

1 **Measurement report: Diurnal and temporal variations of sugar compounds in suburban**  
2 **aerosols from the northern vicinity of Beijing, China: An influence of biogenic and**  
3 **anthropogenic sources**

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18 Key points:

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- 20 1. Autumn time observations of sugar compounds (SCs) in the northern vicinity of Beijing, China.
- 21 2. Influence of natural biogenic emissions on SCs from forest area.
- 22 3. Influence of anthropogenic and bioaerosol on SCs from the Beijing area.
- 23 4. Biomass burning is a significant contributor to SCs.
- 24 5. Biogenic and fungal-microbial emissions are significant sources for mannitol and arabinol.

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27

28 **Abstract**

29            Sugar compounds (SCs) are major water-soluble constituents in atmospheric aerosols. In  
30 this study, we investigated their molecular compositions and abundances in the northern receptor  
31 site (Mangshan) of Beijing, China, to better understand the contributions from biogenic and  
32 anthropogenic sources using a gas chromatography–mass spectrometry technique. The sampling  
33 site receives anthropogenic air mass transported from Beijing by southerly winds, while northerly  
34 winds transport relatively clean air mass from the forest areas. Day- and nighttime variations were  
35 analyzed for anhydrosugars, primary sugars, and sugar alcohols in autumn 2007. We found that  
36 biomass burning (BB) tracers were more abundant in nighttime than daytime, while other SCs  
37 showed different diurnal variations. Levoglucosan was found as a dominant sugar among the SCs  
38 observed, indicating an intense influence of local BB for cooking and space heating at the  
39 surroundings of the Mangshan site. The high levels of arabitol and mannitol in daytime suggest a  
40 significant contribution of locally emitted fungal spores and long-range transported bioaerosols  
41 from the Beijing area. The plant emissions from Mangshan forest park significantly control the  
42 diurnal variations of glucose, fructose, and mannitol. The meteorological parameters (relative  
43 humidity, temperature, and rainfall) significantly affect the concentrations and diurnal variations  
44 of SCs. Sucrose (pollen tracer) showed a clear diurnal variation, peaking in the daytime due to  
45 higher ambient temperature and wind speed, which influences the pollen release from the forest  
46 plants. We found the contribution of trehalose from soil dust in daytime, while microbial and  
47 fungal spores were responsible for nighttime. Anhydrosugar and primary sugars are prime carbon  
48 sources of the Mangshan aerosols. The high ratios of levoglucosan in organic carbon and water  
49 soluble organic carbon in nighttime suggest a significant contribution of BB to organic aerosols at  
50 night. Levoglucosan/mannosan ratios demonstrate that low temperature burning of hardwood is  
51 dominant in Mangshan. The positive matrix factorization analysis concluded that forest  
52 vegetation, fungal species, and local BB are the significant sources of SCs.

53 **Keywords:** Anthropogenic bioaerosols, biomass burning, pollen tracer, fungal tracers, soil dust,  
54 and microbial tracers

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## 56 **1. Introduction**

57 Increased economic growth and massive consumption of fossil fuels from industries emit  
58 anthropogenic gases, aerosols, and biomass burning (BB) products cause severe air pollution in  
59 East Asian countries (Lelieveld et al., 2015; Lin et al., 2014; Kawamura et al., 2013; Li et al.,  
60 2010; Sun et al., 2016). Globally, significant anthropogenic and carbonaceous aerosols are  
61 contributed by China (Cooke et al., 1999, Wang et al., 2007). Beijing is situated in the northern  
62 part of China, with a 20 million people and 5 million motor vehicles. Beijing is one of the largest  
63 polluted cities in East Asia; its air quality deteriorates seriously due to massive emissions of  
64 anthropogenic aerosols from vehicles and industries (Cao et al., 2014; Qiao et al., 2018; Tao et al.,  
65 2017; Wei et al., 2018; Yu et al., 2013). Organic aerosols (OAs) are composed of a complex  
66 mixture of diverse molecules (Xu et al., 2011). They play essential roles in global climate changes  
67 via the modification of radiative forcing and cause a serious negative impact on human health  
68 (Fuzzi et al., 2007). OAs contain various water-soluble organic compounds, which can act as  
69 cloud condensation nuclei (CCN) (Kanakidou et al., 2005).

70 BB is essentially a primary source of OAs, controlling the air quality levels and affecting  
71 the earth's radiative forcing by scattering or absorbing incident solar radiation (Deshmukh et al.,  
72 2019a; Kanakidou et al., 2005; Kanaya et al. 2013; Streets et al., 2003; Sullivan et al., 2008).  
73 There are several kinds of BB, including industrial biofuel burning, open field burning (fires of the  
74 forest, peatlands, and agricultural wastes), and domestic BB burning for house heating and  
75 cooking, which emits BB products into the atmosphere (Akagi et al., 2011). The BB aerosols are  
76 subjected to long-range atmospheric transport once they are emitted into the atmosphere (Verma et  
77 al., 2015). Levoglucosan (1,6-anhydro- $\beta$ -D-glucopyranose) is a pyrolysis product of cellulose and  
78 hemicellulose, which is generally found as major organic constituents in the BB-influenced  
79 aerosols (Simoneit et al., 1999; 2002). Levoglucosan have been reported as a specific tracer for  
80 BB aerosols (Engling et al., 2009).

81 Sugar compounds (SCs) are ubiquitous in the atmosphere from different geographical  
82 locations, including urban, forest, marine, and polar regions (Burshtein et al., 2011; Fu et al.,  
83 2010; Wan et al., 2017). SCs are emitted from algae, microbes, pollen, suspended soil particle, and  
84 associated biota into the atmosphere by various processes, and thus they are termed as primary  
85 biological aerosol particles (PBAPs) (Carvalho et al., 2003; Despres et al., 2012; Elbert et al.,  
86 2007). The detailed study of bio-aerosols has been emphasized in the past decades due to the  
87 global impact of microbes and fungi because they can travel long distances from the source  
88 regions by winds (Burshtein et al., 2011; Brown and Hovmoller, 2002; Yamaguchi et al., 2012).

89 Fungi are essential microbes in the ecosystem, which discharge spores of 8-186 Tg yr<sup>-1</sup> into the  
90 atmospheric environment (Elbert et al., 2007; Heald and Spracklen, 2009). Sugar alcohols like  
91 arabitol and mannitol are enriched in fungal spores; thus, they are considered as specific tracers  
92 (Bauer et al., 2008).

93 Devis et al. (1988, 1990) reported that mannitol was also found in about 70 different  
94 higher plant families. Loescher et al. (1992) reported that mannitol is an important photosynthetic  
95 product converted by biosynthesis in plants. Keller and Matile (1989) also found the arabitol and  
96 mannitol during the increased photosynthesis in growing vegetation. Pollens are the largest  
97 particles that could contribute up to 65% of the PBAPs, which are the significant sources for  
98 sucrose and fructose in the forest aerosols (Manninen et al., 2014; Pacini, 2000). Higher plants  
99 synthesize primary sugars (glucose, fructose, and sucrose) during photosynthesis, which are  
100 circulated by phloem to accumulate in root cells and to develop plant sections (Jaenicke, 2005; Jia  
101 et al., 2010; Pacini, 2000). Cowie et al. (1984) also reported various sugars in terrestrial plant  
102 fruits, flowers, and plant tissues. Bielecki (1995) reported that glucose, fructose, and sucrose are  
103 well-known components of microbes and invertebrates. The plant debris, as well as lichens,  
104 invertebrates, and soil dust, are also recognized as possible sources of primary sugars in the  
105 atmosphere (Medeiros et al., 2006; Rogge et al., 2007; Simoneit et al., 2004).

106 Previous studies analyzed aerosol samples for SCs and discussed several factors to control  
107 their local and global atmospheric levels. Recently, Xu et al. (2020) examined the seasonal  
108 molecular distributions of primary biological aerosols and BB aerosol samples collected from  
109 urban Beijing. They reported a high level of arabitol, mannitol, sucrose, glucose, and fructose in  
110 the vegetation-growing season. Kang et al. (2018) also reported higher concentrations of sugars in  
111 the urban aerosols from Beijing. They suggested a large contribution of coal combustion and  
112 agriculture residue burning under stable meteorological conditions in winter and spring. Verma et  
113 al. (2015, 2018) reported that the atmospheric circulations and long-range transport of organic-  
114 /bio-aerosols from East Asia significantly control the levels and compositions of SCs over the  
115 western North Pacific. The above studies discussed the several factors that affect the  
116 concentrations of SCs in the aerosol samples collected from urban and remote areas.

117 In this study, we conducted analyses of SCs in the aerosol samples collected from the  
118 northern vicinity of Beijing City in 2007. Here, we present comprehensive data sets of  
119 anhydrosugars, primary sugars, and sugar alcohols in the suburban aerosol samples and their  
120 diurnal variations to explain the source variance following the wind patterns in the day- and  
121 nighttime. The positive matrix factorization (PMF) has been applied to clarify the different

122 sources of measured SCs in the aerosol. We present the influence of local meteorology of  
123 sampling site and atmospheric transport from Beijing by southerly winds and Mangshan National  
124 Forest Park by northerly winds on the molecular distributions of SCs. Using the mass  
125 concentration ratio of levoglucosan to mannosan, we explain the relative contribution of hard and  
126 softwood burning to the air quality of Mangshan. This study also discussed carbon contributions  
127 of SCs and BB measured in the Mangshan aerosol samples from different sources.

## 128 **2. Materials and Methods**

### 129 **2.1. Site description and aerosol sample collection**

130 The sampling site (Mangshan: 40.28 N, 116.26 E) is located 40 km north of Beijing. A  
131 detailed description of the sampling site is given in He et al. (2014, 2015). Briefly, Mangshan is  
132 surrounded by urban areas in the south and forest areas with the national park in the north (Fig. 1).  
133 The ambient temperature was higher in daytime (23.9°C) than nighttime (12.1°C), with an average  
134 of 17.8°C during the campaign. The relative humidity (RH) varied significantly from 22.1% to  
135 90.5%, with an average of 51.7% during the study period. The rainfall was observed at midnight  
136 on 15<sup>th</sup> September, the morning of 17<sup>th</sup> to evening 18<sup>th</sup> September, the night of 26<sup>th</sup> September,  
137 and light rain lasted from 4<sup>th</sup> October to the end of the campaign (Fig. 2a). Interestingly, the  
138 sampling site is characterized by a specific wind pattern, i.e., southwest wind (69.9%) prevailed,  
139 followed by northeast wind (23.4%) and southeast wind (6.2%) during the daytime (Fig. 2a). The  
140 northeast wind (99.5%) was dominated at night, which is consistent with the air mass back  
141 trajectories (He et al., 2014) (Fig. 2b). The daytime wind from the southwest direction passed over  
142 Beijing, delivering anthropogenic air mass to the Mangshan site.

143 Detailed descriptions of the total suspended particulate (TSP) samples collected at  
144 Mangshan are given in He et al. (2014, 2015). Briefly, The aerosol samples were collected near  
145 the entrance of Mangshan National Forest Park. The elevation of the sampling location is 187 m  
146 above sea level. A high-volume air sampler (Kimoto-AS810A) at a flow rate of 1.13-1.17 m<sup>3</sup> min<sup>-1</sup>  
147 was used to collect the TSP without cut-off device. In the sampling, no denuder was applied to  
148 remove semi-volatile gases because the filter samples were used to analyze nonvolatile sugar  
149 compounds. However, the levoglucosan partition between the gas and particle phases, but their  
150 concentration was low. The sampling time was rather short due to the day and night sampling.  
151 Therefore, the uncertainty due to the gas phases in the particulate species concentration might be  
152 insignificant. The samples were collected on pre-combusted (450°C for 6 h) quartz fiber filters  
153 (Pallflex 2500QAT-UP, 20 cm × 25cm) from 15<sup>th</sup> September to 5<sup>th</sup> October 2007. After sample

154 collection, the individual filters were placed in pre-combusted glass jars with Teflon-lined screw  
155 caps and stored in a dark, cold room at  $-20^{\circ}\text{C}$  to prevent microbial activity and loss of semi-  
156 volatile organic compounds from the samples. In this study, a total of 58 filter samples were  
157 analyzed. We collected 3h daytime (from 9 to 12, 12 to 15, 15 to 18 h) ( $n=26$ ), 9h daytime (from 9  
158 to 18 h) ( $n=12$ ), and 15h nighttime (from 18 to 9 h) ( $n=20$ ) samples together with four field  
159 blanks. Table S1 shows the details of aerosol sample collection in the Mangshan site.

