



1 **Measurement report: Diurnal and temporal variations of sugar compounds in**
2 **suburban aerosols from the northern vicinity of Beijing, China: An influence of**
3 **biogenic and anthropogenic sources**

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23 **Abstract**

24 Sugar compounds (SCs) are major water-soluble constituents in atmospheric aerosols. In
25 this study, we investigated their molecular compositions and abundances in the northern receptor
26 site (Mangshan) of Beijing, China, to better understand the contributions from biogenic and
27 anthropogenic sources using a gas chromatography–mass spectrometry technique. The sampling
28 site receives anthropogenic air mass transported by southerly winds from Beijing, while relatively
29 clean air mass transported by northerly winds from the forest areas. Day- and nighttime variations
30 were analyzed for anhydrosugars, primary sugars, and sugar alcohols in autumn 2007.
31 Concentrations of overall SCs ranged from 30.8 to 875 ng m⁻³ (avg. 325 ng m⁻³), showing diurnal
32 variations with higher levels in daytime (351 ng m⁻³) than nighttime (276 ng m⁻³). Interestingly,
33 biomass burning (BB) tracers were more abundant in nighttime than daytime, while other SCs
34 showed different diurnal variations. Levoglucosan was found as a dominant sugar among the
35 observed SCs, indicating an intense influence of BB over Mangshan. The levels of fungal tracers
36 (arabitol and mannitol) were higher in daytime than nighttime, suggesting a significant transport of
37 fungal spores and microbes to the receptor site by atmospheric transport from Beijing area. The
38 plant emissions from Mangshan forest park significantly control the diurnal variations of glucose
39 and fructose. The pollen tracer (sucrose) showed a clear diurnal variation, peaking in daytime due to
40 higher ambient temperature and wind speed. We found that soil dust contributes to trehalose in
41 daytime while microorganisms were responsible to its emissions in nighttime. The meteorological
42 parameters (relative humidity, temperature and rainfall) significantly affect the concentrations and
43 diurnal variations of SCs. Positive matrix factorization analysis suggested that local BB and
44 bioaerosols transported from Beijing area were significant sources of SCs. (289)

45 **Keywords:** Anthropogenic bioaerosols, biomass burning, pollen tracer, fungal tracers, soil dust,
46 and microbial tracers

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48 1. Introduction

49 An increased economic growth in East Asian countries causes serious emissions of
50 anthropogenic gas and aerosols, and biomass burning (BB) products (Lin et al., 2014). The rapid
51 urbanization and population growth also contribute to enhance the emission of organic aerosols and
52 bioaerosols into the atmosphere, which comprised of a complex mixture of diverse molecules (Xu
53 et al., 2011). They play essential roles in global climate changes via the modification of radiative
54 forcing and cause a severe negative impact on human health (Fuzzi et al., 2007). Organic aerosols
55 contain water-soluble organic compounds (WSOCs), which can act as cloud condensation nuclei
56 (CCN) (Kanakidou et al., 2005). The BB significantly emits organic aerosols and gases into the
57 atmosphere, controlling the air quality levels (Kanaya et al. 2013; Streets et al., 2003; Sullivan et
58 al., 2008; Deshmukh et al., 2019a). BB is essentially a primary source of organic aerosols, affecting
59 the earth's radiative forcing by scattering or absorbing incident solar radiation (Kanakidou et al.,
60 2005). The BB tracers are subjected to long-range atmospheric transport, once they are emitted to
61 the atmosphere (Verma et al., 2015). Levoglucosan is a thermal decomposition product of cellulose
62 and hemicellulose, which is generally found as one of the major organic constituents in the BB
63 aerosols (Simoneit et al., 1999; 2002).

64 SCs are ubiquitous in the atmosphere from different geographical locations including urban,
65 forest, marine, and polar regions (Wang et al., 2006; Simoneit et al., 2004b; Burshtein et al., 2011;
66 Fu et al., 2010; Wan et al., 2017). SCs are emitted from algae, microbes, pollen, suspended soil
67 particle, and associated biota into the atmosphere (Carvalho et al., 2003). Winds play a vital role in
68 the distribution of microorganisms from their source regions to receptor sites (Brown and
69 Hovmoller, 2002). Thus, the transport of bioaerosols has been emphasized in the past studies
70 (Yamaguchi et al., 2012). Fungi are essential microorganisms in the ecosystem, which mostly emit
71 spores to the environment (Elbert et al., 2007). Fungal spores can travel long distances in the
72 atmosphere (Burshtein et al., 2011). Sugar alcohols like arabitol and mannitol are emitted via the
73 metabolism of microorganisms and fungi, thus they have been proposed as specific tracers for
74 microbial and fungal activities (Bauer et al., 2008; Simoneit et al., 2004b). Bacteria, fungi,
75 invertebrates, and lower plants also emit a significant amount of trehalose (Bielecki, 1995). There
76 are several sources of sugars in the atmosphere, including higher plants and soil dust (Jaenicke,
77 2005). Higher plants synthesize sugars during photosynthesis, which are circulated by phloem to
78 accumulate in root cells and developing plant sections (Pacini, 2000). Various sugars are present in
79 terrestrial plant fruits, flowers, and plant tissues (Cowie et al., 1984).

80 Previous studies analyzed aerosol samples for sugar compounds (SCs) and discussed several



81 factors to control their levels in the local and global atmosphere. Recently, Xu et al. (2020)
82 examined the seasonal molecular distributions of primary biological aerosols and biomass burning
83 aerosol samples collected from urban Beijing. They reported high levels of arabinol, mannitol,
84 sucrose, glucose, fructose in vegetation-growing season. Kang et al. (2018) also reported higher
85 concentrations of sugars in the urban aerosols from Beijing and suggested a large contribution of
86 coal combustion and agriculture residue burning under stable meteorological conditions in winter
87 and spring. Verma et al. (2015, 2018) reported thirteen years (2010-2013) observation on the
88 remote marine aerosols from Chichijima Island. They concluded that long-range transport of
89 organic-/bio-aerosols from East Asia significantly control the levels and compositions of SCs over
90 the North Pacific. The above studies discussed the several factors that affect the concentrations of
91 SCs in the aerosol samples collected from urban and remote areas. Xu et al. (2020) reported
92 multiple sources for arabinol, mannitol, glucose, sucrose and trehalose. Verma et al. (2015, 2018)
93 discussed the atmospheric circulations that significantly affect the seasonal concentrations of SCs.

94 In this study, we conducted the analyses of SCs in the aerosol samples collected from the
95 northern suburb of Beijing. Beijing is one of largest polluted cities in East Asia. Its air quality
96 deteriorates seriously due to rapid economic growth, resulting in massive emissions from vehicles
97 and industries (Cao et al., 2014; Qiao et al., 2018; Tao et al., 2017; Wei et al., 2018; Yu et al.,
98 2013). Here, we present comprehensive data sets of anhydrosugars, primary sugars, and sugar
99 alcohols in the suburban aerosol samples and discuss their diurnal variations to explain the source
100 variance following the wind patterns in day and nighttime. The positive matrix factorization (PMF)
101 has been applied to the measured SCs in the aerosol samples to clarify the influence of local
102 meteorology of sampling site and atmospheric transport from Beijing by southerly winds and
103 Mangshan National Forest Park by northerly winds on the molecular distributions of SCs. Using the
104 mass concentration ratio of levoglucosan to mannosan, we discuss the relative contribution of hard
105 and softwood burning to the air quality of Mangshan. This study also discussed carbon
106 contributions of SCs measured in the Mangshan aerosol samples from different sources.

107 **2. Materials and Methods**

108 **2.1. Site description**

109 The aerosol samples were collected near the entrance of Mangshan National Forest Park
110 (40.28 N, 116.26 E; elevation of 187 m above sea level) from 15th September to 5th October 2007
111 using a high-volume air sampler (Kimoto-AS810A) (He et al. 2014, 2015). The Mangshan site is
112 located at 40 km north of Beijing and is surrounded by urban area in the south and forest areas with



113 the national park in the north (Fig. 1). The ambient temperature was higher in daytime (23.9 °C)
114 than nighttime (12.1 °C) with an average of 17.8 °C during the campaign. The relative humidity
115 (RH) varied significantly from 22.1% to 90.5% with an average of 51.7 % during the study period.
116 Interestingly, a wind pattern was characterized at the sampling site. The southwest wind (69.9%)
117 prevailed during the day, followed by northeast wind (23.4%) and southeast wind (6.2%). The
118 northeast wind (99.5%) was dominated at night, which is consistent with the air mass back
119 trajectories (He et al., 2014). The daytime wind from the southwest direction passed over Beijing,
120 delivering anthropogenic air mass to Mangshan site. The northeast wind carried clean air mass from
121 the forest area at night. The rainfall was observed at the midnight on 15th September, morning of
122 17th to evening 18th September, night of 26th September, and light rain lasted from October 4 to the
123 end of the campaign (Fig. 2).

