

1 **Measurement report: Diurnal and temporal variations of sugar compounds in suburban**
2 **aerosols from the northern vicinity of Beijing, China: An influence of biogenic and**
3 **anthropogenic sources**

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18 Key points:

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- 20 1. Autumn time observations of sugar compounds (SCs) in the northern vicinity of Beijing, China.
- 21 2. Influence of natural biogenic emissions on SCs from forest area.
- 22 3. Influence of anthropogenic and bioaerosol on SCs from the Beijing area.
- 23 4. Biomass burning is a significant contributor to SCs.
- 24 5. Biogenic and fungal-microbial emissions are significant sources for mannitol and arabinol.

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28 **Abstract**

29 Sugar compounds (SCs) are major water-soluble constituents in atmospheric aerosols. In
30 this study, we investigated their molecular compositions and abundances in the northern receptor
31 site (Mangshan) of Beijing, China, to better understand the contributions from biogenic and
32 anthropogenic sources using a gas chromatography–mass spectrometry technique. The sampling
33 site receives anthropogenic air mass transported from Beijing by southerly winds, while northerly
34 winds transport relatively clean air mass from the forest areas. Day- and nighttime variations were
35 analyzed for anhydrosugars, primary sugars, and sugar alcohols in autumn 2007. We found that
36 biomass burning (BB) tracers were more abundant in nighttime than daytime, while other SCs
37 showed different diurnal variations. Levoglucosan was found as a dominant sugar among the SCs
38 observed, indicating an intense influence of local BB for cooking and space heating at the
39 surroundings of the Mangshan site. The high levels of arabitol and mannitol in daytime suggest a
40 significant contribution of locally emitted fungal spores and long-range transported bioaerosols
41 from the Beijing area. The plant emissions from Mangshan forest park significantly control the
42 diurnal variations of glucose, fructose, and mannitol. The meteorological parameters (relative
43 humidity, temperature, and rainfall) significantly affect the concentrations and diurnal variations
44 of SCs. Sucrose (pollen tracer) showed a clear diurnal variation, peaking in the daytime due to
45 higher ambient temperature and wind speed, which influences the pollen release from the forest
46 plants. We found the contribution of trehalose from soil dust in daytime, while microbial and
47 fungal spores were responsible for nighttime. Anhydrosugar and primary sugars are prime carbon
48 sources of the Mangshan aerosols. The high ratios of levoglucosan in organic carbon and water
49 soluble organic carbon in nighttime suggest a significant contribution of BB to organic aerosols at
50 night. Levoglucosan/mannosan ratios demonstrate that low temperature burning of hardwood is
51 dominant in Mangshan. The positive matrix factorization analysis concluded that forest
52 vegetation, fungal species, and local BB are the significant sources of SCs.

53 **Keywords:** Anthropogenic bioaerosols, biomass burning, pollen tracer, fungal tracers, soil dust,
54 and microbial tracers

55

56 **1. Introduction**

57 Increased economic growth and massive consumption of fossil fuels from industries emit
58 anthropogenic gases, aerosols, and biomass burning (BB) products cause severe air pollution in
59 East Asian countries (Lelieveld et al., 2015; Lin et al., 2014; Kawamura et al., 2013; Li et al.,
60 2010; Sun et al., 2016). Globally, significant anthropogenic and carbonaceous aerosols are
61 contributed by China (Cooke et al., 1999, Wang et al., 2007). Beijing is situated in the northern
62 part of China, with a 20 million people and 5 million motor vehicles. Beijing is one of the largest
63 polluted cities in East Asia; its air quality deteriorates seriously due to massive emissions of
64 anthropogenic aerosols from vehicles and industries (Cao et al., 2014; Qiao et al., 2018; Tao et al.,
65 2017; Wei et al., 2018; Yu et al., 2013). Organic aerosols (OAs) are composed of a complex
66 mixture of diverse molecules (Xu et al., 2011). They play essential roles in global climate changes
67 via the modification of radiative forcing and cause a serious negative impact on human health
68 (Fuzzi et al., 2007). OAs contain various water-soluble organic compounds, which can act as
69 cloud condensation nuclei (CCN) (Kanakidou et al., 2005).

70 BB is essentially a primary source of OAs, controlling the air quality levels and affecting
71 the earth's radiative forcing by scattering or absorbing incident solar radiation (Deshmukh et al.,
72 2019a; Kanakidou et al., 2005; Kanaya et al. 2013; Streets et al., 2003; Sullivan et al., 2008).
73 There are several kinds of BB, including industrial biofuel burning, open field burning (fires of the
74 forest, peatlands, and agricultural wastes), and domestic BB burning for house heating and
75 cooking, which emits BB products into the atmosphere (Akagi et al., 2011). The BB aerosols are
76 subjected to long-range atmospheric transport once they are emitted into the atmosphere (Verma et
77 al., 2015). Levoglucosan (1,6-anhydro- β -D-glucopyranose) is a pyrolysis product of cellulose and
78 hemicellulose, which is generally found as major organic constituents in the BB-influenced
79 aerosols (Simoneit et al., 1999; 2002). Levoglucosan have been reported as a specific tracer for
80 BB aerosols (Engling et al., 2009).

81 Sugar compounds (SCs) are ubiquitous in the atmosphere from different geographical
82 locations, including urban, forest, marine, and polar regions (Burshtein et al., 2011; Fu et al.,
83 2010; Wan et al., 2017). SCs are emitted from algae, microbes, pollen, suspended soil particle, and
84 associated biota into the atmosphere by various processes, and thus they are termed as primary
85 biological aerosol particles (PBAPs) (Carvalho et al., 2003; Despres et al., 2012; Elbert et al.,
86 2007). The detailed study of bio-aerosols has been emphasized in the past decades due to the
87 global impact of microbes and fungi because they can travel long distances from the source
88 regions by winds (Burshtein et al., 2011; Brown and Hovmoller, 2002; Yamaguchi et al., 2012).

89 Fungi are essential microbes in the ecosystem, which discharge spores of 8-186 Tg yr⁻¹ into the
90 atmospheric environment (Elbert et al., 2007; Heald and Spracklen, 2009). Sugar alcohols like
91 arabitol and mannitol are enriched in fungal spores; thus, they are considered as specific tracers
92 (Bauer et al., 2008).

93 Devis et al. (1988, 1990) reported that mannitol was also found in about 70 different
94 higher plant families. Loescher et al. (1992) reported that mannitol is an important photosynthetic
95 product converted by biosynthesis in plants. Keller and Matile (1989) also found the arabitol and
96 mannitol during the increased photosynthesis in growing vegetation. Pollens are the largest
97 particles that could contribute up to 65% of the PBAPs, which are the significant sources for
98 sucrose and fructose in the forest aerosols (Manninen et al., 2014; Pacini, 2000). Higher plants
99 synthesize primary sugars (glucose, fructose, and sucrose) during photosynthesis, which are
100 circulated by phloem to accumulate in root cells and to develop plant sections (Jaenicke, 2005; Jia
101 et al., 2010; Pacini, 2000). Cowie et al. (1984) also reported various sugars in terrestrial plant
102 fruits, flowers, and plant tissues. Bielecki (1995) reported that glucose, fructose, and sucrose are
103 well-known components of microbes and invertebrates. The plant debris, as well as lichens,
104 invertebrates, and soil dust, are also recognized as possible sources of primary sugars in the
105 atmosphere (Medeiros et al., 2006; Rogge et al., 2007; Simoneit et al., 2004).

106 Previous studies analyzed aerosol samples for SCs and discussed several factors to control
107 their local and global atmospheric levels. Recently, Xu et al. (2020) examined the seasonal
108 molecular distributions of primary biological aerosols and BB aerosol samples collected from
109 urban Beijing. They reported a high level of arabitol, mannitol, sucrose, glucose, and fructose in
110 the vegetation-growing season. Kang et al. (2018) also reported higher concentrations of sugars in
111 the urban aerosols from Beijing. They suggested a large contribution of coal combustion and
112 agriculture residue burning under stable meteorological conditions in winter and spring. Verma et
113 al. (2015, 2018) reported that the atmospheric circulations and long-range transport of organic-
114 /bio-aerosols from East Asia significantly control the levels and compositions of SCs over the
115 western North Pacific. The above studies discussed the several factors that affect the
116 concentrations of SCs in the aerosol samples collected from urban and remote areas.

117 In this study, we conducted analyses of SCs in the aerosol samples collected from the
118 northern vicinity of Beijing City in 2007. Here, we present comprehensive data sets of
119 anhydrosugars, primary sugars, and sugar alcohols in the suburban aerosol samples and their
120 diurnal variations to explain the source variance following the wind patterns in the day- and
121 nighttime. The positive matrix factorization (PMF) has been applied to clarify the different

122 sources of measured SCs in the aerosol. We present the influence of local meteorology of
123 sampling site and atmospheric transport from Beijing by southerly winds and Mangshan National
124 Forest Park by northerly winds on the molecular distributions of SCs. Using the mass
125 concentration ratio of levoglucosan to mannosan, we explain the relative contribution of hard and
126 softwood burning to the air quality of Mangshan. This study also discussed carbon contributions
127 of SCs and BB measured in the Mangshan aerosol samples from different sources.

128 **2. Materials and Methods**

129 **2.1. Site description and aerosol sample collection**

130 The sampling site (Mangshan: 40.28 N, 116.26 E) is located 40 km north of Beijing. A
131 detailed description of the sampling site is given in He et al. (2014, 2015). Briefly, Mangshan is
132 surrounded by urban areas in the south and forest areas with the national park in the north (Fig. 1).
133 The ambient temperature was higher in daytime (23.9°C) than nighttime (12.1°C), with an average
134 of 17.8°C during the campaign. The relative humidity (RH) varied significantly from 22.1% to
135 90.5%, with an average of 51.7% during the study period. The rainfall was observed at midnight
136 on 15th September, the morning of 17th to evening 18th September, the night of 26th September,
137 and light rain lasted from 4th October to the end of the campaign (Fig. 2a). Interestingly, the
138 sampling site is characterized by a specific wind pattern, i.e., southwest wind (69.9%) prevailed,
139 followed by northeast wind (23.4%) and southeast wind (6.2%) during the daytime (Fig. 2a). The
140 northeast wind (99.5%) was dominated at night, which is consistent with the air mass back
141 trajectories (He et al., 2014) (Fig. 2b). The daytime wind from the southwest direction passed over
142 Beijing, delivering anthropogenic air mass to the Mangshan site.

143 Detailed descriptions of the total suspended particulate (TSP) samples collected at
144 Mangshan are given in He et al. (2014, 2015). Briefly, The aerosol samples were collected near
145 the entrance of Mangshan National Forest Park. The elevation of the sampling location is 187 m
146 above sea level. A high-volume air sampler (Kimoto-AS810A) at a flow rate of 1.13-1.17 m³ min⁻¹
147 was used to collect the TSP without cut-off device. In the sampling, no denuder was applied to
148 remove semi-volatile gases because the filter samples were used to analyze nonvolatile sugar
149 compounds. However, the levoglucosan partition between the gas and particle phases, but their
150 concentration was low. The sampling time was rather short due to the day and night sampling.
151 Therefore, the uncertainty due to the gas phases in the particulate species concentration might be
152 insignificant. The samples were collected on pre-combusted (450°C for 6 h) quartz fiber filters
153 (Pallflex 2500QAT-UP, 20 cm × 25cm) from 15th September to 5th October 2007. After sample

154 collection, the individual filters were placed in pre-combusted glass jars with Teflon-lined screw
155 caps and stored in a dark, cold room at -20°C to prevent microbial activity and loss of semi-
156 volatile organic compounds from the samples. In this study, a total of 58 filter samples were
157 analyzed. We collected 3h daytime (from 9 to 12, 12 to 15, 15 to 18 h) ($n=26$), 9h daytime (from 9
158 to 18 h) ($n=12$), and 15h nighttime (from 18 to 9 h) ($n=20$) samples together with four field
159 blanks. Table S1 shows the details of aerosol sample collection in the Mangshan site.

160 **2.2. Extraction and derivatization of samples**

161 A total of 58 aerosol samples were analyzed for anhydrosugars, primary sugars, and sugar
162 alcohols (Table 1). The sample filters (approximately 21 cm^2) were extracted with a
163 dichloromethane and methanol mixture (2:1) under ultrasonication. Pasteur pipettes packed with
164 pre-combusted quartz wool were used to filter the extracts to remove filter debris. After filtration,
165 the extracts were concentrated in a rotary evaporator under vacuum and dried by nitrogen
166 blowdown. The extracts were reacted with $60\text{ }\mu\text{L}$ of N,O-bis-(trimethylsilyl)trifluoroacetamide
167 (BSTFA) with 1% trimethylsilyl (TMS) chloride in the presence of $10\text{ }\mu\text{L}$ of pyridine at 70°C for
168 three hours to derivatize hydroxyl (OH) and carboxyl (COOH) groups into corresponding
169 trimethylsilyl (TMS) ethers and esters, respectively. After the reaction, n-hexane was used for
170 dilution, and C_{13} n-alkane was added as an internal standard before GC-MS analysis.

171 **2.3. Gas chromatography-mass spectrometry determination of sugar compounds (SCs)**

172 Details of GC-MS operation and identification of SCs are described in Verma et al. (2015,
173 2018). Briefly, GC-MS analyses were performed on Agilent model 6890 gas chromatograph (GC)
174 combined with an Agilent model 5973 mass selective detector (MSD) to determine SCs. The mass
175 spectrometer was operated in the electron ionization (EI) mode at 70 eV with a scan range of m/z
176 $40\text{--}650$. The GC separation was achieved on a DB-5MS fused silica capillary column ($30\text{ m} \times$
177 0.25 mm in diameter, $0.25\text{ }\mu\text{m}$ film thickness) and a split/splitless injector. The GC oven
178 temperature was programmed to maintain at 50°C for 2 min and then to increase from 50 to 120°C
179 at a rate of $15^{\circ}\text{C min}^{-1}$, then from 120 to 305°C at a rate of $5^{\circ}\text{C min}^{-1}$. The final isotherm holds at
180 305°C for 15 min. Helium was used as the carrier gas at a flow rate of 1.0 mL min^{-1} . The sample
181 was injected on a splitless mode at 280°C injector temperature. GC-MS data were acquired and
182 processed with Agilent GC/MSD ChemStation software.

183 The individual compounds (TMS derivatives) were identified by comparing the relative
184 response factors determined by the injection of authentic standards and those reported in the

185 literature and library texts (Claeys et al., 2004). Fragment ions of sugar compounds at 217 and 204
186 were used for quantifications. Total ten sugar compounds, including three anhydrosugars
187 (levoglucosan, galactosan, mannosan), four primary sugars (glucose, fructose, sucrose, trehalose
188 and xylose) and three sugar alcohols (arabitol, mannitol, and inositol), were detected in the
189 Mangshan aerosols. Field blanks were treated as a real sample and analyzed by the procedure used
190 for the real samples. Recoveries for SCs were better than 85% as obtained by the standards spiked
191 to precombusted quartz filter followed by extraction and derivatization. Based on the duplicate
192 analysis, the analytical errors in the concentrations of the detected compounds were obtained to be
193 within 10%. The detection limits of SCs corresponds to ambient concentrations of 150-620 $\mu\text{g L}^{-1}$
194 ¹, which corresponds to ambient concentrations of 15-70 $\mu\text{g m}^{-3}$ under a typical sampling volume
195 of 900 m^3 .

196 **2.4. Chemical analyses of organic carbon (OC), water-soluble organic carbon (WSOC), and** 197 **inorganic ions**

198 The data set and methods for the determination of organic carbon (OC), water-soluble
199 organic carbon (WSOC) and inorganic ion (Ca^{2+}) were reported in He et al. (2015). Briefly, the
200 concentrations of OC were measured using a semi-continuous OC/EC analyzer (Sunset Laboratory
201 Inc., Portland, OR, USA). A punch of the filter ($\Phi 14$ mm) was placed in a quartz boat inside the
202 thermal desorption chamber of the analyzer, and then stepwise heating (IMPROVE) was applied.
203 The oven temperature was programmed as follows: under He, every 2 minutes, the oven
204 temperature was increased starting from 250°C for 2 min, at 450°C for 2 min, and at 550°C for 2
205 min. After that, 550°C was maintained for two minutes under He mixed with 10% O_2 , then at
206 700°C for 2 min and at 870°C for 3.5 min. NDIR detector was used to determine CO_2 generated in
207 the above process (Wang et al., 2005b). The carbon content of the sample that evolves to CO_2
208 between 250 and 700°C was defined as OC.