## 160 **2.2. Extraction and derivatization of samples**

161 A total of 58 aerosol samples were analyzed for anhydrosugars, primary sugars, and sugar  
162 alcohols (Table 1). The sample filters (approximately  $21\text{ cm}^2$ ) were extracted with a  
163 dichloromethane and methanol mixture (2:1) under ultrasonication. Pasteur pipettes packed with  
164 pre-combusted quartz wool were used to filter the extracts to remove filter debris. After filtration,  
165 the extracts were concentrated in a rotary evaporator under vacuum and dried by nitrogen  
166 blowdown. The extracts were reacted with  $60\text{ }\mu\text{L}$  of N,O-bis-(trimethylsilyl)trifluoroacetamide  
167 (BSTFA) with 1% trimethylsilyl (TMS) chloride in the presence of  $10\text{ }\mu\text{L}$  of pyridine at  $70^{\circ}\text{C}$  for  
168 three hours to derivatize hydroxyl (OH) and carboxyl (COOH) groups into corresponding  
169 trimethylsilyl (TMS) ethers and esters, respectively. After the reaction, n-hexane was used for  
170 dilution, and  $\text{C}_{13}$  n-alkane was added as an internal standard before GC-MS analysis.

## 171 **2.3. Gas chromatography-mass spectrometry determination of sugar compounds (SCs)**

172 Details of GC-MS operation and identification of SCs are described in Verma et al. (2015,  
173 2018). Briefly, GC-MS analyses were performed on Agilent model 6890 gas chromatograph (GC)  
174 combined with an Agilent model 5973 mass selective detector (MSD) to determine SCs. The mass  
175 spectrometer was operated in the electron ionization (EI) mode at  $70\text{ eV}$  with a scan range of  $m/z$   
176  $40\text{--}650$ . The GC separation was achieved on a DB-5MS fused silica capillary column ( $30\text{ m} \times$   
177  $0.25\text{ mm}$  in diameter,  $0.25\text{ }\mu\text{m}$  film thickness) and a split/splitless injector. The GC oven  
178 temperature was programmed to maintain at  $50^{\circ}\text{C}$  for 2 min and then to increase from  $50$  to  $120^{\circ}\text{C}$   
179 at a rate of  $15^{\circ}\text{C min}^{-1}$ , then from  $120$  to  $305^{\circ}\text{C}$  at a rate of  $5^{\circ}\text{C min}^{-1}$ . The final isotherm holds at  
180  $305^{\circ}\text{C}$  for 15 min. Helium was used as the carrier gas at a flow rate of  $1.0\text{ mL min}^{-1}$ . The sample  
181 was injected on a splitless mode at  $280^{\circ}\text{C}$  injector temperature. GC-MS data were acquired and  
182 processed with Agilent GC/MSD ChemStation software.

183 The individual compounds (TMS derivatives) were identified by comparing the relative  
184 response factors determined by the injection of authentic standards and those reported in the

185 literature and library texts (Claeys et al., 2004). Fragment ions of sugar compounds at 217 and 204  
186 were used for quantifications. Total ten sugar compounds, including three anhydrosugars  
187 (levoglucosan, galactosan, mannosan), four primary sugars (glucose, fructose, sucrose, trehalose  
188 and xylose) and three sugar alcohols (arabitol, mannitol, and inositol), were detected in the  
189 Mangshan aerosols. Field blanks were treated as a real sample and analyzed by the procedure used  
190 for the real samples. Recoveries for SCs were better than 85% as obtained by the standards spiked  
191 to precombusted quartz filter followed by extraction and derivatization. Based on the duplicate  
192 analysis, the analytical errors in the concentrations of the detected compounds were obtained to be  
193 within 10%. The detection limits of SCs corresponds to ambient concentrations of 150-620  $\text{pg } \mu\text{L}^{-1}$   
194 <sup>1</sup>, which corresponds to ambient concentrations of 15-70  $\text{pg m}^{-3}$  under a typical sampling volume  
195 of 900  $\text{m}^3$ .

#### 196 **2.4. Chemical analyses of organic carbon (OC), water-soluble organic carbon (WSOC), and** 197 **inorganic ions**

198 The data set and methods for the determination of organic carbon (OC), water-soluble  
199 organic carbon (WSOC) and inorganic ion ( $\text{Ca}^{2+}$ ) were reported in He et al. (2015). Briefly, the  
200 concentrations of OC were measured using a semi-continuous OC/EC analyzer (Sunset Laboratory  
201 Inc., Portland, OR, USA). A punch of the filter ( $\Phi 14$  mm) was placed in a quartz boat inside the  
202 thermal desorption chamber of the analyzer, and then stepwise heating (IMPROVE) was applied.  
203 The oven temperature was programmed as follows: under He, every 2 minutes, the oven  
204 temperature was increased starting from 250°C for 2 min, at 450°C for 2 min, and at 550°C for 2  
205 min. After that, 550°C was maintained for two minutes under He mixed with 10%  $\text{O}_2$ , then at  
206 700°C for 2 min and at 870°C for 3.5 min. NDIR detector was used to determine  $\text{CO}_2$  generated in  
207 the above process (Wang et al., 2005b). The carbon content of the sample that evolves to  $\text{CO}_2$   
208 between 250 and 700°C was defined as OC.

209 Aliquots of the filter samples ( $3.14 \text{ cm}^2$ ) were extracted with Milli Q water for the water-  
210 soluble inorganic ion and WSOC measurements. After extraction, one part was used for the  
211 analyses of inorganic ions ( $\text{SO}_4^{2-}$ ,  $\text{NO}_3^-$ ,  $\text{Cl}^-$ ,  $\text{NH}_4^+$ ,  $\text{Na}^+$ ,  $\text{Ca}^{2+}$ ,  $\text{K}^+$  and  $\text{Mg}^{2+}$ ) using an ion  
212 chromatography (IC) system (761 Compact IC, Metrohm, Switzerland). Cations on a Shodex YK-  
213 421 column with 4mM  $\text{H}_3\text{PO}_4$  as eluent and anions were separated on a Shodex SI-90 4E column  
214 with 1.8mM  $\text{Na}_2\text{CO}_3$  and 1.7mM  $\text{NaHCO}_3$  as eluent. The injection loop volume was 200  $\mu\text{l}$ . Both  
215 cations and anions were quantified against a standard calibration curve. Another part of the filtered  
216 water extract was acidified with 1.2 M HCl and purged with pure air to remove dissolved  
217 inorganic carbon and volatile organics. Then WSOC was measured with a carbon analyzer

218 (Shimadzu, TOC-5000). Procedural blanks were carried out in parallel with real samples to  
219 account for any contamination (He et al., 2015).

## 220 **2.5. Positive Matrix Factorization (PMF) Analysis**

221 Positive matrix factorization (PMF) is a powerful statistical tool for resolving the potential  
222 sources contributing to atmospheric particles (Paatero and Tapper, 1994). The measured ambient  
223 concentrations and method detection limits (MDLs) of SCs were used to calculate the  
224 uncertainties. The measured concentrations of SCs below or equal to the MDLs were replaced by  
225 half of the MDL, and associated uncertainties were set at 5/6 of the MDL [(5/6) × MDL] values of  
226 each sample. The geometric mean concentrations were used for missing concentrations, and the  
227 uncertainty of the concentrations greater than the MDL was calculated based on the following  
228 equation:

$$229 \text{ Uncertainty} = \sqrt{(\text{error fraction} \times \text{concentration})^2 + (0.5 \times \text{MDL})^2}$$

230 The error fraction is a user-provided estimation of the analytical uncertainty of the  
231 measured concentration or flux. For example, Han et al. (2017) used an error fraction of 0.2-0.3  
232 for organics and 0.2 for all the species. In this work, the error fraction was set to be 0.3 for all  
233 species. Paatero et al. (2002) and Zhou et al. (2004) reported detailed discussions of the  
234 determination and application of PMF analysis.

## 235 **3. Results and Discussion**

### 236 **3.1. Ambient concentrations and diurnal variations of SCs**

237 We detected a total of ten SCs, including three anhydrosugars, four primary sugars, and three  
238 sugar alcohols in the Mangshan aerosol samples. Figure 3a-c showed the temporal variations and  
239 Table 1 showed minimum, maximum, and average concentrations of anhydrosugars, primary  
240 sugars, and sugar alcohols with a standard deviation. The overall concentrations of SCs varied  
241 from 30.8–875 ng m<sup>-3</sup> (avg. 325 ng m<sup>-3</sup>), which was higher in the daytime (315 ng m<sup>-3</sup>) and lower  
242 at nighttime (276 ng m<sup>-3</sup>), however, we did not observe statistically significant differences  
243 (student t-test, 95% confidence interval, p > 0.05) in their atmospheric abundances. Interestingly,  
244 higher average concentrations of SCs were reported for the aerosol samples collected from at Mt.  
245 Tai (daytime 640 ng m<sup>-3</sup> and nighttime 799 ng m<sup>-3</sup>) in the North China Plain (Fu et al., 2008) than  
246 the Mangshan aerosol. The diurnal concentrations of SCs may be significantly influenced by  
247 vegetation and BB activities in the Mangshan site. SCs are significantly contributed by plant



248 fractions and fungus from the forest area (Zhu et al., 2016). The meteorological parameters also  
249 affect the concentrations of SCs in the forest site (Miyazaki et al., 2012).

250 In addition, anthropogenic aerosols emitted from urban areas are probably transported to the  
251 northern receptor site in daytime by a southerly wind (He et al., 2014; 2015). Therefore, the high  
252 levels of SCs in daytime may be related to the transport of organic and bio-aerosols from urban  
253 regions. The nighttime, the wind direction is shifted to northerly, delivering comparatively clean  
254 air masses from the Mangshan National Forest area to the sampling site. Air mass from the forest  
255 may significantly contribute to nighttime SCs in the Mangshan site. The influence of local sources  
256 and long-range transported aerosols on the SCs will be discussed in sections 3.1.1 to 3.1.3.

### 257 **3.1.1. Ambient concentrations and diurnal variations of anhydrosugars**

258 The average concentrations of anhydrosugars were found  $116 \text{ ng m}^{-3}$ , contributing 31.9%  
259 of overall SCs in the Mangshan aerosols (Table 1). Figure 4a-c shows the temporal variations of  
260 anhydrosugars. They are more abundant in nighttime (avg.  $152 \text{ ng m}^{-3}$ ) than daytime (avg.  $97.1 \text{ ng}$   
261  $\text{m}^{-3}$ ). Levoglucosan ( $100 \text{ ng m}^{-3}$ ) is the most abundant anhydrosugar followed by galactosan ( $10.1$   
262  $\text{ng m}^{-3}$ ) and mannosan ( $6.05 \text{ ng m}^{-3}$ ) detected in Mangshan aerosols. Kang et al. (2018) reported  
263 high levels of levoglucosan (avg.  $110 \text{ ng m}^{-3}$ ) in autumn aerosols from Beijing, China. It is well  
264 known that biofuel burning is the common energy source for cooking and house heating in China  
265 in winter and autumn (Verma et al., 2015), thus the domestic BB activities in the surroundings of  
266 the Mangshan site significantly contribute to the levoglucosan. BB tracers showed significant  
267 positive correlations with each other (levoglucosan and galactosan,  $r = 0.98$ ; levoglucosan and  
268 mannosan,  $r = 0.97$ ; galactosan and mannosan,  $r = 0.98$ ), suggesting their similar sources in the  
269 Mangshan aerosols (Table 2).

270 The levoglucosan concentrations showed significant diurnal variations, which was higher  
271 in nighttime (avg.  $132 \text{ ng m}^{-3}$ ) than daytime (avg.  $83.2 \text{ ng m}^{-3}$ ) (Table 1). A similar diurnal pattern  
272 was also found for the concentrations of galactosan and mannosan. The increased concentrations  
273 of BB tracers were observed during the periods of lower ambient temperature (Figs. 2a, 4a-c). The  
274 higher ambient temperature was recorded in daytime between 09h to 15h during the campaign,  
275 associated with declined BB activities. In this sequence, the nighttime samples were collected  
276 from 18:00h to 09:00h, including peak hours of BB for domestic purpose. Therefore, it is  
277 reasonable to detect higher abundances of BB tracers in the nighttime than daytime. Hence, it is  
278 evident that BB activities were increased at night because of cooking and house heating at cool  
279 night in autumn. In addition, recent studies reported the widespread BB aerosols in the North

280 China Plain, including megacities such as Beijing, Nanjing, Hebei, and Tianjin (Lelieveld et al.,  
281 2015; Kawamura et al., 2013; Li et al., 2010; Sun et al., 2016). Therefore, the atmospheric  
282 transport of BB aerosols from the urban area to the Mangshan site by southerly winds cannot be  
283 excluded. The diurnal variations of levoglucosan may be significantly influenced by the local BB  
284 activities and transported BB aerosols from urban areas, where BB products are generated by  
285 brown coal combustion (Yan et al., 2018).

