



Optical and hygroscopic properties of black carbon influenced by particle microphysics at the top of anthropogenically polluted boundary layer

Shuo Ding¹, Dantong Liu¹, Delong Zhao^{2,3}, Kang Hu¹, Ping Tian^{2,3}, Fei Wang², Ruijie Li², Yichen Chen², Hui He², Mengyu Huang², Deping Ding²

¹Department of Atmospheric Sciences, School of Earth Sciences, Zhejiang University, Hangzhou, China

²Beijing Weather Modification Office, Beijing, China

³Beijing Key Laboratory of Cloud, Precipitation and Atmospheric Water Resources, Beijing, China

Corresponding to: dantongliu@zju.edu.cn



1 Abstract

2 Aerosols at the top of planetary boundary layer (PBL) could modify its atmospheric dynamics by redistributing
3 the solar radiation, and start to be activated to form low-level cloud at this layer. Black carbon (BC), as an
4 aerosol component efficiently absorbing solar radiation, can introduce heating and positive radiative effects at
5 this sensitive layer, especially in the polluted PBL over the continent. This study presents continuous
6 measurements of detailed BC properties at a mountain site locating at the top of polluted PBL over the North
7 China Plain, during seasons with contrast emission structure and meteorology. The pollution level was
8 persistently influenced by local surface anthropogenic emission on daily basis through daytime convective
9 mixing, but the concentration was also enhanced or diluted depending on air mass direction, defined as neutral,
10 polluted and diluted PBL, respectively. Winter was observed to have a higher BC mass fraction (4-8%) than
11 summer (2-7%). By resolving the detailed particle size-resolved mixing state of BC in optical and hygroscopic
12 models, we found enhanced BC mass absorption cross section (MAC_{BC}) for polluted PBL (up to $13 \text{ m}^2\text{g}^{-1}$ at
13 $\lambda=550\text{nm}$), and summer had a higher MAC_{BC} than winter by 5%. The higher BC mass fraction in winter
14 corresponded with a lower single-scattering albedo by 0.03- 0.09 than summer, especially the lowest for
15 diluted winter PBL (0.86 ± 0.02). The water supersaturation (SS) required to activate half number of BC
16 decreased from $0.21\pm 0.08\%$ to $0.1\pm 0.03\%$ for winter diluted and polluted PBL; from $0.22\pm 0.06\%$ to $0.17\pm 0.05\%$
17 for summer. Notably, at the top of anthropogenically polluted PBL in both seasons, the enlarged BC with
18 enhanced absorption capacity could be also efficiently droplet activated, e.g. winter (summer) BC with MAC
19 of 9.84 ± 1.2 (10.7 ± 1) $\text{m}^2\cdot\text{g}^{-1}$ could be half activated at $SS=0.13\pm 0.06\%$ ($0.18\pm 0.05\%$). These BC at the top of
20 the PBL can more directly interact with the free troposphere and be transported to a wider region, exerting
21 important direct and indirect radiative impacts.
22



23 1. Introduction

24 Black carbon aerosol (BC) is strongly shortwave absorbing, wielding important climate warming impact in
25 regional and global scale (Bond et al., 2013; Bond and Bergstrom, 2006). The emission of BC has large
26 regional heterogeneity with higher impact over polluted regions (Ramanathan and Carmichael, 2008). The
27 impacts due to BC heating are importantly determined by its vertical distribution in atmospheric column,
28 which could lead to a more stable or convective planetary boundary layer (PBL) (Ding et al., 2016; Koch and
29 Del Genio, 2010), e.g. an enhanced heating at higher level will depress the development of the layer below
30 (Chung et al., 2002; Ramanathan et al., 2005; Hansen et al., 2005), while the heating can promote the
31 convection above (Mcfarquhar and Wang, 2006; Rudich et al., 2003). Therefore, the properties of BC at the
32 top of PBL are important, which may result in contrast impacts in perturbing atmospheric dynamics. In
33 addition, surface emissions could serve regular cloud condensation nuclei (CCN) on daily basis through
34 daytime convective mixing in the PBL (Bretherton and Wyant, 1997; Wood and Bretherton, 2004). Aerosols
35 can be uplifted to the top of PBL and subsequently activated to incorporate into clouds. However, the way of
36 BC from surface sources to be activated, after the PBL processing during vertical transport, is yet to be
37 explicitly understood.

38 Though ground measurements of BC have been intensively conducted over the polluted North China Plain
39 (NCP) region in last decades (Han et al., 2009; Cheng et al., 2011; Ji et al., 2018; Liu et al., 2019a), they were
40 not able to represent the BC properties at the top of PBL. Additionally, recently conducted series of aircraft
41 measurements provided comprehensive information about vertical distributions of BC over this region
42 (Zhao et al., 2019; Ding et al., 2019a), but the manner of the aircraft measurement could not capture the
43 variation of BC at certain level in sufficient time resolution. The nature of diurnal pattern of PBL means not
44 only the surface concentration of pollutants is influenced by the daily evolution of PBL, but also the pollutants
45 at the top of PBL will have diurnal variation which need stable measurements to be characterized. This is
46 particularly the case if the top of PBL was continuously influenced by surface emissions. However such
47 information is lack at the top of PBL over the polluted NCP region, and many models still reply on surface
48 measurements to estimate the conditions at higher level (Guleria et al., 2014; Srivastava et al., 2012).

49 Microphysical properties of BC importantly determine its optical and hygroscopic properties. For example,
50 the presence of coatings associated with refractory BC (rBC) may enhance its absorbing capacity (Liu et al.,
51 2017), also enhances its hygroscopicity (Liu et al., 2013) and particle size, thereby modifying the potential to
52 be droplet activated (Chuang et al., 2002; Panicker et al., 2016; Ding et al., 2019b). This study is for the first
53 time to characterize the detailed BC microphysics at a mountain site located at the top of PBL, influenced by
54 surface emission on daily basis over the NCP region. We investigated the optical and hygroscopic properties
55 of BC at this level, as influenced by microphysical properties. Such information will support to constrain the
56 impacts of BC in influencing the PBL dynamics and low-level cloud formation over this anthropogenically
57 polluted region.

58

59 2. Site description, meteorology, cluster classification

60 Experiments in this study were performed in winter (Feb. - Mar.) and summer (Jun. - Jul.) of 2019 at a
61 mountain site (115.78°E, 40.52°N, 1344 m), locating on the north of Taihang ridge to the northwest of central
62 Beijing, shown in Fig 1. (a). The site is away from any local primary emissions, but the only sources of
63 pollutants are contributed by lower-level surface emissions and regional transport (the follow-up discussions).

64 Backward trajectories at the site are analyzed using the HYSPLIT 4.0 model (Draxler and Hess, 1998), for
65 every 6-hour during the experimental period. The meteorological field uses the $1^\circ \times 1^\circ$, 3-hourly GDAS1



66 reanalysis product and trajectories back to 48h are calculated. The backward trajectories are used for further
67 cluster analysis, to group the backward trajectories with similar transport pathway, whereby the homogeneity
68 of trajectories is maximized in each cluster, meanwhile the heterogeneity among different clusters is
69 maximized (Makra et al., 2011). This is achieved by analyzing the spatial variance of each trajectory and the
70 total variance in each pre-defined cluster. The iteration next simultaneously calculates and assigns the
71 trajectories to the eventually merged clusters. This analysis has been widely used to identify the main transport
72 pathways of air mass (Markou and Kassomenos, 2010; Philipp, 2009; Jorba et al., 2004; Grivas et al., 2008),
73 and is performed using the built-in module in HYSPLIT model software.

74

75 3. Instrumentation

76 All aerosol measurements were performed behind a PM_{2.5} impactor (BGI SCC1.829), and were dried by a
77 Nafion tube prior to the sampling of the instruments. BC was measured by a single-particle soot photometer
78 (SP2) (DMT Inc. USA). This instrument uses laser-induced technique to incandesce BC-containing particle
79 (Schwarz et al., 2006; Liu et al., 2010). The measured incandescence signal of individual BC particle can be
80 converted to a refractory BC (rBC) mass, which was calibrated using Aquadag® BC particle standard
81 (Acheson Inc.), and a factor of 0.75 was applied to correct for the ambient rBC mass (Laborde et al., 2012).
82 The core size (D_c) of BC is calculated from the measured rBC mass by assuming a material density of BC of
83 1.8 g/cm³ (Bond and Bergstrom, 2006). The scattering cross section of BC is derived by the leading edge only
84 (LEO) method (Gao et al., 2007) on the measured scattering signal of individual BC particle. The entire
85 particle size of BC containing particle (D_p) including coatings is determined by matching the measured
86 scattering cross section with the modelled one using a Mie lookup table (Liu et al., 2014; Taylor et al., 2015).
87 The bulk coating thickness (D_p/D_c) in a given time window is calculated as the cubic root of the total volume
88 of BC-containing particle weighted by the total volume of rBC (Liu et al., 2014):

$$89 \frac{D_p}{D_c} = \sqrt[3]{\frac{\sum_i D_{p,i}^3}{\sum_i D_{c,i}^3}} \quad (1),$$

90 where $D_{p,i}$ and $D_{c,i}$ are the diameters for the i^{th} single particle, respectively. The count (or mass) median
91 diameter (CMD, or MMD) is derived from a number (or mass) size distribution, below and above which size
92 the number (or mass) concentration is equal. Modelled mass absorption cross section (MAC) of individual
93 BC-containing particle is calculated by applying the Mie theory (Bohren, 1998) with a core-shell mixing state
94 assumption. The bulk MAC for a given time window can then be determined by the total MAC of each BC
95 particle divided by the total rBC mass, expressed as:

$$96 MAC = \frac{\sum_i MAC_i \times m_{rBC,i}}{\sum_i m_{rBC,i}} \quad (2),$$

97 where MAC_i and $m_{rBC,i}$ is the MAC and rBC mass for each particle respectively. An example to calculate
98 the MAC from single particle information is shown in Fig. 5f. The absorption coefficient k_{abs} (in Mm⁻¹) is
99 calculated as the MAC (m²·g⁻¹) multiplying the rBC mass concentration (μg·m⁻³) in each size bin, then
100 integrated throughout the size distribution:

$$101 k_{abs} = \sum_i MAC(D_{p,i}, D_{c,i}) m(\log D_{c,i}) \Delta \log D_{c,i} \quad (3),$$

102 where $m(\log D_{c,i})$ is the BC mass concentration at each D_c bin.

