Global methane budget and trend, 2010-2017: complementarity of inverse analyses using in situ (GLOBALVIEWplus CH₄ ObsPack) and satellite (GOSAT) observations

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Abstract

We use satellite (GOSAT) and in situ (GLOBALVIEWplus CH4 ObsPack) observations of atmospheric methane in a joint global inversion of methane sources, sinks, and trends for the 2010-2017 period. The inversion is done by analytical solution to the Bayesian optimization problem, yielding closed-form estimates of information content to assess the consistency and complementarity (or redundancy) of the satellite and in situ datasets. We find that GOSAT and in situ observations are to a large extent complementary, with GOSAT providing a stronger overall constraint on the global methane distributions, but in situ observations being more important for northern mid-latitudes and for relaxing global error correlations between methane emissions and the main methane sink (oxidation by OH radicals). The in-situ-only and the GOSAT-only inversion alone, achieve respectively 113 and 212 independent pieces of information (DOFS) for quantifying mean 2010-2017 anthropogenic emissions on 1009 global model grid elements, and DOFS of 67 and 122 for 2010-2017 emission trends. The joint GOSAT + in situ inversion achieves DOFS of 262 and 161 respectively for mean emissions and trends. The in situ data thus increase the global information content from the GOSAT-only inversion by 20-30%. The in-situ-only and GOSAT-only inversions show consistent corrections to regional methane emissions but are less consistent in optimizing the global methane budget. The joint inversion finds that oil/gas emissions in the US and Canada are underestimated relative to the values reported by these countries to the United Nations Framework Convention on Climate Change (UNFCCC) and used here as prior estimates, while coal emissions in China are overestimated. Wetland emissions in North America are much lower than in the mean WetCHARTs inventory used as prior estimate. Oil/gas emissions in the US increase over the 2010-2017 period but decrease in Canada and Europe. The joint inversion yields a global methane emission of 551 Tg a^{-1} averaged over 2010-2017 and a methane lifetime of 11.2 years against oxidation by tropospheric OH (86% of the methane sink).

1 Introduction

Methane (CH_4) is the second most important anthropogenic greenhouse gas, and plays a central role in atmospheric chemistry as a precursor of tropospheric ozone and a sink of hydroxyl radicals (OH). It is emitted from many natural and anthropogenic sources that are difficult to quantify (Saunois et al., 2020). Atmospheric methane observations from satellites and in situ (surface, tower, shipboard, and aircraft) platforms have been used extensively to infer methane emissions and their trends through inverse analyses (Houweling et al., 2017). But the information from satellite and in situ observations does not always agree (Monteil et al., 2013; Bruhwiler et al., 2017) and is hard to compare because of large differences in observational density, precision, and the actual quantity being measured (Cressot et al., 2014). Here we use an analytical solution to the Bayesian inverse problem to quantitatively compare and combine the information from satellite (GOSAT) and in situ (GLOBALVIEWplus CH_4 ObsPack) observations for estimating global methane sources and their trends over the 2010-2017 period, including contributions from different source sectors and from the methane sink (oxidation by tropospheric OH).

Inverse analyses of atmospheric methane observations using chemical transport models (CTM) provide a formal method for inferring methane emissions and their trends (Brasseur and Jacob, 2017). Global satellite observations of atmospheric methane columns from the shortwave infrared SCIAMACHY and GOSAT instruments have been widely used for this purpose (Bergamaschi et al., 2013; Wecht et al., 2014; Turner et al., 2015; Maasakkers et al., 2019; Miller et al., 2019; Lunt et al., 2019). Other inverse analyses have relied on in situ methane observations that have much higher precision, are more sensitive to surface emissions, and may include isotopic information, but are much sparser (Pison et al, 2009; Bousquet et al., 2011; Miller et al., 2013; Patra et al., 2016; McNorton et al., 2018).

A number of inverse analyses have combined in situ and satellite observations (Bergamaschi et al., 2007, 2009, 2013; Fraser et al., 2013; Monteil et al, 2013; Cressot et al., 2014; Houweling et al., 2014; Alexe et al., 2015; Ganesan et al., 2017; Janardanan et al., 2020), but few of them have compared the information from the two data streams and then mostly qualitatively. Bergamaschi et al. (2009, 2013), Fraser et al. (2014), and Alexe et al. (2015) found that surface and satellite methane observations provided consistent constraints on global methane emissions, but that satellite observations achieved stronger regional constraints in the tropics. No study to our knowledge has compared the ability of satellite and in situ observations to attribute long-term methane trends.

Analytical solution to the inverse problem, as used here, provides closed-form error characterization as part of the solution, and from there allows derivation of the information content from different components of the observing system (Rodgers, 2000). Application to satellite observations has been used to determine where the observations can actually constrain the inverse solution (Turner et al., 2015). The major obstacle to this analytical solution in the past has been the need to construct the Jacobian matrix for the CTM forward model, but this is now readily done using massively parallel computing clusters (Maasakkers et al., 2019). Such a method provides a means to quantify the differences in information

content between different data streams (e.g., satellite vs. in situ) and from there to contribute to the design of a better observing system.

Here we apply satellite observations of atmospheric methane columns from the GOSAT instrument together with an extensive global compilation of in situ observations (including surface, tower, shipboard, and aircraft methane measurements) from the GLOBALVIEWplus CH_4 ObsPack v1.0 data product (Cooperative Global Atmospheric Data Integration Project, 2019), to quantify the global distribution of methane emissions, loss from reaction with OH, and related trends for the 2010-2017 period. We use for this purpose an analytical inversion method that formally characterizes the information content from the two data streams, whether that information is consistent, and whether it is complementary or redundant (Rodgers, 2000; Jacob et al., 2016). Our work provides a comprehensive global perspective on the sources contributing to 2010-2017 methane emissions and trends, as well as a general framework for synthesizing the information from satellite and in situ observations.

2 Methods

Figure 1 summarizes the components of our analytical inversion system, which builds on previous inversions of GOSAT satellite data by Maasakkers et al. (2019) and Zhang et al. (2020a) but adds the in situ observations. We apply observations y from GLOBALVIEWplus observations and/or GOSAT (Section 2.1), with the GEOS-Chem CTM as forward model (Section 2.3), to optimize the state vector x of our inverse problem. The state vector has dimension $n = 3378$ including mean 2010-2017 non-wetland methane emissions on the GEOS-Chem $4° \times 5°$ global grid ($n_1 = 1009$), 2010-2017 linear trends for these emissions on that grid ($n_2 = 1009$), monthly mean wetland methane emissions for individual years in 14 subcontinental regions ($n_3 = 12 \times 8 \times 14 = 1344$), and tropospheric OH concentrations in each hemisphere for individual years ($n_4 = 2 \times 8 = 16$). Section 2.2 describes the prior state vector estimates (x_A) and the prior error covariance matrix (S_A). We derive posterior estimates $\hat{x}$ of the state vector and the associated error covariance matrix $\hat{S}$ by analytical solution to the Bayesian optimization problem (Section 2.4). We present results from three inversions using in situ observations only (in-situ-only inversion), GOSAT observations only (GOSAT-only inversion), and both GOSAT and in situ observations (GOSAT + in situ inversion).

2.1 Methane observations

The GLOBALVIEWplus CH_4 ObsPack v1.0 data product compiled by the National Oceanic and Atmospheric Administration (NOAA) Global Monitoring Laboratory includes worldwide high-accuracy measurements of atmospheric methane concentrations from different observational platforms (surface, tower, shipboard, and aircraft) (Cooperative Global Atmospheric Data Integration Project, 2019). Here we use the ensemble of GLOBALVIEWplus observations for 2010-2017. For surface and tower measurements, we use only daytime (10-16 local time) observations and average them to the corresponding daytime mean values. We exclude outliers at individual sites that depart by more than three standard deviations from the mean. We obtain in this manner 157054 observation data points for the

inversion including 81119 from 103 surface sites, 27433 from 13 towers, 827 from 3 ship cruises, and 47675 from 29 aircraft campaigns. Figure 2a shows the mean methane concentrations in 2010-2017 from the in situ data. The data are relatively dense in North America and western Europe, with also a few sites in China, but otherwise mainly measure background concentrations. The number of available surface and tower observations increases from 10493 in 2010 to 19657 in 2017 with the largest changes in Europe and Canada.

GOSAT is a nadir-viewing satellite instrument launched in 2009 that measures the backscattered solar radiation from a sun-synchronous orbit at around 13:00 local time (Butz et al., 2011; Kuze et al, 2016). Observing pixels are 10-km in diameter and separated by about 250 km along-track and cross-track in normal observation mode, with higher-density data collected in targeted observation modes. Methane is retrieved at the 1.65 µm absorption band. We use dry column methane mixing ratios from the University of Leicester version 9.0 Proxy XCH_4 retrieval (Parker et al., 2020). The retrieval has a single-observation precision of 13 ppb and a regional bias of 2 ppb (Buchwitz et al., 2015). We use GOSAT data for 2010-2017 including 1.6 million retrievals over land as shown in Figure 2b. We do not use glint data over the oceans and data poleward of 60° because of seasonal bias and the potential for large errors (Maasakkers et al., 2019).

2.2 Prior estimates

Table1 summarizes the prior estimates of the mean 2010-2017 methane emissions used for the state vector, and Figure 3 shows the spatial patterns. Natural sources include the ensemble mean of the WetCHARTs inventory version 1.2.1 (Bloom et al., 2017) for wetlands, open fires from the Global Fire Emissions Database version 4s with seasonal and interannual variability (van der Werf et al., 2017), termites from Fung et al. (1991), and seeps from Etiope et al. (2019) with global scaling to 2 Tg a^{-1} from Hmiel et al. (2020). The default anthropogenic emissions are from EDGAR v4.3.2 (Janssens-Maenhout et al., 2019), and are superseded for fugitive fuel emissions (oil, gas, coal) by the Scarpelli et al. (2020) inventory which spatially allocates national emissions reported by countries to the United Nations Framework Convention of Climate Change (UNFCCC). US anthropogenic emissions are further superseded by the gridded version of Inventory of U.S. Greenhouse Gas Emissions and Sinks from the Environmental Protection Agency (EPA GHGI) (Maasakkers et al., 2016). The WetCHARTs wetlands inventory includes seasonal and interannual variability that is optimized in the inversion through correction to the monthly emissions. Seasonality from Zhang et al. (2016) is imposed for rice emissions, and temperature-dependent seasonality is applied to manure emissions (Maasakkers et al., 2016). Other emissions are aseasonal.

We assume a 50% error standard deviation for all anthropogenic and non-wetland natural emissions on the 4° latitude ×5° longitude grid, with no spatial error covariance so that their prior error covariance matrix is diagonal, which is a reasonable assumption for anthropogenic emissions (Maasakkers et al., 2016). We assume $0 \pm 10\%$ a^{-1} as prior estimate for the linear 2010-2017 emission trends on the 4°×5° grid; a sensitivity test using $0 \pm 5\%$ a^{-1} is also performed. The inclusion of linear trends in state vectors

allows us to identify the direction of emission change for each $4^\circ \times 5^\circ$ grid in the 8-year period, but it would not capture high-frequency interannual variability. Prior estimates of monthly mean wetland methane emissions for individual years in 14 subcontinental regions, along with their error covariance matrix, are from the WetCHARTs v1.2.1 inventory ensemble (Bloom et al., 2017). The prior methane emissions total 533 Tg a^{-1}, at the low end of the current top-down estimates (550-594 Tg a^{-1}) for 2008-2017 (Saunois et al., 2020), and this largely reflects the downward revision of global seep emissions by Hmiel et al. (2020).

