

Anonymous Referee #1

Received and published: 21 September 2020

Overall, I recommend publication of the manuscript by Hanson et al. after they have addressed the comments listed below.

Comments

(1) Section 2: Although the chemistry model is described in the earlier publication by Hanson et al. (2019) it would be good to add a paragraph, which summarizes the chemistry treated by the model. We will add a few sentences and reference specific sections (2.1 and S7) and tables (S1) where details are presented in Hanson et al. 2019.

(2) It is mentioned that the binary nucleation experiments yield the lowest values reported so far. The authors should include a figure, where all their measurements (the earlier ones from 2019 and the current ones) are inter-compared with the results from other studies. Currently such a figure is only shown for the experiments with ammonia but not for the nominally binary system.

This figure is now included as Fig. 5b (see included figure) that illustrates well how the results from PhoRF have changed over the course of a few years. A new section '3.5 Nominal Binary Results, Then and Now' was formed from some of the previous text and a new paragraph describing Fig. 5b.

(3) In Figure 7 results from a nucleation and growth model are shown for different sets of thermodynamic data. This model is probably rather complex and therefore evaluation would be beneficial. Evaluation could be performed by using an identical set of thermodynamic data and compare the model output to another model. This could, e.g., be done for the ACDC (Atmospheric Cluster Dynamics Code) model together with the thermodynamic data for H<sub>2</sub>SO<sub>4</sub>-NH<sub>3</sub> nucleation from Ortega et al. (2012). Results for these thermodynamic data using ACDC were presented by Kürten et al. (2016). The verification that we presented in our SC comment we will work up as a section in the Supplement. PhoRF reviewer's comment helped us to see that we had not published such a verification step for the box model; we note that the 2D model had been compared to a commercial CFD result (S2.1.1 of Hanson et al. (2017)).

Further comments

L155 (page 5): Please specify why NO accelerates the H<sub>2</sub>SO<sub>4</sub> production?

NO reacts with HO<sub>2</sub> that is generated in the OH + SO<sub>2</sub> + NO<sub>2</sub> reaction. Will be parenthetically noted in the revised text.

L282 (page 9): Why was the CPC inlet exposed to room air?

To get a reference pulse height for large particles. Because of shifting baselines due to older electronics, we decided to not rely on pulse heights for sizing information. Nonetheless, we traced the main reason for changes in the UCPC vs. DEG system responses to a cylinder changeover.

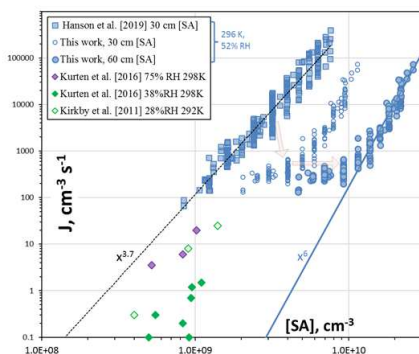


Fig. 5(b) Nucleation rate vs. H<sub>2</sub>SO<sub>4</sub> (SA) concentration. For PhoRF data, nucleation rate given by N<sub>g</sub> divided by an estimated 5 s nucleation time. The [SA] concentration was that calculated at 30 cm, both for the Hanson et al. (2019) data set and for this work (open circles). The filled circles are the nucleation rates for this work but plotted at the 60 cm calculated [SA] concentration. The red arrows show how the system has evolved for data taken at Q<sub>0</sub>=4.25 sccm. Nucleation rates from the CLOUD experiment at 292 and 298 K for nominal binary conditions are also shown.

Additional changes not in direct response to this reviewer's comments (see also the end of response to referee 2.)

In the Supplement there are a few new figures and a couple figures and paragraphs were removed. The new figure S9 is a plot of the mode diameter of the nanoDMA system and the NO<sub>x</sub> measurements over time. It includes some data from when the DEG system was replaced by a second butanol CPC as the DEG UCPC became very noisy at times. Since the measurements in the figure are for particles of 5 nm diameter and larger there are no issues that affect the data presented in this figure.

Comment [D1]: Lines 163-166 in the "changes" indicated PDF manuscript.

Comment [D2]: Lines 481-532.

Comment [D3]: Supplement section S2.1.

Comment [D4]: Line 227.

Comment [D5]: This text has been replaced by lines 463-466 and Supplement sections S1.1.1 and S1.1.2.

Comment [D6]: This new figure has been added for the new version, including model calculations to give a mid-point nucleation rate (i.e. the reference H<sub>2</sub>SO<sub>4</sub> at 60 cm down the reactor.)

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**NOTE: add 5 to all the line numbers referenced in these responses.**

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**Anonymous Referee #2**

Received and published: 27 October 2020

The article presents upgrades to a photolytic flow reactor system to study nucleation of sulfuric acid and presents some new results using the system. Although sulfuric acid water (+base) nucleation has been studied extensively by different teams, there are discrepancies in the results obtained using different measurement equipment, so there is a need for validating earlier studies and improving the measurements. Therefore I think the article is in principle worth publication as measurement report. However, the article needs revision for clarity and more discussion on the uncertainties. The authors believe that the differences in their new results compared to their earlier study (Hanson et al. 2019) is due to improved cleanliness of the system, but since there are no actual measurements of the contaminants this remains speculative. Also, the authors found large discrepancies in the results obtained using two different particle counters. Given the uncertainties, I have doubts how well the results can be compared to other nucleation studies.

We hope our detailed responses to the referee's comments below will alleviate many doubts. In this paragraph we provide an alternate perspective on the effects of the uncertainties they have detailed. Our statements about increased cleanliness are definitely speculative but it is the simplest explanation and a common problem in nucleation experiments. The large discrepancies between counters we have argued are two-fold. In one case, the differences have (i) revealed a source of error or artifact in our DEG system that improves the analysis of the present measurements and (ii) this artifact led us to calculate ion-mediated nucleation with our model and these calculations indicate that IMN could play a significant role at low sulfuric acid abundances. The other discrepancy (the UCPC Np being larger than the DEG Np) seems to be completely traceable to a cylinder changeover event where particle counts were elevated for a week or two. We note that another referee asked that comparisons to previous experiments be included in an additional graph which we are glad to do as it highlights how the results from RhOE have improved. We agree that we should temper our conclusions regarding comparisons with other studies because of the uncertainties in the measurements.

**General comments:**

At first read it was hard to understand the main aim of this measurement report and its connection to the previous study by the same team (Hanson et al. 2019). I think it would be beneficial to state the objectives more clearly in the introduction paragraph.

Excellent point: the previous abstract lacked focus. We have removed a few sentences and now highlight the changes in the results (about an order-of-magnitude or larger in some cases) that have occurred since our initial publication.

**Comment [D7]:** e.g. First sentence of the abstract clearly states the point of this report.

Chapter 3.4.: I would separate the discussion of why DEG-CPC shows considerably higher counts than UCPC (this should be actually discussed a bit more, see my questions below) from the discussion of which nucleation processes affect the UCPC data (r232-276). To me these seem to be two separate issues each deserving their own chapter.

This comment prompted a good, hard, long look at the text and led to extensive reworking of the discussion of the UCPC corrections and a new set of figures comparing it to the DEG system is presented in the Supplement.

**Comment [D8]:** They are now separated into two paragraphs, lines 351-369 (in marked up document).

The summary and conclusions chapter would benefit from shortening and streamlining it. I would concentrate on summarizing what is improved from the 2019 study and what new knowledge that brings, and remove most of the speculation (e.g. related to CLOUD data) that was already discussed in the Results & Discussion part.

We agree and three paragraphs have been removed. We highlight better the changes in the measurements from those presented in our previous publication. As you mention, some recent studies suggest an enhanced collision rate of sulfuric acid molecules (Stolzenburg et al. 2020 but also Halonen et al. 2019) leading to faster growth rates. How much would it affect your results if you include such collision enhancement in your calculations? Can you provide an uncertainty estimation for Fig 7? You note this qualitatively in the conclusions, but maybe this discussion could be moved to results and discussion section and addressed more quantitatively. This is an important point that we had quantitatively skirted in the paper. We have prepared a paragraph summarizing the effects on the model predictions and how the cluster energetics ~~used~~ to be modified to achieve agreement between model and measurements.

Comment [D9]: A new section, 3.8 and table, Table 3, lines 595-734.

Figures: The figure captions and variable names in legends should be revised throughout the article and supplement so that they are self-explanatory. Currently the figures cannot be understood without reading the whole text. E.g. the difference between NH<sub>3</sub>\_S2 and NH<sub>3</sub>\_D52 and meaning of M1 (red squares) in Fig S1 are not clear. It would be helpful if it was made clearer which results are from this study and which are obtained earlier with the same system. We have paid close attention to the captions in the revised version and believe it is clear what has been previously published.

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Detailed comments:

I have several specific questions regarding the particle counting, which need to be clarified as it is one of your main new findings, that the DEG system and UCPC show large discrepancies at low H<sub>2</sub>SO<sub>4</sub> (lower particle concentration).

On p3 r82-85 you write: "While there may be a ~20% undercount in the UCPC results as detailed in the previous paragraph, this may be counteracted somewhat as the UCPC detects more particles than are in the leading edge of the particle size distributions of the DEG system. It is difficult to quantify this amount because the pulse-height response of the instrument depends on the composition of the particles [O'Dowd et al. 2004; Hanson et al. 2002].". The assumed 5% losses for UCPC seem quite low, is there any measurements to characterize the size-dependent losses in the setup used? Does the undercounting depend on the size distribution of the particles you produce, as the two instruments certainly have different detection efficiency curves? Why and how does the composition dependency of pulse-height analysis play a role here, if I understood correctly you use it only to calculate the total concentration?

Our rambling text here led to a misunderstanding of the corrections and we have extensively revised it to make clear the following information. The 20 % undercount was due to not applying an activation efficiency; we have chosen now to instead actually apply the activation efficiency. The 5 % loss that is not (and won't be) applied is due to the sampling through a sharp right angle. Larger losses exist; we did state that the diffusional loss of nanoparticles was calculated using Gormley-Kennedy and in the revision we list typical losses that range up to 60% at 2.7 nm. According to the literature, the two counters have activation efficiencies that are similar down to about 2.5 nm, which is the lower-limit for the comparisons. The composition dependent activation efficiency is an uncertainty and we suppose they are similar whether DEG or butanol fluids; yet we focus on the leading-edge particles with the DEG system whereas that cannot be readily done with the UCPC data.

Comment [D10]: Paragraphs on lines 95-116 have been heavily revised.

How often did you measure the background (zero) of your counters, especially the DEG-counter? I'm asking because if you have even very few background counts from homogenous nucleation of DEG, it would be interpreted as large signal in the MPS

system. How and how often are these instruments calibrated?

The DEG system background count rates are measured every day at the start of the measurements and occasionally again at the end. These instruments have not been calibrated but are operated in accord with the literature sources we reference.

Comment [D11]: A paragraph in the Supplement has been added, S1.1.

On p9 you speculate about different processes that may affect the concentration measured with UCPC. One possibility brought up is particles formed in a charger (r230) and second the direct detection of sulfuric acid clusters (r254). If your CPC uses pulse-height analysis, shouldn't these particles (which are probably very small at detection) be clearly distinguishable from particles formed in a flow tube?

In principle this is true however we think neither are clearly distinguishable from "normal" nanoparticles. The charger ions are at much lower abundance than the neutral particles furthermore the pulse-heights for clusters and particles of 3 nm or smaller become very spread out and do not form identifiable peaks in the distributions.

