Author response to referee comments

Comments to the Author:

I thank the authors for considering my comments and I feel now that the paper is more in-scope regarding the discussion of potential new insights into atmospheric chemistry. I'm more happy for this to be published in ACP, however I do have some comments that I would like to see addressed before publication.

Response: We thank the reviewer for the positive and constructive reviews of our revision. Below, we go through them point by point, with the original remarks in black and our replies in blue. The changes in the revised manuscript are in italic.

General comments:

While the authors discuss in more detail the reasoning behind their method to exclude the larger peaks, I do not think they pay it enough attention to it in the revised manuscript. It is not sufficient to simply demonstrate that a factorisation looks unsatisfactory; the authors should discuss the reasons why it behaves the way it does, and better justify their chosen remedy in this context.

Response: We agree with the reviewer that more discussions should be added in the revised manuscript rather than only in the response. In the revised manuscript, we explained why the mass profiles of several resolved factors look quite similar with inclusion of the large monoterpene-related peaks, which leads to unsatisfactory results. Also we justified that our simple approach (removing the main peaks of the largest signals) worked very well in this study and produced factors that made sense both chemically and through their time variations.

The discussions have been added in the revised manuscript (Lines 155-161):

"Because PMF assumes that the data matrix can be explained by a linear combination of different factors, even a very tiny fraction of these high peaks was split into a factor, they may dominate the mass profile of the factor. As shown in Fig. S1, with the inclusion of $C_6H_8H^+$ and $C_{10}H_{16}H^+$, the mass profiles of several factors were quite similar and dominated by these peaks. Therefore, the major mass bins of these ions were excluded for further PMF analysis, but their corresponding isotopes were retained, effectively downweighting their contributions to the PMF result. This simple approach by removing the main peaks of the largest signals produced factors that made sense both chemically and through their temporal behaviors, lending confidence in the results."

I thank the authors for taking on board my comment exploring rotational ambiguity and the use of FPEAK to explore this is certainly informative in some ways. However, I could question the following concerning the approach:

1. If little difference is found across the range of FPEAK values, then this can simply mean that FPEAK has not been varied across a sufficient range, noting that the effect of varying this parameter will vary dataset to dataset.

Response: We agree with the reviewer that the effect of varying FPEAK values varies dataset to dataset and different ranges of FPEAK have been explored in previous studies (Ulbrich et al., 2009; Zhang et al., 2011). In this study, we follow solutions generally reported in the literature with an FPEAK value between -1 and +1 (Reff et al., 2007). For the low mass ranges of the Landes and SMEAR II dataset, the changes in the factor profiles and time series were very small by varying FPEAK values, indicating that varying FPEAK values from -1 to +1 did not affect the overall results of PMF analysis.

In the revised manuscript, we added in Lines 239-240 that "*For the low mass ranges of the Landes and SMEAR II dataset, the varying FPEAK values did not change the factor profiles and time series much, indicating that varying FPEAK values from -1 to +1 did not affect the overall results of PMF analysis.*"

Ulbrich, I. M., Canagaratna, M. R., Zhang, Q., Worsnop, D. R., and Jimenez, J. L.: Interpretation of organic components from Positive Matrix Factorization of aerosol mass spectrometric data, Atmos. Chem. Phys., 9, 2891-2918, 10.5194/acp-9-2891-2009, 2009.

Zhang, Q., Jimenez, J. L., Canagaratna, M. R., Ulbrich, I. M., Ng, N. L., Worsnop, D. R., Sun, Y. J. A., and Chemistry, B.: Understanding atmospheric organic aerosols via factor analysis of aerosol mass spectrometry: a review, 401, 3045-3067, 10.1007/s00216-011-5355-y, 2011.

Reff, A., Eberly, S. I., and Bhave, P. V.: Receptor Modeling of Ambient Particulate Matter Data Using Positive Matrix Factorization: Review of Existing Methods, J Air Waste Manage, 57, 146-154, 10.1080/10473289.2007.10465319, 2007.

2. The authors seem to basing their choice of FPEAK on the level of separation of mass spectral profiles. I would note that such an approach will tend to favour the positive values of FPEAK, however this does not necessarily make the factors more 'real' because one would expect a certain amount of overlap between the factors in mass spectral space. The authors should explain this better (the assumption that siloxane should only feature in one factor could be considered reasonable, but the reasoning is left largely implicit).

Response: Changing FPEAK changes both the mass profiles and time series of PMF factors simultaneously. In this study, we compared both factor profiles and factor time series to choose a more optimal FPEAK value. For the high mass range of the SMEAR II dataset, the factor time series were similar with FPEAK values from -1 to +1. However, for PMF solutions with positive FPEAK values, the factor profile of monoterpene more oxidized products including organic nitrates contained less traces of siloxanes and was more dominated by the corresponding fingerprint peaks. As discussed in section 4.3, these siloxanes can come from cosmetics and personal care products, and silicone oils used in instrument pumps. The temporal variations of these siloxanes showed high background signals and presented regular diurnal cycles following the variations in ambient temperature, which were significantly different from those of monoterpene more oxidized products. We believe that these siloxanes come from or represent a much different source from monoterpene more oxidized products. Therefore, we chose the PMF solution of FPEAK = +0.6 where siloxanes featured more in one single factor.

As the reviewer suggested, we add more explanations in the revised manuscript (Lines 246-249): *"As discussed in Sect. 4.3, these siloxanes can come from cosmetics and personal care products, and silicone oils used in instrument pumps. The temporal variations of these siloxanes differed significantly from those of monoterpene more oxidized products. After evaluation, we chose the solution with FPEAK = +0.6 for the high mass range of the SMEAR II dataset, where siloxanes feature more in one factor."*

3. There is an explicit recommendation in Paatero (2002) that nonzero FPEAK values should not normally be used for environmental data, so using this as the 'final' factorisation needs to be justified in this context. To be completely clear, my recommendation in the first review was to merely explore (and hopefully quantify) the rotational ambiguity, not necessarily to choose a more optimal FPEAK value.

Response: It has been demonstrated by Hopke (2000) that for real datasets, positive, non-zero FPEAK values generally yield more realistic PMF results. In this study, we explored the rotational ambiguity of PMF results using FPEAK values and found that factorization with positive FPEAK values produce more reasonable PMF solutions. The justification for choosing the PMF solution of FPEAK = +0.6 has been discussed above and added in the revised manuscript.

Hopke, P. K.: A guide to positive matrix factorization, Workshop on UNMIX and PMF as Applied to PM2.5, Research Triangle Park, NC, USA, 2000.

Specific comments:

Line 194: How are the authors defining 'significant'?

Response: After seven factors, the step change in Q/Q_{exp} is lower than 7%. We make the use of "significant" quantitatively in the revised manuscript (Line 198).

Line 195: A more explicit description of the criteria used to choose the number of factors is needed.

Response: To choose the optimum number of factors, we examined multiple criteria including the variations of Q/Q_{exp} vs. varying factor number, the distribution of the scaled residuals for each m/z, sum of the squares of scaled residuals, factor mass profile, factor time series and diurnal cycles, and also signs of split factors, following the procedures proposed by Ulbrich et al. (2009) and Zhang et al. (2011).

We include the description of these criteria in the revised manuscript (Lines 199-201): *"The optimal solution of seven factors was chosen by evaluating the variations of Q/Q_{exp} vs. varying factor number, the distribution of the scaled residuals for each m/z, sum of the squares of scaled residuals, factor mass profile, factor time series and diurnal cycles, and also signs of split factors (Ulbrich et al., 2009; Zhang et al., 2011)."*

Line 493: Regarding the two routes to nitrate formation, the caveat should be included that just because they are contained in the same factor, it does not mean that the same molecules are being produced. With the current wording, a reader could be mislead into thinking this to be the case.

Response: As suggested by the reviewer, the caveat has been included in the revised manuscript (Lines 503-505): *"It should be noted that $C_{10}H_{17}NO_5$ and $C_{10}H_{15}NO_6$ are used as examples because both of them are fingerprint peaks of the factor, but in real environments it may not be the case that these molecules are always produced from the above formation routes."*

Line 517: The end of this section is rather unsatisfactory, leaving the reader wondering whether the observation is linked to chemistry, time of day, insolation, temperature, etc. The authors make speculations regarding day vs night and sunny days, so can these be tested? As a bare minimum, the authors should (1) rule out temperature as a controlling factor because this is strongly anticorrelated with humidity and is known to influence reaction rates, branching ratios, etc., and (2) attempt to see if nocturnal nitrate chemistry could be responsible.

Response: The reason behind the high RH-dependence of sesquiterpene lightly oxidized products cannot be figured out based on the measurements in this study. But as the reviewer suggested, the controlling role of temperature should be ruled out because temperature is strongly anti-correlated with RH and is known to influence terpene emissions and terpene reaction rates. The strong RH-dependence was observed for sesquiterpene lightly oxidized products where sesquiterpene-derived organic nitrates are not included (not observed in this study). Therefore, we do not think nocturnal nitrate chemistry could be the reason.

In the revised manuscript, we add in Lines 528-530 that "*The controlling role of temperature can be ruled out because temperature is strongly anti-correlated with RH and is known to influence terpene emissions and terpene reaction rates.*"

Technical corrections:

Line 33: The comment "These findings highlight the need for further studies to delve into gas-phase atmospheric processes of monoterpenes and sesquiterpenes" verges on pointing out the obvious, as terpene chemistry has been a major topic of atmospheric chemistry for decades.

Response: This comment has been removed.

Line 61: Replace 'revealed' with 'showed'

Response: Done.

Line 64: Remove 'state-of-the-art'. Given that two authors have a financial interest in the instrument, it is important that the language describing it remains objective.

Response: Removed.

Line 468 (and elsewhere): Replace "oxidations" with "oxidation"

Response: Done.

Atmospheric organic vapors in two European pine forests measured by a Vocus PTR-TOF: insights into monoterpene and sesquiterpene oxidation processes

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Abstract.

Atmospheric organic vapors play essential roles in the formation of secondary organic aerosol. Source identification of these vapors is thus fundamental to understand their emission sources and chemical evolution in the atmosphere and their further impact on air quality and climate change. In this study, a Vocus proton-transfer-reaction time-of-flight mass spectrometer (PTR-TOF) was deployed in two forested environments, the Landes forest in southern France and the boreal forest in southern Finland, to measure atmospheric organic vapors, including both volatile organic compounds (VOCs) and their oxidation products. For the first time, we performed binned positive matrix factorization (binPMF) analysis on the complex mass spectra acquired with the Vocus PTR-TOF and identified various emission sources as well as oxidation processes in the atmosphere. Based on separate analysis of low- and high-mass ranges, fifteen PMF factors in the Landes forest and nine PMF factors in the Finnish boreal forest were resolved, showing a high similarity between the two sites. Particularly, terpenes and various terpene reaction products were separated into individual PMF factors with varying oxidation degrees, such as lightly oxidized compounds from both monoterpene and sesquiterpene oxidations, monoterpene-derived organic nitrates, and monoterpene more oxidized compounds. Factors representing monoterpenes dominated the biogenic VOCs in both forests, with less contributions from the isoprene factors and sesquiterpene factors. Factors of the lightly oxidized products, more oxidized products, and organic nitrates of monoterpenes/sesquiterpenes accounted for 8-12% of the measured gas-phase organic vapors in the two forests. Based on the interpretation of the results relating to oxidation processes, further insights were gained regarding monoterpene and sesquiterpene reactions. For example, a strong relative humidity (RH)-dependence was found for the behavior of sesquiterpene lightly oxidized compounds. High concentrations of these compounds only occur at high RH,

yet similar behavior was not observed for monoterpene oxidation products. ~~These findings highlight the need for further studies to delve into gas-phase atmospheric processes of monoterpenes and sesquiterpenes.~~

1 Introduction

Volatile organic compounds (VOCs) and their oxidation products are important contributors to atmospheric secondary organic aerosol (SOA) (Hallquist et al., 2009; Ehn et al., 2014) and new particle formation (Bianchi et al., 2016; Kirkby et al., 2016). Therefore, the identification of these organic vapors and their sources is fundamental for understanding the effects of atmospheric aerosols on climate change and air quality (Schell et al., 2001; Maria et al., 2004). Large amounts of VOCs with varying physicochemical properties are emitted from both biogenic and anthropogenic sources (Friedrich et al., 1999; Kesselmeier et al., 1999), and their oxidation processes in the atmosphere can lead to the formation of thousands of structurally distinct products containing multiple functional groups (Hallquist et al., 2009; Wennberg et al., 2018). Due to the enormous challenge in characterizing these organic vapors at molecular level, knowledge of their sources or formation pathways has remained lacking.

Globally, SOA production from biogenic sources is much larger than that from anthropogenic sources (Tsigaridis and Kanakidou, 2003). As a group of highly reactive gases, typically with one or more C = C double bounds, terpenes make up a major fraction of biogenic VOCs, including isoprene, monoterpenes, and sesquiterpenes (Guenther et al., 1995). In the atmosphere, they react with various oxidants, i.e., hydroxyl radical (OH), ozone (O_3), and nitrate radical (NO_3), and produce a large variety of oxygenated molecules. Isoprene is the most emitted biogenic VOC on the global scale but has a relatively small SOA yield (Ahlberg et al., 2017; McFiggans et al., 2019). Monoterpenes are important sources of SOA (Ehn et al., 2014; Zhang et al., 2018) and their oxidation processes have been found to play important roles in new particle formation (Kirkby et al., 2016; Simon et al., 2020). High ambient concentrations of monoterpenes have been observed in numerous pine forests (Hakola et al., 2012; Noe et al., 2012; Bourtsoukidis et al., 2014). While the concentrations of sesquiterpenes are generally much smaller than those of isoprene and monoterpenes (Sakulyanontvittaya et al., 2008; Sindelarova et al., 2014), sesquiterpenes could contribute significantly to SOA formation because of their reactivity and high aerosol yields (Calogirou et al., 1999; Khan et al., 2017). Previous studies found that sesquiterpenes contributed to the O_3 and OH reactivity in forest environments (Kim et al., 2011; Hellén et al., 2018). The recently developed Vocus proton-transfer-reaction time-of-flight mass spectrometry (PTR-TOF) enables the real-time detection of both VOCs and their oxidation products. With a new chemical ionization source called Vocus, the instrument significantly improves its detection efficiency of product ions compared with conventional PTR instruments (Krechmer et al., 2018). Based on a laboratory comparison of different chemical ionization techniques, Riva et al. (2019) ~~revealed~~ showed that Vocus PTR-TOF is sensitive to a large range of oxygenated VOCs. With the deployment of a Vocus PTR-TOF in the French Landes forest, Li et al. (2020) observed various terpenes and terpene oxidation products, including a range of organic nitrates.

