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The nitrogen budget of laboratory-simulated western U.S. wildfires during the FIREX 2016 FireLab study

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Abstract. Total reactive nitrogen (N_r, defined as all nitrogen-containing compounds except for N_2 and N_2O) was measured by catalytic conversion to NO and detection by NO-O_3 chemiluminescence together with individual N_r species during a series of laboratory fires of fuels characteristic of Western U.S. wildfires, conducted as part of the FIREX FireLab 2016 study. Data from 75 stack fires were analyzed to examine the systematics of nitrogen emissions. The N_r/total-carbon ratios measured in the emissions were compared with fuel and ash N/C ratios and mass to estimate that a mean (±std. dev.) of 0.68 (±0.14) of fuel nitrogen was emitted as N_2 and N_2O. The remaining fraction of N_r was emitted as individual compounds: nitric oxide (NO), nitrogen dioxide (NO_2), nitrous acid (HONO), isocyanic acid (HNCO), hydrogen cyanide (HCN), ammonia (NH_3), and 44 nitrogen-containing volatile organic compounds (NVOCs). The relative difference between the total reactive nitrogen measurement, N_r, and the sum of measured individual N_r compounds had a mean (±std. dev) of 0.152 (±0.098). Much of this "unaccounted" N_r is expected to be particle-bound species, not included in this analysis.

A number of key species, e.g. HNCO, HCN and HONO, were confirmed not to correlate only with flaming or only with smoldering combustion when using modified combustion efficiency (MCE = CO_2/(CO + CO_2)) as a rough indicator. However, the systematic variations of the abundance of these species relative to other nitrogen-containing species were successfully modeled using positive matrix factorization (PMF). Three distinct factors were found for the emissions from combined coniferous fuels, aligning with our understanding of combustion chemistry in different temperature ranges: a combustion factor (Comb-N) (800-1200°C) with emissions of the inorganic compounds NO, NO_2 and HONO, and a minor contribution from organic nitro compounds (R-NO_2); a high-temperature pyrolysis factor (HT-N) (500-800°C) with emissions of HNCO, HCN and nitriles; and a low-temperature pyrolysis factor (LT-N) (<500°C) with mostly ammonia, and NVOCs, with the temperature ranges being based on known combustion and pyrolysis chemistry considerations. The mix of emissions in the PMF factors from the

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chaparral fuels had a slightly different composition: the Comb-N factor was also mostly NO, with small amounts of HNCO, HONO and NH_3, the HT-N factor was dominated by NO_2 and had HONO, HCN, and HNCO, and the LT-N factor was mostly NH_3 with a slight amount of NO contributing. In both cases, the Comb-N factor correlated best with CO_2 emission, while the HT-N factors from coniferous fuels correlated closely with the high temperature VOC factors recently reported by Sekimoto et al., (2018) and the LT-N had some correspondence to the LT-VOC factors. As a consequence, CO_2 is recommended as a marker for combustion N_r emissions, HCN is recommended as a marker for HT-N emissions and the family NH_3/particle ammonium is recommended as a marker for LT-N emissions.

1 Introduction

Wildfires have severe impacts on the chemistry of the atmosphere from local to global scales (Crutzen and Andreae, 1990). A warmer, drier climate in western North America, coupled with policies that have allowed build-up of fuels in forest ecosystems has led to increases in frequency and severity of wildfires in this region (Abatzoglou and Williams, 2016; Westerling et al., 2006). The new strategy for management of wildfire in the U.S. is to allow fire where possible and to fight fire where needed (Lee et al., 2014). The science behind making these decisions and understanding their consequences involves, in part, a better understanding of the emissions from wildfires. The NOAA FIREX (Fire Influence on Regional and Global Environments Experiment) FireLab experiment was conducted in the Fall of 2016, at the U.S. Forest Service Fire Sciences Laboratory in Missoula, Montana, to acquire detailed measurements of particle and gas-phase emissions from fires involving fuels characteristic of the western U.S. (NOAA, 2018). Several aspects of these measurements dealing with VOC species, and individual reactive nitrogen species (N_r, defined as all nitrogen compounds except for N_2 and N_2O) have already been published (Koss et al., 2018; Manfred et al., 2018; Sekimoto et al., 2018; Selimovic et al., 2018; Zarzana et al., 2018), including emissions factors for many of the N_r-species (Koss et al., 2018).

The N_r compounds emitted by natural-convection biomass burning (BB) arise solely from the N in the fuels, since the combustion temperatures are not high enough (<1200°C) to produce NOx from N_2 and O_2 (the so-called Zeldovich or thermal nitrogen cycle) (Lobert and Warnatz, 1993; Taylor et al., 2004; Wotton et al., 2012). The fuel nitrogen cycles that pertain to BB flaming combustion are shown schematically in Figure 1 (Glarborg et al., 2018; Lobert and Warnatz, 1993; Lucassen et al., 2012). N_r compounds are emitted as small molecules, HCN, HNCO and NH_3 resulting from pyrolysis of the fuel, with minor contributions from larger N-containing organic species, especially at lower temperatures. Flame chemistry converts those species to N_2, N_2O, NO, NO_2, and HONO as a result of radical chemistry. It has been recognized for some time that a significant amount of denitrification (conversion of N_r compounds to N_2) occurs due to reactions of NO with NH_i (where i= 1, 2, or 3) or N atoms, as confirmed experimentally (Kuhlbusch et al., 1991). While N atoms are also intermediates in the thermal NO_x cycle and the reaction $N + O_2 \Rightarrow NO + O$ figures in to both the fuel and thermal NO_x cycles, the second reaction of the thermal NO_x cycle, $O + N_2 = NO + N$, is too slow at BB flame temperatures to result in NO_x production (Manion et al., 2015). In addition to the small molecules shown in Figure 1, numerous N_r-compounds are emitted in roughly the following categories: amides, amines, heterocyclic compounds, nitriles, isocyanates, and nitro compounds (Andreae, 2019; Andreae and Merlet, 2001; Koss et al., 2018; Lobert et al., 1991; Lobert et al., 1990; Lobert and Warnatz, 1993;

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Stockwell et al., 2015). These compounds are produced at much lower abundance from fuel pyrolysis and partial reactions with the radical species in Figure 1.

The emissions of N-compounds from BB and wildfires in general have been the subject of considerable research (Akagi et al., 2011; Andreae, 2019; Andreae and Merlet, 2001; Burling et al., 2010; Coggon et al., 2016; Gilman et al., 2015; Kuhlbusch et al., 1991; Lobert et al., 1991; Lobert et al., 1990; Lobert and Warnatz, 1993; McMeeking et al., 2009; Stockwell et al., 2015; Veres et al., 2010; Warneke et al., 2011; Yokelson et al., 2013b; Yokelson et al., 2009). The known N-compounds range in oxidation state from NH_3 to HNO_3 and include N_2 and N_2O. Among the more prominent and important N_r species are: NO_x (NO and NO_2) which is a key player in the atmospheric oxidant cycle; NH_3 which has a major role in particle formation; nitrous acid (HONO) which can be an important radical source; hydrogen cyanide and acetonitrile (HCN, CH_3CN) which are toxic at high concentrations and represent valuable tracers for following fire transport; and isocyanates, isocyanic acid and methyl isocyanate (HNCO, CH_3NCO) which have unique health impacts (Roberts et al., 2011). In addition, nitro (-NO_2), or nitrogen heterocyclic compounds may contribute to so-called brown carbon, aerosol organic compounds exhibiting optical absorption in the near-UV or blue wavelength regions. Wildfire N emissions also have very minor contributions from gas phase nitric acid (HNO_3). Nitric acid is either not efficiently produced by BB or is readily incorporated into aerosol if it is produced in fresh wildfire plumes, as is clear from the absence of HNO_3 enhancements in several studies of BB plumes (Liu et al., 2016; Yokelson et al., 2009) (Alvarado et al., 2010), however nitrate (NO_3^-) has been shown to contribute to aerosol mass particularly for inefficient combustion (May et al., 2014). Flame chemistry is inefficient in forming N_2O, relative to the pathways that form N_2 (Andreae, 2019; Andreae and Merlet, 2001; Griffith et al., 1991; Hao et al., 1991). The modeling of the emissions of these N-compounds on a large scale could benefit from a better understanding of the total budget of these species as a function of fuel nitrogen content and the dependence of the individual species on fuel type and combustion conditions.

The construction of N_r-budgets in this work is made possible by the inclusion of a total reactive nitrogen measurement (termed N_r herein), a method by which all nitrogen compounds besides N_2 and N_2O are converted to NO and detected by NO-O_3 chemiluminescence. This technology has been developed by a number of groups, typically using precious metal or NiCr catalysts that have been shown to convert all N_r compounds to NO (and to some extent NO_2) at high temperatures (750-825°C) (Hardy and Knarr, 1982; Kashihira et al., 1982; Marx et al., 2012; Roberts et al., 1988). There are also commercial instruments that incorporate this technology (see for example Thermo Scientific Model 17i). This technique has been applied to gas phase atmospheric measurements, principally to measure NH_3 by difference techniques (Saylor et al., 2010; Schwab et al., 2007), and has also been used to observe wildfire plumes that have impacted ambient air measurements (Benedict et al., 2017; Prenni et al., 2014). We have recently developed a platinum/molybdenum oxide N_r catalyst system, and confirmed that it quantitatively converts N_r compounds including all particle-bound nitrogen compounds (Stockwell et al., 2018). To our knowledge this technique has not been applied directly to BB emissions before.

This paper will describe the total reactive nitrogen, and individual N_r compound measurements made during the FireLab 2016 experiment. The total N_r measurements will be combined with CO_2, CO, and VOC measurements and fuel, residue and ash C and N content to estimate the amount of N lost to N_2 and N_2O. Fire-integrated N_r will be

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compared to fire-integrated measurements of individual compounds to determine the fraction of unaccounted-for N_r. The systematic behavior of individual N_r species and their fractional contribution to N_r will be examined with respect to fuel type, N content, and combustion processes. A positive matrix factorization (PMF) technique will be used to examine commonalities between fires of different fuels under different conditions and compared to the PMF analysis of the VOC emissions published by Sekimoto et al., (2018). The results will be used to arrive at suggested guidelines that can be used estimate N_r-emissions profiles for fires representative of western North America.

2 Methodology

The FireLab 2016 study involved laboratory burns of fuels mostly characteristic of western North American wildfires, but also some that have global significance such as Indonesian peat and yak dung. The procedures and associated details of the study have been described previously by Selimovic et al., (2018) and will be only briefly summarized here. The detailed data on fuel types, amounts and composition can be found in Table S1, and in the Supplemental section of Selimovic, et al., (2018). The laboratory burns involved fuel samples, ranging in mass from 0.26 to 6.02 kg. Fires were started without the addition of any contaminants, using an electric igniter (a series of NiCr heating elements that were flash-heated electrically), and typically lasted from approximately 5 to 30 minutes. Seventy-five fires were conducted in the configuration where the smoke was directed up the central stack of the facility where it could be sampled simultaneously by all the instruments that measured gas phase species, and some of the particle phase measurements. The sampling platform was about 15 m above the fire and the sampling took place in well-mixed smoke approximately 5s after emission (Christian et al., 2004). Thirty-one additional fires were conducted on most of the same fuels, when the stack was closed and the room was allowed to fill with smoke, permitting sampling to be done over the course of several hours. The following analyses will focus on the "stack" burns, as those measurements had little or no interferences from surfaces, where "room" burns are known to be compromised by the loss of materials, such as NH_3, to the room walls at long sample times (Stockwell et al., 2014). Ash analyses were performed only on the residues from the room burns and those values will be used for the N and C budget calculations, with the assumption that stack and room burns left similar ash considering the combustion conditions were the same for each type of fire. Table 1 lists the compounds and associated techniques used to measure them during the FireLab 2016 study, and describes the grouping of NVOCs measured by PTR-ToF into common categories, e.g. amines, nitriles, etc.