### 286 3.1.2. Ambient concentrations and diurnal variations of primary sugars

287 The fragment of vascular plants contains primary sugars, including glucose, fructose,  
288 sucrose, and trehalose (Medeiros et al., 2006). Primary sugars were found as the most abundant  
289 sugars (avg. 133 ng m<sup>-3</sup>), contributing to 41.8% of the total SCs in Mangshan aerosols (Table 1).  
290 They showed apparent diurnal variations with daytime high (avg. 166 ng m<sup>-3</sup>) and nighttime low  
291 values (avg. 69.4 ng m<sup>-3</sup>) (Figs. 3a-c, 5a-d). Graham et al. (2003) also reported similar diurnal  
292 variations of primary sugars for the Amazon forest aerosols. Sucrose was found as dominant  
293 primary sugars (avg. 58.5 ng m<sup>-3</sup>), accounting for 44% of measured primary sugars in Mangshan  
294 aerosols (Table 1). Pollen was reported as a primary source for sucrose in aerosols collected from  
295 a Texas rural site (Jia et al., 2010). Fu et al. (2012) found high sucrose concentrations up to 1390  
296 ng m<sup>-3</sup> in the aerosols from Jeju Island, South Korea. Therefore, the plant materials, including  
297 pollen spores from the local vegetation of Mangshan National Forest Park, are likely the primary  
298 source of sucrose in the aerosols. Miyazaki et al. (2012) also reported higher sucrose  
299 concentrations in the aerosol samples collected from the Hokkaido deciduous forest.

300 We found a significant diurnal variation of sucrose with higher daytime (82.9 ng m<sup>-3</sup>) than  
301 nighttime (12.3 ng m<sup>-3</sup>). Meteorological parameters such as temperature, rainfall, wind speed, and  
302 solar radiation significantly influence pollen activities and, subsequently, sucrose concentrations  
303 (Verma et al., 2018). Interestingly, an elevated peak of sucrose was observed from 12h to 15h with  
304 higher ambient temperature. In contrast, lower sucrose concentrations were observed from 15h to  
305 9h with lower ambient temperature (Fig. 5a). Daytime increased concentrations of sucrose might  
306 be related to the higher daytime ambient temperature, low RH, and high solar radiation (Miyazaki  
307 et al., 2012). Taylor et al. (2002) reported the influence of the meteorological conditions, i.e.,  
308 strong daytime winds and convective activity, which can result in catapulting of pollen, opening of  
309 pollen-laden flower anthers, and causing enhance entrainment and dispersal of the particles into  
310 the air. Pacini (2000) reported that higher levels of sucrose in daytime coincide with higher counts  
311 of pollen, fern spore, and insect. The positive linear correlations of sucrose with ambient  
312 temperature ( $r = 0.52$ ) and solar radiation ( $r = 0.55$ ) further supported the influence of

313 meteorological parameters in the sucrose concentration (Table 2).

314 Five rain events were recorded during the campaign, i.e., 15<sup>th</sup>, 17<sup>th</sup>, 18<sup>th</sup>, and 26<sup>th</sup>  
315 September, and 1<sup>st</sup> and 5<sup>th</sup> October (Fig. 2a). Pollens are significantly settled down by wet  
316 scavenging during rain events because their sizes are large. A low concentration of sucrose was  
317 found from the beginning of sampling to the morning of 20<sup>th</sup> September and from the afternoon of  
318 26<sup>th</sup> September to the end of the sampling campaign (Fig. 5a). In addition, the increased  
319 concentrations of sucrose were found in the aerosol samples collected from 20<sup>th</sup> to 22<sup>nd</sup> September,  
320 and moderate concentrations were observed after 23<sup>rd</sup> to the evening of 25<sup>th</sup> September during  
321 non-precipitation events. Consequently, the pollens were significantly scavenged during wet  
322 precipitation and washout effect from the atmosphere, resulting in lower sucrose concentrations at  
323 the earlier periods, than later periods. In addition, Rogge et al. (2007) reported that surface soil  
324 dust and unpaved road dust also contribute sucrose in the atmospheric aerosols. However,  
325 insignificant correlations between sucrose and Ca<sup>2+</sup> (daytime,  $r = 0.32$ ; night time,  $r = 0.37$ ) do not  
326 supports soil dust contributions to sucrose in the Mangshan aerosols (Table 2).

327 Glucose was the second dominant primary sugar in the Mangshan aerosols. The average  
328 concentrations of glucose and fructose were observed to be 40.0 ng m<sup>-3</sup> and 20.1 ng m<sup>-3</sup>,  
329 respectively (Table 1, Fig. 5b). The sampling site is characterized by the dense vegetation in the  
330 Mangshan National Forest Park. Therefore, the nectars and fruits of vegetation (Baker et al.,  
331 1998), plant debris (Medeiros et al., 2006) and pollens (Fu et al., 2012) in the forest significantly  
332 contribute to glucose and fructose. The glucose levels are equivalent to that (50.1 ng m<sup>-3</sup>) reported  
333 from the Howland Experimental Forest site in the USA (Medeiros et al., 2006). Glucose and  
334 fructose showed significant diurnal variations, whose concentrations were higher in daytime (44.2  
335 ng m<sup>-3</sup> and 23.9 ng m<sup>-3</sup>, respectively) than nighttime (32.0 ng m<sup>-3</sup> and 12.8 ng m<sup>-3</sup>, respectively)  
336 in Mangshan aerosols (Table 1, Figs. 3b, c; 5b, c). This diurnal variation could be involved with  
337 emissions of pollens, fern spores, and other giant particles by strong winds (Graham et al., 2003;  
338 Pacini, 2000). Similar trends of glucose and fructose were reported in the Amazon forest, being  
339 coincided with plant fragments and insects (Graham et al., 2003). The autumn decay of vascular  
340 plant leaves in the Mangshan forest may have contributed to the levels of glucose and fructose.

341 Although, the daytime southerly winds deliver anthropogenic air masses from megacities  
342 to the sampling site. The daytime winds from the northeast direction (23.4%) also carry air masses  
343 from the forest region, transporting primary sugars to the Mangshan site. However, 99.5% of the  
344 nighttime hours, the wind is shifted to northeasterly, i.e., in forest region (He et al., 2015), but the  
345 emissions of primary sugars at night in the form of plant fragments are lower than in daytime.

346 Because the daytime ambient temperature and solar radiations significantly induce the emissions  
347 of sugar compounds in the forest site (Miyazaki et al., 2012). Therefore, low glucose and fructose  
348 levels were found at nighttime than daytime aerosols at the Mangshan site (Table 1, Fig. 3).  
349 Previous studies have reported lichens (Dahlman et al., 2003) and soil dust (Nolte et al., 2001;  
350 Rogge et al., 2007) as significant sources of both primary sugars. The concentration of glucose  
351 was insignificantly correlated with soil tracer ( $\text{Ca}^{2+}$ ) in day ( $r = 0.02$ ) and nighttime ( $r = 0.27$ ),  
352 denying their soil dust contributions in Mangshan aerosol samples.

353 Trehalose in the environment is significantly controlled by the activities of bacteria, fungi,  
354 yeast, algae, invertebrates, and plant species, as well as suspended soil particles (Medeiros et al.,  
355 2006, Rogge et al., 2007). The average concentration of trehalose was found  $14.3 \text{ ng m}^{-3}$  (Table 1,  
356 Fig. 5d). Yttri et al. (2007) reported higher trehalose concentrations in the aerosol samples  
357 collected from urban ( $29 \text{ ng m}^{-3}$ ) and suburban ( $27 \text{ ng m}^{-3}$ ) than rural ( $3.8 \text{ ng m}^{-3}$ ) areas in  
358 Norway. The above results emphasize that fungi and microbes associated with anthropogenic and  
359 bioaerosols, emitted in the urban and suburban areas, might be responsible for the trehalose  
360 concentration in aerosol samples (Verma et al., 2018). Trehalose showed insignificant diurnal  
361 variation, whose day and night concentrations were observed  $15.3 \text{ ng m}^{-3}$  and  $12.3 \text{ ng m}^{-3}$ ,  
362 respectively, indicating its different emission sources in day and night for Mangshan aerosols (Fig.  
363 3b, c; 5d).

364 The southerly winds might transport fungi and microbes associated with bioaerosols, eject  
365 spores under favorable meteorological conditions (high RH and low temperature) (Jones and  
366 Mitchell et al., 1996). Several microbes and fungi discharge spores at nighttime due to high RH  
367 conditions (Ibrahim et al., 2011; Kim and Xiao, 2005; Malik and Singh, 2004; Sharma and Razak,  
368 2003). Interestingly, trehalose is more significantly correlated with arabitol and mannitol ( $r = 0.76$   
369 and  $0.85$ , respectively) in nighttime than daytime ( $r = 0.49$  and  $0.51$ , respectively) (Table 2),  
370 suggesting that fungal and microbial spores contributed to high levels of trehalose in nighttime.  
371 Hackl et al. (2000) found trehalose as dominant sugar in spring aerosols and proposed it as a tracer  
372 for soil dust particles. Trehalose concentration was more significantly correlated with  $\text{Ca}^{2+}$  ( $r =$   
373  $0.82$ ) in daytime than nighttime ( $r = 0.61$ ), indicating soil dust contribution (Table 2). Therefore,  
374 we hypothesized that winds transported soil particles from the urban area in daytime due to the  
375 active building constructions (He et al., 2015), contributing to the high levels of trehalose in  
376 daytime.

### 377 3.1.3. Ambient concentrations and diurnal variations of sugar alcohols

378 The average concentrations of sugar alcohols were found  $75.8 \text{ ng m}^{-3}$ , contributing 26.4%  
379 of total SCs measured in Mangshan aerosols (Table 1). Sugar alcohols showed clear diurnal  
380 variations in daytime high (avg.  $87.4 \text{ ng m}^{-3}$ ) and nighttime low (avg.  $53.7 \text{ ng m}^{-3}$ ) (Table 1).  
381 Mannitol was found as the dominant sugar alcohol (avg.  $44.1 \text{ ng m}^{-3}$ ), followed by arabitol (avg.  
382  $29.1 \text{ ng m}^{-3}$ ) and inositol (avg.  $2.62 \text{ ng m}^{-3}$ ) (Table 1; Fig. 6a-c). Mannitol and arabitol are  
383 common polyols detected in green algae, lichens, and fungal spores (Bieleski, 1995, Dahlman et  
384 al., 2003; Filippo et al., 2013; Lewis and Smith, 1967; Yttri et al., 2007). Previous studies have  
385 reported that arabitol and mannitol are key components of fungal spores, and thus they are  
386 considered as fungal tracers (Bieleski, 1995; Lewis and Smith, 1967). Several fungal and microbial  
387 species released spores during biological activities into the atmosphere (Dahlman et al., 2003;  
388 Bauer et al., 2008; Filippo et al., 2013). Therefore, the autumn time fungal and microbial species  
389 significantly contribute to arabitol and mannitol in the Mangshan aerosol samples.

390 However, mannitol and arabitol showed a strong positive linear correlation ( $r = 0.81$ ),  
391 which suggested common origins as reported in earlier studies (Fu et al., 2012) (Table 2). In  
392 contrast, the higher concentration of mannitol than arabitol suggested it had sources in addition to  
393 fungal spores in the Mangshan forest site. In this sequence, several previous studies have  
394 confirmed the significance of mannitol in plant photosynthesis (Loescher et al., 1992; Keller and  
395 Matile, 1989; Rumpho et al., 1983). Pashynska et al. (2002) reported that detritus of mature leaves  
396 can emit mannitol into the atmosphere by wind action. Heald and Spracklen (2009) also found a  
397 correlation between the atmospheric water vapor with mannitol concentrations and leaf area index.  
398 They suggested that the activities of the terrestrial biosphere widely affect mannitol concentrations  
399 in the air. Our PMF results also indicated the substantial contribution of mannitol for vegetation  
400 factor (24.8%), which supports that mannitol is attributed by vegetation from the forest area  
401 (section 3.2).

