, with enhancements
299 to amine mixing ratios past this point increasing the nucleation rate marginally (Almeida et al.,
300 2013), while typical concentrations of DMA and other alkylamines vary from zero to a few pptv in
301 the boundary layer (Ge et al., 2011a).

302

303 Barcelona has been shown to contain ppbv levels of ammonia (Pandolfi et al., 2012), arising from
304 both agriculture to the north (Van Damme et al., 2018), and anthropogenic activities such as waste
305 management and traffic, with waste management being the primary ammonia source. Highest
306 ammonia mixing ratios are found in the densely populated old city centre (Reche et al., 2015).
307 Agriculture, waste management, and traffic are also all significant sources of low molecular weight
308 alkylamines, such as DMA (Ábalos et al., 1999; Cadle and Mulawa, 1980; Hutchinson et al., 1982;
309 Ge et al., 2011a), and are likely the source of amines found in this dataset. Activities such as
310 composting and food industry are especially strong sources of trimethylamine (TMA) (Ge et al.,
311 2011a). Although high emission fluxes of TMA are expected in this environment, they are not
312 present in our spectra. The TMA ion has been reported previously with a similar ionisation setup to
313 that utilised in this study (Kürten et al., 2016). On full-event days, the signal for C₂ and C₄ amines
314 has a midday elevation concurrent with peaks to solar radiation (Figure S2), and can help explain
315 the high formation rates we see in this dataset (see section 3.2). The relative strength of these
316 signals are shown in Figure S3, with significantly higher signals attributed to ammonia compared to
317 amines, despite a likely lower sensitivity (Simon et al., 2016).

318

319 HOM concentrations were greatly enhanced during full-event periods (factor of 1.5 higher
320 compared to non-NPF mean), but lower during burst-event periods (factor of 1.2 lower compared to
321 non-NPF mean), implying their necessity for growth. The sources and implications of these HOMs
322 are discussed in section 3.3. Further, concentrations of iodine and DMS-derived acids such as iodic
323 acid (HIO_3) and methanesulphonic acid (MSA) are low ($7.8 \cdot 10^5$ and $3.3 \cdot 10^5 \text{ cm}^{-3}$ respectively),
324 indicating a small influence of oceanic emissions on particle nucleation/growth. Extended box plots
325 as Figures 2 & 3 are presented in Figure S4, and HYSPLIT back trajectories per event in Figure S5.

326

327 **3.2 Mechanisms of New Particle Formation**

328 The correlation between $J_{1,9}$ and concentration of sulphuric acid is plotted in Figure 4. A close
329 relationship between nucleation rates and sulphuric acid concentrations ($R^2 = 0.49$) are consistent
330 with observations globally (Lee et al., 2019). This relationship is not dependent upon condensation
331 sink. These NPF rates have no dependence on other ions as measured by CI-API-ToF, including
332 HIO_3 , MSA, ammonia, amines or HOMs (R^2 for all < 0.1). This is not to say that all of these
333 molecules are not involved in the nucleation process, rather that elevations or reductions to their
334 concentrations during nucleation periods do not have significant impact on nucleation rates. In the
335 example of alkylamines, their gas phase concentration may decrease due to clustering with elevated
336 sulphuric acid, as they cluster at around a 1:1 ratio at high amine mixing ratios (Kürten et al., 2014)
337 (and therefore they will not be detectable as free amines). Further, if amines are present at a few
338 pptv, their mixing ratios are significantly higher than our ambient measured sulphuric acid
339 concentrations, and will be sufficient to accelerate nucleation rates (Almeida et al., 2013).
340 Photochemical losses will also be greater during the periods of highest NPF rate (Ge et al., 2011b).
341 The strength of the relationship between sulphuric acid and nucleation rate has been quantitatively
342 reproduced in chamber studies involving the $\text{H}_2\text{SO}_4\text{-H}_2\text{O-DMA}$, and $\text{H}_2\text{SO}_4\text{-H}_2\text{O-BioOxOrg}$
343 system, accurately reproducing tropospheric observations of nucleation rates (Almeida et al., 2013;
344 Riccobono et al., 2014), although a later revision of the former shows nucleation rates at 278 K

345 exceeding typical tropospheric observations in the presence of high mixing ratios of DMA (Kürten
346 et al., 2018). A comparison between our data and results from the CLOUD chamber is presented in
347 Figure 5; included are the H₂SO₄-H₂O, H₂SO₄-NH₃-H₂O (Kirkby et al., 2011), H₂SO₄-H₂O-DMA
348 (Kürten et al., 2018) and H₂SO₄-BioOxOrg-H₂O systems (Riccobono et al., 2014) – BioOxOrg
349 refers to the oxidation products of pinanediol (C₁₀H₁₈O₂) and OH. Data from these chamber
350 experiments is for 278 K and 38 - 39 % relative humidity. Nucleation rates measured in Barcelona
351 ($J_{1.9} 178 \pm 190 \text{ cm}^{-3} \text{ s}^{-1}$ at $[\text{H}_2\text{SO}_4] 7.1 \cdot 10^6 \pm 2.7 \cdot 10^6 \text{ cm}^{-3}$) are around an order of magnitude lower
352 than that seen for the H₂SO₄-DMA-H₂SO₄ system, but exceed that of the H₂SO₄-BioOxOrg-H₂O
353 system by ~1 order of magnitude, and that of the H₂SO₄-NH₃-H₂O and H₂SO₄-H₂O system multiple
354 orders of magnitude. No dissimilarity is seen between the data points corresponding to full or burst
355 type nucleation, indicating similar mechanisms of formation, despite lower HOM concentrations on
356 burst-event days. Conversely, research in remote boreal environments show that the mechanism of
357 nucleation can modulate dependent upon the H₂SO₄:HOM ratio (Yan et al., 2018). Model studies of
358 sulphuric acid-amine nucleation show a decline in nucleation rate with temperature (Almeida et al.,
359 2013; Olenius et al., 2017), as the evaporation rate of sulphuric acid-amine clusters will increase
360 with temperature (Paasonen et al., 2012). Conversely, evaporation rates of such small clusters, and
361 resultant nucleation rates tend to increase modestly with increases in relative humidity, most
362 pronounced at lower amine concentrations (Almeida et al., 2013; Paasonen et al., 2012). Despite
363 this, high nucleation rates at temperatures nearing 300 K have been reported previously (Kuang et
364 al., 2008; Kürten et al., 2016), although these tend to show a temperature dependence (Yu et al.,
365 2016). No higher-order sulphuric acid clusters, sulphuric acid-base clusters, nor sulphuric acid-
366 HOM clusters were visible in the mass spectral data, likely due to these being below the limit of
367 detection of the instrument (Jokinen et al., 2012), so cluster identity cannot be directly identified.
368 Sulphuric acid trimer stabilisation is dependent upon base abundance (Ortega et al 2012), and
369 conversely, sensitivity of nitrate CI-APi-ToF to sulphuric acid-base clusters is reduced due to the
370 high base content of such clusters (Jen et al., 2016).

371 To further explore the relationship between sulphuric acid clusters and the rate of nucleation, the
372 sulphuric acid dimer:monomer ratio is plotted in Figure 6. The sulphuric acid dimer:monomer ratio
373 is elevated by the presence of gas-phase bases such as DMA, and this elevation is dependent upon
374 both the abundances and proton affinities of such bases (Olenius et al., 2017). Upon charging,
375 evaporation of water and bases from sulphuric acid clusters occurs, and thus these are detected as
376 sulphuric acid dimer (Ortega et al., 2012, 2014). The binding energy of the bisulphate-H₂SO₄ ion is
377 in excess of 40 kcal mol⁻¹ (Curtius et al., 2001), and thus minimal declustering of the dimer is
378 expected within the CI-APi-ToF instrument – however, declustering of higher order sulphuric acid
379 clusters has been shown to be sensitive to voltage tune (Passanati et al., 2019), and this likely
380 extends to the dimer also, and as such discrepancies between sets of results due to instrument setup
381 cannot be ruled out. The ratio of sulphuric acid dimer:monomer is also highly sensitive to
382 condensation sinks, with a difference in dimer concentration of approximately a factor of 4
383 expected at 10⁷ cm⁻³ between 0.001 s⁻¹ (a clean environment) and 0.03 s⁻¹ (condensation sinks
384 during these NPF events measured in this dataset) (Yao et al., 2018) and thus our low
385 dimer:monomer ratio can, in part, be explained by elevated condensation sinks. The dashed line
386 represents the ratio that would be seen due to ion induced clustering (IIC) in the nitrate chemical
387 ionisation system for a 50 ms reaction time (Zhao et al., 2010). The sulphuric acid dimer:monomer
388 ratio seen in the CLOUD H₂SO₄-DMA-H₂O system is plotted, alongside our own data from
389 Barcelona. The ratio from our own data is seen to be much lower than that for the system purely
390 involving DMA as a ternary stabilising species. Similarly, this ratio is lower than for reports of
391 H₂SO₄-DMA-H₂O nucleation in Shanghai (Yao et al., 2018), but is markedly similar to reports in
392 central rural Germany (Kürten et al., 2016). Similar to central Germany, this ratio increases at lower
393 sulphuric acid concentrations to a ratio more similar to the H₂SO₄-DMA-H₂O system. A possible
394 explanation for this is that at higher sulphuric acid concentrations, the concentrations of stronger
395 stabilising bases are insufficient to stabilise all present sulphuric acid, with the higher end of the
396 sulphuric acid concentrations seen in this data roughly equivalent to 1 pptv sulphuric acid (3×10⁷

397 $\text{cm}^{-3} = 1.2 \text{ pptv}$ sulphuric acid). We also cannot account for clustering due to naturally charged
398 sulphuric acid in the atmosphere, but ion concentrations in urban environments tend to be small due
399 to efficient sink processes (Hirsikko et al., 2011). Particle formation plausibly operates by sulphuric
400 acid-amine nucleation involving the measured C_2 and C_4 amines in our data, with nucleation rates
401 hindered relative to those measured in the CLOUD experiments by elevated temperatures, and a
402 decline to the sulphuric acid dimer:monomer ratio indicates that base concentrations may be
403 limited. We cannot rule out an involvement of HOMs in particle formation processes, and, as no
404 higher-order clusters were observed, we cannot establish sulphuric acid-amine nucleation with
405 certainty.

406

407 **3.3 HOMs and Growth**

408 **3.3.1 HOM composition and sources**

409 Barcelona, as a densely populated urban agglomerate, is distinct from the remote conditions under
410 which HOMs have primarily been studied (Bianchi et al., 2016, 2017; Schobesberger et al., 2013;
411 Yan et al., 2016), and is characterised by elevated temperatures, insolation and NO_x mixing ratios,
412 as well as a diverse host of potential precursor VOC. The first of these affects HOM yields
413 significantly, as yields are highly dependent upon temperature (Quéléver et al., 2019; Stolzenburg et
414 al., 2018). Lower temperatures result in slower H-abstractions, which will result in the likelihood of
415 an RO_2 to undergo a different reaction pathway, such as termination with HO_2 to increase (Praske
416 et al., 2018). This is particularly important in this study if there is a large energy barrier for the first
417 or second H-abstraction taking place, as this will determine the number of hydrogen bond donating
418 groups, and therefore whether the NO_3^- CI-APi-ToF is sensitive to a molecule or not. Elevated
419 insolation will result in enhanced photochemistry, and thus more rapid RO_2 formation rates,
420 whereas elevated NO_x will produce more HOM with nitrate ester functionality (Garmash et al.,
421 2019; Rissanen, 2018), which tend towards higher volatilities, and less efficient participation in
422 particle formation (Ehn et al., 2014; Lehtipalo et al., 2018), and growth (Yao et al., 2020).

423

424 Oxygenated volatile organic compounds (OVOC) are defined as species visible in the nitrate CI-API-
425 ToF that do not classify as HOM. Here, the first of the three criteria provided by Bianchi et al. (2019),
426 that HOM must be formed by peroxy radical autoxidation, cannot be applied to define HOM, as
427 knowledge as to whether a molecule is a result of autoxidation requires sound knowledge of the
428 structure of the precursors, oxidants and peroxy radical terminators present, however, the number of
429 molecules observed with $nN = 2$ is around an order of magnitude lower than that for $nN = 1$, where
430 the primary source of multiple nitrogen functionalities would be multiple peroxy radical termination
431 reactions from NO_x , and therefore while multiple generations of oxidation have been shown to occur
432 in aromatics (Garmash et al., 2020), it is a small contributor to the concentration of what is classed as
433 HOM here. The second criterion to define HOM are that they must be formed in the gas phase under
434 atmospherically relevant conditions, which we deem appropriately fulfilled as all CI-API-ToF
435 measurements are of gas phase compounds, and the final criterion is that HOM must contain more
436 than 6 oxygen atoms. To attempt to satisfy these criteria as best possible, the criteria of both
437 containing 6 oxygen atoms and 5 carbon atoms or greater and having an O:C ratio >0.6 is applied.

438

439 The diversified range of HOM precursors in Barcelona will be primarily anthropogenic in origin.
440 Averaged PTR-MS mixing ratios of different VOCs are presented in Figure S6. Figure 7(a) shows
441 HOM concentration plotted against temperature, showing a dependence of HOM concentrations on
442 temperature, with a lesser dependence on global radiation. The precursors for these HOMs are
443 presumed to be largely isoprene, alkylbenzenes, monoterpenes, and PAHs. The mean peak
444 intensities assigned to alkylbenzene derived HOMs are approximately a factor of two higher than
445 those assigned to isoprene and monoterpene oxidation across this entire campaign. In this data these
446 VOC mixing ratios are, with the exception of isoprene, not largely temperature dependent, with
447 many of these HOMs forming under negligible or zero insolation, and therefore very low OH
448 concentrations. These nighttime HOMs will not be derived from the oxidation of aromatics,

449 however, as rates of oxidation of alkylbenzenes by O_3 and NO_3 are negligible (Molteni et al, 2018).
450 These nighttime HOMs will therefore mostly be derived from biogenic emissions which undergo
451 more rapid nocturnal oxidation, and are likely transported from inland by the land breeze during
452 night (Millán, 2014; Querol et al., 2017).