103 the volume ratio between coating and rBC could be obtained from D_c and D_p of individual BC:

$$104 \frac{\varepsilon_{coating}}{\varepsilon_{rBC}} = \left(\frac{D_p}{D_c}\right)^3 - 1 \quad (4),$$

105 where $\varepsilon_{coating}$ and ε_{rBC} is the volume fraction within BC particle. The hygroscopicity parameter of BC



106 (κ_{BCc}) could be calculated with known $\kappa_{coating}$ and $\kappa_{rBC}=0$, based on Zdanovskii–Stokes–Robinson (ZSR) rule
107 (Stokes and Robinson, 1966), expressed as,

$$108 \quad \kappa_{BCc} = \varepsilon_{coating} \times \kappa_{coating} \quad (5),$$

109 In this study, the main focus is the coating abundance distribution across all BC size distribution, but the
110 variation of coating composition is to a less importance, thus a constant $\kappa_{coating}=0.3$ is used to represent typical
111 environment containing aged aerosols (Pringle et al., 2010). An example to calculate κ_{rBC} from single
112 particle information is shown in Fig. 5g.

113 Total aerosol size distribution was measured by a Scanning Mobility Particle Size (SMPS, TSI Inc. Model
114 3936) at mobility diameter 15-700 nm. The total particle mass below diameter 1 μm (PM_{10}) is derived from
115 the SMPS size distribution by assuming a mean particle density of 1.45 g cm^{-3} (Liu et al., 2015; Cross and et
116 al., 2007). The scattering cross-section σ_{sc} (μm^2) of particle at all sizes is calculated by applying Mie
117 calculation assuming a refractive index (RI) of 1.48+0i (Liu et al., 2009) (a positive imaginary RI considering
118 the BC mass fraction is tested to have a minor influence in total scattering within 3%), thereby calculating the
119 scattering coefficient k_{sca} (in Mm^{-1}) of all aerosols:

$$120 \quad k_{sca} = \sum_i \sigma_{sc}(D_i) n(\log D_i) \Delta \log D_i \quad (6),$$

121 where $n(\log D_i)$ represents the particle number concentration at the i^{th} size bin. Single-scattering albedo (SSA)
122 is derived as the quotient of scattering coefficient (k_{sca}) divided by extinction coefficient ($k_{abs} + k_{sca}$).

123

124 4. Results and discussion

125 4.1 Classification of PBL

126 Clustered air masses for both seasons are shown in Fig. 2, and three clusters are classified for each season.
127 For both seasons, Cluster 1 (C1) represents the air masses from the most intensive anthropogenically
128 influenced regions according to the BC emission inventory (Fig. 1b). The seasonal difference is the winter
129 pollution was transported from the west over Shanxi and Hebei province, while was southerly in summer
130 mainly from the North China Plain over Hebei province. Cluster 2 (C2) passed over similar regions in both
131 seasons from the cleaner north. Cluster 3 (C3) both had a longer transport than other clusters, from northwest
132 and northeast in winter and summer respectively. The diurnal variations of PBL height (PBLH) corresponding
133 with periods of each cluster are shown in Fig. 2(d, h), with pronounced diurnal pattern. Apart from C3, winter
134 had a lower PBLH than summer, i.e. the mountain site was slightly above or below the well-developed PBL
135 top in winter and summer, respectively. The consistent diurnal variation of PBL means the mountain site was
136 persistently influenced by the surface sources through daytime convective mixing, when pollutants were
137 transported in the polluted PBL. However, the actual concentration will depend whether the site had been
138 contributed by additional sources or dilution (Fig. 1c).

139 Fig. 3 shows the temporal evolution of BC mass and PM for both seasons classified by air mass clusters,
140 with right panels showing the frequency under each cluster. Fig. 4 shows their diurnal variations. For all
141 experimental period, the concentration level of BC mass generally positively correlated with the PM mass,
142 both indicating the pollution levels. Among the three clusters, C2 had the medium mass concentration, with
143 over 85% frequency of the air masses contributed by local regions ($\pm 1^\circ$ around the measurement site), and
144 additional air masses were from the cleaner northerly direction, similar for both seasons (Fig. 2). It is likely
145 that this cluster was mainly contributed by local emission, and occasionally diluted to some extent by the air
146 mass from less polluted region. Here this cluster is defined as “neutral PBL”, to be discriminated with the
147 other two clusters. In winter C2 showed clear diurnal variation of BC and PM_{10} concentration with significant



148 enhancement in the daytime (Fig. 4b and e) when fully developed PBL (Fig. 2d). This diurnal pattern on the
149 mountain was contrast with the usual surface measurement, when the daytime PBL development could dilute
150 the concentration (Liu et al., 2019a; Han et al., 2009). This opposite trend on the mountain site was consistent
151 with the fully developed PBLH at 12:00-16:00 (Fig. 2d) when the top of PBL reached the mountain site at this
152 time, and pollutants were transported through convective mixing from the surface sources (Fig. 1c). In
153 midnight, the nocturnal depressed PBL trapped the surface pollution towards mountain, and the subsiding
154 cleaner air in free troposphere may have diluted the concentration of pollutants on the mountain site (Sullivan
155 et al., 1998; Bennett et al., 2010). The enhanced concentration from midday in summer was also observed (on
156 the mean, Fig. 4a) but not as pronounced as winter, probably due to some wet removal during the daytime
157 vertical transport from the surface, given the high RH in the summer (Fig. S1).

158 Consistent with the combined back-trajectory and emission analysis above, C1 had the highest BC for both
159 seasons (1.0 ± 0.5 and $0.4 \pm 0.2 \mu\text{g m}^{-3}$ for winter and summer, respectively) and PM mass (23.8 ± 10.3 and
160 $13.4 \pm 9.5 \mu\text{g m}^{-3}$). The concentration of BC mass was enhanced by a factor of 2.8 (1.7) higher than that in C2
161 for winter (summer), with winter having mass concentration frequently exceeding $1 \mu\text{g m}^{-3}$. It clearly shows
162 the diminished diurnal variation compared to the neutral PBL of C2, but all increased throughout the day and
163 night (Fig. 4a and d), particularly for winter. This is because besides the persistent influence of daytime
164 convective mixing as the neutral cluster in C2, C3 cluster had additional contribution from wider polluted
165 regions, hereby defined as “polluted PBL” (Fig. 1c). This contribution by regional transport was not related
166 to a diurnal pattern. The less enhancement in summer suggested the overall lower surface emission of the
167 surrounding regions in warm season.

168 C3 had the lowest pollution level among all air mass clusters with a lower BC and PM mass than C2 by a
169 factor of 2-4. Similarly, low BC ($0.09 \pm 0.03 \mu\text{g m}^{-3}$) and PM₁ mass ($1.8 \pm 0.4 \mu\text{g m}^{-3}$) were for both seasons. C3
170 represented the air masses from regions with low emissions (Fig. 1b). In addition, the faster transport (shown
171 as the longest path for back-trajectories among clusters, Fig. 2c) and the highest PBLH (Fig. 2d) in C3 could
172 efficiently dilute the pollution in the PBL, hereby termed as “diluted PBL”. Compared to C2, C3 can efficiently
173 disperse and reduce the pollutants being vertically transported to the mountain site from the surface, thus with
174 no apparent diurnal pattern of pollution concentrations neither (Fig. 4c).

175 BC mass fraction in PM₁ was overall higher in winter than summer, on average at $5.1 \pm 1.7\%$ and $3.6 \pm 2.3\%$
176 respectively, with both seasons showing a frequency distribution skewed to higher values up to 10% (Fig. 3 k
177 and l). Note that in summer the BC mass fraction among clusters had no discernible differences with only C2
178 showing to be slightly higher. The overall higher BC mass fraction in winter may result from the seasonal
179 variation on emission structure, that the additional primary emissions from heating activities may have
180 introduced more fraction of BC (Chow et al., 2011; Liu et al., 2018). However the lowered BC mass fraction
181 for the polluted PBL in winter (with significant reduction for fraction $>7\%$) may result from the enhanced
182 secondary formation under polluted environment (Volkamer et al., 2006; Hallquist et al., 2009). The
183 comparable BC mass fraction between C1 and C3 in summer (but all lower than neutral PBL) may result from
184 the dominant process of enhanced secondary formation (reducing PM₁) and particle scavenging (reducing BC),
185 for polluted and diluted PBL, respectively. There was a notably higher BC mass fraction at $6.7 \pm 1.5\%$ for the
186 diluted PBL in winter, suggesting a much reduced secondary aerosols in less-polluted environment, and the
187 removal process of BC had been less efficient than other substances (Koch and Del Genio, 2010).