Comment [D12]: See lines 110-116

(r272-285) you note that sometimes the relation between UCPC and DEG measured concentration changes (by several factors). To me it sounds the reason has to be technical, as you also speculate. Isn't there any diagnostics you can use to validate when one of the counters are measuring incorrectly to eliminate this data? Maybe provide a comparison of the UCPC and DEG measurements in the supplement?

We now present a comparison of nanoparticle abundance measurements in the Supplement. The two instruments compare admirably, albeit up to 30% different in one case, over the size range of 3 to 12 nm diameter. Furthermore, we have now traced the anomalously high UCPC measurements to 4 measurement days just after a liquid nitrogen cylinder change. Why this intermittent dust event (or whatever) apparently affected the UCPC instrument more than the DEG system is not known. It may be a clue as to the identity of what caused this (or the other) cylinder-change events.

Comment [D13]: S1.1.1.

Comment [D14]: This month long event is now documented in S1.1.2.

Other substantive changes:

See the Comment attached to Fig. 1 but we overlooked the capillary flow rate variation corrections in our original version. It has now been applied and averages about 10%.

The new Fig 5(b) uses N<sub>g</sub> from the sum of the particles in the DEG-system distribution as was done for the previous data shown in that figure. The H<sub>2</sub>SO<sub>4</sub> concentration at 60 cm as a function of Q4 was taken from a new figure in the supplement, Fig. S11.

# Measurement Report: Sulfuric Acid Nucleation and Experimental Conditions in a Photolytic Flow Reactor

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**Abstract.** Nucleation rates involving sulfuric acid and water measured in a photolytic flow reactor have decreased considerably over a time period of several years. Results show that the system - flow reactor, gas supplies and lines, flow meters, valves, H<sub>2</sub>SO<sub>4</sub> photo-oxidant sources – has reached a baseline stability that yields nucleation information such as cluster free energies. The baseline nucleation rate is punctuated by temporary bursts that in many instances are linked to cylinder changes, delineating this source of potential contaminants. Diagnostics were performed to better understand the system include growth studies to assess H<sub>2</sub>SO<sub>4</sub> levels, chemi-luminescent NO and NO<sub>x</sub> detection to assess the HONO source, and deployment of a second particle detector to assess the nanoparticle detection system. The growth of seed particles show trends consistent with the sizes of nucleated particles and provide an anchor for calculated H<sub>2</sub>SO<sub>4</sub> concentrations. The chemiluminescent detector revealed that small amounts of NO are present in the HONO source, ~10 % of HONO. The second, condensation-type particle counter indicates that the nanoparticle mobility sizing system has a bias at low sulfuric acid levels. The measured and modeled nucleation rates represent upper limits to nucleation in the binary homogeneous system, H<sub>2</sub>SO<sub>4</sub>-H<sub>2</sub>O, as contaminants might act to enhance nucleation rates and ion-mediated nucleation may contribute. Nonetheless, the experimental nucleation rates, that have decreased by an order-of-magnitude or larger since our first publication, extrapolate to some of the lowest reported in experiments with photolytic H<sub>2</sub>SO<sub>4</sub>. Results from experiments with varying water content and with ammonia addition are also presented and have also decreased by an order-of-magnitude from our previous work; revised energetics of clusters in this three-component system are derived which differ from our previous energetics mainly in the 5 acid and larger clusters.

## 1 Introduction

Atmospheric nucleation involving sulfuric acid, water and ammonia is believed to have a large impact on the properties of clouds and their effects on the radiation balance of the atmosphere [Dunne et al. 2016; Coffman and Hegg, 1995]. Since the influence on climate of aerosol particles is well-known to be potentially quite large [IPCC, 2013], the formation of atmospheric particles through nucleation has long been studied with a focus on sulfuric acid [McMurry et al., 2005; Kulmala et al. 2004]. The chemical systems thought to have the largest global impacts on new particle formation are the binary H<sub>2</sub>SO<sub>4</sub>-H<sub>2</sub>O and ternary H<sub>2</sub>SO<sub>4</sub>-H<sub>2</sub>O-NH<sub>3</sub> systems.

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Yet over the decades there is a wide divergence in nucleation rates measured in laboratory experiments, particularly in the binary system - see figure 6 in Zollner et al. [2012] and figure 9 in Hanson et al. [2019]. While many of these discrepancies may be attributable to contamination or other experimental conditions, this leaves nucleation rates for atmospheric conditions somewhat uncertain. Likewise, nucleation rates in the ternary system derived from laboratory experiments have uncertain applicability to the atmosphere due in part to effects of potential contaminant bases.

Recent nucleation results from the CLOUD experiment [Kirkby et al. 2011; Almeida et al. 2013; Ehrhart et al. 2016; Kürten et al. 2016] may best represent binary and ternary nucleation and these results have been parameterized and used in global climate models where their climate effects were found to be significant. Yet these nucleation rates have not been corroborated in independent laboratory investigations. In fact there is disagreement with other recent work, albeit based on extrapolated data and/or theoretical treatments (see Zollner et al. 2012; Hanson et al. 2019; Kürten et al. 2016).

For the binary system, Ehrhart et al. [2016] showed that the CLOUD data at low temperatures agrees well with theoretical nucleation rates from the SAWNUC thermodynamics [Lovejoy et al. 2004]. However, there was poor agreement at temperatures warmer than 270 K. This poor agreement has been attributed to contamination by base molecules that are enhanced at warm chamber temperatures. Certainly contamination is a concern for experimental work in the binary system [e.g. Zollner et al. 2012; Hanson et al. 2019]. Here, we report more measurements in the putative binary system to gain further information on the effects of potential contaminants in our Photolytic Flow Reactor (PhoFR) [Hanson et al. 2019].

We present measurements of NO and HONO + NO<sub>x</sub>, growth measurements, and measurements from a second particle counter that improve knowledge of the conditions in PhoFR.

For the ternary system, Kürten [2019] and Hanson et al. [2019] used cluster models to derive thermodynamic information from the CLOUD and the Photolytic Flow Reactor (PhoFR) results, respectively. Both concluded that the thermodynamics derived by Hanson et al. [2017] for H<sub>2</sub>SO<sub>4</sub>-NH<sub>3</sub> clusters from the bulk flow reactor study of Glasoe et al. [2015] are inaccurate; Hanson et al. [2019] speculated this was due to a cross-contamination by an amine in the ammonia delivery lines.

We present here new ternary system measurements from PhoFR and discuss the effects of potential contaminants. We compare predictions of our cluster model to the CLOUD experimental data and to those of Kürten [2019] and Yu et al. [2020]. Finally, we present the effects of relative humidity (RH) on the nucleation rate in the ternary ammonia-sulfuric acid-water system.

## 2. Experiment.

Nucleation was studied in PhoFR, a vertically-mounted, 120 cm long, 5 cm inner diameter, jacketed, glass flow reactor surrounded by 4 UV (~ 365 nm peak) lamps. The total flow was 2.9 sLpm (273 K and 1 atm, standard L / min), temperature was 296 K, the main gas-source was liquid nitrogen that has ppm-levels of oxygen present, and H<sub>2</sub>SO<sub>4</sub> was produced from photolysis of HONO and subsequent oxidation of SO<sub>2</sub>. Typical flows were: dry nitrogen, Q<sub>2</sub>, 1.33 sLpm; water was introduced as a humidified N<sub>2</sub> flow, Q<sub>3</sub>, 1.5 sLpm; HONO level was set with a flow, Q<sub>4</sub>, of between 2 and 10 sccm (standard

cm<sup>3</sup> per min); the flow of the SO<sub>2</sub> mixture (0.15% in N<sub>2</sub> or 1% in 0.1 O<sub>2</sub>/0.9 N<sub>2</sub>), Q<sub>1</sub>, 4 to 32 sccm. For the majority of the work presented here, there was no added base. For some experiments, diluted base in an N<sub>2</sub> flow of about 35 sccm was added via a port at the top of the flow reactor, as described previously. The base mixing ratios quoted below were calculated from known flow rates and assumed fast mixing and no losses; only small portions of the flow reactor achieved those mixing ratios (note that the simulations included mixing and losses). Particles were detected with a nanoparticle mobility particle sizer (MPS) system (DEG system, Jiang et al. [2011]); this and the complete apparatus are fully described in Hanson et al. [2019]. The Supplement details more of the system's essentials.

A second particle detector, the ultrafine condensation particle counter (UCPC, butanol) of Zollner et al. [2012], was attached to the exit of PhoFR for some experiments. A similar co-deployment of the UCPC and the DEG system in a nucleation experiment found good agreement between them [Glasoe et al. 2015]. The basics of the UCPC is described in Stolzenburg et al. [1991] and we deployed it here with the modifications suggested by Kuang et al. [2012] where the saturator minus condenser temperature was set to 35 C and the condenser and capillary flow rates were 0.45 L/min and 0.06 L/min, respectively. Pulse-height analysis data was collected from which the total particle number concentrations were obtained. Size-dependent losses of nanoparticles in sampling lines and inside the UCPC were accounted for using the Gormley and Kennedy [1948] equations; losses are 20, 40 and 60 % at 8, 4 and 2.7 nm diameter, respectively. For charged NaCl particles, activation efficiency for butanol working fluid is similar to that for diethylene glycol, greater than 80 % for 2.5 nm and larger particles, determined by comparing results from Jiang et al. and Kuang et al. The UCPC sampled flow via one leg of a Swagelok 1/4" tee usually with an additional 0.3 L/min transport flow (total sampling rate of 0.75 L/min). The UCPC data was recorded with the tee placed after the DEG system charger; a limited set of early experiments with the tee upstream of the charger did not show large differences in results. Splitting the sample flow with tube fitting tees probably introduces an additional loss on the order of 5%, depending on the details of the sampling tubulation [Wang et al. 2002]. Since losses due to the tee are small they were not accounted for.

The activation properties of butanol vapor onto nanoparticles composed of sulfuric acid are complicated by potential chemical reactions [O'Dowd et al. 2004; Hanson et al. 2002] and the UCPC's pulse-height analysis sizing capabilities are compromised. Thus, separating the largest particles from the total detected by the UCPC is difficult to do with sufficient certainty. So in principle the analysis of the UCPC data here will be influenced by more particles than are in the leading edge of the particle size distributions of the DEG system. A good relationship between these two data sets is expected when the size distribution has a prominent large particle mode, the so-called leading-edge particles, which is the case for most of the UCPC results presented here.

Growth studies were performed where H<sub>2</sub>SO<sub>4</sub>-dimethylamine nano-particles, formed in the bulk flow reactor of Glasoe et al. [2015], were introduced into PhoFR where they could be exposed to H<sub>2</sub>SO<sub>4</sub>. The growth of the particles was assessed by changes in the particle size distributions upon turning on the UV lights. Specifically, the volume-weighted mean diameters of the leading edge of the size distributions are derived and plotted as a function of HONO. These studies were performed

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over a limited range of HONO concentrations because nucleation of new particles at high HONO levels significantly interferes with the initial size distributions.

The source of HONO is the reaction of HCl with NaONO(s) and, as discussed by Febo et al. [1995], NO can be produced by decomposition of 2 HONO molecules into H<sub>2</sub>O, NO and NO<sub>2</sub>, possibly followed by reaction of NO<sub>2</sub> with HONO giving NO and HNO<sub>3</sub>. A chemiluminescent NO and NO<sub>x</sub> detector (TECO 42b) was periodically used to assess the NO and the NO+NO<sub>2</sub>+HONO levels exiting PhoFR. It was calibrated with a 105 ppb NO in N<sub>2</sub> mixture (AirGas).