2.1 N_r and NO measurements by Chemiluminescence

Total reactive N (N_r) was measured by catalytic conversion to NO, followed by O_3-chemiluminescence using an instrument described previously (Williams et al., 1998). N_r and NO were sampled from inlets inserted adjacent to the inlet-less open-path Fourier transform infrared spectrometer (OP-FTIR) instrument path during the stack burns (Selimovic et al., 2018), and from a platform approximately 4 m off the floor in the middle of the room during the room burns. The catalyst used for the N_r channel, described in detail by Stockwell et al. (2018), consisted of a 11mm I.D. quartz tube, packed with 36 platinum screens, heated to 750°C. This tube was wrapped with high temperature heating tape and insulated inside a 7cm OD stainless steel tube that was fitted to a bulkhead placed through the wall

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of the stack. The N_r channel was diluted by a factor of 5:1 (±3%) using a flow of zero air added immediately downstream of the Pt catalyst assembly. NO was sampled through a 6.3mm O.D. stainless steel inlet tube which was placed through the bulkhead directly into the free air stream of the stack and connected to a 50mm Teflon filter holder immediately outside the stack. The transfer lines for the N_r and NO measurements consisted of 6.35mm O.D, 1mm wall thickness PFA tubing of approximately 20 m in length. N_r and NO data were acquired at 1 s frequency, but the flow rate through each inlet was 1 SL min^{-1}, resulting in residence time in each inlet of 14 s. This time delay was corrected in the data analysis. Any chemical effects of the inlet on the sampled air stream were negligible since the analytes consisted of only NO and NO_2 and those are known to be transmitted by PFA Teflon tubing with essentially no surface effects. However, there were possible effects of the inlets on the temporal features of the measurement through diffusion or turbulent mixing. Those effects were examined through comparison of the temporal variations in the NO signal with the NO measured by the OP-FTIR, and comparison of the N_r signal under smoldering conditions with the NH_3 measured by the OP-FTIR. Both of these comparisons showed that the NO and N_r inlets had effective time constants of 4 seconds, somewhat slower than the diffusive relaxation time assuming solely laminar flow.

The inlet streams were sampled into the NO instrument either directly (NO channel) or after passing through a second catalyst of molybdenum oxide (MoOx) to convert remaining NO_2 to NO. The MoOx catalyst consisted of a molybdenum tube at 350°C to which a small flow of H_2 (0.8%v/v) was added to control the re-dox state of the surface. Both channels of the instrument were "de-tuned" to keep raw photon count rates below 4 MHz, by turning down the O_3 flows and PMT voltages. Calibrations were performed with both a NO standard in N_2 (Scott-Marrin) and 10.1 ppmv standard of HCN in nitrogen (Gasco). The Pt catalyst was dismounted from the stack (or room) every few days and checked for conversion efficiency by the addition of the HCN standard to the inlet. Conversion efficiencies were found to be consistently high (>98%) throughout the entire sampling period (October 5 – November 12, 2016). There were slight background signals (a few tens of ppbv) for both NO and N_r in both the stack and room air prior to and after the burns, and those were subtracted from the fires signals prior to reporting the data. The overall uncertainties in the NO and N_r data were ±10% for each measurement.

2.2 Other measurements

Measurements of individual species during the 2016 FireLab study have been presented in several previous publications. The OP-FTIR measurements were discussed by Selimovic et al., 2018, and the PTR-ToF measurements were discussed by Koss et al., (2018). In addition, some of the calibration methods and GC separation and identifications rely on additional analytical work presented by Sekimoto et al., (2017) and Gilman et al., (2015). We measured the mass and elemental content of the initial fuel and the mass of unburned fuel for all the fires, and we measured the mass and the elemental content of the ash during 21 room burns, which covered all the fuel types discussed.

2.3 PMF Analysis

Trace gas measurements from multiple instruments involved in the FireLab study were combined and analyzed using positive matrix factorization (PMF). PMF is a numerical method that was used in this case to partition

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the compounds involved in a time varying mixture of chemicals into a few groups, or "factors", where a compound can appear in more than one factor. A factor represents a consistent profile of compounds that is representative of one of the sources contributing to the total signal. The sum of all the "factors" then ideally describes the total composition of the measurements, which in this case is the emissions of N_r compounds. By its nature, PMF assumes that the total signal is a linear combination of individual sources that have a consistent composition, the relative contribution of which is represented by the amount of each compound or category found in each factor (Paatero and Tapper, 1994; Ulbrich et al., 2009). We hypothesize that species with dominant fractions in the same factor are related to each other via the same formation processes. With knowledge of factor composition and the amount of each factor at any given time the original emissions measurements can be reconstructed and this approach provides an alternate source of profiles for fire emissions. PMF has also been used by a number of groups to explore how much various source profiles contribute to complex ambient measurements (see for example Ulbrich et al., 2009) and was recently used to analyze PTR-ToF-MS measurements from the FireLab (Sekimoto et al., 2018). Here, PMF was accomplished using the PMF Evaluation Tool v. 2.08A (Ulbrich et al., 2009).

The application of PMF to this data set is different than the instances where it is applied to data from a single instrument in which compound abundances are inherently scaled properly and error estimates are well defined and self-consistent. For example, when applied to mass spectral data from a single instrument, errors can be expected to scale as the square root of ion counts based on fundamental counting statistics (Sekimoto et al., 2018). In this work we are including nitrogen measurements from several instruments, thus we chose to use mixing ratios as the unit of comparison. The error estimates required by the PMF analysis were taken from the reported combined uncertainties: the sum of the detection limit plus the estimated random error of the measured value. The variables that were used in this PMF analysis and their units and corresponding errors are listed in Table 2. Where compound categories are specified (e.g. nitriles), the values were the sum of the measured compounds in that category as listed in the footnotes to Table 1. The data were further adjusted by subtracting the ambient air background before and after the fires, which was a relatively minor adjustment for most compounds and categories. Any negative numbers that resulted were very small compared to the fire emissions, and were set to zero.

Two approaches were taken when performing PMF analyses. The first approach included all individual N-containing compounds together with CO_2, CO, and N_r in the analysis batch (Batch 1), while the second excluded the latter three species (Batch 2). CO_2 and CO were included because of their well-known roles as indicators of flaming and smoldering combustion, respectively, an effect traditionally captured through the use of Modified Combustion Efficiency (MCE) defined as

$$MCE = \Delta CO_2/(\Delta CO_2 + \Delta CO) \quad \text{(Eq. 1)}$$

where ΔCO_2 and ΔCO are the CO_2 and CO levels above the ambient. When CO_2 and CO were included, the carbon species were put into the PMF in units of ppmv, and all the nitrogen species in units of ppbv. This was done because the nitrogen levels were on the order of a percent or less compared to the carbon species. The second analysis batch (Batch 2) involved only the individually-measured nitrogen species and categories listed in Table 1 so that factor

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loadings would be reflective of the nitrogen-only emissions. Batch 2 factors indicate how N_r species are related to each other via combustion chemistry.

We applied PMF to single fire data as well as extended time series that included all fires of a particular fuel type, in-line with the approach laid out by (Sekimoto et al., 2018). By consolidating fuels from a particular vegetation type, the fire to fire variability largely driven by differences in the fuel (e.g. moisture content, structure, quantity) is constrained and the most representative fire conditions are captured. Two fuel groups were analyzed in this way: the western U.S. coniferous ecosystem fuels which included ponderosa pine, lodgepole pine, Douglas fir, Engelmann spruce, and sub-alpine fir and the chaparral ecosystem which was represented by chamise and manzanita. The consolidated time series for the coniferous ecosystems included realistic mixtures, canopy only, and litter only, while duff and rotten logs were analyzed separately, and not included in the timeseries.

3 Results and Discussion

Example timeseries of NO, N_r, ΔCO, ΔCO_2 (CO and CO_2 corrected for their backgrounds) are shown in Figure 2, for a fire burning a sample of ponderosa pine realistic mix (Fire 004). MCE was also plotted in addition to the chemical species. The timeseries for Fire 004 shows a short initial smoldering/distillation phase (MCE 0.7 to 0.8) as heat pyrolyzes the fresh fuel and releases VOCs from exisiting pools in the fuel followed after ignition by a relatively efficient mix of flaming and smoldering combustion (MCE 0.95 to 0.98) and then finally a subsequent period of essentially pure smoldering (MCE ~0.80). The N_r and NO timelines had many features in common because NO is often the most abundant N_r compound (see below). As a result, it is useful to compare the quantities N_r–NO and (N_r-NO)/N_r to the other measures of chemical species or combustion efficiency. As expected, (N_r-NO)/N_r, in Figure 2(c) is anti-correlated with MCE since N_r is primarily NO at high MCE. In addition to the anti-correlation, this non-NO fraction, like its approximate carbon analog CO/CO_2, has a wider dynamic range than MCE and will often suffer less from background variability than carbon-based indices (Yokelson et al., 2013a).

The concentration profiles of the background-corrected measurements of N_r, CO_2, CO, and all the carbon-containing species measured by the FTIR (Selimovic et al., 2018) during the stack burns were integrated over the entire time of the burn to obtain total carbon, termed TC here, and total N_r. The additional carbon species included methane and a number of other gas phase VOCs as well as organic- and black-carbon aerosol. Altogether these carbon species should account for ≥98% of emitted carbon (McMeeking et al., 2009). Total N_r is plotted in Figure 3, versus TC (Figure 3a) and versus nitrogen burned, which is calculated from the %N in the fuel times the mass of fuel consumed (Figure 3b). The points in Figure 3 are colored by the fuel N/C mole % obtained from the elemental analysis of each fuel. The relationship between N_r and TC in panel 3a clusters around the 0.37% line and those points are from fuels most characteristic of the North American biomes impacted by wildfire. There are clear outliers in the correlation of N_r and TC; for example, yak dung and two samples of duff were high due either to the fact that they have high fuel N/C ratios, or they burned with minimal flaming (whole fire MCEs 0.86-0.89), hence experienced less de-nitrification. The fuels that were low in N_r/TC in panel 3a, ponderosa pine rotten log, subalpine fir and excelsior, had low fuel N/C, so when plotted versus nitrogen burned in panel 3b, they cluster with the main group of characteristic fuels, i.e. they are no longer 'outliers' in the distribution.