124 2.2. Sampling details and analytical methodologies

125 We collected 3-h daytime (n=26), 9-h daytime (n=12), and 15-h nighttime (n=20) samples
126 together with four field blanks. After the sampling, filters were individually stored in a glass bottle
127 with a Teflon-lined cap at -20 °C before analysis. Total of 58 aerosol samples were analyzed for
128 anhydrosugars, primary sugars, and sugar alcohols (Table 1). To derivatize hydroxyl (OH) and
129 carboxyl (COOH) groups to trimethylsilyl (TMS) ethers and esters, respectively, the extracts were
130 reacted with 60 µl of N,O-bis-(trimethylsilyl)trifluoroacetamide (BSTFA) with 1% trimethylsilyl
131 (TMS) chloride in the presence of 10 µL of pyridine at 70 °C for three hours. After the reaction, n-
132 hexane was used for dilution, and C₁₃ n-alkane was added as an internal standard before GC-MS
133 analysis. Agilent model 6890 gas chromatograph (GC) combined to Agilent model 5973 mass
134 selective detector (MSD) was used to analyze SCs. The detailed description of GC-MS operation
135 and identification of SCs can be found in Verma et al. (2015, 2018).

136 3. Results and Discussion

137 3.1. Ambient concentrations and diurnal variations of SCs

138 We detected totally ten SCs including three anhydrosugars, four primary sugars and three
139 sugar alcohols in the Mangshan aerosol samples. Figure 3a-j showed the temporal variations of SCs
140 in the Mangshan aerosol samples. Figure 4a-c and Table 1 show minimum, maximum and average
141 concentrations of anhydrosugars, primary sugars and sugar alcohols with a standard deviation. The
142 overall concentrations of SCs varied from 30.8–875 ng m⁻³ (avg. 325 ng m⁻³). They showed distinct
143 diurnal variations with daytime high (34.1–875 ng m⁻³; avg. 351 ng m⁻³) and nighttime low values



144 (30.8 – 759 ng m⁻³; avg. 276 ng m⁻³). Interestingly, higher average concentrations of SCs were
145 reported at Mt. Tai (daytime 640 ng m⁻³ and nighttime 799 ng m⁻³) in the North China Plain (Fu et
146 al., 2008) than those of the Mangshan site from this study.

147 The diurnal variations of SCs may be significantly influenced by the delivery of anthropogenic
148 air masses from megacities such as Beijing, Hebei, and Tianjin, where contaminants are heavily
149 generated by fossil fuel combustion and biomass burning. Anthropogenic aerosols emitted from
150 urban areas are probably transported to the northern receptor site in daytime by a southerly wind.
151 Therefore, the high levels of SCs in daytime may be related with the transport of organic and bio-
152 aerosols from urban regions. In addition, at nighttime the wind direction is shifted to northerly,
153 delivering comparatively clean air masses from the Mangshan National Forest area to the sampling
154 site. Air mass from the forest may significantly contribute nighttime SCs in the Mangshan site.
155 Although transported air mass may have mainly controlled the levels of SCs, contributions of local
156 SCs cannot be excluded.

157 3.1.1. Ambient concentrations and diurnal variations of anhydrosugars

158 Anhydrosugars are important BB tracers because they are emitted from the pyrolysis of
159 cellulose and hemicellulose (Simoneit et al., 1999). Concentrations of anhydrosugars ranged from
160 6.01 to 556 ng m⁻³ (avg. 116 ng m⁻³), contributing 31.9% of overall SCs in the Mangshan aerosols
161 (Table 1). Figure 3a-c shows the temporal variations of anhydrosugars. They are more abundant in
162 nighttime (avg. 152 ng m⁻³) than daytime (avg. 97.1 ng m⁻³). Levoglucosan is most abundant
163 anhydrosugar detected in Mangshan aerosols, whose concentrations are higher in nighttime (5.66–
164 482 ng m⁻³, avg. 132 ng m⁻³) than daytime (1.17–418 ng m⁻³, avg. 83.2 ng m⁻³) (Table 1).
165 Concentrations of galactosan are higher in nighttime (0.69–48.0 ng m⁻³, avg. 13.0 ng m⁻³) than
166 daytime (0.14–45.3 ng m⁻³, avg. 8.53 ng m⁻³). Similarly, mannosan is more abundant in nighttime
167 (0.53–26.1 ng m⁻³, avg. 7.35 ng m⁻³) than daytime (0.13–24.3 ng m⁻³, avg. 5.37 ng m⁻³). Levels of
168 all anhydrosugars in Manshang are lower than those reported from Mt Tai, China (daytime, 391 ng
169 m⁻³ and nighttime, 459 ng m⁻³) (Fu et al., 2008).

170 Biomass burning (BB) tracers showed significant positive correlations each other, i.e.,
171 levoglucosan and galactosan ($r = 0.98$), levoglucosan and mannosan ($r = 0.97$), and galactosan and
172 mannosan ($r = 0.98$), suggesting their similar sources in the Mangshan aerosols (Fig. 5a-c). Figures
173 2 and 3 clearly show that concentrations of BB tracers decreased with increasing ambient
174 temperature. In general BB is common for cooking and heating in East Asian countries (Verma et



175 al., 2015). Hence, it is evident that increased BB activities at nighttime are associated with cool
176 temperature (Fig. 2). The nighttime samples were collected from 18:00h to 09:00h, including peak
177 hours of BB for cooking and space heating. Therefore, it is reasonable to detect higher abundances
178 of BB tracers in nighttime than daytime in the Mangshan site. Warmer ambient temperature was
179 recorded in daytime between 09h to 15h during the campaign, which is associated with declined BB
180 activities. Kang et al. (2018) reported high levels of levoglucosan (avg. 110 ng m^{-3}) in autumn
181 aerosols from Beijing, China. We found a significant enhancement of BB tracers in nighttime,
182 which may be emitted mainly by local BB. However, transport from the North China Plain to the
183 Mangshan site by northerly winds cannot be excluded (Fig. 4).

184 3.1.2. Ambient concentrations and diurnal variations of primary sugars

185 The fragment of vascular plants contains primary sugars (Medeiros et al., 2006).
186 Concentrations of primary sugars ranged from $9.41\text{--}565 \text{ ng m}^{-3}$ (avg. 133 ng m^{-3}) in Mangshan
187 aerosols. Primary sugars were found as the most abundant sugars, contributing to 41.8% of the SCs
188 (Table 1). They showed apparent diurnal variations with daytime high (avg. 166 ng m^{-3}) and
189 nighttime low values (avg. 69.4 ng m^{-3}) (Figs. 3g-j, 4a-c). Graham et al. (2003) also reported
190 similar diurnal variations of primary sugars for the Amazon forest aerosols.

191 Sucrose was found as dominant sugar among primary sugars with a concentration range of
192 $0.02\text{--}474 \text{ ng m}^{-3}$ (avg. 58.5 ng m^{-3}), accounting for 44% of measured primary sugars in Mangshan
193 aerosols. Pollen was reported as dominant sources for sucrose in aerosols collected from a rural site
194 in Texas (Jia et al., 2010). Fu et al. (2012) found high sucrose concentrations up to 1390 ng m^{-3} in
195 the aerosols from Jeju Island, South Korea. We found a significant diurnal variation of sucrose in
196 Mangshan aerosols with higher levels (82.9 ng m^{-3}) in daytime than nighttime (12.3 ng m^{-3}).
197 Interestingly, significantly lower concentrations of sucrose than the Mangshan site were reported in
198 urban aerosols from Nanjing, China, in summer (daytime, 13.2 ng m^{-3} ; nighttime, 9.44 ng m^{-3}) and
199 winter (daytime, 14.5 ng m^{-3} ; nighttime, 10.9 ng m^{-3}) (Wang and Kawamura, 2005). In addition,
200 much lower concentration of sucrose was reported for the aerosols from Mt. Hua (daytime, 3.1 ng
201 m^{-3} ; nighttime, 2.6 ng m^{-3}) in central China (Li et al., 2012). These comparisons indicate that high
202 levels of sucrose are found in forest aerosols rather than urban and anthropogenic aerosols.

