

1 **Measurement report: Diurnal and temporal variations of sugar compounds in suburban**
2 **aerosols from the northern vicinity of Beijing, China: An influence of biogenic and**
3 **anthropogenic sources**

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19 Key points:

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- 21 1. Autumn time observations of sugar compounds (SCs) in the northern vicinity of Beijing, China.
- 22 2. Influence of natural biogenic emissions on SCs from forest area.
- 23 3. Influence of anthropogenic and bioaerosol on SCs from the Beijing area.
- 24 4. Biomass burning is a significant contributor to SCs.
- 25 5. Biogenic and fungal-microbial emissions are essential sources for mannitol and arabitol.

26

27 **Abstract**

28 Sugar compounds (SCs) are major water-soluble constituents in atmospheric aerosols. In
29 this study, we investigated their molecular compositions and abundances in the northern receptor
30 site (Mangshan) of Beijing, China, to better understand the contributions from biogenic and
31 anthropogenic sources using a gas chromatography–mass spectrometry technique. The sampling
32 site receives anthropogenic air mass transported from Beijing by southerly winds, while northerly
33 winds transport relatively clean air mass from the forest areas. Day- and nighttime variations were
34 analyzed for anhydrosugars, primary sugars, and sugar alcohols in autumn 2007. We found that
35 biomass burning (BB) tracers were more abundant in nighttime than daytime, while other SCs
36 showed different diurnal variations. Levoglucosan was found as a dominant sugar among the SCs
37 observed, indicating an intense influence of local BB for cooking and space heating at the
38 surroundings of the Mangshan site. The high levels of arabitol and mannitol in daytime suggest a
39 significant contribution of locally emitted fungal spores and long-range transported bioaerosols
40 from the Beijing area. The plant emissions from Mangshan forest park significantly control the
41 diurnal variations of glucose, fructose, and mannitol. The meteorological parameters (relative
42 humidity, temperature, and rainfall) significantly affect the concentrations and diurnal variations of
43 SCs. Sucrose (pollen tracer) showed a clear diurnal variation, peaking in the daytime due to higher
44 ambient temperature and wind speed, which influences the pollen release from the forest plants. We
45 found the contribution of trehalose from soil dust in daytime, while microbial and fungal spores
46 were responsible for nighttime. Anhydrosugar and primary sugars are prime carbon sources of the
47 Mangshan aerosols. The high ratios of levoglucosan in organic carbon and water soluble organic
48 carbon in nighttime suggest a significant contribution of BB to organic aerosols at night.
49 Levoglucosan/mannosan ratios demonstrate that low temperature burning of hardwood is dominant
50 in Mangshan. The positive matrix factorization analysis concluded that forest vegetation, fungal
51 species, and local BB are the significant sources of SCs.

52 **Keywords:** Anthropogenic bioaerosols, biomass burning, pollen tracer, fungal tracers, soil dust,
53 and microbial tracers

54

55 **1. Introduction**

56 Increased economic growth and massive consumption of fossil fuels from industries emit
57 anthropogenic gases, aerosols, and biomass burning (BB) products cause severe air pollution in East
58 Asian countries (Lelieveld et al., 2015; Lin et al., 2014; Kawamura et al., 2013; Li et al., 2010; Sun
59 et al., 2016). On a global scale, one-fourth of anthropogenic aerosols is contributed by China, and
60 approximately 70% of which emitted from coal burning (Streets et al., 2003). Beijing is situated in
61 the northern part of China, with a 20 million people and 5 million motor vehicles. Beijing is one of
62 the largest polluted cities in East Asia; its air quality deteriorates seriously due to massive emissions
63 of organic aerosols (OAs) from vehicles and industries (Cao et al., 2014; Qiao et al., 2018; Tao et
64 al., 2017; Wei et al., 2018; Yu et al., 2013). OAs are comprised with a complex mixture of diverse
65 molecules (Xu et al., 2011). They play essential roles in global climate changes via the modification
66 of radiative forcing and cause a serious negative impact on human health (Fuzzi et al., 2007). OAs
67 contain various water-soluble organic compounds, which can act as cloud condensation nuclei
68 (CCN) (Kanakidou et al., 2005).

69 BB is essentially a primary source of OAs, controlling the air quality levels and affecting the
70 earth's radiative forcing by scattering or absorbing incident solar radiation (Deshmukh et al., 2019a;
71 Kanakidou et al., 2005; Kanaya et al. 2013; Streets et al., 2003; Sullivan et al., 2008). There are
72 several kinds of BB, including industrial biofuel burning, open field burning (fires of the forest,
73 peatlands, and agricultural wastes), and domestic BB burning for house heating and cooking, which
74 emits BB products into the atmosphere (Akagi et al., 2011). The BB aerosols are subjected to long-
75 range atmospheric transport once they are emitted into the atmosphere (Verma et al., 2015).
76 Levoglucosan (1,6-anhydro- β -D-glucopyranose) is a pyrolysis product of cellulose and
77 hemicellulose, which is generally found as major organic constituents in the BB-influenced aerosols
78 (Simoneit et al., 1999; 2002). Levoglucosan have been reported as a specific tracer for BB aerosols
79 (Engling et al., 2009).

80 Sugar compounds (SCs) are ubiquitous in the atmosphere from different geographical
81 locations, including urban, forest, marine, and polar regions (Burshtein et al., 2011; Fu et al., 2010;
82 Wan et al., 2017). SCs are emitted from algae, microbes, pollen, suspended soil particle, and
83 associated biota into the atmosphere by various processes, and thus they are termed as primary
84 biological aerosol particles (PBAPs) (Carvalho et al., 2003; Despres et al., 2012; Elbert et al.,
85 2007). The detailed study of bio-aerosols has been emphasized in the past decades due to the global
86 impact of microbes and fungi because they can travel long distances from the source regions by
87 winds (Burshtein et al., 2011; Brown and Hovmoller, 2002; Yamaguchi et al., 2012). Fungi are

88 essential microbes in the ecosystem, which discharge spores of 8-186 Tg yr⁻¹ into the atmospheric
89 environment (Elbert et al., 2007; Heald and Spracklen, 2009). Sugar alcohols like arabitol and
90 mannitol are enriched in fungal spores; thus, they are considered as specific tracers (Bauer et al.,
91 2008).

92 Devis et al. (1988, 1990) reported that mannitol was also found in about 70 different higher
93 plant families. Loescher et al. (1992) reported that mannitol is an important photosynthetic product
94 converted by biosynthesis in plants. Keller and Matile (1989) also found the arabitol and mannitol
95 during the increased photosynthesis in growing vegetation. Pollens are the largest particles that
96 could contribute up to 65% of the PBAPs, which are the significant sources for sucrose and fructose
97 in the forest aerosols (Manninen et al., 2014; Pacini, 2000). Higher plants synthesize primary sugars
98 (glucose, fructose, and sucrose) during photosynthesis, which are circulated by phloem to
99 accumulate in root cells and to develop plant sections (Jaenicke, 2005; Jia et al., 2010; Pacini,
100 2000). Cowie et al. (1984) also reported various sugars in terrestrial plant fruits, flowers, and plant
101 tissues. Bielecki (1995) reported that glucose, fructose, and sucrose are well-known components of
102 microbes and invertebrates. The plant debris, as well as lichens, invertebrates, and soil dust, are also
103 recognized as possible sources of primary sugars in the atmosphere (Medeiros et al., 2006; Rogge et
104 al., 2007; Simoneit et al., 2004).

105 Previous studies analyzed aerosol samples for SCs and discussed several factors to control
106 their local and global atmospheric levels. Recently, Xu et al. (2020) examined the seasonal
107 molecular distributions of primary biological aerosols and BB aerosol samples collected from urban
108 Beijing. They reported a high level of arabitol, mannitol, sucrose, glucose, and fructose in the
109 vegetation-growing season. Kang et al. (2018) also reported higher concentrations of sugars in the
110 urban aerosols from Beijing. They suggested a large contribution of coal combustion and
111 agriculture residue burning under stable meteorological conditions in winter and spring. Verma et
112 al. (2015, 2018) reported that the atmospheric circulations and long-range transport of organic-/bio-
113 aerosols from East Asia significantly control the levels and compositions of SCs over the western
114 North Pacific. The above studies discussed the several factors that affect the concentrations of SCs
115 in the aerosol samples collected from urban and remote areas.

116 In this study, we conducted analyses of SCs in the aerosol samples collected from the
117 northern vicinity of Beijing City in 2007. Here, we present comprehensive data sets of
118 anhydrosugars, primary sugars, and sugar alcohols in the suburban aerosol samples and their diurnal
119 variations to explain the source variance following the wind patterns in the day- and nighttime. The
120 positive matrix factorization (PMF) has been applied to clarify the different sources of measured

121 SCs in the aerosol. We present the influence of local meteorology of sampling site and atmospheric
122 transport from Beijing by southerly winds and Mangshan National Forest Park by northerly winds
123 on the molecular distributions of SCs. Using the mass concentration ratio of levoglucosan to
124 mannosan, we explain the relative contribution of hard and softwood burning to the air quality of
125 Mangshan. This study also discussed carbon contributions of SCs and BB measured in the
126 Mangshan aerosol samples from different sources.

127 **2. Materials and Methods**

128 **2.1. Site description and aerosol sample collection**

129 The sampling site (Mangshan: 40.28 N, 116.26 E) is located 40 km north of Beijing. A
130 detailed description of the sampling site is given in He et al. (2014, 2015). Briefly, Mangshan is
131 surrounded by urban areas in the south and forest areas with the national park in the north (Fig. 1).
132 The ambient temperature was higher in daytime (23.9°C) than nighttime (12.1°C), with an average
133 of 17.8°C during the campaign. The relative humidity (RH) varied significantly from 22.1% to
134 90.5%, with an average of 51.7% during the study period. The rainfall was observed at midnight on
135 15th September, the morning of 17th to evening 18th September, the night of 26th September, and
136 light rain lasted from 4th October to the end of the campaign (Fig. 2a). Interestingly, the sampling
137 site is characterized by a specific wind pattern, i.e., southwest wind (69.9%) prevailed, followed by
138 northeast wind (23.4%) and southeast wind (6.2%) during the daytime (Fig. 2a). The northeast wind
139 (99.5%) was dominated at night, which is consistent with the air mass back trajectories (He et al.,
140 2014) (Fig. 2b). The daytime wind from the southwest direction passed over Beijing, delivering
141 anthropogenic air mass to the Mangshan site.

142 Detailed descriptions of the total suspended particulate (TSP) samples collected at
143 Mangshan are given in He et al. (2014, 2015). Briefly, The aerosol samples were collected near the
144 entrance of Mangshan National Forest Park. The elevation of the sampling location is 187 m above
145 sea level. TSP (size <100 μm) samples were collected on pre-combusted (450°C for 6 h) quartz
146 fiber filters (Pallflex 2500QAT-UP, 20 cm \times 25cm) from 15th September to 5th October 2007. A
147 high-volume air sampler (Kimoto-AS810A) at a flow rate of 1.13-1.17 m³ min⁻¹ was used to collect
148 the TSP samples. After sample collection, the individual filters were placed in pre-combusted glass
149 jars with Teflon-lined screw caps and stored in a dark, cold room at -20°C to prevent microbial
150 activity and loss of semi-volatile organic compounds from the samples. In this study, a total of 58
151 filter samples were analyzed. We collected 3h daytime (from 9 to 12, 12 to 15, 15 to 18 h) (n=26),

152 9h daytime (from 9 to 18 h) (n=12), and 15h nighttime (from 18 to 9 h) (n=20) samples together
153 with four field blanks. Table S1 shows the details of aerosol sample collection in the Mangshan site.