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With the benefit of the ~~state of the art~~ capabilities of Vocus PTR-TOF to detect hundreds to thousands of molecules, a great challenge arises to analyze the complex dataset where emission sources and atmospheric physical and chemical processes are mixed together. The characteristic analysis based solely on individually identified compounds cannot give the full picture of the measurements. Factor analytical techniques, e.g., positive matrix factorization (PMF), have been utilized to extract information from mass spectrometer data by resolving co-varying signals with common sources or atmospheric processes into a single factor (Paatero and Tapper, 1994). For example, PMF analysis has been widely applied by the research community using aerosol mass spectrometer to identify multiple primary organic aerosol sources and SOA aging processes (Lanz et al., 2007; Ulbrich et al., 2009; Zhang et al., 2011). Yan et al. (2016) successfully applied PMF to unit-mass-resolution (UMR) nitrate ion-based chemical ionization mass spectrometer (NO_3^- CIMS) data to differentiate mainly monoterpene highly oxygenated organic molecules (HOMs) formed from different formation pathways in the boreal forest. The application of PMF to high-resolution (HR) NO_3^- CIMS data by Massoli et al. (2018) identified more HOM factors at an isoprene-dominated forest site in Alabama, USA. Recently, the mass spectral binning combined with PMF (binPMF) was proposed as a novel and simple method for analyzing high-resolution mass spectra datasets (Zhang et al., 2019a). This approach divides the full mass spectra into small bins as input data to PMF, thus avoiding the time-consuming and complicated peak identification. Zhang et al. (2019b) further applied binPMF to sub-ranges of ambient NO_3^- CIMS mass spectra and separated more meaningful factors related to chemical processes yielding HOMs.

In this work, we present the first factor analysis on Vocus PTR-TOF datasets to identify and apportion the contribution of different sources and formation pathways to atmospheric organic vapors. The measurements were conducted in two forest ecosystems in Europe, the French Landes forest and the boreal forest in southern Finland. Due to orders of magnitude differences in the signal intensities of ions between lower mass range and higher mass range, we divided the mass spectra into two sub-ranges (50–200 Th and 201–320 Th) and performed binPMF analysis on these ranges separately. While the UMR analysis loses all possible HR details and the HR peak identification introduces high uncertainties due to the complexity of overlapping peaks, the binPMF method includes as much of the HR information as possible in a robust way. The resolved factors were linked to possible sources or chemistry processes by examining their mass profiles, time series, diurnal cycles, and correlation with molecular markers. Comparison were discussed among different factors and also between the two forests for the common sources apportioned. Based on the interpretation of the resolved factors, further insights were provided regarding the atmospheric processes of monoterpenes and sesquiterpenes.

2 Materials and methods

2.1 Site description and measurement period

The measurement data were obtained during summertime in two forest environments in Europe, the Landes forest in southwestern France and the boreal forest research station SMEAR (Station for Measuring Forest Ecosystem-Atmosphere Relations) II in southern Finland. The field campaign in the Landes forest was conducted from 8 to 20 July 2018 as part of the

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Characterization of Emissions and Reactivity of Volatile Organic Compounds in the Landes forest (CERVOLAND) campaign. An overview of the Vocus PTR-TOF measurements in the Landes forest has been presented earlier by Li et al. (2020). The ambient observations at the SMEAR II station were performed during 18 June – 18 July 2019.

The Landes forest (44º29'N, 0º57'W) is the largest man-made pine forest in Europe, mainly filled with maritime pine trees
100 (*Pinus pinaster Aiton*). The sampling site is situated at the European Integrated Carbon Observation System (ICOS) station at Bilos. The nearest urban area of the Bordeaux metropole is around 40 km to the northwest. A more detailed description of the measurement site can be found elsewhere (Kammer et al., 2018; Bsaibes et al., 2020; Li et al., 2020). Ambient meteorological parameters, including temperature, relative humidity (RH), wind speed and direction, solar radiation, and pressure, and mixing ratios of nitrogen oxides (NO and NO_2) and O_3 were continuously monitored at the station throughout the campaign.

105 The SMEAR II station (61º51'N, 24º17'E) is located in a boreal mixed-coniferous forest in Hyytiälä, southern Finland (Hari and Kulmala, 2005). The site is dominated by a rather homogeneous Scots pine (*Pinus sylvestris L.*) stand and represents a rural background measurement station. The nearest large city Tampere, located about 60 km to the southwest, has approximately 200 000 inhabitants. The station is equipped with extensive facilities to measure forest ecosystem-atmosphere interactions. Ambient meteorological parameters (i.e., global radiation, UVA, UVB, temperature, RH, pressure, and wind
110 speed and direction), mixing ratios of various trace gases (i.e., carbon dioxide, carbon monoxide, sulfur dioxide, NO_x, and O_3), and particle concentration and size distribution, are continuously measured at the station.

2.2 Instrumentation

A Vocus PTR-TOF was deployed in both forest ecosystems to characterize atmospheric organic vapors. Equipped with a new chemical ionization source with a low-pressure reagent-ion source and focusing ion-molecule reactor (FIMR), the Vocus PTR-
115 TOF is able to measure organic vapors with a wide range of volatilities (Krechmer et al., 2018; Riva et al., 2019; Li et al., 2020). A quadrupole radio frequency (RF) field inside the FIMR focuses ions to the central axis and improves the detection efficiency of product ions. Compared with conventional PTR instruments, the sensitivity and detection efficiency of Vocus PTR-TOF are significantly improved (detection limit < 1 pptv). With a high water mixing ratio (10% v/v–20% v/v) in the FIMR, the instrument shows no humidity dependence for sensitivity. More instrumental details have been provided elsewhere
120 (Krechmer et al., 2018; Li et al., 2020).

During both campaigns, we operated the Vocus ionization source at a pressure of 1.5 mbar. Sample air was drawn in through a ~1-m-long PTFE tubing (10 mm o.d., 8 mm i.d.). A sample air flow of 4.5 L min^{-1} was used to reduce inlet wall losses and sampling delay. Around 100–150 ccm of this flow was sampled into the Vocus and the remainder was directed to the exhaust. The mass resolving power of the long TOF mass analyzer was 12 000 – 13 000 m Δm^{-1} during our measurements.
125 Data were recorded with a time resolution of 5 s. During the campaign in the Landes forest, background checks were automatically performed every hour using ultra-high-purity nitrogen (UHP N_2). The instrument was calibrated twice a day using a mixture of terpenes (α-/β-pinene+limonene; *p*-cymene). For measurements at the SMEAR II station, background measurements by injection of zero air using a built-in active carbon filter and quantitative calibrations with a multi-component

standard cylinder were automatically conducted every three hours. All the *m/z* ratios mentioned in this work include the
130 contribution from the charger ion (H$^+$, mass of 1 Th) unless stated otherwise.

2.3 binPMF data preparation and analysis

As described by Zhang et al. (2019a), binPMF divides the mass spectra into small bins and then takes advantage of PMF
analysis to separate different sources or formation processes. The binPMF allows utilization of the high-resolution information
of the complex mass spectra without the time-consuming and potentially error-prone steps of peak identification and peak
135 fitting before the factorization. Selected peaks of interest can be analyzed after binPMF, based on the output factors. PMF
assumes that factor profiles are constant and unique, and the measured signal of a chemical component is a linear combination
of different factors. This approach does not require a priori information about the factors. The detailed working principle of
PMF has been provided in numerous previous studies (Paatero and Tapper, 1994; Zhang et al., 2011; Yan et al., 2016).

To prepare the data and error matrices for PMF input, the Vocus PTR-TOF data were processed using the software
140 package "Tofware" (v3.2.0; Tofwerk), which runs in the Igor Pro environment (WaveMetrics, OR, USA). The detailed data
processing routines have been presented elsewhere (Stark et al., 2015). Signals were averaged over 30 min for data processing.
Unlike traditional UMR or HR fitting of the mass spectra, in binPMF analysis, the mass spectra were divided into small bins
after mass calibration. Due to the greater mass resolving power of the TOF mass analyzer compared with former binPMF
studies (Zhang et al., 2019a, 2019b), a bin width of 0.01 Th was applied in this study. At a nominal mass *N*, signals between
145 *N*-0.15 and *N*+0.35 Th were included for binning. The error matrix was calculated to include uncertainty from counting
statistics following Poisson distribution and instrument electronic noise, as described by Yan et al. (2016) and Zhang et al.
(2019a). The electric noise was estimated as the median of the standard deviation of binned noise signals between two nominal
masses, with noise range between *N*+0.4 and *N*+0.6 Th.

Figure 1 shows the average mass spectra of the measurements in the Landes forest as an example. Since the signal intensity
150 of larger molecules is generally much lower than that of low-mass molecules, we divided the mass spectra into two sub-ranges,
the low mass range (51–200 Th) and the high mass range (201–320 Th). Factor analysis was separately performed on these
two sub-ranges using an Igor-based open-source PMF Evaluation Tool (PET; http://cires1.colorado.edu/jimenez-
group/wiki/index.php/PMF-AMS_Analysis_Guide#PMF_Evaluation_Tool_Software). We ran the PMF up to ten factors for
both sub-ranges. For the low mass range of 51–200 Th, the signals at *m/z* 81 Th ($C_6H_8H^+$, monoterpene fragment) and 137 Th
155 ($C_{10}H_{16}H^+$, monoterpenes) were markedly higher than the others. Because PMF assumes that the data matrix can be explained
by a linear combination of different factors, even a very tiny fraction of these high peaks is split into a factor, they may
dominate the mass profile of the factor. As shown in Fig. S1, w~~W~~ith the inclusion of $C_6H_8H^+$ and $C_{10}H_{16}H^+$~~these ions~~, the mass
profiles of several factors were quite similar and dominated by these peaks ~~(Fig. S1)~~. Therefore, the major mass bins of these
ions were excluded for further PMF analysis, but their corresponding isotopes were retained, effectively downweighting their
160 contributions to the PMF result. This simple approach by removing the main peaks of the largest signals produced factors that
made sense both chemically and through their temporal behaviors, lending confidence in the results. However, to quantify the

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relative contribution of different factors, the signals of these removed mass bins were counted back into their corresponding factors. More details can be found in Sect. 4.4.

2.4 Estimation of OH and NO$_3$ radicals

165   The OH concentration was calculated by scaling the measured UVB radiation intensity with the empirically derived factors from Petäjä et al. (2009) and Kontkanen et al. (2016):

$$[OH]_{proxy} = \left(\frac{8.4 \times 10^{-7}}{8.6 \times 10^{-10}} UVB^{0.32}\right)^{1.92}$$

Measurements of NO$_3$ concentration is challenging. The concentration of NO$_3$ radical was calculated by assuming a steady state between its production from O$_3$ and NO$_2$ and its removal by oxidation reactions and losses in the atmosphere.

170   Details can be found in Allan et al. (2000) and Peräkylä et al. (2014).

3 Dataset overview

Figure 2 shows the temporal behaviors of temperature, global radiation, concentrations of O$_3$ and NO$_x$, and concentrations of isoprene and monoterpenes in the Landes forest and at the SMEAR II station. In the Landes forest, the weather was mainly sunny during the observation period (global radiation > 400 W m^{-2}), indicating strong photochemical activity. The air mass in

175   the forest was largely influenced by local sources, with wind speeds below canopy lower than 1 m s^{-1} over the whole campaign. The O$_3$ concentration fluctuated dramatically between day and night, with the average daytime concentration peaking up to 50 ppb and the average nighttime level falling below 2 ppb (Li et al., 2020). The low O$_3$ concentration at night was probably to some extent caused by its titration by monoterpenes (Fig. 2a; Kammer et al., 2018, 2020). The Landes forest is known for strong monoterpene emissions (Simon et al., 1994). During our measurements, the average mixing ratios of isoprene and

180   monoterpenes were 0.6 ppb and 6.0 ppb, respectively. More details about this dataset can be found in Li et al. (2020). All data in the Landes forest are reported in local time and all data at the SMEAR II station in Finnish winter time (both equal UTC time + 2).

During the measurements at the SMEAR II station, 84% (26 out of 31) of the days had strong photochemistry (global radiation > 400 W m^{-2}), with the rest being cloudy days. The diurnal variation in O$_3$ concentration was not as dramatic as that

185   in the Landes forest. In the daytime, the O$_3$ concentration sometimes reached up to 50 ppb. At night, the O$_3$ level still largely remained high, above 20 ppb, in contrast to the observations in the Landes forest. A possible explanation is less nighttime O$_3$ consumption by terpenes at the SMEAR II station. On average, the mixing ratios of isoprene and monoterpenes were 0.2 ppb and 0.8 ppb, respectively, during the measurements, much lower than those in the Landes forest.

4 Results and discussion

4.1 Choice of PMF solution and factor interpretation

To interpret the PMF results, the most critical decision is to choose the best number of factors. More factors introduce more degrees of freedom to explain variations in the data, but too many factors may cause splitting of real factors and lead to mathematical artifacts without physical meaning (Ulbrich et al., 2009). The factor interpretation results in this work are summarized in Table 1. In the factor name, L means the Landes forest and S means the SMEAR II station.

For the low mass range of the Landes forest dataset, the Q/Q_{exp} varied from 15.5 to 6.0 for two to ten factors (Q is the total sum of the squares of the scaled residuals for PMF solutions). The larger Q/Q_{exp} indicates underestimation of the errors or high residuals for some bins that cannot be simply modeled by the solution (Ulbrich et al., 2009). After seven factors, increasing the factor number does not significantly decrease the Q/Q_{exp} (step change < 7%). The optimal solution of seven factors was chosen by evaluating the variations of Q/Q_{exp} vs. varying factor number, the distribution of the scaled residuals for each m/z, sum of the squares of scaled residuals, factor mass profile, factor time series and diurnal cycles, and also signs of split factors (Ulbrich et al., 2009; Zhang et al., 2011). ~~after a detailed evaluation following the procedures proposed by Ulbrich et al. (2009) and Zhang et al. (2011).~~ Figure S2 shows the distribution of scaled residuals as a function of m/z. For some bins the residuals are still high (the scaled residuals as high as ±200). The seven factors include Factor L1 closely related to the $C_4H_8H^+$ ion, Factor L2 attributed to a plume event occurring on a single night during the campaign, Factor L3 mainly containing lightly oxidized compounds with six or seven carbon atoms ("C_6" or "C_7"), Factor L4 representing monoterpenes, Factor L5 indicative of isoprene and its oxidation products, Factor L6 identified as unknown source with large contributions from unknown peaks, and Factor L7 dominated by monoterpene lightly oxidized compounds. The direct comparison of the mass spectra, time series, and diurnal cycles of six-factor and eight-factor solutions are shown in Fig. S3 and Fig. S4. In the six-factor case, the $C_4H_8H^+$ ion-related factor cannot be separated. With eight-factor results, the factor representing isoprene and its oxidation products is split into two components with similar time series. For the high mass range of the Landes forest dataset, the Q/Q_{exp} decreased from 2.5 to 0.9 for two to ten factors. After evaluation, we choose the eight-factor solution to explain the data. The Q/Q_{exp} value of the eight-factor solution was 1.1 and the decreasing trend in Q/Q_{exp} obviously slowed down after eight factors. The distribution of scaled residuals as a function of m/z for the eight-factor solution is shown in Fig. S5. The eight factors are interpreted as Factor L8 dominated by lightly oxygenated compounds containing 13 carbon atoms ("C_{13}"), Factor L9 attributed to a plume event occurring on a single night during the campaign, Factor L10 mainly related to sesquiterpene lightly oxidized compounds, Factor L11 representing more oxidized products mainly from monoterpene oxidation~~s~~, Factor L12 indicating sesquiterpenes, Factor L13 largely composed of monoterpene-derived organic nitrates, Factor L14 mainly containing oxidized compounds with twelve, fourteen or sixteen carbon atoms ("C_{12}", "C_{14}" or "C_{16}")and Factor L15 as unknown source largely contributed by siloxane compounds. Figure S6 and Figure S7 display the mass spectra, time series, and daily variations of seven-factor and nine-factor solutions. In the seven-factor case, monoterpene more oxidized products and monoterpene-derived organic nitrates are mixed together into a single factor. However, in the nine-factor solution,

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the unknown factor mainly composed of siloxane compounds is split into two factors with similar mass profiles and similar diurnal trends.