453

454 Operating under the assumption that C_5 , C_6 , C_7 , C_8 , and C_9 HOMs primarily arise from isoprene,
455 benzene toluene, C_2 -alkylbenzene C_3 -alkylbenzene oxidation respectively (Massoli et al., 2018;
456 Molteni et al., 2018; Wang et al., 2017), HOM signals plotted against parent VOC concentration
457 indicate their dependence upon that VOC. Here, a C_7 HOM is thought to follow the formula C_7H_8 -
458 $12O_{5-10}N_{0-2}$. We have plotted HOM concentrations against VOC concentrations in Figure 7(b). C_{10}
459 HOMs are not included in these analyses as these may primarily arise from $C_{10}H_{12-14}$ alkylbenzene,
460 or monoterpene oxidation. HOM concentration appears mostly independent of VOC concentration,
461 with the exception of isoprene, for which emissions are highly temperature dependent, and thus this
462 is likely a function of the effect of temperature on HOM formation (Figure 7(a)). A lack of
463 correlation between other VOCs and their HOMs confirms that this relationship between HOMs
464 and temperature is not a function of enhanced VOC emission fluxes from, for example, evaporation,
465 except in the instance of isoprene. Fragmented monoterpene oxidation products will also contribute
466 to the total number of C_9 HOMs, and similarly, other VOCs can fragment upon oxidation. However,
467 these results indicate that HOM concentrations are elevated by temperature, and operate quite
468 independent of precursor VOC concentration.

469

470 DBE as calculated by equation 3 is equal to the number of pi bonds and rings within a molecule.
471 Benzene, toluene, and similar aromatics have $DBE = 4$, naphthalene = 7 and monoterpenes = 3.
472 DBE can be used as an indicator of sources when considering HOM in bulk. Saturation mass
473 concentration as calculated by equation 4 can help describe capacity of a molecule to both condense
474 onto newly formed particles and participate in nucleation. Figure 8 shows concentrations of HOMs

475 and other oxygenated organic molecules binned to the nearest integer $\text{Log}_{10}(C^*)(300\text{ K})$, coloured
476 by DBE. Mean ion signals per carbon number are shown in Figure S7. Most measured molecules
477 fall into the SVOC class ($0.3 < C^*(300\text{ K}) < 300\ \mu\text{g m}^{-3}$) which will mostly exist in equilibrium
478 between gas and particle phase. Highest SVOC concentrations arise from fingerprint molecules for
479 isoprene oxidation under high NO_x ($\text{C}_5\text{H}_{10}\text{N}_2\text{O}_8$) (Brean et al., 2019), and oxidation of small
480 alkylbenzenes ($\text{C}_7\text{H}_8\text{O}_5$, $\text{C}_8\text{H}_{10}\text{O}_5$). LVOC and ELVOC ($3 \cdot 10^{-5} < C^* < 0.3\ \mu\text{g m}^{-3}$ and $3 \cdot 10^{-9} <$
481 $C^*(300\text{ K}) < 3 \cdot 10^{-5}\ \mu\text{g m}^{-3}$ respectively) have a greater contribution from molecules with higher
482 DBE, i.e., $\text{C}_{10}\text{H}_{10}\text{O}_8$ arising most likely from PAH oxidation (Molteni et al., 2018), and $\text{C}_{10}\text{H}_{15}\text{O}_7\text{N}$,
483 a common molecule arising from monoterpene oxidation in the presence of NO_x . The contribution
484 of molecules with carbon number ≤ 9 to these LVOC is modest, and ELVOCs are entirely
485 comprised of molecules with carbon numbers ≥ 10 , and is dominated by DBEs of 8 and 4,
486 attributable to PAH and monoterpene oxidation respectively. No molecules classed as ultra-low
487 volatility organic compounds (ULVOC, $C^*(300\text{ K}) < 3 \cdot 10^{-9}\ \mu\text{g m}^{-3}$) were observed in our data, and
488 thus any pure HOM nucleation is unlikely.

489

490 3.3.2 HOMs and NPF

491 As shown in Figure 3, an elevated HOM concentration appears to be a necessary condition for particle
492 growth past 10 nm during NPF events. These days are associated with elevated temperatures, solar
493 radiation, higher ozone, and lower $\text{NO}:\text{NO}_2$ ratio. HIO_3 is also significantly higher on burst-event
494 days. A recent study in a remote environment reports growth rates matching condensation rates
495 without accounting for aqueous phase chemistry (Mohr et al., 2019). From 2D-VBS volatility
496 calculations discussed in the previous section, it is shown that LVOC and ELVOC measured in
497 Barcelona plausibly arise from the oxidation of aromatics (particularly PAHs in the case of ELVOC)
498 and monoterpenes. Calculated growth rates according to the method of Nieminen et al. (2010) are
499 presented in Figure S8 for both $\text{GR}_{1.9-5}$ and GR_{5-20} . Best agreement for GR_{5-20} is when condensation
500 of SVOC, LVOC, ELVOC, MSA, HIO_3 and H_2SO_4 is considered, and best agreement for $\text{GR}_{1.9-5}$ is

501 seen for condensation of all these except SVOC. The uncertainties in this method are large, and
502 assumptions of irreversible condensation of SVOC onto particles of 5 nm likely lead to
503 overestimations; however, these results indicate an essential role of the condensation of organic
504 compounds to produce high growth rates observed in urban environments.

505

506 Figure 9 shows three mass-defect plots for a non-event period, full-event period, and burst-event
507 period. The non-event day included in Figure 8 was characterised by lower solar radiation and
508 temperatures than average, so lower signals for oxygenated species are seen due to weaker
509 photochemistry (i.e., OH[•] concentration), and slower autoxidation due to slower H-shift reactions
510 (Frege et al., 2018; Quéléver et al., 2019; Stolzenburg et al., 2018). The full-event day sees
511 enhancements to smaller OVOCs and HOMs compared to the non-event day, especially around
512 150-200 m/Q, which contains peaks corresponding to dicarboxylic acids and isoprene oxidation
513 products. Some of the largest peaks in the mass spectra correspond to formulae seen arising from
514 the enhanced OH[•] oxidation of alkylbenzenes (such as C₇H₇NO₆) (Molteni et al., 2018; Wang et al.,
515 2017). Larger HOMs see a less significant enhancement to smaller alkylbenzene derived HOMs.
516 The presence of larger, unidentified HOMs >400 m/Q is enhanced during full-events, these peaks
517 will comprise the largest compounds, most likely of class ELVOC, arising from the oxidation of
518 large VOCs, or RO₂[•]-RO₂[•] accretion reactions, and thus, we likely underpredict ELVOC
519 concentrations and resultant impacts on particle growth in Figure S8. These unidentified peaks
520 >400 m/Q are both more numerous and larger during full-event periods, with a factor of two
521 difference in total peak area. The burst-event day has significantly lower concentrations of OVOCs
522 and HOMs, and to a lesser degree, their nitrogen containing counterparts (N-OVOCs and N-
523 HOMs), with significantly fewer compounds >400 m/Q. The most significant difference between
524 full and burst-event days is in the SVOCs, accounting for a factor of two difference in
525 concentration. The sulphur containing acids all have similar peak areas to the full-event day. These
526 elevations to condensable OVOCs and HOMs on particle formation days with growth are consistent

527 with particle composition data as measured by ACSM (Figure S9). Particle composition on full-
528 event days shows an elevation to organic mass concentration in the late evening and night around
529 when new particles from NPF will reach sizes detectable by the ACSM (~75 nm, Ng et al., 2011).
530 Organic mass between 16:00 – 23:00 is $3.5 \mu\text{g m}^{-3}$ on burst-event days, versus $7.8 \mu\text{g m}^{-3}$ on full-
531 event days.

532

533 **4. CONCLUSIONS**

534 We show new particle formation rates in Barcelona are linearly dependent upon the sulphuric acid
535 concentrations, and while formation rates far exceed that of $\text{H}_2\text{SO}_4\text{-BioOxOrg-H}_2\text{O}$ nucleation, they
536 fall short of those of $\text{H}_2\text{SO}_4\text{-DMA-H}_2\text{O}$ nucleation at 278 K, as does the sulphuric acid
537 dimer:monomer ratio, possibly explained by cluster evaporation due to high temperatures in
538 summertime Barcelona (303 K during events), and limited pools of gas-phase amines. These results
539 are similar to reports of nucleation rates in rural Germany (Kürten et al., 2016). As no higher-order
540 clusters were directly measured, we cannot determine nucleation mechanisms with certainty, and an
541 involvement of HOMs in nucleation is plausible.

542

543 High concentrations of OVOCs and HOMs were measured by CI-APi-ToF. Of these, the SVOC
544 arose from mostly isoprene and alkylbenzene oxidation, whereas LVOC and ELVOC arose from
545 alkylbenzene, monoterpene and PAH oxidation together, with a dependence of their concentration
546 on temperature. Concentrations of species associated with coastal and oceanic sources such as MSA
547 and HIO_3 were low. High HOM signals are seen to be a necessary condition for new particle growth
548 past 10 nm, with the most significant difference between days with and without particle growth
549 being SVOC concentrations (factor of 2 difference), while modelled growth rates from
550 condensation of these organic compounds, alongside H_2SO_4 , MSA and HIO_3 were shown to match
551 growth rates within measurement error. Thus, oxidation of traffic derived alkylbenzenes and PAHs,

552 and to a lesser degree, isoprene and monoterpene emissions is a significant determinant of new
553 particle growth in this environment.

554

555 These results are consistent with extensive chamber and flow tube studies on particle formation
556 from sulphuric acid, amines and HOMs, and further, nucleation rates relative to sulphuric acid are
557 similar to many tropospheric observations. Barcelona is representative of many Mediterranean
558 urban environments, with moderate pollution, influence of shipping emissions, and high insolation,
559 and the present study reveals the complexity of NPF mechanisms in these environments.

560

561 **DATA AVAILABILITY**

562 Data supporting this publication are openly available from the UBIRA eData repository at
563 <https://doi.org/10.25500/edata.bham.00000434>

564

565 **AUTHOR CONTRIBUTIONS**

566 RMH and XQ conceived the study, JB and DCSB carried out the CI-APi-TOF and related
567 measurements with assistance from AA and MCM. The VOC measurements were proposed by NM
568 and collected by BT-R. JB wrote the first draft of the manuscript which was enhanced by
569 contributions from the co-authors.

570

571 **COMPETING INTERESTS**

572 The authors have no conflict of interests.

573 **ACKNOWLEDGEMENTS**

574 Financial assistance from the Spanish Ministry of Science, Innovation and Universities and
575 Competitiveness and FEDER funds under the project HOUSE (CGL2016-78594-R), and by the
576 Generalitat de Catalunya (AGAUR 2017 SGR41) is gratefully acknowledged. MCM acknowledges

577 the Ramón y Cajal Fellowship awarded by the Spanish Ministry. Financial support of the UK
578 scientists by the Natural Environment Research Council through the National Centre for Atmospheric
579 Science is also acknowledged (R8/H12/83/011).

580

581 **FIGURE LEGENDS:**

582

583 **Figure 1:** Average SMPS contour plots for (a) non-event days, (b) full-event days and (c) burst-
584 event days.

585

586 **Figure 2:** Box plots for days of non-event, full-event and burst-event, showing (a) condensation
587 sink, (b) temperature, and (c) global radiation from hourly data. “Full-event” and “burst-
588 event” include data across the entire day.

589

590 **Figure 3:** Box plots for days of non-event, full-event and burst-event, showing (a) sulphuric acid,
591 (b) C₂ and C₄ amines, as clustered with the nitrate dimer and trimer, and (c) summed
592 HOM concentration from C₅₊ from hourly data. Units for ammonia + amines are
593 normalised counts, as no calibration was performed. Event days include data across the
594 full event day.

595

596 **Figure 4:** Formation rate (J_{1.9}) plotted against sulphuric acid monomer concentration, coloured by
597 condensation sink. Circles represent burst-events, squares represent full events. Data is
598 for hourly averages across NPF periods, typically within the hours 08:00 – 16:00. Slope
599 of the line = $4.9 \cdot 10^{-5} \text{ s}^{-1}$. Error bars represent systematic uncertainties on [H₂SO₄] and
600 J_{1.9}.

601

602 **Figure 5:** Formation rate plotted against sulphuric acid monomer concentration for data collected
603 from Barcelona. Tan circles represent burst-events, purple squares represent full events.
604 as well as that for the H₂SO₄-H₂O (blue inverted triangles), H₂SO₄-NH₃-H₂O (yellow
605 inverted triangles), H₂SO₄-DMA-H₂O (pink triangles), and H₂SO₄-BioOxOrg-H₂O
606 (brown diamonds) systems from the CLOUD chamber (Kürten et al., 2018 Kirkby et al.,
607 2011; Riccobono et al., 2014). CLOUD chamber experiments were performed at 278 K
608 and 38 – 39 % RH. Data is for hourly averages across NPF periods, typically within the
609 hours 08:00 – 16:00. Error bars represent systematic uncertainties on [H₂SO₄] and J_{1.9}.

610

611 **Figure 6:** Sulphuric acid dimer concentration plotted against monomer concentration, showing
612 burst-event periods (tan circles), full event periods (purple squares), non-event periods
613 (green inverted triangles), and the ratio of sulphuric acid dimer:monomer in the CLOUD
614 chamber for the H₂SO₄-H₂O-DMA system (pink triangles) (Almeida et al., 2013).
615 Dashed line represents the dimer concentration produced by ion induced clustering in the
616 chemical ionization unit (Zhao et al., 2010). CLOUD chamber experiments were
617 performed at 278 K and 38 – 39 % RH. Data is for hourly averages across NPF periods,
618 typically within the hours 08:00 – 16:00. Error bars represent systematic uncertainties on
619 [H₂SO₄] and [(H₂SO₄)₂].

620

621 **Figure 7:** Influencing factors on HOM concentration, showing (a) C₅₋₁₀ HOM concentration plotted
622 against temperature, coloured by global radiation. Ellipsis shows 95% confidence on a
623 multivariate t-distribution. (b) HOM concentration by carbon number plotted against
624 parent VOC mixing ratio. These are segregated by carbon number/VOC, i.e. C₇ HOMs
625 plotted against toluene, under the assumption that toluene oxidation is the main producer
626 of C₇ HOMs. Time for both plots is of hourly time resolution.

627

628 **Figure 8:** Concentrations of all oxygenated organic molecules and HOMs binned to integer
629 Log₁₀(C*) values, coloured by DBE.