Prior monthly 3-D fields of global tropospheric OH concentrations are taken from a GEOS-Chem simulation with full chemistry (Wecht et al., 2014) that yields a methane lifetime $\tau_{CH_4}^{OH}$ due to oxidation by tropospheric OH of 10.6 years and an inter-hemispheric OH ratio (North to South) of 1.16. The methane lifetime is consistent with the value of 11.2±1.3 years inferred from methylchloroform observations (Prather et al., 2012), while the inter-hemispheric OH ratio lies between the observed range of 0.97±0.12 (Patra et al., 2014) and the recent multi-model estimates of 1.3±0.1 (Zhao et al., 2019). We assume no interannual variability in this prior OH field. We assume 10% as prior error standard deviation for the hemispheric OH concentrations in individual years, based on Holmes et al. (2013), and also conduct a sensitivity test assuming 5%. Corrections to OH in the inversion are applied as a hemispheric scaling factor for individual years, without changing the spatial or temporal pattern of the original fields. Zhang et al. (2018) conducted methane inversions with twelve different OH fields from the ACCMIP model ensemble (Naik et al., 2013) and found no significant difference in results with the GEOS-Chem OH fields used here except for two outlier models.

2.3 Forward Model

We use the GEOS-Chem 12.5.0 (http://geos-chem.org) global CTM (Bey et al., 2001; Wecht et al., 2014; Maasakkers et al., 2019) as forward model to simulate atmospheric methane concentrations and their sensitivity to the state vector elements. The model is driven by MERRA-2 reanalysis meteorological fields from the NASA Global Modeling and Assimilation Office (GMAO) (Gelaro et al., 2017). The methane sink is computed within the model from 3-D tropospheric oxidant fields including OH (optimized in the inversion), Cl atoms (Wang et al., 2019), 2-D stratospheric oxidant fields (Murray et al., 2012), and soil uptake (Murguia-Flores et al., 2018). We conduct GEOS-Chem model simulations for 2010-2017 at global $4^\circ \times 5^\circ$ resolution with 47 vertical layers extending to the mesosphere.

GEOS-Chem has excessive methane in the high-latitudes stratosphere, a flaw common to many models (Patra et al., 2011) especially at coarse model resolution. Following Zhang et al. (2020), we compute correction factors to GEOS-Chem stratospheric methane subcolumns as a function of season and equivalent latitude to match the measurements from the solar occultation ACE-FTS v3.6 instrument (Waymark et al., 2014; Koo et al., 2017). As shown in Zhang et al. (2020), the correction can be up to 10% at high latitudes during winter and spring. We apply the correction factors before the inversion to avoid wrongly attributing this model transport bias to methane emissions and loss. Figure S1 shows that the

systematic differences in the posterior scaling factors of non-wetland emissions with vs. without bias correction are more prominent at the northern high latitudes, as also shown in Stanevich et al. (2020), but the global total emissions only differ by 1%.

Initial GEOS-Chem methane concentrations on January 1, 2010 are adjusted to have unbiased zonal mean relative to GOSAT observations for January 2010, and we find that the resulting model values are also unbiased relative to the GLOBALVIEWplus in situ observations in January 2010. In this manner, model discrepancies with observations over the 2010-2017 period can be attributed to model errors in emissions or OH over that period, instead of error in initial conditions. We archive model methane dry mixing ratios at each location and time of the in situ and GOSAT datasets for 2010-2017.

As forward model F for the inversion, GEOS-Chem relates the state vector x to the atmospheric concentrations y as $y = F(x)$ (Fig.1). The simulation of observations with the prior estimates of state vectors (x_A) in 2010-2017 diagnoses systematic errors in comparison to observations that enable improved estimate of the state vector through the inversion. In addition, the random component of the discrepancy can be used to estimate the observation error (sum of instrument error, representation error, and forward model error) in the Bayesian optimization problem using the residual error method (Heald et al., 2004). The method assumes that the systematic component of the model bias $(\overline{y - F(x_A)})$ for individual years, where the overbar denotes the temporal average in a 4°× 5° grid cell (for GOSAT) or for an observation platform (for in situ observations), is to be corrected in the inversion, while the residual term $(\varepsilon_0 = y - F(x_A) - \overline{y - F(x_A)})$ represents the random observation error. Here we applied this method to construct the observation error covariance matrix S_o from the statistics of ε_0. For in-situ observations, we derive ε_0 separately for the ensemble of background surface sites (Dlugokencky et al., 1994), non-background sites, tower sites, shipboard measurements, and aircraft measurements, while for GOSAT observations ε_0 is calculated for each 4°× 5° grid cell.

We find that the mean standard deviation of the random observation error (ε_0) for the GLOBALVIEWplus in situ data averages 36 ppbv (20 and 45 ppbv for background and non-background surface observations, 68 ppbv for tower observations, 10 ppbv for shipboard observations, 24 ppbv for aircraft observations), compared to 13 ppbv for GOSAT. The observation error for in situ observations is dominated by the forward model error while for GOSAT it is dominated by the instrument error. The forward model error is higher for surface concentrations near source regions than for columns or other in situ observations measuring background, because the amplitude of methane variability is much higher (Cusworth et al., 2018) and more challenging for a model at 4°× 5° resolution to capture. We assume that S_o is diagonal in the absence of better objective information, but in fact some error correlation between different observations could be expected to arise from transport and source aggregation errors in the forward model. This is considered by introducing a regularization factor γ in the minimization of the cost function for the inversion (Section 2.4).

2.4 Analytical Inversion

The Bayesian solution to the state vector optimization problem assuming Gaussian prior and observation errors involves minimizing the cost function $J(x)$:

$$J(x) = (x - x_A)^T S_A^{-1}(x - x_A) + \gamma(y - F(x))^T S_O^{-1}(y - F(x)) \ (1),$$

where x is the state vector, x_A denotes the prior estimate of x, S_A is the prior error covariance matrix, y is the observation vector, $F(x)$ represents the GEOS-Chem simulation of y, S_O is the observation error covariance matrix, and γ is a regularization factor. The need for γ in $J(x)$ is to avoid giving excessive weighting to observations, due to the likely underestimation of S_O when unknown error correlations are not included in its construction (Zhang et al., 2018; Maasakkers et al., 2019). γ here plays the same role as the regularization parameter in Tikhonov methods (Brasseur and Jacob, 2017) and reflects our inability to properly quantify the magnitude of errors.

Minimization of the cost function in equation (1) has an analytical solution if the forward model is linear (Rodgers, 2000). The optimization of methane emission and its trends is strictly linear by design because we use prescribed monthly 3-D OH fields as described in Section 2.2. There is some non-linearity regarding the optimization of OH, because the sensitivity of the methane concentration to changes in OH concentrations depends on the methane concentration through first-order loss, but we assume that the variability of methane concentration is sufficiently small that this non-linearity is negligible (we verify this assumption below). We thus express the GEOS-Chem forward model as $y = Kx + c$, where $K = \partial y/\partial x$ represents the Jacobian matrix and c is an initialization constant. We construct the Jacobian matrix K explicitly by conducting GEOS-Chem simulations with each element of the state vector perturbed separately. For the linear emission trend elements, this is done by perturbing the 2010-2017 emission trend in each grid cell from 0% (the best prior estimate) to 10% a^{-1}; for OH, this is done by perturbing yearly hemispheric OH fields by 20% without modifying the spatial or seasonal distribution. Comparison of the resulting Jacobian matrix to GEOS-Chem as $F(x) – Kx - c$ shows a negligible residual difference of 2±3 ppb, verifying the assumption of linearity.

Minimizing the Bayesian cost function by solving $dJ(x)/dx = 0$ yields closed-form expressions for the posterior estimate of the state vector $\hat{x}$ with error covariance matrix $\hat{S}$:

$$\hat{x} = x_A + G(y - Kx_A) \ (2),$$

$$\hat{S} = (\gamma K^T S_O^{-1} K + S_A^{-1})^{-1} \ (3),$$

where G is the gain matrix,

$$\mathbf{G} = \frac{\partial \hat{x}}{\partial y} = (\gamma K^T S_O^{-1} K + S_A^{-1})^{-1} \gamma K^T S_O^{-1} \ (4).$$

From the posterior error covariance matrix one can derive the averaging kernel matrix describing the sensitivity of the posterior estimate to the true state:

$$A = \frac{\partial \hat{x}}{\partial x} = I_n - \hat{S}S_A^{-1} \quad (5).$$

The trace of **A** quantifies the degrees of freedom for signal (DOFS), which represents the number of pieces of independent information gained from the observing system for constraining the state vector (Rodgers, 2000).

We choose the value of the regularization parameter γ in order to avoid overfitting to the observations when the number m of observations is much larger than the number n of state vector elements, and the error covariance of the observations cannot be properly quantified. Overfitting would be implied by a highly unlikely departure of the posterior solution from the prior estimate, which can be indicated from the posterior cost function. For a given state vector element i, the expected value of $(x_i - x_{Ai})^2$ is the prior error variance σ_{Ai}^2. For an n-dimensional state vector with a diagonal prior error covariance matrix, the state component J_A of the cost function is the sum of n random normal elements

$$J_A(x) = (x - x_A)^T S_A^{-1}(x - x_A) = \sum_n \frac{(x_i - x_{Ai})^2}{\sigma_{Ai}^2} \quad (6),$$

and its pdf is given by the Chi-square distribution with n degrees of freedom (n=3378 in this case), with an expected value of n and a standard deviation of $\sqrt{2n}$. One can apply the same reasoning to the observation component J_O of the posterior cost function,

$$J_O(x) = (y - Kx)^T S_O^{-1}(y - Kx) = \sum_m \frac{(y_i - Kx_i)^2}{\sigma_{Oi}^2} \quad (7),$$

whose pdf follows a chi-square distribution with m degrees of freedom. However, this component is less sensitive to the choice of γ because of the large random error component for individual observations.

Figure 4 shows the dependences of $J_A(\hat{x})$ and $J_O(\hat{x})$ on the choice of the regularization parameter γ, for the in situ and GOSAT observations. The in situ observations are sufficiently sparse that γ = 1 (no regularization) is expected. In the case of GOSAT, however, γ = 1 would yield $J_A(\hat{x}) = 6n \gg n \pm \sqrt{2n}$ which indicates overfitting, while γ = 0.1 yields $J_A(\hat{x}) \approx n$ which is the expected value and is used here. This can be explained by the high observation density of GOSAT, such that error correlation between individual observations through the forward model may be expected and would have a large effect on the solution. Maasakkers et al. (2019) found that $\gamma = 0.05$ and $\gamma = 0.1$ gave similar solutions in their global inversions of GOSAT data. We also conduct sensitivity tests using $\gamma = 0.5$ for in situ observations and $\gamma = 0.05$ for GOSAT observations.