For the SMEAR II dataset, the optimal solutions of five-factor and four-factor are chosen for the low and high mass ranges, respectively. The Q/Q_{exp} varied from 7.2 to 2.5 for two to ten factors in the low mass range and from 2.0 to 1.0 for two to ten factors in the high mass range. The five factors for the low mass range are identified as Factor S1 - $C_4H_8H^+$ ion-related, Factor S2 - monoterpenes, Factor S3 - lightly oxidized compounds with six to nine carbon atoms, Factor S4 - isoprene and its oxidation products, and Factor S5 - monoterpene lightly oxidized compounds. The mass spectra, time series, and diurnal profiles of the four-factor and six-factor solutions for the low mass range are presented in Fig. S8 and Fig. S9. For the four-factor solution, monoterpene lightly oxidized products are not separated as a single factor and mixed into the others. In the six-factor case, the factor indicative of monoterpene lightly oxidized products is split into two factors. The four factors for the high mass range include Factor S6 - sesquiterpene lightly oxidized products, Factor S7 - sesquiterpenes, Factor S8 - more oxidized compounds, and Factor S9 - unknown source. The direct comparison of the mass spectra, time series, and diurnal variations of three-factor and five-factor solutions are shown in Fig. S10 and Fig. S11. The three-factor solution does not identify a factor representing sesquiterpenes. In the five-factor case, the factor of unknown source mainly contributed by siloxane compounds is split into two factors with similar mass profiles.

The rotational freedom of the PMF solutions was explored through the use of the FPEAK parameters. For each of the optimal solutions, we varied the FPEAK values between -1 and +1 with the step of 0.2. For the low mass ranges of the Landes and SMEAR II dataset, the varying FPEAK values did not change the factor profiles and time series much, indicating that varying FPEAK values from -1 to +1 did not affect the overall results of PMF analysis. For the high mass range of the Landes measurements, we saw variations especially in the factor profiles by varying FPEAK values. But after a detailed evaluation, we found no evidence that solutions with FPEAK values away from zero were preferable. However, for the high mass range of the SMEAR II measurements, the solutions with positive values of FPEAK worked better than that with FPEAK = 0 in terms of factor profiles. The factor time series were similar when FPEAK values varied. But for the factor profiles with positive FPEAK values, the factor of monoterpene more oxidized products including organic nitrates contained less traces of siloxanes and showed elevated fractions of the corresponding fingerprint peaks (Fig. S12). As discussed in Sect. 4.3, these siloxanes can come from cosmetics and personal care products, and silicone oils used in instrument pumps. The temporal variations of these siloxanes differed significantly from those of monoterpene more oxidized products. After evaluation, we chose the solution with FPEAK = +0.6 for the high mass range of the SMEAR II dataset, where siloxanes feature more in one factor.

4.2 Source identification in the Landes forest

Figure 3 and Figure 4 illustrate the factor profiles, time series, and diurnal variations of the seven factors resolved in the low mass range. For the high mass range, the mass spectra of the five factors are shown in Fig. 5, and their time series and daily variations in Fig. 6. The high-resolution peak fitting was further performed on the mass profile to identify the fingerprint peaks

in the factors. Fingerprint peaks are defined largely based on their distribution in the factors rather than their absolute intensity
255   in the mass profile. The correlation map of each factor with various compounds is shown in Fig. S13.

4.2.1 Low mass range (51–200 Th)

Factor L1: $C_4H_8H^+$ ion-related

Factor L1 shows irregular diurnal variations with spiky peaks in the time series (Fig. 4b). The major bins that are largely
distributed into this factor are $C_4H_8H^+$ and $C_4H_{10}O_2H^+$. Factor L1 closely correlates with these fingerprint peaks. Considering
260   the high signal intensity of $C_4H_8H^+$ ion and its large contribution to this factor, we name Factor L1 as $C_4H_8H^+$ ion-related.

Factor L2: A plume event

Factor L2 is identified as a plume event occurring on a single night during the campaign. As shown in Fig. 4a, the time
series of this factor are characterized by much higher intensities at midnight of 9 July 2018 than over the other days. Fingerprint
peaks in this factor are aromatic compounds such as $C_6H_6H^+$, $C_7H_6H^+$, and $C_6H_6OH^+$. Factor L2 is well correlated with benzene
265   and phenol ($r^2 = 0.88$; Fig. S13).

Factor L3: C_6 and C_7 lightly oxidized products

The diurnal cycle of Factor L3 exhibits a small morning peak at 9:00 and significantly elevated intensities during
nighttime, peaking at around 22:00 (Fig. 4b). As illustrated in the mass profile of Factor L3, this factor is mainly composed of
lightly oxidized compounds containing six or seven carbon atoms such as $C_6H_{10}OH^+$, $C_7H_{10}OH^+$, $C_6H_{10}O_2H^+$, and $C_7H_{12}O_2H^+$.

270   *Factor L4: monoterpenes*

The mass profile of Factor L4 is dramatically characterized by a monoterpene peak ($^{13}CC_9H_{16}H^+$) and its major fragments
(i.e., $^{13}CC_5H_8H^+$ and $C_7H_8H^+$) inside the instrument. As shown in Fig. 4b, the diurnal variation of this factor follows a similar
pattern to that of monoterpenes (Li et al., 2020). The signal intensity of the factor starts to increase at 20:00, peaks at midnight,
and then drops to around the detection limit during daytime. Monoterpene emissions are mainly influenced by temperature
275   (Hakola et al., 2006; Kaser et al., 2013). Therefore, with the continuous emissions of monoterpenes and the shallow boundary
layer at night, the signal intensities of monoterpenes are observed to be elevated. The signal of $C_{10}H_{16}OH^+$ is also mostly
resolved into this factor. $C_{10}H_{16}O$ could be primary emissions of oxygenated monoterpenes or monoterpene oxidation products
(Kallio et al., 2006; McKinney et al., 2011). Previous ambient observation has demonstrated that the atmospheric behavior of
$C_{10}H_{16}O$ has high similarity to that of monoterpenes (Li et al., 2020).

280   *Factor L5: isoprene and its oxidation products*

The marker peaks in Factor L5 are highly dominated by isoprene and its major oxidation products in the atmosphere, i.e.,
$C_5H_8H^+$ and $C_4H_6OH^+$ (Wennberg et al., 2018). Isoprene emissions strongly depend on light intensity (Monson and Fall, 1989;
Kaser et al., 2013) and generally show high concentrations in the day. Similarly, the daily variations of Factor L5 display
maximum signal during daytime and minima at night.

285   *Factor L6: unknown source*

Factor L6 is characterized by increased signals in the afternoon. The major peaks in its factor profile are $C_6H_4O_2H^+$, $C_6H_6O_3H^+$, and numerous unidentified peaks with negative mass defect. As this factor is clearly separated as a single source with high signals during our observations and the molecule markers remain unidentified, we name this factor as an unknown source.

290 *Factor L7: monoterpene lightly oxidized products*

Fingerprint peaks in this factor are monoterpene oxidation products with oxygen number from one to three, such as $C_9H_{14}OH^+$, $C_{10}H_{14}OH^+$, $C_{10}H_{16}O_2H^+$, and $C_{10}H_{16}O_3H^+$. This factor displays clear morning and evening peaks, similar to the behavior of these lightly oxidized compounds (Li et al., 2020).

4.2.2 High mass range (201–320 Th)

295 *Factor L8: C_{13} lightly oxidized products*

The mass profile of Factor L8 is characterized by high peaks of lightly oxidized compounds containing 13 carbon atoms, like $C_{13}H_{18}O_2H^+$ and $C_{13}H_{20}O_3H^+$. Similar to C_6 and C_7 lightly oxidized compounds, this factor shows a morning peak at 9:00 and an evening peak at around midnight (Fig. 6b).

Factor L9: A plume event

300 Factor L9 is characterized with much higher intensities on a single night (9 July 2018) during the campaign (Fig. 6a). Fingerprint peaks in the mass profile of Factor L9 are numerous unidentified ions. The time series of Factor L9 correlate tightly with aromatic compounds C_6H_6 and C_6H_6O ($r^2 = 0.75$).

Factor L10: sesquiterpene lightly oxidized products

The fingerprint peaks identified in this factor are $C_{15}H_{22}OH^+$, $C_{15}H_{24}OH^+$, $C_{15}H_{22}O_2H^+$, $C_{15}H_{24}O_2H^+$, and $C_{15}H_{24}O_3H^+$,
305 which are typical reaction products from sesquiterpene oxidations (Fu et al., 2009; Yee et al., 2018). The signal intensity of this factor is generally high during nighttime, but shows another morning peak at 8:00. In addition to the production from sesquiterpene oxidation processes, $C_{15}H_{22}O$ and $C_{15}H_{24}O$ can be oxygenated sesquiterpene alcohols and aldehydes directly emitted from vegetation (Kännaste et al., 2014).

Factor L11: monoterpene more oxidized products

310 The mass spectrum of this factor is mainly characterized by more oxidized compounds from monoterpene oxidations such as $C_{10}H_{16}O_4H^+$, $C_{10}H_{14}O_5H^+$, $C_{10}H_{16}O_5H^+$, and $C_{10}H_{16}O_6H^+$. As shown in Fig. S13, the time series of Factor L11 show good correlations with these compounds. Compared with monoterpene lightly oxidized compounds, the diurnal cycle of this factor shows a broad daytime distribution peaking between 14:00 and 20:00, caused by strong and complex photochemical reactions during the day.

315 *Factor L12: sesquiterpenes*

The mass spectra of Factor L12 are clearly dominated by a big single peak of $C_{15}H_{24}H^+$, indicating the influence of sesquiterpenes. Sesquiterpene emissions from plants are found to exhibit a strong dependence on temperature (Duhl et al., 2008). Therefore, similar to the diurnal cycle of Factor L4, this factor shows prominently enhanced signals during nighttime.

Factor L13: monoterpene-derived organic nitrates

320   The signal intensity of this factor starts to increase in the early morning (around 7:00) and presents a distinct morning peak at 9:00. In addition, a much smaller evening peak is observed at 21:00. The daily variations of this factor are quite similar to those of monoterpene-derived organic nitrates measured in the Landes forest (Li et al., 2020). Consistently, the major peaks in the factor profile are $C_{10}H_{15}NO_4H^+$, $C_{10}H_{15}NO_5H^+$, $C_9H_{13}NO_6H^+$, and $C_{10}H_{15}NO_6H^+$, indicating the dominant contribution of organic nitrates formed from monoterpene oxidation processes.

325 *Factor L14: C_{12}, C_{14} or C_{16} lightly oxidized compounds*

      The mass profile of Factor L14 is characterized with distinct peaks of C_{12}, C_{14} or C_{16} lightly oxidized compounds, i.e., $C_{12}H_{26}O_3H^+$, $C_{14}H_{26}O_2H^+$, $C_{16}H_{30}O_2H^+$, and $C_{16}H_{30}O_3H^+$. The time series of Factor L14 correlate very well with those of $C_{12}H_{26}O_3$ ($r^2 = 0.83$), characterized with enhanced signals during daytime and low intensities at night (Fig. 6b). $C_{12}H_{26}O_3$ has been found during the photooxidation of dodecane (Zhang et al., 2014).

330 *Factor L15: unknown source*

      The mass profile of Factor L15 is predominantly characterized by high cyclic volatile methyl siloxanes (VMSs) peaks and some unidentified peaks (Fig. 5). The major cyclic VMSs are protonated D3 siloxane, D4 siloxane, and their H_3O^+ cluster ions, which have been widely used in cosmetics and personal care products (Buser et al., 2013; Yucuis et al., 2013). The diurnal cycle of this factor shows a bit higher intensity during daytime but also big background signals at night. A similar factor has

335 also been identified at the SMEAR II station. More detailed discussions can be found in Sect. 4.3.2.

4.3 Source identification in the southern Finnish boreal forest

The factor profiles, time series, and diurnal cycles of the five-factor solution for the low mass range are presented in Fig. 7 and Fig. 8. Figure 9 and Figure 10 present the mass spectra, time series, and daily variations of the four factors identified in the higher mass range at the SMEAR II station. The correlation coefficients among each factor and various fingerprint compounds

340 can be found in Fig. S14.

4.3.1 Low mass range (51–200 Th)

Factor S1: $C_4H_8H^+$ ion-related

      Similar to the source identification in the Landes forest, a factor related to $C_4H_8H^+$ ion is clearly resolved at the SMEAR II station. The major peaks in this factor are $C_4H_8H^+$, $C_4H_{12}O_2H^+$, and $C_4H_{14}O_3H^+$.

345 *Factor S2: monoterpenes*

      A factor representing monoterpenes is also identified at the SMEAR II station, with fingerprint peaks of $^{13}CC_5H_8H^+$, $C_7H_{10}H^+$, and $^{13}CC_9H_{16}H^+$. Monoterpenes undergo some degree of fragmentation within PTR instruments, and $C_6H_8H^+$ and $C_7H_{10}H^+$ have been observed to be the major fragments of monoterpenes (Tani et al., 2003; Tani, 2013; Kari et al., 2018). The signal intensity of monoterpenes at the SMEAR II station is much lower than that in the Landes forest.

350 *Factor S3: C_6-C_9 lightly oxygenated compounds*

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The mass profile of Factor S3 is characterized by lightly oxygenated compounds with carbon atoms varying from six to nine (C_6-C_9) such as $C_6H_{10}OH^+$, $C_6H_{12}OH^+$, $C_7H_{10}OH^+$, $C_8H_{14}OH^+$, and $C_9H_{18}OH^+$. The signal intensity of this factor shows high peaks at night and low appearance during daytime. These lightly oxygenated molecules can be directly emitted from anthropogenic and biogenic sources or come from oxidation processes of various VOC precursors (Conley et al., 2005; Pandya et al., 2006; Rantala et al., 2015; Hartikainen et al., 2018). For instance, $C_7H_{10}O$ has been found from direct soil emissions (Abis et al., 2020) or oxidation processes of 1,2,4-trimethyl benzene (Mehra et al., 2020). Therefore, we expect the molecules in this factor to be either directly emitted or as oxidation products of forest emissions.