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266 The points in Figure 3a are all lower than the corresponding fuel N/C mole ratio, due to the denitrification
267 chemistry, shown in Figure 1, and verified in lab studies described by Kuhlbusch et al. (1991), and the production of
268 N_2O which is also not measured by the N_r technique. The sum of N_2 and N_2O produced in the fires can be estimated
269 from the difference between the fuel N/C and the N_r/Total C emitted and the data on C and N content remaining in
270 the ash. The mass balance equations used for these calculations are detailed in the Supplemental Materials.
271 The distribution of the N lost to N_2 and N_2O is shown in Figure 4. Chemical analyses were not done for all
272 fuels during the stack burns, and the analysis above assumes that the ash residues and ash/burned fuel ratios from the
273 stack burns were well represented by those for the same fuels used in the room burns, for which mass yields and
274 chemical analyses were done. Data are missing for fuels that did not have a corresponding ash analysis. The median
275 fraction of N lost to N_2 and N_2O for ash-corrected fires was 0.70, and the mean (±standard deviation) was 0.68 (±0.14).
276 This fuel-based estimate is uncertain by approximately 25% because of the above assumptions concerning the
277 applicability of the ash analyses from the room burns and because fuel moisture corrections were assumed to apply to
278 all of the materials burned, foliage vs. woody biomass (see SI for details). The emission of N_2O relative to N_2 is
279 approximately 10% or less for a wide range of fuels (Andreae, 2019). Assuming the N remainder in our work is at
280 least 90% N_2 gives values that are somewhat higher than the N_2 values reported by Kuhlbusch et al., (1991) where N_2
281 accounted for 36% of fuel N burned in flaming stage fires. A closer inspection of Kuhlbusch et al., (1991) showed a
282 range of N_2 yields of 40-54% at highest MCEs of 0.94-0.97. Possible reasons for these differences are that the
283 Kuhlbusch et al., (1991) fires were limited to grasses, hay, and pine needles, and the fires were confined to a closed
284 container and so may not have experienced the convection and turbulence of typical biomass fires. In addition, the
285 fires analyzed in our work were somewhat weighted towards the full canopy and higher temperature burning fuels,
286 since ash analyses were not done for peat, dung and many of the "litter" samples, all of which tend to burn less
287 efficiently. Goode et al., (1999) estimated an N_2 emission of 45±5% for MCE values of 0.95 in grass and surface fuels.
288 The range of values determined in our work overlap with these literature values, but are on average higher. It should
289 be noted that such loss of reactive nitrogen can have implications for ecosystem N budgets, as discussed by Kuhlbusch
290 et al., (1991).
291 The composition of the N that does not get converted to N_2 or N_2O is of intense importance in determining
292 atmospheric impacts of fires. Emission factors for all the individual N_r compounds identified in our work have been
293 compiled and reported in previous publications (Koss et al., 2018; Selimovic et al., 2018), so this paper will focus on
294 the N_r budget. The balance of N_r budget for Fire 047, sub-alpine fir realistic mix, is shown in Figure 5, in which the
295 timelines of N_r, NO, N_r-NO, sumN, and NVOC are plotted along with MCE and (N_r-NO)/N_r. The quantity sumN is
296 the sum of all other non-NO compounds, and NVOC is the subset of sumN that are organic nitrogen compounds
297 measured by the PTR-ToF, as listed in Table 1. This fire had a mixture of flaming and smoldering combustion
298 throughout the fire as indicated by MCE and nitrogen profiles (panel (d)). The comparison of N_r-NO with the sumN
299 in panel (b) shows that much of the N is accounted for. The major contributors to sumN for this fire were HNCO,
300 HCN, HONO, NO_2, and NH_3, while NVOC was a very small contributor to sumN (panel (b)). Note that while HNO_3
301 is measurable by FTIR with good sensitivity, no HNO_3 signals were observed above detection limit, which was a few
302 ppbv. Panel (c) shows the residual left after NO and sumN are subtracted from N_r, corresponding to an integrated

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amount of 15.6 ±8% of N_r. This residual is reasonable considering typical published particle N_r measurements (Akagi et al., 2012; Akagi et al., 2011; Liu et al., 2017; May et al., 2014), and consistent with there being some particle N_r from flaming, which are most likely organic nitrates or nitro-organics, and particle ammonium from smoldering with potassium or ammonium nitrate potentially accounting for substantial N_r.

In contrast to the above, the nitrogen emissions from Fire 050, yak dung, are shown in Figure 6. This fuel produced mostly smoldering emissions as exemplified by the low NO levels relative to N_r (panel a), and the low MCEs observed (panel d). The sum of N_r species was somewhat correlated with the quantity N_r-NO, but was substantially lower, and the residual N_r unaccounted for by the gas-phase measurements was 33.9 ±16% of N_r (panel c). The majority of sumN was represented by HCN and NH_3, with acetonitrile (CH_3CN) higher than any of the other inorganics, HNCO, NO_2 or HONO. The NVOCs were also a larger fraction of N_r-NO than in the case of Fire 047 shown above, a feature that implies that more semi-volatile organic compounds, SVOC, survive these types of fires and could make a proportionally higher contribution to the N_r budget in this fire relative to Fire 047. FireLab results of particle organic carbon measurements (Jen et al., 2019) and field measurements in environments with a lot of dung burning (Jayarathne et al., 2018; Stockwell et al., 2016a) are consistent with a higher EF for particle organic carbon and by extension particle NVOC compounds. The quantity (N_r-NO)/N_r was relatively high and had less dynamic range than for fires with more flaming combustion like Fire 047.

The fire-integrated measurements of inorganic and NVOC species are listed in the Supplemental section as ratios to N_r for each stack fire (Table S1). The summary of all the fire integrated X_i/N_r fractions (where X_i is the N_r species or quantity) is given in Table 3 for all the fires for which we have a complete set of measurements (43 fires). In general, NO was the major species followed by NH_3, and the other inorganic N_r species, NO_2, HNCO, HONO, and HCN had individual contributions of 4.3 to 9.4 %. NVOC species were less than 5% of N_r on average. The unaccounted-for N_r, defined as (N_r-NO-sumN)/N_r had a median value of 14.3% and a mean (±std. dev.) of 15 (±10)%. Overall, 85% of N_r was accounted for by the gas phase measurements. The distribution of whole fire N_r residuals is plotted as a histogram in Figure 7. We expect the residual N_r was composed of either semi- or low-volatility compounds, or particle-bound N_r compounds, which we know are converted efficiently by the N_r catalyst (Stockwell et al., 2018) but not detected by the instruments included in this analysis. Along these lines, there is some indication that the residual has a systematic variation with whole fire MCE, with higher residuals (up to 30%) observed at lower MCEs and higher (N_r – NO)/N_r (see Figure S1 a&b), which would be consistent with higher EF for SVOC at low MCE (Jen et al., 2019) and particle N_r having a higher contribution from NO_3^- (May et al., 2014), and perhaps particle ammonium or reduced-N_r compounds. In general, there is more particulate organic material emitted from fires at low MCE (Jen et al., 2019), so we would expect more particle N at low MCE to go along with that.

3.1 Systematic dependences of N_r composition on combustion processes.

The features noted in fires shown above, as well as the anti-correlation of MCE and (N_r-NO)/N_r lead to the question of whether there are systematic dependences in N_r-compound composition on fire stage that can be used to formally classify and/or potentially predict the relative emissions of N_r compounds. MCE has been used as a rough indicator of the relative amounts of flaming and smoldering combustion in a fire, with high MCE (~99%) being “pure”

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flaming, low MCE (~80%) being "pure smoldering," and an MCE of ~0.9 being roughly equal amounts of both (Sect. 2.1.1 in Akagi et al., 2011). It should be understood that "smoldering" in this framework is a lumped term that includes all non-flame processes such as pyrolysis, glowing, and distillation, while flames cannot exist without these processes producing gaseous fuel to support them (Yokelson et al., 1996). In addition, "pure flaming" is essentially the efficient oxidation of smoldering products before they enter the atmosphere. However, for MCE to predict flaming and smoldering N_r species well, the variable fuel N must be considered. For instance, NO_x is clearly produced by flaming based on its temporal profile, but fire-integrated EF_{NOx} may not correlate with MCE due to variable fuel N. In these cases, EF_{NOx}/fuel N or $\Delta NH_3/\Delta NOx$ may still correlate (or anti-correlate) well with MCE (e.g. Fig. 4 in Burling et al., 2010 or Yokelson et al., 1996). Finally, the flame chemistry involving NH_3, HNCO, and HCN both produces and destroys NO in a fashion that does not conserve N_r. This chemistry is explored in Figure 8 in which NOx, NH_3, HNCO, HCN, HONO, and CH_3CN ratios to N_r are plotted vs real-time MCE for Fire 047 as a typical example for fires that have a substantial range of MCEs (e.g. from 0.8 to above 0.98). The relationship between NH_3/N_r and MCE confirms that NH_3 is primarily a smoldering emission and NO_x/N_r increases with increasing MCE in a non-linear fashion that confirms it is primarily a flaming compound. Such a non-linear dependence has also been seen for other flaming-related quantities such as Elemental Carbon/TC or EF_{HCl} (Christian et al 2003; Stockwell et al., 2014). Most importantly, the variations of $HNCO/N_r$, HCN/N_r, $HONO/N_r$, and CH_3CN/N_r versus MCE don't arise dominantly from either regime as these are species that are likely produced by multiple pathways (e.g. "incomplete flaming", pyrolysis, possibly glowing). By "incomplete" flame chemistry we mean the production of incompletely oxidized products in flames such as the complex system of reactions shown in Fig. 1. These reactions involving HNCO, HCN and NH_3 both produce and destroy NO, while HONO is produced from reactions of NO and NO_2 that are faster at slightly lower temperatures, for example the three-body association reaction of NO with OH radical (Manion et al., 2015). Variable turbulence in the turbulent diffusion flames that are characteristic of open BB likely contributes to varying temperatures, and therefore, varying amounts of incomplete oxidation of the fuel N (Shaddix et al., 1994).

The complexity of the dependence of N_r speciation on combustion chemistry suggests that MCE is an insufficient model to use for applying lab results to real-world fire emissions (Stockwell et al., 2016a; Yokelson et al., 2013b). Accordingly, we employed the positive matrix factorization (PMF) method (see Methodology section) that has been used by a number of groups to probe the sources contributing to complex mixtures (see for example Ulbrich et al., 2009 Sekimoto et al., 2018). Our PMF results showed several general features, irrespective of the inclusion or exclusion of CO_2, CO and N_r. The emissions were best fit by three factors (with approximate descriptive names justified below and prime species): (1) a combustion (flaming) factor (abbreviated Comb-N), (2) a high temperature pyrolysis factor (HT-N), and (3) a low temperature pyrolysis factor, (LT-N). We use these terms in part to harmonize our discussion with the VOC results discussed by Sekimoto 2018. An example timeseries for the PMF analysis of a coniferous fuel with just the N_r species included is shown in Figure 9 for a realistic mix of lodgepole pine (Fire 063), and Figure S2 shows the consolidated time series of all coniferous fuels fit using just the N_r species. The three factors successfully describe the majority of the N_r-emissions where the difference between the measured and calculated mass is on average 5.1% for coniferous fuels and 4.6% for chaparrals as indicated in Table 4.