The analytical solution to the Bayesian optimization problem, as done here, has several advantages relative to the more commonly used variational (numerical) solution. (1) It finds the true minimum in the cost function, rather than an approximation that may be sensitive to the choice of initial estimate. (2) It identifies the information content of the inversion and the ability to constrain each state vector element. (3) It enables a range of sensitivity analyses, modifying the prior estimates, modifying the error covariance matrices, adding/subtracting observations, etc. at minimal computational cost. We will make use of these

317 advantages in comparing the ability of the in-situ-only, GOSAT-only, and GOSAT + in situ inversions,
318 and to test how choices in cost-function construction affect our conclusions including changing the
319 regularization parameter γ, changing the prior error estimates, and using different types of in-situ
320 observations. Our analysis will focus on results from the base inversions with the default settings, but we
321 will use results from the sensitivity inversions to address specific issues.

322

323 A requirement of the analytical approach is that the Jacobian matrix be explicitly constructed, requiring
324 $n + 1$ forward model runs. Building the Jacobian matrix for the 3378 state vectors in this 8-year period
325 study requires about one million core hours (8 cores $\times$ 36 hours per simulation $\times$ 3378 simulations).
326 However, this construction is readily done in parallel on high-performance computing clusters.

327

328 Our inversion returns posterior emission estimates and their temporal trends on a $4°\times 5°$ grid for non-
329 wetland emissions, and monthly mean wetland emissions for individual years in 14 subcontinental regions.
330 We cannot separate individual sectors within a $4°\times5°$ grid cell because they will all have the same response
331 function (Jacobian column). However, we can aggregate results spatially and by sector in a way that
332 retains the error covariance of the solution (Maasakkers et al., 2019). Consider a reduced state vector $\boldsymbol{x_{red}}$
333 representing a linear combination of the original state vector elements that may be a sum over a particular
334 region or the globe, and may be weighted by the contributions from individual sectors following the prior
335 distribution. The linear transformation from the posterior full-dimension state vector $\widehat{\boldsymbol{x}}$ to the reduced
336 state vector $\widehat{\boldsymbol{x}}_{red}$ is defined by a summation matrix $\boldsymbol{W}$

337 $$\widehat{\boldsymbol{x}}_{red} = \boldsymbol{W}\widehat{\boldsymbol{x}} \quad (8).$$

338

339 The posterior error covariance and averaging kernel matrices for the reduced state vector can then be
340 calculated as:

341 $$\widehat{\boldsymbol{S}}_{red} = \boldsymbol{W}\widehat{\boldsymbol{S}}\boldsymbol{W}^T \quad (9),$$
342 $$\boldsymbol{A}_{red} = \boldsymbol{W}\boldsymbol{A}\boldsymbol{W}^* \quad (10),$$

343 where $\boldsymbol{W}^* = \boldsymbol{W}^T(\boldsymbol{W}\ \boldsymbol{W}^T)^{-1}$ (Calisesi et al., 2005). $\widehat{\boldsymbol{S}}_{red}$ provides a means to determine error
344 correlations between aggregates of quantities optimized by the inversion, e.g., between global methane
345 emissions and global OH concentrations. $\boldsymbol{A}_{red}$ provides a means to determine the ability of the inversion
346 to constrain an aggregated term (e.g., emissions from a particular sector).

347

348 **3. Results and discussion**
349 **3.1 Ability to fit the in situ and GOSAT data**
350 We will present results from three different inversions for 2010-2017: (1) using only in situ observations
351 (in-situ-only inversion), (2) using only GOSAT observations (GOSAT-only inversion), and (3) using both
352 GOSAT and in situ observations (GOSAT + in situ inversion). Here we first evaluate the ability of these
353 different inversions to fit the in situ and GOSAT observations, including when the data are not used in the
354 inversion (consistency check). This is done by conducting GEOS-Chem simulations with posterior values
355 for the state vectors and comparing to observations.

356

Figures 5 and 6 show the resulting comparisons for the in situ observations, arranged by type of platform (Fig.5), and by latitude bands and months (panels (a)-(d) in Fig.6). The model simulation with prior estimates shows a 30-60 ppb low bias for all in situ platforms growing with time. The in-situ-only inversion effectively corrects this bias and its trend, and also significantly improves the correlations across all platforms. The GOSAT-only inversion performs comparably in correcting the 2010-2017 trend for the independent in-situ data (Fig.6c) and bias for background observations (e.g. aircraft observations in the Southern Hemisphere (Fig.S2)), but there is a low bias at northern mid-latitudes reflecting surface and tower data in North America and Europe. As we will see, the in situ observations are important for optimizing emissions in these regions.

Figure 6 also compares the fits to the GOSAT observations (panels (e)-(h)). The GOSAT-only inversion corrects the bias and trend in the prior simulation at all latitudes. The in-situ-only inversion corrects the trends, but biases low to the GOSAT observations by about 10 ppbv with larger bias in the Southern Hemisphere due to the sparsity of in situ observation there. The comparison suggests that in situ and GOSAT observations are largely consistent for informing the global methane change but also have some complementarity for the inversion. The GOSAT + in situ joint inversion shows good agreement with both the in situ and GOSAT observations.

Figure 7a further evaluates the global methane growth rate as determined by the methane budget imbalance for individual years in 2010-2017 from the three inversions. The observed methane growth rate inferred from the NOAA sites (https://www.esrl.noaa.gov/gmd/ccgg/trends_ch4/, last access: 20 June 2020) averages 7.2 ± 2.8 ppb a^{-1} over the period, peaking in 2014, and overall accelerating with higher growth in 2015-2017 than in 2010-2013. We find that all posterior simulations show comparable mean methane growth rate (7.7 ± 3.7 ppb a^{-1} for in-situ-only inversion, 8.8 ± 2.2 ppb a^{-1} for GOSAT-only inversion, and 8.3 ± 1.8 ppb a^{-1} for the GOSAT + in situ inversion). However, the in-situ-only inversion overestimates the increasing trend in the methane growth rate, largely driven by the year 2017, and fails to fit its interannual variability. This may reflect the heavy weighting of the in situ observations toward northern mid-latitudes. GOSAT observations in the inversion do much better in capturing the observed methane interannual variability and trend. Adding in situ observations to GOSAT observations provides a better fit in 2015 than GOSAT-only inversion but has an insignificant effect in other years. Zhang et al. (2020a) interpreted the trend and interannual variability in the GOSAT-only inversion as due to a combination of anthropogenic emissions, wetlands, and OH concentrations.

3.2 Anthropogenic methane emissions

Figure 8 shows the averaging kernel sensitivities (diagonal elements of the averaging kernel matrix) and posterior scaling factors for the non-wetland emissions (dominated by anthropogenic emissions) in the in-situ-only, GOSAT-only, and GOSAT + in situ joint inversions. The DOFS (trace of the averaging kernel matrix) quantify the number of independent pieces of information from the inversion, starting from 1009

395 unknowns for anthropogenic emissions (Figure 1). The DOFS are 113 for the in-situ-only inversion, 212

396 for the GOSAT-only inversion, and 262 for the GOSAT + in situ joint inversion. The higher DOFS from

397 the joint inversion indicate that the satellite and in situ observations have complementarity but also some

398 redundancy. Strict complementarity would imply a DOFS of 325=113+212. We find that 75% of the in

399 situ information is at northern mid-latitudes (30-60°N, DOFS=82, calculated as the sum of averaging

400 kernel sensitivities in that latitude band) where the observations are densest, with another 9% (DOFS=10)

401 at 60-90°N. GOSAT provides more information than do in situ observations at northern mid-latitudes

402 (DOFS=96) and dominates in the tropics (DOFS=105). This dominance of satellites for informing

403 methane sources in the tropics has been pointed out in previous studies (Bergamaschi et al., 2013; Monteil

404 et al., 2013; Fraser et al., 2013; Alexe et al., 2015). We find that the DOFS from the in-situ-only inversion

405 observations are mostly (85%) from the surface and tower measurements (Fig.S3).

406

407 We investigate further the inversion results for northern mid-latitudes where most of the information of

408 in situ observations is contained including for the US, Canada, Europe, and China. Table 2 gives the

409 optimization of anthropogenic methane emissions (calculated as the difference between total non-wetland

410 emissions and the non-wetland natural emissions) in these regions. Figure 9 shows the optimization by

411 source sectors, assuming that (1) the partitioning between sectors of non-wetland emissions in individual

412 grid cells is correct in the prior inventory (this does not assume that the prior distribution of sectoral

413 emissions is correct), (2) the scaling factors are to be applied equally to all sectors in a grid cell. These

414 assumptions are adequate when the sectors are spatially separated but are more prone to error when they

415 spatially overlap. Figure 9 also shows the averaging kernel sensitivities of emission sectors (diagonal

416 terms of $\boldsymbol{A}_{red}$ derived from Equations (8) and (10)), measuring the ability of the inversion to optimize

417 different emissions sectors, and the DOFS for each inversion summed over the region. Wetland methane

418 emissions are optimized separately as will be discussed in Section 3.3.

419

420 Inspection of the DOFS shows that the in situ observations are more effective than GOSAT for optimizing

421 US anthropogenic methane emissions (DOFS=41 vs. DOFS=22) and this applies to all sectors (Figure 9).

422 The averaging kernel sensitivities panel in Figure 9 shows that US results from the joint GOSAT + in situ

423 inversion are mostly determined by the in situ observations. The joint GOSAT + in situ inversion increases

424 anthropogenic US emissions from 28 Tg a^{-1} in the prior EPA GHGI to 36 Tg a^{-1}, with most of the increase

425 driven by oil/gas sources in the central US. Averaging kernel sensitivity for major sectors is large (0.63-

426 0.93), indicating that the posterior estimates are mostly determined by the observations rather than by the

427 prior estimates. The underestimate of oil/gas emissions in the EPA GHGI has been reported before in

428 local observations and higher-resolution inversions (Miller et al., 2013; Turner et al., 2015; Alvarez et al.,

429 2018; Cui et al., 2019; Maasakkers et al., 2020).

430

431 The in situ observations are also more effective than GOSAT in optimizing anthropogenic methane

432 emissions in Canada (DOFS=21 vs. DOFS=6), particularly in Alberta where oil/gas emissions are high

433 (Fig.8). This reflects in part our exclusion of GOSAT data poleward of 60°N. Oil/gas emissions in Canada

increase by a factor of 2 in the GOSAT + in situ inversion to 4.5 Tg a^{-1} compared to UNFCCC prior estimate, with an averaging kernel sensitivity of 0.57 (Fig.9). Total anthropogenic emissions increase from 5 Tg a^{-1} to 8 Tg a^{-1}.