154 2.2. Extraction and derivatization of samples

155 A total of 58 aerosol samples were analyzed for anhydrosugars, primary sugars, and sugar
156 alcohols (Table 1). The sample filters (approximately 21 cm²) were extracted with a
157 dichloromethane and methanol mixture (2:1) under ultrasonication. Pasteur pipettes packed with
158 pre-combusted quartz wool were used to filter the extracts to remove filter debris. After filtration,
159 the extracts were concentrated in a rotary evaporator under vacuum and dried by nitrogen
160 blowdown. The extracts were reacted with 60 µl of N,O-bis-(trimethylsilyl)trifluoroacetamide
161 (BSTFA) with 1% trimethylsilyl (TMS) chloride in the presence of 10 µL of pyridine at 70°C for
162 three hours to derivatize hydroxyl (OH) and carboxyl (COOH) groups into corresponding
163 trimethylsilyl (TMS) ethers and esters, respectively. After the reaction, n-hexane was used for
164 dilution, and C₁₃ n-alkane was added as an internal standard before GC-MS analysis.

165 2.3. Gas chromatography-mass spectrometry determination of sugar compounds (SCs)

166 Details of GC-MS operation and identification of SCs are described in Verma et al. (2015,
167 2018). Briefly, GC-MS analyses were performed on Agilent model 6890 gas chromatograph (GC)
168 combined with an Agilent model 5973 mass selective detector (MSD) to determine SCs. The mass
169 spectrometer was operated in the electron ionization (EI) mode at 70 eV with a scan range of *m/z*
170 40–650. The GC separation was achieved on a DB-5MS fused silica capillary column (30 m × 0.25
171 mm in diameter, 0.25 µm film thickness) and a split/splitless injector. The GC oven temperature
172 was programmed to maintain at 50°C for 2 min and then to increase from 50 to 120°C at a rate of
173 15°C min⁻¹, then from 120 to 305°C at a rate of 5°C min⁻¹. The final isotherm holds at 305°C for 15
174 min. Helium was used as the carrier gas at a flow rate of 1.0 mL min⁻¹. The sample was injected on
175 a splitless mode at 280°C injector temperature. GC-MS data were acquired and processed with
176 Agilent GC/MSD ChemStation software.

177 The individual compounds (TMS derivatives) were identified by comparing the relative
178 response factors determined by the injection of authentic standards and those reported in the
179 literature and library texts (Claeys et al., 2004). Fragment ions of sugar compounds at 217 and 204
180 were used for quantifications. Total ten sugar compounds, including three anhydrosugars
181 (levoglucosan, galactosan, mannosan), four primary sugars (glucose, fructose, sucrose, trehalose
182 and xylose) and three sugar alcohols (arabitol, mannitol, and inositol), were detected in the

183 Mangshan aerosols. Field blanks were treated as a real sample and analyzed by the procedure used
184 for the real samples. Recoveries for SCs were better than 85% as obtained by the standards spiked
185 to precombusted quartz filter followed by extraction and derivatization. Based on the duplicate
186 analysis, the analytical errors in the concentrations of the detected compounds were obtained to be
187 within 10%. The detection limits of SCs corresponds to ambient concentrations of 150-620 pg μL^{-1} ,
188 which corresponds to ambient concentrations of 15-70 pg m^{-3} under a typical sampling volume of
189 900 m^3 . The data of water-soluble organic carbon (WSOC), organic carbon (OC), and inorganic
190 ions (Ca^{2+}) were reported in He et al. (2015).

191 2.4. Positive Matrix Factorization (PMF) Analysis

192 Positive matrix factorization (PMF) is a powerful statistical tool for resolving the potential
193 sources contributing to atmospheric particles (Paatero and Tapper, 1994). The measured ambient
194 concentrations and method detection limits (MDLs) of SCs were used to calculate the uncertainties.
195 The measured concentrations of SCs below or equal to the MDLs were replaced by half of the
196 MDL, and associated uncertainties were set at 5/6 of the MDL $[(5/6) \times \text{MDL}]$ values of each
197 sample. The geometric mean concentrations were used for missing concentrations, and the
198 uncertainty of the concentrations greater than the MDL was calculated based on the following
199 equation:

$$200 \quad \text{Uncertainty} = \sqrt{(\text{error fraction} \times \text{concentration})^2 + (0.5 \times \text{MDL})^2}$$

201 The error fraction is a user-provided estimation of the analytical uncertainty of the measured
202 concentration or flux. For example, Han et al. (2017) used an error fraction of 0.2-0.3 for organics
203 and 0.2 for all the species. In this work, the error fraction was set to be 0.3 for all species. Paatero et
204 al. (2002) and Zhou et al. (2004) reported detailed discussions of the determination and application
205 of PMF analysis.

206 3. Results and Discussion

207 3.1. Ambient concentrations and diurnal variations of SCs

208 We detected a total of ten SCs, including three anhydrosugars, four primary sugars, and three
209 sugar alcohols in the Mangshan aerosol samples. Figure 3a-c showed the temporal variations and
210 Table 1 showed minimum, maximum, and average concentrations of anhydrosugars, primary
211 sugars, and sugar alcohols with a standard deviation. The overall concentrations of SCs varied from
212 30.8–875 ng m^{-3} (avg. 325 ng m^{-3}), showed insignificant diurnal variations. Interestingly, higher
213 average concentrations of SCs were reported for the aerosol samples collected from at Mt. Tai

(daytime 640 ng m⁻³ and nighttime 799 ng m⁻³) in the North China Plain (Fu et al., 2008) than the Mangshan aerosol. The diurnal concentrations of SCs may be significantly influenced by vegetation and BB activities in the Mangshan site. SCs are significantly contributed by plant fractions and fungus from the forest area (Zhu et al., 2016). The meteorological parameters also affect the concentrations of SCs in the forest site (Miyazaki et al., 2012).

In addition, anthropogenic aerosols emitted from urban areas are probably transported to the northern receptor site in daytime by a southerly wind (He et al., 2014; 2015). Therefore, the high levels of SCs in daytime may be related to the transport of organic and bio-aerosols from urban regions. The nighttime, the wind direction is shifted to northerly, delivering comparatively clean air masses from the Mangshan National Forest area to the sampling site. Air mass from the forest may significantly contribute to nighttime SCs in the Mangshan site. The influence of local sources and long-range transported aerosols on the SCs will be discussed in sections 3.1.1 to 3.1.3.

3.1.1. Ambient concentrations and diurnal variations of anhydrosugars

The average concentrations of anhydrosugars were found 116 ng m⁻³, contributing 31.9% of overall SCs in the Mangshan aerosols (Table 1). Figure 4a-c shows the temporal variations of anhydrosugars. They are more abundant in nighttime (avg. 152 ng m⁻³) than daytime (avg. 97.1 ng m⁻³). Levoglucosan (100 ng m⁻³) is the most abundant anhydrosugar followed by galactosan (10.1 ng m⁻³) and mannosan (6.05 ng m⁻³) detected in Mangshan aerosols. Kang et al. (2018) reported high levels of levoglucosan (avg. 110 ng m⁻³) in autumn aerosols from Beijing, China. It is well known that biofuel burning is the common energy source for cooking and house heating in China in winter and autumn (Verma et al., 2015), thus the domestic BB activities in the surroundings of the Mangshan site significantly contribute to the levoglucosan. BB tracers showed significant positive correlations with each other (levoglucosan and galactosan, $r = 0.98$; levoglucosan and mannosan, $r = 0.97$; galactosan and mannosan, $r = 0.98$), suggesting their similar sources in the Mangshan aerosols (Table 2).

The levoglucosan concentrations showed significant diurnal variations, which was higher in nighttime (avg. 132 ng m⁻³) than daytime (avg. 83.2 ng m⁻³) (Table 1). A similar diurnal pattern was also found for the concentrations of galactosan and mannosan. The increased concentrations of BB tracers were observed during the periods of lower ambient temperature (Figs. 2a, 4a-c). The higher ambient temperature was recorded in daytime between 09h to 15h during the campaign, associated with declined BB activities. In this sequence, the nighttime samples were collected from 18:00h to 09:00h, including peak hours of BB for domestic purpose. Therefore, it is reasonable to

246 detect higher abundances of BB tracers in the nighttime than daytime. Hence, it is evident that BB
247 activities were increased at night because of cooking and house heating at cool night in autumn. In
248 addition, recent studies reported the widespread BB aerosols in the North China Plain, including
249 megacities such as Beijing, Nanjing, Hebei, and Tianjin (Lelieveld et al., 2015; Kawamura et al.,
250 2013; Li et al., 2010; Sun et al., 2016). Therefore, the atmospheric transport of BB aerosols from
251 the urban area to the Mangshan site by southerly winds cannot be excluded. The diurnal variations
252 of levoglucosan may be significantly influenced by the local BB activities and transported BB
253 aerosols from urban areas, where BB products are generated by brown coal combustion (Yan et al.,
254 2018).

255 3.1.2. Ambient concentrations and diurnal variations of primary sugars

256 The fragment of vascular plants contains primary sugars, including glucose, fructose,
257 sucrose, and trehalose (Medeiros et al., 2006). Primary sugars were found as the most abundant
258 sugars (avg. 133 ng m^{-3}), contributing to 41.8% of the total SCs in Mangshan aerosols (Table 1).
259 They showed apparent diurnal variations with daytime high (avg. 166 ng m^{-3}) and nighttime low
260 values (avg. 69.4 ng m^{-3}) (Figs. 3a-c, 5a-d). Graham et al. (2003) also reported similar diurnal
261 variations of primary sugars for the Amazon forest aerosols. Sucrose was found as dominant
262 primary sugars (avg. 58.5 ng m^{-3}), accounting for 44% of measured primary sugars in Mangshan
263 aerosols (Table 1). Pollen was reported as a primary source for sucrose in aerosols collected from a
264 Texas rural site (Jia et al., 2010). Fu et al. (2012) found high sucrose concentrations up to 1390 ng
265 m^{-3} in the aerosols from Jeju Island, South Korea. Therefore, the plant materials, including pollen
266 spores from the local vegetation of Mangshan National Forest Park, are likely the primary source of
267 sucrose in the aerosols. Miyazaki et al. (2012) also reported higher sucrose concentrations in the
268 aerosol samples collected from the Hokkaido deciduous forest.

269 We found a significant diurnal variation of sucrose with higher daytime (82.9 ng m^{-3}) than
270 nighttime (12.3 ng m^{-3}). Meteorological parameters such as temperature, rainfall, wind speed, and
271 solar radiation significantly influence pollen activities and, subsequently, sucrose concentrations
272 (Verma et al., 2018). Interestingly, an elevated peak of sucrose was observed from 12h to 15h with
273 higher ambient temperature. In contrast, lower sucrose concentrations were observed from 15h to
274 9h with lower ambient temperature (Fig. 5a). Daytime increased concentrations of sucrose might be
275 related to the higher daytime ambient temperature, low RH, and high solar radiation (Miyazaki et
276 al., 2012). Taylor et al. (2002) reported the influence of the meteorological conditions, i.e., strong
277 daytime winds and convective activity, which can result in catapulting of pollen, opening of pollen-
278 laden flower anthers, and causing enhance entrainment and dispersal of the particles into the air.

279 Pacini (2000) reported that higher levels of sucrose in daytime coincide with higher counts of
280 pollen, fern spore, and insect. The positive linear correlations of sucrose with ambient temperature
281 ($r = 0.52$) and solar radiation ($r = 0.55$) further supported the influence of meteorological
282 parameters in the sucrose concentration (Table 2).