630

631 **Figure 9:** Mass defect plots for (a) non-event, (b) full-event, and (c) burst-event periods, data
632 taken from 10:00 – 15:00 on the days 11/07/2018, 16/07/2018 and 15/07/2018
633 respectively. Size corresponds to mass spectral peak area. Ions are coloured according
634 to identified chemical composition. *Blue* points correspond to HOMs containing all
635 organic species with ≥ 5 carbon atoms and ≥ 6 oxygen atoms, and an O:C ratio of >0.6 .
636 *Purple* points correspond to the same but for species containing 1-2 nitrogen atoms.
637 Species not meeting this HOM criteria were classed generally as OVOCs which are
638 coloured *brown*, with the nitrogen containing OVOCs coloured *orange*. Sulphur acids
639 (*red*) include ions HSO_4^- , CH_3SO_3^- and SO_5^- , as well as the sulphuric acid dimer. Iodine
640 acids (*green*) contains both IO_3^- and I^- (the latter presumably deprotonated hydrogen
641 iodide). Unidentified points are left uncoloured.
642

643

644 **REFERENCES**

645

646 Ábalos, M., Bayona, J. M. and Ventura, F.: Development of a solid-phase microextraction GC-NPD
647 procedure for the determination of free volatile amines in wastewater and sewage-polluted waters,
648 *Anal. Chem.*, 71(16), 3531–3537, doi:10.1021/ac990197h, 1999.

649

650

651 Almeida, J., Schobesberger, S., Kürten, A., Ortega, I. K., Kupiainen-Määttä, O., Praplan, A. P.,
652 Adamov, A., Amorim, A., Bianchi, F., Breitenlechner, M., David, A., Dommen, J., Donahue, N. M.,
653 Downard, A., Dunne, E., Duplissy, J., Ehrhart, S., Flagan, R. C., Franchin, A., Guida, R., Hakala, J.,
654 Hansel, A., Heinritzi, M., Henschel, H., Jokinen, T., Junninen, H., Kajos, M., Kangasluoma, J.,
655 Keskinen, H., Kupc, A., Kurtén, T., Kvashin, A. N., Laaksonen, A., Lehtipalo, K., Leiminger, M.,
656 Leppä, J., Loukonen, V., Makhmutov, V., Mathot, S., McGrath, M. J., Nieminen, T., Olenius, T.,
657 Onnela, A., Petäjä, T., Riccobono, F., Riipinen, I., Rissanen, M., Rondo, L., Ruuskanen, T., Santos,
658 F. D., Sarnela, N., Schallhart, S., Schnitzhofer, R., Seinfeld, J. H., Simon, M., Sipilä, M., Stozhkov,
659 Y., Stratmann, F., Tomé, A., Tröstl, J., Tsagkogeorgas, G., Vaattovaara, P., Viisanen, Y., Virtanen,
660 A., Vrtala, A., Wagner, P. E., Weingartner, E., Wex, H., Williamson, C., Wimmer, D., Ye, P., Yli-
661 Juuti, T., Carslaw, K. S., Kulmala, M., Curtius, J., Baltensperger, U., Worsnop, D. R., Vehkamäki,
662 H. and Kirkby, J.: Molecular understanding of sulphuric acid-amine particle nucleation in the
663 atmosphere, *Nature*, 502(7471), 359–363, doi:10.1038/nature12663, 2013.

664

665 Bianchi, F., Tröstl, J., Junninen, H., Frege, C., Henne, S., Hoyle, C. R., Molteni, U., Herrmann, E.,
666 Bukowiecki, N., Chen, X., Duplissy, J., Gysel, M., Hutterli, M., Kangasluoma, J., Kontkanen, J.,
667 Manninen, H. E., Münch, S., Peräkylä, O., Petäjä, T., Rondo, L., Williamson, C., Weingartner, E.,
668 Worsnop, D. R., Kulmala, M., Dommen, J. and Baltensperger, U.: New particle formation in the free
669 troposphere : A question of chemistry and timing, *Science* (80-), 5456(May), 1–11, 2016.

670

671 Bianchi, F., Garmash, O., He, X., Yan, C., Iyer, S., Rosendahl, I., Xu, Z., Rissanen, M. P., Riva, M.,
672 Taipale, R., Sarnela, N., Petäjä, T., Worsnop, D. R., Kulmala, M., Ehn, M. and Junninen, H.: The role
673 of highly oxygenated molecules (HOMs) in determining the composition of ambient ions in the boreal
674 forest, *Atmos. Chem. Phys.*, 17(22), 13819–13831, doi:10.5194/acp-17-13819-2017, 2017.

675

676 Bianchi, F., Kurtén, T., Riva, M., Mohr, C., Rissanen, M. P., Roldin, P., Berndt, T., Crouse, J. D.,
677 Wennberg, P. O., Mentel, T. F., Wildt, J., Junninen, H., Jokinen, T., Kulmala, M., Worsnop, D. R.,
678 Thornton, J. A., Donahue, N., Kjaergaard, H. G. and Ehn, M.: Highly Oxygenated Organic Molecules
679 (HOM) from Gas-Phase Autoxidation Involving Peroxy Radicals: A Key Contributor to Atmospheric
680 Aerosol, *Chem. Rev.*, 2019.

681

682 Bousiotis, D., Dall'osto, M., Beddows, D. C. S., Pope, F. D. and Harrison, R. M.: Analysis of new
683 particle formation (NPF) events at nearby rural, urban background and urban roadside sites, *Atmos.*
684 *Chem. Phys.*, 19, 5679–5694, doi:10.5194/acp-19-5679-2019, 2019.

685

686 Brean, J., Harrison, R. M., Shi, Z., Beddows, D. C. S., W Joe, A. F. and Nicholas Hewitt, C.:
687 Observations of highly oxidised molecules and particle nucleation in the atmosphere of Beijing,
688 *Atmos. Chem. Phys. Discuss.*, (March), 1–35, doi:10.5194/acp-2019-156, 2019.

689

690 Brines, M., Dall'Osto, M., Beddows, D. C. S., Harrison, R. M., and Querol, X.: Simplifying aerosol
691 size distributions modes simultaneously detected at four monitoring sites during SAPUSS, *Atmos.*
692 *Chem. Phys.*, 14, 2973–2986, <https://doi.org/10.5194/acp-14-2973-2014>, 2014.

693

694 Brines, M., Dall'Osto, M., Beddows, D. C. S., Harrison, R. M., Gómez-Moreno, F., Núñez, L.,
695 Artíñano, B., Costabile, F., Gobbi, G. P., Salimi, F., Morawska, L., Sioutas, C., and Querol, X.:

696 Traffic and nucleation events as main sources of ultrafine particles in high-insolation developed world
697 cities, *Atmos. Chem. Phys.*, 15, 5929–5945, <https://doi.org/10.5194/acp-15-5929-2015>, 2015.

698

699 Cadle, S. H. and Mulawa, P. A.: Low-molecular-weight aliphatic amines in exhaust from catalyst-
700 equipped cars, *Environ. Sci. Technol.*, 14(6), 718–723, doi:10.1021/es60166a011, 1980.

701

702 Carnerero, C., Pérez, N., Petäjä, T., Laurila, T. M., Ahonen, L. R., Kontkanen, J., Ahn, K. H.,
703 Alastuey, A. and Querol, X.: Relating high ozone, ultrafine particles, and new particle formation
704 episodes using cluster analysis, *Atmos. Environ. X*, 4(May), 1–20, doi:10.1016/j.aeaoa.2019.100051,
705 2019.

706

707 Cohen, A. J., Brauer, M., Burnett, R., Anderson, H. R., Frostad, J., Estep, K., Balakrishnan, K.,
708 Brunekreef, B., Dandona, L., Dandona, R., Feigin, V., Freedman, G., Hubbell, B., Jobling, A., Kan,
709 H., Knibbs, L., Liu, Y., Martin, R., Morawska, L., Pope, C. A., Shin, H., Straif, K., Shaddick, G.,
710 Thomas, M., van Dingenen, R., van Donkelaar, A., Vos, T., Murray, C. J. L. and Forouzanfar, M. H.:
711 Estimates and 25-year trends of the global burden of disease attributable to ambient air pollution: an
712 analysis of data from the Global Burden of Diseases Study 2015, *Lancet*, 389(10082), 1907–1918,
713 doi:10.1016/S0140-6736(17)30505-6, 2017.

714

715 Cubison, M. J. and Jimenez, J. L.: Statistical precision of the intensities retrieved from constrained
716 fitting of overlapping peaks in high-resolution mass spectra, *Atmos. Meas. Tech.*, 8(6), 2333–2345,
717 doi:10.5194/amt-8-2333-2015, 2015.

718

719 Curtius, J., Froyd, K. D. and Lovejoy, E. R.: Cluster ion thermal decomposition (I): Experimental
720 kinetics study and ab initio calculations for $\text{HSO}_4^-(\text{H}_2\text{SO}_4)_x(\text{HNO}_3)_y$, *J. Phys. Chem. A*, 105(48),
721 10867–10873, doi:10.1021/jp0124950, 2001.

722

723 Dal Maso, M., Kulmala, M., Riipinen, I., Wagner, R., Hussein, T., Aalto, P. P. and Lehtinen, K. E.
724 J.: Formation and growth of fresh atmospheric aerosols: Eight years of aerosol size distribution data
725 from SMEAR II, Hyytiälä, Finland, *Boreal Environ. Res.*, 10(5), 323–336, 2005.

726

727 Dall’Osto, M., Querol, X., Alastuey, A., O’Dowd, C., Harrison, R. M., Wenger, J. and Gómez-
728 Moreno, F. J.: On the spatial distribution and evolution of ultrafine particles in Barcelona, *Atmos.*
729 *Chem. Phys.*, 13(2), 741–759, doi:10.5194/acp-13-741-2013, 2013.

730

731 Donahue, N. M., Epstein, S. A., Pandis, S. N. and Robinson, A. L.: Atmospheric Chemistry and
732 Physics A two-dimensional volatility basis set: 1. organic-aerosol mixing thermodynamics, *Atmos.*
733 *Chem. Phys.*, 11, 3303–3318, doi:10.5194/acp-11-3303-2011, 2011.

734 Ehn, M., Thornton, J. A., Kleist, E., Sipilä, M., Junninen, H., Pullinen, I., Springer, M., Rubach, F.,
735 Tillmann, R., Lee, B., Lopez-Hilfiker, F., Andres, S., Acir, I.-H., Rissanen, M., Jokinen, T.,
736 Schobesberger, S., Kangasluoma, J., Kontkanen, J., Nieminen, T., Kurtén, T., Nielsen, L. B.,
737 Jørgensen, S., Kjaergaard, H. G., Canagaratna, M., Maso, M. D., Berndt, T., Petäjä, T., Wahner, A.,
738 Kerminen, V.-M., Kulmala, M., Worsnop, D. R., Wildt, J. and Mentel, T. F.: A large source of low-
739 volatility secondary organic aerosol, *Nature*, 506(7489), 476–479, doi:10.1038/nature13032, 2014.

740

741 Elm, J., Myllys, N. and Kurtén, T.: What is Required for Highly Oxidized Molecules to Form Clusters
742 with Sulphuric Acid?, *J. Phys. Chem. A*, 121(23), 4578–4587, doi:10.1021/acs.jpca.7b03759, 2017.

743

744 Frege, C., Ortega, I. K., Rissanen, M. P., Praplan, A. P., Steiner, G., Heinritzi, M., Ahonen, L.,
745 Amorim, A., Bernhammer, A. K., Bianchi, F., Brilke, S., Breitenlechner, M., Dada, L., Dias, A.,
746 Duplissy, J., Ehrhart, S., El-Haddad, I., Fischer, L., Fuchs, C., Garmash, O., Gonin, M., Hansel, A.,
747 Hoyle, C. R., Jokinen, T., Junninen, H., Kirkby, J., Kürten, A., Lehtipalo, K., Leiminger, M., Lee

748 Mauldin, R., Molteni, U., Nichman, L., Petäjä, T., Sarnela, N., Schobesberger, S., Simon, M., Sipilä,
749 M., Stolzenburg, D., Tomé, A., Vogel, A. L., Wagner, A. C., Wagner, R., Xiao, M., Yan, C., Ye, P.,
750 Curtius, J., Donahue, N. M., Flagan, R. C., Kulmala, M., Worsnop, D. R., Winkler, P., Dommen, J.
751 and Baltensperger, U.: Influence of temperature on the molecular composition of ions and charged
752 clusters during pure biogenic nucleation, *Atmos. Chem. Phys.*, 18(1), 65–79, doi:10.5194/acp-18-65-
753 2018, 2018.

754

755 Garmash, O., Rissanen, M. P., Pullinen, I., Schmitt, S., Kausiala, O., Tillmann, R., Zhao, D., Percival,
756 C., Bannan, T. J., Priestley, M., Hallquist, Å. M., Kleist, E., Kiendler-Scharr, A., Hallquist, M.,
757 Berndt, T., McFiggans, G., Wildt, J., Mentel, T. F., and Ehn, M.: Multi-generation OH oxidation as
758 a source for highly oxygenated organic molecules from aromatics, *Atmos. Chem. Phys.*, 20, 515–
759 537, <https://doi.org/10.5194/acp-20-515-2020>, 2020.

760

761 Ge, X., Wexler, A. S. and Clegg, S. L.: Atmospheric amines - Part I. A review, *Atmos. Environ.*,
762 45(3), 524–546, doi:10.1016/j.atmosenv.2010.10.012, 2011a.

763

764 Ge, X., Wexler, A. S. and Clegg, S. L.: Atmospheric amines - Part II. Thermodynamic properties and
765 gas/particle partitioning, *Atmos. Environ.*, 45(3), 561–577, doi:10.1016/j.atmosenv.2010.10.013, 2011b.

766

767 Glasoe, W. A., Volz, K., Panta, B., Freshour, N., Bachman, R., Hanson, D. R., McMurry, P. H. and
768 Jen, C.: Sulphuric acid nucleation: An experimental study of the effect of seven bases, *J. Geophys.*
769 *Res. Atmos.*, 175(120), 1933–1950, doi:10.1038/175238c0, 2015.

770

771 Graus, M., Müller, M. and Hansel, A.: High resolution PTR-TOF: Quantification and Formula
772 Confirmation of VOC in Real Time, *J. Am. Soc. Mass Spectrom.*, 21(6), 1037–1044,
773 doi:10.1016/j.jasms.2010.02.006, 2010.