Factor S4: isoprene and its oxidation products

At the SMEAR II station, a factor largely composed of isoprene and its oxidation products is also resolved. The outstanding peaks in the factor profile are $C_5H_8H^+$, $C_4H_6OH^+$, $C_4H_8O_2H^+$, and $C_5H_8O_2H^+$. The signal intensity of this factor is around ten times lower than that of Factor L5 measured in the Landes forest. Similar to previous isoprene observations at the sampling site (Hakola et al., 2012), this factor shows a broad daytime peak and low signals at night.

Factor S5: monoterpene lightly oxidized products

Similar to Factor L7 identified in the Landes forest, this factor is characterized by major peaks of monoterpene lightly oxidized compounds, as shown in Fig. 7. The signal intensity of this factor starts to increase at 20:00 and presents an obvious morning peak at 7:00.

4.3.2 High mass range (201-320 Th)

Factor S6: sesquiterpene lightly oxidized products

This factor is identified as sesquiterpene lightly oxidized compounds with high peaks of $C_{14}H_{22}OH^+$, $C_{14}H_{24}OH^+$, $C_{15}H_{22}OH^+$, and $C_{15}H_{24}OH^+$, similar to Factor L10 in the Landes forest. The time series of this factor show strong correlations with the lightly oxidized products of sesquiterpenes (Fig. S14; $r^2 > 0.88$).

Factor S7: sesquiterpenes

Similar to Factor L12 in the Landes forest, this factor is characterized by the big peak of $C_{15}H_{24}H^+$, demonstrating the dominance of sesquiterpenes in the factor. Figure S14 shows that this factor closely correlates with monoterpenes and sesquiterpenes, with r^2 being 0.73 and 0.85, respectively. Compared with the identification of Factor L12, representing sesquiterpenes in the Landes forest, the signal intensity of this factor at the SMEAR II station is approximately three times lower. Including the lower signals of monoterpenes and isoprene, the results indicate weaker biogenic VOC emissions in the Hyytiälä boreal forest than in the Landes forest.

Factor S8: monoterpene more oxidized products including organic nitrates

Factor S8 is mainly composed of more oxidized compounds, particularly from monoterpene oxidation processes, including monoterpene-derived organic nitrates. The major peaks are shown in Fig. 9. Mixed with monoterpene-derived organic nitrates, this factor of more oxidized compounds displays a small morning peak at 8:00 and generally high signals during daytime (Fig. 10).

Factor S9: unknown source

385   The marker peaks of Factor S9 are mainly high cyclic volatile methyl siloxanes (VMSs) and unidentified compounds (Fig. 9), i.e., protonated D3 siloxane, D4 siloxane, and their H_3O^+ cluster ions. In addition to cosmetics and personal care products, siloxanes can also be emitted by silicone oils (Schweigkofler et al., 1999), which have been widely used in instrument pumps (Gonvers et al., 1985). In this study, the temporal behaviors of Factor S9 are contributed by high background signals and present a very regular diurnal cycle with higher signal intensities during daytime and lower ones at night, which basically

390   follow the variations in ambient temperature. Therefore, we speculate that Factor S9 is mainly caused by emissions from silicone oil pumps used by several instruments in the container, and these emissions are influenced by daily temperature changes.

4.4 Comparison among different factors

The monoterpene factor and sesquiterpene factor correlate very well with each other at both sites (Fig. 11; $r^2 = 0.69$ in the

395   Landes forest and $r^2 = 0.59$ at the SMEAR II station). The emissions of monoterpenes and sesquiterpenes are both strongly influenced by temperature. Their signals peak at night with the effect of the shallow boundary layer. In the daytime, the low signals of the monoterpene and sesquiterpene factors are likely a combination of enhanced atmospheric mixing after sunrise and the rapid photochemical consumption of monoterpenes and sesquiterpenes. The signal of monoterpene factor is around 15 times higher than that of sesquiterpene factor at the SMEAR II station while it is around 60 times in the Landes forest. Previous

400   studies found that sesquiterpene emissions from pines, spruces, and birches under normal conditions were 5-15% of total monoterpene emissions by mass (Rinne et al., 2009 and references therein), in line with our observations.

In the Landes forest, a factor of C_6 and C_7 lightly oxidized products (Factor L3) was resolved in the low mass range and a factor representative of C_{13} lightly oxidized products (Factor L7) was identified in the high mass range. Interestingly, these two factors show a close correlation with each other ($r^2 = 0.64$). The C_6 oxygenated compounds have been observed during

405   the oxidation processes of benzene and C_7 oxygenated compounds from toluene oxidations (Sato et al., 2012; Zaytsev et al., 2019). These compounds can also be directly emitted from biogenic or anthropogenic sources (Conley et al., 2005; Pandya et al., 2006; Rantala et al., 2015). The temporal behaviour of Factor L7 is similar to that of Factor L3, indicating potentially similar formation pathways of these lightly oxygenated compounds. Therefore, the C_{13} oxidized compounds are speculated to be produced through the dimer formation mechanisms of C_6 and C_7 species (Valiev et al., 2019). In addition, $C_{13}H_{20}O_3$ can be

410   direct emissions of methyl jasmonate (Meja), which is a typical green leaf volatile used in plant-plant communications for defensive purposes (Cheong and Choi, 2003). But considering the close correlation between Factor L3 and Factor L7, we conclude that these C13 lightly oxidized compounds are formed from atmospheric oxidation processes, not direct plant emissions.

Monoterpene lightly oxidized products and sesquiterpene lightly oxidized products were resolved as individual factors at

415   both sites (Factor L7 vs. Factor L10 in the Landes forest and Factor S5 vs. Factor S6 at the SMEAR II station). While the diurnal variations of monoterpene lightly oxidized products are similar to those of sesquiterpene lightly oxidized products,

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their time series do not follow very well with each other, suggesting the probably different formation pathways or different factors influencing the atmospheric processes of monoterpenes and sesquiterpenes. More discussions can be found in Sect. 4.6.

420   In this study, the source apportionment analysis was performed separately on two subranges of the mass spectra. It can happen that the same factor is identified in both subranges. For example, both Factor L2 and Factor L9 are defined as the plume event during the measurements. The time series of Factor L2 and Factor L9 show a high correlation coefficient of 0.93 and correlate tightly with aromatic compounds, indicating the major influence of anthropogenic sources. As mentioned above, the air masses in the Landes forest were relatively stable during our observations with wind speed below canopy < 1 m s^{-1}.
425   Therefore, the influence of long-range regional transport on the atmosphere in the forest is expected to be minor. We speculate that the plume event is a result of local anthropogenic disturbances favored by the lower boundary layer height at night.

4.5 Comparison between the two forests

To give an overview of the source distributions in the two forest ecosystems, we calculated the mass fraction of each factor based on their average signal intensities. We acknowledge that it is not a perfect method to quantify the contributions of various
430   sources and formation processes. The sensitivities of different VOCs measured by the PTR instruments may vary by a factor of 2-3 (Sekimoto et al., 2017; Yuan et al., 2017). The uncertainties can come from the challenge to convert the signal intensity to atmospheric concentrations because of problematic calibrations, especially given that many unknown molecules exist in the mass spectra. The major bins at m/z 81 Th and 137 Th, which were initially excluded to perform PMF analysis, were counted into their corresponding factors. For example, the signals of the discarded bins at m/z 81 Th and 137 Th were estimated by
435   multiplying their isotope signals by the corresponding scale number and added to the factor representing monoterpenes. The average mass fractions of various PMF factors in total measured organic vapors are shown in Fig. 12.

While the atmospheric environment and ecosystem processes differ markedly in the Landes forest and the southern Finnish boreal forest, the results of this study reveal similar biogenic sources and oxidation processes in these forest environments. For instance, the biogenic VOCs at the two sites are both dominated by monoterpenes, with the average fractions
440   of 29% in the Landes forest and at the SMEAR II station. These two forests are both characterized by pine trees, with dominant emissions of α-pinene and β-pinene (Riba et al., 1987; Simon et al., 1994; Hellén et al., 2018). According to the PMF results, isoprene and its major oxidation products in these environments (mainly C_4H_6O) contribute 14% and 21% in the two ecosystems, respectively. Factors indicative of sesquiterpenes are identified in the high mass range at both sites. The average contribution of sesquiterpenes (0.5% in the Landes forest and 1.7% at the SMEAR II station) is much smaller than that of
445   monoterpenes and isoprene. Factors of the lightly oxidized products, more oxidized products, and organic nitrates of monoterpenes/sesquiterpenes in total contribute 8% and 12% of the measured organic vapors in the Landes forest and at the SMEAR II station, respectively.

The factor related to $C_4H_8H^+$ ion was resolved at both sites and contributes 10% in the Landes forest and 16% at the SMEAR II station. According to the discussions by Li et al. (2020), the observation of $C_4H_8H^+$ in the Landes forest can be

450 attributed to several sources. For instance, the protonated butene may contribute to the $C_4H_8H^+$ signal, which is emitted by biogenic or anthropogenic sources (Hellén et al., 2006; Zhu et al., 2017). Another possible explanation is that the $C_4H_8H^+$ ion is produced during the fragmentation of many VOCs in the PTR instruments (Pagonis et al., 2019). The green leaf volatiles (GLV) have been found to fragment at *m/z* 57 Th inside the PTR instruments, which are a group of six-carbon aldehyde, alcohols and their esters released by plants. Furthermore, butanol can easily lose an OH during the PTR source ionization and

455 produce prominent $C_4H_8H^+$ peaks (Spanel and Smith, 1997). Therefore, the condensation particle counters (CPCs) using butanol for aerosol measurements at the site could also be an important source of $C_4H_8H^+$ ions, although the exhaust air from these instruments has been filtered using charcoal denuder. At the SMEAR II station, the bivariate polar plot where the concentrations of air pollutants are shown as a function of WS and WD indicates that high signals of $C_4H_8H^+$ generally occur when the wind comes from the north (Fig. S15). Located in the north of the measurement container is a particle measurement

460 cottage with several CPCs inside using butanol. A previous study at this station also found that $C_4H_8H^+$ signals detected by PTR-TOF mainly come from butanol used by aerosol instruments (Schallhart et al., 2018). Therefore, it is expected that Factor S1 at the SMEAR II station is mainly contributed by butanol fragmentation inside the instrument where butanol comes from nearby aerosol instruments.

Figure 13 presents the comparison of the mass spectra of the common sources identified at both sites, with the *x* and *y*

465 axis showing the mass fraction of different bins in the factor profile. The scattering in the plots is mainly caused by mass bins with much lower mass fractions. However, the dominant bins with high mass contributions in the factor profiles generally correlate well and are located close to the 1:1 line. It shows that the mass spectra of the common sources match well in these two forests and the sources and processes are indeed similar despite the quite different regions the forests are in.

4.6 Insights into terpene oxidation processes

470 Terpenes undergo varying degrees of oxidations in the atmosphere and produce a large variety of organic compounds with different volatilities (Donahue et al., 2012; Ehn et al., 2014). With the sub-range PMF analysis performed in this study, terpene reaction products with varying oxidation degrees are successfully separated. The sources of monoterpene lightly oxidized products, sesquiterpene lightly oxidized products, monoterpene more oxidized compounds, and monoterpene-derived organic nitrates are identified in both forests with distinct characteristics. These factors account for 8-12% of the measured organic

475 vapors in the two forests. It provides a great opportunity to gain insights into terpene oxidation processes. Because some environmental parameters, for example, measurements of UVB to estimate OH concentration, are not available in the Landes forest, the results from SMEAR II station are presented as follows.

4.6.1 Monoterpene oxidations

The oxidation processes of monoterpenes at the SMEAR II station have been investigated by several previous studies, mostly

480 based on the highly oxidized compounds. Utilizing non-negative matrix factorization analysis on iodide-adduct CIMS data at the SMEAR II station, Lee et al. (2018) found that the gas-phase organic species subgroup of $C_{6-10}H_yO_{\geq 7}$ showed distinct

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daytime diel trends. Yan et al. (2016) conducted source apportionment of HOMs at the SMEAR II station and separated various HOM formation pathways, such as monoterpene ozonolysis and monoterpene oxidation initiated by NO_3 radical. In this study, three types of monoterpene reaction products were detected: monoterpene lightly oxidized compounds, monoterpene more oxidized compounds, and monoterpene-derived organic nitrates. The latter two were not clearly separated into different factors at the SMEAR II station due to the similarities in their overall time trends. For example, the time series of $C_{10}H_{15}NO_5H^+$ correlate well with those of $C_{10}H_{16}O_4H^+$ and $C_{10}H_{16}O_5H^+$ ($r^2 > 0.61$).

Consistent with previous observations, monoterpene more oxidized products (i.e., $C_{10}H_{16}O_4$ and $C_{10}H_{14}O_5$) have a broad high distribution throughout the day due to the active photochemical processes during daytime. Monoterpene-derived organic nitrates (i.e., $C_{10}H_{17}NO_4$, $C_{10}H_{15}NO_5$, and $C_9H_{13}NO_6$) are mainly characterized by a distinct morning peak at around 8:00, approximately 2 h after the NO peak. But their intensities are also elevated at night. PMF analysis of NO_3^- CIMS dataset observed similar diurnal variations of terpene organic nitrates factor at a forest site in the southeastern US (Massoli et al., 2018). Compared with β-pinene and most other monoterpenes, the overall organic nitrate yield from α-pinene + NO_3 is rather low (Fry et al., 2014; Kurtén et al., 2017). Laboratory studies found that using iodide-adduct FIGAERO-HR-ToF CIMS, $C_{10}H_{15}NO_6$ is the most abundant organic nitrate in both gas- and particle-phase measurements of α-pinene + NO_3 reactions (Nah et al., 2016). Boyd et al. (2015) mainly detected $C_{10}H_{17}NO_4$, $C_{10}H_{15}NO_5$, $C_{10}H_{17}NO_5$, and $C_{10}H_{15}NO_6$ with iodide-adduct CIMS from the α-pinene + NO_3 system. Using $C_{10}H_{17}NO_5$ and $C_{10}H_{15}NO_6$ as the examples, we checked their correlations with the products of [OH] × [monoterpenes], [O_3] × [monoterpenes], and [NO_3] × [monoterpenes] in different periods of the day (Fig. 14; Fig. S16). Comparatively, $C_{10}H_{17}NO_5$ and $C_{10}H_{15}NO_6$ correlate better with the products of [OH] × [monoterpenes] and [O_3] × [monoterpenes] during daytime (9:00~18:00). However, for the product of [NO_3] × [monoterpenes], its correlation coefficients with $C_{10}H_{17}NO_5$ and $C_{10}H_{15}NO_6$ are a bit higher at night (20:00 to 4:00 of the next day). These results indicate that monoterpene-derived organic nitrates can be mainly formed by the NO_3-initiated oxidations at night, but in daytime by the OH and O_3-initiated oxidations followed by NO termination of the RO_2. It should be noted that $C_{10}H_{17}NO_5$ and $C_{10}H_{15}NO_6$ are used as examples because both of them are fingerprint peaks of the factor, but in real environments it may not be the case that these molecules are always produced from the above formation routes.