283 Five rain events were recorded during the campaign, i.e., 15th, 17th, 18th, and 26th September,
284 and 1st and 5th October (Fig. 2a). Pollens are significantly settled down by wet scavenging during
285 rain events because their sizes are large. A low concentration of sucrose was found from the
286 beginning of sampling to the morning of 20th September and from the afternoon of 26th September
287 to the end of the sampling campaign (Fig. 5a). In addition, the increased concentrations of sucrose
288 were found in the aerosol samples collected from 20th to 22nd September, and moderate
289 concentrations were observed after 23rd to the evening of 25th September during non-precipitation
290 events. Consequently, the pollens were significantly scavenged during wet precipitation and
291 washout effect from the atmosphere, resulting in lower sucrose concentrations at the earlier periods,
292 than later periods. In addition, Rogge et al. (2007) reported that surface soil dust and unpaved road
293 dust also contribute sucrose in the atmospheric aerosols. However, insignificant correlations
294 between sucrose and Ca^{2+} (daytime, $r = 0.32$; night time, $r = 0.37$) do not supports soil dust
295 contributions to sucrose in the Mangshan aerosols (Table 2).

296 Glucose was the second dominant primary sugar in the Mangshan aerosols. The average
297 concentrations of glucose and fructose were observed to be 40.0 ng m^{-3} and 20.1 ng m^{-3} ,
298 respectively (Table 1, Fig. 5b). The sampling site is characterized by the dense vegetation in the
299 Mangshan National Forest Park. Therefore, the nectars and fruits of vegetation (Baker et al., 1998),
300 plant debris (Medeiros et al., 2006) and pollens (Fu et al., 2012) in the forest significantly
301 contribute to glucose and fructose. The glucose levels are equivalent to that (50.1 ng m^{-3}) reported
302 from the Howland Experimental Forest site in the USA (Medeiros et al., 2006). Glucose and
303 fructose showed significant diurnal variations, whose concentrations were higher in daytime (44.2
304 ng m^{-3} and 23.9 ng m^{-3} , respectively) than nighttime (32.0 ng m^{-3} and 12.8 ng m^{-3} , respectively) in
305 Mangshan aerosols (Table 1, Figs. 3b, c; 5b, c). This diurnal variation could be involved with
306 emissions of pollens, fern spores, and other giant particles by strong winds (Graham et al., 2003;
307 Pacini, 2000). Similar trends of glucose and fructose were reported in the Amazon forest, being
308 coincided with plant fragments and insects (Graham et al., 2003). The autumn decay of vascular
309 plant leaves in the Mangshan forest may have contributed to the levels of glucose and fructose.

310 Although, the daytime southerly winds deliver anthropogenic air masses from megacities to
311 the sampling site. The daytime winds from the northeast direction (23.4%) also carry air masses

312 from the forest region, transporting primary sugars to the Mangshan site. However, 99.5% of the
 313 nighttime hours, the wind is shifted to northeasterly, i.e., in forest region (He et al., 2015), but the
 314 emissions of primary sugars at night in the form of plant fragments are lower than in daytime.
 315 Because the daytime ambient temperature and solar radiations significantly induce the emissions of
 316 sugar compounds in the forest site (Miyazaki et al., 2012). Therefore, low glucose and fructose
 317 levels were found at nighttime than daytime aerosols at the Mangshan site (Table 1, Fig. 3).
 318 Previous studies have reported lichens (Dahlman et al., 2003) and soil dust (Nolte et al., 2001;
 319 Rogge et al., 2007) as significant sources of both primary sugars. The concentration of glucose was
 320 insignificantly correlated with soil tracer (Ca^{2+}) in day ($r = 0.02$) and nighttime ($r = 0.27$), denying
 321 their soil dust contributions in Mangshan aerosol samples.

322 Trehalose in the environment is significantly controlled by the activities of bacteria, fungi,
 323 yeast, algae, invertebrates, and plant species, as well as suspended soil particles (Medeiros et al.,
 324 2006, Rogge et al., 2007). The average concentration of trehalose was found 14.3 ng m^{-3} (Table 1,
 325 Fig. 5d). Yttri et al. (2007) reported higher trehalose concentrations in the aerosol samples collected
 326 from urban (29 ng m^{-3}) and suburban (27 ng m^{-3}) than rural (3.8 ng m^{-3}) areas in Norway. The
 327 above results emphasize that fungi and microbes associated with anthropogenic and bioaerosols,
 328 emitted in the urban and suburban areas, might be responsible for the trehalose concentration in
 329 aerosol samples (Verma et al., 2018). Trehalose showed insignificant diurnal variation, whose day
 330 and night concentrations were observed 15.3 ng m^{-3} and 12.3 ng m^{-3} , respectively, indicating its
 331 different emission sources in day and night for Mangshan aerosols (Fig. 3b, c; 5d).

332 The southerly winds might transport fungi and microbes associated with bioaerosols, eject
 333 spores under favorable meteorological conditions (high RH and low temperature) (Jones and
 334 Mitchell et al., 1996). Several microbes and fungi discharge spores at nighttime due to high RH
 335 conditions (Ibrahim et al., 2011; Kim and Xiao, 2005; Malik and Singh, 2004; Sharma and Razak,
 336 2003). Interestingly, trehalose is more significantly correlated with arabitol and mannitol ($r = 0.76$
 337 and 0.85 , respectively) in nighttime than daytime ($r = 0.49$ and 0.51 , respectively) (Table 2),
 338 suggesting that fungal and microbial spores contributed to high levels of trehalose in nighttime.
 339 Hackl et al. (2000) found trehalose as dominant sugar in spring aerosols and proposed it as a tracer
 340 for soil dust particles. Trehalose concentration was more significantly correlated with Ca^{2+} ($r =$
 341 0.82) in daytime than nighttime ($r = 0.61$), indicating soil dust contribution (Table 2). Therefore, we
 342 hypothesized that winds transported soil particles from the urban area in daytime due to the active
 343 building constructions (He et al., 2015), contributing to the high levels of trehalose in daytime.

344 3.1.3. Ambient concentrations and diurnal variations of sugar alcohols

345 The average concentrations of sugar alcohols were found 75.8 ng m^{-3} , contributing 26.4%
346 of total SCs measured in Mangshan aerosols (Table 1). Sugar alcohols showed clear diurnal
347 variations in daytime high (avg. 87.4 ng m^{-3}) and nighttime low (avg. 53.7 ng m^{-3}) (Table 1).
348 Mannitol was found as the dominant sugar alcohol (avg. 44.1 ng m^{-3}), followed by arabitol (avg.
349 29.1 ng m^{-3}) and inositol (avg. 2.62 ng m^{-3}) (Table 1; Fig. 6a-c). Mannitol and arabitol are common
350 polyols detected in green algae, lichens, and fungal spores (Bieleski, 1995, Dahlman et al., 2003;
351 Filippo et al., 2013; Lewis and Smith, 1967; Yttri et al., 2007). Previous studies have reported that
352 arabitol and mannitol are key components of fungal spores, and thus they are considered as fungal
353 tracers (Bieleski, 1995; Lewis and Smith, 1967). Several fungal and microbial species released
354 spores during biological activities into the atmosphere (Dahlman et al., 2003; Bauer et al., 2008;
355 Filippo et al., 2013). Therefore, the autumn time fungal and microbial species significantly
356 contribute to arabitol and mannitol in the Mangshan aerosol samples.

357 However, mannitol and arabitol showed a strong positive linear correlation ($r = 0.81$), which
358 suggested common origins as reported in earlier studies (Fu et al., 2012) (Table 2). In contrast, the
359 higher concentration of mannitol than arabitol suggested its addition sources than fungal spores in
360 the Mangshan forest site. In this sequence, several previous studies have confirmed the significance
361 of mannitol in plant photosynthesis (Loescher et al., 1992; Keller and Matile, 1989; Rumpho et al.,
362 1983). Pashynska et al. (2002) reported that detritus of mature leaves can emit mannitol into the
363 atmosphere by wind action. Heald and Spracklen (2009) also found a correlation between the
364 atmospheric water vapor with mannitol concentrations and leaf area index. They suggested that the
365 activities of the terrestrial biosphere widely affect mannitol concentrations in the air. Our PMF
366 results also indicated the substantial contribution of mannitol for vegetation factor (24.8%), which
367 supports that mannitol is attributed by vegetation from the forest area (section 3.2).

368 In addition, the meteorological parameters, including high RH and temperature affect the
369 fungal and bacterial activities (Kim and Xiao, 2005; Sharma and Razak, 2003). The maximum
370 growth of fungi and bacteria was observed at 92–100% RH (Ibrahim et al., 2011). Interestingly, the
371 concentrations of arabitol and mannitol gradually increased after the end of precipitation, following
372 the increases in ambient temperature and RH (Figs. 2a, 6a, b). Miyazaki et al. (2012) also discussed
373 the increased contributions of arabitol and mannitol with daytime ambient temperature and solar
374 radiation in the aerosol samples collected from the forest area. Similar temporal trends and positive
375 linear correlations were observed between arabitol ($r = 0.69$) and mannitol ($r = 0.57$) with RH,
376 which supports the above phenomenon for Mangshan aerosols (Table 2). Therefore, we propose
377 that a favorable meteorological condition in autumn increases the emissions of fungal spores and

378 fragments of forest vegetation, which may be responsible for arabitol and mannitol contributions in
379 the Mangshan aerosols.

380 The diurnal variation of mannitol and arabitol were characterized by higher in the daytime
381 (51.7 ng m^{-3} and 32.5 ng m^{-3} , respectively) than nighttime (29.6 ng m^{-3} and 22.5 ng m^{-3} ,
382 respectively) (Fig. 3b, c). Yamaguchi et al. (2012) reported that fungal spores and bacterial cells
383 associated with bioaerosols could be transported long distances. The Mangshan site receives
384 significant anthropogenic and bioaerosols from Beijing City by southerly winds. Therefore, the
385 daytime plant activities, influenced by solar radiation and ambient temperature and the long-range
386 transport of fungal spores from megacities (Beijing) by southwest winds govern the diurnal
387 variation of sugar alcohols in the Mangshan atmosphere. On the other hand, lower concentrations in
388 nighttime can be explained by the clean air mass transport by mountain breeze from the Mangshan
389 National Forest area.

390 **3.2. Source apportionment of SCs**

391 To investigate the source apportionment of SCs, positive matrix factorization (PMF)
392 analysis was performed for the measured aerosol samples using tracer compounds for
393 anhydrosugars, primary sugars, and sugar alcohols. It is essential to select a suitable number of
394 factor solutions in the PMF analysis. Based on the possible sources of SCs, four to six factor
395 solutions were run in PMF model. In the four-factor solutions, the SCs, including arabitol,
396 mannitol, and trehalose, were merged in a single factor; this might underestimate the soil dust
397 sources. The SCs, including glucose, fructose trehalose, arabitol, and mannitol, were distributed in
398 more than four factors; it might be overestimated the number of factor solutions according to
399 possible sources of SCs. Therefore, a total of five interpretable factor solutions were characterized
400 by the enrichment of each tracer compound to be significant to categorize the origins of individual
401 sugars, which reproduced more than 95% of SCs.

402 These five-factor solutions were preferred based on minimum robust and true Q values
403 (goodness of fit parameters) of the base runs, which observed 3103 and 3505, respectively. In each
404 bootstrap run, the concentrations and percentages of tracers were close to those of base-run results.
405 The PMF results of SCs indicate a stability because no significant changes were found between Q
406 values and factor profiles of F_{peak} rotation runs compared with the base run. PMF results show a
407 good correlation between the values of observed and predicted (modeled) concentrations in scatter
408 plot, indicating that the model very well fits the individual sugar species. These results support the
409 perfect rationality of the source apportionment (Figure S-1). The time series plot of observed and

410 predicted concentration (modeled) also shown that the model well fits the observed data set (Figure
411 S-2). Figures 6 and 7 show the factor profile resolved by PMF analysis of the Mangshan aerosol
412 samples. The percentages of each component are summed for factors 1 to 5 to be calculated as
413 100%.