774

775 Guo, S., Hu, M., Zamora, M. L., Peng, J., Shang, D., Zheng, J., Du, Z., Wu, Z., Shao, M., Zeng, L.,
776 Molina, M. J. and Zhang, R.: Elucidating severe urban haze formation in China., *Proc. Natl. Acad.*
777 *Sci. U. S. A.*, 111(49), 17373–8, doi:10.1073/pnas.1419604111, 2014.

778

779 Harrison, R. M., Rob Mackenzie, A., Xu, H., Alam, M. S., Nikolova, I., Zhong, J., Singh, A., Zeraati-
780 Rezaei, S., Stark, C., Beddows, D. C. S., Liang, Z., Xu, R. and Cai, X.: Diesel exhaust nanoparticles
781 and their behaviour in the atmosphere, *Proc. R. Soc. A Math. Phys. Eng. Sci.*, 474(2220),
782 doi:10.1098/rspa.2018.0492, 2018.

783

784 Henschel, S., Querol, X., Atkinson, R., Pandolfi, M., Zeka, A., Tertre, A. L., Analistis, A.,
785 Katsouyanni, K., Chanel, O., Pascal, M., Bouland, C., Haluza, D., Medina, S., and Goodman, P. G.:
786 Ambient air SO₂ patterns in 6 European cities, *Atmos. Environ.*, 79, 236–247,
787 <https://doi.org/10.1016/j.atmosenv.2013.06.008>, 2013.

788

789 Hirsikko, A., Nieminen, T., Gagné, S., Lehtipalo, K., Manninen, H. E., Ehn, M., Hörrak, U.,
790 Kerminen, V.-M., Laakso, L., McMurry, P. H., Mirme, A., Mirme, S., Petäjä, T., Tammet, H.,
791 Vakkari, V., Vana, M. and Kulmala, M.: Atmospheric ions and nucleation: a review of observations,
792 *Atmos. Chem. Phys.*, 11, 767–798, doi:10.5194/acp-11-767-2011, 2011.

793

794 Hutchinson, G. L., Mosier, A. R. and Andre, C. E.: Ammonia and Amine Emissions from a Large
795 Cattle Feedlot, *J. Environ. Qual.*, 11(2), 288–293, 1982.

796

797 Hyttinen, N., Kupiainen-Määttä, O., Rissanen, M. P., Muuronen, M., Ehn, M. and Kurtén, T.:
798 Modeling the Charging of Highly Oxidized Cyclohexene Ozonolysis Products Using Nitrate-Based

799 Chemical Ionization, *J. Phys. Chem. A*, 119(24), 6339–6345, doi:10.1021/acs.jpca.5b01818, 2015.
800

801 IPCC, 2013: Climate Change 2013: The Physical Science Basis. Contribution of Working Group I to
802 the Fifth Assessment Report of the Intergovernmental Panel on Climate Change, edited by V. B. and
803 P. M. M. Stocker, T.F., D. Qin, G.-K. Plattner, M. Tignor, S.K. Allen, J. Boschung, A. Nauels, Y.
804 Xia, Cambridge University Press, Cambridge., 2014.
805

806 Jen, C. N., Zhao, J., McMurry, P. H., and Hanson, D. R.: Chemical ionization of clusters formed from
807 sulfuric acid and dimethylamine or diamines, *Atmos. Chem. Phys.*, 16, 12513–12529,
808 <https://doi.org/10.5194/acp-16-12513-2016>, 2016.
809

810 Jokinen, T., Sipilä, M., Junninen, H., Ehn, M., Lönn, G., Hakala, J., Petäjä, T., Mauldin, R. L.,
811 Kulmala, M. and Worsnop, D. R.: Atmospheric sulphuric acid and neutral cluster measurements using
812 CI-API-TOF, *Atmos. Chem. Phys.*, 12(9), 4117–4125, doi:10.5194/acp-12-4117-2012, 2012.
813

814 Jokinen, T., Sipilä, M., Kontkanen, J., Vakkari, V., Tisler, P., Duplissy, E.-M., Junninen, H.,
815 Kangasluoma, J., Manninen, H. E., Petäjä, T., Kulmala, M., Worsnop, D. R., Kirkby, J., Virkkula, A.
816 and Kerminen, V.-M.: Ion-induced sulphuric acid–ammonia nucleation drives particle formation in
817 coastal Antarctica, *Sci. Adv.*, 4(11), eaat9744, doi:10.1126/sciadv.aat9744, 2018.
818

819 Junninen, H., Ehn, M., Petäjä, Luosujärvi, L., Kotiaho, T., Kostianen, R., Rohner, U., Gonin, M.,
820 Fuhrer, K., Kulmala, M. and Worsnop, D. R.: A high-resolution mass spectrometer to measure
821 atmospheric ion composition, *Atmos. Meas. Tech.*, 3(4), 1039–1053, doi:10.5194/amt-3-1039-2010,
822 2010.
823

824 Kerminen, V. M., Chen, X., Vakkari, V., Petäjä, T., Kulmala, M. and Bianchi, F.: Atmospheric new
825 particle formation and growth: Review of field observations, *Environ. Res. Lett.*, 13(10),
826 doi:10.1088/1748-9326/aadf3c, 2018.
827

828 Kirkby, J., Curtius, J., Almeida, J., Dunne, E., Duplissy, J., Ehrhart, S., Franchin, A., Gagné, S., Ickes,
829 L., Kürten, A., Kupc, A., Metzger, A., Riccobono, F., Rondo, L., Schobesberger, S., Tsagkogeorgas,
830 G., Wimmer, D., Amorim, A., Bianchi, F., Breitenlechner, M., David, A., Dommen, J., Downard, A.,
831 Ehn, M., Flagan, R. C., Haider, S., Hansel, A., Hauser, D., Jud, W., Junninen, H., Kreissl, F., Kvashin,
832 A., Laaksonen, A., Lehtipalo, K., Lima, J., Lovejoy, E. R., Makhmutov, V., Mathot, S., Mikkilä, J.,
833 Minginette, P., Mogo, S., Nieminen, T., Onnela, A., Pereira, P., Petäjä, T., Schnitzhofer, R., Seinfeld,
834 J. H., Sipilä, M., Stozhkov, Y., Stratmann, F., Tomé, A., Vanhanen, J., Viisanen, Y., Vrtala, A.,
835 Wagner, P. E., Walther, H., Weingartner, E., Wex, H., Winkler, P. M., Carslaw, K. S., Worsnop, D.
836 R., Baltensperger, U. and Kulmala, M.: Role of sulphuric acid, ammonia and galactic cosmic rays in
837 atmospheric aerosol nucleation, *Nature*, 476(7361), 429–435, doi:10.1038/nature10343, 2011.
838

839 Kirkby, J., Duplissy, J., Sengupta, K., Frege, C., Gordon, H., Williamson, C., Heinritzi, M., Simon,
840 M., Yan, C., Almeida, J., Trostl, J., Nieminen, T., Ortega, I. K., Wagner, R., Adamov, A., Amorim,
841 A., Bernhammer, A. K., Bianchi, F., Breitenlechner, M., Brilke, S., Chen, X., Craven, J., Dias, A.,
842 Ehrhart, S., Flagan, R. C., Franchin, A., Fuchs, C., Guida, R., Hakala, J., Hoyle, C. R., Jokinen, T.,
843 Junninen, H., Kangasluoma, J., Kim, J., Krapf, M., Kurten, A., Laaksonen, A., Lehtipalo, K.,
844 Makhmutov, V., Mathot, S., Molteni, U., Onnela, A., Perakyla, O., Piel, F., Petaja, T., Praplan, A. P.,
845 Pringle, K., Rap, A., Richards, N. A. D., Riipinen, I., Rissanen, M. P., Rondo, L., Sarnela, N.,
846 Schobesberger, S., Scott, C. E., Seinfeld, J. H., Sipilä, M., Steiner, G., Stozhkov, Y., Stratmann, F.,
847 Tomé, A., Virtanen, A., Vogel, A. L., Wagner, A. C., Wagner, P. E., Weingartner, E., Wimmer, D.,
848 Winkler, P. M., Ye, P., Zhang, X., Hansel, A., Dommen, J., Donahue, N. M., Worsnop, D. R.,
849 Baltensperger, U., Kulmala, M., Carslaw, K. S. and Curtius, J.: Ion-induced nucleation of pure
850 biogenic particles, *Nature*, 533(7604), 521–526, doi:10.1038/nature17953, 2016.

851
852 Kuang, C., McMurry, P. H., McCormick, A. V. and Eisele, F. L.: Dependence of nucleation rates on
853 sulfuric acid vapor concentration in diverse atmospheric locations, *J. Geophys. Res. Atmos.*, 113(10),
854 1–9, doi:10.1029/2007JD009253, 2008.

855
856 Kulmala, M., Dal Maso, M., Mäkelä, J. M., Pirjola, L., Väkevä, M., Aalto, P., Miiikkulainen, P.,
857 Hämeri, K. and O’Dowd, C. D.: On the formation, growth and composition of nucleation mode
858 particles, *Tellus, Ser. B Chem. Phys. Meteorol.*, 53(4), 479–490, doi:10.1034/j.1600-0889.2001.d01-
859 33.x, 2001.

860
861 Kulmala, M., Petäjä, T., Nieminen, T., Sipilä, M., Manninen, H. E., Lehtipalo, K., Dal Maso, M.,
862 Aalto, P. P., Junninen, H., Paasonen, P., Riipinen, I., Lehtinen, K. E. J., Laaksonen, A. and Kerminen,
863 V.-M.: Measurement of the nucleation of atmospheric aerosol particles, *Nat. Protoc.*, 7(9), 1651–
864 1667, doi:10.1038/nprot.2012.091, 2012.

865
866 Kulmala, M., Kerminen, V. M., Petäjä, T., Ding, A. J. and Wang, L.: Atmospheric gas-to-particle
867 conversion: Why NPF events are observed in megacities?, *Faraday Discuss.*, 200, 271–288,
868 doi:10.1039/c6fd00257a, 2017.

869
870 Kürten, A., Jokinen, T., Simon, M., Sipilä, M., Sarnela, N., Junninen, H., Adamov, A., Almeida, J.,
871 Amorim, A., Bianchi, F., Breitenlechner, M., Dommen, J., Donahue, N. M., Duplissy, J., Ehrhart, S.,
872 Flagan, R. C., Franchin, A., Hakala, J., Hansel, A., Heinritzi, M., Hutterli, M., Kangasluoma, J.,
873 Kirkby, J., Laaksonen, A., Lehtipalo, K., Leiminger, M., Makhmutov, V., Mathot, S., Onnela, A.,
874 Petäjä, T., Praplan, A. P., Riccobono, F., Rissanen, M. P., Rondo, L., Schobesberger, S., Seinfeld, J.
875 H., Steiner, G., Tomé, A., Tröstl, J., Winkler, P. M., Williamson, C., Wimmer, D., Ye, P.,
876 Baltensperger, U., Carslaw, K. S., Kulmala, M., Worsnop, D. R. and Curtius, J.: Neutral molecular
877 cluster formation of sulfuric acid–dimethylamine observed in real time under atmospheric conditions,
878 *Proc. Natl. Acad. Sci.*, 111(42), 15019–15024, doi:10.1073/pnas.1404853111, 2014.

879
880 Kürten, A., Münch, S., Rondo, L., Bianchi, F., Duplissy, J., Jokinen, T., Junninen, H., Sarnela, N.,
881 Schobesberger, S., Simon, M., Sipilä, M., Almeida, J., Amorim, A., Dommen, J., Donahue, N. M.,
882 Dunne, E. M., Flagan, R. C., Franchin, A., Kirkby, J., Kupc, A., Makhmutov, V., Petäjä, T., Praplan,
883 A. P., Riccobono, F., Steiner, G., Tomé, A., Tsagkogeorgas, G., Wagner, P. E., Wimmer, D.,
884 Baltensperger, U., Kulmala, M., Worsnop, D. R. and Curtius, J.: Thermodynamics of the formation
885 of sulfuric acid dimers in the binary (H₂SO₄-H₂O) and ternary (H₂SO₄-H₂O-NH₃) system, *Atmos.*
886 *Chem. Phys.*, 15(18), 10701–10721, doi:10.5194/acp-15-10701-2015, 2015.

887
888 Kürten, A., Bergen, A., Heinritzi, M., Leiminger, M., Lorenz, V., Piel, F., Simon, M., Sitals, R.,
889 Wagner, A. C. and Curtius, J.: Observation of new particle formation and measurement of sulphuric
890 acid, ammonia, amines and highly oxidized organic molecules at a rural site in central Germany,
891 *Atmos. Chem. Phys.*, 16(19), 12793–12813, doi:10.5194/acp-16-12793-2016, 2016.

892
893 Kürten, A., Li, C., Bianchi, F., Curtius, J., Dias, A., Donahue, N. M., Duplissy, J., Flagan, R. C.,
894 Hakala, J., Jokinen, T., Kirkby, J., Kulmala, M., Laaksonen, A., Lehtipalo, K., Makhmutov, V.,
895 Onnela, A., Rissanen, M. P., Simon, M., Sipilä, M., Stozhkov, Y., Tröstl, J., Ye, P. and McMurry, P.
896 H.: New particle formation in the sulfuric acid-dimethylamine-water system: Reevaluation of
897 CLOUD chamber measurements and comparison to an aerosol nucleation and growth model, *Atmos.*
898 *Chem. Phys.*, 18(2), 845–863, doi:10.5194/acp-18-845-2018, 2018.

899
900 Kurtén, T., Noppel, M., Vehkamäki, H., Salonen, M. and Kulmala, M.: Quantum chemical studies of
901 hydrate formation of H₂SO₄ and HSO₄⁻, *Boreal Environ. Res.*, 12(3), 431–453, 2007.