In situ and GOSAT observations show comparable ability in optimizing the total anthropogenic emissions in Europe (DOFS=16~18). They agree that prior anthropogenic methane emissions are too high in northern Europe but disagree in southern Europe. Averaging kernel sensitivities from the insitu-only inversion are slightly weaker than for the US and Canada because of the lower density of in situ sites. The Integrated Carbon Observation system (ICOS) network (https://www.icos-cp.eu/, last access: 17 July 2020) has increased substantially the number of available methane observations in Europe since 2017 so that future inversions should expect a stronger constraint from in situ observations. Total European anthropogenic emissions decrease from 27 Tg a^{-1} to 23 Tg a^{-1} in the GOSAT + in situ joint inversion, with decreases for all sectors but this may reflect the inability of our 4°× 5° resolution to effectively separate emission sectors.

The only other region where in situ observations provide significant information is China, though the corresponding DOFS=13 is less than for GOSAT (DOFS=22). Both inversions agree that emissions must be greatly decreased from the prior estimate, and the joint inversion (DOFS=28) has stronger power in doing so. The posterior 2010-2017 Chinese anthropogenic emission is 43 Tg a^{-1} in the joint inversion, compared to 63 Tg a^{-1} in the prior estimate. Our results agree with a recent study by Janardanan et al. (2020), which also used GOSAT and surface observations to estimate a mean 2011-2017 anthropogenic methane emission in China of 46±9 Tg a^{-1}. The downward correction is mainly driven by a 40% decrease in coal emissions from 19 Tg a^{-1} to 11 Tg a^{-1} (Fig. 9). Previous inversions using the EDGAR inventory (>20 Tg a^{-1}) as prior estimate found a similar correction (Alexe et al., 2015; Thompson et al., 2015; Turner et al., 2015; Maasakkers et al., 2019; Miller et al., 2019). In our case, the prior estimate of coal emissions (19 Tg a^{-1}) is the value reported by China to the UNFCCC and we find that it is still too high. A recent inventory by Sheng et al. (2019) gives a coal emission estimate of 15 Tg a^{-1} for China in 2010-2016.

3.3 Wetland methane emissions

The inversion optimizes wetland emissions for the 14 regions of Figure 3 and for 96 individual months covering 2010-2017, amounting to 1344 state vector elements. Results from the in-situ-only, GOSAT-only, and GOSAT + in situ inversions yield DOFS of 221, 183, and 301 respectively. In situ observations provide more information for boreal wetlands while GOSAT dominates for tropical wetlands.

Zhang et al. (2020a) give a detailed analysis of GOSAT-only inversion results for tropical wetlands. Here we analyzed further the boreal/temperate North America wetlands, where in situ observations provide significant added information (Figure 10). Both in situ and GOSAT observations agree that the prior WetCHARTs emissions are too high. The posterior estimates from the GOSAT + in situ inversion are 4.5 and 2.0 Tg a^{-1} for boreal and temperate North America, respectively, compared to 12.8 and 6.9 Tg a^{-1} in

WetCHARTs. Posterior boreal wetland CH_4 emissions for North America are on the lower end but within the WetCHARTs estimates (WetCHARTs models range 3~33 Tg a^{-1}); however, posterior temperate CH_4 emissions for North America are outside the WetCHARTs range (3~12 Tg a^{-1}). The correction for boreal North America is particularly large in May-June, which can potentially be attributed to suppression of wetland emissions by either snow cover (Pickett-Heaps et al., 2011) or frozen soils (Zona et al., 2016). The WetCHARTs emission overestimate for temperate North America (mainly coastal wetlands in the eastern US) has been reported before from inversions using aircraft data (Sheng et al., 2018) and GOSAT data (Maasakkers et al., 2020).

3.4 Anthropogenic methane emission trends

Figure 11 presents the 2010-2017 trends (% a^{-1}) of anthropogenic methane emissions from the three inversions, and the corresponding averaging kernel sensitivities. The GOSAT + in situ inversion has a DOFS = 161 for quantifying the spatial distribution of the trends. Most of that information is from GOSAT (DOFS = 122) but in situ observations add significant information. Information from in situ observations is concentrated in the US, Canada, Europe, and China. Table 2 summarizes the trends for the four regions. Figure 12 shows the trends disaggregated by sectors, using the same procedure as for Figure 9.

In situ observations provide stronger constraints than GOSAT on anthropogenic emission trends in the US (DOFS=29 vs. DOFS=12). They agree on the upward trend in the eastern US as also found in Maasakkers et al. (2020) which used GOSAT in a high resolution inversion to interpret methane trends in the US in 2010-2015. However, they show opposite trends (positive trend from in-situ-only inversion but negative from GOSAT-only inversion) in total emissions and in the central south US (Table 2, Fig. 11). The GOSAT + in situ joint inversion (DOFS=31) estimates that US anthropogenic methane emissions increased by 0.4 Tg a^{-1} a^{-1} (1.1% a^{-1}) from 2010 to 2017, with the largest contribution from oil/gas emissions (0.3 Tg a^{-1} a^{-1}, 2.5% a^{-1}). This posterior trend is much smaller than previous studies showing large increases in US oil/gas emissions (2.1–4.4 Tg a^{-1} a^{-1}) inferred from ethane/propane levels (Franco et al., 2016; Hausmann et al., 2016; Helmig et al., 2016), but is more consistent with a recent study by Lan et al. (2019) of 0.3±0.1 Tg a^{-1} a^{-1} in 2006-2015 based on long-term in situ measurements. The inversion also reveals rising emissions from oil/gas in the central south US, including the Permian Basin which is currently the largest oil-producing basin in the US (Zhang et al., 2020b).

We find that anthropogenic emissions in Canada decrease over the 2010-2017 period by 0.2 Tg a^{-1} a^{-1} (2.5% a^{-1}) in the GOSAT + in situ joint inversion, mostly driven by oil/gas emissions in Alberta and livestock emissions (Figs. 11-12). Anthropogenic emissions in Europe decrease by 0.4 Tg a^{-1} a^{-1} (1.7 % a^{-1}).

All three inversions show increases of Chinese anthropogenic methane emissions over 2010-2017 by 0.1-0.4 Tg a^{-1} a^{-1} (0.3-0.9% a^{-1}), but the spatial patterns and source attributions are different. The largest difference is for coal mining emissions in the North China Plain, where in situ observations indicate a

512 decrease of -0.8 Tg a^{-1} a^{-1} while GOSAT shows an increase of 0.1 Tg a^{-1} a^{-1}. A previous GOSAT inversion
513 study found a large increase of coal mining emissions in China over 2010-2015 (Miller et al., 2019).
514 However, a recent bottom-up inventory estimates that Chinese coal emission peaked in 2012 and
515 decreased afterward, leading to no significant overall trend for 2010-2016 (Sheng et al., 2019). Our
516 inversion assumes linear trends in emissions over 2010-2017 but that may not be appropriate for China.

3.5 Global methane budget for 2010-2017

519 Table 1 shows the optimized global anthropogenic emissions from different sectors as determined by the
520 joint GOSAT + in situ inversion. Corrections to the global prior estimates are mostly determined by
521 GOSAT (Fig. 8). They include upward corrections to livestock and rice methane emissions, and
522 downward correction to the coal mining emissions driven by overestimation in China. The joint inversion
523 also estimates a global increase in anthropogenic emissions by 1.7±0.6 Tg a^{-1} a^{-1} (0.5% a^{-1}) in 2010-2017,
524 dominantly driven by trends in the tropics (Fig. 11).

526 A number of previous studies have analyzed surface observations to interpret global methane budgets and
527 trends (Dlugokencky et al., 2009; Bruhwiler et al., 2014; Houweling et al., 2017). As shown in Figure 6,
528 our in-situ-only inversion can fit the GOSAT observations of global methane distribution and trend,
529 indicating that the in situ data provide useful information on the global budget. Here we examine whether
530 this information adds to that from GOSAT. For this purpose and following Maasakkers et al. (2019), we
531 collapse the full state vector to a reduced state vector ($\widehat{x}_{red}$) that contains global mean methane emissions
532 and OH as elements, and derive the associated error covariance matrix ($\widehat{S}_{red}$) as introduced in Section
533 2.4.

535 Figure 13 shows the joint probability density functions (PDFs) of the mean anthropogenic methane
536 emissions and methane lifetime against oxidation by tropospheric OH from the three inversions. There is
537 strong negative correlation (r=-0.72) between the optimization of methane emissions and OH in the
538 GOSAT-only inversion, and somewhat less in the in-situ-only inversion (r=-0.53), although the posterior
539 error variance is larger due to the lower data density as indicated by the axes of the ellipses. A sensitivity
540 inversion using only the surface and tower measurements in the in-situ-only inversion yields r=-0.37
541 (Fig.13b). It indicates that in situ observations, in particular surface and tower measurements, are more
542 effective than the satellite observations in constraining methane emissions independently from the sink
543 by OH. A likely reason is that surface measurements in source regions are more sensitive to methane
544 emissions than are column measurements. We also find that the in-situ-only inversion yields a larger
545 interannual variability of posterior OH concentrations and thus methane lifetime than the GOSAT-only
546 inversion (Fig.7b and Fig.S4). This is because the number and location of the observations varies from
547 year to year, particularly for aircraft campaigns and ship cruises., due to the heterogeneous spatial and
548 temporal distribution of the in situ observations.

550 Comparison of the posterior PDFs between the GOSAT-only and in-situ-only inversions implies that the

two are inconsistent in optimizing global methane budgets, since the 99% probability contours do not overlap (Fig.13a). A possible cause is that the posterior error covariance matrix underestimates the actual error variance due to its assumption of independent identically distributed (IID) observational errors (Brasseur and Jacob, 2017), and this would particularly affect the global budget which sums emission results for individual grid cells. Remarkably, the solution from the GOSAT + in situ joint inversion is more in agreement with in situ observations than GOSAT, and does not lie between these two solutions. Inspection of Figure 6c shows that the GOSAT-only inversion is biased low relative to in situ observations at northern mid-latitudes and biased high in the southern hemisphere, implying that both emissions and OH concentrations are too low. On the other hand, Figure 6f indicates either underestimation of emissions or overestimation of OH concentrations in the in-situ-only inversion, and the former one is more likely as GOSAT measurements used here are over land which should be more sensitive to emissions than OH loss. Ingestion of both observations in the GOSAT + in situ inversion thus enhances both the methane emissions and OH concentrations compared to the in-situ-only and GOSAT-only inversion to correct these biases. It also narrows the posterior error of mean anthropogenic emissions and methane lifetime against tropospheric OH by 20% and 50% compared to the GOSAT-only and in-situ-only inversions, respectively (Fig. 13a). Thus we find that the GOSAT and in situ observations are complementary in quantifying the global budget.

Table 3 summarizes the global mean methane budget in 2010-2017. The GOSAT + in situ joint inversion estimates a total methane emission of 551 Tg a^{-1}, of which 371 Tg a^{-1} are anthropogenic, and a total sink of 529 Tg a^{-1}. The total emission is within the 550-594 Tg a^{-1} range of top-down estimates but lower than the 594-881 Tg a^{-1} range of bottom-up estimates reported for the 2008-2017 decade by the Global Carbon Project (Saunois et al., 2020). Our joint inversion yields a methane lifetime against OH oxidation of 11.2 years, consistent with the observationally-based estimate of 11.2±1.3 years (Prather et al., 2012), and pushes the northern to southern hemispheric OH ratio (1.06 in GOSAT + in situ inversion versus 1.16 in prior estimate) closer to the values of 0.97±0.12 inferred from methyl chloroform observations (Patra et al., 2014).