414 Factor 1 is characterized by the high contribution of glucose (80.2%) followed by fructose
415 (69.6%), mannitol (24.8%), and inositol (15.1%) (Fig. 7a). Glucose and fructose are highly water-
416 soluble SCs present in the leaves and bark of plants (Graham et al., 2003). High concentrations of
417 glucose and fructose have been reported in vascular plants and phytoplankton by Cowie and Hegdes
418 (1984). The dominant glucose and fructose in the Mangshan aerosol samples collected in autumn
419 are rational as leaf senescence and decay results in both primary sugars being released into the
420 atmosphere during the fall season. We found an excellent correlation between glucose and fructose
421 ($r = 0.94$) in the Mangshan aerosols (Table 2), indicating the similar vegetation sources for both
422 sugar species in autumn (Baker et al., 1998; Burshtein et al., 2011; Pacini, 2000). Higher
423 concentrations of glucose and fructose in the aerosol samples collected during the autumn season
424 are reasonable because leaf senescence and decay result in an increased emission of primary sugars
425 into the atmosphere.

426 Several studies have reported that plant species significantly contribute to mannitol in the
427 atmosphere (Burshtein et al., 2011; Devis et al., 1988; 1990). Miyazaki et al. (2014) also found a
428 significant amount of trehalose, mannitol, and arabitol in the aerosol samples collected from the
429 forest and concluded their origin from the terrestrial plants within the forest. Significant positive
430 linear correlations of mannitol with fructose in daytime ($r = 0.79$) and nighttime ($r = 0.86$) further
431 denote that abundance of mannitol is due to the decay of plant leaves in autumn (Table 2).
432 Therefore, we conclude that the contributions of mannitol is from both vegetation and fungal spores
433 in the Mangshan aerosol samples. Hence mannitol showed the presence in factor 1. Vegetations
434 contribute to SCs during the campaign. Therefore, factor 1 can be termed as a vegetation factor due
435 to the high abundances of glucose, fructose, and mannitol.

436 Factor 2 is dominated by high loading of trehalose (80.2%), followed by mannitol (29.7%),
437 glucose (19.8%), and arabitol (18.2%) (Fig. 7b). The contribution of trehalose to soil dust has been
438 reported in several studies from different locations around the world, suggesting trehalose as a
439 tracer for the surface soil (Jia et al., 2010; Medeiros et al., 2006). In addition, previous studies
440 reported that bacteria and other microbes in the soil are also an essential source of trehalose (Rogge
441 et al., 2007). Trehalose is significantly correlated with arabitol ($r=0.58$) and mannitol ($r=0.58$), and
442 Ca^{2+} ($r=0.70$), demonstrating its microbial and soil dust origin. Therefore, factor 2 can be termed as

443 microbial and soil dust factor.

444 Factor 3 is characterized by levoglucosan (82.2%), galactosan (77%), and mannosan
445 (73.6%) (Fig. 7c). Previous studies have reported that these SCs are associated with BB aerosols
446 (Fraser and Lakshmanan, 2000; Graham et al., 2002; Simoneit, 2002). Simoneit et al. (1999)
447 reported that the pyrolysis of cellulose and hemicellulose emitted levoglucosan, galactosan and
448 mannosan. These sugar species are major organic components emits in the atmosphere by BB
449 activities (Simoneit et al., 2002). The BB influenced aerosols are enriched with levoglucosan,
450 mannosan, and galactosan (Nolte et al., 2001; Medeiros et al., 2006). The domestic BB for cooking
451 and house heating due to low ambient temperature and field burning of agricultural residues occur
452 in East Asia (Verma et al., 2015). The PMF results are very well supported by the fact that
453 anhydrosugars are associated with BB (Simoneit et al., 1999). Therefore, factor 3 can be termed as
454 a BB factor due to the high abundance of BB products.

455 Factor 4 is dominated by high loading of sucrose (90%), followed by inositol (36.9%) and
456 fructose (11.7%) (Fig. 7d). Sucrose plays a crucial role in the plant blossoming process as the
457 dominant sugar compound of pollen grains (Pacini, 2000). Several studies also reported that sucrose
458 is abundant sugar species found in airborne pollen grains and flowering plants (Fu et al., 2012;
459 Graham et al., 2003; Medeiros et al., 2006; Pacini, 2000). Therefore, sucrose is reported as an
460 excellent tracer for airborne pollen spores (Pacini, 2000). Thus factor 4 is termed as pollen factor
461 due to the high loading of sucrose.

462 Factor 5 is characterized by a higher contribution of arabitol (61.5%) followed by mannitol
463 (39.3%) and inositol (15.3%) (Fig. 7e). Sugar species contributing to factor 5 are associated with
464 fungal spores (Bauer et al., 2008). Various fungi and microbes emit spores, which are tracers for the
465 arabitol and mannitol; therefore, both sugars are considered as specific tracers of fungal activities
466 (Medeiros et al., 2006; Rogge et al., 2007). Thus, factor 5 is termed as a fungal factor due to the
467 high loading of arabitol and mannitol. Overall, the average contributions of each factor to measured
468 SCs were estimated by PMF analyses (Fig. 8), in which BB was found to account for 27% of
469 measured SCs. The vegetation and microbial and soil dust sources equally contribute (21%) to total
470 SCs. The fungal spores and pollen spores contribute 16% and 15% of total SCs, respectively.
471 Finally, biomass burning emissions from the local areas and megacities via long-range atmospheric
472 transport were identified as an important source for the Mangshan aerosols.

473 3.3. Contributions of sugar compounds to WSOC and OC

474 The contribution of carbon content of measured SCs varied from 14.1-371 ng m⁻³ (av. 145

475 ng m⁻³) in daytime and 12.8-322 ng m⁻³ (av. 117 ng m⁻³) in nighttime, accounting for 0.83% and
 476 0.91% of OC, respectively (Fig. 9a, b). The mean carbon contents of anhydrosugars showed clear
 477 diurnal variation with higher nighttime values (67.1 ng m⁻³) than daytime (42.7 ng m⁻³), accounting
 478 for 0.43 % and 0.22 % of OC, respectively. These results suggest that BB significantly contributed
 479 to Mangshan aerosols. However, the carbon contents of primary sugars showed opposite diurnal
 480 variations; higher (68.5 ng m⁻³) in daytime than nighttime (28.3 ng m⁻³), accounting for 0.41 % and
 481 0.28 % of OC, respectively (Fig. 9a, b). This study suggests that the daytime emissions of primary
 482 sugars from local vegetation and the decay of plant leaf in forest significantly contribute to OC. The
 483 carbon concentration contributed by sugar alcohols showed insignificant diurnal variations i.e. 34.6
 484 ng m⁻³ in daytime and 21.3 ng m⁻³ in nighttime, accounting for 0.20 % and 0.19 % of OC,
 485 respectively. This result indicates multiple carbon sources of sugar alcohols in day and night. In
 486 addition, contributions of anhydrosugars, primary sugars, and sugar alcohols to WSOC were similar
 487 to those of OC in Mangshan aerosols.

488 Based on the PMF analysis, we found five sources for SCs measured in Mangshan aerosols.
 489 The different tracer compounds were used to calculate carbon contents: biomass burning-C (i.e.,
 490 levoglucosan, galactosan, mannosan), vegetation-C (glucose, fructose), fungal-C (arabitol,
 491 mannitol), pollen-C (sucrose), and microbial-soil-C (trehalose) (Fig. 9c, d). Among the five sources,
 492 biomass burning-C was found as the largest carbon contributor to Mangshan aerosols (36.7%),
 493 followed by fungal-C (23.7%), vegetation-C (19.7%), pollen-C (14.2%), and microbial-soil-C
 494 (4.84%). Biomass burning-C accounted for 1.38% and 0.43% at night, while 0.57% and 0.22% in
 495 daytime for WSOC and OC, respectively. The BB for cooking and space heating in winter and
 496 autumn seasons are common in central China (Akagi et al., 2011), which should increase the
 497 nighttime levels of Biomass burning-C at the Mangshan site. However, the carbon contribution by
 498 vegetation and fungal sources are similar during day and nighttime for the Mangshan aerosols.
 499 Pollen-C accounted for 0.20% and 0.07% of OC in daytime and nighttime, respectively. Higher
 500 pollen activities are key sources for the high daytime levels of pollen-C in the forest site (Taylor et
 501 al., 2002).

502 3.4. Contribution of levoglucosan to OC and WSOC

503 We calculated the mass concentration ratios of levoglucosan to OC (Lev/OC) and WSOC
 504 (Lev/WSOC) to evaluate the contributions of BB and anthropogenic emissions to Mangshan
 505 aerosols (Fig. 9a-c). Fossil fuel combustion and BB emit WSOC and OC. They are also secondarily
 506 produced by photochemical oxidation of volatile organic compounds in the atmosphere (Wang et
 507 al., 2005; Deshmukh et al., 2019b). Coal combustion and vehicle exhaust can contribute to the high

508 levels of OC and WSOC in aerosols (Xu et al., 2020). Levoglucosan, a dominant constituent of BB
509 products, has been considered as an excellent tracer of BB (Simoneit, 2002; Kuo et al., 2011).

510 Average Lev/OC ratio (5.69×10^{-3}) was lower than that of Lev/WSOC (1.66×10^{-2}) in
511 Mangshan samples (Fig. 10a). Yan et al. (2018) reported similar ratios of Lev/OC (4.0×10^{-3}) and
512 Lev/WSOC (1.6×10^{-2}) for coal combustion, suggesting a significant carbon contribution to
513 Mangshan aerosols from coal combustions in the industrial areas via long range transport.
514 Interestingly, we found a substantial diurnal variation of Lev/OC and Lev/WSOC ratios. The
515 average Lev/OC and Lev/WSOC ratios are several times higher in nighttime (8.48×10^{-3} and
516 2.70×10^{-2} , respectively) than daytime (4.21×10^{-3} and 1.11×10^{-2} , respectively) (Fig. 10b, c). These
517 results indicate that BB contributed substantially to the Mangshan organic aerosols in nighttime.
518 Moreover, the correlations of levoglucosan with OC and WSOC are stronger in nighttime ($r = 0.81$
519 and 0.70 , respectively) than daytime ($r = 0.45$ and 0.40 , respectively), demonstrating the dominance
520 of BB-derived aerosols in the nighttime Mangshan samples (Table 2).

521 In addition, WSOCs are derived from various emission sources. We propose that secondary
522 organic aerosols constitute a significant fraction of WSOC and OC in daytime Mangshan aerosols.
523 The photochemical oxidation of organic precursors emitted from fossil fuel combustion in
524 industries and vehicular exhausts also contributes to secondary production of WSOC and OC in
525 daytime (He et al., 2015), suggesting that emissions from the urban Beijing area may significantly
526 influence the daytime levels of Mangshan aerosols. He et al. (2015) proposed a possible
527 contribution of photochemical formation of secondary organic aerosols to atmospheric WSOC and
528 OC in north China. Nevertheless, the photochemical degradation of levoglucosan by OH radicals
529 under ultraviolet radiations and high temperatures (Hennigan et al., 2010) may play a key role in
530 lowering the ratios of Lev/OC and Lev/WSOC in daytime Mangshan aerosols.