902
903 Kurtén, T., Loukonen, V., Vehkamäki, H. and Kulmala, M.: Amines are likely to enhance neutral and
904 ion-induced sulphuric acid-water nucleation in the atmosphere more effectively than ammonia,
905 *Atmos. Chem. Phys.*, 8(14), 4095–4103, doi:10.5194/acp-8-4095-2008, 2008.
906
907
908 Lee, S. H., Gordon, H., Yu, H., Lehtipalo, K., Haley, R., Li, Y. and Zhang, R.: New Particle Formation
909 in the Atmosphere: From Molecular Clusters to Global Climate, *J. Geophys. Res. Atmos.*,
910 doi:10.1029/2018JD029356, 2019.
911
912 Lehtinen, K. E. J., Dal Maso, M., Kulmala, M. and Kerminen, V. M.: Estimating nucleation rates
913 from apparent particle formation rates and vice versa: Revised formulation of the Kerminen-Kulmala
914 equation, *J. Aerosol Sci.*, 38(9), 988–994, doi:10.1016/j.jaerosci.2007.06.009, 2007.
915 Lehtipalo, K., Yan, C., Dada, L., Bianchi, F., Xiao, M., Wagner, R., Stolzenburg, D., Ahonen, L. R.,
916 Amorim, A., Baccarini, A., Bauer, P. S., Baumgartner, B., Bergen, A., Bernhammer, A.-K.,
917 Breitenlechner, M., Brilke, S., Buchholz, A., Mazon, S. B., Chen, D., Chen, X., Dias, A., Dommen,
918 J., Draper, D. C., Duplissy, J., Ehn, M., Finkenzeller, H., Fischer, L., Frege, C., Fuchs, C., Garmash,
919 O., Gordon, H., Hakala, J., He, X., Heikkinen, L., Heinritzi, M., Helm, J. C., Hofbauer, V., Hoyle, C.
920 R., Jokinen, T., Kangasluoma, J., Kerminen, V.-M., Kim, C., Kirkby, J., Kontkanen, J., Kürten, A.,
921 Lawler, M. J., Mai, H., Mathot, S., Mauldin, R. L., Molteni, U., Nichman, L., Nie, W., Nieminen, T.,
922 Ojdanic, A., Onnela, A., Passananti, M., Petäjä, T., Piel, F., Pospisilova, V., Quéléver, L. L. J.,
923 Rissanen, M. P., Rose, C., Sarnela, N., Schallhart, S., Schuchmann, S., Sengupta, K., Simon, M.,
924 Sipilä, M., Tauber, C., Tomé, A., Tröstl, J., Väisänen, O., Vogel, A. L., Volkamer, R., Wagner, A.
925 C., Wang, M., Weitz, L., Wimmer, D., Ye, P., Ylisirniö, A., Zha, Q., Carslaw, K. S., Curtius, J.,
926 Donahue, N. M., Flagan, R. C., Hansel, A., Riipinen, I., Virtanen, A., Winkler, P. M., Baltensperger,
927 U., Kulmala, M. and Worsnop, D. R.: Multicomponent new particle formation from sulphuric acid,
928 ammonia, and biogenic vapors, *Sci. Adv.*, 4(12), eaau5363, doi:10.1126/sciadv.aau5363, 2018.
929 Massoli, P., Stark, H., Canagaratna, M. R., Krechmer, J. E., Xu, L., Ng, N. L., Mauldin, R. L., Yan,
930 C., Kimmel, J., Misztal, P. K., Jimenez, J. L., Jayne, J. T. and Worsnop, D. R.: Ambient
931 Measurements of Highly Oxidized Gas-Phase Molecules during the Southern Oxidant and Aerosol
932 Study (SOAS) 2013, *ACS Earth Sp. Chem.*, 2(7), 653–672,
933 doi:10.1021/acsearthspacechem.8b00028, 2018.
934
935 Mikkonen, S., Romakkaniemi, S., Smith, J. N., Korhonen, H., Petäjä, T., Plass-Duelmer, C., Boy, M.,
936 McMurry, P. H., Lehtinen, K. E. J., Joutsensaari, J., Hamed, A., Mauldin, R. L., Birmili, W., Spindler,
937 G., Arnold, F., Kulmala, M. and Laaksonen, A.: A statistical proxy for sulphuric acid concentration,
938 *Atmos. Chem. Phys.*, 11(21), 11319–11334, doi:10.5194/acp-11-11319-2011, 2011.
939
940 Millán, M.: Extreme hydrometeorological events and climate change predictions in Europe, *J.*
941 *Hydrol.*, 518(PB), 206–224, doi:10.1016/j.jhydrol.2013.12.041, 2014.
942
943 Miller, M. R., Raftis, J. B., Langrish, J. P., McLean, S. G., Samutrtai, P., Connell, S. P., Wilson, S.,
944 Vesey, A. T., Fokkens, P. H. B., Boere, A. J. F., Krystek, P., Campbell, C. J., Hadoke, P. W. F.,
945 Donaldson, K., Cassee, F. R., Newby, D. E., Duffin, R. and Mills, N. L.: Inhaled Nanoparticles
946 Accumulate at Sites of Vascular Disease, *ACS Nano*, 11(5), 4542–4552,
947 doi:10.1021/acsnano.6b08551, 2017.
948
949 Minguillón, M. C., Brines, M., Pérez, N., Reche, C., Pandol, M., Fonseca, A. S., Amato, F., Alastuey,
950 A., Lyasota, A., Codina, B., Lee, H., Eun, H., Ahn, K. and Querol, X.: New particle formation at
951 ground level and in the vertical column over the Barcelona area, , 165, 118–130,
952 doi:10.1016/j.atmosres.2015.05.003, 2015.
953

954

955 Minguillón, M.C., Pérez, N., Marchand, N., Bertrand, A., Temime-Roussel, B., Agrios, K., Szidat,
956 S., van Drooge, B.L., Sylvestre, A., Alastuey, A., Reche, C., Ripoll, A., Marco, E., Grimalt, J.O.,
957 Querol, X.: Secondary organic aerosol origin in an urban environment. Influence of biogenic and fuel
958 combustion precursors. *Faraday Discuss.*, 189, 337-359, 2016.

959

960 Mohr, C., Thornton, J. A., Heitto, A., Lopez-Hilfiker, F. D., Lutz, A., Riipinen, I., Hong, J., Donahue,
961 N. M., Hallquist, M., Petäjä, T., Kulmala, M. and Yli-Juuti, T.: Molecular identification of organic
962 vapors driving atmospheric nanoparticle growth, *Nat. Commun.*, 10(1), 1–7, doi:10.1038/s41467-
963 019-12473-2, 2019.

964

965 Møller, K. H., Tram, C. M. and Kjaergaard, H. G.: Side-by-Side Comparison of Hydroperoxide and
966 Corresponding Alcohol as Hydrogen-Bond Donors, *J. Phys. Chem. A*, 121(15), 2951–2959,
967 doi:10.1021/acs.jpca.7b01323, 2017.

968

969 Molteni, U., Bianchi, F., Klein, F., El Haddad, I., Frege, C., Rossi, M. J., Dommen, J. and
970 Baltensperger, U.: Formation of highly oxygenated organic molecules from aromatic compounds,
971 *Atmos. Chem. Phys.*, 18(3), 1909–1921, doi:10.5194/acp-18-1909-2018, 2018.

972

973 Ng, N. L., Herndon, S. C., Trimborn, A., Canagaratna, M. R., Croteau, P. L., Onasch, T. B., Sueper,
974 D., Worsnop, D. R., Zhang, Q., Sun, Y. L. and Jayne, J. T.: An Aerosol Chemical Speciation Monitor
975 (ACSM) for routine monitoring of the composition and mass concentrations of ambient aerosol,
976 *Aerosol Sci. Technol.*, 45(7), 770–784, doi:10.1080/02786826.2011.560211, 2011.

977

978 Olenius, T., Halonen, R., Kurtén, T., Henschel, H., Kupiainen-Määttä, O., Ortega, I. K., Jen, C. N.,
979 Vehkamäki, H. and Riipinen, I.: New particle formation from sulfuric acid and amines: Comparison
980 of monomethylamine, dimethylamine, and trimethylamine, *J. Geophys. Res.*, 122(13), 7103–7118,
981 doi:10.1002/2017JD026501, 2017.

982

983 Olin, M., Kuuluvainen, H., Aurela, M., Kalliokoski, J., Kuittinen, N., Isotalo, M., Timonen, H. J.,
984 Niemi, J. V., Rönkkö, T., and Dal Maso, M.: Traffic-originated nanocluster emission exceeds
985 H₂SO₄-driven photochemical new particle formation in an urban area, *Atmos. Chem. Phys.*, 20, 1–
986 13, <https://doi.org/10.5194/acp-20-1-2020>, 2020.

987

988 Ortega, I. K., Olenius, T., Kupiainen-Määttä, O., Loukonen, V., Kurtén, T., and Vehkamäki, H.:
989 Electrical charging changes the composition of sulfuric acid–ammonia/dimethylamine clusters,
990 *Atmos. Chem. Phys.*, 14, 7995–8007, <https://doi.org/10.5194/acp-14-7995-2014>, 2014.

991

992 Paasonen, P., Olenius, T., Kupiainen, O., Kurtén, T., Petäjä, T., Birmili, W., Hamed, A., Hu, M.,
993 Huey, L. G., Plass-Duelmer, C., Smith, J. N., Wiedensohler, A., Loukonen, V., McGrath, M. J.,
994 Ortega, I. K., Laaksonen, A., Vehkamäki, H. and Kulmala, M.: On the formation of sulphuric acid
995 – Amine clusters in varying atmospheric conditions and its influence on atmospheric new
996 particle formation, *Atmos. Chem. Phys.*, 12(19), 9113–9133, doi:10.5194/acp-12-9113-2012, 2012.

997

998 Pandolfi, M., Amato, F., Reche, C., Alastuey, A., Otjes, R. P., Blom, M. J. and Querol, X.: Summer
999 ammonia measurements in a densely populated Mediterranean city, *Atmos. Chem. Phys.*, 12(16),
1000 7557–7575, doi:10.5194/acp-12-7557-2012, 2012.

1001

1002 Passananti, M., Zapadinsky, E., Zanca, T., Kangasluoma, J., Mylly, N., Rissanen, M. P., Kurtén, T.,
1003 Ehn, M., Attoui, M. and Vehkamäki, H.: How well can we predict cluster fragmentation inside a mass
1004 spectrometer?, *Chem. Commun.*, 55(42), 5946–5949, doi:10.1039/c9cc02896j, 2019.

1005

1006
1007 Penner, J. E., Xu, L. and Wang, M.: Satellite methods underestimate indirect climate forcing by
1008 aerosols., *Proc. Natl. Acad. Sci. U. S. A.*, 108(33), 13404–13408, doi:10.1073/pnas.1018526108,
1009 2011.
1010
1011 Praske, E., Otkjær, R. V., Crounse, J. D., Hethcox, J. C., Stoltz, B. M., Kjaergaard, H. G. and
1012 Wennberg, P. O.: Atmospheric autoxidation is increasingly important in urban and suburban North
1013 America, *Proc. Natl. Acad. Sci.*, 115(1), 64–69, doi:10.1073/pnas.1715540115, 2018.
1014
1015 Quéléver, L. L. J., Kristensen, K., Normann Jensen, L., Rosati, B., Teiwes, R., Daellenbach, K. R.,
1016 Peräkylä, O., Roldin, P., Bossi, R., Pedersen, H. B., Glasius, M., Bilde, M. and Ehn, M.: Effect of
1017 temperature on the formation of highly oxygenated organic molecules (HOMs) from alpha-pinene
1018 ozonolysis, *Atmos. Chem. Phys.*, 19, 7609–7625, doi:10.5194/acp-19-7609-2019, 2019.
1019 Querol, X., Gangoiti, G., Mantilla, E., Alastuey, A., Minguillón, M. C., Amato, F., Reche, C., Viana,
1020 M., Moreno, T., Karanasiou, A., Rivas, I., Pérez, N., Ripoll, A., Brines, M., Ealo, M., Pandolfi, M.,
1021 Lee, H. K., Eun, H. R., Park, Y. H., Escudero, M., Beddows, D., Harrison, R. M., Bertrand, A.,
1022 Marchand, N., Lyasota, A., Codina, B., Olid, M., Udina, M., Jiménez-Esteve, B., Jiménez-Esteve, B.
1023 B., Alonso, L., Millán, M. and Ahn, K. H.: Phenomenology of high-ozone episodes in NE Spain,
1024 *Atmos. Chem. Phys.*, 17(4), 2817–2838, doi:10.5194/acp-17-2817-2017, 2017.
1025
1026 Reche, C., Viana, M., Karanasiou, A., Cusack, M., Alastuey, A., Artiñano, B., Revuelta, M. A.,
1027 López-Mahía, P., Blanco-Heras, G., Rodríguez, S., Sánchez de la Campa, A. M., Fernández-
1028 Camacho, R., González-Castanedo, Y., Mantilla, E., Tang, Y. S. and Querol, X.: Urban NH₃ levels
1029 and sources in six major Spanish cities, *Chemosphere*, 119, 769–777,
1030 doi:10.1016/j.chemosphere.2014.07.097, 2015.
1031
1032 Riccobono, F., Schobesberger, S., Scott, C., Dommen, J., Ortega, I., Rondo, L., Almeida, J., Amorim,
1033 A., Bianchi, F., Breitenlechner, M., David, A., Downard, A., Dunne, E., Duplissy, J., Ehrhart, S.,
1034 Flagan, R., Franchin, A., Hansel, A., Junninen, H., Kajos, M., Keskinen, H., Kupc, A., Kürten, A.,
1035 Kvashin, A., Laaksonen, A., Lehtipalo, K., Makhmutov, V., Mathot, S., Nieminen, T., Onnela, A.,
1036 Petäjä, T., Praplan, A., Santos, F., Schallhart, S., Seinfeld, J., Sipilä, M., Van Spracklen, D., Stozhkov,
1037 Y., Stratmann, F., Tomé, A., Tsagkogeorgas, G., Vaattovaara, P., Viisanen, Y., Vrtala, A., Wagner,
1038 P., Weingartner, E., Wex, H., Wimmer, D., Carslaw, K., Curtius, J., Donahue, N., Kirkby, J.,
1039 Kulmala, M., Worsnop, D. and Baltensperger, U.: Oxidation products of biogenic emissions
1040 contribute to nucleation of atmospheric particles., *Science (80-.)*, 344(6185), 717–721,
1041 doi:10.1126/science.1243527, 2014.
1042
1043 Rissanen, M. P.: NO₂ Suppression of Autoxidation-Inhibition of Gas-Phase Highly Oxidized Dimer
1044 Product Formation, *ACS Earth Sp. Chem.*, 2(11), 1211–1219,
1045 doi:10.1021/acsearthspacechem.8b00123, 2018.
1046
1047 Rose, C., Zha, Q., Dada, L., Yan, C., Lehtipalo, K., Junninen, H., Mazon, S. B., Jokinen, T., Sarnela,
1048 N., Sipilä, M., Petäjä, T., Kerminen, V.-M., Bianchi, F. and Kulmala, M.: Observations of biogenic
1049 ion-induced cluster formation in the atmosphere, *Sci. Adv.*, 4(4), 5218, doi:10.1126/sciadv.aar5218,
1050 2018.
1051
1052 Schervish, M. and Donahue, N. M.: Peroxy radical chemistry and the volatility basis set, *Atmos.*
1053 *Chem. Phys.*, 20(2), 1183–1199, doi:10.5194/acp-20-1183-2020, 2020.
1054
1055 Schobesberger, S., Junninen, H., Bianchi, F., Lönn, G., Ehn, M., Lehtipalo, K., Dommen, J., Ehrhart,
1056 S., Ortega, I. K., Franchin, A., Nieminen, T., Riccobono, F., Hutterli, M., Duplissy, J., Almeida, J.,
1057 Amorim, A., Breitenlechner, M., Downard, A. J., Dunne, E. M., Flagan, R. C., Kajos, M., Keskinen,

1058 H., Kirkby, J., Kupc, A., Kürten, A., Kurtén, T., Laaksonen, A., Mathot, S., Onnela, A., Praplan, A.
1059 P., Rondo, L., Santos, F. D., Schallhart, S., Schnitzhofer, R., Sipilä, M., Tomé, A., Tsagkogeorgas,
1060 G., Vehkamäki, H., Wimmer, D., Baltensperger, U., Carslaw, K. S., Curtius, J., Hansel, A., Petäjä,
1061 T., Kulmala, M., Donahue, N. M. and Worsnop, D. R.: Molecular understanding of atmospheric
1062 particle formation from sulphuric acid and large oxidized organic molecules., *Proc. Natl. Acad. Sci.*
1063 *U. S. A.*, 110(43), 17223–17228, doi:10.1073/pnas.1306973110, 2013.