188

189 4.2 Particle microphysics of BC

190 BC core size showed no variation among all PBL types in both seasons. The mass (count) median diameter
191 of BC core is 203 ± 12 nm (106 ± 7 nm) and 167 ± 11 nm (85 ± 5 nm) for winter and summer respectively. The BC



192 core size could be used to discern emission sources, such as BC from solid fuel burning tended to have larger
193 MMD compared to those from traffic sources (Liu et al., 2014). The larger MMD in winter than summer was
194 consistent with previous observation over this region (Ding et al., 2019a;Hu et al., 2020), which may indicate
195 additional sources from residential solid fuel burning for heating in cold season.

196 The coatings relative to BC, as reflected by D_p/D_c , was positively correlated with the pollution level of the
197 PBL (represented by PM_{10} mass loading) in winter (Fig. 5a), and in summer from neutral to polluted PBL (Fig.
198 5b). The polluted PBL in winter had the highest D_p/D_c (1.76 ± 0.3), in line with the most polluted condition.
199 For both neutral and diluted PBL, winter had higher BC coatings than summer. Apart from some periods in
200 summer, the diluted PBL occasionally showed higher D_p/D_c up to 1.8, comparable with the high end of that
201 for polluted PBL.

202 The warmer temperature in summer may have caused a more likely tendency for some semi-volatile particle
203 species to evaporate, hereby reduced coatings on BC than in winter even at the same pollution level, which is
204 consistent with previous ground studies over this region (Liu et al., 2019a). The highest coatings were
205 associated with the most polluted condition, implying that the regional transport of air mass from wider
206 polluted region may have advected both primary emissions and gas-precursors, during when some gas
207 partitioning processes may have occurred on BC during transport and caused the high coatings. Under clean
208 environment, summer showed remarkably higher coatings than winter, which may result from a more intense
209 solar radiation received at the mountain site in summer, where photochemical reactions may have caused
210 significant formation of secondary particulate matter (Xu et al., 2017; Bao et al., 2018; Liu et al., 2019b).

211 The detailed size distributions of uncoated and coated BC are shown in Fig. 5 c-e. Here three typical
212 examples are chosen to represent the polluted and diluted PBL in winter (case 1 and case 2), and polluted PBL
213 in summer (case 3). Comparing between case 1 and case 2, it showed the polluted period had caused increased
214 size for all particle from peak diameter (d) less than $d=30$ nm to the accumulation mode of $d=120$ nm. Coated
215 BC diameter peaked from 150 to 220 nm. It means the condensation process had occurred on all particles and
216 enlarged their sizes including BC. The coated BC in most polluted PBL has significantly extended its coated
217 size up to $d=400-700$ nm. Note that for the clean PBL in winter, the coated BC was populated at $d=150$ nm,
218 leading to a higher BC number fraction of 22% at this size. For polluted PBL in summer, peaking diameter of
219 all particles ($d=450$ nm) was lower than winter, and coated BC also peaked at a smaller size ($d=153$ nm). Fig.
220 S3 shows almost consistent coated BC sizes in summer (120-130 nm), which was smaller than winter because
221 of the smaller core size (though a higher D_p/D_c).

222 Detailed core-size resolved mixing states, expressed in a space of coated versus uncoated BC diameter, are
223 shown in Fig. 5 f-h, corresponding to the three cases above. $D_p=D_c$ indicates no coatings on BC and coating
224 increases when D_p is larger than D_c . For polluted PBL in winter, most BC population lied between 100-200
225 nm core diameter and 200-300 nm coated diameter, in contrast with the diluted PBL that a significant fraction
226 of uncoated BC and coated diameter at 100-140 nm. Summer polluted PBL showed reduced fraction of bare
227 BC but a range of coating thickness for coated BC. Such analysis in mapping the size-resolved mixing state
228 of single BC particle on the space of uncoated-coated size, can resolve the optical (contour in Fig. 5f) and
229 hygroscopic (contour in Fig. 5g) properties that BC particle microphysics could influence, which will be
230 discussed next.

231

232 4.3 Optical properties of BC

233 Fig. 6 a-b and Fig. S4 show the mass absorption cross section (MAC) of uncoated and coated BC for different
234 PBL types in both seasons. Summer showed systematically higher MAC ($7.2 \pm 0.1 \text{ m}^2 \cdot \text{g}^{-1}$) than winter (6.6 ± 0.3
235 $\text{m}^2 \cdot \text{g}^{-1}$) for uncoated BC, because the BC core was smaller in summer than winter (Fig. S2 and Fig. S3). The



236 MAC for coated BC was largely modulated by coatings, showing positive correlation with PM_{10} in winter
237 ($MAC_{BC,coated}=0.09\times PM_{10}+8.7$, $r=0.79$) and summer ($MAC_{BC,coated}=0.12\times PM_{10}+9.5$, $r=0.5$), corresponding with
238 the diluted to polluted PBL types. The enhanced absorption efficiency of coated BC with increased pollution
239 level was consistent with previous studies over this region (Zhang et al., 2018; Ding et al., 2019a). The mean
240 $MAC_{BC,coated}$ for polluted PBL in summer ($11\pm 1\text{ m}^2\cdot\text{g}^{-1}$) was higher than that in winter ($10.7\pm 0.9\text{ m}^2\cdot\text{g}^{-1}$) by
241 2%. The neutral and diluted PBL also had higher MAC in summer than winter (by 8% and 22% respectively).
242 The overall higher absorption efficiency of BC in summer (especially in diluted PBL) was due to the smaller
243 core size, leading to a higher baseline MAC for uncoated BC by 10% than winter. This effect prevailed the
244 even lower coating amount and associated absorption enhancement in summer (Fig. 5b). Notably, the MAC_{BC}
245 for diluted summer PBL had been enhanced by both reduced BC core size and increased coatings (Fig. 5b)
246 reaching almost equivalent MAC compared to winter neutral PBL. This means at the top of the clean PBL in
247 summer, there was a BC layer with high absorbing capacity (in contrast to winter with much lower MAC),
248 which may efficiently absorb the strong solar radiation.

249 The single-scattering albedo at $\lambda=550\text{nm}$ (SSA_{550}) in winter was systematically lower for all PBL types (Fig.
250 6 c-d and Fig. S4), lowered by 0.06, 0.05 and 0.08 than summer for diluted, neutral and polluted PBL,
251 respectively. The decreased SSA_{550} was in line with the increased BC mass fraction (Fig. 3), and also influenced
252 by the absorbing efficiency. Our winter results here were comparable to previous aircraft measurements over
253 Beijing region in cold season of 2016, with SSA ranging from 0.8 to 0.98 from clean to heavily polluted period
254 in the PBL (Tian et al., 2020). For the diluted PBL in winter, SSA_{550} could reach as low as 0.86 ± 0.02 when
255 BC mass fraction reached $7\pm 1.5\%$ (Fig. 3k). This was within the range of that study in clean period at
256 $SSA=0.8-0.85$ in the PBL, and the neutral and polluted PBL were within similar range in transition and fully
257 polluted PBL at $SSA=0.85-0.98$, suggesting the clusters here may represent different development stages of
258 pollution events. The lower SSA in winter PBL than summer tended to induce radiative effects towards more
259 positive effect (Hansen et al., 1981; Haywood and Ramaswamy, 1998) at the top of PBL, in particular for the
260 clean winter PBL.

261

262 4.4 Hygroscopic properties of BC

263 The droplet activation of BC is determined by the particle size and hygroscopicity, where the coatings play
264 roles in enlarging entire BC size (Dusek et al., 2006a; Dusek et al., 2006b) and increasing its hygroscopicity
265 (Liu et al., 2013). Both factors can be obtained by our observation of size-solved mixing state of BC. The
266 coated BC diameter based on number concentration (coated CMD) is shown in Fig. 7 a, b. The CMD of coated
267 BC in winter polluted PBL was $0.19\pm 0.05\ \mu\text{m}$, which were the largest among PBL types. Generally, coated
268 BC CMD was in positive correlation with the PM_{10} , in winter fitted as $CMD=0.003\times PM_{10}+0.14$ ($r=0.69$) and in
269 summer as $CMD=0.001\times PM_{10}+0.11$ ($r=0.73$). The coated BC CMD in winter PBL was larger than summer by
270 14 nm, 35 nm and 56 nm for diluted, neutral and polluted PBL, respectively.