402 In addition, the meteorological parameters, including high RH and temperature affect the  
403 fungal and bacterial activities (Kim and Xiao, 2005; Sharma and Razak, 2003). The maximum  
404 growth of fungi and bacteria was observed at 92–100% RH (Ibrahim et al., 2011). Interestingly,  
405 the concentrations of arabitol and mannitol gradually increased after the end of precipitation,  
406 following the increases in ambient temperature and RH (Figs. 2a, 6a, b). Miyazaki et al. (2012)  
407 also discussed the increased contributions of arabitol and mannitol with daytime ambient  
408 temperature and solar radiation in the aerosol samples collected from the forest area. Similar  
409 temporal trends and positive linear correlations were observed between arabitol ( $r = 0.69$ ) and  
410 mannitol ( $r = 0.57$ ) with RH, which supports the above phenomenon for Mangshan aerosols

411 (Table 2). Therefore, we propose that a favorable meteorological condition in autumn increases  
412 the emissions of fungal spores and fragments of forest vegetation, which may be responsible for  
413 arabitol and mannitol contributions in the Mangshan aerosols.

414 The diurnal variation of mannitol and arabitol were characterized by higher in the daytime  
415 ( $51.7 \text{ ng m}^{-3}$  and  $32.5 \text{ ng m}^{-3}$ , respectively) than nighttime ( $29.6 \text{ ng m}^{-3}$  and  $22.5 \text{ ng m}^{-3}$ ,  
416 respectively) (Fig. 3b, c). Yamaguchi et al. (2012) reported that fungal spores and bacterial cells  
417 associated with bioaerosols could be transported long distances. The Mangshan site receives  
418 significant anthropogenic and bioaerosols from Beijing City by southerly winds. Therefore, the  
419 daytime plant activities, influenced by solar radiation and ambient temperature and the long-range  
420 transport of fungal spores from megacities (Beijing) by southwest winds govern the diurnal  
421 variation of sugar alcohols in the Mangshan atmosphere. On the other hand, lower concentrations  
422 in nighttime can be explained by the clean air mass transport by mountain breeze from the  
423 Mangshan National Forest area.

### 424 **3.2. Source apportionment of SCs**

425 To investigate the source apportionment of SCs, positive matrix factorization (PMF)  
426 [software version 5.0 \(Environmental Protection Agency, USA\)](#) was used. The PMF analysis was  
427 performed for the measured aerosol samples using tracer compounds for anhydrosugars, primary  
428 sugars, and sugar alcohols. It is essential to select a suitable number of factor solutions in the PMF  
429 analysis. Based on the possible sources of SCs, four to six factor solutions were run in PMF  
430 model. In the four-factor solutions, the SCs, including arabitol, mannitol, and trehalose, were  
431 merged in a single factor; this might underestimate the soil dust sources. In six factor solutions,  
432 the SCs, including glucose, fructose trehalose, arabitol, and mannitol, were distributed in more  
433 than four factors; it might be overestimated the number of factor solutions according to possible  
434 sources of SCs. Therefore, a total of five interpretable factor solutions were characterized by the  
435 enrichment of each tracer compound to be significant to categorize the origins of individual  
436 sugars, which reproduced more than 95% of SCs.

437 These five-factor solutions were preferred based on minimum robust and true Q values  
438 (goodness of fit parameters) of the base runs, which observed 3103 and 3505, respectively. In each  
439 bootstrap run, the concentrations and percentages of tracers were close to those of base-run results.  
440 The PMF results of SCs indicate a stability because no significant changes were found between Q  
441 values and factor profiles of  $F_{\text{peak}}$  rotation runs compared with the base run. PMF results show a  
442 good correlation between the values of observed and predicted (modeled) concentrations in scatter

443 plot, indicating that the model very well fits the individual sugar species. These results support the  
444 perfect rationality of the source apportionment (Figure S-1). The time series plot of observed and  
445 predicted concentration (modeled) also shown that the model well fits the observed data set  
446 (Figure S-2). The time series plots of the factors solutions determined by PMF were similar to the  
447 temporal plots of the concentration of sugar species of the factor composition (Figure S-3). The  
448 numbers of factors were reduced if the pair of factors was strongly correlated. The composition of  
449 each factor was also checked; none of the pair of factors were found with similar composition. We  
450 also investigated the change in factor profile with positive and negative values of  $f_{peak}$  for the  
451 chosen solution in the PMF analysis. Figures 6 and 7 show the factor profile resolved by PMF  
452 analysis of the Mangshan aerosol samples. The percentages of each component are summed for  
453 factors 1 to 5 to be calculated as 100%.

454 Factor 1 is characterized by the high contribution of glucose (80.2%) followed by fructose  
455 (69.6%), mannitol (24.8%), and inositol (15.1%) (Fig. 7a). Glucose and fructose are highly water-  
456 soluble SCs present in the leaves and bark of plants (Graham et al., 2003). High concentrations of  
457 glucose and fructose have been reported in vascular plants and phytoplankton by Cowie and  
458 Hegdes (1984). The dominant glucose and fructose in the Mangshan aerosol samples collected in  
459 autumn are rational as leaf senescence and decay results in both primary sugars being released into  
460 the atmosphere during the fall season. We found an excellent correlation between glucose and  
461 fructose ( $r = 0.94$ ) in the Mangshan aerosols (Table 2), indicating the similar vegetation sources  
462 for both sugar species in autumn (Baker et al., 1998; Burshtein et al., 2011; Pacini, 2000). Higher  
463 concentrations of glucose and fructose in the aerosol samples collected during the autumn season  
464 are reasonable because leaf senescence and decay result in an increased emission of primary  
465 sugars into the atmosphere.

466 Several studies have reported that plant species significantly contribute to mannitol in the  
467 atmosphere (Burshtein et al., 2011; Devis et al., 1988; 1990). Miyazaki et al. (2014) also found a  
468 significant amount of trehalose, mannitol, and arabitol in the aerosol samples collected from the  
469 forest and concluded their origin from the terrestrial plants within the forest. Significant positive  
470 linear correlations of mannitol with fructose in daytime ( $r = 0.79$ ) and nighttime ( $r = 0.86$ ) further  
471 denote that abundance of mannitol is due to the decay of plant leaves in autumn (Table 2).  
472 Therefore, we conclude that the contributions of mannitol is from both vegetation and fungal  
473 spores in the Mangshan aerosol samples. Hence mannitol showed the presence in factor 1.  
474 Vegetations contribute to SCs during the campaign. Therefore, factor 1 can be termed as a  
475 vegetation factor due to the high abundances of glucose, fructose, and mannitol.

476 Factor 2 is dominated by high loading of trehalose (80.2%), followed by mannitol (29.7%),  
477 glucose (19.8%), and arabitol (18.2%) (Fig. 7b). The contribution of trehalose to soil dust has been  
478 reported in several studies from different locations around the world, suggesting trehalose as a  
479 tracer for the surface soil (Jia et al., 2010; Medeiros et al., 2006). In addition, previous studies  
480 reported that bacteria and other microbes in the soil are also an essential source of trehalose  
481 (Rogge et al., 2007). Trehalose is significantly correlated with arabitol ( $r=0.58$ ) and mannitol  
482 ( $r=0.58$ ), and  $\text{Ca}^{2+}$  ( $r=0.70$ ), demonstrating its microbial and soil dust origin. Therefore, factor 2  
483 can be termed as microbial and soil dust factor.

484 Factor 3 is characterized by levoglucosan (82.2%), galactosan (77%), and mannosan  
485 (73.6%) (Fig. 7c). Previous studies have reported that these SCs are associated with BB aerosols  
486 (Fraser and Lakshmanan, 2000; Graham et al., 2002; Simoneit, 2002). Simoneit et al. (1999)  
487 reported that the pyrolysis of cellulose and hemicellulose emitted levoglucosan, galactosan and  
488 mannosan. These sugar species are major organic components emits in the atmosphere by BB  
489 activities (Simoneit et al., 2002). The BB influenced aerosols are enriched with levoglucosan,  
490 mannosan, and galactosan (Nolte et al., 2001; Medeiros et al., 2006). The domestic BB for  
491 cooking and house heating due to low ambient temperature and field burning of agricultural  
492 residues occur in East Asia (Verma et al., 2015). The PMF results are very well supported by the  
493 fact that anhydrosugars are associated with BB (Simoneit et al., 1999). Therefore, factor 3 can be  
494 termed as a BB factor due to the high abundance of BB products.

495 Factor 4 is dominated by high loading of sucrose (90%), followed by inositol (36.9%) and  
496 fructose (11.7%) (Fig. 7d). Sucrose plays a crucial role in the plant blossoming process as the  
497 dominant sugar compound of pollen grains (Pacini, 2000). Several studies also reported that  
498 sucrose is abundant sugar species found in airborne pollen grains and flowering plants (Fu et al.,  
499 2012; Graham et al., 2003; Medeiros et al., 2006; Pacini, 2000). Therefore, sucrose is reported as  
500 an excellent tracer for airborne pollen spores (Pacini, 2000). Thus factor 4 is termed as pollen  
501 factor due to the high loading of sucrose.

502 Factor 5 is characterized by a higher contribution of arabitol (61.5%) followed by mannitol  
503 (39.3%) and inositol (15.3%) (Fig. 7e). Sugar species contributing to factor 5 are associated with  
504 fungal spores (Bauer et al., 2008). Various fungi and microbes emit spores, which are tracers for  
505 the arabitol and mannitol; therefore, both sugars are considered as specific tracers of fungal  
506 activities (Medeiros et al., 2006; Rogge et al., 2007). Thus, factor 5 is termed as a fungal factor  
507 due to the high loading of arabitol and mannitol. Overall, the average contributions of each factor  
508 to measured SCs were estimated by PMF analyses (Fig. 8), in which BB was found to account for



509 27% of measured SCs. The vegetation and microbial and soil dust sources equally contribute  
510 (21%) to total SCs. The fungal spores and pollen spores contribute 16% and 15% of total SCs,  
511 respectively. Finally, biomass burning emissions from the local areas and megacities via long-  
512 range atmospheric transport were identified as an important source for the Mangshan aerosols.

### 513 3.3. Contributions of sugar compounds to WSOC and OC

514 The contribution of carbon content of measured SCs varied from 14.1-371 ng m<sup>-3</sup> (av.  
515 145 ng m<sup>-3</sup>) in daytime and 12.8-322 ng m<sup>-3</sup> (av. 117 ng m<sup>-3</sup>) in nighttime, accounting for 0.83%  
516 and 0.91% of OC, respectively (Fig. 9a, b). The mean carbon contents of anhydrosugars showed  
517 clear diurnal variation with higher nighttime values (67.1 ng m<sup>-3</sup>) than daytime (42.7 ng m<sup>-3</sup>),  
518 accounting for 0.43 % and 0.22 % of OC, respectively. These results suggest that BB significantly  
519 contributed to Mangshan aerosols. However, the carbon contents of primary sugars showed  
520 opposite diurnal variations; higher (68.5 ng m<sup>-3</sup>) in daytime than nighttime (28.3 ng m<sup>-3</sup>),  
521 accounting for 0.41 % and 0.28 % of OC, respectively (Fig. 9a, b). This study suggests that the  
522 daytime emissions of primary sugars from local vegetation and the decay of plant leaf in forest  
523 significantly contribute to OC. The carbon concentration contributed by sugar alcohols showed  
524 insignificant diurnal variations i.e. 34.6 ng m<sup>-3</sup> in daytime and 21.3 ng m<sup>-3</sup> in nighttime,  
525 accounting for 0.20 % and 0.19 % of OC, respectively. This result indicates multiple carbon  
526 sources of sugar alcohols in day and night. In addition, contributions of anhydrosugars, primary  
527 sugars, and sugar alcohols to WSOC were similar to those of OC in Mangshan aerosols.

528 Based on the PMF analysis, we found five sources for SCs measured in Mangshan  
529 aerosols. The different tracer compounds were used to calculate carbon contents: biomass burning-  
530 C (i.e., levoglucosan, galactosan, mannosan), vegetation-C (glucose, fructose), fungal-C (arabitol,  
531 mannitol), pollen-C (sucrose), and microbial-soil-C (trehalose) (Fig. 9c, d). Among the five  
532 sources, biomass burning-C was found as the largest carbon contributor to Mangshan aerosols  
533 (36.7%), followed by fungal-C (23.7%), vegetation-C (19.7%), pollen-C (14.2%), and microbial-  
534 soil-C (4.84%). Biomass burning-C accounted for 1.38% and 0.43% at night, while 0.57% and  
535 0.22% in daytime for WSOC and OC, respectively. The BB for cooking and space heating in  
536 winter and autumn seasons are common in central China (Akagi et al., 2011), which should  
537 increase the nighttime levels of Biomass burning-C at the Mangshan site. However, the carbon  
538 contribution by vegetation and fungal sources are similar during day and nighttime for the  
539 Mangshan aerosols. Pollen-C accounted for 0.20% and 0.07% of OC in daytime and nighttime,  
540 respectively. Higher pollen activities are key sources for the high daytime levels of pollen-C in the  
541 forest site (Taylor et al., 2002).