The 2D model of PhoFR developed in Hanson et al. [2019] was also used here. Flow was assumed to be fully developed laminar (parabolic velocity profile) and the full suite of photolytic reactions were included along with H<sub>2</sub>SO<sub>4</sub> condensation to and evaporation from its molecular clusters.

In short, the photo-chemistry is initiated by photolysis of HONO yielding OH and NO, and H<sub>2</sub>SO<sub>4</sub> is formed promptly after OH reacts with SO<sub>2</sub>. HO<sub>2</sub> is also a product and it can react with NO generating another OH radical. See sections 2.1 and S7 of Hanson et al. [2019] for the detailed photochemistry. Cluster chemistry is tracked in detail with evaporation rates determined by thermodynamics for clusters up to ten H<sub>2</sub>SO<sub>4</sub> molecules; clusters larger than this were not allowed to evaporate. Note that water molecules were not tracked in the clusters but were allowed to affect their size (thus kinetics) and evaporation rates (thus thermodynamics); this quasi-unary approach for simulating the binary system is similar to that of Lovejoy et al. [2004] and Yu [2005].

Note that base (ammonia or dimethylamine) molecules can be included in the simulations in a quasi-binary approach to these ternary systems. Molecular clusters up to ten H<sub>2</sub>SO<sub>4</sub> molecules containing up to ten base molecules could be simulated; for most simulations presented here, only clusters up to 3 base molecules were needed (i.e., results were within a few percent of simulations when up to 6 base molecules were included.) The binary (quasi-unary) free energies were allowed to significantly influence the ammoniated cluster free energies for the larger clusters (see Supplement); thus water has an effect on NH<sub>3</sub>-H<sub>2</sub>SO<sub>4</sub> quasi-binary free energies. Further details including kinetic rates, diffusion coefficients, cluster thermodynamics and other assumptions are presented in the Supplement.

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### 3. Results and Discussion

#### 3.1 System stability, HONO source evaluation

A long time series of data (particle number density, N<sub>p</sub>, vs. time) is shown in Figure 1 for a set of flow rates that give standard experimental conditions: 52 % RH, 296 K, and a putative [HONO] of 5x10<sup>11</sup> cm<sup>-3</sup>. The previously published N<sub>p</sub> data from May and June 2018 [Hanson et al. 2019] are also included. The data after Nov 2018 are generally much lower than data previous to this time which we believe is due to an improvement in the cleanliness of the system.

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The data in Figure 1 reveal a long-term stability of the system but with sporadic changes that might be due to contaminants etc.: the spikes and variability in N<sub>p</sub> are often associated with cylinder changeovers (+ symbols; those in bold correlate with spikes). Despite repeated flushing of the gas-delivery lines, room air and perhaps dust may get introduced into PhoFR during a changeover; degassing of the exit of the dewar's valve and the regulator port that are exposed to the atmosphere for

200 weeks while at the supplier (Airgas) may play a role. A stainless steel filter was installed on the flow meter manifold in late August 2019 but spikes still appeared after that. Nightmare compounds such as diamines could be responsible as they can lead to a single particle for each molecule [Jen et al., 2016]: a  $10^{-15}$  mole fraction could lead to spike of  $10^4 \text{ cm}^{-3}$ . Yet for all of 2019 there is a 'floor' of about  $1000 \text{ particles cm}^{-3}$ . The floor of the 2019 data is about 3 % of the average of the particle concentrations in early 2018, a significant drop that we attribute to the depletion or elimination of a source of base molecules within the flow. Extrapolating this floor in the DEG data to lower sulfuric acid, the PhoFR data are now more or less in agreement with the photolytic data from the CLOUD consortium but still somewhat larger than our earlier (Zollner et al. 2012) bulk flow reactor data. We present data below that indicates this floor is heavily influenced by a background level of particles in the DEG system, with the UCPC seeing much lower  $N_p$ . Nonetheless, this data reveals that the stability of the system over the time periods of years is suitable for studying nucleation and growth for the binary (sulfuric acid - water) system and that the effects of contaminants can be discerned. See the Supplement for plots of  $N_p$  and  $D_p$  vs.  $Q_4$  binned into four time periods.

Temporal variability in the floor is still evident which may be influenced by variations in experimental conditions such as the oxidant source. To assess the HONO source, we measured NO and  $\text{NO}_x$  in the effluent of PhoFR and small levels of NO (and also probably  $\text{NO}_2$ ) were present, roughly 10 % of the detected level of  $\text{NO}_x$  ( $\text{NO} + \text{NO}_2 + \text{HONO}$ ): this data as a function of  $Q_4$ , the HCl-laden flow over the  $\text{NaONO(s)}$  powder, are shown in Figure 2.

The total amount of  $\text{NO}_x$  produced in the HONO source is well-represented by a linear dependence on the flow rate  $Q_4$  of the HCl-laden flow. However, NO is not linear with  $Q_4$  and it has a higher relative variability than  $\text{NO}_x$ . Fluctuating NO impurity levels will affect the  $\text{H}_2\text{SO}_4$  production rate (see below) and thus  $\text{H}_2\text{SO}_4$  concentrations, contributing to the variability of  $N_p$ . On average, the fraction of  $\text{NO}_x$  that is NO ranges from 0.13 at  $Q_4 = 2.5 \text{ sccm}$  but decreases to 0.07 at  $Q_4 = 10 \text{ sccm}$ . A decreased decomposition of HONO at higher flow rates is expected as the  $\text{NaONO}$ -vessel (since Nov-2019, a 5 ml pear-shaped glass vial) is more rapidly flushed; in fact, Febo et al. [1995] recommend flows of 100's of mL per min to minimize the HONO decomposition rate; they used dry gas to dilute water vapor that also helped slow HONO loss. In our system for studying oxidation of organic compounds, this change was instituted and an additional flow of 50 sccm dry air caused measured NO levels to decrease to a few percent of the  $\text{NO}_x$  level (Hanson et al. "Growth with SOA", in preparation, 2020).

We used our model to assess how HONO decomposition affects  $\text{H}_2\text{SO}_4$  levels and thus  $N_p$ . Three scenarios were run where  $\text{NO}_x$  was attributed to (1) 100% HONO, (2) 80 % HONO / 10 % NO / 10 %  $\text{NO}_2$ , or (3) 85% HONO / 10 % NO / 5 %  $\text{HNO}_3$ . Scenario (3) is a situation where 10% of the HONO decomposes to 5% NO and  $\text{NO}_2$ , followed by  $\text{NO}_2$  reacting with HONO giving an additional 5% NO and  $\text{HNO}_3$  ( $\text{HNO}_3$  is not included in  $\text{NO}_x$ , there could be a 4<sup>th</sup> scenario, 89.5% HONO and 10.5%  $\text{NO}_x$ .) A revised photolysis rate of  $4.0 \times 10^{-4} \text{ s}^{-1}$  was used for HONO (see the next section.)

The centerline  $\text{H}_2\text{SO}_4$  is shown in Figure 3 as a function of axial distance down the reactor for these three scenarios. Compared to the 100 % HONO simulations, those with 15-20% HONO decomposition and 10 % NO lead to substantial increases in the rate of oxidation of  $\text{SO}_2$ , boosting  $\text{H}_2\text{SO}_4$  production by up to a factor of two due to OH generated in the

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235 | reaction of NO with HO<sub>2</sub>. The particle sizes would also be affected, where the total exposure of a particle to H<sub>2</sub>SO<sub>4</sub> along  
the length of PhoFR can be approximately doubled at 10% NO levels; the main reason that the photolysis rate for HONO  
deduced here is 50 % of that in our previous work (see the next section). Further discussion of the modeling and how the  
calculated H<sub>2</sub>SO<sub>4</sub> from our previous work compares to the present calculations are contained in the Supplement. Not only  
does a small amount of NO accelerate H<sub>2</sub>SO<sub>4</sub> production but these simulations also reveal that calculated H<sub>2</sub>SO<sub>4</sub> levels are  
240 | sensitive to small fluctuations (6%) in HONO levels when NO is present: average H<sub>2</sub>SO<sub>4</sub> in scenario (3) is 20 % higher than  
that in scenario (2).  
An effect on N<sub>p</sub> due to varying SO<sub>2</sub> mixing ratios was found in the measurements presented by Hanson et al. [2019] but this  
effect is now no longer experimentally present. Standard photochemistry with 100% HONO entering PhoFR predicted a  
small dependence, but interestingly the inclusion of a few ppbv of NO from the HONO source (Supplement and see below)  
245 | increases the predicted dependence on SO<sub>2</sub>. The system's increased cleanliness (or perhaps change in the type of  
contaminant) and the decreased experimental dependence on SO<sub>2</sub> level may be related. The Supplement has more discussion  
and presents experimental and theoretical N<sub>p</sub> as a function of SO<sub>2</sub>.

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### 3.2 Growth of nano-particles produced in bulk flow reactor

A series of experiments were conducted where particles generated in a bulk flow reactor (BFR, [Zollner et al. 2012; Glasoe  
250 | et al. 2015]) were introduced into PhoFR where they could be grown with photochemical H<sub>2</sub>SO<sub>4</sub>. BFR flow was 1.5 sLpm  
and enough dimethylamine was introduced into BFR's base addition port to induce particle number densities sufficient for  
the growth measurements (the no-loss amine level was typically 15 pptv). Total flow in PhoFR was maintained at 2.9 sLpm  
and relative humidity and SO<sub>2</sub> concentrations were controlled as before as well as HONO where Q<sub>4</sub> was varied from 1.5 to 6  
255 | sccm. The amount of dimethylamine in the initial particles is not known and it is assumed to have no effect on the  
photochemical growth of the nanoparticles. A set of size distributions from the DEG system as well as the results of three  
other growth experiments, one with ammonia swapped in for dimethylamine, are presented in the Supplement.  
Two sets of growth experiments are shown in Fig. 4, a plot of the volume-mean diameter of the leading edge particles vs. Q<sub>4</sub>,  
the flow of the HCl-laden flow through the NaONO(s) vessel. The diamonds are with the UV lights on, the circles are lights  
off (plotted at Q<sub>4</sub> = 0; HONO level did not affect the initial size of the particles from BFR.) For one set of conditions the  
260 | BFR particles were shut off so that the particles nucleated in PhoFR could be measured without interference from the seed  
particles and their sizes are shown as the orange square and green triangle. The solid red line is representative of the  
leading-edge diameters from our previous work with PhoFR [Hanson et al. 2019].

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The increase in diameter of the external nano-particles has roughly the same dependence on Q<sub>4</sub> as do particles nucleated in  
PhoFR. The size of the nucleated particles vs. Q<sub>4</sub> has been variable but it has not changed significantly over time; the data  
265 | from Hanson et al. [2019] compares well with the present nucleation results, see the Supplement. This indicates that growth  
conditions are similar in PhoFR whether particles are pre-formed or are nucleated there (at least for those at the leading  
edge). The good correlations between Q<sub>4</sub>, the leading edge size of nucleated particles, and calculated H<sub>2</sub>SO<sub>4</sub> [Hanson et al.

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2019] is supported here. Since knowledge of the photo-chemistry in PhoFR has improved dramatically, this correlation is explored further in the next paragraph.