4.6.2 Sesquiterpene oxidations

The lightly oxygenated compounds from sesquiterpene reactions present a big morning peak and elevated signal intensities at night, similar to the diurnal variations of monoterpene lightly oxidized products. Hellén et al. (2018) showed that at the SMEAR II station, O_3 oxidation dominated the first step of sesquiterpene reactions for the whole year. It has also been observed in central Amazonia that sesquiterpenes contributed the highest to total O_3 reactivity although sesquiterpene concentrations were much lower than those of monoterpenes and isoprene (Yee et al., 2018). At the SMEAR II station, emissions of sesquiterpenes are dominated by β-caryophyllene (Hellén et al., 2018). Photooxidation of β-caryophyllene in the chamber experiments resulted in high aerosol yield and is expected to strongly influence SOA formation (Jaoui et al., 2013). Using the mass spectrometric techniques, Jokinen et al. (2016) observed the production of highly oxidized organic compounds from β-

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515 caryophyllene ozonolysis, i.e., monomers $C_{15}H_{24}O_{7,9,11}$ and $C_{15}H_{22}O_{9,11}$, and dimers $C_{29}H_{46}O_{12,14,16}$ and $C_{30}H_{46}O_{12,14,16}$. However, due to the instrumental limitation, only the lightly oxidized products from sesquiterpene reactions were identified in this study.

Interestingly, a strong RH-dependence was observed for the correlations between sesquiterpene lightly oxidized compounds and the product of [OH] × [sesquiterpenes] or [O$_3$] × [sesquiterpenes]. These products represent the oxidation rates
520 of sesquiterpenes with OH radical and O$_3$. As shown in Fig. 15, the corresponding correlation coefficients vary significantly with RH. In addition, the signal intensities of sesquiterpene lightly oxidized products also show high dependence on RH. At lower RH (RH<40%), the signal intensities of sesquiterpene lightly oxidized products are relatively low and correlate closely with the product of [OH] × [sesquiterpenes] and [O$_3$] × [sesquiterpenes]. The high signal intensities of sesquiterpene lightly oxidized products occur when RH>70% but the correlation between sesquiterpene lightly oxidized compounds and the product
525 of [OH] × [sesquiterpenes] or [O$_3$] × [sesquiterpenes] is more scattered. Such high RH-dependence was not observed for monoterpene lightly oxidized compounds (Fig. S17). These findings have not been observed by previous studies and the reasons behind remain unclear. High-RH conditions typically occur during nights with temperature inversion (Zha et al., 2018), while RH below 40% generally only occurs at the station during sunny days. The controlling role of temperature can be ruled out because temperature is strongly anti-correlated with RH and is known to influence terpene emissions and terpene reaction
530 rates. Future studies are needed to dig deep into the atmospheric processes of sesquiterpenes and monoterpenes.

5 Concluding remarks

In this study, we conducted Vocus PTR-TOF measurements in two forest environments and performed binPMF analysis on these complex mass spectra. In addition to VOC species, Vocus PTR-TOF is able to measure large amounts of oxygenated VOCs with enhanced detection efficiency. According to the results in this work, factor analysis on Vocus PTR-TOF mass
535 spectra separated VOC precursors and their reaction products with varying oxidation degrees into different factors. These factors showed distinct characteristics in the atmosphere. Comparatively, the conventional PTR instruments or gas chromatograph-mass spectrometry (GC-MS) largely detect VOC precursors of low-mass molecules (Dewulf et al., 2002; de Gouw et al., 2007). Previous source apportionment studies on these datasets mainly identified primary biogenic and anthropogenic emission sources (Vlasenko et al., 2009; Patokoski et al., 2014; Baudic et al., 2016; Debevec et al., 2017; Sarkar
540 et al., 2017; Wang et al., 2020). Recently, factorization methods have been applied on NO$_3^-$ CIMS dataset to identify various atmospheric formation pathways of HOMs (Yan et al., 2016; Massoli et al., 2018; Zhang et al., 2019b). Here, for the first time, source apportionment of Vocus PTR-TOF data identified various primary emission sources and secondary formation pathways of atmospheric organic vapors, highlighting the novelty of Vocus PTR-TOF in measuring both VOCs and oxygenated VOCs. The relative abundances of organic precursors, the lightly oxidized products, and the more oxidized products can be utilized
545 by modellers to evaluate simulation output, improve model performance, and provide new perspectives to understand gas-phase physicochemical processes.

Compared with VOC species, VOC reaction products are generally present in much smaller amounts in the atmosphere. Therefore, utilizing a sub-range PMF analysis, or other similarly weighting method, is particularly important for Vocus PTR-TOF observations, where several orders of magnitude differences are expected between VOC precursors and their oxidation

550 products. Compared with the low mass range, the average contributions of the high mass range in total signals are significantly smaller, 2% and 9%, in the Landes forest and at the SMEAR II station, respectively. However, the identified factors in the high mass range, such as sesquiterpenes, sesquiterpene lightly oxidized products, monoterpene-derived organic nitrates, and more oxidized compounds, can provide crucial insights into atmospheric physicochemical processes. For example, we found that the correlations between sesquiterpene lightly oxidized compounds and the products of $[OH] \times [sesquiterpenes]$ or $[O_3] \times$

555 $[sesquiterpenes]$ show strong dependences on RH. High signal intensities of sesquiterpene lightly oxidized compounds only occur at high-RH conditions. Such high RH-dependence was not observed for monoterpene lightly oxidized compounds.

To summarize, this study successfully performed binPMF analysis on sub-ranges of mass spectrometry dataset acquired with a Vocus PTR-TOF in two European forest ecosystems, the Landes forest and a southern Finnish boreal forest. Both primary emission sources and secondary oxidation processes of organic vapors were identified in the two environments,

560 particularly for terpenes and their reaction products with varying oxidation degrees (including organic nitrates). Factors of the lightly oxidized products, more oxidized products, and organic nitrates of monoterpenes/sesquiterpenes accounted for 8-12% of the measured gas-phase organic vapors in the two forests. Further interpretations show a strong RH-dependence for the behaviour of sesquiterpene lightly oxidized products but not for that of monoterpene lightly oxidized products, for which the reasons behind need more investigations in the future.

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Table 1. Summary of source identification results for the two forest sites (L, Landes; S, SMEAR II).

|	Factor name	Possible source/chemistry	Fingerprint molecules
Landes forest			
	Factor L1	$C_4H_8H^+$ ion-related	$C_4H_8H^+$, $C_4H_{10}O_2H^+$
	Factor L2	A plume event	$C_6H_6H^+$, $C_7H_6H^+$, $C_6H_6OH^+$, unidentified peaks
	Factor L3	C_6 and C_7 lightly oxidized products	$C_6H_{10}OH^+$, $C_7H_{10}OH^+$, $C_6H_{10}O_2H^+$, $C_7H_{12}O_2H^+$,
	Factor L4	Monoterpenes	$C_6H_8H^+$, $C_7H_{10}H^+$, $C_{10}H_{16}H^+$
	Factor L5	Isoprene and its oxidation products	$C_5H_8H^+$, $C_4H_6OH^+$, $C_4H_6O_3H^+$
	Factor L6	Unknown source	$C_6H_4O_2H^+$, $C_6H_6O_3H^+$, unidentified peaks with negative mass defect
	Factor L7	Monoterpene lightly oxidized products	$C_9H_{14}OH^+$, $C_{10}H_{14}OH^+$, $C_{10}H_{16}O_2H^+$, $C_{10}H_{16}O_3H^+$
	Factor L8	C_{13} lightly oxidized products	$C_{13}H_{18}O_2H^+$, $C_{13}H_{20}O_3H^+$
	Factor L9	A plume event	Unidentified peaks
	Factor L10	Sesquiterpene lightly oxidized products	$C_{15}H_{22}OH^+$, $C_{15}H_{24}OH^+$, $C_{15}H_{22}O_2H^+$, $C_{15}H_{24}O_2H^+$, $C_{15}H_{24}O_3H^+$
	Factor L11	Monoterpene more oxidized products	$C_{10}H_{16}O_4H^+$, $C_{10}H_{14}O_5H^+$, $C_{10}H_{16}O_5H^+$, $C_{10}H_{16}O_6H^+$
	Factor L12	Sesquiterpenes	$C_{15}H_{24}H^+$
	Factor L13	Monoterpene-derived organic nitrates	$C_{10}H_{15}NO_4H^+$, $C_{10}H_{15}NO_5H^+$, $C_9H_{13}NO_6H^+$, $C_{10}H_{15}NO_6H^+$
	Factor L14	C_{12}, C_{14} or C_{16} lightly oxidized products	$C_{12}H_{26}O_3H^+$, $C_{14}H_{26}O_2H^+$, $C_{16}H_{30}O_2H^+$
	Factor L15	Unknown source	D3 siloxane, D4 siloxane, unidentified peaks
SMEAR II			
	Factor S1	$C_4H_8H^+$ ion-related	$C_4H_8H^+$, $C_4H_{12}O_2H^+$, $C_4H_{14}O_3H^+$
	Factor S2	Monoterpenes	$C_6H_8H^+$, $C_7H_{10}H^+$, $C_{10}H_{16}H^+$
	Factor S3	C_6-C_9 lightly oxygenated compounds	$C_6H_{10}OH^+$, $C_6H_{12}OH^+$, $C_7H_{10}OH^+$, $C_8H_{14}OH^+$, $C_9H_{12}O_2H^+$
	Factor S4	Isoprene and its oxidation products	$C_5H_8H^+$, $C_4H_6OH^+$
	Factor S5	Monoterpene lightly oxidized products	$C_9H_{14}OH^+$, $C_{10}H_{14}OH^+$, $C_{10}H_{16}OH^+$, $C_{10}H_{16}O_2H^+$, $C_{10}H_{16}O_3H^+$
	Factor S6	Sesquiterpene lightly oxidized products	$C_{14}H_{22}OH^+$, $C_{14}H_{24}OH^+$, $C_{15}H_{22}OH^+$, $C_{15}H_{24}OH^+$
	Factor S7	Sesquiterpenes	$C_{15}H_{24}H^+$
	Factor S8	Monoterpene more oxidized products including organic nitrates	$C_{10}H_{16}O_4H^+$, $C_{14}H_{22}O_3H^+$, $C_{15}H_{24}O_3H^+$, $C_{10}H_{17}NO_3H^+$, $C_9H_{13}NO_6H^+$
	Factor S9	Unknown source	D3 siloxane, D4 siloxane, unidentified peaks