531 3.5. Mass concentration ratios of levoglucosan/mannosan

532 The mass concentration ratios of levoglucosan and mannosan (Lev/Man) were calculated to
533 better characterize the emissions sources of BB tracers (softwood vs. hardwood) in the Mangshan
534 site. Figure 10d represents the variations of Lev/Man ratios for overall, day- and nighttime periods.
535 The Lev/Man ratios have been used to distinguish the hardwood (angiosperm) and softwood
536 (gymnosperm) burning in the ice core record from the Russian Far East (Kawamura et al., 2012).
537 Hardwood contains 55–65% cellulose and 20–30% hemicellulose (Klemm et al., 2005).
538 Levoglucosan and mannosan are derived from the thermal decomposition of cellulose and
539 hemicelluloses, respectively (Simoneit, 2002). Levoglucosan is thermally more stable than
540 mannosan and galactosan (Kuo et al., 2011). Hence, a lower Lev/Man ratio is associated with

541 softwood burning, whereas a higher ratio is associated with hardwood burning (Engling et al., 2006,
542 2009). However, we found insignificant diurnal variations of Lev/Man ratios between night (9.33-
543 25.9, avg. 15.8) and daytime aerosols (0.90-23.3, avg. 13.6). Likewise, comparable Lev/Man ratios
544 (9-13 for PM₁₀ and 10-13 for PM_{2.5}) were reported for aerosol samples from Tanzania, where wood
545 and charcoal are primary fuels used for domestic cooking and heating (Mkoma and Kawamura,
546 2013). Interestingly, wheat straws and lignite are used in China for domestic cooking and house
547 heating, which may also contribute to levoglucosan and mannosan in the Mangshan aerosols.

548 Different Lev/Man ratios were reported in the chamber and controlled field experiments,
549 e.g., 4-22 for conifer and savanna grass burning (Iinuma et al., 2007), and 41.6 for rice straw and
550 and 55.7 cereal straw burning (Engling et al., 2009; Zhang et al., 2007). Kuo et al. (2011) reported
551 higher emissions of levoglucosan during high-temperature flaming (27.5-52.3) compared to low-
552 temperature smoldering (2.43-3.08). Hence, it is not easy to differentiate hardwood and softwood
553 burning based on Lev/Man ratios alone. Several studies reported a high Lev/Man ratio for both
554 softwood and hardwood burning. Thus, there may exist some other factors that significantly control
555 the Lev/Man ratios. Yan et al. (2018) found a significant contribution of levoglucosan in coal
556 combustion with Lev/Man ratio of 7.2. The variations of Lev/Man ratios in Mangshan may be
557 significantly influenced by several factors, i.e., flaming vs. smoldering, duration of biomass
558 burning, coal combustion, and hardwood vs. softwood burning. The moderate Lev/Man ratios in
559 autumn aerosols from Mangshan suggest that low temperature smoldering processes of hardwood
560 contribute to levels of levoglucosan and mannosan. However, the contribution of coal combustions
561 for house heating could not be excluded.

562 **4. Summary and Conclusions**

563 Anhydrosugars, primary sugars, and sugar alcohols were detected with distinct diurnal
564 variations in suburban aerosol samples collected at the Mangshan site in the northern vicinity of
565 Beijing. The wind patterns indicate that daytime air masses were transported from urban Beijing to
566 Mangshan, while clean air masses were delivered in nighttime from the Mangshan National Forest
567 Park. Daytime air masses from urban Beijing significantly influence the air quality of the northern
568 forest region. We observed the highest abundance of primary sugars, followed by anhydrosugars
569 and sugar alcohols. Local emissions from the forest plants and fungal species are the main
570 contributors to the primary sugars and sugar alcohols in the Mangshan aerosols. The meteorological
571 parameter significantly influenced the levels of SCs in the Mangshan samples. We observed a
572 significant influence of enhanced ambient temperature and solar radiation on the pollen rupture and
573 increased RH on fungal and microbial growth. This study suggested the source variation for

574 trehalose, i.e., local microbes at night and soil dust particles transported from Beijing areas by
575 southerly wind in daytime. We found that vegetation and fungal spores are not a specific source of
576 glucose and mannitol, respectively. Both sugars may have multiple sources in the forest aerosols.

577 PMF results concluded the contributions of 36% from vegetation (21% vegetation factor and
578 15% pollen factor) and 37% from microbial and fungal species (21% microbial soil dust and 16%
579 fungal factor). The BB activities for domestic cooking and space heating in north China contributed
580 higher organic carbon at nighttime (0.43%) than in daytime (0.22%). Therefore, local BB seriously
581 affected the air quality of the Mangshan site. Lev/Man ratio suggested that low temperature
582 smoldering burning of hardwood is the main source for BB aerosols. SCs were recognized as a
583 significant aerosol component at Mangshan, northern suburbs of Beijing. SCs can influence the air
584 quality and thus climate because they are essential components of organic aerosols on a global
585 scale. This study of SCs at Mangshan demonstrates that ambient levels of SCs are highly sensitive
586 to the emissions of anthropogenic and biogenic aerosols. Higher contribution of levoglucosan to
587 SCs demonstrated a significant BB activity around the Mangshan site in north China.

588

589 **Data availability.** Raw data are available on request by contacting the corresponding author.

590 **Author contributions.** This research was designed YK, KK and ZW. Laboratory measurements
591 were performed by FY with a support of PF. The paper was prepared by SKV and KK.

592 **Competing interests.** The authors declare that they have no conflict of interest.

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891 **Figure Captions**

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893 Figure 1. Geographical location of Mangshan, China. The map was downloaded from © Google
894 Maps 2019.

895 Figure 2. (a) The meteorological parameters at Mangshan during sampling periods, (b) Fractions of
896 local wind directions at Mangshan site, north of Beijing, China.

897 Figure 3. Concentrations (ng m^{-3}) of sugar compound (a) overall, (b) daytime and (c) nighttime in
898 aerosol samples from Mangshan during September-October 2007 (The error bars denote the
899 standard deviation).

900 Figure. 4. Temporal variations in the concentrations (ng m^{-3}) of anhydrosugars in the Mangshan
901 aerosol samples collected for September-October 2007. (Solid circle represents nighttime
902 samples collected from 18:00 to 09:00 hours. Hollow circle represents daytime samples).

903 Figure. 5. Temporal variations in the concentrations (ng m^{-3}) of primary sugars in the Mangshan
904 aerosol samples collected for September-October 2007. (Solid circle represents nighttime
905 samples collected from 18:00 to 09:00 hours. Hollow circle represents daytime samples).

906 Figure. 6. Temporal variations in the concentrations (ng m^{-3}) of sugar alcohols in the Mangshan
907 aerosol samples collected for September-October 2007. (Solid circle represents nighttime
908 samples collected from 18:00 to 09:00 hours. Hollow circle represents daytime samples).

909 Figure 7. PMF analyses of sugar compounds in Mangshan aerosols based on the autumn 2007 data
910 set.

911 Figure 8. Source contributions to sugar compounds from various sources based on PMF analyses.

912 Figure 9. The concentrations and relative contributions of the carbon content of anhydrosugars,
913 primary sugars and sugar alcohols to the carbon concentrations of measured sugar compounds,
914 water-soluble organic carbon (WSOC) and organic carbon (OC) fraction of Mangshan aerosols
915 (a = daytime and b = nighttime). The concentrations and relative contribution of the carbon
916 content of five sources of sugar compounds to total sugar compounds measured, WSOC and
917 OC fraction of Mangshan aerosols (c = daytime and d = nighttime).

918 Figure 10. Mass concentrations ratio of carbon contents of (a) levoglucosan (Lev) to organic carbon
919 (OC) and water soluble organic carbon (WSOC), (b) levoglucosan (Lev) to organic carbon
920 (OC) daytime and night time, (c) levoglucosan (Lev) to water soluble organic carbon (WSOC)
921 daytime and night time, (d) average levoglucosan to mannosan ratios (Lev/Man) in the
922 Mangshan aerosol samples for autumn 2007.