We examine in Figure 13b the sensitivity of the global methane budget optimization to the choice of different regularization parameter γ (and therefore observation error $\boldsymbol{S_O}$) and prior error of methane emission trends and OH concentrations. We find that reducing γ or prior errors of trend and OH by 50% yields consistent estimates of anthropogenic emissions and OH concentrations as compared to the default inversion, with differences within 3%. Decreasing the weighting of observations in the inversion (i.e. assuming larger observation error) enlarges the posterior error and pushes the posterior estimates closer to the prior estimates. Assuming a lower prior error for OH concentration from 10% to 5% results in lower methane lifetime (closer to the prior) and higher emissions, and also reduces the error correlation between the optimization of methane emissions and OH, while assuming a lower prior error for non-wetland emission trends leads to an opposite effect. Our results are consistent with Maasakkers et al. (2019), which shows that different assumptions of error distribution and magnitude tin their analyses have relatively

590   small results. We also find that having the shipboard and aircraft measurements in the in-situ-only
591   inversion pushes the estimate to be more consistent with the GOSAT-only inversion (Fig.13b), implying
592   that the shipboard and aircraft measurements by emphasizing the methane in the remote atmosphere play
593   a similar role as satellite measurements in global methane budget optimization.

594

595   **4 Conclusions**

596   We quantified and attributed global sources, sinks, and trends of atmospheric methane for 2010-2017 by
597   inversions of GOSAT satellite data and the GLOBALVIEWplus in situ methane observations from surface
598   sites, towers, ships, and aircraft. The inversions use an analytical solution to Bayesian optimization
599   problem including closed-form error covariance matrices from which the detailed information content of
600   the inversion can be derived. We conduct inversions using GOSAT and in situ data separately and
601   combined. In this manner we are able to quantify the consistency and complementarity (or redundancy)
602   of the satellite and in situ observations.

603

604   We find that the GOSAT-only inversion can generally fit the in situ data and the in-situ-only inversion
605   can generally fit the GOSAT data, indicating consistency between the two data sets. However, the
606   GOSAT-only inversion has difficulty fitting the in situ observations in source regions (US and Europe),
607   while -the in-situ-only inversion cannot reproduce the interannual variability of the methane growth rate
608   due to the heavy weighting of in situ data to northern mid-latitudes. The GOSAT + in situ inversion shows
609   the best fit to the ensemble of observations.

610

611   GOSAT and in situ observations have complementarity in constraining global emissions. GOSAT
612   provides stronger constraints than in situ observations for the tropics, while in situ observations are more
613   important in the US, Canada, Europe, and northern China where observations are most dense. The
614   GOSAT-only and in-situ-only inversions also show consistent corrections to regional methane emissions
615   in the US, Europe, and China. The joint GOSAT + in situ inversion indicates large underestimates of
616   oil/gas emissions in the US and Canada, and large overestimates of coal emissions in China, relative to
617   the national inventories reported to the United Nations Framework Convention on Climate Change
618   (UNFCCC) and used here as prior estimates for our inversions. Emissions from boreal wetlands are
619   overestimated in the mean WetCHARTs inventory used as prior estimate, particularly in May-June when
620   snow cover and frozen soils inhibit methane emission.

621

622   Our inversions indicate increasing trends in US anthropogenic emissions driven by oil/gas production but
623   decreasing trends in Canada (oil/gas) and Europe. Joint inversion of GOSAT + in situ data shows a weak
624   decreasing trend in Chinese coal emissions for 2010-2017, consistent with a recent bottom-up inventory
625   (Sheng et al., 2019).

626

627   We find that GOSAT and in situ observations are also complementary in constraining the global methane
628   budget. While the global budget information relies more on GOSAT observations, information from the

in situ observations at northern mid-latitudes avoids the large error correlations between methane emissions and sink from OH and also corrects the underestimation of both emission and OH in the GOSAT-only inversion. Our joint GOSAT + in situ inversion yields global methane emissions and loss of 551 and 529 Tg a^{-1} a^{-1} averaged over 2010-2017, and a methane lifetime of 11.2 years.

Our study presents a framework to integrate satellite and in situ data in analytical inversions. We conclude that on the basis of the present observation system, in situ and satellite observations are complementary for constraining global methane budgets and regional emissions. Satellite observations of atmospheric methane are presently expanding with the new availability of global daily data from the TROPOMI instrument launched in October 2017 (Hu et al., 2018). This will call for re-evaluating the role of in situ observations for constraining regional and global methane budgets, as can be done with the methods presented here. In situ observations will in any case continue to play a critical role for documenting long-term trends of methane with consistent calibration, for observation of oceanic and polar regions where satellites have limited capability, for high-frequency measurements in source regions giving insight into the magnitude and intermittency of local emissions, and for independent validation of satellite-based inversions.

Data availability

The GLOBALVIEWplus CH$_4$ ObsPack v1.0 data product is available at https://www.esrl.noaa.gov/gmd/ccgg/obspack/data.php?id=obspack_ch4_1_GLOBALVIEWplus_v1.0_ 2019-01-08 (last access: July 17, 2020). The GOSAT proxy satellite methane observations are available at https://doi.org/10.5285/18ef8247f52a4cb6a14013f8235cc1eb (last access: July 17, 2020). Modeling data can be accessed by contacting the corresponding authors Xiao Lu (xiaolu@g.harvard.edu) and Yuzhong Zhang (zhangyuzhong@westlake.edu.cn).

Author contributions

XL and DJJ designed the study. XL and YZZ conducted the modeling and data analyses with contributions from JDM, MPS, LS, ZQ, TRS, HON, RMY, and JXS. AA contributed to the GLOBALVIEWplus CH$_4$ ObsPack v1.0 data product. RJP and HB contributed to the GOSAT satellite methane retrievals. AAB and SM contributed to the WetCHARTs wetland emission inventory and its interpretation. XL and DJJ wrote the paper with input from all authors.

Competing interests

The authors declare that they have no conflict of interest.

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Table 1. Global sources and sinks of atmospheric methane, 2010-2017[a].

	Prior [b]	Posterior [c]
Total sources [Tg a⁻¹]	533	551
Natural Sources		
Wetlands	161	148
Open fires	14	16
Termites	12	14
Seeps	2	2
Anthropogenic sources		
Livestock	117	136
Oil	42	40
Natural gas	25	30
Coal mining	31	23
Rice cultivation	38	44
Wastewater	37	42
Landfills	30	31
Other Anthropogenic	25	25
Total Sinks [Tg a⁻¹]	540	529
Tropospheric OH	468	456
Stratospheric loss [d]	33	33
Soil uptake [d]	34	34
Tropospheric Cl [d]	5	5

[a] 8-year mean values for 2010-2017.

[b] Prior natural source estimates (2000-2017 means) are from Bloom et al. (2017) for wetlands, Etiope et al. (2019) and Hmiel et al. (2020) for seeps, Fung et al. (1991) for termite emissions, van der Werf et al. (2017) for open fire emissions. Prior anthropogenic source estimates for 2012 are from EDGAR v4.3.2 (Janssens-Maenhout et al., 2017) except from Scarpelli et al. (2020) for fuel exploitation (oil, gas, coal), and are overwritten for the US with the gridded EPA inventory of Maasakkers et al. (2016). The prior tropospheric OH concentration field is from Wecht et al. (2014) and yields a methane lifetime against oxidation by tropospheric OH of 10.6 years.

[c] From the joint inversion of GOSAT and in situ data

[d] These minor sinks are not optimized by the inversion.

981 **Table 2.** Anthropogenic methane emissions and trends, 2010-2017 [a]

Inversions	In-situ-only inversion	GOSAT-only inversion	GOSAT+in situ inversion
US [b] (prior: 28 Tg a^{-1})			
Posterior (Tg a^{-1})	35	31	36
2010-2017 trend (Tg a^{-1} a^{-1})	0.5	-0.1	0.4
Canada (prior: 5 Tg a^{-1})			
Posterior (Tg a^{-1})	8	5	8
2010-2017 trend (Tg a^{-1} a^{-1})	-0.2	-0.0	-0.2
Europe [c] (prior: 27 Tg a^{-1})			
Posterior (Tg a^{-1})	28	17	23
2010-2017 trend (Tg a^{-1} a^{-1})	0.1	-0.6	-0.4
China (prior: 63 Tg a^{-1})			
Posterior (Tg a^{-1})	45	46	43
2010-2017 trend (Tg a^{-1} a^{-1})	0.3	0.4	0.1

982 [a] Posterior estimates of mean 2010-2017 emissions and trends for the in-situ-only, GOSAT-only, and GOSAT + in situ
983 joint inversions.
984 [b] Including contiguous US and Alaska.
985 [c] Europe is defined as west of 30°E, excluding Russia.
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989 **Table 3.** Optimized global methane budget, 2010-2017.

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Inversions	In-situ-only inversion	GOSAT-only inversion	GOSAT+in situ inversion
Total sources [Tg a^{-1}]	515	504	551
Anthropogenic [a]	359	333	371
Seeps, termites	15	15	16
Open fires	15	16	16
Wetlands	126	140	148
Total sinks [Tg a^{-1}]	496	480	529
Tropospheric OH[b]	423	408	456
Other losses [c]	73	72	73
Mean imbalance [Tg a^{-1}]	19	24	22

991   [a] See Table 1 for sectoral breakdown from the joint inversion.

992   [b] Methane lifetime against oxidation by tropospheric OH is 11.2±0.1 years in the GOSAT + in situ inversion.

993   [c] Soils, stratosphere, and oxidation by tropospheric Cl.