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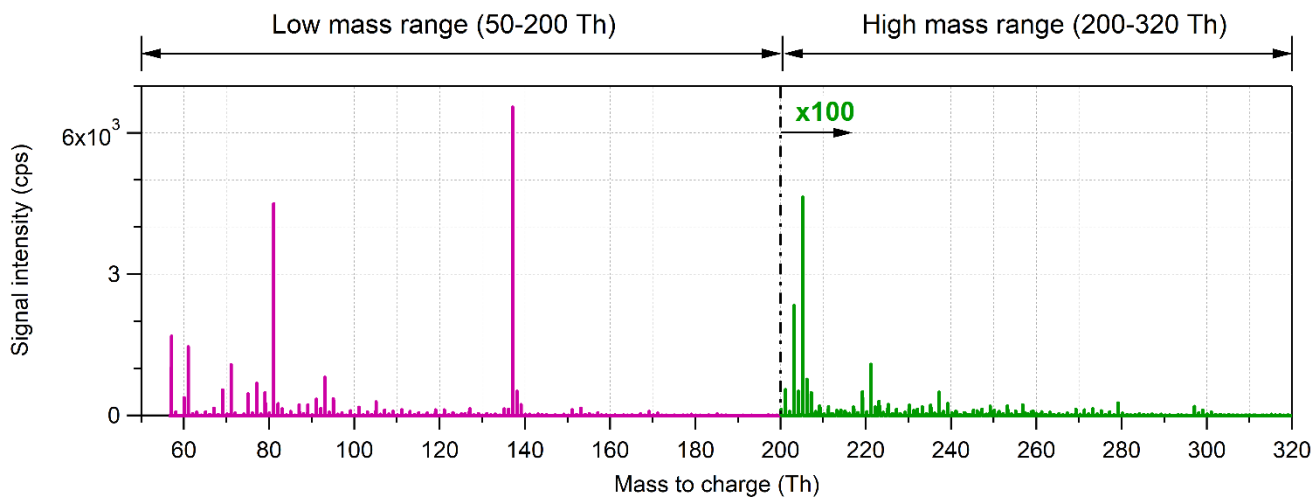
Figure 1. Average mass spectrum measured in the Landes forest. The mass spectrum is divided into two sub-ranges for further source identification analysis. The intensity scale is shown 100-fold for the high mass range (201-320 Th).
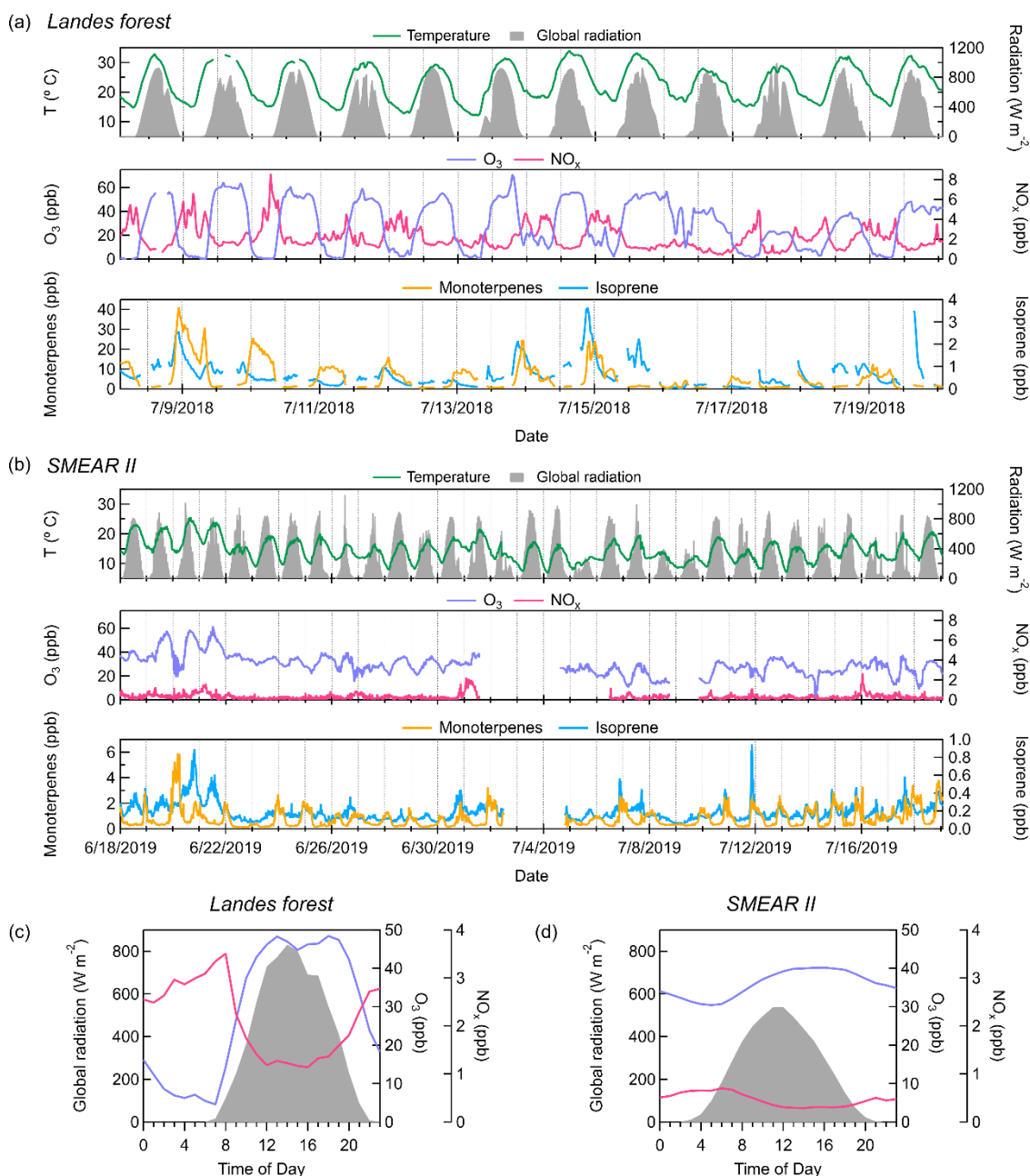
Figure 2. Time series of temperature, global radiation, concentrations of O₃ and NOₓ, and mixing ratios of monoterpenes and isoprene throughout the measurements (a) in the Landes forest and (b) at the SMEAR II station. Average diurnal cycles of global radiation, O₃ concentration (in blue), and NOₓ concentration (in pink) (c) in the Landes forest and (d) at the SMEAR II station. All parameters, except monoterpenes and isoprene, are shown in the same y-axis scale for the two sites. Monoterpene and isoprene concentrations are much lower at the SMEAR II station than in the Landes forest.
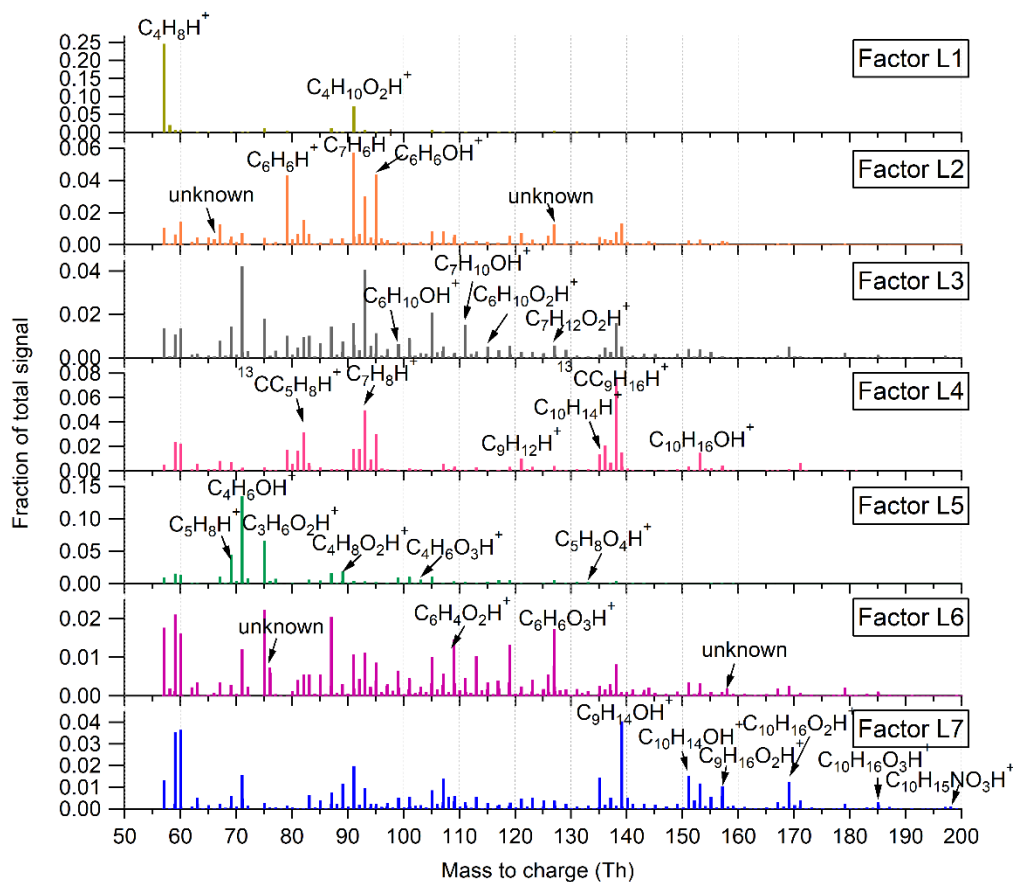
Figure 3. Mass profiles of the seven factors resolved in the low mass range in the Landes forest. Fingerprint peaks identified by high-resolution peak fitting are shown in the mass spectra.
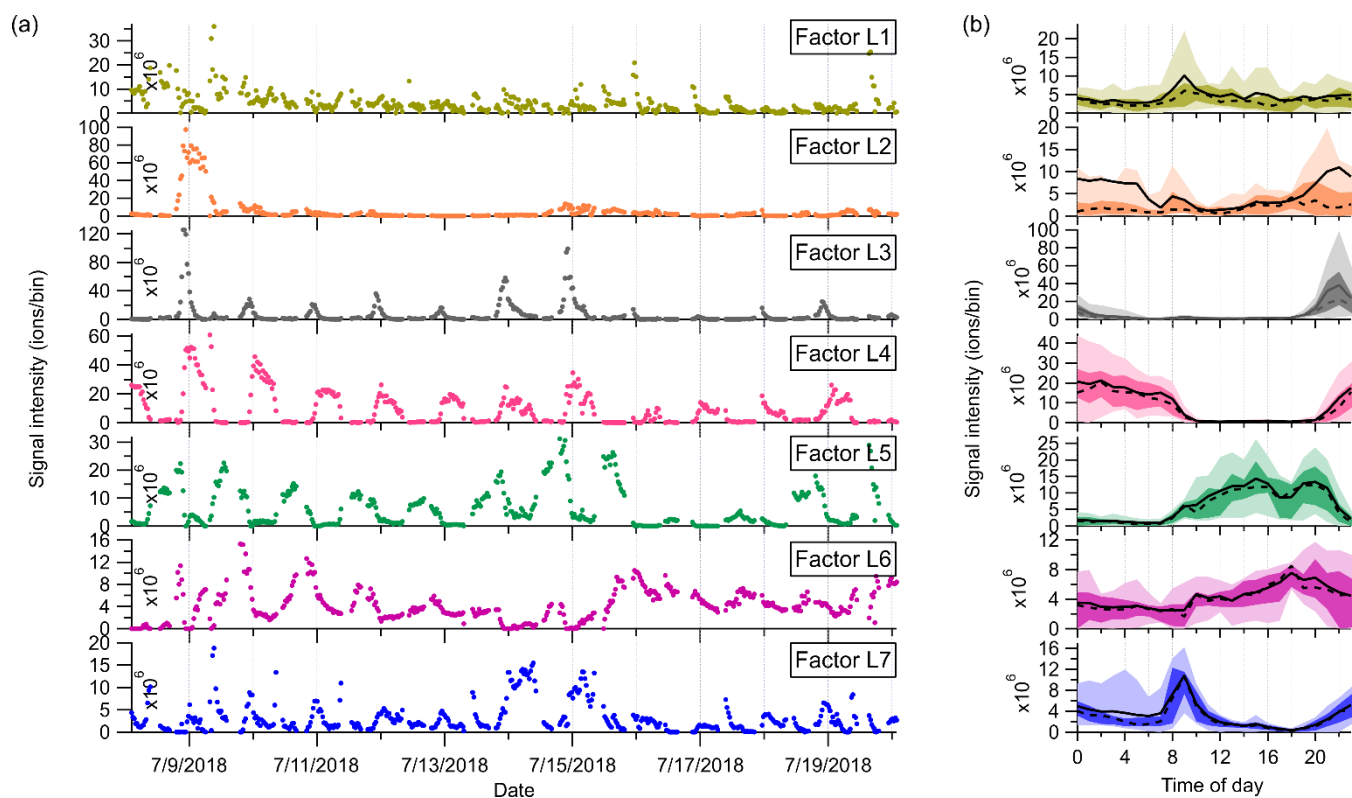
Figure 4. (a) Time series and (b) diurnal variations of the seven factors identified in the low mass range in the Landes forest. The solid and dashed lines in the diurnal plots show the mean and median values, respectively, and the shaded area shows 10th, 25th, 75th, and 90th percentiles.
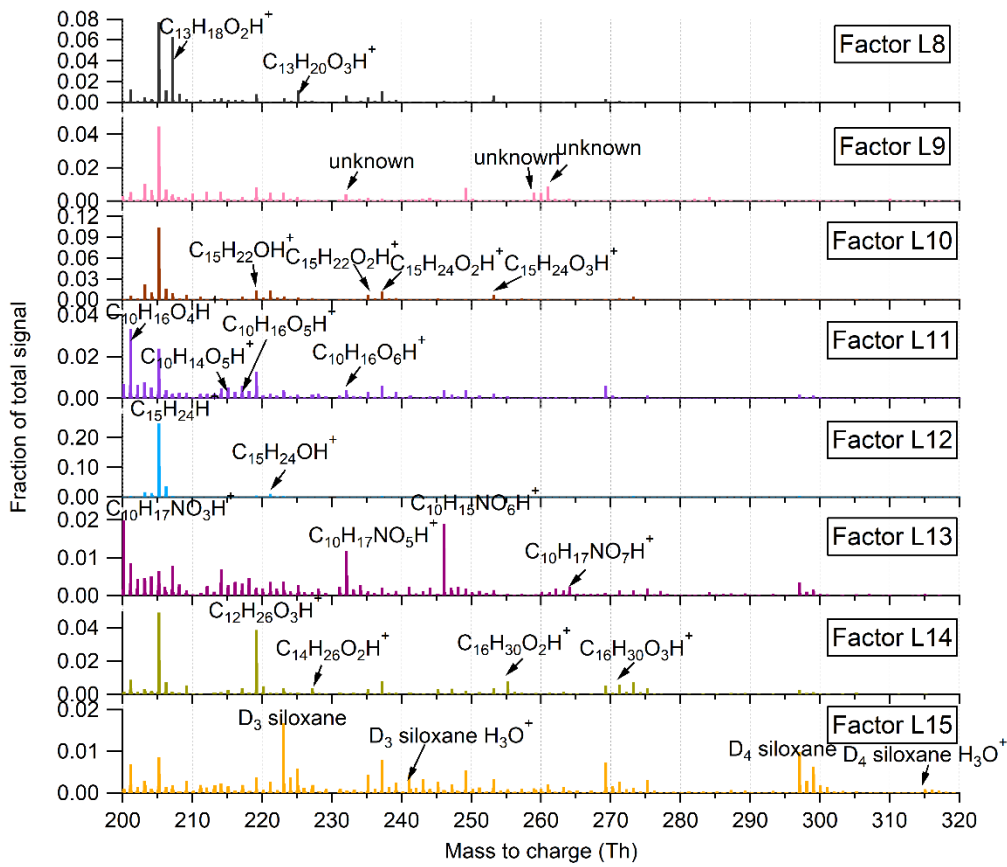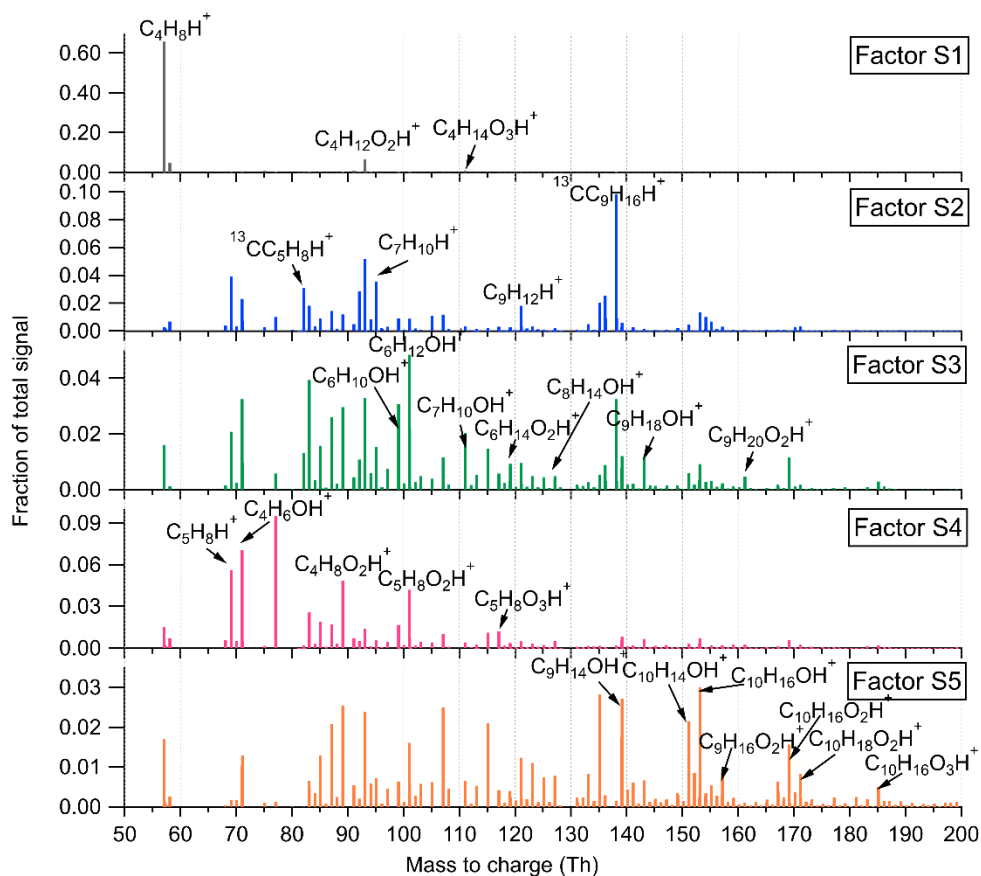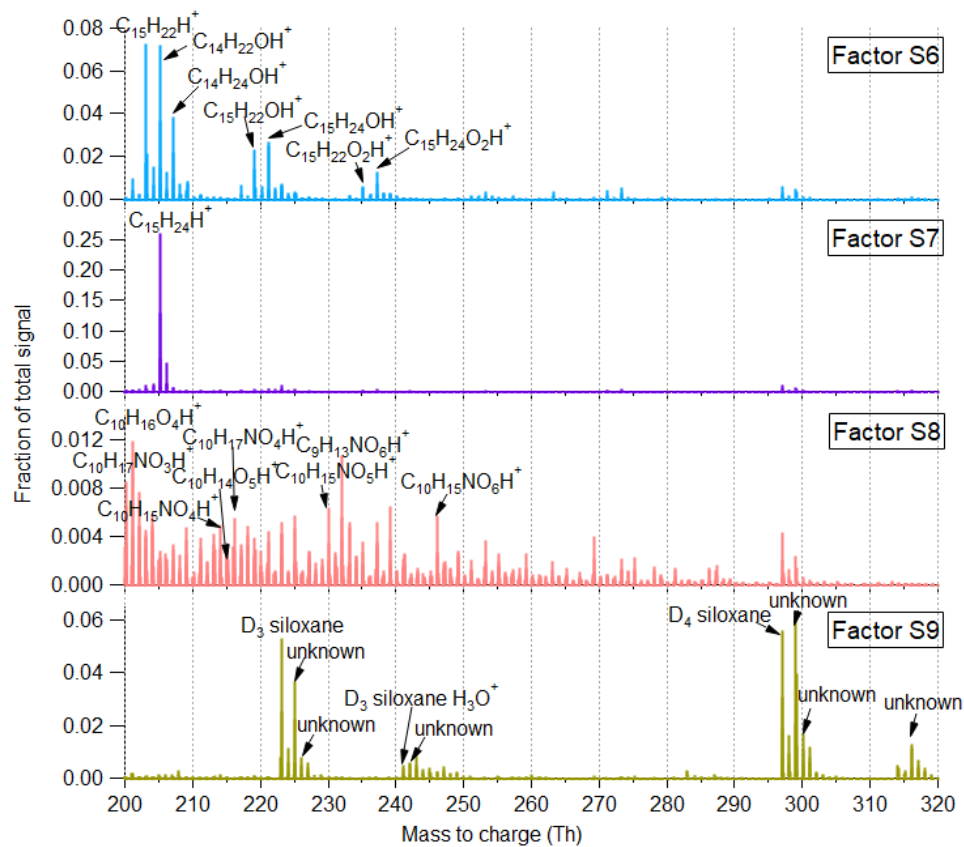
Figure 5. Mass profiles of the eight factors resolved in the high mass range in the Landes forest, with major fingerprint peaks labeled in the mass spectra.
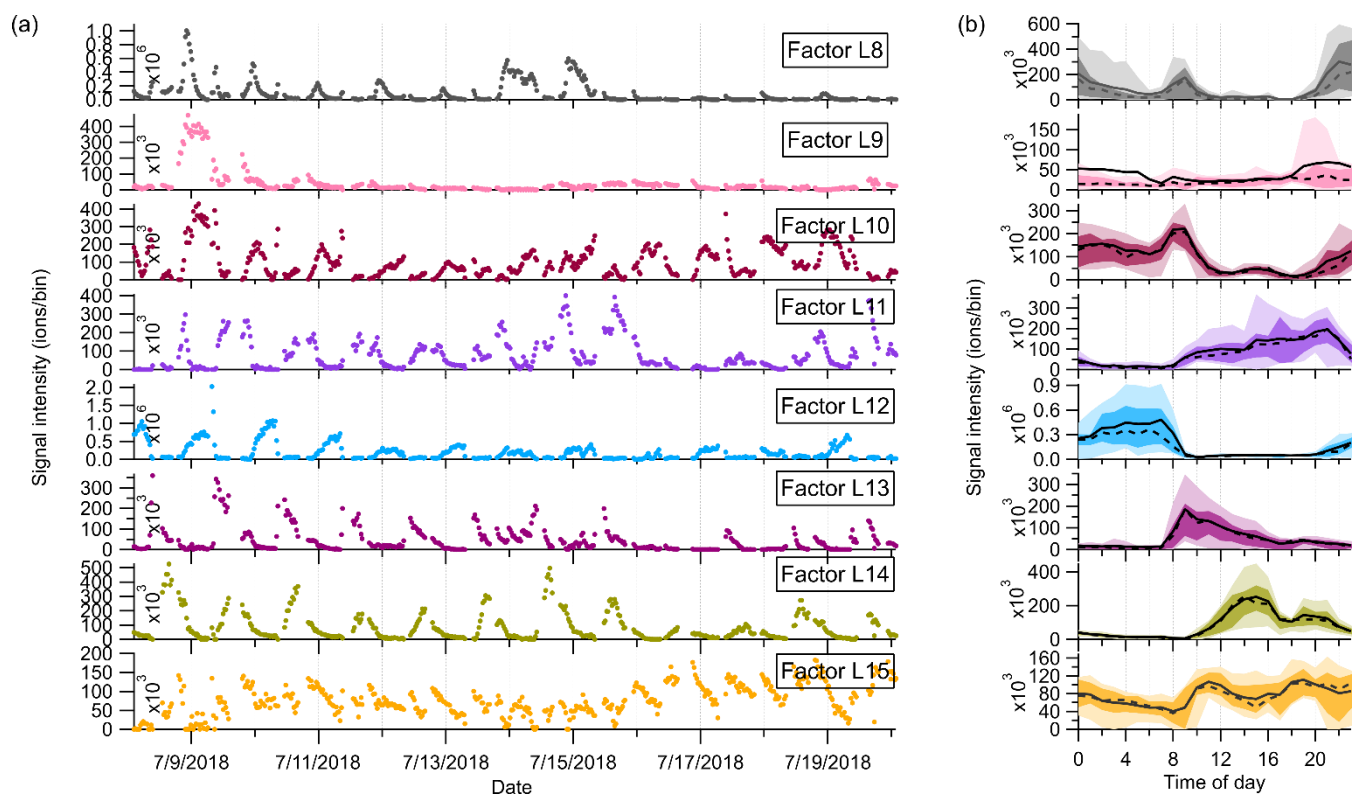
Figure 6. (a) Time series and **(b)** diurnal trends of the eight factors resolved in the high mass range in the Landes forest. The solid and dashed lines in the diurnal plots show the mean and median values, respectively, and the shaded area shows 10th, 25th, 75th, and 90th percentiles.