The increase in the size of the particles is directly linked to their exposure to  $\text{H}_2\text{SO}_4$ , and we analyze the data at  $Q_4 = 4.2$  sccm in a simplified manner to obtain an average  $\text{H}_2\text{SO}_4$ . Fig. 4 shows a change in diameter for the initial 2-to-3 nm particles of 6.5 nm (+/- 1 nm) and the total residence time along the flow reactor axis is 22.7 s (from  $Z = 0$  to 125 cm and assuming an axial flow velocity of 5.5 cm/s for fully developed laminar flow.) In the Supplement we adapt the growth equation of Verheggen and Mozurkewich (2002) by augmenting the first term with a  $(1 + d_{\text{SA}}/D_p)^2$  factor [Lehtinen and Kulmala, 2003] accounting for the size of the  $\text{H}_2\text{SO}_4$  molecule with diameter  $d_{\text{SA}}$ . This augmentation factor can be viewed as reconciling gas-kinetic molecule-cluster collision rates (e.g., McMurry [1980]) with particle-molecule collision rates (e.g. condensational growth rates of Fuchs and Sutugin [1970]). The simplification is to assume constant values for the size-dependent terms in equation (S4) (see Supplement for a detailed derivation.) The observed growth at  $Q_4 = 4.2$  sccm requires an average  $\text{H}_2\text{SO}_4$  concentration of  $8.6 \times 10^9 \text{ cm}^{-3}$  for the 22.7 s the particles spend in PhoFR. The Supplement includes changes in the size-dependent terms in the growth calculation and the resulting average  $\text{H}_2\text{SO}_4$  is within 5 % of this value.

### 3.3 HONO photolysis rate

This average  $\text{H}_2\text{SO}_4$  concentration along with the measured  $\text{NO}_x$  and NO was used to estimate the HONO photolysis rate: the photolysis rate was varied until the growth-increment-derived average  $\text{H}_2\text{SO}_4$  concentration of  $8.6 \times 10^9 \text{ cm}^{-3}$  was achieved. Firstly, because our earlier UV absorption measurements (supplement of Hanson et al. [2019]) showed the probable presence of  $\text{NO}_2$ , we assume  $\text{NO}_2$  produced in the HONO self-reaction did not react further with HONO: the measured  $\text{NO}_x$  is taken to be 80 % HONO, 10 % NO and 10 %  $\text{NO}_2$ . The blue profile in Fig. 3 was calculated with a photolysis rate of  $4 \times 10^{-4} \text{ s}^{-1}$  and yields an average on-axis  $[\text{H}_2\text{SO}_4]$  of  $8.6 \times 10^9 \text{ cm}^{-3}$ . This HONO photolysis rate of  $4.0 \times 10^{-4} \text{ s}^{-1}$  is  $\frac{1}{2}$  of that from ( $8 \times 10^{-4} \text{ s}^{-1}$ ) derived in our previous work. Note that the previous work assumed zero NO was entering with HONO and a reaction between  $\text{HO}_2$  and  $\text{SO}_2$  was posited which lead to a calculated average  $\text{H}_2\text{SO}_4$  of  $7.4 \times 10^8 \text{ cm}^{-3}$  for  $Q_4 = 4.2$  sccm, only about 14 % less than the present calculations yield. The supplement presents a comparison of these two  $\text{H}_2\text{SO}_4$  calculations as well as the potential effect of the radial gradients in  $\text{H}_2\text{SO}_4$ .

We have made two important assumptions regarding  $\text{H}_2\text{SO}_4$  molecules colliding with particles: that the mass accommodation coefficient is unity and there is no van der Waals enhancement of the collision rate. These are reasonable assumptions but are not known with certainty; if not true the estimated photolysis rate could be significantly impacted. We note that some recent work [Stolzenburg et al. 2020] suggests that the  $\text{H}_2\text{SO}_4$  collision rate with particles is affected by van der Waals forces but we note that the uncertainty is large. In section 3.8 we present some consequences on cluster free energies if there is an enhanced collision rate.

With this new knowledge (NO entering with HONO) and the new estimate for the HONO photolysis rate and assuming no enhanced van der Waals forces, the calculated  $\text{H}_2\text{SO}_4$  concentration at the midpoint of the reactor ( $Z = 60 \text{ cm}$ ,  $R = 0 \text{ cm}$ ) is  $8.8 \times 10^9 \text{ cm}^{-3}$  for  $Q_4 = 4.2$  sccm. Note that even though the fraction of  $\text{NO}_x$  that is NO depends on  $Q_4$ , the relationship

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between  $Q_4$  and  $\text{H}_2\text{SO}_4$  in PhoFR is linear although not strictly proportional (see the Supplement for model calculated dependencies.)

### 3.4 UCPC vs. DEG

Shown in Fig. 5(a) are the results of nucleation studies in PhoFR with total particle number densities,  $N_p$ , measured with both the DEG system and the UCPC plotted vs.  $Q_4$ , the flow through the HONO generation system. The two counters are in near agreement for large values of  $Q_4$  but at 4.2 sccm and below there is a discrepancy that grows as  $Q_4$  decreases, reaching over a factor of ten at  $Q_4 = 2.5$  sccm. A potential cause is particles formed in ion-mediated processes, either with ambient ionization in the flow reactor or in the charger for the DEG system:  $\text{H}_2\text{SO}_4$  vapor carried into the charger yields  $\text{HSO}_4^-$  ions that can grow larger by accumulating  $\text{H}_2\text{SO}_4$  molecules, perhaps stabilized with water molecules or impurity bases. Either source of these ions could lead to significant artifact  $N_p$  in the DEG system analysis because large correction factors are applied ( $\sim 100$ ) because it is assumed that small neutral particles must be charged.

On the other hand, for 3 nm diameter or larger particles, the UCPC counts particles with roughly equal efficiency whether charged or not. Binary ion-mediated nucleation in PhoFR under ambient ionization conditions, leads to predicted  $N_p$  (rate times 5 s; Yu et al. [2020], blue diamonds, and Merikanto et al. [2016], open diamonds) that suggests our UCPC data in Fig. 5(a) was affected by ion processes, and significantly at low  $Q_4$  where the neutral nucleation rate is expected to be low. For the ion-mediated theoretical predictions [Yu et al. 2020; Merikanto et al. 2016],  $\text{H}_2\text{SO}_4$  concentration was taken to be  $(Q_4/\text{sccm}) \times 2.2 \times 10^9 \text{ cm}^{-3}$ . They compare favorably to the lowest UCPC  $N_p$ , and it is possible that an ambient ionization rate of 2 ion pairs  $\text{cm}^{-3} \text{ s}^{-1}$  (as used here) produced a level of 10 particles per  $\text{cm}^3$  in PhoFR. The peach line is  $N_p$  from simulations of neutral nucleation using a quasi-unary thermodynamic scheme NH3\_D52, see below. The blue triangles are the simulated  $N_p$  of large negative ions (9  $\text{H}_2\text{SO}_4$  molecules plus bi-sulfate) using quasi-unary thermodynamics deduced from Froyd et al. [2003] (see Supplement). The Froyd et al.  $\text{HSO}_4^-(\text{H}_2\text{SO}_4)_n$  thermodynamics supports a role for ion processes in our experiment, albeit a less active one than is suggested by the theoretical work of Yu et al. and Merikanto et al.

How does this affect the interpretation of the  $N_p$  vs.  $Q_4$  results for the nominal binary system? Firstly, the DEG  $N_p$  at low  $Q_4$  should be disregarded; in addition, the UCPC  $N_p$  at low  $Q_4$  should be considered to be affected by ion-mediated nucleation. The dependence of nucleation rates on  $\text{H}_2\text{SO}_4$  levels ( $Q_4$ ) supports these assertions. A power dependency of  $N_p$  on  $Q_4^6$  is shown in the figure. It is representative of the high  $Q_4$  data and extrapolation to low  $Q_4$  highlights that other nucleation processes are in effect at low  $Q_4$ , even for the UCPC. This is based on the expectation that the power relationship should steepen, at least not be weakened, as  $\text{H}_2\text{SO}_4$  decreases. This is because the critical cluster must increase in size to have a decreased evaporation rate as  $\text{H}_2\text{SO}_4$  decreases. Evaporation rates generally decrease as cluster size increases, as it does in the liquid drop model and in quantum chemistry calculations. Some experiments also show such behavior (e.g. Ball et al. [1999], Wyslouzil et al. [1991]) as does the simulated  $N_p$  labeled D52.

The bias in the low  $Q_4$  data for the UCPC could be due to a source of particles other than the ion-mediated process discussed above. For example, dimethylamine at a mole fraction of  $1 \times 10^{-16}$  leads to simulated  $N_p$  of  $800 \text{ cm}^{-3}$  at  $Q_4 = 4.2$  sccm (see

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Supplement). Another possibility is that the UCPC detects  $\text{H}_2\text{SO}_4$  and clusters to a small degree:  $[\text{H}_2\text{SO}_4]$  in the  $10^9 \text{ cm}^{-3}$  range but detected at an extremely small efficiency, say  $10^{-8}$ , could still result in 10s of Hz of count rates (thus 10s of particles /  $\text{cm}^3$ ). Other clusters such as the dimer and trimer, although likely present at lower levels,  $\sim 10^5 \text{ cm}^{-3}$ , might also give significant counts if efficiencies climb rapidly with size.

The UCPC  $N_p$  is usually lower than the DEG  $N_p$  as demonstrated in Figs. 5(a) and 6(b) and in the Supplement. Yet there was a period of time (see the Supplement) where both had elevated count rates and in fact the UCPC  $N_p$  was at times greater than the DEG  $N_p$ . During this time, the data in Fig. 6(a) for measurements with  $\sim 650 \text{ ppt NH}_3$  added were taken where UCPC  $N_p$  are as much as a factor of three greater than the  $N_p$  in the leading edge of the DEG system's size distribution. The reason for this behavior is not known. Further discussion of this time period is presented in the Supplement.

### 3.5 Nominal binary nucleation results, then and now

Assuming that the UCPC  $N_p$  at  $Q_4 = 4.2 \text{ sccm}$  in Fig. 5(a) is indicative of the binary system, an estimate for the binary nucleation rate  $J_{\text{bin}}$  is  $50 \text{ cm}^{-3} \text{ s}^{-1}$ , converting  $N_p$  to  $J_{\text{bin}}$  by dividing by an estimated nucleation time of 5 s, the transit time over a 30 cm length of the reactor centered at 60 cm. This assumes that the majority of the large particles are formed near the midpoint of the reactor. Particles are also likely to form downstream of this region but they will be somewhat smaller than the leading edge particles. Nevertheless, the sulfuric acid concentration for this  $J_{\text{bin}}$  we take to be the simulated value at 60 cm and on centerline,  $[\text{H}_2\text{SO}_4] = 8.8 \times 10^9 \text{ cm}^{-3}$ .

This so-called midpoint nucleation rate is a metric that has limited use but does serve as a crude point of comparison. A nucleation rate of  $50 \text{ cm}^{-3} \text{ s}^{-1}$  at  $8.8 \times 10^9 \text{ cm}^{-3} \text{ H}_2\text{SO}_4$  (52 % RH, 296 K) is lower than some recent binary system (nominal) results: the Zollner et al. [2012] bulk flow reactor result for these conditions (as extrapolated in Hanson et al. [2019]) is  $300 \text{ cm}^{-3} \text{ s}^{-1}$  and the SAWNUC predicted rate is  $230 \text{ cm}^{-3} \text{ s}^{-1}$  [Ehrhart et al. 2016], extrapolated from  $J_{\text{bin}} = 3 \text{ cm}^{-3} \text{ s}^{-1}$  at  $5 \times 10^9 \text{ cm}^{-3}$ , 292 K and 38 % RH. See the Supplement for model predicted  $N_p$  from SAWNUC and from two 52% binary system thermodynamics used here.