203 Meteorological parameters such as temperature, rainfall, wind speed, and solar radiation, are
204 associated with the concentration levels of sucrose in the forest site. The pollen activities are
205 significantly affected by the local meteorology (Verma et al., 2018). Interestingly, Figure 3d shows



206 a high peak of sucrose from 02h to 15h with higher ambient temperature, whereas lower
207 concentrations of sucrose were observed from 15h to 09h with lower ambient temperature. High
208 daytime concentrations of sucrose might be related to the increased daytime ambient temperature,
209 low RH, and high solar radiation. Wickmen (1994) reported that the meteorological conditions
210 resulted in the catapulting pollen into the air and the opening of pollen-laden flower anthers, under
211 the influence of strong daytime winds and convective activity, causing enhanced entrainment and
212 dispersal of the particles. Pacini (2000) also reported that higher levels of sucrose in daytime
213 coincide with higher counts of pollen, fern spore, and insect. The significant correlations of sucrose
214 with ambient temperature ($r = 0.52$) and solar radiation ($r = 0.55$) further supported the association
215 of meteorological parameters with the sucrose concentrations of Mangshan aerosols (Fig. 5d, e).

216 Five rain events were recorded during the campaign, i.e., 15th, 17th, 18th and 26th September,
217 and 1st and 5th October (Fig. 3). Pollens are significantly settled down by wet scavenging during
218 rain events because their sizes are large. Low sucrose concentration was found from the beginning
219 of sampling to the morning of 20th September and from the afternoon of 26th September to the end
220 of the sampling campaign (Fig. 3d). In addition, the increased concentrations of sucrose were found
221 in the aerosol samples collected from 20th to 22nd September and moderate concentrations observed
222 after 23rd to the evening of 25th September during non-precipitation events. Consequently, the
223 pollens might be significantly scavenged from the atmosphere during wet precipitation (washout
224 effect). The wet scavenging resulted in lower sucrose concentrations at the earlier periods, while
225 high sucrose concentrations were found during later periods.

226 The concentrations of glucose ranged from 1.86 to 297 ng m⁻³ (avg. 40.0 ng m⁻³). Similarly,
227 a wide range of fructose (1.72-117 ng m⁻³, avg. 20.1 ng m⁻³) was observed in Mangshan aerosols
228 (Table 1). The levels of glucose are equivalent to that (50.1 ng m⁻³) reported from the Howland
229 Experimental Forest site in USA (Medeiros et al., 2006). The high concentrations of these sugars
230 suggested that they might be associated with pollen/fern spores, insects, and other plant materials
231 from the background forest region. The nectars and fruits of subtropical and tropical plants are
232 reported as key sources for these primary sugars (Baker et al., 1998). Several studies have reported
233 lichens and soil dust as major sources of both primary sugars (Dahlman et al., 2003; Nolte et al.,
234 2001). The local vegetation and autumn decay of vascular plant leaves in the Mangshan forest
235 might have contributed to high levels of these primary sugars.

236 Glucose and fructose showed significant diurnal variations, whose concentrations were
237 higher in daytime (44.2 ng m⁻³ and 23.9 ng m⁻³, respectively) than nighttime (32.0 ng m⁻³ and 12.8



238 ng m^{-3} , respectively) in Mangshan aerosols (Fig. 3e, f; 4b, c; Table 1). Similar diurnal variations but
239 lower concentrations were reported in Mt. Hua aerosols collected from the Guanzhong Plain,
240 central China in January (daytime 6.2 ng m^{-3} and 6.3 ng m^{-3} , respectively; nighttime 4.8 ng m^{-3} and
241 4.5 ng m^{-3} , respectively), during a typical burning season of biofuel and coal for house heating (Li
242 et al., 2012). Similar trends of glucose and fructose were reported in urban Nanjing aerosols
243 (daytime 16.9 ng m^{-3} and 18.9 ng m^{-3} , respectively, and nighttime 21.6 ng m^{-3} and 30.9 ng m^{-3} ,
244 respectively) (Wang and Kawamura, 2005). On the other hand, higher concentrations with different
245 diurnal variations (daytime 49.5 ng m^{-3} and 16.7 ng m^{-3} , respectively; nighttime, 74.7 ng m^{-3} and
246 21.3 ng m^{-3} , respectively) have been reported for Mt. Tai aerosols collected from the North China
247 Plain (Fu et al., 2008). Interestingly, both primary sugars were significantly enriched in forest
248 aerosol samples compared to urban aerosols.

249 The vegetation and pollen/fern spores from the local forest area of Mangshan significantly
250 contributed to the high levels of glucose and fructose. The Mangshan site is located at the northern
251 receptor of anthropogenic air mass from megacities by southerly winds. However, the southerly
252 winds also cover the major area of the vegetation region, including the northeast direction (23.4%)
253 in daytime (He et al., 2014), which may transport the primary sugars (glucose and fructose)
254 abundantly to the Mangshan site. Although the wind direction shifted to northeasterly (99.5%) in
255 nighttime, sugar emissions from the vegetation sources may have decreased due to lower ambient
256 temperature, which may be a possible reason for low levels of glucose and fructose in nighttime
257 aerosols at Mangshan site.

258 Trehalose in the environment is significantly controlled by the activities of bacteria, fungal
259 spores, yeast, algae, invertebrates, and plant species as well as suspended soil particles (Verma et
260 al., 2018). The concentrations of trehalose ranged from 0.06 to 39.5 ng m^{-3} (avg. 14.3 ng m^{-3} , see
261 Table 1), which are lower than those of TSP samples collected from urban Guangzhou during
262 summer (67.8 ng m^{-3}) and autumn (42.7 ng m^{-3}) (Ma et al., 2009). Interestingly, higher trehalose
263 concentrations have been reported in urban (29 ng m^{-3}) and suburban (27 ng m^{-3} and 30 ng m^{-3})
264 aerosol samples than rural aerosols in Norway (3.8 ng m^{-3}) (Yttri et al., 2007). The above results
265 emphasizes that fungi and microbes associated with anthropogenic and bioaerosols, which are
266 vigorously emitted in the urban and suburban areas, may be responsible for the high levels of
267 trehalose (Verma et al., 2018).



268 Trehalose showed insignificant diurnal variation, whose concentrations ranged from 0.06–
269 34.4 ng m⁻³ (avg. 15.3 ng m⁻³) in daytime and 0.87–39.5 ng m⁻³ (avg. 12.3 ng m⁻³) in nighttime
270 (Fig. 4b, c). The similar levels of trehalose in day- and nighttime indicated its multiple sources in
271 the Mangshan aerosols. The southerly winds might transport anthropogenic aerosols associated with
272 microorganisms, which emit trehalose at nighttime due to low temperature and high RH.
273 Interestingly, the nighttime concentration of trehalose was significantly covaried with arabitol and
274 mannitol ($r = 0.76$ and 0.85 , respectively) (Fig. 5f, g), suggesting that fungal and microbial species
275 contributed to trehalose in nighttime. The metabolic activities of microorganisms increase in
276 autumn with favorable metrological conditions to emit trehalose (Verma et al. 2018). In addition,
277 Hackl et al. (2000) found trehalose as dominant sugar in spring aerosols and proposed it as a tracer
278 for soil dust particles. Daytime wind transports soil particles from the urban area due to the active
279 building constructions. The transported soil particles may contribute to the high levels of trehalose
280 in daytime at the receptor site of Mangshan. The trehalose concentration was significantly
281 correlated with Ca ($r = 0.82$) in daytime, indicating a significant contribution of soil dust to the site
282 (Figs. 3g, 5h).

283 3.1.3. Ambient concentrations and diurnal variations of sugar alcohols

284 The concentrations of sugar alcohols ranged from 8.53–259 ng m⁻³ (avg. 75.8 ng m⁻³),
285 contributing 26.4% of total SCs in Mangshan aerosols (Table 1). Sugar alcohols showed distinct
286 diurnal variations, with daytime high (avg. 87.4 ng m⁻³) and nighttime low (avg. 53.7 ng m⁻³)
287 concentrations. The sugar alcohols are emitted by the metabolic activities of fungi and microbes
288 into the atmosphere (Filippo et al., 2013). The mannitol and arabitol are key components of fungal
289 spores, and thus they are termed as fungal tracers (Bielecki, 1995). Mannitol was found as a
290 dominant sugar alcohol in the Mangshan aerosols, whose concentrations ranged from 4.19–182 ng
291 m⁻³ (avg. 44.1 ng m⁻³) followed by arabitol (3.89–72.2 ng m⁻³, avg. 29.1 ng m⁻³). The inositol was
292 found as a minor sugar alcohol with an average concentration of 2.62 ng m⁻³. During precipitation
293 events, concentrations of arabitol, mannitol, and inositol decreased (Fig. 3h-j). Interestingly, their
294 concentrations increased gradually after the end of precipitation, following the increases in ambient
295 temperature and RH (Figs. 2, 3h-j).