209 Aliquots of the filter samples (3.14 cm^2) were extracted with Milli Q water for the water-
210 soluble inorganic ion and WSOC measurements. After extraction, one part was used for the
211 analyses of inorganic ions (SO_4^{2-} , NO_3^- , Cl^- , NH_4^+ , Na^+ , Ca^{2+} , K^+ and Mg^{2+}) using an ion
212 chromatography (IC) system (761 Compact IC, Metrohm, Switzerland). Cations on a Shodex YK-
213 421 column with 4mM H_3PO_4 as eluent and anions were separated on a Shodex SI-90 4E column
214 with 1.8mM Na_2CO_3 and 1.7mM NaHCO_3 as eluent. The injection loop volume was 200 μl . Both
215 cations and anions were quantified against a standard calibration curve. Another part of the filtered
216 water extract was acidified with 1.2 M HCl and purged with pure air to remove dissolved
217 inorganic carbon and volatile organics. Then WSOC was measured with a carbon analyzer

218 (Shimadzu, TOC-5000). Procedural blanks were carried out in parallel with real samples to
219 account for any contamination (He et al., 2015).

220 **2.5. Positive Matrix Factorization (PMF) Analysis**

221 Positive matrix factorization (PMF) is a powerful statistical tool for resolving the potential
222 sources contributing to atmospheric particles (Paatero and Tapper, 1994). The measured ambient
223 concentrations and method detection limits (MDLs) of SCs were used to calculate the
224 uncertainties. The measured concentrations of SCs below or equal to the MDLs were replaced by
225 half of the MDL, and associated uncertainties were set at 5/6 of the MDL [(5/6) × MDL] values of
226 each sample. The geometric mean concentrations were used for missing concentrations, and the
227 uncertainty of the concentrations greater than the MDL was calculated based on the following
228 equation:

$$229 \text{ Uncertainty} = \sqrt{(\text{error fraction} \times \text{concentration})^2 + (0.5 \times \text{MDL})^2}$$

230 The error fraction is a user-provided estimation of the analytical uncertainty of the
231 measured concentration or flux. For example, Han et al. (2017) used an error fraction of 0.2-0.3
232 for organics and 0.2 for all the species. In this work, the error fraction was set to be 0.3 for all
233 species. Paatero et al. (2002) and Zhou et al. (2004) reported detailed discussions of the
234 determination and application of PMF analysis.

235 **3. Results and Discussion**

236 **3.1. Ambient concentrations and diurnal variations of SCs**

237 We detected a total of ten SCs, including three anhydrosugars, four primary sugars, and three
238 sugar alcohols in the Mangshan aerosol samples. Figure 3a-c showed the temporal variations and
239 Table 1 showed minimum, maximum, and average concentrations of anhydrosugars, primary
240 sugars, and sugar alcohols with a standard deviation. The overall concentrations of SCs varied
241 from 30.8–875 ng m⁻³ (avg. 325 ng m⁻³), which was higher in the daytime (315 ng m⁻³) and lower
242 at nighttime (276 ng m⁻³), however, we did not observe statistically significant differences
243 (student t-test, 95% confidence interval, p > 0.05) in their atmospheric abundances. Interestingly,
244 higher average concentrations of SCs were reported for the aerosol samples collected from at Mt.
245 Tai (daytime 640 ng m⁻³ and nighttime 799 ng m⁻³) in the North China Plain (Fu et al., 2008) than
246 the Mangshan aerosol. The diurnal concentrations of SCs may be significantly influenced by
247 vegetation and BB activities in the Mangshan site. SCs are significantly contributed by plant

248 fractions and fungus from the forest area (Zhu et al., 2016). The meteorological parameters also
249 affect the concentrations of SCs in the forest site (Miyazaki et al., 2012).

250 In addition, anthropogenic aerosols emitted from urban areas are probably transported to the
251 northern receptor site in daytime by a southerly wind (He et al., 2014; 2015). Therefore, the high
252 levels of SCs in daytime may be related to the transport of organic and bio-aerosols from urban
253 regions. The nighttime, the wind direction is shifted to northerly, delivering comparatively clean
254 air masses from the Mangshan National Forest area to the sampling site. Air mass from the forest
255 may significantly contribute to nighttime SCs in the Mangshan site. The influence of local sources
256 and long-range transported aerosols on the SCs will be discussed in sections 3.1.1 to 3.1.3.

257 **3.1.1. Ambient concentrations and diurnal variations of anhydrosugars**

258 The average concentrations of anhydrosugars were found 116 ng m^{-3} , contributing 31.9%
259 of overall SCs in the Mangshan aerosols (Table 1). Figure 4a-c shows the temporal variations of
260 anhydrosugars. They are more abundant in nighttime (avg. 152 ng m^{-3}) than daytime (avg. 97.1 ng
261 m^{-3}). Levoglucosan (100 ng m^{-3}) is the most abundant anhydrosugar followed by galactosan (10.1
262 ng m^{-3}) and mannosan (6.05 ng m^{-3}) detected in Mangshan aerosols. Kang et al. (2018) reported
263 high levels of levoglucosan (avg. 110 ng m^{-3}) in autumn aerosols from Beijing, China. It is well
264 known that biofuel burning is the common energy source for cooking and house heating in China
265 in winter and autumn (Verma et al., 2015), thus the domestic BB activities in the surroundings of
266 the Mangshan site significantly contribute to the levoglucosan. BB tracers showed significant
267 positive correlations with each other (levoglucosan and galactosan, $r = 0.98$; levoglucosan and
268 mannosan, $r = 0.97$; galactosan and mannosan, $r = 0.98$), suggesting their similar sources in the
269 Mangshan aerosols (Table 2).

270 The levoglucosan concentrations showed significant diurnal variations, which was higher
271 in nighttime (avg. 132 ng m^{-3}) than daytime (avg. 83.2 ng m^{-3}) (Table 1). A similar diurnal pattern
272 was also found for the concentrations of galactosan and mannosan. The increased concentrations
273 of BB tracers were observed during the periods of lower ambient temperature (Figs. 2a, 4a-c). The
274 higher ambient temperature was recorded in daytime between 09h to 15h during the campaign,
275 associated with declined BB activities. In this sequence, the nighttime samples were collected
276 from 18:00h to 09:00h, including peak hours of BB for domestic purpose. Therefore, it is
277 reasonable to detect higher abundances of BB tracers in the nighttime than daytime. Hence, it is
278 evident that BB activities were increased at night because of cooking and house heating at cool
279 night in autumn. In addition, recent studies reported the widespread BB aerosols in the North

280 China Plain, including megacities such as Beijing, Nanjing, Hebei, and Tianjin (Lelieveld et al.,
281 2015; Kawamura et al., 2013; Li et al., 2010; Sun et al., 2016). Therefore, the atmospheric
282 transport of BB aerosols from the urban area to the Mangshan site by southerly winds cannot be
283 excluded. The diurnal variations of levoglucosan may be significantly influenced by the local BB
284 activities and transported BB aerosols from urban areas, where BB products are generated by
285 brown coal combustion (Yan et al., 2018).

286 3.1.2. Ambient concentrations and diurnal variations of primary sugars

287 The fragment of vascular plants contains primary sugars, including glucose, fructose,
288 sucrose, and trehalose (Medeiros et al., 2006). Primary sugars were found as the most abundant
289 sugars (avg. 133 ng m⁻³), contributing to 41.8% of the total SCs in Mangshan aerosols (Table 1).
290 They showed apparent diurnal variations with daytime high (avg. 166 ng m⁻³) and nighttime low
291 values (avg. 69.4 ng m⁻³) (Figs. 3a-c, 5a-d). Graham et al. (2003) also reported similar diurnal
292 variations of primary sugars for the Amazon forest aerosols. Sucrose was found as dominant
293 primary sugars (avg. 58.5 ng m⁻³), accounting for 44% of measured primary sugars in Mangshan
294 aerosols (Table 1). Pollen was reported as a primary source for sucrose in aerosols collected from
295 a Texas rural site (Jia et al., 2010). Fu et al. (2012) found high sucrose concentrations up to 1390
296 ng m⁻³ in the aerosols from Jeju Island, South Korea. Therefore, the plant materials, including
297 pollen spores from the local vegetation of Mangshan National Forest Park, are likely the primary
298 source of sucrose in the aerosols. Miyazaki et al. (2012) also reported higher sucrose
299 concentrations in the aerosol samples collected from the Hokkaido deciduous forest.

300 We found a significant diurnal variation of sucrose with higher daytime (82.9 ng m⁻³) than
301 nighttime (12.3 ng m⁻³). Meteorological parameters such as temperature, rainfall, wind speed, and
302 solar radiation significantly influence pollen activities and, subsequently, sucrose concentrations
303 (Verma et al., 2018). Interestingly, an elevated peak of sucrose was observed from 12h to 15h with
304 higher ambient temperature. In contrast, lower sucrose concentrations were observed from 15h to
305 9h with lower ambient temperature (Fig. 5a). Daytime increased concentrations of sucrose might
306 be related to the higher daytime ambient temperature, low RH, and high solar radiation (Miyazaki
307 et al., 2012). Taylor et al. (2002) reported the influence of the meteorological conditions, i.e.,
308 strong daytime winds and convective activity, which can result in catapulting of pollen, opening of
309 pollen-laden flower anthers, and causing enhance entrainment and dispersal of the particles into
310 the air. Pacini (2000) reported that higher levels of sucrose in daytime coincide with higher counts
311 of pollen, fern spore, and insect. The positive linear correlations of sucrose with ambient
312 temperature ($r = 0.52$) and solar radiation ($r = 0.55$) further supported the influence of

313 meteorological parameters in the sucrose concentration (Table 2).

314 Five rain events were recorded during the campaign, i.e., 15th, 17th, 18th, and 26th
315 September, and 1st and 5th October (Fig. 2a). Pollens are significantly settled down by wet
316 scavenging during rain events because their sizes are large. A low concentration of sucrose was
317 found from the beginning of sampling to the morning of 20th September and from the afternoon of
318 26th September to the end of the sampling campaign (Fig. 5a). In addition, the increased
319 concentrations of sucrose were found in the aerosol samples collected from 20th to 22nd September,
320 and moderate concentrations were observed after 23rd to the evening of 25th September during
321 non-precipitation events. Consequently, the pollens were significantly scavenged during wet
322 precipitation and washout effect from the atmosphere, resulting in lower sucrose concentrations at
323 the earlier periods, than later periods. In addition, Rogge et al. (2007) reported that surface soil
324 dust and unpaved road dust also contribute sucrose in the atmospheric aerosols. However,
325 insignificant correlations between sucrose and Ca²⁺ (daytime, $r = 0.32$; night time, $r = 0.37$) do not
326 supports soil dust contributions to sucrose in the Mangshan aerosols (Table 2).

327 Glucose was the second dominant primary sugar in the Mangshan aerosols. The average
328 concentrations of glucose and fructose were observed to be 40.0 ng m⁻³ and 20.1 ng m⁻³,
329 respectively (Table 1, Fig. 5b). The sampling site is characterized by the dense vegetation in the
330 Mangshan National Forest Park. Therefore, the nectars and fruits of vegetation (Baker et al.,
331 1998), plant debris (Medeiros et al., 2006) and pollens (Fu et al., 2012) in the forest significantly
332 contribute to glucose and fructose. The glucose levels are equivalent to that (50.1 ng m⁻³) reported
333 from the Howland Experimental Forest site in the USA (Medeiros et al., 2006). Glucose and
334 fructose showed significant diurnal variations, whose concentrations were higher in daytime (44.2
335 ng m⁻³ and 23.9 ng m⁻³, respectively) than nighttime (32.0 ng m⁻³ and 12.8 ng m⁻³, respectively)
336 in Mangshan aerosols (Table 1, Figs. 3b, c; 5b, c). This diurnal variation could be involved with
337 emissions of pollens, fern spores, and other giant particles by strong winds (Graham et al., 2003;
338 Pacini, 2000). Similar trends of glucose and fructose were reported in the Amazon forest, being
339 coincided with plant fragments and insects (Graham et al., 2003). The autumn decay of vascular
340 plant leaves in the Mangshan forest may have contributed to the levels of glucose and fructose.

341 Although, the daytime southerly winds deliver anthropogenic air masses from megacities
342 to the sampling site. The daytime winds from the northeast direction (23.4%) also carry air masses
343 from the forest region, transporting primary sugars to the Mangshan site. However, 99.5% of the
344 nighttime hours, the wind is shifted to northeasterly, i.e., in forest region (He et al., 2015), but the
345 emissions of primary sugars at night in the form of plant fragments are lower than in daytime.

346 Because the daytime ambient temperature and solar radiations significantly induce the emissions
347 of sugar compounds in the forest site (Miyazaki et al., 2012). Therefore, low glucose and fructose
348 levels were found at nighttime than daytime aerosols at the Mangshan site (Table 1, Fig. 3).
349 Previous studies have reported lichens (Dahlman et al., 2003) and soil dust (Nolte et al., 2001;
350 Rogge et al., 2007) as significant sources of both primary sugars. The concentration of glucose
351 was insignificantly correlated with soil tracer (Ca^{2+}) in day ($r = 0.02$) and nighttime ($r = 0.27$),
352 denying their soil dust contributions in Mangshan aerosol samples.

353 Trehalose in the environment is significantly controlled by the activities of bacteria, fungi,
354 yeast, algae, invertebrates, and plant species, as well as suspended soil particles (Medeiros et al.,
355 2006, Rogge et al., 2007). The average concentration of trehalose was found 14.3 ng m^{-3} (Table 1,
356 Fig. 5d). Yttri et al. (2007) reported higher trehalose concentrations in the aerosol samples
357 collected from urban (29 ng m^{-3}) and suburban (27 ng m^{-3}) than rural (3.8 ng m^{-3}) areas in
358 Norway. The above results emphasize that fungi and microbes associated with anthropogenic and
359 bioaerosols, emitted in the urban and suburban areas, might be responsible for the trehalose
360 concentration in aerosol samples (Verma et al., 2018). Trehalose showed insignificant diurnal
361 variation, whose day and night concentrations were observed 15.3 ng m^{-3} and 12.3 ng m^{-3} ,
362 respectively, indicating its different emission sources in day and night for Mangshan aerosols (Fig.
363 3b, c; 5d).

364 The southerly winds might transport fungi and microbes associated with bioaerosols, eject
365 spores under favorable meteorological conditions (high RH and low temperature) (Jones and
366 Mitchell et al., 1996). Several microbes and fungi discharge spores at nighttime due to high RH
367 conditions (Ibrahim et al., 2011; Kim and Xiao, 2005; Malik and Singh, 2004; Sharma and Razak,
368 2003). Interestingly, trehalose is more significantly correlated with arabitol and mannitol ($r = 0.76$
369 and 0.85 , respectively) in nighttime than daytime ($r = 0.49$ and 0.51 , respectively) (Table 2),
370 suggesting that fungal and microbial spores contributed to high levels of trehalose in nighttime.
371 Hackl et al. (2000) found trehalose as dominant sugar in spring aerosols and proposed it as a tracer
372 for soil dust particles. Trehalose concentration was more significantly correlated with Ca^{2+} ($r =$
373 0.82) in daytime than nighttime ($r = 0.61$), indicating soil dust contribution (Table 2). Therefore,
374 we hypothesized that winds transported soil particles from the urban area in daytime due to the
375 active building constructions (He et al., 2015), contributing to the high levels of trehalose in
376 daytime.