### 542 3.4. Contribution of levoglucosan to OC and WSOC

543 We calculated the mass concentration ratios of levoglucosan to OC (Lev/OC) and WSOC  
544 (Lev/WSOC) to evaluate the contributions of BB and anthropogenic emissions to Mangshan  
545 aerosols (Fig. 9a-c). Fossil fuel combustion and BB emit WSOC and OC. They are also  
546 secondarily produced by photochemical oxidation of volatile organic compounds in the  
547 atmosphere (Wang et al., 2005a; Deshmukh et al., 2019b). Coal combustion and vehicle exhaust  
548 can contribute to the high levels of OC and WSOC in aerosols (Xu et al., 2020). Levoglucosan, a  
549 dominant constituent of BB products, has been considered as an excellent tracer of BB (Simoneit,  
550 2002; Kuo et al., 2011).

551 Average Lev/OC ratio ( $5.69 \times 10^{-3}$ ) was lower than that of Lev/WSOC ( $1.66 \times 10^{-2}$ ) in  
552 Mangshan samples (Fig. 10a). Yan et al. (2018) reported similar ratios of Lev/OC ( $4.0 \times 10^{-3}$ ) and  
553 Lev/WSOC ( $1.6 \times 10^{-2}$ ) for coal combustion, suggesting a significant carbon contribution to  
554 Mangshan aerosols from coal combustions in the industrial areas via long range transport.  
555 Interestingly, we found a substantial diurnal variation of Lev/OC and Lev/WSOC ratios. The  
556 average Lev/OC and Lev/WSOC ratios are several times higher in nighttime ( $8.48 \times 10^{-3}$  and  
557  $2.70 \times 10^{-2}$ , respectively) than daytime ( $4.21 \times 10^{-3}$  and  $1.11 \times 10^{-2}$ , respectively) (Fig. 10b, c). These  
558 results indicate that BB contributed substantially to the Mangshan organic aerosols in nighttime.  
559 Moreover, the correlations of levoglucosan with OC and WSOC are stronger in nighttime ( $r =$   
560  $0.81$  and  $0.70$ , respectively) than daytime ( $r = 0.45$  and  $0.40$ , respectively), demonstrating the  
561 dominance of BB-derived aerosols in the nighttime Mangshan samples (Table 2).

562 In addition, WSOCs are derived from various emission sources. We propose that  
563 secondary organic aerosols constitute a significant fraction of WSOC and OC in daytime  
564 Mangshan aerosols. The photochemical oxidation of organic precursors emitted from fossil fuel  
565 combustion in industries and vehicular exhausts also contributes to secondary production of  
566 WSOC and OC in daytime (He et al., 2015), suggesting that emissions from the urban Beijing area  
567 may significantly influence the daytime levels of Mangshan aerosols. He et al. (2015) proposed a  
568 possible contribution of photochemical formation of secondary organic aerosols to atmospheric  
569 WSOC and OC in north China. Nevertheless, the photochemical degradation of levoglucosan by  
570 OH radicals under ultraviolet radiations and high temperatures (Hennigan et al., 2010) may play a  
571 key role in lowering the ratios of Lev/OC and Lev/WSOC in daytime Mangshan aerosols.

### 572 3.5. Mass concentration ratios of levoglucosan/mannosan

573 The mass concentration ratios of levoglucosan and mannosan (Lev/Man) were calculated  
574 to better characterize the emissions sources of BB tracers (softwood vs. hardwood) in the

575 Mangshan site. Figure 10d represents the variations of Lev/Man ratios for overall, day- and  
576 nighttime periods. The Lev/Man ratios have been used to distinguish the hardwood (angiosperm)  
577 and softwood (gymnosperm) burning in the ice core record from the Russian Far East (Kawamura  
578 et al., 2012). Hardwood contains 55–65% cellulose and 20–30% hemicellulose (Klemm et al.,  
579 2005). Levoglucosan and mannosan are derived from the thermal decomposition of cellulose and  
580 hemicelluloses, respectively (Simoneit, 2002). Levoglucosan is thermally more stable than  
581 mannosan and galactosan (Kuo et al., 2011). Hence, a lower Lev/Man ratio is associated with  
582 softwood burning, whereas a higher ratio is associated with hardwood burning (Engling et al.,  
583 2006, 2009). However, we found insignificant diurnal variations of Lev/Man ratios between night  
584 (9.33-25.9, avg. 15.8) and daytime aerosols (0.90-23.3, avg. 13.6). Likewise, comparable Lev/Man  
585 ratios (9-13 for PM<sub>10</sub> and 10-13 for PM<sub>2.5</sub>) were reported for aerosol samples from Tanzania,  
586 where wood and charcoal are primary fuels used for domestic cooking and heating (Mkoma and  
587 Kawamura, 2013). Interestingly, wheat straws and lignite are used in China for domestic cooking  
588 and house heating, which may also contribute to levoglucosan and mannosan in the Mangshan  
589 aerosols.

590 Different Lev/Man ratios were reported in the chamber and controlled field experiments,  
591 e.g., 4-22 for conifer and savanna grass burning (Iinuma et al., 2007), and 41.6 for rice straw and  
592 and 55.7 cereal straw burning (Engling et al., 2009; Zhang et al., 2007). Kuo et al. (2011) reported  
593 higher emissions of levoglucosan during high-temperature flaming (27.5-52.3) compared to low-  
594 temperature smoldering (2.43-3.08). Hence, it is not easy to differentiate hardwood and softwood  
595 burning based on Lev/Man ratios alone. Several studies reported a high Lev/Man ratio for both  
596 softwood and hardwood burning. Thus, there may exist some other factors that significantly  
597 control the Lev/Man ratios. Yan et al. (2018) found a significant contribution of levoglucosan in  
598 coal combustion with Lev/Man ratio of 7.2. The variations of Lev/Man ratios in Mangshan may be  
599 significantly influenced by several factors, i.e., flaming vs. smoldering, duration of biomass  
600 burning, coal combustion, and hardwood vs. softwood burning. The moderate Lev/Man ratios in  
601 autumn aerosols from Mangshan suggest that low temperature smoldering processes of hardwood  
602 contribute to levels of levoglucosan and mannosan. However, the contribution of coal  
603 combustions for house heating could not be excluded.

#### 604 **4. Summary and Conclusions**

605 Anhydrosugars, primary sugars, and sugar alcohols were detected with distinct diurnal  
606 variations in suburban aerosol samples collected at the Mangshan site in the northern vicinity of  
607 Beijing. The wind patterns indicate that daytime air masses were transported from urban Beijing to

608 Mangshan, while clean air masses were delivered in nighttime from the Mangshan National Forest  
609 Park. Daytime air masses from urban Beijing significantly influence the air quality of the northern  
610 forest region. We observed the highest abundance of primary sugars, followed by anhydrosugars  
611 and sugar alcohols. Local emissions from the forest plants and fungal species are the main  
612 contributors to the primary sugars and sugar alcohols in the Mangshan aerosols. The  
613 meteorological parameter significantly influenced the levels of SCs in the Mangshan samples. We  
614 observed a significant influence of enhanced ambient temperature and solar radiation on the pollen  
615 rupture and increased RH on fungal and microbial growth. This study suggested the source  
616 variation for trehalose, i.e., local microbes at night and soil dust particles transported from Beijing  
617 areas by southerly wind in daytime. We found that vegetation and fungal spores are not a specific  
618 source of glucose and mannitol, respectively. Both sugars may have multiple sources in the forest  
619 aerosols.

620 PMF results concluded the contributions of 36% from vegetation (21% vegetation factor  
621 and 15% pollen factor) and 37% from microbial and fungal species (21% microbial soil dust and  
622 16% fungal factor) of total measured SCs. The BB activities for domestic cooking and space  
623 heating in north China contributed higher organic carbon at nighttime (0.43%) than in daytime  
624 (0.22%). Therefore, local BB seriously affected the air quality of the Mangshan site. Lev/Man  
625 ratio suggested that low temperature smoldering burning of hardwood is the main source for BB  
626 aerosols. SCs were recognized as a significant aerosol component at Mangshan, northern suburbs  
627 of Beijing. SCs can influence the air quality and thus climate because they are essential  
628 components of organic aerosols on a global scale. This study of SCs at Mangshan demonstrates  
629 that ambient levels of SCs are highly sensitive to the emissions of anthropogenic and biogenic  
630 aerosols. Higher contribution of levoglucosan to SCs demonstrated a significant BB activity  
631 around the Mangshan site in north China.

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633 **Data availability.** Raw data are available on request by contacting the corresponding author.

634 **Author contributions.** This research was designed YK, KK and ZW. Laboratory measurements  
635 were performed by FY with a support of PF. The paper was prepared by SKV and KK.

636 **Competing interests.** The authors declare that they have no conflict of interest.

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973 **Figure Captions**

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975 Figure 1. Geographical location of Mangshan, China. The map was downloaded from © Google  
976 Maps 2019.

977 Figure 2. (a) The meteorological parameters at Mangshan during sampling periods, (b) Fractions  
978 of local wind directions at Mangshan site, north of Beijing, China.

979 Figure 3. Concentrations ( $\text{ng m}^{-3}$ ) of sugar compound (a) overall, (b) daytime and (c) nighttime in  
980 aerosol samples from Mangshan during September-October 2007 (The error bars denote the  
981 standard deviation).

982 Figure. 4. Temporal variations in the concentrations ( $\text{ng m}^{-3}$ ) of anhydrosugars in the Mangshan  
983 aerosol samples collected for September-October 2007. (Solid circle represents nighttime  
984 samples collected from 18:00 to 09:00 hours. Hollow circle represents daytime samples).

985 Figure. 5. Temporal variations in the concentrations ( $\text{ng m}^{-3}$ ) of primary sugars in the Mangshan  
986 aerosol samples collected for September-October 2007. (Solid circle represents nighttime  
987 samples collected from 18:00 to 09:00 hours. Hollow circle represents daytime samples). **Y-**  
988 **axis shows temporal variations in the concentrations ( $\mu\text{g m}^{-3}$ ) of  $\text{Ca}^{2+}$ .**

989 Figure. 6. Temporal variations in the concentrations ( $\text{ng m}^{-3}$ ) of sugar alcohols in the Mangshan  
990 aerosol samples collected for September-October 2007. (Solid circle represents nighttime  
991 samples collected from 18:00 to 09:00 hours. Hollow circle represents daytime samples).

992 Figure 7. PMF analyses of sugar compounds in Mangshan aerosols based on the autumn 2007 data  
993 set.

994 Figure 8. Source contributions to sugar compounds from various sources based on PMF analyses.

995 Figure 9. The concentrations and relative contributions of the carbon content of anhydrosugars,  
996 primary sugars and sugar alcohols to the carbon concentrations of measured sugar  
997 compounds, water-soluble organic carbon (WSOC) and organic carbon (OC) fraction of  
998 Mangshan aerosols (a = daytime and b = nighttime). The concentrations and relative  
999 contribution of the carbon content of five sources of sugar compounds to total sugar  
000 compounds measured, WSOC and OC fraction of Mangshan aerosols (c = daytime and d =  
001 nighttime).

002 Figure 10. Mass concentrations ratio of carbon contents of (a) levoglucosan (Lev) to organic  
003 carbon (OC) and water soluble organic carbon (WSOC), (b) levoglucosan (Lev) to organic  
004 carbon (OC) daytime and night time, (c) levoglucosan (Lev) to water soluble organic carbon  
005 (WSOC) daytime and night time, (d) average levoglucosan to mannosan ratios (Lev/Man) in  
006 the Mangshan aerosol samples for autumn 2007.