296 Significantly low levels of mannitol and arabitol were reported for remote sites, including
297 the North Pacific (0.01-0.23 ng m⁻³ and 0.01-0.19 ng m⁻³) and North Atlantic (0.01-0.06 ng m⁻³ and
298 0.006-0.07 ng m⁻³) (Fu et al., 2011). These remote locations are far away from the continent and
299 anthropogenic activities. The moderate levels of mannitol and arabitol compared to the Mangshan



300 site were reported at Gosan, Jeju Island (11 ng m^{-3} and 12 ng m^{-3} , respectively) (Fu et al., 2012),
301 Cape Hedo, Okinawa Island (13.0 ng m^{-3} and 12.3 ng m^{-3} , respectively) (Zhu et al., 2015) and
302 Chichijima Island (18.2 ng m^{-3} and 15.8 ng m^{-3} , respectively) (Verma et al., 2018) in autumn. These
303 islands are situated in the outflow region of East Asia with the emissions of bio- and anthropogenic
304 aerosols from megacities. These continental sources might affect the levels of trehalose, arabitol,
305 and mannitol over the remote ocean (Verma et al., 2018).

306 Mannitol and arabitol showed a strong correlation ($r = 0.81$) (Fig. 5i), suggesting similar
307 sources of both sugar alcohols in the Mangshan site. The significant diurnal co-variance of mannitol
308 and arabitol is characterized by higher concentrations in daytime (51.7 ng m^{-3} and 32.5 ng m^{-3} ,
309 respectively) than nighttime (29.6 ng m^{-3} and 22.5 ng m^{-3} , respectively) (Fig. 4b, c). The lower
310 concentrations than Mangshan were reported for aerosol samples from Mt. Tai in the North China
311 Plain in daytime (77.8 ng m^{-3} and 52.5 ng m^{-3} , respectively) and nighttime (83.9 ng m^{-3} and 56.4 ng
312 m^{-3} , respectively) (Fu et al., 2008). Additionally, the diurnal variations of fungal tracers may be
313 significantly influenced by local meteorological parameters. RH well correlated with mannitol ($r =$
314 0.57) and arabitol ($r = 0.69$) (Fig. 5j, k). The temporal trends of arabitol, mannitol and inositol
315 correlated with RH. Therefore, we hypothesized that increased daytime concentrations of arabitol
316 and mannitol (Fig. 3h, i) are associated with increased emissions of yeast and fungal spores that are
317 transported from megacities to the receptor site. On the other hand, lower concentrations can be
318 explained by clean air mass delivered by mountain breeze from the Mangshan National Forest area
319 in nighttime.

320 3.2. Positive matrix factorization (PMF) analysis

321 To investigate the source apportionment of SCs, positive matrix factorization (PMF)
322 analysis was performed for the studied aerosol samples using tracer compounds for anhydrosugars,
323 primary sugars and sugar alcohols. The PMF analysis resulted five interpretable factors, which
324 reproduced more than 95% of SCs. Paatero et al. (2002) and Zhou et al. (2004) discussed the
325 determination and applications of PMF in details. Figures 6 and 7 show the factor profile resolved
326 by PMF analysis of the Mangshan aerosol samples. The percentages of each component are
327 summed for factors 1 to 5 to be calculated as 100%.

328 Factor 1 is characterized by high contribution of glucose (80.2%) followed by fructose
329 (69.6%), mannitol (24.8%) and inositol (15.1%) (Fig. 6a). Glucose and fructose are highly water-
330 soluble SCs that are present in the leaves and bark of plants (Graham et al., 2003). High



331 concentrations of glucose and fructose have been reported in vascular plants and phytoplankton by
332 Cowie and Hegdes (1984). The dominant glucose and fructose in the Mangshan aerosol samples
333 collected in autumn are rational as leaf senescence and decay result both primary sugars being
334 released into the atmosphere during the fall season. We found an excellent correlation between
335 glucose and fructose ($r = 0.94$) in the Mangshan aerosols (Fig. 5l), indicating the similar vegetation
336 sources for both sugar species in autumn (Baker et al., 1998; Burshtein et al., 2011; Pacini, 2000).
337 The SCs emitted from vegetation during the campaign significantly contributed to factor 1.
338 Therefore, factor 1 can be termed as a vegetation factor due to the high abundances of both sugar
339 species.

340 Factor 2 is dominated by high loading of trehalose (80.2%), followed by mannitol (29.7%),
341 galactosan (19.8%), and arabitol (18.2%) (Fig. 6b). The contribution of trehalose to soil dust has
342 been reported in several studies from different locations around the world, suggesting trehalose as a
343 tracer for the surface soil (Jia et al., 2010; Medeiros et al., 2006). In addition, previous studies
344 reported that bacteria and other microbes in the soil are also an important source for trehalose
345 (Rogge et al., 2007). The significant contributions of arabitol and mannitol indicated microbial
346 sources of trehalose for factor 2. Trehalose is significantly correlated with fungal spore tracers such
347 as arabitol ($r=0.58$) and mannitol ($r=0.58$), and Ca ($r=0.70$) (Fig. 5m-o); the latter is a primary soil
348 dust tracer. The above discussions suggest multiple sources of trehalose. Therefore, factor 2 can be
349 termed as microbial and soil dust factor.

350 Factor 3 is characterized by levoglucosan (82.2%), galactosan (77%), and mannosan
351 (73.6%) (Fig. 6c). Previous studies have reported that these sugars are associated with BB aerosols
352 (Simoneit, 2002). The PMF results are very well supported by the fact that anhydrosugars are
353 associated with BB in the Mangshan site. Factor 4 is dominated by high loading of sucrose (90%),
354 followed by inositol (36.9%) and fructose (11.7%) (Fig. 6d). Sucrose, fructose and inositol are
355 associated with airborne pollen spores so that factor 4 is termed as pollen factor (Pacini, 2000;
356 Verma et al., 2018). Sucrose plays a crucial role in the plant blossoming process because it is the
357 dominant sugar compound of pollen grains (Pacini, 2000).

358 Factor 5 is characterized by a higher contribution of arabitol (61.5%) than mannitol (39.3%)
359 and inositol (15.3%) (Fig. 6e). Sugar species that contribute to factor 5 are associated with fungal
360 spores (Bauer et al., 2008). Arabitol and mannitol are considered as suitable tracers for fungal
361 activities (Medeiros et al., 2006). Therefore, factor 5 is termed as a fungal factor due to the high
362 loading of arabitol and mannitol. Overall, the average contributions of each factor to measured SCs
363 were estimated by PMF analyses (Fig. 7), in which BB was found to account 27% of measured SCs.



364 The vegetation and microbial plus soil dust sources equally contribute (21%) to total SCs. The
365 fungal spores and pollen spores contribute 16% and 15% of total SCs, respectively. Finally,
366 biomass burning emissions from the local areas and megacities via long-range atmospheric
367 transport were identified as a severe source for the Mangshan aerosols.

368 3.3. Contributions of sugar compounds to WSOC and OC

369 The mean carbon content of measured sugar compounds varied from 14.1-371 ng m⁻³ (av.
370 145 ng m⁻³) in daytime and 12.8-322 ng m⁻³ (av. 117 ng m⁻³) in nighttime, accounting for 0.83% and
371 0.91% of OC, respectively (Fig. 8a, b). The mean carbon contents of anhydrosugars showed clear
372 diurnal variation with higher values in nighttime (67.1 ng m⁻³) than daytime (42.7 ng m⁻³),
373 accounting for 0.43 % and 0.22 % of OC, respectively. These results suggest that biomass burning
374 significantly contributed to Mangshan aerosols. However, the carbon contents of primary sugars
375 showed opposite diurnal variations; higher (68.5 ng m⁻³) in daytime than nighttime (28.3 ng m⁻³)
376 accounting for 0.41 % and 0.28 % of OC, respectively. (Fig. 8a, b). This study suggests that the
377 daytime emissions of primary sugars from local vegetation and the decay of plant leaf in forest
378 significantly contribute to OC. We found lower levels of sugar alcohols in daytime (34.6 ng m⁻³)
379 and nighttime (21.3 ng m⁻³), accounting for 0.20 % and 0.19 % of OC, respectively. This result
380 indicates multiple carbon sources of sugar alcohols in day and night. In addition, contributions of
381 anhydrosugars, primary sugars, and sugar alcohols to WSOC were similar to those of OC in
382 Mangshan aerosols.