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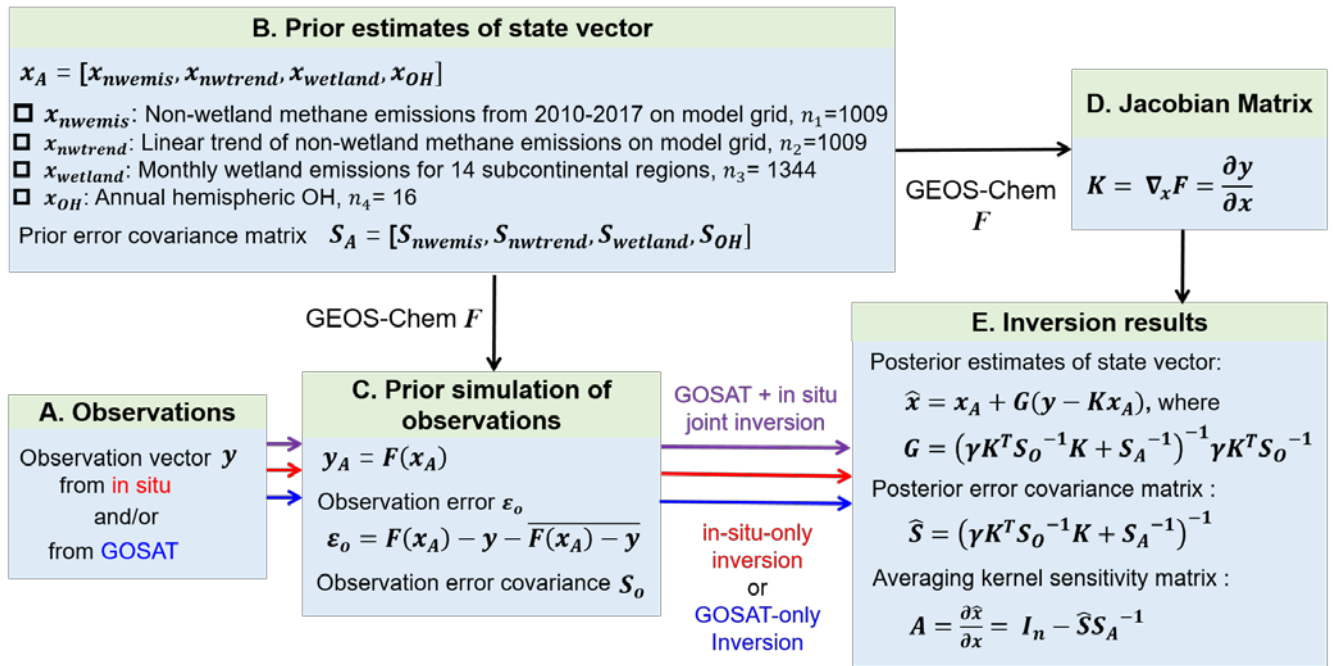
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997 **Figure 1.** Analytical inversion framework. The inversion is applied to GOSAT and GLOBALVIEWplus
998 in situ observations for 2010-2017. GEOS-Chem is the chemical transport model (CTM) used as forward
999 model for the inversion. γ is a regularization factor in the Bayesian cost function (see text).
1000
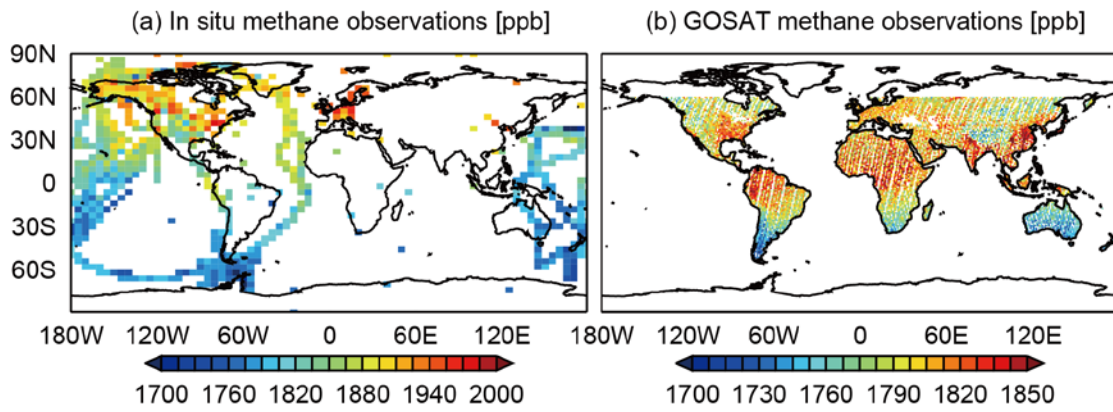
Figure.2 Mean 2010-2017 methane observations from the GLOBALVIEWplus ObsPack data product and from GOSAT. The GLOBALVIEWplus in situ data are local dry mixing ratios and are averaged over the 4°×5° model grid for visibility. The GOSAT data are dry column mixing ratios on a 1°×1° grid from the University of Leicester version 9 Proxy XCH4 retrieval (Parker et al., 2020), excluding observations over oceans and poleward of 60°N. Note the difference in color scale between panels.
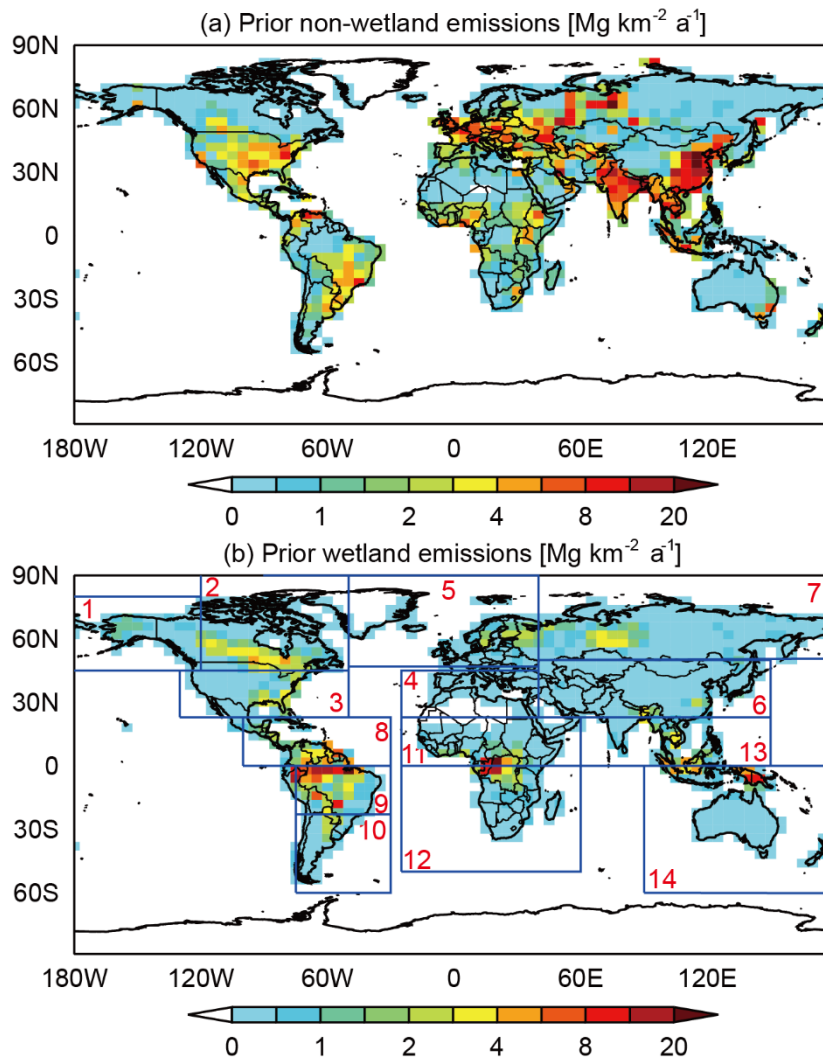
Figure 3. Prior estimates of mean 2010-2017 methane emissions. The top panel shows the non-wetland emissions on the 4°×5° grid used for the inversion. The bottom panel shows the wetland emissions and the 14 subcontinental wetland regions used for the inversion following Bloom et al. (2017).
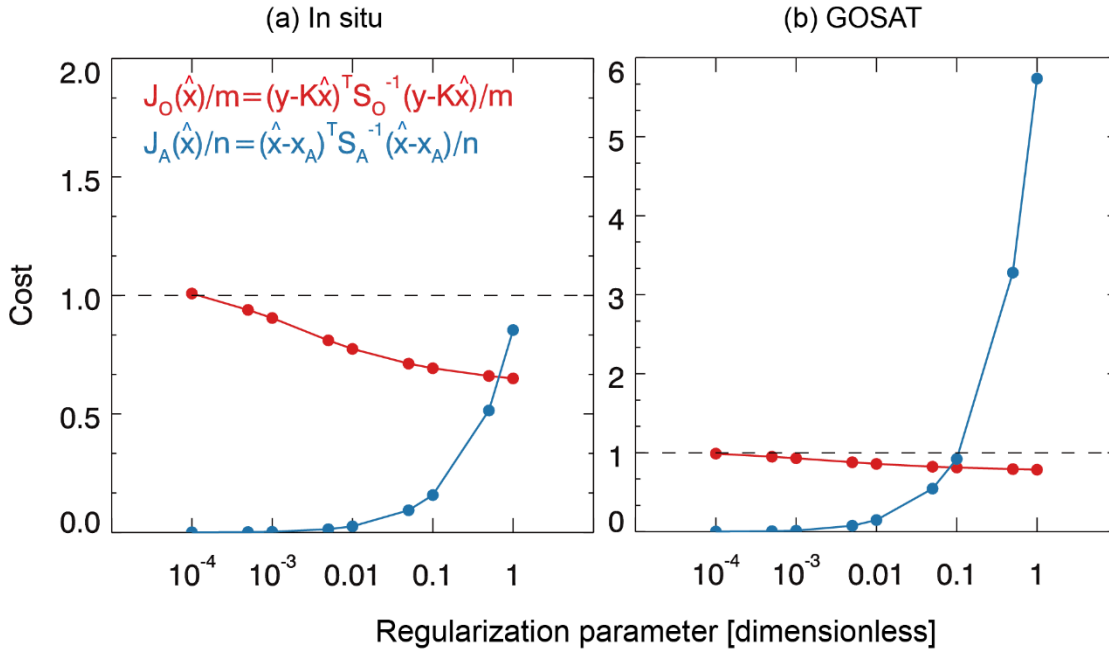
Figure 4. Optimization of the regularization parameter γ in the Bayesian cost function (Equation (1)). The figure shows the posterior observation component $J_O(\hat{x}) = (y - K\hat{x})^T S_O^{-1}(y - K\hat{x})$ and the posterior state component $J_A(\hat{x}) = (\hat{x} - x_A)^T S_A^{-1}(\hat{x} - x_A)$ for the insitu-only and GOSAT-only inversions.

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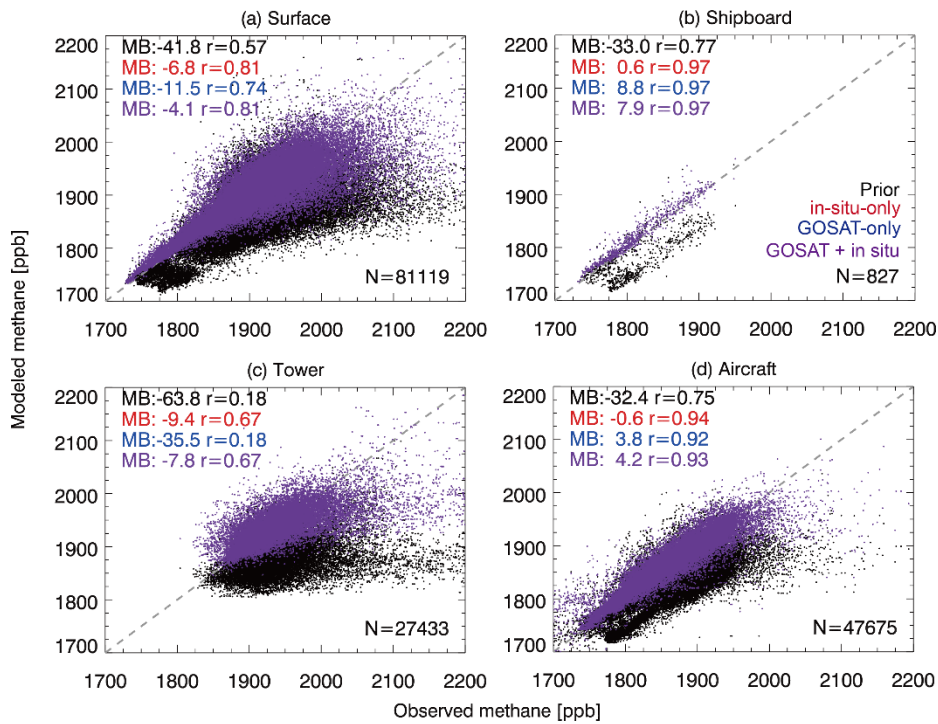
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1026 **Figure 5.** Ability of the inversions to fit the in situ methane observations. Panels (a)-(d) compare the
1027 surface, tower, shipboard, and aircraft observations in 2010-2017 to the GEOS-Chem simulation using
1028 the prior (black) and posterior estimates of methane emissions and OH concentrations from the in-situ-
1029 only inversion (red, dots not shown), GOSAT-only inversion (blue dots not shown), and GOSAT + in situ
1030 joint inversion (purple). The numbers (N) of observations from each platform, the mean bias (MB), and
1031 the correlation coefficients (*r*) between the observed and simulated values are shown inset.