1064
1065 Simon, M., Heinritzi, M., Herzog, S., Leiminger, M., Bianchi, F., Praplan, A., Dommen, J., Curtius,
1066 J. and Kurten, A.: Detection of dimethylamine in the low pptv range using nitrate chemical ionization
1067 atmospheric pressure interface time-of-flight (CI-API-TOF) mass spectrometry, *Atmos. Meas. Tech.*,
1068 9(5), 2135–2145, doi:10.5194/amt-9-2135-2016, 2016.

1069
1070 Sipilä, M., Sarnela, N., Jokinen, T., Henschel, H., Junninen, H., Kontkanen, J., Richters, S.,
1071 Kangasluoma, J., Franchin, A., Peräkylä, O., Rissanen, M. P., Ehn, M., Vehkamäki, H., Kurten, T.,
1072 Berndt, T., Petäjä, T., Worsnop, D., Ceburnis, D., Kerminen, V. M., Kulmala, M. and O’Dowd, C.:
1073 Molecular-scale evidence of aerosol particle formation via sequential addition of HIO₃, *Nature*,
1074 537(7621), 532–534, doi:10.1038/nature19314, 2016.

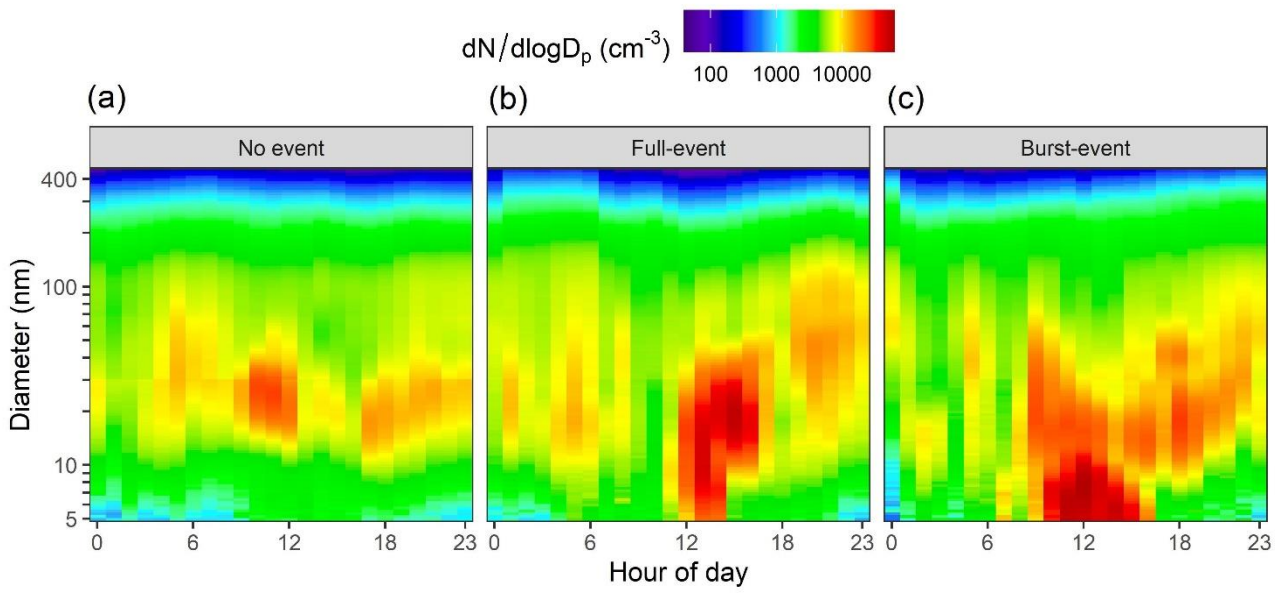
1075
1076 Stolzenburg, D., Fischer, L., Vogel, A. L., Heinritzi, M., Schervish, M., Simon, M., Wagner, A. C.,
1077 Dada, L., Ahonen, L. R., Amorim, A., Baccarini, A., Bauer, P. S., Baumgartner, B., Bergen, A.,
1078 Bianchi, F., Breitenlechner, M., Brilke, S., Buenrostro Mazon, S., Chen, D., Dias, A., Draper, D. C.,
1079 Duplissy, J., El Haddad, I., Finkenzeller, H., Frege, C., Fuchs, C., Garmash, O., Gordon, H., He, X.,
1080 Helm, J., Hofbauer, V., Hoyle, C. R., Kim, C., Kirkby, J., Kontkanen, J., Kürten, A., Lampilahti, J.,
1081 Lawler, M., Lehtipalo, K., Leiminger, M., Mai, H., Mathot, S., Mentler, B., Molteni, U., Nie, W.,
1082 Nieminen, T., Nowak, J. B., Ojdanic, A., Onnela, A., Passananti, M., Petäjä, T., Quéléver, L. L. J.,
1083 Rissanen, M. P., Sarnela, N., Schallhart, S., Tauber, C., Tomé, A., Wagner, R., Wang, M., Weitz, L.,
1084 Wimmer, D., Xiao, M., Yan, C., Ye, P., Zha, Q., Baltensperger, U., Curtius, J., Dommen, J., Flagan,
1085 R. C., Kulmala, M., Smith, J. N., Worsnop, D. R., Hansel, A., Donahue, N. M. and Winkler, P. M.:
1086 Rapid growth of organic aerosol nanoparticles over a wide tropospheric temperature range, *Proc.*
1087 *Natl. Acad. Sci.*, 201807604 [online] Available from:
1088 <http://www.pnas.org/lookup/doi/10.1073/pnas.1807604115>, 2018.

1089
1090 Tröstl, J., Chuang, W. K., Gordon, H., Heinritzi, M., Yan, C., Molteni, U., Ahlm, L., Frege, C.,
1091 Bianchi, F., Wagner, R., Simon, M., Lehtipalo, K., Williamson, C., Craven, J. S., Duplissy, J.,
1092 Adamov, A., Almeida, J., Bernhammer, A. K., Breitenlechner, M., Brilke, S., Dias, A., Ehrhart, S.,
1093 Flagan, R. C., Franchin, A., Fuchs, C., Guida, R., Gysel, M., Hansel, A., Hoyle, C. R., Jokinen, T.,
1094 Junninen, H., Kangasluoma, J., Keskinen, H., Kim, J., Krapf, M., Kürten, A., Laaksonen, A., Lawler,
1095 M., Leiminger, M., Mathot, S., Möhler, O., Nieminen, T., Onnela, A., Petäjä, T., Piel, F. M.,
1096 Miettinen, P., Rissanen, M. P., Rondo, L., Sarnela, N., Schobesberger, S., Sengupta, K., Sipilä, M.,
1097 Smith, J. N., Steiner, G., Tomé, A., Virtanen, A., Wagner, A. C., Weingartner, E., Wimmer, D.,
1098 Winkler, P. M., Ye, P., Carslaw, K. S., Curtius, J., Dommen, J., Kirkby, J., Kulmala, M., Riipinen, I.,
1099 Worsnop, D. R., Donahue, N. M. and Baltensperger, U.: The role of low-volatility organic compounds
1100 in initial particle growth in the atmosphere, *Nature*, 533(7604), 527–531, doi:10.1038/nature18271,
1101 2016.

1102
1103 Van Damme, M., Clarisse, L., Whitburn, S., Hadji-Lazaro, J., Hurtmans, D., Clerbaux, C. and
1104 Coheur, P. F.: Industrial and agricultural ammonia point sources exposed, *Nature*, 564(7734), 99–
1105 103, doi:10.1038/s41586-018-0747-1, 2018.

1106
1107 Wang, S., Wu, R., Berndt, T., Ehn, M. and Wang, L.: Formation of Highly Oxidized Radicals and
1108 Multifunctional Products from the Atmospheric Oxidation of Alkylbenzenes, *Environ. Sci. Technol.*,
1109 51(15), 8442–8449, doi:10.1021/acs.est.7b02374, 2017.

1110
1111 Yan, C.: The role of H₂SO₄-NH₃ anion clusters in ion-induced aerosol nucleation mechanisms in the
1112 boreal forest, *Atmos. Chem. Phys.*, (April), 1–20, doi:10.1029/2001JD001100, 2018.
1113
1114 Yan, C., Nie, W., Äijälä, M., Rissanen, M. P., Canagaratna, M. R., Massoli, P., Junninen, H., Jokinen,
1115 T., Sarnela, N., Häme, S. A. K., Schobesberger, S., Canonaco, F., Yao, L., Prévôt, A. S. H., Petäjä,
1116 T., Kulmala, M., Sipilä, M., Worsnop, D. R. and Ehn, M.: Source characterization of highly oxidized
1117 multifunctional compounds in a boreal forest environment using positive matrix factorization, *Atmos.*
1118 *Chem. Phys.*, 16(19), 12715–12731, doi:10.5194/acp-16-12715-2016, 2016.
1119
1120 Yan, C., Nie, W., Vogel, A. L., Dada, L., Lehtipalo, K., Stolzenburg, D., Wagner, R., Rissanen, M.
1121 P., Xiao, M., Ahonen, L., Fischer, L., Rose, C., Bianchi, F., Gordon, H., Simon, M., Heinritzi, M.,
1122 Garmash, O., Roldin, P., Dias, A., Ye, P., Hofbauer, V., Amorim, A., Bauer, P. S., Bergen, A.,
1123 Bernhammer, A.-K., Breitenlechner, M., Brilke, S., Buchholz, A., Mazon, S. B., Canagaratna, M. R.,
1124 Chen, X., Ding, A., Dommen, J., Draper, D. C., Duplissy, J., Frege, C., Heyn, C., Guida, R., Hakala,
1125 J., Heikkinen, L., Hoyle, C. R., Jokinen, T., Kangasluoma, J., Kirkby, J., Kontkanen, J., Kürten, A.,
1126 Lawler, M. J., Mai, H., Mathot, S., Mauldin, R. L., Molteni, U., Nichman, L., Nieminen, T., Nowak,
1127 J., Ojdanic, A., Onnela, A., Pajunoja, A., Petäjä, T., Piel, F., Quéléver, L. L. J., Sarnela, N., Schallhart,
1128 S., Sengupta, K., Sipilä, M., Tomé, A., Tröstl, J., Väisänen, O., Wagner, A. C., Ylisirniö, A., Zha, Q.,
1129 Baltensperger, U., Carslaw, K. S., Curtius, J., Flagan, R. C., Hansel, A., Riipinen, I., Smith, J. N.,
1130 Virtanen, A., Winkler, P. M., Donahue, N. M., Kerminen, V.-M., Kulmala, M., Ehn, M. and Worsnop,
1131 D. R.: Size-dependent influence of NO_x on the growth rates of organic aerosol particles, *Sci. Adv.*,
1132 6(22), eaay4945, doi:10.1126/sciadv.aay4945, 2020.
1133
1134 Yao, L., Garmash, O., Bianchi, F., Zheng, J., Yan, C., Kontkanen, J., Junninen, H., Mazon, S. B.,
1135 Ehn, M., Paasonen, P., Sipilä, M., Wang, M., Wang, X., Xiao, S., Chen, H., Lu, Y., Zhang, B., Wang,
1136 D., Fu, Q., Geng, F., Li, L., Wang, H., Qiao, L., Yang, X., Chen, J., Kerminen, V. M., Petäjä, T.,
1137 Worsnop, D. R., Kulmala, M. and Wang, L.: Atmospheric new particle formation from sulphuric acid
1138 and amines in a Chinese megacity, *Science* (80-.), 361(6399), 278–281,
1139 doi:10.1126/science.aao4839, 2018.
1140
1141 Yli-Juuti, T., Pajunoja, A., Tikkanen, O. P., Buchholz, A., Faiola, C., Väisänen, O., Hao, L., Kari, E.,
1142 Peräkylä, O., Garmash, O., Shiraiwa, M., Ehn, M., Lehtinen, K. and Virtanen, A.: Factors controlling
1143 the evaporation of secondary organic aerosol from α -pinene ozonolysis, *Geophys. Res. Lett.*, 44(5),
1144 2562–2570, doi:10.1002/2016GL072364, 2017.
1145
1146 Yu, H., Zhou, L., Dai, L., Shen, W., Dai, W., Zheng, J., Ma, Y. and Chen, M.: Nucleation and growth
1147 of sub-3 nm particles in the polluted urban atmosphere of a megacity in China, *Atmos. Chem. Phys.*,
1148 16(4), 2641–2657, doi:10.5194/acp-16-2641-2016, 2016.
1149
1150 Zhang, R., Khalizov, A., Wang, L., Hu, M. and Xu, W.: Nucleation and growth of nanoparticles in
1151 the atmosphere, *Chem. Rev.*, 112(3), 1957–2011, doi:10.1021/cr2001756, 2012.
1152
1153 Zhao, J., Eisele, F. L., Titcombe, M., Kuang, C. and McMurry, P. H.: Chemical ionization mass
1154 spectrometric measurements of atmospheric neutral clusters using the cluster-CIMS, *J. Geophys.*
1155 *Res.*, 115(D8), 1–19, doi:10.1029/2009jd012606, 2010.
1156
1157