377 3.1.3. Ambient concentrations and diurnal variations of sugar alcohols

378 The average concentrations of sugar alcohols were found 75.8 ng m^{-3} , contributing 26.4%
379 of total SCs measured in Mangshan aerosols (Table 1). Sugar alcohols showed clear diurnal
380 variations in daytime high (avg. 87.4 ng m^{-3}) and nighttime low (avg. 53.7 ng m^{-3}) (Table 1).
381 Mannitol was found as the dominant sugar alcohol (avg. 44.1 ng m^{-3}), followed by arabitol (avg.
382 29.1 ng m^{-3}) and inositol (avg. 2.62 ng m^{-3}) (Table 1; Fig. 6a-c). Mannitol and arabitol are
383 common polyols detected in green algae, lichens, and fungal spores (Bieleski, 1995, Dahlman et
384 al., 2003; Filippo et al., 2013; Lewis and Smith, 1967; Yttri et al., 2007). Previous studies have
385 reported that arabitol and mannitol are key components of fungal spores, and thus they are
386 considered as fungal tracers (Bieleski, 1995; Lewis and Smith, 1967). Several fungal and microbial
387 species released spores during biological activities into the atmosphere (Dahlman et al., 2003;
388 Bauer et al., 2008; Filippo et al., 2013). Therefore, the autumn time fungal and microbial species
389 significantly contribute to arabitol and mannitol in the Mangshan aerosol samples.

390 However, mannitol and arabitol showed a strong positive linear correlation ($r = 0.81$),
391 which suggested common origins as reported in earlier studies (Fu et al., 2012) (Table 2). In
392 contrast, the higher concentration of mannitol than arabitol suggested it had sources in addition to
393 fungal spores in the Mangshan forest site. In this sequence, several previous studies have
394 confirmed the significance of mannitol in plant photosynthesis (Loescher et al., 1992; Keller and
395 Matile, 1989; Rumpho et al., 1983). Pashynska et al. (2002) reported that detritus of mature leaves
396 can emit mannitol into the atmosphere by wind action. Heald and Spracklen (2009) also found a
397 correlation between the atmospheric water vapor with mannitol concentrations and leaf area index.
398 They suggested that the activities of the terrestrial biosphere widely affect mannitol concentrations
399 in the air. Our PMF results also indicated the substantial contribution of mannitol for vegetation
400 factor (24.8%), which supports that mannitol is attributed by vegetation from the forest area
401 (section 3.2).

402 In addition, the meteorological parameters, including high RH and temperature affect the
403 fungal and bacterial activities (Kim and Xiao, 2005; Sharma and Razak, 2003). The maximum
404 growth of fungi and bacteria was observed at 92–100% RH (Ibrahim et al., 2011). Interestingly,
405 the concentrations of arabitol and mannitol gradually increased after the end of precipitation,
406 following the increases in ambient temperature and RH (Figs. 2a, 6a, b). Miyazaki et al. (2012)
407 also discussed the increased contributions of arabitol and mannitol with daytime ambient
408 temperature and solar radiation in the aerosol samples collected from the forest area. Similar
409 temporal trends and positive linear correlations were observed between arabitol ($r = 0.69$) and
410 mannitol ($r = 0.57$) with RH, which supports the above phenomenon for Mangshan aerosols

411 (Table 2). Therefore, we propose that a favorable meteorological condition in autumn increases
412 the emissions of fungal spores and fragments of forest vegetation, which may be responsible for
413 arabitol and mannitol contributions in the Mangshan aerosols.

414 The diurnal variation of mannitol and arabitol were characterized by higher in the daytime
415 (51.7 ng m^{-3} and 32.5 ng m^{-3} , respectively) than nighttime (29.6 ng m^{-3} and 22.5 ng m^{-3} ,
416 respectively) (Fig. 3b, c). Yamaguchi et al. (2012) reported that fungal spores and bacterial cells
417 associated with bioaerosols could be transported long distances. The Mangshan site receives
418 significant anthropogenic and bioaerosols from Beijing City by southerly winds. Therefore, the
419 daytime plant activities, influenced by solar radiation and ambient temperature and the long-range
420 transport of fungal spores from megacities (Beijing) by southwest winds govern the diurnal
421 variation of sugar alcohols in the Mangshan atmosphere. On the other hand, lower concentrations
422 in nighttime can be explained by the clean air mass transport by mountain breeze from the
423 Mangshan National Forest area.

424 **3.2. Source apportionment of SCs**

425 To investigate the source apportionment of SCs, positive matrix factorization (PMF)
426 software version 5.0 (Environmental Protection Agency, USA) was used. The PMF analysis was
427 performed for the measured aerosol samples using tracer compounds for anhydrosugars, primary
428 sugars, and sugar alcohols. It is essential to select a suitable number of factor solutions in the PMF
429 analysis. Based on the possible sources of SCs, four to six factor solutions were run in PMF
430 model. In the four-factor solutions, the SCs, including arabitol, mannitol, and trehalose, were
431 merged in a single factor; this might underestimate the soil dust sources. In six factor solutions,
432 the SCs, including glucose, fructose trehalose, arabitol, and mannitol, were distributed in more
433 than four factors; it might be overestimated the number of factor solutions according to possible
434 sources of SCs. Therefore, a total of five interpretable factor solutions were characterized by the
435 enrichment of each tracer compound to be significant to categorize the origins of individual
436 sugars, which reproduced more than 95% of SCs.

437 These five-factor solutions were preferred based on minimum robust and true Q values
438 (goodness of fit parameters) of the base runs, which observed 3103 and 3505, respectively. In each
439 bootstrap run, the concentrations and percentages of tracers were close to those of base-run results.
440 The PMF results of SCs indicate a stability because no significant changes were found between Q
441 values and factor profiles of F_{peak} rotation runs compared with the base run. PMF results show a
442 good correlation between the values of observed and predicted (modeled) concentrations in scatter

443 plot, indicating that the model very well fits the individual sugar species. These results support the
444 perfect rationality of the source apportionment (Figure S-1). The time series plot of observed and
445 predicted concentration (modeled) also shown that the model well fits the observed data set
446 (Figure S-2). The time series plots of the factors solutions determined by PMF were similar to the
447 temporal plots of the concentration of sugar species of the factor composition (Figure S-3). The
448 numbers of factors were reduced if the pair of factors was strongly correlated. The composition of
449 each factor was also checked; none of the pair of factors were found with similar composition. We
450 also investigated the change in factor profile with positive and negative values of f_{peak} for the
451 chosen solution in the PMF analysis. Figures 6 and 7 show the factor profile resolved by PMF
452 analysis of the Mangshan aerosol samples. The percentages of each component are summed for
453 factors 1 to 5 to be calculated as 100%.

454 Factor 1 is characterized by the high contribution of glucose (80.2%) followed by fructose
455 (69.6%), mannitol (24.8%), and inositol (15.1%) (Fig. 7a). Glucose and fructose are highly water-
456 soluble SCs present in the leaves and bark of plants (Graham et al., 2003). High concentrations of
457 glucose and fructose have been reported in vascular plants and phytoplankton by Cowie and
458 Hegdes (1984). The dominant glucose and fructose in the Mangshan aerosol samples collected in
459 autumn are rational as leaf senescence and decay results in both primary sugars being released into
460 the atmosphere during the fall season. We found an excellent correlation between glucose and
461 fructose ($r = 0.94$) in the Mangshan aerosols (Table 2), indicating the similar vegetation sources
462 for both sugar species in autumn (Baker et al., 1998; Burshtein et al., 2011; Pacini, 2000). Higher
463 concentrations of glucose and fructose in the aerosol samples collected during the autumn season
464 are reasonable because leaf senescence and decay result in an increased emission of primary
465 sugars into the atmosphere.

466 Several studies have reported that plant species significantly contribute to mannitol in the
467 atmosphere (Burshtein et al., 2011; Devis et al., 1988; 1990). Miyazaki et al. (2014) also found a
468 significant amount of trehalose, mannitol, and arabitol in the aerosol samples collected from the
469 forest and concluded their origin from the terrestrial plants within the forest. Significant positive
470 linear correlations of mannitol with fructose in daytime ($r = 0.79$) and nighttime ($r = 0.86$) further
471 denote that abundance of mannitol is due to the decay of plant leaves in autumn (Table 2).
472 Therefore, we conclude that the contributions of mannitol is from both vegetation and fungal
473 spores in the Mangshan aerosol samples. Hence mannitol showed the presence in factor 1.
474 Vegetations contribute to SCs during the campaign. Therefore, factor 1 can be termed as a
475 vegetation factor due to the high abundances of glucose, fructose, and mannitol.

476 Factor 2 is dominated by high loading of trehalose (80.2%), followed by mannitol (29.7%),
477 glucose (19.8%), and arabitol (18.2%) (Fig. 7b). The contribution of trehalose to soil dust has been
478 reported in several studies from different locations around the world, suggesting trehalose as a
479 tracer for the surface soil (Jia et al., 2010; Medeiros et al., 2006). In addition, previous studies
480 reported that bacteria and other microbes in the soil are also an essential source of trehalose
481 (Rogge et al., 2007). Trehalose is significantly correlated with arabitol ($r=0.58$) and mannitol
482 ($r=0.58$), and Ca^{2+} ($r=0.70$), demonstrating its microbial and soil dust origin. Therefore, factor 2
483 can be termed as microbial and soil dust factor.

484 Factor 3 is characterized by levoglucosan (82.2%), galactosan (77%), and mannosan
485 (73.6%) (Fig. 7c). Previous studies have reported that these SCs are associated with BB aerosols
486 (Fraser and Lakshmanan, 2000; Graham et al., 2002; Simoneit, 2002). Simoneit et al. (1999)
487 reported that the pyrolysis of cellulose and hemicellulose emitted levoglucosan, galactosan and
488 mannosan. These sugar species are major organic components emits in the atmosphere by BB
489 activities (Simoneit et al., 2002). The BB influenced aerosols are enriched with levoglucosan,
490 mannosan, and galactosan (Nolte et al., 2001; Medeiros et al., 2006). The domestic BB for
491 cooking and house heating due to low ambient temperature and field burning of agricultural
492 residues occur in East Asia (Verma et al., 2015). The PMF results are very well supported by the
493 fact that anhydrosugars are associated with BB (Simoneit et al., 1999). Therefore, factor 3 can be
494 termed as a BB factor due to the high abundance of BB products.

495 Factor 4 is dominated by high loading of sucrose (90%), followed by inositol (36.9%) and
496 fructose (11.7%) (Fig. 7d). Sucrose plays a crucial role in the plant blossoming process as the
497 dominant sugar compound of pollen grains (Pacini, 2000). Several studies also reported that
498 sucrose is abundant sugar species found in airborne pollen grains and flowering plants (Fu et al.,
499 2012; Graham et al., 2003; Medeiros et al., 2006; Pacini, 2000). Therefore, sucrose is reported as
500 an excellent tracer for airborne pollen spores (Pacini, 2000). Thus factor 4 is termed as pollen
501 factor due to the high loading of sucrose.

502 Factor 5 is characterized by a higher contribution of arabitol (61.5%) followed by mannitol
503 (39.3%) and inositol (15.3%) (Fig. 7e). Sugar species contributing to factor 5 are associated with
504 fungal spores (Bauer et al., 2008). Various fungi and microbes emit spores, which are tracers for
505 the arabitol and mannitol; therefore, both sugars are considered as specific tracers of fungal
506 activities (Medeiros et al., 2006; Rogge et al., 2007). Thus, factor 5 is termed as a fungal factor
507 due to the high loading of arabitol and mannitol. Overall, the average contributions of each factor
508 to measured SCs were estimated by PMF analyses (Fig. 8), in which BB was found to account for

509 27% of measured SCs. The vegetation and microbial and soil dust sources equally contribute
510 (21%) to total SCs. The fungal spores and pollen spores contribute 16% and 15% of total SCs,
511 respectively. Finally, biomass burning emissions from the local areas and megacities via long-
512 range atmospheric transport were identified as an important source for the Mangshan aerosols.

513 3.3. Contributions of sugar compounds to WSOC and OC

514 The contribution of carbon content of measured SCs varied from 14.1-371 ng m⁻³ (av.
515 145 ng m⁻³) in daytime and 12.8-322 ng m⁻³ (av. 117 ng m⁻³) in nighttime, accounting for 0.83%
516 and 0.91% of OC, respectively (Fig. 9a, b). The mean carbon contents of anhydrosugars showed
517 clear diurnal variation with higher nighttime values (67.1 ng m⁻³) than daytime (42.7 ng m⁻³),
518 accounting for 0.43 % and 0.22 % of OC, respectively. These results suggest that BB significantly
519 contributed to Mangshan aerosols. However, the carbon contents of primary sugars showed
520 opposite diurnal variations; higher (68.5 ng m⁻³) in daytime than nighttime (28.3 ng m⁻³),
521 accounting for 0.41 % and 0.28 % of OC, respectively (Fig. 9a, b). This study suggests that the
522 daytime emissions of primary sugars from local vegetation and the decay of plant leaf in forest
523 significantly contribute to OC. The carbon concentration contributed by sugar alcohols showed
524 insignificant diurnal variations i.e. 34.6 ng m⁻³ in daytime and 21.3 ng m⁻³ in nighttime,
525 accounting for 0.20 % and 0.19 % of OC, respectively. This result indicates multiple carbon
526 sources of sugar alcohols in day and night. In addition, contributions of anhydrosugars, primary
527 sugars, and sugar alcohols to WSOC were similar to those of OC in Mangshan aerosols.

528 Based on the PMF analysis, we found five sources for SCs measured in Mangshan
529 aerosols. The different tracer compounds were used to calculate carbon contents: biomass burning-
530 C (i.e., levoglucosan, galactosan, mannosan), vegetation-C (glucose, fructose), fungal-C (arabitol,
531 mannitol), pollen-C (sucrose), and microbial-soil-C (trehalose) (Fig. 9c, d). Among the five
532 sources, biomass burning-C was found as the largest carbon contributor to Mangshan aerosols
533 (36.7%), followed by fungal-C (23.7%), vegetation-C (19.7%), pollen-C (14.2%), and microbial-
534 soil-C (4.84%). Biomass burning-C accounted for 1.38% and 0.43% at night, while 0.57% and
535 0.22% in daytime for WSOC and OC, respectively. The BB for cooking and space heating in
536 winter and autumn seasons are common in central China (Akagi et al., 2011), which should
537 increase the nighttime levels of Biomass burning-C at the Mangshan site. However, the carbon
538 contribution by vegetation and fungal sources are similar during day and nighttime for the
539 Mangshan aerosols. Pollen-C accounted for 0.20% and 0.07% of OC in daytime and nighttime,
540 respectively. Higher pollen activities are key sources for the high daytime levels of pollen-C in the
541 forest site (Taylor et al., 2002).

542 3.4. Contribution of levoglucosan to OC and WSOC

543 We calculated the mass concentration ratios of levoglucosan to OC (Lev/OC) and WSOC
544 (Lev/WSOC) to evaluate the contributions of BB and anthropogenic emissions to Mangshan
545 aerosols (Fig. 9a-c). Fossil fuel combustion and BB emit WSOC and OC. They are also
546 secondarily produced by photochemical oxidation of volatile organic compounds in the
547 atmosphere (Wang et al., 2005a; Deshmukh et al., 2019b). Coal combustion and vehicle exhaust
548 can contribute to the high levels of OC and WSOC in aerosols (Xu et al., 2020). Levoglucosan, a
549 dominant constituent of BB products, has been considered as an excellent tracer of BB (Simoneit,
550 2002; Kuo et al., 2011).

551 Average Lev/OC ratio (5.69×10^{-3}) was lower than that of Lev/WSOC (1.66×10^{-2}) in
552 Mangshan samples (Fig. 10a). Yan et al. (2018) reported similar ratios of Lev/OC (4.0×10^{-3}) and
553 Lev/WSOC (1.6×10^{-2}) for coal combustion, suggesting a significant carbon contribution to
554 Mangshan aerosols from coal combustions in the industrial areas via long range transport.
555 Interestingly, we found a substantial diurnal variation of Lev/OC and Lev/WSOC ratios. The
556 average Lev/OC and Lev/WSOC ratios are several times higher in nighttime (8.48×10^{-3} and
557 2.70×10^{-2} , respectively) than daytime (4.21×10^{-3} and 1.11×10^{-2} , respectively) (Fig. 10b, c). These
558 results indicate that BB contributed substantially to the Mangshan organic aerosols in nighttime.
559 Moreover, the correlations of levoglucosan with OC and WSOC are stronger in nighttime ($r =$
560 0.81 and 0.70 , respectively) than daytime ($r = 0.45$ and 0.40 , respectively), demonstrating the
561 dominance of BB-derived aerosols in the nighttime Mangshan samples (Table 2).