383 Based on the PMF analysis, we found five sources for SCs measured in Mangshan aerosols.
384 The following tracer compounds were used to calculate carbon contents: biomass burning (i.e.,
385 levoglucosan, galactosan, mannosan; BB-C), vegetation (glucose, fructose; VG-C), fungi (arabitol,
386 mannitol; FG-C), pollen (sucrose; PL-C) and microbial/soil (trehalose; MbS-C) (Fig. 8c, d). Among
387 the five sources, BB-C was found as the largest contributor to Mangshan aerosols (36.7%) followed
388 by FG-C (23.7%), VG-C (19.7%), PL-C (14.2%), and MbS-C (4.84%). In addition, BB in nighttime
389 accounted for 1.38% and 0.43% of WSOC and OC, respectively (Fig. 8c, d), although the levels are
390 lower in daytime (0.57% and 0.22%, respectively). The BB for cooking and space heating are very
391 common in central China (Verma et al., 2015), which should increase the nighttime levels of BB-C
392 at the Mangshan site. However, the emissions from VG-C and FG-C sources are similar during day
393 and nighttime for the Mangshan aerosols. PL-C accounted for 0.20 % and 0.07 % of OC in daytime
394 and nighttime, respectively. Higher pollen activities are key sources for the daytime high levels of
395 PL-C in the forest site (Verma et al., 2018).



396 3.4. Contribution of levoglucosan to OC and WSOC

397 We calculated the mass concentration ratio of levoglucosan to OC (Lev/OC) and WSOC
398 (Lev/WSOC) to evaluate the contribution of BB and anthropogenic emissions to Mangshan
399 aerosols. Fossil fuel combustion and BB primarily emit WSOC and OC. They are also secondarily
400 produced by photochemical oxidation of volatile organic compounds in the atmosphere (Wang et
401 al., 2005; Huang et al., 2006; Deshmukh et al., 2019b). The coal combustion and vehicle exhaust
402 can highly contribute to the levels of OC and WSOC in aerosols (Xu et al., 2020). Levoglucosan, a
403 dominant constituent of BB products has been considered as an excellent tracer of BB (Kuo et al.,
404 2011).

405 Lev/OC average ratio (5.69×10^{-3}) was lower than Lev/WSOC (1.66×10^{-2}) in Mangshan
406 samples (Fig. 9a). Interestingly, we found a strong diurnal variation of Lev/OC and Lev/WSOC
407 ratios in the nighttime samples. The average ratios are higher in nighttime (Lev/OC: 8.48×10^{-3} ,
408 Lev/WSOC: 2.70×10^{-2}) than daytime (Lev/OC: 4.21×10^{-3} , Lev/WSOC: 1.11×10^{-2}) (Fig. 9b, c).
409 These results indicate that BB contributed substantially to organic carbon in the nighttime aerosols.
410 Our results also denote that secondary production of OC and WSOC from BB-derived organic
411 precursors was crucial during nighttime at the Mangshan site. We presume that the nighttime
412 contribution of BB in the site is due to the cooking and space heating in central China at night.
413 Moreover, the correlations of levoglucosan with OC and WSOC are stronger in nighttime ($r = 0.81$
414 and 0.70 , respectively) than daytime ($r = 0.45$ and 0.40 , respectively), indicating the dominance of
415 BB-derived organic carbon during nighttime in the Mangshan aerosol samples.

416 In addition, water-soluble organic compounds are derived from various emission sources.
417 We propose that secondary organic aerosols can constitute a significant fraction of WSOC and OC
418 in daytime Mangshan aerosols. The photochemical oxidation of organic precursors emitted by fossil
419 fuel combustion of industries and vehicular exhausts also contributes to WSOC and OC in daytime
420 (He et al., 2015), suggesting that emissions from the urban Beijing area may significantly influence
421 the daytime levels of Mangshan aerosols. He et al. (2015) suggested a possible contribution of
422 photochemical formation of secondary organic aerosols to atmospheric WSOC and OC in central
423 China. Nevertheless, the photochemical degradation of levoglucosan by OH radicals under
424 ultraviolet radiations and high temperatures (Hennigan et al., 2010) might play a key role for the
425 lower ratios of Lev/OC and Lev/WSOC in daytime than nighttime in the Mangshan aerosols.

426 3.5. Mass concentration ratios of levoglucosan/mannosan

427 The mass concentration ratios of levoglucosan and mannosan (Lev/Man) were calculated to
428 better characterize the emissions sources of BB tracers (softwood vs. hardwood) in the Mangshan



429 site. Figure 10 represents the variations of Lev/Man ratios for overall, day- and nighttime periods.
430 The Lev/Man ratios have been used to distinguish the burning of hardwood (angiosperm) and
431 softwood (gymnosperm) in the ice core record from the Russian Far East (Kawamura et al., 2012).
432 Hardwood contains 55–65% of cellulose and 20–30% hemicellulose (Klemm et al., 2005).
433 Levoglucosan and mannosan are derived from thermal decomposition of cellulose and
434 hemicelluloses, respectively (Simoneit, 2002). Levoglucosan is thermally more stable than
435 mannosan and galactosan (Kuo et al., 2011). Hence, lower Lev/Man ratio is associated with
436 softwood burning whereas higher ratio with hardwood burning (Engling et al., 2006, 2009).
437 However, we found insignificant diurnal variations of Lev/Man ratios between night (9.33–25.9,
438 avg. 15.8) and day (0.90–23.3, avg. 13.6) in Mangshan. Interestingly, wheat straws and lignite are
439 used in China for domestic cooking and house heating. Such biofuel combustion might also
440 contribute to the high levels of levoglucosan and mannosan in the Mangshan aerosols.

441 Likewise, comparable Lev/Man ratios (9–13 for PM₁₀ and 10–13 for PM_{2.5}) were reported
442 for aerosol samples from Tanzania, where wood and charcoal are primary fuels used for domestic
443 cooking and heating (Mkoma and Kawamura, 2013). Slightly higher Lev/Man ratios were observed
444 in springtime (15–40) aerosols from Jeju Island (Simoneit et al., 2004b). Engling et al. (2013)
445 reported a high Lev/Man ratio (40) for aerosol samples from south-central Taiwan. Furthermore, a
446 high Lev/Man ratio (46) was reported in field burning for wheat straw (Fu et al., 2008). Tsai et al.
447 (2013) reported low Lev/Man ratios of 5.73–7.69 and 14.1–14.9 for urban/industrial aerosols during
448 non-episodic pollution and episodic pollution events, respectively. The low Lev/Man ratios for
449 urban aerosols are reasonable because biofuel and coal are often used for industrial purposes, while
450 biomass burning is more common in rural and suburban areas for household heating and cooking.

451 In addition, there are several biomass combustion studies by chambers and controlled field
452 experiments. Iinuma et al. (2007) reported lower Lev/Man ratios for the burning of conifer (4, 6.5,
453 4.7) than savanna grasses (21.7, 22.7). High Lev/Man ratios were reported for rice straw burning
454 (41.6) (Sheesley et al., 2003) and for cereal straw burning (55.7) (Zhang et al., 2007). Similarly,
455 Schmidl et al. (2008) reported lower Lev/Man ratios (3–5) for softwood than hardwood (14–15)
456 burning by a laboratory chamber experiment. Engling et al. (2009) found higher Lev/Man ratios of
457 40–55 in the controlled field burning of cereal and rice straws. Tsai et al. (2010) reported Lev/Man
458 ratio of 12.5 for fresh sandalwood burning and 13.5 for aged sandalwood burning. As discussed
459 above, it is not easy to differentiate hardwood and softwood burning base on Lev/Man ratios
460 because several studies reported a high Lev/Man ratio for both softwood and hardwood burning.
461 Thus, there may exist some other factors that significantly control the Lev/Man ratios.



462 Additionally, Kuo et al. (2011) reported a significant effect of burning temperature on the
463 emissions of levoglucosan and mannosan: they observed higher levels of levoglucosan at high
464 temperature. The mannosan production increases up to 200 °C. Compared to low-temperature
465 smoldering, higher emissions of levoglucosan was observed during high-temperature flaming (Gao
466 et al., 2003). Engling et al. (2006) reported higher Lev/Man ratios for the flaming (27.5-52.3) than
467 smoldering (2.43-3.08). Notably, higher temperatures (≥ 300 °C) and longer combustion time may
468 result in higher Lev/Man ratios (Kuo et al., 2011). The variations of Lev/Man ratios in the
469 Mangshan aerosols may be significantly influenced by several parameters such as burning
470 temperature associated with the flaming or smoldering process, duration of biomass burning, and
471 hardwood or softwood burning. In spite of the above discussions, moderate Lev/Man ratios in
472 autumn aerosols from Mangshan suggest that softwood burning contributes to the levels of
473 levoglucosan and mannosan. However, the contribution of a hardwood burning from house heating
474 associated with the smoldering process or low temperature could not be excluded.