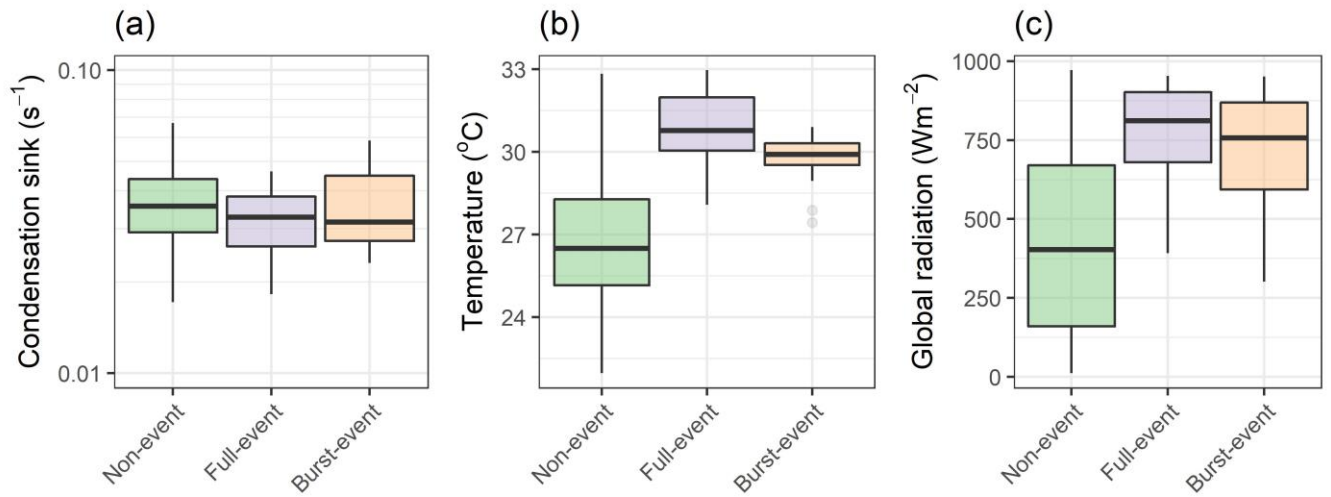


1159

1160 **Figure 1:** Average SMPS contour plots for (a) non-event days, (b) full-event days and (c) burst-event
1161 days.

1162

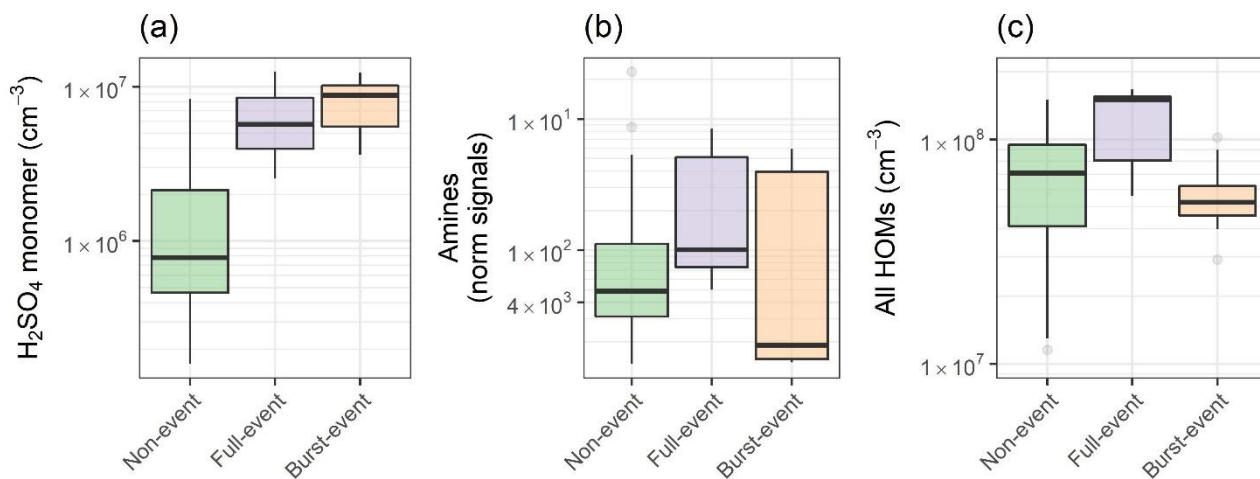
1163



1165

1166 **Figure 2:** Box plots for days of non-event, full-event and burst-event, showing (a) condensation sink,
 1167 (b) temperature, and (c) global radiation from hourly data. “Full-event” and “burst-event” include
 1168 data across the entire day.

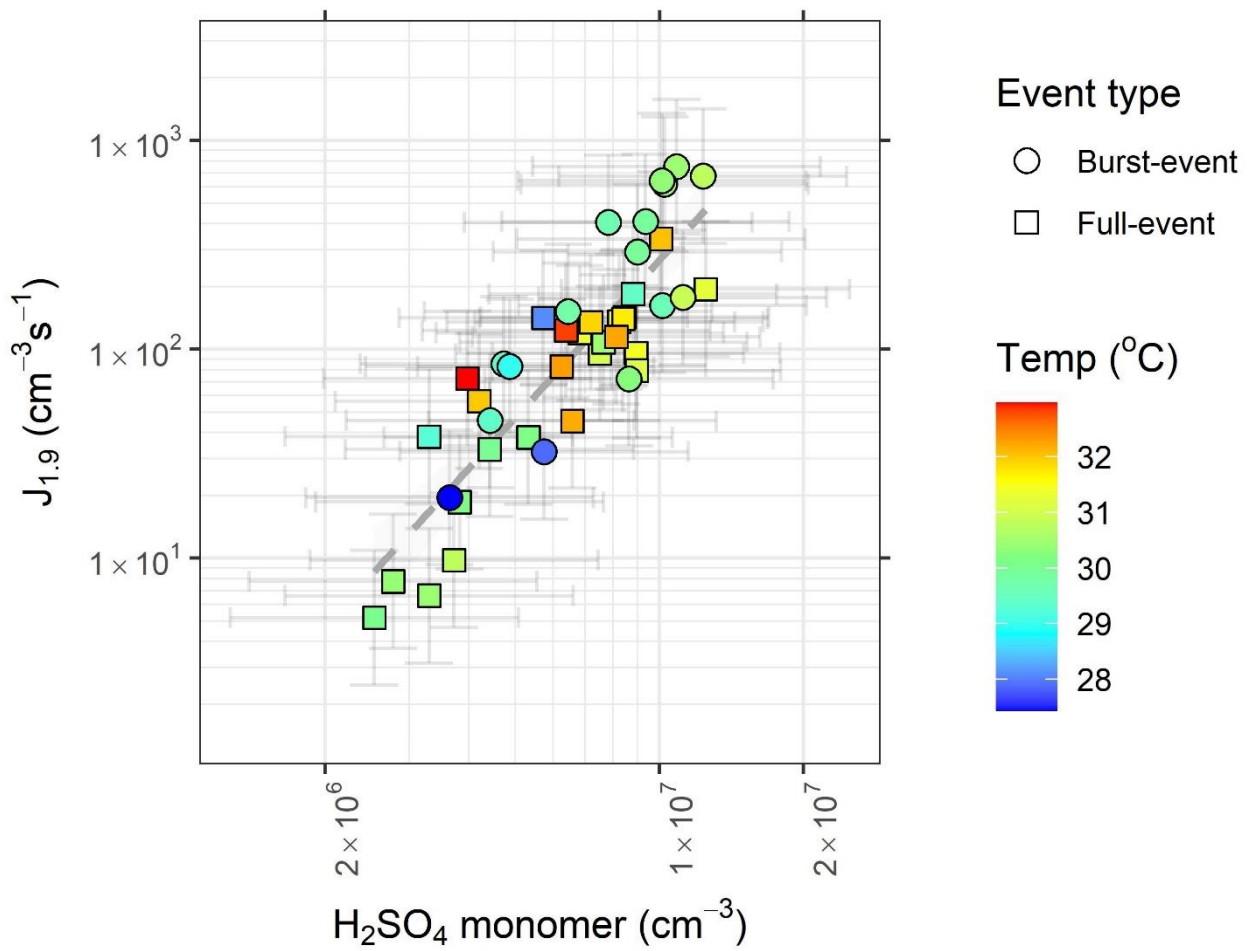
1169



1171

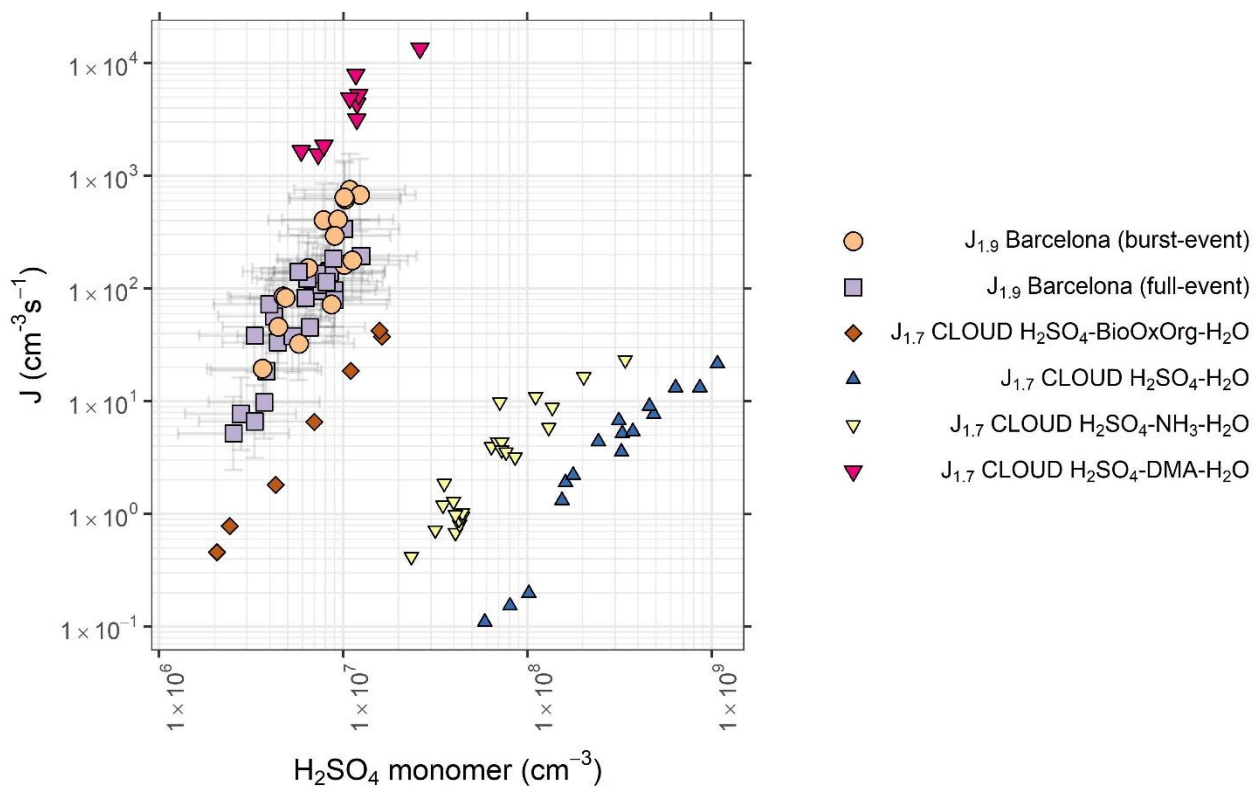
1172 **Figure 3:** Box plots for days of non-event, full-event and burst-event, showing (a) sulphuric acid, (b)
 1173 C_2 and C_4 amines, as clustered with the nitrate dimer and trimer, and (c) summed HOM concentration
 1174 from C_{5+} from hourly data. Units for ammonia + amines are normalised counts, as no calibration was
 1175 performed. Event days include data across the full event day.