271 The hygroscopicity parameter of BC-containing particle (κ_{BCc}) at the size of CMD is positively correlated
272 with coating content. In winter κ_{BCc} was at 0.15 ± 0.02 , 0.20 ± 0.03 and 0.24 ± 0.03 for diluted, neutral and
273 polluted PBL respectively, while the corresponding values were 0.19 ± 0.03 , 0.20 ± 0.03 and 0.22 ± 0.03 in
274 summer, consistent with the reduced D_p/D_c from polluted PBL to diluted PBL for both seasons. κ_{BCc} could be
275 fitted as $\kappa_{BCc}=0.003\times PM_{10}+0.17$ ($r=0.77$) and $\kappa_{BCc}=0.004\times PM_{10}+0.17$ ($r=0.49$) for winter and summer
276 respectively. Note that summer diluted PBL had significantly higher κ (0.19 ± 0.03) than winter for the same
277 PBL type.

278 After obtaining the CMD of coated BC and the corresponding κ_{BCc} , a critical water superstation (SS) could
279 be derived from the Köhler model (Fig. S5) for BC at the size of CMD to be activated. By assuming that all
280 of the population larger than CMD could also be consequently activated (because of the larger particle size



281 and higher hygroscopicity), the SS obtained above is thus the lower estimate for half of the number population
282 of BC to be activated, termed as SS_{half} . In line with the increased coated BC size and particle hygroscopicity,
283 SS_{half} decreased with increased pollution level, from $0.21 \pm 0.08\%$ to $0.1 \pm 0.03\%$ for winter diluted to polluted
284 PBL; from $0.22 \pm 0.06\%$ to $0.17 \pm 0.05\%$ for summer. This highlights the lowest possible SS required to activate
285 BC in polluted winter PBL, and for the same PBL type, summer will need a higher SS, apart from occasionally
286 some lower SS_{half} for summer diluted PBL. This potential CCN ability of BC is derived from the physical
287 properties of BC itself, but the actual activation of BC depends on the ambient superstation condition which
288 is determined by the size distribution of existing droplets and other aerosols competing CCN (Pruppacher et
289 al., 1998; Mcfiggans et al., 2005). The results here generally consistent with a previous surface measurement
290 of BC CCN-activation in urban Nanjing as constrained by size-resolved compositions (Wu et al., 2019), when
291 an activation fraction of 33% at $SS=0.1\%$ was found. Previous study using flight measurements over NCP
292 region (Ding et al., 2019a) found a $SS=0.08\%$ required to activate half of the BC number in heavy pollution
293 condition, consistent with the polluted PBL here.

294

295 5. Conclusion

296 By performing continuous measurement on a mountain site located at the top of planetary boundary layer
297 (PBL) over the north China Plain region in winter and summer, the optical and hygroscopic properties of BC
298 were investigated. We identified three types of PBL, all persistently influenced by surface anthropogenic
299 emission on daily basis through daytime convective mixing, but could be either enhanced or diluted subject
300 to received air masses. By investigating the detailed microphysical properties of BC, this study provides a
301 clear picture of optical and hygroscopic characteristics of BC at the top of anthropogenically influenced PBL.
302 Highlighted information includes higher BC mass fractions in winter than summer, corresponding with a lower
303 single-scattering albedo by 0.05-0.08, especially the lowest for diluted winter PBL (0.86 ± 0.02); both mass
304 absorbing efficiency and CCN ability of BC are positively correlated with the pollution level of PBL, due to
305 enhanced coating content under more polluted environment, e.g. from diluted to polluted PBL, coating content
306 increased by 39% (11%), absorbing efficiency increased by 31% (10%), and the water supersaturation in
307 activating half number of BC decreased by 53% (26%) for winter and summer respectively. It clearly
308 demonstrates that BC with higher coating content could be efficiently incorporated into liquid clouds, and
309 meanwhile these BC had high absorbing capacity, which means these highly-absorbing BC may have great
310 potentials for in-cloud heating (Jacobson, 2012; Nenes and et al., 2002).

311 Compared to surface measurement, the results here are more directly linked to the aerosol properties closer
312 to the condensation level which is subsequently CCN-activated. BC located at this layer more importantly
313 determines its heating impacts due to receiving a stronger solar radiation. Rather than being subject to
314 significant scavenging processes of low-level emissions, BC transported to the lower free troposphere may be
315 transported to a wider region (Weinzierl, 2008; Yang et al., 2018), exerting regional direct and indirect radiative
316 impacts.

317 Acknowledgment

318 This research was supported by the National Key Research and Development Program of China
319 (2016YFA0602001), National Natural Science Foundation of China (41875167, 41675038).

320

321 There are no potential conflicts of interest.

322 Data for this study are available from the file sharing link (<https://pan.baidu.com/s/1YqnLHZly24URgULZ-HIIOW>) using extracting code u457.

324

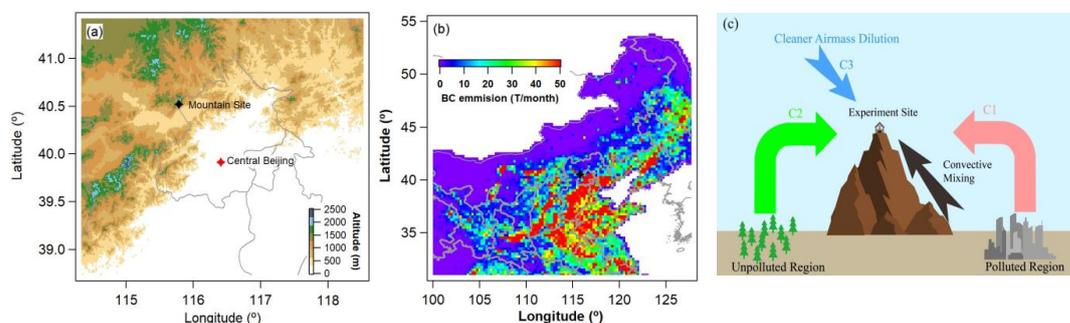


325 **Author Contributions**

326 DL, DZ and DD led and designed the study. SD, DL, DZ, KH, PT, RL, YC, FW, KB, HH, and MH set up and
327 conducted the experiment. SD, DL and KH contributed to the data process. SD, DL wrote the paper.

328

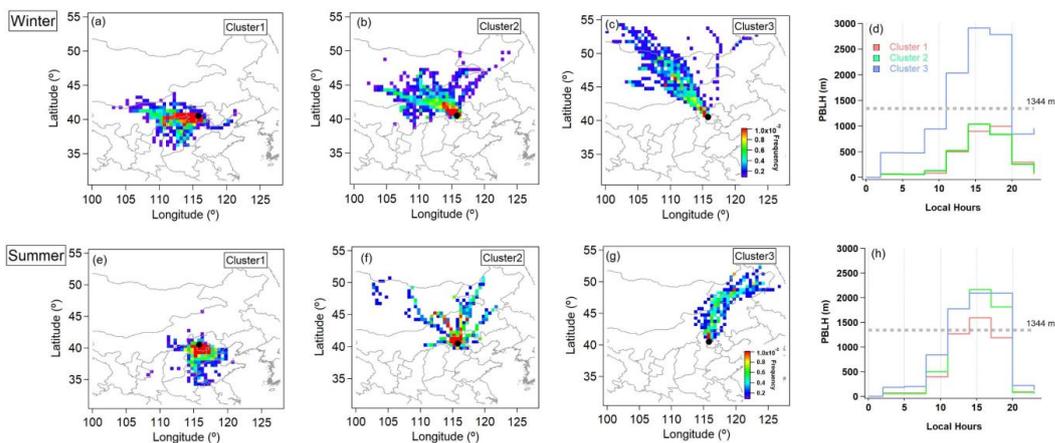
329 **Figures and captions**



330

331 Fig. 1. Experimental site descriptions. (a) The location of the experimental site and central Beijing, marked with
332 black and red star respectively, where the color bar denotes the terrain height. (b) the monthly BC emission
333 inventory in China (Li et al., 2017). (c) Schematic illustration for different types of PBL defined in this study.