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The 'loadings' of the three different factors, i.e. the contribution of compounds to each factor, for coniferous fuels are shown in Figure 10(a), and the distribution of a given compound or compound class amongst the three factors is shown in Figure 10(b) as normalized fraction. Normalized fraction is equal to the PMF-determined contribution of a compound to a factor, divided by the sum of the contribution of the compound to all three factors. The Comb-N factor contained NO, NO_2, and HONO, the HT-N factor had mostly HCN, HNCO, nitriles, with contributions from NO_2 and nitro compounds, and the LT-N factor contained NH_3, amines, amides and heterocyclics. Within the Comb-N factor there is some evidence that the relative amounts of HONO and NOx depend on fuel moisture. For example, the ratio HONO/NOx for whole fires shows some correlation with needle moisture in coniferous fires that were canopy fuels (Foliage and small woody biomass), as shown in Figure S3. This may be due to flame process the interconvert NOx and HONO in the presence of water vapor of OH (see Figure 1).

Literature values from studies where flame temperature was measured are typically in the range of 1100 – 1200 °C (Taylor et al., 2004; Wotton et al., 2012), so we would assume that would constitute the upper range of our Comb-N factor. The radical chemistry involving HCN, HNCO and NH_3 starts to shut down below about 800-900°, according to the modeling of Glarborg et al., (2018), so we set 800°C as a lower limit for the Comb-N factor. The HT-N factor species are known to be produced by the intense pyrolysis of fuel N_r compounds (Hansson et al., 2004; Liu et al., 2018; Ren et al., 2010), which for these compounds becomes important at temperatures around 500-600°C. Accordingly, we estimate the temperature range for the HT-N factor at 500 – 800°C. The remaining LT-N factor results from mild pyrolysis and pertains to fire conditions of roughly 500°C and below, and was dominated by NH_3, amines, amides and some of the more complex organics (Koss et al., 2018). The names and temperature ranges are approximate and likely include processes that occur inside flames as part of the flame proper, as turbulent diffusive flames are highly variable in space and time.

It is useful to explore the correlation of N-PMF factors with other fire indicators to determine relationships for parameterizing N_r emissions together with carbon and VOC emissions. The Comb-N factor for coniferous fuels, which consisted of NO_x and HONO, would be expected to correlate with CO_2 but not as well with MCE since the latter includes an indicator of incomplete combustion. The timeseries of Comb-N along with CO_2 and with MCE for Fire 037 (ponderosa pine), are plotted in Figure 11. As expected they show an excellent correlation of Comb-N with CO_2 (R^2=0.942) since all the species are flaming compounds, but non-linear correlation of Comb-N with MCE (R^2=0.363) since the latter factors in a smoldering compound (CO), similar to the NOx/N_r vs. MCE plot for Fire 047 in Figure 8. The excellent correlation of Comb-N with CO_2 is a broadly applicable result, the R^2 parameters for all the fires shown in Figure S2 had an average of 0.898, and ranged from 0.806 to 0.966. As a consequence, we can conclude that CO_2 would be the best tracer for Comb-N in many western U.S. ecosystems where conifers predominate, provided ambient CO_2 backgrounds can be properly accounted for as described by Yokelson et al., (2013a).

Our Comb-N factor did not correspond to the high temperature VOC factor (HT-VOC) found by Sekimoto et al., (2018), because combustion produces NOx, and HONO, but almost none of the compounds classified as VOCs survive these conditions. However, the HT-N and HT-VOC factors were well correlated for many fires. An example of this is shown in Figure 12 for Fire 037, a sample that was broadly representative of ponderosa pine (i.e. canopy and litter). This result can be rationalized by the fact that while HT-VOC factors have large contributions from many more

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compounds that the N compounds measured here, they also have large contributions (>85%) from HCN, HNCO, and HONO, (in other words >85% of HCN, HNCO and HONO are found in the HT-VOC factor). Since the HT-N factors are also heavily weighted by HCN and HNCO, it is reassuring that both of these PMF analyses have independently identified these species as important contributors to the HT fire regime. The R^2 correlation coefficients of HT factors for the coniferous shown in Figure S2 averaged 0.866 and ranged from 0.419 to 0.959. As a consequence of this correlation, we can conclude that HCN is the best marker for the HT-N and HT-VOC factors in most western U.S. wildfires, since HCN is essentially inert on the timescales of fire plumes.

The correlations of LT-N and LT-VOC factors were not particularly high for most of the coniferous fires shown in Figure S2. The average R^2 was 0.427 with a range of between 0.072 and 0.827. The reasons for this lack of correlation are not clear, as NH_3, amines and amides appear predominantly in both LT factors, and the absolute concentrations of NH_3 are usually quite high in these fires relative to VOCs (Sekimoto et al., 2018). However, the LT-VOC factor includes many more compounds with a variety of functional groups not found in the LT-N factor, so it appears that the VOC and N compounds have sufficiently different pyrolysis chemistry that the LT factors do not show much correlation. We conclude that NH_3 (and particle NH_4^+) will be the best marker for the LT-N factor in western U.S. coniferous wildfires, but the LT-VOC chemistry might not be captured reliably by this marker.

The emissions from burning chaparral fuels (manzanita and chamise) collected at two sites in California were also analyzed as a group and yielded three separate factors in a fashion similar to the coniferous fuels (see Figure S4 for the PMF timeline). The chaparral factors had slightly different composition (Figure S5), the combustion factor was mostly NO, with small amounts of HNCO, HONO and NH_3, the high temperature factor was dominated by NO_2 and included HONO, HCN, and HNCO, and the low temperature factor was mostly NH_3 with a slight amount of NO contributing. The NVOC species were found in both the medium and low temperature factors.

There was less similarity between the Comb-N factor and CO_2 emissions for chaparral fuels compared to those found for coniferous fuels, with an average correlation coefficient (R^2) of 0.689, with a range from 0.244 and 0.950. As a result, there may not be a simple conserved tracer for the combustion factor of these fuel types, however total odd nitrogen (NOy) which is NOx and all the compounds that are produced from NOx in the troposphere, may be useful as it is a reasonably conserved tracer in the absence of wet or dry deposition of particles. Correlation coefficients between the HT-N and HT-VOC factors were on average $R^2 = 0.551$, with a range 0.047-0.911. The correlations between LT-N and LT-VOC factors were in the same range for chaparral fuels as for coniferous, average $R^2 = 0.447$, range 0.028-0.827.

There were some fuels that do not sustain flaming combustion well, specifically duff, Yak dung and Indonesian peat. These fires exhibited little or no NO emission commensurate with minimal flaming combustion. Instead the emissions were mostly the pyrolysis products NH_3, (0.22 – 0.53 N_r fraction), and HCN (up to 0.32 N_r fraction for peat). It was also apparent that these fires also had unaccounted for N_r, close to, or just over 0.30 (Table S1). The distribution of N_r compounds in the one peat fire that we measured (Fire 055) is in line with those reported for fires measured in situ which showed relatively high EFs for HCN and NH_3 (Stockwell et al., 2016b; Stockwell et al., 2015).

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449 The application of our N_r emissions results to real-world fires will depend somewhat on the nature of the
450 information available on a particular fire, or fire complex. As a good starting point, or in the absence of detailed N
451 and C analyses of fuels, a N_r/C ratio of 0.37% appears to capture most of the fires studied in this work. The Nr can be
452 apportioned according to the results summarized in Table 3. Adjustments to those fractions can be made either by
453 scaling slightly by average MCE, with the knowledge that intermediate species (HT-N pyrolysis species) such as HCN
454 and HNCO do not scale in the simple manner that NH_3 and NO_x do. If measurements of marker compounds are
455 available then CO_2, HCN, and the sum $NH_3 + NH_4^+$ can be used for the combustion, high-temperature pyrolysis, and
456 low-temperature pyrolysis factors respectively.
457
458 **4 Conclusions**
459 Seventy-five stack fire experiments were conducted during the FIREX FireLab experiments in Fall, 2016. A
460 range of fuels characteristic of the western U.S. was burned under conditions and in mixtures meant to represent
461 authentic wildfire conditions, as closely as is possible in the laboratory. Total reactive nitrogen (N_r = all N-containing
462 compounds except N_2 and N_2O) was measured along with a suite of N-containing compounds in order to obtain a
463 budget for N_r-emissions and to examine relationships between fuels, combustion conditions, and emissions chemistry.
464 Natural convection wildfires do not burn hot enough to produce NO_x from N_2 and O_2, so all N_r emissions
465 come from the fuel N. Almost all of the fires representative of North American ecosystems had emissions that clustered
466 around a N_r/C ratio of 0.37%, which can serve as a starting point for scaling emissions from these ecosystems
467 Comparison total N_r and total carbon emissions with the N/C ratios of both the original fuel and remaining ash allowed
468 us to estimate that an average of 68% (±14%) of the fuel nitrogen ends up as N_2 and N_2O. This loss of nitrogen can
469 be used to estimate how much fuel nitrogen ends up as N_r, which is a crucial aspect of fire plume chemistry since the
470 photo chemistry many fire plumes is NOx-limited, and NH_3 is an important contributor to particle chemistry. Of the
471 remaining N emitted as N_r, approximately 85% (±10%) was accounted for by individually measured gas-phase species,
472 while the rest was most likely particle-bound NH_4^+ and NO_3^-, with a smaller contribution from low-volatility species
473 or other species such as cyanogen (Lobert and Warnatz, 1993), that were not quantified by the instruments for
474 individual measurements we used in this study.
475 The individual N_r species composition normalized to Total N_r, to account for fuel N variability, correlated
476 with flaming versus vs. smoldering combustion as indicated by modified combustion efficiency (MCE) for some
477 species (e.g. NH_3, NO_x). Other species, such as HCN and HNCO, peaked at intermediate MCE values. However,
478 positive matrix factorization (PMF) showed that all the measured N_r emissions from the main two categories of fuels,
479 conifers and chaparral, grouped into three mixtures (factors), roughly attributed to temperature: combustion (NOx,
480 HONO), high temperature (HNCO, HCN, nitriles), and low temperature (NH_3, amines, amides). Chemical kinetic and
481 pyrolysis considerations set the temperature ranges for these regimes at approximately 800-1200°C, 500-800°C and
482 <500°C respectively.
483 This paper connects mechanistic aspects of N combustion chemistry to the budget of N_r emissions from
484 biomass burning. The emission composition measurements detailed here give useful information concerning what the
485 initial conditions will be in actual fire plumes. These results suggest that for coniferous fuels characteristic of the

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western U.S. CO_2 is the best marker for flaming combustion, HCN is the best marker for high temperature pyrolysis processes, and NH_3/NH_4^+ is the best marker for low temperature pyrolysis processes. The HT-N and HT-VOC pyrolysis factors showed high degree of correlation especially for coniferous fuels, which can simplify how these different classes of emissions can be estimated. Results of future field intensives can be combined with this emissions information to refine these recommendations on how to put N_r-chemistry into the modeling frameworks needed to predict fire plume chemistry and impacts.