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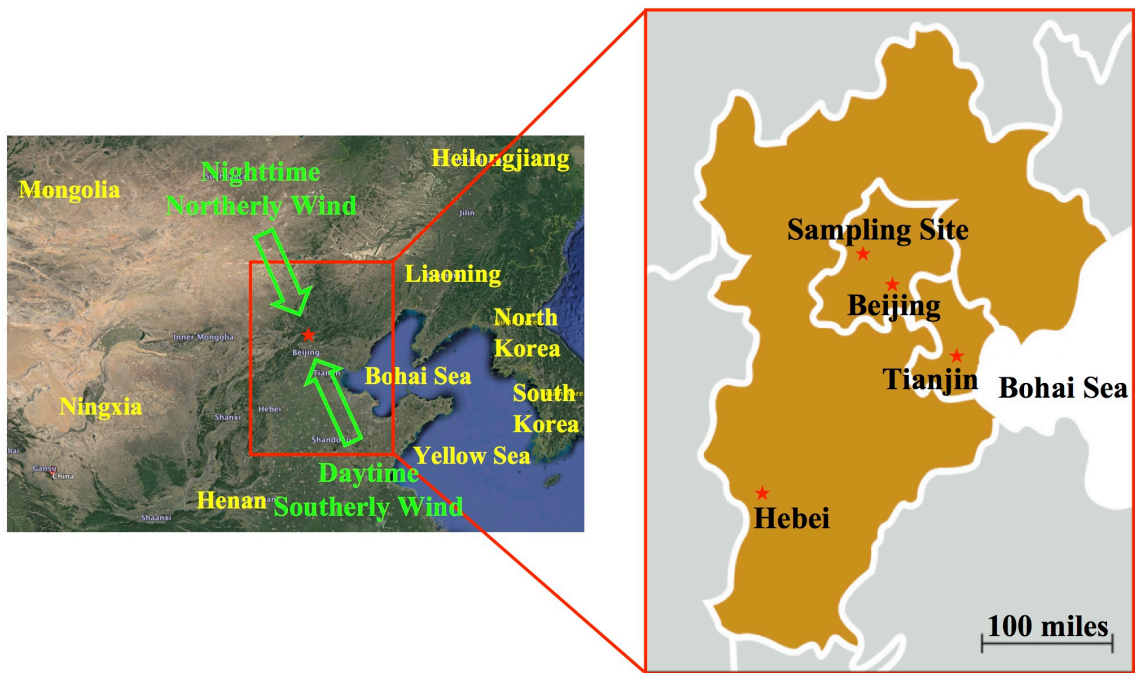
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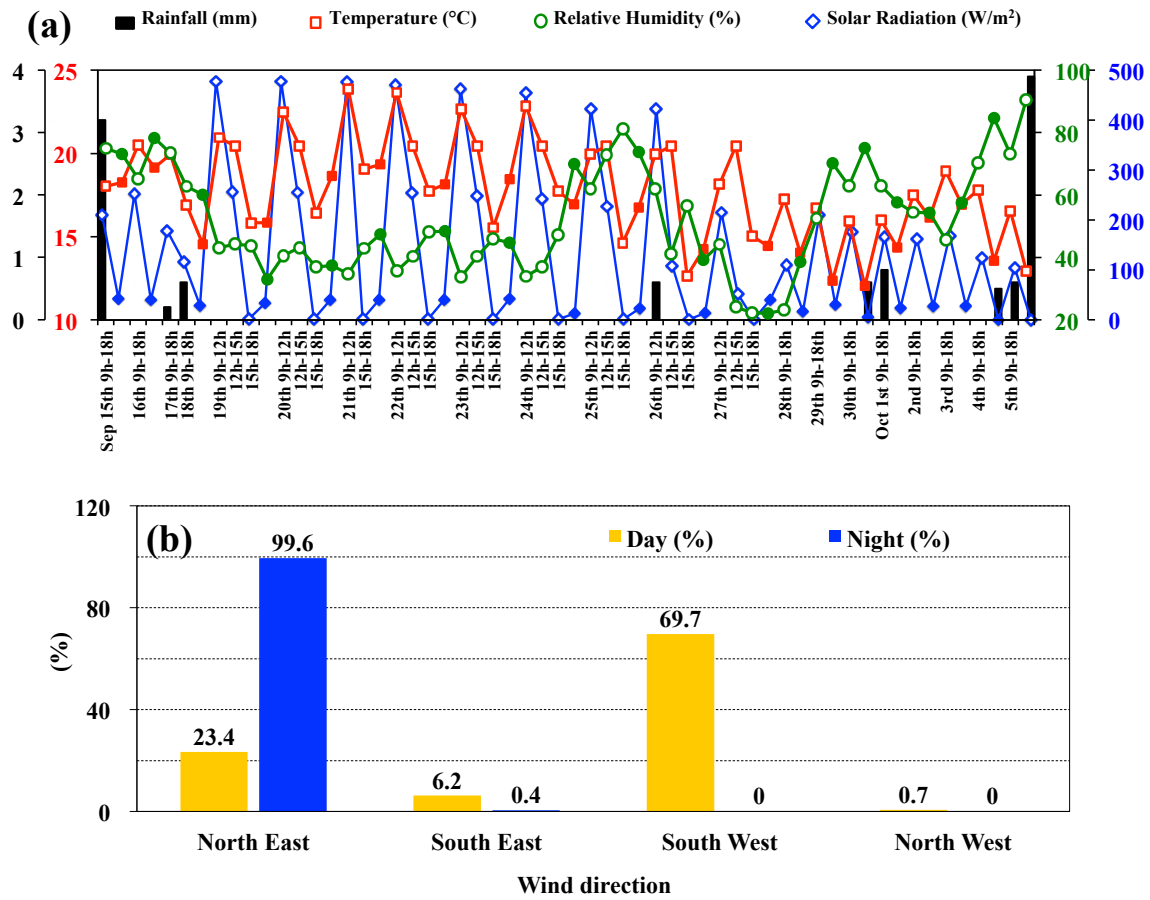
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Fig. 2.