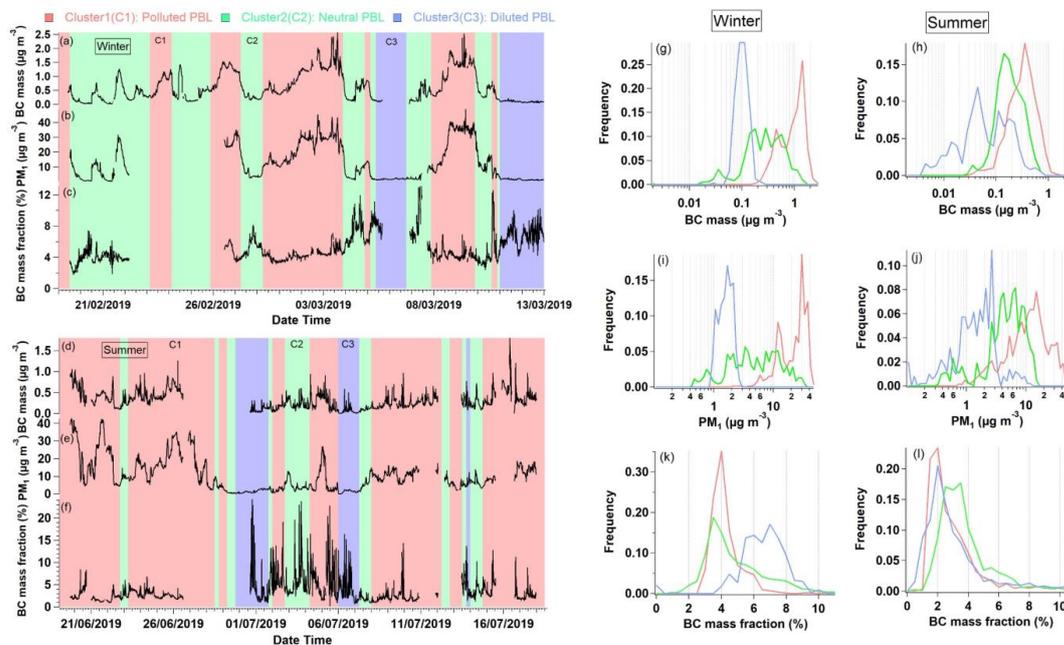
334



335

336

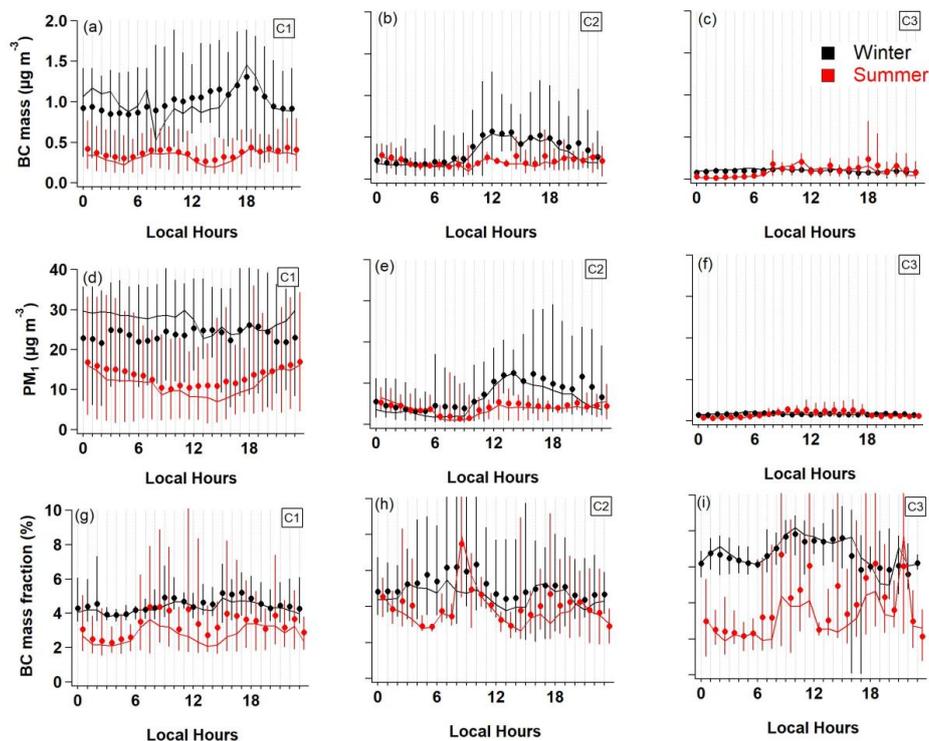
337 Fig. 2. The clustered backward trajectories from HYSPLIT model in both seasons: (a)-(c) for winter and (e)-
338 (g) for summer, colored by occurrence frequency in each geographic grid. (d) and (h) are the diurnal variation
339 of the height of PBL for the three clusters in both seasons, with the dashed line denoting the mountain site
340 altitude.
341



342

343

344 Fig. 3. Time series of BC mass (a)(d), PM_{10} (b)(e), BC mass fraction (c)(f) at both seasons, shaded by the
345 periods in identified three clusters, by red, green and blue corresponding to cluster 1 (C1), cluster 2 (C2),
346 cluster 3 (C3) respectively. Frequency histograms of BC mass (g)(h), PM_{10} (i)(j), and BC mass fraction (k)(l)
347 for each cluster in both seasons.

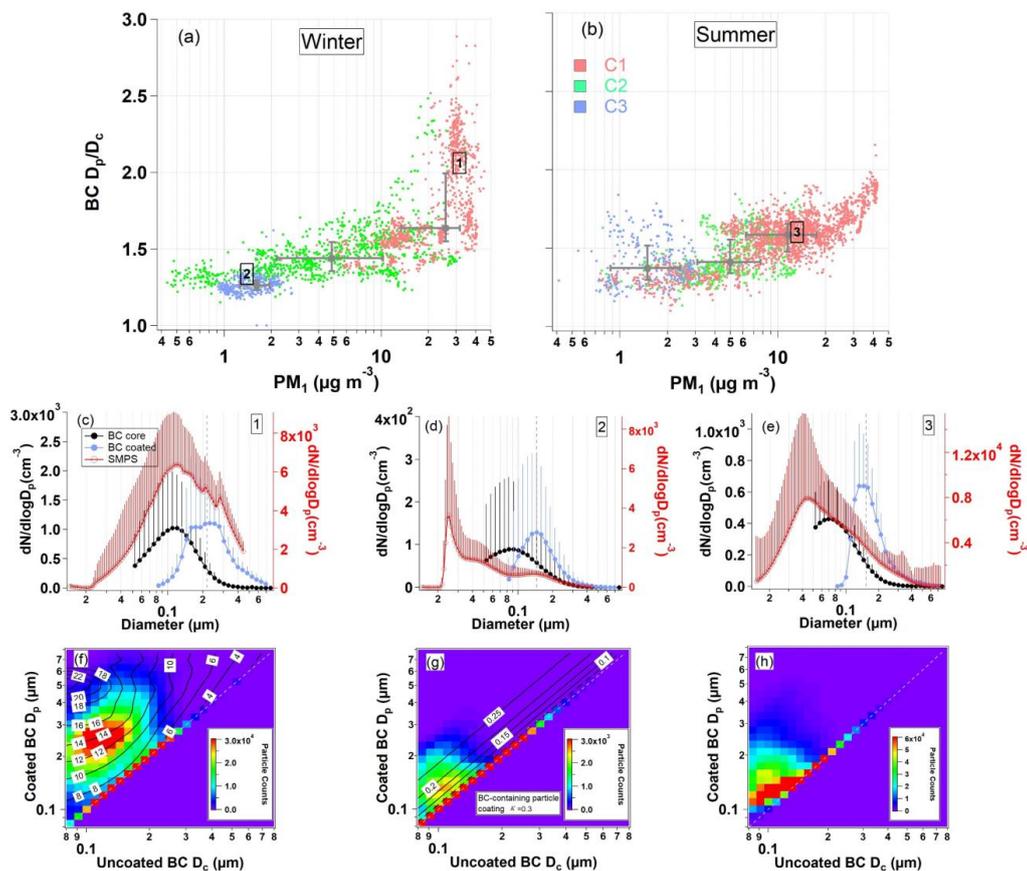


348

349

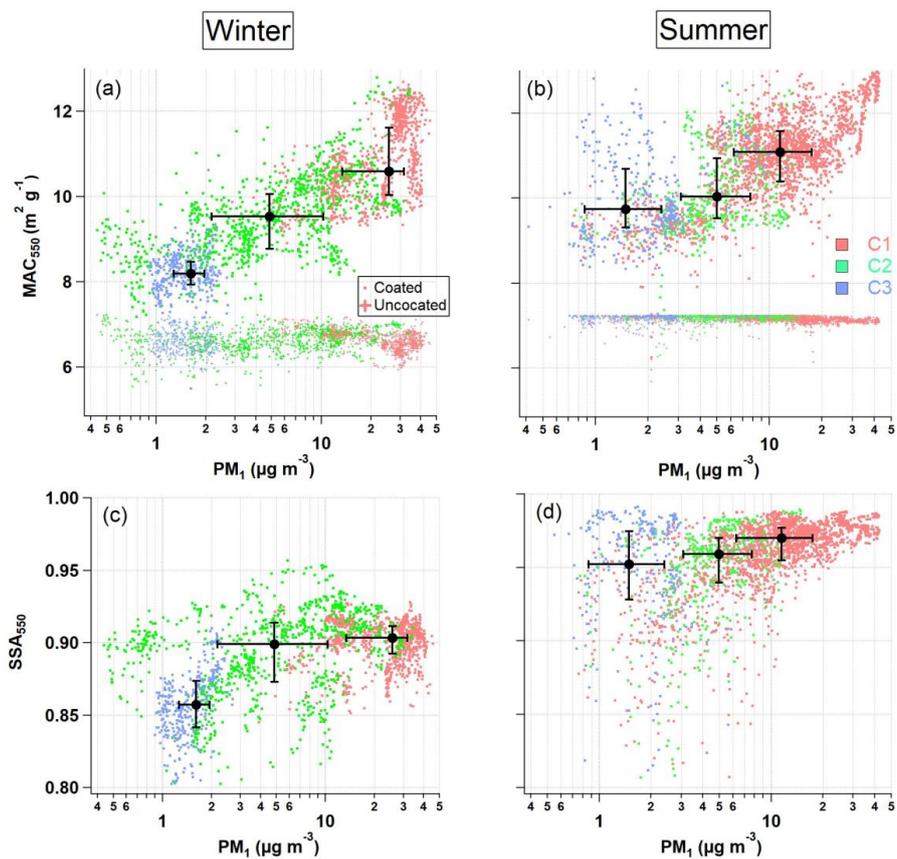
350 Fig. 4. Diurnal variation of BC mass (a-c), PM_{10} (d-f) and BC mass fraction (g-i) for the three PBL types in
351 both seasons, where black and red denote the winter and summer respectively. The solid circles, lines and
352 whiskers denote the mean, median, 25th, 75th percentile respectively.