Data availability

The FIREX Firelab 2016 data are available at: https://esrl.noaa.gov/csd/groups/csd7/measurements/2016firex/FireLab/DataDownload/. The descriptions of the measurements can be found here: https://esrl.noaa.gov/csd/groups/csd7/measurements/2016firex/FireLab/dataidtable.html. The complete ash analyses are available on request.

Author Contributions

JMR, RY, CW and JdG designed the research. The measurements were conducted by JMR, CS, CW, RJY, JdG, YL, VS, ARK, KS, MMC, BY, KJZ, SSB, CS, and SHD. All authors contributed to the discussion and interpretation of the results and writing the paper.

Competing interests

Joost de Gouw worked as a consultant for Aerodyne Research during part of the preparation phase of this paper.

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Any mention of brand names or manufacturers is for information purposes only and does not constitute an endorsement.

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Table 1. Nitrogen compounds observed in the FIREX FireLab 2016 Study.

Compound/Class	Importance	Measurement Method	References
Total Reactive N	Total available for atmospheric reactions	Catalytic Conversion NO/O_3 chemiluminescence	Stockwell et al., 2018
Nitric Oxide	Major "flaming stage" product, oxidant production	NO/O_3 chemiluminescence OP-FTIR	Williams et al., 1998 Selimovic et al., 2018
Nitrogen Dioxide	Atmospheric oxidant production	OP-FTIR, ACES	Stockwell et al., 2014, Min et al., 2016, Zarzana, et al., 2018
Nitrous Acid	HO_x radical source	OP-FTIR, ACES	Stockwell et al., 2014, Min et al., 2016, Zarzana et al., 2018
Nitric Acid	Particle precursor	OP-FTIR	Yokelson et al 2009, McMeeking et al 2009
Hydrogen Cyanide	Flame chemistry, Atmospheric tracer, Toxicity	OP-FTIR, PTR-ToF-MS	Selimovic, et al., 2018 Koss, et al., 2018
Isocyanic Acid	Flame chemistry, Toxicity, Health effects	PTR-ToF-MS	Koss, et al., 2018
Ammonia	Major "smoldering stage" product, Main atmospheric base, Particle formation	OP-FTIR	Selimovic, et al., 2018
NVOCs: Amides[1] Amines[2] Heterocyclics[3] Nitriles[4] Nitro compds[5]	Brown carbon, Toxicity, Tracers	PTR-ToF-MS, GC/MS, I^- CIMS	Koss, et al., 2018 Gilman et al., 2015 Lerner et al., 2017 Lee et al., 2014

1). Ethylamine, methanimine, propeneamine, sulfinylmethanamine, trimethylamine, buteneamines
2). Formamide, acetamide, methylmaleimide
3). C_2-pyrroles, dihydropyridine, ethynylpyrrole, methylpyridine, methylpyrrole, pyridinealdehyde, 4-pyrindinol, vinylpyridine
4). Acetonitrile, acrylonitrile, benzonitrile, butanenitrile, butynenitrile, benzoacetonitrile, C_7acrylonitrile, C_8-nitriles, heptylnitrile, furancarbonitrile, methylbenzoacetonitrile, pentylnitriles, propanenitrile, propynenitrile, butenenitrile, methylisocyanate.
5). Butenenitrates, nitrobenzene, nitroethane, nitroethene, nitrofuran, nitromethane, nitropropanes, nitrotoluene.

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Table 2. Compounds and compound classes used in the PMF analyses and their corresponding errors.

Compound or Class	unit	Batch 1	Batch 2	Estimated error
CO_2	ppmv	X		20% + 2 ppmv
CO	ppmv	X		20% + 0.002 ppmv
N_r	ppbv	X		10% + 1 ppbv
NH_3	ppbv	X	X	5% + 2 ppbv
NO	ppbv	X	X	10% + 1 ppbv
NO_2	ppbv	X	X	10% + 0.2 ppbv
HONO	ppbv	X	X	20% + 1 ppbv
HCN	ppbv	X	X	15% + 0.2 ppbv
HNCO	ppbv	X	X	15% + 0.2 ppbv
Nitriles	ppbv	X	X	20% + 0.2 ppbv
Amines	ppbv	X	X	20% + 0.2 ppbv
Amides	ppbv	X	X	20% + 0.2 ppbv
Nitro-compounds	ppbv	X	X	20% + 0.2 ppbv
Heterocyclics	ppbv	X	X	20% + 0.2 ppbv

Table 3. Summary of X_i/N_r Measurements for all Stack Burns[1]

Quantity	**Average ±(std dev) %**
NO/N_r	34.5 (16.6)
NO_2/N_r	9.4 (6.2)
$HNCO/N_r$	6.0 (2.9)
$HONO/N_r$	4.5 (2.2)
HCN/N_r	4.3 (2.3)
NH_3/N_r	19.3 (6.7)
$NVOC/N_r$	4.3 (2.8)
$(N_r\text{-sumN})/N_r$	15.2 (9.8)

1). Not every measurement was available for every fire, consequently the values do not add up to exactly 100%.

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Table 4. Residuals of the PMF analyses by fuel, as percent of total signal

Fuel	Total Number	Component	Fire Number	Residual (%), avg (stdev)
Ponderosa Pine	9	Realistic (mix)	Fire 37,59,72	3.8 (± 1.4)
		Canopy (pure)	Fire 19[a],39	
		Litter (pure)	Fire 38	
Lodgepole Pine	5	Realistic	Fire 07[a],58,63	5.1 (±3.1)
		Canopy	Fire 40	
		Litter	Fire 41	
Douglas Fir	4	Realistic	Fire 14[a],57	6.8 (±3.1)
		Canopy	Fire 18	
		Litter	Fire 43[a]	
SubAlpine Fir	5	Realistic	Fire 47,67	6.6 (±2.3)
		Canopy	Fire 15,23	
		Litter	Fire 51[a]	
Engelmann Spruce	2	Realistic	Fire 08[a]	3.1 (±1.9)
		Canopy	Fire 25	
Chamise (San Dimas, CA)	2	Canopy	Fire 24,29	4.4 (±2.7)
Chamise (North Mountain, CA.)	2	Canopy	Fire 27,32	4.2 (±1.0)
Manzanita San Dimas, CA)	2	Canopy	Fire 30,33	4.8 (±2.1)
Manzanita (North Mountain, CA.))	2	Canopy	Fire 28	5.1

a-Excluded from Batch 2

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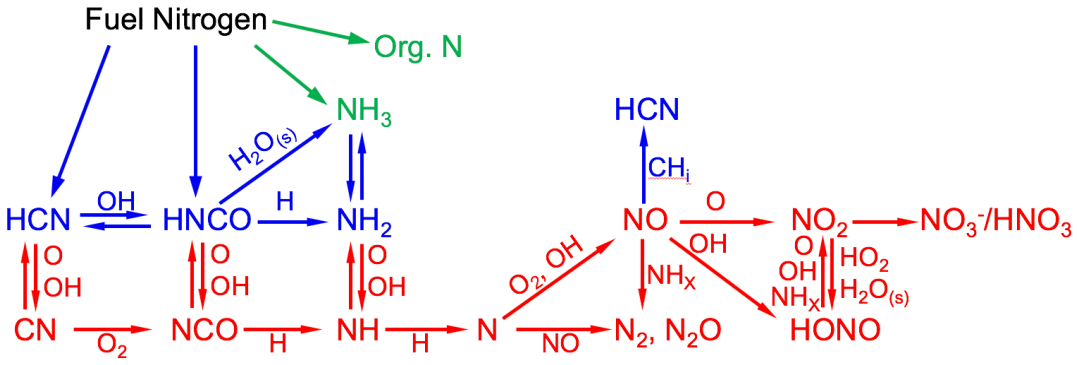

Figure 1. Schematic of the combustion chemistry of the small molecules that are emitted from BB and represent sources and sinks of reactive nitrogen (N_r), adapted from (Glarborg et al., 2018; Lobert and Warnatz, 1993; Lucassen et al., 2012). $H_2O_{(s)}$ denotes the combination of H_2O and a surface to facilitate the reaction. Red color indicates the highest temperature (combustion) processes, blue indicates intermediate temperature processes and green indicates the lowest temperature processes.

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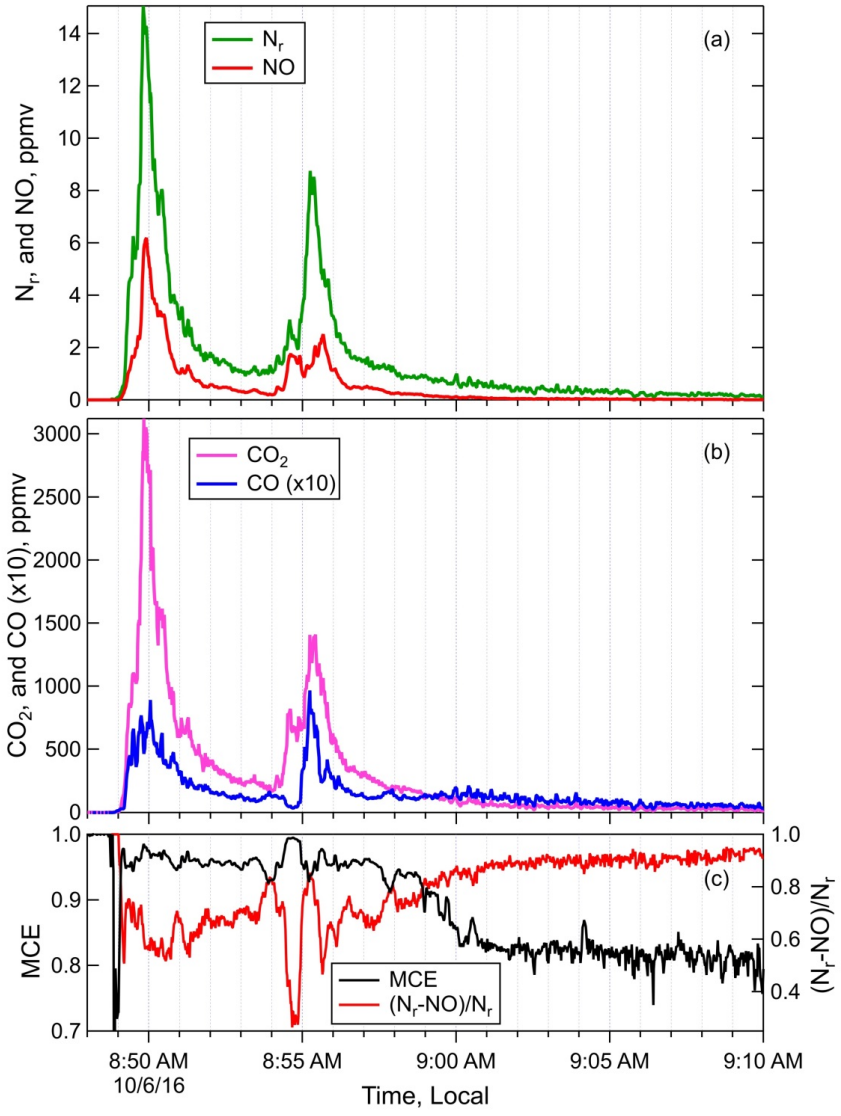