562 In addition, WSOCs are derived from various emission sources. We propose that
563 secondary organic aerosols constitute a significant fraction of WSOC and OC in daytime
564 Mangshan aerosols. The photochemical oxidation of organic precursors emitted from fossil fuel
565 combustion in industries and vehicular exhausts also contributes to secondary production of
566 WSOC and OC in daytime (He et al., 2015), suggesting that emissions from the urban Beijing area
567 may significantly influence the daytime levels of Mangshan aerosols. He et al. (2015) proposed a
568 possible contribution of photochemical formation of secondary organic aerosols to atmospheric
569 WSOC and OC in north China. Nevertheless, the photochemical degradation of levoglucosan by
570 OH radicals under ultraviolet radiations and high temperatures (Hennigan et al., 2010) may play a
571 key role in lowering the ratios of Lev/OC and Lev/WSOC in daytime Mangshan aerosols.

572 3.5. Mass concentration ratios of levoglucosan/mannosan

573 The mass concentration ratios of levoglucosan and mannosan (Lev/Man) were calculated
574 to better characterize the emissions sources of BB tracers (softwood vs. hardwood) in the

575 Mangshan site. Figure 10d represents the variations of Lev/Man ratios for overall, day- and
576 nighttime periods. The Lev/Man ratios have been used to distinguish the hardwood (angiosperm)
577 and softwood (gymnosperm) burning in the ice core record from the Russian Far East (Kawamura
578 et al., 2012). Hardwood contains 55–65% cellulose and 20–30% hemicellulose (Klemm et al.,
579 2005). Levoglucosan and mannosan are derived from the thermal decomposition of cellulose and
580 hemicelluloses, respectively (Simoneit, 2002). Levoglucosan is thermally more stable than
581 mannosan and galactosan (Kuo et al., 2011). Hence, a lower Lev/Man ratio is associated with
582 softwood burning, whereas a higher ratio is associated with hardwood burning (Engling et al.,
583 2006, 2009). However, we found insignificant diurnal variations of Lev/Man ratios between night
584 (9.33-25.9, avg. 15.8) and daytime aerosols (0.90-23.3, avg. 13.6). Likewise, comparable Lev/Man
585 ratios (9-13 for PM₁₀ and 10-13 for PM_{2.5}) were reported for aerosol samples from Tanzania,
586 where wood and charcoal are primary fuels used for domestic cooking and heating (Mkoma and
587 Kawamura, 2013). Interestingly, wheat straws and lignite are used in China for domestic cooking
588 and house heating, which may also contribute to levoglucosan and mannosan in the Mangshan
589 aerosols.

590 Different Lev/Man ratios were reported in the chamber and controlled field experiments,
591 e.g., 4-22 for conifer and savanna grass burning (Iinuma et al., 2007), and 41.6 for rice straw and
592 and 55.7 cereal straw burning (Engling et al., 2009; Zhang et al., 2007). Kuo et al. (2011) reported
593 higher emissions of levoglucosan during high-temperature flaming (27.5-52.3) compared to low-
594 temperature smoldering (2.43-3.08). Hence, it is not easy to differentiate hardwood and softwood
595 burning based on Lev/Man ratios alone. Several studies reported a high Lev/Man ratio for both
596 softwood and hardwood burning. Thus, there may exist some other factors that significantly
597 control the Lev/Man ratios. Yan et al. (2018) found a significant contribution of levoglucosan in
598 coal combustion with Lev/Man ratio of 7.2. The variations of Lev/Man ratios in Mangshan may be
599 significantly influenced by several factors, i.e., flaming vs. smoldering, duration of biomass
600 burning, coal combustion, and hardwood vs. softwood burning. The moderate Lev/Man ratios in
601 autumn aerosols from Mangshan suggest that low temperature smoldering processes of hardwood
602 contribute to levels of levoglucosan and mannosan. However, the contribution of coal
603 combustions for house heating could not be excluded.

604 **4. Summary and Conclusions**

605 Anhydrosugars, primary sugars, and sugar alcohols were detected with distinct diurnal
606 variations in suburban aerosol samples collected at the Mangshan site in the northern vicinity of
607 Beijing. The wind patterns indicate that daytime air masses were transported from urban Beijing to

608 Mangshan, while clean air masses were delivered in nighttime from the Mangshan National Forest
609 Park. Daytime air masses from urban Beijing significantly influence the air quality of the northern
610 forest region. We observed the highest abundance of primary sugars, followed by anhydrosugars
611 and sugar alcohols. Local emissions from the forest plants and fungal species are the main
612 contributors to the primary sugars and sugar alcohols in the Mangshan aerosols. The
613 meteorological parameter significantly influenced the levels of SCs in the Mangshan samples. We
614 observed a significant influence of enhanced ambient temperature and solar radiation on the pollen
615 rupture and increased RH on fungal and microbial growth. This study suggested the source
616 variation for trehalose, i.e., local microbes at night and soil dust particles transported from Beijing
617 areas by southerly wind in daytime. We found that vegetation and fungal spores are not a specific
618 source of glucose and mannitol, respectively. Both sugars may have multiple sources in the forest
619 aerosols.

620 PMF results concluded the contributions of 36% from vegetation (21% vegetation factor
621 and 15% pollen factor) and 37% from microbial and fungal species (21% microbial soil dust and
622 16% fungal factor) of total measured SCs. The BB activities for domestic cooking and space
623 heating in north China contributed higher organic carbon at nighttime (0.43%) than in daytime
624 (0.22%). Therefore, local BB seriously affected the air quality of the Mangshan site. Lev/Man
625 ratio suggested that low temperature smoldering burning of hardwood is the main source for BB
626 aerosols. SCs were recognized as a significant aerosol component at Mangshan, northern suburbs
627 of Beijing. SCs can influence the air quality and thus climate because they are essential
628 components of organic aerosols on a global scale. This study of SCs at Mangshan demonstrates
629 that ambient levels of SCs are highly sensitive to the emissions of anthropogenic and biogenic
630 aerosols. Higher contribution of levoglucosan to SCs demonstrated a significant BB activity
631 around the Mangshan site in north China.

632

633 ***Data availability.*** Raw data are available on request by contacting the corresponding author.

634 ***Author contributions.*** This research was designed YK, KK and ZW. Laboratory measurements
635 were performed by FY with a support of PF. The paper was prepared by SKV and KK.

636 ***Competing interests.*** The authors declare that they have no conflict of interest.

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References:

- Akagi, S. K., Yokelson, R. J., Wiedinmyer, C., Alvarado, M. J., Reid, J. S., Karl, T., Crounse, J. D., and Wennberg, P. O.: Emission factors for open and domestic biomass burning for use in atmospheric models, *Atmos. Chem. Phys.*, 11, 4039–4072, <https://doi.org/10.5194/acp-11-4039-2011>.
- Baker, H. G., Baker, I., and Hodges, S. A.: Sugar composition of nectars and fruits consumed by birds and bats in the tropics and subtropics, *Biotropica*, 30, 559–586, <https://doi.org/10.1111/j.1744-7429.1998.tb00097.x>, 1998.
- Bauer, H., Schueller, E., Weinke, G., Berger, A., Hitzenberger, R., Marr, I. L., and Puxbaum, H.: Significant contributions of fungal spores to the organic carbon and to the aerosol mass balance of the urban atmospheric aerosol, *Atmos. Environ.*, 42, 5542–5549, <https://doi.org/10.1016/j.atmosenv.2008.03.019>, 2008.
- Bieleski, R. L.: Onset of phloem export from senescent petals of Daylily, *Plant Physiol.*, 109, 557–565, <https://doi.org/10.1104/pp.109.2.557>, 1995.
- Brown, J. K. M., and Hovmoller, M. S.: Epidemiology – Aerial dispersal of pathogens on the global and continental scales and its impact on plant disease, *Science*, 297, 537–541, <https://doi.org/10.1126/science.1072678>, 2002.
- Burshtein, N., Lang-Yona, N., and Rudich, Y.: Ergosterol, arabitol and mannitol as tracers for biogenic aerosols in the eastern Mediterranean, *Atmos. Chem. Phys.*, 11, 829–839, <https://doi.org/10.5194/acp-11-829-2011>, 2011.
- Cao, C., Jiang, W., Wang, B., Fang, J., Lang, J., Tian, G., Jiang, J. K., and Zhu, T. F.: Inhalable microorganisms in Beijing's PM_{2.5} and PM₁₀ pollutants during a severe smog event, *Environ. Sci. Technol.*, 48, 1499–1507, <https://doi.org/10.1021/es4048472>, 2014.
- Carvalho, A., Pio, C., and Santos, C.: Water-soluble hydroxylated organic compounds in German and Finnish aerosols, *Atmos. Environ.*, 37, 1775–1783, [https://doi.org/10.1016/S1352-2310\(03\)00066-9](https://doi.org/10.1016/S1352-2310(03)00066-9), 2003.
- Claeys, M., Graham, B., Vas, G., Wang, W., Vermeylen, R., Pashynska, V., Cafmeyer, J., Guyon, P., Andreae, M. O., Artaxo, P., and Maenhaut, W.: Formation of secondary organic aerosols through photooxidation of isoprene, *Science*, 303, 1173–1176, <https://doi.org/10.1126/science.1092805>, 2004.
- Cooke, W.F., Liousse, C., Cachier, H., and Feichter, J.: Construction of a 1° X 1° fossil fuel emission data set for carbonaceous aerosol and implementation and radiative impact in the ECHAM4 model, *J. Geophys. Res.-Atmos.*, 104, 22137–22162, <https://doi.org/10.1029/1999JD900187>, 1999.
- Cowie, G. L., and Hedges, J. I.: Carbohydrate sources in a coastal marine-environment, *Geochim. Cosmochim. Acta*, 48, 2075–2087, [https://doi.org/10.1016/0016-7037\(84\)90388-0](https://doi.org/10.1016/0016-7037(84)90388-0), 1984.
- Dahlman, L., Persson, J., Nasholm, T. and Palmqvist, K.: Carbon and nitrogen distribution in the green algal lichens *Hypogymnia physodes* and *Platismatia glauca* in relation to nutrient supply, *Planta*, 217, 41–48, <https://doi.org/10.1007/s00425-003-0977-8>, 2003.
- Davis, J. M., and Loescher, W. H.: [¹⁴C]-Assimilate translocation in the light and dark in celery (*Apium graveolens*) leaves of different ages, *Physiol. Plant*, 79, 656–662, <https://doi.org/10.1111/j.1399-3054.1990.tb00040.x>, 1990.
- Davis, J. M., Fellman, J. K., and Loescher, W. H.: Biosynthesis of sucrose and mannitol as a function of leaf age in celery (*Apium graveolens* L.), *Plant Physiol.*, 186, 129–133, <https://doi.org/10.1104/pp.86.1.129>, 1988.
- Deshmukh, D.K., Haque, M.M., Kim, Y., and Kawamura, K.: Organic tracers of fine aerosol particles in central Alaska: summertime composition and sources, *Atmos. Chem. Phys.*, 19, 14009–14029, <https://doi.org/10.5194/acp-19-14009-2019>, 2019a.
- Deshmukh, D.K., Kawamura, K., Gupta, T., Haque, M.M., Zhang, Y.L., Singh, D.K., and Tsai,

- 695 Y.I.: High loadings of water-soluble oxalic acid and related compounds in PM_{2.5} aerosols in
696 Eastern Central India: influence of biomass burning and photochemical processing, *Aerosol*
697 *Air Qual. Res.*, 19, 2625-2644, <https://doi.org/10.4209/aaqr.2019.10.0543>, 2019b.
- 698 Després, V. R., Huffman, J. A., Burrows, S. M., Hoose, C., Safatov, A. S., Buryak, G., Fröhlich-
699 Nowoisky, J., Elbert, W., Andreae, M. O., Pöschl, U., and Jaenicke, R.: Primary biological
700 aerosol particles in the atmosphere: a review, *Tellus B*, 64, 15598, <https://doi.org/10.3402/tellusb.v64i0.15598>, 2012.
- 702 Elbert, W., Taylor, P. E., Andreae, M. O., and Pöschl, U.: Contribution of fungi to primary
703 biogenic aerosols in the atmosphere: wet and dry discharged spores, carbohydrates, and
704 inorganic ions, *Atmos. Chem. Phys.*, 7, 4569-4588, <https://doi.org/10.5194/acp-7-4569-2007>,
705 2007.
- 706 Engling, G., Carrico, C. M., Kreidenweis, S. M., Collett, J. L., Day, D. E., Malm, W. C., Lincoln,
707 E., Hao, W. M., Iinuma, Y., and Herrmann, H.: Determination of levoglucosan in biomass
708 combustion aerosol by high-performance anion-exchange chromatography with pulsed
709 amperometric detection, *Atmos. Environ.*, 40, S299-S311,
710 <https://doi.org/10.1016/j.atmosenv.2005.12.069>, 2006.
- 711 Engling, G., Lee, J. J., Tsai, Y. W., Lung, S. C. C., Chou, C. C. K., and Chan, C. Y.: Size-resolved
712 anhydrosugar composition in smoke aerosol from controlled field burning of rice straw,
713 *Aerosol Sci. Tech.*, 43, 662-672, <https://doi.org/10.1080/02786820902825113>, 2009.
- 714 Filippo, P. D., Pomata, D., Riccardi, C., Buiarelli, F., and Perrino, C.: Fungal contribution to size-
715 segregated aerosol measured through biomarkers, *Atmos. Environ.*, 64, 132-140,
716 <https://doi.org/10.1016/j.atmosenv.2012.10.010>, 2013.
- 717 Fraser, M. P., and Lakshmanan, K.: Using levoglucosan as a molecular marker for the long-range
718 transport of biomass combustion aerosols, *Environ. Sci. Technol.*, 34(21), 4560-4564,
719 <https://doi.org/10.1021/es991229l>, 2000.
- 720 Fu, P. Q., Kawamura, K., Kobayashi, M., and Simoneit, B. R. T.: Seasonal variations of sugars in
721 atmospheric particulate matter from Gosan, Jeju Island: Significant contributions of airborne
722 pollen and Asian dust in spring, *Atmos. Environ.*, 55, 234-239,
723 <https://doi.org/10.1016/j.atmosenv.2012.02.061>, 2012.
- 724 Fu, P. Q., Kawamura, K., Okuzawa, K., Aggarwal, S. G., Wang, G. H., Kanaya, Y., and Wang, Z.
725 F.: Organic molecular compositions and temporal variations of summertime mountain
726 aerosols over Mt. Tai, North China Plain, *J. Geophys. Res.-Atmos.*, 113D19107,
727 <http://doi:10.1029/2008jd009900>, 2008.
- 728 Fu, P. Q., Kawamura, K., Pavuluri, C. M., Swaminathan, T., and Chen, J.: Molecular
729 characterization of urban organic aerosol in tropical India: Contributions of primary emissions
730 and secondary photooxidation, *Atmos. Chem. Phys.*, 10, 2663-2689, <http://doi:10.5194/acp-10-2663-2010>, 2010.
- 732 Fuzzi, S., Decesari, S., Facchini, M. C., Cavalli, F., Emblico, L., Mircea, M., Andreae, M. O.,
733 Trebs, I., Hoffer, A., Guyon, P., Artaxo, P., Rizzo, L. V., Lara, L. L., Pauliquevis, T., Maen-
734 haut, W., Raes, N., Chi, X. G., Mayol-Bracero, O. L., Soto-Garcia, L. L., Claeys, M.,
735 Kourtchev, I., Rissler, J., Swietlicki, E., Tagliavini, E., Schkolnik, G., Falkovich, A. H.,
736 Rudich, Y., Fisch, G., and Gatti, L. V.: Overview of the inorganic and organic composition of
737 size-segregated aerosol in Rondonia, Brazil, from the biomass-burning period to the onset of
738 the wet season, *J. Geophys. Res.-Atmos.*, 112, D01201,
739 <https://doi.org/10.1029/2005jd006741>, 2007.
- 740 Graham, B., Guyon, P., Taylor, P. E., Artaxo, P., Maenhaut, W., Glovsky, M. M., Flagan, R. C.,
741 and Andreae, M. O.: Organic compounds present in the natural Amazonian aerosol:
742 Characterization by gas chromatography-mass spectrometry, *J. Geophys. Res.-Atmos.*,
743 108(D24), <http://doi:10.1029/2003jd003990>, 2003.