353



354
 355
 356
 357
 358
 359
 360
 361
 362
 363

Fig. 5. Size-resolved mixing state of BC. The bulk relative coating thickness (D_p/D_c) as a function of PM_{10} in winter (a) and summer (b) for the three PBL types, with solid circles, whiskers denoting the median, 25th, 75th percentiles respectively. Three typical periods, as marked as 1-3 in (a) and (b), are extracted for size distribution analysis. (c),(d) and (e) are the corresponding number size distribution of all particles, uncoated and coated BC for period 1-3 respectively. The bottom panels are coated BC diameter as a function of uncoated BC diameter, colored by number density of single particle, where (f) and (g) are mapped with contour lines numbered by the MAC and κ_{BCc} respectively.

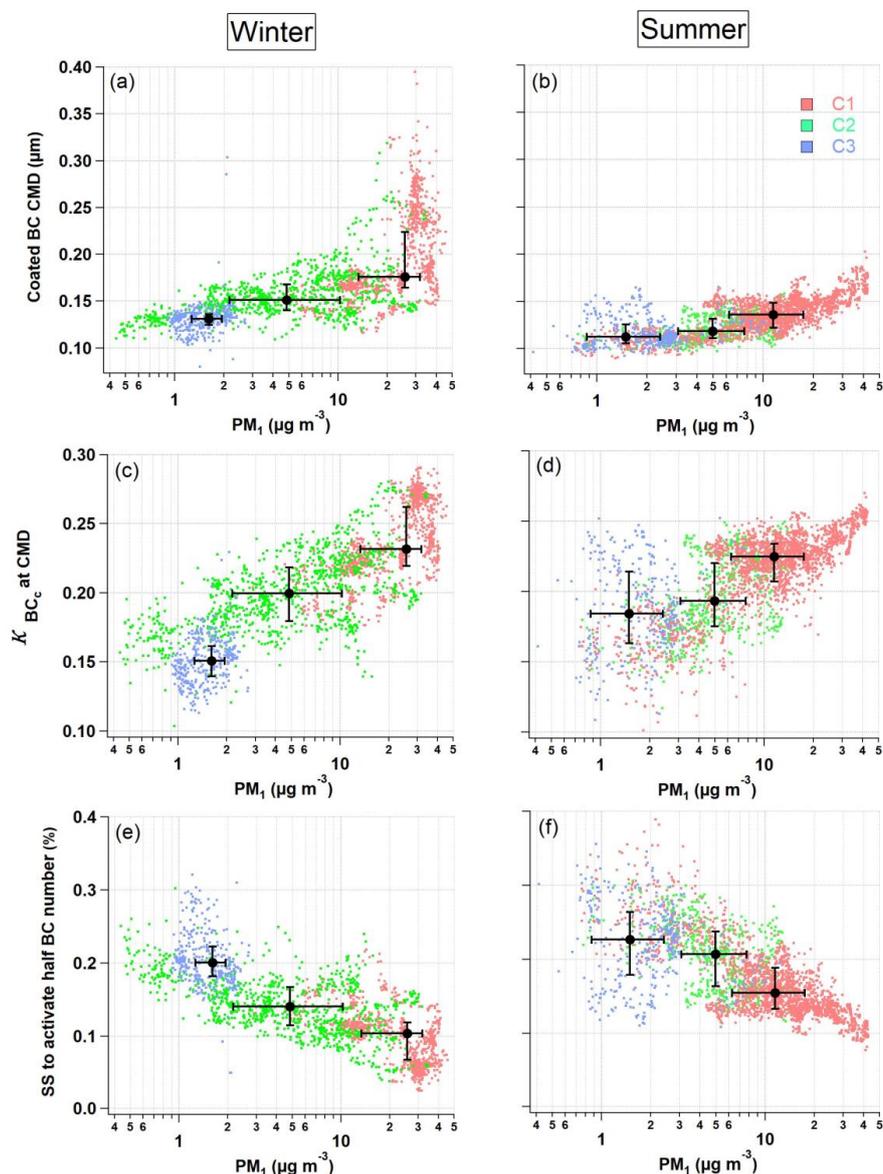


364

365

366 Fig. 6. Optical properties of BC for the three PBL types in both seasons. (a)(b) are mass absorption cross
367 section at $\lambda=550$ nm (MAC_{550}), with dot and plus markers denoting coated and uncoated BC respectively.
368 (c)(d) are single-scattering albedo at $\lambda=550$ nm (SSA_{550}). In each panel, the solid circles and whiskers denote
369 the median, 25th and 75th percentile respectively.

370



371
372
373
374
375
376
377
378

Fig. 7. Hygroscopic properties of BC for the three PBL types in both seasons. (a)(b) are the count median diameter of coated BC; (c)(d) are the κ_{BCc} assuming $\kappa_{\text{coating}}=0.3$; (e)(f) are the supersaturation (SS) to activate half of the BC number population. In each panel, the solid circles and whiskers denote the median, 25th and 75th percentile respectively.

379 Reference

- 380 Bao, F., Li, M., Zhang, Y., Chen, C., and Zhao, J.: Photochemical Aging of Beijing Urban PM2.5: HONO Production, Environmental
381 Science & Technology, 52, 6309-6316, 10.1021/acs.est.8b00538, 2018.
- 382 Bennett, L. J., Weckwerth, T. M., Blyth, A. M., Geerts, B., Miao, Q., and Richardson, Y.: Observations of the Evolution of the



- 383 Nocturnal and Convective Boundary Layers and the Structure of Open-Celled Convection on 14 June 2002, *Monthly Weather*
384 *Review*, 138, 2589-2607, 2010.
- 385 Bohren, C. F., and Huffman, D. R.: Absorption and Scattering by a Sphere, in: *Absorption and Scattering of Light by Small Particles*,
386 82-129, 1998.
- 387 Bond, T. C., and Bergstrom, R.: Light Absorption by Carbonaceous Particles: An Investigative Review, *Aerosol Science and*
388 *Technology*, 40, 27-67, 2006.
- 389 Bond, T. C., Doherty, S. J., Fahey, D. W., Forster, P. M., Berntsen, T., DeAngelo, B. J., Flanner, M. G., Ghan, S., Kärcher, B., Koch,
390 D., Kinne, S., Kondo, Y., Quinn, P. K., Sarofim, M. C., Schultz, M. G., Schulz, M., Venkataraman, C., Zhang, H., Zhang, S., Bellouin,
391 N., Guttikunda, S. K., Hopke, P. K., Jacobson, M. Z., Kaiser, J. W., Klimont, Z., Lohmann, U., Schwarz, J. P., Shindell, D., Storelvmo,
392 T., Warren, S. G., and Zender, C. S.: Bounding the role of black carbon in the climate system: A scientific assessment, *Journal of*
393 *Geophysical Research: Atmospheres*, 118, 5380-5552, 10.1002/jgrd.50171, 2013.
- 394 Bretherton, C. S., and Wyant, M. C.: Moisture Transport, Lower-Tropospheric Stability, and Decoupling of Cloud-Topped Boundary
395 Layers, *Journal of the Atmospheric Sciences*, 54, 148-167, 1997.
- 396 Cheng, Y., Su, H., Rose, D., Gunthe, S. S., Berghof, M., Wehner, B., Achtert, P., Nowak, A., Takegawa, N., and Kondo, Y.: Size-
397 resolved measurement of the mixing state of soot in the megacity Beijing, China: diurnal cycle, aging and parameterization,
398 *Atmospheric Chemistry and Physics*, 12, 4477-4491, 2011.
- 399 Chow, J. C., Watson, J. G., Lowenthal, D. H., Antony Chen, L. W., and Motallebi, N.: PM_{2.5} source profiles for black and organic
400 carbon emission inventories, *Atmospheric Environment*, 45, 5407-5414, <https://doi.org/10.1016/j.atmosenv.2011.07.011>, 2011.
- 401 Chuang, C. C., Penner, J. E., Prospero, J. M., Grant, K. E., Rau, G. H., and Kawamoto, K.: Cloud susceptibility and the first aerosol
402 indirect forcing: Sensitivity to black carbon and aerosol concentrations, *Journal of Geophysical Research*, 107, 4564, 2002.
- 403 Chung, C. E., Ramanathan, V., and Kiehl, J. T.: Effects of the south Asian absorbing haze on the northeast monsoon and surface-air
404 heat exchange, *Journal of Climate*, 15, 2462-2476, 2002.
- 405 Cross, E. S., and et al.: Laboratory and ambient particle density determinations using light scattering in conjunction with aerosol
406 mass spectrometry, *Aerosol Sci. Technol.*, 41, 343, 2007.
- 407 Ding, A., Huang, X., Nie, W., Sun, J., Kerminen, V. M., Petaja, T., Su, H., Cheng, Y., Yang, X., and Wang, M.: Enhanced haze
408 pollution by black carbon in megacities in China, *Geophysical Research Letters*, 43, 2873-2879, 2016.
- 409 Ding, S., Liu, D., Zhao, D., Hu, K., Tian, P., Zhou, W., Huang, M., Yang, Y., Wang, F., and Sheng, J.: Size-Related Physical Properties
410 of Black Carbon in the Lower Atmosphere over Beijing and Europe, *Environmental Science & Technology*, 53, 11112-11121, 2019a.
- 411 Ding, S., Zhao, D., He, C., Huang, M., He, H., Tian, P., Liu, Q., Bi, K., Yu, C., Pitt, J., Chen, Y., Ma, X., Chen, Y., Jia, X., Kong, S.,
412 Wu, J., Hu, D., Hu, K., Ding, D., and Liu, D.: Observed Interactions Between Black Carbon and Hydrometeor During Wet
413 Scavenging in Mixed-Phase Clouds, *Geophysical Research Letters*, 46, 8453-8463, 10.1029/2019gl083171, 2019b.
- 414 Draxler, R. R., and Hess, G.: An overview of the HYSPLIT_4 modelling system for trajectories, *Aust. Meteorol. Mag.*, 47, 295,
415 1998.
- 416 Dusek, U., Frank, G., Hildebrandt, L., Curtius, J., Schneider, J., Walter, S., Chand, D., Drewnick, F., Hings, S. S., and Jung, D.: Size
417 matters more than chemistry for cloud-nucleating ability of aerosol particles, *Science*, 312, 1375-1378, 2006a.
- 418 Dusek, U., Reischl, G. P., and Hitznerberger, R.: CCN activation of pure and coated carbon black particles, *Environmental Science*
419 *& Technology*, 40, 1223-1230, 2006b.
- 420 Gao, R. S., Schwarz, J. P., Kelly, K. K., Fahey, D. W., Watts, L. A., Thompson, T. L., Spackman, J. R., Slowik, J. G., Cross, E. S.,
421 and Han, J.: A Novel Method for Estimating Light-Scattering Properties of Soot Aerosols Using a Modified Single-Particle Soot
422 Photometer, *Aerosol Science and Technology*, 41, 125-135, 2007.
- 423 Grivas, G., Chaloulakou, A., and Kassomenos, P.: An overview of the PM₁₀ pollution problem, in the Metropolitan Area of Athens,
424 Greece. Assessment of controlling factors and potential impact of long range transport, *Science of The Total Environment*, 389, 165-
425 177, <https://doi.org/10.1016/j.scitotenv.2007.08.048>, 2008.
- 426 Guleria, R. P., Kuniyal, J. C., Dhyani, P. P., Joshi, R. C., and Sharma, N. L.: Impact of aerosol on surface reaching solar irradiance
427 over Mohal in the northwestern Himalaya, India, *Journal of Atmospheric and Solar-Terrestrial Physics*, 108, 41-49, 2014.
- 428 Hallquist, M., Wenger, J. C., Baltensperger, U., Rudich, Y., Simpson, D., Claeys, M., Dommen, J., Donahue, N. M., George, C., and
429 Goldstein, A. H.: The formation, properties and impact of secondary organic aerosol: current and emerging issues, *Atmospheric*