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832 **Figure 2. Timelines of the N_r, NO (panel a), ΔCO_2, ΔCO (panel b), MCE and $(N_r\text{-NO})/N_r$**
833 **(panel c) measured during Fire 004, a ponderosa pine realistic mix sample. Note that ΔCO is**
834 **plotted at x10 the measured abundance for clarity.**

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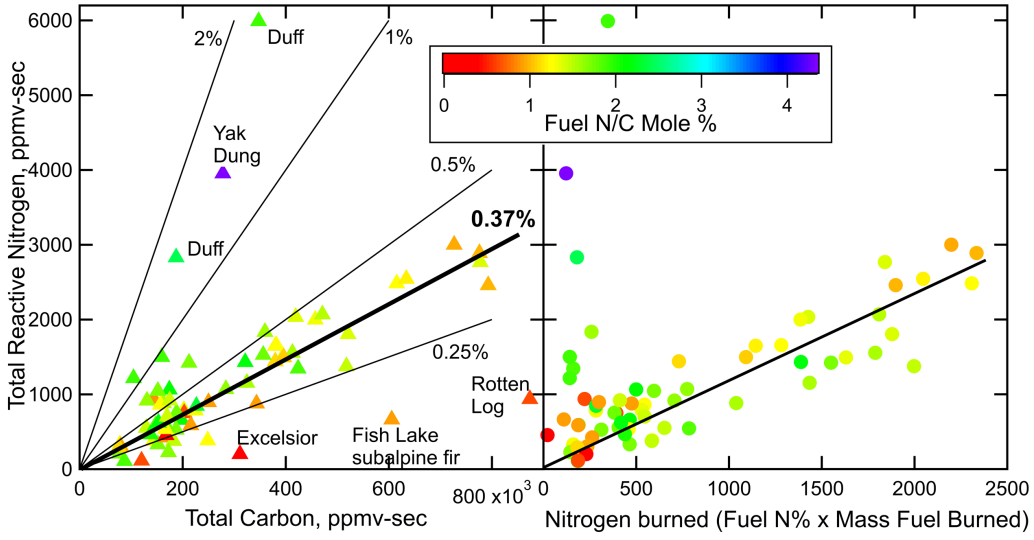

Figure 3. Integrated N_r versus integrated Total Carbon (panel a), and versus nitrogen burned based on fuel nitrogen content and mass of fuel burned (panel b). The points are colored by fuel nitrogen to carbon ratio.

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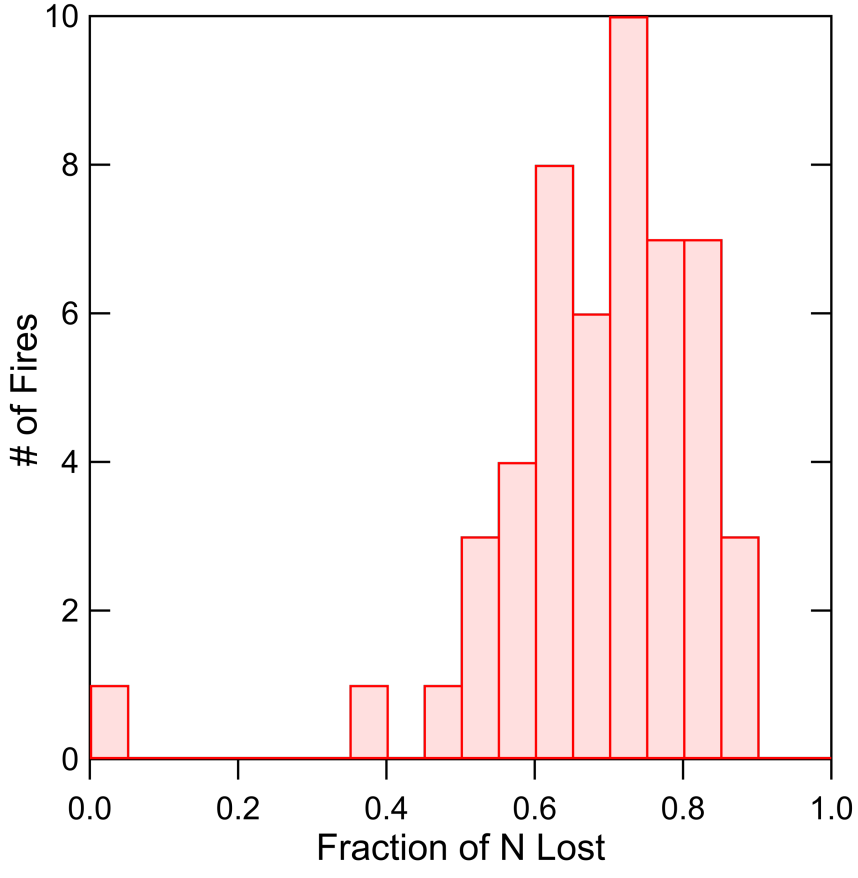

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849 **Figure 4. The histogram of the fraction of N loss to N_2 and N_2O estimated from the mass**
850 **balance analysis described in the Supplemental Materials (52 burns).**

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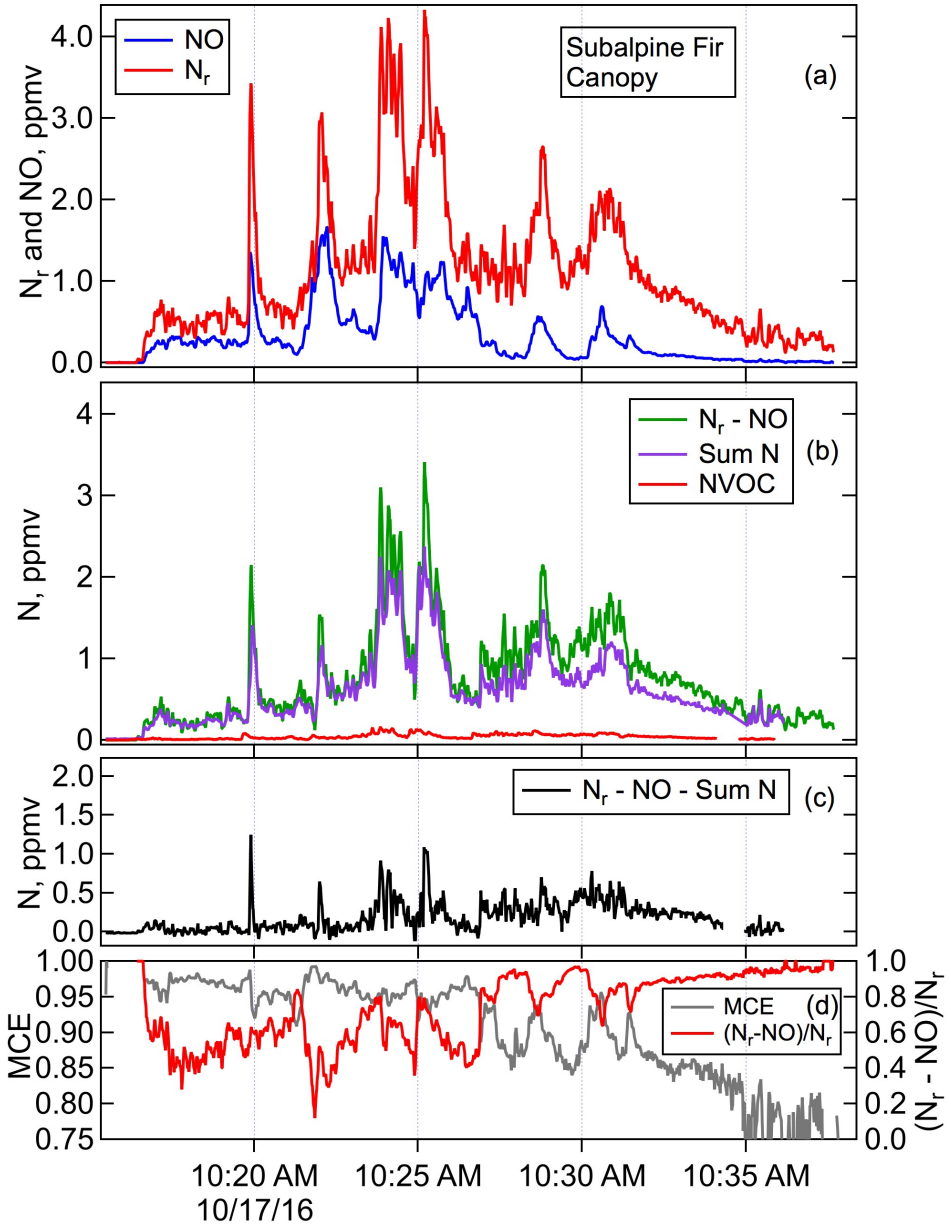

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853 **Figure 5. Timelines of N_r and NO (panel a), N_r – NO, the sum of all measures N species except**
854 **for NO (panel b), residual of N_r minus all measured N species (N_r – NO – Sum N, panel c),**
855 **and MCE and (N_r – NO)/N_r (panel d) for Fire047, subalpine fir realistic mix.**
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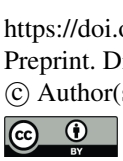

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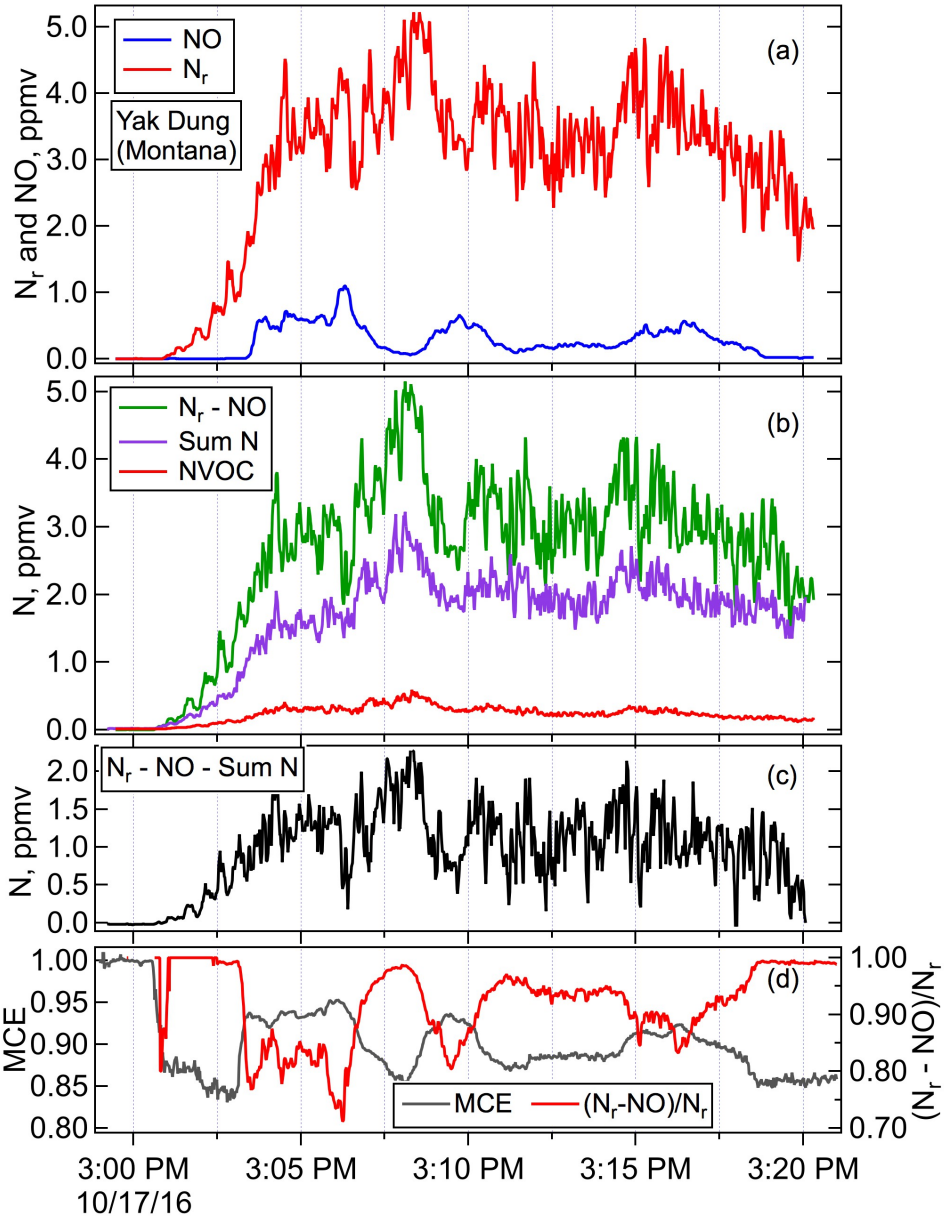