- 744 Graham, B., Mayol-Bracero, O. L., Guyon, P., Roberts, G. C., Decesari, S., Facchini, M. C.,
745 Artaxo, P., Maenhaut, W., Koll, P., and Andreae, M. O.: Water-soluble organic compounds in
746 biomass burning aerosols over Amazonia 1. Characterization by NMR and GC-MS, *J.*
747 *Geophys. Res.*, 107(D20), 8047, <https://doi.org/10.1029/2001JD000336>, 2002.
- 748 Hackl, E., Bachmann, G., and Zechmeister-Boltenstern, S.: Soil microbial biomass and
749 rhizosphere effects in natural forest stands, *Phyton.-Ann. Rei. Bot. A*, 40, 83-90, 2000.
- 750 Han, F., Kota, S. H., Wang, Y., & Zhang, H.: Source apportionment of PM_{2.5} in Baton Rouge,
751 Louisiana during 2009–2014. *Science of the Total Environment*, 586, 115–126.
752 <https://doi.org/10.1016/j.scitotenv.2017.01.189>, 2017.
- 753 He, N., Kawamura, K., Kanaya, Y., and Wang, Z. F.: Diurnal variations of carbonaceous
754 components, major ions, and stable carbon and nitrogen isotope ratios in suburban aerosols
755 from northern vicinity of Beijing, *Atmos. Environ.*, 123, 18-24,
756 <https://doi.org/10.1016/j.atmosenv.2015.10.052>, 2015.
- 757 He, N., Kawamura, K., Kanaya, Y., and Wang, Z.: Diurnal variations of carbonaceous
758 components, major ions, and stable carbon and nitrogen isotope ratios in suburban aerosols
759 from northern vicinity of Beijing, *Atmos. Environ.*, 123, 18-24.
760 <http://dx.doi.org/10.1016/j.atmosenv.2015.10.052>, 2014.
- 761 Heald, C. L., and Spracklen, D. V.: Atmospheric budget of primary biological aerosol particles
762 from fungal spores, *Geophys. Res. Lett.*, 36, L09806, <https://doi.org/10.1029/2009GL037493>,
763 2009.
- 764 Hennigan, C. J., Sullivan, A. P., Collett Jr., J. L., and Robinson, A. L.: Levoglucosan stability in
765 biomass burning particles exposed to hydroxyl radicals, *Geophys. Res. Lett.*, 37, L09806,
766 <https://doi.org/10.1029/2010gl043088>, 2010.
- 767 Ibrahim, M., Rabah, A. B., Liman, B., and Ibrahim, N. T.: Effect of temperature and relative
768 humidity on the growth of *Helminthosporium fulvum*, *Nigerian J. Basic Appli. Sci.*, 19, 127–
769 129, <https://doi.org/10.4314/njbas.v19i1.69357>, 2011.
- 770 Iinuma, Y., Brüggemann, E., Gnauk, T., Müller, K., Andreae, M. O., Helas, G., Parmar, R., and
771 Herrmann, H.: Source characterization of biomass burning particles: The combustion of
772 selected European conifers, African hardwood, savanna grass, and German and Indonesian
773 peat, *J. Geophys. Res.-Atmos.*, 112(D8), <http://doi:10.1029/2006jd007120>, 2007.
- 774 Jaenicke, R.: Abundance of cellular material and proteins in the atmosphere, *Science*, 308, 73,
775 <http://doi:10.1126/science.1106335>, 2005.
- 776 Jia, Y. L., Clements, A. L., and Fraser, M. P.: Saccharide composition in atmospheric particulate
777 matter in the southwest US and estimates of source contributions, *J. Aerosol Sci.*, 41, 62-73,
778 <https://doi.org/10.1016/j.jaerosci.2009.08.005>, 2010.
- 779 Jones, E. B. G., and Mitchell, J. I.: Biodiversity of marine fungi, in *biodiversity: international*
780 *biodiversity seminar*, edited by: Cimerman, A. and Gunde-Cimerman, N., 31–42, Ljubljana:
781 National Inst. Chemistry and Slovenia National Commission for UNESCO, 1996.
- 782 Kanakidou, M., Seinfeld, J. H., Pandis, S. N., Barnes, I., Dentener, F. J., Facchini, M. C., Van
783 Dingenen, R., Ervens, B., Nenes, A., Nielsen, C. J., Swietlicki, E., Putaud, J. P., Balkanski,
784 Y., Fuzzi, S., Horth, J., Moortgat, G. K., Winterhalter, R., Myhre, C. E. L., Tsigaridis, K.,
785 Vignati, E., Stephanou, E. G., and Wilson, J.: Organic aerosol and global climate modelling: a
786 review, *Atmos. Chem. Phys.*, 5, 1053–1123, <https://doi.org/10.5194/acp-5-1053-2005>, 2005.
- 787 Kanaya Y., Akimoto H., Wang Z.-F., Pochanart P., Kawamura K., Y. Liu, J. Li, Y. Komazaki, H.
788 Irie, X.-L. Pan, F. Taketani, K. Yamaji, H. Tanimoto, S. Inomata, S. Kato, J. Suthawaree, K.
789 Okuzawa, G. Wang, S. G. Aggarwal, P. Q. Fu, Y. Wang, and Zhuang G., Overview of the Mt.
790 Tai Experiments (MTX2006) in central East China in June 2006: studies of significant
791 regional air pollution, *Atmos. Chem. Phys.*, 13, 8265–8283, <https://doi.org/10.5194/acp-13-8265-2013>, 2013.

- 793 Kang, M., Ren, L., Ren, H., Zhao, Y., Kawamura, K., Zhang, H., Wei, L., Sun, Y., Wang, Z., Fu,
794 P. Q.: Primary biogenic and anthropogenic sources of organic aerosols in Beijing, China:
795 Insights from saccharides and n-alkanes, *Environ. Pollut.*, 243, 1579-1587,
796 <https://doi.org/10.1016/j.envpol.2018.09.118>, 2018.
- 797 Kawamura, K., Izawa, Y., Mochida, M., and Shiraiwa, T.: Ice core records of biomass burning
798 tracers (levoglucosan and dehydroabietic, vanillic and p-hydroxybenzoic acids) and total
799 organic carbon for past 300 years in the Kamchatka Peninsula, Northeast Asia, *Geochim.*
800 *Cosmochim. Acta*, 99, 317-329, <https://doi.org/10.1016/j.gca.2012.08.006>, 2012.
- 801 Kawamura, K., Tachibana, E., Okuzawa, K., Aggarwal, S. G., Kanaya, Y., Wang, Z. F.: High
802 abundances of water-soluble dicarboxylic acids, ketocarboxylic acids and α -dicarbonyls in the
803 mountaintop aerosols over the north China plain during wheat burning season, *Atmos. Chem.*
804 *Phys.* 13 (16), 8285-8302, <https://doi.org/10.5194/acp-13-8285-2013>, 2013.
- 805 Keller, F., and Matile, P.: Storage of sugars and mannitol in petioles of celery leaves, *New Phytol.*
806 113:291–299, <https://doi.org/10.1111/j.1469-8137.1989.tb02406.x>, 1989.
- 807 Kim, V. K. and Xiao, C. L.: Influence of culture media and environmental factors on mycelial
808 growth and picnidial production of *Sphaeropsis pyriputrescens*, *Mycologia*, 97, 25–32,
809 <https://doi.org/10.3852/mycologia.97.1.25>, 2005.
- 810 Klemm, D., Heublein, B., Fink, H. P., and Bohn, A.: Cellulose: Fascinating biopolymer and
811 sustainable raw material, *Angew. Chem. Int. Ed.*, 44, 3358-3393,
812 <https://doi.org/10.1002/anie.200460587>, 2005.
- 813 Kuo, L. J., Louchouart, P., and Herbert, B. E.: Influence of combustion conditions on yields of
814 solvent-extractable anhydrosugars and lignin phenols in chars: Implications for
815 characterizations of biomass combustion residues, *Chemosphere*, 85, 797-805,
816 <https://doi.org/10.1016/j.chemosphere.2011.06.074>, 2011.
- 817 Lelieveld, J., Evans, J.S., Fnais, M., Giannadaki, D., Pozzer, A.: The contribution of outdoor air
818 pollution sources to premature mortality on a global scale, *Nature*, 525 (7569), 367-371.
819 <https://doi.org/10.1038/nature15371>, 2015.
- 820 Lewis, D. and Smith, D.: Sugar alcohols (polyols) in fungi and green plants, 1. Distribution,
821 physiology and metabolism, *New Phytol.*, 66, 143–184, <https://doi.org/10.2307/2430328>, 1967.
- 822 Li, W. J., Shao, L. Y., and Buseck, P. R.: Haze types in Beijing and the influence of agricultural
823 biomass burning, *Atmos. Chem. Phys.* 10 (17), 8119-8130, [https://doi.org/10.5194/acp-10-](https://doi.org/10.5194/acp-10-8119-2010)
824 [8119-2010](https://doi.org/10.5194/acp-10-8119-2010), 2010.
- 825 Lin, J., Pan, D., Davis, S. J., Zhang, Q., He, K., Wang, C., Streets, D. G., Wuebbles, D. J., and
826 Guan, D.: China's international trade and air pollution in the United States, *Proc. Natl. Acad.*
827 *Sci. U.S.A.*, 111, 1736–1741, <https://doi.org/10.1073/pnas.1312860111>, 2014.
- 828 Loescher, W. H., Tyson, R. H., Everard, J. D., Redgwell, R. J., and Bieleski, R. L.: Mannitol
829 synthesis in higher plants: evidence for the role and characterization of a NADPH-Dependent
830 mannose 6-Phosphate reductase, *Plant Physiol.*, 98(4), 1396–1402, [https://doi:](https://doi.org/10.1104/pp.98.4.1396)
831 [10.1104/pp.98.4.1396](https://doi.org/10.1104/pp.98.4.1396), 1992.
- 832 Malik, V. K. and Singh, S.: Effect of temperature and relative humidity on teliospore germination
833 in *Ustilago hordei*, *J. Mycol. Plant Pathol.*, 34, 410–411,
834 <https://doi.org/10.4314/njbas.v19i1.69357>, 2004.
- 835 Manninen, H. E., Bäck, J., Sihto Nissilä, S. L., Huffman, J. A., Pessi, A. M., Hiltunen, V., Aalto,
836 P. P., Hidalgo Fernández, P. J., Hari, P., and Saarto, A.: Patterns in airborne pollen and other
837 primary biological aerosol particles (PBAP), and their contribution to aerosol mass and
838 number in a boreal forest, *Boreal Environ. Res.*, 19, 383–405,
839 <https://hdl.handle.net/10138/165208>, 2014.

- 840 Medeiros, P. M., Conte, M. H., Weber, J. C., and Simoneit, B. R. T.: Sugars as source indicators
841 of biogenic organic carbon in aerosols collected above the Howland Experimental Forest,
842 Maine, *Atmos. Environ.*, 40, 1694–1705, <https://doi.org/10.1016/j.atmosenv.2005.11.001>,
843 2006.
- 844 Miyazaki, Y., Fu, P. Q., Ono, K., Tachibana, E., and Kawamura, K.: Seasonal cycles of water-
845 soluble organic nitrogen aerosols in a deciduous broadleaf forest in northern Japan, *J. Geophys.*
846 *Res. Atmos.*, 119, 1440–1454, <https://doi.org/10.1002/2013JD020713>, 2014.
- 847 Miyazaki, Y., Jung, J., Fu, P. Q., Mizoguchi, Y., Yamanoi, K., and Kawamura, K.: Evidence of
848 formation of submicrometer water-soluble organic aerosols at a deciduous forest site in
849 northern Japan in summer, *J. Geophys. Res.*, 117, D19213,
850 <https://doi.org/10.1029/2012JD018250>, 2012.
- 851 Mkoma, S. L., and Kawamura, K.: Molecular composition of dicarboxylic acids, ketocarboxylic
852 acids, a-dicarbonyls and fatty acids in atmospheric aerosols from Tanzania, East Africa during
853 wet and dry seasons, *Atmos. Chem. Phys.*, 13, 2235–2251, <https://doi.org/10.5194/acp-13-2235-2013>, 2013.
- 855 Nolte, C. G., Schauer, J. J., Cass, G. R., and Simoneit, B. R. T.: Highly polar organic compounds
856 present in wood smoke and in the ambient atmosphere, *Environ. Sci. Technol.*, 35(10), 1912-
857 1919, <https://doi.org/10.1021/es001420r>, 2001.
- 858 Paatero, P., and Tapper, U.: Positive matrix factorization - a nonnegative factor model with
859 optimal utilization of error-estimates of data values, *Environmetrics*, 5, 111-126,
860 <https://doi.org/10.1002/env.317005020>, 1994.
- 861 Paatero, P., Hopke, P. K., Song, X. H., and Ramadan, Z.: Understanding and controlling rotations
862 in factor analytic models, *Chemom. Intell. Lab. Syst.*, 60, 253–264,
863 [https://doi.org/10.1016/S0169-7439\(01\)00200-3](https://doi.org/10.1016/S0169-7439(01)00200-3), 2002.
- 864 Pacini, E.: From anther and pollen ripening to pollen presentation, *Plant Syst. Evol.*, 222, 19-43,
865 <https://doi.org/10.1007/BF00984094>, 2000.
- 866 Pashynska, V., Vermeylen, R., Vas, G., Maenhaut, W., and Claeys, M.: Development of a gas
867 chromatographic/ion trap mass spectrometric method for the determination of levoglucosan and
868 saccharidic compounds in atmospheric aerosols. Application to urban aerosols, *J. Mass*
869 *Spectrom.*, 37, 1249–1257, <https://doi.org/10.1002/jms.391>, 2002.
- 870 Qiao, X., Ying, Q., Li, X., Zhang, H., Hu, J., Tang, Y., and Chen, X.: Source apportionment of
871 PM_{2.5} for 25 Chinese provincial capitals and municipalities using a source-oriented
872 community multiscale air quality model, *Sci. Total Environ.*, 612, 462-471,
873 <https://doi.org/10.1016/j.scitotenv.2017.08.272>, 2018.
- 874 Rogge, W. F., Medeiros, P. M., and Simoneit, B. R. T.: Organic marker compounds in surface
875 soils of crop fields from the San Joaquin Valley fugitive dust characterization study, *Atmos.*
876 *Environ.*, 41, 8183-8204, <https://doi.org/10.1016/j.atmosenv.2007.06.030>, 2007.
- 877 Rumpho, M. E., Edwards, G. E., and Loescher, W. H.: A pathway for photosynthetic carbon flow
878 to mannitol in celery leaves. Activity and localization of key enzymes, *Plant Physiol.*, 73:869–
879 873, <https://doi.org/10.1104/pp.73.4.869>, 1983.
- 880 Sharma, R. and Rajak, R. C.: Keratinophilic fungi: Nature's keratin degrading machines,
881 *Resonance*, 8, 28–30, <https://doi.org/10.1007/BF02837919>, 2003.
- 882 Simoneit, B. R. T., Kobayashi, M., Mochida, M., Kawamura, K., Lee, M., Lim, H. J., Turpin, B.
883 J., and Komazaki, Y.: Composition and major sources of organic compounds of aerosol
884 particulate matter sampled during the ACE-Asia campaign, *J. Geophys. Res.-Atmos.*, 109,
885 D19S10, <https://doi.org/10.1029/2004jd004598>, 2004.
- 886 Simoneit, B. R. T., Schauer, J. J., Nolte, C. G., Oros, D. R., Elias, V. O., Fraser, M. P., Rogge, W.
887 F., and Cass, G. R.: Levoglucosan, a tracer for cellulose in biomass burning and atmospheric

888 particles, *Atmos. Environ.*, 33, 173-182, [https://doi.org/10.1016/S1352-2310\(98\)00145-9](https://doi.org/10.1016/S1352-2310(98)00145-9),
889 1999.

890 Simoneit, B. R. T.: Biomass burning - A review of organic tracers for smoke from incomplete
891 combustion, *Appl. Geochem.*, 17, 129-162, [https://doi.org/10.1016/S0883-2927\(01\)00061-0](https://doi.org/10.1016/S0883-2927(01)00061-0),
892 2002.

893 Streets, D. G., Yarber, K. F., Woo, J. H., and Carmichael, G. R.: Biomass burning in Asia: Annual
894 and seasonal estimates and atmospheric emissions, *Global Biogeochem. Cycle*, 17, 1099,
895 <http://doi:10.1029/2003GB002040>, 2003.