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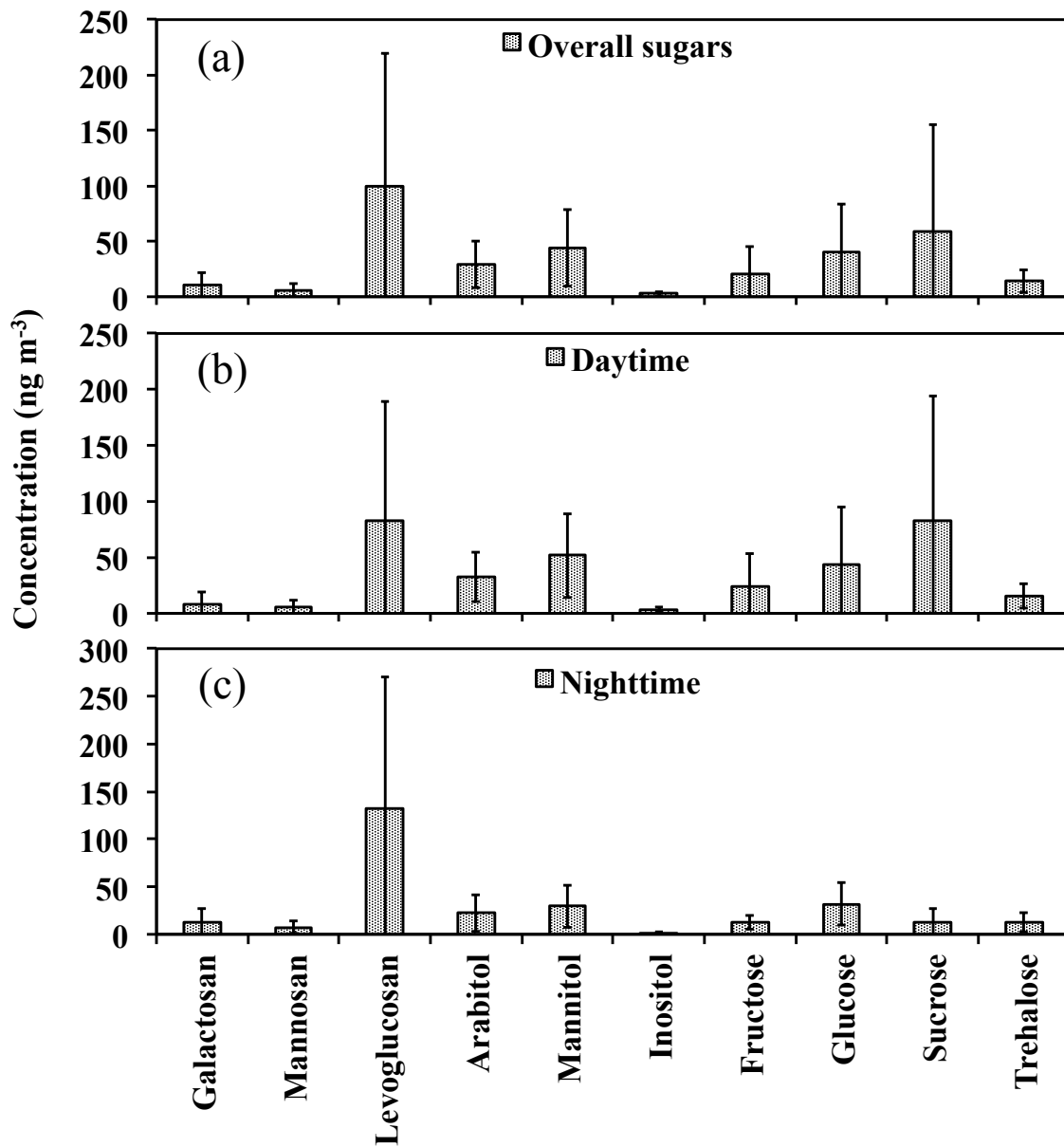
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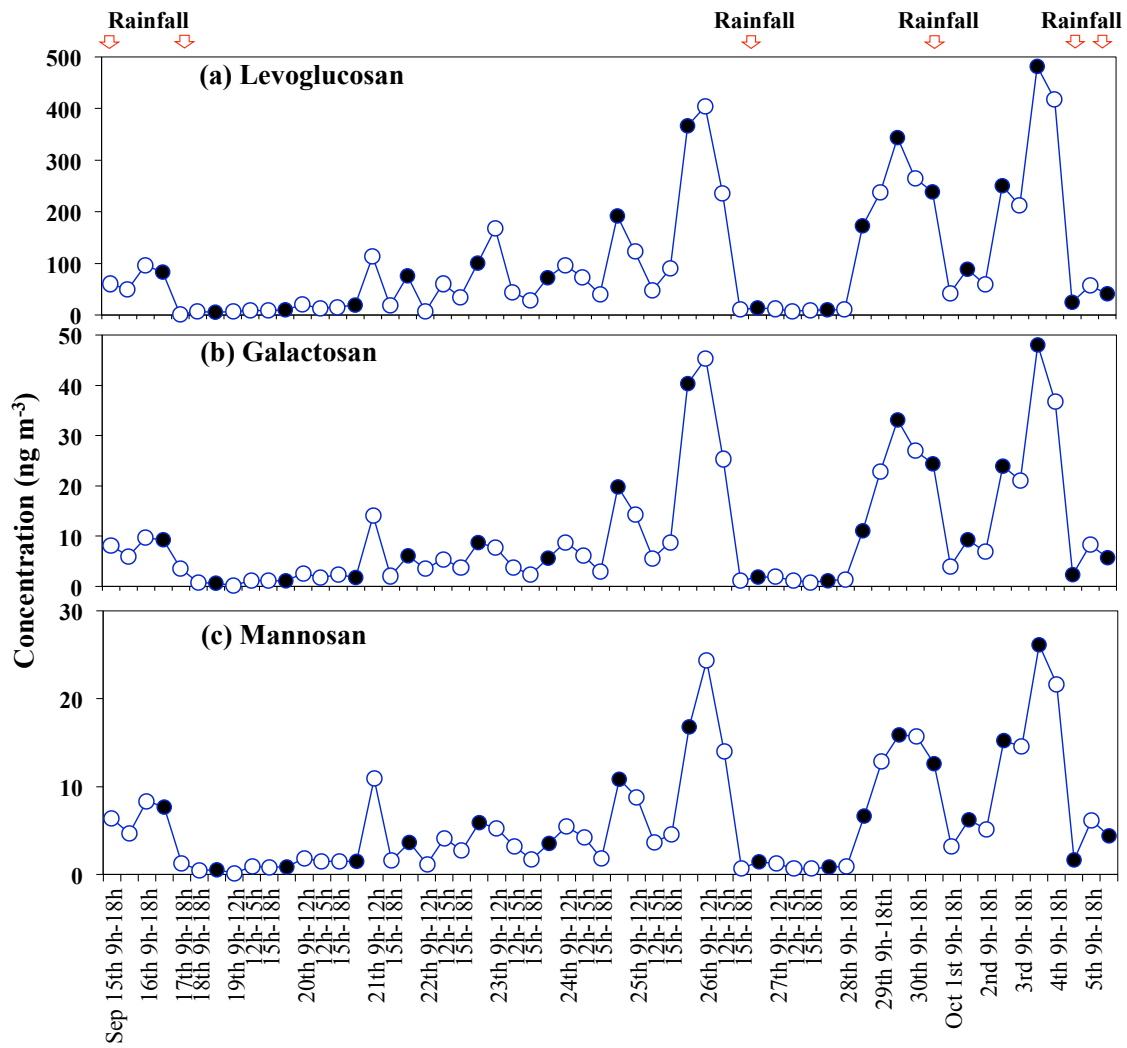
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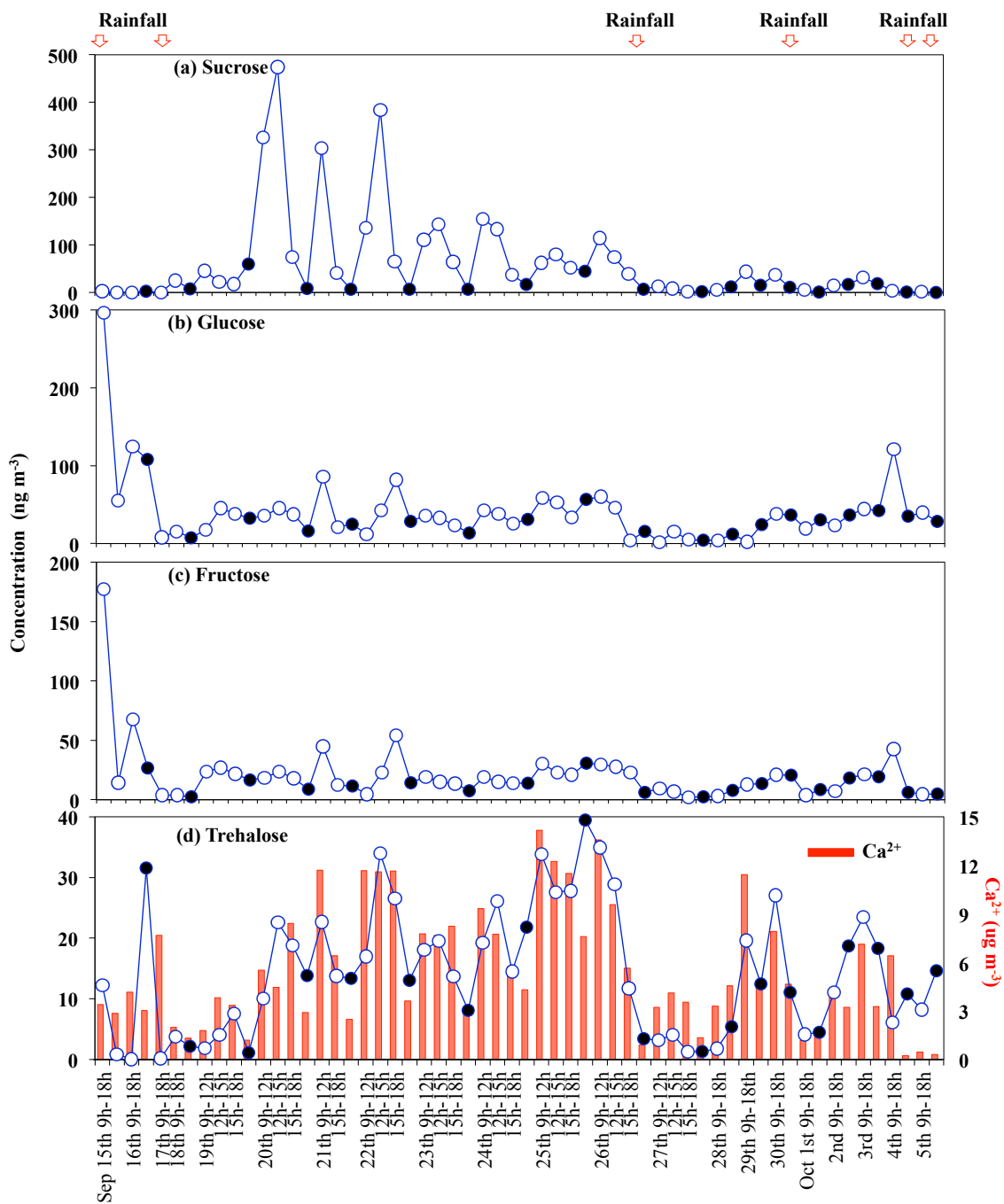
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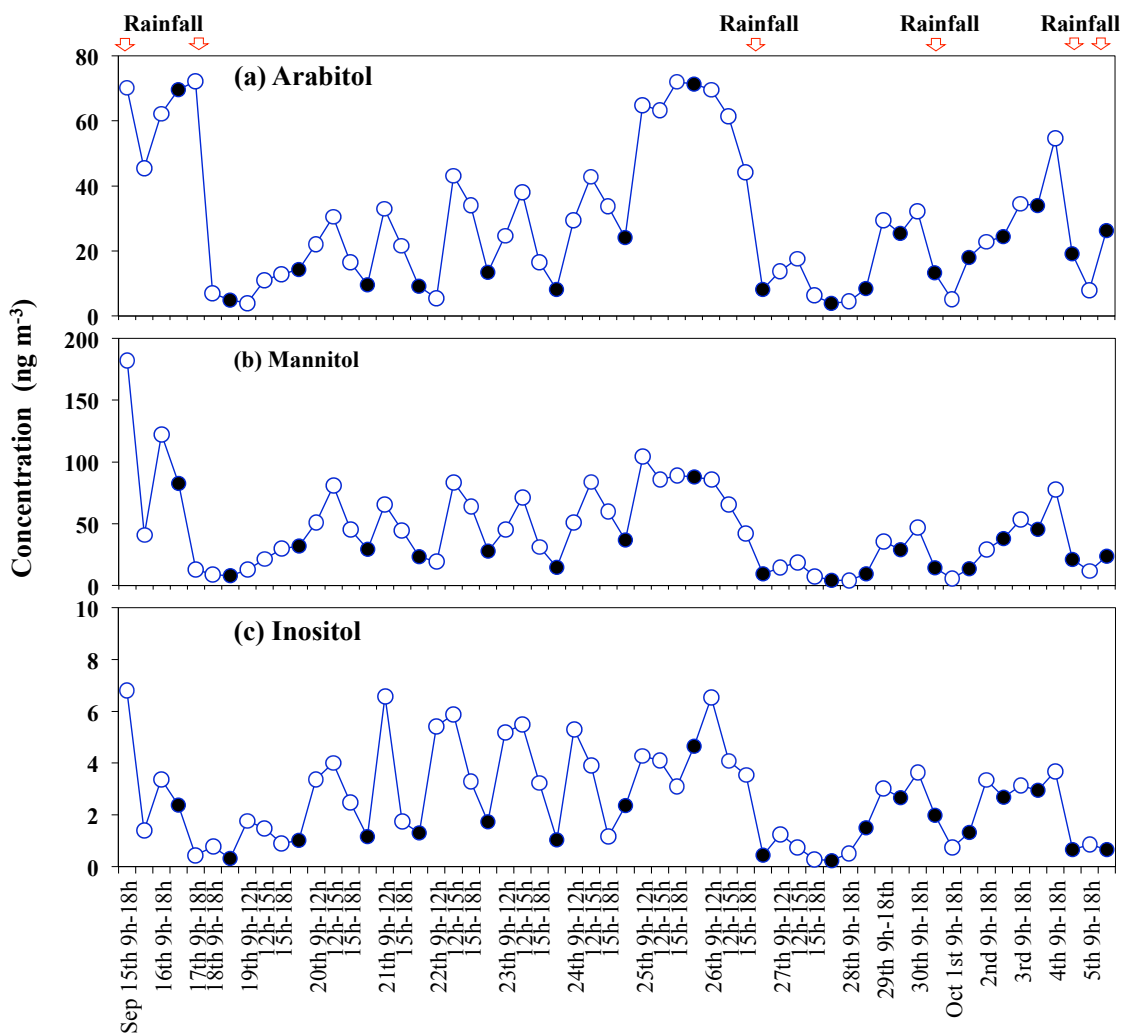
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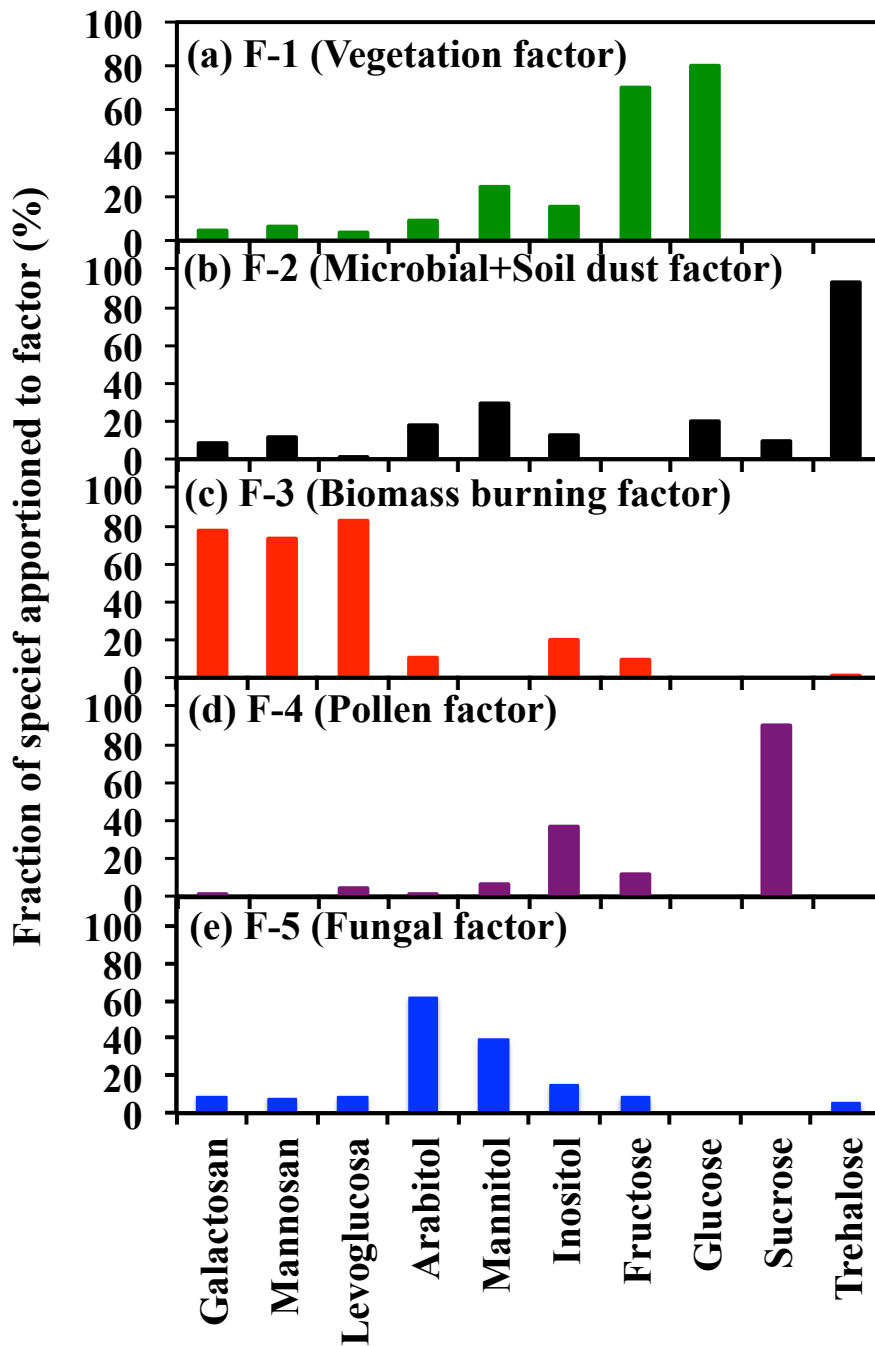
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Fig. 7.