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859
860 **Figure 6. Timelines of N_r and NO (panel a), N_r-NO, the sum of all measured N_r species except**
861 **for NO (panel b), and residual of N_r minus all measured N species (N_r-NO-Sum N, panel c)**
862 **and MCE and (N_r-NO)/N_r (panel d) for Fire050, Montana yak dung.**
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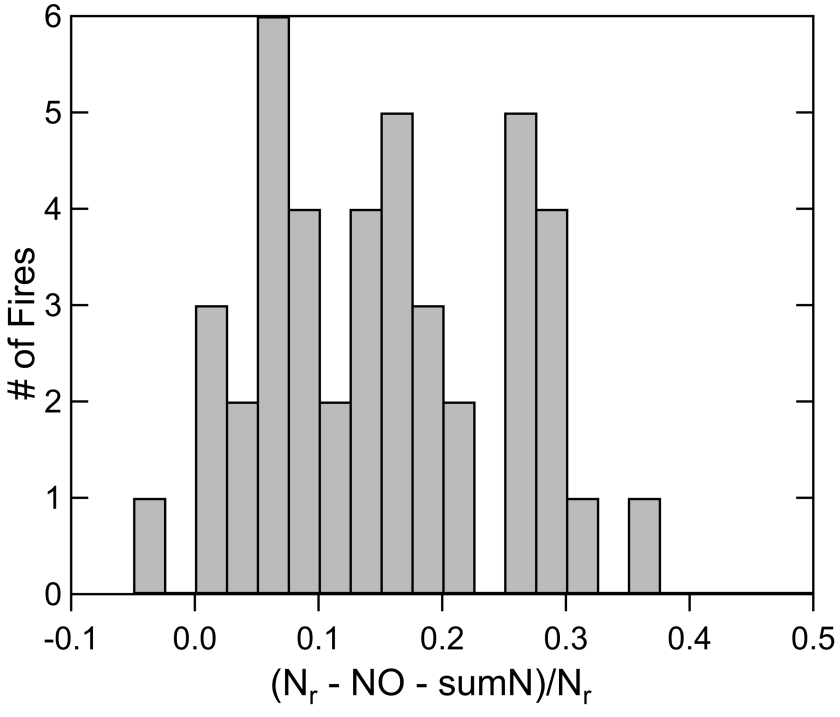

Figure 7. A histogram of the residual N for all the stack fires during the 2016 FireLab study for which there are FTIR, ACES and PTR-ToF measurements (n=43). The median is 0.143, and the mean (±std dev) was 0.15 (±0.10).

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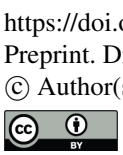

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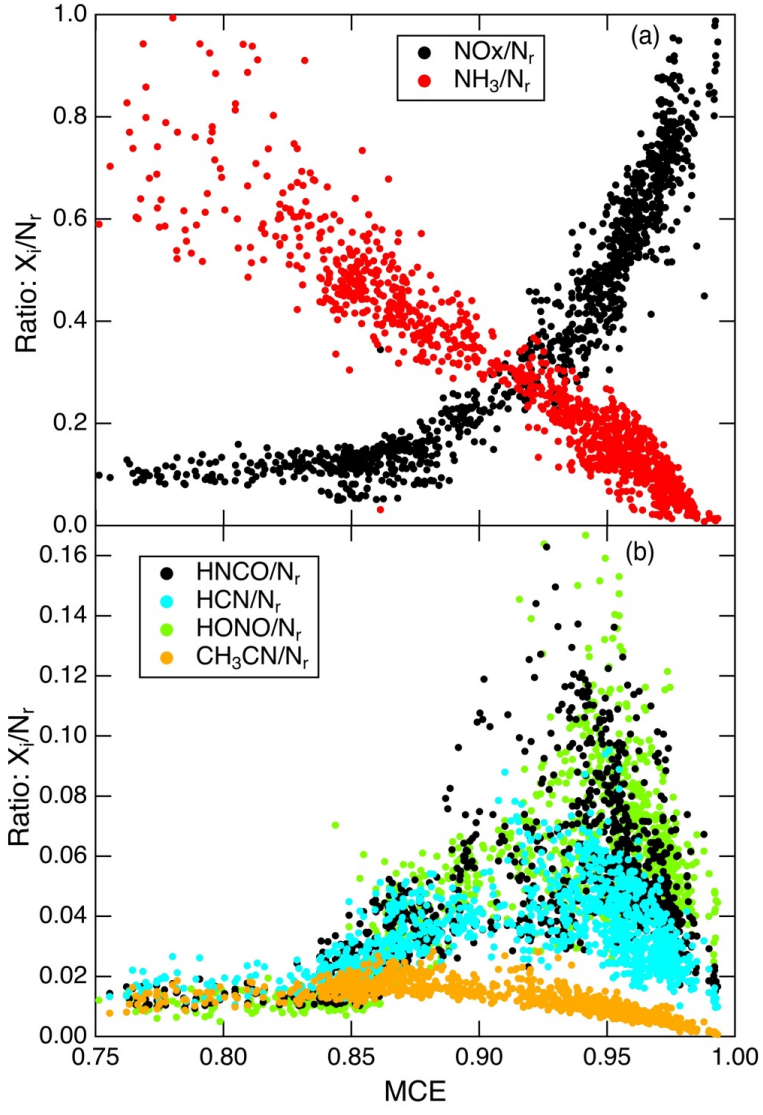

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878 **Figure 8. The relationships between NO_x/N_r and NH_3/N_r vs MCE (panel a), and the**
879 **$HNCO/N_r$, HCN/N_r, $HONO/N_r$, and CH_3CN/N_r vs MCE (panel b) for Fire 047.**
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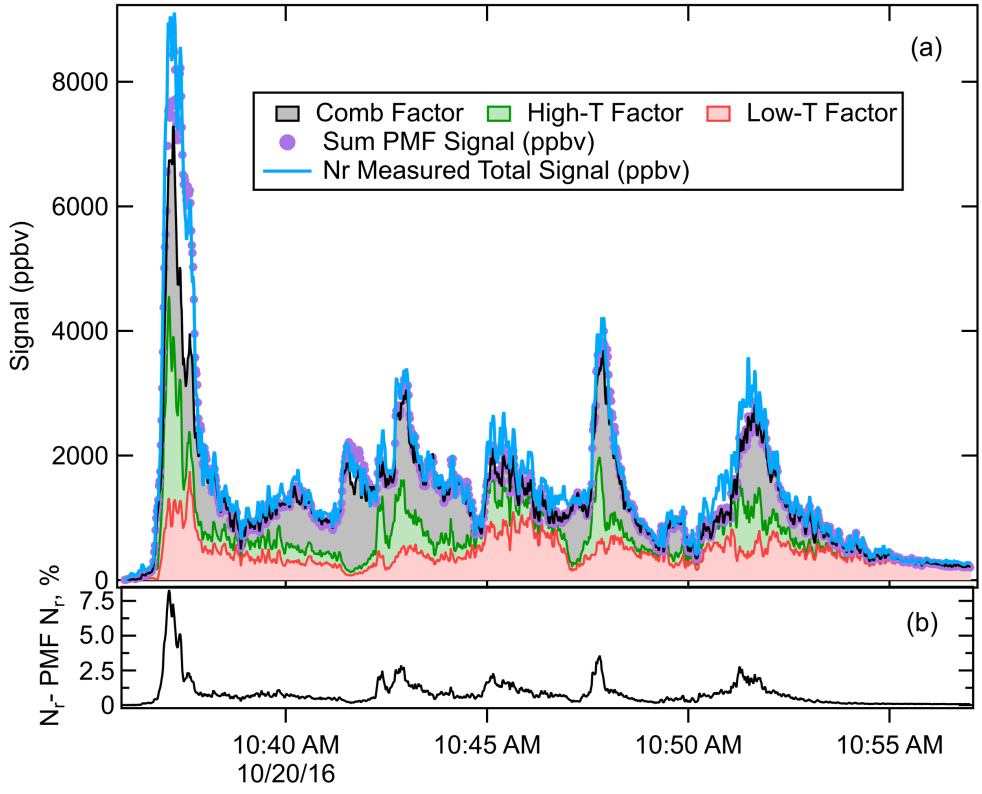

883
884 **Figure 9. Pane (a), the measured N_r signal for Fire 063 (lodgepole pine) (blue line), the sum**
885 **of the signal reconstructed by the PMF (purple points) and the three PMF factors:**
886 **combustion (grey), high temperature (green) and low temperature (red), plotted in a stacked**
887 **fashion (i.e. added on top of one another). Panel (b) the "residual" of the PMF fit consisting**
888 **of the measured N_r signal minus the N_r signal reconstructed by the PMF, as a percentage of**
889 **the N_r signal.**