896 Sullivan, A. P., Holden, A. S., Patterson, L. A., McMeeking, G. R., Kreidenweis, S. M., Malm,
897 W. C., Hao, W. M., Wold, C. E., and Collett Jr., J. L.: A method for smoke marker
898 measurements and its potential application for determining the contribution of biomass
899 burning from wildfires and prescribed fires to ambient PM_{2.5} organic carbon, *J. Geophys.
900 Res.*, 113, D22302, <http://doi:10.1029/2008JD010216>, 2008.

901 Sun, Y., Jiang, Q., Xu, Y., Ma, Y., Zhang, Y., Liu, X., Li, W., Wang, F., Li, J., Wang, P., Li, Z.:
902 Aerosol characterization over the north China plain: haze life cycle and biomass burning
903 impacts in summer, *J. Geophys. Res.* 121 (5), 2508-2521.
904 <https://doi.org/10.1002/2015JD024261>, 2016.

905 Tao, J., Zhang, L. M., Cao, J. J., and Zhang, R. J.: A review of current knowledge concerning
906 PM_{2.5} chemical composition, aerosol optical properties and their relationships across China,
907 *Atmos. Chem. Phys.*, 17, 9485-9518, <https://doi.org/10.5194/acp-17-9485-2017>, 2017.

908 Taylor, P. E., Flagan, R. C., Valenta, R., and Glovsky, M. M.: Release of allergens as respirable
909 aerosols: A link between grass pollen and asthma, *J. Allergy Clin. Immunol.*, 109, 51-56,
910 <https://doi.org/10.1067/mai.2002.120759>, 2002.

911 Verma, S. K., Kawamura, K., Chen, J., and Fu, P. Q.: Thirteen years of observations on primary
912 sugars and sugar alcohols over remote Chichijima Island in the western North Pacific, *Atmos.
913 Chem. Phys.*, 18, 81-101, <https://doi.org/10.5194/acp-18-81-2018>, 2018.

914 Verma, S. K., Kawamura, K., Chen, J., Fu, P. Q., and Zhu, C.: Thirteen years observation of
915 biomass-burning organic tracers over Chichijima Island in the western North Pacific: An
916 outflow region of Asian aerosols, *J. Geophys. Res. Atmos.*, 120,
917 <http://doi:10.1002/2014JD022224>, 2015.

918 Wan, X., Kang, S., Li, Q., Rupakheti, D., Zhang, Q., Guo, J., Chen, P., Tripathi, L., Rupakheti,
919 M., Panday, A. K., Wang, W., Kawamura, K., Gao, S., Wu, G., and Cong, Z.: Organic
920 molecular tracers in the atmospheric aerosols from Lumbini, Nepal, in the northern Indo-
921 Gangetic Plain: influence of biomass burning, *Atmos. Chem. Phys.*, 17, 8867-8885,
922 <https://doi.org/10.5194/acp-17-8867-2017>, 2017.

923 Wang, G., and Kawamura, K.: Molecular characteristics of urban organic aerosols from Nanjing: a
924 case study of a mega-city in China, *Environ. Sci. Technol.*, 39, 7430-7438,
925 <https://doi.org/10.1021/es051055+>, 2005a.

926 Wang, G., Kawamura, K., Zhao, X., Li, Q., Dai, Z., and Niu, H.: Identification, abundance, and
927 seasonal variation of anthropogenic organic aerosols from a mega-city in China, *Atmos.
928 Environ.*, 41, 407-416, <https://doi.org/10.1016/j.atmosenv.2006.07.033>, 2007.

929 Wang, H., Kawamura, K., and Shooter, D.: Carbonaceous and ionic components in wintertime
930 atmospheric aerosols from two New Zealand cities: implications for solid fuel combustion,
931 *Atmos. Environ.*, 39, 5865-5875, <https://doi.org/10.1016/j.atmosenv.2005.06.031>, 2005b.

932 Wei, L. F., Yue, S. Y., Zhao, W. Y., Yang, W. Y., Zhang, Y. J., Ren, L. J., Han, X. K., Guo, Q. J.,
933 Sun, Y. L., Wang, Z. F., and Fu, P. Q.: Stable sulfur isotope ratios and chemical compositions
934 of fine aerosols (PM_{2.5}) in Beijing, China, *Sci. Total Environ.* 633, 1156-1164,
935 <https://doi.org/10.1016/j.scitotenv.2018.03.153>, 2018.

- 936 Xu, S., Ren, L., Lang, Y., Hou, S., Ren, H., Wei, L., Wu, L., Deng, J., Hu, W., Pan, X., Sun, Y.,
937 Wang, Z., Su, H., Cheng, Y., and Fu, P. Q.: Molecular markers of biomass burning and primary
938 biological aerosols in urban Beijing: size distribution and seasonal variation, *Atmos. Chem.*
939 *Phys.*, 20, 3623–3644, <https://doi.org/10.5194/acp-20-3623-2020>, 2020.
- 940 Xu, W. Y., Zhao, C. S., Ran, L., Deng, Z. Z., Liu, P. F., Ma, N., Lin, W. L., Xu, X. B., Yan, P.,
941 He, X., Yu, J., Liang, W. D., and Chen, L. L.: Characteristics of pollutants and their
942 correlation to meteorological conditions at a suburban site in the North China Plain, *Atmos.*
943 *Chem. Phys.*, 11, 4353–4369, <https://doi.org/10.5194/acp-11-4353-2011>, 2011.
- 944 Yamaguchi, N., Ichijo, T., Sakotani, A., Baba, T., and Nasu, M.: Global dispersion of bacterial
945 cells on Asian dust, *Sci. Rep.*, 2(1), 525, <http://doi:10.1038/srep00525>, 2012.
- 946 Yan, C., Zheng, M., Sullivan, A. P., Shen, G., Chen, Y., Wang, S., Zhao, B., Cai, S., Desyaterik,
947 Y., Li, X., Zhou, T., Gustafsson, O., and Collett, J. L.: Residential coal combustion as a
948 source of levoglucosan in China, *Environ. Sci. Technol.* 2018, 52, 3, 1665–1674
949 <http://doi:10.1021/acs.est.7b05858>, 2018.
- 950 Yttri, K. E., Dye, C., and Kiss, G.: Ambient aerosol concentrations of sugars and sugar-alcohols at
951 four different sites in Norway, *Atmos. Chem. Phys.*, 7, 4267–4279,
952 <https://doi.org/10.5194/acp-7-4267-2007>, 2007.
- 953 Yu, L., Wang, G., Zhang, R., Zhang, L., Song, Y., Wu, B. B., Li, X. F., An, K., and Chu, J. H.:
954 Characterization and source apportionment of PM_{2.5} in an urban environment in Beijing,
955 *Aerosol Air Qual. Res.*, 13, 574–583, <https://doi.org/10.4209/aaqr.2012.07.0192>, 2013.
- 956 Zhang, Y. X., Shao, M., Zhang, Y. H., Zeng, L. M., He, L. Y., Zhu, B., Wei, Y. J., and Zhu, X. L.:
957 Source profiles of particulate organic matters emitted from cereal straw burnings, *J. Environ.*
958 *Sci.-China*, 19, 167–175, [https://doi.org/10.1016/S1001-0742\(07\)60027-8](https://doi.org/10.1016/S1001-0742(07)60027-8), 2007.
- 959 Zhou, L., Kim, E., Hopke, P. K., Stanier, C. O., and Pandis, S.: Advanced factor analysis on
960 Pittsburgh particle size-distribution data special issue of aerosol science and technology on
961 findings from the fine particulate matter supersites program, *Aerosol Sci. Technol.*, 38, 118–
962 132, <https://doi.org/10.1080/02786820390229589>, 2004.
- 963 Zhu, C., Kawamura, K., Fukuda, Y., Mochida, M., and Iwamoto, Y.: Fungal spores overwhelm
964 biogenic organic aerosols in a midlatitudinal forest, *Atmos. Chem. Phys.*, 16, 7497–7506,
965 <https://doi.org/10.5194/acp-16-7497-2016>, 2016.

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973 **Figure Captions**

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975 Figure 1. Geographical location of Mangshan, China. The map was downloaded from © Google
976 Maps 2019.

977 Figure 2. (a) The meteorological parameters at Mangshan during sampling periods, (b) Fractions
978 of local wind directions at Mangshan site, north of Beijing, China.

979 Figure 3. Concentrations (ng m^{-3}) of sugar compound (a) overall, (b) daytime and (c) nighttime in
980 aerosol samples from Mangshan during September-October 2007 (The error bars denote the
981 standard deviation).

982 Figure. 4. Temporal variations in the concentrations (ng m^{-3}) of anhydrosugars in the Mangshan
983 aerosol samples collected for September-October 2007. (Solid circle represents nighttime
984 samples collected from 18:00 to 09:00 hours. Hollow circle represents daytime samples).

985 Figure. 5. Temporal variations in the concentrations (ng m^{-3}) of primary sugars in the Mangshan
986 aerosol samples collected for September-October 2007. (Solid circle represents nighttime
987 samples collected from 18:00 to 09:00 hours. Hollow circle represents daytime samples). Y-
988 axis shows temporal variations in the concentrations ($\mu\text{g m}^{-3}$) of Ca^{2+} .

989 Figure. 6. Temporal variations in the concentrations (ng m^{-3}) of sugar alcohols in the Mangshan
990 aerosol samples collected for September-October 2007. (Solid circle represents nighttime
991 samples collected from 18:00 to 09:00 hours. Hollow circle represents daytime samples).

992 Figure 7. PMF analyses of sugar compounds in Mangshan aerosols based on the autumn 2007 data
993 set.

994 Figure 8. Source contributions to sugar compounds from various sources based on PMF analyses.

995 Figure 9. The concentrations and relative contributions of the carbon content of anhydrosugars,
996 primary sugars and sugar alcohols to the carbon concentrations of measured sugar
997 compounds, water-soluble organic carbon (WSOC) and organic carbon (OC) fraction of
998 Mangshan aerosols (a = daytime and b = nighttime). The concentrations and relative
999 contribution of the carbon content of five sources of sugar compounds to total sugar
000 compounds measured, WSOC and OC fraction of Mangshan aerosols (c = daytime and d =
001 nighttime).

002 Figure 10. Mass concentrations ratio of carbon contents of (a) levoglucosan (Lev) to organic
003 carbon (OC) and water soluble organic carbon (WSOC), (b) levoglucosan (Lev) to organic
004 carbon (OC) daytime and night time, (c) levoglucosan (Lev) to water soluble organic carbon
005 (WSOC) daytime and night time, (d) average levoglucosan to mannosan ratios (Lev/Man) in
006 the Mangshan aerosol samples for autumn 2007.

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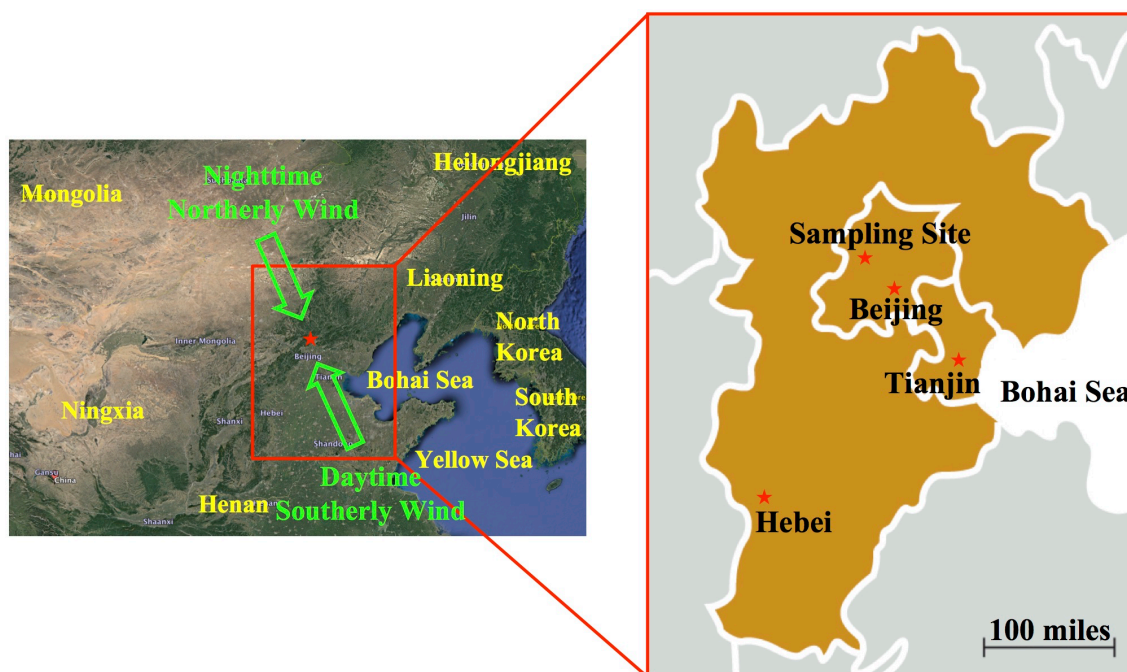
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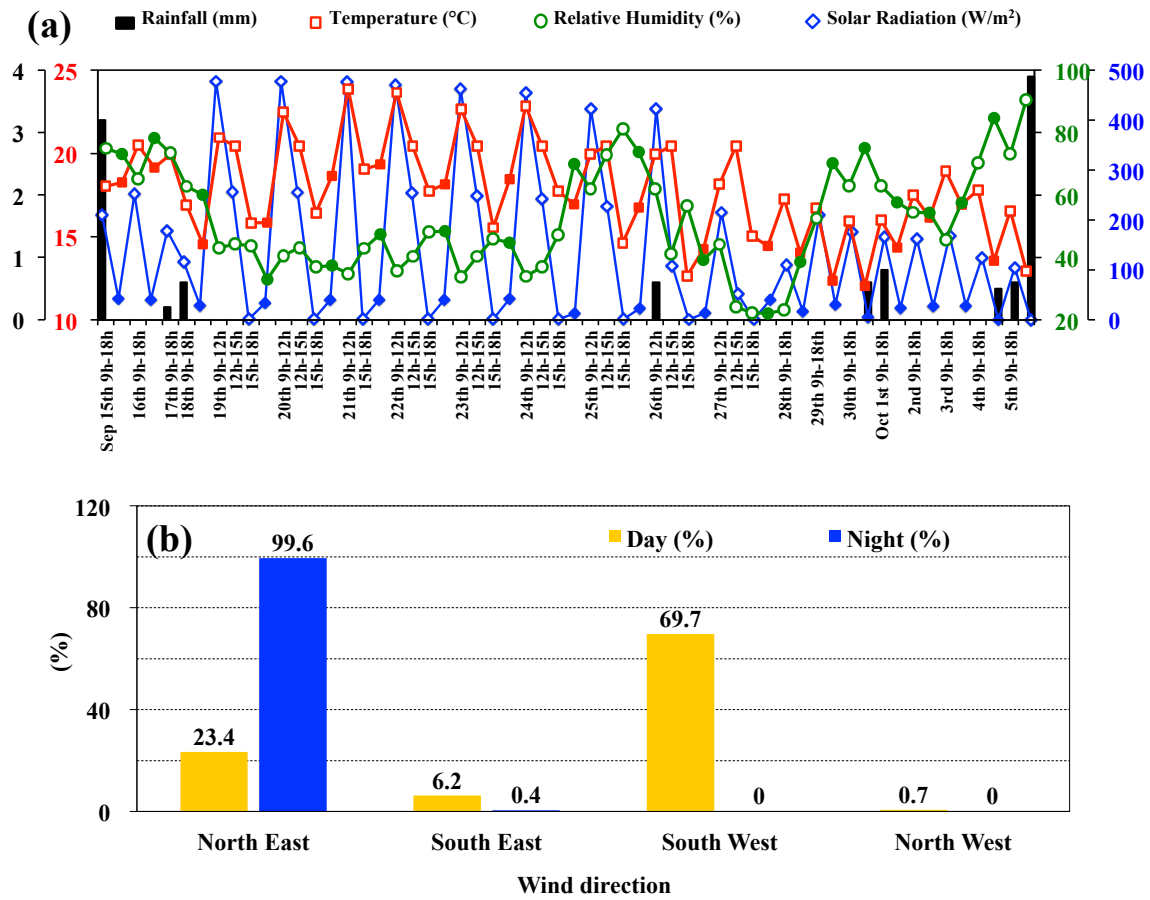
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Fig. 2.