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Figure 6. Ability of the inversions to fit the in situ methane observations and GOSAT satellite observations. Panels (a)-(d) show the monthly time series of the differences between observed and simulated in situ methane concentrations averaged over different latitude bands from 2010 to 2017. Panels (e)-(h) are the same as panels (a)-(d) but for GOSAT methane concentrations.

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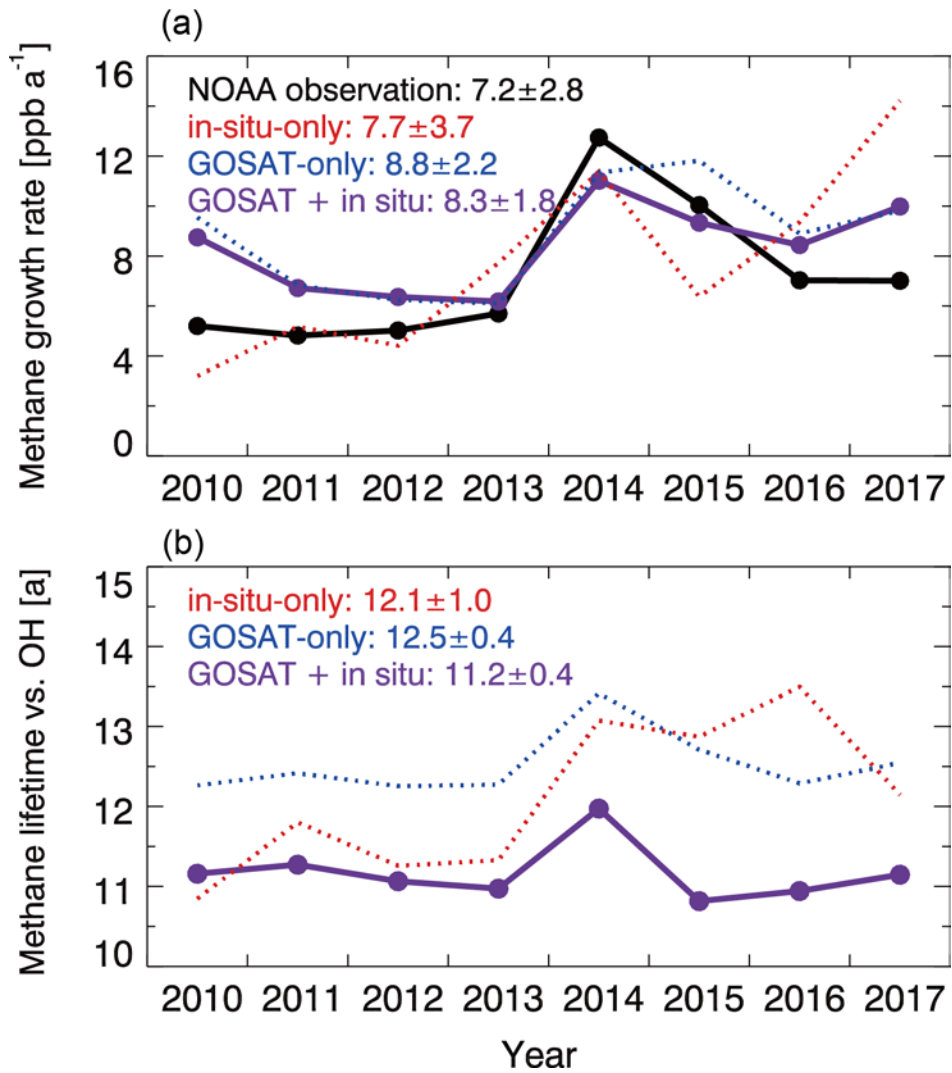
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1044 **Figure 7.** (a) Annual global growth rate of atmospheric methane, 2010-2017. Results from our three
1045 different inversions (in-situ-only, GOSAT-only, GOSAT + in situ) are compared to the observed growth
1046 rates inferred from the NOAA surface observational network
1047 (https://www.esrl.noaa.gov/gmd/ccgg/trends_ch4/, last access: 20 June, 2020). Mean annual growth rates
1048 and standard deviations from the different inversions are shown inset. (b). Methane lifetime against
1049 oxidation by tropospheric OH, 2010-2017, from the three different inversions. Mean lifetime and standard
1050 deviations are shown inset. The methane lifetime in the prior estimate is 10.6 years.
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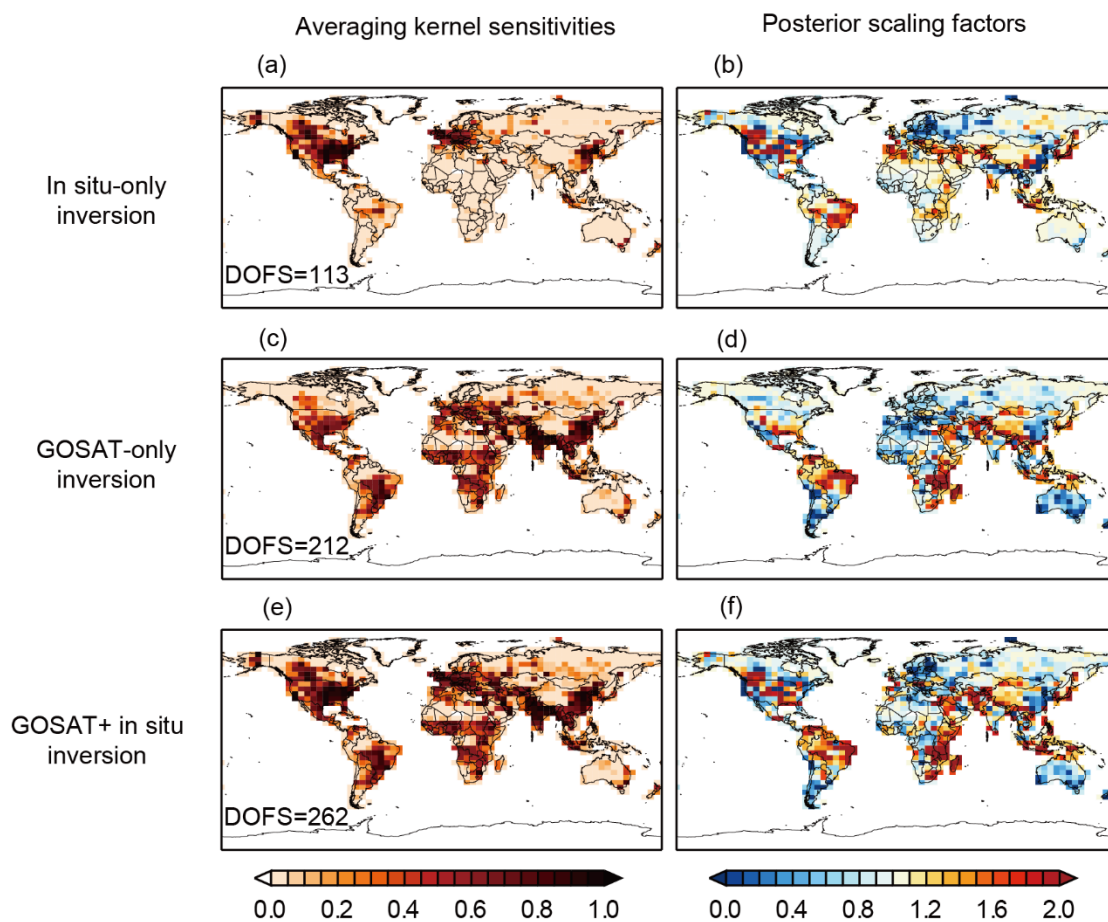
Figure 8. Optimization of mean 2010-2017 non-wetland (mainly anthropogenic) emissions. The in-situ-only inversion uses in situ observations, the GOSAT-only inversion uses GOSAT satellite observations, and the GOSAT + in situ inversion uses both. The left panels show the averaging kernel sensitivities (diagonal elements of the averaging kernel matrix) for each inversion, with the degrees of freedom for signal (DOFS, defined as the trace of the averaging kernel matrix) given inset. The right panels show the correction factors to the prior emissions (Figure 3a). Wetland emissions are corrected separately (see text).
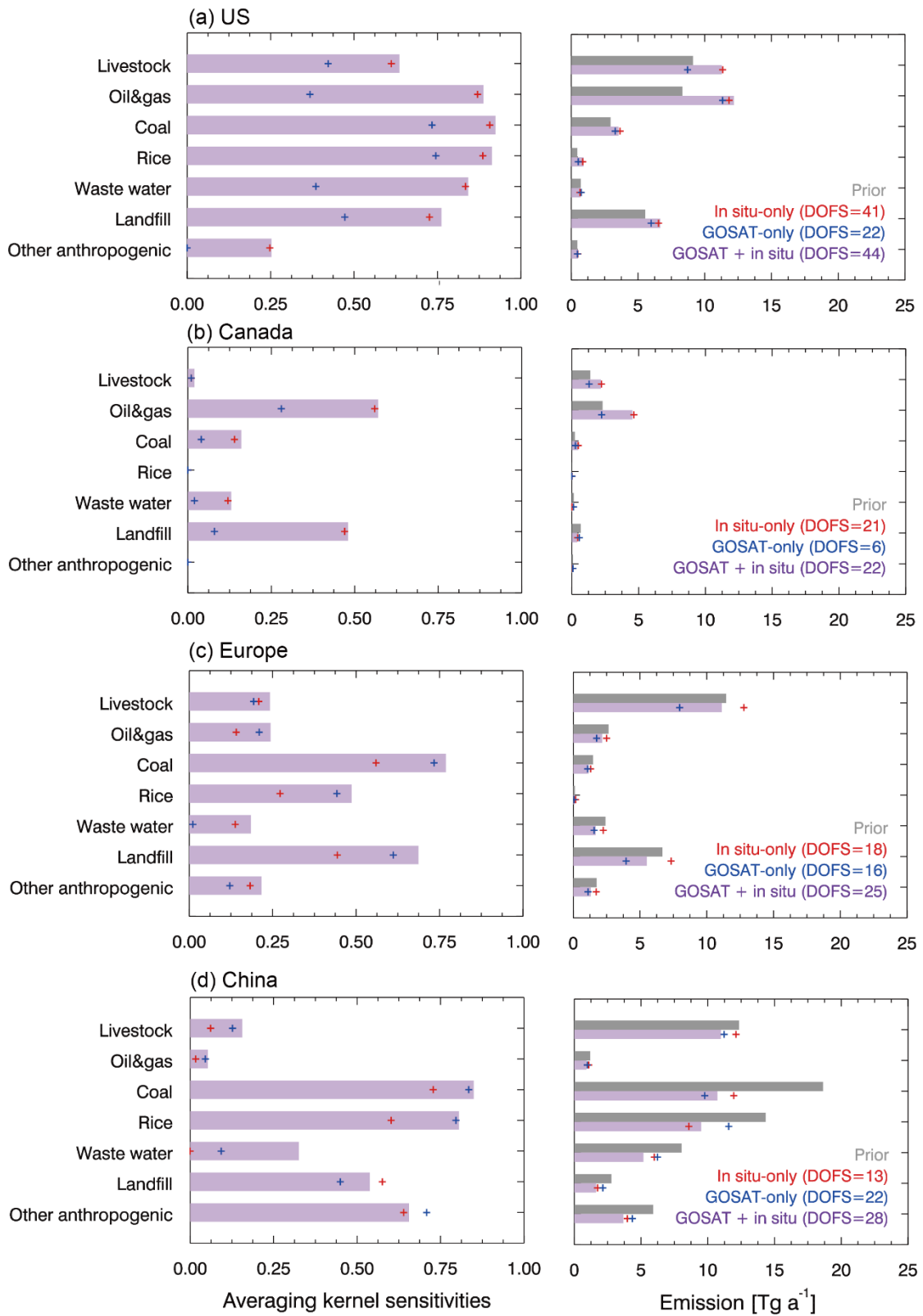
Figure 9. Optimization of anthropogenic methane emissions by source sectors in the in-situ-only, GOSAT-only, and GOSAT + in situ inversions. The left panel shows the averaging kernel sensitivities for each emission sector (see text for description), the right panel shows the emissions. Europe is defined as west of 30°E, which excludes Russia.
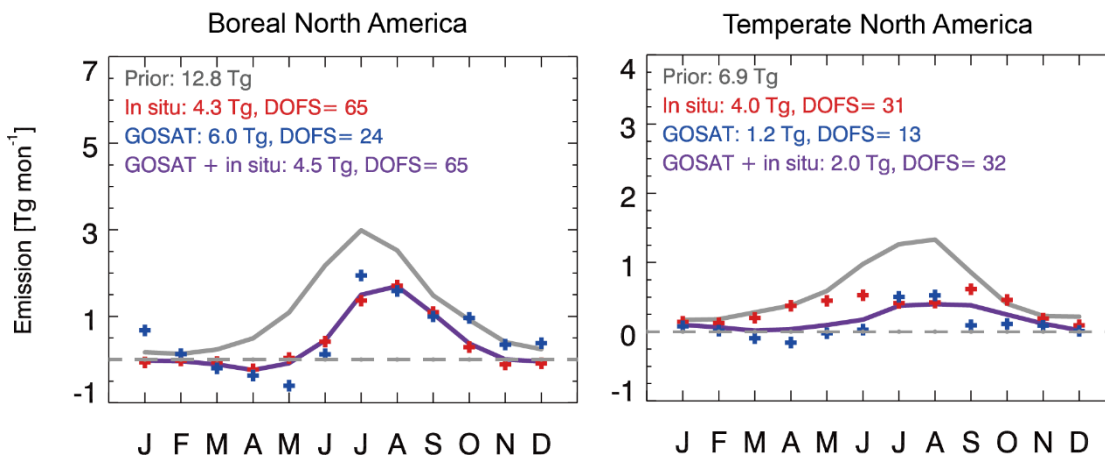
Figure 10. Wetland emissions in boreal and temperate North America (regions 2 and 3 of Figure 3). Prior and posterior estimates of the monthly mean wetland emissions averaged over 2010-2017 from different inversions are shown. Annual mean emissions and the degree of freedom for signal (DOFS) for monthly emissions in individual years are shown inset. Note differences in scale between panels. Negative emissions are allowed statistically by the inversion but are likely not physical.