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Figure 7. Mass profiles of the five factors identified in the low mass range at the SMEAR II station, with fingerprint peaks shown in the mass spectra.
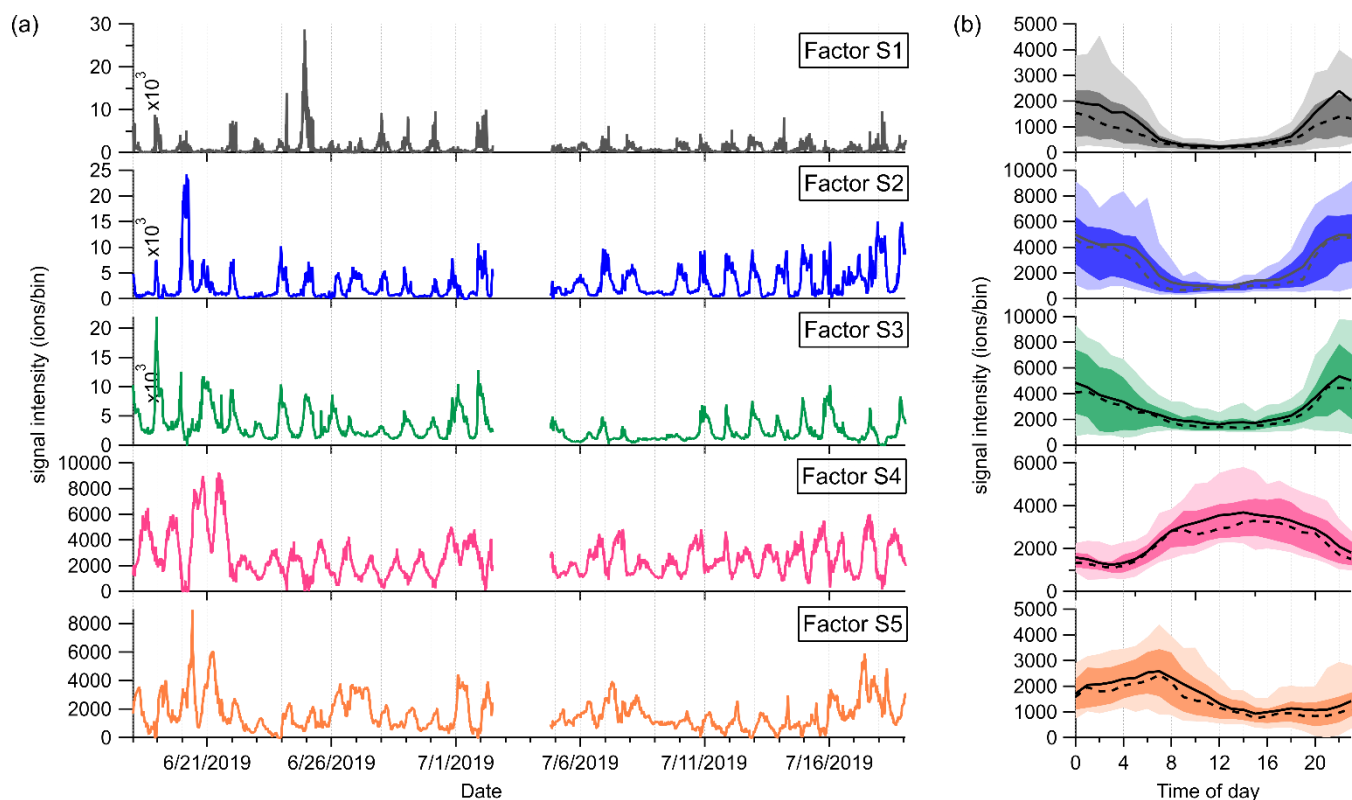
Figure 8. (a) Time series and (b) diurnal cycles of the five factors in the low mass range at the SMEAR II station. The solid and dashed lines in the diurnal plots show the mean and median values, respectively, and the shaded area shows 10th, 25th, 75th, and 90th percentiles.