- 430 Chemistry and Physics, 9, 5155-5236, 2009.
- 431 Han, S., Kondo, Y., Oshima, N., Takegawa, N., Miyazaki, Y., Hu, M., Lin, P., Deng, Z. Z., Zhao, Y., and Sugimoto, N.: Temporal
432 variations of elemental carbon in Beijing, *Journal of Geophysical Research*, 114, 2009.
- 433 Hansen, J., Johnson, D. W., Lacis, A. A., Lebedeff, S., Lee, P., Rind, D., and Russell, G. L.: Climate impact of increasing atmospheric
434 carbon dioxide, *Science*, 213, 957-966, 1981.
- 435 Hansen, J., Sato, M., Ruedy, R., Nazarenko, L., Lacis, A., Schmidt, G. A., Russell, G., Aleinov, I., Bauer, M., Bauer, S., Bell, N.,
436 Cairns, B., Canuto, V., Chandler, M., Cheng, Y., Del Genio, A., Faluvegi, G., Fleming, E., Friend, A., Hall, T., Jackman, C., Kelley,
437 M., Kiang, N., Koch, D., Lean, J., Lerner, J., Lo, K., Menon, S., Miller, R., Minnis, P., Novakov, T., Oinas, V., Perlwitz, J., Perlwitz,
438 J., Rind, D., Romanou, A., Shindell, D., Stone, P., Sun, S., Tausnev, N., Thresher, D., Wielicki, B., Wong, T., Yao, M., and Zhang,
439 S.: Efficacy of climate forcings, *Journal of Geophysical Research: Atmospheres*, 110, 10.1029/2005jd005776, 2005.
- 440 Haywood, J. M., and Ramaswamy, V.: Global sensitivity studies of the direct radiative forcing due to anthropogenic sulfate and
441 black carbon aerosols, *Journal of Geophysical Research*, 103, 6043-6058, 1998.
- 442 Hu, K., Zhao, D., Liu, D., Ding, S., Tian, P., Yu, C., Zhou, W., Huang, M., and Ding, D.: Estimating radiative impacts of black
443 carbon associated with mixing state in the lower atmosphere over the northern North China Plain, *Chemosphere*, 252, 126455,
444 <https://doi.org/10.1016/j.chemosphere.2020.126455>, 2020.
- 445 Jacobson, M. Z.: Investigating cloud absorption effects: global absorption properties of black carbon, tar balls, and soil dust in
446 clouds and aerosols, *J. Geophys. Res.: Atmos.*, 117, 2012.
- 447 Ji, D., Yan, Y., Wang, Z., He, J., Liu, B., Sun, Y., Gao, M., Li, Y., Cao, W., Cui, Y., Hu, B., Xin, J., Wang, L., Liu, Z., Tang, G., and
448 Wang, Y.: Two-year continuous measurements of carbonaceous aerosols in urban Beijing, China: Temporal variations,
449 characteristics and source analyses, *Chemosphere*, 200, 191-200, <https://doi.org/10.1016/j.chemosphere.2018.02.067>, 2018.
- 450 Jorba, O., Perez, C., Rocadenbosch, F., and Baldasano, J. M.: Cluster analysis of 4 day back trajectories arriving in the Barcelona
451 Area (Spain) from 1997 to 2002, *Journal of Applied Meteorology*, 43, 887-901, 2004.
- 452 Koch, D., and Del Genio, A.: Black carbon semi-direct effects on cloud cover: review and synthesis, *Atmos. Chem. Phys.*, 10, 7685,
453 2010.
- 454 Laborde, M., Crippa, M., Tritscher, T., Juranyi, Z., Decarlo, P. F., Temimeroussel, B., Marchand, N., Eckhardt, S., Stohl, A., and
455 Baltensperger, U.: Black carbon physical properties and mixing state in the European megacity Paris, *Atmospheric Chemistry and
456 Physics*, 13, 5831-5856, 2012.
- 457 Liu, D., Flynn, M., Gysel, M., Targino, A. C., Crawford, I., Bower, K. N., Choularton, T. W., Juranyi, Z., Steinbacher, M., and
458 Hüglin, C.: Single particle characterization of black carbon aerosols at a tropospheric alpine site in Switzerland, *Atmospheric
459 Chemistry and Physics*, 10, 7389-7407, 2010.
- 460 Liu, D., Allan, J., Whitehead, J., Young, D., Flynn, M., Coe, H., McFiggans, G., Fleming, Z. L., and Bandy, B.: Ambient black
461 carbon particle hygroscopic properties controlled by mixing state and composition, *Atmospheric Chemistry and Physics*, 13, 2015-
462 2029, 10.5194/acp-13-2015-2013, 2013.
- 463 Liu, D., Allan, J. D., Young, D. E., Coe, H., Beddows, D., Fleming, Z. L., Flynn, M. J., Gallagher, M. W., Harrison, R. M., Lee, J.,
464 Prevot, A. S. H., Taylor, J. W., Yin, J., Williams, P. I., and Zotter, P.: Size distribution, mixing state and source apportionment of
465 black carbon aerosol in London during wintertime, *Atmospheric Chemistry and Physics*, 14, 10061-10084, 10.5194/acp-14-10061-
466 2014, 2014.
- 467 Liu, D., Whitehead, J. D., Alfarrá, M. R., Reyesvillegas, E., Spracklen, D. V., Reddington, C. L., Kong, S., Williams, P. I., Ting, Y.,
468 and Haslett, S. L.: Black-carbon absorption enhancement in the atmosphere determined by particle mixing state, *Nature Geoscience*,
469 10, 184-188, 2017.
- 470 Liu, D., Joshi, R., Wang, J., Yu, C., Allan, J. D., Coe, H., Flynn, M., Xie, C., Lee, J. D., and Squires, F.: Contrasting physical
471 properties of black carbon in urban Beijing between winter and summer, *Atmospheric Chemistry and Physics*, 19, 6749-6769, 2019a.
- 472 Liu, D., Zhao, D., Xie, Z., Yu, C., Chen, Y., Tian, P., Ding, S., Hu, K., Lowe, D., and Liu, Q.: Enhanced heating rate of black carbon
473 above the planetary boundary layer over megacities in summertime, *Environmental Research Letters*, 14, 124003, 2019b.
- 474 Liu, P., Zhao, C., Liu, P., Deng, Z., Huang, M., Ma, X., and Tie, X.: Aircraft study of aerosol vertical distributions over Beijing and
475 their optical properties, *Tellus Series B-chemical & Physical Meteorology*, 61, 756-767, 2009.
- 476 Liu, Y., Yan, C., and Zheng, M.: Source apportionment of black carbon during winter in Beijing, *Science of The Total Environment*,