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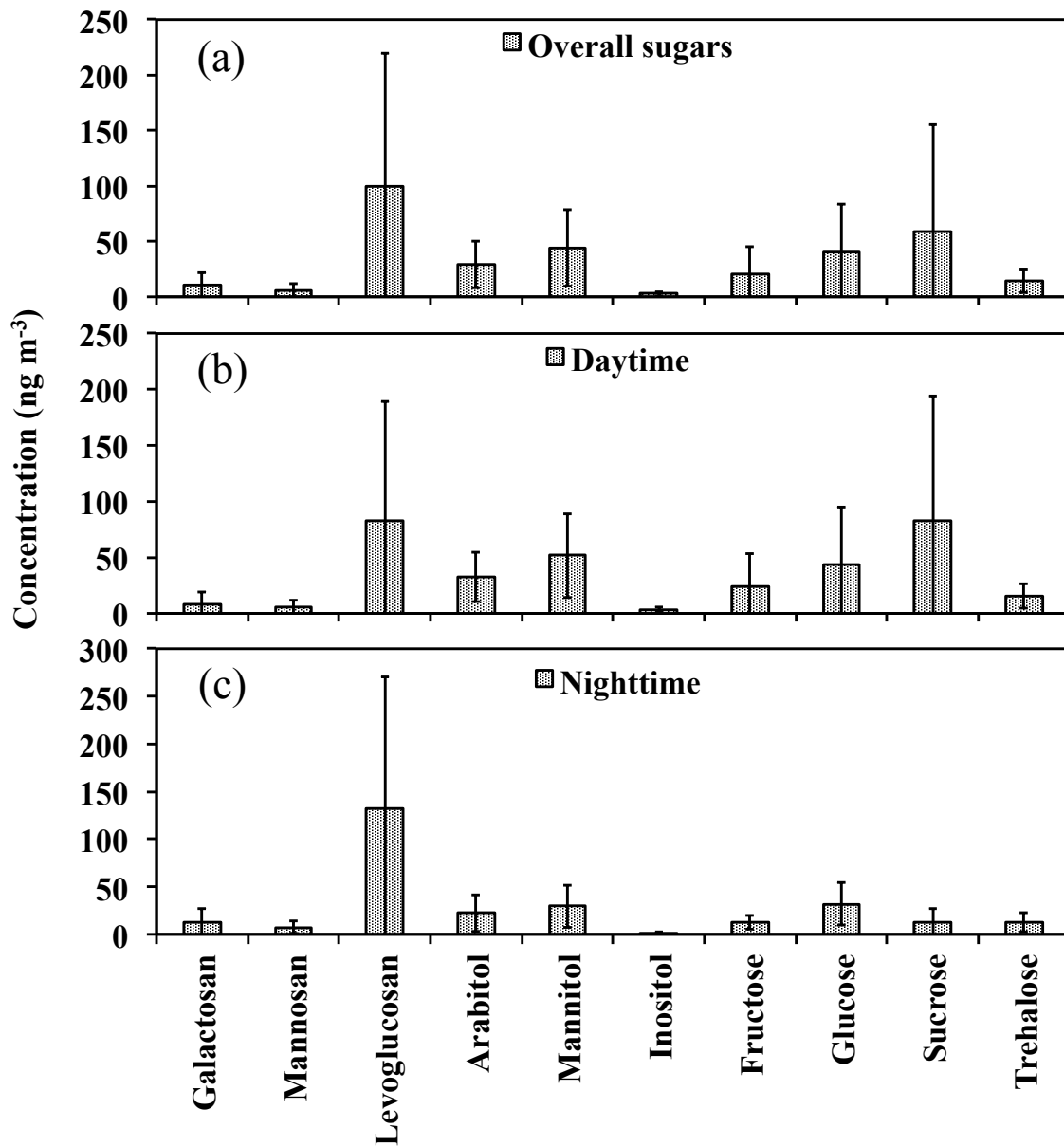
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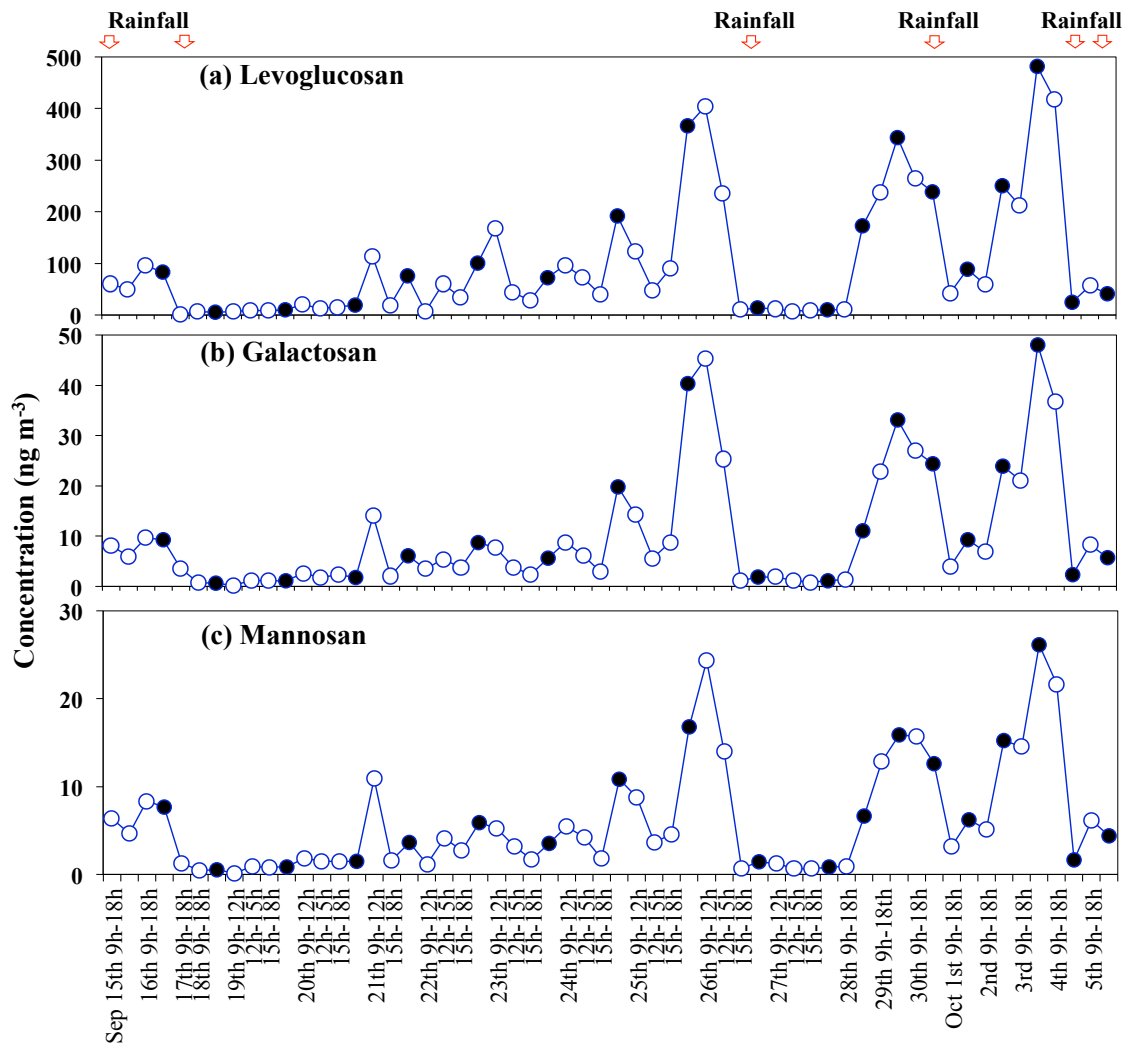
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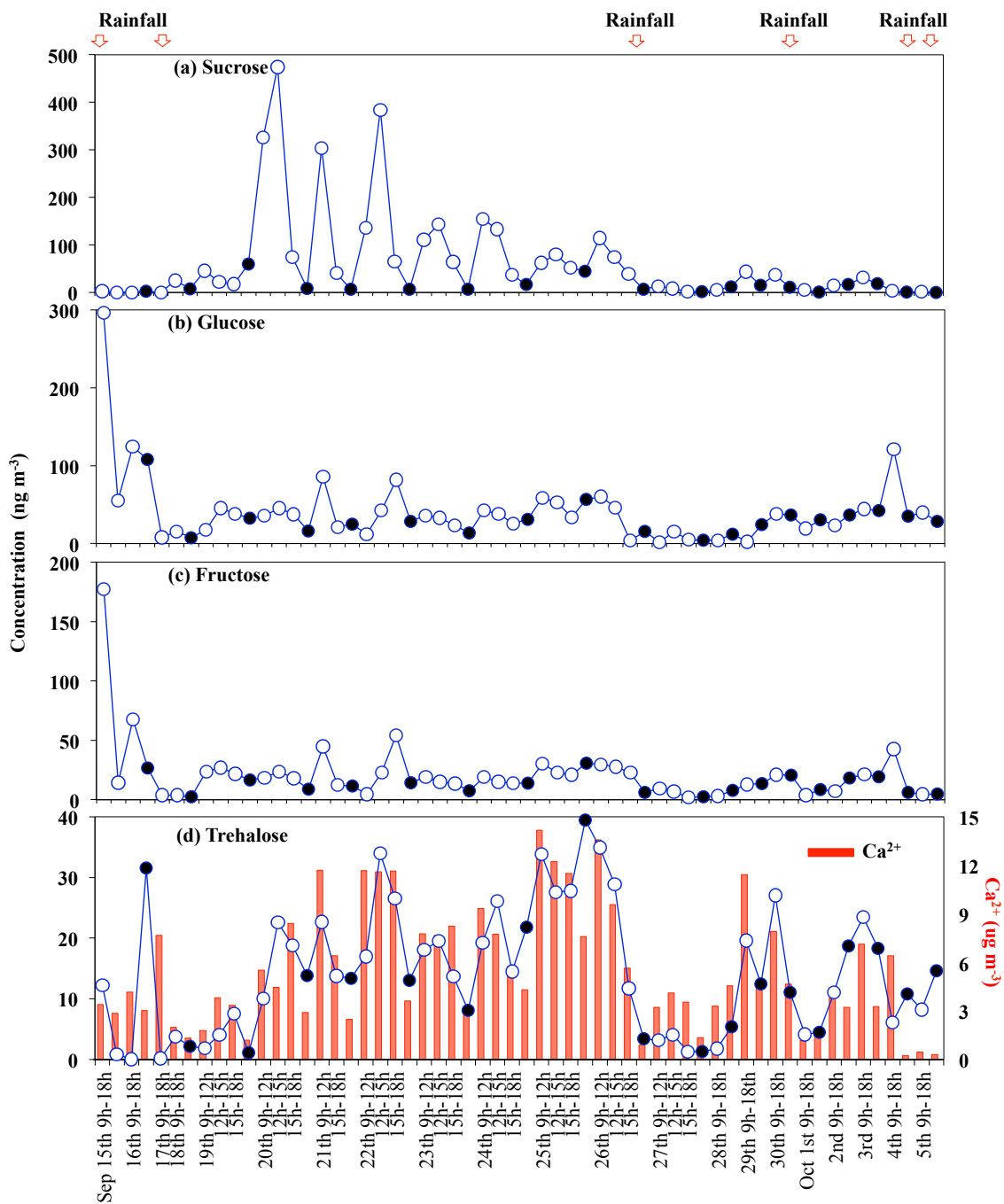
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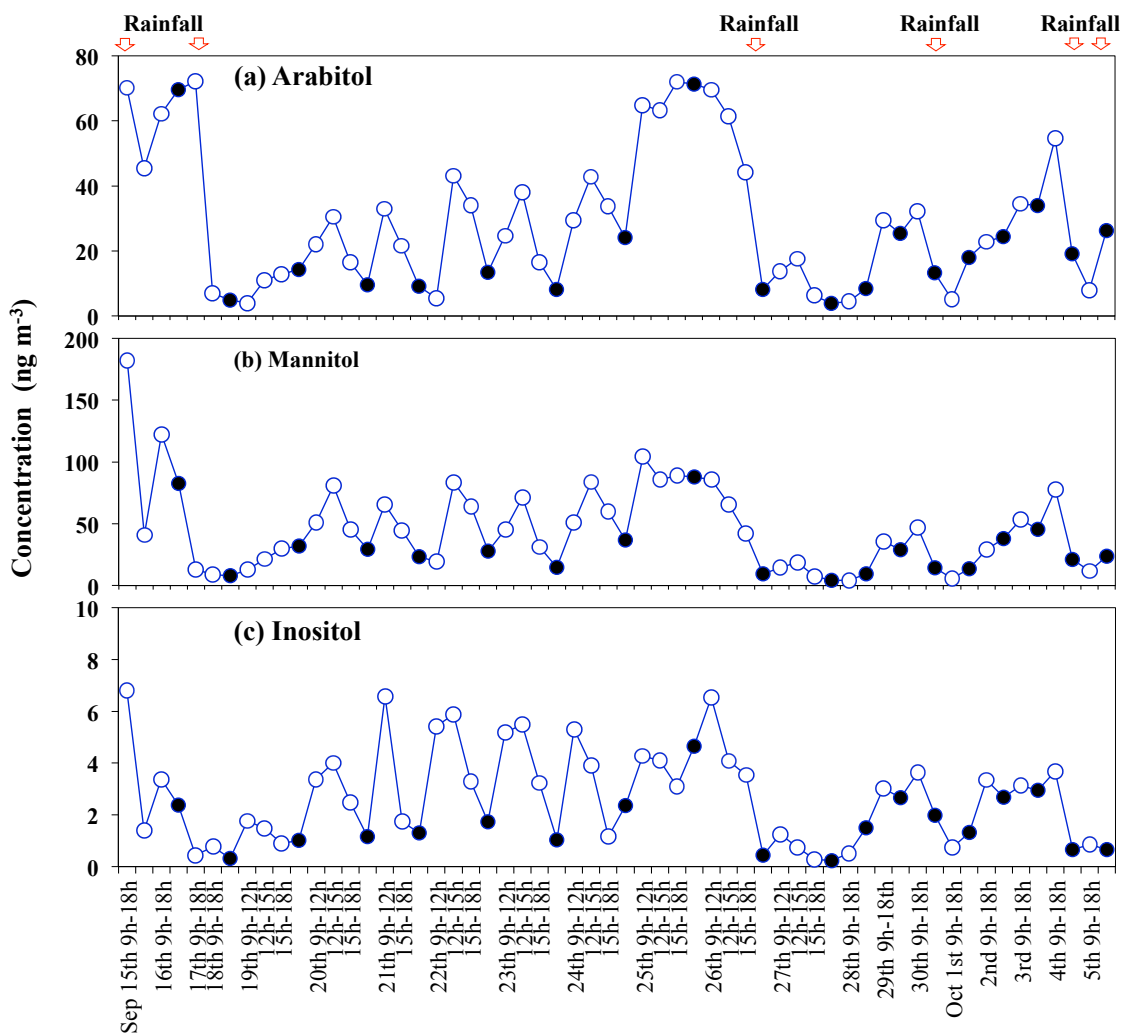
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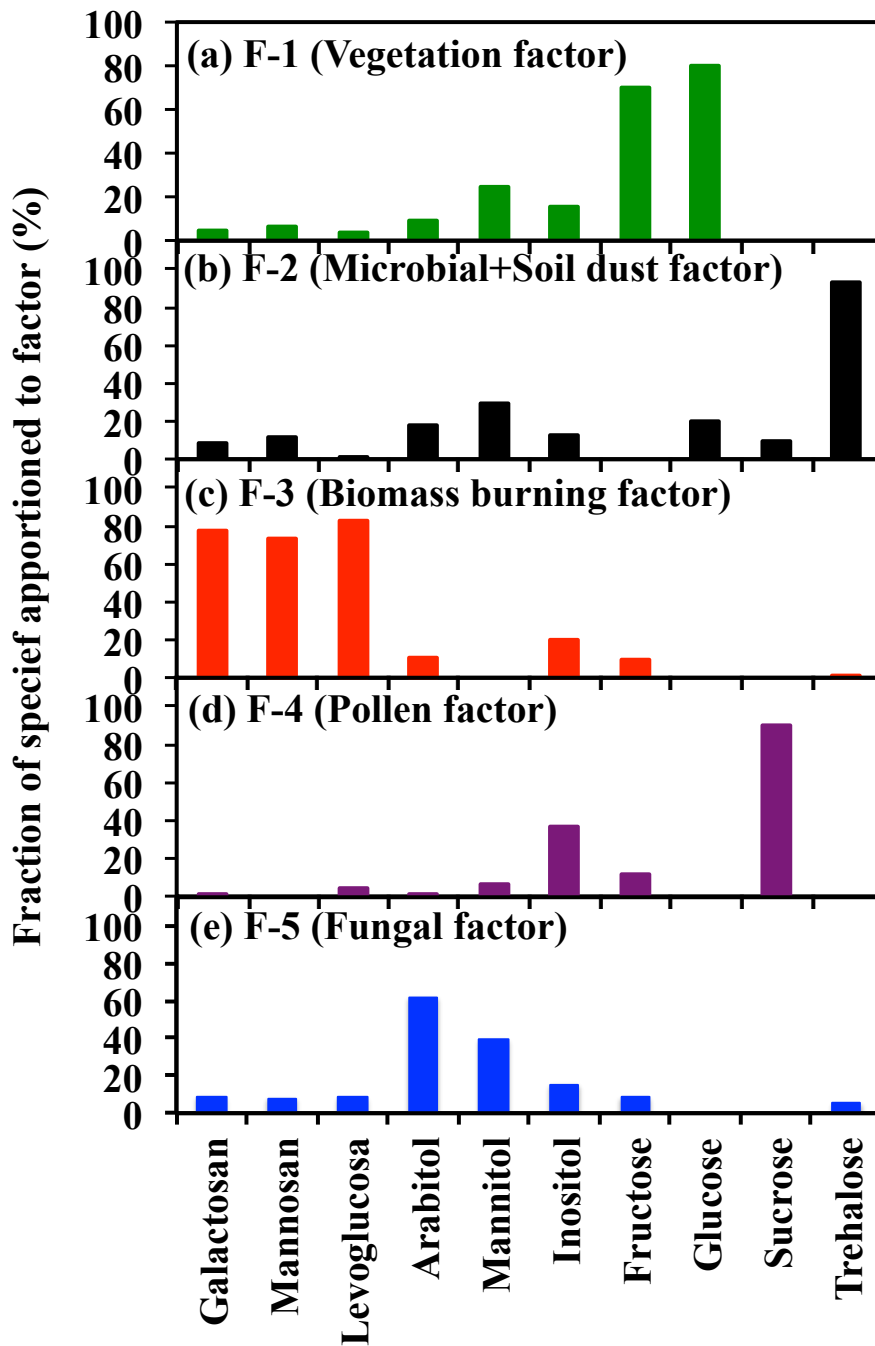
Fig. 6.