Figure 9. Mass profiles of the four factors resolved in the high mass range at the SMEAR II station. The fingerprint peaks are labeled in the mass spectra.
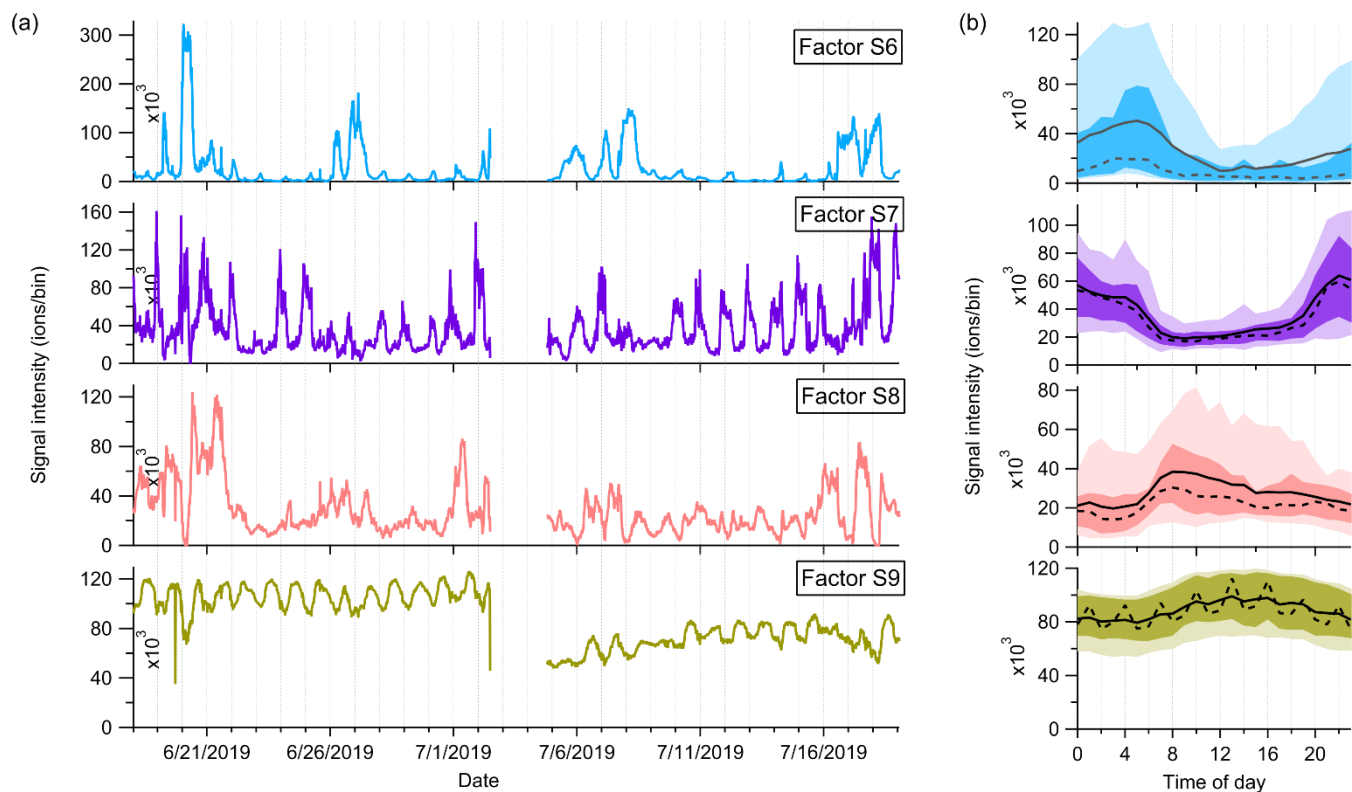
Figure 10. (a) Time series and **(b)** daily trends of the four factors in the high mass range at the SMEAR II station. The solid and dashed lines in the diurnal plots show the mean and median values, respectively, and the shaded area shows 10th, 25th, 75th, and 90th percentiles.
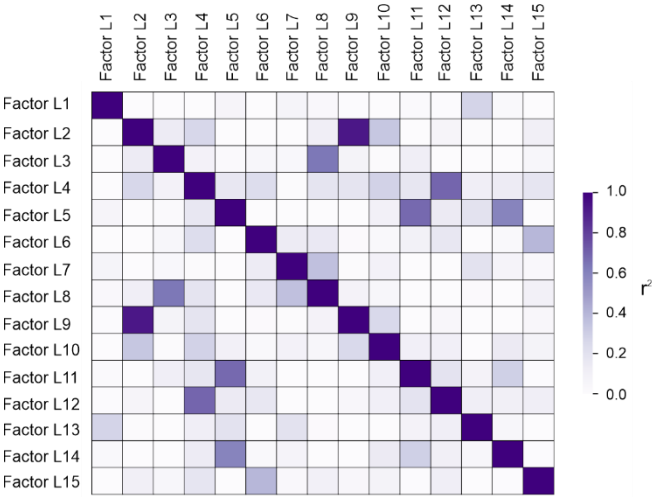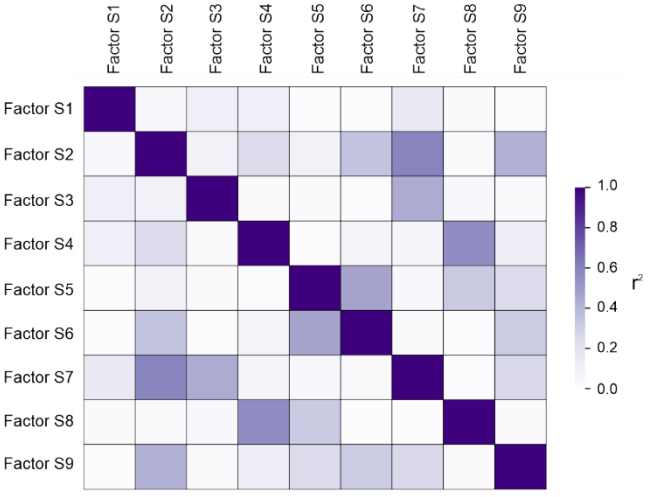
975　Figure 11. The correlations among various factors identified (a) in the Landes forest and (b) at the SMEAR II station, with the color representing the correlation coefficients (r^2).
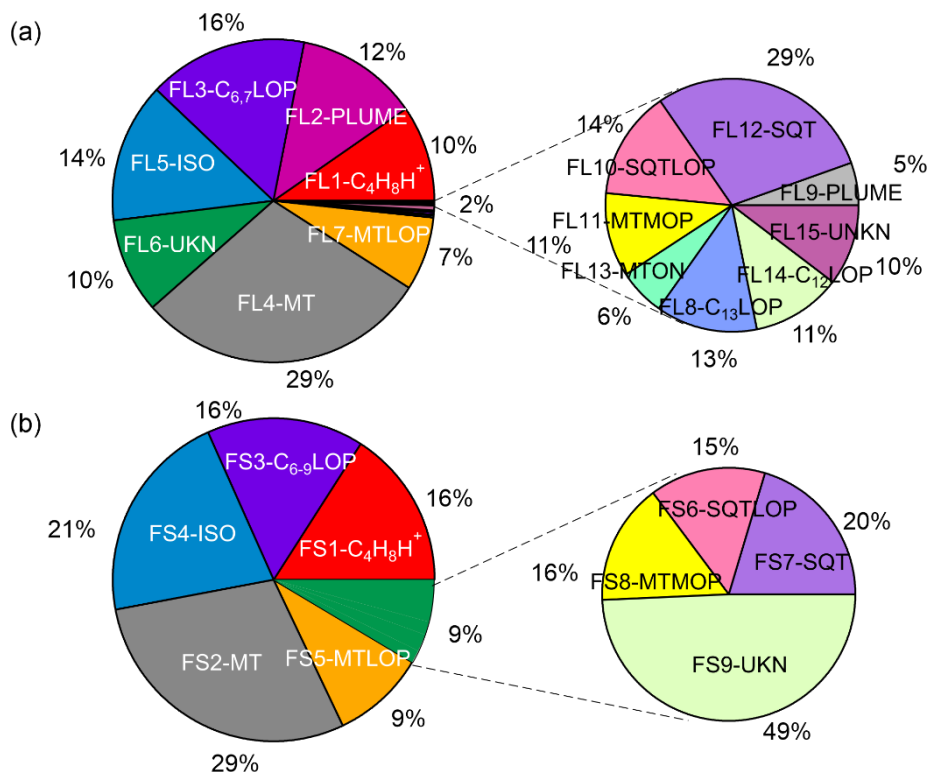
Figure 12. Average mass contributions of various identified factors in total measured organic vapors (a) in the Landes forest and (b) at the SMEAR II station. The common sources apportioned at both sites are presented in the same color in (a) and (b). FL: factors in the Landes forest; FS: factors at the SMEAR II station; $C_4H_8H^+$, $C_4H_8H^+$ ion-related; PLUME: a plume event; $C_{6,7}LOP$: C_6 and C_7 lightly oxidized products; $C_{6-9}LOP$: C_6-C_9 lightly oxygenated compounds; MT: monoterpenes; ISO: isoprene and its oxidation products; UKN: unknown source; MTLOP: monoterpene lightly oxidized products; $C_{12}LOP$: C_{12}, C_{14}, or C_{16} lightly oxidized products; $C_{13}LOP$: C_{13} lightly oxidized products; SQT: sesquiterpenes; MTON: monoterpene-derived organic nitrates; SQTLOP: sesquiterpene lightly oxidized products; MTMOP: monoterpene more oxidized products.
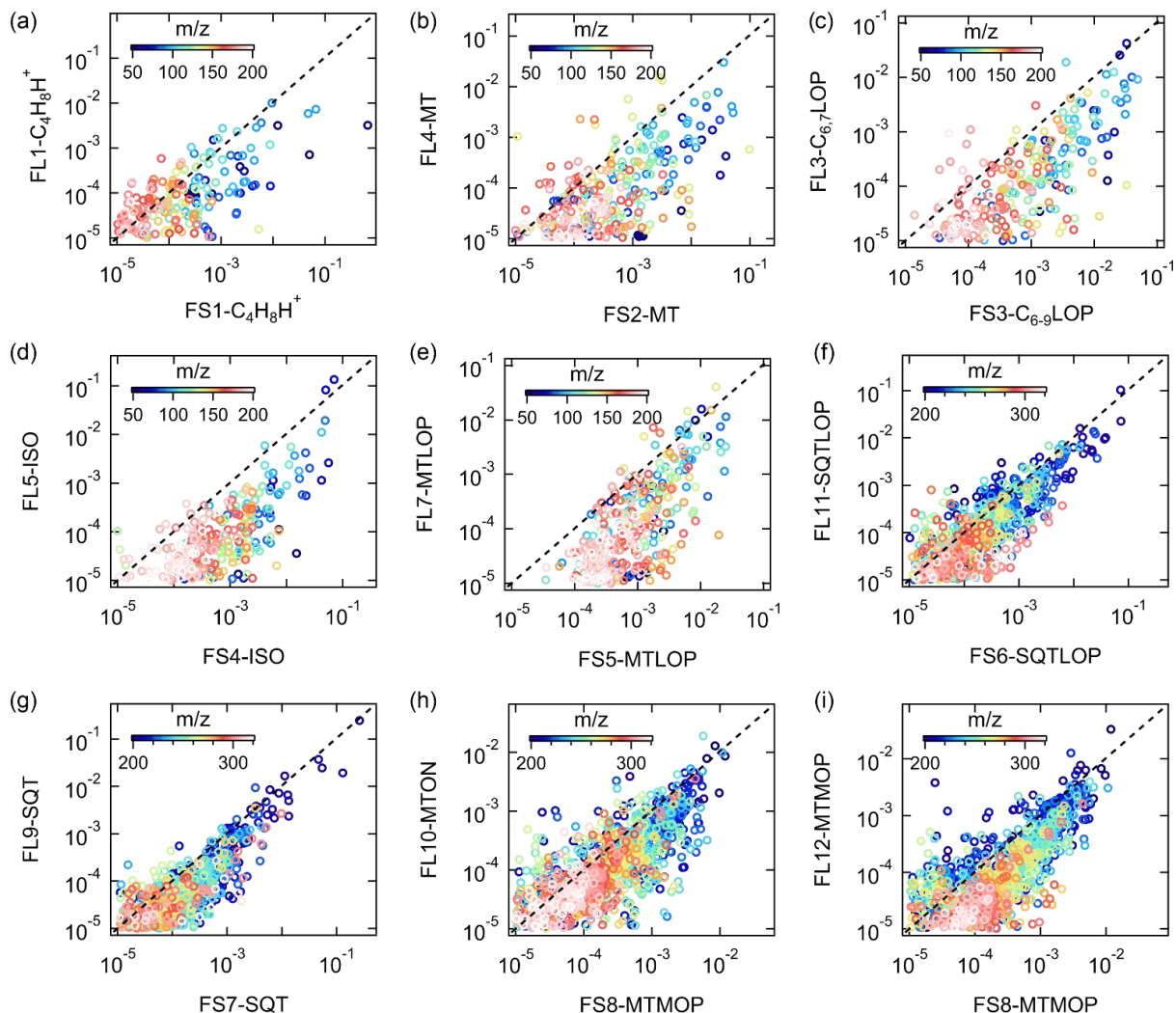
Figure 13. Comparison between factor profiles of the common sources apportioned in the Landes forest and at the SMEAR II station. The x- and y-axis show the fraction of each bin in the mass spectra of the factors. FL: factors in the Landes forest; FS: factors at the SMEAR II station; $C_4H_8H^+$, $C_4H_8H^+$ ion-related; $C_{6,7}$LOP: MT: monoterpenes; C_6 and C_7 lightly oxidized products; C_{6-9}LOP: C_6-C_9 lightly oxygenated compounds; ISO: isoprene and its oxidation products; MTLOP: monoterpene lightly oxidized products; C_{13}LOP: SQTLOP: sesquiterpene lightly oxidized products; SQT: sesquiterpenes; MTON: monoterpene-derived organic nitrates; MTMOP: monoterpene more oxidized products.
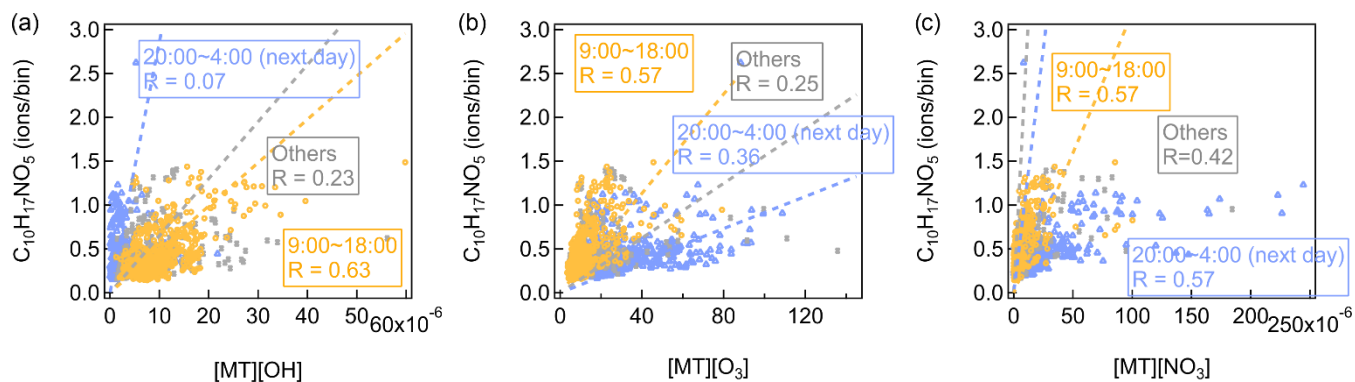
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Figure 14. Scatter plots of $C_{10}H_{17}NO_5$ versus the product of (a) [OH] × [monoterpenes], (b) [O₃] × [monoterpenes], and (c) [NO₃] × [monoterpenes]. Different colours represent different periods of the day.
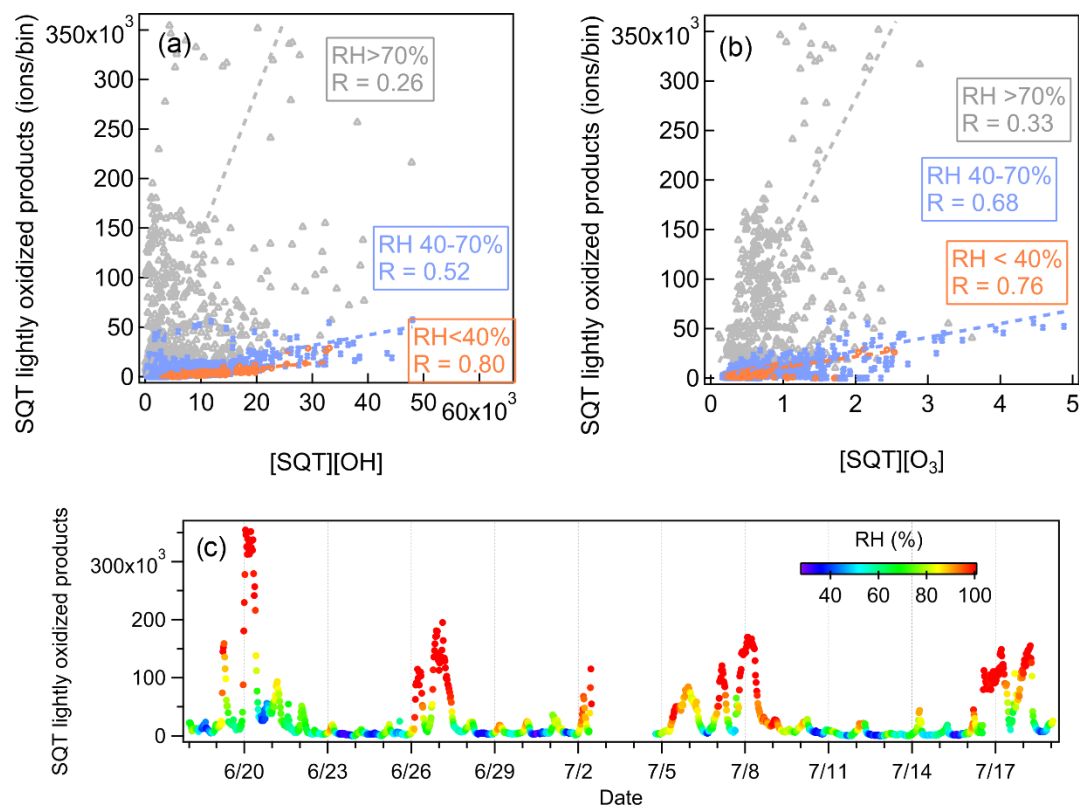
1000

Figure 15. Scatter plots of sesquiterpene lightly oxidized products versus the product of (a) [OH] × [sesquiterpenes], and (b) [O₃] × [sesquiterpenes]. Different colours indicate different ranges of RH. (c) Time series of sesquiterpene lightly oxidized products colored by RH.

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