The overall changes in the DEG system measurements in PhoFR are presented in Figure 5(b), a plot of experimental nucleation rates plotted as a function of sulfuric acid concentration. The previous data from PhoFR (Hanson et al. [2019]) are the blue squares, the present data (Dec. 2018 to Feb 2019 excluding measurements just after cylinder changeovers) are plotted as the blue open and filled circles and data from the CLOUD experiment are plotted as diamonds. A midpoint nucleation rate metric is consistent with our assertion that the reactor has gotten cleaner, as the onset of nucleation has likely moved down the reactor with increased  $\text{H}_2\text{SO}_4$ . Since the reference point has increased between then and now, the present measurements are plotted with both open (reference point at 30 cm, as in Hanson et al. [2019]) and filled (60 cm) blue circles; the arrows in the plot illustrate the overall evolution of particle formation in PhoFR. The relationship between [SA] at 60 cm and  $Q_4$  was taken from the model results (plot shown in the Supplement.) These estimates reflect the changes in the experiment and illustrate that the present experimental nucleation rates are some of the lowest measured in photolytic

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- Deleted: ; integrating over the size distribution reduces the discrepancy to about a factor of 2. We note that the On another occasion, we noticed that the UCPC  $N_p$  decreased considerably during the course of a day's experiment. We think these discrepancies may be related to occasional exposure of the UCPC sampling line to room air, to get the reference pulse heights for large particles, that might lead to variable amounts of cluster/particle formation along the sampling line.
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systems, assuming extrapolations with the  $[\text{H}_2\text{SO}_4]^6$  relationship hold true. The 2D model results using the NH3\_D52 thermodynamics (zero ammonia added yields binary nucleation) are also plotted in Fig. 5(b) and support such a relationship.

The early bulk studies of Wyslouzil et al. Viisanen et al. and Ball et al. report nucleation rates, extrapolated to the present conditions, that are lower than the present work; consideration of the uncertainties in the sulfuric acid concentrations mitigates these differences as discussed in Zollner et al. [2012]. The Zollner et al. [2012] results for 40 % RH fall nearly on top of the power 6 dependency line in Fig. 5b. Comparisons with the recent photolytic studies by Yu et al. [2017] and Tiszenkel et al. [2019] are complicated by the assumed presence of amines evidenced by high nucleation rates at very low  $[\text{H}_2\text{SO}_4]$ . Note that the 292 K CLOUD binary system measurements reported by Ehrhart et al. [2016] are much higher than the SAWNUC predictions; it was believed to be affected by residual ammonia thus not considered representative of the putative binary system. We discuss the effect of ammonia on the 292 K CLOUD data below.

### 3.6 Addition of base

Ammonia was added to the system in a port at the top of PhoFR just above the illuminated section, as described in Hanson et al. [2019]. The results from both counters are shown in Fig. 6(a) where  $N_p$  is plotted versus the  $\text{NH}_3$  mixing ratio for  $Q_4 = 3.0$  sccm (interpolated data). Upon addition of ammonia to the flow reactor, nucleation increases and  $N_p$  climbs at a roughly squared power dependence on ammonia. We performed simulations with  $\text{NH}_3\text{-H}_2\text{SO}_4$  thermodynamics for molecular clusters; two different sets of free energies were used and results are shown as the dashed lines in Fig. 6(a). To illustrate an effect of a contaminant or other particle formation process not included in our simulations, also plotted are net  $N_p$  (x and + symbols) where  $N_p$  for zero added  $\text{NH}_3$  was subtracted from the data.

The model results plotted as the green dashed line use the thermodynamics that fit our 2018 data at 52 % RH (NH3\_52, Hanson et al. [2019]); it leads to significant over-prediction probably due to slightly over-strong binding energies. This scheme was developed from data that was influenced by an arguably higher level of contaminants in the system at that time (March to June 2018) compared to now. It may have also been affected by not including any HONO decomposition nor initial NO among other differences in the assumed photochemistry that led to calculated  $\text{H}_2\text{SO}_4$  on average about 14 % lower than the present calculations yield (see Supplement). The new set of  $\text{NH}_3\text{-H}_2\text{SO}_4$  cluster thermodynamics for 52 % RH was developed to better model the current data. The model results using this new set (NH3\_D52) are plotted as the dashed orange line; they secure the lower edge of the range of data and exhibit a similar power dependency, 1.7, on  $\text{NH}_3$  to that exhibited in the data.

Comparison to our earlier measurements (Hanson et al. 2019), encompassed by the gray quadrilateral (for  $Q_4 = 2.1$  to 4.2 sccm) in Fig. 6(a), reveal that the measurements at the highest ammonia levels have some overlap while the low ammonia results are disparate. The previous measurements have a near-unity power dependency on  $\text{NH}_3$  which is also exhibited in many other experiments [Kirkby et al., 2011; Kürten et al., 2016; Benson et al., 2011]. The current measurements have a power dependency on  $\text{NH}_3$  near 2 (for net  $N_p$ ) which is also exhibited in theoretical predictions (such as Kürten et al. [2016].)

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**Deleted:** The UCPC vs. DEG  $N_p$  relationship has not been consistent over time. Data shown in Figs. 5 and 6(b) show that the UCPC  $N_p$  is generally lower than the DEG system  $N_p$ . Yet in recent data, e.g. the Fig. 6(a) data for measurements with ~ 650 ppt  $\text{NH}_3$ , added, UCPC  $N_p$  are as much as a factor of three greater than the  $N_p$  in the leading edge of the DEG system's size distribution; integrating over the size distribution reduces the discrepancy to about a factor of 2. On another occasion, we noticed that the UCPC  $N_p$  decreased considerably during the course of a day's experiment. We think these discrepancies may be related to occasional exposure of the UCPC sampling line to room air/the reference that might lead to variable amounts of cluster/particle formation along the sampling line. ¶

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635 Another test of the  $\text{NH}_3\text{-H}_2\text{SO}_4$  cluster free energies is the variation of  $N_p$  with  $\text{H}_2\text{SO}_4$ . Shown in Fig. 6(b) are measured  $N_p$  from the particle counters vs.  $Q_4$  with 360 pptv  $\text{NH}_3$  added. Note the large decrease in the current  $N_p$  from that in June 2018 (230 pptv  $\text{NH}_3$ ) [Hanson et al. 2019]. The modeled  $N_p$  traverses the current data but shows a larger dependence on  $Q_4$  ( $Q_4^{5.7}$ ) than is exhibited in the data ( $Q_4^{3.8}$ ). This could be due to a deficiency of these thermodynamics and/or the presence of a contaminant that more heavily influences the measurements at low  $N_p$  and  $Q_4$ .

640 A synergistic effect between amines and ammonia [Yu et al. 2012; Glasoe et al. 2015] may have affected the measurements if contaminant amine compounds were present. Although it is difficult to ascertain such an effect in our data it might provide an explanation for the large scatter in the data in Fig. 6(a) that does not seem to decrease at the highest  $\text{NH}_3$ . This effect would be of a different nature than the contaminant effect discussed in the previous paragraph. This tightly-wound story of contaminants is speculative, considering the lack of direct knowledge of the contaminant and the little-understood synergistic effect that is just recently receiving theoretical scrutiny [Temelso et al.; Wang et al. 2018; Myllys et al. 2019]. 645 The lower envelope of the  $N_p$  vs.  $\text{NH}_3$  data in Fig. 6(a) is adequately represented by the model and the data in Fig. 6(b) indicate that the effects of other processes probably affected the measurements the most at low  $N_p$ , using the model results as a gauge. The high  $N_p$  data may very well represent the  $\text{NH}_3\text{-H}_2\text{SO}_4$  system at 52% RH. This is supported by the Glasoe et al. finding that the synergistic effect dissipates as the ammonia abundance increases beyond a few hundred pptv  $\text{NH}_3$ .

650 Clearly, the thermodynamics for the binary system ( $\text{H}_2\text{O-H}_2\text{SO}_4$ ) are important for interpreting experimental results. They provide a reference for assessing whether a contaminant has significantly affected a measured nucleation rate and also for determining the effect of added base on nucleation rates. Please see the Supplement for more discussion of the binary system experiments and a comparison of free energies.

While the effects of contaminants in PhoFR have decreased since our earlier measurements were published, the episodic spikes in  $N_p$  (Fig. 1) indicate an intermittent appearance of something that boosts a nucleation process other than the binary process. Since we cannot ensure this other process has no effect at high  $Q_4$ , the  $\text{NH}_3\text{\_D52}$  thermodynamics must be 655 considered a phenomenological description; at zero ammonia it represents a limit on the pure binary system at 52% RH. Nonetheless, ammonia's influence on cluster energetics is similar in these two sets of  $\text{NH}_3\text{-H}_2\text{SO}_4\text{-52\% RH}$  free energies ( $\text{NH}_3\text{\_52}$  and  $\text{NH}_3\text{\_D52}$ ). For example, the step-wise free energies for clusters up to 5 acid and 3 ammonia molecules differ 660 by less than 1 kcal/mol. In the supplement is a comparison plot of the step-wise free energy changes for acid addition for both free energies.

We argue that the  $\text{NH}_3\text{\_D52}$  free energy landscape can be ~~considered nearly free of the effect of a contaminant and of ion-~~ mediated processes because these rogue processes operated most significantly at low ammonia and  $Q_4$ . Then  $\text{NH}_3\text{\_D52}$  free energies can be used in box model simulations to, in effect, extrapolate our data to the experimental conditions of other data 665 sets. This was done previously [Hanson et al. 2019, Fig. 10] using the  $\text{NH}_3\text{\_52}$  free energies to compare to some  $\text{NH}_3\text{-H}_2\text{SO}_4$  experimental results [Berndt et al. 2010; Dunne et al. 2016]. Part of that plot is reproduced in Fig. 7 focusing on the  $[\text{H}_2\text{SO}_4]=1.5\times 10^8\text{ cm}^{-3}$  data.  $\text{NH}_3\text{\_D52}$  free energies result in nucleation rates about an order of magnitude below those calculated with  $\text{NH}_3\text{\_52}$ . Our conclusions change little from 2019: the Berndt et al. data (2 to 70 parts per million  $\text{NH}_3$ ) and

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670 the CLOUD data at  $> 200$  pptv  $\text{NH}_3$  stand apart from all the previous work in this system and are most consistent with data from PhoFR.

The predicted ternary homogeneous nucleation rates of Yu et al. [2020], J(THN) in Fig. 7, are quite different from the current predictions (which are thought to be free of ionization effects) and both of these differ from the neutral ternary system rates predicted by Kürten [2019] (dashed red line). When ionization rates characteristic of ambient levels due to galactic cosmic rays (GCR) are present, the Yu et al. [2020] ternary ion-mediated nucleation rates (TIMN) approximate the CLOUD data. Since Kürten et al. [2019] derived neutral cluster free energies to explain the neutral rates measured in CLOUD, they are unsurprisingly close to Yu et al.'s TIMN predictions. Yet the purported chemistries of these two schemes are quite different.