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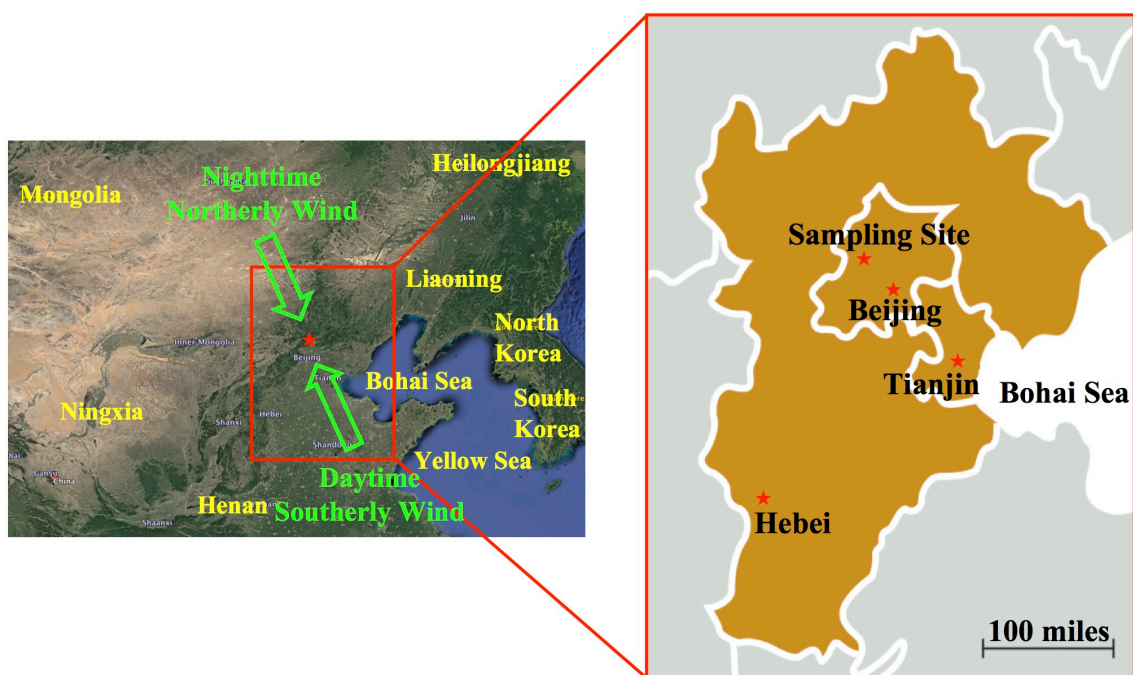
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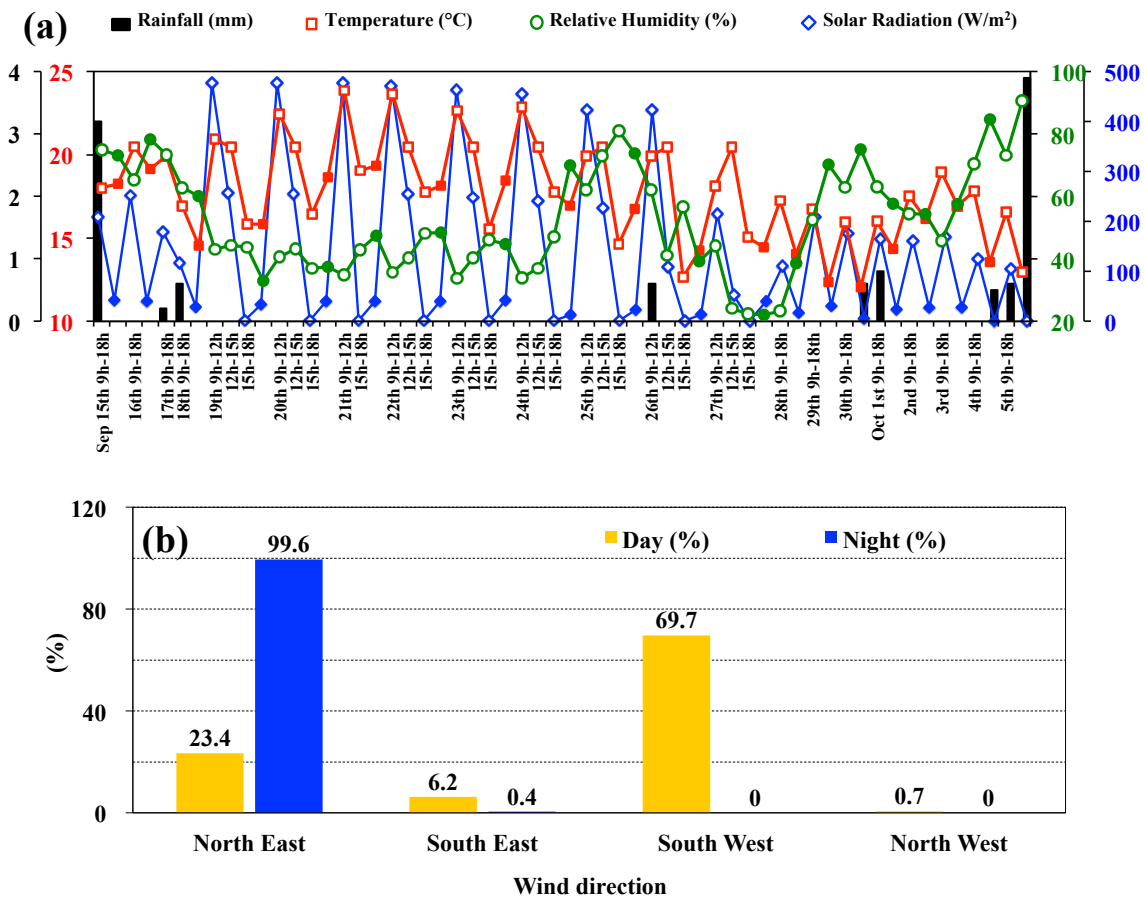
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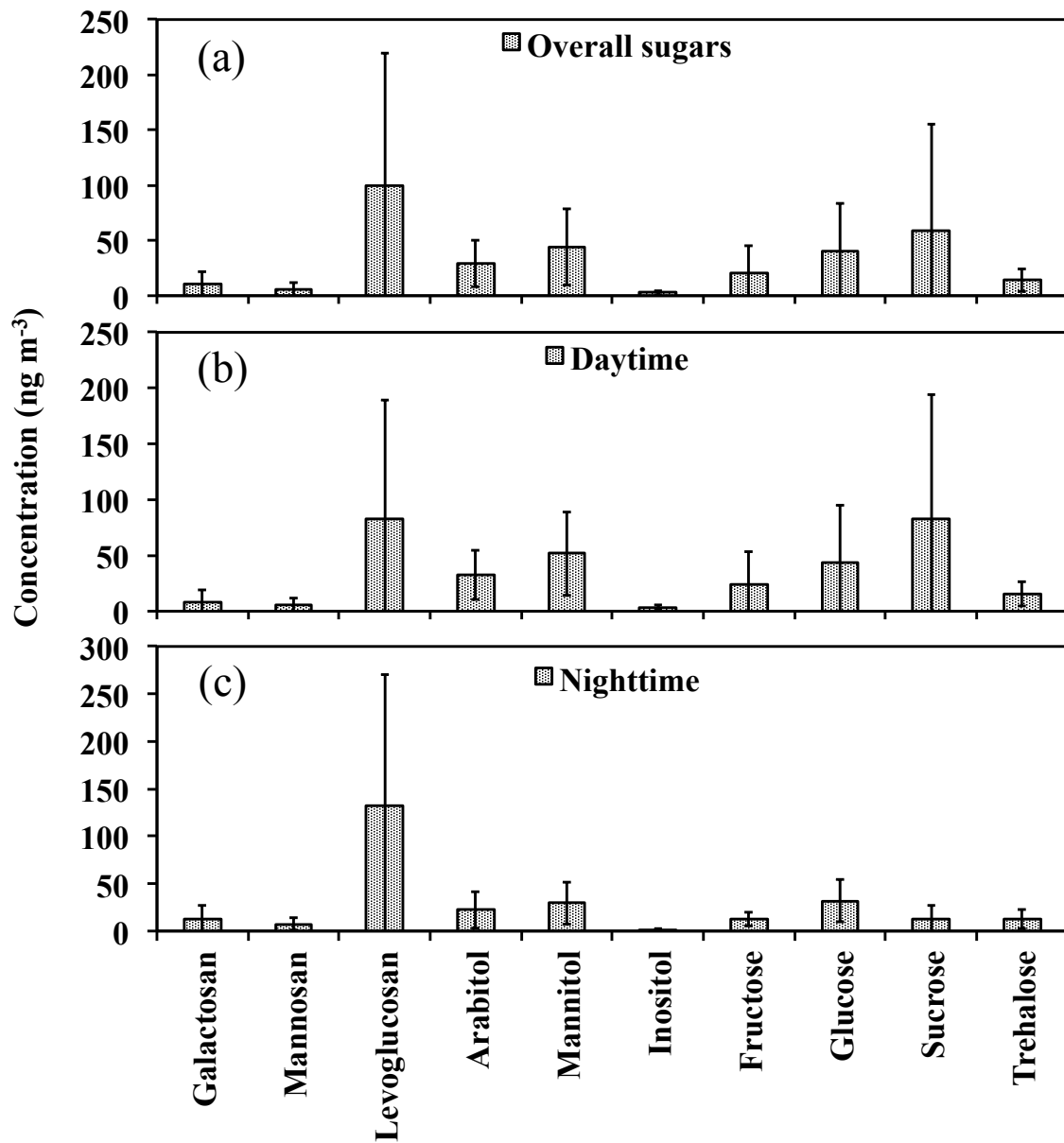
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Fig. 3.



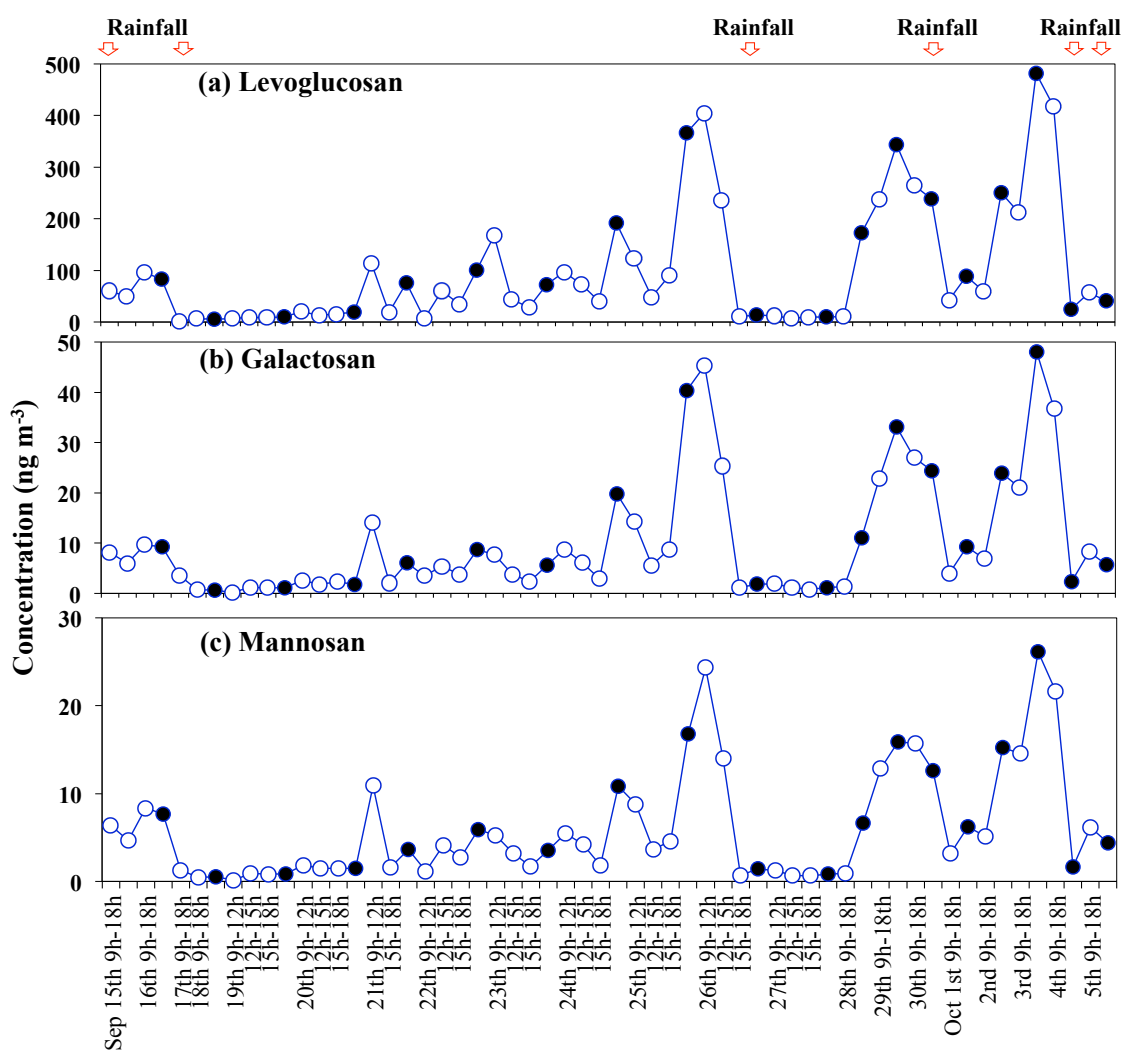
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Fig. 5.