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1076 **Figure 11.** Same as Figure 8 but for optimization of non-wetland (mainly anthropogenic) emission trends
1077 (% a⁻¹) in 2010-2017.
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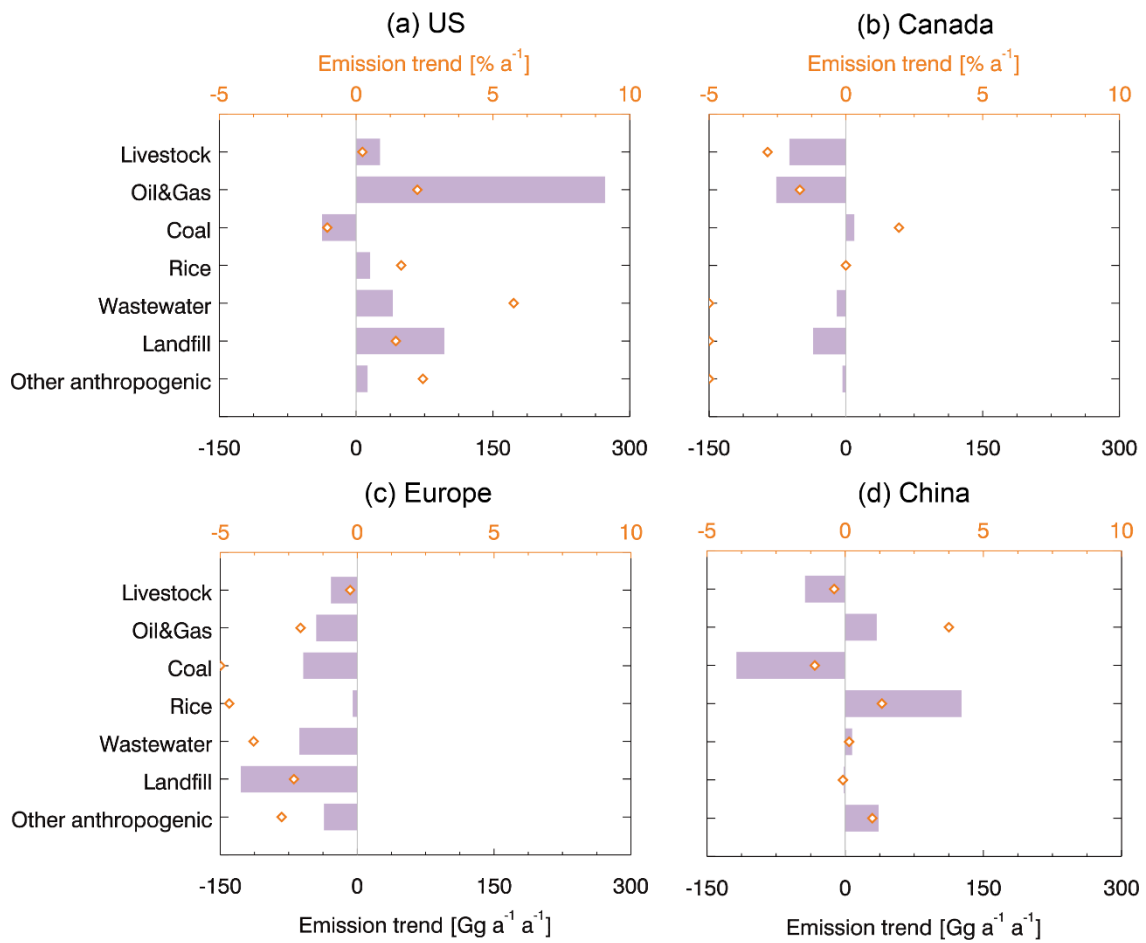
Figure 12. Optimization by sector of regional anthropogenic methane emission trends in 2010-2017. Bars and diamonds represent trends in Gg a^{-1} a^{-1} (bottom axis) and % a^{-1} (top axis) over the 2010-2017 period from the GOSAT + in situ joint inversion.
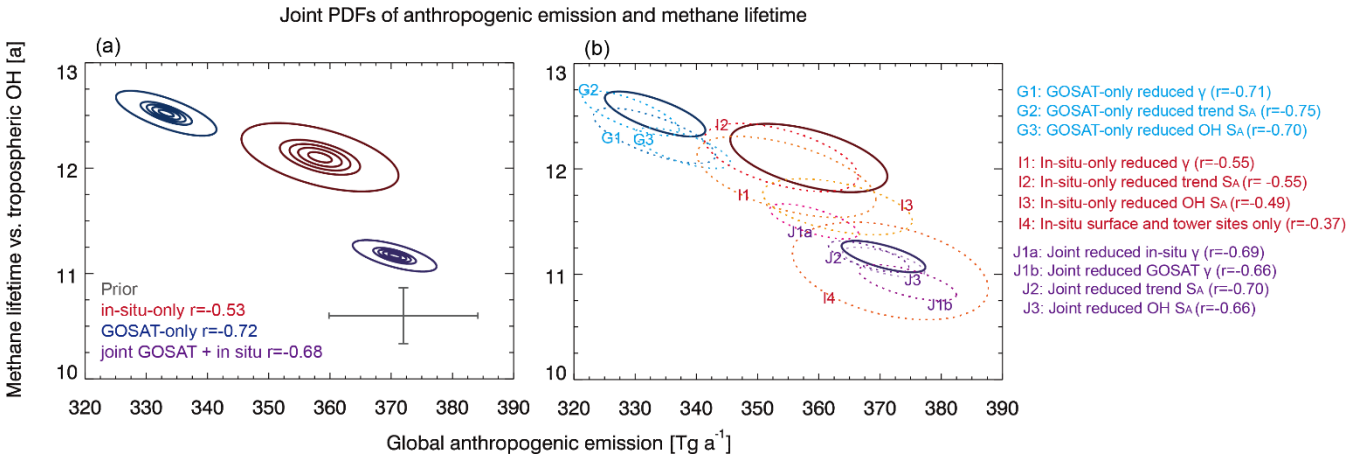
Figure 13. Joint probability density functions (PDFs) of global mean anthropogenic methane emission and methane lifetime against oxidation by tropospheric OH optimized by different inversions. Panel (a) shows the results from the prior and the three base inversions. The prior estimates are shown in grey with bars representing the prior error standard deviation. The thick contours show probabilities of 0.99 (outermost), 0.7, 0.5, 0.3, and 0.1 (innermost). The error correlation coefficients are given inset. Panel (b) shows the 0.99 probability contours from the three base inversions along with the same contours for ten additional sensitivity inversions using reduced values of the regularization parameter γ (0.05 instead of 0.1 for GOSAT, 0.5 instead of 1 for in situ); reduced errors for the methane emission trends on the 4°×5° grid (5% a^{-1} instead of 10% a^{-1}); reduced errors on annual hemispheric mean OH concentrations (5% instead of 10%); or surface and tower data only in the in-situ-only inversion.