- 477 618, 531-541, <https://doi.org/10.1016/j.scitotenv.2017.11.053>, 2018.
- 478 Liu, Z., Hu, B., Ji, D., Wang, Y., Wang, M., and Wang, Y.: Diurnal and seasonal variation of the PM_{2.5} apparent particle density in
479 Beijing, China, *Atmospheric Environment*, 120, 328-338, 2015.
- 480 Makra, L., Matyasovszky, I., Guba, Z., Karatzas, K., and Anttila, P.: Monitoring the long-range transport effects on urban PM₁₀
481 levels using 3D clusters of backward trajectories, *Atmospheric Environment*, 45, 2630-2641,
482 <https://doi.org/10.1016/j.atmosenv.2011.02.068>, 2011.
- 483 Markou, M. T., and Kassomenos, P.: Cluster analysis of five years of back trajectories arriving in Athens, Greece, *Atmospheric*
484 *Research*, 98, 438-457, <https://doi.org/10.1016/j.atmosres.2010.08.006>, 2010.
- 485 Mcfarquhar, G. M., and Wang, H.: Effects of aerosols on trade wind cumuli over the Indian Ocean: Model simulations, *Quarterly*
486 *Journal of the Royal Meteorological Society*, 132, 821-843, 2006.
- 487 Mcfiggans, G., Artaxo, P., Baltensperger, U., Coe, H., Facchini, M. C., Feingold, G., Fuzzi, S., Gysel, M., Laaksonen, A., and
488 Lohmann, U.: The effect of physical and chemical aerosol properties on warm cloud droplet activation, *Atmospheric Chemistry and*
489 *Physics*, 6, 2593-2649, 2005.
- 490 Nenes, A., and et al.: Black carbon radiative heating effects on cloud microphysics and implications for the aerosol indirect effect
491 2. Cloud microphysics, *J. Geophys. Res.: Atmos.*, 107, 4605, 2002.
- 492 Panicker, A. S., Pandithurai, G., Safai, P. D., and Prabha, T. V.: Indirect forcing of black carbon on clouds over northeast India,
493 *Quarterly Journal of the Royal Meteorological Society*, 142, 2968-2973, [10.1002/qj.2878](https://doi.org/10.1002/qj.2878), 2016.
- 494 Philipp, A.: Comparison of principal component and cluster analysis for classifying circulation pattern sequences for the European
495 domain, *Theoretical and Applied Climatology*, 96, 31-41, 2009.
- 496 Pringle, K. J., Tost, H., Pozzer, A., Pöschl, U., and Lelieveld, J.: Global distribution of the effective aerosol hygroscopicity parameter
497 for CCN activation, *Atmospheric Chemistry and Physics*, 10, 5241-5255, [10.5194/acp-10-5241-2010](https://doi.org/10.5194/acp-10-5241-2010), 2010.
- 498 Pruppacher, H. R., Klett, J. D., and Wang, P. K.: *Microphysics of Clouds and Precipitation*, *Aerosol Science and Technology*, 28,
499 381-382, [10.1080/02786829808965531](https://doi.org/10.1080/02786829808965531), 1998.
- 500 Ramanathan, V., Chung, C. E., Kim, D., Bettge, T. W., Buja, L., Kiehl, J. T., Washington, W. M., Fu, Q., Sikka, D., and Wild, M.:
501 Atmospheric brown clouds: Impacts on South Asian climate and hydrological cycle, *Proceedings of the National Academy of*
502 *Sciences of the United States of America*, 102, 5326-5333, 2005.
- 503 Ramanathan, V., and Carmichael, G.: Global and regional climate changes due to black carbon, *Nat. Geosci.*, 1, 221, 2008.
- 504 Rudich, Y., Sagi, A., and Rosenfeld, D.: Influence of the Kuwait oil fires plume (1991) on the microphysical development of clouds,
505 *Journal of Geophysical Research*, 108, [10.1029/2003JD003472](https://doi.org/10.1029/2003JD003472), 2003.
- 506 Schwarz, J. P., Gao, R. S., Fahey, D. W., Thomson, D. S., Watts, L. A., Wilson, J. C., Reeves, J. M., Darbeheshti, M., Baumgardner,
507 D., and Kok, G. L.: Single-particle measurements of midlatitude black carbon and light-scattering aerosols from the boundary layer
508 to the lower stratosphere, *Journal of Geophysical Research*, 111, 2006.
- 509 Srivastava, A., Ram, K., Pant, P., Hegde, P., and Joshi, H.: Black carbon aerosols over Manora Peak in the Indian Himalayan foothills:
510 implications for climate forcing, *Environmental Research Letters*, 7, 014002, 2012.
- 511 Stokes, R. H., and Robinson, R. A.: Interactions in Aqueous Nonelectrolyte Solutions. I. Solute-Solvent Equilibria, *The Journal of*
512 *Physical Chemistry*, 70, 2126-2131, 1966.
- 513 Sullivan, P. P., Moeng, C., Stevens, B., Lenschow, D. H., and Mayor, S. D.: Structure of the Entrainment Zone Capping the
514 Convective Atmospheric Boundary Layer, *Journal of the Atmospheric Sciences*, 55, 3042-3064, 1998.
- 515 Taylor, J. W., Allan, J. D., Liu, D., Flynn, M., Weber, R., Zhang, X., Lefer, B. L., Grossberg, N., Flynn, J., and Coe, H.: Assessment
516 of the sensitivity of core / shell parameters derived using the single-particle soot photometer to density and refractive index,
517 *Atmospheric Measurement Techniques*, 8, 1701-1718, [10.5194/amt-8-1701-2015](https://doi.org/10.5194/amt-8-1701-2015), 2015.
- 518 Tian, P., Liu, D., Zhao, D., Yu, C., Liu, Q., Huang, M., Deng, Z., Ran, L., Wu, Y., Ding, S., Hu, K., Zhao, G., Zhao, C., and Ding,
519 D.: In situ vertical characteristics of optical properties and heating rates of aerosol over Beijing, *Atmospheric Chemistry and Physics*,
520 20, 2603-2622, [10.5194/acp-20-2603-2020](https://doi.org/10.5194/acp-20-2603-2020), 2020.
- 521 Volkamer, R., Jimenez, J. L., Martini, F. S., Dzepina, K., Zhang, Q., Salcedo, D., Molina, L. T., Worsnop, D. R., and Molina, M. J.:
522 Secondary organic aerosol formation from anthropogenic air pollution: Rapid and higher than expected, *Geophysical Research*
523 *Letters*, 33, 2006.



- 524 Weinzierl, B.: Radiatively-driven processes in forest fire and desert dust plumes, DLR Deutsches Zentrum für Luft- und Raumfahrt
525 e.V. - Forschungsberichte, 2008.
- 526 Wood, R., and Bretherton, C. S.: Boundary Layer Depth, Entrainment, and Decoupling in the Cloud-Capped Subtropical and
527 Tropical Marine Boundary Layer, *Journal of Climate*, 17, 3576-3588, 2004.
- 528 Wu, Y., Liu, D., Wang, J., Shen, F., Chen, Y., Cui, S., Ge, S., Wu, Y., Chen, M., and Ge, X.: Characterization of Size-Resolved
529 Hygroscopicity of Black Carbon-Containing Particle in Urban Environment, *Environmental Science & Technology*, 53, 14212-
530 14221, [10.1021/acs.est.9b05546](https://doi.org/10.1021/acs.est.9b05546), 2019.
- 531 Xu, W., Han, T., Du, W., Wang, Q., Chen, C., Zhao, J., Zhang, Y., Li, J., Fu, P., Wang, Z., Worsnop, D. R., and Sun, Y.: Effects of
532 Aqueous-Phase and Photochemical Processing on Secondary Organic Aerosol Formation and Evolution in Beijing, China,
533 *Environmental Science & Technology*, 51, 762-770, [10.1021/acs.est.6b04498](https://doi.org/10.1021/acs.est.6b04498), 2017.
- 534 Yang, X., Xu, J., Bi, F., Zhang, Z., Chen, Y., He, Y., Han, F., Zhi, G., Liu, S., and Meng, F.: Aircraft measurement over the Gulf of
535 Tonkin capturing aloft transport of biomass burning, *Atmospheric Environment*, 182, 41-50,
536 <https://doi.org/10.1016/j.atmosenv.2018.03.020>, 2018.
- 537 Zhang, Y., Zhang, Q., Cheng, Y., Su, H., Li, H., Li, M., Zhang, X., Ding, A., and He, K.: Amplification of light absorption of black
538 carbon associated with air pollution, *Atmospheric Chemistry and Physics*, 18, 9879-9896, 2018.
- 539 Zhao, D., Huang, M., Tian, P., He, H., Lowe, D., Zhou, W., Sheng, J., Wang, F., Bi, K., Kong, S., Yang, Y., Liu, Q., Liu, D., and
540 Ding, D.: Vertical characteristics of black carbon physical properties over Beijing region in warm and cold seasons, *Atmospheric*
541 *Environment*, 213, 296-310, <https://doi.org/10.1016/j.atmosenv.2019.06.007>, 2019.

542