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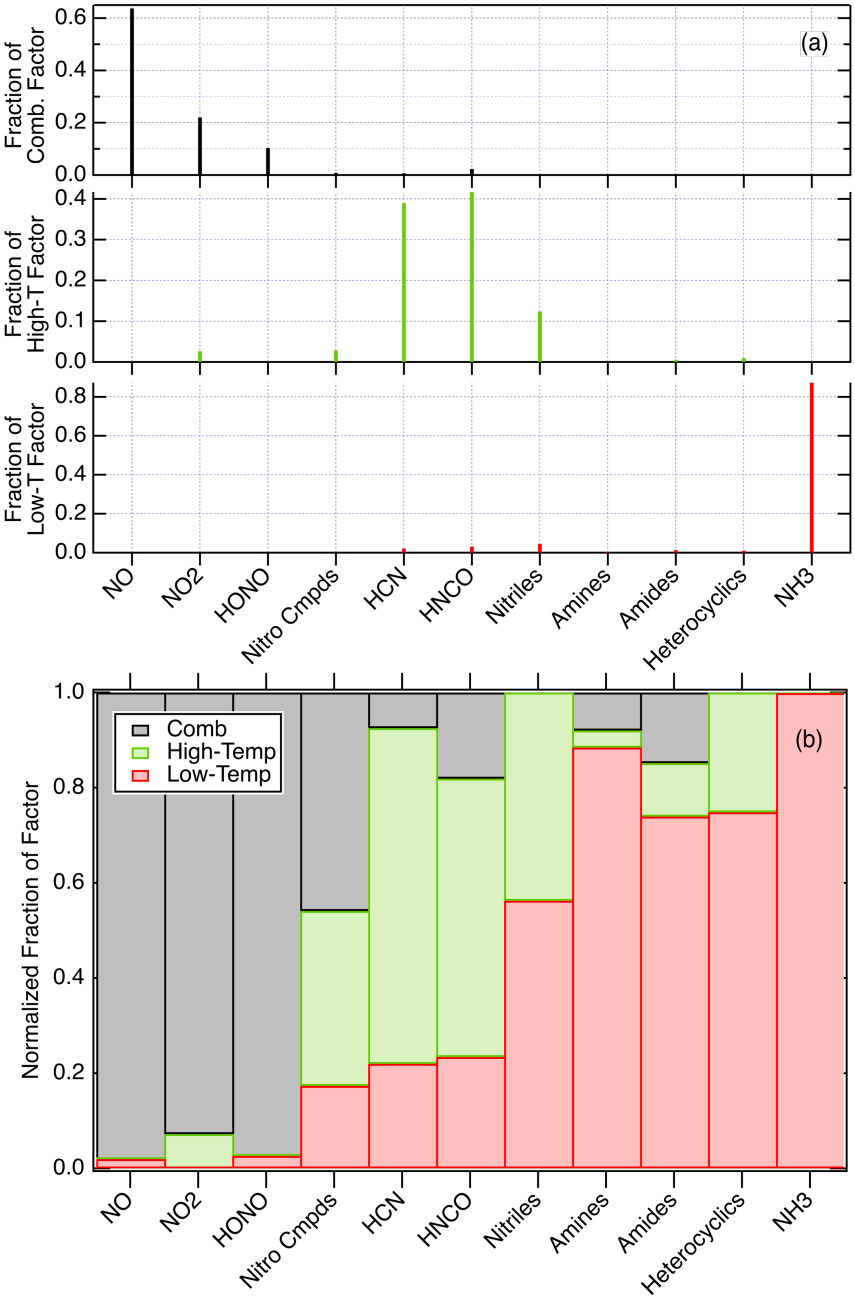

Figure 10. The contributions of nitrogen species to the factors that simulate the emissions from coniferous fuels shown in Figure S2 (panel a), and the fraction of each compound or class found in each factor (panel b).

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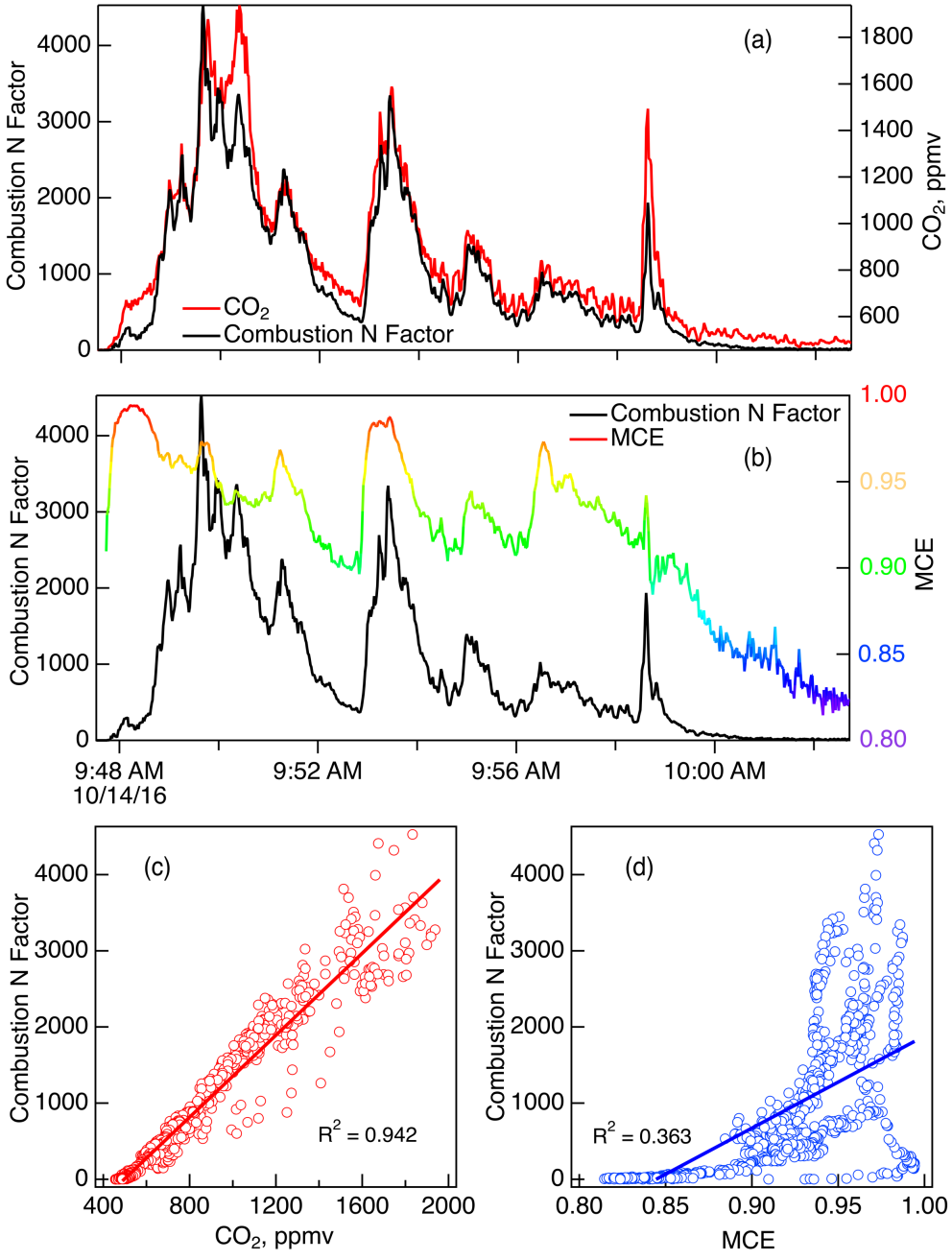

Figure 11. Comparisons of the N-PMF combustion factor (Comb-N) with CO_2 (Panel a) and MCE (Panel b) for Fire 037 (ponderosa pine). Panel (c) shows the scatter plot of the Comb-N factor versus CO_2 and panel (d) shows the scatter plot of Comb-N factor versus MCE.

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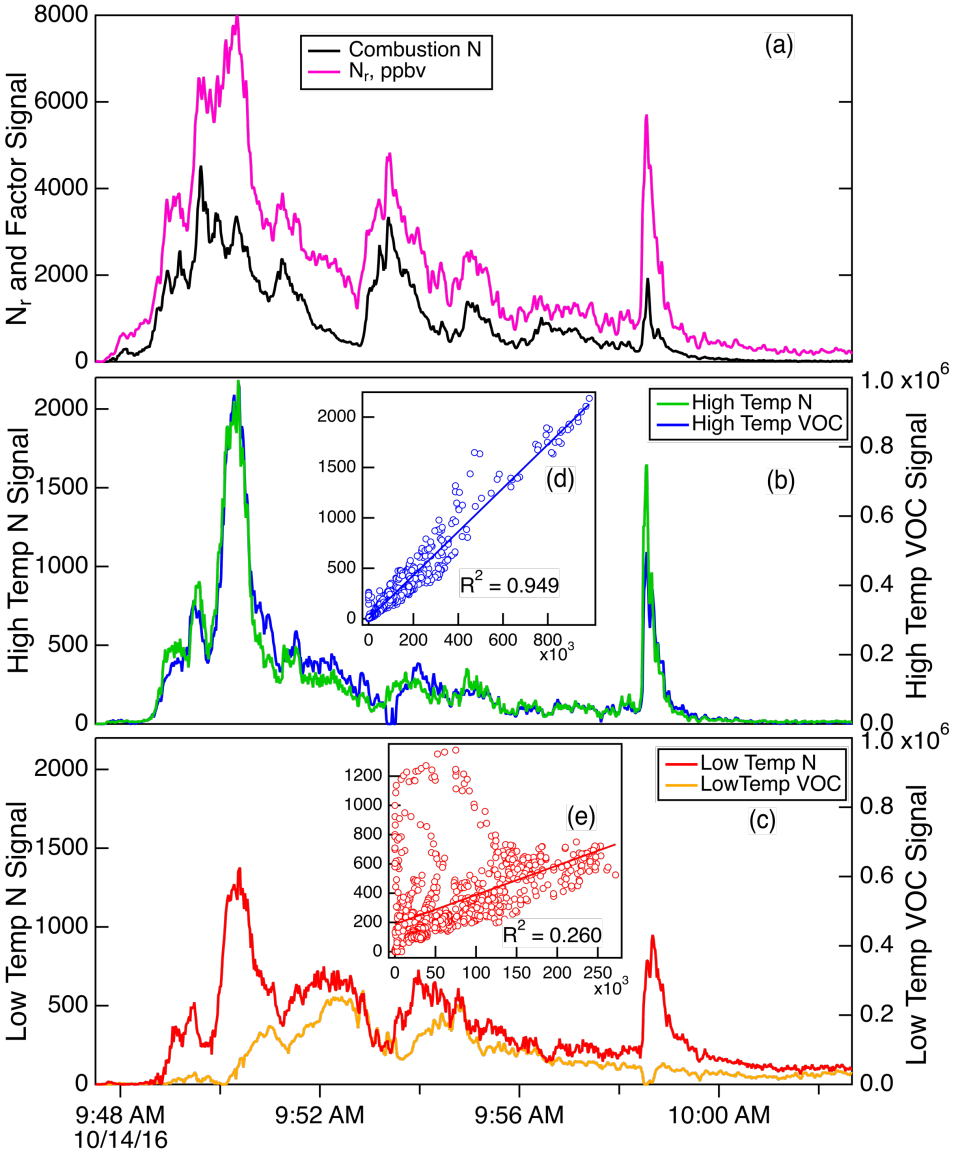

Figure 12. Details of the PMF factors for Fire 037 (ponderosa pine). Panel (a) shows the total N_r signal (magenta) and the Comb-N factor (black), panel (b) shows the HT-N factor (green) and HT- VOC factor (blue), and panel (c) shows the LT-N factor (red) and LT-VOC factor (orange). The insets (panel d) show the correlation of the two HT factors and the correlation between the two LT factors.