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Fig. 7.



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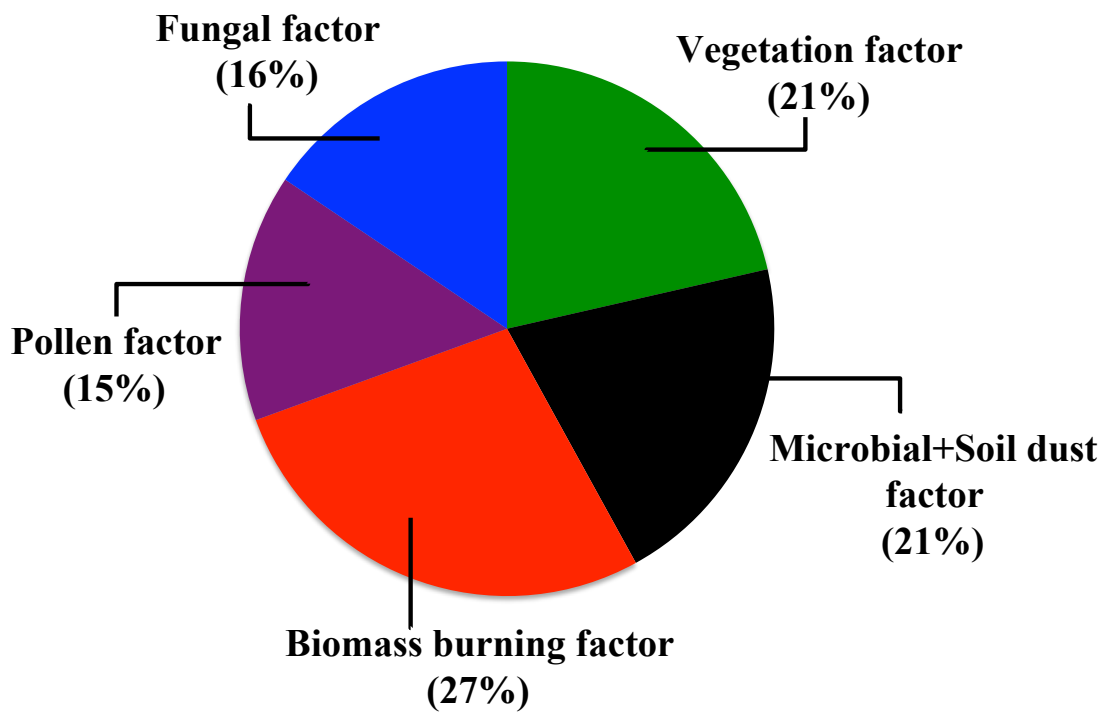
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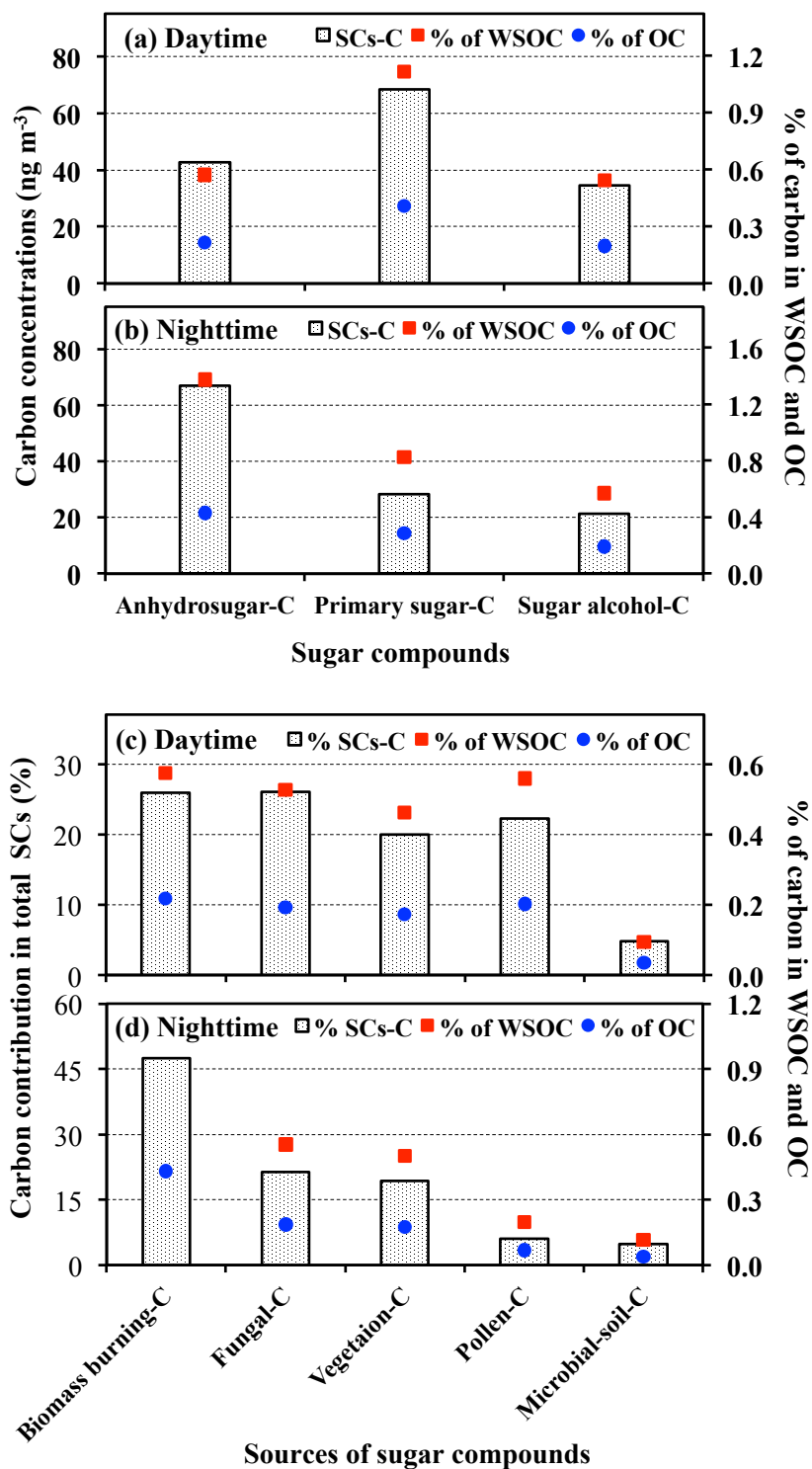
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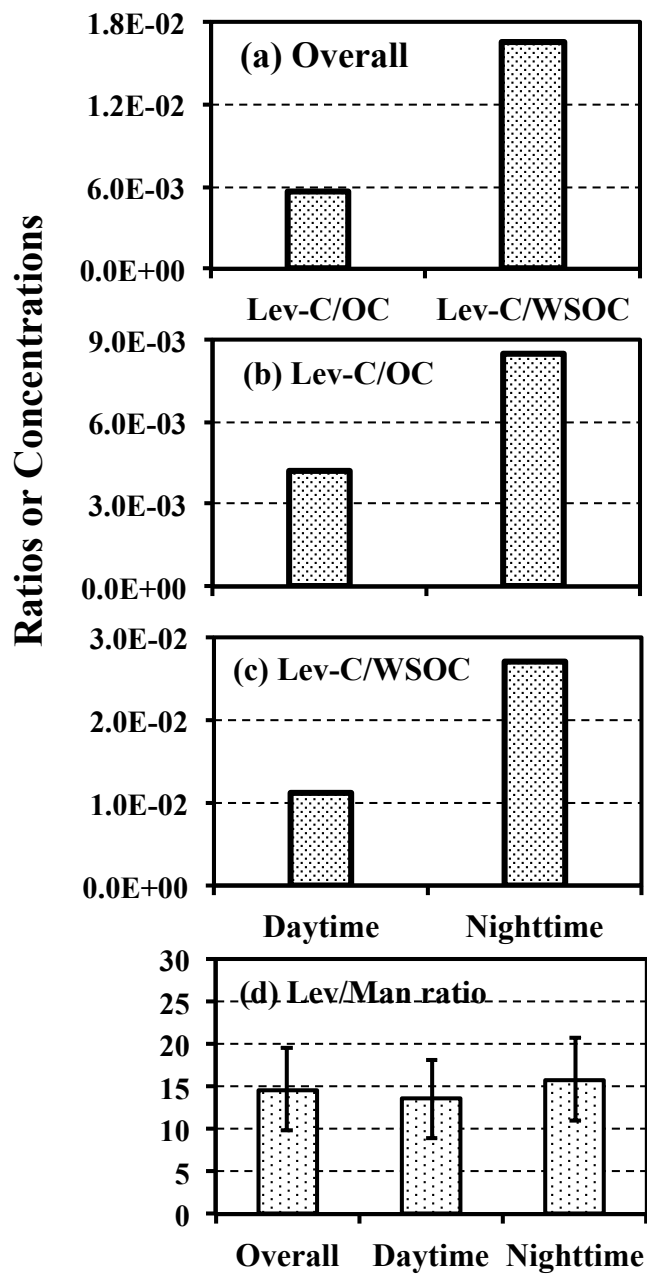
Fig. 9.



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Fig. 10.



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Table 1. Minimum, maximum, average and standard deviations of concentrations of sugar compounds in aerosol samples (TSP) from Mangshan, China.

Sugar Compounds	Overall				Daytime (n = 38)				Nighttime (n = 20)			
	Min	Max	Avg.	S.D.	Min	Max	Avg.	S.D.	Min	Max	Avg.	S.D.
Anhydrosugars												
Galactosan	0.14	48.0	10.1	11.9	0.14	45.3	8.53	10.5	0.69	48.0	13.0	14.0
Mannosan	0.13	26.1	6.05	6.33	0.13	24.3	5.37	6.01	0.53	26.1	7.35	6.87
Levogluconan	1.17	482	100	119	1.17	418	83.2	106	5.66	482	132	138
Sugar alcohols												
Arabitol	3.89	72.2	29.1	21.5	3.99	72.2	32.5	22.0	3.89	71.3	22.5	19.4
Mannitol	4.19	182	44.1	34.5	4.19	182	51.7	37.5	4.40	87.7	29.6	22.3
Inositol	0.23	6.8	2.62	1.81	0.27	6.80	3.14	1.90	0.23	4.65	1.62	1.09
Primary sugars												
Fructose	1.72	177	20.1	24.6	1.72	177	23.9	29.3	2.64	30.9	12.8	7.67
Glucose	1.86	297	40.0	43.4	1.86	297	44.2	50.8	4.52	108	32.0	22.8
Sucrose	0.02	474	58.5	96.5	0.02	474	82.9	112	0.04	60.1	12.3	15.1
Trehalose	0.06	39.5	14.3	10.5	0.06	34.9	15.3	10.6	0.87	39.5	12.3	10.2
Anhydrosugars	6.01	556	116	137	6.01	476	97.1	122	6.88	556	152	159
Primary sugars	9.41	565	133	125	9.41	565	166	141	10.5	172	69.4	43.0
Sugar alcohols	8.53	259	75.8	54.7	9.09	259	87.4	57.5	8.53	164	53.7	41.9
Total Sugars	30.8	875	325	232	34.1	875	351	240	30.8	759	276	212
Anhydrosugars (%)			31.9				24.6				45.7	
Primary sugars (%)			41.8				47.3				31.3	
Sugar alcohols (%)			26.4				28.1				23.0	

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Table 2. Statistical summary of correlations among the chemical species and meteorological variables in aerosol samples collected at a forest site in northern Japan

Linear regression	Correlation coefficient	p value	Significance of correlation at P value < 0.05
Overall (n = 58)			
Levoglucon vs. Galactosan	0.98	< 0.05	Significant
Levoglucon vs. Mannosan	0.97	< 0.05	Significant
Mannosan vs. Galactosan	0.98	< 0.05	Significant
Sucrose vs. Temperature	0.52	< 0.05	Significant
Sucrose vs. Solar radiation	0.55	< 0.05	Significant
Arabitol vs. Mannitol	0.81	< 0.05	Significant
Arabitol vs. RH	0.69	< 0.05	Significant
Mannitol vs. RH	0.57	< 0.05	Significant
Glucose vs. Fructose	0.94	< 0.05	Significant
Trehalose vs. Arabitol	0.58	< 0.05	Significant
Trehalose vs. Mannitol	0.58	< 0.05	Significant
Trehalose vs. Ca ²⁺	0.70	< 0.05	Significant
Daytime (n = 38)			
Sucrose vs. Ca ²⁺	0.32	> 0.05	Not significant
Glucose vs. Ca ²⁺	0.02	> 0.05	Not significant
Trehalose vs. Arabitol	0.49	< 0.05	Significant
Trehalose vs. Mannitol	0.51	< 0.05	Significant
Trehalose vs. Ca ²⁺	0.81	< 0.05	Significant
Fructose vs. Mannitol	0.79	< 0.05	Significant
Levoglucon vs. OC	0.45	< 0.05	Significant
Levoglucon vs. WSOC	0.40	< 0.05	Significant
Nighttime (n = 20)			
Sucrose vs. Ca ²⁺	0.37	> 0.05	Not significant
Glucose vs. Ca ²⁺	0.27	> 0.05	Not significant
Trehalose vs. Arabitol	0.76	< 0.05	Significant
Trehalose vs. Mannitol	0.85	< 0.05	Significant
Trehalose vs. Ca ²⁺	0.61	< 0.05	Significant
Fructose vs. Mannitol	0.86	< 0.05	Significant
Levoglucon vs. OC	0.81	< 0.05	Significant
Levoglucon vs. WSOC	0.70	< 0.05	Significant

The data of Ca²⁺, OC and WSOC are adapted from He et al. (2015).

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