One point to add to the comparisons depicted in Fig. 7 is that the putative binary data from CLOUD, thought to be due to a 4 pptv  $\text{NH}_3$  impurity, departs significantly (many orders of magnitude) from simulated data using  $\text{NH}_3\_D52$  for 4 pptv  $\text{NH}_3$ . The decreasing discrepancy with  $\text{NH}_3\_D52$  as ammonia increases, gray quadrilateral encompassing the CLOUD data, is reminiscent of the relationship in Fig. 6(a) of our earlier data to the present data. This may indicate a similar situation in the two experimental apparatus where a low-level contaminant (that is now after years of use apparently much lower in PhoFR) that is relatively potent leads to particle formation rates well above what is expected. A synergistic effect may also be in play here with a strong nucleator such as dimethylamine: our box model calculation using DMA\_I thermodynamics [Hanson et al. 2017, 2019] and dimethylamine at  $10^{-3}$  pptv yields a nucleation rate of  $10^{-3} \text{ cm}^{-3} \text{ s}^{-1}$  for the conditions in Fig. 7. A plausible scenario is that the synergistic effect of 4 pptv  $\text{NH}_3$  boosts this to  $10^{-2} \text{ cm}^{-3} \text{ s}^{-1}$  and that the synergistic effect increases with  $\text{NH}_3$  level (or the exceedingly low level of dimethylamine increases slightly as  $\text{NH}_3$  is increased). By sheer abundance,  $\text{NH}_3$  could come to dominate the amino-containing moieties in the CLOUD mass spectra (e.g. Schobesberger et al. 2015; Almeida et al. 2013).

### 3.7 Variation of water content

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Relative humidity was varied by adjusting the ratio of humidified to dry  $\text{N}_2$  flows ( $Q_3$ , flow through the humidifier, and  $Q_2$  dry  $\text{N}_2$  flow) maintaining a constant total flow rate of 2.9 sLpm. Representative data for a constant HONO level (given by  $Q_4=5.25$  sccm) vs.  $Q_3$  are plotted in Fig. 8 along with the fit to our previous data at  $Q_4=4.2$  sccm [Hanson et al. 2019]. Despite a large decrease in  $N_p$  at a given RH between then and now, the variation of  $N_p$  vs. humidity has a similar dependence on RH as we observed previously (Hanson et al., 2019): a power dependence of 4.75 is shown in the figure. For two sets of data, the DEG system  $N_p$  (leading edge) are also shown with good agreement at high  $N_p$  while a bias looks to be present in the DEG results at low RH and  $N_p$ .

There is a considerable level of variability in the UCPC data that may reflect changes in a contaminant in PhoFR. Yet there is a sensitivity to the temperature of the room (monitored on the surface of the glass cone at the top of the reactor,  $T_{\text{cone}}$ ): on the 22<sup>nd</sup> of April, the  $Q_3=2.3$  sLpm UCPC data had  $T_{\text{cone}} = 23.6$  C for the  $5 \times 10^4 \text{ cm}^{-3}$  data point and it was 25 C for the  $6 \times 10^3$

cm<sup>-3</sup> data. The Supplement has a temperature sensitivity example where time series of  $N_p$  and  $T_{\text{cone}}$  are shown. Another set of RH dependency data for  $Q_4 = 4.2$  sccm is also shown in the supplement.

Over a comparable range of RH, Zollner et al. [2012] reported a dependence of  $N_p$  on RH to the 5-to-9 power, with the higher value for RH > 50%. There are significant differences between these two experiments that might affect RH dependencies. For example, Zollner et al. had a bulk source for H<sub>2</sub>SO<sub>4</sub> and a heated mixing region so that the 6 sLpm total flow and the H<sub>2</sub>SO<sub>4</sub> it was carrying underwent cooling as particles nucleated. Thus the Zollner et al. H<sub>2</sub>SO<sub>4</sub> underwent a wide range of hydration due to the 17 K cooling. This change in hydration affects its rate of diffusion and thus loss to the wall; cooling is generally 2 K or less in the present work and there is a steadier hydration level for H<sub>2</sub>SO<sub>4</sub>. Nevertheless, the scatter in the data shown in Fig. 8 precludes putting much significance on the  $RH^{4.75}$  dependence and its comparison to the Zollner et al. RH dependency.

The main conclusion is that  $N_p$  is much lower at all RH compared to our previous work [Hanson et al, 2019]. An ion-mediated process may play a small role that affects measurements at low RH and thus  $N_p$ , but since most of the  $N_p$  are much greater than the  $\sim 10$  cm<sup>-3</sup> expected for this process, the results are dominated by changes in water vapor. But a contaminant may have affected the present data as was postulated for the previous data [Hanson et al. 2019] where a level of 0.6 pptv methylamine was consistent with the  $N_p$  vs.  $Q_4$  data. If methylamine is also present here we estimate that its level has dropped to 0.1 pptv (using the squared power dependence on methylamine found for nucleation rate [Glasoe et al., 2015].) Since the identity and the source of the contaminant are largely unknown we can only speculate here but we point out that a level of dimethylamine of  $10^{-4}$  pptv is shown in the supplement to be consistent with the variation of  $N_p$  with SO<sub>2</sub>. In terms of RH dependencies, primary alkylamines such as methylamine may be influenced by water content while dimethylamine is not (at least at the 2 pptv level, Hanson et al. [2019].)

We varied RH with added ammonia (120 pptv) to see how water influences nucleation in the NH<sub>3</sub>-H<sub>2</sub>O-H<sub>2</sub>SO<sub>4</sub> system. These data,  $N_p$  vs.  $Q_3$ , are plotted in Fig. 9 (orange symbols) along with the Feb-14 data without ammonia added from Fig. 8.

There is a significant dependence of  $N_p$  on RH for the added ammonia case. The enhancing effect of NH<sub>3</sub> (ratio of  $N_p$  with added NH<sub>3</sub> to the nominal binary case) appears to be less at high RH: a factor of  $\sim 4$  at  $Q_3 = 2$  and 2.35 slpm while it is a factor of ten at  $Q_3 = 0.65$  and 1.1 sLpm. This may reflect the overwhelming abundance of H<sub>2</sub>O vs. NH<sub>3</sub>: clusters take what there is more of by opportunity (entropy) rather than by binding energy (enthalpy). H<sub>2</sub>SO<sub>4</sub> cluster affinity for NH<sub>3</sub> is greater than it is for H<sub>2</sub>O, but the stepwise free energy difference is less than a few kcal/mol for clusters containing more than 5 H<sub>2</sub>SO<sub>4</sub> molecules.

Previous work showed similar behavior. Ball et al. showed a larger enhancement due to added ammonia for measurements at 5 % RH than at 15 % RH. Benson et al. [2009] also reported ammonia enhancement factors that increased with decreasing RH but were also dependent on other experimental conditions, nonetheless: for 20 ppbv NH<sub>3</sub> and an H<sub>2</sub>SO<sub>4</sub> concentration of  $1.2 \times 10^9$  cm<sup>-3</sup>, the enhancement factor increased from 20 to 400 when the RH decreased from 8 to 4 %.

Although CLOUD experiments (Kürten et al [2016]; Kirkby et al. [2011]; Almeida et al. [2013]) did not systematically investigate an RH dependence for the ternary system, Dunne et al. explored the effect of RH on ammonia-sulfuric acid

nucleation in a global climate model and saw an effect of up to 34 % in the number density of 3 nm particles at cloud height. The Dunne et al. ad hoc RH dependency is shown in Fig. 9 normalized to our measured  $N_p$  at  $Q_3=1.05$  sLpm (the green dashed line also brushes our model calculated  $N_p$ , NH3\_D52, at  $Q_3=1.50$  sLpm.) It describes the observed increase in particle numbers with RH fairly well but has difficulties below about 25 % RH.

### 3.8 van der Waals Interactions.

To demonstrate the potential influence of enhanced collision rates of hydrated  $H_2SO_4$  molecules with clusters, the model was run with collision rates doubled and the photolysis rates halved. These two changes would result in equivalent growth conditions for the external particles as discussed in section 3.2 as  $H_2SO_4$  concentrations decrease by a factor of two but their uptake rate onto particles is doubled. A doubling of the collision rates is consistent with the findings of Stolzenburg et al. [2020] but we note an alternate analysis that resulted in a smaller effect was put forth in the referee comments. The effects on calculated  $H_2SO_4$ , the 6 acid cluster, and the number of particles for binary nucleation at  $Q_4=4.2$  sccm are displayed in Table 1.

The first row presents results using the hard sphere collision rates and thermodynamics NH3\_D52 as presented in Fig. 5(a) where the simulated  $N_p$  is  $96 \text{ cm}^{-3}$ . Using the same thermodynamics but halving the HONO photolysis rate while doubling all collision rates, the simulated  $N_p$  drops by 95 % (second row). Therefore, at these lower sulfuric acid levels, the thermodynamics must be strengthened to increase nucleation rates. Coincidentally, using our previous thermodynamics, NH3\_52 [Hanson et al. 2019] predicted  $N_p$  increases to  $125 \text{ cm}^{-3}$ . The point here is that the main effect of enhanced collision rates is that the cluster thermodynamics must be strengthened by a few kcal/mol (see the Supplement) to get simulation-experiment agreement. The other effect is that the mid-point nucleation rate metric must be altered: the effective  $H_2SO_4$  concentration would decrease by about 50 %.

Table 1. Effect of doubling forward rate coefficients due to enhanced van der Waals interactions on clusters in binary homogeneous nucleation. The two thermodynamic schemes for binary conditions are plotted in the Supplement.

$k_{\text{phot}}$	factor	$[H_2SO_4]_{\text{ave}}$	Thermo.	SA	$N_p$
$4.0 \times 10^{-4}$	1	$8.6 \times 10^9$	NH3_D52	2.6	96
$2.0 \times 10^{-4}$	2	$4.8 \times 10^9$	NH3_D52	0.15	5.5
$2.0 \times 10^{-4}$	2	$4.8 \times 10^9$	NH3_52	2.5	125

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#### 4. Summary and Conclusions

The overarching goal is to improve the accuracy of nucleation rates calculated for atmospheric conditions. Which we believe we are doing—improving the measurements, identifying specific effects of contaminants, refining the thermodynamics of clusters—and then testing the results of our approach against previous work through detailed comparisons.

We have a better understanding of the chemistry within PhoFR. We detected small amounts of NO produced in the source via the HONO self-reaction which significantly affects the photochemistry. We postulated that hard-to-detect amounts (e.g. sub- $10^{-15}$  mole fraction dimethylamine) of amines affected the measurements but their influence has decreased over time.

The deployment of two different types of particle detectors yielded information on nucleation processes other than the binary  $\text{H}_2\text{SO}_4\text{-H}_2\text{O}$  and ternary  $\text{NH}_3\text{-H}_2\text{SO}_4\text{-H}_2\text{O}$  systems. With the long-term measurements from PhoFR indicating the decrease of something like a contaminant, the potential role of ion-mediated nucleation processes [Yu et al., 2020; Lovejoy et al. 2004] was uncovered. We were able to show that ion-mediated nucleation can play a significant role at low nucleation rates in the binary system, whether ambient ionization in the flow reactor or potentially within the charger of the nano-MPS (DEG system).

The nucleation rates in the ammonia-sulfuric acid-water system measured in PhoFR have also decreased since our 2018 results [Hanson et al. 2019], consistent with the postulated decrease in an amine-type of contaminant. The free energy scheme for  $\text{NH}_3\text{-H}_2\text{SO}_4$  clusters developed here has ammonia's influence on the putative binary system about the same as in our previous work [Hanson et al. 2019] for clusters up to 5  $\text{H}_2\text{SO}_4$  molecules. Thus the cluster free energies with 6 or more sulfuric acid molecules were the most consequential. These changes in the largest clusters were sufficient to lead to agreement of simulations with experimental.