475 **4. Summary and Conclusions**

476 Anhydrosugars, primary sugars and sugar alcohols were detected with distinct diurnal
477 variations in suburban aerosol samples collected at the Mangshan site in the northern vicinity of
478 Beijing. The wind patterns indicate that daytime air masses were transported from urban Beijing to
479 Mangshan, while nighttime air masses were delivered from the Mangshan National Forest Park.
480 Daytime air masses from urban Beijing significantly influence the air quality of northern forest
481 region. The aerosol concentrations were characterized by the peaks of anhydrosugars in daytime.
482 Levoglucosan was observed as a dominant anhydrosugar among sugar compounds (SCs) due to
483 potential biomass burning (BB) in Mangshan. Levoglucosan showed clear diurnal variation with
484 higher concentrations in nighttime than daytime. BB for cooking and space heating significantly
485 contributed to the high levels of levoglucosan at the suburban site.

486 Sucrose was found as the second most abundant sugar with higher concentrations in
487 daytime. These results suggested a significant influence of enhanced ambient temperature and solar
488 radiation for the pollen rupture in daytime. The insignificant diurnal variation of trehalose indicated
489 multiple sources, i.e., microbial activities at night due to high RH, while soil dust particles
490 transported from Beijing areas due to a southerly wind in daytime. The maximum concentrations of
491 sugar alcohols were observed in daytime. Arabitol and mannitol are delivered from fungal and
492 microbial sources in daytime due to favorable meteorological conditions (high RH and
493 temperature). PMF analyses concluded that local BB and long-range transport of bioaerosols from
494 Beijing seriously affected the air quality of the Mangshan site. The mean contribution of BB tracers



495 to OC was estimated to be 0.43% in nighttime due to domestic cooking and space heating in central
496 China. Although the levels are lower (0.22%) in daytime, Lev/Man ratios obtained in the aerosols
497 suggest that low temperature or smoldering burning of softwood/hardwood was the main source.

498 SCs were recognized as a significant component of aerosols from Mangshan, northern
499 suburbs of Beijing. SCs can influence the air quality and possibly climate because they are essential
500 components of organic aerosols on a global scale. This study of SCs at Mangshan demonstrates that
501 ambient levels of SCs are highly sensitive to the emissions of anthropogenic and biogenic aerosols.
502 Higher contribution of BB product (levoglucosan) to SCs showed a significant BB activity around
503 the Mangshan site in north China.

504

505 **Data availability.** Raw data are available on request by contacting the corresponding author.

506 **Author contributions.** This research was designed YK, KK and ZW. Laboratory measurements
507 were performed by FY with a support of PF. The paper was prepared by SKV and KK.

508 **Competing interests.** The authors declare that they have no conflict of interest.

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757 **Figure Captions**

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759 Figure 1. Geographical location of Mangshan, China. The map was downloaded from © Google
760 Maps 2019.

761 Figure 2. The meteorological parameters at Mangshan during sampling periods.

762 Figure 3. Temporal variations in the concentrations (ng m^{-3}) of sugar compounds in the Mangshan
763 aerosol samples collected for September-October 2007.

764 Figure 4. Concentrations (ng m^{-3}) of sugar compound (a) overall, (b) daytime and (c) nighttime in
765 aerosol samples from Mangshan during autumn 2007 (The error bars denote the standard
766 deviation).

767 Figure 5. Linear regression analysis between mass of sugar components (a) levoglucosan and
768 galactosan, (b) levoglucosan and mannosan, (c) galactosan and mannosan, (d) sucrose and
769 ambient temperature, (e) sucrose and solar radiation, (f) trehalose and arabitol at nighttime, (g)
770 trehalose and mannitol at nighttime, (h) trehalose and Ca^{2+} at daytime, (i) arabitol and
771 mannitol, (j) relative humidity and arabitol, (k) relative humidity and mannitol, (l) glucose and
772 fructose, (m) trehalose and arabitol, (n) trehalose and mannitol, and (o) trehalose and Ca^{2+}
773 in the aerosol samples collected from Mangshan, China.

774 Figure 6. PMF analyses of sugar compounds in Mangshan aerosols based on the autumn 2007 data
775 set.

776 Figure 7. Source contributions to sugar compounds from various sources based on PMF analyses.

777 Figure 8. The concentrations and relative contributions of the carbon content of anhydrosugars (AS-
778 C), primary sugars (PS-C) and sugar-alcohols (SA-C) to the carbon concentrations of measured
779 sugar compounds (SCs-C), water-soluble organic carbon (WSOC) and organic carbon (OC)
780 fraction of Mangshan aerosols (a = daytime and b = nighttime). The concentrations and relative
781 contribution of the carbon content of five sources of sugar compounds to total sugar
782 compounds measured, WSOC and OC fraction of Mangshan aerosols (c = daytime and d =
783 nighttime).

784 Figure 9. Mass concentrations ratio of levoglucosan (Lev) to organic carbon (OC) and water soluble
785 organic carbon (WSOC) in the Mangshan aerosol samples for autumn 2007.

786 Figure 10. Average levoglucosan to mannosan ratios (Lev/Man) in the Mangshan aerosol samples
787 for autumn 2007.

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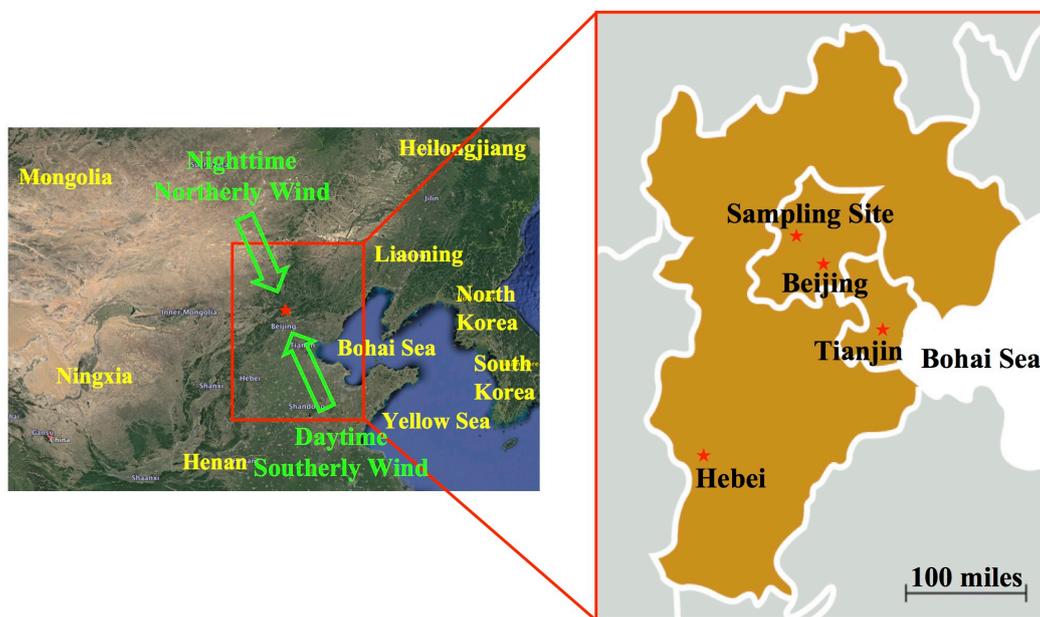
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792 Fig. 1.



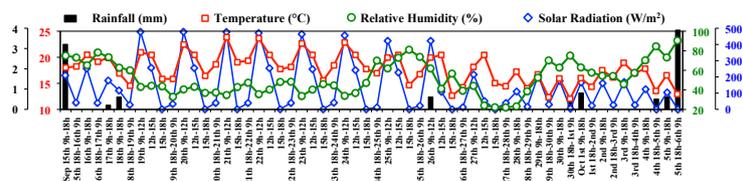
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797 Fig. 2.