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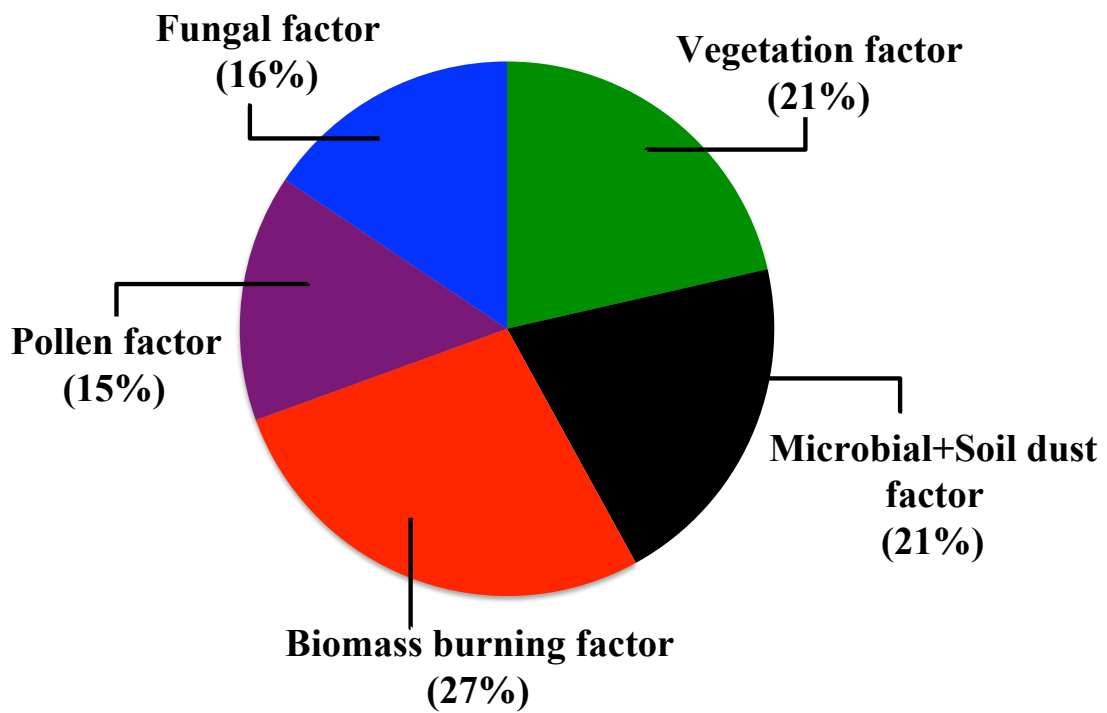
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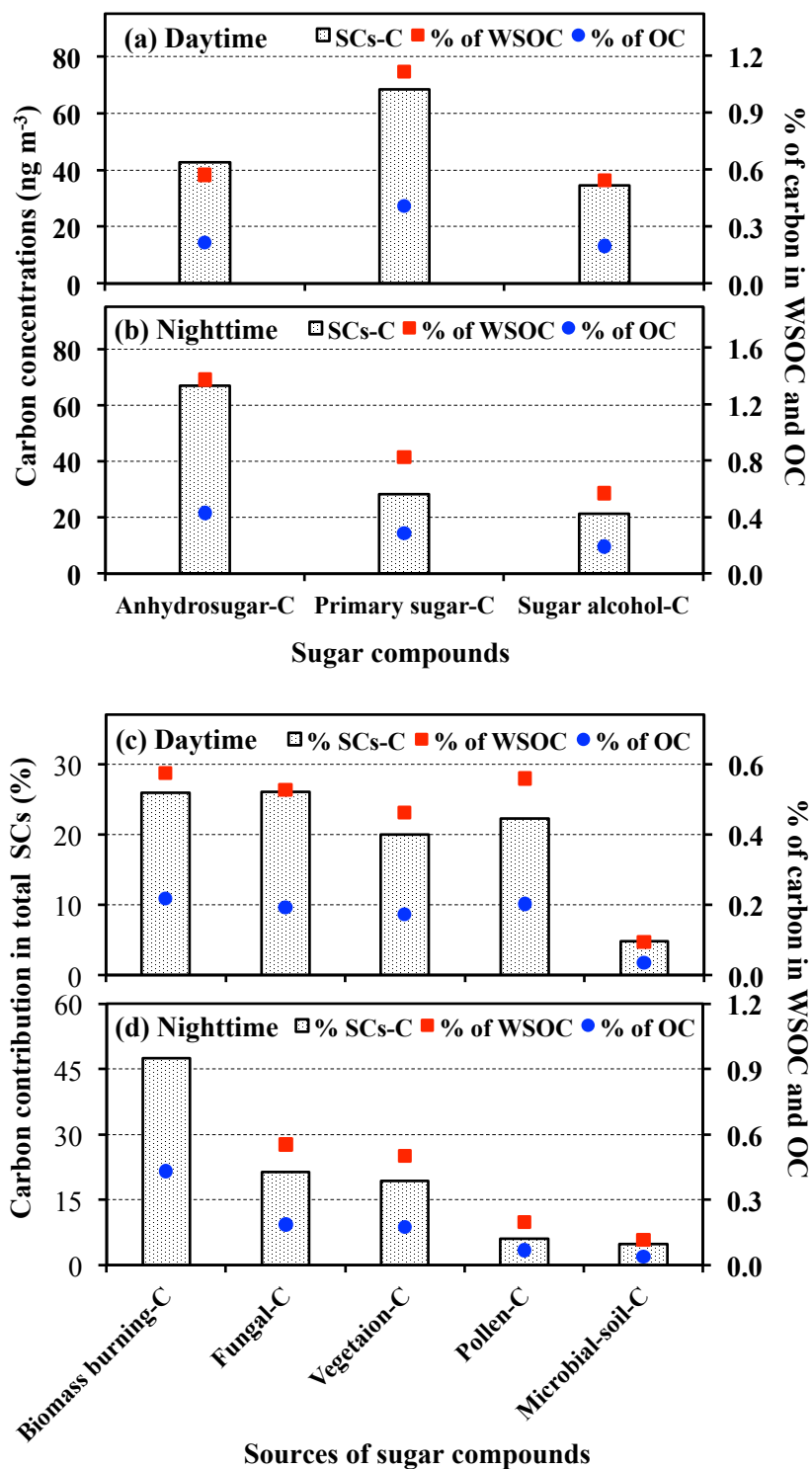
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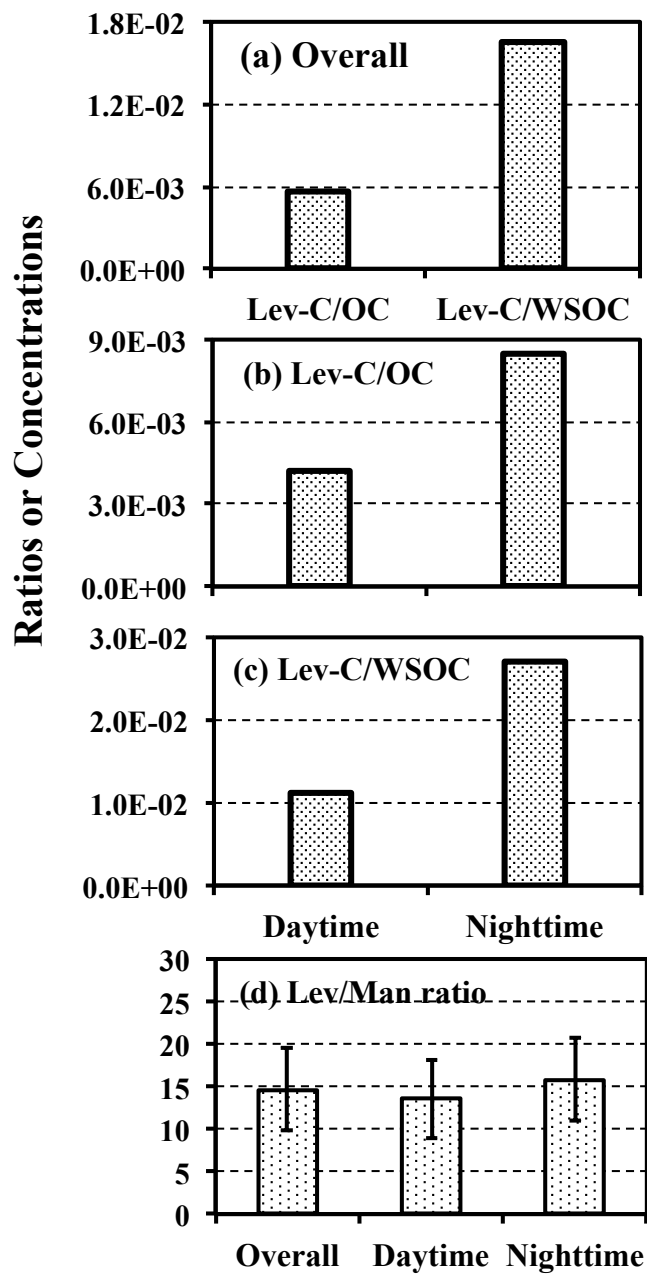
Fig. 9.



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Fig. 10.



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Table 1. Minimum, maximum, average and standard deviations of concentrations of sugar compounds in aerosol samples (TSP) from Mangshan, China.

Sugar Compounds	Overall				Daytime (n = 38)				Nighttime (n = 20)			
	Min	Max	Avg.	S.D.	Min	Max	Avg.	S.D.	Min	Max	Avg.	S.D.
<b>Anhydrosugars</b>												
Galactosan	0.14	48.0	10.1	11.9	0.14	45.3	8.53	10.5	0.69	48.0	13.0	14.0
Mannosan	0.13	26.1	6.05	6.33	0.13	24.3	5.37	6.01	0.53	26.1	7.35	6.87
Levogluconan	1.17	482	100	119	1.17	418	83.2	106	5.66	482	132	138
<b>Sugar alcohols</b>												
Arabitol	3.89	72.2	29.1	21.5	3.99	72.2	32.5	22.0	3.89	71.3	22.5	19.4
Mannitol	4.19	182	44.1	34.5	4.19	182	51.7	37.5	4.40	87.7	29.6	22.3
Inositol	0.23	6.8	2.62	1.81	0.27	6.80	3.14	1.90	0.23	4.65	1.62	1.09
<b>Primary sugars</b>												
Fructose	1.72	177	20.1	24.6	1.72	177	23.9	29.3	2.64	30.9	12.8	7.67
Glucose	1.86	297	40.0	43.4	1.86	297	44.2	50.8	4.52	108	32.0	22.8
Sucrose	0.02	474	58.5	96.5	0.02	474	82.9	112	0.04	60.1	12.3	15.1
Trehalose	0.06	39.5	14.3	10.5	0.06	34.9	15.3	10.6	0.87	39.5	12.3	10.2
Anhydrosugars	6.01	556	116	137	6.01	476	97.1	122	6.88	556	152	159
Primary sugars	9.41	565	133	125	9.41	565	166	141	10.5	172	69.4	43.0
Sugar alcohols	8.53	259	75.8	54.7	9.09	259	87.4	57.5	8.53	164	53.7	41.9
Total Sugars	30.8	875	325	232	34.1	875	351	240	30.8	759	276	212
Anhydrosugars (%)			31.9				24.6				45.7	
Primary sugars (%)			41.8				47.3				31.3	
Sugar alcohols (%)			26.4				28.1				23.0	

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Table 2. Statistical summary of correlations among the chemical species and meteorological variables in aerosol samples collected at a forest site in northern Japan

Linear regression	Correlation coefficient	p value	Significance of correlation at P value < 0.05
<b>Overall (n = 58)</b>			
Levogluconan vs. Galactosan	0.98	< 0.05	Significant
Levogluconan vs. Mannosan	0.97	< 0.05	Significant
Mannosan vs. Galactosan	0.98	< 0.05	Significant
Sucrose vs. Temperature	0.52	< 0.05	Significant
Sucrose vs. Solar radiation	0.55	< 0.05	Significant
Arabitol vs. Mannitol	0.81	< 0.05	Significant
Arabitol vs. RH	0.69	< 0.05	Significant
Mannitol vs. RH	0.57	< 0.05	Significant
Glucose vs. Fructose	0.94	< 0.05	Significant
Trehalose vs. Arabitol	0.58	< 0.05	Significant
Trehalose vs. Mannitol	0.58	< 0.05	Significant
Trehalose vs. Ca <sup>2+</sup>	0.70	< 0.05	Significant
<b>Daytime (n = 38)</b>			
Sucrose vs. Ca <sup>2+</sup>	0.32	> 0.05	Not significant
Glucose vs. Ca <sup>2+</sup>	0.02	> 0.05	Not significant
Trehalose vs. Arabitol	0.49	< 0.05	Significant
Trehalose vs. Mannitol	0.51	< 0.05	Significant
Trehalose vs. Ca <sup>2+</sup>	0.81	< 0.05	Significant
Fructose vs. Mannitol	0.79	< 0.05	Significant
Levogluconan vs. OC	0.45	< 0.05	Significant
Levogluconan vs. WSOC	0.40	< 0.05	Significant
<b>Nighttime (n = 20)</b>			
Sucrose vs. Ca <sup>2+</sup>	0.37	> 0.05	Not significant
Glucose vs. Ca <sup>2+</sup>	0.27	> 0.05	Not significant
Trehalose vs. Arabitol	0.76	< 0.05	Significant
Trehalose vs. Mannitol	0.85	< 0.05	Significant
Trehalose vs. Ca <sup>2+</sup>	0.61	< 0.05	Significant
Fructose vs. Mannitol	0.86	< 0.05	Significant
Levogluconan vs. OC	0.81	< 0.05	Significant
Levogluconan vs. WSOC	0.70	< 0.05	Significant

The data of Ca<sup>2+</sup>, OC and WSOC are adapted from He et al. (2015).

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