The effect of water on nucleation rates was explored and progress was made in decreasing the upper limit nucleation rates for the binary system. The free energies for the  $\text{NH}_3$ -free clusters in the  $\text{NH}_3\text{-D52}$  scheme presented here are within a few kcal/mole of the modified liquid drop model (SAWNUC) of Lovejoy et al. [2004]. The effect of water when ammonia was present is substantial and indicates that there may be a synergistic effect involving water on the cluster's free energies (and possibly a methylamine contaminant at the 0.1 pptv level.) Interestingly, the Dunne et al. ad hoc equation reasonably explains our measured relative humidity effect at 120 pptv  $\text{NH}_3$  and  $\sim 1 \times 10^{10} \text{ cm}^{-3}$   $\text{H}_2\text{SO}_4$ . Theoretical work on the effect of water in this ternary system is sparse but initial reports on the effect of water on the small clusters ( $\sim 3$   $\text{H}_2\text{SO}_4$  molecules) suggests small effects, even a decrease in nucleation as RH increases (see Fig. 9(b) of Henschel et al. [2016]). As we have argued here, it is likely that the effect of water in this ternary system is greatest in larger clusters, those with 6 or more  $\text{H}_2\text{SO}_4$  molecules.

The model's HONO photolysis rate was tuned to give  $\text{H}_2\text{SO}_4$  concentrations that explained the results of the nanoparticle growth studies. Collision rates of SA molecules (whether hydrated or not) with clusters were assumed to be free of van der Waals interactions and the mass accommodation coefficient  $\alpha$  was assumed to be unity [Hanson, 2005]. If van der Waals

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interactions are important and  $\alpha$  is unity then there is a lower level of SA in the flow reactor than the model was tuned to and the thermodynamics are not valid; binding energies are too small. On the other hand, if van der Waals interactions are not important and  $\alpha$  is less than unity, then our thermodynamics have binding energies that are too strong. Future work on the growth of larger nano-particles may shed light on this issue. Also, measurements of H<sub>2</sub>SO<sub>4</sub> uptake efficiency onto particles smaller than the 140 nm diameter particles that Hanson [2005] studied will provide additional information on the efficacy of van der Waals interactions on collision rates.

The next steps for PhoFR measurements include measuring the particle growth of externally generated nanoparticles as a function of the initial nanoparticle size and composition. These measurements will help tease out information on whether van der Waals interactions are important for uptake of H<sub>2</sub>SO<sub>4</sub>. Coupled with new measurements of the H<sub>2</sub>SO<sub>4</sub> uptake coefficient on sulfuric acid particles at sizes smaller than the 140 nm diameter particles used in Hanson [2005], these interactions can be further delineated. We also plan to investigate the effect on nucleation rates of adding methylamine to PhoFR.

**Data and Code availability:** Spreadsheets for the data in the figures are available by request. The code for the simulation of the flow reactor and the box model are written in Delphi and can be made available upon request.

**Author contribution:** SM and DH designed the experiments, carried them out, and analyzed the data. JK and DH developed the model code and performed the simulations. DH, JK and MW evaluated the photo-chemistry and interpreted the nucleation results. DH prepared the manuscript with contributions from JK and MW.

**Deleted:** We are working on phenomenological free energy schemes for methylamine-H<sub>2</sub>SO<sub>4</sub> clusters from the BFR results of Glasoe et al. These will be used to compare to measurements in this system and will also be used to estimate a potential level of methylamine contaminant that could have affected our nominal binary system measurements. We are also working on a phenomenological free energy landscape for H<sub>2</sub>O-H<sub>2</sub>SO<sub>4</sub> clusters that can bridge the relative humidity results, effectively providing a basis for our quasi-unary thermodynamics as well as providing composition information. ¶

## References

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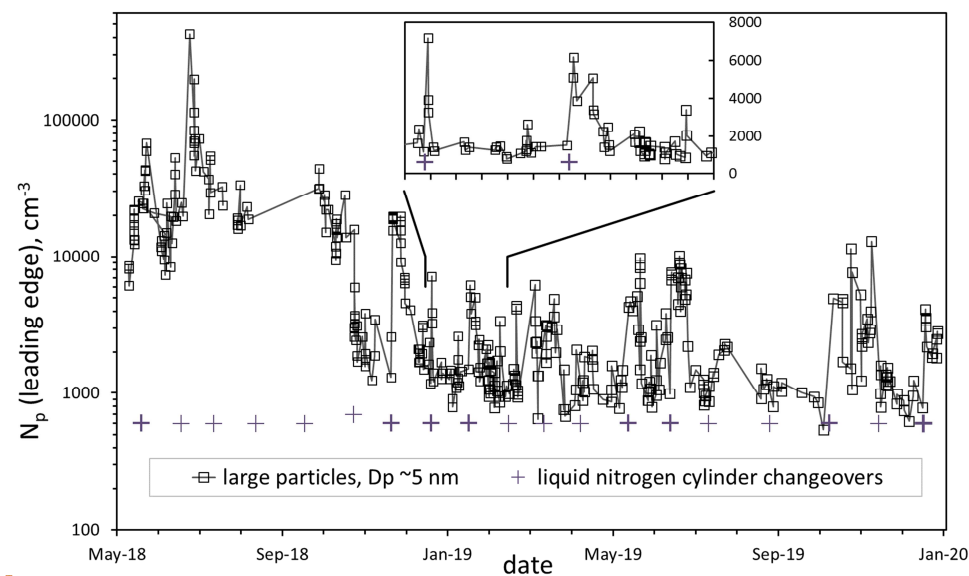


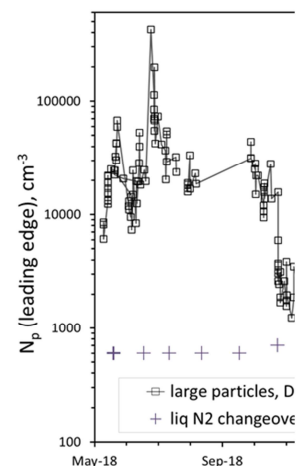
Figure 1: Particle number density in the leading edge of the DEG system distribution over time for standard conditions ( $Q_4= 4.2$  sccm, RH 52%, 296 K). Leading edge refers to all particles described by a log-normal distribution of the largest particles (roughly 5 nm mode diameter for these conditions). The approximately monthly liquid nitrogen cylinder changeovers are indicated by the + symbols; **bolded indicates associated spike in  $N_p$** . The inset shows a 60 day period of data on a linear scale.

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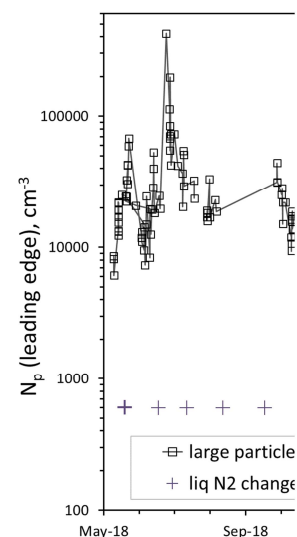
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Comment [R7]: The DEG capillary flow rate was monitored but was not applied for much of the data that was submitted in the first version. The corrections ranged from -38 to +8 % with an average of -10%.

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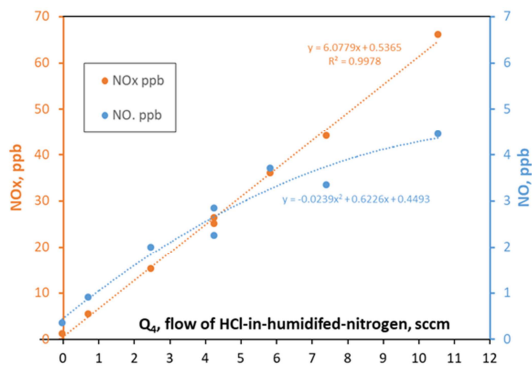


Fig. 2.  $\text{NO}_x$  (left axis) and  $\text{NO}$  (right axis) vs. flow rate  $Q_4$  of HCl-laden flow over  $\text{NaONO(s)}$ . Assuming that the  $\text{NO}_x$  level is related one-to-one to  $\text{HCl}$ , the  $\text{HCl}$  mixing ratio in flow  $Q_4$  is 17.6 ppmv. Note: sccm = standard  $\text{cm}^3$  per min.  $\text{NO}_x$  here includes HONO.

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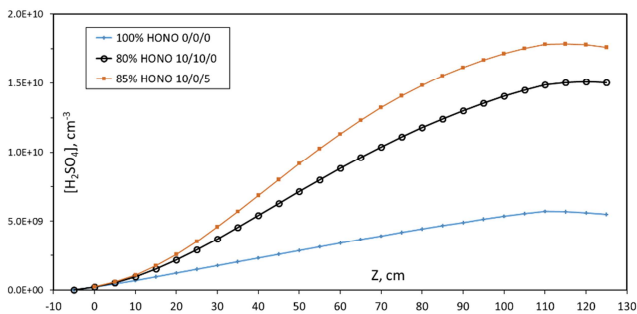
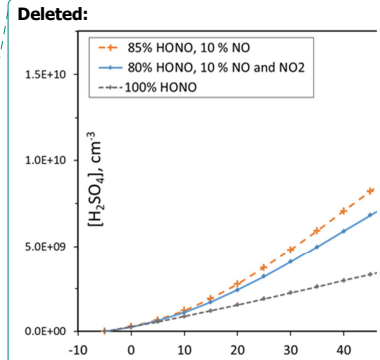


Fig. 3. Simulated centerline  $\text{H}_2\text{SO}_4$  vs. axial distance for three different HONO decomposition scenarios. Parameters: HONO photolysis rate of  $4.0 \times 10^{-4} \text{ s}^{-1}$ , 17 ppmv  $\text{NO}_x$  (+ HONO) in  $Q_4$  and no reaction between  $\text{HO}_2$  and  $\text{SO}_2$ . These simulations were run for  $Q_4 = 4.2 \text{ sccm}$ . The 100% level of HONO is a concentration of  $5.79 \times 10^{11} \text{ cm}^{-3}$  or ~25 ppbv.



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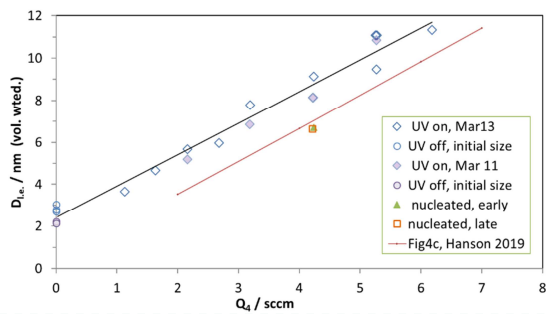


Fig. 4. Growth studies of externally produced nanoparticles,  $D_{1e}$  vs.  $Q_4$ .  $D_{1e}$  is the volume-weighted diameter of the leading edge of the size distribution. The nanoparticles initial size is indicated by the data plotted at  $Q_4 = 0$  sccm. The red line labeled Fig4c is from Hanson et al. [2019].