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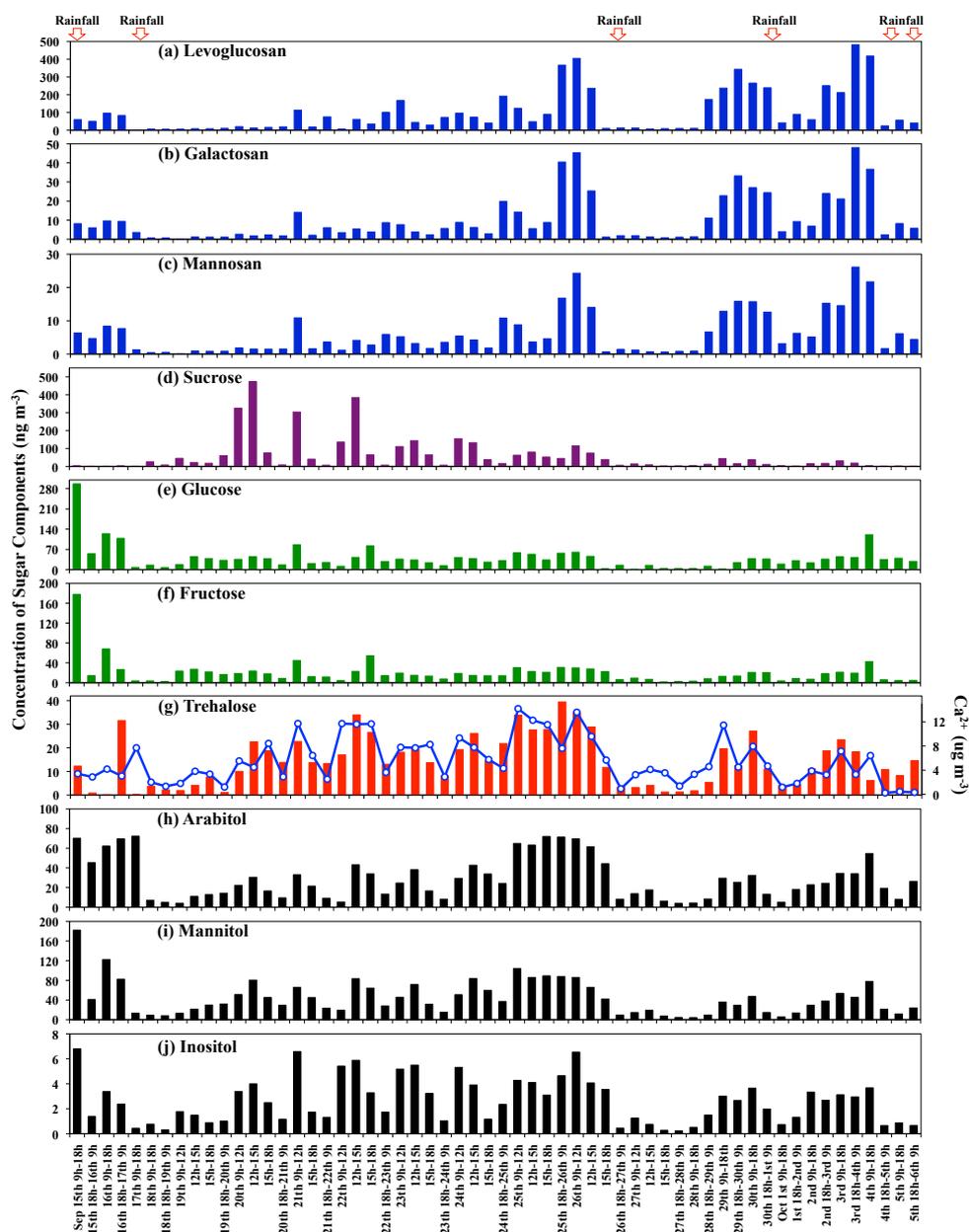


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805 Fig. 3.

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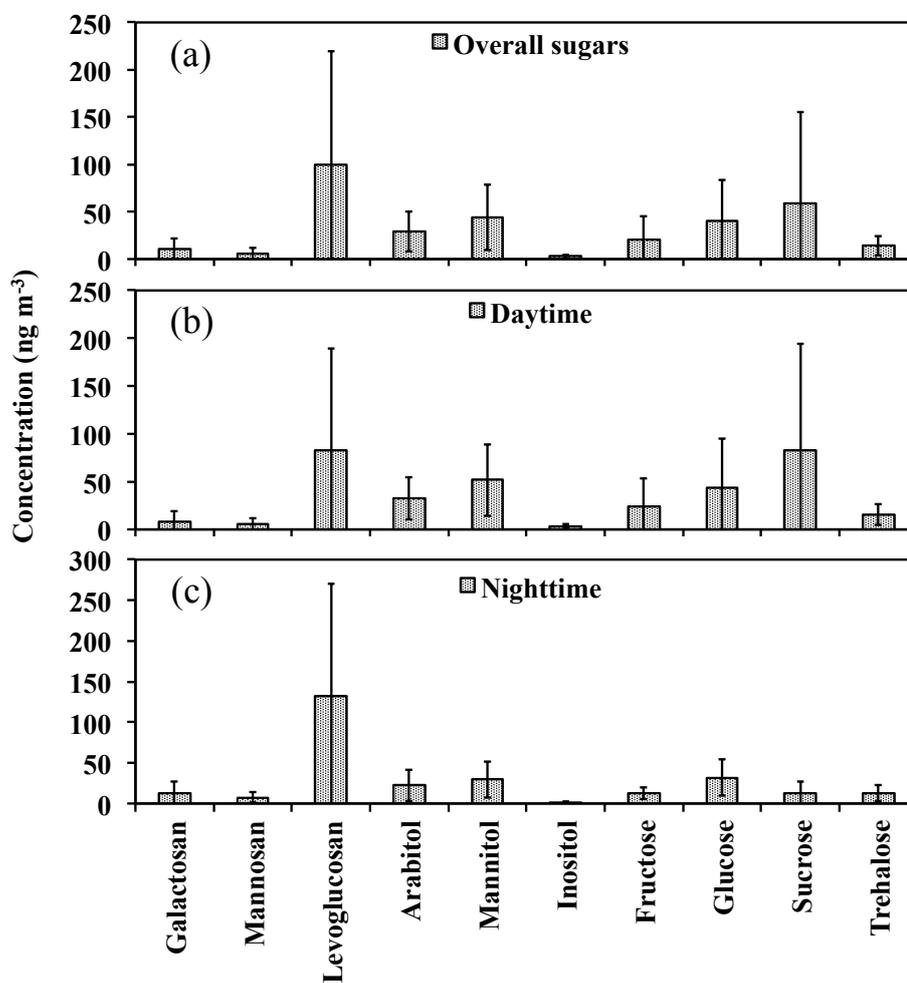


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812 Fig. 4.



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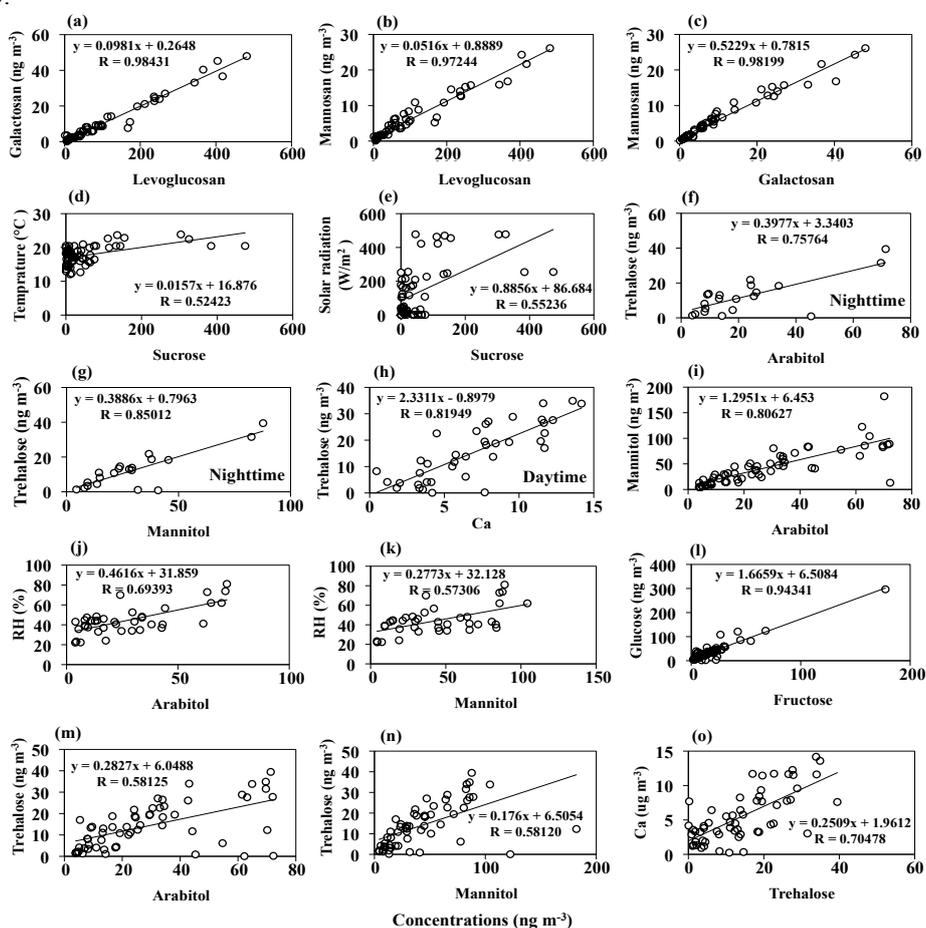
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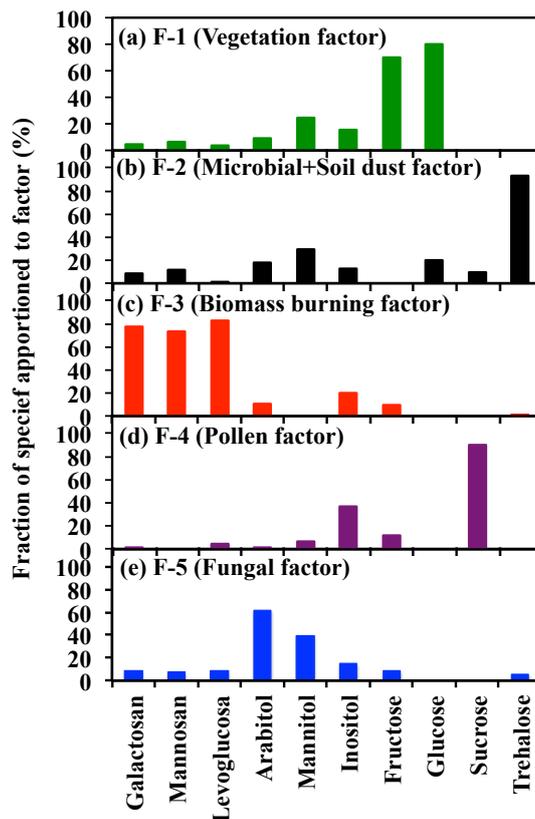
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 831 Fig. 5.