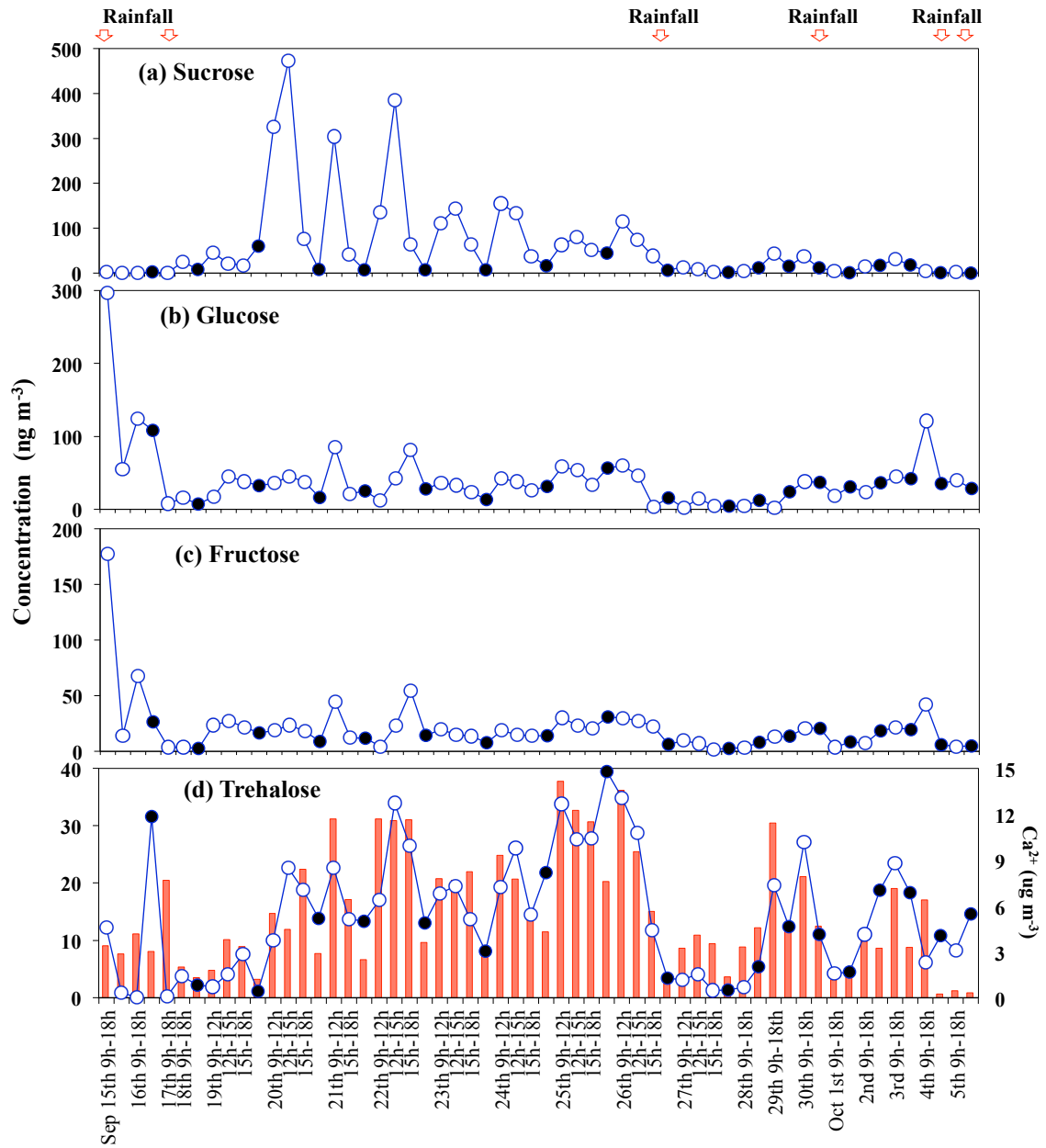
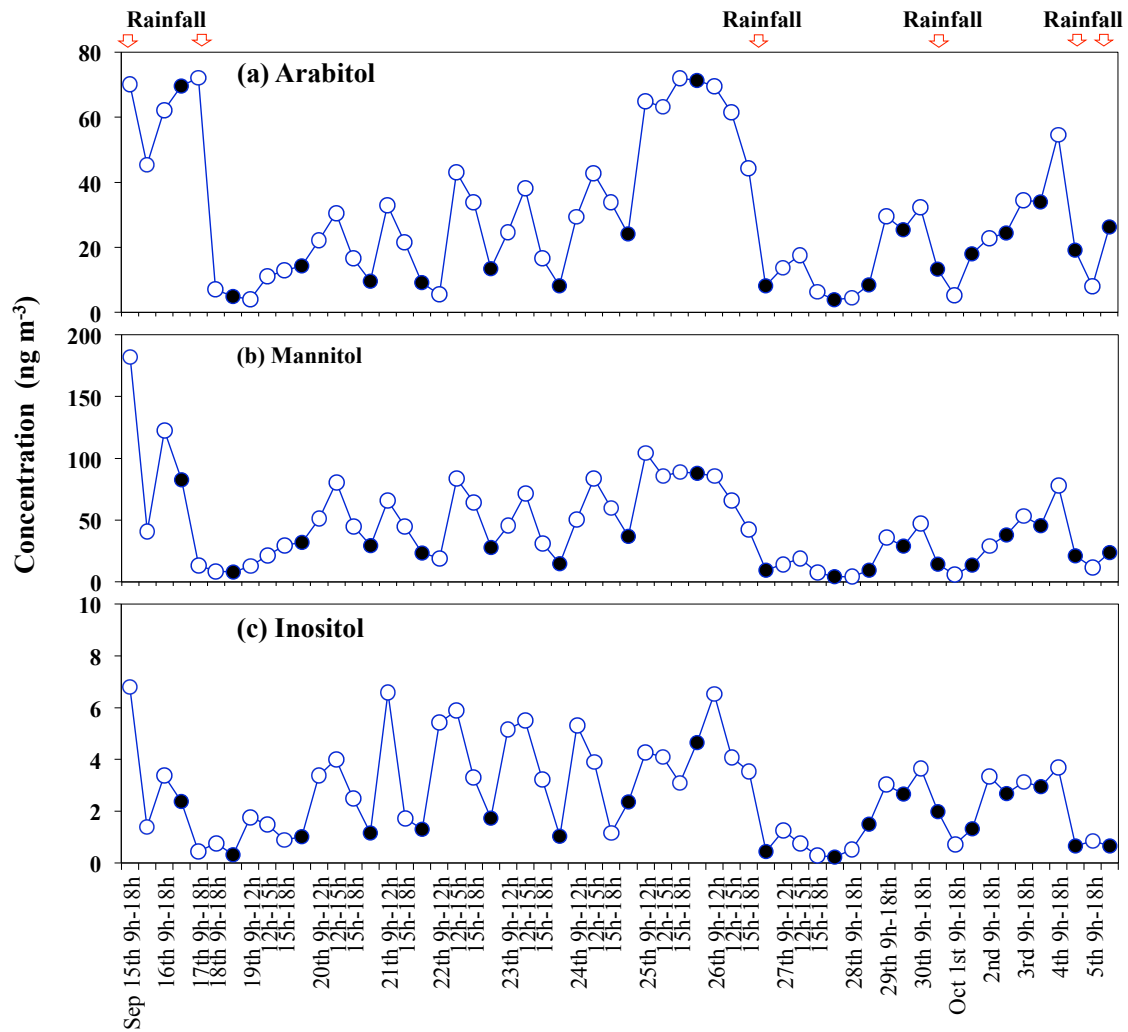
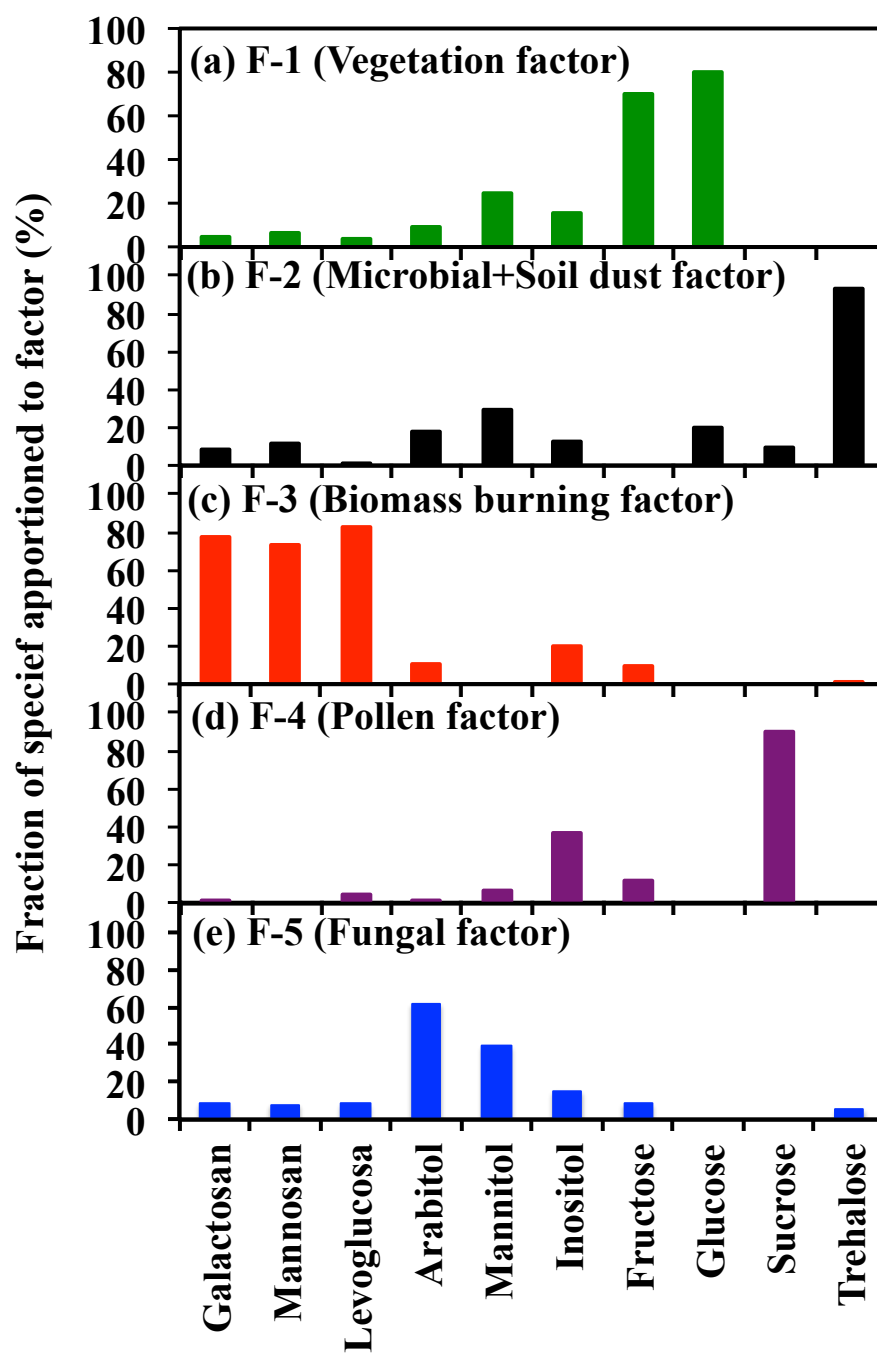


Fig. 6.



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Fig. 7.