1176



1178

1179 **Figure 4:** Formation rate ($J_{1.9}$) plotted against sulphuric acid monomer concentration, coloured by
 1180 condensation sink. Circles represent burst-events, squares represent full events. Data is for hourly
 1181 averages across NPF periods, typically within the hours 08:00 – 16:00. Slope of the line = $4.9 \cdot 10^{-5} \text{ s}^{-1}$
 1182 ¹. Error bars represent systematic uncertainties on $[\text{H}_2\text{SO}_4]$ and $J_{1.9}$

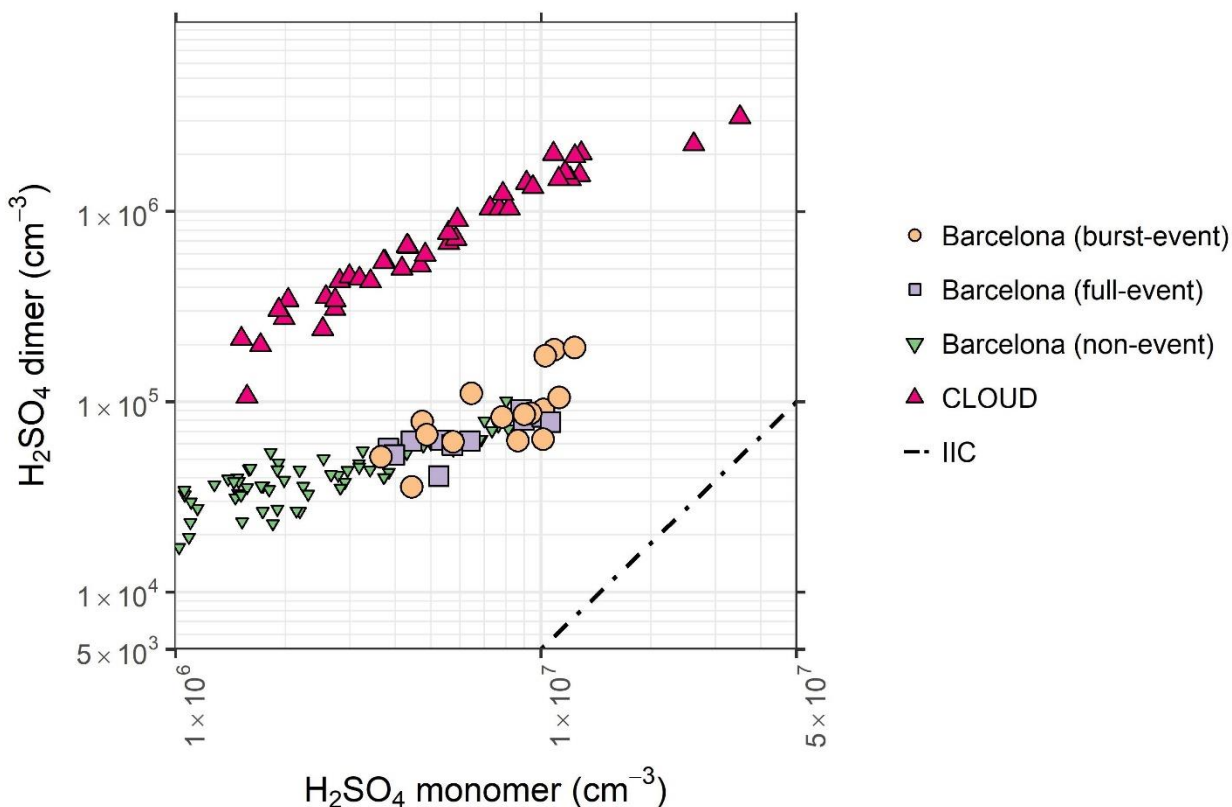


1183

1184 **Figure 5:** Formation rate plotted against sulphuric acid monomer concentration for data collected
 1185 from Barcelona. Tan circles represent burst-events, purple squares represent full events. as well as
 1186 that for the H₂SO₄-H₂O (blue inverted triangles), H₂SO₄-NH₃-H₂O (yellow inverted triangles),
 1187 H₂SO₄-DMA-H₂O (pink triangles), and H₂SO₄-BioOxOrg-H₂O (brown diamonds) systems from the
 1188 CLOUD chamber (Kürten et al., 2018 Kirkby et al., 2011; Riccobono et al., 2014). CLOUD
 1189 chamber experiments were performed at 278 K and 38 – 39 % RH. Data is for hourly averages
 1190 across NPF periods, typically within the hours 08:00 – 16:00. Error bars represent systematic
 1191 uncertainties on [H₂SO₄] and J_{1.9}.

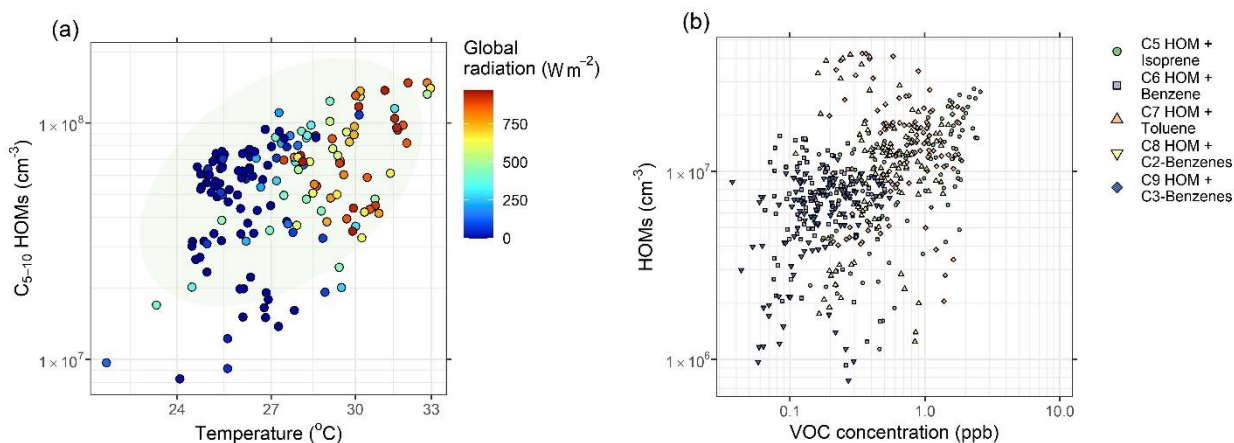
1192

1193



1194
 1195 **Figure 6:** Sulphuric acid dimer concentration plotted against monomer concentration, showing burst-
 1196 event periods (tan circles), full event periods (purple squares), non-event periods (green inverted
 1197 triangles), and the ratio of sulphuric acid dimer:monomer in the CLOUD chamber for the H_2SO_4 -
 1198 H_2O -DMA system (pink triangles) (Almeida et al., 2013). Dashed line represents the dimer
 1199 concentration produced by ion induced clustering in the chemical ionization unit (Zhao et al., 2010).
 1200 CLOUD chamber experiments were performed at 278 K and 38 – 39 % RH. Data is for hourly
 1201 averages across NPF periods, typically within the hours 08:00 – 16:00. Error bars represent systematic
 1202 uncertainties on $[\text{H}_2\text{SO}_4]$ and $[(\text{H}_2\text{SO}_4)_2]$.

1203

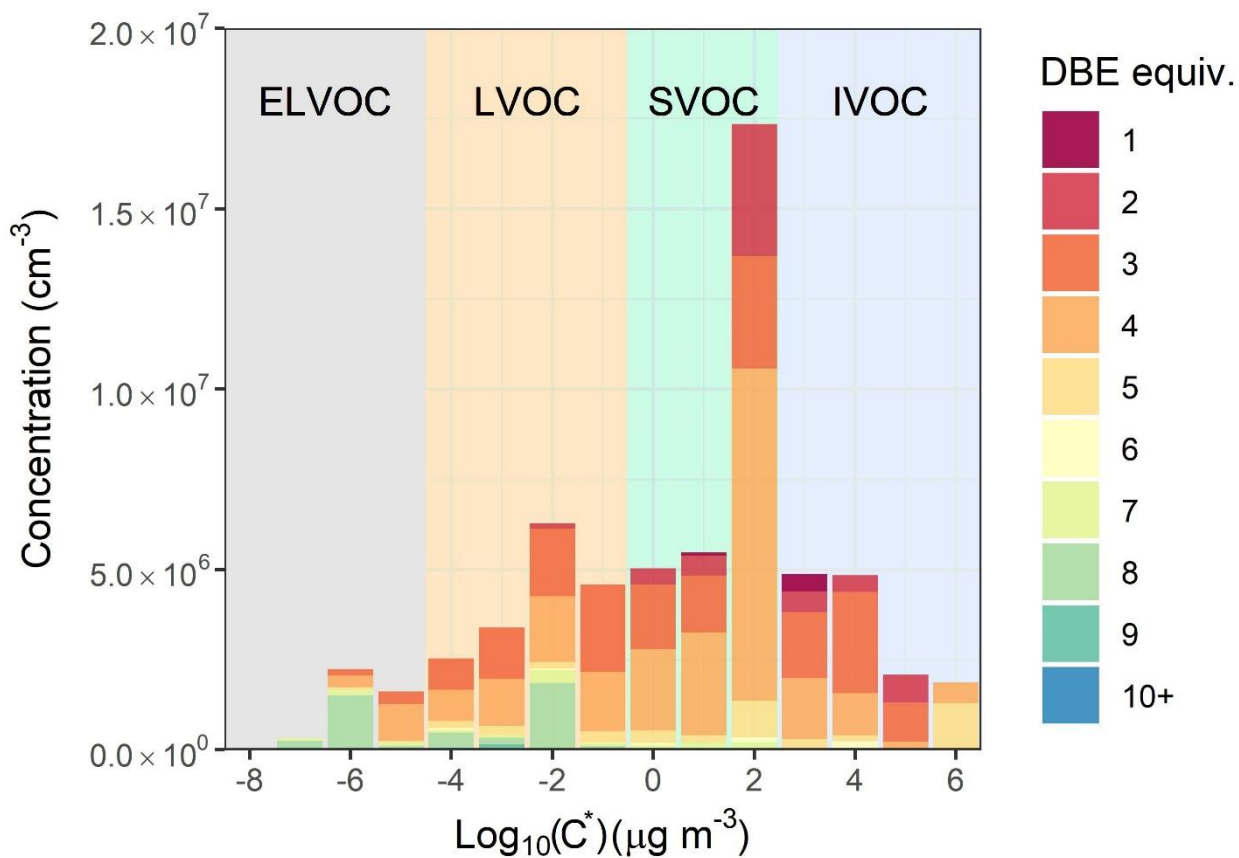


1205

1206 **Figure 7:** Influencing factors on HOM concentration, showing (a) C₅₋₁₀ HOM concentration plotted
 1207 against temperature, coloured by global radiation. Ellipsis shows 95% confidence on a multivariate
 1208 t-distribution. (b) HOM concentration by carbon number plotted against parent VOC mixing ratio.
 1209 These are segregated by carbon number/VOC, i.e. C₇ HOMs plotted against toluene, under the
 1210 assumption that toluene oxidation is the main producer of C₇ HOMs. Time for both plots is of hourly
 1211 time resolution.

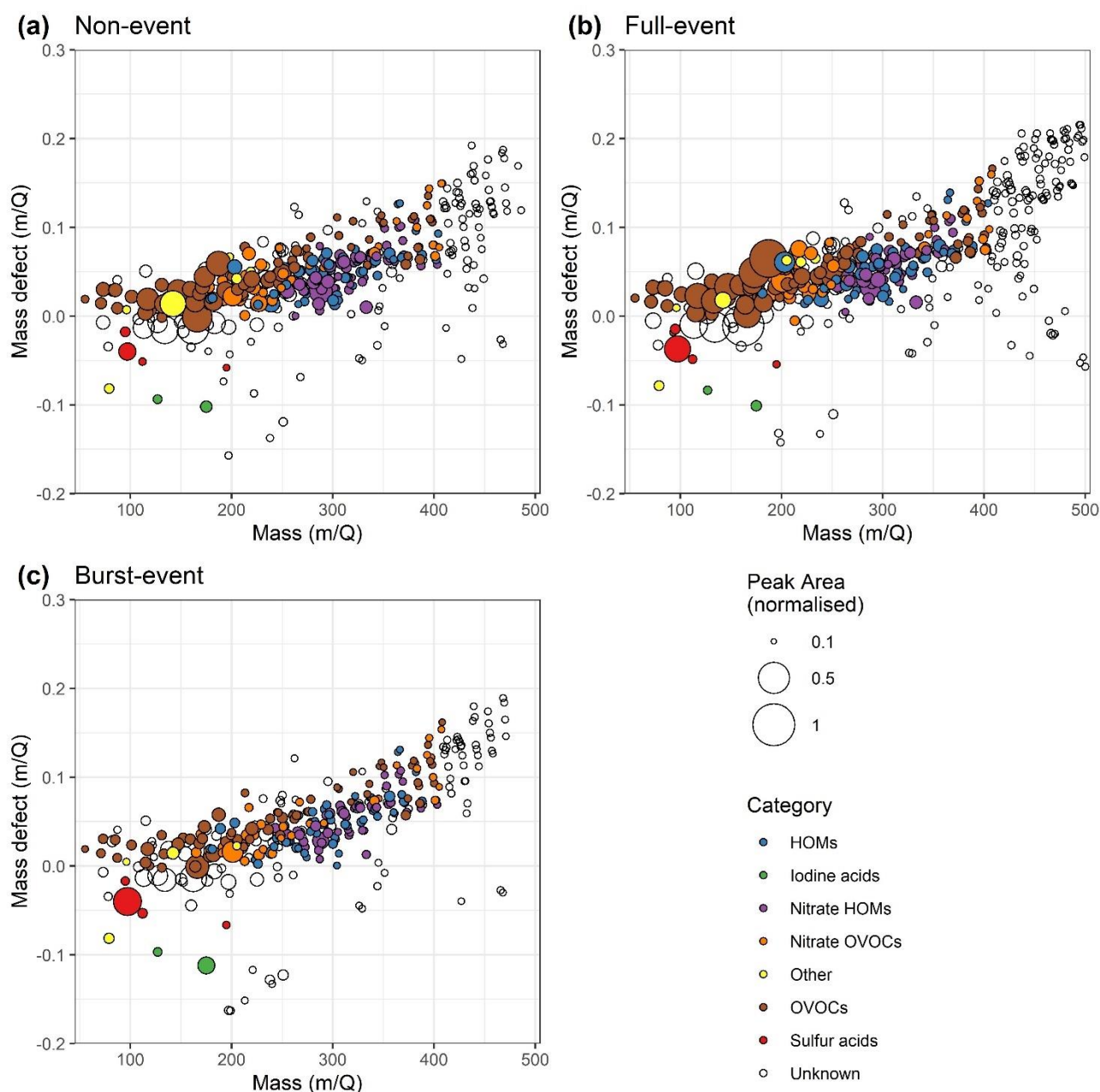
1212

1213



1214

1215 **Figure 8:** Concentrations of all oxygenated organic molecules and HOMs binned to integer $\text{Log}_{10}(C^*)$
1216 values, coloured by DBE.



1217

1218 **Figure 9:** Mass defect plots for (a) non-event, (b) full-event, and (c) burst-event periods, data taken
 1219 from 10:00 – 15:00 on the days 11/07/2018, 16/07/2018 and 15/07/2018 respectively. Size
 1220 corresponds to mass spectral peak area. Ions are coloured according to identified chemical
 1221 composition. *Blue* points correspond to HOMs containing all organic species with ≥ 5 carbon atoms
 1222 and ≥ 6 oxygen atoms, and an O:C ratio of >0.6 . *Purple* points correspond to the same but for species
 1223 containing 1-2 nitrogen atoms. Species not meeting this HOM criteria were classed generally as
 1224 OVOCs which are coloured *brown*, with the nitrogen containing OVOCs coloured *orange*. Sulphur
 1225 acids (*red*) include ions HSO_4^- , CH_3SO_3^- and SO_5^- , as well as the sulphuric acid dimer. Iodine acids
 1226 (*green*) contains both IO_3^- and I^- (the latter presumably deprotonated hydrogen iodide). Unidentified
 1227 points are left uncoloured.

1228