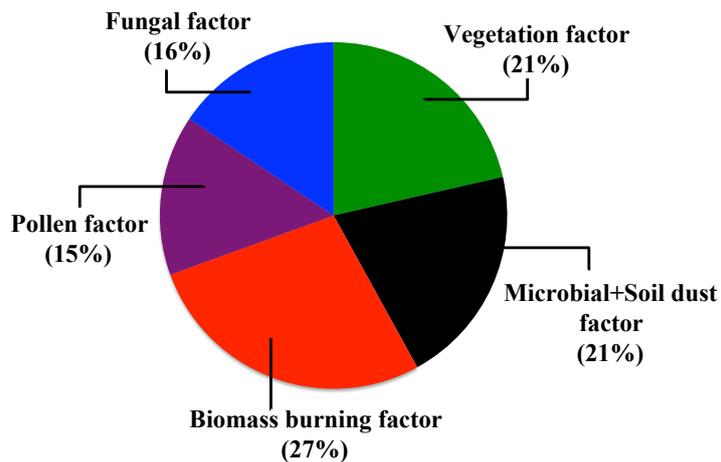
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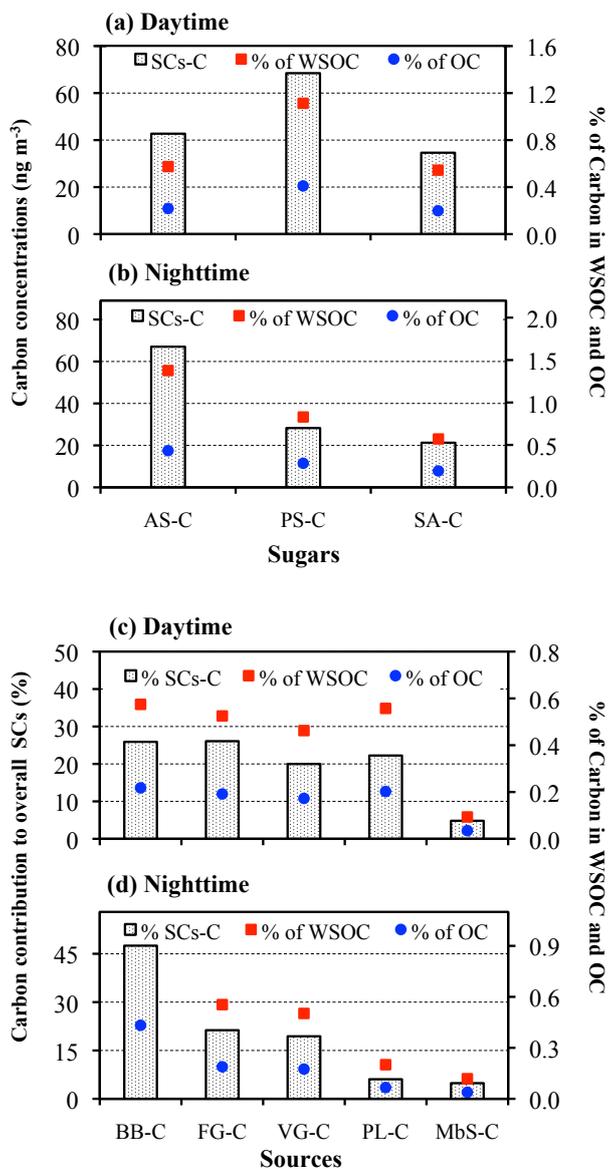
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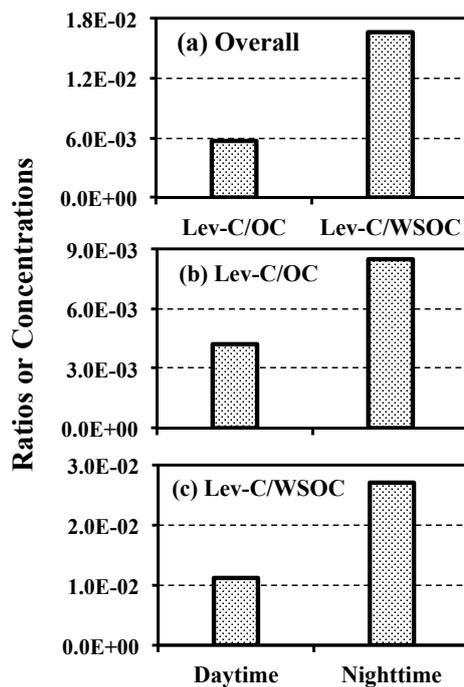


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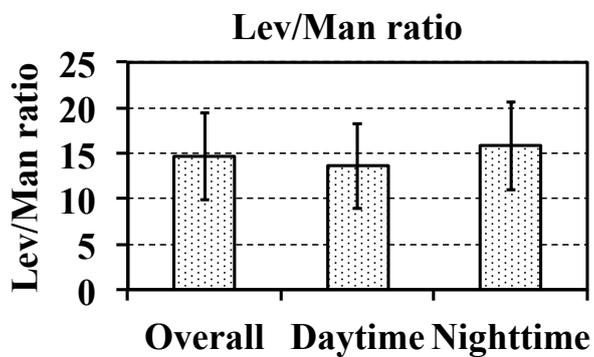
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868 Fig. 9.



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875 Fig. 10.
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883 Table 1. Minimum, maximum, average and standard deviations of concentrations of sugar
 884 compounds in aerosol samples (TSP) from Mangshan, China.
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	Overall				Daytime (n = 38)				Nighttime (n = 20)			
	Min	Max	Avg.	S.D.	Min	Max	Avg.	S.D.	Min	Max	Avg.	S.D.
Anhydrosugars												
Galactosan	0.14	48.0	10.1	11.9	0.14	45.3	8.53	10.5	0.69	48.0	13.0	14.0
Mannosan	0.13	26.1	6.05	6.33	0.13	24.3	5.37	6.01	0.53	26.1	7.35	6.87
Levogluconan	1.17	482	100	119	1.17	418	83.2	106	5.66	482	132	138
Sugar alcohols												
Arabitol	3.89	72.2	29.1	21.5	3.99	72.2	32.5	22.0	3.89	71.3	22.5	19.4
Mannitol	4.19	182	44.1	34.5	4.19	182	51.7	37.5	4.40	87.7	29.6	22.3
Inositol	0.23	6.8	2.62	1.81	0.27	6.80	3.14	1.90	0.23	4.65	1.62	1.09
Primary sugars												
Fructose	1.72	177	20.1	24.6	1.72	177	23.9	29.3	2.64	30.9	12.8	7.67
Glucose	1.86	297	40.0	43.4	1.86	297	44.2	50.8	4.52	108	32.0	22.8
Sucrose	0.02	474	58.5	96.5	0.02	474	82.9	112	0.04	60.1	12.3	15.1
Trehalose	0.06	39.5	14.3	10.5	0.06	34.9	15.3	10.6	0.87	39.5	12.3	10.2
Anhydrosugars	6.01	556	116	137	6.01	476	97.1	122	6.88	556	152	159
Primary sugars	9.41	565	133	125	9.41	565	166	141	10.5	172	69.4	43.0
Sugar alcohols	8.53	259	75.8	54.7	9.09	259	87.4	57.5	8.53	164	53.7	41.9
Total Sugars	30.8	875	325	232	34.1	875	351	240	30.8	759	276	212
Anhydrosugars (%)			31.9				24.6				45.7	
Primary sugars (%)			41.8				47.3				31.3	
Sugar alcohols (%)			26.4				28.1				23.0	

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