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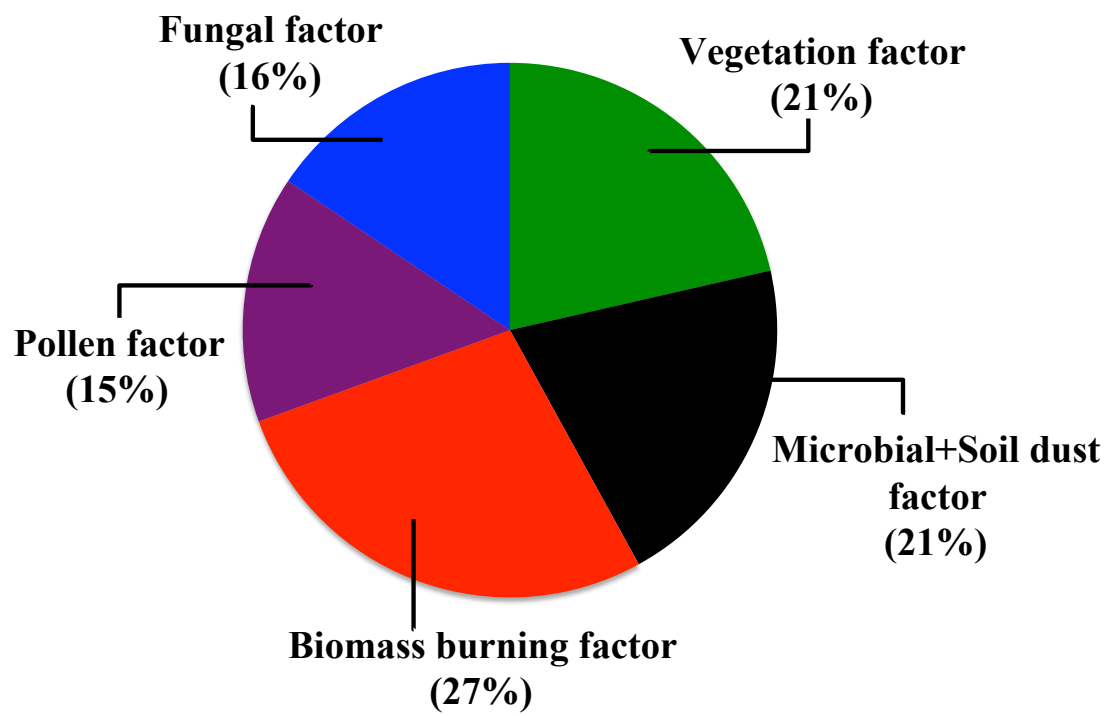
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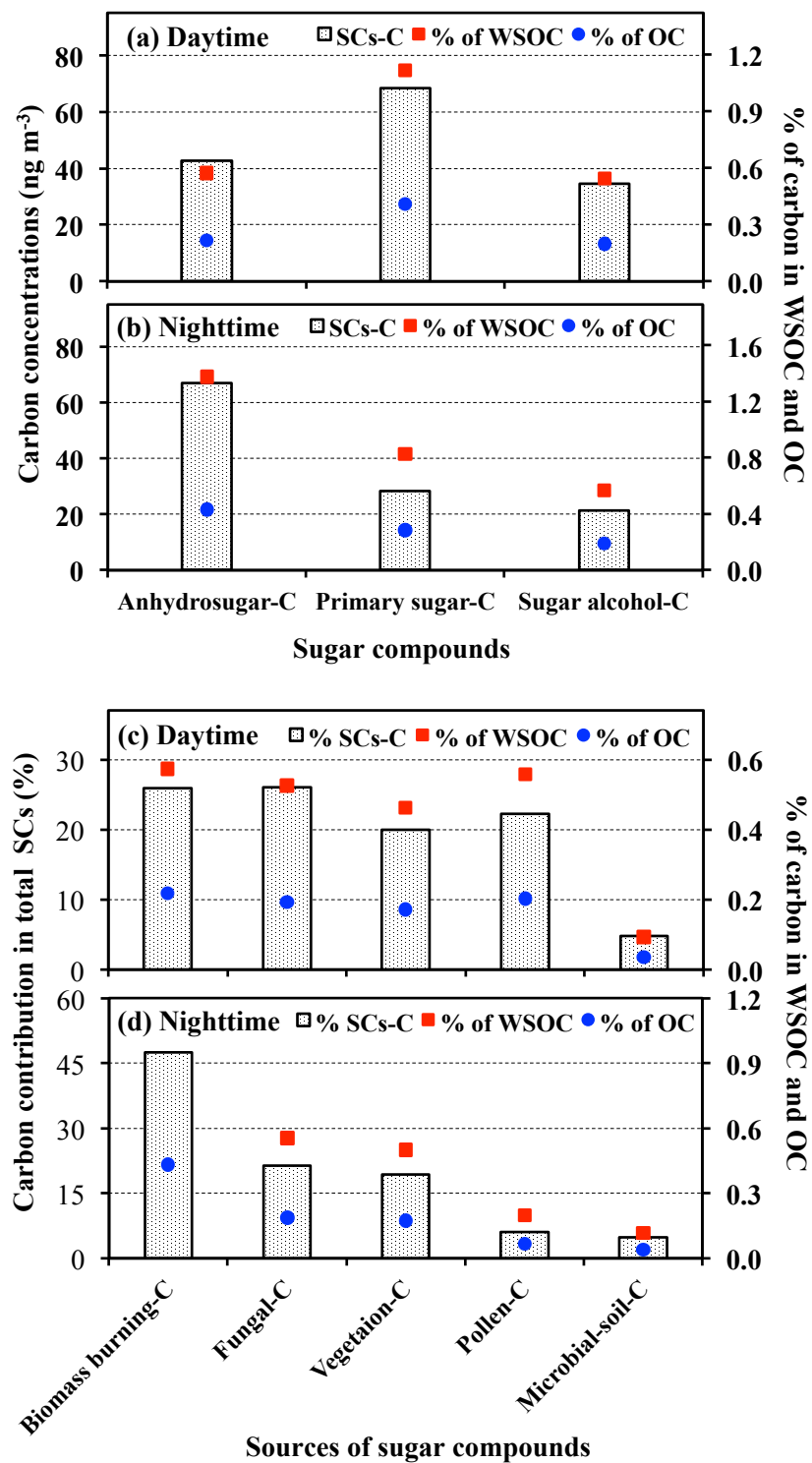
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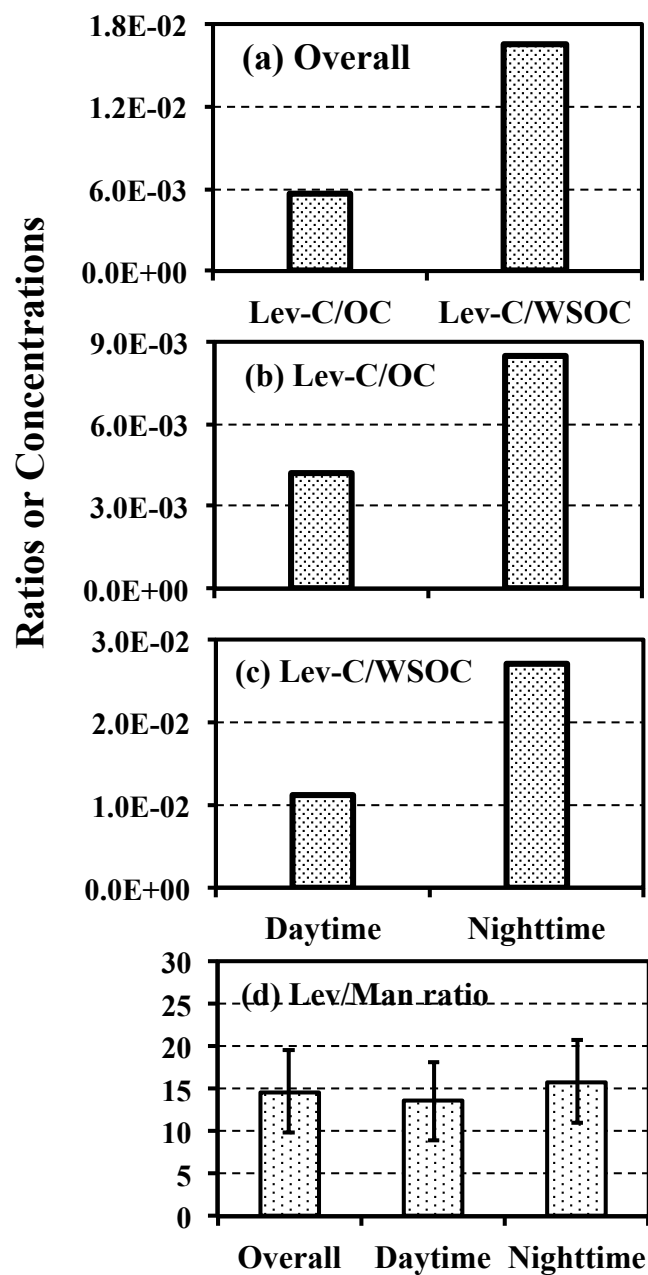


Table 1. Minimum, maximum, average and standard deviations of concentrations of sugar compounds in aerosol samples (TSP) from Mangshan, China.

Sugar Compounds	Overall				Daytime (n = 38)				Nighttime (n = 20)			
	Min	Max	Avg.	S.D.	Min	Max	Avg.	S.D.	Min	Max	Avg.	S.D.
Anhydrosugars												
Galactosan	0.14	48.0	10.1	11.9	0.14	45.3	8.53	10.5	0.69	48.0	13.0	14.0
Mannosan	0.13	26.1	6.05	6.33	0.13	24.3	5.37	6.01	0.53	26.1	7.35	6.87
Levogluconan	1.17	482	100	119	1.17	418	83.2	106	5.66	482	132	138
Sugar alcohols												
Arabitol	3.89	72.2	29.1	21.5	3.99	72.2	32.5	22.0	3.89	71.3	22.5	19.4
Mannitol	4.19	182	44.1	34.5	4.19	182	51.7	37.5	4.40	87.7	29.6	22.3
Inositol	0.23	6.8	2.62	1.81	0.27	6.80	3.14	1.90	0.23	4.65	1.62	1.09
Primary sugars												
Fructose	1.72	177	20.1	24.6	1.72	177	23.9	29.3	2.64	30.9	12.8	7.67
Glucose	1.86	297	40.0	43.4	1.86	297	44.2	50.8	4.52	108	32.0	22.8
Sucrose	0.02	474	58.5	96.5	0.02	474	82.9	112	0.04	60.1	12.3	15.1
Trehalose	0.06	39.5	14.3	10.5	0.06	34.9	15.3	10.6	0.87	39.5	12.3	10.2
Anhydrosugars	6.01	556	116	137	6.01	476	97.1	122	6.88	556	152	159
Primary sugars	9.41	565	133	125	9.41	565	166	141	10.5	172	69.4	43.0
Sugar alcohols	8.53	259	75.8	54.7	9.09	259	87.4	57.5	8.53	164	53.7	41.9
Total Sugars	30.8	875	325	232	34.1	875	351	240	30.8	759	276	212
Anhydrosugars (%)			31.9				24.6				45.7	
Primary sugars (%)			41.8				47.3				31.3	
Sugar alcohols (%)			26.4				28.1				23.0	

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Table 2. Statistical summary of correlations among the chemical species and meteorological variables in aerosol samples collected at a forest site in northern Japan.

Linear regression	Correlation coefficient	p value	Significance of correlation at P value < 0.05
Overall (n = 58)			
Levogluconan vs. Galactosan	0.98	< 0.05	Significant
Levogluconan vs. Mannosan	0.97	< 0.05	Significant
Mannosan vs. Galactosan	0.98	< 0.05	Significant
Sucrose vs. Temperature	0.52	< 0.05	Significant
Sucrose vs. Solar radiation	0.55	< 0.05	Significant
Arabitol vs. Mannitol	0.81	< 0.05	Significant
Arabitol vs. RH	0.69	< 0.05	Significant
Mannitol vs. RH	0.57	< 0.05	Significant
Glucose vs. Fructose	0.94	< 0.05	Significant
Trehalose vs. Arabitol	0.58	< 0.05	Significant
Trehalose vs. Mannitol	0.58	< 0.05	Significant
Trehalose vs. Ca ²⁺	0.70	< 0.05	Significant
Daytime (n = 38)			
Sucrose vs. Ca ²⁺	0.32	> 0.05	Not significant
Glucose vs. Ca ²⁺	0.02	> 0.05	Not significant
Trehalose vs. Arabitol	0.49	< 0.05	Significant
Trehalose vs. Mannitol	0.51	< 0.05	Significant
Trehalose vs. Ca ²⁺	0.81	< 0.05	Significant
Fructose vs. Mannitol	0.79	< 0.05	Significant
Levogluconan vs. OC	0.45	< 0.05	Significant
Levogluconan vs. WSOC	0.40	< 0.05	Significant
Nighttime (n = 20)			
Sucrose vs. Ca ²⁺	0.37	> 0.05	Not significant
Glucose vs. Ca ²⁺	0.27	> 0.05	Not significant
Trehalose vs. Arabitol	0.76	< 0.05	Significant
Trehalose vs. Mannitol	0.85	< 0.05	Significant
Trehalose vs. Ca ²⁺	0.61	< 0.05	Significant
Fructose vs. Mannitol	0.86	< 0.05	Significant
Levogluconan vs. OC	0.81	< 0.05	Significant
Levogluconan vs. WSOC	0.70	< 0.05	Significant

The data of Ca²⁺, OC and WSOC are adapted from He et al. (2015).

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