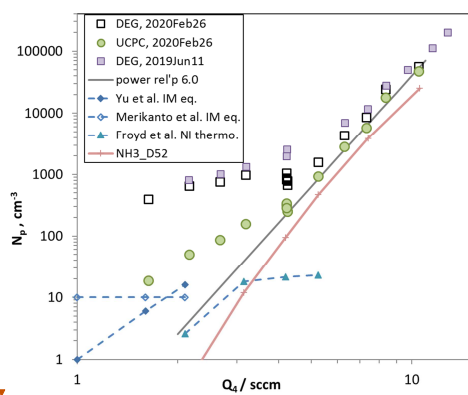
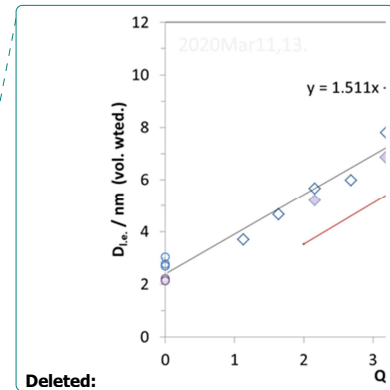
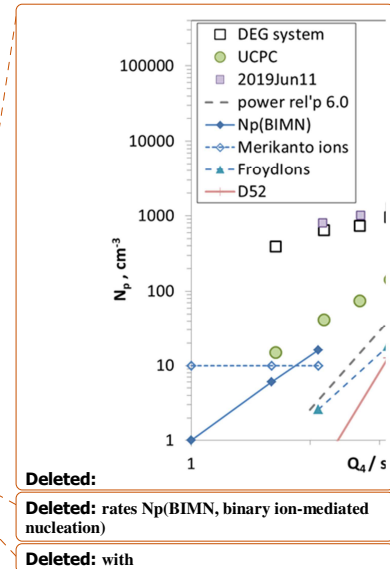


Fig. 5(a).  $N_p$  vs.  $Q_4$ , two different particle counting systems. DEG system refers to the nano-MPS followed by a diethyleneglycol CPC ( $N_p$  for the leading edge) and UCPC is a butanol-based detector. DEG system data from 2019 are also shown.  $N_p$  from ion-mediated (IM) nucleation from Yu et al. [2020] (BIMN, binary ion-mediated nucleation) and from Merikanto et al. [2016] (rates multiplied by 5 s nucleation time.) Also shown are  $N_p$  calculated in our 2D model for negative ions (NI, thermodynamics from Froyd et al. 2003, blue diamonds) and for neutral nucleation using NH3\_D52 thermodynamics (peach +). A power dependency of  $Q_4^6$  is also shown that describes the upper half of the UCPC data.



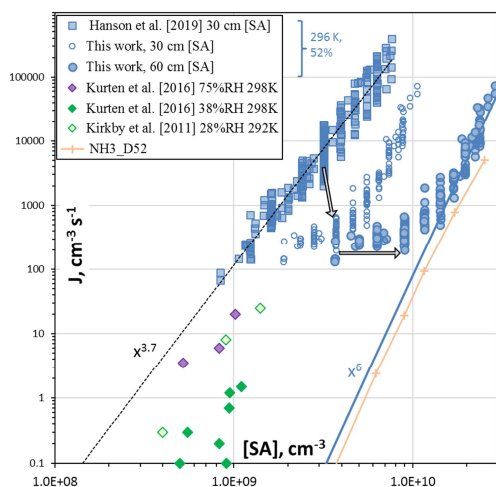


Fig. 5(b). Nucleation rate vs.  $\text{H}_2\text{SO}_4$  (SA) concentration. Nucleation rate given by total number of particles  $\geq 2.3$  nm in the DEG distribution divided by an estimated 5 s nucleation time. The [SA] concentration calculated at 30 cm was used to plot both the Hanson et al. [2019] data set and this work (open circles). The filled circles are the nucleation rates for this work but plotted at the 60 cm calculated [SA] concentration. The arrows indicate the system's evolution for results at  $Q_4=4.25$  sccm. Nucleation rates from the CLOUD experiment at 292 and 298 K for nominal binary conditions are also shown. The nucleation rate calculated using the 2D model with NH3\_D52 thermodynamics, is also plotted (+) at 60 cm [SA].

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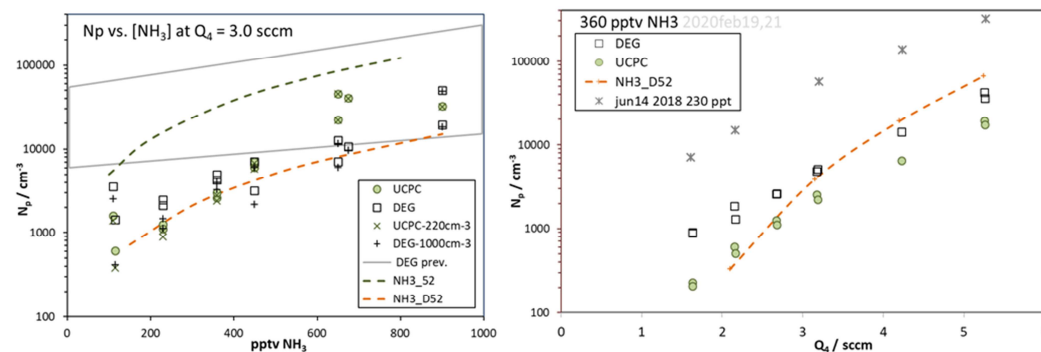


Fig. 6.  $N_p$  vs. (a) ammonia level and (b)  $Q_4$ . (a) Experimental data for the two particle detection systems, minus the zero-added base  $N_p$ , are plotted as the squares (DEG) and diamonds (M1) and the gross  $N_p$  are also shown. Model predicted  $N_p$  (binary  $N_p$  are negligible) are shown as the green (NH3\_52) and orange (NH3\_D52) dashed lines. Clusters up to 10 acid molecules and either 3 or 6 ammonia molecules were simulated; runs with 6 ammonia molecules were within 3 % of runs with 3. Supplement has further details of the model runs. Gray quadrilateral "DEG prev." encompasses data from Fig. 5 of Hanson et al. [2019]. (b) Symbols as in (a) with  $\text{NH}_3$  present at 360 pptv. Asterisks are previous data [Hanson et al., 2019] from 2018Jun14 at 230 pptv  $\text{NH}_3$ .

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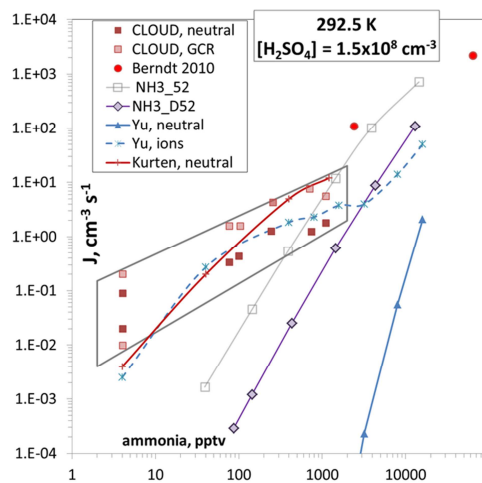


Fig. 7. Nucleation rate  $J$  versus ammonia level for  $[H_2SO_4]=1.5 \times 10^8 \text{ cm}^{-3}$  and 292.5 K. Berndt et al. 50% RH and CLOUD at 38 % RH (Dunne et al. [2016], also included are low ionization rate measurements, “GCR”; the data at 4 pptv  $NH_3$  were for  $H_2SO_4$  up to  $3 \times 10^8 \text{ cm}^{-3}$ ). Box model calculations (described in Hanson et al. [2019]) using two different thermodynamic schemes for 52 % RH are also shown (NH3\_D52 this work; NH3\_52 presented previously, Hanson et al. [2019]). Red solid line from Kürten et al. 2019. Ammonia nucleation from Yu et al. [2020] where neutral is the ternary system (THN) equation and ‘ions’ is the ternary system (TIMN) equation with an ion pair production rate of  $2 \text{ cm}^3 \text{ s}^{-1}$ .

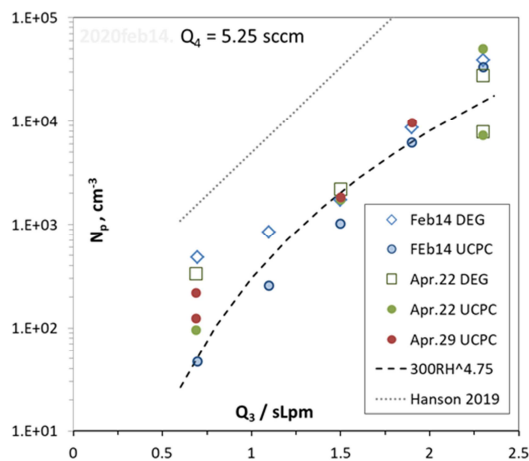


Fig. 8 Variation of  $N_p$  with relative humidity, set by the flow through the humidifier,  $Q_3$  (dates in 2020.) RH ranges from 24 to 81 % for the range of  $Q_3$  in the figure. HONO level was kept constant,  $Q_4 = 5.25 \text{ sccm}$ . A power dependency on RH of 4.75 is shown as the dashed line. The thin dotted line is the exponential fit to the RH-dependency found in Hanson et al. [2019] at  $Q_4=4.2 \text{ sccm}$ .

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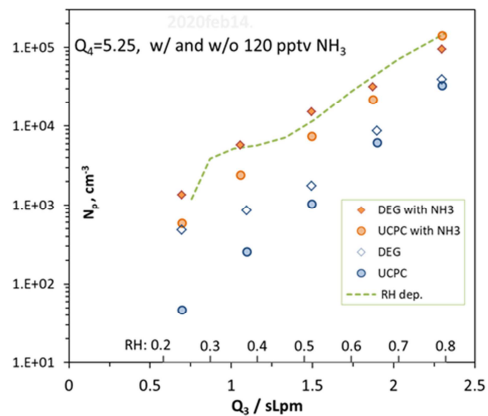


Fig 9. Variation of  $N_p$  with  $Q_3$ , a proxy for RH; with and without 120 pptv  $\text{NH}_3$  added.  $Q_4 = 5.25$  sccm and  $N_p$  from both particle counters are shown. RH varies from 23 to 81 % over the range of  $Q_3$  investigated. The Dunne et al. *ad hoc* RH dependency is shown, normalized to our measurements at  $Q_3 = 1.05$  sLpm.

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We presented evidence that the CLOUD data for 292 K and warmer were probably affected by a contaminant or non-binary system process such as occurs for our low  $N_p$  data. We contend that 4 pptv  $NH_3$  can't be the complete reason and another contaminant or other processes than the ternary system may be operating in this CLOUD data set – the model results for this amount of  $NH_3$  suggest a much smaller nucleation rate than was reported.

We suggest that  $NH_3$  at single digit levels is too little to significantly affect our data even when the apparatus is at its cleanest. Intermittent contaminants at levels that affect the measurements do appear, especially on cylinder changeovers. We suppose inadvertent dust particles can be introduced into PhoFR or onto surfaces along the gas-supply lines (the liquid nitrogen cylinder port's exposure to room air can also be a factor). Note that when the effect of this contaminant had dissipated, we derived a set of free energies from the data that helped us present limiting nucleation rates for the binary and ternary systems, allowing us to make conclusions regarding contaminants or other nucleation processes.

Our room temperature results provide a more stringent test of the binary system thermodynamics than do the CLOUD chamber results at low temperatures where the influence of relative humidity and thermodynamics are lessened compared to results at room temperature.

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